# The collapse of a keystone forage fish species

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## Abstract

Capelin are the linchpin of the Northwest Atlantic ecosystem as they are the primary energy conduit from lower to higher trophic levels. According to acoustic monitoring surveys conducted by Canada and the former USSR, the Newfoundland (NAFO Division 2J3KL) capelin stock suffered an order of magnitude decline in biomass in 1990. This collapse was concomitant with drastic changes in the ecosystem during the late 1980s and early 1990s, and the capelin collapse was considered a key signal in the identification of a regime shift that occurred in the early 1990s. However, while more than a dozen studies have provided evidence supporting a capelin stock collapse, there is also literature that supports a non-collapse hypothesis of capelin. The non-collapse hypothesis purports that rather than collapsing in 1991 the capelin stock either (1) changed its migratory patterns with the timing of the acoustic survey remaining constant, leading to a spatio-temporal mismatch between the survey and the stock, or (2) capelin became less migratory and remained inshore year round, therefore being largely underestimated by the offshore surveys. We tested the collapse and non-collapse hypotheses using multiple independent datasets, which included both fishery-dependent (inshore commercial catch) and fishery-independent (spring and fall acoustic and fall bottom trawl surveys, oceanography cruises, capelin larval indices, aerial surveys, predator diet and behavior) data, and diverse statistical methods. The hypothesis of a non-collapse of the capelin stock was rejected by our analyses. The weight of evidence approach led us to conclude that the Newfoundland capelin stock suffered a population collapse in 1990 with minimal recovery over the subsequent 25 years.

## Introduction

Forage fish play crucial roles in many ecosystems, transferring the bulk of the energy from lower trophic levels to large vertebrate predators. Typically, they are small shoaling species that are characterized by short life expectancy and rapid but variable growth driven by environmental factors. Forage fish species often exhibit ‘boom and bust’ population dynamics, i.e. their abundances change rapidly and substantially and undergo phases of extremely high and extremely low abundances (Soutar & Issacs 1969, Schwartzlose et al. 1999, Chavez et al. 2003, Pikitch et al. 2012, Pikitch et al. 2014). Forage fish species can experience prolonged periods of ‘bust’ dynamics. For example, the Norwegian spring-spawning Atlantic herring (*Clupea harengus*) stock collapsed in the late 1960s after a pulse of overfishing and remained at very low levels until the late 1980s (Toresen & Østvedt 2000, Dragesund et al. 2008, Skagseth et al. 2015); while sardine (*Sardinops sagax*) and anchovy (*Engraulis* spp.) have decade-scale regimes of high and low abundances where populations thrived for 20 to 30 years and then disappeared for similar periods (Schwartzlose et al. 1999, Chavez et al. 2003, Barange et al. 2009).

Capelin (*Mallotus villosus*) is the focal forage fish species in ecosystems of the North Atlantic Ocean (Templeman 1948, Jangaard 1974, Vilhjálmsson 1994, Carscadden et al. 2001). The three most important capelin populations in the North Atlantic are in the Barents Sea, off the coast of Iceland, and along the Newfoundland and Labrador (Canada) continental shelf. The Barents Sea capelin stock experienced four collapses over the past 4 decades: the mid- to late-1980s, the mid-1990s, the mid-2000s, and the mid-2010s. The size of the stock fluctuated between 3 and 7 million tonnes during the ‘boom’ phase and around 200 thousand tonnes during the ‘bust’ phase. There is general agreement that ecosystem changes were the driving force behind these ‘boom-bust’ dynamics (Gjøsæter et al. 2009). The Icelandic capelin stock underwent similar dynamics, with three ‘bust’ phases over the past 4 decades: the early 1980s, the early 1990s, and most of the 2000s. The size of the stock was around 1.5-2 million tonnes during the ‘boom’ phase and between 100-500 thousand tonnes during ‘bust’ phase (ICES 2017). The first two ‘bust’ phases were due to a combination of poor recruitment and the stock being easily available to the fishing fleet, while the most recent ‘bust’ phase was likely associated with a climate-related shift in distribution (Pálsson et al. 2012, Carscadden et al. 2013).

Fisheries and Oceans Canada (DFO) is responsible for the assessment of the Newfoundland and Labrador (NL) capelin stock. DFO concluded that the stock experienced an order of magnitude decline in 1990 with minimal recovery during the past two decades (DFO 1994, Miller 1994, 1997, DFO 2008, 2010, 2013, 2015, 2018). The size of the stock fluctuated between 2 and 6 million tonnes between 1982 and 1991, and between 25 and 900 thousand tonnes during the ensuing period (DFO 2015). This decline was concomitant with drastic changes in the ecosystem during the late 1980s and early 1990s (deYoung & Rose 1993, Gomes et al. 1995, Montevecchi & Myers 1997, Lilly et al. 2000, Rice 2002, Rose 2007, Koen-Alonso et al. 2010, Hammill et al. 2011, Pedersen et al. 2017), including major changes in the biology and ecology of capelin, such as delayed and protracted spawning, changes in their geographical and vertical distribution, and declines in somatic condition and size and age at maturity (Frank et al. 1996, Carscadden & Nakashima 1997, Carscadden et al. 2001, Mowbray 2002, Nakashima & Wheeler 2002, DFO 2010). The collapse of capelin is considered a key signal in the identification of the regime shift that occurred in the early 1990s.

The non-collapse hypothesis postulates that the capelin stock in 1991 did not collapse but that the offshore acoustic surveys have failed to detect large capelin aggregations since 1991 because of a spatio-temporal mismatch between the surveys and the stock (Frank et al. 2016). Specifically, the non-collapse hypothesis states that post-1991 either (1) capelin changed their migratory patterns while the timing of the acoustic survey remained constant, leading to a spatio-temporal mismatch between the survey and the stock or (2) capelin became non-migratory in 1990-91 and are therefore undetected by the offshore surveys. Support for the non-collapse hypothesis was based on changes in the biology of capelin post-1991 (e.g., distribution and demography), re-analysis of the offshore research surveys (multi-species bottom trawl and acoustic), and the response of various components of the ecosystem [e.g., zooplankton, Atlantic cod (*Gadus morhua*), seabirds, seals] to the large-scale changes that have occurred since the early 1990s (Frank et al. 2016). The objective of this paper is to assess the empirical support for the hypothesis of capelin stock collapse (DFO 2015) versus that of non-collapse (Frank et al. 2016) using all available data.

**Methods**

To test the hypotheses of collapse and non-collapse of the capelin stock in NAFO Divisions 2J3KL (hereafter Div. 2J3KL; Fig. 1), we applied the weight of evidence approach using multiple, independent data sets and diverse statistical methods (e.g., triangulation, sensu Munafò & Davey Smith 2018). In some sections, new data and analyses were presented to test alternative hypotheses. Each section concludes by weighting evidence in support for each alternative hypothesis. Once this was completed for all sections, the weight of evidence approach was used to determine which hypothesis was best supported by the combination of previously-published results and additional analyses based on independent data sets.

## Capelin

### Offshore capelin distribution: acoustic surveys

The capelin collapse hypothesis is supported by the fall and spring offshore acoustic surveys conducted by Canada and the former USSR that found a sudden decrease in capelin biomass in the fall of 1990 in Div. 2J3KL (e.g., Miller & Lilly 1991, Bakanev 1992, Miller 1992, 1993, 1994, Mowbray 2014). The non-collapse hypothesis purports that Canada and the USSR offshore acoustic surveys found low capelin biomasses in the fall of 1991 and 1992 in Div. 2J3KL because capelin did not migrate offshore starting in the fall of 1990 (Frank et al. 2016).

From 1982 to 1992, Canada conducted fall (October) acoustic surveys for capelin in Div. 2J3K (Fig. 1) (e.g., Miller and Carscadden 1984, Miller and Lilly 1991). The fall acoustic surveys targeted the maturing portion of the stock during winter feeding migrations to provide estimates of the size and number of maturing fish being recruited to the fishery the following year (Mowbray 2014). The capelin stock in Div. 2J3K was also acoustically surveyed by the former USSR in November from 1982-1992 (Bakanev 1992; Fig. 1). The USSR fall acoustic surveys were conducted approximately a month later than the Canadian surveys and consistently estimated higher capelin abundances, which may have been due to timing, as capelin start aggregating into large overwintering shoals in November (Winters 1995). In 1990, the Canadian fall acoustic survey reported a very low capelin biomass (96 thousand tonnes, < 5% of the biomass surveyed in 1989) while the USSR acoustic survey reported the smallest biomass since 1984 (631 thousand tonnes) (Winters 1995). Both the USSR and Canadian acoustic surveys reported record low capelin biomass in the fall of 1991 and 1992 (16-55 thousand tonnes) (Winters 1995). The decrease in capelin biomass in both surveys corresponded with very few capelin being detected off the coast of Labrador and a southward shift in stock distribution to the southern portion of Div. 3K and northern portion of Div. 3L (Miller & Lilly 1991, Miller 1992). A Canadian expanded fall survey (Div. 2J3KL) in 1993-94 was conducted to determine if the ‘missing’ capelin could be located. However, the expanded fall acoustic survey confirmed the findings of the 1991-92 fall surveys that offshore capelin biomass was low and characterized by a southward change in distribution (Miller 1994, 1995). Post-1991, capelin is only acoustically surveyed in the spring in Div. 3L by Canada (1996, 1999-2005, 2007-2015, 2017-18) with the discontinuation of the other three acoustic surveys in the early 1990s (Canada fall acoustic survey in 1994; USSR fall and spring acoustic surveys in 1992 and 1994, respectively).

While the offshore acoustic surveys provided strong evidence of a capelin collapse, they did not survey the inshore area. The non-collapse hypothesis uses unquantified densities of overwintering maturing capelin and large schools of immature capelin observed during winter surveys conducted in Trinity Bay in 1967-68 (Winters 1970) as evidence that significant densities of capelin can inhabit the inshore year-round. DFO tested the hypothesis of year-round residency of capelin in the inshore by conducting seasonal acoustic surveys in Trinity Bay (September and October 2003; January, June and September 2004-05) and expanding the annual offshore spring acoustic survey into Trinity Bay (1999-2005, 2007-13, 2017) ([Fig. 1](#Ref514161259); see supplementary section for details on methods). Seasonally, capelin densities were low in Trinity Bay in January and May, and the maximum mean density of capelin was observed in June (10,000 kg km-2), which corresponded with the start of the spawning period when capelin were highly aggregated inshore ([Fig. 2](#Ref514161259) a). In September and October, capelin densities were low once again (Fig. 2 a). There was also a distinct seasonal pattern in the age and maturity composition inshore. In January, overwintering fish were composed of ~70% immature age-1 and age-2 fish ([Fig. 2](#Ref514161271) b, c); the relative contribution of older fish increased through the spring as maturing age-2 and age-3 fish migrated into Trinity Bay ([Fig. 2](#Ref514161271) b, c); and by October, immature age-1 fish dominated the inshore area, strongly suggesting that spent mature fish either died or left the Bay (Fig. 2 b). The seasonal surveys found no evidence of a large inshore, non-migratory capelin stock. In agreement with this finding, an inshore acoustic survey in January 2000 for overwintering cod from Conception Bay to Notre Dame Bay found concentrations of juvenile capelin (O’Driscoll and Rose 2001) but few older fish (G.A. Rose, unpublished data).

With little evidence of a change in capelin migratory behavior inshore post-1991, we ... We assumed that the portion of the capelin stock available to each of the five acoustic surveys (spring and fall offshore Canada and USSR; and Trinity Bay inshore) has remained constant through time, plus or minus inter-annual variability due to changes in migration routes such as those observed in Icelandic capelin stocks (Olafsdottir & Rose 2012). If this is true, then the trends observed in each survey are theoretically generated by the same population process. Using this logic, we fitted a state-space Ricker-logistic model to the acoustic estimates of capelin biomass from the five acoustic surveys conducted in Div. 2J3KLNO. Under this classic discrete population model, the expected biomass of individuals in year *y* is modeled as a function of the biomass in year *y-1*,

Where *r* is interpreted as the intrinsic growth rate of the population, *K* is the carrying capacity of the environment, and is the process error which is assumed to follow a normal distribution with a mean of 0 and a standard deviation of σ. All surveys are assumed to be ‘observing’ the same latent process and, as such, survey indices are modeled as a function of ,

Here are catchability parameters that adjust stock biomass to the scale of the surveys and represents the observation error of each survey which is assumed to follow a normal distribution with a mean of 0 and a standard deviation of . Under this formulation, all catchability parameters could not be estimated freely, therefore the catchability parameter of the Canadian spring acoustic survey was fixed to 1 because it is the longest running time series and it also has a history of obtaining the largest population estimates. The model was constructed and fit using TMB (Kristensen et al. 2016). The model provides a good fit to each of the five survey indices and strongly supports a dramatic collapse in capelin biomass from 1990 to 1991 (Fig. 3). While this model does not rule out the possibility that the population became non-migratory and remained inshore (i.e. the catchability of all acoustic surveys dropped dramatically in 1990), it is telling that the standard deviation around the inshore acoustic survey of Trinity Bay is nearly six times that estimated for the spring acoustic survey (1.26 vs. 0.22, respectively). This means that, in the context of a Ricker-logistic population model, the Trinity Bay acoustic survey is providing little information on the population dynamics of capelin. If the majority of the capelin population were inshore, then one would expect inshore indices to be much more influential. Moreover, capelin biomass estimates from the inshore Trinity Bay acoustic survey in the spring are typically a small fraction of the estimates from the offshore (~10% of the spring acoustic survey). It is, therefore, unlikely that the 3-6 million tonnes (Mt) of ‘missing’ capelin reside along the coast of NL.

In summary, while the fall and spring acoustic surveys in Div. 2J3KLNO support the collapse of the capelin stock, they cannot refute the hypothesis that capelin are non-migratory post-1991 as inshore areas were not systematically surveyed. The lack of significant inshore aggregations of capelin outside of the peak spawning period during seasonal and annual spring acoustic surveys does, however, provide support for the hypothesis of a capelin stock collapse in the early 1990s.

**Offshore capelin distribution: annual multi-species bottom-trawl surveys**

The non-collapse hypothesis used the fall bottom trawl survey (FBTS) data to point to a westerly, inshore shift in the center of capelin concentration in 1996-2010 compared to 1985-1995 (Frank et al. 2016). However, the center of concentration of capelin using the FBTS presence/absence data from 1985-1995 found inshore distributions occurred in three high abundance years (1986-1988, Fig S2 in Frank et al. 2016). This variability is likely related to the poor catchability of capelin in the Engel otter trawl (used from 1978-1994), which was designed for harvesting commercial groundfish like flatfish and Atlantic cod. In 1995, the gear used for sampling in the FBTS was changed to a Campelen 1800 shrimp trawl, which improved the catchability of capelin in the survey. A similar center of gravity analysis using only post-1995 FBTS data showed a southerly shift in capelin distribution with a recent shift to the northwest in 2011-14 (DFO 2015). However, neither of these analyses accounted for inter-annual changes in capelin spatial distribution due to FBTS sampling effort nor considered the uncertainty around the center of gravity estimates.

To address the abovementioned issues and to test the non-collapse hypothesis of a shoreward shift in capelin distribution post-1991, we revisited the center of gravity analysis of the FBTS data using the approach described in Thorson et al. (2016). We used the VAST package in R (Thorson & Barnett 2017) to fit a geostatistical delta-generalized linear mixed model to estimate the spatial and temporal distribution of capelin (Thorson et al. 2016). The advantages of this approach are that it accounts for inter-annual changes in the spatial distribution of sampling effort and offers a means of estimating the standard error of the center of gravity metric, which provides perspective on the significance of distributional shifts. Like DFO (2015), we focused on the post-1995 period when the catchability of capelin improved with the use of a Campelen 1800 shrimp trawl. Our geostatistical analysis does not support the hypothesis that capelin shifted their distribution towards the inshore post-1995 and there is no evidence of an easterly or westerly movement in the center of gravity of capelin ([Fig. 4](#Ref514161325)). Instead, the center of gravity of capelin remains > 100 km offshore and demonstrates pronounced shifts in the north-south axis ([Fig. 4](#Ref514161325)).

Like the acoustic surveys, the FBTS has poor survey coverage of the inshore. It is, therefore, possible that significant aggregations of capelin could go unnoticed. Nonetheless, simple back-of-the-envelope calculations indicate that it is unrealistic to assume that the 3 to 6 Mt of capelin that are ‘missing’ in the offshore surveys are now residing in the inshore. Even though the inshore strata are inconsistently covered by the annual FBTS with an inshore area of ~35,000 to ~71,000 km2 remaining un-surveyed each year, the minimum density for 3 to 6 Mt of capelin hiding in these inshore waters would have to be between ~41,000 to ~170,000 kg km-2, uniformly distributed throughout the un-surveyed area. It is unlikely that such densities of capelin would go unnoticed by the FBTS; furthermore, if they were present, then the center of gravity would be oriented towards the inshore during years when more inshore strata were surveyed (1996-1998, 2000-2002, 2004-2006, 2010, 2013), which it is not (Fig. 4). These inshore density estimates also appear unrealistic in the context of existing inshore acoustic estimates of capelin density. Specifically, the maximum mean density of capelin observed in the Trinity Bay acoustic survey in June was 10,000 kg km-2 , and the maximum mean density of capelin in Trinity Bay observed outside the spawning period was only 40 kg km-2 (Fig. 2 a). Overall, this analysis indicates it is unlikely that the capelin stock is currently non-migratory and has remained inshore since 1991.

There has been a bias for increased catches of capelin in the FBTS post-1995 not only due to a change in sampling gear but also due to the increased proportion of capelin biomass in the trawl zone (bottom 4 m of the water column) post-1991 (Mowbray 2002), likely in response to a decline in the risk of Atlantic cod predation that may drive capelin into the pelagic zone (Rose 1993, McQuinn 2009). Furthermore, when capelin densities are low, capelin are found in closer association with the bottom and diel vertical migration is less pronounced compared to when capelin densities are high (Mowbray 2002). Due to the inherent biases in the FBTS data, we also considered other data sources to investigate the center of distribution of capelin post-1991 (sensu Jech & McQuinn 2016). Juvenile capelin surveys using an International Young Gadoid Pelagic Trawl (IYGPT) in the northeastern bays and the offshore from 1994-99 found centers of distribution of capelin juveniles on the northern Grand Banks and along the northeast coast, but not in the bays, of Newfoundland (Anderson et al. 2002) which is consistent with capelin migrating to their nursery areas in the offshore. This independent study of juvenile capelin distribution supports our center of gravity analysis using the FBTS data.

In summary, capelin distribution moved in the north-south rather than east-west axis post-1995 based on a center of gravity of analysis that accounts for both the inter-annual changes in the spatial distribution of sampling effort and the uncertainty around the center of gravity estimates. While this center of gravity analysis supports the capelin collapse hypothesis, there remain inherent biases in using the bottom-trawl data to quantify pelagic species demographics. However, the impossibility of the ‘missing’ 3-6 Mt of capelin (up to 170,000 kg km-2, uniformly distributed throughout the un-surveyed area) remaining undetected in the inshore strata post-1991 when there are hundreds of inshore harvesters on the water with echo sounders provides support for the capelin collapse hypothesis.

### Residence time of capelin concentrations during offshore acoustic surveys in Div. 3L

The capelin non-collapse hypothesis purports that the delay in the timing of capelin spawning post-1991 led to a mismatch in capelin availability to the spring acoustic survey, which has resulted in the spring acoustic survey severely underestimating capelin biomass offshore (Frank et al. 2016). The Canadian regional stock assessment process supports the capelin collapse hypothesis by concluding that the stock experienced an order of magnitude decline in the early 1990s and has remained at historic lows for the past 25 years based on the offshore fall and spring acoustic surveys (DFO 1994, Miller 1994, 1997, DFO 2008, 2010, 2013, 2015, 2018).

Since 1991, capelin spawning has been persistently delayed on average by four weeks (DFO 2018). Meanwhile, the spring acoustic survey has been fixed spatially and temporally since the 1980s, implying changes in how the spring acoustic survey perceives migrating capelin post-1991. However, it is important to note that the spring acoustic survey, which encompasses a capelin nursery area, was primarily designed to survey the non-migratory, immature portion of the stock, rather than the spawning migration (Mowbray 2014). All age classes acoustically surveyed are included in the annual index of capelin abundance, but the spring acoustic survey does not target capelin spawning migrations, and, therefore, cannot be used to derive a proxy for spawning stock biomass.

While the delay in spawning time should not have a direct effect on the acoustic abundance index derived from the spring acoustic survey, the earlier observed maturation of capelin post-1991 may have an effect on the capelin abundance index produced by the spring acoustic survey. The age-2 portion of the stock is the main component being surveyed and the proportion of maturing age-2 capelin has increased since 1991 (4% pre-1991 compared to 37-79% post-1991) (Mowbray 2014; DFO 2018). Earlier maturation could alter the internal structure of the stock with inter-annual variability in the proportion of age-2s starting their adult migration patterns and would, therefore, be unavailable to the survey. However, there has been high internal consistency in the spring acoustic survey, with the index of abundance for the age-3 cohort being strongly correlated with the index of abundance for the age-2 cohort from the previous year (DFO 2018), which implies that the spring acoustic survey can meaningfully capture relative changes in the overall stock, regardless of migrating/non-migrating fractions.

If there was a significant inshore capelin population post-1991, we would expect to see this change in population demographics reflected in both the offshore acoustic survey and the inshore commercial catch. For the inshore commercial fishery pre-1991, mature age-2 capelin was a negligible component of the fishery (< 5% of total catch) (DFO 2018). Post-1991, the contribution of mature age-2 capelin increased to almost half of commercial inshore catches (DFO 2018), which supports the non-collapse hypothesis of non-migratory fish maturing early (Frank et al. 2016). However, the population demographics from catches offshore refute the non-migratory hypothesis as the proportion of age-2 and age-3 fish sampled offshore has remained consistent pre- and post-1991 at ~60% and ~28% of the catch, respectively (DFO 2018). Furthermore, the increased proportion of mature age-2 fish in the inshore post-1991catch can be explained by the increase in proportion of age-2 fish maturing offshore and undergoing a spawning migration, rather than an absence of migration per se.

In summary, neither the persistent delay in spawning time nor earlier maturation age of capelin post-1991 seem to significantly affect the ability of the spring acoustic survey to provide a relative index of capelin abundance. Data from the spring acoustic survey supports the capelin collapse hypothesis.

### Independent indices of inshore capelin abundance

The non-collapse hypothesis postulated that strong correlations among independent inshore indices of capelin abundance post-1991 in addition to observed minimal changes in these indices between the late 1980s and the early 1990s was supportive of a stable stock (Carscadden and Nakashima 1997, Frank et al. 2016). In contrast, the offshore fall and spring acoustic surveys indicated a stock collapse (Miller 1994, 1997, DFO 2008, 2010, 2013, 2015, 2018, Mowbray 2014).

Two inshore indices collected by DFO during the 1980s and 1990s were an aerial abundance index and inshore commercial catch rates. The aerial survey was designed to estimate capelin spawning stock biomass based on the area of capelin schools near spawning beaches in Div. 3L (Nakashima 1997). The aerial survey commenced in 1982, and initially followed four defined survey tracks in Conception and Trinity Bays during a fixed period of mid-June to early July (Carscadden et al. 1994). Protracted spawning post-1991 violated a key assumption of the aerial surveys: all schools must arrive at the same time in each bay to form a single spawning peak. Protracted spawning from early July to mid-August in 1991-93 resulted in multimodal capelin spawning peaks that were covered with variable success by the aerial survey (Nakashima 1996). For example, in 1993, the peak spawning period was adequately surveyed in Conception Bay, but two spawning peaks in Trinity Bay, based on the egg deposition index, were missed (Nakashima 1996). In 1996, aerial coverage was at its lowest since 1991 due to poor weather and technical problems (Nakashima 1997). In 1997, the geographical coverage of the aerial survey was reduced to two transects in the inner areas of Trinity and Conception Bays (Anon 1998). While the estimated aerial abundance index in 1997 was fourth highest in the series, there were concerns that the limited geographical coverage of the aerial survey did not accurately reflect the status of the stock (Anon 1998). Six of the eight years of aerial data post-1991 did not adequately cover peak spawning times (1991-93), had poor weather and technical difficulties (1996), and had reduced geographical coverage (1997-98). The aerial survey was discontinued in 1999.

The inshore commercial catch rate data have not been included in the capelin stock assessment process since 1993 due to changes in management regulations post-1991. Due to the small sizes of spawning capelin post-1991, management regulations introduced a size criterion of 50 count/kg to reduce dumping of undersized capelin (Carscadden & Nakashima 1997). This size criterion effectively closed the fishery in 1994 and 1995. From 1996, the size criterion was removed but management regulations to reduce discarding of small, unmarketable capelin resulted in fishing effort being concentrated to a few days when capelin were highly available (Anon 1998). Post-1991, the inter-annual variability in participation in the fishery due to fish quality, a massive decline in fishing effort compared to the 1980s (mean effort from 1991 through 1997 declines substantially compared to mean effort from 1980 to 1990 – less than ¼ of the effort in the trap fishery and less than 1/6 of the effort in the purse seine fishery; this was in part due to the introduction of the size criterion to reduce capelin discards, mean annual catches from 1991 through 1997 in the purse seine and trap fisheries falling to between a quarter and a half of their mean annual catches from 1980 through 1990, the mean number of fishers participating in the fisheries dropping by more than half and market forces resulted in overall fishing effort being reduced and concentrated it to only times when capelin were highly available resulting in high catch rates and rendered the inshore catch rate index useless as an indicator of stock abundance (Anon 1998).

In summary, due to changes in capelin biology and management measures post-1991, neither of the inshore indices provide reliable data on spawning stock biomass. These indices cannot be used to support or refute either hypothesis.

### Inshore recruitment index

Capelin larval production in Trinity Bay did not decrease appreciably post-1991, which supports the non-collapse hypothesis (Frank et al. 2016). DFO collects two larval indices in Trinity Bay: an emergent larval index (3-10 mm SL) in a nearshore area and a late-larval index (10-30 mm SL) (see Nakashima & Mowbray 2014 for more details). Given the persistently late capelin spawning since 1991, it is fair to compare late-larval densities in August in years pre-1991 and September in years post-1991 (Nakashima & Mowbray 2014). Late-larval densities during the 2000s were consistently lower and more variable than during the 1980s: average late-larval density in Trinity Bay in September 2002-15 was 30.9 m-2 (SD: 26.9, range 6.73-96.95 m-2) while in August 1982-86 was 48.8 m-2 (SD: 15.1, range 33.2-73.6 m-2). In 12 of the 14 years in the 2000s, average late-larval densities in September were below the average August larval densities in the 1980s.

If the bulk of capelin biomass was residing inshore year-round and was not available to the annual offshore spring acoustic survey, then we would expect an absence of relationship between larval production and the age-2 abundance index from the spring acoustic survey. However, post-1991, the emergent larval index was related to the age-2 abundance index from the spring acoustic survey, which was lagged by 2 years in order to compare survival in the same cohort (Murphy et al. 2018). This suggests that the offshore acoustic survey tracks inshore larval productivity, which supports the capelin collapse hypothesis.

In summary, the positive, significant relationship between the emergent larval index and the offshore age-2 abundance index post-1991 supports previous research that identified early larval survival as an important driver of capelin recruitment (Frank & Leggett 1982, Leggett et al. 1984, Dalley et al. 2002). This significant relationship between two fishery-independent inshore and offshore indices post-1991 provides support for the capelin collapse hypothesis.

### Ecosystem response

### Temporal dynamics of cod weight-at-age and condition

From the early 1980s to the late 1990s, Atlantic (also known as Northern) cod’s condition and weight-at-age declined markedly in the northern portion of its range (Div. 2J) with similar declining trends in Div. 3K, while condition indices remained relatively unchanged in the southern portion of Atlantic cod’s range (Div. 3L). Furthermore, weight-at-age of 4- and 5-year old Atlantic cod saw a small increase in the mid-1990s in Div. 3L that later subsided. This ephemeral increase in weight-at-age in Div. 3L during the mid-1990s was used as support for the non-collapse hypothesis (Frank et al. 2016).

The existence of spatial structure in condition traits of Atlantic cod has been documented extensively in the scientific literature (e.g., Lilly et al. 2005, Neville et al. 2018, Rose & Rowe 2018), with gradients from north to south in growth (length-at-age) and condition indices (liver, gutted and total body mass) (Buren et al. 2014b, Morgan et al. 2017). Atlantic cod weight-at-age and liver condition indices have never been spatially homogenous due to the species complex inhabiting distinct ecosystem production units in Div. 2J3KLNO (e.g., Lilly 2005, Koen-Alonso et al. 2013, Morgan et al. 2017). These units are characterized by distinct marine communities and food web systems (Pepin et al. 2010, 2012, 2014, Koen-Alonso et al. 2013, NAFO 2014). Therefore, the non-homogenous traits of Atlantic cod from Labrador (Div. 2J) to the southern Grand Banks (Div. 3NO) are typical of this stock complex and cannot be used to support the non-collapse hypothesis.

Changes in spatial overlap between the collapsed stocks of Atlantic cod and capelin post-1991 may have exacerbated the spatial differences in Atlantic cod condition indices (Rose et al. 2000). Historically, Atlantic cod was a dominant fish predator on the NL Shelf, with capelin being its primary prey (Winters & Carscadden 1978, Lilly 1987, Lilly 1991). During and post-1991, capelin shifted its fall distribution from having two distinct aggregations, one in the northwest (Div. 2J3K) and one in the southeast (Div. 3L, at the northern slope of the Grand Banks) to having only one in the southeast (Lilly & Davis 1993, Miller 1994), with hypothesized excursions of capelin onto the Flemish Cap and the Scotian Shelf (Frank et al. 1996); more recent genetic work using microsatellite markers fails to support the hypothesized excursion of capelin to the Scotian Shelf (Kenchington et al. 2015). Coincidently, Atlantic cod moved southward on the northeast Newfoundland Shelf in the late 1980s and early 1990s and aggregated within a small area on the northern Grand Banks and in the Bonavista Corridor (Rose 1993, Rose et al. 2000). A hypothesis proposed to explain the observed shift in Atlantic cod’s distribution is that they moved in response to the southerly distribution of capelin post-1991 (Rose et al. 2000). Atlantic cod’s weight-at-age and liver condition worsened in northerly areas where there was no spatial overlap between Atlantic cod and capelin, and remained relatively stable in southerly areas, where the collapsed Atlantic cod stock overlapped with capelin.

In summary, the spatial structure of Atlantic cod condition indices is explained by the distinct ecosystem production units this stock complex inhabits. Since 1991, Atlantic cod condition worsened in northerly areas where there was no spatial overlap with capelin, and remained relatively stable in southerly areas where the collapsed Atlantic cod stock overlapped with the capelin stock. The observed change in Atlantic cod distribution and condition indices post-1991 supports the capelin collapse hypothesis.

**Harp seal population trends and diet**

A large number of starving harp seals was observed following the collapse of capelin in the Barents Sea in the mid-1980s (Haug & Nilssen 1995). The absence of an obvious response in Northwest Atlantic harp seals (*Pagophilus groenlandicus*) to the collapse of the capelin stock in 1991 was considered support for the non-collapse hypothesis (Frank et al. 2016). However, there are significant differences between the two regions. In the Barents Sea, the collapse of capelin during the mid-1980s occurred when the stocks of other important forage fish, including Atlantic herring and Arctic cod (*Boreogadus saida*), were severely depleted (Hamre 1994, Hop & Gjøsæter 2013). Therefore, the effects of the capelin collapse were amplified and reached several taxa including seals, seabirds and Atlantic cod (Hamre 1994). Barents Sea capelin declined in 1992-93 without a similar ‘invasion’ of starving seals. This was likely due to the availability of alternative prey (Atlantic herring and Arctic cod, Nilssen et al. 1998). In the Northwest Atlantic, alternate prey was still available during the early 1990s (Lilly et al. 1994, Bourne et al. 2015). This was reflected in Arctic cod and Atlantic herring being the dominant prey items in harp seal diet in the early 1990s (Stenson 2012).

While Northwest Atlantic harp seals did not show catastrophic mortalities post-1991, they have been impacted by the decline in capelin. Until the late 1970s, pregnancy rates were consistently around 85%. Since then, pregnancy rates have been highly variable (ranging from ~20% to 75%) with an overall declining trend (Stenson et al. 2014, 2016). In addition, late-term abortions have become a regular occurrence since the late 1980s (Stenson et al. 2016). Stenson et al. (2016) found that while the general decline in harp seal fecundity reflected density-dependent processes associated with increased population size, including the late-term abortion rates in their model allowed them to explain the large inter-annual variability in pregnancy rates. Changes in the abortion rates, in turn, were found to be influenced by ice cover in late January and capelin biomass. Buren et al. (2014a) found that capelin abundance is correlated with ice conditions, suggesting that late January ice conditions reflect changes in environmental conditions that influence many prey species. While higher catches in the Canadian commercial seal hunt from 1996-2008 contributed to reductions in the rate of harp seal population growth, lower pregnancy rates also had a major impact on the dynamics of this population. This is evident in the past decade where the commercial catches have declined but there has not been a concomitant increase in harp seal population abundance.

In summary, we cannot conclude that the absence of starving seals post-1991 indicates that capelin biomass remained stable and did not collapse. However, a declining trend in pregnancy rates and an increase in late-term abortions, which was related to capelin abundance, suggest capelin abundance has been a limiting factor in harp seals fecundity during the past three decades.

### Seabird population trends and diets

Populations of common murres (*Uria aalge*), Atlantic puffins (*Fratercula arctica*) and northern gannets (*Morus bassanus*) off eastern Newfoundland have increased during the last three decades (Chardine et al. 2003). Given that capelin is an important prey item for these predators, particularly during the breeding season, increases in their populations are inconsistent with a collapse in the capelin stock (Frank et al. 2016). However, the population increase of common murres post-1991 was associated with major reductions in adult mortality due to the coincident closure of the Atlantic cod fishery. The removal of thousands of gillnets from inshore areas during the 1990s and 2000s resulted in a significant reduction in bycatch mortality (Regular et al. 2013). Reductions in common murre adult mortality associated with ship-sourced oil pollution and hunting also decreased during this same period (Wilhelm et al. 2009). The cumulative effects of these reductions in common murre adult mortality would have overweighed negative population effects associated with bottom-up prey base reductions. Increases in populations of Atlantic puffins and northern gannets were associated with the same cumulative effects.

Throughout the 1990s, common murres on Funk Island almost exclusively fed capelin to their chicks during the breeding season. This was proposed as support for the capelin non-collapse hypothesis (Frank et al. 2016). However, consistently high abundances of capelin at annually persistent spawning sites within seabird foraging ranges allowed for the high percentage of capelin in parental deliveries (Davoren et al. 2012, Davoren 2013). Fish stocks in general and pelagic stocks in particular contract their geographic range during periods of rapid population decline (Winters & Wheeler 1985, Burgess et al. 2017). This pattern has been described for several finfish and shellfish populations (Prince et al. 2008, Wilberg et al. 2009), including Atlantic cod (Rose & Kulka 1999) and Northwest Atlantic herring stocks (Winters & Wheeler 1985). Capelin’s center of distribution moved southward during the early 1990s, i.e. closer to the vicinity of seabird colonies in Newfoundland’s northeast coast. Therefore, a higher proportion of capelin in common murre’s diet is not inconsistent with the collapse hypothesis, whereby the range of the collapsed capelin stock overlapped with the foraging range of breeding colonies of common murres.

Capelin represented a large proportion of northern gannet’s diet from 1990-2004 (20 – 100 %) in contrast to pre-1990 (<12%, Montevecchi 2007). This change in northern gannet diet was proposed as support for the non-collapse hypothesis (Frank et al. 2016). However, this change in diet was due to the cold water intrusion that occurred during the 1990s which precluded northern gannet’s preferred large pelagic warm-water prey (mackerel *Scomber scombrus*, Atlantic saury *Scomberesox saurus* and short-finned squid *Illex illecebrosus*) from moving into the region (Montevecchi & Myers 1997, Montevecchi 2007). In addition, capelin was a minor prey item in seabird diets during the 1990s in Labrador (Bryant & Jones 1999, Baillie & Jones 2004). This is consistent with the hypothesis of a collapsed capelin stock, hyper-aggregated in the southern portion of its range.

In summary, the trends in seabird abundance do not provide support for either the collapse or non-collapse hypothesis as other variables, such as removal of gill nets in the inshore area, had a larger impact on seabird survival. Seabird dietary information does not provide support for either hypothesis, but is consistent with a collapsed, hyper-aggregated capelin stock.

### Zooplankton response: *Calanus finmarchicus* abundance

Given the magnitude of the capelin collapse, a significant increase in their main copepod prey, *Calanus finmarchicus*, may have been expected(Frank et al. 2016). To test this hypothesis, the continuous plankton recorder (CPR) data were used to estimate *C. finmarchicus* densities pre- and post-1991 in the NL region (Frank et al. 2016). No significant difference in *C. finmarchicus* densities pre- and post-1991 was found. However, the usefulness of CPR data for the Northwest Atlantic has been questioned. Head and Pepin (2010) noted that only two years between 1960-1978 had more than 8 months of observations over the Grand Banks sections of the CPR sampling, which was the section used to test the collapse hypothesis (Frank et al. 2016), and inconsistencies in the course of the survey tracks from ships-of-opportunity resulted in uneven sampling of different water masses (Pepin et al. 2011). Furthermore, there was a substantial reduction in CPR mileage towed in the 1980s with a contraction of monitoring to 20° W in the eastern Atlantic (Reid et al. 2003). During this decade, monitoring science fell out of favor (Reid et al. 2003). Other researchers have looked at the CPR data as a potential data source for historical productivity in the NL region, but the large CPR data gap in the region from 1979-1990 precluded its use in a recent capelin study (Mullowney et al. 2016).

Since 1999, DFO has run the Atlantic Zonal Monitoring Program (AZMP) in three regions in eastern Canada: Scotian Shelf, Gulf of St. Lawrence, and NL. The AZMP program collects physical, chemical and biological data at stations along 14 oceanographic transects during offshore spring and fall cruises. While the AZMP data cannot be used to directly support or reject the hypothesis of a capelin stock collapse due to its commencement in the late 1990s, the AZMP data has been used to support the hypothesis that bottom-up processes are driving capelin survival in the NL region (Buren et al. 2014a, Obradovich et al. 2014, Murphy et al. 2018).

**Physical variability**

The non-collapse hypothesis used an analysis of ocean climate to show a change in environmental conditions ~ 5 years after the proposed capelin collapse and concluded that the physical evidence was weak for an environmental driver of capelin collapse (Frank et al. 2016). The authors contend that the transition from cold to warm conditions in 1996 is the distinguishing signal of their PC1 analysis (Frank et al. 2016). However, elsewhere in the literature, 1991, not 1996, has been identified as climatologically important due to its strong cold anomaly (e.g., Drinkwater 1996, Colbourne et al. 2014, 2015, 2016) and biologically important due to the dramatic regime shift in the North Atlantic ecosystem in 1991 with the collapse of Atlantic cod, capelin and other finfish species and an increase in shellfish biomass (Gomes et al. 1995, Lilly et al. 2000, Rice 2002, Koen-Alonso et al. 2010, Hammill et al. 2011, Buren et al. 2014a, Pedersen et al. 2017); seabird dietary shifts from warm- to cold-water pelagic prey (Montevecchi & Myers 1992, Montevecchi & Myers 1997, Montevecchi 2007); and shifts in groundfish diet (Dawe et al. 2012).

In summary, based on the extensive published literature on the regime shift in the Northwest Atlantic (e.g., Drinkwater 1996, Buren et al. 2014a, Pedersen et al. 2017), the weight of evidence approach suggests that the cold-water anomaly of the early 1990s was the physical driver of capelin collapse.

## Discussion

Numerous sources of primary and government literature concluded that the NL capelin stock suffered an order of magnitude decline in the early 1990s (DFO 1994, Miller 1994, 1997, Rose & O'Driscoll 2002, Davoren & Montevecchi 2003, Rose 2007, DFO 2008, 2010, 2013, Buren et al. 2014a, Mullowney & Rose 2014, DFO 2015, Murphy et al. 2018). In contrast, Frank et al. (2016) postulated that the capelin stock did not suffer a collapse but rather experienced a dramatic change in phenology post-1991 and became non-migratory. We used the weight of evidence approach to evaluate the empirical support for the hypothesis of a capelin stock collapse using multiple, independent lines of enquiry with diverse statistical methods (e.g., triangulation, sensu Munafò & Davey Smith 2018). The weight of evidence approach led us to conclude that the Div. 2J3KL capelin stock suffered a bottom-up, climate-driven population collapse in 1991 with minimal recovery in the subsequent 25 years.

There are two alternative explanations for the non-collapse hypothesis: (1) a spatio-temporal mismatch between the spring acoustic survey and capelin phenology; and (2) a change in biology of capelin from a highly migratory stock to one that resides inshore year-round. The first hypothesis was rejected by both Frank et al. (2016) and our analyses. We tested the second non-collapse hypothesis by using multiple independent datasets, which included both fishery-dependent (inshore commercial catch) and fishery-independent (spring and fall offshore acoustic surveys, an inshore acoustic survey, FBTS, larval indices, predator diet, predator behavior), and diverse statistical methods.

Using a Ricker-logistic population model, we found strong coherence among the five acoustic surveys with all the surveys indicating a collapse in the capelin population in the early 1990s. Furthermore, the Trinity Bay inshore acoustic survey provided little information on the population dynamics of capelin compared to the four offshore acoustic surveys, which is supported by the Trinity Bay inshore acoustic surveys in May providing annual estimates of the inshore capelin population as ~10% of the capelin biomass offshore in most years. The center of gravity approach using the FBTS data (Thorson et al. 2016) found no evidence of inter-annual longitudinal movements of capelin post-1995, but rather that the stock’s center of gravity moved latitudinally. However, only a limited number of inshore strata are surveyed by the offshore acoustic surveys and FBTS. If we considered all of the inshore strata not surveyed by the FBTS, there would need to be a minimum of 41,000 kg km-2 of capelin uniformly distributed in these inshore strata to compensate for the ‘missing’ 3-6 Mt capelin from the offshore. Seasonal inshore acoustic surveys in Trinity Bay found a maximum of 10,000 kg km-2 in June, and the inshore capelin densities were a fraction of this outside of the peak spawning period.

The lack of adult capelin in the inshore area outside of the spawning period was also corroborated with predator diet and behavior data. Consumption of capelin by Atlantic cod in the inshore (1996-2003) was highly prevalent in June compared to January (Sherwood et al. 2007); common murres exhibited a temporal shift towards later breeding in the late 1990s, which corresponded with the later inshore arrivals of capelin in the common murres’ foraging range (Davoren & Montevecchi 2003); dietary shifts in four seabird species (great shearwater *Ardent* *gravis*, sooty shearwater *Ardenna grisea*, herring gull *Larus argentatus*, great black-backed gull *Larus marinus*) and humpback whale (*Megaptera novaeangliae*) throughout the summer was associated with dramatic shifts in inshore capelin abundance associated with the spawning migration (Gulka et al. 2017). Furthermore, it is highly unlikely that 3-6 Mt of capelin inshore would have been missed by both DFO and harvesters since 1991, given the presence of hundreds of active inshore fishing vessels equipped with echo sounders over much of the northeast coast of Newfoundland.

Using the weight of evidence approach, the majority of the independent data sources examined support the hypothesis of a collapsed capelin stock. We found strong internal and external consistency in trends across multiple sources of data. For instance, the spring acoustic survey shows strong cohort tracking within the survey and between recruits enumerated by an independent larval survey, and there is agreement across independent acoustic surveys conducted in the offshore. All acoustic surveys indicate that the stock collapsed in the early 1990s and subsequent surveys and data, both fisheries dependent and independent, have failed to prove the existence of millions of tonnes of non-migratory capelin along the coast of NL.

## Figure captions

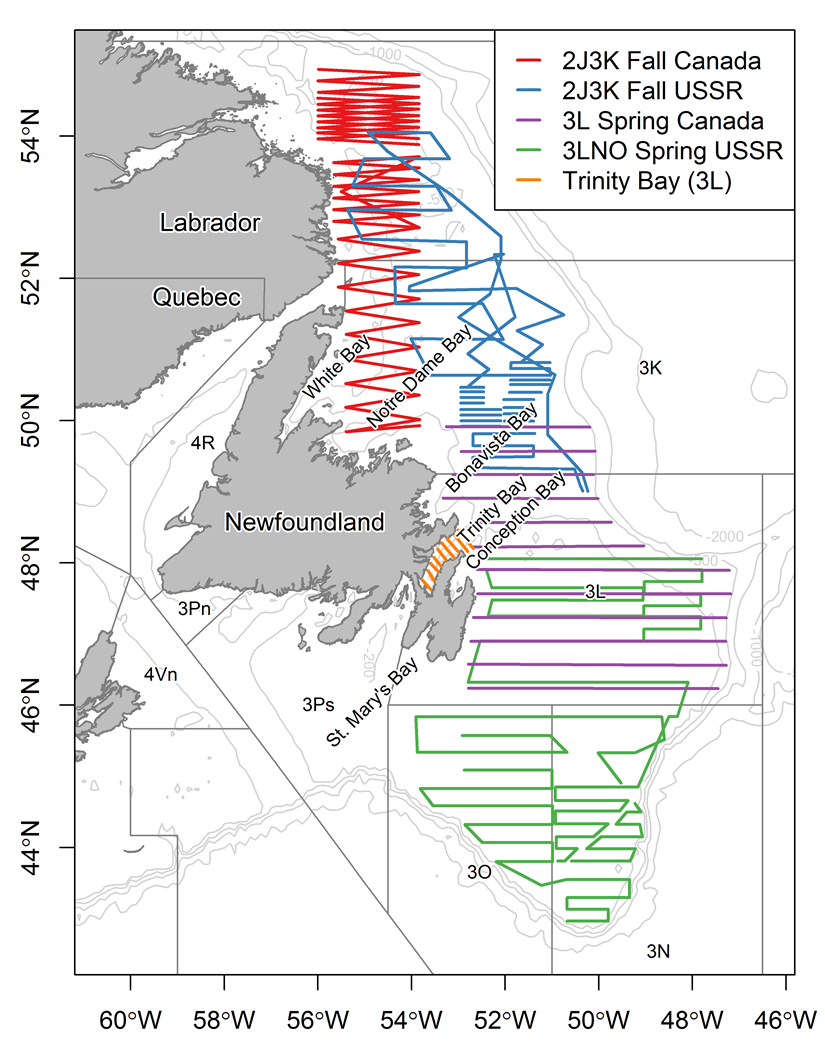
Fig. 1. Capelin stock area in NAFO Divisions 2J3KL including the embayments of Newfoundland, Canada. Included are the acoustic survey tracks conducted in May 2018 by Canada in Div. 3L (offshore) and Trinity Bay (inshore) (DFO, unpublished data); June 1991 by the former USSR in Div. 3LNO (see Bakanev 1992 for more details); October 1983 by Canada in Div. 2J3K (see Miller and Carscadden 1983 for more details); and November 1991 by the former USSR in Div. 2J3K (see Bakanev 1992 for more details).

Fig. 2. (a) Capelin biomass estimated from the seasonal inshore acoustic survey in Trinity Bay, NL, Canada. The May values are for the main portion of Trinity Bay only, while the other months surveyed the entire bay, including the arms and headland (note the log scale); and (b) capelin maturity stage composition (n = 5319) and (c) capelin age composition (n=864) sampled in the seasonal inshore acoustic surveys in Trinity Bay in 2003-05. S/R is spent/recovering, Mat. is maturing, and Imm. is immature.

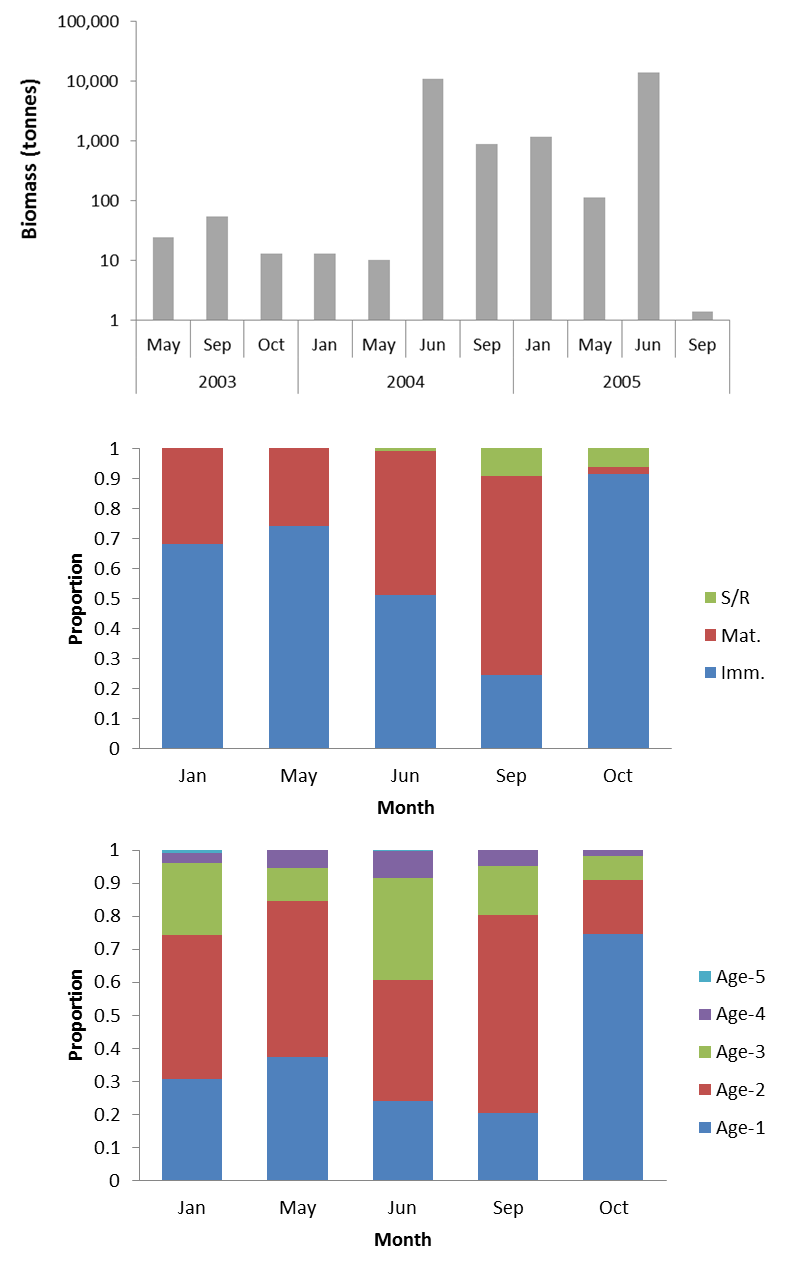
Fig. 3. (a) Trend in capelin biomass estimated using a state-space Ricker-logistic model fit to (b) trends in five acoustic survey indices: Div. 2J3K Fall Canada (1982-92), Div. 2J3K Fall USSR (1982-92), Div. 3L Spring Canada (1982-92, 1996, 1999-2005, 2007-15, 2017), Div. 3LNO Spring USSR (1982-94), and Trinity Bay (inshore Div. 3L; 1999-2005, 2007-13, 2017). Shaded area in (a) represents 95% confidence intervals of the biomass estimates, the lines in (b) indicate model fits to each survey index and the vertical lines in (b) indicate 95% confidence intervals of the index. Note the log scale.

Fig. 4. Center of gravity analysis using the VAST package in R (Thorson et al. 2016, Thorson & Barnett 2017) using data from the fall bottom-trawl survey in NAFO Divisions 2J3KL (Newfoundland and Labrador, Canada; 1995-2017) to fit a geostatistical delta-generalized linear mixed model to estimate the spatial and temporal distribution of capelin. Annual center of gravity estimates are connected by lines through time, and approximate 95% confidence intervals around these estimates are indicated by the dotted black line. The red area indicates areas not covered by the survey and the light pink (cream) area indicates inshore strata that are poorly covered by the fall bottom-trawl survey.

**Fig. 1**

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**Fig. 2**

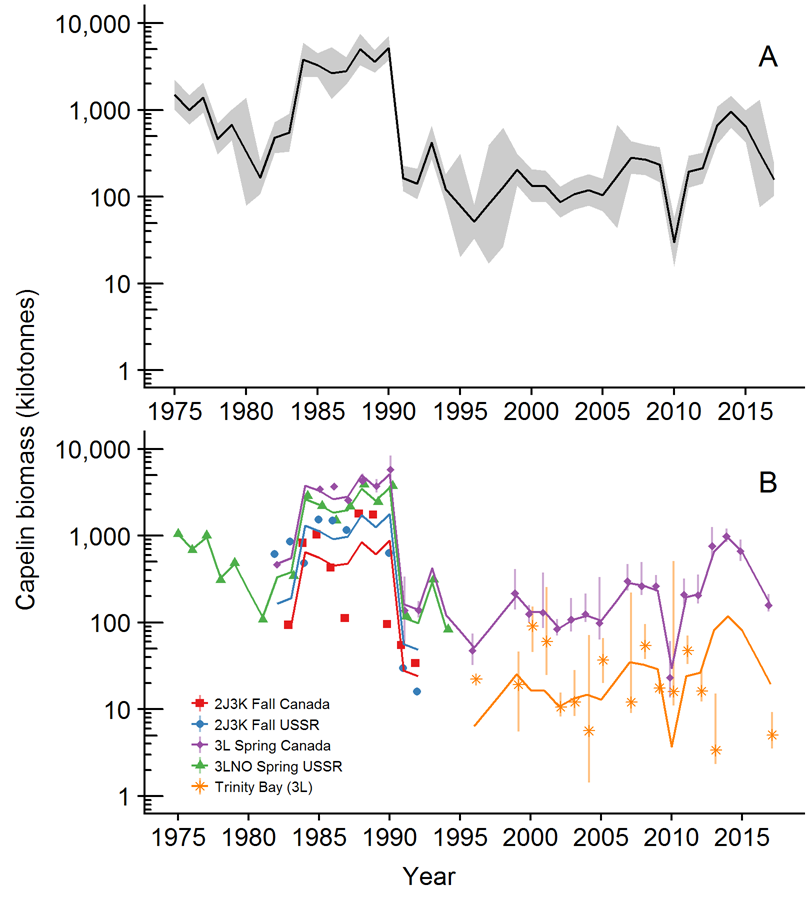
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a

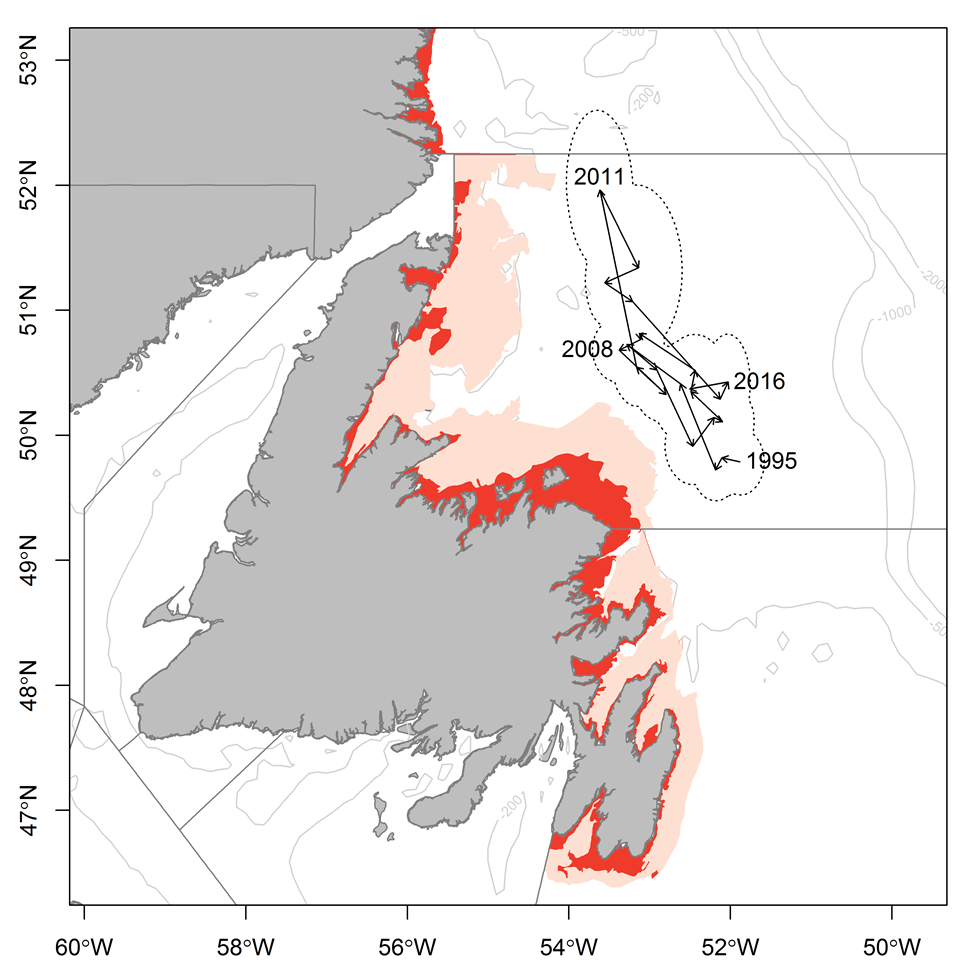
b

c

**Fig. 3**

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**Fig. 4**

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## Supplementary material

### Trinity Bay seasonal inshore acoustic surveys (2003-2005)

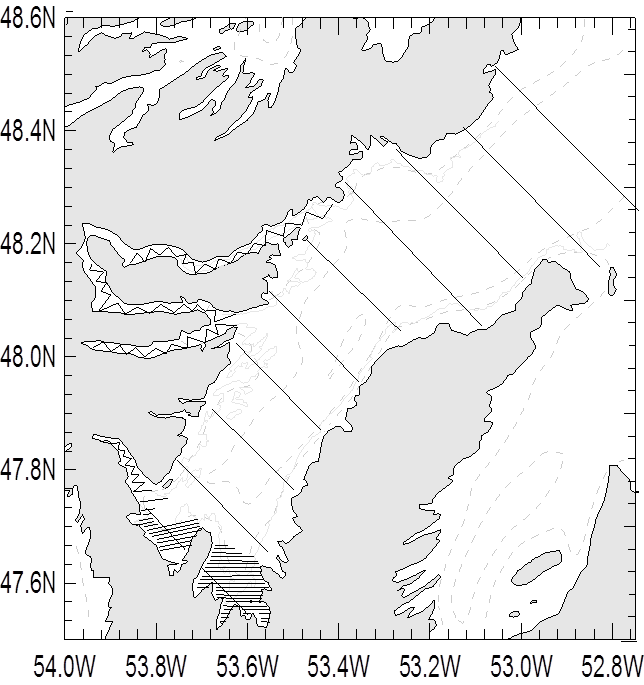
Inshore seasonal acoustic surveys (January, June, September 2003-05) in Trinity Bay, Canada were conducted from a 23 m inshore research vessel (CCGS Shamook) using a calibrated EK500 echo-sounder with a towed 38 kHz transducer. Surveys followed a fixed transect design and covered both the main portion of Trinity Bay as well as the four arms (Suppl. Fig 1). When acoustic targets were encountered, sampling was conducted using bottom and midwater trawls to target the portion of the water column where the acoustic signal occurred. Samples were obtained from most aggregations. Length, sex and maturity stage were recorded for all fish sampled and ages determined for two fish per sex per 0.5 cm interval.

Spatial patterns in age composition were similar to those patterns reported by Winters (1970) with older, larger capelin overwintering in the main portion of the bay while juvenile capelin were more prevalent in the inner arms. In all months except June, capelin were aggregated along the sides of the trench around 200 m depth, whereas in June capelin were present in the arms and in shallower water closer to shore at the bottom of the bay.

**Supplementary Figures**

Fig. S1 Inshore seasonal survey in Trinity Bay, Newfoundland, Canada (2003-2005) of acoustic transects (solid lines) and 100, 200 and 500 depth contours (dashed lines).

Fig. S1



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