

# Recruitment rate estimation in waterfowl using a non-invasive method

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## Abstract

The response of a waterfowl population to a harvest pressure depends on its capacity to renew. The recruitment, i.e. the number of young adults reproducing for the first time, is a key indicator to describe the renewal of a population, and therefore an essential tool in population management. The recruitment rate, defined inhere as the number of recruits produced per breeder, is even more informative because it is independent of the breeding population size and allows comparisons over time and between species. The proportion of young adults in a waterfowl population, which is a main step in estimating recruitment rate, is often estimated from game-hunting samples. However, this proportion is not accessible in years without harvest, or with a low harvest rate. Moreover, the age-structure in the harvest samples, commonly called hunting bag, does not necessarily reflect the underlying age-structure of the population. It is often skewed towards juveniles and can lead to an overestimation of the recruitment rate. In waterfowl, adult males usually display brighter colours than females and immatures from both sexes. This dichromatism can be characterized and monitored from count surveys, a non-invasive method. Such information can be used to infer the proportion of young adults, and consequently the recruitment rate. In using two populations of ruddy duck, *Oxyura oxyura*, this study develops a Bayesian method to estimate the recruitment rate from count data. The results are compared to recruitment rate estimates from samples and are discussed by testing their consistency to support the observed population growth rates. The results suggest that the counting method is a versatile tool to estimate the recruitment rate if the species is monitored during the appropriate time window. For the relevant waterfowl species, these results argue for considering a two-category protocol in winter count surveys.

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# 1 Introduction

## *Definitions*

One defines the recruitment rate of a breeding population as the average number of recruits produced per each breeder (Flint, 2015). This parameter might sometimes be called “annual fertility” (Koons et al., 2014), “reproductive rate” (Cooch et al., 2014), “reproductive success” (Etterson et al., 2011), or “productivity” (Johnson et al., 1987). If one considers a species reaching its sexual maturity as early as its first year, all individuals are mature at the reproduction season. The recruitment rate is thus the ratio of the number of young adults at the reproduction season of year  $t$ , i.e. the recruits, to the number of all individuals at the reproduction season of the previous year,  $t - 1$ . The declination for species with age at first reproduction higher than 1 follows the same process (see Robertson, 2008).

The recruitment rate parameter is complex because it is composed of two main sub-parameters : the fecundity followed by the juvenile survival (Etterson et al., 2011). Both sub-parameters can be split into other sub-parameters. The fecundity can be defined by the product of the nesting success, the average clutch size, and the hatching success. The juvenile survival can be defined as the product of the post-hatching survival and the post-fledging survival. The recruitment rate as defined in the present study is thus a parameter integrating multiple life steps from the egg to the young adult attempting to breed for its first time (Koons et al., 2017).

## *The use of recruitment rate in population management*

Population conservation, harvesting, and control require tools in population management to reach targets such as a population size for example (Shea & NCEAS Working Group on Population Management, 1998). To do so, a first “blind” strategy is to test iteratively management rules to reach progressively an optimum. Such approach can be time-consuming given the intrinsic variability of the population response to its changing environment. It might also suffer from a rejection of the stakeholders because of the lack of transparency of an arbitrary process (Williams & Madsen, 2013). Another approach is to try to predict the growth rate of the population as a function of the harvest rate and other environment inputs. Such modelling approach can provide clues on the recovering time of a population upon a variety of management rules (Otis, 2006). These decision-support tools are then key inputs to status on the best management rules to implement. The growth rate of a population depends on the adult survival and the recruitment rate, see Equation (8) in Section 2. The adult survival is generally estimated by mark recapture/recovery methods (Lebreton, 2001). If the harvest pressure behaves as an additive mortality to the natural mortality for adults (e.g. Iverson et al., 2013), it is quite straightforward to predict its impact of the adult survival. According to Blums & Clark (2004), the recruitment for diving ducks is poorly related to fecundity, but depends mostly on juvenile survival, which is driven by weather conditions. Studies on other birds concludes also on the weak correlation between fecundity and recruitment (e.g. Murray, 2000). The difficulty to track juvenile survival (Schmidt et al., 2008) demonstrates the interest to focus on the recruitment rate to predict population growth rate over long time periods. The key to predict the growth rate of a population is then to estimate the average value and the variability of the recruitment rate, potentially influenced by the harvest pressure (Anders & Marshall, 2005).

## *The variability of the recruitment rate*

A waterfowl species that is released in a favourable habitat will naturally expand (Malthus, 1872). The growth of such a population reflects that the recruitment rate parameter is higher than the adult mortality rate. This expansion is limited by the carrying capacity of the habitat (Sayre, 2008). When a waterfowl species is endemic of an ecosystem and evolves in stable environmental conditions, the population size varies around the carrying capacity because the individuals compete for space or/and food (Nummi et al., 2015), which prevents the population from expanding. This competition commonly induces a lower fecundity or/and a higher juvenile mortality. The population renewal is slowed because the recruitment rate parameter is affected by density-dependent effects (Gunnarsson et al., 2013). For an endemic species evolving in stable environmental conditions, a balance between the recruitment rate and the adult mortality rate explains the stabilization of the number of individuals (Flint, 2015).

Hunting waterfowl induces adult and juvenile mortality (Bellrose, 1980). Intuitively, one expects that such a hunted population would observe lower adult survival and recruitment rate and then would rapidly deplete to its extinction. However, hunting waterfowl has a long history, and most of these species have persisted over time (Cooch et al., 2014). So paradoxically, hunting immature individuals does not always induce a decrease of recruitment rate. The heterogeneity of quality among immature individuals might explain this by inducing a compensation process if the harvest selects mostly individuals with a low reproductive potential (Gimenez et al., 2017; Lindberg et al., 2013). The recruitment rate can even be enhanced because the harvest pressure reduces the densities and potentially the competition for space and food (Nummi et al., 2015; Péron et al., 2012). In conclusion, a limited exploitation of a population observing heterogeneity and/or density-dependent effect might increase or at least not affect the recruitment rate. A balance between adult mortality and recruitment rate can be theoretically reached if the population is exploited to a constant and moderate harvest rate (Tsikliras & Froese, 2018). The size of a newly exploited population in stable environmental conditions should thus reach a different equilibrium, which is expected to be lower than the size without exploitation.

The fecundity and the juvenile survival, the two components of the recruitment rate, are very sensitive to weather conditions (Blums & Clark, 2004; Folliot et al., 2017). The environment conditions are thus an additional source of variability of the recruitment rate. The response of the recruitment rate to a harvest pressure is thus not straightforward. It is noisy because of its sensitivity to environmental factors. It also depends directly on the intensity of harvest on the immature individuals and indirectly on the mitigation of the density-dependent effects.

*What part of the recruitment rate gradient is necessary for population management?*

In population management, a common question for a threaten or an invasive species is to determine the measure that will allow to reach a targetted population size/density in a limited time period (Shea & NCEAS Working Group on Population Management, 1998). For such question, the carrying capacity of the habitat is theoretically not reached. So a manager should mostly focus on collecting knowledge on the maximum recruitment rate and its variability by removing the potential density-dependent effects occurring only for the highest theoretical densities (Péron, 2013, p. @Eraud2021).

*What is the ideal biological model to explore the gradient of the recruitment rate?*

Hunting waterfowl has a long history, but the biological effects of long-term moratorium have been poorly surveyed (Martínez-Abraín et al., 2013). So there are few opportunities to explore the response of the recruitment rate to a gradient of pressure for such species. Alien vertebrate species are often introduced during unintentional or intentional release events (Saul et al., 2016). They observe commonly a first period of colonisation on new territories without management measures. Since its presence might impact the balance of the colonized ecosystem, it happens that managers take strong control measures to restrict the growth of these populations or to eradicate them (Oficialdegui et al., 2020). Such species are then a good biological model to overcome density-dependent effects and explore the effects of a harvest gradient on recruitment rate.

*How to estimate recruitment rate?*

The recruitment rate parameter of a population can be assessed first by evaluating the fecundity of the breeding population and then by monitoring the survival from the egg life stage to the first-breeding event. To do so, it is necessary to monitor nests in habitats representative of the managed population (Blums & Clark, 2004). The survival the chick to its recruitment can be done by capture mark recapture method quite tricky to set up [Blums & Clark (2004); Schmidt2008]. Even if such approach provides valuable information on the dynamics of the recruitment and can highlight the bottlenecks, the fieldwork required to estimate both components is too time-consuming to generalize such approach for waterfowl population management.

In waterfowl, it is generally impossible to differentiate immatures from adults without examining birds in hand. Aging individuals from samples coming from hunting bags is thus another possible approach. Such approach suffers from two caveats. First, it is impossible to assess the recruitment rate when there is no harvest whereas it is necessary to predict what would happen if a moratorium seems required to allow a population to recover, or to calibrate a priori the harvest effort necessary to control an invasive species

(Smith et al., 2005). Secondly, like any predation action, hunting is selective and provides a biased picture of the population age structure in favor of the immatures (Bellrose, 1980; Fox et al., 2014). The hunting bag analysis might then lead to overestimate the recruitment rate and consequently the maximum sustainable harvest rate, with adverse consequences on population management (Fox et al., 2014).

In waterfowl, it is common that adult males display brighter colors than females when immatures from both sexes display cover-up plumage similar to adult females (Johnson, 1999). Knowing sex ratios, this discrepancy between adult and immature males might be used to estimate the proportion of immature in a waterfowl population just by counting them from distance (Smith et al., 2001). Such approach is promising because it requires only to count the proportion of birds of the two plumage kinds to infer the recruitment rate. Unlike the two previous methods, this one is non-invasive and requires a little time investment, which allows to collect long time series and consequently to track recruitment rate variability.

#### *Methodology of this study*

From two populations of ruddy duck, an species introduced in Europe in the 40's, this study develops a Bayesian framework to infer recruitment rate from count data. A comparison to the results of the common approach using samples from hunting bags aims at checking if the recruitment rate variability is well tracked and if the hypothesis of the recruitment rate overestimation by sampling methods is consistent with the results. Using two other approaches using average population growth rates, one testes the consistency of the order of magnitude of the recruitment rate estimates from the counting method.

## 2 Materials & methods

### 2.1 Biological model

The ruddy duck is a diving duck species introduced in the United Kingdom in the 40's (Gutiérrez-Expósito et al., 2020). The first reproduction was observed in the 60's. The population rapidly grew to reach more than 5000 individuals spread over the entire country. A new population set up in France from the 90's, likely because of an arrival of few individuals from the United Kingdom. This population set up around a single wintering spot, the Grand Lieu Lake (geographical coordinates: 47.09°N, 1.67°W). Since this species is considered as a threat to the white-headed duck because of the risk of genetic introgression (Muñoz-Fuentes et al., 2007), an European plan of eradication has been adopted and control measures were taken in both countries (Gutiérrez-Expósito et al., 2020).

Just before the reproduction period, i.e. in winter, one can distinguish from shore individuals with male-like plumage from female-like individuals (Figure 1). It is noticeable that the apparent proportion of males in the counts from shore is far below the proportion of males in the samples from the hunting bag (Figure 2).



Figure 1: Typical observation of a ruddy duck flock in winter: 4 male-like individuals with a white cheek and a black cap, 10 female-like individuals with a whitish cheek and a dark stripe across it, 3 unidentified individuals © Jay McGowan - 3 February 2013 - Tompkins, New York, United States

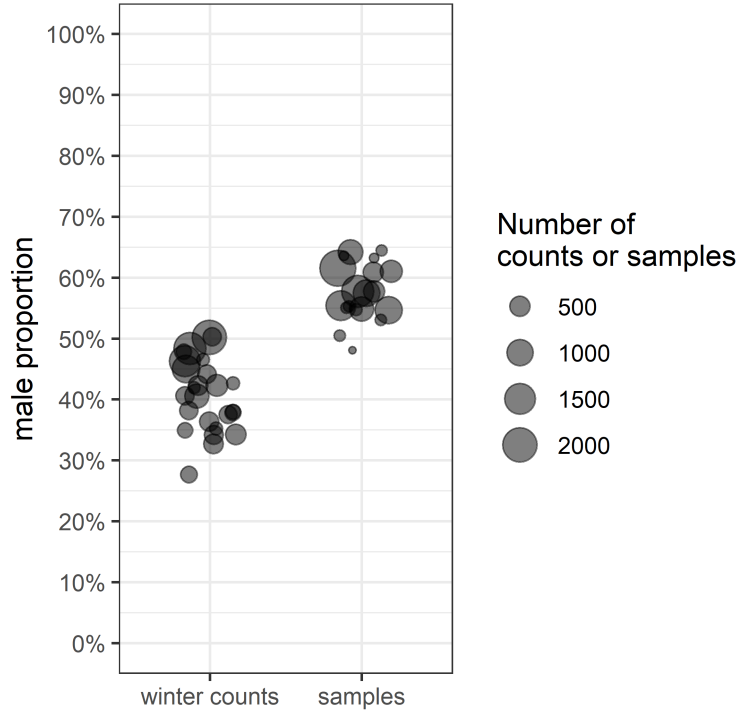


Figure 2: Apparent male proportion in winter counts vs male proportion in samples; for the first category, one point corresponds to the proportion of male-like plumage in a population counted in winter; for the second category, the male proportion is estimated from all ruddy ducks sampled over one year

Since sex identification of samples is very accurate **ref required**, the counting method misses the identification of a significant part of males in the population. In many waterfowl species, immature individuals of both sexes display cover-up plumage similar to adult female plumage **ref required**. Consequently, the counts in winter allows adult males to be distinguished from all other individuals (adult females, immature females, and immature males). This property can be valuable to estimate the recruitment with a few assumptions. One assumes in this study that the immature individuals have a equally balanced sex ratio (Bellrose et al., 1961; Blums & Mednis, 1996; Wood et al., 2021)

Ruddy duck populations from the UK and from France follow similar demographic histories (Figure 1). They grew freely during some years in both countries until 1999. Limited control measures have been then applied from 1999 to 2005 in the UK and from 1999 to 2018 in France. In both countries, these management measures led to stop the expansion of the ruddy ducks. From 2005 in the UK and from 2018 in France, higher control effort led to a rapid depletion in both populations.



Figure 3: Evolution of the size of the two Ruddy duck populations in light of the data time series;

Such biological model is thus suited to study the response of the harvest pressure on recruitment rate.

## 2.2 Data time series

The best picture stands just before the breeding period, so generally in winter for waterfowl.

The French and the UK wintering populations have been monitored since the presence of the first individuals was validated.

$PR_{i,t}$  estimation by the counting method

## 2.3 Modelling

Name	Class	Description
$SM$	Data	Total number of adult males sampled
$SF$	Data	Total number of adult females sampled
$PM$	Parameter	Proportion of males in the adult part of population at its equilibrium
$S_{i,t}$	Data	Number of individuals sampled in population $i$ in year $t$
$SJ_{i,t}$	Data	Number of juveniles sampled in population $i$ in year $t$
$CM_{i,t}$	Data	Cumulated number of type-male individuals counted in population $i$ in year $t$

Name	Class	Description
$C_{i,t}$	Data	Cumulated number of individuals counted in population $i$ in year $t$
$Cmax_{i,t}$	Data	Maximum number of individuals counted in population $i$ in year $t$
$PR_{i,t}$	Parameter	Proportion of recruits in population $i$ in year $t$
$R_{i,t}$	Parameter	Number of recruits in population $i$ in year $t$
$P_{i,t}$	Parameter	Number of recruits in population $i$ in year $t$ per breeder in year $t - 1$
$N0_j$	Parameter	Intercept of the regression model, the $j$ index refers to a restricted time period for a single population
$SD_j$	Parameter	Standard deviation of the regression model, the $j$ index refers to a restricted time period for a single population
$\lambda_j$	Parameter	Population growth rate (in $year^{-1}$ ), the $j$ index refers to a restricted time period for a single population
$S_j$	Parameter	Adult survival rate (in $year^{-1}$ ), the $j$ index refers to a restricted time period for a single population

The proportion of males in the adult part of a population at its equilibrium is inferred from the samples. Indeed, the available French samples were insufficient to get Since the . Age in samples to 2003 to 2020 for the United Kingdom and

$$PM \sim \text{Beta}(SM, SF) \quad (1)$$

$PR_{i,t}$  estimation by the counting method

$$CM_{i,t} \sim \text{Binom}(PM.(1 - PR_{i,t}), C_{i,t}) \quad (2)$$

$PR_{i,t}$  estimation by the sampling method

$$SJ_{i,t} \sim \text{Binom}(PR_{i,t}, S_{i,t}) \quad (3)$$

From  $PR_{i,t}$  estimates,  $R_{i,t}$  and  $P_{i,t}$  are expressed as:

$$R_{i,t} = PR_{i,t}.Cmax_{i,t} \quad (4)$$

$$P_{i,t} = \frac{PR_{i,t}.Cmax_{i,t}}{Cmax_{i,t-1}} \quad (5)$$

The index  $j$  refers to a single population  $i$  and a restricted time period for  $t$ . The evolution of the population size can be related to the time to infer the population growth rate. It is a linear regression on the logarithm scale:

$$\log(Cmax_{i,t}) \sim \text{Norm}(N0_j + \log(\lambda_j).t, SD_j) \quad (6)$$

$$S_j = \lambda_j - \overline{P_{i,t}} \quad (7)$$

(R Core Team, 2020) R package Nimble (de Valpine et al., 2017)

Bayesian framework is efficient to propagate error through the parameters

We generated three chains of length 25000, discarding the first 12500 samples as burn- in. Convergence of all chains was assessed using the Gelman and Rubin convergence diagnostic ( $R < 1.1$ , Gelman and Rubin 1992). We fit the models using NIMBLE (de Valpine et al. 2017) run from R (R Core Team 2019).

## 2.4 Comparison with the sampling method

## 3 Result validation using survival analysis

The adult survival is commonly assessed by capture mark recapture and published in the literature. Since the survival can also be estimated from both the counting and the sampling methods, the consistency of the two approaches is discussed by comparing the survival estimates with the literature values.

Name	Description
$N_t$	Number of breeders in year $t$
$D_t$	Number of breeders dead after reproduction in year $t$
$R_t$	Number of recruits in year $t$
$S_t$	Adult survival rate, i.e. proportion of breeder in year $t - 1$ still alive in year $t$
$P_t$	Number of recruits in year $t$ produced per breeder in year $t - 1$
$\lambda_t$	Growth rate of the population between the reproduction period of year $t - 1$ and year $t$

$$\begin{aligned}
N_t &= N_{t-1} - D_{t-1} + R_t \\
N_t &= N_{t-1} - (1 - S_t).N_{t-1} + P_t.N_{t-1} \\
N_t &= S_t.N_{t-1} + P_t.N_{t-1} \\
\frac{N_t}{N_{t-1}} &= S_t + P_t \\
\lambda_t &= S_t + P_t
\end{aligned} \tag{8}$$

The population size in year  $t$  is equal to the number of breeders that survived over year  $t - 1$  plus the offspring produced in year  $t - 1$  that survived until the reproduction period of year  $t$ , i.e. the recruitment in year  $t$ . The growth rate of a population is thus the sum of the recruitment rate and the adult survival rate.

The proportion of males in the adult part of a population at its equilibrium is inferred from the samples.

Indeed, the available French samples were insufficient to get Since the . Age in samples to 2003 to 2020 for the United Kingdom and

develop the counting method :

2 ways to check the method :

sampling method? Ok but few data as poorly harvested

realistic survival? if  $> 1$ , estimated recruitment rate is too low to support the maximum growth rate in the range of similar species in litterature -> satisfying

blabla



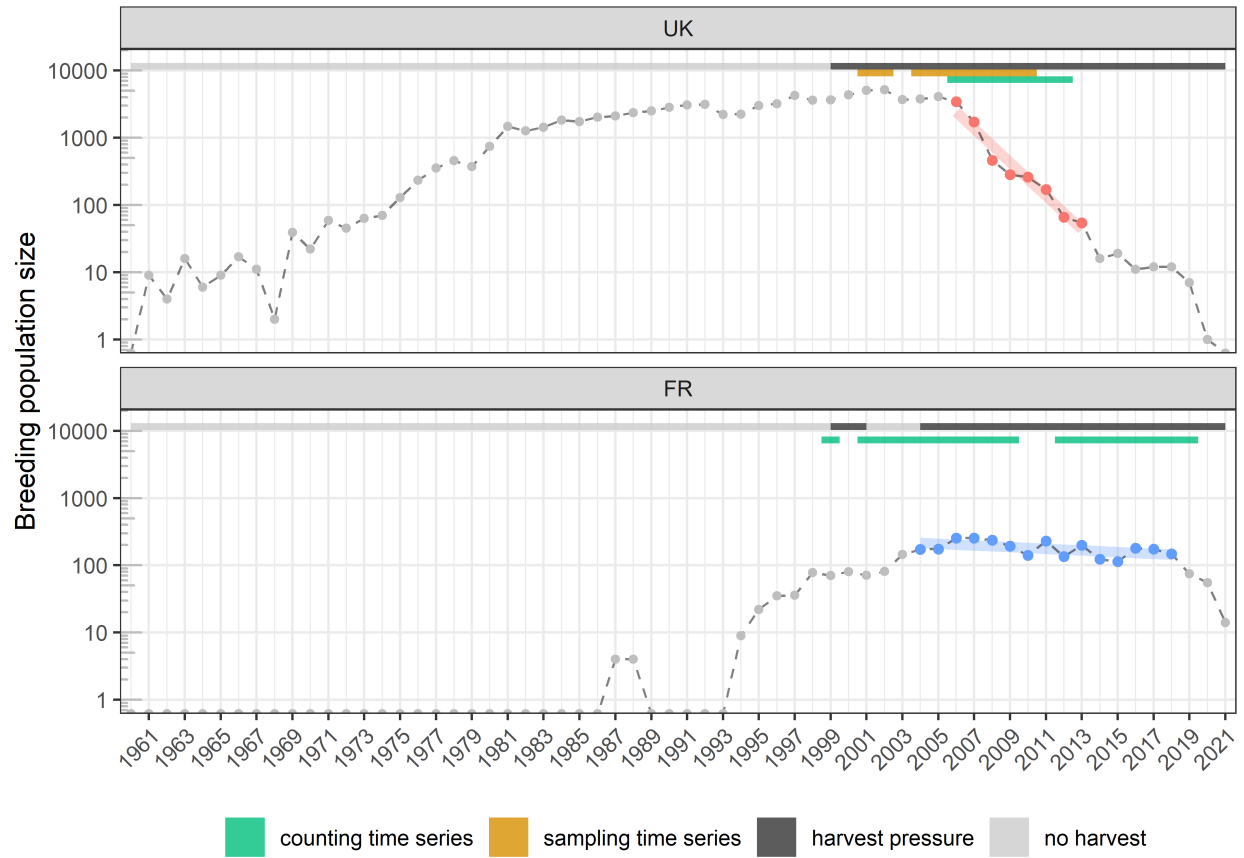


Figure 4: Evolution of the size of the two Ruddy duck populations in light of the harvest pressure; on time periods with data time series, population growth rates can be estimated (slopes in red for the UK and in blue for France)

### 3.1 Result validation using maximum recruitment rate proxy

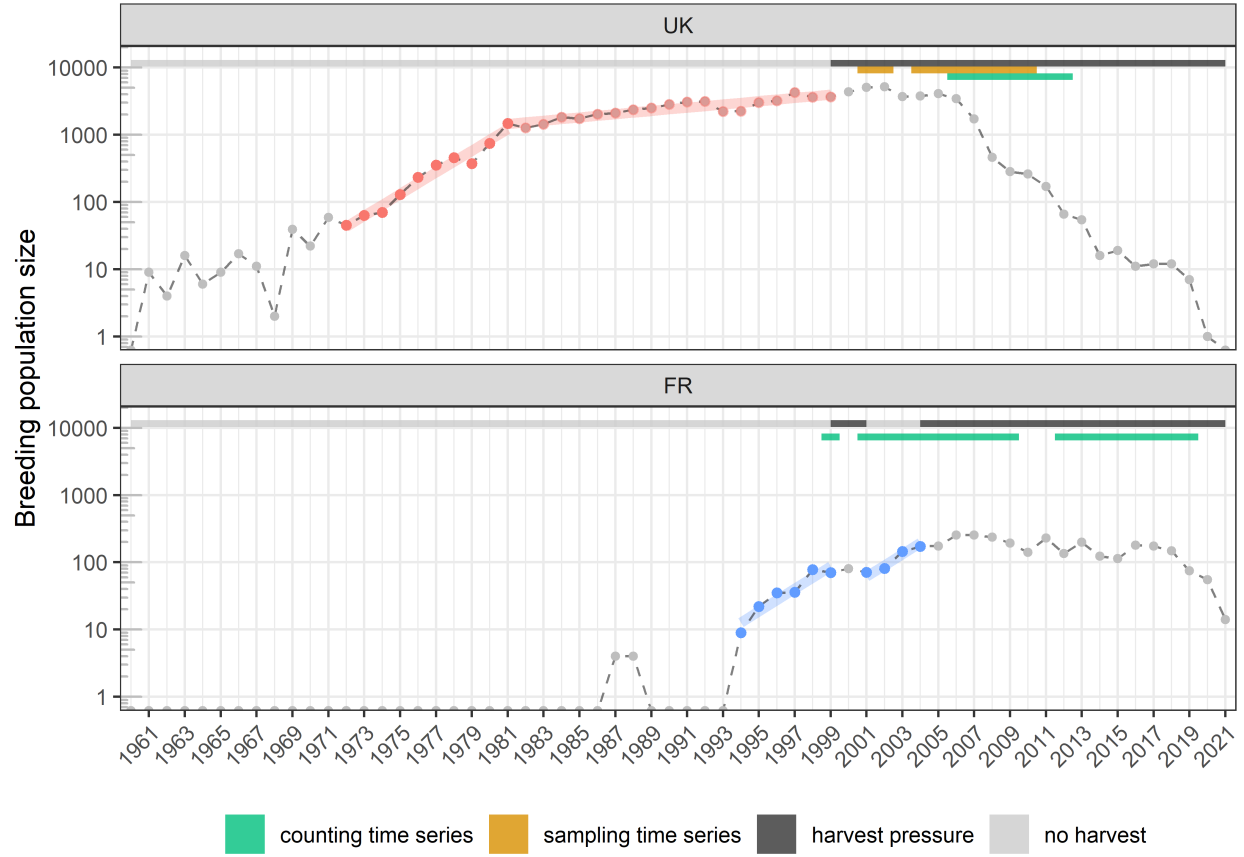


Figure 5: Evolution of the size of the two Ruddy duck populations in light of the harvest pressure; on time periods without harvest, maximum population growth rates can be estimated (slopes in red for the UK and in blue for France)

## 4 Results

### 4.1 recruitment rate variability

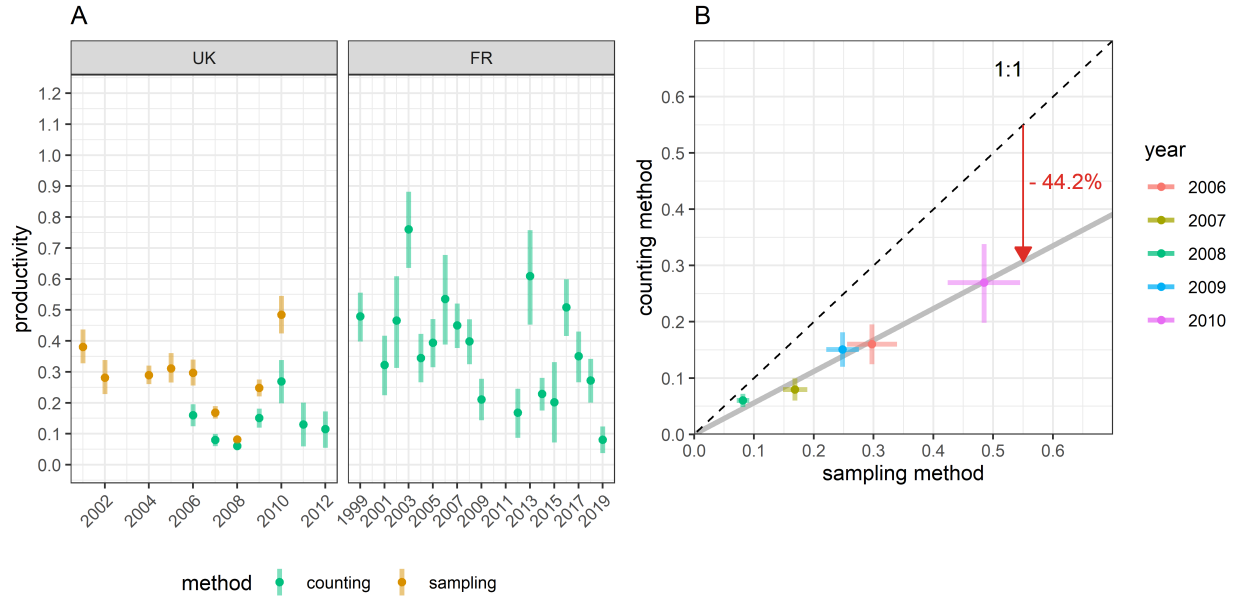


Figure 6: (A) Estimates of recruitment rate, i.e. average number of recruits produced per breeder; (B) Comparison of the recruitment rate values between the two estimation methods; bars define the 95% confidence intervals

quite stable when harvest rate is limited (France 2000), but is lower and noisy when harvest rate is huge and the population depleted.

## 4.2 Result validation using survival analysis

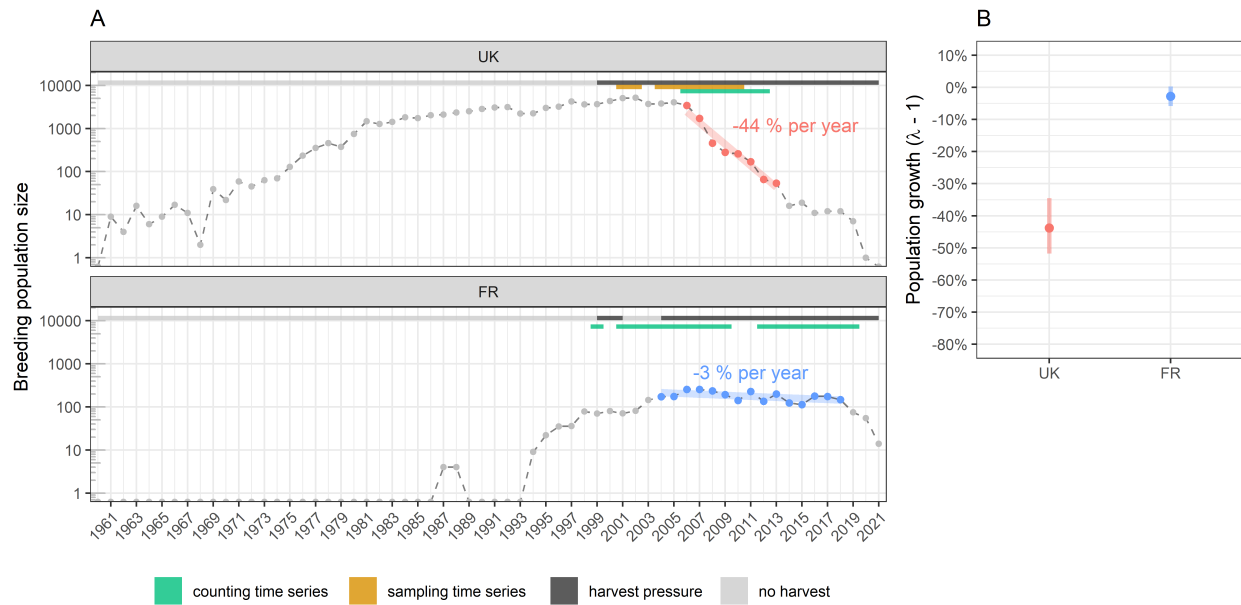


Figure 7: (A) Estimates of the population growth rate on time intervals with data availability. This aims at producing survival estimates to discuss the reliability of the methods to estimate recruitment rate; (B) Growth rate estimates; bars define the 95% confidence intervals

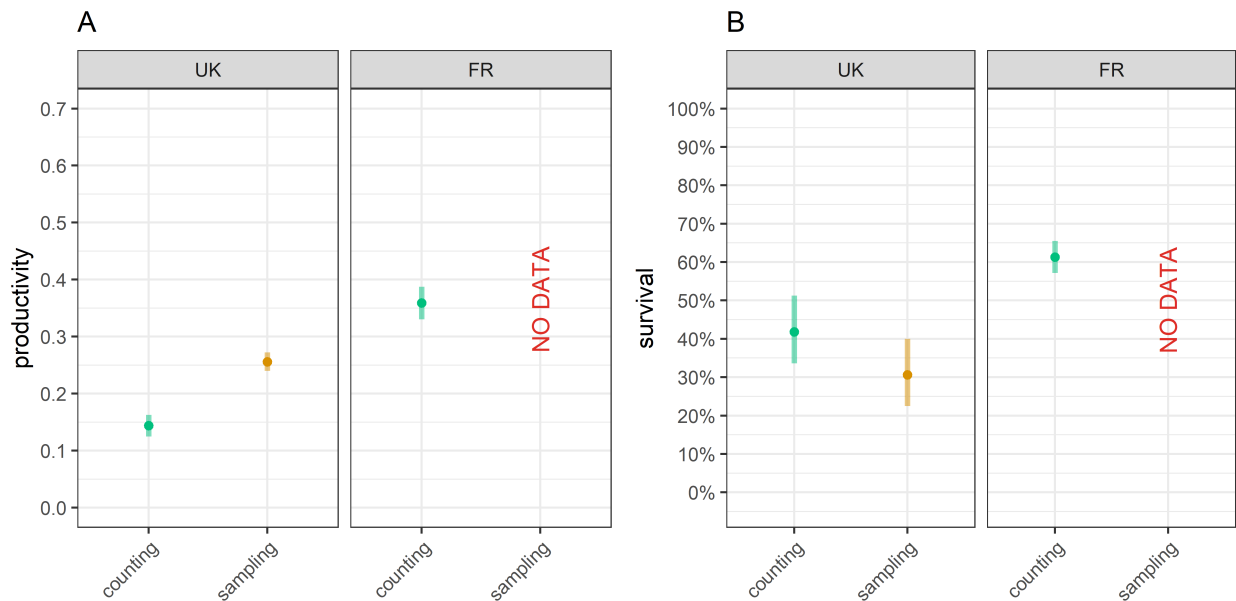


Figure 8: (A) Average recruitment rate estimates on the time intervals defined for survival analysis; (B) Average survival estimates deduced from growth rate and recruitment rate estimates on the time intervals; bars define the 95% confidence intervals

### 4.3 Result validation using maximum recruitment rate proxy

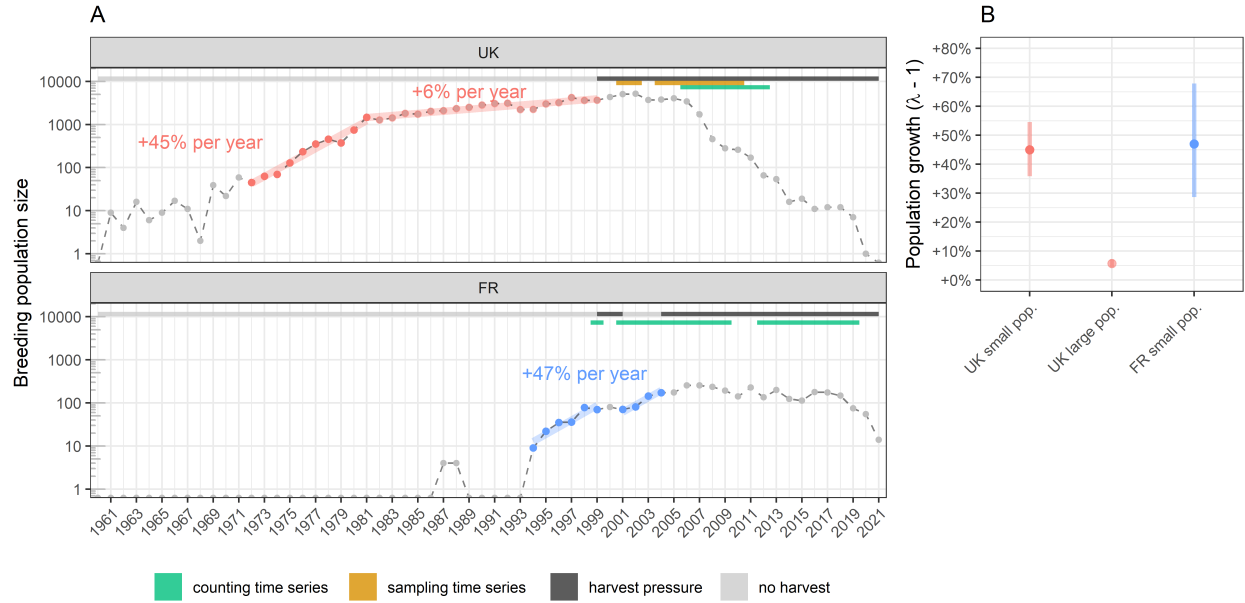


Figure 9: (A) Estimates of the population growth rate without harvest; two estimates were produced for the UK population because the growth rate dropped significantly when its size reached more than 1000 individuals; (B) Growth rate estimates; bars define the 95% confidence intervals

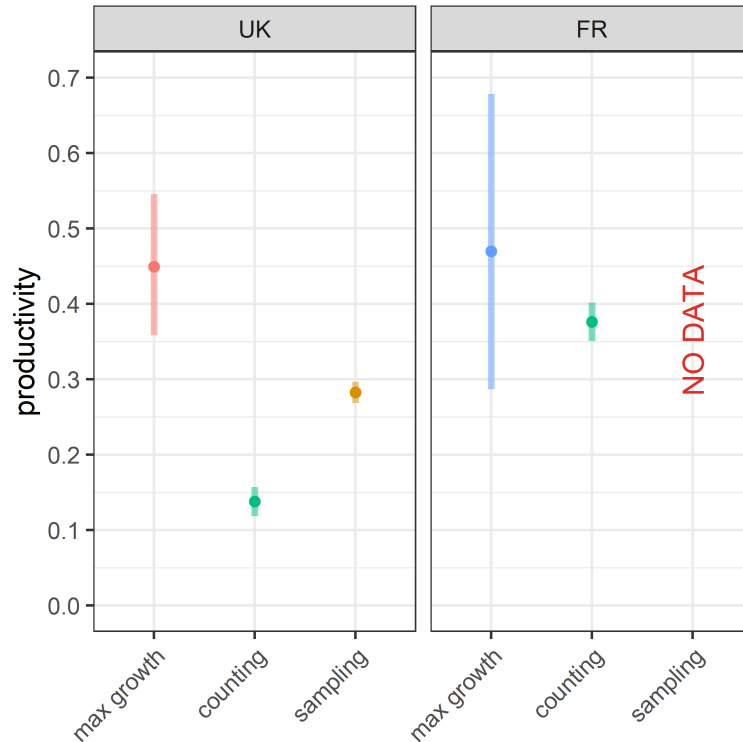


Figure 10: Comparison between a proxy of maximum recruitment rate and the average recruitment rate estimated following *counting* and *sampling* methods; the proxy is estimated during the maximum growth periods by combining the growth rate to the hypothesis of a 100% adult survival (*max growth*); this hypothesis implies that the true maximum of recruitment rate is likely higher; bars define the 95% confidence intervals

## 5 Discussion

According to Blums & Clark (2004), the recruitment for diving ducks is not related to the fecundity, but depends mostly on the juvenile survival, which is driven by weather conditions. Studies on other birds also conclude of the poor correlation between fecundity and recruitment (e.g. Murray, 2000).

age proportion  $\rightarrow [0, 1]$  recruitment rate  $[@, + \infty]$  donc plus sensitive

recruitment rate the most sensitive parameter to density dependent effect (Koons et al., 2014)

Hypothesis of time invariant sex ratios in immature and adult: limits

For hunted duck species in North America, the mortality rate on first-year immatures compared to adults is higher by a factor ranging from 1.4 to 2 (Bellrose 1980).

recruitment rate: Stabilité grosse pop, variabilité petite pop.

Au dela du fait que la croissance de la pop est la somme de la survie et de la recruitment rate, intéressant de parler de la variability de la survie vs variabilité de la recruitment rate pour parler de risk management.

recruitment rate: drop from 2009

one considers harvest as an impact only adult survival, not on age structure, but we should, oversimplistics blabla

Nichols & Williams (2006)  $\rightarrow$  il faut compter de la façon dont on demande

Taper des jeunes c'est pas forcément très utile car density dependence

If the maximum recruitment rate is not reached under optimal conditions (maximum population size not reached, and limited harvest pressure), it is a signal that one should explore the intermediate life stages before recruitment to provide a diagnostic on the changes of the environment conditions that have affected the recruitment rate.

Condition of application to other species.

## Aknowledgements

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## References

- Anders, A. D., & Marshall, M. R. (2005). Increasing the accuracy of productivity and survival estimates in assessing landbird population status. *Conservation Biology*, 19(1), 66–74.
- Bellrose, F. C. (1980). *Ducks, geese & swans of North America: 3rd. Ed.* Stackpole Books. <https://books.google.fr/books?id=VpcWzgEACAAJ>
- Bellrose, F. C., Scott, T. G., Hawkins, A. S., & Low, J. B. (1961). Sex ratios and age ratios in North American ducks. *Illinois Natural History Survey Bulletin*, 27(6), 391–486.
- Blums, P., & Clark, R. G. (2004). Correlates of lifetime reproductive success in three species of European ducks. *Oecologia*, 140(1), 61–67. <https://doi.org/10.1007/s00442-004-1573-8>
- Blums, P., & Mednis, A. (1996). Secondary sex ratio in Anatinae. *The Auk*, 113(2), 505–511. <https://doi.org/10.2307/4088920>
- Cooch, E. G., Guillemain, M., Boomer, G. S., Lebreton, J.-D., & Nichols, J. D. (2014). The effects of harvest on waterfowl populations. *Wildfowl*, 220–276.
- de Valpine, P., Turek, D., Paciorek, C. J., Anderson-Bergman, C., Lang, D. T., & Bodik, R. (2017). Programming with models: Writing statistical algorithms for general model structures with NIMBLE. *Journal of Computational and Graphical Statistics*, 26(2), 403–413. <https://doi.org/10.1080/10618600.2016.1172487>
- Eraud, C., Devaux, T., Villers, A., Johnson, F. A., & Francesiaz, C. (2021). Popharvest: An R package to assess the sustainability of harvesting regimes of bird populations. *Ecology and Evolution*. <https://doi.org/10.1002/ece3.8212>
- Etterson, M. A., Ellis-Felege, S. N., Evers, D., Gauthier, G., Grzybowski, J. A., Mattsson, B. J., Nagy, L. R., Olsen, B. J., Pease, C. M., Burg, M. P. van der, & Potvien, A. (2011). Modeling fecundity in birds: Conceptual overview, current models, and considerations for future developments. *Ecological Modelling*, 222(14), 2178–2190. <https://doi.org/10.1016/j.ecolmodel.2010.10.013>
- Flint, P. (2015). Population dynamics of sea ducks: Using models to understand the causes, consequences, evolution, and management of variation in life history characteristics. In J.-P. L. Savard, D. V. Derksen, D. Esler, & J. M. Eadie (Eds.), *Ecology and conservation of North American sea ducks* (Vol. 46, pp. 63–96). CRC Press.

- Folliot, B., Caizergues, A., Barbotin, A., & Guillemain, M. (2017). Environmental and individual correlates of common pochard (*Aythya ferina*) nesting success. *European Journal of Wildlife Research*, 63(4). <https://doi.org/10.1007/s10344-017-1126-1>
- Fox, A. D., Clausen, K. K., Dalby, L., Christensen, T. K., & Sunde, P. (2014). Age-ratio bias among hunter-based surveys of Eurasian wigeon *Anas penelope* based on wing vs. Field samples. *Ibis*, 157(2), 391–395. <https://doi.org/10.1111/ibi.12229>
- Jimenez, O., Cam, E., & Gaillard, J.-M. (2017). Individual heterogeneity and capture-recapture models: What, why and how? *Oikos*, 127(5), 664–686. <https://doi.org/10.1111/oik.04532>
- Gunnarsson, G., Elmberg, J., Pöysä, H., Nummi, P., Sjöberg, K., Dessborn, L., & Arzel, C. (2013). Density dependence in ducks: A review of the evidence. *European Journal of Wildlife Research*, 59(3), 305–321. <https://doi.org/10.1007/s10344-013-0716-9>
- Gutiérrez-Expósito, C., Pernollet, C., Adriaens, T., & Henderson, I. (2020). Ruddy duck (*Oxyura jamaicensis* Gmelin, 1789). In *Invasive birds: Global trends and impacts* (pp. 200–205). CABI.
- Iverson, S. A., Reed, E. T., Hughes, R. J., & Forbes, M. R. (2013). Age and breeding stage-related variation in the survival and harvest of temperate-breeding Canada geese in Ontario. *The Journal of Wildlife Management*, 78(1), 24–34. <https://doi.org/10.1002/jwmg.636>
- Johnson, D. H., Sparling, D. W., & Cowardin, L. M. (1987). A model of the productivity of the mallard duck. *Ecological Modelling*, 38(3-4), 257–275. [https://doi.org/10.1016/0304-3800\(87\)90100-1](https://doi.org/10.1016/0304-3800(87)90100-1)
- Johnson, K. P. (1999). The evolution of bill coloration and plumage dimorphism supports the transference hypothesis in dabbling ducks. *Behavioral Ecology*, 10(1), 63–67. <https://doi.org/10.1093/beheco/10.1.63>
- Koons, D. N., Arnold, T. W., & Schaub, M. (2017). Understanding the demographic drivers of realized population growth rates. *Ecological Applications*, 27(7), 2102–2115. <https://doi.org/10.1002/eap.1594>
- Koons, D. N., Gunnarsson, G., Schmutz, J. A., & Rotella, J. J. (2014). Drivers of waterfowl population dynamics : From teal to swans. *Wildfowl, Special Issue 4*, 169–191. <http://wildfowl.wwt.org.uk/index.php/wildfowl/article/view/2606>
- Lebreton, J.-D. (2001). The use of bird rings in the study of survival. *Ardea*, 89(1), 85–100.
- Lindberg, M. S., Sedinger, J. S., & Lebreton, J.-D. (2013). Individual heterogeneity in black brant survival and recruitment with implications for harvest dynamics. *Ecology and Evolution*, 3(12), 4045–4056. <https://doi.org/10.1002/ece3.767>
- Malthus, T. R. (1872). *An essay on the principle of population*.
- Martínez-Abrán, A., Viedma, C., Gómez, J. A., Bartolomé, M. A., Jiménez, J., Genovart, M., & Tenan, S. (2013). Assessing the effectiveness of a hunting moratorium on target and non-target species. *Biological Conservation*, 165, 171–178. <https://doi.org/10.1016/j.biocon.2013.06.009>
- Muñoz-Fuentes, V., Vilà, C., Green, A. J., Negro, J. J., & Sorenson, M. D. (2007). Hybridization between white-headed ducks and introduced ruddy ducks in Spain. *Molecular Ecology*, 16(3), 629–638. <https://doi.org/10.1111/j.1365-294x.2006.03170.x>
- Murray, B. G. (2000). Measuring annual reproductive success in birds. *The Condor*, 102(2), 470–473. <https://doi.org/10.1093/condor/102.2.470>
- Nichols, J. D., & Williams, B. K. (2006). Monitoring for conservation. *Trends in Ecology & Evolution*, 21(12), 668–673. <https://doi.org/10.1016/j.tree.2006.08.007>
- Nummi, P., Holopainen, S., Rintala, J., & Pöysä, H. (2015). Mechanisms of density dependence in ducks: Importance of space and per capita food. *Oecologia*, 177(3), 679–688. <https://doi.org/10.1007/s00442-014-3133-1>
- Oficialdegui, F. J., Delibes-Mateos, M., Green, A. J., Sánchez, M. I., Boyero, L., & Clavero, M. (2020). Rigid laws and invasive species management. *Conservation Biology*. <https://doi.org/10.1111/cobi.13481>



- Otis, D. L. (2006). Mourning dove hunting regulation strategy based on annual harvest statistics and banding data. *The Journal of Wildlife Management*, 70(5), 1302–1307. [https://doi.org/10.2193/0022-541x\(2006\)70%5B1302:mdhrsb%5D2.0.co;2](https://doi.org/10.2193/0022-541x(2006)70%5B1302:mdhrsb%5D2.0.co;2)
- Péron, G. (2013). Compensation and additivity of anthropogenic mortality: Life-history effects and review of methods. *Journal of Animal Ecology*, 82(2), 408–417. <https://doi.org/10.1111/1365-2656.12014>
- Péron, G., Nicolai, C. A., & Koons, D. N. (2012). Demographic response to perturbations: The role of compensatory density dependence in a north american duck under variable harvest regulations and changing habitat. *Journal of Animal Ecology*, 81(5), 960–969. <https://doi.org/10.1111/j.1365-2656.2012.01980.x>
- R Core Team, (2020). *R: A language and environment for statistical computing*. R Foundation for Statistical Computing. <https://www.R-project.org/>
- Robertson, G. J. (2008). Using winter juvenile/adult ratios as indices of recruitment in population models. *Waterbirds*, 31(sp2), 152–158. <https://doi.org/10.1675/1524-4695-31.sp2.152>
- Saul, W.-C., Roy, H. E., Booy, O., Carnevali, L., Chen, H.-J., Genovesi, P., Harrower, C. A., Hulme, P. E., Pagad, S., Pergl, J., & Jeschke, J. M. (2016). Assessing patterns in introduction pathways of alien species by linking major invasion data bases. *Journal of Applied Ecology*, 54(2), 657–669. <https://doi.org/10.1111/1365-2664.12819>
- Sayre, N. F. (2008). The genesis, history, and limits of carrying capacity. *Annals of the Association of American Geographers*, 98(1), 120–134. <https://doi.org/10.1080/00045600701734356>
- Schmidt, K. A., Rush, S. A., & Ostfeld, R. S. (2008). Wood thrush nest success and post-fledging survival across a temporal pulse of small mammal abundance in an oak forest. *Journal of Animal Ecology*, 77(4), 830–837. <https://doi.org/10.1111/j.1365-2656.2008.01378.x>
- Shea, K., & NCEAS Working Group on Population Management. (1998). Management of populations in conservation, harvesting and control. *Trends in Ecology & Evolution*, 13(9), 371–375. [https://doi.org/10.1016/s0169-5347\(98\)01381-0](https://doi.org/10.1016/s0169-5347(98)01381-0)
- Smith, C. M., Goudie, R. I., & Cooke, F. (2001). Winter age ratios and the assessment of recruitment of harlequin ducks. *Waterbirds: The International Journal of Waterbird Biology*, 24(1), 39. <https://doi.org/10.2307/1522241>
- Smith, G. C., Henderson, I. S., & Robertson, P. A. (2005). A model of ruddy duck *Oxyura jamaicensis* eradication for the UK. *Journal of Applied Ecology*, 42(3), 546–555. <https://doi.org/10.1111/j.1365-2664.2005.00977.x>
- Tsikliras, A. C., & Froese, R. (2018). Maximum sustainable yield. In *Encyclopedia of ecology (second edition)* (Vol. 1, pp. 108–115). Elsevier. <https://doi.org/10.1016/b978-0-12-409548-9.10601-3>
- Williams, J. H., & Madsen, J. (2013). Stakeholder perspectives and values when setting waterbird population targets: Implications for flyway management planning in a european context. *PLOS ONE*, 8(11), e81836. <https://doi.org/10.1371/journal.pone.0081836>
- Wood, K. A., Brides, K., Durham, M. E., & Hearn, R. D. (2021). Adults have more male-biased sex ratios than first-winter juveniles in wintering duck populations. *Avian Research*, 12(1). <https://doi.org/10.1186/s40657-021-00286-1>

## A Supplement

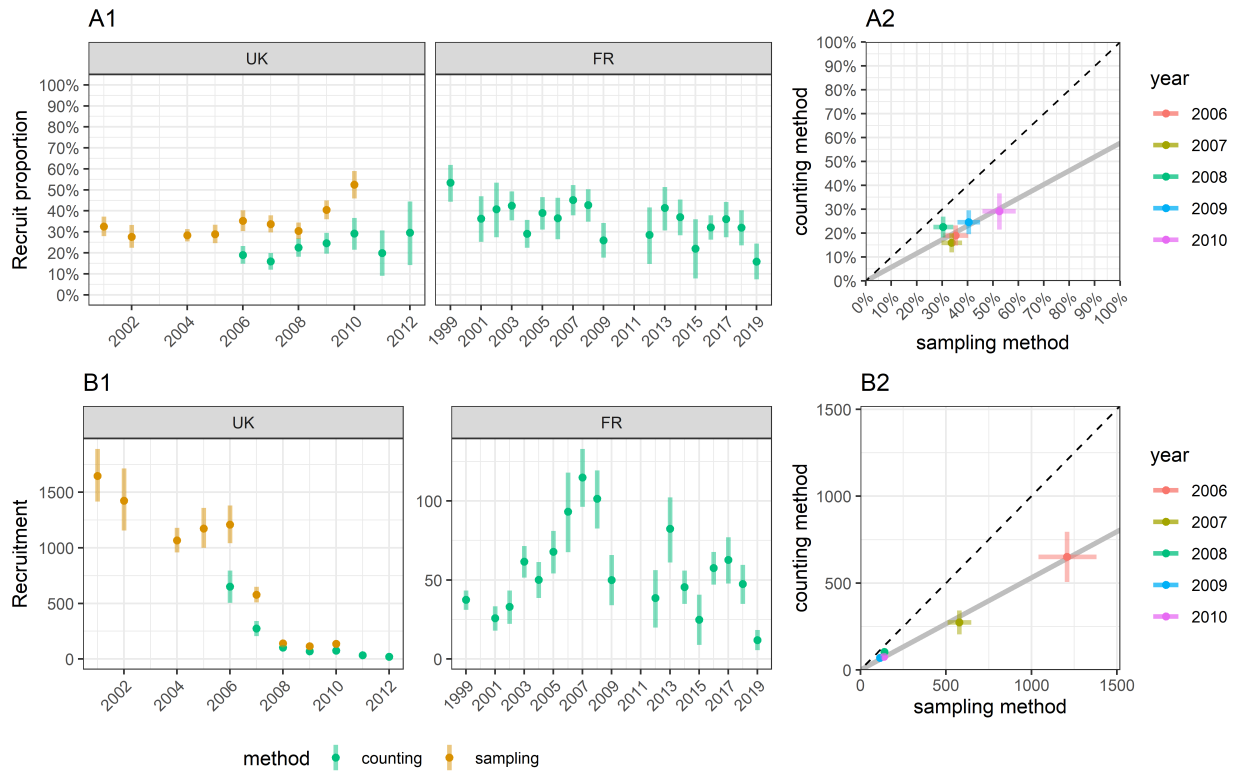


Figure 11: (A1) Estimates of the number of recruits, i.e. the young adults reproducing for their first time; (B1) estimates of the proportion of recruits in the population at the reproduction period; (A2) & (B2) comparison of the values between the two estimation methods; bars define the 95% confidence intervals