Running head: PLANT-POLLINATOR NETWORK ASSEMBLY

# The temporal assembly of plant-pollinator networks following restoration

Lauren C. Ponisio<sup>1,2</sup>, Marilia P. Gaiarsa<sup>3</sup>, Claire Kremen<sup>1</sup>

- Department of Environmental Science, Policy, and Management University of California, Berkeley 130 Mulford Hall Berkeley, California, USA 94720
- Department of Entomology
   University of California, Riverside
   417 Entomology Bldg.
   Riverside, California, USA
   92521
- 3. Departamento de Ecologia Universidade de Sao Paulo Sao Paulo, SP, Brazil 05508-900

1 Abstract

TO BE RE-WRITTEN The structure of networks is related to ability of communities to
maintain function in the face of species extinction. Understanding network structure and how
it relates to network disassembly, therefore, is a priority for system-level conservation biology.
We explore the assembly of plant-pollinator communities on native plant restorations in the
Central Valley of California.

- 7 Keywords: changing points, temporal networks, hedgerows, species interactions, preferential at-
- 8 tachment, mutualisms

# 9 Introduction

Global change has created a severe biodiversity crisis, and as species are lost, so are their interactions (Dunn *et al.*, 2009; Barnosky *et al.*, 2011). Because mutualistic interactions are essential
for maintaining the diversity their component guilds of species, these systems are particularly at
risk from coextinction cascades. The nature of these coextinction cascades depends on the interaction patterns within a community (Memmott *et al.*, 2004; Rezende *et al.*, 2007; Bascompte &
Stouffer, 2009). Recovering the lost biodiversity and interactions through ecological restoration
has become increasingly imperative, and a key restoration aim is to facilitate assembly of robust
interaction networks (Menz *et al.*, 2010). We know little, however, about how to re-assemble interacting communities through restoration, or the process of ecological network assembly more
generally.

The mostly widely explored mechanism of network assembly, preferential attachment (Barabási & 20 Albert, 1999), predicts that a new species is more likely to interact with species that are already 21 well-connected ("the rich-get-richer" principle, Barabási & Albert, 1999). In pollination systems 22 — a particularly ubiquitous mutualism (Ollerton et al., 2011; Klein et al., 2007) — some studies have found support for this mechanism of assembly. Investigating the day-to-day, temporal assembly of a plant-pollinator network within a season, Olesen et al. (2008) found that new species tended to interact with already well-connected species, likely because these species are either more 26 abundant or more temporally persistent (Olesen et al., 2008). In addition, using a space-for-time 27 substitution to study primary succession along a glacier foreland, Albrecht et al. (2010) found 28 some indication assembly was occurring through preferential attachment. Network nestedness, a 29 pattern of interactions where a generalist core interacts with both specialist and generalist species, 30 increased as the community aged (Albrecht et al., 2010). Increasing nestedness could result from a 31 process like preferential attachment where specialist species attach to the well-connected, general-32 ist core. In addition, non-successional temporal dynamics also suggest a stable core of generalists

- persists despite high turnover of peripheral species (Fang & Huang, 2012; Díaz-Castelazo et al.,
- <sup>35</sup> 2010; Alarcón et al., 2008).
- In contrast to the ordered network build-up described by preferential attachment, assembly can be
- punctuated by significant reorganizations of interactions (Peel & Clauset, 2014). For example, as
- new species are added, resident species change their interaction partners to minimize competition,
- or become extinct. Such significant reorganizations of interactions, or changing points, have been
- observed in networks responding to abrupt shifts in the behavior of interacters (Peel & Clauset,
- 41 2014). No studies, however, have examined whether changing points occur during ecological
- network assembly.
- 43 Understanding network assembly is particularly relevant to ecological restoration, which is essen-
- 44 tially 'applied succession' (Parker, 1997, e.g.,). In pollination systems, the time since an area was
- restored has been shown to effect the structure of networks (Forup et al., 2008a,b; Devoto et al.,
- 46 2012), suggesting interactions are evolving as the community develops. Understanding the mech-
- 47 anisms of network assembly will help to guide the restoration of particular communities.
- Facilitating effective restoration of networks is particularly imperative in areas where species in-
- teractions provide essential ecosystem services, such as crop pollination. In intensively managed
- <sub>50</sub> agricultural landscapes, the demand for pollination services is the greatest (Kremen, 2008). How-
- ever, honey bees, managed extensively around the world to provide crop pollination, are in global
- decline (Neumann & Carreck, 2010; van Engelsdorp et al., 2009). In addition, native pollina-
- tors, which have the capacity to provide sufficient crop pollination (Kremen et al., 2002; Winfree
- et al., 2007; Kremen et al., 2004), are in short supply because these landscapes make poor habitats
- 55 for pollinator populations (Kremen et al., 2002). To ensure provision the continued provision of
- ecosystem services and curb biodiversity loss, effective restoration of pollinators and their interac-
- 57 tions in agricultural landscapes is critical.
- To promote pollinator services in agriculture, farmers are increasingly turning to the habitat restora-

tion technique of planting strips of native plants along farm edges (hedgerows) to help provide
habitat for pollinators without removing arable land from production. Hedgerows have been
shown to augment the richness, abundance and trait diversity of pollinators in agricultural landscapes(Morandin & Kremen, 2013; M'Gonigle *et al.*, 2015; Kremen & M'Gonigle, 2015; Ponisio *et al.*, 2015). In addition, hedgerows promote the persistence and colonization of floral resource
specialists (M'Gonigle *et al.*, 2015). Little is known however, about the assembly of the network
following hedgerow restoration.

Using a long-term data-set of plant-pollinator communities assembling following hedgerow restoration in the highly simplified and intensively managed agricultural landscape of California's Central 67 Valley, we explore the process of network development. We first determine whether the mecha-68 nism underlying network assembly is a smooth build up of interactions as would be predicted by 69 preferential attachment, or punctuated by significant reorganizations of interactions (i.e., changing 70 points). Even with changing points in interaction organization, networks could still be assembling 71 via preferential attachment if the network reorganizations were primarily driven the by peripheral, 72 temporally variable species while a stable, well-connected core of species still persists. We thus 73 examine whether the species are most variable in their network position — and thus important contributors network reorganizations — are less persistent and connected species. Lastly, we examine 75 whether networks are assembling toward predictable interaction patterns, and the ramifications for the robustness of the networks to species extinction and perturbation.

## 78 Materials & Methods

## 79 Study sites and collection methods

We surveyed plant-pollinator interaction networks of assembling hedgerows (N=5), as well as in two types of non-assembling communities to serve as controls: unrestored, weedy field margins 81 (N=19) and established hedgerows (greater than 10 years since planting, N=29). The sites were 82 located in the Central Valley of California in Yolo, Colusa and Solano Counties. This area is comprised of intensively managed agriculture – primarily monocultures of conventional row crops, vineyards and orchards. Hedgerows we planted along field margins where they do not remove valuable land from production, and are ca. 3-6m wide and approximately 350m long and border large (ca. 30-hectare) crop fields. Hedgerows consist of native, perennial, shrub and tree plantings including Rosa californica, Cercis occidentalis, Ceanothus spp., Heteromeles arbutifolia, Sambucus mexicana, Eriogonum spp., Baccharis spp., Salvia spp. and others (Kremen & M'Gonigle, 2015; M'Gonigle et al., 2015). The mean distance between monitoring sites was 15 km, and the minimum distance between sites of the same type sampled in the same year was 2 km. The en-91 tire area surveyed spanned almost 300 km<sup>2</sup>. The crop fields adjacent to all sites were similarly managed as intensive, high-input monoculture.

Monitoring began in 2006 and continued through 2015. Sampling of the assembling hedgerows

began the year before the area was restored. For logistical reasons, no sampling of assembling

hedgerows was conducted in 2010. Sites were sampled between two and five times per year (Ta-

bles S1-S3). In each round of sampling, the order in which sites were sampled was randomized.

Surveys were conducted under sunny conditions when the temperature was above 21°C and wind

99 speed was below 2.5 meters/second.

Flower-visitors to plants in hedgerows and unrestored controls were netted for one hour of active search time (the timer was paused when handling specimens). Honeybees (*Apis mellifera*) were

not collected because their abundance is determined largely by the placement of hives throughout
the region by bee-keepers. All other insect flower visitors that touched the reproductive parts of
the flower were collected; however, here we focus only on wild bees and syrphids (representing
49 and 19 percent of records, respectively), the most abundant and effective pollinators in the
system (C. Kremen, A. Klein and L. Morandin, unpublished data). Bee and syrphid specimens
were identified to species (or morpho-species for some bee specimens in the genera *Nomada* and *Sphecodes*) by expert taxonomists.

Quantitative networks were generative for each site through time. To account for the unequal number of surveys between years, the mean of the interactions between a pair of plants and pollinators across surveys within a year was used as a representation of the frequency of interactions.

# 112 Change point analysis

#### 113 Identifying change points

We employed a change point detection method (Peel & Clauset, 2014) to identify fundamental 114 changes in large-scale pattern of interactions of plants and pollinators. A change point is caused by 115 a merge, split, fragmentation or formation of communities (also called modules or compartments). Change point detection methods have yet to be generalized to quantitative networks, so for this 117 analysis we focused on qualitative (binary) networks. Following Peel & Clauset (2014), we first 118 defined a probability distribution over the networks using the generalized hierarchical random graph model (GHRG). The GHRG model is able to capture both assortative and disassortative community structure patterns at all scales in the network (Peel & Clauset, 2014). A network G is composed of vertices V and edges  $E \subseteq VV$ . The GHRG model decomposes the N vertices into a 122 series of nested groups, the relationships among which are represented by the dendrogram T. The 123 tips of T are the vertices of G, and the probability that two vertices u and v connect is given by the parameter  $p_r$ . The probability distribution of the network G thus defined as:

$$P(G|T, pr) = p_r^{E_r} (1 - p_r)^{N_r - E_r}$$
(1)

Where  $E_r$  is the observed number of edges between vertices with the common ancestor r, and  $N_r$  is the total possible edges.

Using Bayesian posterior inference and techniques from phylogenetic tree reconstruction, we fit the GHRG model to the networks (Peel & Clauset, 2014). This is accomplished by using a Markov Chain Monte Carlo (MCMC) procedure to first sample the posterior distribution of bipartitions, from which a consensus tree is derived (Peel & Clauset, 2014). We used  $\beta$  distributions with the hyperparameters  $\alpha = \beta = 1$  to define priors for  $p_r$ .

Once the GHRG model has been fit to the networks, we determine whether a change point occurred between two time slices. To detect a change point, we compare the fit of two models – one where a change point had occurred between two networks, and one where no change occurred – using posterior Bayes factors. We chose a sliding window of length, w, of four, within which to find change points. Larger windows allow for more gradual changes, and four was the maximum possible with our maximum of nine years of data. Lastly, to calculate a p-value for the Bayes factors, we use parametric bootstrapping to numerically estimate the null distribution (Peel & Clauset, 2014). The change point analysis was carried out using code published online by L. Peel. Analyses we conducted in Python 3.4.

We next test whether the change points occurring in maturing hedgerows were a component of the assembly process or a product of environmental shifts that lead to network reorganizations in all types of communities. We model the number of change points as successes and the total number of years each site was sampled as trails, and use a generalized linear model with Binomial error to test whether the probability of a change point occurring varied by site type. For the non-assembling

hedgerows and weedy field margins, only sites with five or greater years of sampling was included in this analysis. All statistical analysis were conducted in R 3.2.3 (R Core Team, 2015).

#### 149 Characteristics of species that contribute to change points

To further elucidate the nature of the change points, we examine the characteristics of the species that contributed the reorganization of interactions. Some species remain in relatively similar network positions through time, whereas others are more variable in their role and thus contribute more strongly to network reorganization. We test the that more persistent species with the highest degree are the most stable in their network position, as would be expected is the networks were assembling via preferential attachment.

We calculate species persistence as the proportion of the surveys a pollinator was observed. Species 156 observed consistently within and between years are thus maximally persistent. Weighted species 157 degree is calculated from interaction observations observed in more extensive data-set from Yolo 158 County (approx. 18000 interaction records) that included both the data included in this study and 159 additional data from sites where we collected flower visitors using the same methods (M'Gonigle 160 et al., 2015; Ponisio et al., 2015). To represent the the variability of species within networks, 161 we computed the coefficient of variation of weighted closeness at each site through time. Close-162 ness describes the centrality of a species in the network by calculating path lengths to other vertices 163 (species) in the graph. We used linear mixed models to test whether the variability of species close-164 ness values was related to the persistence or degree of that species (Bates et al., 2014; Kuznetsova et al., 2014). We included random effects for species, as well as site. We focused on the pollinator species because the hedgerow flowers are planted and thus are not directly assembling. Because 167 degree and persistence were strongly correlated, ( $\rho = 0.84$ , p-value  $< 2 * 10^{-16}$ ), each explana-168 tory variable was included in the model separately. Because a linear increase in closeness, as might 169 be expected with assembly by preferential attachment, would lead to a high variability in closeness scores, we also test whether closeness increases through time.

# 72 Temporal changes in interaction patterns

#### 73 Network structure

The changing points in network structure may contribute to the reorganization of the assembling networks into predictable interaction patterns. Pollination networks exhibit two main structural 175 patterns — modularity (e.g., Olesen et al., 2007) and nestedness (e.g., Bascompte et al., 2006, 176 2003). In modular networks, interactions are insular, occurring within separate groups or "mod-177 ules" more often than between modules. Communities in the network may fragment as the network 178 assembles, enhancing modularity. Conversely, nested networks are like pyramid of interactions, 179 where there are some species that interact with many species, other species that interact with a 180 subset of those species, and so on. If species entering the network tend to interact with the gen-181 eralist base of the network pyramid (i.e., via preferential attachment), nestedness would increase 182 through time. Lastly, if the network is accumulating specialist species or species are beginning 183 to limit their interaction niche breath as the network assembles, this would lead to an increase in 184 the network-level specialization (Blüthgen et al., 2006). To test whether network modularity, nest-185 edness or specialization changed linearly with assembly, we used linear mixed models with the descriptive network metrics as the response variable, year of assembly as the explanatory variable, and random effects of site and year.

We use NODF (weighted or unweighted) to evaluate network nestedness (Almeida-Neto *et al.*, 2008). NODF evaluates whether species with fewer partners interact with subsets of partners with which more connected species interact (Almeida-Neto *et al.*, 2008). To estimate modularity, we use a hierarchical clustering algorithm (Newman & Girvan, 2004; Csardi & Nepusz, 2006). We

calculated standardized z-scores so that nestedness and modularity metrics could be compared between communities. The z-scores were calculated by generating an ensemble of 999 randomly 194 assembled communities, subtracting the mean of the statistic calculated across these communities 195 from the observed value, and then dividing by the standard deviation. To assemble random com-196 munities, we reshuffled the interactions between species but fixed the total number of interactions, 197 species and the distribution of the interaction frequencies (Galeano et al., 2009). Lastly, Network 198 specialization was measured using H2, which estimate the deviation of the observed interaction 199 frequency between plants and pollinators from a null expectation where all partners interact in 200 proportion to their abundances (Blüthgen et al., 2006). It ranges from 0 for generalized networks 201 to 1 for specialized networks. 202

#### Network structure

The changing points in network structure may contribute to the reorganization of the assembling networks into predictable interaction patterns. Pollination networks exhibit two main structural 205 patterns — modularity (e.g., Olesen et al., 2007) and nestedness (e.g., Bascompte et al., 2006, 206 2003). In modular community interactions are insular, occurring within separate groups or "modules" more often than between modules. Communities in the network may fragment as the network 208 assembles, enhancing modularity. Conversely, nested networks are like pyramid of interactions, 209 where there are some species that interact with many species, other species that interact with a sub-210 set of those species, and so on. If species entering the network tend to interact with the generalist 211 base of the network pyramid, nestedness would increase through time. Alternatively, if the net-212 work is accumulating specialist species or if species are beginning to limit their interaction niche 213 breath as the network assembles, this would lead to an increase in the network-level specialization 214 (Blüthgen et al., 2006) and nestedness would decrease through time. To test whether network mod-215 ularity, nestedness or specialization changed linearly with assembly, we used linear mixed models with the descriptive network metrics as the response variable, year of assembly as the explanatory variable, and random effects of site and year.

#### Network robustness

Lastly, we test the changes in interaction patterns associated with network assembly affect the 220 robustness of the network to species loss and to cascading effects. Following Memmott et al. 221 (2004), we simulate the extinction of plant species the subsequent extinction cascades of pollinator 222 species. Because the reproduction of plant species if facilitated by active restoration efforts, it is 223 unlikely the extinction of pollinators would affect plant populations in the hedgerows. However, 224 plants ceasing to bloom, for example in response to drought, will likely affect the pollinators that 225 depend on them. Plants species were eliminated based on their degree or abundance, and the 226 number of pollinators that secondarily went extinct is calculated. The area below the extinction 227 curve is used as a measure of network robustness (Dormann et al., 2008).

We also explored how the robustness to cascading effects changed as community assembled, using algebraic connectivity— the second smallest eigenvalue of the Laplacian matrix (Fiedler, 1973)— as a proxy for network robustness (e.g. Gaiarsa et al., submitted). Algebraic connectivity was first used to describe spectral properties of complex graphs other than ecological networks, and it is related to how difficult it is to turn a network into completely disconnected groups of nodes, or species (Fiedler, 1973; Costa *et al.*, 2007), and also to flows in . The larger the algebraic connectivity, the more robust a network is to perturbations (e.g. Gaiarsa et al., submitted).

# 236 Results

Over eight years and 747 samples, we collected and identified 19,547 wild bees and syrphids comprising 173 species from 50 genera. We observed 1,521 unique interactions between plants and pollinators.

## **Change point analysis**

## Identifying change points

The majority (76%) of the sites tests underwent at least one significant reorganization of interactions (Fig. 1). All five of the assembling hedgerows experienced changing points, whereas only 40% and 81% of non-assembling hedgerows and field margins, respectively, underwent significant interaction reorganizations. Assembling hedgerows experienced significantly more changing points than the non-assembling networks (difference in the odds ratios between assembling and non-assembling networks, 3.316, 95% CI [1.314, 8.572], *p*-value= 0.0117). Network assembly following restoration is thus punctuated by more interaction reorganizations than would be expected by environmental shifts that would effect assembling and non-assembling networks equally.

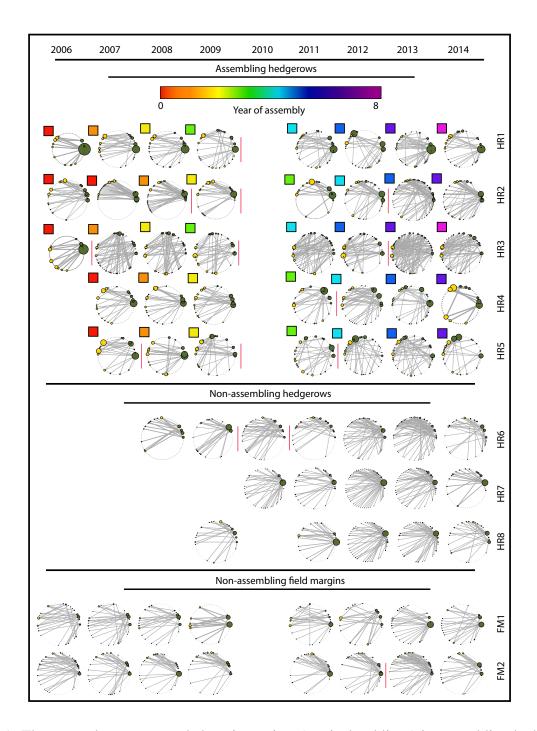


Figure 1: The network structure and changing points (vertical red lines) in assembling hedgerows and a representative sample of non-assembling hedgerows and weedy field margins. In each network, plants and pollinators are represented by green and yellow circles, respectively, weighted by their degree. Each species has has a consistent position in the network across years. In the assembling hedgerows, colored circles in the corner of each network represent the years post restoration.

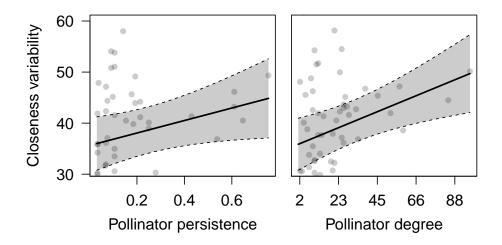


Figure 2: The coefficient of variation of network position, as represented by closeness, plotted against pollinator persistence and degree. Points represents means for each species across sites. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.

### **Characteristics of species that contribute to change points**

- 251 In contradiction to the predictions of assembly by preferential attachment, both pollinator persis-
- 252 tence and degree were positively related to network position variability. (ADD STATS IF KEEP-
- 253 ING RESULT).

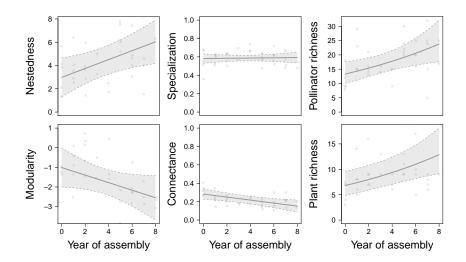


Figure 3: The change in nestedness, modularity, specialization and connectance as the networks assemble. The left panels represent z-scores. Scores greater than  $\sim 2$  or less than  $\sim -2$  are significantly more or less structured than randomly assembled networks. Points are the value for each site at each year of assembly. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.

# Temporal changes in interaction patterns

#### Network structure

Network nestedness significantly increased with assembly (estimate of the slope of nestedness through time  $\pm$  standard error of the estimate,  $1.834\pm0.6142$ , p-value=0.022, Fig. 3). Modularity decreased (Fig. 3), though the slope was not significantly different from zero (estimate of the slope of modularity through time  $\pm$  standard error of the estimate,  $-0.524\pm0.295$ , p-value=0.124). Specialization remained relatively constant (estimate of the slope of specialization through time  $\pm$  standard error of the estimate,  $0.003\pm0.015$ , p-value=0.827).

## Network robustness

- 263 Assembly did not effect the robustness of the networks to species extinction when species where re-
- $_{264}$  moved incrementally by degree (estimate of the slope of robustness through time  $\pm$  standard error
- of the estimate,  $6*10^{-5} \pm 4*10^{-3}$ , p-value=0.987) or abundance (0.001  $\pm$  0.003, p-value=0.65,
- 266 Fig. 4).
- <sup>267</sup> In contrast, the robustness of networks to perturbation, as measured by the algebraic connectivity
- of the network, increased as the network assembled (estimate of the slope of robustness through
- time  $\pm$  standard error of the estimate, 0.6814  $\pm$  0.272, p-value=0.042, Fig. 4).

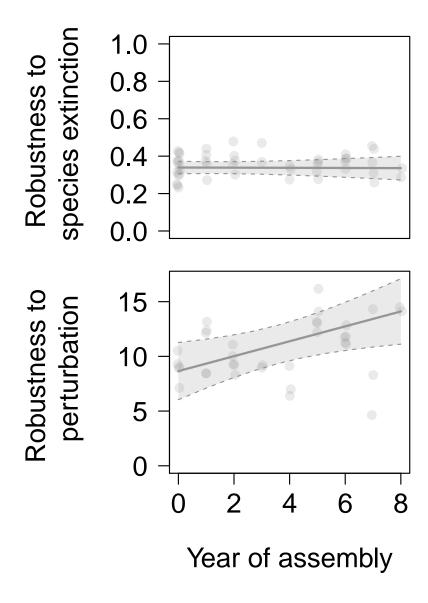


Figure 4: The robustness of networks to species extinction and perturbation. The robustness to species extinction is measured by incrementally removing species by degree, through removing species by abundance did not yield qualitatively different results. Points are the value for each site at each year of assembly. The solid line indicates the mean slope estimate and the dashed lines are the 95% confidence intervals around the estimate.

## **Discussion**

293

We show that the temporal assembly of plant-pollinator networks following restoration is a highly 271 dynamic process where patterns of interactions undergo several significant reorganizations. These 272 network organizations are unlikley to be a project of environmental forces because the changing 273 points in non-assembling communities are both less frequent and in different time points. The 274 reorganizations are thus a product of the assembly process. There were no consistent trends in when change points occurred within assembling hedgerows or 276 across all sites, except many sites had changing points between year 2009 and 2011 (Fig. 1). In 277 California, 2011 marked the beginning of a multi-year drought. In the assembling hedgerows were 278 not sampled in 2010, so disentangling whether the changing points are due to skipping a year of 279 assembly or the drought is not impossible. Interestingly, most of the assembling hedgerows did 280 not undergo a significant reorganization of interactions immediatly the hedgerow was planted (i.e., 281 the transition from weedy field margin to hedgerow). This result is consistent with other studies 282 finding that the gains in species richness of both the plant and pollinators takes several years post-283 restoration (Kremen and M'Gonigle, in prep). Surprisingly, the species that were most variable in their network position and thus contributed 285 most to network reorganizations where species with the highest degrees (i.e., most generalized) 286 and persistence. For example, the five most ubiquitous species in our study landscape — Halictus 287 ligatus, Halictus tripartitus, Lasioglossum (Dialictus) incompletum, and Toxomerus marginatus — 288 where also the only species that changed what module (i.e., community) they were present in across years in all of the assembling hedgerows. Because species degree and persistence were strongly correlated, it is difficult to disentable the causal mechanism for why species with their chatacterists 29 are so variable in their network position. Generalized species may be best able to exploit the limited 292 floral resources the intensivly managed agriculture landscape, and thus also most persistent. These adaptable species will take advanatge of the most advantage resources, which my depend the other pollinator species that are present and the state of the floral resource. More persistent species also have longer phenologies, so they have the opportunity to visit many different flowers, resulting in a high degree. Either way, our results suggest that given the opportunity and ability to use different resources, species will often their network positions.

The frequent changing points in network organization, dynamic nature of the location of species in networks, and turnover of species and nodes all point to an assembly mechanism other than 300 preferential attachment. Nestedness did increase with years post restoration, as would be expected 30 if colonizing, specialist species attached to a central, generalist core. With preferential attachment, 302 however, we could also expect connectance and specialization to increase, and see neither. The 303 increase in nestedness is thus more likely a consequence of the enhancement of species richness 304 through time DO WE HAVE CITATIONS FOR THIS? In addition, though the colonization and 305 persistence of floral specialists increases with assembly (M'Gonigle et al., 2015), this must be ac-306 cumpanied changes in the diet breat of the present species in order for network-level specialization 307 to remain constant through time. 308

Preferential attachment is the most widely accepted theory explaining network assembly, but few 309 ecological studies have been conducted that test for this mechanism (Olesen et al., 2008). other 310 citations?? Though Olesen et al. (2008) found some evidence for an assmbly process like pre-311 frential attachment governming the temporal build up of plant-pollinator networks within a season, 312 the frequent reorganizations of the hedgerow communities suggest a very different mechanism is at 313 play. Plant-pollinator networks in general are highly dynamic, with high turnover of species and in-314 teractions both within and between seasons (Burkle & Alarcón, 2011). Though our non-assembling 315 communities experience fewer network reorganizations than the assembling hedgerows, 82% of 316 field marings and 40% of mature hedgerows underwent at least one changing point in network 317 structure. Few alternative models to perferential attachment that predict these network reorganiza-318

tions have been explored, however.

Complementarity between traits like tounge length and corolla depth as well as body and flower impose some biophysical limits to the interactions between plants and pollintors (Vázquez *et al.*, 2009; Stang *et al.*, 2009, 2006; Santamaría & Rodríguez-Gironés, 2007). Similarly, the phenologies of plants and pollinators must also overlap (Vázquez *et al.*, 2009; Vázquez *et al.*, 2009).

324 (Valdovinos et al., 2013)

The changes in network patterns with assembly also increased the robustness of the networks to perturbation, but not to plant species loss.

The increase in nestedness has been related to an increase in the robustness of communities to 327 species extinctions because the core of generalist species would act as a buffer against species loss 328 (REF). Similarly, communities with greater modularity would be more robust in general, because 329 modules would act as a buffer against perturbation spreading and the pervasive effects of species 330 loss. We expected that as communities assemble and species richness increase, communities would 331 become increasingly robust to species loss. however, we found that robustness to species loss did 332 not change through time. This is interestingly from the conservation biology point of view, because 333 even early assembling communities seem to be as robust as more mature communities, despite the 334 richness insurance hypothesis. However, communities are not only subjected to species loss, but 335 also to other types of perturbations that can cascade through the network. Gaiarsa et al. (submitted) suggest using algebraic connectivity to explore how vulnerable ecoogical communities might be to 337 cascading effects. Our results indicate that more mature communities are more robust to cascading effects than early on assemblages, and that this result is related to species richness. This migth be related to XX....

# 41 Acknowledgments

We would like to thank Paulo Guimarães Leto Peel and Aaron Clauset for their invaluable discussions and for help with the change point analysis. We thank the growers and land owners that allowed us to work on their property. We also greatly appreciate the identification assistance of expert taxonomists Martin Hauser, Robbin Thorp and Jason Gibbs. This work was supported by funding from the Army Research Office (W911NF-11-1-0361 to CK), the Natural Resources Conservation Service (CIG-69-3A75-12-253, CIG-69-3A75-9-142, CIG-68-9104-6-101 and WLF-69-7482-6-277 to The Xerces Society), the National Science Foundation (DEB-0919128 to CK), The U.S. Department of Agriculture (USDA-NIFA 2012-51181-20105 to Michigan State University). Funding for LCP was provided by an NSF Graduate Research Fellowship and the USDA NIFA Graduate Fellowship. Funding for MPG was provided by São Paulo Research Foundation (FAPESP, grant 2013/13319-5).

## References

- Alarcón, R., Waser, N.M. & Ollerton, J. (2008). Year-to-year variation in the topology of a plant–pollinator interaction network. *Oikos*, 117, 1796–1807.
- Albrecht, M., Riesen, M. & Schmid, B. (2010). Plant–pollinator network assembly along the chronosequence of a glacier foreland. *Oikos*, 119, 1610–1624.
- Almeida-Neto, M., Gumarães, P., Gumarães, P.R., Loyola, R.D. & Ulrich, W. (2008). A consistent metric for nestedness analysis in ecological systems: reconciling concept and measurement.

  Oikos, 117, 1227–1239.
- Barabási, A.L. & Albert, R. (1999). Emergence of scaling in random networks. *science*, 286, 509–512.

- Barnosky, A.D., Matzke, N., Tomiya, S., Wogan, G.O., Swartz, B., Quental, T.B., Marshall, C.,
- McGuire, J.L., Lindsey, E.L., Maguire, K.C. et al. (2011). Has the earth/'s sixth mass extinction
- already arrived? *Nature*, 471, 51–57.
- Bascompte, J., Jordano, P., Melián, C.J. & Olesen, J.M. (2003). The nested assembly of plant-
- animal mutualistic networks. *Proc. Natl. Acad. Sci. USA*, 100, 9383–9387.
- Bascompte, J., Jordano, P. & Olesen, J.M. (2006). Asymmetric coevolutionary networks facilitate
- biodiversity maintenance. *Science*, 312, 431–433.
- Bascompte, J. & Stouffer, D.B. (2009). The assembly and disassembly of ecological networks.
- *Phil. Trans. R. Soc. B*, 364, 1781.
- Bates, D., Maechler, M., Bolker, B. & Walker, S. (2014). lme4: Linear mixed-effects models using
- Eigen and S4. R package version 1.1-7.
- Blüthgen, N., Menzel, F. & Blüthgen, N. (2006). Measuring specialization in species interaction
- networks. *BMC Ecol.*, 6, 9.
- Burkle, L.A. & Alarcón, R. (2011). The future of plant-pollinator diversity: Understanding in-
- teraction networks across time, space, and global change. American Journal of Botany, 98,
- <sub>378</sub> 528.
- <sup>379</sup> Costa, L.d.F., Rodrigues, F.A., Travieso, G. & Villas Boas, P.R. (2007). Characterization of com-
- plex networks: A survey of measurements. Advances in physics, 56, 167–242.
- Csardi, G. & Nepusz, T. (2006). The igraph software package for complex network research.
- InterJournal, Complex Systems, 1695.
- Devoto, M., Bailey, S., Craze, P. & Memmott, J. (2012). Understanding and planning ecological
- restoration of plant–pollinator networks. *Ecology letters*, 15, 319–328.

- Díaz-Castelazo, C., Guimaraes Jr, P.R., Jordano, P., Thompson, J.N., Marquis, R.J. & Rico-Gray,
- V. (2010). Changes of a mutualistic network over time: reanalysis over a 10-year period. Ecol-
- оду, 91, 793–801.
- Dormann, C., Gruber, B. & Fründ, J. (2008). Introducing the bipartite package: analysing ecolog-
- ical networks. *R News*, 8, 8.
- Dunn, R.R., Harris, N.C., Colwell, R.K., Koh, L.P. & Sodhi, N.S. (2009). The sixth mass coextinc-
- tion: are most endangered species parasites and mutualists? Proceedings of the Royal Society
- of London B: Biological Sciences, 276, 3037–3045.
- van Engelsdorp, D., Evans, J., Saegerman, C., Mullin, C., Haubruge, E., Nguyen, B., Frazier, M.,
- Frazier, J., Cox-Foster, D., Chen, Y., Underwood, R., Tarpy, D. & Pettis, J. (2009). Colony
- collapse disorder: A descriptive study. *Plos One*, 4, 1–17.
- Fang, Q. & Huang, S.Q. (2012). Relative stability of core groups in pollination networks in a
- biodiversity hotspot over four years. *PloS one*, 7, e32663.
- Fiedler, M. (1973). Algebraic connectivity of graphs. Czechoslovak mathematical journal, 23,
- <sup>399</sup> 298–305.
- Forup, M., Henson, K., Craze, P. & Memmott, J. (2008a). The restoration of ecological interac-
- tions: plant-pollinator networks on ancient and restored heathlands. J. Appl. Ecol., 45, 742–752.
- 402 Forup, M.L., Henson, K.S., Craze, P.G. & Memmott, J. (2008b). The restoration of ecological
- interactions: plant–pollinator networks on ancient and restored heathlands. Journal of Applied
- 404 Ecology, 45, 742–752.
- Galeano, J., Pastor, J.M. & Iriondo, J.M. (2009). Weighted-interaction nestedness estimator (wine):
- a new estimator to calculate over frequency matrices. Environmental Modeling & Software, 24,
- 407 1342–1346.

- Klein, A., Vaissière, B., Cane, J., Steffan-Dewenter, I., Cunningham, S., Kremen, C. & Tscharntke,
- T. (2007). Importance of pollinators in changing landscapes for world crops. *Proc. R. Soc. B*,
- 410 274, 303–313.
- Kremen, C. (2008). Bee Pollinators in Agricultural Ecosystems, New York: Oxford University
- Press, chap. Crop Pollination Services From Wild Bees, pp. 10–26.
- Kremen, C. & M'Gonigle, L.K. (2015). Small-scale restoration in intensive agricultural landscapes
- supports more specialized and less mobile pollinator species. *J. Appl. Ecol.*, 52, 602–610.
- Kremen, C., Williams, N., Bugg, R., Fay, J. & Thorp, R. (2004). The area requirements of an
- ecosystem service: crop pollination by native bee communities in California. *Ecol. Lett.*, 7,
- 417 1109–1119.
- 418 Kremen, C., Williams, N. & Thorp, R. (2002). Crop pollination from native bees at risk from
- agricultural intensification. *Proc. Natl. Acad. Sci. U.S.A.*, 99, 16812–16816.
- Kuznetsova, A., Bruun Brockhoff, P. & Haubo Bojesen Christensen, R. (2014). *ImerTest: Tests*
- for random and fixed effects for linear mixed effect models (lmer objects of lme4 package). R
- package version 2.0-11.
- 423 Memmott, J., Waser, N.M. & Price, M.V. (2004). Tolerance of pollination networks to species
- extinctions. *Proc. R. Soc. Lond. B*, 271, 2605–2611.
- Menz, M., Phillips, R., Winfree, R., Kremen, C., Aizen, M., Johnson, S. & Dixon, K. (2010).
- Reconnecting plants and pollinators: challenges in the restoration of pollination mutualisms.
- 427 *Trends Plant Sci.*, 16, 4–12.
- M'Gonigle, L., Ponisio, L., Cutler, K. & Kremen, C. (2015). Habitat restoration promotes pollina-
- tor persistence and colonization in intensively-managed agriculture. Ecol. Appl., 25, 1557–1565.

- Morandin, L. & Kremen, C. (2013). Hedgerow restoration promotes pollinator populations and exports native bees to adjacent fields. *Ecol. Appl.*, 23, 829–839.
- Neumann, P. & Carreck, N.L. (2010). Honey bee colony losses. J. Api. Res., 49, 1–6.
- Newman, M.E.J. & Girvan, M. (2004). Finding and evaluating community structure in networks.
- *Phys. Rev. E*, 69, 026113.
- Olesen, J.M., Bascompte, J., Dupont, Y. & Jordano, P. (2007). The modularity of pollination networks. *Proc. Natl. Acad. Sci. USA*, 104, 19891–19896.
- Olesen, J.M., Bascompte, J., Elberling, H. & Jordano, P. (2008). Temporal dynamics in a pollination network. *Ecology*, 89, 1573.
- Ollerton, J., Winfree, R. & Tarrant, S. (2011). How many flowering plants are pollinated by animals? *Oikos*, 120, 321–326.
- Parker, V.T. (1997). The scale of successional models and restoration objectives. *Restoration ecology*, 5, 301–306.
- Peel, L. & Clauset, A. (2014). Detecting change points in the large-scale structure of evolving networks. *arXiv preprint arXiv:1403.0989*.
- Ponisio, L.C., M'Gonigle, L.K. & Kremen, C. (2015). On-farm habitat restoration counters biotic homogenization in intensively managed agriculture. *Global change biology*.
- R Core Team (2015). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna, Austria.
- Rezende, E.L., Lavabre, J.E., Guimarães, P.R., Jordano, P. & Bascompte, J. (2007). Non-random coextinctions in phylogenetically structured mutualistic networks. *Nature*, 448, 925–928.

- Santamaría, L. & Rodríguez-Gironés, M.A. (2007). Linkage rules for plant–pollinator networks:
- Trait complementarity or exploitation barriers? *PLoS Biol.*, 5, e31.
- Stang, M., Klinkhamer, P., Waser, N.M., Stang, I. & van der, M.E. (2009). Size-specific interaction
- patterns and size matching in a plant-pollinator interaction web. Ann. Bot., 103.
- Stang, M., Klinkhamer, P.G. & Van Der Meijden, E. (2006). Size constraints and flower abundance
- determine the number of interactions in a plant–flower visitor web. *Oikos*, 112, 111–121.
- <sup>457</sup> Valdovinos, F.S., Moisset de Espanés, P., Flores, J.D. & Ramos-Jiliberto, R. (2013). Adaptive
- foraging allows the maintenance of biodiversity of pollination networks. *Oikos*, 122, 907–917.
- <sup>459</sup> Vázquez, D.P., Blüthgen, N., Cagnolo, L. & Chacoff, N.P. (2009). Uniting pattern and process in
- plant-animal mutualistic networks: a review. *Annals of Botany*, 103, 1445–1457.
- <sup>461</sup> Vázquez, D.P., Chacoff, N.P. & Cagnolo, L. (2009). Evaluating multiple determinants of the
- structure of plant-animal mutualistic networks. *Ecology*, 90, 2039–2046.
- Winfree, R., Williams, N., Dushoff, J. & Kremen, C. (2007). Native bees provide insurance against
- ongoing honey bee losses. *Ecol. Lett.*, 10, 1105–1113.