Emissions reductions from harvested wood products and management residuals

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1 California Forest Management Emissions Profile

Forest management activities in California produce logs for lumber markets while maintaining and enhancing forest health. In addition to merchantable logs, these harvest activities produce logging residuals that are either left in the stand to decompose, or piled and burned as directed by forest practice rules (California Forest Practice Rules, Article 7 § 917.2). By utilizing residual wood biomass produced from these forest management activities, there is potential to reduce emissions from greenhouse gas (GHG) and other climate pollutants.

Currently, the majority of biomass produced from forest management activities end up in residual biomass material that is either left in the woods to decompose, or aggregated at a landing where it is eventually burned. When considered alongside the accumulation of woody material from historic fire suppression activity, thire exists a heightened volume of combustible woody material in excess of historic reference conditions which elevates the risk of damaging wildfire in much of California's forestland. Common practice for fuel load management in California forests include prescribed natural fire and sanitation pile burning. However, as demonstrated by previous studies, prescribed natural fire is often only an effective tool for reducing fuel loading and maintaining fire resilient landscapes when coupled with mechanical treatment to remove biomass (Stephens et al 2009). Open pile burning of residual biomass can result in higher emissions of strong radiative forcing agents (black carbon) and criteria air pollutants (PM, NOX) when compared to controlled combustion in biomass power plants with modern emissions control technology.

Ultimately, any combustion or decomposition of residual material results in emissions of greenhouse gases (GHG), criteria air pollutants (CAP), and short-lived climate pollutants (SLCP). Even in the absence of forest management activity, atmospheric emissions are produced from stochastic processes such as wildfire, pest, and disease outbreaks. As such, utilization strategies are necessary to reduce the air quality impact of common forestry practice. Furthermore, the opportunity cost of residual biomass use in bioenergy and applications weigh in favor of alternative utilization strategies. Figure 1 presents an overview of emissions and emissions reduction pathways for wood from California's forests.

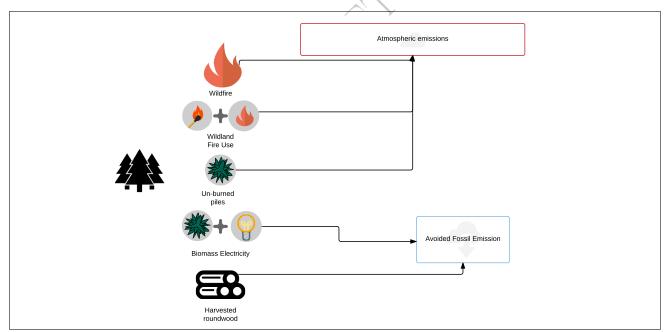


Figure 1: Overview of fates of wood resulting from harvest and mortality in California forests. Note that time is not represented in this figure.

The focus of this analysis is on deriving emissions associated with **management activity**. This report does not assess greenhouse gas emissions from pest or disease induced mortality, which is estimated at approximately 34 MMT CO2e annually in California forests Christensen et al. (2016). Emissions from mortality are indirectly related to management activities just as wildfire is, and must be accounted for to comprehensively evaluate the climate impacts of harvesting.

To estimate emissions affected by management activity, it is first necessary to identify the following steps:

1. Estimate CO2 equivalent emissions from burning forest management residuals using criteria pollutant and GHG emissions inventory published by the California Air Resources Board (CARB)

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- 2. Estimate the volume and fate of wood removed, left in the forest, and burned as a result of direct anthropogenic management activities.
- 3. Establish life-cycle displacement factors (DF) for all utilized wood and apply DF to harvested wood to obtain an aggregate estimate.

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Quantifying the climate effects of wood products and forest management residuals is important to the development of the Forest Climate Plan $(FCP)^1$ as well as efforts underway by the California Board of Forestry and CalFire to meet the intent of AB 1504 $(2010)^2$. To inform these efforts, this report provides estimates of the following:

- GHG and SLCP emissions produced from the combustion or decomposition of logging residuals.
- GHG emissions reductions from the use of wood products harvested in the state.

Estimates are based empirical data and reflect past forest management activities. It is **critical** to note that the empirical data used in this analysis reflect point-in-time measures that are affected by a dynamic system of climate, growth, and mortality as well as macroeconomic and policy forces. This analysis may provide insight into opportunities to more effectively utilize woody biomass residuals from current forest management activities to reduce emissions.

1.1 Key Findings

- Baseline GHG and SLCP emissions from burning of forest management residuals can be estimated and should be considered in any forest management emissions baseline.
- Total emissions from pile burning of forest management residuals (including SLCP and GHG components) extrapolated from CARB emissions inventory are estimated at 1.16 MTCO2e
- Wood harvested in California in 2012 resulted in avoided emissions of 2.29 MMTCO2e
- Logging residuals not used in bioenergy production contributed annual emissions of:
 - XXX MMTCO2e resulting from anthropogenic burning of logging residuals
 - XXX MMTCO2e resulting from decomposition of logging residuals left unburned
- Un-utilized slash from non-commercial management activities on National Forest System lands contributed emissions of XXX MMTCO2e
- Forest Inventory and Analysis re-sample data has been used in the southeast to quantify removals resulting from non-commercial management activity and could be used for this purpose in California
- The Prescribed Fire Information Reporting System (PFIRS) may be a useful tool for quantifying emissions from pile burns and prescribed fire. It is a requirement that prescribed fires and pile burns on National Forest System Lands are reported through PFIRS. However, California Air Quality Management Districts are not required to report emissions through this system at this time. Therefore, it is not possible to associate burns in the PFIRS with commercial harvest activities.

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¹The Forest Climate Action Team (FCAT) was assembled in August of 2014 with the primary purpose of developing a Forest Carbon Plan by the end of 2016. FCAT is comprised of Executive level members from many of the State's natural resources agencies, state and federal forest land managers, and other key partners directly or indirectly involved in California forestry. FCAT is under the leadership of CAL FIRE, Cal-EPA, and The Natural Resources Agency.

 $^{^2\}mathrm{AB}\text{-}1504$ Forest resources: carbon sequestration. (2009-2010)



2 Estimating CO2 Equivalent Emissions from In-Forest Biomass Combustion

The California Air Resources Board (CARB) reports on emissions from forest biomass burning in current Greenhouse Gas (GHG) and Criteria Air Pollutant (CAP) emissions inventories. Both are necessary resources for establishing aggregate annual climate-forcing emissions (Figure 2). The GHG inventory captures gasses with radiative forcing properties including CO2 and CH4, but does not capture elemental carbon or black carbon (BC) emissions which also have strong radiative forcing properties (Table 1). The California Air Resources Board (2015, 2016) CAP report captures SLCP emissions from wildfire (80.52 MMTCO2e) and prescribed fire (3.66 MMTCO2e) from which black carbon emissions may be estimated. However, no reference in the CAP report is made to the source of these SLCP estimates. When viewed in aggregate, a comprehensive reporting of total climate impact from anthropogenic burning may be estimated.

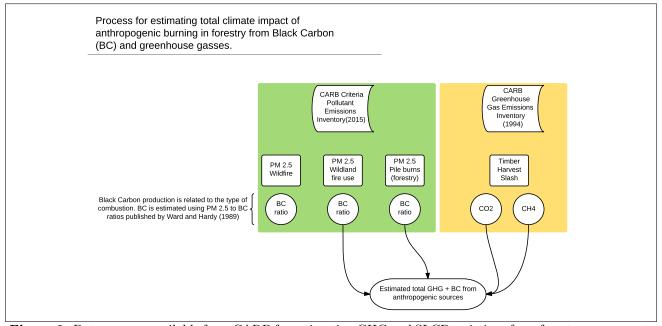


Figure 2: Data sources available from CARB for estimating GHG and SLCP emissions from forest management.

Table 1. Italige of Global Warning I ofendal (GWI) varies for Diack Carbon.								
GWP_{20}	$\text{GWP}\sigma_{20}$	GWP_{100}	$\text{GWP}\sigma_{100}$	GWP_{500}	$\mathrm{GWP}\sigma_{500}$	Source		
2200.0	888.82	633.33	255.41	193.33	77.67	Fuglestvedt2010		
3200.0		900.0				CaliforniaAirResourcesBoard2015		

Table 1: Range of Clobal Warming Potential (CWP) values for Black Carbon

2.1 Estimating Black Carbon Emissions from Biomass Burning

Black carbon is not directly reported by statewide emissions summaries. The 2015 Criteria air pollutant (CAP) emissions estimates published by the California Air Resources Board reports on particulate matter (PM 2.5) of which black carbon is a component. CARB reports annual PM 2.5 emissions in tons/day as shown in 2.

Source	$PM 2.5 (t y^{-1})$
ALL VEGETATION	137630.15
FOREST MANAGEMENT	5480.51
WILDLAND FIRE USE (WFU)	6802.43

Black Carbon is a fraction of the Total Carbon (TC) component of PM 2.5. Using estimated ratios of smoldering to flaming combustion for hand/machine piled burns, prescribed natural fire and wildfire from Ward and Hardy (1989), Black Carbon emissions can be calculated from PM 2.5 with the following equations Eq. (1)

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Table 2: Emissions of PM 2.5 in 2015 as reported by CARB

Source	PM 2.5 (t)
ALL VEGETATION	137630.15
FOREST MANAGEMENT	5480.51
WILDLAND FIRE USE (WFU)	6802.43

$$BC = (PM_{2.5} \times F \times TC_f \times BC_f) + (PM_{2.5} \times S \times TC_s \times BC_s)$$
(1)

where:

BC = Black Carbon (mass units)

 $PM_{2.5} = PM_{2.5}$ (mass units)

F =Percent of combustion in flaming phase

 $TC_f = \text{Total Carbon fraction of } PM_{2.5} \text{ for flaming phase}$

 $BC_f = Black Carbon fraction of Total Carbon for flaming phase$

S =Percent of combustion in smoldering phase

 TC_s = Total Carbon fraction of $PM_{2.5}$ for smoldering phase

 $BC_s = \text{Black Carbon fraction of Total Carbon for smoldering phase}$

Based on Ward and Hardy (1989) and Jenkins et al. (1996), the following ratios are used herein.

Table 3: Peter, I feel like there is too much going on in this table. Factors used for calculating Black Carbon (BC) emissions from the three primary combustion sources. BC is a fraction of Total Carbon (TC) which is a

fraction of total PM 2.5. Coefficients of variation (C_v) are reported here as well.

Source	$\mathrm{BC_f}\ \mathrm{t^{\text{-}1}}\ \mathrm{PM}$	$\mathrm{TC_f^{Cv}}$ t ⁻¹ PM	$\mathrm{BC_f^{Cv}}$ t ⁻¹ TC	$\mathrm{BC_s}$ t ⁻¹ PM 2.5	$\mathrm{TC_{s}^{Cv}}$ t ⁻¹ PM	$\mathrm{BC_{s}^{Cv}}$ t ⁻¹ TC
Pile Burn	0.046904	0.09	0.45	0.01624	0.01	0.49
Prescribed	0.08016309	0.0733	0.5833	0.020944	0.08	0.29
Wildfire	0.05870124	0.0867	0.4467	0.0228641	0.06	0.338
			y			
Source	$\mathrm{BC_f}\ \mathrm{t^{1}}\ \mathrm{PM}$	$\mathrm{TC_f^{Cv}}$ t ⁻¹ PM	$\mathrm{BC_f^{Cv}}$ t ⁻¹ TC	$\mathrm{BC_s}$ t ⁻¹ PM 2.5	$TC_{\rm s}^{\rm Cv}$ t ⁻¹ PM	$\mathrm{BC_{s}^{Cv}}$ t ⁻¹ TC
Source Pile Burn	$\frac{{ m BC_f\ t^{-1}\ PM}}{0.046904}$	${\rm TC_f^{Cv}\ t^{-1}\ PM} \over 0.09$	$\frac{\rm BC_f^{Cv}\ t^{-1}\ TC}{0.45}$	${ m BC_s} \ { m t^{-1}} \ { m PM} \ 2.5$ 0.01624	$TC_{\rm s}^{\rm Cv} \ {\rm t}^{\text{-1}} \ {\rm PM} \ 0.01$	$\frac{\rm BC_{s}^{Cv}\ t^{-1}\ TC}{0.49}$
		1	i	-	ь	

To arrive at a rough estimate of BC emissions based on PM2.5, the following steps are taken:

- 1. Determine the amount of PM2.5 produced in the flaming and smoldering phases of combustion for each type (piles, prescribed, wildfire). Ratios from Ward and Hardy (1989), 3 are used.
- 2. Define 1000 normal probability distributions using the coefficient of variation from Table 3 to determint the percent of PM2.5 comprised of carbonaceous material (TC) and percent of TC comprised of black carbon (BC). ???Give estimates and coefficient of variation estimates provided by Ward and Hardy (1989), tables 2 and 3.???
- 3. Estimate annual BC emissions reg:tab:carb $_{bc}$ based on probability distributions defined in step 2.

Source	$PM 2.5 (t y^{-1})$	$BC (t y^{-1})$	GWP (t y^{-1})
ALL VEGETATION	137630.15	11225.85	35922719.54
FOREST MANAGEMENT	5480.51	346.06	1107396.54
WILDLAND FIRE USE (WFU)	6802.43	687.77	2200877.13

Given the variance in baseline assumptions of BC volume 3, it is critical to identify the minimum and maximum range of probable emissions volumes. The following plot represents estimates of total BC emissions resulting from combustion of biomass in the CARB CAP emissions categories reflecting woody biomass combustion in wildfire, pile burning, and prescribed natural fire.

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Table 4: Annual black carbon			
Source	$PM 2.5 (t y^{-1})$	$BC (t y^{-1})$	GWP (t y^{-1})
ALL VEGETATION	137630.15	11225.85	35922719.54
FOREST MANAGEMENT	5480.51	346.06	1107396.54
WILDLAND FIRE USE (WFU)	6802.43	687.77	2200877.13

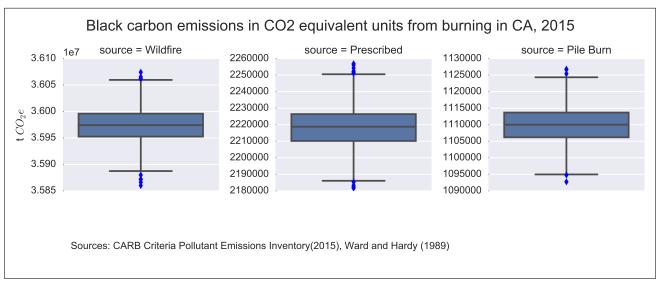


Figure 3: Short-lived climate pollution from open burning of biomass as reported by CARB criteria pollutant emissions inventory.

2.2 Estimating GHG Emissions from Biomass Burning

To estimate GHG emissions from pile burning, we can use the ratio of PM2.5 to CO2 and to CH4 ?? from the Piled Fuels Emissions Calculator. The following ratios are then used to estimate GHG emissionsn from CARB-reported PM emissions.

Pile Biomass (t)	Consumed Biomass (t)	PM2.5 (t)	CO2(t)	CH4 (t)
1.360178	1.224161	0.008263	2.0366	0.0034

In addition, the [[http://www.arb.ca.gov/cc/inventory/archive/tables/net_co2_flux_2007-11-19.pdf][CARB greenhouse gas emissions inventory]] estimates emissions from wildfire and slash burning(<- peter, how is this different to pile burns?) up through 2004 (Table 5).

Table 5: Annual GHG Emissions estimated from CARB GHG emissions inventory Source Category Average annual emissions 1994-2004 MMTCO $_{2e}$ Forest and rangeland fires 2.0194

Timber harvest slash 0.155266666666667

2.3 Estimating Total Emissions from Biomass Burning

To arrive at an estimate of total CO2 equivalent (CO2e) emissions in 2015 from burning of forest management residuals using published CARB estimates, we combine the CO2 emissions reported for 2004 in the LULUC <- where is this from??? Biodegradable Carbon Emissions and Sinks with black carbon emissions extrapolated from the CARB Criteria Air Pollutant Emissions inventory estimates. The time discrepancy between the 2004 and 2015 is acknowledged as an irreconcilable source of uncertainty in this estimation. Further model-based estimation could be used to derive a ratio of GHG to PM using the USFS CONSUME model. This analysis does however show that a baseline of substantial emissions from forest management residuals has been reported in CARB emissions inventories and should be recognized as a baseline condition. We find that a rough estimate of CO2e emissions from pile burning annual approaches 1 Mt CO2e ??.

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Table 6: Estimated average annual CO2 equivilant emissions by source and emission type.

$\mathrm{sc}_{\mathrm{cat}}$	avg(mmtco2e)
Forest and rangeland fires	2.0194
Timber harvest slash	0.1552666666666667

	${ m Mt~CO2e}$	Source
0	0.17	CO2 pile burning
1	0.99	CO2e BC pile burning
2	1.16	Total Mt CO2e

BC emissions in terms of CO2e has not been included in any GHG emissions inventory published by CARB. Remove?

3 Estimating Emissions Impact from Utilization of Harvested Wood

Wood harvested from California's forests are utilized in a variety of construction, landscaping, and consumer products. During the manufacture of these products, this wood is fractionated through a multi-stage process of harvesting, processing, and utilization to reside in several residual biomass fates (below) for which the time horizon of carbon return to the atmosphere varies widely.

Logging Residuals Tops, limbs, and sub-merchantable material produced from harvest activities in the woods

Processing (Mill) Residuals Sawdust, shavings, bark, and off cuts from primary and secondary manufacturing.

Construction Debris Fraction of wood used in construction or finished products that are not integratrated into its final form.

Demolition Wood used in construction that has reached the end of its useful life.

Each category has multiple potential fates which can greatly influence the net emissions impact attributed to the initial forest management activity. While wood products used in construction or finished products may sequester carbon in a stable environment for a long period, residues sent to landfills or left in the woods as slash emit climate forcing gasses. These wood residues may be directed towards alternative product streams (i.e., chip, power and heat generation) with controlled combustion peter - I changed the meaning here. The fate of each of these pools is determined by a highly dynamic political and economic system. To understand how policy decisions will impact the fate and subsequent climate impact of harvested wood products, a detailed process model is necessary. ??

3.1 Disposition of Harvested Wood in California.

To provide a rough estimate of the fate of annually harvested roundwood material, we apply what we know about milling efficiency improvements, logging utilization rates, and construction use efficiency to historical production volumes.

3.1.1 Logging Residues

According to Morgan and Spoelma (2008), logging residues produced from sawlog harvest can be estimated using a factor of 0.0302 (+/-.0123 @95%CI) times the total cubic sawlog volume delivered to a mill. Unfortunately, we cannot say how logging residue production has changed over time in California. Simmons and Morgan (2014) found that logging utilization has decreased in Idaho from 1990 to 2011 by 72%. For the purpose of this analysis, we will assume that similar changes have occurred in California timber harvesting. We then estimate a logging residue production factor for years before 1990 based on the following equation. We assume 1990 residue ratios for all years prior.

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$$Vlr_x = Vrw_x \left(\eta_{04} + (\eta_{o4}\eta_\Delta)\right)$$
 Where:
$$Vrw_x = \text{Rundwood volume harvested in year } x$$

$$\eta_{04} = \mathcal{N}(0.0302, 0.0123) \text{ ratio of logging residues to roundwood harvested in CA, 2004}$$

$$\eta_\Delta = 0.72 \text{ (percent change in efficiency over time period)}$$

For logging residue production factors between 1990 and 2004, we calculate logging residues as a function of the percent change in logging residual ratios estimated for Idaho Simmons and Morgan (2014) applied to the known logging residual ratio reported by Morgan and Spoelma (2008). To reflect the uncertainty in the estimate provided by Morgan and Spoelma (2008), we calculate the logging residual using a randomly selected value from a normal probability distribution defined by the estimate and upper and lower bounds of the 95% confidence interval provided:

$$Vlr_x = Vrw_x \left(\eta_{04} + \left(\eta_{04} \left((Y_1 - x) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right) \right)$$
 Where:
$$Vrw_x = \text{Rundwood volume harvested in year } x$$
 ag residues to roundwood harvested in CA, 2004

 $\eta_{04}=\mathcal{N}(0.0302,0.0123)$ ratio of logging residues to roundwood harvested in CA, 2004 $Y_1=2004 \text{ (year for which logging residual estimate available for CA)}$ x=year for which logging residues are calculated $\eta_{\Delta}=0.72 \text{ (percent change in logging residue ratio over time period)}$ $Y_{\Delta}=21 \text{ (number of years over which logging residue ratio decreased)}$

Logging residual volume in years following 2004 are calculated as follows:

$$Vlr_x = Vrw_x \left(\eta_{04} - \left(\eta_{04} \left((x-Y_1)\frac{\eta_\Delta}{Y_\Delta}\right)\right)\right)$$
 Where:
$$Vrw_x = \text{Rundwood volume harvested in year } x$$

$$\eta_{04} = \mathcal{N}(0.0302, 0.0123) \text{ ratio of logging residues to roundwood harvested in CA, 2004}$$

$$Y_1 = 2004 \text{ (year for which logging residue stimate available for CA)}$$

$$x = \text{year for which logging residues are calculated}$$

$$\eta_\Delta = 0.72 \text{ (percent change in logging residue ratio over time period)}$$

$$Y_\Delta = 21 \text{ (number of years over which logging residue ratio decreased)}$$

3.1.2 Processing Residues

Milling efficiency has increased by roughly 14% in California in the period between 1970 and 2006 Keegan et al. (2010). For this analysis we assume a continuous improvement such that for years prior to 1970, milling efficiency in year x is calculated as:

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$$Vmr_x = Vrw_x \left(\eta_{70} - \left((Y_1 - x) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$

Where:

 $Vrw_x = \text{Rundwood volume harvested in year } x$ $\eta_{70} = 0.42 \text{ (milling efficiency in 1970)}$ $Y_1 = 1970 \text{ (earliest year mill efficiency available for)}$ x = year for which milling residues are calculated $\eta_{\Delta} = 0.06 \text{ (increase in milling efficiency from 1970-2011)}$ $Y_{\Delta} = 41 \text{ (number of years overwhihc milling efficiency increased)}$

For years after 1970, milling efficiency for year x is calculated as:

$$Vmr_x = Vrw_x \left(\eta_{70} + \left((x - Y_1) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$
 Where:
$$Vrw_x = \text{Rundwood volume harvested in year } x$$

$$\eta_{70} = 0.42 \text{ (milling efficiency in 1970)}$$

$$Y_1 = 1970 \text{ (earliest year mill efficiency available for)}$$

$$x = \text{year for which milling residues are calculated}$$

$$\eta_{\Delta} = 0.06 \text{ (increase in milling efficiency from 1970-2011)}$$

3.1.3 Construction Residues

To estimate annualized construction waste material, we use ratios of finished wood products to construction debris and demolition debris referenced in McKeever (2004). This data from McKeever is sparse and should be considered unreliable for years other than those for which it is reported. Construction debris was estimated in 2002 as approximately 15% of the total wood used in construction. Demolition debris from wood produced annually from wood grown on California forestland is outside of the scope of this report.

 $Y_{\Delta} = 41$ (number of years overwhile milling efficiency increased)

Table ?? presents ten year average estimates of logging and milling residuals, finished lumber, and construction debris based on Board of Equalization (BOE) roundwood harvest volumes. / this is kinda random? no mention of BOE data earlier?/

Table 7: Ten-year average logging and mill residual estimates based on BOE harvest volumes in Million Cubic Feet (MCF). RW:Roundwood harvested, LR: Logging residues, MR: Mill Residues, FL: Finished Lumber, CD: Construction Debris

10-year start	10-year end	RW	LR	MR	FL	CD
1978	1988	681.701	30.6277	299.522	382.179	57.3269
1979	1989	680.582	38.4338	300.229	380.353	57.0529
1980	1990	681.083	31.1597	301.528	379.555	56.9333
1981	1991	681.601	31.5936	302.612	378.989	56.8483
1982	1992	686.631	39.3647	305.606	381.025	57.1538
1983	1993	695.872	34.9048	310.422	385.451	57.8176
1984	1994	678.459	36.7081	303.4	375.059	56.2589
1985	1995	657.737	31.0119	294.892	362.845	54.4267
1986	1996	631.918	28.5312	284.093	347.825	52.1738
1987	1997	600.752	26.7997	270.919	329.833	49.4749
1988	1998	560.495	30.0402	253.572	306.923	46.0384
1989	1999	518.282	18.8804	235.308	282.975	42.4462

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Table 7: Ten-year average logging and mill residual estimates based on BOE harvest volumes in Million Cubic Feet (MCF). RW:Roundwood harvested, LR: Logging residues, MR: Mill Residues, FL: Finished Lumber, CD: Construction Debris

10-year start	10-year end	RW	LR	MR	FL	CD
1990	2000	477.206	20.8067	217.442	259.764	38.9645
1991	2001	436.798	15.4676	199.72	237.078	35.5618
1992	2002	411.648	15.5598	188.838	222.81	33.4214
1993	2003	389.756	13.9915	179.386	210.37	31.5555
1994	2004	370.287	14.3281	171.013	199.274	29.8912
1995	2005	360.411	14.1414	166.982	193.429	29.0143
1996	2006	349.131	12.3767	162.271	186.86	28.0291
1997	2007	338.319	11.4827	157.756	180.563	27.0845
1998	2008	321.14	10.8187	150.231	170.909	25.6364
1999	2009	299.649	11.0332	140.54	159.109	23.8663
2000	2010	283.222	9.79385	133.256	149.966	22.4949
2001	2011	271.892	7.2886	128.347	143.545	21.5318
2002	2012	266.945	7.36299	126.396	140.549	21.0823
2003	2013	266.193	5.82813	126.488	139.705	20.9558
2004	2014	262.901	6.77406	125.34	137.561	20.6341

3.2 Emissions from un-utilized logging residues

From logging residuals not used in bioenergy, emmisions are produced from combustion or biological decomposition of the material over time. To calculate the ratio of burned to decompsed logging residues, I first calculate the total biomass volume of pile burned forest management residuals, then compare with total residues as reported by the TPO to find the difference. C02e emissions are then independly derived for both conversion streams of logging residuals. / forremove utilize the CARB estimate of annual PM2.5 emissions produced from forest management with the Consume fire behavior model to extrapolate total

1. Estimate biomass from PM2.5:

To estimate total biomass from PM2.5, I assume 90% consumption of biomass in piles and use the relationship of pile tonnage to PM emissions calculated using the Piled Fuels Biomass and Emissions Calculator provided by the Washington State Department of Natural Resources. This calculator is based on the Consume fire behavior model published by the US Forest Service. The ratio of PM2.5 to unburned tonnage of biomass used below is 0.00605508984853. Ratio of PM2.5 to consumed fuel is 0.00672787321276.

Table 8: Forest biomass burned in piles based on ARB-reported PM2.5 emissions in the 'Forest Management' category using a ratio of 164.610674089 ton biomass per ton PM2.5.

YEAR	PM2.5 (t)	Pile-Burned Biomass (t)
2000	5474.31	901129.28
2005	5474.31	901129.28
2010	5474.31	901129.28
2012	5477.3	901621.96
2015	5480.51	902150.69

Total emissions resulting from **pile burned** forest management residuals can then be derived for the two greenhouse gasses produced from pile burning (CO2, CH4) and from BC:

2. Emissions from decomposition of un-utilized forest management residuals:

Un-utilized residual biomass not consumed in pile burns decomposes over time resulting in methane and carbon dioxide inconsistent emissions. To provide a full picture of the emissions from residual material produced from commercial timber harvesting in California, we must account for decomposition of unutilized logging residuals left on-site that are not burned. To establish the fraction of logging residue that is left to decompose, residues burned and used in bioenergy are subtracted from the total reported by the TPO:

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$$LR_d = LR - LR_{piles} - LR_{bio}$$

where:

 $LR_d =$ Logging residuals subject to an erobic decomposition

LR = Total logging residue reported by TPO

 $LR_{piles} =$ Logging residues combusted in anthropogenic pile burns

 $LR_{bio} =$ Logging residues used to produce bioenergy

To calculate the GHG emissions from decomposition of piles, we use the following equation.

$$CO_{2}e_{decomp} = (LR_{d} \times C_{LR} \times CO2_{ratio}) + (LR_{d} \times C_{LR} \times CH_{4ratio} \times GWP_{CH_{4}})$$

where:

 CO_2e_{decomp} = Carbon dioxide equivalent emissions from decomposition of logging slash

 $C_{LR} = \text{Carbon fraction of biomass: } 0.5$

 $CO2_{ratio}$ = Fraction of carbon released as CO_2 : 0.61

 $CH_{4ratio} = Fraction of carbon released as <math>CH_4$: 0.09

 GWP_{CH_4} = Global warming potential of methane: 56

3.3 Emissions from non-commercial management residuals

The Timber Products Output (TPO) in California does not report wood volume produced from non-commercial management activities. This includes management activities such as pre-commercial thinning, sanitation thinning, and fuels reduction thinning. To estimate the volume of material produced from these activities we use the following sources:

- 1. **Public lands:** The USFS Forest Service Activity Tracking System (FACTS) reports management activities conducted on National Forest System Lands. To ensure estimates of biomass volume using FACTS are not duplicative of reported volume in the TPO a series of filters are applied to the FACTS attributes to identify only non-commercial management activities.
- 1. Forest Service Activity Tracking System (FACTS)

Data from TPO does not account for forest management activities that do not result in commercial products (timber sales, biomass sales). The USFS reports Hazardous Fuels Treatment (HFT) activities as well as Timber Sales (TS) derived from the FACTS database. I use these two data sets to estimate the number of acres treated that did not produce commercial material (sawlogs or biomass) and where burning was not used. The first step is to eliminate all treatments in the HFT data set that included timber sales. I accomplish this by eliminating all rows in the HFT data set that have identical FACTS_ID fields in the TS dataset. I further filter the HFT dataset by removing any planned but not executed treatments (nbr_units1 >0 below - nbr_units1 references NBR_UNITS_ACCOMPLISHED in the USFS dataset, see metadata for HFT here), and use text matching in the 'ACTIVITY' and 'METHOD' fields to remove any rows that contain reference to 'burning' or 'fire'. Finally, we remove all rows that that reference 'Biomass' in the method category as it is assumed that this means material was removed for bioenergy.I use a range of 10-35 BDT/acre to convert acres reported in FACTS to volume. The following table presents descriptive statistics for estimates of residual unutilized wood biomass on an annual basis in million cubic feet.

	$nf\backslash_n$	$\mathrm{nf}\backslash_{\mathrm{lr}}$	$\mathrm{opriv}\backslash_{\mathrm{lr}}$	$\mathrm{fi}\backslash_{\mathrm{lr}}$	$\mathrm{opub}\backslash_{\mathrm{lr}}$
count	11	4	4	4	4
mean	12.0194	17.7	28.95	66.425	2.4
std	4.68948	5.07346	16.1593	6.07639	1.79444
\min	2.37421	11.2	11.2	59.6	0.3
25%	8.92407	15.025	19.525	62.225	1.275
50%	13.3557	18.5	27.75	66.85	2.5
75%	14.5349	21.175	37.175	71.05	3.625
\max	17.8532	22.6	49.1	72.4	4.3

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(a) **Private industrial timber lands:** CalFIRE's Forest Practice Geographical Information System. **TODO**

3.4 Avoided Emissions from Wood Product Displacement Factors

For each product application, wood may be substituted by a range of other materials. For example, in residential construction, precast concrete and structural steel framing are competitive alternatives to wood. This choice of materials has a profound impact on GHG emissions in the construction sector and is expressed as a displacement factor (DF). A displacement factor quantifies the amount of emissions reduction achieved per unit of wood used. The displacement factors published in (Sathre and O'Connor, 2010) and used in this analysis are based on the following emission reduction sources:

- 1. **Reduced emissions from manufacturing:** Wood products require less total energy than to manufacture than products made from alternative materials.
- 2. Avoided process emissions: Production of wood alternatives such as cement are associated with substantial CO2 emissions.
- 3. Carbon storage in products: Carbon in harvested wood is drawn from the atmosphere through photosynthesis and will remain fixed through the useful life of the wood product.
- 4. Carbon storage in forests: Forests producing wood continue to grow. It is assumed that forests producing wood in California are managed to sustain forest growth (not converted to non-forest land uses).
- 5. Avoided fossil fuel emissions due to bioenergy substitution: Logging and milling residuals used to produce energy avoid emissions from fossil energy sources in the energy sector.
- 6. Carbon dynamics in landfills: A fraction of carbon from wood deposited in landfills remains in semipermanent storage. The remainder is converted to methane through biological decomposition in the landfill. Capture and use of the methane as an energy source, in turn reduces emissions from fossil energy sources.

A meta analysis conducted by (Sathre and O'Connor, 2010) compared empirical analysis from 21 international studies and found an average emissions reduction of 2.1 tons of carbon (3.9 t CO2e) per ton of dry wood used. While studies ranged substantially around the average, the authors found that the majority of published displacement factors ranged between 1 and 3 tC/t dry wood.

3.5 Displacement Factors Applied to Timber Products Output

To evaluate the climate impact of harvested wood in California, I used harvested roundwood estimates from the Timber Products Output (TPO) database³. I used two estimates of the DF applied to the harvested wood reported in the TPO based on whether logging residuals were used in bioenergy or left in the woods (to decompse or burn).

Figure 4 reflects the flow of wood from Californias forest to its fate in-use and is the frame of reference for the following analysis.

I applied displacement factors reported by Sathre and O'Connor (2010) to the reported harvest volumes from the TPO database.

The following references are used to arrive at an average displacement factor of 2.625 tCO2e/t finished wood product for harvested roundwood without logging residue utilization.

For harvested roundwood with logging residue utilization the following studies are used. I used an average of the DF reported here of 3.243 tCO2e/t finished wood product.

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Timber Products Output Reporting Tool http://srsfia2.fs.fed.us/php/tpo_2009/tpo_rpa_int1.php



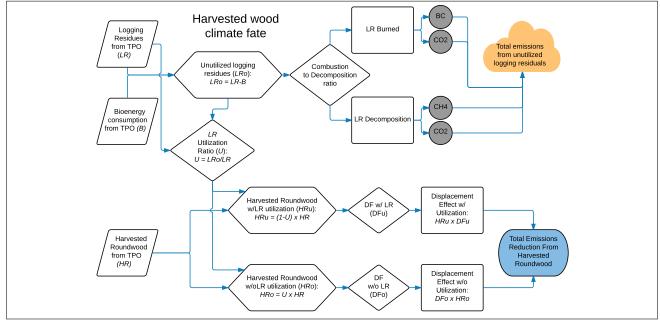


Figure 4: Wood flows from timber harvest in California

Table 9: Wood displacement factor without residue utilization

reference	displacement factor
Eriksson et al. (2007)	1.7
Eriksson et al. (2007)	2.2
Salazar and Meil (2009)	4.9
Werner et al. (2005)	1.7
	Y

The TPO reports values in terms of roundwood harvested for products, but the displacement factors presented in Sathre and O'Connor are in terms of tons of carbon in wood products. Therefore we must assume a milling efficiency to convert TPO volume estimates to finished wood product volume. I assumed a milling efficiency of 0.5.

Further, TPO is reported in cubic feet and the DF implies a mass unit. To convert cubic meters to a mass unit, we used the average wood density of harvested volume in California weighted by species as reported in Mciver et al. (2012). The resulting weighted average wood density used here is **27.94** lbs/cuft.

Using the McIver and Morgan, we determine the percent of harvested wood used in bioenergy feedstocks. From personal communications with Chelsea McIver, all bioenergy feedstock reported is sourced in-woods (ie, not mill residues).

Table 10: Wood discplacement factor with residue utilization

reference	displacement factor
Eriksson et al. (2007)	1.9
Eriksson et al. (2007)	2.5
Gustavsson et al. (2006)	4
Gustavsson et al. (2006)	5.6
Gustavsson et al. (2006)	2.2
Gustavsson et al. (2006)	3.3
Pingoud et al. (2001)	3.2

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Table 11: % volume of wood diverted to Bioenergy use by year

	year	bioenergy % of harvest
0	2000	0.024
1	2006	0.036
2	2012	0.082

	Ownership	Roundwood Products	Logging Residues	Year
0	National Forest	72.4	20.7	2012
1	Other Public	16.2	3.4	2012
2	Forest Industry	328.9	72.4	2012
3	Other Private	53	11.2	2012
4	National Forest	52.8	16.3	2006
5	Other Public	1.1	0.3	2006
6	Forest Industry	274.3	59.6	2006
7	Other Private	139.2	33.2	2006
8	National Forest	90.8	22.6	2000
9	Other Public	5.2	1.6	2000
10	Forest Industry	372.5	70.6	2000
11	Other Private	159.4	49.1	2000
12	National Forest	132.1	11.2	1994
13	Other Public	24.7	4.3	1994
14	Forest Industry	396.1	63.1	1994
15	Other Private	174.7	22.3	1994

In addition to the TPO, the California Board of Equalization (BOE) also reports historic timber harvest volumes. Comparing between years where both sources report data, the BOE database reports on average, 8% less volume than the TPO (Table 12) database. This is reasonable considering that:

- 1. BOE data may be under-reported, as there may be a financial incentive to reduce tax burden
- 2. BOE does not include volume harvested from native American tribal lands in the state

Table 12: Total annual harvest reported by Mciver et al. (2012) and California Board of Equalization.

year	McIver, et. al. (2012) MMBF	BOE MMBF	BOE/M&M
1978	4606.0	4491	0.98
1979	4044.0	3991	0.99
1980	3478.0	3164	0.91
1981	2832.0	2672	0.94
1982	2488.0	2318	0.93
1983	3638.0	3358	0.92
1984	3701.0	3546	0.96
1985	4093.0	3818	0.93
1986	4416.0	4265	0.97
1987	4667.0	4500	0.96
1988	4847.0	4670	0.96
1989	4699.0	4424	0.94
1990	4264.0	4021	0.94
1991	3439.0	3195	0.93
1992	3192.0	2973	0.93
1993	3041.0	2871	0.94
1994	2814.0	2316	0.82
1995	2520.0	2306	0.92
1996	2515.0	2273	0.9
1997	2640.0	2400	0.91
1998	2420.0	2091	0.86
1999	2429.0	2144	0.88
		Continued	on next page

Continued on next page

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Table 12: Total annual harvest reported by Mciver et al. (2012) and California Board of Equalization.

year	McIver, et. al. (2012) MMBF	BOE MMBF	$\mathrm{BOE}/\mathrm{M\&M}$
2000	2244.0	1966	0.88
2001	1801.0	1603	0.89
2002	1691.73	1690	1.0
2003	1667.95	1663	1.0
2004	1704.0305	1706	1.0
2005	1738.5	1725	0.99
2006	1960.35	1631	0.83
2007	1759.6	1626	0.92
2008	1476.0745	1372	0.93
2009	911.19	805	0.88
2010	1302.38	1161	0.89
2011	1432.5	1288	0.9
2012	1421.3	1307	0.92

// move to appendix?//The TPO reports harvest from tribal lands, which produces an average 0.74% of the total annual harvest in the state for the 37 years of parallel data. For this analysis we used TPO data to include harvest volume from tribal lands.

Table 13: Annual harvest by ownership from Mciver et al. (2012) (MCF)

1947 0.0 0.0 569.85 0.0 1948 0.0 0.0 735.29 0.0 1949 0.0 0.0 698.53 0.0 1950 0.0 0.0 808.82 0.0 1951 0.0 0.0 900.74 0.0 1952 2.57 113.79 808.82 4.78 1953 3.31 117.65 977.94 2.76 1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01	year	State	Federal	Private	Tribal
1949 0.0 0.0 698.53 0.0 1950 0.0 0.0 808.82 0.0 1951 0.0 0.0 900.74 0.0 1952 2.57 113.79 808.82 4.78 1953 3.31 117.65 977.94 2.76 1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 34	1947	0.0	0.0	569.85	0.0
1950 0.0 0.0 900.74 0.0 1951 0.0 0.0 900.74 0.0 1952 2.57 113.79 808.82 4.78 1953 3.31 117.65 977.94 2.76 1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 <t< td=""><td>1948</td><td>0.0</td><td>0.0</td><td>735.29</td><td>0.0</td></t<>	1948	0.0	0.0	735.29	0.0
$\begin{array}{cccccccccccccccccccccccccccccccccccc$	1949	0.0	0.0	698.53	0.0
1952 2.57 113.79 808.82 4.78 1953 3.31 117.65 977.94 2.76 1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43	1950	0.0		808.82	0.0
1953 3.31 117.65 977.94 2.76 1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82	1951	0.0		900.74	0.0
1954 2.94 141.54 880.51 4.6 1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35	1952	2.57)113.79		4.78
1955 2.57 191.73 906.25 6.07 1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25	1953	3.31	117.65	977.94	2.76
1956 4.41 206.99 862.13 5.33 1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17	1954	2.94	141.54	880.51	4.6
1957 4.96 170.59 801.47 6.62 1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8	1955	2.57	191.73	906.25	6.07
1958 5.51 208.27 821.69 6.99 1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07	1956	4.41	206.99	862.13	5.33
1959 4.96 279.6 788.6 9.19 1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35	1957	4.96	170.59	801.47	6.62
1960 5.15 250.37 680.15 8.82 1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 <td>1958</td> <td>5.51</td> <td>208.27</td> <td>821.69</td> <td>6.99</td>	1958	5.51	208.27	821.69	6.99
1961 5.33 259.74 707.72 10.11 1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 <td>1959</td> <td>4.96</td> <td>279.6</td> <td>788.6</td> <td>9.19</td>	1959	4.96	279.6	788.6	9.19
1962 6.25 259.01 744.49 8.64 1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1978 5.15	1960	5.15	250.37	680.15	8.82
1963 4.04 311.76 678.31 9.93 1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15	1961		259.74	707.72	10.11
1964 4.6 348.16 643.38 9.01 1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78	1962	6.25	259.01	744.49	
1965 5.7 363.05 591.91 9.74 1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1963	4.04	311.76	678.31	9.93
1966 5.88 360.85 545.96 8.27 1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1964	4.6	348.16	643.38	9.01
1967 6.43 355.51 562.5 7.54 1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1965	5.7	363.05	591.91	9.74
1968 8.82 440.44 542.28 14.52 1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1966	5.88	360.85	545.96	8.27
1969 7.35 372.61 529.41 9.93 1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1967	6.43	355.51	562.5	7.54
1970 6.25 345.4 481.62 5.15 1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1968	8.82	440.44	542.28	14.52
1971 7.17 383.09 476.1 12.87 1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1969		372.61	529.41	9.93
1972 6.8 411.58 591.91 12.13 1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82				481.62	5.15
1973 6.07 371.69 516.54 9.38 1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1971	7.17		476.1	
1974 7.35 322.79 525.74 9.38 1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1972	6.8	411.58	591.91	12.13
1975 6.43 287.87 498.16 3.31 1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1973		371.69	516.54	9.38
1976 7.35 348.53 507.35 6.99 1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1974		322.79	525.74	
1977 5.15 323.35 544.12 6.99 1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82	1975	6.43	287.87	498.16	3.31
1978 5.15 332.35 509.19 8.64 1979 4.78 321.32 417.28 8.82					
1979 4.78 321.32 417.28 8.82					
1980 3.68 279.04 356.62 7.72	1979	4.78		417.28	8.82
	1980	3.68	279.04	356.62	7.72

Continued on next page

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Table 13: Annual harvest by ownership from Mciver et al. (2012) (MCF)

year	State	Federal	Private	Tribal
1981	2.76	201.65	316.18	4.04
1982	7.72	173.9	275.74	1.47
1983	7.9	313.42	347.43	2.57
1984	6.25	288.05	386.03	3.86
1985	6.62	339.52	406.25	0.92
1986	5.33	365.26	441.18	4.96
1987	7.72	364.89	485.29	7.54
1988	5.7	403.68	481.62	2.57
1989	6.8	373.53	483.46	2.02
1990	4.41	283.09	496.32	2.57
1991	6.99	248.35	376.84	4.41
1992	4.23	190.99	391.54	5.88
1993	6.25	137.32	415.44	2.39
1994	3.12	152.02	362.13	2.76
1995	7.35	101.1	354.78	2.94
1996	10.11	86.4	365.81	2.39
1997	8.64	101.65	375.0	2.76
1998	4.78	83.46	356.62	2.94
1999	0.0	97.24	349.26	0.0
2000	3.49	63.42	345.59	1.84
2001	2.94	56.07	272.06	1.84
2002	0.18	31.38	279.41	2.5
2003	0.18	28.85	277.57	3.29
2004	0.18	20.78	292.28	3.05
2005	0.18	43.66	275.74	1.95
2006	0.74	41.61	318.01	2.37
2007	0.18	58.57	264.71	3.55
2008	0.18	37.7	233.46	2.48
2009	0.18	30.37	136.95	0.72
2010	0.18	49.89	189.34	1.79
2011	0.18	55.42	207.72	2.1
2012	5.13	37.39	218.75	1.49

To use the TPO data to estimate emissions reductions using the DF, we apply a conversion factor of **5.44** MCF/MMBF. This is an approximation as the actual sawlog conversion factor varies with average harvested log size, which has changed over time.

Using the ratio of logging residuals consumed by bioenergy (mciver), to the total logging residuals reported in the TSP, we can calculated the harvest volume the ratio of harvest volume to logging residuals used in bioenergy, we calculated based on the ratio of reported consumption of logging residuals in bioenergy by Mciver et al. to the total logging residuals reported in the TPO. Mciver et al. report bioenergy consumption from 2000 forward. For years previous, we use the average bioenergy consumption from 2000 - 2012. These results assume bioenergy consumption throughout the reporting years. Bioenergy use of residuals did not begin until the late 1970. Further analysis is necessary to modify these results to reflect the development of the bioenergy industry.

To calculate the total emissions reduction resulting from California's timber harvest, we apply the appropriate displacement factor (with or without logging residual utilization) to the commensurate fraction of harvested roundwood. The results are shown in the following chart.

Contribution of the varios ownership categories to the aggregate is shown in Figure 6.

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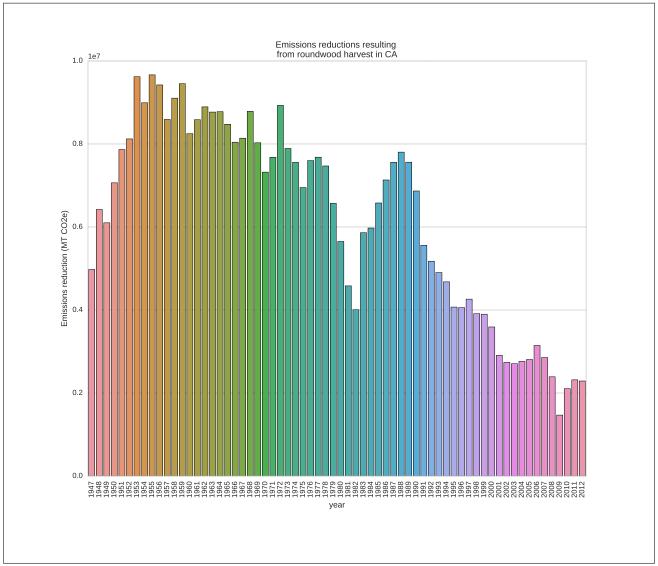


Figure 5: Historical emissions reductions resulting from harvested roundwood using displacement factors from (Sathre and O'Connor, 2010) applied to TPO data.

4 Further Questions

This analysis is a first step towards a broader analysis of the climate impacts of harvested wood in California. The following are key questions which follow from this analysis.

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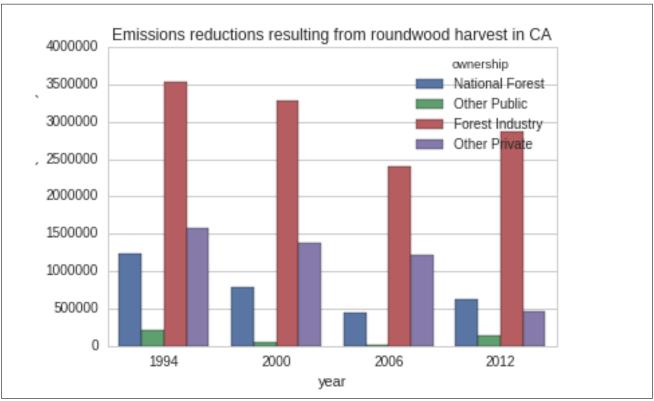


Figure 6: Historical emissions reductions by ownership for selected years resulting from harvested roundwood using displacement factors from (Sathre and O'Connor, 2010) applied to TPO data.

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