**PRZM5-VVWM**

**For Use with PWC 3**

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# PRZM5-VVWM

PRZM5 and VVWM have been compiled into a single program. This allows for more efficient implementation especially when transferring data between PRZM5 and VVWM. PRZM-VVWM is designed for batch runs with large numbers of scenarios and introduces the concept of *schemes*, which allow for entering multiple application schemes in a single batch run. This program is significantly changed in the way that it handles inputs, but mechanistically it is nearly the same as the previous version of PRZM and VVWM. Chemical inputs, Scenario and Water body inputs are now stored in separate files, which facilitates runs with multiple water bodies and fields.

Minor program changes

CAM are renumbered.

Manning’s N eliminated

Initial water content is set to Max water capacity. No longer an input.

Application efficiencies are now internally calculated based on drift

Additional groundwater degradation profiles are provided, including linear, exponential, stepped ramp.

Application window

Application schemes

Rain restriction options

# Input Files for PRZM5-VVWM

PRZM5-VVWM requires 4 text-based input files:

* PRZM5VVWM input file. This is a text file. It can be named anything and have any extension although \*.inp is often used. Data in this input file are delimited by commas or spaces.
* Scenario File (sc2 or csv file) , this can either be a scn2 legacy file(s) or a newly implemented multi-scenario csv file. It contains all the intrinsic properties of the field.
* Watershed File (\*.wat), this file contains al the information to describe the watershed extrinsic properties and the physical intrinsic properties of the waterbody.
* Meteorological data file (\*.wea). This file contains the daily weather records used in the simulation. It is associated with the field scenario and is specified in the scenario file.

## PRZM5VVWM Input File

The subroutine *read\_inputfile* handles all the main input read. Input values in the file are delimited by commas or spaces. Text that includes blanks should be inside quotation marks. The word *Line* in the text below refers to lines containing input data that are read in by the program and the following terms are used to describe the input type:

TEXT is any character combination (e.g., Jack, George E. Romero, c:\youraunt\your uncle\)

TEXT IN QUOTES is any character combination (e.g., “Jack”, “George E. Romero”, “c:\youraunt\your uncle\”)

LOGICAL: TRUE or FALSE, or just T or F

INTEGER: Whole number without a decimal point

REAL: Number with or without scientific notation (e.g., 23.1, 73.m, 41.2345e-63)

The following lines of input data are contained in the input file:

**Line 1:**

Ancillary Information, not used in program

[TEXT]

For example, a date or a descriptive title.

**Line 2:**

working\_directory

[TEXT]

Working directory, the directory where all output will be delivered. Path should end ending with a backslash “\” (e.g., c:\me\mythings\chemicals\).

**Line 3**:

family\_name

[TEXT].

Family name for output. This is a text identifier for output files. File names for each run will contain this text. Typically, it may be useful to use the chemical name (e.g., hydroxyquinone, 2-4-5 Trioxin)

**Line 4:** .

weatherfiledirectory

[TEXT]

Full path for the meteorological data file. Path should end with “\”. Note that the meteorological file name is specified in the scenario file. The weather file name in the scenario file will be appended to this path.

**Line 5:**

is\_koc, is\_freundlich, is\_nonequilibrium

[LOGICAL, LOGICAL, LOGICAL]

is\_koc: TRUE if the value in Line 6 is a Koc, False if the value is Kd, TRUE

is\_nonequilibrium: TRUE to use nonequilibrium (field only, not waterbody)

is\_freundlich: TRUE to use Freundlich isotherms (field only, not waterbody)

**Line 6:**

nchem

[INTEGER]

1= parent only, 2 = calculate parent and daughter, 3 = calculate parent, daughter, and granddaughter

**Line 7:**

N\_f\_input(1), N\_f\_input(2), N\_f\_input(3) [REAL, REAL, REAL]

Freundlich coefficients for each nchem from line 6. [(mg/kg)/(mg/L) N\_f\_input]

**Line 8:**

N\_f\_input(1), N\_f\_input(2), N\_f\_input(3)

[REAL, REAL, REAL]

Freundlich exponents corresponding to the coefficient in Line 7.

**Line 9:**

k\_f\_2\_input(1), k\_f\_2\_input(2), k\_f\_2\_input(3)

[REAL, REAL, REAL]

Freundlich coefficients for each nchem from line 6 for the nonequilibrium region.

[(mg/kg)/(mg/L) N\_f\_input]

**Line 10:**

N\_f\_2\_input(1), N\_f\_2\_input(2), N\_f\_2\_input(3)

[REAL, REAL, REAL]

Freundlich exponents corresponding to the coefficient in Line 9 for the nonequilibrium region.

**Line 11:**

**k2(1), k2(2), k2(3)**

[REAL, REAL, REAL]

Mass transfer coefficient between equilibrium and nonequilibrium regions, referenced to region 2 (per day)

**Line 12:**

lowest\_conc, number\_subdelt

[REAL, INTEGER]

lowest\_conc: concentration below which isotherms become linear. (mg/L) This prevents numerical difficulties at low concentrations.

number\_subdelt – the chemical transport can be divided into sub-daily time steps to alleviate numerical problems that might occur when using nonlinear isotherms. This specifies how many sub daily steps should occur.

**Line 13:**

water\_column\_halflife\_input(1), water\_column\_halflife\_input(2), water\_column\_halflife\_input(3), xAerobic(1), xAerobic(2)

[REAL, REAL, REAL**,** REAL, REAL, REAL]

water\_column\_halflife\_input(1 to 3): water column halflives for parent, daughter, granddaughter (days)

xAerobic(1): moles of daughter produced per mole parent degraded

xAerobic(2): moles of granddaughter produced per mole daughter degraded

**Line 14**:

water\_column\_ref\_temp(1),water\_column\_ref\_temp(2),water\_column\_ref\_temp(3)

[REAL, REAL, REAL].

The temperature (°C) at which the rates in Line 13 were derived.

**Line 15:**

benthic\_halflife\_input(1), benthic\_halflife\_input(2), benthic\_halflife\_input(3), xBenthic(1), xbenthic(2) [REAL, REAL, REAL, REAL, REAL].

benthic\_halflife\_input(1 to 3): bethic region halflives for parent, daughter, granddaughter (days)

xBenthic(1): moles of daughter produced per mole parent degraded

xBenthic(2): moles of granddaughter produced per mole daughter degraded

**Line 16:**

benthic\_ref\_temp(1), benthic\_ref\_temp(2),benthic\_ref\_temp(3)

[REAL, REAL, REAL]

The temperature (°C) at which the rates in Line 15 were derived.

**Line 17**

photo\_halflife\_input(1), photo\_halflife\_input(2), photo\_halflife\_input(3), xPhoto(1), xphoto(2)

[REAL, REAL, REAL, REAL, REAL]

Near surface aqueous photolysis half life. Applied only to the water body.

photo\_halflife\_input (1 to 3): bethic region halflives for parent, daughter, granddaughter (days)

xPhoto(1): moles of daughter produced per mole parent degraded

xPhoto(2): moles of granddaughter produced per mole daughter degraded

**Line 18**

rflat(1), rflat(2), rflat(3)

[REAL, REAL, REAL]

rflat(1 to 3): latitude at which the rates in Line 17 represented (e.g., latitude mimicked by the lab study).

**Line 19**

hydrolysis\_halflife\_input(1), hydrolysis\_halflife\_input(2), hydrolysis\_halflife\_input(3), xhydro(1), xhydro(2)

[REAL, REAL, REAL, REAL, REAL]

Hydrolysis halflife. Applied to the water body and to groundwater.

hydrolysis \_halflife\_input (1 to 3): bethic region halflives for parent, daughter, granddaughter (days)

xhydro(1): moles of daughter produced per mole parent degraded

xhydro(2): moles of granddaughter produced per mole daughter degraded

**Line 20**

soil\_degradation\_halflife\_input (1), soil\_degradation\_halflife\_input (2), soil\_degradation\_halflife\_input (3), xsoil(1), xsoil(2) , is\_total\_degradation

[REAL, REAL, REAL, REAL, REAL, LOGICAL]

Soil degradation halflife. This is the total system halflife, unless the option

xsoil(1): moles of daughter produced per mole parent degraded

xsoil(2): moles of granddaughter produced per mole daughter degraded

is\_total\_degradation: True if degradation occurs on sorbed and aqueous phases, False if degradation occurs only in the aqueous phase

**Line 21**

soil\_ref\_temp(1), soil\_ref\_temp(2), soil\_ref\_temp(3)

[REAL, REAL, REAL]

The temperature (°C) at which the rates in Line 20 were derived.

**Line 22**

plant\_pesticide\_degrade\_rate(1), plant\_pesticide\_degrade\_rate(2), plant\_pesticide\_degrade\_rate(3), foliar\_formation\_ratio\_12, foliar\_formation\_ratio\_23

[REAL, REAL, REAL, REAL, REAL]

plant\_pesticide\_degrade\_rate(1 to 3):

foliar\_formation\_ratio\_12: moles of daughter produced per mole parent degraded

foliar\_formation\_ratio\_23: moles of granddaughter produced per mole daughter degraded

**Line 23**

plant\_washoff\_coeff(1), plant\_washoff\_coeff(2), plant\_washoff\_coeff(3)

[REAL, REAL, REAL]

plant\_washoff\_coeff(1 to 3) Fraction of pesticide washed off foliage per cm of rainfall [per cm] for parent, daughter, and granddaughter

**Line 24**

mwt(1), mwt(2), mwt(3)

[REAL, REAL, REAL]

Molecular weight for parent, daughter, and granddaughter

**Line 25**

vapor\_press(1), vapor\_press(2),vapor\_press(3)

[REAL, REAL, REAL]

Vapor pressure for parent, daughter, and granddaughter [torr]

**Line 26**

solubilty(1), solubilty(2), solubilty(3)

[REAL, REAL, REAL]

Solubility for parent, daughter, and granddaughter [mg/L]

**Line 27**

Henry\_unitless(1),Henry\_unitless(2),Henry\_unitless(3)

[REAL, REAL, REAL]

Henrys’s Law Coefficient for parent, daughter, and granddaughter [volumetric concentration ratio, unitless]

**Line 28**

DAIR(1),DAIR(2),DAIR(3)

[REAL, REAL, REAL]

Diffusion coefficient in air for parent, daughter, and granddaughter [cm2/day]

**Line 29**

Heat\_of\_Henry(1), Heat\_of\_Henry(2), Heat\_of\_Henry(3)

[REAL, REAL, REAL]

Enthapy of phase transfer from dissolved in water to dissolved in air [J/mole]

**Line 30**

Q\_10

[REAL]

A multiplier of change in degradation for every 10°C increase in temperature.

**Line 31**

is\_constant\_profile

[LOGICAL]

If set to TRUE, then the degradation throughout the vertical soil profile is constant.

**Line 32**

is\_ramp\_profile, ramp1, ramp2, ramp3

[LOGICAL, REAL, REAL, REAL]

If set to TRUE, then the degradation throughout the vertical soil profile is constant

is\_ramp\_profile

ramp1 first plateau depth (cm), surface degradation rate is constant to this depth

ramp2 ramp depth as measured from surface (cm), depth at which the second plateau starts

ramp3 value of the second plateau (fraction from zero to one).

**Line 33**

is\_exp\_profile, exp\_profile1, exp\_profile2

[LOGICAL, REAL, REAL]

If set to TRUE, then the degradation declines exponentially

exp\_profile1, exponent

exp\_profile2, asymptote

**START OF SCHEME SECTION**

**Line 34**

number\_of\_schemes: Number of schemes

**(Line numbers vary from here on)**

The following entries describe the schemes, this section will repeat depending upon the number of schemes.

**Scheme Line 1**

scheme\_number, scheme\_name

[INTEGER, TEXT]

scheme\_number: used to identify scheme, only used to help users identify output files

scheme\_name: any words to help describe the scheme

**Scheme Line 2**

app\_reference\_point\_schemes

[INTEGER]

app\_reference\_point\_schemes: 0=absolute date, 1=emergence, 2=maturity, 3=removal

**Scheme Line 3**

num\_apps\_in\_schemes

[INTEGER]

num\_apps\_in\_schemes: number of applications in a year

**Scheme Line Group 4: Lines Number = Number of Applications**

**Application Information (number of lines = num\_apps\_in\_scheme)**

Date, rate, method, depth, split, drift, driftfactor, periodicity, lag

[TEXT, REAL, INTEGER, REAL, REAL, REAL,REAL, INTEGER, INTEGER]

Date: If relative date is selected, then date can be in form of number of an integer number of days relative to emergence, maturity, or harvest. If absolute date is selected, then it should be in form of ”Month/day”. Also, date can be in form of “Month/Day/Year”. If the latter is the case, then the application occurs only on that specific day. If year is not specified, then the application will occur on every year of the simulation.

Rate: chemical application (kg/ha)

Method

1 = Ground Application, chemical distributed linearly decreasing to zero chemical at 4 cm. Depth inputs have no effect.

2 = foliar application. Chemical applied above crop canopy which captures chemical in proportions to canopy areal cover. Chemical not captured is distributed into the soil linearly decreasing to zero chemical at 4 cm.

3 = soil applied, user-defined incorporation depth, uniform with depth.

4 = soil applied; chemical incorporated entirely into depth specified by user

5 = T-Band application, user-defined incorporation depth . Use *split* input variable below to define fraction of chemical to be applied in top 2 cm, remainder of chemical will be uniformly incorporated between 2 cm and the user-defined depth.

6 = soil applied, user-defined incorporation depth, linearly decreasing with depth.

7 = soil applied, user-defined incorporation depth, linearly increasing with depth.

8 = linear foliar based on crop canopy, chemical reaching the soil surface incorporated to the depth given by depth.

Depth: incorporation depth for methods without default depth (cm)

Split: For T-band applications, this variable represents the fraction of chemical that is incorporated into the top 2 cm.

Drift: integer number that represents the index of the drift value. Drift values are indexed in the waterbody input file or are hard coded for default waterbodies. See drift section.

Driftfactor: a real number used to adjust the preset drift values. This could be used for example as an adjustment to account for spray drift buffers.

Periodicity: the repeat period for the pesticide applications. For example, 1 indicates applications occur every year, 2 indicates applications every second year, and 3 indicate applications occur every 3rd year.

Lag: Applications begin after the number of lag years from the start of the simulations. Zero indicates that applications start on the first year of the simulation (i.e/, no lag)

**Scheme Line 5: Application Window Information**

is\_app\_window, app\_window\_span, app\_window\_step

[LOGICAL, INTEGER, INTEGER]

is\_app\_window: TRUE = Do multiple runs with window. FALSE = do not

app\_window\_span: Number of days across which the pesticide could be applied

app\_window\_step: Step size across the window (e.g., 1 = every day, 2 = every other day)

**Scheme Line 6: Application Window Information**

is\_adjust\_for\_rain\_schemes,rain\_limit\_schemes,optimum\_application\_window\_schemes,

intolerable\_rain\_window\_schemes,min\_days\_between\_apps\_schemes

[LOGICAL, REAL, INTEGER, INTEGER, INTEGER]

is\_adjust\_for\_rain\_schemes: TRUE = use the rain restriction routine

rain\_limit\_schemes: amount of rain to trigger restriction (cm)

optimum\_application\_window\_schemes: window specifying the span of days over which the adjusted application could be applied

intolerable\_rain\_window\_schemes: number of days looking forward when an intolerable rain should not occur

min\_days\_between\_apps\_schemes: specify the legal (or practical) minimum interval between applications

**Scheme Line 7**

number\_of\_scenarios

[INTEGER]

number\_of\_scenarios: number of SCN2 scenario files to be used in the scheme

**Scheme Group of Lines 8**

scenario\_names

[TEXT]

scenario\_names: list of full path and SCN2 scenario file names (one per line)

**Scheme Line 9**

is\_batch\_scenario

[LOGICAL]

is\_batch\_scenario: TRUE if scenarios will be read in from a CSV format file

**Scheme Line 10**

scenario\_batchfile

[TEXT]

**Water Shed/Body Section**

**WS Line 1**

ERFLAG

[INTEGER]

ERFLAG: flag to indicate erosion model, 1= MUSLE, 2 = MUST 3 = MUSS. Any other number = erosion.

**WS LINE 2** UNUSED

**WS LINE 3**  UNUSED

**WS LINE 4**  UNUSED

**WS LINE 5**  UNUSED

**WS LINE 6**  UNUSED

**WS LINE 7**

adjust\_cn

[LOGICAL]

adjust\_cn: TRUE = adjust CN according to soil water, FALSE = curve number is constant

**WS LINE 8**

itsapond, itsareservoir, itsother

[LOGICAL, LOGICAL, LOGICAL]

Itsapond:  do a standard USEPA pond simulation

Itsareservoir: do a standard USEPA reservoir simulation

Itsother: do other water bodies as read from external files

**WS LINE 9**

num\_special\_waterbodies

[INTEGER]

num\_special\_waterbodies: number of special water bodies to be simulated and which a re listed below

**WS LINE 10**

waterbody\_names

[TEXT]

waterbody\_names: list of full path and file names for the special waterbodies, one per line

OUTPUT SECTION

**OS Line 1**

is\_runoff\_output  [LOGICAL], output water runoff

**OS Line 2**

is\_erosion\_output [LOGICAL], output erosion

**OS Line 3**

is\_runoff\_chem\_output [LOGICAL], output chemical mass in runoff

**OS Line 4**

is\_erosion\_chem\_output [LOGICAL], output chemical mass in eroded solids

**OS Line 5**

is\_conc\_bottom\_output [LOGICAL], output chemical concentration at bottom compartment

**OS Line 6**

is\_daily\_volatilized\_output

**OS Line 7**

is\_daily\_chem\_leached\_output, leachdepth [LOGICAL, REAL], output chemical mass at specified depth, depth

**OS Line 8**

is\_chem\_decayed\_part\_of\_soil\_output, decay\_start, decay\_end [LOGICAL, REAL, REAL], output chemical decayed at specified depth interval, top depth, bottom depth

**OS Line 9**

is\_chem\_in\_all\_soil\_output

**OS Line 10**

is\_chem\_in\_part\_soil\_output, fieldmass\_start, fieldmass\_end

OS Line 11

is\_chem\_on\_foliage\_output

OS Line 12

is\_precipitation\_output

OS Line 13

is\_evapotranspiration\_output

OS Line 14

is\_soil\_water\_output

OS Line 15

is\_irrigation\_output

OS Line 16

is\_infiltration\_at\_depth\_output,infiltration\_point

OS Line 17

is\_infiltrated\_bottom\_output

OS Line 18

is\_waterbody\_info\_output [LOGICAL] output water body time series files

OS Line 19 UNUSED

OS Line 20

OS Line 21

OS Line 22

OS Line 23

OS Line 24

OS Line 25

OS Line 26 extra\_plots [INTEGER]

OS Line 27 PLNAME, chem\_id, MODE, ARG, ARG2,CONST

[TEXT, INTEGER, INTEGER, INTEGER, INTEGER, REAL] these are the series of output variables brought in from the old PRZM5 and previous PRZM versions. One line for each output variable.

PLNAME: Can be any of the text in Table 3.5

chem\_id: 0=not chemical output, 1= parent, 2= daughter, 3=granddaughter

MODE: values as specified in table xx

Arg: compartment number of interests, or first compartment of a rage of compartments

Arg2: last compartment of a rage of compartments, otherwise enter ARG (or any integer)

CONST: output will be multiplied by this number

## Field Property Input File

**Scenario Line 1:**  PFAC, SFAC, min\_evap\_depth

PFAC: Factor to convert pan evaporation values to evapotranspiration. Older *DVF* files used pan factor. Newer *WEA* files have evapotranspiration, so this value should be 1 for WEA files.

SFAC: snowmelt factor in cm per °C above freezing.

min\_evap\_depth: minimum depth from which evaporation is extracted year-round (cm).

**Scenario Line 2:** ERFLAG

ERFLAG: flag to calculate erosion where:

ERFLAG = 0, No Erosion calculated

ERFLAG = 2, MUSLE

ERFLAG = 4, MUSS

**Scenario Line 3:** USLEK, USLELS, USLEP, AFIELD, IREG, SLP, HL

USLEK: Universal Soil Loss Equation (K) of soil erodibility.

USLELS: Universal Soil Loss Equation (LS) topographic factor.

USLEP: Universal Soil Loss Equation (P) practice factor.

AFIELD: area of field or plot in hectares.

IREG: location of NRCS 24-hour hyetograph.

SLP: land slope (%).

HL: hydraulic length (m).

**Scenario Line 4:** NUSLEC, use\_usleyears

NUSLEC: number of different crops in the simulation.

use\_usleyears logical input to specify the years for each event (T) or use same events every year (F)

**RECORD 5:** DUSLEC, GMUSLEC, GYUSLEC, USLEC, MNGN, CN\_2

DUSLEC day of hydrologic event property change

GMUSLEC month of hydrologic event property change

GYUSLEC year of hydrologic event property change

USLEC: soil loss cover management factors (C value).

MNGN: Manning’s roughness coefficient N; only needed if using velocity method for erosion; enter 0 or some other dummy value if lag method is used (see Record A3).

CN\_2: runoff curve number of antecedent moisture condition 2.

**RECORD 6:** num\_crop\_periods

num\_crop\_periods: total number of cropping periods for entire simulation to be read in Record 7

**RECORD 7:** emd, emm, emy, mad, mam, may, had, ham, hay, root\_depth, max\_caopy\_cover, max\_canopy\_height, max\_canopy\_holdup, foliar\_disposition

Repeat this record up to NCPDS.

emd: integer day of crop emergence.

emm: integer month of crop emergence.

emy: integer year of crop emergence.

mad: integer day of crop maturation.

mam: integer month of crop maturation.

may: integer year of crop maturation.

had: integer day of crop harvest.

ham: integer month of crop harvest.

hay: integer year of crop harvest.

max\_root\_depth: maximum rooting depth of the crop (cm).

max\_caopy\_cover: maximum areal coverage of the canopy (percent).

max\_canopy\_height: maximum canopy height at maturity date (cm)

max\_canopy\_holdup: maximum interception storage of the crop (cm).

foliar\_disposition = condition for disposition of foliar pesticide after harvest. 1 = surface applied, 2 = complete removal, 3 = left alone. Required if CAM=2.

**RECORD 8:** irflag, is\_temperature\_simulated

irflag: irrigation flag. 0 = no irrigation simulated, 1 = year-round, 2 = during crop period only.

is\_temperature\_simulated: soil temperature simulation flag ( Logical) True or False

**RECORD 9:** IRTYP, FLEACH, PCDEPL, max\_irrig, UserSpecifiesDepth, User\_ irrig\_depth

Only if irflag = 1 or 2 (see record 13).

IRTYP: type of irrigation:

1 = flood irrigation,

2 = not used

3 = over canopy,

4 = under canopy sprinkler,

5 = over canopy without runoff generation,

6 = over canopy, user-defined rates, with runoff generation,

7 = over canopy, user-defined rates, without runoff generation.

FLEACH: leaching factor as a fraction of irrigation water depth.

PCDEPL: fraction of available water capacity at which irrigation is applied.

Usually ~0.45 – 0.55; PRZM accepts values between 0.0 and 0.9

max\_irrig: maximum daily irrigation rate (cm day-1).

UserSpecifiesDepth user specifies the depth (TRUE), or irrigate according to root depth (FALSE).

User\_ irrig\_depth user specified irrigation depth (cm) if UserSpecifiesDepth is TRUE, otherwise 0

**RECORD 10:** ALBEDO(12), EMMISS,

Only if is\_temperature\_simulated = True

ALBEDO: monthly values of soil surface albedo (12 values).

EMMISS: reflectivity of soil surface to longwave radiation (fraction).

**RECORD 11:** uWind\_Reference\_Height, Height\_stagnant\_air\_layer\_cm

uWind\_Reference\_Height: height of wind speed measurement above the soil surface (m).

Height\_stagnant\_air\_layer\_cm: thickness of boundary at soil surface (formerly hard coded at 0.5 cm)

**RECORD 12:** BBT(12)

Only if is\_temperature\_simulated = True

BBT: average monthly values of bottom boundary soil temperatures in °C (12 values).

**RECORD 13:**  Q\_10, TBASE

Only if tempflag = 1.

Q\_10: factor for rate increase when temperature increases by 10°C. If Q\_10 is set equal to zero, then PRZM will not simulate degradation change with temperature.

TBASE: temperature during the test of biodegradation.

**RECORD 14:**  NHORIZ

NHORIZ: total number of horizons (minimum of 1).

**RECORD 15:** HORIZN, thickness, Num\_Delx, dispersion\_input, bd\_input, theta\_zero\_input, fc\_input, wp\_input, oc\_input, sand\_input, clay\_input, soil\_temp\_input

One line for each horizon (total lines = NHORIZ).

HORIZN: dummy place holder, horizon number if wanted (used only for input file clarity).

thickness: thickness of the horizon (cm).

num\_delx: number of compartments in the horizon (new PRZM5 parameter).

dispersion\_input: pesticide hydrodynamic solute dispersion coefficient for each NCHEM.

bd\_input: bulk density if BDFLAG=0 or mineral density if BDFLAG=1 (g/ cm3).

theta\_zero\_input: initial soil water content in the horizon (cm3/cm3).

max\_water: max water content potentially held during the time step in the horizon (cm3/cm3).

min\_water: lowest water content in the horizon (cm3/cm3).

oc\_input: organic carbon in the horizon (percent).

sand\_input: sand content in the horizon. Required if tempflag = 1; otherwise set to 0.0 (%).

clay\_input: clay content in the horizon. Required if tempflag = 1; otherwise set to 0.0 (%).

soil\_temp\_input: initial temperature of the horizon (°C). Required only if is\_temperature\_simulated = True; otherwise set to 0.0 or any numeric place holder

**RECORD 16:** runoff\_extract\_depth, runoff\_decline, runoff\_effic

runoff\_extract\_depth soil depth over which runoff interacts with soil (cm).

runoff\_decline exponential decline coefficient of runoff interaction (cm-1).

runoff\_effic fraction of total runoff that interacts with soil.

**RECORD 17:**  erosion\_depth, erosion\_decline, erosion\_effic

erosion\_depth soil depth over which erosion interacts with soil (cm).

erosion\_decline exponential decline coefficient of erosion interaction (cm-1).

erosion\_effic fraction of total erosion that interacts with soil.

## Waterbody/Watershed Input File

**Waterbody Line 1:**

## Meteorological Data File

PRZM5 requires the use of a meteorological file, specified in the PRZM5 input file (Section 2.2) and read by ReadInputs.f90. Information on daily precipitation, pan evaporation, temperature, wind speed, and solar radiation is included in each record of the meteorological file. This version of PRZM5 can read two types of weather files: a newer *wea* format as described in Fry et al. (2016), and the older format *dvf* files. The wea files are simpler and date is comma delimited. The wea files have the following structure:

Month, Day, Year, Precipitation (cm/d), ET0 (cm/d), temperature (°C), wind speed (cm/s), solar radiation (La/d). Typical consecutive lines in a wea file could look like this:

01,01, 2008, 2.32, 0.10, 15, 3.55, 86.14

01,02 ,2008, 0.00, 0.20, 16, 3.55, 152.77

Each line represents one day of weather. The lines must be ordered in chronological order with no missing days. The PRZM5 simulation will start on the first day in the weather file and end on the last day in the weather file.

# Parameter Estimation

Parameters used in the PRZM5 code are listed for reference.

**AFIELD** - The erosion area or plot size in hectares (ha). Area only impacts the erosion routine. Elsewhere all output is on a per-area basis and area has no impact on results.

**ALBEDO** - Soil surface albedo. To simulate soil temperatures, ALBEDO values must be specified for each month. As the surface conditions change, the ALBEDO values change accordingly. Values for some natural surface conditions are provided in **Table 3.1**.

**max\_root\_depth** – Effective root depth, not necessarily the maximum depth. PRZM requires this parameter in centimeters to estimate the measurement of root depth from the land surface for purposes of estimating evapotranspiration. For ranges on specific root depths, consult a current version of USDA Usual Planting and Harvesting Dates or the local Cooperative Extension Service.

Table 3.1 Albedo Factors of Natural Surfaces for Solar Radiation (Brutsaert, 1982; van Wijk, 1963)

| Surface | Reflectivity |
| --- | --- |
| Fresh Dry Snow | 0.80-0.90 |
| Clean, Stable Snow Cover | 0.60-0.75 |
| Old and Dirty Snow Cover | 0.30-0.65 |
| Dry Salt Cover | 0.50 |
| Lime | 0.45 |
| White Sand, Lime | 0.30-0.40 |
| Quartz Sand | 0.35 |
| Granite | 0.15 |
| Dark Clay, Wet | 0.02-0.08 |
| Dark Clay, Dry | 0.16 |
| Sand, Wet | 0.09 |
| Sand, Dry | 0.18 |
| Sand, Yellow | 0.35 |
| Bare Fields | 0.12-0.25 |
| Wet Plowed Field | 0.05-0.14 |
| Newly Plowed Field | 0.17 |
| Grass, Green | 0.16-0.27 |
| Grass, Dried | 0.16-0.19 |
| Grass, High Density | 0.18-0.20 |
| Prairie, Wet | 0.22 |
| Prairie, Dry | 0.32 |
| Stubble Fields | 0.15-0.17 |
| Grain Crops | 0.10-0.25 |
| Alfalfa, Lettuce, Beets, Potatoes | 0.18-0.32 |
| Coniferous Forest | 0.10-0.15 |
| Deciduous Forest | 0.15-0.25 |
| Forest with Melting Snow | 0.20-0.30 |
| Yellow Leaves (fall) | 0.33-0.36 |
| Desert, Dry Soils | 0.20-0.35 |
| Desert, Midday | 0.15 |
| Desert, Low Solar Altitude | 0.35 |

**min\_evap\_depth** - This value represents the minimum depth that is accessible for evapotranspiration for soil evaporation. The evapotranspiration depth specified by min\_evap\_depth applies when root depth is less than min\_evap\_depth. Otherwise, the evapotranspiration depth is equal to the root depth. Values of min\_evap\_depth are around 5 cm.

**APPEFF** - Application efficiency of pesticide application (TAPP). **application\_rate** will be multiplied by APPEFF to calculate the effective rate of application.

**application\_rate** - Target application rate for pesticide(s). For each pesticide and each application date, the amount of pesticide is entered in kg active ingredient per ha. Typical rates are included on the product's registration label. Actual rates used in the model are reduced by the application efficiency (APPEFF).

**bd\_input** - Soil bulk density. This value is required in the basic chemical transport equations of PRZM5 and is also used to estimate moisture saturation values. Values can be found in the USDA Soil Data Mart.

**BBT** - Bottom boundary soil temperatures. BBT values for each month must be specified. The BBT for shallow core depths will vary significantly with time throughout the year. For deep cores, BBT will be relatively constant. The average temperature of shallow groundwater is displayed in **Figure 3.2** (Available at: <http://www.epa.gov/athens/learn2model/part-two/onsite/tempmap.html>).



Figure 3.2 Average shallow groundwater temperatures in the United States (from <http://www.epa.gov/athens/learn2model/part-two/onsite/tempmap.html> derived from Collins [1925]).

**CAM** - Chemical application model flag. This flag specifies how the pesticide is applied to soil or foliage. The CAM schemes are:

CAM = 1, Below crop application (no crop interception). The distribution of chemical below surface is Chemical is inversely proportional to depth to 4 cm, where the distribution declines to zero.

CAM = 2, above crop application in which the crop foliage captures pesticide in proportion to its areal coverage. Pesticide that directly hits ground follows CAM 1 incorporation.

CAM = 3 which is used for uniform incorporation into the soil to a user-specified depth.

CAM = 4, which incorporates the chemical directly to the user-specified depth.

CAM = 5, which approximates T-Band application to a user-defined incorporation depth. Variable DRFT should be used to define the fraction of chemical to be applied in the top 2 cm. The remainder of the chemical is uniformly incorporated between 2 cm and the user-defined depth.

CAM = 6 results in linearly decreasing incorporation to a user defined depth.

CAM = 7 results in linearly increasing incorporation to a user defined depth.

CAM =100 Similar to CAM 2 but the user can control the ground distribution depth with the DEPI parameter.

**max\_canopy\_holdup** - The maximum interception storage of the crop (cm). This parameter estimates the amount of rainfall that is intercepted by a fully developed plant canopy and retained on the plant surface.

The PRZM 3.12 manual stated that z range of 0.1 to 0.3 for a dense crop canopy was reported by Knisel (1980); however, reference to these values could not be found in Knisel (1980). The PRZM 3.12 manual gives a table of values for CINTCP, but the source is unknown. Alternatively, CINTCP can be calculated from a simple crop interception relationship cited by several (Dickinson, 1984; Brisson et al. 1998; Kozak et al, 2007; Giante et al., 2009), which assumes a 0.02 cm covering of the canopy according to:

(3-1)

Where LAI is the leaf area index.

**CN** - Runoff curve numbers of antecedent runoff condition II, as defined by the Natural Resources Conservation Service NEH-4 (NRCS, 2003).

**max\_canopy\_cover -** This is the maximum areal crop coverage. PRZM estimates crop ground cover to a maximum value, COVMAX, by linear interpolation between emergence and maturity dates. As a crop grows, its ground cover increases and captures proportionally more pesticide from above canopy applications. Similarly, rainfall storage capacity increases. For most crops, the maximum coverage will be on the order of 80% to 100%.

**DAIR** - Vapor phase diffusion coefficient. When Henry's Law constant (HENRYK) is greater than zero, vapor phase diffusion is used to calculate equilibrium between vapor and solution phases. Jury et al. (1983b) concluded that the diffusion coefficient will not show significant variations for different pesticides at a given temperature; they recommended using a constant value of 0.43 m2/day (4300 cm2/day) for all pesticides.

**DGRATE** - Vapor phase degradation rate constant(s). Pesticides are degraded by different mechanisms, and at different rates, depending upon whether they are in vapor, liquid, or sorbed phase (Streile, 1984).

**DISP** - Dispersion coefficient of the pesticide(s). For root zone transport, this will be dominated by hydrodynamic dispersion (not by molecular diffusion), and thus PRZM5 no longer groups this parameter with chemical properties. Note that in most practical cases, the temporal and spatial discretization scheme in PRZM5 will already create dispersion of the magnitudes observed in the field (Young and Carleton, 2012), and thus DISP can be set to zero.

**DKW112, DKW113, DKW123, DKS112, DKS113, DKS123** - Molar transformation ratio from a parent chemical (1 or 2) to a degradate chemical (2 and/or 3) for dissolved (DKW) and adsorbed (DKS) phase residues.

**DRIFT**

**DRIFT FACTOR**

**DSRATE** - Sorbed phase degradation rate constant for parent and degradates.

**DWRATE** - Solution phase degradation rate constant for parent and degradates.

**EMMISS** - Infrared Emissivity (unitless). Most natural surfaces have an infrared emissivity between 0.9 and 0.99. Values for all natural surfaces are not well known but are usually close to unity. Specific values of EMMISS for some natural surfaces are given in **Table 3.2**.

Table 3.2 Emissivity Values for Natural Surfaces at Normal Temperatures (van Wijk, 1963; Brutsaert, 1982).

| Natural Surface | Emissivity |
| --- | --- |
| Leaves | 0.94-0.98 |
| Water | 0.95 |
| Snow (old) | 0.97 |
| Snow (fresh) | 0.99 |

**ENPY** - Enthalpy of vaporization. This parameter is used in the temperature correction equation for Henry's Law constant. See USEPA EPISUITE for estimates.

**ERFLAG** - Flag to designate which erosion routine should be used.

**erosion\_depth–** the maximum soil depth that erosion can access (De in equation 4-86)

**erosion\_decline–** specifies the shape of the erosion extraction profile (Ke in equation 4-86)

**erosion\_effic–** the fraction of erosion that interacts with the soil (Fe in equation 4-86)

**FLEACH** - The leaching factor as a fraction of irrigation water depth. This factor is used to specify the amount of water added by irrigation to leach salts from saline soil and is defined as a fraction of the amount of water required to meet the soil water deficit. For instance, a value of 0.25 indicates that 25% extra water is added to meet the soil water deficit.

**height\_stagnant\_air\_layer\_cm:** Specifies the boundary condition for the soil surface-to-air interface.

**HENRYK** - Henry's constant is a ratio of the pesticide equilibrium concentration in the air to its concentration in water (unitless).

**canopy\_height** - Maximum canopy height of the crop at maturation (cm). Canopy height increases during crop growth resulting in pesticide flux changes in the plant compartment. Users should have site-specific information on canopy\_height since it varies with climate, crop species, and environmental conditions. General ranges for different crops are listed in **Table 3.3**. This parameter is used in the volatilization routine.

Table 3.3 Maximum Canopy Height at Crop Maturation

| Crop | canopy\_height (cm) | Reference |
| --- | --- | --- |
| Barley | 20 - 50 | Szeicz et al. (1969) |
| Grain Sorghum | 90 - 110 | Smith et al. (1978) |
| Alfalfa | 10 - 50 | Szeicz et al. (1969) |
| Corn | 80 - 300 | Szeicz et al. (1969) |
| Potatoes | 30 - 60 | Szeicz et al. (1969) |
| Soybeans | 90 - 110 | Smith et al. (1978) |
| Sugarcane | 100 - 400 | Szeicz et al. (1969) |

**IREG** - NRCS rainfall distribution region. For time period May 1 to September 15, IREG is used in the time of concentration calculation of peak flow. For the rest of the year, IREG = 2. See **Figure 3.3** for appropriate region.



Figure 3.3 Approximate geographic boundaries for NRCS rainfall distributions, from TR-55 (NRCS, 1986).

**IRFLAG** - Flag to simulate irrigation. If irrigation is desired, the user has a choice of applying water for the whole year or during a cropping period whenever a specified deficit exists.

**IRTYP** - Specifies the type of irrigation used.

**foliar\_disposition** - Flag indicating the disposition of pesticide remaining on foliage after harvest. This flag only applies if CAM = 2. If foliar\_disposition = 1, the pesticide remaining on foliage is converted to a surface application and distributed in the soil uniformly to 4 cm. If foliar\_disposition = 2, the pesticide remaining on foliage is completely removed after harvest. If foliar\_disposition = 3, the pesticide remaining on foliage is retained as surface residue and continues to undergo decay and washoff.

i**s\_temperature\_simulated** - Flag for soil temperature simulation. This flag allows a user to specify soil temperatures (BBT) for shallow core depths. For deep cores (CORED), temperatures will remain relatively constant.

**KD** - Pesticide soil-water distribution coefficient. The user can enter KD directly if KDFLAG = 0 (see PCMC and SOL) or allow the model to calculate KD automatically (KDFLAG = 1).

**num\_applications** - Number of pesticide applications. This is the total number of application dates specified during the simulation.

**NCHEM** - Number of chemicals in the simulation. PRZM5 allows up to three chemicals to be specified. Using more than one chemical (i.e., NCHEM=3) indicates either a parent-degradate relationship or multiple separate chemicals.

**num\_crop\_periods** - Number of cropping periods. This value is entered as a sum of all cropping dates from the beginning simulation date to the ending simulation date.

**NHORIZ** - Total number of horizons. PRZM5 allows the user to specify how many horizons are to be simulated within the core depth (CORED). The horizon should serve as a distinct morphologic zone, generally described by layers (i.e., surface, subsurface, substratum) according to soil pedon descriptions or soil interpretation records, if available.

**NPLOTS** - Number of time series outputs. PRZM5 can report several output variables (PLNAME) to a time series file. NPLOTS specifies how many are written in a single simulation.

**num\_delx** - The number of compartments in a soil layer (or horizon). The thickness of each compartment within a horizon is calculated as THKNS/Num\_delx. This compartment thickness affects spatial resolution and numerical dispersion. As num\_delx increases, the compartment thickness decreases, allowing a more precise placement of pesticide in the soil horizon and a higher resolution of output. The compartment thickness will also directly affect the dispersion of the pesticide in the soil profile. The dispersivity coefficient will be effectively

(3-2)

where α= dispersivity (cm)

Δx = compartment thickness (cm)

T = dispersivity induced by the temporal discretization (cm)

The value of T is not under the user’s control and is a function of the rainfall that occurs during a time step (1 day). T is on the order of 2 cm for the typical U.S. rainfall in agricultural areas (Young and Carleton, 2012).

Note: Num\_delx is a new PRZM5 parameter. In previous PRZM versions, the number of compartments was indirectly determined by a specified compartment thickness and horizon thickness. When the compartment thickness used by the program was not equal to the input compartment thickness, unexpected program behavior resulted. Using Num\_delx, (an integer) fixes this problem.

**OC** - Percent of soil organic carbon (OC). OC is conventionally related to soil organic matter as %OC = %OM/1.724.

**PCDEPL** - Fraction of available water capacity where irrigation is triggered. The moisture level where irrigation is required is defined by the user as a fraction of the available water capacity. This fraction will depend upon the soil moisture holding characteristics, the type of crop planted, and regional agricultural practices. In general, PCDEPL should range between 0.0 and 0.6, where a value of 0.0 indicates that irrigation begins when soil moisture drops to wilting point, and 0.6 indicates the more conservative practice of irrigating at 60 percent of the available water capacity.

**PFAC** - The pan factor is a dimensionless number used to convert daily pan evaporation to daily potential evapotranspiration. Pan factors generally range from 0.60 to 0.80. See **Figure 3.4** for specific regions of the United States.

Pan evaporation correction factors

Figure 3.4 Pan evaporation correction factors (from U.S. Weather Bureau).

**plant\_washoff\_coeff** - Foliar washoff coefficient in units of fraction removed per cm of rainfall. Washoff from plant surfaces is modeled with consideration for rainfall, foliar mass of pesticide, and extraction ability. Values are variable and depend upon the crop, pesticide properties, and application method. Smith and Carsel (1984) suggest that a value of 0.10 is suitable for most pesticides.

**plant\_pesticide\_degrade\_rate** - Foliage pesticide first-order decay rate. This parameter acts exactly like plant\_volatilization\_rate except that a degradate may be produced with plant\_pesticide\_degrade\_rate**.**

**PLNAME** - Name of output variable. When creating time series output, PLNAME specifies the variable in **Table 3.5** for which that output data are written.

| **Table 3.5 Variable Designations for Output Files** | | | |
| --- | --- | --- | --- |
| **PLNAME** | **Description** | **Units** | **Compartment Required \*** |
| INTS | Water held on crop canopy | cm | -- |
| PRCP | Precipitation | cm | -- |
| TETD | Actual evapotranspiration from entire profile | cm | -- |
| IRRG | Irrigation | cm | -- |
| SWTR | Soil water in compartment | cm | ARG, ARG2 |
| THET | Soil Water Fraction (Multicompartment requests are not useful for compartments of unequal size) | cm/cm | ARG, ARG2 |
| SLET | Evaporation from soil compartment | cm | ARG, ARG2 |
| SNOP | Snowpack | cm | -- |
| THRF | Water flow through canopy | cm | -- |
| SNOF | Snowfall | cm | -- |
| CEVP | Canopy evaporation | cm | -- |
| TETD | Evapotranspiration | cm | -- |
| RUNF | Runoff | cm | -- |
| ESLS | Eroded solids | 1000 kg | -- |
| INFL | Vertical water flow at a given depth. Use only with modes TSER or TCUM (one compartment only to make sense ) | cm | ARG |
| STMP | Soil Temperature | °C | ARG, ARG2 |
| CHGT | Crop height | cm | no |
|  |  |  |  |
| RFLX | Pesticide in runoff | g/cm2 | -- |
| EFLX | Pesticide in erosion | g/cm2 | -- |
| FPVL | Foliar pesticide volatilization flux | g/cm2 | -- |
| VFLX | Pesticide volatilized from surface | g/cm2 | -- |
| COFX | Pesticide lost from bottom boundary | g/cm2 | -- |
| DKFX | Pesticide decayed in compartment | g/cm2 | ARG, ARG2 |
| UFLX | Pesticide uptake into plant from compartment | g/cm2 | ARG, ARG2 |
| WFLX | Pesticide washoff from canopy | g/cm2 | -- |
| TUPX | Total pesticide uptake by plant | g/cm2 | -- |
| TDKF | Pesticide decay in entire soil profile |  |  |
| MASS | Pesticide mass in equilibrium region for compartment | g/cm2 | ARG, ARG2 |
| MAS2 | Pesticide mass in nonequilibrium region for compartment | g/cm2 | ARG, ARG2 |
| DCON | Pesticide concentration in pore water | mg/l | ARG, ARG2 |
| ACON | Pesticide concentration sorbed to soil | mg/kg | -- |
| GCON | Pesticide concentration in gas | mg/l | -- |
| TCON | Total pesticide concentration in soil | mg/kg | -- |
| TPST | Total pesticide in compartment | g/cm2 | ARG, ARG2 |
| FPST | Pesticide on foliage | g/cm2 | -- |
| SPST | Pesticide in soil water | g/cm2 | ARG, ARG2 |
| TPAP | Total Pesticide application | g/cm2 | -- |
| FPDL | Foliar Pesticide degradation | g/cm2 | -- |

\*For single-compartment output which requires only one argument (ARG), also enter the value for ARG into the ARG2 field to enable proper reading of the inputs.

**plant\_volatilization\_rate** - Foliage pesticide first-order volatilization rate (or any other form of degradation that does not produce a degradate. This process acts in parallel with plant\_pesticide\_degrade\_rate.

**foliar\_formation\_ratio\_12, foliar\_formation\_ratio\_23** – Ratio of formation of degradate1 to parent and degradate 2 to degradate 1, respectively

**Q\_10** - Factor for rate increase when temperature increases by 10°C. Set to 2 for doubling of microbial degradation rate every 10 degrees.

**max\_irrig** - Maximum sprinkler application rate. max\_irrig is u is defined as a depth (cm) of water delivered per day.

**runoff\_extract\_depth –** the maximum soil depth that runoff can access (*Dr* in equation 4-82)

**runoff\_decline –** specifies the shape of the runoff extraction profile (*Kr* in equation 4-82)

**runoff\_effic –** the fraction of runoff that interacts with the soil (*Fr* in equation 4-82)

**SFAC** - The snowmelt factor is used to calculate snowmelt rates in relation to temperature. PRZM5 considers snow to be any precipitation that falls when the air temperature is below 0°C. In areas where climatology prevents snowfall, SFAC should be set to 0.0. Involved snow melt models exist but for the ways in which PRZM5 is typically used a single value of 0.274 should be adequate (USDA, 2004).

**SLP** - Slope of hydraulic flow path.

**SPT** - Initial soil temperature profile. To simulate the soil temperature profile, initial SPT values for each soil horizon must be specified. For quicker stabilization, temperatures in the lower profile should be similar to the bottom boundary temperature, and temperatures in the upper profile should correspond to the average air temperature at the simulation start dates.

**Tband\_top** – For T-Band applications (CAM = 7), this value represents the fraction of the chemical application incorporated into the top 2 cm.

**max\_water, min\_water** – The maximum and minimum water levels for the tipping bucket concept of water transport. Suitable surrogates could be field capacity and wilting point or the water capacity at -33 and -1500 KPa, respectively, as found in many soil data bases. If data are not available, several empirical relations can be used, such as those of Rawls (1983) (See PRZM3 Manual).

**theta\_zero\_input** - Initial water content of the soil. This value provides the model with a starting calculation for moisture. This parameter will have little impact for simulations that last for more than a couple days. Use the same value used for max\_water.

**thickness** - Thickness of the horizon. This value is the depth (cm) of the horizon specified (HORIZN) in relation to core depth (CORED).

**UPTKF** - Plant uptake efficiency factor for transpiration stream. The product of UPTKF and the soil pore water concentration represents the effective concentration of water that would flow into the plant at the evapotranspiration flow rate. A value of 1.0 for UPTKF results in an estimated uptake equal to the product of the transpiration rate and the dissolved-phase soil concentration. UPTKF is somewhat analogous to the short-term (24 to 48 hr) transpiration stream concentration factor (TSCF) of Briggs et al. (1982). Briggs et al (1982) proposed an empirical relationship as follows:

UPTKF ≈ TSCF = 0.784 exp [-(log *Kow* - 1.78)2 / 2.44] (3-3)

Briggs’ laboratory results for UPTKF ranged from 0.11 to 0.94 for the 17 of the 18 pesticides that were tested. One problem with the UPTKF concept in PRZM is that the effective uptake can continue indefinitely, regardless of how much pesticide is implied to be held up in the plant. Another consideration is that PRZM does not distinguish evaporation from transpiration, so all plant uptake is based on the evapotranspiration rate. Caution should be used when implementing UPTKF for simulations that occur for more than just a few days and where evaporation is a sizable portion of the evapotranspiration rate.

**USLEC** - The universal soil loss cover management factor (C value). Values for USLEC are dimensionless and range from 0.001 (well managed) to 1.0 (fallow or tilled condition). Up to 32 values may be entered for the year dependent on crop growth and tillage operations. Specific values can be calculated via Wischmeier and Smith (1978). Generalized values are provided in the PRZM 3.12 Manual.

**USLEK** - The universal soil loss equation (K) of soil erodibility. This is a soil-specific parameter developed by the USDA. Specific values can be obtained from the USDA Soil Data Mart.

**USLELS** - The universal soil loss equation (LS) topographic factor. This is a slope length and steepness parameter developed by the USDA. The value is dimensionless and can be estimated from Wischmeier and Smith (1978).

**USLEP** - The universal soil loss equation (P) practice factor. This value is developed by the USDA to describe conservative agricultural practices. Values are dimensionless and range from 0.10 (extensive practices) to 1.0 (no supporting practices). Specific values can be estimated from Wischmeier and Smith (1978) and Stewart (1975).

**uWind\_Reference\_Height** - Height of wind speed measuring instrument. The wind speed anemometer is usually fixed at 10 meters (30 feet) above the ground surface. This height may differ at some weather stations such as at a class A station, where the anemometer may be attached to the evaporation pan. The correct value can be obtained from the meteorological data reports for the station whose data are in the simulation.

# PRZM5 Theory

## Overview

PRZM5 is a 1-dimensional hydrology, heat and solute transport model, developed primarily for agricultural pesticide simulations. The hydrologic component for calculating runoff and erosion is based on the NRCS curve number method and the Universal Soil Loss Equation, respectively. Water balances are maintained with consideration for runoff, evapotranspiration, irrigation, and precipitation. Daily precipitation, temperature, wind speed, and pan evaporation are supplied as model inputs. Vertical water movement is simulated by a capacity model (or tipping bucket) concept in which vertical water movement is always downward and occurs when a soil compartment is filled to a maximum capacity.

Dissolved, adsorbed, and vapor-phase concentrations of pesticide in the soil are calculated by considering the processes of runoff, erosion, degradation, volatilization, foliar wash off, removal by plant uptake, leaching, dispersion, and sorption. The vertical transport is solved by a finite difference solution. The time step is daily.

## Description of the Algorithms

The current processes simulated by PRZM5 are:

* Crop Growth
* Irrigation
* Precipitation & Snowmelt
* Runoff
* Canopy Water Interception
* Evaporation
* Leaching
* Erosion
* Soil Temperature
* Chemical Application & Foliar Washoff
* Chemical Runoff & Vertical Transport in Soil
* Chemical Volatilization

## Crop Growth

Crop size is assumed to increase proportionally with time from its emergence date to its maturity date. Crop size in PRZM5 refers to the canopy coverage and root depth. PRZM5 inputs are the maximum fraction of areal coverage of the canopy and the maximum root depth. The maximum root depth and the maximum canopy coverage occur at the date of crop maturity (as specified by the user). Root depth and canopy coverage remain at maximum until harvest date (as specified by the user), at which time canopy coverage and root depth are reset to zero. This concept is depicted in **Figure 4.1**. Note that *harvest* does not necessarily refer to harvest of the crop, and special consideration should be given to crops that do not lose their canopy at harvest (e.g., apple trees whose PRZM-defined "harvest" does not refer to the time of fruit harvest, but instead refers to the time of leaf fall).



Figure 4.1 Depiction of the crop growth routine in PRZM5. Crop size increases linearly from emergence date to maturity date, where it remains at maximum size until the harvest date.

## Irrigation

There are two types of irrigation available—over-canopy and under-canopy irrigation. Over-canopy irrigation is applied from above the canopy and may result in pesticide washoff if pesticide foliar application is used. Under-canopy irrigation is applied to the ground and will not wash off foliar pesticide.

Irrigation is controlled automatically, based on the soil moisture deficit and the precipitation in the meteorological file. PRZM5 and previous versions of PRZM do not allow irrigation to occur on a day with precipitation. The moisture deficit is calculated for the soil profile from the surface to a relevant depth. In PRZM5, this relevant depth can be set to the root depth or to a user-specified depth. The soil moisture deficit is calculated as follows:

(4-1)

D = total soil moisture deficit for the relevant soil depth (cm)

N = number of compartments for the relevant soil depth

θmax,i = max capacity for compartment *i*

θi = soil moisture in compartment *i*

Δzi = thickness of compartment *i*

PRZM5 has four irrigation options:

* Type 1: This type of irrigation applies just enough water to satisfy the soil moisture deficit. The water is applied directly into the top surface layer and will not generate runoff. There is no maximum limit to the amount of water that can be applied. (Previous versions of PRZM called this “flood irrigation”.)
* Type 3: This type of irrigation applies water above the canopy. Its effect on hydrology and pesticide transport is identical to precipitation. The model attempts to satisfy the soil moisture deficit, the canopy holdup, and any additional leaching requirements, but the amount is limited by a user input (**max\_irrig**).
* Type 4: This type of irrigation applies water under the canopy. Its effect on hydrology is identical to precipitation, except that the crop canopy has no impact on the water accounting. The model attempts to satisfy the soil moisture deficit and any additional leaching requirements, but the amount is limited by a user input (**max\_irrig**).
* Type 6: This type of irrigation applies water above the canopy and applies a user-specified amount (**max\_irrig**) that is not based on the soil moisture deficit. Its effect on hydrology and pesticide transport is identical to precipitation.

## Precipitation & Snowmelt

Incoming precipitation is first partitioned between snow and rain, depending upon temperature. Air temperatures below 0°C produce snow and may result in the accumulation of a snowpack. Snow accumulation melts when temperatures are above 0°C. PRZM uses the degree day method to calculate the amount of snow melt as a function of temperature (USDA, 2004). The daily rate of snowmelt is estimated by the following:

(4-2)

Where M = melted snow

Cm = snowmelt factor (cm/°C/day)

T = temperature (°C)

SP = snow accumulation (cm)

## Runoff

The runoff calculation partitions the precipitation between infiltrating water and surface runoff according to the NRCS Curve Number (CN) Method (NRCS, 2003). The CN method estimates runoff from precipitation as follows:

(4-3)

Where Q = runoff (cm)

P = precipitation (cm)

S = potential maximum retention (cm)

S is related to the soil type, crop cover, and management practices and is calculated by tabulated values for CN as follows:

(4-4)

In this implementation of the CN method, rain, irrigation water, and snowmelt are treated as having the same effect as precipitation. Thus, precipitation in the above equations is the sum of rain, snowmelt, and irrigation. The CN used for daily runoff calculations is adjusted on a daily basis according to the soil moisture.

Young and Carleton (2006) documented the method used in PRZM 3.12 and found it disagreed with the methods described in previous PRZM manuals. The subsequent discussion here clarifies the way that PRZM5 (as well as previous PRZM versions) calculate the curve number adjustments. In PRZM, the user enters the CN, which represents average antecedent conditions (CNII). PRZM then determines the associated low and high CNs (CNI and CNIII, respectively) from CN tables given in NRCS (2003). PRZM then calculates the average soil moisture in the top 10 cm of soil for each day, and calculates an adjusted CN based on this soil moisture.

As described in Young and Carleton (2006), PRZM uses the following definitions to make these calculations:

* CNI occurs when the average soil moisture content is zero in the top 10 cm of soil.
* CNII occurs when the soil moisture content in the top 10 cm of soil is equal to a representative halfway point between field capacity and wilting point.
* CNIII occurs when soil moisture content rises to a value equal to the sum of field capacity plus wilting point.

When the PRZM-calculated soil moisture falls between these values, PRZM uses linear interpolation to arrive at a CN. The scheme is depicted in **Figure 4.2**. As shown, PRZM CNs are somewhat restricted in variability and will never reach the CNI or CNIII values.

PRZM scheme for CN adjustments.  Solid lines reflect the range of CN adjustments possible.  WP is wilting point; FC is field capacity.  Dotted lines are used in the PRZM calculations, but these CNs may not be attained in PRZM.

Figure 4.2 PRZM scheme for CN adjustments. Solid lines reflect the range of CN adjustments possible. WP is wilting point; FC is field capacity. Dotted lines are used in the PRZM calculations, but these CNs may not be attained in PRZM.

**New Note for PRZM5:** Unlike previous versions of PRZM, PRZM5 is more in line with the NRCS (2003) method. PRZM5 does not adjust the initial abstraction (the value 0.2S) or use throughfall instead of precipitation in the runoff equation. These events are already accounted for implicitly in the NRCS CN method and making such adjustments would introduce false precision. Thus, PRZM5 maintains the original intent and limitations of the NRCS CN method.

## Canopy Water Interception

After the runoff is calculated, excess water (the difference between rain/overhead irrigation and runoff) is used to satisfy the canopy capacity. Although in reality canopy capture would occur before runoff occurs, the CN method already has canopy capture implicitly included, and thus runoff is predetermined for a given precipitation event. Canopy water holdup is given by:

(4-5)

Where Ct/2 = canopy water at end of precipitation event, but prior to evaporation (cm)

Ct-1 = canopy water from previous time step (cm)

Cmax = maximum water that the canopy can hold (cm)

P = precipitation (cm)

I = overhead irrigation (cm)

Q = runoff (cm)

## Evaporation

Daily potential evapotranspiration is calculated from the daily pan evaporation in the meteorological file and the pan coefficient. This product, known as the *free water surface evaporation*, is a good representation of potential evapotranspiration from an adequately watered natural surface like soil or vegetation (NOAA, 1982).

(4-6)

Where ETp = potential evapotranspiration (cm)

Kp = pan coefficient (unitless)

Ep = pan evaporation (cm)

The potential evapotranspiration is used first on the plant canopy and then on the soil. The canopy water content at the end of the time step (Ct) is equal to the amount of water added from equation (4-5) minus the potential evaporation.

(4-7)

If the potential evaporation has not been satisfied by the canopy, then the remaining potential evaporation is applied to the soil. The evapotranspiration is satisfied preferentially towards the surface in a proportional manner, and it is also satisfied preferentially according to available water in a proportional manner. This distribution of uptake preferences roughly mimics the natural processes of plant uptake and root structure (USDA, 1991). The function of evapotranspiration with depth and available water can be described by the following equations:

 (4-8)

where U is defined by a depth adjustment and a moisture adjustment:

 (4-9)

where *d*maxi is the depth that evapotranspiration occurs on day i

*dx* is the depth at position x.

Available water is defined as:

(4-10)

Where W*soil*,(0:*x*) = soil water in the spatial range 0:x

W*wp*,(0:*x*) = wilting point in the spatial range 0:x

As an example, Figure 4.3a shows a moisture distribution hypothetically taken as distributed in a linearly increasing manner to the bottom of the evaporation zone (25 cm). Figure 4.3b shows the contributions of depth and available water (the two parenthetical fractions in equation 4-9). Figure 4.3c shows the combined effect of available water and depth on the distribution that the model will use to extract water for evapotranspiration. Note that previous PRZM 3 documentation seemed to state that evapotranspiration is satisfied by sequentially removing water from the surface downward; however, the real mechanism that PRZM uses is the one presented below:



**Figure 4.3 Evaporation Preference Distribution.** a) A hypothetical moisture distribution with depth. b) The individual distributions of depth and moisture influence on ET reduction. c) The distribution of ET with depth due to the combined effects of depth and availability of water.

The potential ET first consumes any available canopy held water. If the canopy water does not satisfy the demand, then the water in the soil is used to attempt to satisfy the demand. The remaining ET demand is distributed in a linearly decreasing manner through the surface soils to a maximum specified depth and also is proportional to the amount of available water at any location in the evaporation zone, with the distribution through the soil profile described by the following equations:

 (4-11)

where

 (4-12)

and

 (4-13)

where Xet = the maximum depth for evapotranspiration action.

This type of distribution ensures that water is taken from locations where it is easiest to obtain—that is, water is preferentially taken from the depths nearer the surface and preferentially from depths that contain greater amounts of available water. The available water is not permitted to go below zero. As with previous PRZM versions, PRZM5 applies an additional constraint: if the soil moisture is less that 0.6 of the available water, then the available soil ET is reduced in proportion down to WP (where it is zero).

If is less than 0.6, then the evapotranspiration is effectively reduced

 (4-14)

## Leaching

Vertical water movement in PRZM5 is approximated by a capacity model also known as a "tipping bucket" approach. Water content in any soil compartment is determined by continuity. The soil water content at any point in the soil column is first calculated from the amount of infiltrating water from the above layer as:

(4-15)

Where θi,t+1 = the soil water content of *i* at the end of the current time step (cm3/cm3)

θi,t = the soil water content of *i* at the start of the current time step (cm3/cm3)

*vi*= the water velocity entering compartment *i* from the above compartment (cm/day)

ETi = the evapotranspiration at depth *i* (cm/day)

Δt = time step (day)

Δzi = compartment *i* thickness (cm)

If the soil water content exceeds the max capacity, then the excess water is used as flow for the next compartment, and the flow into the next compartment is calculated as:

(4-16)

Where θi,max is the maximum water capacity of layer i

In the case of excess water, the water content for compartment *i* is set to the value for maximum capacity:

(4-17)

This accounting continues throughout the soil column.

The upper boundary condition is defined by the infiltration into the surface. This flux boundary condition is the result of inputs from of precipitation, irrigation, and snowmelt. The velocity of water into the surface (*v0*) is as follows:

(4-18)

Where P = precipitation (other than snow) (cm/day)

I = Irrigation (cm/day)

M = Melted snow (cm/day)

C = Canopy captured water (cm/day)

R = Runoff (cm/day)

## Erosion

PRZM5 retains two of the three erosion options available from previous PRZM versions; although the remaining two options are still somewhat problematic regarding quality assurance. MUSLE (Williams, 1975) has been challenged for not being mathematically sound (Kinnell, 2004), and MUSS lacks any published documentation other than a conference abstract (according to the abstract, MUSS is for “small watersheds”, but the definition of *small* is undefined). The third option, MUST, was eliminated because of lack of use and documentation. Nevertheless, these models have been used for regulatory assessments in the past, and thus MUSLE and MUSS have been retained to maintain continuity with previous assessments. MUSLE and MUSS options are formulated as follows:

MUSLE: (4-19)

MUSS: (4-20)

Where *Xe* = the event soil loss (metric tons per ha day-1)

*Vr* = volume of event (daily) runoff (mm)

*qp* = peak storm runoff (mm/h)

*A* = field size (ha)

*K* = soil erodibility factor (dimensionless)

*LS* = length-slope factor (dimensionless)

*C* = soil cover factor (dimensionless)

*P* = conservation practice factor (dimensionless)

The peak storm runoff value (*qp)* is calculated using the Graphical Peak Discharge Method as described in NRCS (1986).

(4-21)

Where qp = peak storm runoff (ft3/s)

*qu* = unit peak discharge (csm/in)

A = drainage area (mi2)

Q = runoff (in)

*Fp* = pond or swamp adjustment factor

For areas without swamps or ponds, Fp is equal to 1.0 (a value preset within PRZM). The unit peak discharge is calculated by NRCS (1986) as follows:

(4-22)

Where Tc = time of concentration (hr)

C0, C1, C2 = coefficients from NRCS (1986), given in **Table 4.1** below

Rainfall intensity is assumed to occur according to design storm distributions (Type I, IA, II, and III) as given by NRCS (1986). **Figure 3.3** shows a map of the rainfall type distributions in the U.S.

Table 4.1 Coefficients for the equation 4-22 (NRCS, 1986).

| Rainfall Type | Ia/P | C0 | C1 | C2 |
| --- | --- | --- | --- | --- |
| I | 0.10 | 2.30550 | -0.51429 | -0.11750 |
| 0.20 | 2.23537 | -0.50387 | -0.08929 |
| 0.25 | 2.18219 | -0.48488 | -0.06589 |
| 0.30 | 2.10624 | -0.45695 | -0.02835 |
| 0.35 | 2.00303 | -0.40769 | 0.01983 |
| 0.40 | 1.87733 | -0.32274 | 0.05754 |
| 0.45 | 1.76312 | -0.15644 | 0.00453 |
| 0.50 | 1.67889 | -0.06930 | 0.0 |
| IA | 0.10 | 2.03250 | -0.31583 | -0.13748 |
| 0.20 | 1.91978 | -0.28215 | -0.07020 |
| 0.25 | 1.83842 | -0.25543 | -0.02597 |
| 0.30 | 1.72657 | -0.19826 | 0.02633 |
| 0.50 | 1.63417 | -0.0910 | 0.0 |
| II | 0.10 | 2.55323 | -0.61512 | -0.16403 |
| 0.30 | 2.46532 | -0.62257 | -0.11657 |
| 0.35 | 2.41896 | -0.61594 | -0.08820 |
| 0.40 | 2.36409 | -0.59857 | -0.05621 |
| 0.45 | 2.29238 | -0.57005 | -0.02281 |
| 0.50 | 2.20282 | -0.51599 | -0.01259 |
| III | 0.10 | 2.47317 | -0.51848 | -0.17083 |
| 0.30 | 2.39628 | -0.51202 | -0.13245 |
| 0.35 | 2.35477 | -0.49735 | - 0.11985 |
| 0.40 | 2.30726 | -0.46541 | -0.11094 |
| 0.45 | 2.24876 | -0.41314 | -0.11508 |
| 0.50 | 2.17772 | - 0.36803 | -0.09525 |

### Time of Concentration: Velocity Method

The time of concentration (*Tc*) is defined as the time it takes water to flow from the furthest point in the watershed to a point of interest within the watershed; *Tc* is a function of basin shape, topography, and surface cover. *Tc* is calculated by summing the travel time for various flow segments within the watershed (NRCS, 1986). As with previous PRZM versions, PRZM5 is configured to have two flow segments: 1) sheet flow for the first 100 meters, and 2) shallow concentrated flow (unpaved) for the remaining portion of the hydraulic length. Under the sheet flow segment, *Tc* is calculated as:

(4-23)

Where s = slope (ft/ft)

P = precipitation

*N* = Manning’s roughness coefficient for the watershed

*L* = hydraulic flow length (m)

*P* = daily precipitation (cm)

*s* = slope of the hydraulic grade line (land slope, m/m)

*a*, *b* = unit conversion factors

The equation for shallow concentrated flow is derived from Manning’s equation assuming a roughness coefficient, *N,* of 0.05 and a hydraulic radius of 0.2 (Soil Conservation Service, 1986). The average velocities for estimating travel time for shallow concentrated flow are as follows:

Unpaved V = 16.1345s0.5 (4-24)

Paved V = 20.3282 s0.5 (4-25)

Where V= average velocity (ft/s)

s = slope of hydraulic grade line (watercourse slope, ft/ft)

These two equations are based on the solution of Manning’s equation with different assumptions for N (Manning’s roughness coefficient) and r (hydraulic radius, ft). For unpaved areas, N is 0.05 and r is 0.4; for paved areas, N is 0.025 and r is 0.2.

### Time of Concentration: Watershed Lag Method

This method is specified in NEH-4 and is simpler and less parameter intensive than the previous PRZM method. Additionally, the previous method was based on cha5racteristics of urban watersheds as specified by TR-55. Such characteristics may not be representative of agricultural runoff. The Watershed Lag method does not require a Manning’s N value and is more appropriate for use in broad-scale assessments where precise site-specific characterization is less meaningful.

(4-26)

L = hydraulic length (ft)

slp = slope (%)

S = max potential retention

=

cn = curve number

## Soil Temperature

Soil temperature is modeled in PRZM5 to correct for temperature effects on volatilization and degradation. The objective of the soil temperature model is to provide a scientifically sound and usable approach to predict with reasonable accuracy the daily average soil temperatures at the soil surface and (in and below) the root zone, using basic soil physical and thermal properties and daily climatic measurements taken at weather stations. PRZM5 calculates the soil temperature profile using inputs of air temperature, solar radiation, surface albedo, wind velocity, evaporation, soil water content, and soil physical properties, by the methods of de Vries (1963), Hanks et al. (1971), van Bavel and Hillel (1975), and Thibodeaux (1979, 1996).

Several models are available to predict soil temperatures under various soil surface conditions, but there are restrictions to their general use as they either require large databases that are not available or are site specific. Existing soil temperature models form two general groups: (1) process-oriented models, which require detailed information about soil and surface characteristics, initial and boundary conditions, and inputs, and (2) semi- or non-process-oriented models, which often use weather station information and soil temperature information at one depth to develop empirical relationships. The PRZM3 manual further describes the key characteristics of the soil temperature models previously reviewed.

In PRZM5, daily bottom boundary temperature (BBT) and soil surface albedo (ABSOIL) values are first calculated by interpolating between neighboring BBT and albedo monthly values. Surface albedo is estimated from crop canopy albedo (COVER) and soil surface albedo, with an evaporation correction for the canopy albedo.

### Thermal Diffusivity

If the user does not provide the thermal conductivity and heat capacity as inputs, the thermal diffusivity of the soil compartment (quotient of thermal conductivity () and heat capacity per unit volume (C), /C) is estimated using the methods of de Vries (1963), where soil water content changes with time and depth (L). For each soil compartment, PRZM5 estimates the corrected total volume (1-porosity, accounting for volume fractions of sand, clay, and organic carbon), and the volume fraction (Xvol) of each soil constituent from weight percentages (i.e., sand, clay, organic carbon):

(4-26)

(4-27)

(4-28)

(4-29)

Where

Sand = mass fraction of sand

Clay = mass fraction of clay

Xsand = volume fraction of sand

oc = mass fraction of organic carbon

Xoc = volume fraction of organic carbon

Xother = volume fraction of other solid constituents (not sand or oc)

θ = porosity (unitless)

ρb = bulk density [g/cm3]

The water and air content of the soil pores are also estimated. (Note that these volume estimations are redundant calculations since PRZM5 already calculates these values in other parts of the program; future versions should reduce these redundancies.) The water content Xwater is reset to the wilting point when the water content of the soil is less than wilting point. The water content is set to to the porosity when the initial water content is greater than porosity. The air cvolume fraction (XAir) is then calculated as the difference between the porosity () and the water content.

When the water content is greater than the field capacity of the soil, an additional parameter (G), the depolarization factor of an ellipsoid, is estimated as follows:

G = [0.333 - (4-30)

The apparent thermal conductivity of the air-filled pores (ALAMDA, a), accounting for moisture movement, is calculated as the sum of the apparent thermal conductivities due to normal heat conduction (AIRLMD) and vapor movement (VAPLMD).

a = AIRLMD + VAPLMD (4-31)

Otherwise, if the water content is less than the maximum capacity, G and a are calculated as:

G = 0.013 + (4-32)

a = AIRLMD + (4-33)

In the simplest case, PRZM5 calculates the soil thermal conductivity (THCOND,  [cal / cm d ◦C]) based on the theory developed by H.C. Burger (1915), as discussed in de Vries (1963):

(4-34)

Where k1 represents the ratio of the average (by space) temperature gradient in the granules to the corresponding gradient in the medium:

(4-35)

k1 is calculated in PRZM5, assuming that the axes (a, b, c) of ellipsoidal granules are randomly oriented:

(4-36)

Volumetric heat capacity per unit volume (VHTCAP) of the soil layer (cal / cm3 ◦C) is then calculated as:

VHTCAP = 0.46 ( Xsand + Xoc ) + 0.60 Xother) + Xwater (4-37)

Where 0.46 cal/g ◦C represents the average specific heat for mineral soils at 10◦C, and 0.60 cal/g ◦C is the average specific heat for organic soils.

The diffusion coefficient (or diffusivity, DIFFCO, in cm2/d) is finally calculated as:

DIFFCO = (L) / VHTCAP(L) (4-38)

### Upper Boundary Temperature

In PRZM5, the upper boundary temperature is estimated using an energy balance at the air-soil interface. From Thibodeaux (1996), the air density (AIRDEN), air (g/cm3) is first calculated as:

air = (-0.0042 T + 1.292) \* 1e -3 (4-39)

Then the heat transfer coefficient (HTC) at the air-surface interface (cm/d) is estimated as:

HTC = (4-40)

Where vonKarman = von Karman constant used in boundary layer meteorology (0.39 or 0.40)

WIND = Wind speed (cm/d)

WIND Ref Height = Wind reference height (m)

D = Zero displacement height (m)

Zo = Roughness length (m)

The Energy Balance equation at the air-soil interface is summarized as:

q13 = qlw + qe + qc + qs (4-41)

Where q13 = Sensible heat flux into the soil column (cal/cm2 d)

qlw = Net longwave radiation flux (cal/cm2 d)

qe = Evaporative (water) heat flux(cal/cm2 d)

qc = Sensible heat flux (conduction) between surface and air (cal/cm2 d)

qs = Shortwave solar radiation (cal/cm2 d)

These heat fluxes are calculated individually in PRZM5 as:

qlw,atm = EMISS \* 0.936e-5 \* T2 \* 11.7e-8 (4-42)

qlw,soil = EMISS \* 11.7e-8 (4-43)

qe = 580.0 \* EVAP \* 1.0 (4-44)

qc = air \* 0.2402 \* HTC (4-45)

qs = (1- ABSOIL) \* SOLRAD (4-46)

The 4th order equation for upper boundary temperature (UBT) is then solved using the Newton-Raphson method, in terms of soil surface temperature using the calculated heat fluxes. The soil temperature profile is then calculated, given the upper boundary, bottom boundary, and initial temperatures.

### Temperature Dependent Degradation

In PRZM5, there is a correction for temperature-dependent degradation based on the *Q10* equation (similar to Arrhenius equation):

(4-47)

Where *FT* = correction factor for biodegradation rate based on the actual temperature

Q\_10= factor for rate increase when temperature increases by 10◦C

*T* = actual soil temperature

*TBASE* = temperature during the test of biodegradation

## Chemical Application & Foliar Washoff

PRZM5 includes the following application methods:

* Below Crop (or Ground) – Pesticide is distributed to 4 cm, linearly decreasing with depth.
* Above Crop (Foliar) - Linear application to foliage based on crop canopy, which varies during the growing season. Any non-intercepted pesticide is treated as a below crop application (see above)
* Uniform below crop - Uniform incorporation into the soil to a user-specified depth.
* T-Band - Recommended for T-Band granular application. User defines the fraction of chemical to be applied in the top 2 cm, and the remainder of the chemical is applied uniformly between 2 cm and a user-specified incorporation depth.
* @Depth - Residues are incorporated entirely into a single compartment at a user-specified depth.
* Increasing to a depth - Pesticide is distributed to user-specified depth, linearly increasing with depth.
* Decreasing to a depth - Pesticide is distributed to user-specified depth, linearly decreasing with depth.

Note: If DEPI = 0, or DEPI < depth of the first (top) surface soil layer, the chemical reaching the soil surface is distributed into the first surface soil layer.

## Chemical Application Timing with Respect to Rainfall

Because farmers are unlikely to apply chemicals to fields when a large storm is reliably predicted, the PRZM5 has an optional algorithm to adjust the application timing to avoid such events. This algorithm will check the weather and scan forward for predictable events. A predictable event is determined by both the size of the event and the number of days looking forward into the future. For example, a farmer is unlikely to apply chemicals on a day when weather reports indicate that a 5-inch rainfall event will occur the following day. The built-in algorithm works as shown in Figure 4.4, where the program scans forward and backward for a most likely application day. In this figure, an original application day has been selected as indicated by the black hash mark. The program then scans the weather file forward and searches whether a hard rain (variable named rain\_limit in PRZM5) occurs within the acceptable window as indicated by the yellow shaded range (intolerable\_rain\_window in PRZM5). If a hard rain occurs within the range, then the program proposes a new application date that is one day forward from the original date (orange hash mark in figure). If the hard rain still occurs within the acceptable window (yellow shaded range), then the program proposes a date one day before the original date. If a hard rain still occurs in the unacceptable range, then the program proposes a new date 2 days forward from the original date. This forward and backward checking occurs until a satisfactory date is found. If no date is found within the user a user-specified maximum window (optimum\_application\_window in PRZM5), then the program uses the original date regardless of rainfall conditions.



Figure 4.4. Algorithm for choosing application date

## Chemical Processes on Canopy

Chemical degradation on the canopy is modeled as a first order process and solved with analytical solutions

Pesticide foliar washoff in PRZM5 is calculated in the same manner as in PRZM3 (although the PRZM3 manual did not coincide with the PRZM3 code). In PRZM3, residues from washoff are vertically distributed in the soil in proportion to the available space in the soil to a depth of 2 cm. To perform this calculation, the program estimates the available pore space in each compartment down to 2 cm. It then allocates the washed off pesticide among the compartments in proportion to the available space.

## Chemical Runoff & Vertical Transport in Soil

### Transport Model

PRZM5 simulates chemical changes to mass in the soil by runoff, erosion, volatilization, degradation and several other processes. The PRZM5 solute simulation is a classic vertical solute transport model with advection, dispersion, and dissipation, where the dissipation processes include degradation, and removal by runoff, erosion, and plant uptake. Additionally, PRZM5 can handle nonlinear isotherms and nonequilibrium sorption. The core mass balance PRZM5 relationships are as follows:

(4-48)

(4-49)

Where

C = concentration in water (mass/volume)

k2 = sorbed-phase referenced mass transfer coefficient (time-1)

S = sorbed concentration (mass/mass)

Se = region 2 sorbed concentration that would be at equilibrium with Region 1(mass/mass)

S2 = region 2 sorbed concentration (mass/mass)

G = concentration in gas phase (mass/volume)

V = volume of water in the compartment (volume)

Vg = volume of gas in the compartment (volume)

M = mass of soil in compartment (mass)

m2 =mass in nonequilibrium region

µw = first-order degradation coefficient for water phase (time-1)

µs = first-order coefficient for sorbed phase (time-1)

µg = first-order degradation coefficient for gas phase (time-1)

q= runoff flow intensity associated with depth z (volume/time/length)

Ez = eroded sediment flow intensity associated with depth z (mass/time/length)

QI = infiltration flow (vertical) at depth z (volume/time)

D = dispersion coefficient (length2/time)

A = cross section area (length2)

θ = water volume to total volume ratio, or porosity (volume/volume)

θg = gas volume to total volume ratio, (volume/volume)

Δz = compartment length in vertical dimension (length)

z = vertical dimension (length)

Sorption in the equilibrium and nonequilibrium regions is modeled with Freundlich isotherms, defined as follows:

(4-50)

(4-51)

Where KF = Freundlich coefficient for equilibrium region (mass/mass)/(mass/volume)N

N = Freundlich exponent for equilibrium region (dimensionless)

KF2 = Freundlich coefficient for nonequilibrium region (mass/mass)/(mass/volume)N2

N2 = Freundlich exponent for nonequilibrium region (dimensionless)

#### Boundary Conditions

The upper boundary conditions for aqueous transport are no advective chemical flux and no aqueous dispersive chemical flux. For the gas-phase, the condition is a zero concentration boundary that lies above a fixed gas layer. These conditions are described by the following equations:

(4-52)

(4-53)

(4-54)

Where u = boundary layer thickness (see volatilization).

The lower boundary is zero concentration for both gas and aqueous phase

(4-55)

(4-56)

#### Split-Operation Solution

The equilibrium region and the nonequilibrium region are solved by splitting the operations. In the first step, movement of chemical from the equilibrium region to the nonequilibrium region is calculated from an analytical solution. In the second step, the advection dispersion degradation step is calculated numerically.

#### Region 2 Mass Transfer between Equilibrium and Nonequilibrium Sites

In the first step of the split operation, an analytical solution is used to determine the mass transferred between the region 1 (equilibrium region) and region 2 (nonequilibrium region). This step is solved explicitly with the currently known values of all parameters. The differential equation describing that a mass change in Region 1 is equal to same change in Region 2 except with a sign difference:

(4-57)

or

(4-58)

For this step, the relation between Se and C is approximated as being constant and equal to

(4-59)

(4-60)

(4-61)

(4-62)

These equations can then be rearranged into the generic form

 (4-63)

 (4-64)

Equations (47) and (48) have the solution:

 (4-65)

 (4-66)

where

 (4-67)

 (4-68)

 (4-69)

 (4-70)

#### Sub-daily Time Step

We can also define an effective linearized sorption coefficient that can facilitate solutions later:

(4-71)

Where Kd,C is the effective sorption coefficient (volume/mass).

Gas partitioning is modeled with a Henry's Law Coefficient

(4-72)

where H = Henry's law coefficient (dimensionless, volumetric reference)

Making substitutions, dividing by total volume, and noting that θ and Kd,C vary with time gives:

(4-73)

With an overall degradation term defined as:

(4-74)

Where

= overall degradation (time-1)

I = area-normalized infiltration flow at depth z (length/time), and

ρb = bulk density (mas/volume).

µI = area normalized depth-dependent runoff dissipation intensity (time-1)

µR = area normalized depth-dependent runoff dissipation intensity (time-1)

µE = area normalized depth-dependent erosion dissipation intensity (time-1)

Discretizing the differential equation with a fully implicit temporal scheme gives

(4-76)

#### Tridiagonal Coefficients

The series of equations represented by Equation (4-75) can be presented in the following form, which represents a tridiagonal matrix:

(4-77)

Where the coefficients are defined as

(4-78)

(4-79)

(4-80)

(4-81)

#### Predictor Corrector Method

When the Freundlich exponent is not equal to 1, an estimate must be made for the effective Kd at time = i+1. This is accomplished by a predictor-corrector method similar to Douglas and Jones (1963). In this procedure, the current value of the effective Kd (at t = i) is used to predict a future concentration at a half time step (t = i+1/2) and the associated future effective Kd (t = i+1/2). This half-step future Kd is used in the corrector step to make a full time step for concentration at t = i+1. To ensure numerical robustness, PRZM5 allows for calculation at time steps less than a day (see below).

#### Subdaily time steps

Situation with high degrees of nonlinearity nonequilibrium may lead to numerical problems like stiff equations. For such cases, use of smaller time steps may be warranted for proper calculations. PRZM5 allows specification of sub-daily time steps by the user (see Record A3). When a subdaily time step is used, the main transport loop proceeds at the smaller specified time steps. Hydrologic calculations (e.g, runoff) are simply direct fraction of total daily flow that are in proportion to the subdaily time step and are not impacted by these time steps. Output is still recorded at the daily time steps.

### Runoff Extraction of Chemical

In PRZM5, the runoff includes a portion that interacts with the soil and a portion that does not. The portion that interacts with the soil is conceptualized as having a flow profile that decreases exponentially as depth increases. This is mathematically the near equivalent of the PRZM3 conceptualization. While PRZM3 invoked an abstract concept of partial concentrations that interact with runoff in a nonequilibrium manner, PRZM5 maintains equilibrium conditions between the runoff and soil in a manner similar to that conceptualized by Ahuja et al. (1981, 1983). The runoff flow distribution beneath the surface is represented by a subsurface flow intensity distribution described by:

(4-82)

Where *q(z)* = the runoff intensity at depth z (cm runoff/cm depth)

*q0* = the hypothetical flow intensity at the surface (cm runoff/cm depth)

*Kr* = the decline coefficient describing the decrease in flow with depth (cm-1)

z = depth (cm)

The flow is assumed to be constrained to a subsurface depth of Dr. (In PRZM3, the value of *Dr* was hard-coded to be 2 cm.) Furthermore, PRZM5 invokes the concept that not all the runoff flow moves through the subsurface, and a portion will flow above the surface without interacting with the subsurface chemical. The fraction that interacts with the subsurface is *Fr*, and 1-*Fr* is bypass flow flowing above the surface that does not contribute to the equilibrium extraction of the chemical in runoff. The total flow moving laterally through the subsurface is equal to the integral of the subsurface flow distribution from the surface to the depth *Dr*:

(4-83)

Where *F* = the fraction of runoff traveling through the soil (unitless)

*Qr* = runoff (cm)

*Dr* = maximum depth of runoff interaction (cm)

Equation 4-83 can be solved to determine the surface runoff intensity as follows:

(4-84)

Substituting 4-84 into 4-82 gives the intensity as a function of depth z:

(4-85)

Young and Fry (2019) calibrated this model using watershed-scale pesticide runoff data. The calibrations resulted in the extraction profile shown in Figure 4.5 with parameters of: F = 0.19, Kr = 1.4 cm-1, and Dr = >8 cm.



**Figure 4.5** Depiction of the PRZM5 effective runoff extraction profile using parameters calibrated by Young and Fry (2019).

### Erosion Extraction of Chemical

In PRZM5, the extraction of pesticide by eroded sediment is conceptualized differently than in PRZM3. In PRZM3, the pesticide associated with erosion was taken from the topmost compartment regardless of its size. This result was undesirable, as pesticide transport off the field became a strong function of the top compartment size; thus, pesticide transport could be dramatically reduced by decreasing the compartment size. To eliminate artificial relationships, PRZM5 allows the user to specify the interaction in a manner similar to the runoff extraction relationship. In a derivation analogous to the runoff extraction routine above, the erosion intensity with depth is estimated as follows:

(4-86)

Where *E(z)* = the erosion intensity at depth z (kg /cm)

*Me* = mass of eroding solids (kg)

*Ke* = the decline coefficient describing the decrease in erosion with depth (cm-1)

*De* = maximum depth erosion interaction (cm).

*Fe* = fraction of erosion interacting with soil/pesticide.

## Chemical Volatilization

Chemical volatilization was not included in the original PRZM release. PRZM3 attempted to incorporate volatilization for the first time to simulate vapor-phase pesticide transport within soil/plant compartments, but the routine never functioned properly. In PRZM5, the general PRZM3 chemical volatilization routine has been retained, but a more functional version will need to be developed. The theory behind the PRZM3 chemical volatilization routine is described here, from section 6.3.6 of the PRZM3 Manual, including several key processes (**Figure 4.6**):

* Vapor-phase movement of pesticide through soil profile
* Boundary layer transfer at soil-air interface
* Vertical diffusion of pesticide vapor within the plant canopy
* Pesticide mass transfer between plant (leaves) and surrounding atmosphere
* Soil temperature effects on pesticide volatilization

Depiction of volatilization processes.

**Figure 4.6 Pesticide vapor and volatilization processes.**

### Soil Vapor Phase and Volatilization Flux

#### Surface boundary condition

The initial volatilization rate after a pesticide is incorporated into the soil depends on the vapor pressure of a pesticide at the surface. As the pesticide concentration changes at the surface, volatilization may also become increasingly dependent on the rate of movement of the pesticide to the soil surface (Jury et al., 1983a, 1983b).

One boundary layer model used in PRZM3 determines pesticide volatilization based on molecular diffusion through a stagnant surface boundary layer (Jury et al., 1983a, 1983b). If the diffusion rate through the stagnant layer matches the upward flux to the soil surface, without the surface concentration building up, then the stagnant layer does not act as a barrier to loss. Conversely, if the diffusion rate is less than the flux to the surface, then the concentration at the surface will be greater than zero.

The pesticide volatilization flux from the soil profile is estimated as:

(4-87)

Where J1 = volatilization flux from soil (g/day)

Da = molecular diffusivity of the pesticide in air (cm2/day)

A = cross-sectional area of soil column (cm2)

d = height of stagnant air boundary layer (cm), assumed to be 0.5 cm (Wagenet and Biggar, 1987)

Cg,1 = vapor phase concentration in surface soil layer (g/cm3)

Cg,d = vapor phase concentration above the stagnant air boundary layer (g/cm3) (zero if soil surface is bare; positive if plant canopy exists)

#### Volatilization Flux through Plant Canopy

The pesticide volatilization flux through the plant canopy is calculated by Fick’s First Law of diffusion:

(4-88)

Where Jz(z) = pesticide volatilization flux at height z (g m-2 s-1)

dP/dz = pesticide concentration gradient (g m-2)

Kz(z) = vertical diffusivity at height z (m2 s-1), as a function of meteorology

Based on Fick’s First Law, pesticide concentrations at two or more heights can be used to estimate the pesticide gradient and subsequent flux. To estimate vertical diffusivity, additional meteorological information would be needed. However, PRZM3 circumvents these data requirements by using a relationship for Kz, which is a function of height within the canopy (Mehlenbacher and Whitfield, 1977), so only the pesticide concentration gradient is needed for estimating Jz(z). Kz is calculated at various heights within the plant canopy as follows:

(4-89)

(4-90)

(4-91)

Where Kz(z) = thermal eddy diffusivity at height z (m2 s-1)

Kz(zch) = thermal eddy diffusivity at canopy height zch (m2 s-1)

zch = top of canopy height (m)

zo = surface roughness length (m)

D = zero plane displacement height (m)

k = von Karman’s constant, 0.4

U\* = friction velocity (m s-1)

h = stability function for sensible heat

m(m) = integrated momentum stability parameter

m = stability function for momentum

Uch = wind at canopy height (m s-1)

In agricultural applications, the canopy height (zch) is used as the reference height for calculating U\*. The user must input the wind speed and height at which the measurement was made. Since PRZM3 assumes neutrally stable atmospheric conditions (where m(m) = 0), the wind speed at canopy height (Uch) is calculated as follows:

(4-92)

Where Ur = wind speed (m s-1) at zr, retrieved from meteorological file

zch = top of canopy height (m)

Dch = zero plane displacement height (m) associated with canopy

zo,ch = roughness length (m) associated with canopy

zr = reference height (m), assumed equal to 10.0 (for open flat terrain)

Dr = zero plane displacement height (m) associated with measurement, assumed equal to 0.0 (for open flat terrain)

zo,r = roughness length (m) associated with measurement, assumed equal to 0.03 (for open flat terrain)

Table 6.2 of the PRZM 3 Manual provides aerodynamic parameters (i.e., reference heights, zero plane displacements, and roughness lengths) commonly used (Burns et al., 2005). PRZM3 assumes the open flat terrain conditions for its wind speed calculations, but the user may specify a reference height in the PRZM input file.

For short crops (i.e., lawns), zo adequately describes the total roughness length, and no zero-plane displacement is needed (D = 0).

For tall crops, zo is related to canopy height (zch) by:

(4-93)

D is calculated, as zo (for tall crops) is not an adequate description of the total roughness length. For a wider range of crops and heights (0.02 m < zch < 25 m), the following equation is used (Stanhill, 1969):

(4-94)

In PRZM3 when zch ≤ 5 cm, D = 0 and zo is set to the value given by equation (4-48) evaluated at zch = 0.05 m. Once zo and D have been estimated, U\* can be calculated if the stability parameters (m and h) are known. These two stability parameters are closely related to the Richardson number (Ri), which is a measure of the rate of conversion of convective turbulence to mechanical turbulence (or the relationship between the temperature gradient and wind speed). Based on Thibodeaux (1996), Ri (typically -2.0 ≤ Ri < 0.2) is calculated as follows:

Ri = = (4-95)

Where

Tmean = mean temperature at defined level (K) =

g = acceleration of gravity, 9.8 m/s2 (86400 s/d)2 = 7.32e10 m/d2

v1 = wind speed at the soil surface (m/d)

v2 = wind speed at the top of the canopy (m/d)

T1  = air temperature on the soil surface (◦C)

T2 = ambient air temperature (◦C)

The sign of Ri indicates the atmospheric condition, and its magnitude reflects the degree of influence:

For Ri > 0.003, stable conditions and little vertical mixing

For |Ri| < 0.003, neutral stability conditions

For Ri < -0.003, unstable conditions and convective mixing

To relate the atmospheric stability parameters to Ri, Arya (1988) proposed using the Ri to calculate a dimensionless height (z):

(4-96)

The stability functions for momentum (m) and sensible heat (m) are then calculated as follows:

(4-97)

(4-98)

The integrated momentum stability parameter (m) is finally given by Thibodeux (1996):

(4-99)

#### Resistance Approach for Estimating Volatilization Flux from Soil

To calculate the volatilization flux from the soil, a resistance-type approach is used. For pre-plant pesticides and time periods just after emergence and post-harvest, transport by volatilization from plant surfaces is much less than vapor phase transport by other mechanisms. When plant leaves are not a significant source or sink of pesticide vapor, the resistances for the whole plant compartment is estimated as follows (Mehlenbacher and Whitfield, 1977):

(4-100)

Where

(4-101)

(4-102)

R = total vertical transfer resistance (day cm-1)

Rbd = boundary layer resistance (day cm-1)

Rpc = plant canopy resistance (day cm-1)

The flux is then calculated as follows:

(4-103)

Where Jpc = volatilization flux from plant canopy (g cm-2 day-1)

Cg,l = pesticide vapor concentration in top soil layer (g cm-3)

If the plant canopy acts as a significant source or sink, another approach should be taken, as described in the following section.

#### Volatilization Flux from Plant Surfaces

Based on Stamper et al. (1979), PRZM3 uses first-order kinetics to calculate volatilization flux from plant leaf surfaces, where the user inputs a first-order rate constant for volatilization. The plant leaf volatilization is estimated as follows:

(4-104)

Where Jpl = volatilization flux from the leaf (g cm-2 day-1)

M = foliar pesticide mass (g cm-2)

Kf = first-order volatilization rate (day-1)

The average pesticide concentration in the plant canopy is also estimated:

(4-105)

Where Cg\* = average concentration in air between ground surface and plant canopy height (g cm-3)

R0.5 = canopy resistance at one half of the canopy height to top of the canopy

#### Degradation in Soil

Adjustment for implicit temporal scheme. Use of analytical correction to give exact amount of degradation when soil degradation alone is in place. Soil degradation happens all the time regardless of rain, so it dominates usually so correction only done here. Other dissipation mechanisms are still numeric.

# Waterbody Calculations (VVWM)

## 1 Introduction

USEPA’s Office of Pesticide Programs (OPP) uses computer models to estimate pesticide exposure in surface waters resulting from pesticide applications to agricultural fields. These models are used to simulate pesticide applications to agricultural fields, the subsequent fate and transport in surface waters, and ultimately, estimated environmental concentrations (EECs) that are both protective and scientifically defensible. Using historical meteorological data from the region specified in the risk assessment, PRZM (Carsel *et al.*, 1997) calculates daily runoff and spray drift fluxes from “standard” fields over a simulation period (typically 30 years). These standard fields are parameterized to represent particular crops and regions of the United States (e.g., corn grown in Ohio). Another model, the Variable Volume Water Body Model (VVWM), simulates standard water bodies that receive pesticides from the standard fields. VVWM simulates the USEPA standard water bodies (i.e., farm pond and index reservoir) as well as user-defined water bodies. The VVWM also allows for variations in water body volume on a daily basis due to runoff, precipitation, and evaporation. Temperature, wind speeds, and pesticide dissipation processes are also allowed to vary daily.

## 2 The Varying Volume Water Body Model

### 2.1 Conceptualization and Mathematics

The VVWM is conceptualized in Figure 1 and consists of two regions: a water column and a benthic region. Each individual region is completely mixed and at equilibrium with all phases in that region, with equilibrium described by a linear isotherm. The two regions are coupled by a turbulent-mixing, first-order, mass-transfer process. As Figure 1 also shows, the pond volume may vary by inputs of precipitation and runoff and by outputs of evaporation and overflow.

Conceptual model of the standard water body

**Figure 1. Graphic of the standard water body showing inputs, outputs, and transformation processes.**

The mathematics are solved by daily piecewise analytic solutions. The temporal resolution is one day because daily inputs are readily acquired (i.e., runoff, rainfall, and evaporation data are 24-hour totals), and regulatory needs seldom require finer resolution. The water body volumes and flow rates are also daily values, consistent with the input data resolution. For the analytic solution, water body properties are held constant each day, but may vary from day to day.

All individual dissipation processes (e.g., metabolism, hydrolysis, and volatilization) are represented as first-order in concentration, as described later. On any given day, solute mass in the water body is described by two differential equations, namely a mass balance on the water column:

 (1)

and a mass balance on the benthic region:

 (2)

Where

B = benthic flow rate of sediment, [kg/s]

c1 = aqueous concentration in water column, [kg/ m3]

c2 = aqueous concentration in benthic region, [kg/ m3]

Csed = concentration of suspended sediment in water column = msed\_1/v1 [kg/m3]

CDOC = concentration of DOC in water column = mDOC/v1, [kg/m3]

Cbio = concentration of biota in water column = mbio/v1, [kg/m3]

msed\_1 = mass of suspended sediment in water column, [kg]

mDOC\_1 = mass of DOC in water column, [kg]

mbio\_1 = mass of suspended biota in water column, [kg]

msed\_2 = mass of suspended sediment in water column, [kg]

mDOC\_2 = mass of DOC in benthic region, [kg]

mbio\_2 = mass of biota in benthic region, [kg]

ssed\_1 = sorbed concentration on suspended sediment in water column, [kg/ kg]

sDOC\_1 = sorbed concentration on suspended DOC in water column, [kg/ kg]

sbio\_1 = sorbed concentration on suspended biota in water column, [kg/ kg]

ssed2 = sorbed pesticide concentration on benthic sediment, [kg/ kg]

sDOC\_2 = sorbed pesticide concentration on benthic DOC, [kg/ kg]

sbio\_2 = sorbed pesticide concentration on benthic biota, [kg/ kg]

v1 = volume of water in region 1 on the specific day, [m3]

v2 = volume of water in region 2, [m3]

Q = volumetric flow rate of water out of water column, [m3/s]

α = first-order water-column-to-benthic mass transfer coefficient, [m3/s]

μhydr = 1st order hydrolysis rate coefficient, [s-1]

μphoto =1st order photolysis rate coefficient, [s-1]

μvol = effective 1st order volatilization rate coefficient, [s-1]

μbio\_a1=1st order aqueous-phase metabolic degradation rate coefficient in water column, [s-1]

μbio\_sed1 = 1st order sediment-sorbed metabolic degradation rate coefficient in water column, [s-1]

μbio\_bio1 = 1st order biota-sorbed metabolic degradation rate coefficient in water column, [s-1]

μbio\_DOC1 = 1st order DOC-sorbed metabolic degradation rate coefficient in water column, [s-1]

μbio\_a2 =1st order aqueous-phase metabolic degradation rate coefficient in benthic region, [s-1]

μbio\_sed2 = 1st order sediment-sorbed metabolic degradation rate coefficient in benthic region, [s-1]

μbio\_bio2 = 1st order biota-sorbed metabolic degradation rate coefficient in benthic region, [s-1]

μbio\_DOC2 = 1st order DOC-sorbed metabolic degradation rate coefficient in benthic region, [s-1]

The following assumptions are made: (1) suspended matter in the water column has negligible volume, (2) hydrolysis, photolysis, and volatilization act only on dissolved species, (3) within a single region (water column or benthic), the rate coefficient for biological metabolism is the same for both dissolved and sorbed forms of pesticide (e.g., μbio\_1 = μbio\_a1 = μbio\_sed1 = μbio\_DOC1 = μbio\_biota1, and μbio\_2 = μbio\_a2 = μbio\_sed2 = μbio\_DOC2 = μbio\_biota2), (4) the hydrolysis rate coefficient in the benthic region is the same as that in the water column, (5) linear isotherm equilibrium exists within each region among all sorbed species. With these assumptions, we can rewrite equations (1) and (2) in a simpler form as follows:

 (3)

 (4)

1. where

 (5)

 (6)

 (7)

 (8)

where fw1 and fw2 are the fractions of solute in the aqueous phase within the water column and benthic regions, respectively, as defined by the following equations:

 (9)

 (10)

and where Ksed\_1, Kbio\_1, KDOC\_1 are the linear isotherm partitioning coefficients for suspended sediments, biota, and DOC in the water column, and Ksed\_2, Kbio\_2, KDOC\_2 are the linear isotherm partitioning coefficients for sediments, biota, and DOC in the benthic region (all with units of m3/kg).

The term, fw1, varies daily depending on the volume of the water body (v1) as described below in *Section 2.6* *Daily Piecewise Calculations*. We assume that the mass of sediment, biota, and DOC remain constant. However, this assumption has very little impact on the model output since partitioning to these species is insignificant, except when given extremely high partitioning coefficients.

Given a set of initial conditions, equations (3) and (4) completely describe the standard water bodies. It is clear that there are only four parameters that influence the concentration—Γ1, Γ2, Ω, and Θ. Γ1 is the effective overall degradation rate in the water column, [s-1]. Γ2 is the effective overall degradation rate in the benthic region, [s-1]. Ω is a mass transfer coefficient describing transfer between the benthic and water column, [s-1]. Θ is the ratio of solute holding capacity in the benthic region to that in the water column, [unitless]. The sections that follow describe the details of the components of these equations with respect to the standard water bodies.

***2.2 Solute Holding Capacity Ratio (Θ)***

The solute holding capacity ratio (Θ) is the ratio of solute holding capacity in the benthic region to the solute capacity in the water column, as defined by equation (8). The individual partitioning coefficients (Kd\_sed, Kd\_biota, and Kd\_DOC) used in equation (8) are generally not directly measured for a pesticide assessment. To account for these unknown coefficients, the standard water bodies use various estimation means that relate the various partitioning coefficients to the organic carbon partitioning coefficient (Koc), which is usually known in a pesticide assessment process.

For the sediment, the partitioning coefficient is directly proportional to Koc, with the constant of proportionality being the amount of organic carbon in the sediment, which is set to standard values for the standard water bodies (see Table 1). The fraction of organic carbon (foc) is assumed to be the same in the benthic and water column. The sediment partitioning coefficients can thus be determined from the following equation:

 (11)

where Koc = organic carbon partitioning coefficient, [mL/g]

foc = fraction of organic carbon in sediment [unitless]

Note that the units of the coefficients in equations (1) to (10) are all given in *s.i*. form, which is maintained throughout this document. However, for some fundamental parameters such as Koc, which is usually presented in units of mL/g, common units and conversion factors are used.

The partitioning coefficients for DOC are determined from the default empirical relationships described in the EXAMS documentation (Burns, 2000). The VVWM incorporates the notion of Burns (2000) that benthic DOC has higher partitioning characteristics than water column DOC for standard water bodies:

 (12)

 (13)

The partitioning coefficients for biota are also determined from default empirical relations described in the EXAMS documentation:

 (14)

By inserting equations (11) through (14) into equation (8) and substituting specific values from Table 1 into equation (8), the solute holding capacity (Θ) can be written as a function of solely Koc, as presented in Figure 2 for both the standard pond and reservoir.

1. ***2.3 Effective Water Column Dissipation (Γ1)***

The overall dissipation rate in the water column (Γ1), as defined in equation (5) is the sum of contributions from hydrologic washout and degradation by mechanisms of biological metabolism, photolysis, and hydrolysis. The specific methods and assumptions that are used in the VVWM to determine these individual first-order dissipation processes are described below.

1. **2.3.1 Hydrologic Washout **

The first term in equation (5), Q/v1, represents the effective first-order dissipation rate due to flow moving pesticide out of the water body. Flow out of the water body only occurs if meteorological conditions produce enough water inflow to cause the water body to overflow (see *Section 2.6 Daily Piecewise Calculations*). The washout term acts on all forms of pesticide (aqueous dissolved and sorbed to suspended matter), as is apparent from equation (1) and the definitions for Xsed, Xbio, and XDOC. This means that the settling of suspended solids is not explicitly considered in the VVWM, and pesticides in both dissolved and suspended sorbed forms can flow out of the reservoir.

Flow is obtained from an input file or entered as a constant baseflow. The input file provides a daily flow and is typically generated by the PRZM model as a zts file (see section 6.22) Baseflow is additive to any flow from the zts file.

1. **2.3.2 Metabolism (μbio\_1)**

In the registration process of pesticides, an estimate of the aqueous degradation rate under aerobic conditions is supplied by the registrant. Such estimates are derived from laboratory tests following standard EPA-approved protocols, which are typically conducted in aqueous/sediment systems at 20 to 25° C. These tests generally do not differentiate between degradation occurring on the dissolved and sorbed forms of the pesticide; an overall degradation rate is generally all that is available. Therefore, the VVWM treats the sorbed-phase and aqueous-phase degradation rates as the same, which makes both equal to the overall rate.

As temperature varies in a water body, the USEPA has established a standard for temperature adjustments of the aerobic metabolism rate when regulating pesticides as follows:

 (15)

where μ25 = laboratory measured aerobic metabolism rate at 25°C, [s-1]

T = temperature of modeled water body, [°C]

Tref = temperature at which laboratory study was conducted, [°C]

This temperature adjustment doubles the metabolism rate for every 10°C rise in temperature, and halves the rate for every 10°C decrease. Air temperature is taken from the meteorological data that corresponds to the crop/location scenario being simulated. The VVWM uses the previous 30-day average temperature and adjusts the temperature daily. (Note: EXAMS made temperature adjustments on a monthly calendar basis, which required tracking of the Gregorian calendar).

1. **2.3.3 Hydrolysis (μhydr\_1)**

The hydrolysis rate is directly obtained from experimental measurements, as supplied by pesticide registrant data submissions. In the VVWM, the effective hydrolysis rate is the experimentally determined overall hydrolysis rate from tests conducted at the pH of interest. In a typical USEPA assessment, the pH is 7 (Note: Because pH is not included explicitly in the VVWM, the appropriate input is the overall hydrolysis rate, not the specific neutral-, base-, or acid-catalyzed hydrolysis rate coefficients, as in EXAMS).

Unlike the metabolism rate, temperature adjustments of the hydrolysis rate are not made by the VVWM. Temperature-dependent hydrolysis characterizations are not generally made for the registration process, and the USEPA has not adopted a standard adjustment for temperature effects on hydrolysis. Therefore, the hydrolysis rate is as follows:

 (16)

where μoverall, pH = laboratory-measured overall hydrolysis rate at pH of interest, [s-1].

The VVWM uses the assumption that hydrolysis acts only on dissolved species. Therefore, the effective hydrolysis rate is reduced by the fraction of total pesticide that is present in dissolved aqueous form (fw1), as defined in equation (9) and implemented in equation (5).

**2.3.4 Photolysis (μphoto)**

Photolysis rates are derived from standard laboratory tests following USEPA-approved protocols. These tests are designed to estimate the photodegradation rate for near-surface conditions at a specific latitude and under clear-sky conditions. The VVWM adopts the methods given by EXAMS (Burns 1997, 2000) to account for latitude adjustments, light attenuation, and cloud cover:

 (17)

where flat = latitude adjustment factor, [unitless]

fcloud = cloudiness adjustment factor, [unitless]

fatten = attenuation factor to absorption, [unitless]

μmeasured = measured near-surface photolysis rate coefficient at reference latitude and clear

atmospheric conditions [sec-1]

Although cloudiness does not affect the current standard water bodies (fcloud is set to a standard value of 1), fcloud is included here for the purposes of formality and because it may be considered in future versions.

The latitude of the standard water body varies, depending on the desired location in the U.S. where the pesticide assessment is being made. The effect that latitude has on incident light is accounted for by the latitude adjustment factor (flat), which the VVWM adopts from EXAMS (Burns, 2000). Full details of the reasoning behind flat can be found in the EXAMS documentation, and only the resulting equation is given here:

 (18)

where Lref = reference latitude at which the measured photolysis rate was determined, [degrees]

Lsim = latitude of the simulated scenario, [degrees]

The light attenuation factor (fatten) described by Burns (2000) has also been adopted; the full details are available in the EXAMS documentation:

 (19)

where Dfac = EXAMS-defined distribution factor default value = 1.19, [unitless]

d1 = depth of water column, [m]

a = total absorption coefficient, [m-1]

The absorption coefficient (a) is calculated from EXAMS default conditions—that is, from the spectral absorption coefficient assuming that the wavelength of maximum absorption occurs at 300 nm:

(20)

where CDOC, CSed have been previously defined under equation (1), and CCHL is the chlorophyll concentration [mg/L].

Temperature effects are not considered in the above equations, except when the water temperature is 0°C or below. Photolysis is inhibited, as in EXAMS. Temperature effects are not considered since the USEPA generally does not receive temperature dependent data for the registration process and has not adopted a standard temperature adjustment for photolysis.

**2.3.5 Volatilization (μvolatilization)**

The VVWM uses a two-film model for volatilization calculations and all of the default volatilization assumptions as described in the EXAMS documentation (Burns, 2000). The concentration of a pesticide in the atmosphere is assumed to be negligible, and thus volatilization becomes a first-order dissipation process. The overall volatilization rate coefficient is expressed as follows:

 (21)

where A = surface area of water column, [m2]

kvol = volatilization exchange coefficient, [m/s]

and the volatilization exchange coefficient comprises liquid-phase and gas-phase resistances:

 (22)

where kw = liquid-phase resistance [m/s]

ka = gas-phase resistance, [m/s]

H = Henry’s law constant (m3atm/mol)

R = the universal gas constant (8.206 x 10-5 m3atm/mol/K)

T= temperature (K)

The VVWM uses the EXAMS methods of referencing the liquid exchange resistance of pesticides to the liquid resistance of oxygen, and uses molecular weight as the sole surrogate for molecular diffusivity variations among compounds. Further details can be found in the EXAMS documentation (Burns, 2000), but the resulting relationship is as follows:

 (23)

where kO2 = oxygen exchange constant at 20°C, [m/s]

MW = molecular weight of pesticide, [g/mol]

The oxygen exchange constant is determined from the empirical relationship of Banks (1975). Adjustments are also made for temperatures other than 20°C. Note that although EXAMS uses a reference temperature of 20°C for the Banks (1975) relationship, it is not clear from Banks (1975) what the actual reference temperature should be. Schwarzenbach et al. (1992) used a 10°C reference for the same relationship. Until further clarified, a 20°C reference temperature is used. For wind velocities (vwind) less than 5.5 m/s, kO2 is calculated as:

 (24)

and for wind velocities greater than or equal to 5.5 m/s, kO2 is:

 (25)

where u10 = wind velocity at 10 m above water surface [m/s].

Wind speeds measured at 10 m above the surface are read from the meteorological files. To convert to wind speeds at a different height, the following equation is used:

 (26)

where z0 is the boundary roughness height, which is assumed to be 1 mm for the standard water bodies. Given a wind speed (measured at 10 m) from the meteorological file, the equivalent wind speed at 0.1 m is:

 (27)

In the VVWM, wind speed varies on a daily basis, unlike in EXAMS where the average monthly wind speed varies on a monthly basis.

The gas-phase resistance is referred to as water vapor resistance, and an empirical relationship based on a linear regression of laboratory-derived data from Liss (1973) relates the water vapor exchange velocity to wind speed:



where ka,H2O = the water vapor exchange velocity (m/s)

u0.1 = wind speed velocity measured at 0.1 m above the surface (m/s)

The exchange rate of a pesticide is then related to the exchange rate of water by:

 (28)

where α (not to be confused with the alpha in equation 1 and 2) is a value that depends on the conceptual model believed to describe volatilization and ranges from 0.5 for the surface renewal model to 1.0 for the stagnant film model (Cusler,1984 ; Schwarzenbach et al., 1993). The VVWM uses a value of 1.0 for α; thus, implying a stagnant film model. However, some laboratory data suggest that α may be better represented with a value of 0.67 (Mackay and Yuen, 1983). The diffusion coefficient of the pesticide is related to the diffusion coefficient of water by the common approximate relationship (e.g., Schwarzenbach et al., 1993):

 (29)

Substituting (29) into (28) gives:

 (30)

The resulting relationship is:

 (31)

The Henry’s Law constant is generally not available from pesticide registration submissions, so it is approximated in the VVWM from vapor pressure and solubility. The Henry’s Law constant also is not adjusted for temperature, as this information is not supplied in the pesticide registration, and OPP has not adopted a standard temperature adjustment factor. The resulting relationship is:

 (32)

where vp = vapor pressure [torr]

Sol = solubility [mg/L]

***2.4 Effective Benthic Region Dissipation (Γ2)***

The overall benthic dissipation in the VVWM, as defined in equation (6), is affected by biodegradation and hydrolysis and sediment flow. As with the water column, OPP assumes that biodegradation in the benthic region affects all forms of pesticide (both dissolved and sorbed forms) and that hydrolysis affects only aqueous dissolved forms (see equation 6 and definition of fw2). Dissipation by sediment flow is caused by the sediemt flowing in and out of the benthic compartment. The benthic compartment has a fixed size, so any sediment entering must displace an equal amount of sediment that exits the compartment along with any sorbed pesticide.

1. **2.4.1 Benthic Hydrolysis (μhydr\_2)**

In the current standard water bodies, the pH of the entire system (benthic and water column) are held at a constant pH of 7, although a subsequent paper will suggest using scenario-specific pH values. Benthic hydrolysis is assumed to occur at the same rate as hydrolysis in the water column; the previous discussion of hydrolysis in the water column applies to the benthic region:

 (33)

1. **2.4.2 Benthic Metabolism (μbio\_2)**

In the VVWM, benthic metabolism is assumed to occur under anaerobic conditions. Therefore, anaerobic metabolism rates are derived from laboratory tests following standard EPA-approved protocols. These studies are typically conducted in aqueous/sediment systems at 20 - 25°C. As with water column metabolism, OPP assumes that sorbed-phase degradation occurs at the same rate as aqueous-phase degradation, and temperature effects on metabolism are handled in the same way. Thus, the effective rate is the following:

 (34)

where μmeasured = laboratory measured anaerobic metabolism rate at Tref

T = temperature of modeled water body [°C]

Tref = temperature at which anaerobic laboratory study was conducted [°C].

1. **2.4.3 Benthic Sediment Flow**

The last term in equation 6 represents the dissipation rate due to sediment flow through the benthic zone. Sediment flow-through is necessary to maintain a mass balance on the sediment in the benthic region. All sediment that flows into the benthic region must displace an equal amount of sediment that exits the region and carries sorbed pesticide with it.

1. ***2.5 Mass Transfer Coefficient (Ω)***

The mass transfer coefficient (Ω) defined in equation (7) is an overall coefficient that includes all means of pesticide exchange between the water column and benthic regions. This coefficient includes exchange through the aqueous phase as well as by mixing of sediments between the two compartments. The physical process of this combined mixing is assumed to be completely described by a first-order mass transfer coefficient (α). The parameter α is referenced to the aqueous phase, but implicitly includes exchange due to mixing of sediments as well as aqueous exchange. In compartment modeling, it is unnecessary to explicitly model the individual exchange mechanisms (as EXAMS does) since all phases of pesticide within a compartment are at equilibrium. Therefore, the concentration of a pesticide in any given form (aqueous or sorbed) dictates the concentration of the other forms of the pesticide.

In the VVWM, the α term is based upon parameters and assumptions given in the EXAMS documentation. Although not explicitly presented as such, EXAMS uses a boundary layer model to exchange pesticide mass between the water column and benthic regions. EXAMS defines the parameter DSP, which represents a Fickian-type dispersion coefficient in the benthic sediment. This dispersion coefficient acts on the total concentration within the benthic region, implying that sediment-sorbed pesticide moves through the benthic region at the same rate as dissolved-phase pesticide (e.g., via bioturbation). The rate of mass change in the benthic region is approximated under steady state conditions across a boundary layer of constant thickness:

 (35)

where M2 = total pesticide mass in benthic region

A = area of benthic/water column interface, [m2]

1. D = effective overall dispersion coefficient in benthic media (includes both sorbed and
2. dissolved phases), [m2/s]; DSP in EXAMS
3. Δx = thickness of boundary layer, [m]

 = total partition coefficient for total concentrations, [unitless]

1. CT1 = total concentration in water column, [kg/m3]

CT2 = total concentration in benthic region, [kg/m3]

The total concentrations in the water column and benthic regions are calculated as follows:

 (36)

1.  (37)

where c1 and v1 are the aqueous-phase concentration and the aqueous volume, as previously defined under equation (1). Σ(m1Kd1) and Σ(m2Kd2) are short-hand notation for the sum of all solid masses and the respective Kds presented under equation (1) for the water column and benthic regions, respectively; VT1 and VT2 are the total volumes of the water column and benthic region, respectively, which include both the water and the solids volumes. The total pesticide mass in the benthic region is expressed as follows:

 (38)

The total partitioning coefficient is defined as the ratio of CT2 to CT1 when the system is at equilibrium:

 (when benthic region is at equilibrium with water column) (39)

By substituting in the definitions of CT1 and CT2 from equations (36) and (37) and recognizing that at equilibrium c1 = c2, the total partitioning coefficient becomes:

 (40)

Substituting equations (36) to (40) into equation (35) yields the following:

 (41)

Comparing equation (41) with equation (2), we can see that:

 (42)

and that Ω is:

 (43)

where D = overall water column -to-benthic dispersion coefficient (m2/s)

1. Δx = boundary layer thickness (m)
2. A = area of water body (m2)

D in the above equation is set to a constant (Table 1) for the USEPA standard pond. The value of D was originally chosen to be on the order of Fickian-type dispersion coefficients in sediments, as observed in field studies reported in the EXAMS documentation. Although equation (42) implies a mechanistic meaning to α, it is difficult to adequately transform Fickian-type dispersion coefficients into first-order mass transfer coefficients for finite volume compartments, and it is equally difficult to define a boundary layer thickness, especially when there is sediment and aqueous mixing. The EXAMS documentation suggested that the boundary layer thickness be equal to the distance between the center of the water column and the center of the benthic region; however, the actual boundary layer thickness is difficult to estimate and is likely more related to benthic animal life and associated turbulence than to water column depth.

Attempting to model the benthic mass transfer parameter as a function of water column depth would be speculative, so the VVWM currently maintains a constant thickness.

1. ***2.6 Daily Piecewise Calculations***

Because we retain an analytical solution, the VVWM is solved in a daily piecewise fashion, in which the volume of the water column changes at the beginning of the day and remains constant for the duration of that day. Mass is conserved in the water column by recalculating a new beginning day concentration with any volume change.

1. **2.6.1 Volume Calculations**
2. The volume of the water column aqueous phase is calculated from daily runoff, precipitation, and evaporation for any day as follows:
3.  for 0 < v1 < vmax (44)
4. where v0 = the aqueous volume of the previous day (m3)
5. R = daily runoff into the water body (m3)
6. P = daily direct precipitation on water body (m3)
7. E = daily evaporation of runoff (m3)
8. S = daily seepage = 0 (neglected) (m3)
9. Daily runoff is taken from the PRZM model output. Daily precipitation and evaporation are taken from the meteorological file. Seepage at this time is not considered, as in EXAMS. If the newly calculated volume (v1) is greater than vmax, then the volume for the day is set to vmax, and the excess water is used in the calculation of washout. The minimum water volume is zero, but it is set to an actual minimum to prevent numerical difficulties associated with calculations involving infinity and zero. There also may be some practical physical lower boundary appropriate for the minimum volume, such as those associated with soil water holding capacity, water tables, and refilling practices of pond owners. These factors need to be explored further.
10. **2.6.2 Initial Conditions**

Initial concentrations are determined by the pesticide mass inputs from PRZM and spray drift. PRZM gives daily outputs for pesticide mass associated with aqueous-phase runoff and erosion solids. Pesticide mass in aqueous-phase runoff and from spraydrift are delivered to water bodies are delivered to the water column as an initial daily condition. For chemical associated with eroded solids, the VVWM allows two options: fixed or varying. With the fixed option, eroded chemical is distributed with a constant user-defined frac. For example, 90% of all eroded pesticide is delivered directly to the benthic region and 10% is delivered to the water column. This was the way that EXAMS had handled the eroded chemical distribution. With the varying option, the incoming eroded pesticide is first equilibrated in the water column and any pesticide remaining on the eroded sediment is delivered directly to the sediment with the remaining pesticide delivered to the water column. For this case the distribution will vary depending on how much eroded sediment enters the waterbody (described more fully below).

1. ***2.6.2.1 Fixed Fraction of Eroded Pesticide***

For the fixed-fraction option of eroded pesticide, the initial concentrations, upon addition of new pesticide mass, are expressed as follows:

 (45)

 (46)

where Mrunoff = mass of pesticide entering water body via runoff (kg)

Merosion = mass of pesticide entering water body via erosion (kg)

Mdrift = mass of pesticide entering water body via spray drift (kg)

C10,prior = aqueous concentration in water column before new mass additions and with consideration for any volume changes for the day (kg/m3)

C20,prior = aqueous concentration in benthic region before new mass additions (kg/m3)

v1, prior = the water column volume from the previous day (m3)

fw1,prior = fw1 from the previous day

Xfixed = fractional initial distribution (between water column and benthic region) of the pesticide associated with eroded solids as it enters the water body

1. ***2.6.2.2 Initial Distribution of Eroded Pesticide***

The basis for the variable option to handle eroded pesticide in VVWM, is to maintain equilibrium in the water column when eroded sediment enters. With this option, the eroded pesticide and the eroded soil mass is equilibrated with the water column. Any chemical mass that remains sorbed to the eroded sediment is then delivered to the benthic region. With the remaining chemical mass staying in the water column. This then establishes the daily initial condition. Thus, the initial conditions are:

 (47)

 (48)

The term *Xvariable* in equations above represents the distribution of eroded pesticide between the water column and the benthic region when the eroded pesticide enters the water body.

** (49)

Where merosion = mass of eroded sediment entering the waterbody during an erosion event (kg)

1. ***2.7 Analytical Solution***
2. Equations (3) and (4) along with the initial conditions represent the two equations describing the standard water bodies. These equations are in the following form:

 (50)

 (51)

where









Equations (47) and (48) have the solution:

 (52)

 (53)

where

 (54)

 (55)

 (56)

 (57)

Average concentrations can be determined over any interval in which all parameters remain constant. In the VVWM, parameters change on a daily basis, so the average water column concentration is expressed as follows:

 (58)

where C1,avg = average water column concentration of time from t1 to t2 [kg/m3]

t1 = beginning of the time interval considered [s-1], (zero for our case of daily estimates)

t2 = end of the time interval considered [s-1], (86,400 seconds for our case of daily estimates)

1. **3 The USEPA Standard Water Bodies**

All parameters in the above equations, except for the pesticide-specific parameters, have standard values set by the USEPA for the standard farm pond and index reservoir scenarios (Table 1). Many of these values have no documentation and simply have evolved over many years of repeated, unquestioned use. Table 2 shows how the parameters in the VVWM simplify and replace previous EXAMS parameters and expressions, and Table 3 lists the original EXAMS standard parameters. The VVWM also gives the option to define a custom-sized water body.

1. **Table 1. Standard Parameter Values for the VVWM.**

| **Parameter** | **Units** | **Farm Pond Values** | **Index Reservoir Values** | **Notes** |
| --- | --- | --- | --- | --- |
| v1 | m3 | 20,000 | 144,000 | water column volume |
| v2 | 1. m3 | 1. 249.8 | 1,314 | aqueous benthic volume(a) |
| A | m2 | 10,000 | 52,555 | surface area, calculated (v1/d1) |
| d1 | 1. m | 1. 2.0 | 2.74 | water column depth |
| d2 | 1. m | 1. 0.05 | 0.05 | benthic depth |
| msed\_1 | 1. kg | 1. 600 | 4,320 | 1. based on suspended solids concentration of 30 mg/L (see Csed\_1) |
| 1. mbio\_1 | kg | 8.0 | 57.60 | based on biota concentration of 0.4 mg/L |
| mDOC\_1 | kg | 100 | 720 | based on DOC concentration of 5 mg/L |
| foc | 1. — | 0.04 | 0.04 | fraction of organic carbon (water column and benthic) |
| 1. msed\_2 | 1. kg | 1. 6.752 x 105 | 3.552 x 106 | 1. (b) |
| mbio\_2 | kg | 0.0600 | 0.3156 | 1. (c) |
| mDOC\_1 | kg | 1.249 | 6.570 | 1. (d) |
| pH |  | 1. 7 | 7 |  |
| CCHL | mg/L | 0.005 | 0.005 | chlorophyll concentration |
| CDOC | mg/L | 5 | 5 | DOC concentration |
| Csed\_1 | mg/L | 30 | 30 | suspended solids concentration |
| Cbio | mg/L | 0.4 | 0.4 | biomass concentration |
|  |  |  |  |  |
| D | m2/s | 8.33 x 10-9 | 8.33 x 10-9 | sediment dispersion coefficient |
| Δx | m | 1.02 | 1.39 | benthic/water column boundary layer thickness |
| VT2 |  | 1. 500 | 2,630 | total volume of benthic region (d1 x A) |

(a) calculated from: VOL2\*BULKD\*(1.-100./PCTWA)

(b) calculated from: (BULKD)(VOL2)(100000)/PCTWA (see Table 2)

(c) calculated from: BNMAS\*AREA\*.001(see Table 2)

1. (d) calculated from: DOC\*v2/1000

**Table 2. VVWM Equivalents of EXAMS Parameters.**

| **VVWM Parameters** | **Expressed in Terms of EXAMS Parameters** |
| --- | --- |
| m1  [kg] | (SUSED)(VOL1) (10-3) |
| m2 [kg] |  |
| v1 [m3] | VOL1 |
| v2 [m3] | \* |
| Q [m3/s] | STFLO (3600 s/hr) |
| μA1 [s-1] | (KBACW1)(BACPL)/(3600s/hr) |
| μS1 [s-1] | (KBACW2)(BACPL)/(3600s/hr) |
| μA2 [s-1] |  |
| μS2 [s-1] |  |
| Ω [s-1] |  |
| Kd1 [m3/kg] | (KOC)(FROC)(10-3 m3/L) |
| Kd2 [m3/kg] | (KOC)(FROC)(10-3 m3/L) |

\*Assumes that the density of water is 1,000 kg/m3

**Table 3. EXAMS Standard Parameters. The parameters here are for reference only and are for those who wish to understand EXAM definitions. The equivalent parameters in the VVWM are defined differently.**

| **EXAMS Parameter** | | **EXAMS Value for Standard Pond** | **EXAMS Value for Standard Drinking Water Reservoir** |
| --- | --- | --- | --- |
| PRBEN | — | 0.5 | 0.5 |
| PCTWA | — | 137 | 137 |
| BULKD | g/mL | 1.85 | 1.85 |
| FROC | — | 0.04 | 0.04 |
| CHARL | m | 1.05 |  |
| DSP | m2/hr | 3.00 x 10-5 | 3.00 x 10-5 |
| AREA | m2 | 10000 | 52600 |
| VOL1 | m3 | 20,000 | 144,000 |
| VOL2 | m3 | 500 | 2,630 |
| DEPTH1 | m | 2 | 2.74 |
| SUSED | mg/mL | 30 | 0.005 |
| CHL | mg/L | 0.005 | 0.005 |
| DOC1 | mg/L | 5.0 mg/L | 5.0 mg/L |
| DOC2 | mg/L | 5.0 mg/L | 5.0 mg/L |
| LAT |  | 34 | 39.1 |
| BNMAS | g/m2 | 0.006 | 1. 0.006 |
| BNBAC1 | -- | -- | -- |
| BNBAC2 | cfu/100g | 37 | 37 |
| BACPL1 | cfu/mL | 1 | 1 |
| BACPL2 | — | -- |  |
| DFAC | — | 1.19 | 1.19 |
| WIND | m/s | metfile | metfile |
| STFLO | m3/hr | 0 | Average daily rainfall (from 36 years of data) |
| TCEL | °C | monthly avg | monthly avg |

1. ***3.1 Farm Pond***

The standard farm pond, representing a highly vulnerable exposure scenario, is a pond located at the edge of a pesticide-treated field. The pond dimensions (1 ha area by 2 m depth), originally based on a Georgian farm pond size, are in accordance with USDA guidance for pond construction for an appropriately-sized pond fed by a 10-ha watershed—that is, approximately 2 acres of drainage per acre-ft of storage in central Georgia (USDA, 1982). In the farm pond, inflow is assumed to exactly balance evaporative losses (leaching is not modeled). Table 1 gives some of the standard parameters for the pond.

1. ***3.2 Index Reservoir***

The index reservoir represents a natural or artificial lake fed by perennial and ephemeral streams, varying in flow due to precipitation, evaporation, and runoff from the surrounding watershed and groundwater discharge. The reservoir is a potential drinking water source that may be affected by pesticide runoff, spray drift, and leaching to groundwater. The reservoir is a fixed volume water body with outflow equated to runoff that enters the reservoir. Table 1 gives some of the standard parameters for the index reservoir.

1. ***3.3 Custom Water Body***

A custom water body also can be defined in the VVWM with specific dimensions, including the field area [m2], water body area [m2], initial depth [m], maximum depth [m], and hydraulic length [m]. The custom water body can be of varying volume, or of constant volume with (or without) flow through. This third option allows for greater flexibility in evaluating pesticide fate and transport in a non-standard receiving water body.

1. **4 VVWM Evaluations**
2. ***4.1 Solute Holding Capacity Ratio Sensitivity***
3. As Figure 2 shows, the standard index reservoir has a lower solute holding capacity ratio than the standard pond, and this is due to the greater water column depth of the reservoir. The point where Θ is equal to 1 represents the Koc for which the solute capacity in the benthic region is equal to that in the water column. For the pond, equal capacities occur at Koc of 730 mL/g, and for the reservoir, the equal capacities occur at 1,000 mL/g. Of course, the water column and benthic regions are not at equilibrium, so the actual distribution of solute will not be evenly split between benthic and water column at these Koc values. These values and Figure 2, however, give some physical insight into how the standard water bodies can potentially distribute solute.
4. It is also of interest to examine the relative significance of the individual media within each region with regard to the distribution of solute among them. Figure 3 shows the relative capacities of the individual media (aqueous and sorbed to biota, DOC, and suspended sediment) within the water column as a function of Koc. Up to a Koc value of ~10,000 mL/g, only the water phase is significant. Up to Koc values of 100,000, biota partitioning is not significant, and at a Koc value of 100,000, the combined capacities of all sorbed species amounts to less than 20 percent of the total water column capacity. It can also be seen that, for the standard water bodies, DOC and suspended sediments have nearly equal capacities for solute.
5. Figure 4 shows the relative capacities for the benthic region. For the benthic region of the standard water bodies, DOC and biota partitioning are not significant at any Koc value; the relative fractions for DOC and biota are on the order of 10-7 to 10-5, which cannot be seen in the Koc range shown (Figure 4). At a Koc of about 9 mL/g, solute is evenly distributed between the pore-water-dissolved fraction and the sediment-sorbed fraction. At Koc values above 1,000 mL/g, the vast majority of solute in the benthic region is sorbed to sediment.

Solute holding capacity as a function of Koc

**Figure 2. Solute holding capacity as a function of Koc for the USEPA standard water bodies.**

Relative solute holding capacity of components in water column

**Figure 3. Relative solute holding capacity of individual components in water column.**

Relative solute holding capacity of components in benthic layer

**Figure 4. Relative solute holding capacity of individual components in benthic region.**

1. ***4.2 Washout and Overflow Sensitivity***

Figures 5 and 6 show how the VVWM overflow modification affects pesticide dissipation in the standard pond and standard reservoir, respectively. The effective dissipation half-life due to washout of a pesticide is shown for a range of typical annual average runoff flow rates as determined from OPP’s standard scenarios. This figure only gives an idea of the potential long-term effect of the VVWM washout addition. Short-term effects will be quite variable since washout is calculated on a daily basis, and during overflow events, the effective half-life may differ greatly from long-term averages.

Washout halflife as a function of flow rate for a water body the size of the USEPA standard pond.

**Figure 5. Effective half-life of pesticide due to washout in the standard pond as currently parameterized (1 hA area, 2 m deep). Range of flow rates are for the current standard field size (10 hA).**

Washout halflife as a function of flow rate for a water body the size of the USEPA standard reservoir.

**Figure 6. Effective half-life of pesticide due to washout in the standard reservoir as currently parameterized (5.26 hA, 2 m deep). Range of flow rates are for the current standard field size (10 hA).**

1. ***4.3 Photolysis Sensitivity***

With the above considerations, the effective photolysis rate in the standard water bodies only depends on the laboratory-measured photolysis rate, the latitude of the water body, and the reference latitude of the measured photolysis rate. The effective photolysis rate can be written in terms of these parameters. For the farm pond, the effective rate is calculated from the following equation:

 (52)

Values for the standard water bodies are given in Table 1. Given the values for standard water bodies in Table 1(a = 42.096 m-1), fatten = 0.009981 for the farm pond, fatten = 0.007286 for the reservoir. and *flat* =s 0.804 for 34°.

From equation (52) for a standard farm pond at latitude of 34° and with a reference laboratory latitude of 0°, the effective aqueous-phase photolysis rate is 124 times lower than the measured laboratory rate. For the standard reservoir at the same latitude, the rate is 170 times less than the laboratory determined value. As with hydrolysis, photolysis is assumed to act upon only dissolved forms of pesticide; therefore, the overall effective hydrolysis rate is further reduced by the factor fw in equation (5).

A plot of the inverse of equation (52) shows its effect on the half-life as given in Figure 7. This figure shows that depth is nearly proportional to the increase in half-life at the scale shown. A closer look at depth in Figure 8 shows that the direct proportional relationship begins at about 0.02 m, indicating that the photolysis has fully attenuated by this depth. Further increases in half-life are simply due to the greater amount of volume in the water column.

Photolysis sensitivity with depth of water body

**Figure 7. The effect of depth on the effective half-life due to photolysis, showing the almost proportional linear relationship of half-life with depth.**

Photolysis sensitivity with depth of water body: small scale depths

**Figure 8. Smaller scale depth figure, showing that reductions in photolysis half-life become proportional (linear) with depth after about 0.02 m.**

## Volatilization

1. The effect that wind speed has on effective half-life is given in Figure 9 for the standard pond. The figure shows that wind speed variations will have an increasingly dramatic effect as Henry’s law constant is reduced. The use of daily wind speeds in the VVWM thus has significant short-term implications (acute concentrations) for low Henry’s law compounds.
2. Volatilization as calculated by the VVWM is relatively insensitive to changes in temperature because OPP has not adopted a temperature adjustment standard for the Henry’s Law coefficient and volatilization data (as a function of temperature) required for registration. Thus, OPP currently assumes that the Henry’s Law coefficient is constant regardless of temperature.

Sensitivity of volatilization half life to wind speed and Henrys' Law constant

**Figure 9. Effect of Henry’s Law Constant and wind speed (measured at 6m) on effective volatilization half-life of aqueous phase. MW= 100, Temp = 25 °C.**

Sensitivity of volatilization half life to temperature (excluding effects of temperature on and Henrys' Law constant)

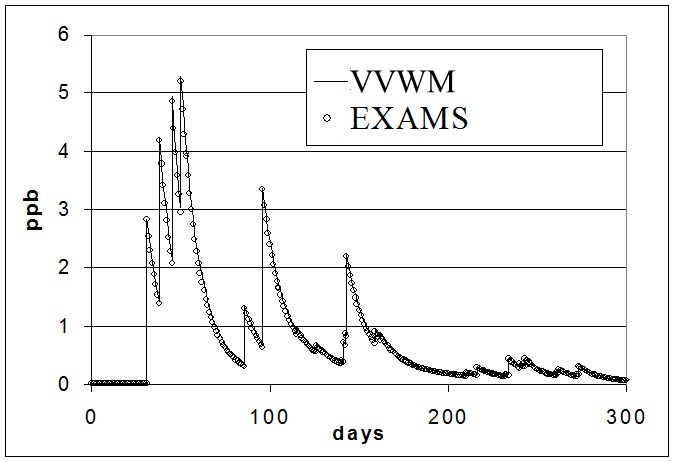
**Figure 10. Effect of Henry’s Law Constant and temperature on effective volatilization half-life of aqueous phase. The lack of temperature sensitivity is a result of not considering the effect of temperature on Henry’s Law Constant. Wind speed = 1 m/s, MW=100.**

Comparison of the volatilization between VVWM and EXAMS

**Figure 11. Comparison of the volatilization mechanisms of the VVWM and EXAMS for conditions: solubility = 100 mg/L, MW=100, vapor pressure = 0.1 torr, Koc = 1 mL/g, wind speed = 1 m/s, temperature = 25o C, and an input mass of 0.02 kg to the water column. A constant volume condition was used for the VVWM.**

1. **5 Testing and Comparison of VVWM Solution with EXAMS**

Individual processes of the VVWM analytical solution were tested by comparing the output with that of EXAMS. For these tests, a constant volume condition was imposed on the VVWM, so that only the processes common to both EXAMS and the VVWM were tested. Individual processes were tested by either zeroing out all other dissipation or making them insignificant, and using a single initial aqueous-phase input. The results from a test of the volatilization routine are shown in Figure 11. Here the analytical solution for volatilization in the VVWM is captured and correctly formulated. Other processes such as hydrolysis, photolysis, metabolism, and benthic mass transfer were tested in a similar manner, and all tested equally well. Combined processes with multiple inputs, including spray drift, erosion, and runoff, as read from PRZM output files, were also tested. An example is given in Figure 12, which shows excellent agreement with EXAMS, and further verifies the proper formulation of the processes within the VVWM.



**Figure 12. Comparison of VVWM with EXAMS for the following conditions: MW = 100, solubility = 100 mg/L, vapor pressure = 0.01 torr, aerobic half-life = 10 days, anaerobic half-life = 100 days, Koc = 100 mL/g, wind speed = 1 m/s, temperature = 25 °C, and arbitrarily selected PRZM input fluxes. A constant volume condition was used for the VVWM.**

1. ***6.2 Input Files***
2. **6.2.1 General Input File**

The input file is a text file with the structure given in Table 4. For lines that hold multiple inputs, the data is separated by a comma or space.

**Table 4. General Input File Format.**

| Line | **Fortran Variable Name** | **Type** | **Description** |
| --- | --- | --- | --- |
| 1 | output filename | character(256) | Full path and name of main output file (less suffix). This establishes the base name and location of the output files.  This also specifies the name of the \*.zts file that will be read for the mass and water flow. This input file must be named *outputfilename*.zts where *outputfilename* is the string defined by the variable outputfilename. |
| 2 | UNUSED |  |  |
| 3 | nchem | integer | 1 = parent only, 2 = parent and degradate, 3= parent, degradate 1, degradate 2 (sequential) |
| 4 | is\_koc | logical | Establishes whether the sorption coefficient is Koc or Kd; True = Koc , False = Kd |
| 5 | koc\_all(i) | real | Sorption coefficient (mL/g); the number of values should match nchem |
| 6 | aer\_aq\_all(i) | real | Water column degradation half-life (days); the number of values should match nchem |
| 7 | temp\_ref\_aer\_all(i) | real | Reference temperature for water column degradation; the number of values should match nchem |
| 8 | anae\_aq\_all(i) | real | Benthic degradation half-life (days); the number of values should match nchem |
| 9 | temp\_ref\_anae\_all(i) | real | Reference temperature for benthic degradation; the number of values should match nchem |
| 10 | photo\_all(i) | real | Photolysis half-life (days); the number of values should match nchem |
| 11 | RFLAT\_all(i) | real | Reference latitude for photolysis; the number of values should match nchem |
| 12 | hydro\_all(i) | real | Hydrolysis half-life (days); the number of values should match nchem |
| 13 | UNUSED |  |  |
| 14 | UNUSED |  |  |
| 15 | UNUSED |  |  |
| 16 | MWT(i) | real | Molecular Weight; the number of values should match nchem |
| 17 | VAPR\_all(i) | real | Vapor Pressure (torr); the number of values should match nchem |
| 18 | SOL\_all(i) | real | Solubility (mg/L); the number of values should match nchem |
| 19 | xAerobic(i) | real | Molar Conversion Factor for water column degradation; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 20 | xBenthic(i) | Real | Molar Conversion Factor for benthic degradation; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 21 | xPhoto(i) | Real | Molar Conversion Factor for photolysis; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 22 | xHydro(i) | real | Molar Conversion Factor for hydrolysis; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 23 | UNUSED |  |  |
| 24 | UNUSED |  |  |
| 25 | UNUSED |  |  |
| 26 | Henry\_unitless(i) |  | Unitless Henry’s Law constant; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 27 | Heat\_of\_Henry(i) |  | Enthalpy of phase transformation from aqueous solution to air solution; the number of values should match (nchem-1): parent to degradate 1, degradate 1 to degradate 2 |
| 28 | QT | real | Q10 factor by which degradation increases for every 10 °C rise in temperature. |
| 29 | scenario\_id | Character(50) | Text to describe the field scenario. Used for naming output files. |
| 30 | metfilename | Character(256) | Full path and file name of the meteorological file. |
| 31 | Text\_latitude |  | Latitude of scenario |
| 32 | UNUSED |  |  |
| 33 | UNUSED |  |  |
| 34 | burialflag | logical | If set to .TRUE. This will activate pesticide removal by sediment mass balance. Sediment flow in will equal sediment flow out. |
| 35 | UNUSED |  |  |
| 36 | UNUSED |  |  |
| 37 | UNUSED |  |  |
| 38 | UNUSED |  |  |
| 39 | D\_over\_dx | real | Mass transfer coefficient (m/s) as defined by D/Δx in Eqn . 46 |
| 40 | is\_calc\_prben, PRBEN | Logical, real | TRUE means use the varying distribution method, FALSE means use fixed distribution; Xd in equation 40 and 41, is dummy number if is\_calc\_prben set to TRUE |
| 41 | benthic\_depth | real | Depth of benthic region (m) |
| 42 | porosity | real | Porosity of benthic region (--) |
| 43 | bulk\_density | real | Bulk density of benthic region (g/mL). Mass of solids per total volume. |
| 44 | FROC2 | real | Fraction of organic carbon on sediment in benthic region. |
| 45 | DOC2 | real | Concentration of dissolved organic carbon in benthic region (mg/L) |
| 46 | BNMAS | real | Areal concentration of biosolids in benthic region (g/m2) |
| 47 | DFAC | real | Photolysis parameter defined in eqn. 23 |
| 48 | SUSED | real | Suspended solids concentration in water column (mg/L) |
| 49 | CHL | real | Chlorophyll concentration in water column (mg/L) |
| 50 | FROC1 | real | Fraction of organic carbon on suspended sediment in water column. |
| 51 | DOC1 | real | Concentration of dissolved organic carbon in water column (mg/L) |
| 52 | PLMAS | real | Concentration of biosolids in water column (mg/L) |
| 53 | UNUSED |  |  |
| 54 | UNUSED |  |  |
| 55 | direct\_crapping | real | For FDA, this is continuous daily applied mass directly to waterbody (kg) |
| 56 | napp | integer | Number of spray drift events that will be used to apply pesticide mass to pond |
| 57 | appdate\_sim\_ref(i) | integer | Dates of spray drift events reference to days of the simulation (first day of simulation = 1) |
| 58 | simtypeflag | integer | Flag to identify the type of water body: 1= User defined parameters; 2=USEPA Pond; 3=USEPA Reservoir; 4 = constant vol no flow; 5 = constant vol w/flow |
| 59 | afield | real | Area of adjacent runoff producing field. This is used to convert area-normalized pesticide mass in the mass-input file to actual mass (m2). |
| 60 | area | real | Area of water body (m2). |
| 61 | depth\_0 | real | Depth at which the input concentrations of physical parameters (e.g., suspended solids, CHL., etc) were measured. |
| 62 | depth\_max | real | Maximum depth that water can rise before overflow (m). |
| 63 | spray(i) | real | Mass of pesticide (kg) delivered from spray drift corresponding to dates of appdate\_sim\_ref(i) |
| 64 | flow\_averaging | integer | Number of days that are used to average the influent water flow. If = 0, then the flow rate to be used in the program is the average flow rate of the entire simulation. |
| 65 | baseflow | real | Provids an additional constant flow through the waterbody m3/s |
| 66 | Cropped fraction | real | Holds the Fraction of Cropped Area. Of the watershed. Only used so that it is recorded in the output. Program does not use these values for calculations |
| 67 | is\_hed\_files\_made | logical | TRUE will produce the files specified in lines 75 to 80 |
| 68 | is\_add\_return\_frequency, additional\_return\_frequency | logical, real | Alternative return frequency to be used in addition to the 10 year return default for output |
| 69 | outputfile\_parent\_daily | text | Output file name |
| 70 | outputfile\_deg1\_daily | text | Output file name |
| 71 | outputfile\_deg2\_daily | text | Output file name |
| 72 | outputfile\_parent\_analysis | text | Output file name |
| 73 | outputfile\_deg1\_analysis | text | Output file name |
| 74 | outputfile\_deg2\_analysis | text | Output file name |
| 75 | outputfile\_parent\_deem | text | Output file name |
| 76 | outputfile\_deg1\_deem | text | Output file name |
| 77 | outputfile\_deg2\_deem | text | Output file name |
| 78 | outputfile\_parent\_calendex | text | Output file name |
| 79 | outputfile\_deg1\_calendex | text | Output file name |
| 80 | outputfile\_deg2\_calendex | text | Output file name |
| 81 | outputfile\_parent\_esa | text | Output file name |
| 82 | outputfile\_deg1\_esa | text | Output file name |
| 83 | outputfile\_deg2\_esa | text | Output file name |

1. **6.2.2 ZTS Input File**

The ZTS file contains daily mass inputs, water flows, and sediment deliveries. The ZTS file is automatically created by the PRZM model or it may be manually created. It must be named as:

***inputfilename*.zts**

where *inputfilename* is the same as that used above for the Input File and likewise specifies the full path and name of the file. The ZTS file has a format as shown in Table 5. Each line (except the first three) represents the daily values for each input variable. Data on each line may be separated by a space or comma. The number of data lines in the file must correspond to the number of days in the meteorological file.

**Table 5. ZTS File Format.**

| **Line #** | **Data** |
| --- | --- |
| 1 | not read |
| 2 | not read |
| 3 | not read |
| 4 | X, X, X, Q, B, MRp, MEp, MR1, ME1, MR2, ME2 |
| .  .  . | .  .  . |
| N | X, X, X, Q, B, MRp, MEp, MR1, ME1, MR2, ME2 |

Where

N refers to the last line in the ZTS file. It corresponds to the number of records in the meteorological file.

X is dummy data that is not used, but must be in place. In a PRZM-generated ZTS file, these are the year, month, and day values.

Q is the daily water per field area that flows into the water body (cm/ha/day). This is used for calculating washout and volume changes of the water body if these options are chosen.

B is the daily solids that enter the water body (tonnes/day) and is used for sediment mass balance if that option is chosen.

MRp is mass of pesticide per field area entering water body by runoff (g/cm2/day)

MEp is mass of pesticide per field area entering water body by erosion (g/cm2/day)

If degradate 1 is being simulated (nchem >1), then the following would be entered:

MR1 is mass of degradate 1 per field area entering water body by runoff (g/cm2/day)

ME1 is mass of degradate 1 per field area entering water body by erosion (g/cm2/day)

If degradate 2 is being simulated (nchem =2), then the following would be entered:

MR2 is mass of degradate 2 per field area entering water body by runoff (g/cm2/day)

ME2 is mass of degradate 2 per field area entering water body by erosion (g/cm2/day)

1. **6.2.3 Weather Files**
2. ***6.2.3.1 Old DVF weather files***

The meteorological file is specified in line 30 of the input file. This file has the same formatting as that required by PRZM. These files must have the extension *dvf*. The fortran formatting for each line is:

1X, 3I2, 4F10.0

With the input variable of: MM, MD, MY, PRECIP, PEVP, TEMP, WIND

where

MM = meteorological month

MD = meteorological day

MY = meteorological year

PRECIP = precipitation (cm/day)

PEVP = pan evaporation data (cm/day)

TEMP = temperature (°C)

WIND = wind speed (cm/sec)

Example Partial Meteorological File:

010161 0.00 0.30 9.5 501.6 240.3

010261 0.10 0.21 6.3 368.0 244.3

010361 0.00 0.28 3.5 488.3 303.0

The meteorological file determines the simulation time. The simulation will start at the first date and end with the last date in this file. Dates must be continuous in the file. The file does not have to start or end on any particular calendar date; the program accepts partial years.

1. ***6.2.3.1 New WEA weather files***

VVWM accept files formatted per Fry et al. (2016). These files must have the extension *wea*.

1. ***6.3 Output Files***
2. **6.3.1 Regulatory Summary Output File**

A summary file that contains USEPA regulatory values for concentration is produced for each chemical simulated and is named:

***outputfilename*\_*scenario\_ID*\_*waterbodytext*\_*Parent-Degradate*.txt**

where

*outputfilename* - as specified in Line 1 of input file.

*scenario\_ID* - as specified in Line 29 of input file.

*waterbodytext* - Depending on the water body simulated, this will be "Custom", "Pond", or "Reservoir" if *simtypeflag* (Input Line 57) = 1, 2, or 3, respectively

*Parent-Degradate* - This will be "Parent", "Degradate1", or "Degradate2" and indicates which of the products are contained in the file.

The content of this output file contains frequency occurrences of concentrations (e.g., 1-in-10 year values). These values are calculated to approximate the return frequency of concentrations and are based on calendar years (Jan 1 to Dec 31). Chronic concentrations calculated by averaging days that may span across years can be problematic in that chronic concentrations could be longer than peak concentrations for the same year. This is because the method used here, which may be accurate for short term weather events, is only an approximation for events that last several days like the chronic concentrations.

The frequency of return is calculated by separating concentrations into calendar years. For chronic concentrations, each daily concentration is calculated with a backward average, and thus some days may include influences from previous calendar years. For each calendar year, the highest concentration for that year is determined. These concentrations are then ordered from high to low and then the appropriate percentile is taken. The position in the order is calculated by:

Where R is the return years

N is the number of years

For example, if the return frequency is 10 years and there are 30 years in the simulation, then

The value used would be the one ranked in the 27.9 position in the order, with linear interpolation used to address the fractional part.

1. **6.3.2 Daily Values Output File**

An output file that contains the daily values for water body depth, water column concentration, and benthic pore water concentration is created with the name:

***outputfilename*\_*scenario\_ID*\_*waterbodytext*\_*Parent-Degradate\_daily*.txt**

# Computer Implementation

PRZM5 is written with the standards of Fortran 2003. Subroutines have been grouped into modules with intuitive names (e.g., Erosion, Pesticide Application, and Irrigation). This should facilitate locating relevant routines during maintenance. The Main Program (PRZM5) is in the file PRZM5.f90. The main calls the input file read routines (PRZMRD\_PRZM5) and then sets up a daily time loop. The daily time loop reads a line from the meteorological file and then runs a day of the hydrology, plant growth, and transport routines. Daily output is delivered to a \*.zts file, where desired output is specified by the user in the input file. A file that records run status and errors is created as kecho.prn. All files are delivered to the default working directory.

To run PRZM5, the command line is as follows:

prompt> PRZM5.exe *inputfilepath*

Where the *inputfilepath* argument is optional. If *inputfilepath* is specified, then it will be the default path for the input file and where the output will be delivered. If unspecified, then the default working directory is used. The input file must be named "PRZM5.inp".

**Example command:** the following command from a command prompt (C:>) would run the PRZM5.exe file that is in the *C:\models\* folder and use the *PRZM5.inp* file that is in the *C:\models\testrun\* directory:

C:> c:\models\Przm5.exe c:\models\testrun\

If spaces are used in the path name, then the argument must be enclosed in quotation marks, as follows:

C:> c:\models\Przm5.exe “c:\My Models\testrun\”

***6.1 Executable and the Command Line***

Running the VVWM requires the executable and three input files: a general input file, a “ZTS” file, and a meteorological file. The executable is run from a command line with the following command:

**fortranvvwm.exe “*inputfilename”***

where *fortranvvwm.exe* is the name of the executable, and *inputfilename* is a command line argument that specifies the path and name of the **General Input File**. For example,

C:\> fortranvvwm.exe “C:\My Documents\Test\MyFirstInputFile.txt”

In this case, the fortranvvwm.exe file is located on the C: directory and the input file is named MyFirstInputFile.txt and located in the C:\My Documents\Test\ directory. Note: Quotation marks around the command line argument are necessary if there are any blank spaces in the argument.

# References

Arya, S.P.S., 1988. Momentum and heat exchanges with homogeneous surfaces. Pages 157-181, in *Introduction to Micrometeorology*. Academic Press, San Diego, California, USA.

Blakey J.F., 1966. Temperature of surface waters in the conterminous United States. US Geological Survey Hydrologic Investigations Atlas HA-235, Washington, DC.

Brutsaert, W., 1982. Evaporation into the Atmosphere: Theory, History, and Applications. D. Reidel Publishing Co., Dordrecht, Holland.

Burns, L.A., L.A. Suárez, and L. Prieto, 2005. United States Meteorological Data: Daily and

Hourly Files to Support Predictive Exposure Modeling. U.S. Environmental Protection

Agency, Athens, Georgia.

Collins, W.D., 1925. Temperature of water available for industrial use in the United States: U.S. Geol. Survey Water-Supply, Paper 520-F, p. 97-104.

de Vries, D.A., 1963. Thermal properties of soils, Chapter 7 in Physics of Plant Environment, edited by W. R. van Wijk, p. 210-235.

Dickenson, RE (1984) modeling evapotranspiration for the three-dimensional global climate models. In Climate Processes and Climate Sensitivity” Geophysical Monograph 29, Vol 5

Douglas J.J. and B.F. Jones, 1963. On predictor-corrector methods for nonlinear parabolic equations. J. SIAM, 11, 195-204.

Fry, M.M., Rothman, G., Young, D.F., and Thurman, N., 2016. Daily gridded weather for exposure modeling, *Environmental Modelling & Software*, 82, 167-173, doi.org/10.1016/j.envsoft.2016.04.008

Hanks, R. J., D. D. Austin, and W. T. Ondrechen, 1971. Soil temperature estimation by a numeric method. Soil Sci. Soc. Am. J., 35(5), p. 665-667.

Jury, W. A., R. Grover, W. F. Spencer, and W. J. Farmer, 1983a. Behavior assessment model for

trace organics in soil: I. Model description. Journal of Environmental Quality **12**:558-

564.

Jury, W. A., W. F. Spencer, and W. J. Farmer, 1983b. Use of models for assessing relative volatility, mobility, and persistence of pesticides and other trace organics in soil systems. Pages 1-43 in J. Saxena, editor. Hazard Assessment of Chemicals. Academic Press, New York, New York, USA.

Kilmer, V. J., 1982. Handbook of Soils and Climate in Agriculture. CRC Press, Inc., Boca Raton,

Florida, USA.

Kinnell, P.I.A., 2004. Letter to the Editor on “The Mathematical Integrity of Some Universal Soil Lass Equation Variants”, Soil Science Society of America Journal, 68(1), 336-337.

Knisel, W. G., R. A. Leonard, and F. M. Davis, 1994. GLEAMS: Version 2.10 -- Part I: Nutrient Component Documentation. U.S. Department of Agriculture, Agricultural Research Service, Tifton, Georgia, USA, http://www.tifton.uga.edu/sewrl/Gleams/glmspub.htm

Mehlenbacher, L. A., and D. W. A. Whitfield, 1977. Modelling thermal eddy diffusivity at canopy height. Boundary-Layer Meteorology 12:153-170.

NOAA, 1982. Evaporation Atlas for the Contiguous 48 United States, NOAA Technical Report NWS 33, U.S. Dept of Commerce, National Oceanic and Atmospheric Administration, Washington DC.

NRCS, 1986. Urban Hydrology for Small Watersheds, TR-55, Natural Resources Conservation Service, United States Department of Agriculture, Washington DC.

NRCS, 2003. National Engineering Handbook Section 4: Hydrology. Natural Resources Conservation Service, United States Department of Agriculture, Washington DC.

Rosenberg, N. J., 1974. Microclimate: The Biological Environment. Wiley Interscience, New York, New York, USA.

Soil Conservation Service, 1986. Urban Hydrology for Small Watersheds. Technical Release TR-55, U.S. Department of Agriculture Soil Conservation Service, Washington, DC, USA.

Stamper, J. H., H. N. Nigg, and J. C. Allen, 1979. Organophosphate insecticide disappearance from leaf surfaces: an alternative to first-order kinetics. Environmental Science & Technology 13:1402-1405.

Stanhill, G., 1969. A simple instrument for the field measurement of turbulent diffusion flux. Journal of Applied Meteorology 8:509-513.

Stewart, B. A., D. A. Woolhiser, W. H. Wischmeier, J. H. Caro, and M. H. Frere, 1975. Control of Water Pollution from Cropland: Volume I - A Manual for Guideline Development. EPA-600/2-75-026a, U.S. Environmental Protection Agency, Environmental Research Laboratory, Athens, Georgia, USA.

Stewart, B. A., D. A. Woolhiser, W. H. Wischmeier, J. H. Caro, and M. H. Frere, 1976. Control of Water Pollution from Cropland: Volume II: - An Overview. EPA-600/2-75-026b, U.S. Environmental Protection Agency, Environmental Research Laboratory, Athens, Georgia, USA.

Thibodeaux, L. J., 1996. Environmental Chemodynamics: Movement of Chemicals in Air, Water, and Soil, 2nd ed., John Wiley & Sons, Inc. New York, NY, p. 246-250, 424-428.

USDA, 1986. Urban Hydrology for Small Watersheds, TR-55. United States Department of Agriculture, Technical Release 55. Natural Resources Conservation Service.

USDA, 1991. Soil-Plant-Water Relationships, in *National Engineering Handbook, Section 15 Irrigation*. *NEH* 15-1, 2nd Edition. United States Dept. of Agriculture.

USDA, 2004. *National Engineering Handbook,* NEH -4 Part 630*,* Chapter 11 Snowmelt. Natural Resources Conservation Service, United States Dept. of Agriculture.

van Bavel, C. H. M. and D. I. Hillel, 1976. Calculating potential and actual evaporation from a bare soil surface by simulation of concurrent flow of water and heat, *Agricultural Meteorology*, 17, p. 453-476.

van Wijk, W. R., editor, 1963. Physics of Plant Environment. North-Holland Publishing Co., Amsterdam, The Netherlands.

Wagenet, R. J., and J. L. Hutson, 1987. Leaching Estimation and Chemistry Model. Cornell University, Ithaca, New York, USA.

Williams, J. R., 1975. Sediment yield prediction with universal equation using runoff energy factor. Pages 244-252 in Present and Prospective Technology for Predicting Sediment Yields and Sources (ARS-S-40). U.S. Department of Agriculture Sedimentation Laboratory, Oxford, Mississippi, USA.

Wischmeier, W. H., and D. D. Smith, 1978. Predicting rainfall erosion losses - a guide to conservation planning. *Agriculture Handbook 53*7, U.S. Department of Agriculture, Washington, DC, USA.

Young, D.F. and J. N. Carleton, 2012. Estimating dispersivity, mass recovery and water hold up in field-scale leaching studies by use of a capacity model, Abstract H11B-1173, presented at 2012 Fall Meeting, AGU, San Francisco, CA, 3-7 Dec.

Young, D.F. and J.N. Carleton, 2006. Implementation of a Probabilistic Curve Number Method in the PRZM Runoff Model. *Environmental Modelling & Software*, 21(8), 1172-1179.

Young, D.F. and M.M. Fry, 2019. Field-scale evaluation of pesticide uptake into runoff using a mixing cell and a non-uniform uptake model. *Environmental Modelling & Software*. 122, 104055.

1. **8 References**

Banks, R. B., 1975. Some Features of Wind Action on Shallow Lakes*. Journal of the Environmental Engineering Division., ASCE*. 101(EE5), 813-827.

Burns, L.A., Cline, D.M., and Lassiter, R.P., 1982. Exposure Analysis Modeling System (EXAMS): User Manual and System Documentation. EPA-600/3-82-023, U.S. EPA.

Burns, L.A., 1997. Exposure Analysis Modeling System (EXAMS II) Users Guide to Version 2.97.5, EPA/600R-97/047, U.S. EPA

Carsel, R., J. Imhoff, P. Hummel, J. Cheplick, and A. Donigan, 1997. PRZM 3.1 User’s Manual, National Exposure Research Lab, Office of Research and Development, U.S. Environmental Protection Agency, Athens, Georgia.

Burns, L.A., 1985. Models for predicting the fate of synthetic chemicals in aquatic ecosystems, in: *Validation and Predictability of Laboratory Methods for Assessing the Fate and Effects of Contaminants in Aquatic Ecosystems*, ASTM STP 865, T.P. Boyle, Ed., American Society of Testing Materials, Philadelphia, pp 176-190.

Cusler, E.L., 1984. *Diffusion: Mass Transfer in Fluid Systems*, Cambridge University Press, New York

Fry, M.M., Rothman, G., Young, D.F., and Thurman, N., 2016. Daily gridded weather for exposure modeling, *Environmental Modelling & Software*, 82, 167-173, doi.org/10.1016/j.envsoft.2016.04.008

Liss, P.S., 1973. Processes of Gas Exchange Across an Air-Water Interface. *Deep Sea Research*, 20(3), 221-238.

Shwarzenbach, R.P., Gschwend, P.M., and Dieter, D.M., 1993. *Environmental Organic Chemistry*, John Wiley & Sons, New York.