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# The temporal assembly of plant-pollinator networks following restoration

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Abstract

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The ability of communities to maintain function in the face of species extinction is related to the structure of networks. Understanding network structure and how it relates to network assembly, therefore, is a priority for system-level conservation biology. Using a long-term data set comprising nearly 20,000 pollinator visitation records, we explore the assembly of plantpollinator communities at native plant restorations in the Central Valley of California. Across years, species are highly dynamic in their network position, causing assembly to be punctuated by significant reorganizations of interactions. In contrast, the non-assembling networks did not restructure as frequently. Across all communities, Pollinator species were opportunistic in the flowers they visited. Indeed, the most persistent and generalized species were also the most variable in their network positions. High species and interaction turnover was ubiquitous across assembling and non-assembling communities, though unique interactions turnover at higher rates in assembling hedgerows as the networks continually reorganize. Nestedness of assembling networks also increased through time, but there was no increase in robustness to simulated plant extinctions. The sensitivity of networks to cascading perturbations, however, increased as the communities assembled, at least partially due to accumulating species richness. We elucidate some of the mechanisms underlying plant-pollinator network assembly and restoration, while providing further evidence that hedgerows are a valuable tool for promoting species conservation and ecosystem provision in agricultural areas.

20 community assembly, change points, specialization, nestedness, modularity, bipartite, preferential
21 attachment

# 22 Introduction

Global change has created a severe biodiversity crisis, and as species are lost, so are their interactions (Dunn et al., 2009; Barnosky et al., 2011). Because mutualistic interactions are essential 24 for maintaining the diversity their component guilds, these systems are particularly at risk from 25 coextinction cascades. The nature of these cascades will depend on the interaction patterns within a community (Memmott et al., 2004; Rezende et al., 2007; Bascompte & Stouffer, 2009; Thébault & Fontaine, 2010). To safeguard function it has become increasingly imperative to aid the recovery of lost biodiversity and interactions through ecological restoration, and a key restoration aim is to facilitate assembly of robust interaction networks (Menz et al., 2010). We know little, however, about how to re-assemble interacting communities through restoration, or the process of ecological network assembly more generally. Preferential attachment, the most widely explored mechanism of network assembly, (Barabási & 33 Albert, 1999), predicts that species entering a network are more likely to interact with species that are already well-connected ("the rich-get-richer" principle, Barabási & Albert, 1999). In pollination systems — a particularly ubiquitous mutualism (Ollerton et al., 2011; Klein et al., 2007) some studies have found support for this assembly mechanism. Investigating the day-to-day, temporal assembly of a plant-pollinator network within a season, Olesen et al. (2008) found that new species tended to interact with already well-connected species, potentially because these species 39 are either more abundant or more temporally persistent. In addition, using a space-for-time sub-40 stitution to study primary succession along a glacier foreland, Albrecht et al. (2010) also found 41 some evidence that assembly occurred through preferential attachment. Network nestedness, a pattern of interactions where a core formed by generalist species that interact with both specialist 43 and generalist species, increased as the community aged (Albrecht et al., 2010). Thus, an increase in nestedness could result from preferential attachment process where specialist species attach to the well-connected, generalist core.

In contrast to the network build-up described by preferential attachment, significant reorganizations of interactions can punctuate assembly (Peel & Clauset, 2014). Such significant reorganizations of interactions, or network changing points, are observed in social networks responding to abrupt shifts in the behavior of interactors (Peel & Clauset, 2014). In ecological communities, such shifts 50 may occur if, as new species colonize, resident species change their interaction partners to optimize 51 their foraging effort. In plant-pollinator communities, theory predicts pollinators optimize their use 52 of floral resources to reduce interspecific competition and improve resource-use efficiency (Pyke, 53 1984; Valdovinos et al., 2010, 2013; Albrecht et al., 2010; Blüthgen et al., 2007). No studies, 54 however, have examined whether changing points occur during ecological network assembly, and 55 how these changes relate to the species behavior.

Understanding network assembly is particularly relevant to ecological restoration, which is essen-57 tially 'applied succession' (e.g., Parker, 1997). In pollination systems, the time since an area was 58 restored has been shown to affect the structure of networks (Forup et al., 2008; Devoto et al., 2012), 59 suggesting interactions are changing as the community develops. Understanding the mechanisms of network assembly will help to guide community restoration. Facilitating network restoration 61 is especially imperative in areas where species interactions provide essential ecosystem services, 62 such as crop pollination. In intensively managed agricultural landscapes, the demand for pollina-63 tion services is the greatest (Kremen, 2008). However, honey bees, managed extensively around the world to provide crop pollination, are in global decline (Neumann & Carreck, 2010; van Engelsdorp et al., 2009). In addition, native pollinators, which are capable of providing sufficient crop pollination (Kremen et al., 2002; Winfree et al., 2007; Kremen et al., 2004), are in short supply because these landscapes make poor habitats for pollinator populations (Kremen et al., 2002). To ensure the continued provision of ecosystem services and curb biodiversity loss, it is critical to restore pollinators and their interactions in agricultural landscapes. To promote pollinator services in agriculture, farmers may chose to plant strips of native plants along farm edges (hedgerows) to help provide habitat for pollinators without removing arable land from production. Hedgerows augment the richness, abundance and trait diversity of pollinators in agricultural landscapes (Morandin & Kremen, 2013; M'Gonigle *et al.*, 2015; Kremen & M'Gonigle, 2015; Ponisio *et al.*, 2016). In addition, hedgerows promote the persistence and colonization of floral resource specialists (M'Gonigle *et al.*, 2015). It is important to further understand how these new species are being incorporated into the network as the community assembles, or the consequences for interaction patterns and robustness.

We explore the process of network development using a nine year dataset of plant-pollinator communities assembly following hedgerow restoration in the highly simplified and intensively managed agricultural landscape of California's Central Valley. We first determine whether the mecha-81 nism underlying network assembly is a build up of interactions as would be predicted by preferen-82 tial attachment, or punctuated by significant reorganizations of interactions (i.e., network changing 83 points). Even with changing points in interaction organization, networks could still be assembling 84 via preferential attachment if the network reorganizations were primarily driven the by peripheral, 85 temporally variable species while a stable, well-connected core of species persist. We test whether 86 the species that are most variable in their network position — and thus important contributors to 87 network reorganizations — are less persistent and connected species. To further explore the mech-88 anisms underlying the temporal dynamics of the networks, we examine patterns in the species and interaction temporal turnover. Lastly, we investigate whether networks are assembling toward predictable interaction patterns, and the ramifications for the robustness of the networks to species extinction and cascading perturbations.

## Materials & Methods

## 94 Study sites and collection methods

We surveyed plant-pollinator interaction networks of independent assembling hedgerows communities (N=5), as well as in two types of non-assembling communities to serve as controls: unrestored, weedy field margins (N=19) and mature hedgerows (greater than 10 years since planting, 97 N=29). The sites were located in the Central Valley of California in Yolo, Colusa and Solano Counties. This area is composed of intensively managed agriculture — primarily monocultures of conventional row crops, vineyards and orchards. Hedgerows we planted along field margins where they do not remove valuable land from production. Plantings are between are ca. 3-6 m wide and approximately 350 m long, bordering large (ca. 30-hectare) crop fields. Hedgerows consist of native, perennial, shrub and tree plantings including Rosa californica, Cercis occidentalis, Cean-103 othus spp., Heteromeles arbutifolia, Sambucus mexicana, Eriogonum spp., Baccharis spp., Salvia 104 spp. and others (Fig. S1 Menz et al., 2010; Kremen & M'Gonigle, 2015; M'Gonigle et al., 2015). 105 The mean distance between monitoring sites was 15 km, and the minimum distance between sites 106 of the same type sampled in the same year was 2 km. The entire area surveyed spanned almost 107 300 km<sup>2</sup>. The crop fields adjacent to all sites were similarly managed as intensive, high-input 108 monoculture. 109

Monitoring began in 2006 and continued through 2014. Surveys of the assembling hedgerows began the year before the area was restored. For logistical reasons, no sampling of assembling hedgerows was conducted in 2010. Sites were sampled between two and five times per year (Tables S1-S3). In each round of sampling, the order in which sites were sampled was randomized. Surveys were conducted under sunny conditions when the temperature was above 21°C and wind speed was below 2.5 meters/second.

Flower-visitors to plants in hedgerows and unrestored controls were netted for one hour of active

search time (the timer was paused when handling specimens). Honeybees (*Apis mellifera*) were not collected because their abundance is determined largely by hive placement by bee-keepers.

All other insect flower visitors that touched the reproductive parts of the flower were collected; however, here we focus only on wild bees and syrphids — the most abundant and effective pollinators in the system (representing 49 and 19 percent of records, respectively, C. Kremen, A. Klein and L. Morandin, unpublished data). Bee and syrphid specimens were identified to species (or morpho-species for some bee specimens in the genera *Nomada* and *Sphecodes*) by expert tax-

Quantitative networks were generated for each site through time. To account for the unequal number of surveys between years (Tables S1-S3), we use the mean of the interactions between a pair of plants and pollinators across surveys within a year to represent interaction frequency.

## 128 Change point analysis

### 129 Identifying change points

We employed a change point detection method (Peel & Clauset, 2014) to identify fundamental 130 reorganizations in large-scale interaction patterns. A change point is caused by a merge, split, fragmentation or formation of modules (also called compartments). Change point detection methods 132 have yet to be generalized to quantitative networks, so for this analysis we focused on qualitative 133 (binary) networks. Following Peel & Clauset (2014), we first defined a probability distribution 134 over the networks using the generalized hierarchical random graph model (GHRG). The GHRG model is able to capture both assortative and disassortative structure patterns at all scales in the network (Peel & Clauset, 2014). A network G is composed of vertices V and edges E. The GHRG 137 model decomposes the N vertices into a series of nested groups, the relationships among which are 138 represented by the dendrogram T. The tips of T are the vertices of G, and the probability that two vertices u and v connect is given by the parameter  $p_r$ . The probability distribution of the network G thus defined as:

$$P(G|T, pr) = p_r^{E_r} (1 - p_r)^{N_r - E_r}$$
(1)

Where  $E_r$  is the observed number of edges between vertices with the common ancestor r, and  $N_r$  is the total possible edges.

Using Bayesian posterior inference and techniques from phylogenetic tree reconstruction, we fit the GHRG model to the networks (Peel & Clauset, 2014). This is accomplished by using a Markov chain Monte Carlo (MCMC) procedure to first sample the posterior distribution of bipartitions, from which a consensus tree is derived (Peel & Clauset, 2014). We use  $\beta$  distributions with the hyperparameters  $\alpha = \beta = 1$  to define priors for  $p_r$ .

Once the GHRG model has been fit to the networks, we determine whether a change point occurred between two time slices. To detect a change point, we compare the fit of two models — one 150 where a change point occurred between two networks, and one where no change occurred — 151 using posterior Bayes factors. We chose a sliding window of length, w, of four, within which to 152 find change points. Larger windows allow for more gradual changes, and four was the maximum 153 possible with our maximum of eight years of data. Lastly, to calculate a p-value for the Bayes 154 factors, we use parametric bootstrapping to numerically estimate the null distribution (Peel & 155 Clauset, 2014). We employed code published online by L. Peel for the change point analysis. 156 Analyses were conducted in Python 3.4. 157

We next test whether the change points occurring in maturing hedgerows were a component of
the assembly process or a product of environmental shifts that lead to network reorganizations
in all types of communities. We model the number of change points as successes and the total
number of years each site was sampled as trails, and use a generalized linear model with Binomial

error to test whether the probability of a change point occurring varied by site type. We used standard techniques to determine whether the assumptions of the models were met for this and all subsequent models. For the non-assembling hedgerows and weedy field margins, only sites with five or greater survey years were included in this analysis (N=11). All statistical analysis were conducted in R 3.2.3 (R Core Team, 2015).

#### Characteristics of species that contribute to change points

To further elucidate the nature of the change points, we examine the characteristics of the species that contributed to interaction reorganization. Some species remain in relatively similar network positions through time, whereas others are more variable in their position and thus contribute more strongly to network reorganization. We test whether the more persistent species with the highest degree (number of different interaction partners) are the most stable in their network positions, as would be expected if the networks were assembling via preferential attachment.

We calculate species persistence as the proportion of surveys in which a plant or pollinator is ob-174 served. Species observed consistently within and between years are thus maximally persistent. 175 Weighted species degree is calculated from interaction observations from an extensive dataset 176 from Yolo County (approx. 18000 interaction records) that included both the data included in 177 this study and additional data from sites where we collected flower visitors using the same meth-178 ods (M'Gonigle et al., 2015; Ponisio et al., 2016). To represent network position variability, we 179 computed the coefficient of variation of weighted closeness centrality (Freeman, 1978) at each site through time. Closeness centrality describes the centrality of a species in the network by calculating path lengths to other vertices (species) in the graph (Freeman, 1978). The shorter the mean 182 path length to other species, the higher is the closeness centrality. We use linear mixed models to 183 test whether the species closeness variability (log) is related to the persistence or degree of that 184 species (Bates et al., 2014; Kuznetsova et al., 2014). We included random effects for species, as well as site. Because the degree and persistence of pollinators were strongly correlated, ( $\rho = 0.84$ , p-value  $< 2*10^{-16}$ ), we include each explanatory variable in the model separately. Plant degree and persistence were not significantly correlated, but we use the same models as we did for the pollinators for consistency. An approximately logarithmic increase in closeness centrality, as would be expected with assembly by preferential attachment, would also lead to a high variability in closeness scores, so we test whether log closeness centrality increases through time.

## Species and interaction turnover

Reorganizations of network structure can be the result of species turnover or species changing 193 their interaction partners (i.e., re-wiring). To better understand the mechanisms underlying the 194 temporal dynamics of the assembling networks, we examined patterns of species and interaction 195 turnover. For example, assembling networks may have higher rates of pollinator turnover than 196 non-assembling communities because new pollinator species are colonizing and establishing themselves (M'Gonigle et al., 2015). Similarly, if pollinators try to maximize their foraging efficiency 198 based on the species present, interactions may turnover more quickly than in established communi-199 ties. In addition, at assembling hedgerows, unvisited plants may appear to "colonize" the networks 200 as they establish new interactions with pollinators. 201

To estimate the temporal species and interaction turnover, we use an approach similar to calculating spatial  $\beta$ -diversity. Instead of calculating the variation in community composition across sites within a year, however, we estimated turnover across years at a site. We first calculated the pairwise dissimilarity of plants, pollinators and interactions between years within each site using the Chao dissimilarity estimator that incorporates abundances, while also accounting for unobserved records (Chao *et al.*, 2005). Dissimilarity estimates can be affected by the total number of species and individuals sampled at a site (e.g., Ponisio *et al.*, 2016). For example, the probability that two sites do not share any species is higher when there are few individuals at those sites. Following

Ponisio *et al.* (2016), we use null models that constrained species richness to estimate the deviation of the observed dissimilarity from that, which would be expected under a random community assembly process. With the corrected dissimilarity values, we then calculated the multivariate dispersion of community composition across years (Anderson *et al.*, 2011). In order to test whether assembling hedgerows had more species and interactions turnover than non-assembling communities, the species and interaction temporal turnover estimates were included as response variables in a linear mixed model with site type as an explanatory variable along with site as a random effect (Bates *et al.*, 2014; Kuznetsova *et al.*, 2014).

Though species may turnover across years, some groups of species may essentially replace each 218 other if they fill similar roles in the network, occupying the same network position and interact-219 ing with similar species. At non-assembling communities, species turnover may overestimate the 220 temporal changes in the networks if the interactions occurring in one year are similar to those in 221 the next year when they are weighted by the similarity of their constituent species (Fig. 1). We 222 develop a method to examine the temporal turnover of interactions with weightings based on their 223 similarity. We followed Ahn et al. (2010) algorithm to hierarchically cluster all the interactions 224 (edges) across sites and years based on their similarity, and build a dendrogram. The interaction 225 similarity is based how may plants and pollinators (vertices) two edges share (Ahn et al., 2010; 226 Kalinka & Tomancak, 2011). The more species edges shared in common, the shorter the branch 227 length between them on the dendrogram. We next calculated the temporal turnover of interactions 228 weighted by their similarity, as approximated by "phylogenetic" distance (Graham & Fine, 2008; 229 Kembel et al., 2010). When then use linear models to test whether the weighted turnover of inter-230 actions varied between assembling and non-assembling networks (Bates et al., 2014; Kuznetsova 231 et al., 2014).

# Temporal changes in interaction patterns

### Network structure

Any changing points in network structure may contribute to the reorganization of the assembling 235 networks into predictable interaction patterns. Pollination networks exhibit two main structural 236 patterns — modularity (e.g., Olesen et al., 2007) and nestedness (e.g., Bascompte et al., 2006, 237 2003). In modular networks, interactions are insular, occurring within separate groups or "mod-238 ules" more often than between modules. Modules in the network may fragment as the network 239 assembles, enhancing modularity. Conversely, nested networks are like a pyramid of interactions, where there are some species that interact with many species, other species that interact with a 241 subset of those species, and so on. If species entering the network tend to interact with the generalist base of the network pyramid as would be expected with preferential attachment, nestedness would increase through time. The connectance — the proportion of observed out of possible interactions — would decrease as new, specialist species, preferentially attach to the core. Lastly, the 245 overall level of network specialization may change as the community assembles. Network-level 246 specialization will increase if specialist species colonize the network or species begin to limit their 247 interaction niche breath as the network assembles (Blüthgen et al., 2006). 248 To evaluate network nestedness, we used the estimator weighted NODF (Almeida-neto et al.,

2008). NODF evaluates whether species with fewer partners interact with subsets of partners with which more connected species interact (Almeida-neto *et al.*, 2008). To estimate modularity, we use a hierarchical clustering algorithm (Newman & Girvan, 2004; Csardi & Nepusz, 2006). We evaluate network specialization with the metric H2, which estimates the deviation of the observed interaction frequency between plants and pollinators from a null expectation where all partners interact in proportion to their abundances (Blüthgen *et al.*, 2006). It ranges from zero for generalized networks to one for specialized networks. We calculated standardized *z*-scores so that nested-

ness, modularity and specialization metrics could be compared across communities. The *z*-scores were calculated by generating an ensemble of 999 randomly assembled communities, subtracting the mean of the statistic calculated across these communities from the observed value, and then dividing by the standard deviation. To assemble random communities, we reshuffled the interactions between species but fixed the total number of interactions, species and interaction frequency distributions (Galeano *et al.*, 2009).

To test whether network modularity, nestedness, connectance or specialization changed linearly with assembly, we use linear mixed models with the descriptive network metrics as the response variable, year of assembly as the explanatory variable, and random effects of site and year. The number of species in a network affects affects the patterns of interaction possible, so we also examined the change in plant and pollinator species richness through time. We employ generalized linear mixed models with Poisson error to model richness. We scaled explanatory variables.

## 269 Network robustness

Lastly, we test whether the changes in interaction patterns associated with network assembly affect 270 the robustness of the network to species loss and cascading perturbations. Following Memmott 271 et al. (2004), we simulate plant species extinction and the subsequent extinction cascades of polli-272 nator species. Because the reproduction of plant species if facilitated by active restoration efforts, it is unlikely the extinction of pollinator species would affect plant populations in the hedgerows. However, plants ceasing to bloom, for example in response to drought, will likely affect the pollina-275 tors that depend on them. We eliminate plants species based on their degree or abundance, and then 276 calculate the number of pollinators that secondarily went extinct. The area below the extinction 277 curve is an estimate of network robustness (Memmott et al., 2004; Dormann et al., 2008). 278

We also explored how the robustness to cascading perturbations changed as community assembled,

using algebraic connectivity — the second smallest eigenvalue of the Laplacian matrix (Fiedler, 1973) — as a proxy for network robustness (Gaiarsa et al., submitted). Algebraic connectivity relates to how difficult it is to turn a network into completely disconnected groups of species (Costa *et al.*, 2007), and has implications to how integrated is the dynamics of the elements of the networks (Gibert *et al.*, 2013). The larger the algebraic connectivity, the more robust a network is to cascading perturbations, and the easier it is to break the community into isolated groups of species (Gaiarsa et al., submitted).

## 287 Results

Over eight years and 747 samples, we collected and identified 19,547 wild bees and syrphids comprising 173 species from 50 genera. We observed 1,521 unique interactions between plants and pollinators.

# **Change point analysis**

#### 292 Identifying change points

The majority (76%) of the sites underwent at least one significant interaction reorganization (Fig. 2, 3). All five of the assembling hedgerows experienced network changing points, whereas only 40% and 81% of non-assembling hedgerows and field margins, respectively, underwent significant interaction reorganizations. Assembling hedgerows had significantly more changing points than the non-assembling networks (estimate of the difference in the odds ratios between assembling and non-assembling networks, 3.316, 95% CI [1.314, 8.572], *p*-value= 0.0117). Network assembly following restoration is thus punctuated by more interaction reorganizations than would be expected by environmental shifts alone that would effect assembling and non-assembling networks

301 equally.

#### 302 Characteristics of species that contribute to change points

In contradiction to the predictions of assembly by preferential attachment, both pollinator per-303 sistence and degree were positively related to network position variability (Fig. 4, estimate of 304 the slope of closeness centrality variability and persistence  $\pm$  standard error of the estimate, 305  $0.653 \pm 0.225$ , p-value=0.009; slope of closeness centrality variability and degree,  $0.008 \pm 0.002$ , 306 p-value=0.002). In addition, plant persistence and degree were not significantly related to net-307 work position variability (Fig. 4, estimate of the slope of closeness variability and persistence 308  $\pm$  SE,  $-2.063 \pm 3.091$ , p-value=0.5; slope of closeness variability and degree,  $0.0018 \pm 0.002$ , 309 p-value=0.3). The variability of species network position was not the result of closeness linearly 310 increasing through time, and, in fact, plant and pollinator closeness decreased slightly through time 311 (Fig. S2, estimate of the slope of closeness through time  $\pm$  SE, pollinators:  $-0.0003 \pm 0.00005$ , p-312 value= $2.7 * 10^{-12}$ ; plants  $-0.007 \pm 0.001$ , p-value= $1.4 * 10^{-6}$ ). Through statistically significant, 313 the slopes are so slight they may not be biologically significant.

## Species and interaction turnover

The rates of plant, pollinator and interaction temporal turnover were similar across assembling hedgerows, non-assembling hedgerows and field margins, though mature hedgerows has marginally significantly less pollinator turnover than field margins (Fig. 5, estimate  $\pm$  SE of the difference in turnover between field margins and mature hedgerows,  $-0.0498 \pm 0.026$ , p-value=0.058). When interactions where weighted by their similarity, both assembling and mature hedgerows had higher rates of turnover than field margins (Fig. 5, estimate  $\pm$  SE of the difference in turnover between field margins and assembling hedgerows,  $0.115 \pm 0.027$ , p-value=0.0002; field margins hedgerows,  $0.115 \pm 0.027$ , p-value=0.0002; field margins hedgerows,  $0.115 \pm 0.027$ , p-value=0.0002; field margins

gins and mature hedgerows,  $0.082 \pm 0.024$ , p-value=0.002). The weighted interaction turnover at assembling hedgerows, however, was not significantly higher than in non-assembling, mature hedgerows.

# Temporal changes in interaction patterns

#### 7 Network structure

Network nestedness significantly increased with assembly (Fig. 6, estimate of the slope of nested-328 ness through time  $\pm$  SE, 1.834  $\pm$  0.6142, p-value=0.022). All of the networks were significantly 329 nested (z-scores > 2, Fig. 6). Modularity decreased (Fig. 6), though the slope was not significantly 330 different from zero (estimate of the slope of modularity through time  $\pm$  standard error of the es-331 timate,  $-0.524 \pm 0.295$ , p-value=0.124). In addition, none of the networks were significantly 332 modular (z-scores < 2, Fig. 6). Connectance decreased as the community assembled (Fig. 6, esti-333 mate of the slope of connectance through time  $\pm$  standard error of the estimate,  $-0.0434 \pm 0.0152$ , 334 p-value=0.03). Specialization also decreased, though the slope was only marginally significantly 335 different from zero (estimate of the slope of specialization through time  $\pm$  SE,  $-0.926 \pm 0.450$ , 336 p-value=0.078). Most communities were more generalized than expected when interactions were 337 randomized (Fig. 6). 338 Both plant and pollinator species richness increased through time (Fig. 6, estimate of the slope of 339 richness through time  $\pm$  SE, pollinators: 0.193  $\pm$  0.0729, p-value=0.008; plants: 0.212  $\pm$  0.0653, 340 p-value=0.001). Unsurprisingly, pollinator species are colonizing and persisting at the assembling 341 hedgerows. Plant species richness is based on the flowers actually visited by pollinators and not the 342 presence of a particular plant species at a site. Thus, though some new plant species may establish themselves in the hedgerows, the increase in plant richness in the networks is likely due to plants attracting visitors, as they offer better rewards.

### Network robustness

Assembly did not effect the robustness of the networks to species extinction when species where removed incrementally by degree (estimate of the slope of robustness through time  $\pm$  SE, 6 \*  $10^{-5} \pm 4 * 10^{-3}$ , p-value=0.987) or abundance (0.001  $\pm$  0.003, p-value=0.65). In contrast, the sensitivity of networks to cascading perturbations, as measured by the algebraic connectivity of the network, increased as the network assembled (Fig. 7, estimate of the slope of sensitivity to cascading perturbations through time  $\pm$  SE, 0.6814  $\pm$  0.272, p-value=0.042).

## 3 Discussion

We show that the temporal assembly of plant-pollinator networks following restoration is a highly 354 dynamic process where interactions often undergo significant reorganizations, the so called chang-355 ing points. If these network reorganizations were a product of environmental forces alone, we 356 would expect to observe the same changing points at the same periods, consistently across all sites. However, network changing points in non-assembling communities are less frequent, and there are few consistent trends in when change points occurred across all sites. Several sites had 359 network changing points between years 2009 and 2011 (Fig. 2). In California, 2011 marked the 360 beginning of a multi-year drought. The assembling hedgerows were not sampled in 2010, so dis-361 entangling whether the changing points are due to skipping a year of assembly or the drought is 362 not possible. Interestingly, most assembling hedgerows did not undergo a significant interaction 363 reorganization immediately after a hedgerow was planted (i.e., the transition from weedy field 364 margin to hedgerow). This result is consistent with the finding that in our study system, hedgerow 365 restoration takes several years to have an impact on the plant-pollinator communities (Kremen and 366

367 M'Gonigle, in prep).

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In addition to finding multiple network organization changing points during assembly, the way in 368 which these reorganizations occur was different from what is expected by preferential attachment. Based on a preferential attachment process, we expect that the most persistent and high degree 370 species would remain stable in the network core during assembly (Barabási & Albert, 1999). Sur-37 prisingly, however, we encountered the opposite pattern. For example, the four most ubiquitous species in our study landscape — Halictus ligatus, Halictus tripartitus, Lasioglossum (Dialictus) 373 incompletum, and Toxomerus marginatus — were the only species that changed which module they were a member across years in all the assembling hedgerows. Because species degree and 375 persistence were strongly correlated, it is difficult to disentangle the causal mechanism for why 376 species with those characteristics are so variable in their network position. Generalized species 377 may be able to better exploit the limited floral resources in the intensively managed agriculture 378 landscape, and thus also be the most persistent (in ant-plant mutualisms, Díaz-Castelazo et al., 379 2010). More persistent species usually have longer phenologies, so they can visit many different 380 flowers, resulting in a higher degree (Vázquez et al., 2009; Fort et al., 2016). Either way, our 381 result suggests that adaptable species can change their network position to utilize the most ad-382 vantageous floral resources available, which may depend on the other pollinator species that are 383 present, and the state of the plant community (MacLeod et al., 2016; Gómez & Zamora, 2006; 384 Waser et al., 1996a). Thus given the opportunity and ability to use different resources, species will 385 often change their network positions (MacLeod et al., 2016). 386 Interestingly, though assembling hedgerows had more network reorganizations than non-assembling 387 communities, pollinator species and interaction turnover occurred at similar rates across site types. 388 Assembling hedgerows have higher turnover than non-assembling field margins only when inter-389

actions were weighted by their similarity. This is likely because though species and interactions

are turning over at the field margins, species and interactions that fill similar roles in the network

are replacing each other. In contrast, at the assembling hedgerows, unique interactions are turning over as the networks continually reorganize. Non-assembling mature hedgerow communities, 393 however, had similar rates of weighted interaction turnover as assembling hedgerows but also the 394 lowest pollinator turnover. Pollinator communities at mature hedgerows may be generally more 395 stable, but rare and/or specialized pollinators could generate this pattern if they entered a com-396 munity, formed unique interactions with plants that did not previously share pollinators, but did 397 not persist in the networks. These species would not contribute strongly to network reorganiza-398 tion or species turnover, but would enhance weighted interaction turnover. Mature hedgerows may 399 thus both support more stable pollination communities and likely ecosystem services, while also 400 providing resources for rare and/or specialized species (Kremen & M'Gonigle, 2015; M'Gonigle 401 et al., 2015). 402

When we explore the how the network-level interaction patterns changed through time, we found 403 that nestedness did increase as the community assembled, as would be expected if colonizing, 404 specialist species preferentially attached to a central, generalist core (Albrecht et al., 2010). In 405 addition, connectance decreased, as would be expected if the network is being colonized by spe-406 cialist species and the overall mean number of interactions per species did not change. However, 407 the frequent changing points in network organization, dynamic nature of species positions in the 408 networks, and turnover of species and interactions all point to an assembly mechanism other than 409 preferential attachment. The stable level of network-level specialization through the assembly 410 process may be due to the increased colonization of specialized species (M'Gonigle et al., 2015) 411 accompanied by an increase in the diet breath of resident species. This would be expected if res-412 ident species were able to minimize their foraging time by expanding their diet breath as plant diversity increases with hedgerow maturation (Waser et al., 1996b; Pyke, 1984; Blüthgen et al., 2007; Albrecht et al., 2010). Such a change in pollinator behavior would also explain the increase in network nestedness. Because so many mechanisms give rise to the same patterns of interaction, additional tests are necessary to assess the contribution of different mechanisms to community 418 assembly.

Interestingly, however, the changes in network patterns with assembly the robustness of hedgerow communities to species loss did not change with assembly. This is particularly surprising given the observed increase in nestedness, which is often associated with an enhanced in robustness to extinction (Memmott *et al.*, 2004). Perhaps assembling hedgerows have yet to reach sufficient levels of nestedness to realize the benefits nestedness confers. Nestedness of the assembling hedgerows, however, did not asymptote within the eight years following restoration that the sites were surveyed, so hedgerow networks may eventually reach sufficient levels of nestedness to gain the robustness advantage.

Contrary to the general restoration goals, the susceptibility of the networks to cascading perturba-427 tions increased as the communities assembled. Because network vulnerability to cascading perturbations, as measured by algebraic connectivity, is correlated with species richness, the increase and 429 plant and pollinator richness following restoration is at least partially responsible for the increase 430 to cascading effects. Connectance is also positively related to algebraic connectivity (Gibert et al., 431 2013), but because we observed a decrease in connectance, topological characteristics of the net-432 works beyond species richness and connectance are needed to explain the increased sensitivity to 433 perturbations spreading. These hedgerows were designed to provide floral resources to the largest 434 number of pollinators across the growing season (Menz et al., 2010). The generalized nature of 435 the floral community may explain why the networks tended to be more generalized than expected 436 if interactions were randomly distributed across species (Fig. 6). In addition, the design of the 437 hedgerow plantings may have facilitated the emergence of a single, highly connected module in 438 all of the networks (see 3 for examples). This network configuration results in short path lengths 439 (the distance between species in a network based on their shared partners), and thus, a perturbation in one species can more easily spread to other species. In order to promote more resilient communities, future restoration efforts should explore designing floral communities to promote more interaction partitioning using, for example, algorithms to optimize different network properties based on prior knowledge of pollinator floral preferences (M'Gonigle *et al.*, 2016), and on desired network architectures that renders them more robust both to species loss and to cascading effects.

Plant-pollinator networks, in general, are highly dynamic, with high turnover of species and interactions both within and between seasons (Burkle & Alarcón, 2011). Though our non-assembling communities experience fewer network reorganizations than the assembling hedgerows, 82% of 449 field margins and 40% of mature hedgerows underwent at least one changing point in network 450 structure. Pollinators are also highly opportunistic (Petanidou et al., 2008; Vázquez, 2005; Al-451 brecht et al., 2010), though trait complementarity such as tongue length and corolla depth impose 452 some biophysical limits to the interactions between plants and pollinators (Vázquez et al., 2009; 453 Vázquez et al., 2009; Stang et al., 2009, 2006; Santamaría & Rodríguez-Gironés, 2007). Such 454 opportunism may buffer plant-pollinator communities from global change (e.g., Ramos-Jiliberto 455 et al., 2012; Kaiser-Bunbury et al., 2010), but our limited understanding of the assembly of these 456 communities impedes making such predictions (Vázquez et al., 2009; Burkle & Alarcón, 2011). 457 Unlike in the broader food web literature, we have few assembly models of mutualistic network as-458 sembly (Valdovinos et al., 2013; Nuismer et al., 2013; Guimarães et al., 2011). In addition, the few 459 developed models often borrow their mechanisms from competitive interactions, leading to inac-460 curate biological assumptions (Holland et al., 2006). We need further development of mechanistic 461 models of mutualistic systems to generate testable predictions, along with empirical exploration 462 of network assembly. Plant-pollinator communities and mutualisms broadly are vital for biodi-463 versity maintenance and essential ecosystem service provision. We must therefore understand the 464 processes underlying their assembly to facilitate restoration and conservation.

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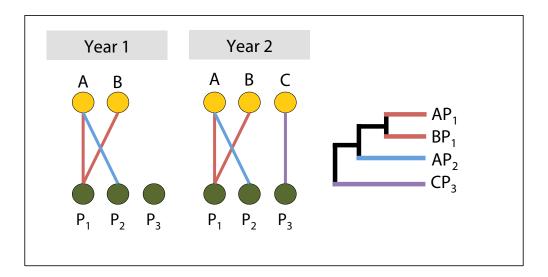


Figure 1: Diagram illustrating the analysis to examine the temporal turnover of interactions weighted based on their similarity. A, B and C are animal species, and Ps are plant species. The dendrogram depicts the interaction similarity across years based on the number of shared constituent species.

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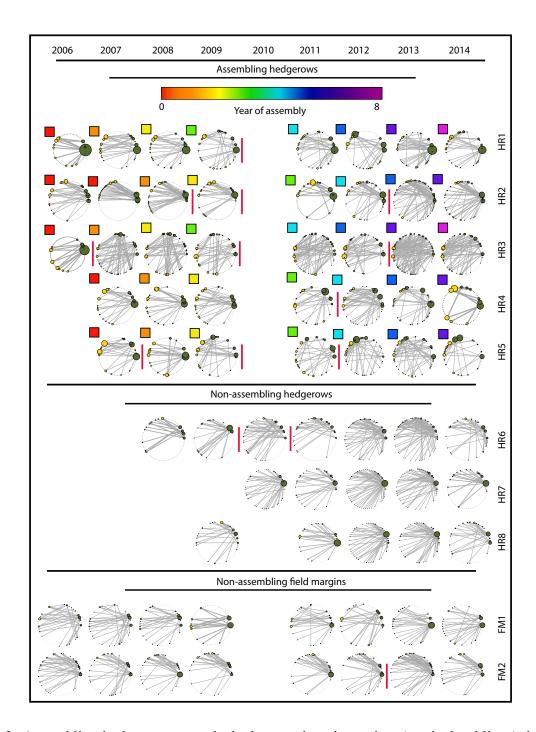


Figure 2: Assembling hedgerow networks had more changing points (vertical red lines) than non-assembling hedgerows and weedy field margins (a representative sample of non-assembling sites are depicted here). In each network, plants and pollinators are represented by green and yellow circles, respectively, weighted by their degree. Each species has a consistent position in the network across years. In the assembling hedgerows, colored squares in the corner of each network represent the year of assembly. Before hedgerows are installed (when they are still field margins) the year of assembly is zero.

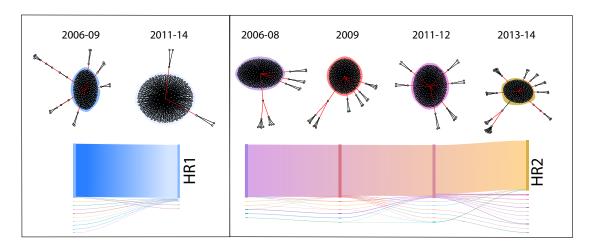


Figure 3: The community (i.e., module or compartment) membership between network changing points. Two representative assembling hedgerows are depicted. In the top panel, species are grouped by community. The bottom panels visualize the flow of species between communities between changing points.

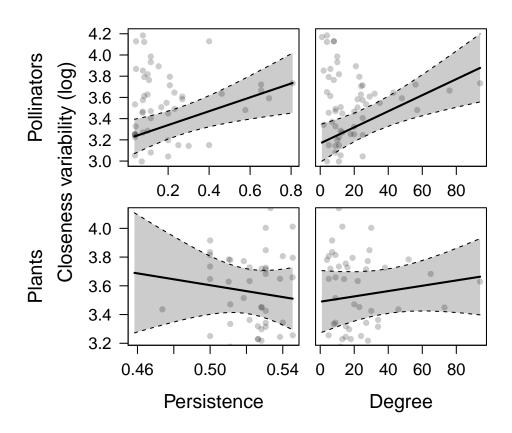


Figure 4: The variation coefficient of network position, as represented by closeness, plotted against pollinator persistence and degree. Persistence and degree were positively related to network position variability in pollinators, but unrelated in plants. Points represent means for each species across sites. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.

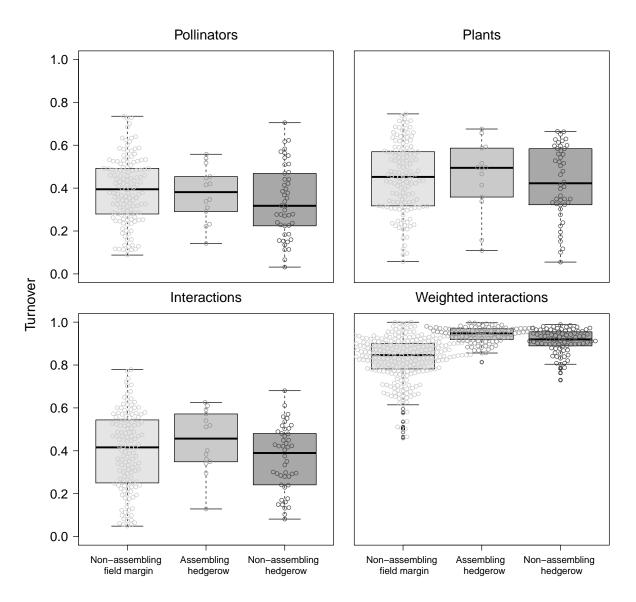


Figure 5: The species, interaction and weighted interactions turnover of plant-pollinator networks at non-assembling field margins sites, assembling hedgerows, and non-assembling, mature hedgerows. Rates of species and interaction turnover were similar between site types, though mature hedgerows has marginally significantly less pollinator turnover. However, when interactions where weighted by their similarity, both hedgerow types had higher turnover that unrestored field margins. Boxplots represent medians (black horizontal line) first and third quartiles (box perimeter) and extremes (whiskers).

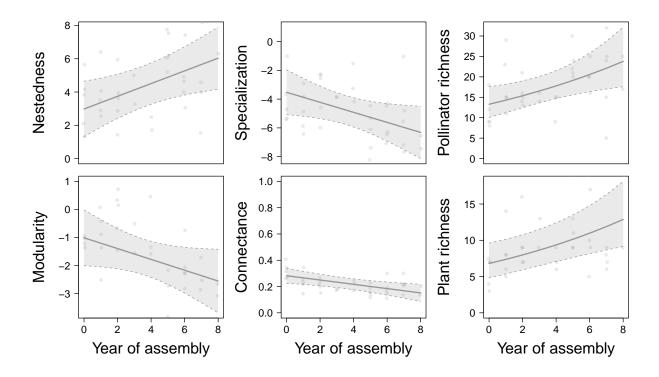


Figure 6: Nestedness, plant richness and pollinator richness increased as the networks assembled. Specialization and modularity remained consistent across years, while connectance decreased. The nestedness, modularity and specialization scores represent z-scores. Scores greater than  $\sim 2$  or less than  $\sim -2$  are significantly more or less structured than randomly assembled networks. Points are the metric value for each site at each year of assembly. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.

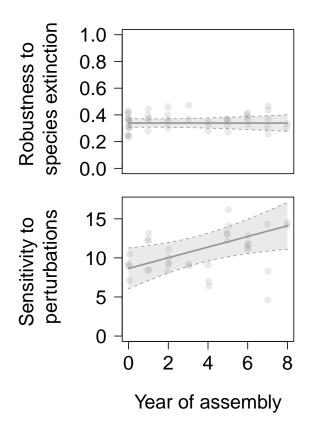


Figure 7: The robustness of networks to species extinction did not change with network assembly, but the sensitivity to cascading perturbations increased. The robustness to species extinction is measured by incrementally removing species by degree, though removing species by abundance did not yield qualitatively different results. The robustness of networks to cascading perturbations is measured as the algebraic connectivity, the second smallest eigenvalue of the Laplacian matrix. Points are the value for each site at each year of assembly. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.