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# Intro

* The vegetation response to water deficit was assessed using remote sensing information and dendrocronologial data.

### Sequías

Although summer drought is a characteristic feature of the Mediterranean climate (Lionello 2012), an increase in the frequency and severity of drought events have been recorded for the Mediterranean region (Hoerling et al. 2012), particularly for southern Europe (Vicente-Serrano et al. 2014, Spinoni et al. 2015, Stagge et al. 2017) where a trend towards drier summers has been reported for the last three decades (Spinoni et al. 2017a). Además las proyecciones apuntan a que este incremento en la frecuencia continuará (Spinoni et al. 2017b).

In the Iberian Peninsula, major drought episodies were recorded in 1981, 1995, 2000 and 2005 (Vicente-Serrano et al. 2014). The 2004/2005 and 2011/2012 hydrological years are considered two of the worst drought periods recorded in the Iberian Peninsula, particularly in the southern sector (García-Herrera et al. 2007, Trigo et al. 2013, Gouveia et al. 2015). These events were extreme in both its magnitude and spatial extent (Gouveia and Trigo 2014).

notas sequia

* the drought events have been longest and most severe in the period 1991-2010 for mediterranean are of Southern Europe
* Seasonally, drought frequency is projected to increase everywhere in Europe for both scenarios in spring and summer, especially over southern Europe, and less intensely in autumn (Spinoni et al. 2017b) Aumento de la frecuencia de sequías en spring and summer desde 1950 hasta 2014 (Spinoni et al. 2017a) Both for frequency and severity, the evolution towards drier conditions is more relevant in the last three decades over Mediterranean area in summer, an increase in the drought severity in the Iberian Peninsula has been observed in the last decades (Vicente-Serrano et al. 2014).

Althought several works have reported these two years as some of the worst drought events, we characterised the drought at several spatio-temporal scales in the study area. From a long-term perspective, we compare the accumulated monthly precipitation at a meteorological station (Granada, Base Aérea) during the hydrological years 2004-2005 and 2011-2012 with the average of accumulated monthly precipitation for the period 1950-2015.

### Extreme sequias y ecologia

* ojo hablar de que la sequía es un proceso natural, pero que se está viendo aumentado en los últimos años:

Drought is a natural phenomenon that occurs when water availability is significantly below normal levels over a long period and the supply cannot meet the existing demand.

ver intro Ecosphere: Drought affects ecological systems across every climatic zone worldwide and is exacerbated by climate change and increasing anthropogenic water demands (Mishra and Singh 2010). Characterized by below-normal precipitation (Dai 2011), meteorological drought results from complex interactions between the atmosphere and hydrologic processes within the biosphere. Unlike aridity, which is a permanent feature of climate (Wilhite 1992), drought is a temporary extreme event (Palmer 1965, Mishra and Singh 2010) that can persist for extended time periods (months to years; Mishra and Singh 2010). Drought can cause significant changes in ecosystem productivity and water dynamics, and it is one of the most economically and ecologically disruptive extreme events affecting millions of people globally (Dai 2011).

Extreme climate events (e.g. droughts) severely affect forests and grasslands throuhg changes in plant physiology, phenology and carbon allocation (Ummenhofer and Meehl 2017) (incluir citas para los efectos de la sequía sobre la vegetación).

No solo efectos aislados, sino también en conjunción con otros factores, sobre todo en el mediterráneo donde convergen muchos de los factores que pueden interaccionar:

* La sequía es un factor crucial a tener en cuenta, ya que es además de los efectos que puede tener de forma aislada, se ha visto que además presenta muchas interacciones con otros factores, siendo por tanto un factor crucial ($IMPROVE$ ver (Doblas-Miranda et al. 2017)):
* Some interactions alter the effects of a single factor, as drought enhances or decreases the effects of atmospheric components on plant ecophysiology
* Drought and land use changes, among others, alter water resources and lead to land degradation, vegetation regeneration decline, and expansion of forest diseases.
* Climate change, and especially drought, emerges as a crucial factor in most of the reviewed interactions and therefore it should be considered when it comes to designing and applying international management policies
* Drought should be considered when designing and applying management policies.

Además la sequía se espera que tengan …Droughts are most likely to have the largest and most long-lasting impacts globally due to large indirect and lagged impacts and long recovery especially for forest ecosystems (ver 18 en Ummenhofer and Meehl (2017)). Así por ejemplo se ha visto que todos los biomas presentan una vulnerabilidad similar … En una revision sobre la vulnerabilidad del sistema de trasnporte en plantas al embolismo inducido por sequía, ha mostrado que una convergencia de la vulnerabilidad de los bosques a la sequía, mostrando que todos los biomas son igualmente vulnerables a los fallos hidráulicos independientemente del régimen de precipitación (Choat et al. 2012, <doi:10.1038/nature11688>)

– Existen evidencias que sugieren que muchos bosques son vulnerables a eventos climáticos extremos … (Zhang) y esto puede ser especialmente relevante para especies situadas en el rear edge (completar)

Una vez dicho esto…

* Como la dendro ayuda a evaluar las sequías (ver Gazol 2017)
* Remote sensing:
  + Uso de Remote sensing para estudiar la sequía –> Leer Zhang et al. (2013) y escribir algo. Drought monitoring using remote-sensing approach was originally applied to agriculture. Several remote-sensing derived indices have been used to study the drought effects on vegetation …
  + Leer también a AghaKouchak et al. (2015)

Climate change projections indicate that extreme events will become more common in the future (IPCC 2013), making it important that we understand how ecosystems respond to these events and the potential feedbacks to radiative forcing.

### 

Uso de NDVI como estimador de la NPP:

* The NDVI properties have allowed the use of this information for estimating the Net Primary Production (NPP) (Goward and Dye, 1987; Running et al., 2004; Hasenauer et al., 2012). Different studies have already found a strong relationship between NPP and radial growth (e.g., Granier et al., 2008; Babst et al., 2013, 2014a, 2014b; Vicente-Serrano et al., 2015), albeit with significant differences, particularly those related to species, sites and environmental conditions.
* Ver Gilaber et al. 2017 <http://www.mdpi.com/2072-4292/9/3/193>

$NOTA$: NDVI sirve para estimar la producción primaria neta. Existen diferentes estudio que han evaluado el efecto de la sequía sobre la producción primaria neta utilizando NDVI.

### Aims

$IMPROVE$

In this study we combined remote sensing information and dendroecological methods to evaluate the drought impacts in both greenees and growth of *Q. pyrenaica* forests in Sierra Nevada. Specifically,

The aims of this work were:

* To quantify how two extreme drought events influenced the greenness and radial growth of *Q. pyreancia* forests in their rear edge,
* to analyze the resilience of these forests to successive extreme drought events,
* and to explore differences in the resilience metrics between populations located in contrasting slopes within the rear edge of the distribution of this species.

Do northern and southern populations of Pyrenean oak forests differ in their resistance, resilience or recovery to extreme drought events?

## References

# Materials and methods

## Species and study site

The Pyrenean oak (*Quercus pyrenaica* Willd.) forests extend through south-western France and the Iberian Peninsula (Franco 1990) reaching its southern limit in north of Morocco. In the Iberian Peninsula these forests live under meso-supramediterranean and mesotemperate areas and subhumid, humid and hyperhumid ombroclimate (Rivas-Martínez et al. 2002) living on siliceous soils, or soils poor in basic ions (Vilches de la Serna 2014). *Q. pyrenaica* requires between 650 and 1200 mm of annual precipitation and a summer minimal precipitation between 100 and 200 mm (Martínez-Parras and Molero-Mesa 1982, García and Jiménez 2009), with summer rainfall being a key factor in the distribution of the species (Gavilan et al. 2007, Río et al. 2007).

This species reaches its southernmost European limit at Sierra Nevada, a high-mountain range located in southern Spain (37°N, 3°W) with elevations of between 860 m and 3482 m *a.s.l.*. The climate is Mediterranean, characterized by cold winters and hot summers, with pronounced summer drought (July-August). There are eight oak patches (2400 Has) identified (**Figure 1**) in this mountain range, ranging between 1100 and 2000 m *a.s.l.* and generally associated to major river valleys. Sierra Nevada is considered a glacial refugia for deciduous *Quercus* species during glaciation [Brewer et al. (2002); Olalde et al. (2002); RodriguezSanchez2010] and these populations are considered as a rear edge of the habitat distribution, which is important in determining habitat responses to expected climate change (Hampe and Petit 2005).

The populations of Pyrenean oak forests at Sierra Nevada are considered relict forests (Melendo and Valle 2000, Vivero et al. (2000)), undergoing intensive anthropic use in the last few decades (Camacho-Olmedo et al. 2002, Valbuena-Carabaña et al. 2010, Valbuena-Carabaña and Gil 2017). In fact, the status of conservation of this species for southern Spain is “Vulnerable” (Vivero et al. 2000). The relict presence of this species in Sierra Nevada is related both to its genetic resilience as well as to its high intraspecific genetic diversity (Valbuena-Carabaña et al. 2010). However, they are also expected to suffer the impact of climate change, due to their climate requirements (wet summers). Thus, simulations of the climate change effects on this habitat forecast a reduction in suitable habitats for Sierra Nevada (Benito et al. 2011).

## Drought episodes

* :red\_circle: Meter aquí algunos datos de sequia, similar a lo planteado por Gazol

:red\_circle: duda aqui Varias referencias hablan de los años 2005 y 2012 como extremadamente secos. Pero habría que hacer alguna referencia y/o análisis. Tengo dudas de si hemos de analizar (e incluir) que efectivamente los años 2005 y 2012 fueron caracterizados por un extrema sequía, por lo que habría que incluyendo referencia a apéndice

* O quizá un apartado llamado Drought episodes (similar a esto <https://www.nature.com/articles/srep28269>)

## Greenness data

To characterize the vegetation greenness of *Q. pyrenaica* we used the *Enhanced Vegetation Index* (EVI) derived from MOD13Q1 product obtained by the *Moderate Resolution Imaging Spectroradiometer* (MODIS) sensor (Didan 2015). EVI and NDVI (*Normalized Difference Vegetation Index*) are the most common greenness vegetation indices. We used EVI instead of NDVI because EVI is more sensitive to changes in high-biomass areas (a serious shortcoming of NDVI); EVI reduces the influence of atmospheric conditions on vegetation index values, and EVI corrects for canopy background signals (Huete et al. 2002, Cabello et al. 2012, Krapivin et al. 2015).

MODIS EVI Data from Collection 6 were obtanied using Google Earth Engine platform for the period 2000 - 2016. EVI data consits of 16-day maximun value composite images (23 per year) of the EVI value with a spatial resolution of 250 m x 250 m. We selected the pixels covering the distribution of *Q. pyrenaica* forests in Sierra Nevada (*n* = 928 pixels). A data filtering was applied to select EVI valid values. The filtering was done using quality flags and VI Usefulness Indices accompanying the EVI data. We filter out those values affected by high content of aerosols, clouds, snow and shadows, following the filtering recomendations of EVI data for mountain regions (Reyes-Díez et al. 2015).

After the filter out process, we built the annual EVI profile for each pixel and then computed the EVI’s annual mean values and the EVI anomaly for each pixel for the period 2000 - 2016. EVI mean was chosen since it is highly stable under the use of any filter (Reyes-Díez et al. 2015) and because it showed high significant correlations with annual (r = 0.81) and seasonal EVI values ( = 0.76 and = 0.88).

To explore the effect of drought events on greenness we calculated the EVI standardized anomaly () pixel-by-pixel, since it minimizes biases in the evaluation of anomalies, providing more information about the magnitude of the anomalies (Samanta et al. 2012, Gao et al. 2016). For each pixel we annually averaged all the EVI valid values, and then the standardized anomaly was computed as:

where is the EVI standardized anomaly for the year ; the annual mean value of EVI for the year ; the average of the annual EVI values for the period of reference (all except year), and the standard deviation for the reference period. Each pixel was categorized according the EVI standardized anomalies as “greening” (EVI standardized anomalies greater than + 1), “browning” (EVI standardized anomalies less than - 1) or “no changes” (EVI standardized anomalies between − 1 and + 1) (Samanta et al. 2010, 2012).

## Field sampling and dendrochronological methods

### Tree sampling

Sampling was carried during autumn of 2016. Trees were sampled at two locations in contrasting slopes of Sierra Nevada: San Juan (SJ; northern site) and Cáñar (CA; southern site) (Figure 1; Table 1). For the southern site two elevations were sampled: CA-Low and CA-High. All the sites were oak monospecific and representatives of two of the three the population’s cluster identified for the specie in this mountain range (Pérez-Luque et al. 2015). In each site between 15 and 20 dominant trees were randomly selected. Two cores of 5 mm of diameter were taken from each at breast heigth (1.3 m) using an increment borer. Diameter at breast height (DBH) and total height were measured using a girth tape and a Vertex IV ultrasonic hypsometer (Haglöf, Sweden) respectively. We assessed stand competition of the target tree by recording all neighboring living trees with dbh > 7.5 cm in a circle of 10 m radi. The species, dbh and total height were annotated. We also measured distance and azimuth to the target tree. (:red\_circle: indices competencia??)

### Dendroecological analyses

Tree cores were air dried, glued onto wooden mounts and sanded. Annual radial growth (ring width, RW) were measured with a LINTAB measuring device (Rinntech, Heidelberg, Germany) coupled to a stereomicroscope, with an accuracy of 0.01 mm. Individual ring series were visually and statistically cross-dated with TSAP software (Rinntech, Heidelberg, Germany), using the statistics Gleichläufigkeit (GLK), t-value and the crossdating index (CDI). Validation of the cross-dating was done using COFECHA software (Holmes 1983).

The growth trends were analyzed at different time scales. To study the response of growth to the inter-annual variability of climate (short-term response) we used pre-whitened residual chronologies calculated from ratios between raw growth measurements and individual cubic splines with a 50 % frequency cutoff at 30 years (Fritts 1976). Tree-ring width series were standardized and detrended using the dplR (Bunn 2008, 2010) package. Residual site chronologies were obtained computing the biweight robust mean of all prewhitened growth indices for the trees of the same site (Cook and Kairukstis 1990). The statistical quality of each chronology was checked via expressed population signal (Wigley et al. 1984). A threshold value of EPS > 0.85 was used to determine the cutoff year of the time span that could be considered reliable.

The long-term growth response was analyzed using basal area increment (hereafter BAI, ). BAI represents a more accurate indicator of growth than ring-width, since it removes variation in growth attributable to increasing stem circumference (Biondi and Qeadan 2008). We used measured dbh and raw nondetrened ring-widths to compute BAI by subtracting twice the annual ring width from the annual diameter, starting from the measured diameter outside the bark (Piovesa et al. 2008). We used the following equation:

where is the radius of the tree and is the year of tree-ring formation. We calculated a mean BAI serie for each individual tree. Site BAI chronologies were obtained by averaging individual tree BAI time series.

Statistical descriptive parameters (Fritts 1976) including the mean, standard deviation, first-order autocorrelation of raw series, the mean sensitivity (a measure of the year-to-year variability) and the mean correlation between individual series of residual ring-width indices, were also calculated for each site chronology.

una frase conectora We also explored the climate-growth relationships and the disturbance …

### Climate and growth

Climate data were obtained from the European Daily High-Resolution Observational Gridded Dataset (E-OBS v16) (Haylock et al. 2008). Monthly precipitation and mean, minimum and maximum temperatures were obtained at a 0.25 x 0.25 º resolution for the 1950-2016 period. We selected grid cells covering each sampled sites. Data were downloaded and preprocess using the climate4R bundle (<http://www.meteo.unican.es/en/climate4R>) (Cofiño et al. 2018, Frías et al. 2018). We also used the Standardized Precipitation-Evapotranspiration Index (SPEI), a multiscalar drought index that incorporates both precipitation and temperature (Vicente-Serrano et al. 2010, Beguería et al. 2014). SPEI values for the period 1961-2014 were obtained with a spatial resolution of 1.1 km from the Drought indices dataset for Spain database (<http://monitordesequia.csic.es/>), a high resolution database of drought indices for Spain (Vicente-Serrano et al. 2017). We select a temporal scale of 6 months.

The relationships between residual site chronologies and the climatic variables were assessed by a bootstrapped Pearson’s correlation estimate using the treeclim package (Zang and Biondi 2015). The bootstrapped confidence intervals were used to estimate the significance (p < 0.05) of the correlation coefficients.

### Disturbance analyses

Disturbance chronologies were built using tree-ring width to identify abrupt and sustained increases (releases) or decreases (supressions) in radial growth (Nowacki and Abrams 1997). By using a temporal window of reasonable length, this approach filters out the response to short-term changes in temperature and precipitation and minimizes the long-term growth trends (Nowacki and Abrams 1997, Fraver and White 2005).

Growth changes (GC) were calculated for the individual tree-ring series using a 10-year running window as either positive (PGC) or negative (NGC) growth changes:

where is preceding 10-year median and is subsequent 10-year median. We used medians since they are mores robust estimator of central tendency than means (Rubino and McCarthy 2004, Camarero et al. 2011). Site disturbance chronologies were constructed by averaging the individual disturbances series annually. A minimum of 25% was established to depict canopy disturbances, and we considered a stand-wise disturbance when more than 50% of the individual trees display growth changes above 25 %.

## Resilience

To evaluate the effects of the drougth events on greeennes and tree growth (BAI) we used three resilience indices proposed by Lloret et al. (2011): resilience (*Rs*), resistance (*Rt*), and recovery (*Rc*). The resistance index (*Rt*) quantifies the severity of the impact of the disturbance in the year it occurred. It is estimated as the ratio between the performance during and befor the disturbance:

Resistance (*Rt*) = Drought / Predrought

The Recovery index (*Rc*) is the ability to recover from disturbance relative to its severity, and it is estimated as the ratio between performance after and during disturbance:

Recovery (*Rc*) = Postdrought / Drought

The Resilience index (*Rs*) is the capacity to reach pre-disturbance performance levels, and it is estimated as the ratio between the performance after and before disturbance:

Resilience (*Rs*) = Postdrought / Predrought

We computed the values of these indices for tree growth (BAI) and greenness during each drought event. We considered 2005 and 2012 as singles drought events. The predrought and postdrought values of each target variable (*i.e.*: BAI or EVI) were computed as the mean value during a period of three years before and after the drougth event respectively. A period of three years was chosen because we found similar results comparing periods of two, three and four years (DUDA: incluir tabla de coeficientes y/o gráfica?? como suplement, see Gazol 2017).

## Statistical analysis

We tested for significant differences between drought events (2005 and 2012) and oak population (northern and southern slopes) for each of the resilience indices. Robust two-way ANOVAs were used beacuse original and log-transformed data both did not match the assumptions of normality and homogeneity of variance (Wilcox 2012). Robust measures of central tendency (M-estimator based on Huber’s Psi) were used since they were close to mean value in all cases (Wilcox 2012). When running the robust ANOVA test, data were boostrapped 3000 times and trimmed automatically to control the potential influence of outliers (Field et al. 2012, Wilcox 2012). Post-hoc differences were assessed pairwise using a similar boostrap test. All the robust ANOVA and post-hoc tests were carried out using the WRS2 (Mair et al. 2017) and rcompanion (Mangiafico 2017) R packages. The level of significance was set at 0.05 and adjusted for multiple comparisons. All analyses were carried out in R software (R Core Team 2017) using several packages (see Supplementary S)

# Results

## Vegetation Greenness

Trends analysis showed that 78.9 % of the pixels of *Q. pyrenaica* forests experienced an EVI positive trend for the 2000-2016 period, of which 31.67 % were significant trends. The strongest trends were observed in southwestern populations (Figura S2). Vegetation greenness of *Quercus pyrenaica* forests were lower during the 2005 and 2012 year than the greenness observed for the reference period (Figure 2a). The lowest values for EVI standardized anomalies were recorded in 2005 being singnificantly lower (-2.285 ± 0.029) than 2012 (-0.418 ± 0.029) (LSMEANS, t.ratio = -45.358; p\_value < 0.0001), particularly for northern populations (Figure 2b). Reduction in annual EVI mean was considerably higher in northern populations than in southern ones during the 2005 drought (Figure 2b).

According to the standardized anomalies, *Q. pyrenaica* forests suffered a browning during 2005 drought event (99.36 % and 79.37 % of the pixels for northern and southern populations respectively), however no changes were observed for the 2012 drought (Figure S3).

## Responses to drougth: greenness

*Q. pyrenaica* forest showed significantly lower resistance to 2005 drought event than to 2012 one [2005: 0.858 (0.853-0.863); 2012: 0.943 (0.939 - 0.947); table 2; Figure 3a, F = 799.86, p <0.0001]. The 2005 drought reduced the greenness of oak to 85.8 % while the 2012 reduced 94.3 % (tabla rs1 quizás, duda en esto). Southern populations showed significantly higher values of resistance to drought than northern ones, except for 2012 where non-significant differences were recorded (Figure 3a).

The oak forests recovered their greenness more rapidly after the 2005 drought than after 2012. In the period after 2005 drought, greeness achieved was 112 % (Rc = 1.12) and after 2012 105.7 % (Rc = 1.057). A similar recovery after the 2005 and 2012 drought event was observed for southern populations (p = 0.2453; Figure 3a), whilst the northern populations showed a significantly greater recovery after the 2005 drought than after the 2012 drought.

Resilience values were significantly higher for the 2012 drought event than for 2005, although both values were close to 1 indicating that greenness level was rather similar after each disturbance event. The southern populations showed higher resilience values than the northern ones, although they were not significantly different for 2005 drought (p = 0.036).

Figura R2. Interaction plot resilience metrics

## Responses to drougth: tree-growth (BAI)

Although no significant differences were observed in the resistance of the oak growth to both drought events (F = 6.0189; p = 0.019; table 2; Figure 3b), the 2005 drought reduced growth more than 2012 did (Rt = 0.721 and 0.819 respectively). Northern population showed resistance values lower than southern ones, especially for 2005 drought event where the growth was reduced to 44.5 % of the preceding period.

Oak forests recovered their growth more rapidly after 2012 than after 2005 staying at levels of Rc < 1. Norhtern populations showed signigicantly higher values of recovery than southern ones, which even not reach recovery values of Rc = 1.

Significantly higher values of resilience of growth were observerd for the 2012 drought event than for the 2005 one. We found no differences of resilience values for growth between northern and southern populations (F = 1.31; p = 0.404). For 2005 drought event the resilience of northern population was lower than southern ones, but opposite pattern was found for the 2012 drought event, with northern population showing values greater than unity.

# Climate and tree

Tree-growth exhibited a significant response to water availability more than to the temperatures. Cumulative precipitation of the hydrological year (September of the previous year to August of the growth year) influenced positively the tree-growth for the both northern and southern populations (FIG ). Precipitation of previous december was also positively correlated with tree growth in the northern population and in the highest location of the southern population. Hydrological, Spring and Summer SPEI showed a strong positive correlation with tree-growth, specially for northern population. Spring Maximum temperature was the most significant limitation of the tree-growth for the southern population while maximum temperatures of the previous August did for the northern one. Conversely, mean, minimum and maximum temperatures of current September positively influenced the tree-growth in the northern population.

# Disturbance

En las poblaciones del norte osbervamos dos eventos de PGC por encima del 25 e incluso el 50 %. El primero de ellos desde 1940 hasta 1950, mientras que el segundo entre 1994 y 2001. Estos periodos se alternan con periodos de NGC: previo 1940, 1960 - 1991, y quizá aunque no lo vemos en al actualidad (?? duda. Esto es especular un poco). Esto responde creo que a:

1. actividad minera y tratamientos forestales. La actividad minera en la zona fue intensa (aunque a veces intermitente, entre 1890 y 1957 (cuando se clausuró); y previamente desde 1858 tras los trabajos de Amalio Maestre). Hay costancia de la existencia de hornos de fundición de mineral en alguna de las minas (La Probadora).
2. actividades forestales:

* Se conoce la existencia de un proyecto redactado en 1943 de repoblación forestal de la cabecera del río Genil (he pedido los mapas).
* Ver actividades forestales (bd proyectos de gestión)

En el trabajo de (Gea-Izquierdo and Cañellas 2014) se observa una liberación posterior a 1950 (QPUY9) (entre 1950 y 1960 aprox.) y luego una supresión hasta mediados de 1990. Esto coincide con lo que hemos observado en nuestra chrono.

Respecto a las poblaciones del sur, destaca en la caH una primera liberación en torno a 1829-1830 y 1837-1840; y luego predominan los eventos de supresión que se alternan con algunos de liberación (muy débiles) hastas 1950. Posteriormente se alternan liberaciones-supresiones débiles. Destaca una liberación en torno a 2000-2003 (relación con clima??? posterior a sequía 1999???)

En caL, observamos un periodod de limitación hasta 1885 aprox. Luego se alternan al igual que antes periodos de liberación con otos de supresión con señales más fuertes que en caL, y al igual que antes destaca con mas intensidad si cabe, una liberación entre 2000 y 2003 (antrópico o climática ??)

Duda: ocurre en los dos sitios caL y caH, entonces, es aumento de crecimiento debido a actuación forestal? o debido a clima??

(Gea-Izquierdo and Cañellas 2014) encontró un pico de disturbance en torno a 1860 en QUPY10 (cañar) sugiriendo que la baja densidad en esos sitios fue presumible el resultado de un thinning (aclareo). El patrón de supresión que observa para este sitio va en la linea de lo que nosotros encontramos (supresiones que alcanzan a veces el 25 % y otras no)

En cuanto a la señal a nivel de sitio, para el Norte, vemos que tanto las supresiones como las liberaciones se observan para mas del 50 % de los árboles. En el caso del sur, vemos que algunas supresiones se observan para mas del 50 % de los árboles, mientras que las liberaciones observadas en torno a los 2000 se observan para mas del 50 % de los árboles

notas para discusión

En ambos casos (evi y bai) la sequía de 2005 y de 2012 redujeron el greenness y el crecimiento. Además en ambos casos la sequía de 2005 redujo mas la variable, sin embargo, para el greenness si existen diferencias entre ambos eventos de sequía, mientras que para el crecimiento no. Por tanto podemos decir que se ha reducido el greennes pero para el crecimiento esto está mas amortiguado.

Al explorar la resistencia por sitio y drought se observa una curiosidad. La resistencia de los robledales de CaLow se mantiene casi igual. La de la cara norte (SJ) sufre mucho en 2005 y luego se recupera un poco. La de caH presenta menor resistencia en 2012 que en 2005 (y eso llama la atención)

## References

# Discussions notes

## Greenness

* Aproximadamete el x% de los robledales mostraron un incremento en los valores de productividad (greenees) … (relacionar con tendencias EVI y ontologias).
* tendencias ontologias
* Si analizamos el plot de trajectorias (evi medio anual), vemos como en 2005 y 2012 se observó un browning para los valores medios de EVI. En ambos casos, se observó un patrón homogéneo de browing, aunque en 2012 el browning fue mucho menor.

## Dos cronos en Cañar

* Gea-Izquierdo and Cañellas (2014) muestrea solo en CA-Low (QUPY10) y obtiene resultados similares a CA-Low.
* En tan poco espacio hay diferencias enormes entre las dos cronos
* *Las tendencias de BAI de Cáñar baja altitud son diferentes a las del norte, las cuales se parecen mucho a las que suelen encontrarse en la mayoría de sitios. Puedes mirar por ejemplo las tendencias de las líneas negras (de más de 100 años, sobre todo) en la Fig 2 del artículo de Ecosystems… fíjate que QUPY3 se parece más a QUPY10… y QUPY3 era la única dehesa (en parte) y uno de los sitios más secos (por otro lado, aunque frío). En fin, todo esto es algo especulativo, pero sirva como exploración de tus cronos, y para mostrar que me parece muy interesante la diferencia entre Cáñar alto y bajo (y nuevo).*

## Reducción crecimiento y sequías

Se observó una gran reducción de crecimiento en 1994-1995, 1999 y 2005 para Pinus nigra en el SE de Andalusia (Sánchez-Salguero et al. 2013) coincidiendo con años de sequía para esa zona. La máxima reducción en crecimiento fue observada para 2005, siendo significativamente mayor que para el resto de años (Sánchez-Salguero et al. 2013).

## Relaciones crecimiento-clima en otras especies

## References

## Table 1. Dendro summary statistics

## Table 2. Robust ANOVAS (F-value)

## Table 2

|  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- |
|  | rc F | rc p | rs F | rs p | rt F | rt p | variable |
| Disturb | 29.55 | 0.001 | 44.31 | 0.001 | 6.02 | 0.019 | BAI |
| Site | 53.08 | 0.001 | 1.31 | 0.534 | 59.25 | 0.001 | BAI |
| Disturb X Site | 4.4 | 0.134 | 30.01 | 0.001 | 32.24 | 0.001 | BAI |
| Disturb | 311.99 | 0.001 | 207.18 | 0.001 | 799.87 | 0.001 | EVI |
| Site | 105.41 | 0.001 | 29.82 | 0.001 | 153.22 | 0.001 | EVI |
| Disturb X Site | 364.31 | 0.001 | 6.14 | 0.014 | 234.7 | 0.001 | EVI |

### Figure 1

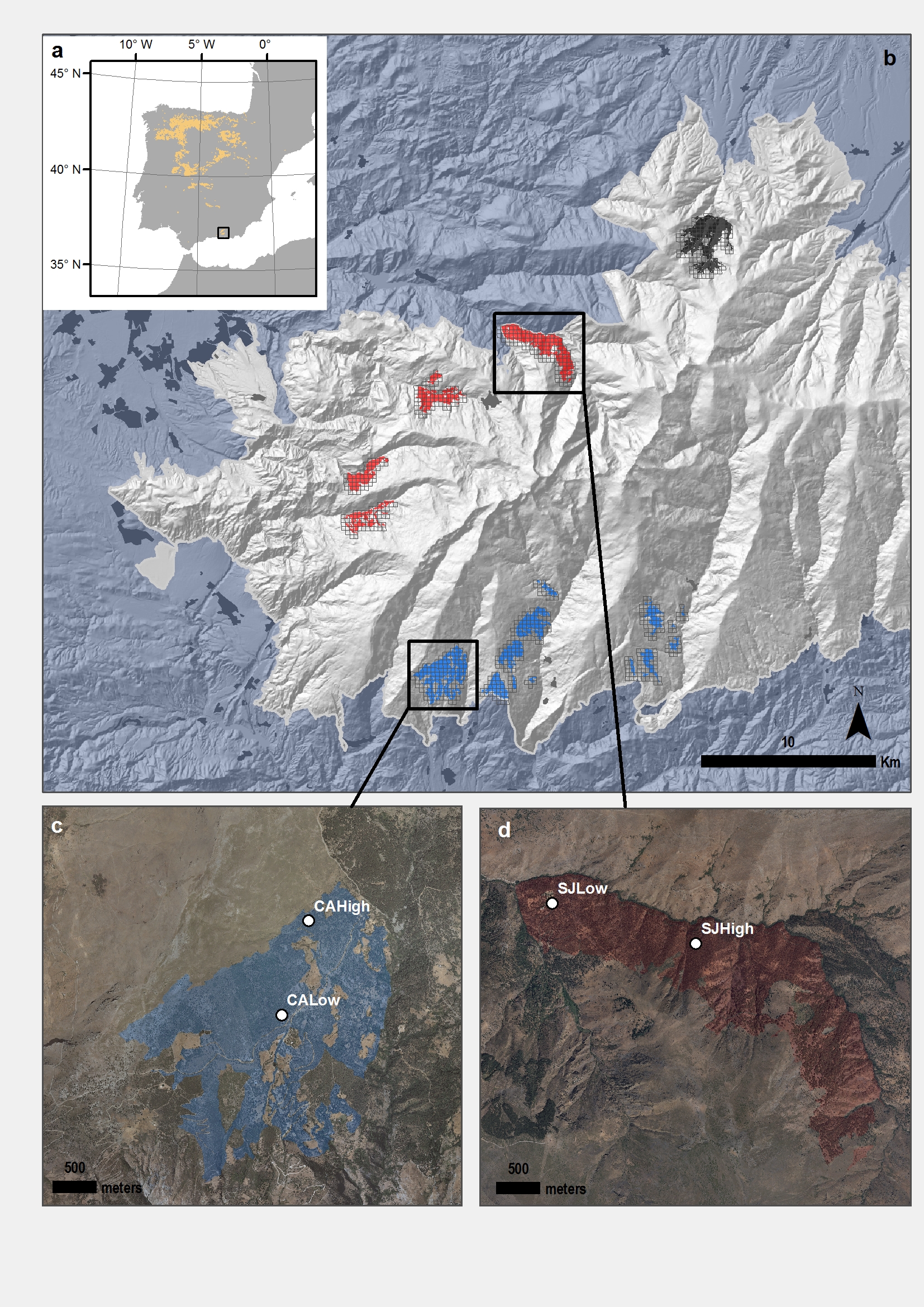


Figure 1. Distribution of *Quercus pyrenaica* forests in Iberian Peninsula (a) and in Sierra Nevada mountain range, where three cluster of oak populations have been identified (Pérez-Luque et al. 2015) (showed in different colour) (b). A grid of with the MODIS pixels for each population is shown (see material and methods). Detailed location of the sampling sites: northern (San Juan, SJ) (c) and southern ones (Cáñar: CALow and CAHigh)(d). Colour Orthophotography of 2009 from Regional Ministry of the Environment, Regional Government of Andalusia.

### Figure 2

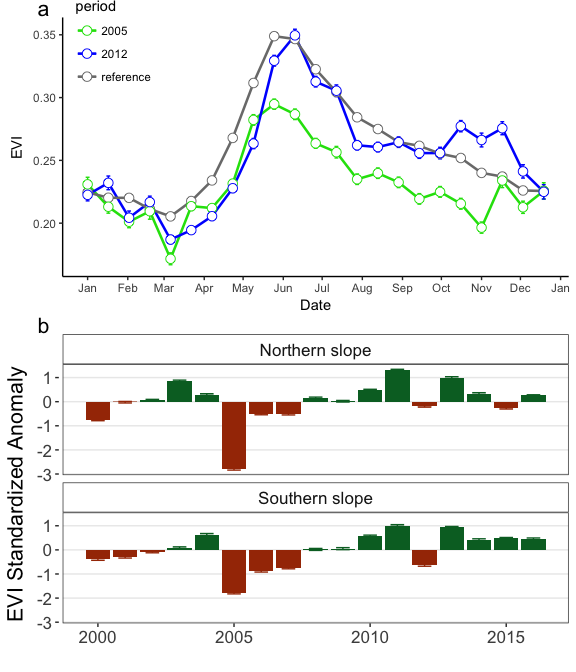


Figure 2. Comparison of EVI profile for the reference period (*gray*) and during the 2005 (*green*) and 2012 (*blue*) drought events (a). EVI standardized anomaly () during the period 2000-2016 for northern and southern populations (b).

### Figure 3

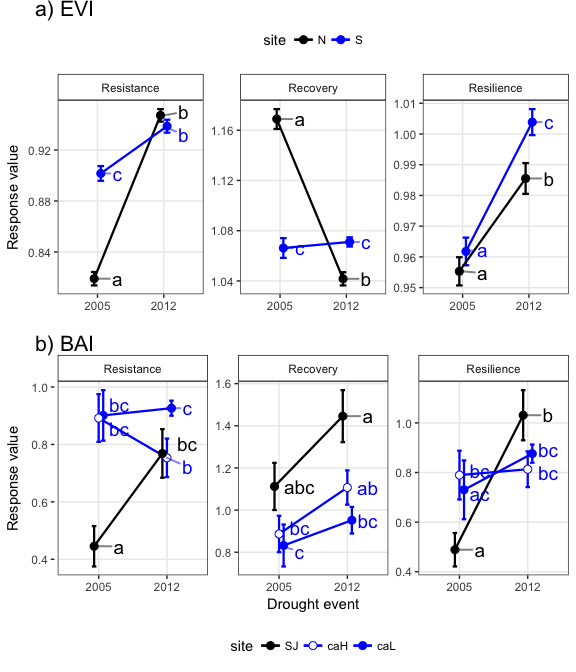
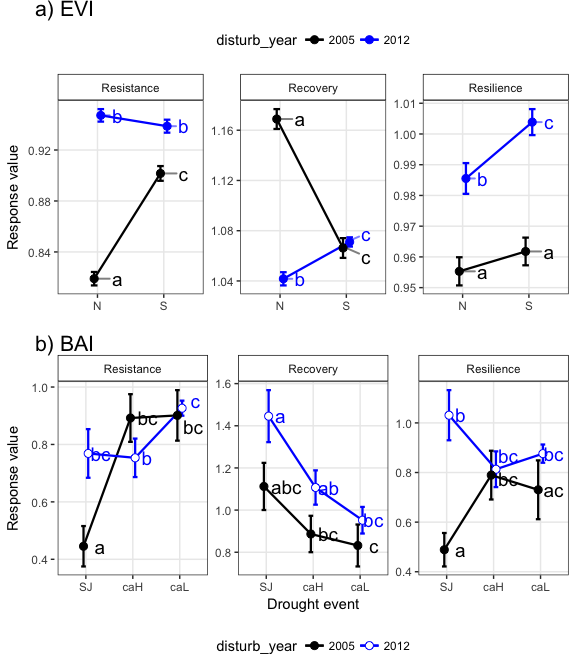


Figure 3. Response of northern (*black*) and southern (*blue*) populations of *Q. pyrenaica* forests to drougth in terms of resistance, recovery and resilience of greenness (EVI; a) and tree growth (BAI; b) for the years 2005 and 2012.

### Figure 4



# Supplementary

## Supplementary 1

Listado de pkgs usados

## Suplementary 2

## Figure S2

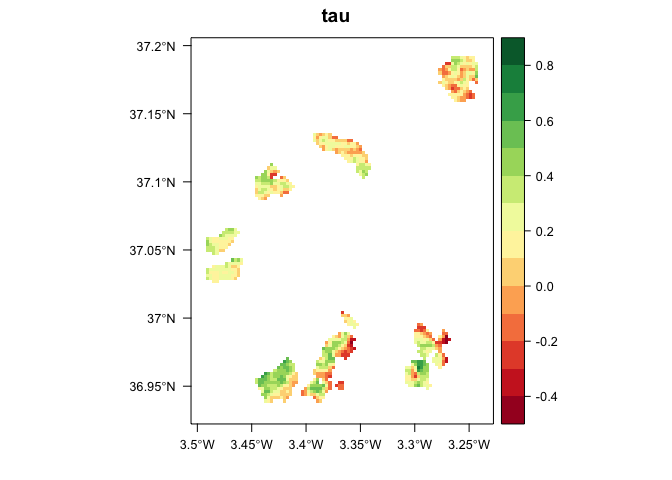


Figure S2. Spatial distribution of the EVI-annual trends for the period 2000-2016. Trends () with the most positive values (green) indicate positive trend.

## Figure S3

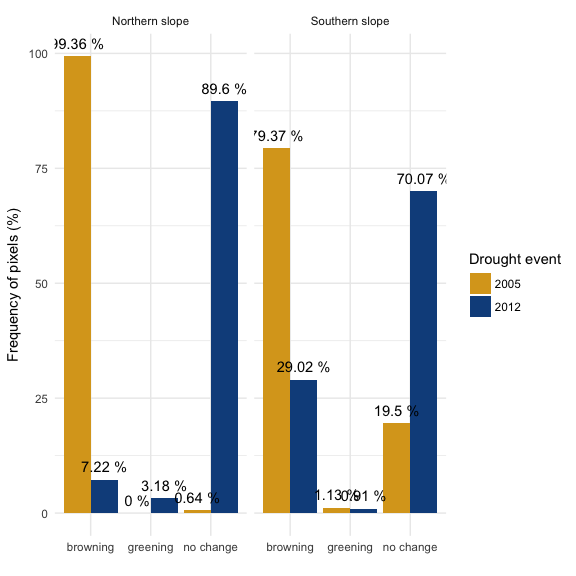


Figure S3. Percentage of pixels showing browning, greenning or no-changes during the 2005 and 2012 droguht events according to EVI standardized anomalies.

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