

Emissions reductions from harvested wood products and management residuals

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1 California Forest Management Emissions Profile

Forest management activities in California produce logs for lumber markets while maintaining and enhancing forest health. In addition to merchantable logs, these harvest activities produce logging residuals that are either left in the stand to decompose, or pile burned as directed by forest practice rules (California Forest Practice Rules, Article 7 § 917.2). The extent to which these residuals are utilized avoiding open combustion or decomposition materially impacts emissions of GHG and SLCP..

The majority of biomass produced from forest management activities end up as residual biomass material that is either left in the woods to decompose, or aggregated at a landing where it is eventually burned. When considered alongside the accumulation of woody material from historic fire suppression activity, there exists a heightened volume of combustible woody material in excess of historic reference conditions, elevating the risk of damaging wildfire in much of California's forestland. Currently, common practice for fuel load management in California forests involves prescribed natural fire and sanitation pile burning. However, these approaches have limitations as previous studies have demonstrated that prescribed natural fire is often only effective at reducing fuel loading and maintaining fire resilient landscapes when it is coupled with mechanical treatment to remove biomass Stephens et al. (2009). Open pile burning of residual biomass can also result in greater emissions of strong radiative forcing agents (black carbon) and criteria air pollutants such as Particulate Matter (PM) and Oxides of Nitrogen (NOX) when compared to controlled combustion in biomass power plants with modern emissions control technology.

Ultimately, any combustion or decomposition of residual material results in emissions of GHG, Criteria Air Pollutants (CAP), and SLCP. In the absence of forest management activity, atmospheric emissions are produced from stochastic processes such as wildfire, pest, and disease outbreaks. Utilization strategies are necessary to reduce the air quality impact of common forestry practice. The high climate opportunity cost¹ of open burning residual biomass in forestry and agriculture favor diverting residual material streams towards alternative utilization strategies such as biomass power. Figure 1 presents an overview of emissions and emissions reduction pathways for wood from California's forests.

The focus of this analysis is on deriving the net Carbon Dioxide Equivalents (CO₂e) emissions and emissions reductions associated with **management activity**. This report does not assess greenhouse gas emissions from pest or disease induced mortality, which is estimated at approximately 34 MMT of CO₂e annually in California forests Christensen et al. (2016). Emissions from mortality are indirectly related to management activities just as wildfire is, and must be accounted for to comprehensively evaluate the climate impacts of harvesting. To estimate emissions and reductions by forest management activity, it is necessary to take the following steps:

1. Estimate CO₂e emissions from burning forest management residuals using the California Air Resources Board (CARB), CAP and GHG emissions inventories.
2. Estimate the volume and fate of wood that is removed, left in the forest, and burned as a result of direct anthropogenic management activities.
3. Apply life-cycle Displacement Factor (DF) for all utilized wood and apply DF to harvested wood to obtain an aggregate estimate to estimate emissions reductions from wood use.

1.1 Report Objectives

Quantifying the climate effects of wood products and forest management residuals is important to the development of the Forest Climate Plan (FCP)² as well as efforts underway by the California Board of Forestry (BOF) and CalFire to meet the intent of AB 1504 (2010)³. To inform these efforts, this report provides estimates of the following:

- Statewide GHG and SLCP emissions produced from the combustion or decomposition of logging residuals.

¹Climate opportunity cost is used in this context to refer to the aggregate emissions of particulate and gasses with strong radiative forcing properties associated with open pile or broadcast burning.

²The Forest Climate Action Team (FCAT) was assembled in August of 2014 with the primary purpose of developing a Forest Carbon Plan by the end of 2016. FCAT is comprised of Executive level members from many of the States natural resources agencies, state and federal forest land managers, and other key partners directly or indirectly involved in California forestry. FCAT is under the leadership of CAL FIRE, Cal-EPA, and The Natural Resources Agency.

³AB-1504 Forest resources: carbon sequestration.(2009-2010)

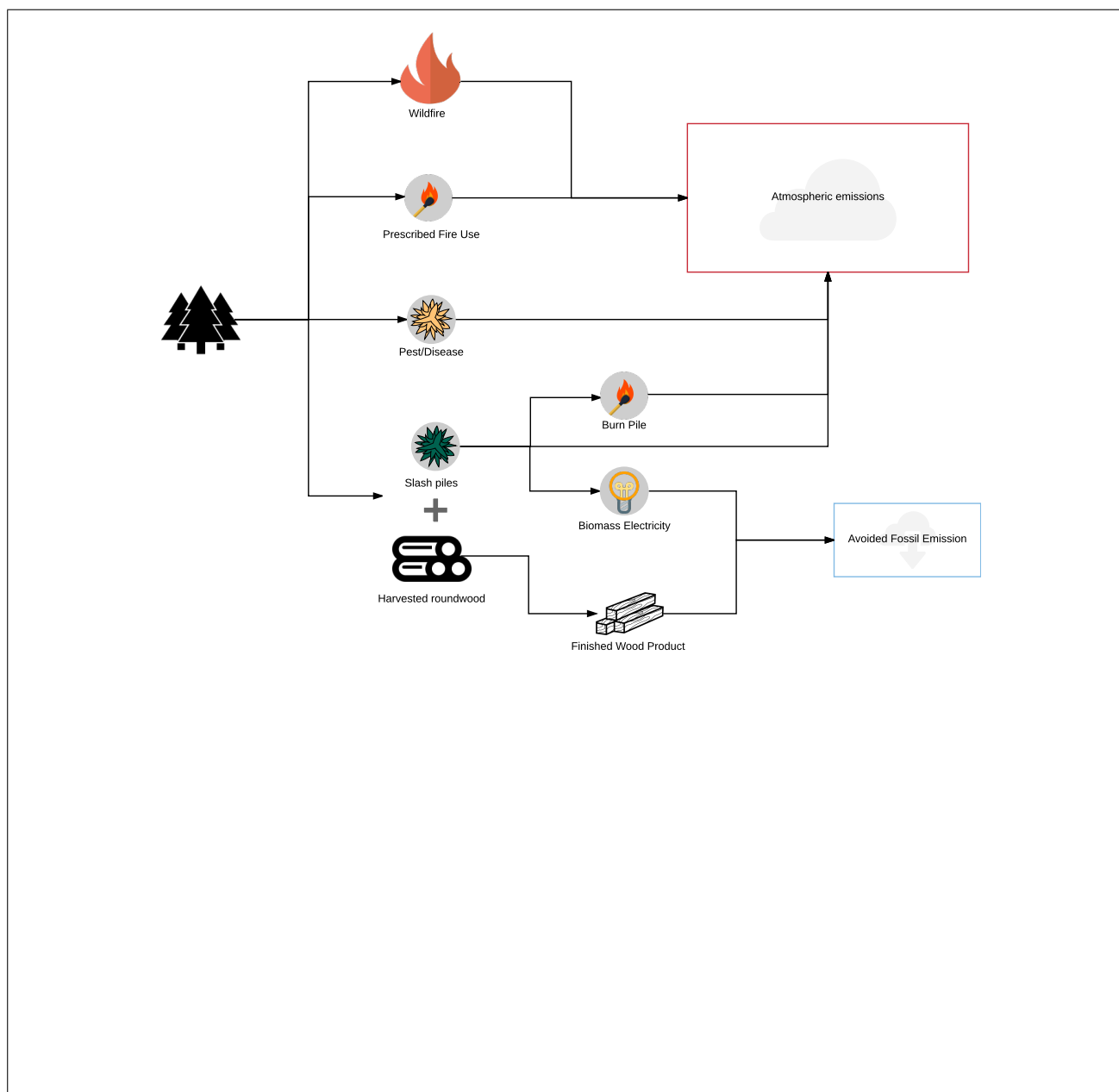


Figure 1: Overview of fates of wood resulting from harvest and mortality in California forests. Note that time is not represented in this figure.

- Statewide GHG emissions reductions from the use of wood products harvested in the state.

Estimates are based empirical data and reflect past forest management activities. It is **critical** to note that the empirical data used in this analysis reflect point-in-time measures that are affected by a dynamic system of climate, growth, and mortality as well as macroeconomic and policy forces. This analysis may provide insight into opportunities to more effectively utilize woody biomass residuals from current forest management activities to reduce emissions.

1.2 Key Findings

- Baseline GHG and SLCP emissions from burning of forest management residuals can be estimated and should be considered in any forest management emissions baseline.
- Total annual emissions from pile burning of forest management residuals (including SLCP and GHG components) extrapolated from CARB emissions inventory are estimated at **2.5 MMTCO_{2e}**
- Wood harvested in California in 2012 resulted in avoided emissions of **2.29 MMTCO_{2e}**

- Logging residuals not used in bioenergy production contributed annual emissions of:
 - **XXX MMTCO₂e** resulting from anthropogenic burning of logging residuals
 - **XXX MMTCO₂e** resulting from decomposition of logging residuals left unburned
- Un-utilized slash from non-commercial management activities on National Forest System lands contributed emissions of XXX MMTCO₂e
- Forest Inventory and Analysis re-sample data has been used in the southeast to quantify removals resulting from non-commercial management activity and could be used for this purpose in California
- The Prescribed Fire Information Reporting System (PFIRS) may be a useful tool for quantifying emissions from pile burns and prescribed fire. It is a requirement that prescribed fires and pile burns on National Forest System Lands are reported through PFIRS. However, California Air Quality Management Districts are not required to report emissions through this system at this time. Therefore, it is not possible to associate burns in the PFIRS with commercial harvest activities.
- Brown or Organic Carbon (OC) carbon has stronger radiative absorption than BC and is associated with biomass burning. Accounting for anthropogenic production of OC would be included in emissions baselines against which alternative utilization (energy) should be measured against.

2 Estimating CO₂ Equivalent Emissions from In-Forest Biomass Combustion

The CARB reports on emissions from in-forest biomass combustion with current GHG and CAP emissions inventories. Both are necessary resources for establishing aggregate annual climate-forcing emissions (Figure 2). The GHG inventory captures gasses with radiative forcing properties including CO₂ and CH₄, but does not capture elemental carbon or BC emissions which also have strong radiative forcing properties (Table 1). The California Air Resources Board (2015, 2016) CAP report captures SLCP emissions from wildfire (80.52 MMTCO₂e) and prescribed fire (3.66 MMTCO₂e) from which black carbon emissions may be estimated. However, no reference in the CAP report is made to the source of these SLCP estimates. When viewed in aggregate, a comprehensive reporting of total climate impact from anthropogenic burning may be estimated.

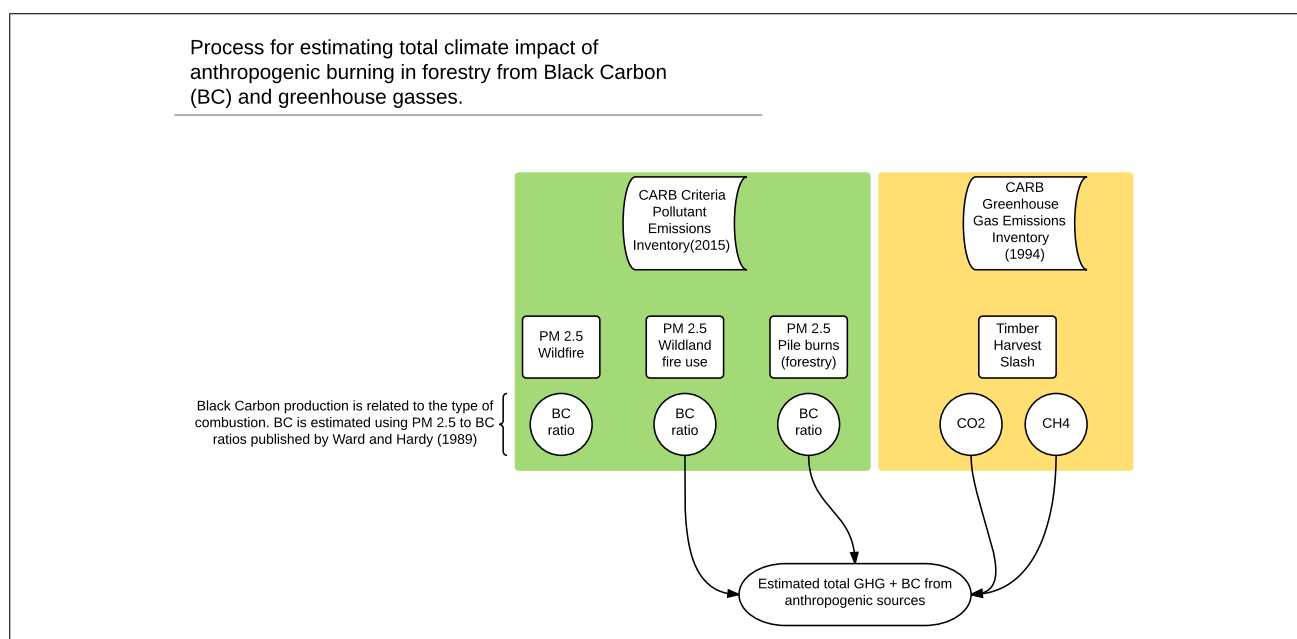


Figure 2: Data sources available from CARB for estimating GHG and SLCP emissions from forest management.

The GHG inventory captures gasses with radiative forcing properties including CO₂ and CH₄, but does not capture elemental carbon or BC emissions which also have strong radiative forcing properties (Table 1).

The California Air Resources Board (2015, 2016) CAP report captures SLCP emissions from wildfire (80.52 MMTCO₂e) and prescribed fire (3.66 MMTCO₂e) from which BC emissions may be estimated. However, no

2.1 Estimating Black Carbon Emissions from Biomass Burning

GWP ₂₀	GWP _{σ20}	GWP ₁₀₀	GWP _{σ100}	GWP ₅₀₀	GWP _{σ500}	Source
2200.0	888.82	633.33	255.41	193.33	77.67	Fuglestvedt2010
3200.0		900.0				CaliforniaAirResourcesBoard2015

Table 1: Range of Global Warming Potential(GWP) values for Black Carbon.

reference in the CAP report is made to the source of these SLCP estimates. When viewed in aggregate, a comprehensive estimate of total climate impact from anthropogenic burning may be made.

2.1 Estimating Black Carbon Emissions from Biomass Burning

Black Carbon (BC) is not directly reported by statewide emissions summaries. BC is a fraction of the Total Carbon (TC) component measured in particulate matter (PM 2.5). PM emissions are published annually by CARB (Criteria air pollutant (CAP) emissions estimates). By using the 2015 CAP emissions estimates shown in Table ?? with estimated ratios of smoldering to flaming combustion for hand/machine piled burns, prescribed natural fire and wildfire from Ward and Hardy (1989), Black Carbon emissions can be calculated from PM 2.5 with Eq. (1)

Source (CARB nomenclature)	Description	PM 2.5 (t y ⁻¹)
ALL VEGETATION	Wildfire	137630.15
FOREST MANAGEMENT	Pile burning	5480.51
WILDLAND FIRE USE (WFU)	Prescribed natural fire	6802.43

Using the 2015 CAP emissions estimates shown in Table ?? with estimated ratios of smoldering to flaming combustion for hand/machine piled burns, prescribed natural fire and wildfire from Ward and Hardy (1989), BC emissions can be estimated from PM 2.5 using equation (1)

$$BC = (PM_{2.5} \times F \times TC_f \times BC_f) + (PM_{2.5} \times S \times TC_s \times BC_s) \quad (1)$$

where:

BC = Black Carbon (mass units)

$PM_{2.5}$ = $PM_{2.5}$ (mass units)

F = Percent of combustion in flaming phase

TC_f = Total Carbon fraction of $PM_{2.5}$ for flaming phase

BC_f = Black Carbon fraction of Total Carbon for flaming phase

S = Percent of combustion in smoldering phase

TC_s = Total Carbon fraction of $PM_{2.5}$ for smoldering phase

BC_s = Black Carbon fraction of Total Carbon for smoldering phase

The ratio of smoldering to flaming combustion behavior for each biomass burning scenario means that each has a different BC Δ PM ratio. To arrive at a rough estimate of BC emissions based on PM2.5, ratios from Ward and Hardy (1989) and Jenkins et al. (1996) ratios in Table 2 are used herein.

Source	BC _f t ⁻¹ PM	TC _f ^{Cv} t ⁻¹ PM	BC _f ^{Cv} t ⁻¹ TC	BC _s t ⁻¹ PM2.5	TC _s ^{Cv} t ⁻¹ PM	BC _s ^{Cv} t ⁻¹ TC
Pile Burn	0.046904	0.09	0.45	0.01624	0.01	0.49
Prescribed	0.08016309	0.0733	0.5833	0.020944	0.08	0.29
Wildfire	0.05870124	0.0867	0.4467	0.0228641	0.06	0.338

Table 2: Factors used for calculating BC emissions from the three primary combustion sources. BC is a fraction of Total Carbon (TC) which is a fraction of total PM 2.5. Coefficients of variation (C_v) are reported here as well.

Source	BC _f t ⁻¹ PM	TC _f ^{Cv} t ⁻¹ PM	BC _f ^{Cv} t ⁻¹ TC	BC _s t ⁻¹ PM2.5	TC _s ^{Cv} t ⁻¹ PM	BC _s ^{Cv} t ⁻¹ TC
Pile Burn	0.046904	0.09	0.45	0.01624	0.01	0.49
Prescribed	0.08016309	0.0733	0.5833	0.020944	0.08	0.29
Wildfire	0.05870124	0.0867	0.4467	0.0228641	0.06	0.338

We then estimate annual BC emissions (Table 3).

Source	PM2.5 (t y ⁻¹)	BC (t y ⁻¹)	CO2e (t y ⁻¹)
ALL VEGETATION	137630.15	11225.85	35922719.54
FOREST MANAGEMENT	5480.51	346.06	1107396.54
WILDLAND FIRE USE (WFO)	6802.43	687.77	2200877.13

Table 3: Annual black carbon emissions calculated from CARB volumes

Given the variance in BC production from smoldering ($\pm 49\%$) and flaming ($\pm 45\%$) phases (Table 2), actual emissions of BC may vary substantially depending on combustion. In addition to these estimates Chow et al. (2010) provides an alternative source for estimates of BC and OC emissions in the state in 2006. Further work is necessary to evaluate the impacts of OC on the net CO2e emissions from pile burning. Pokhrel et al. (2016) estimated the absorptive properties of OC to be 1.5 - 2.5 that of BC. Chow et al. (2010) estimated that on 2006 29,530 Mt of OC was emitted from wildfires.

2.2 Estimating GHG Emissions from Biomass Burning

The CARB GHG emissions inventory resolved to combustion source (piles, prescribed, etc.) for forests and rangelands has not been updated since 2004. To provide a comparable estimate of GHG emissions from pile burning we use the ratio of PM to the net CO2e emissions from all GHG species produced from the Piled Fuels Emissions Calculator (CONSUME model equations) sources two approaches are taken. As PM is reported in the CAP for pile burning we can apply this ratio to estimate GHG emissions for the same time period.

To estimate GHG emissions from **pile burning**, we use the ratio of PM 2.5 to CO2 and to CH4 from the Piled Fuels Emissions Calculator. These ratios are then applied to CARB-reported PM emissions to estimate GHG emissions (Table ??).

Pile Biomass (t)	Consumed Biomass (t)	PM2.5 (t)	CO2 (t)	CH4 (t)
1.360178	1.224161	0.008263	2.0366	0.0034

$$PM\Delta GHG = \frac{PM2.5}{CO2 + (CH4 \times CH4_{GWP})} \quad (2)$$

2015 PM2.5 (t)	PM Δ GHG	2015 GHG MMTCO2e
5480.51	3.8947021 (-3)	1.4071705

GHG emissions from **wildfire and prescribed fire** are difficult to estimate at present but the CARB GHG emissions inventory provided estimates for years between 1994 and 2004 (Table 4).

Source Category	Average annual emissions 1994-2004 MMt CO2e
Forest and rangeland fires	2.0194
Timber harvest slash	0.1552666666666667

Table 4: Annual GHG Emissions estimated from CARB GHG emissions inventory

2.3 Estimating Total Emissions from Biomass Burning

To arrive at an estimate of total CO2 equivalent (CO2e) emissions for 2015, we combine BC emissions estimates from the CARB Criteria Air Pollutant Emissions Inventory with GHG emissions estimates reported in the

CARB GHG Emissions Inventory and the CONSUME model. The time discrepancy between the most recent GHG Emissions Inventory in 2004 and the current year is acknowledged as an irreconcilable source of uncertainty in this estimation. The use of the COMSUME model to predict pile burning emissions is likely more reliable and as pile burning is a central focus of this analysis we confidently use the emissions estimated here for pile burning throughout. Model-based estimation could be used to derive a ratio of GHG to PM using the USFS CONSUME model. Overall, this analysis demonstrates that substantial emissions from forest management residuals have been reported by CARB emissions inventories and that such inventories could be utilized to establish a baseline condition (Table ??).

MMt CO ₂ e	Source
1.4071705	CO ₂ e pile burning
1.1073965	CO ₂ e BC pile burning
2.514567	Total MMt CO₂e

3 Estimating Emissions Impact from Utilization of Harvested Wood

Wood harvested from California's forests are utilized in a variety of construction, landscaping, and consumer products. During the manufacture of these products, this wood is fractionated through a multi-stage process of harvesting, processing, and utilization to reside in several residual biomass fates (below).

Logging Residuals Tops, limbs, and sub-merchantable material produced from harvest activities in the woods. These residuals may be left on site to naturally decompose or be combusted through controlled pile burning or wildfire.

Processing (Mill) Residuals Sawdust, shavings, bark, and off cuts from primary and secondary manufacturing. These residuals may be directed towards alternative product streams (i.e. wood pellet, wood chip, power and heat generation) or sent to a landfill.

Construction Debris Fraction of wood used in construction or finished products that are not integrated into its final form. These residuals are most commonly sent to a landfill.

Demolition Wood used in construction that has reached the end of its useful life. These residuals are most commonly sent to a landfill.

These biomass fates have widely variable time horizons for carbon return to the atmosphere and can greatly influence the net emissions impact attributed to the initial forest management activity. While wood products used in construction, finished products, or other stable environments may sequester carbon for a long period, residues sent to landfills or left in the woods as slash emit climate forcing gasses to the atmosphere. Some of these wood residues may be redirected towards alternative controlled combustion applications (i.e., pellet production, power and heat generation) to avoid emissions.

Ultimately the fate of these pools are determined by a highly dynamic political and economic system. To understand how policy decisions will impact the fate and subsequent climate impact of harvested wood products, a detailed process model is necessary to track the distribution of harvested wood material. Figure ??

3.1 Disposition of Harvested Wood in California.

To provide a rough estimate of the fate of annually harvested roundwood material, we estimate the volume of wood biomass residing in Logging, Processing, and Construction residuals. To estimate current values, we apply known milling efficiency improvements, logging utilization rates, and construction use efficiency to historical production volumes.

3.1.1 Logging Residues

According to Morgan and Spoelma (2008), logging residues produced from sawlog harvest can be estimated using a factor of 0.0302 (+/-0.0123 @95%CI) times the total cubic sawlog volume delivered to a mill. Simmons and

Morgan (2014) found that logging utilization has decreased in Idaho from 1990 to 2011 by 72%. Unfortunately, we cannot say how logging residue production has changed over time in California. For the purpose of this analysis, we will assume that similar changes have occurred in California timber harvesting.

We estimate logging residue production factor for years before 1990 based on the following equation. We assume 1990 residue ratios for all years prior.

$$Vlr_x = Vrw_x (\eta_{04} + (\eta_{04}\eta_{\Delta}))$$

Where:

Vrw_x = Roundwood volume harvested in year x

$\eta_{04} = \mathcal{N}(0.0302, 0.0123)$ ratio of logging residues to roundwood harvested in CA, 2004

$\eta_{\Delta} = 0.72$ (percent change in efficiency over time period)

For logging residue production factors between 1990 and 2004, we calculate logging residues by adjusting the logging residual ratio reported by Morgan and Spoelma (2008) with the percent change in logging residual ratios estimated for Idaho by Simmons and Morgan (2014). To reflect the uncertainty in the estimate provided by Morgan and Spoelma (2008), we calculate the logging residual using a randomly selected value from a normal probability distribution defined by the estimate and upper and lower bounds of the 95% confidence interval provided:

$$Vlr_x = Vrw_x \left(\eta_{04} + \left(\eta_{04} \left(Y_1 - x \right) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$

Where:

Vrw_x = Roundwood volume harvested in year x

$\eta_{04} = \mathcal{N}(0.0302, 0.0123)$ ratio of logging residues to roundwood harvested in CA, 2004

$Y_1 = 2004$ (year for which logging residual estimate available for CA)

x = year for which logging residues are calculated

$\eta_{\Delta} = 0.72$ (percent change in logging residue ratio over time period)

$Y_{\Delta} = 21$ (number of years over which logging residue ratio decreased)

Logging residual volume in years following 2004 are calculated as follows:

$$Vlr_x = Vrw_x \left(\eta_{04} - \left(\eta_{04} \left(x - Y_1 \right) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$

Where:

Vrw_x = Roundwood volume harvested in year x

$\eta_{04} = \mathcal{N}(0.0302, 0.0123)$ ratio of logging residues to roundwood harvested in CA, 2004

$Y_1 = 2004$ (year for which logging residual estimate available for CA)

x = year for which logging residues are calculated

$\eta_{\Delta} = 0.72$ (percent change in logging residue ratio over time period)

$Y_{\Delta} = 21$ (number of years over which logging residue ratio decreased)

3.1.2 Processing Residues

Milling efficiency has increased by roughly 14% in California in the period between 1970 and 2006 Keegan et al. (2010). For this analysis we assume a continuous improvement such that for years prior to 1970, milling efficiency in year x is calculated as:

$$Vmr_x = Vrw_x \left(\eta_{70} - \left((Y_1 - x) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$

Where:

Vrw_x = Rundwood volume harvested in year x

$\eta_{70} = 0.42$ (milling efficiency in 1970)

$Y_1 = 1970$ (earliest year mill efficiency available for)

x = year for which milling residues are calculated

$\eta_{\Delta} = 0.06$ (increase in milling efficiency from 1970-2011)

$Y_{\Delta} = 41$ (number of years over which milling efficiency increased)

For years after 1970, milling efficiency for year x is calculated as:

$$Vmr_x = Vrw_x \left(\eta_{70} + \left((x - Y_1) \frac{\eta_{\Delta}}{Y_{\Delta}} \right) \right)$$

Where:

Vrw_x = Rundwood volume harvested in year x

$\eta_{70} = 0.42$ (milling efficiency in 1970)

$Y_1 = 1970$ (earliest year mill efficiency available for)

x = year for which milling residues are calculated

$\eta_{\Delta} = 0.06$ (increase in milling efficiency from 1970-2011)

$Y_{\Delta} = 41$ (number of years over which milling efficiency increased)

3.1.3 Construction Residues

To estimate annualized construction waste material, we apply the ratio of construction and demolition debris to finished wood products (McKeever (2004)) to Board of Equalization (BOE) roundwood harvest volumes. In 2002, construction debris was estimated as approximately 15% of the total wood used in construction. Of note is that the data from McKeever is sparse and should be considered unreliable for years other than those for which it is reported.

3.1.4 Demolition Debris

Debris from wood produced from wood grown on California forestland is outside of the scope of this report.

3.1.5 Harvested Wood Residue Summary

The following Table ?? presents ten year average estimates of logging and milling residuals, finished lumber, and construction debris based on Board of Equalization (BOE) roundwood harvest volumes.

3.2 Emissions from Un-Utilized Residues

Residuals not utilized in bioenergy applications or sent to a landfill eventually produce emissions through combustion or biological decomposition of the material over time. Most of these residues originate from logging activity. To calculate CO₂e emissions from un-utilized residues, I first estimate the total volume of biomass pile burned in forests using the CARB estimate of PM. Then, by comparing total volume of pile burned and bioenergy diverted biomass against the total biomass volume from the Timber Products Output (TPO), I resolve the remaining biomass volume as emitted through decomposition.

3.2.1 Estimate Pile-Burned Biomass Emissions from PM2.5:

To estimate total biomass from PM2.5, I assume 90% consumption of biomass in piles and use the relationship of pile tonnage to PM emissions as calculated from the Piled Fuels Biomass and Emissions Calculator provided by the Washington State Department of Natural Resources. This calculator is based on the Consume fire behavior model published by the US Forest Service. The ratio of PM to unburned tonnage of biomass used below is 0.00605508984853. Ratio of PM2.5 to consumed fuel is 0.00672787321276.

YEAR	PM2.5 (t)	Pile-Burned Biomass (t)
2000	5474.31	901129.28
2005	5474.31	901129.28
2010	5474.31	901129.28
2012	5477.3	901621.96
2015	5480.51	902150.69

Table 5: Forest biomass burned in piles based on ARB-reported PM2.5 emissions in the 'Forest Management' category using a ratio of 164.610674089 ton biomass per ton PM2.5.

Total emissions resulting from **pile burned** forest management residuals can then be derived for the two greenhouse gasses produced from pile burning (CO₂, CH₄) and from BC (Table ??).

3.2.2 Emissions from Decomposition of un-utilized forest management residuals:

Un-utilized residual biomass not consumed in pile burns decomposes over time resulting in CH₄ and CO₂ emissions. To provide a full picture of the emissions from residual material produced from commercial timber harvesting in California, we must account for decomposition of unutilized logging residuals left on-site that are not burned. To establish the fraction of logging residue that is left to decompose, residues burned and used in bioenergy are subtracted from the total reported by the TPO:

$$LR_d = LR - LR_{piles} - LR_{bio}$$

where:

LR_d = Logging residuals subject to anerobic decomposition

LR = Total logging residue reported by TPO

LR_{piles} = Logging residues combusted in anthropogenic pile burns

LR_{bio} = Logging residues used to produce bioenergy

To calculate the GHG emissions from decomposition of piles, we use the following equation.

$$CO_2e_{decomp} = (LR_d \times C_{LR} \times CO_2ratio) + (LR_d \times C_{LR} \times CH_4ratio \times GWP_{CH_4})$$

where:

CO_2e_{decomp} = Carbon dioxide equivalent emissions from decomposition of logging slash

C_{LR} = Carbon fraction of biomass: 0.5

CO_2ratio = Fraction of carbon released as CO₂: 0.61

CH_4ratio = Fraction of carbon released as CH₄: 0.09

GWP_{CH_4} = Global warming potential of methane: 56

3.3 Emissions of Residuals from non-commercial management Activity

The Timber Products Output (TPO) in California does not report wood volume produced from non-commercial management activities. This includes management activities such as pre-commercial thinning, sanitation thinning, and fuels reduction thinning. To estimate the volume of material produced from these activities we use the following sources:

1. **Public lands:** The USFS Forest Service Activity Tracking System (FACTS) reports management activities conducted on National Forest System Lands. To ensure estimates of biomass volume using FACTS are not duplicative of reported volume in the TPO a series of filters are applied to the FACTS attributes to identify only non-commercial management activities.

1. Forest Service Activity Tracking System (FACTS)

Data from TPO does not account for forest management activities that do not result in commercial products (timber sales, biomass sales). The USFS reports Hazardous Fuels Treatment (HFT) activities as well as Timber Sales (TS) derived from the FACTS database. I use these two data sets to estimate the number of acres treated that did not produce commercial material (sawlogs or biomass) and where burning was not used. The first step is to eliminate all treatments in the HFT data set that included timber sales. I accomplish this by eliminating all rows in the HFT data set that have identical **FACTS_ID** fields in the TS dataset. I further filter the HFT dataset by removing any planned but not executed treatments (**nbr_units1** >0 below – **nbr_units1** references **NBR_UNITS_ACCOMPLISHED** in the USFS dataset, see metadata for HFT here), and use text matching in the 'ACTIVITY' and 'METHOD' fields to remove any rows that contain reference to 'burning' or 'fire'. Finally, we remove all rows that reference 'Biomass' in the method category as it is assumed that this means material was removed for bioenergy. I use a range of 10-35 BDT/acre (mean 22.5) to convert acres reported in FACTS to volume. The following table presents descriptive statistics for estimates of residual unutilized wood biomass on an annual basis in million cubic feet.

	NFNC	NFC	P	FI	OP
count	11	4	4	4	4
mean	12.0194	17.7	28.95	66.425	2.4
std	4.68948	5.07346	16.1593	6.07639	1.79444
min	2.37421	11.2	11.2	59.6	0.3
25%	8.92407	15.025	19.525	62.225	1.275
50%	13.3557	18.5	27.75	66.85	2.5
75%	14.5349	21.175	37.175	71.05	3.625
max	17.8532	22.6	49.1	72.4	4.3

- (a) **Private industrial timber lands:** CalFIRE's Forest Practice Geographical Information System. **TODO**

3.4 Avoided Emissions from Wood Product Displacement Factors

For each product application, wood may be substituted by a range of other materials. For example, in residential construction, precast concrete and structural steel framing are competitive alternatives to wood. This choice of materials has a profound impact on GHG emissions in the construction sector and is expressed as a displacement factor (DF). A displacement factor quantifies the amount of emissions reduction achieved per unit of wood used. The displacement factors published in (Sathre and O'Connor, 2010) and used in this analysis are based on the following emission reduction sources:

1. **Reduced emissions from manufacturing:** Wood products require less total energy than to manufacture than products made from alternative materials.
2. **Avoided process emissions:** Production of wood alternatives such as cement are associated with substantial CO₂ emissions.
3. **Carbon storage in products:** Carbon in harvested wood is drawn from the atmosphere through photosynthesis and will remain fixed through the useful life of the wood product.
4. **Carbon storage in forests:** Forests producing wood continue to grow. It is assumed that forests producing wood in California are managed to sustain forest growth (not converted to non-forest land uses).
5. **Avoided fossil fuel emissions due to bioenergy substitution:** Logging and milling residuals used to produce energy avoid emissions from fossil energy sources in the energy sector.

6. **Carbon dynamics in landfills:** A fraction of carbon from wood deposited in landfills remains in semi-permanent storage. The remainder is converted to methane through biological decomposition in the landfill. Capture and use of the methane as an energy source, in turn reduces emissions from fossil energy sources.

A meta analysis conducted by (Sathre and O'Connor, 2010) compared empirical analysis from 21 international studies and found an average emissions reduction of 2.1 tons of carbon (3.9 t CO₂e) per ton of dry wood used. While studies ranged substantially around the average, the authors found that the majority of published displacement factors ranged between 1 and 3 tC/t dry wood.

//** Displacement Factors Applied to Timber Products Output

To evaluate the climate impact of harvested wood in California, I used harvested roundwood estimates from the Timber Products Output (TPO) database⁴. I used two estimates of the DF applied to the harvested wood reported in the TPO based on whether logging residuals were used in bioenergy or left in the woods (to decompose or burn).

Figure 3 reflects the flow of wood from California's forest to its fate in-use and is the frame of reference for the following analysis.

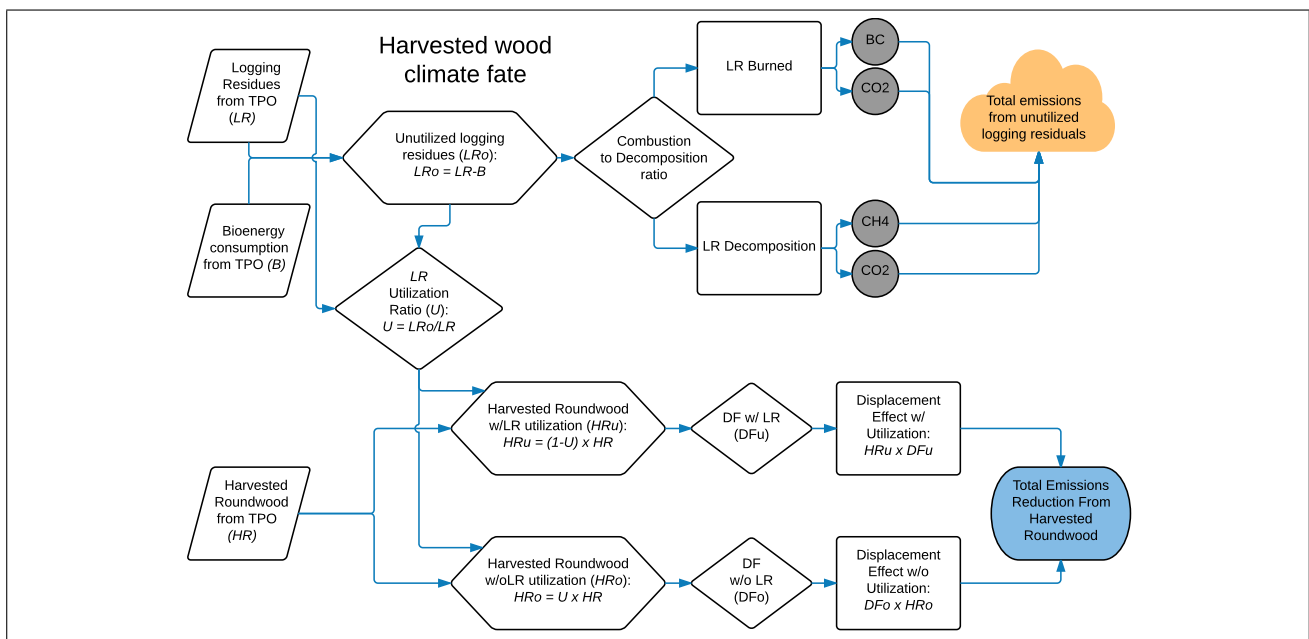


Figure 3: Wood flows from timber harvest in California

I applied displacement factors reported by Sathre and O'Connor (2010) to the reported harvest volumes from the TPO database.

The following references are used to arrive at an average displacement factor of **2.625** tCO₂e/t finished wood product for harvested roundwood without logging residue utilization.

reference	displacement factor
Eriksson et al. (2007)	1.7
Eriksson et al. (2007)	2.2
Salazar and Meil (2009)	4.9
Werner et al. (2005)	1.7

Table 6: Wood displacement factor without residue utilization

For harvested roundwood with logging residue utilization the following studies are used. I used an average of the DF reported here of **3.243** tCO₂e/t finished wood product.

The TPO reports values in terms of roundwood harvested for products, but the displacement factors presented in Sathre and O'Connor are in terms of tons of carbon in wood products. Therefore we must assume

⁴Timber Products Output Reporting Tool http://srsfia2.fs.fed.us/php/tpo_2009/tpo_rpa_int1.php

reference	displacement factor
Eriksson et al. (2007)	1.9
Eriksson et al. (2007)	2.5
Gustavsson et al. (2006)	4
Gustavsson et al. (2006)	5.6
Gustavsson et al. (2006)	2.2
Gustavsson et al. (2006)	3.3
Pingoud et al. (2001)	3.2

Table 7: Wood displacement factor with residue utilization

a milling efficiency to convert TPO volume estimates to finished wood product volume. I assumed a milling efficiency of 0.5.

Further, TPO is reported in cubic feet and the DF implies a mass unit. To convert cubic meters to a mass unit, we used the average wood density of harvested volume in California weighted by species as reported in Mciver et al. (2012). The resulting weighted average wood density used here is **27.94 lbs/cuft**.

Using the McIver and Morgan, we determine the percent of harvested wood used in bioenergy feedstocks. From personal communications with Chelsea McIver, all bioenergy feedstock reported is sourced in-woods (ie, not mill residues).

	year	bioenergy % of harvest
0	2000	0.024
1	2006	0.036
2	2012	0.082

Table 8: % volume of wood diverted to Bioenergy use by year

	Ownership	Roundwood Products	Logging Residues	Year
0	National Forest	72.4	20.7	2012
1	Other Public	16.2	3.4	2012
2	Forest Industry	328.9	72.4	2012
3	Other Private	53	11.2	2012
4	National Forest	52.8	16.3	2006
5	Other Public	1.1	0.3	2006
6	Forest Industry	274.3	59.6	2006
7	Other Private	139.2	33.2	2006
8	National Forest	90.8	22.6	2000
9	Other Public	5.2	1.6	2000
10	Forest Industry	372.5	70.6	2000
11	Other Private	159.4	49.1	2000
12	National Forest	132.1	11.2	1994
13	Other Public	24.7	4.3	1994
14	Forest Industry	396.1	63.1	1994
15	Other Private	174.7	22.3	1994

In addition to the TPO, the California Board of Equalization (BOE) also reports historic timber harvest volumes. Comparing between years where both sources report data, the BOE database reports on average, 8% less volume than the TPO (Table 9) database. This is reasonable considering that:

1. BOE data may be under-reported, as there may be a financial incentive to reduce tax burden
2. BOE does not include volume harvested from native American tribal lands in the state

year	McIver, et. al. (2012) MMBF	BOE MMBF	BOE/M&M
1978	4606.0	4491	0.98
1979	4044.0	3991	0.99
1980	3478.0	3164	0.91

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year	McIver, et. al. (2012) MMBF	BOE MMBF	BOE/M&M
1981	2832.0	2672	0.94
1982	2488.0	2318	0.93
1983	3638.0	3358	0.92
1984	3701.0	3546	0.96
1985	4093.0	3818	0.93
1986	4416.0	4265	0.97
1987	4667.0	4500	0.96
1988	4847.0	4670	0.96
1989	4699.0	4424	0.94
1990	4264.0	4021	0.94
1991	3439.0	3195	0.93
1992	3192.0	2973	0.93
1993	3041.0	2871	0.94
1994	2814.0	2316	0.82
1995	2520.0	2306	0.92
1996	2515.0	2273	0.9
1997	2640.0	2400	0.91
1998	2420.0	2091	0.86
1999	2429.0	2144	0.88
2000	2244.0	1966	0.88
2001	1801.0	1603	0.89
2002	1691.73	1690	1.0
2003	1667.95	1663	1.0
2004	1704.0305	1706	1.0
2005	1738.5	1725	0.99
2006	1960.35	1631	0.83
2007	1759.6	1626	0.92
2008	1476.0745	1372	0.93
2009	911.19	805	0.88
2010	1302.38	1161	0.89
2011	1432.5	1288	0.9
2012	1421.3	1307	0.92

Table 9: Total annual harvest reported by Mciver et al. (2012) and California Board of Equalization.

// move to appendix?//The TPO reports harvest from tribal lands, which produces an average 0.74% of the total annual harvest in the state for the 37 years of parallel data. For this analysis we used TPO data to include harvest volume from tribal lands.

year	State	Federal	Private	Tribal
1947	0.0	0.0	569.85	0.0
1948	0.0	0.0	735.29	0.0
1949	0.0	0.0	698.53	0.0
1950	0.0	0.0	808.82	0.0
1951	0.0	0.0	900.74	0.0
1952	2.57	113.79	808.82	4.78
1953	3.31	117.65	977.94	2.76
1954	2.94	141.54	880.51	4.6
1955	2.57	191.73	906.25	6.07
1956	4.41	206.99	862.13	5.33
1957	4.96	170.59	801.47	6.62
1958	5.51	208.27	821.69	6.99
1959	4.96	279.6	788.6	9.19
1960	5.15	250.37	680.15	8.82
1961	5.33	259.74	707.72	10.11
1962	6.25	259.01	744.49	8.64
1963	4.04	311.76	678.31	9.93

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year	State	Federal	Private	Tribal
1964	4.6	348.16	643.38	9.01
1965	5.7	363.05	591.91	9.74
1966	5.88	360.85	545.96	8.27
1967	6.43	355.51	562.5	7.54
1968	8.82	440.44	542.28	14.52
1969	7.35	372.61	529.41	9.93
1970	6.25	345.4	481.62	5.15
1971	7.17	383.09	476.1	12.87
1972	6.8	411.58	591.91	12.13
1973	6.07	371.69	516.54	9.38
1974	7.35	322.79	525.74	9.38
1975	6.43	287.87	498.16	3.31
1976	7.35	348.53	507.35	6.99
1977	5.15	323.35	544.12	6.99
1978	5.15	332.35	509.19	8.64
1979	4.78	321.32	417.28	8.82
1980	3.68	279.04	356.62	7.72
1981	2.76	201.65	316.18	4.04
1982	7.72	173.9	275.74	1.47
1983	7.9	313.42	347.43	2.57
1984	6.25	288.05	386.03	3.86
1985	6.62	339.52	406.25	0.92
1986	5.33	365.26	441.18	4.96
1987	7.72	364.89	485.29	7.54
1988	5.7	403.68	481.62	2.57
1989	6.8	373.53	483.46	2.02
1990	4.41	283.09	496.32	2.57
1991	6.99	248.35	376.84	4.41
1992	4.23	190.99	391.54	5.88
1993	6.25	137.32	415.44	2.39
1994	3.12	152.02	362.13	2.76
1995	7.35	101.1	354.78	2.94
1996	10.11	86.4	365.81	2.39
1997	8.64	101.65	375.0	2.76
1998	4.78	83.46	356.62	2.94
1999	0.0	97.24	349.26	0.0
2000	3.49	63.42	345.59	1.84
2001	2.94	56.07	272.06	1.84
2002	0.18	31.38	279.41	2.5
2003	0.18	28.85	277.57	3.29
2004	0.18	20.78	292.28	3.05
2005	0.18	43.66	275.74	1.95
2006	0.74	41.61	318.01	2.37
2007	0.18	58.57	264.71	3.55
2008	0.18	37.7	233.46	2.48
2009	0.18	30.37	136.95	0.72
2010	0.18	49.89	189.34	1.79
2011	0.18	55.42	207.72	2.1
2012	5.13	37.39	218.75	1.49

Table 10: Annual harvest by ownership from Mciver et al. (2012) (MCF)

To use the TPO data to estimate emissions reductions using the DF, we apply a conversion factor of **5.44** MCF/MMBF. This is an approximation as the actual sawlog conversion factor varies with average harvested log size, which has changed over time.

Using the ratio of logging residuals consumed by bioenergy (mciver), to the total logging residuals reported

in the TSP, we can calculate the harvest volume the ratio of harvest volume to logging residuals used in bioenergy, we calculated based on the ratio of reported consumption of logging residuals in bioenergy by Mciver et al. to the total logging residuals reported in the TPO. Mciver et al. report bioenergy consumption from 2000 forward. For years previous, we use the average bioenergy consumption from 2000 – 2012. These results assume bioenergy consumption throughout the reporting years. Bioenergy use of residuals did not begin until the late 1970. Further analysis is necessary to modify these results to reflect the development of the bioenergy industry.

To calculate the total emissions reduction resulting from California's timber harvest, we apply the appropriate displacement factor (with or without logging residual utilization) to the commensurate fraction of harvested roundwood. The results are shown in the following chart.

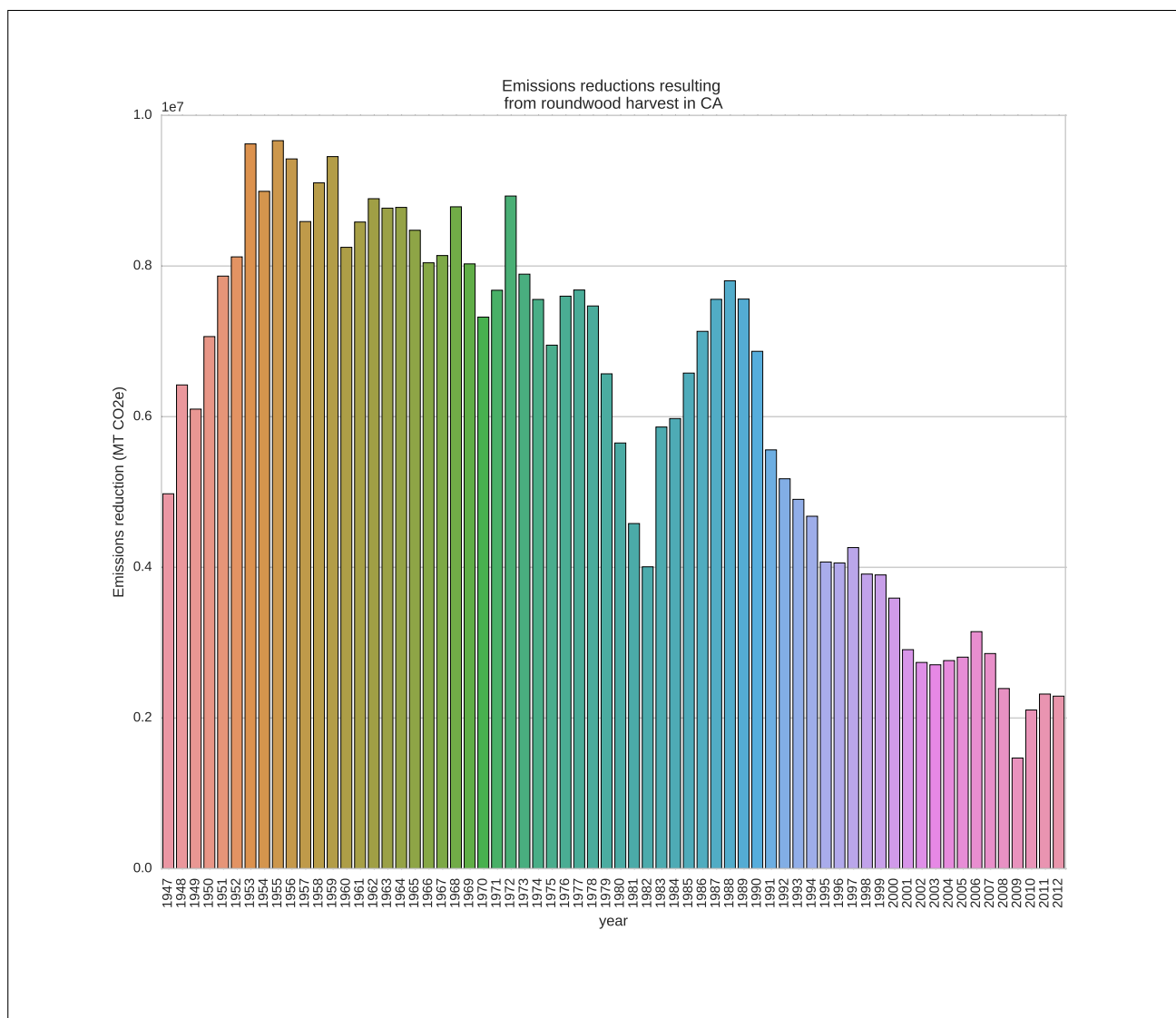


Figure 4: Historical emissions reductions resulting from harvested roundwood using displacement factors from (Sathre and O'Connor, 2010) applied to TPO data.

Contribution of the various ownership categories to the aggregate is shown in Figure 5.

4 Further Questions

This analysis is a first step towards a broader analysis of the climate impacts of harvested wood in California. The following are key questions which follow from this analysis.

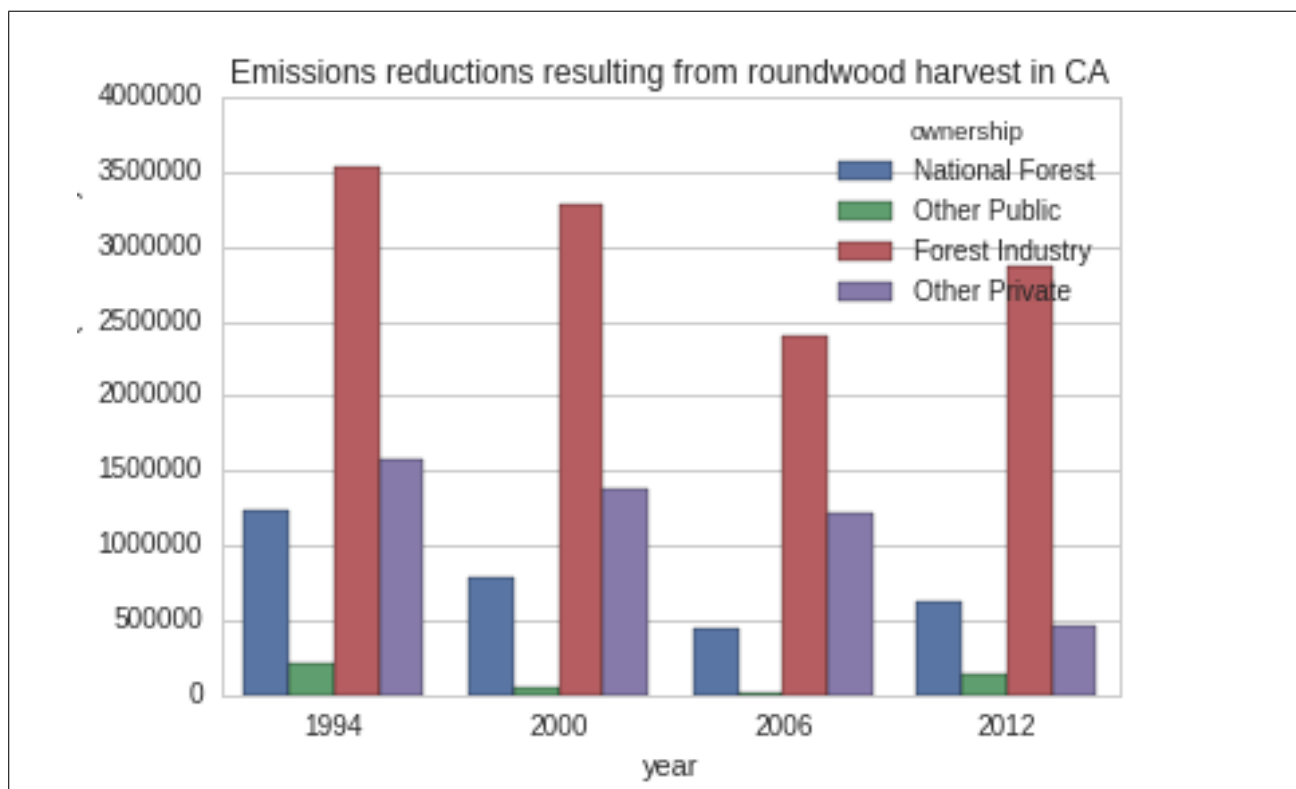


Figure 5: Historical emissions reductions by ownership for selected years resulting from harvested roundwood using displacement factors from (Sathre and O'Connor, 2010) applied to TPO data.

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