

A comparison of design-based and model-based approaches for finite population spatial data.

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Abstract

This is the abstract.

1. Introduction

There are two general approaches for using data to make frequentist statistical inferences about a population: design-based and model-based. When data cannot be collected for all units in a population (population units), data are collected on a subset of the population units. This subset is called a sample. In the design-based approach, inferences about the underlying population are informed via a probabilistic process assigning some population units to the sample. Alternatively, in the model-based approach, inferences are made from specific assumptions about the underlying process generating the data. Each paradigm has a deep historical context (Sterba, 2009) and its own set of benefits and drawbacks (Hansen et al., 1983).

Though the design-based and model-based approaches apply to statistical inference in a broad sense, we focus on comparing these approaches for spatial data. We define spatial data as data that incorporates the specific locations of the population units into either the design or estimation process. De Gruijter and Ter Braak (1990) give an early comparison of design-based and model-based approaches for spatial data, quashing the belief that design-based approaches could not be used for spatially correlated data. Since then, there have been several general comparisons between design-based and model-based approaches for spatial data (Brus and De Gruijter, 1997; Brus, 2020; Ver Hoef, 2002; Ver Hoef, 2008; Wang et al., 2012). Cooper (2006) reviews the two approaches in an ecological context before introducing a “model-assisted” variance estimator that combines aspects from each approach. In addition to Cooper (2006), there has been substantial research and development into estimators that use both design and model-based principles (see e.g., Sterba (2009), Cicchitelli and Montanari (2012), Chan-Golston et al. (2020) for a Bayesian approach).

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Though comparisons between design-based and model-based approaches to spatial data have been studied, no numerical comparison has been made between design-based approaches that incorporate spatial locations and model-based approaches. In this manuscript, we compare design-based approaches that incorporate spatial locations to model-based approaches for spatial data. We focus on finite populations, but these comparisons generalize to infinite populations as well. A finite population contains a finite number of population units; an example is lakes (treated as a whole with the lake centroid representing location) in the contiguous United States. An infinite population contains an infinite number of population units; an example is locations within a single lake. The rest of the manuscript is organized as follows. In Section 2, we introduce and compare several sampling and estimation procedures of the design-based and model-based approaches for finite population spatial data. In Section 3, we use a simulation approach to study the behavior and performance of both approaches. In Section 4, we use both approaches to analyze real data consisting of mercury concentration from lakes in the contiguous United States. And in Section 5, we end with a discussion and provide directions for future research.

2. Background

The design-based and model-based approaches incorporate randomness in fundamentally different ways. In this section, we describe the role of randomness and its effects on subsequent inferences. We then discuss specific inference methods of the approaches for spatial data.

2.1. Comparing Design-Based and Model-Based Approaches

The design-based approach assumes the population is fixed. Randomness is incorporated via the selection of units in a sampling frame according to a sampling design. A sampling frame is the set of all units available to be sampled. A sampling design assigns a positive probability of inclusion (inclusion probability) to each unit in the sampling frame. Some examples of commonly used sampling designs include simple random sampling, stratified random sampling, and cluster sampling. If a sampling design selects units from the sampling frame while ignoring their spatial locations, we call them “Independent Random Sampling” (IRS) designs. If a sampling design selects units from the sampling frame while incorporating their spatial locations, we call them spatially balanced designs. Spatially balanced designs can be obtained using the Generalized Random Tessellation Stratified (GRTS) algorithm (Stevens and Olsen, 2004), which we discuss in more detail in Section 2.2. The design-based approach combines the randomness of the sampling design and the data collected via the sample to estimate fixed, unknown parameters (e.g., means and totals) of a population.

Treating the data as fixed and incorporating randomness through the sampling design yields estimators having very few other assumptions. Confidence intervals for these types of estimators are typically derived using limiting arguments that incorporate all possible randomizations of sampling units selected via the

sampling design. Means and totals, for example, are asymptotically normally distributed (normal) by the Central Limit Theorem (under some assumptions). If we repeatedly sample the surface, then 95% of all 95% confidence intervals constructed from a procedure with appropriate coverage will contain the true, fixed mean. Särndal et al. (2003) and Lohr (2009) provide thorough reviews of the design-based approach.

The model-based approach assumes the data are a random realization of a data-generating stochastic process. Randomness is incorporated through distributional assumptions on this process. Strictly speaking, randomness need not be incorporated through random sampling, though Diggle et al. (2010) warn against preferential sampling. Preferential sampling occurs when the process generating the data locations and the process being modeled are not independent of one another. To guard against preferential sampling, model-based approaches often still implement random sampling.

Instead of estimating fixed but unknown parameters like a mean or total (as in the design-based approach), the goal of model-based inference in the spatial context is often to predict a realized variable, or value. For example, suppose the realized mean of all population units is the value of interest. Instead of *estimating* a fixed, unknown mean, we are *predicting* the value of the mean, a random variable. Prediction intervals are then derived using assumptions of the data generating process. If we repeatedly generate the response values from the same spatial process and sample, then 95% of all 95% prediction intervals constructed from a procedure with appropriate coverage will contain their respective realized means. Cressie (1993) and Schabenberger and Gotway (2017) provide reviews of model-based approaches for spatial data. A visual comparison of the design-based and model-based assumptions is provided in Figure 1 (Ver Hoef (2002) and Brus (2020) provide similar figures).

2.2. Spatially Balanced Design and Analysis

The design-based approach can be used to select samples that are “well-spread” in space, or spatially balanced. Spatially balanced samples are useful because parameter estimates from these samples tend to vary less than parameter estimates from samples that are not spatially balanced (Barabesi and Franceschi, 2011; Benedetti et al., 2017; Grafström and Lundström, 2013; Robertson et al., 2013; Stevens and Olsen, 2004; Wang et al., 2013). The first spatially balanced sampling algorithm that saw widespread use was the Generalized Random Tessellation Stratified (GRTS) algorithm (Stevens and Olsen, 2004). To quantify the spatial balance of a sample, Stevens and Olsen (2004) proposed loss metrics based on Voroni polygons. After the GRTS algorithm was developed, several other spatially balanced sampling algorithms have emerged, including the Local Pivotal Method (Grafström et al., 2012; Grafström and Matei, 2018), Spatially Correlated Poisson Sampling (Grafström, 2012), Balanced Acceptance Sampling (Robertson et al., 2013), Within-Sample-Distance Sampling (Benedetti and Piersimoni, 2017), and Halton Iterative Partitioning Sampling (Robertson et al., 2018). In this manuscript, we use the Generalized Random Tessellation Stratified (GRTS) algorithm to select spatially balanced samples sampling because the

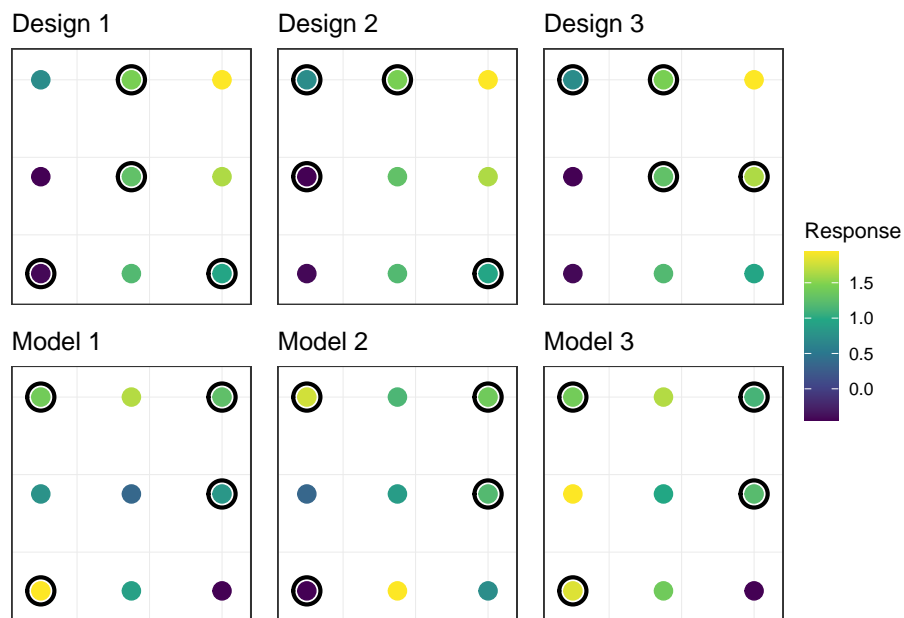


Figure 1: A comparison of sampling under the design-based and model-based frameworks. Points circled are those that are sampled. In the top row, we have one fixed population, and three random samples of size four. The response values at each site are fixed, but we obtain different estimates for the mean response because the randomly sampled sites vary from sample to sample. In the bottom row, we have three realizations of the same spatial process sampled at the same locations. The spatial process generating the response values has a single mean, but the realized mean is different in each of the three panels.

algorithm has several attractive properties. It accommodates finite and infinite sampling frames. It accommodates equal, unequal, and proportional (to size) inclusion probabilities. It accommodates legacy (historical) sampling (Foster et al., 2017). It accommodates a minimum distance between units in a sample. It accommodates replacement units in a sample, which are units that can be sampled in place of an original unit that can no longer be sampled. Lastly, it can be implemented using the `spsurvey` R package (Dumelle et al. (2021)).

The GRTS algorithm samples from finite and infinite populations by utilizing a mapping between two-dimensional and one-dimensional space. The units in the two-dimensional sampling frame are divided into cells using a hierarchical address. This hierarchical address is then used to map the units from two-dimensional space to a one-dimensional line where each unit’s line length equals its inclusion probability. A systematic sample is conducted on the line and linked back to a unit in two-dimensional space, which results in the desired sample. Stevens and Olsen (2004) provides further details.

After selecting a spatially balanced sample using the GRTS algorithm (i.e., a GRTS sample), data are collected and used to estimate population parameters. To unbiasedly estimate population means and totals from sample data, one can use the Horvitz-Thompson estimator (Horvitz and Thompson, 1952). If τ is a population total, the Horvitz-Thompson estimate of τ , denoted by $\hat{\tau}_{ht}$, is given by

$$\hat{\tau}_{ht} = \sum_{i=1}^n Z_i \pi_i^{-1}, \quad (1)$$

where Z_i is the value of the i th unit in the sample and π_i is the inclusion probability of the i th unit in the sample. An estimate of the population mean can be obtained by dividing $\hat{\tau}_{ht}$ by number of population units.

While the Horvitz-Thompson estimator is unbiased for population means and totals, it is also important to quantify the uncertainty in these estimates. Horvitz and Thompson (1952) and Sen (1953) provide variance estimators for $\hat{\tau}_{ht}$, but these estimators have two drawbacks. First, they rely on calculating π_{ij} , the probability that unit i and unit j are both in the sample – this quantity can be challenging if not impossible to calculate analytically. Second, these estimators ignore the spatial locations of the units in the sampling frame. To address these two drawbacks simultaneously, Stevens and Olsen (2003) proposed the local neighborhood variance estimator. The local neighborhood variance estimator does not rely on π_{ij} and incorporates spatial locations – for technical details see Stevens and Olsen (2003). Stevens and Olsen (2003) show the local neighborhood variance estimator tends to reduce the estimated variance of $\hat{\tau}$ compared to variance estimators ignoring spatial locations, yielding narrower confidence intervals for τ .

2.3. Finite Population Block Kriging

Finite Population Block Kriging (FPBK) is a model-based approach that expands the geostatistical Kriging framework to the finite population setting (Ver

161 Hoef, 2008). Instead of developing inference based on a specific sampling design,
 162 we assume the data are generated by a spatial process. Ver Hoef (2008) gives
 163 details on the theory of FPBK, but some of the basic principles are summarized
 164 below. Let $\mathbf{z} \equiv \{z(s_1), z(s_2), \dots, z(s_N)\}$ be an $N \times 1$ response vector at locations
 165 s_1, s_2, \dots, s_N that can be measured at the N population units. Suppose
 166 we want to predict some linear function of the response variable, $f(\mathbf{z}) = \mathbf{b}'\mathbf{z}$,
 167 where \mathbf{b}' is a $1 \times N$ vector of weights. For example, if we want to predict the
 168 population total across all population units, then we would use a vector of 1's
 169 for the weights.

170 We often only have a sample of the N population units. Denoting quantities
 171 that are part of the sampled population units with a subscript s and quantities
 172 that are part of the unsampled population units with subscript u , let

$$\begin{pmatrix} \mathbf{z}_s \\ \mathbf{z}_u \end{pmatrix} = \begin{pmatrix} \mathbf{X}_s \\ \mathbf{X}_u \end{pmatrix} \boldsymbol{\beta} + \begin{pmatrix} \boldsymbol{\delta}_s \\ \boldsymbol{\delta}_u \end{pmatrix}, \quad (2)$$

173 where \mathbf{X}_s and \mathbf{X}_u are the design matrices for the sampled and unsampled
 174 population units, respectively, and $\boldsymbol{\beta}$ is the parameter vector of fixed effects.

Let $\boldsymbol{\delta} \equiv [\boldsymbol{\delta}_s \boldsymbol{\delta}_u]'$, where $\boldsymbol{\delta}_s$ and $\boldsymbol{\delta}_u$ are random errors for the sampled and
 unsampled population units, respectively. We assume $E(\boldsymbol{\delta}) = \mathbf{0}$ and that there
 is spatial correlation in $\boldsymbol{\delta}$ that can be modeled using a covariance function.
 It is common to assume the covariance function is second-order stationary
 and isotropic (Cressie, 1993), and that the spatial covariance decreases as
 the separation between population units increases. Many spatial covariance
 functions exist, but the primary function we use throughout the simulations and
 applications in this manuscript is the exponential covariance function: the i, j th
 element of the matrix $\text{cov}(\boldsymbol{\delta})$ is

$$\text{cov}(\delta_i, \delta_j) = \begin{cases} \sigma_1^2 \exp(-h_{i,j}/\phi) & h_{i,j} > 0 \\ \sigma_1^2 + \sigma_2^2 & h_{i,j} = 0 \end{cases}, \quad (3)$$

175 where σ_1^2 is dependent random error variance measuring coarse-scale (correlated)
 176 variability, σ_n^2 is the independent random error variance measuring fine-scale
 177 (independent) variability, ϕ is the range parameter measuring the distance-decay
 178 rate of the correlation, and $h_{i,j}$ is the Euclidean distance between population
 179 units i and j . Often σ_1^2 and σ_2^2 are called the partial sill and nugget, respectively.
 180 Any spatial covariance function could be used in the place of the exponential,
 181 including functions that allow for non-stationarity or anisotropy (Chiles and
 182 Delfiner, 1999, pp. 80–93).

183 With the above model formulation, the Best Linear Unbiased Predictor
 184 (BLUP) for $f(\mathbf{b}'\mathbf{z})$ and its prediction variance can be computed. While details
 185 of the derivation are in (Ver Hoef, 2008), we note here that the predictor and
 186 its variance are both moment-based, meaning that they do not rely on any
 187 distributional assumptions.

188 We note that we only use FPBK in this paper in order to focus more on
 189 comparing the design-based and model-based approaches. Other methods, such
 190 as k-nearest-neighbors (Fix and Hodges, 1951; Ver Hoef and Temesgen, 2013),

random forest (Breiman, 2001), Bayesian models (Chan-Golston et al., 2020), among others, could also be used to obtain predictions for a mean or total from spatially correlated responses of a finite population. We choose to use FPBK because it is faster than a Bayesian approach and it was developed with theoretically-based variance estimators of means and totals for spatial data, whereas random forests and k-nearest-neighbors use ad-hoc variance estimators in most cases (Ver Hoef and Temesgen, 2013); additionally, FBPK outperformed the other methods in most scenarios.

3. Numerical Study

We used a numerical simulation study to investigate performance of four design-analysis combinations, summarized in Table 1.

	Design	Model
IRS	IRS-Design	IRS-Model
GRTS	GRTS-Design	GRTS-Model

Table 1: Types of Sampling Design and Analysis combinations considered in the simulation study. The rows give the two types of sampling designs while the columns give the two types of analyses.

We used a crossed design with the simulation parameters given in Table 2 for a total of 36 scenarios. All scenarios used exponential correlation with a $\sqrt{4}/3$ for $N = 900$ response values simulated on the unit square in either random locations (Layout = Random) or gridded locations (Layout = Gridded). The mean for the spatial process generating the response was set to zero.

For the lognormal scenarios, the response values were simulated using the specified correlation parameters using a normal distribution and were subsequently exponentiated. A total variance of 0.6931 and a mean of 0 on the normal scale yielded a total variance of 2 and a mean of 1.414 after exponentiation. Therefore, when the model-based methods were used for lognormal response, the correlation was mis-specified. We chose to simulate values with a lognormal distribution so that we could test the model-based analysis approach with a mis-specified model and so that we could test both analysis approaches on data that exhibits a large amount of skewness.

Sample Size (n)	50	100	200
Layout	Random	Gridded	-
Proportion of Dependent Error	0	0.5	0.9
Response Type	Normal	Lognormal	-

Table 2: Simulation parameters. Total variability for all scenarios was 2.

There were 2000 simulation trials for each of the 36 parameter combinations. In each trial, response values were generated from a spatial process with the specified parameters, and a GRTS sample and an IRS sample were selected.

219 For the GRTS sample, the design-based approach using the local neighborhood
 220 variance (GRTS-Design) and a model-based approach were applied (GRTS-
 221 Model). For the IRS sample, the design-based approach using the simple random
 222 sample variance (IRS-Design) and a model-based approach were applied (IRS-
 223 Model).

224 The GRTS algorithm and the local neighborhood variance estimator are
 225 available in the **R** package `spsurvey` (Dumelle et al., 2021). FPBK can be
 226 readily performed in **R** with the `sptotal` package (Higham et al., 2021). We
 227 use `sptotal` for both the simulation analysis and the application, estimating
 228 parameters with Restricted Maximum Likelihood (REML).

Figure 2 shows the relative efficiency of the four approaches from Table 1
 using the random location layout, where “IRS-Design” is the baseline:

$$\text{EFF} = \frac{\text{rMS(P)E of approach}}{\text{rMS(P)E of IRS-Design}},$$

229 where rMS(P)E is the root-mean-squared error (design-based) or the root-
 230 mean-squared-prediction error (model-based). When there is no spatial correla-
 231 tion (top row), the four approaches have approximately equal rMSPE, even when
 232 the assumptions of the model-based approaches are violated. So, using GRTS
 233 or using a spatial model does not result in much, if any, loss in efficiency even
 234 if the response variable is not spatially correlated. When there is high spatial
 235 correlation (bottom row), the GRTS-Model approach tends to perform best, but
 236 difference in relative efficiency between GRTS-Model and GRTS-Design is small.
 237 In the lognormal, high partial sill settings (bottom-right facet), GRTS-Design
 238 outperforms IRS-Model by a large margin, suggesting that the design decision
 239 (whether to use IRS or GRTS) is more important than the analysis decision
 240 (whether to analyze using model assumptions or not).

241 Unsurprisingly, Figure 2 also shows that, when the assumptions for GRTS-
 242 Model are satisfied, the approach outperforms GRTS-Design. However, even
 243 when the model that generates the data is different than the model used to fit the
 244 data, as in the lognormal response, the model-based approach still outperforms
 245 the design-based approach when there is a high amount of spatial correlation.
 246 Additionally, as the sample size increases, IRS-Design performs relatively worse
 247 compared to the other approaches. These conclusions were similar to those
 248 observed when the data were gridded.

249 We also studied 95% interval coverage among the approaches. The design-
 250 based 95% confidence intervals and model-based 95% prediction intervals are
 251 constructed using the normal distribution. Justification for the design-based
 252 intervals lies in the asymptotic normality of totals via the Central Limit Theorem,
 253 and justification for the model-based intervals lies in the normality assumption
 254 of the errors. Figure 3 shows the 95% interval coverage for each of the four
 255 approaches in the random location layout. All four approaches have somewhat
 256 similar interval coverage in all settings, with GRTS-Design having slightly lower
 257 coverage when the response is normal.

258 In the normal response settings, all approaches have coverage around 95%.
 259 This is expected, as the intervals are also based on the normal distribution In the

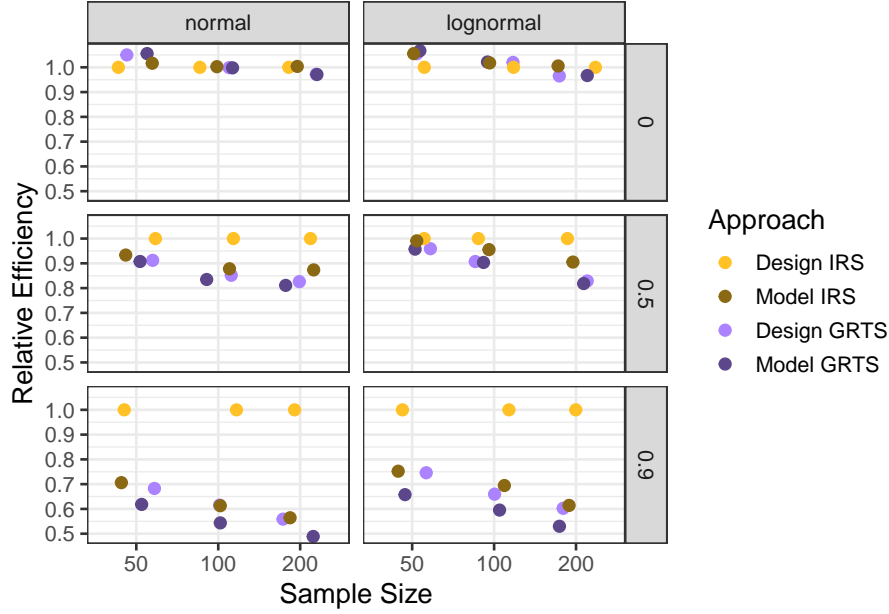


Figure 2: Relative Efficiency of the four design-analysis approaches. The plot is faceted by the type of response on the columns and the partial-sill to total-variance ratio on the rows.

lognormal response settings, however, all approaches have coverage below 95%. This is also expected, as the intervals are still based on the normal distribution. In the lognormal response settings, interval coverage increases both as the sample size increases and as the strength of spatial dependence increases. This suggests that the larger the sample size and the stronger the spatial dependence, the more resistant these intervals are to departures from normality of the data. These conclusions were similar to those observed when the data were gridded.

In addition to rMS(P)E and interval coverage, we also recorded average bias. The average bias is nearly zero for all approaches in all scenarios, so we omit a visualization of the results here. The supplementary material contains tables with mean bias, rMS(P)E , and interval coverage for all 36 simulation scenarios.

4. Application

The Environmental Protection Agency (EPA), states, and tribes periodically conduct National Aquatic Research Surveys (NARS) in the United States to assess the water quality of various bodies of water. We will use the 2012 National Lakes Assessment (NLA), which measures various aspects of lake health and quality in lakes in the contiguous United States, to obtain an interval for mean mercury concentration. Although we know the true mean mercury concentration values for the 986 lakes from the 2012 NLA, we will explore

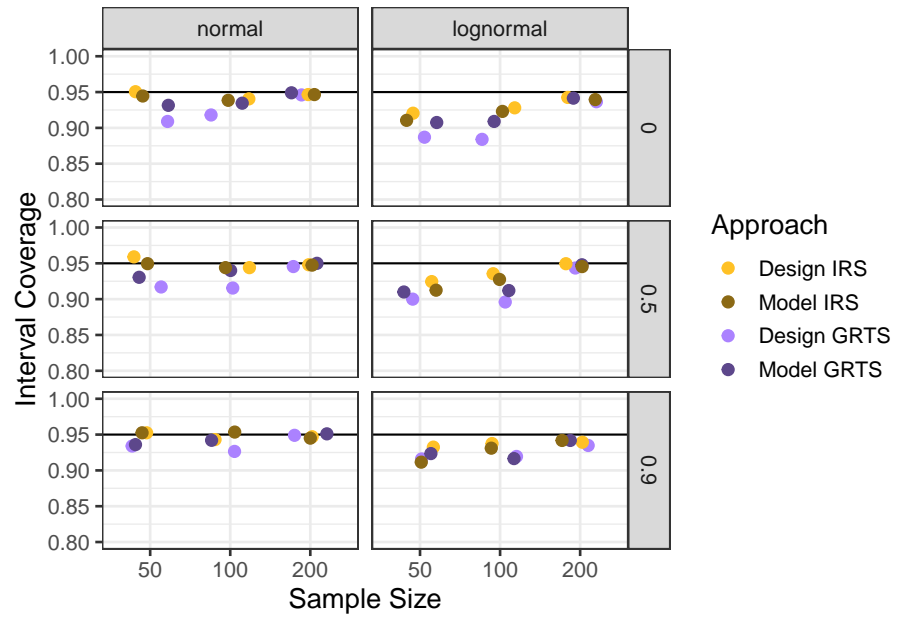


Figure 3: Coverage of the four design-analysis approaches. All confidence intervals are normal-based and have a nominal confidence level of 0.95, marked with a horizontal line. The plot is faceted by the type of response on the columns and the partial-sill to total-variance ratio on the rows.

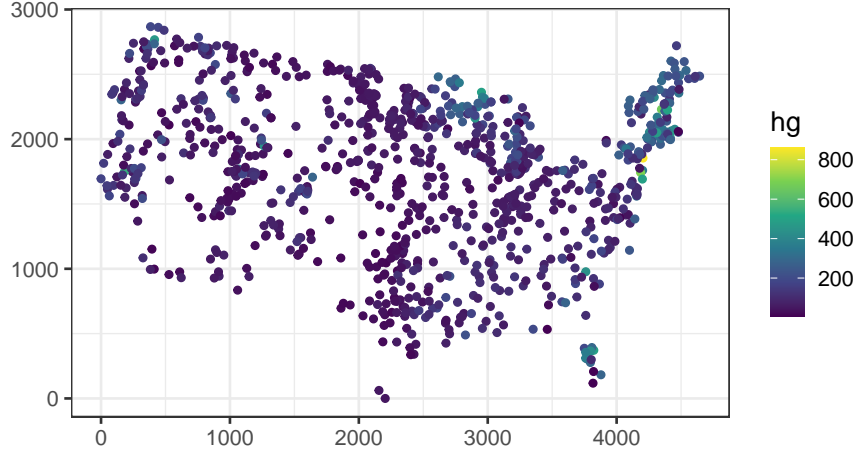


Figure 4: Population distribution of mercury concentration for 986 lakes in the contiguous United States.

whether or not we obtain an adequately precise estimate for the realized mean mercury concentration if we sample only 100 of the 986 lakes.

Figure 4 shows that mercury concentration is right-skewed, with most lakes having a low value of mercury concentration but a few having a much higher concentration. Mercury concentration exhibits some spatial correlation, with high mercury concentrations in lakes in the northeast and north central United States. The realized mean mercury concentration in the 986 lakes is 103.2 ng / g.

Approach	Estimate	SE	95% LB	95% UB
IRS-Design	112.7	8.8	95.4	129.9
IRS-Model	110.5	7.9	95.0	125.9
GRTS-Design	101.8	6.1	89.8	113.7
GRTS-Model	102.3	5.9	90.8	113.9

Table 3: Application of design-based and model-based approaches to the NLA data set on mercury concentration. The true mean concentration is 103.2 ng / g.

Table 3 shows the application of a design-based analysis of an IRS sample, a model-based analysis of an IRS sample, a design-based analysis of a GRTS sample, and a model-based analysis of a GRTS sample. For all four analyses, the true realized mean mercury concentration is within the bounds of the 95% intervals.

291 However, we should not generalize the results of this particular realization to
292 any other data set or even to other potential samples of this data set.

293 But, we do note a couple of patterns. The design-based IRS analysis shows
294 the largest standard error: a likely reason is that this is the only approach that
295 does not incorporate any spatial information regarding mercury concentration
296 across the contiguous United States. We also see that both approaches using
297 the GRTS sample have a lower standard error than the both approaches using
298 the IRS sample. We would expect this to be the case for most samples because
299 mercury concentration exhibits spatial patterning, so a spatially balanced sample
300 should usually yield a lower standard error.

301 5. Discussion

302 The design-based and model-based approaches to inference are fundamentally
303 different paradigms by which to samples are selected and data are analyzed. The
304 design-based approach incorporates randomness through sampling to estimate
305 a population parameter. The model-based approach incorporates randomness
306 through distributional assumptions to predict the realized value of a random
307 variable. Though these approaches have often been compared in the literature
308 both from theoretical and analytical perspectives, our contribution lies in study-
309 ing them in a spatial context while implementing spatially balanced sampling.
310 Aside from the theoretical differences described, a few analytical findings are
311 particularly notable: the design decision (GRTS vs IRS) seems much more
312 important than the analysis decision (design-based vs model-based); independent
313 of the analysis approach, there is no reason to prefer IRS over GRTS for spatial
314 data – GRTS tends to perform at least as well as IRS when there is no spatial
315 correlation and increasingly better than IRS as the strength of spatial correlation
316 increases; the gap in relative efficiency between GRTS-design and GRTS-model
317 widens as the strength of spatial correlation increases; and when the data are
318 skewed, interval coverage for all approaches improves both as the sample size
319 increases and as the strength of correlation increases.

320 There are several benefits and drawbacks of the design-based and model-based
321 approaches for spatial data, some of which we have not yet discussed but are
322 worthy of consideration in future research. Design-based approaches are often
323 computationally efficient, while model-based estimation of covariance parameters
324 can be computationally burdensome, especially for likelihood-based methods such
325 as REML that rely on inverting a covariance matrix. The design-based approach
326 also more naturally handles binary data, free from the more complicated logistic
327 regression formulation commonly used to handle binary data in a model-based
328 approach. The model-based approach, however, can more naturally quantify the
329 relationship between covariates (predictor variables) and the response variable.
330 The model-based approach also yields estimated spatial covariance parameters,
331 which help better understand the process of study. Model selection is also
332 possible using model-based approaches and criteria such as likelihood ratio tests
333 or AIC (Akaike, 1974). Model-based approaches are capable of more efficient
334 small-area estimation than design-based approaches by leveraging distributional

assumptions in areas with few observed sites. Model-based approaches can also compute site-by-site predictions at unobserved locations and use them to construct informative visualizations. The benefits and drawbacks of both approaches, alongside our theoretical and analytical comparisons, should be heavily considered when choosing among them. This is especially true from an analysis perspective, as we found that using a spatially balanced sampling algorithm benefits both design-based and model-based analyses.

Data and Code Availability

This manuscript has a supplementary R package that contains all of the data and code used. Instructions for download are available at <https://github.com/michaeldumelle/DvMsp>.

Supplementary Material

In the supplementary material, we provide summary statistics for all 36 simulation scenarios.

References

- Akaike, H., 1974. A new look at the statistical model identification. *IEEE transactions on automatic control* 19, 716–723.
- Barabesi, L., Franceschi, S., 2011. Sampling properties of spatial total estimators under tessellation stratified designs. *Environmetrics* 22, 271–278.
- Benedetti, R., Piersimoni, F., 2017. A spatially balanced design with probability function proportional to the within sample distance. *Biometrical Journal* 59, 1067–1084.
- Benedetti, R., Piersimoni, F., Postiglione, P., 2017. Spatially balanced sampling: A review and a reappraisal. *International Statistical Review* 85, 439–454.
- Breiman, L., 2001. Random forests. *Machine learning* 45, 5–32.
- Brus, D., De Gruijter, J., 1997. Random sampling or geostatistical modelling? Choosing between design-based and model-based sampling strategies for soil (with discussion). *Geoderma* 80, 1–44.
- Brus, D.J., 2020. Statistical approaches for spatial sample survey: Persistent misconceptions and new developments. *European Journal of Soil Science*.
- Chan-Golston, A.M., Banerjee, S., Handcock, M.S., 2020. Bayesian inference for finite populations under spatial process settings. *Environmetrics* 31, e2606.
- Chiles, J.-P., Delfiner, P., 1999. *Geostatistics: Modeling Spatial Uncertainty*. John Wiley & Sons, New York.
- Cicchitelli, G., Montanari, G.E., 2012. Model-assisted estimation of a spatial population mean. *International Statistical Review* 80, 111–126.
- Cooper, C., 2006. Sampling and variance estimation on continuous domains. *Environmetrics: The official journal of the International Environmetrics Society* 17, 539–553.

374 Cressie, N., 1993. Statistics for spatial data. John Wiley & Sons.

375 De Gruijter, J., Ter Braak, C., 1990. Model-free estimation from spatial samples:
376 A reappraisal of classical sampling theory. *Mathematical geology* 22, 407–415.

377 Diggle, P.J., Menezes, R., Su, T., 2010. Geostatistical inference under prefer-
378 ential sampling. *Journal of the Royal Statistical Society: Series C (Applied*
379 *Statistics)* 59, 191–232.

380 Dumelle, M., Kincaid, T.M., Olsen, A.R., Weber, M.H., 2021. Spsurvey: Spatial
381 sampling design and analysis.

382 Fix, E., Hodges, J.L., 1951. Discriminatory analysis, nonparametric discrimina-
383 tion: Consistency properties. *USAF School of Aviation Medicine*.

384 Foster, S.D., Hosack, G.R., Lawrence, E., Przeslawski, R., Hedge, P., Caley,
385 M.J., Barrett, N.S., Williams, A., Li, J., Lynch, T., others, 2017. Spatially
386 balanced designs that incorporate legacy sites. *Methods in Ecology and*
387 *Evolution* 8, 1433–1442.

388 Grafström, A., 2012. Spatially correlated poisson sampling. *Journal of Statistical*
389 *Planning and Inference* 142, 139–147.

390 Grafström, A., Lundström, N.L., 2013. Why well spread probability samples are
391 balanced. *Open Journal of Statistics* 3, 36–41.

392 Grafström, A., Lundström, N.L., Schelin, L., 2012. Spatially balanced sampling
393 through the pivotal method. *Biometrics* 68, 514–520.

394 Grafström, A., Matei, A., 2018. Spatially balanced sampling of continuous
395 populations. *Scandinavian Journal of Statistics* 45, 792–805.

396 Hansen, M.H., Madow, W.G., Tepping, B.J., 1983. An evaluation of model-
397 dependent and probability-sampling inferences in sample surveys. *Journal of*
398 *the American Statistical Association* 78, 776–793.

399 Higham, M., Ver Hoef, J., Frank, B., Dumelle, M., 2021. Sptotal: Predicting
400 totals and weighted sums from spatial data.

401 Horvitz, D.G., Thompson, D.J., 1952. A generalization of sampling without
402 replacement from a finite universe. *Journal of the American statistical*
403 *Association* 47, 663–685.

404 Lohr, S.L., 2009. Sampling: Design and analysis. Nelson Education.

405 Robertson, B., Brown, J., McDonald, T., Jaksons, P., 2013. BAS: Balanced
406 acceptance sampling of natural resources. *Biometrics* 69, 776–784.

407 Robertson, B., McDonald, T., Price, C., Brown, J., 2018. Halton iterative
408 partitioning: Spatially balanced sampling via partitioning. *Environmental*
409 *and Ecological Statistics* 25, 305–323.

410 Särndal, C.-E., Swensson, B., Wretman, J., 2003. Model assisted survey sampling.
411 Springer Science & Business Media.

412 Schabenberger, O., Gotway, C.A., 2017. Statistical methods for spatial data
413 analysis. CRC press.

414 Sen, A.R., 1953. On the estimate of the variance in sampling with varying
415 probabilities. *Journal of the Indian Society of Agricultural Statistics* 5, 127.

416 Sterba, S.K., 2009. Alternative model-based and design-based frameworks for
417 inference from samples to populations: From polarization to integration.
418 *Multivariate behavioral research* 44, 711–740.

419 Stevens, D.L., Olsen, A.R., 2003. Variance estimation for spatially balanced
420 samples of environmental resources. *Environmetrics* 14, 593–610.

421 Stevens, D.L., Olsen, A.R., 2004. Spatially balanced sampling of natural re-
422 sources. *Journal of the american Statistical association* 99, 262–278.

423 Ver Hoef, J., 2002. Sampling and geostatistics for spatial data. *Ecoscience* 9,
424 152–161.

425 Ver Hoef, J.M., 2008. Spatial methods for plot-based sampling of wildlife
426 populations. *Environmental and Ecological Statistics* 15, 3–13.

427 Ver Hoef, J.M., Temesgen, H., 2013. A comparison of the spatial linear model
428 to nearest neighbor (k-NN) methods for forestry applications. *PloS one* 8,
429 e59129.

430 Wang, J.-F., Jiang, C.-S., Hu, M.-G., Cao, Z.-D., Guo, Y.-S., Li, L.-F., Liu, T.-J.,
431 Meng, B., 2013. Design-based spatial sampling: Theory and implementation.
432 *Environmental modelling & software* 40, 280–288.

433 Wang, J.-F., Stein, A., Gao, B.-B., Ge, Y., 2012. A review of spatial sampling.
434 *Spatial Statistics* 2, 1–14.