#### The Statistical Mechanics of Biodiversity

by

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#### Abstract

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Invasive brag; forbearance.

I dedicate this thesis			
my Ph.D: to my mot Linden Schneider;	her Judy Rominger and to the kolea (P		

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## Introduction

 $\operatorname{stub}$ 

## Chapter 1

# Punctuated non-equilibrium and niche conservatism explain biodiversity fluctuations through the Phanerozoic

#### 1.1 Background

Biodiversity has not remained constant nor followed a simple trajectory through geologic time (Raup et al., 1982; Sepkoski, 1984; Gilinsky, 1994; Liow & Stenseth, 2007; Alroy et al., 2008; Alroy, 2010). Instead, it has been marked by fluctuations in the number of extant taxa, both positive in the case of net origination or negative in the case of net extinction. Major events, such as adaptive radiations and mass extinctions have received special attention (Benton, 1995; Erwin, 1998), but fluctuations of all sizes are ubiquitous (Sepkoski, 1984; Alroy et al., 2008; Quental & Marshall, 2013).

Several approaches have been taken to study the complex trajectory of paleo-biodiversity ranging from the hypothesis that biological systems self-organize to the brink of critical phase-transitions (Bak & Sneppen, 1993; Solé et al., 1997) to invocations of non-linear environmental perturbations (Newman & Roberts, 1995) and escalatory co-evolutionary interactions (Vermeij, 1987). New data and analyses have not supported any of these hypotheses at the scale of the entire Phanerozoic marine invertebrate fauna (Kirchner & Weil, 1998; Madin et al., 2006; Alroy et al., 2008). Other studies have modeled the mean trend in diversity as tracking a potentially evolving equilibrium (Sepkoski, 1984; Alroy et al., 2008; Alroy, 2010; Rabosky, 2009) and yet ignore the potential role of stochasticity and non-equilibrium dynamics in producing observed patterns (Erwin, 2012; Liow & Stenseth, 2007; Quental & Marshall, 2013). As such, we still lack a synthetic theory of evolving biodiversity through the fossil record. Here we use a simple model of evolution in an abstract niche space derived from universal non-equilibrium processes to predict, with great accuracy, the complex distribution of pervasive diversity fluctuations throughout the marine Phanerozoic.

Despite the heterogeneity of explanations of Phanerozoic biodiversity, consensus has

emerged on three properties of macroevolution: *i*) gross ecological and life history attributes of clades (i.e. groups of related species descending from a common ancestor) are often maintained, a phenomenon known as niche conservatism (Roy et al., 2009; Hopkins et al., 2014); *ii*) long periods of niche conservatism are interrupted by adaptive diversification and exploration of new ecological niche space (Eldredge & Gould, 1972; Newman et al., 1985; Hopkins et al., 2014); and *iii*) as a consequence of the interaction between their life history characteristics and the dynamics of the environments they inhabit (Vrba, 1983) different clades experience different rates of morphological evolution, speciation and extinction (Simpson, 1953; Sepkoski, 1984; Holman, 1989; Gilinsky, 1994).

Observed bursts of adaptive radiation leading to novel morphologies in the fossil record led Eldredge and Gould to their hypothesis of punctuated equilibrium (Eldredge & Gould, 1972). Here we show that this punctuation is actually akin to the "super statistical theory" of non-equilibrium dynamics in statistical physics (Beck & Cohen, 2003). Super-statistics (Beck & Cohen, 2003) proposes that non-equilibrial systems can be decomposed into locally equilibrial sub-systems. The distribution of equilibria across sub-systems determines the dynamics of the complete system (Beck & Cohen, 2003). When these sub-systems are superimposed the resulting system can no longer be described by a single equilibrial model. In the context of macroevolution we propose that a clade with conserved life history characteristics corresponds to a locally equilibrial sub-system. If a certain region of niche space can only contain a finite diversity of taxa (Simpson, 1953; Gavrilets & Vose, 2005; Rabosky, 2009; Price et al., 2014) then diversity within clades should fluctuate stochastically about this equilibrium due to random origination and extinction. The magnitude of these macroevolutionary rates should be a function of the life history and ecological characteristics that define that region of niche space. Larval type (Jablonski, 2008), body plan (Erwin, 2012), body size (Harnik, 2011), range size (Harnik, 2011; Foote et al., 2008) and substrate preference (Hopkins et al., 2014) have all been shown to influence such rates. Thus different regions of niche space, and the clades occupying them, will experience different magnitudes of stochastic fluctuation in diversity. As clades occasionally split to fill new regions of niche space their punctuated diversification determines the non-equilibrium nature of the entire biota. Here we show that these properties of macroevolution are sufficient to explain the complex fluctuations of marine invertebrate diversity through the Phanerozoic.

In statistical mechanics, local sub-systems can be defined by a simple statistical parameter  $\beta$  often corresponding to inverse temperature. In the context of macroevolution we define  $\beta$  as the inverse variance of a homogenous origination-extinction process, which will capture all relevant information about a clade's diversification under such a process. The limit distribution of time averaged fluctuations in clade k's diversity through time should be approximately Gaussian with variance  $1/\beta_k$  (supplement section ??). We posit that just as in statistical systems, non-equilibrium dynamics can arise from the mixing of the dynamics of many locally equilibrial subsystems. For marine Phanerozoic diversity this corresponds to mixing the dynamics of many clades, all being described by their unique  $\beta_k$  values. Three exemplar dynamics taken from a bias-corrected (see methods section) aggregation of the Paleobiology Database (PBDB) (Alroy et al., 2008) are shown in Figure 1.1. To predict

the super-statistical behavior of the entire marine invertebrate Phanerozoic fauna we must integrate over all possible local equilibria that each clade could experience. The distribution of  $\beta_k$  values describes the probability that a given clade, chosen at random, will occupy a region of niche space characterized by that inverse variance value. The form of this stationary distribution of  $\beta_k$  values could shed interesting light on the biological processes that lead different clades to different equilibria, as discussed below. Figure 1.1 shows the shape of this stationary distribution estimated from bias-corrected PBDB (Alroy et al., 2008) data.

To uncover the super-statistical nature of the marine invertebrate Phanerozoic fauna we analyze the distribution of diversity fluctuations using two canonical databased of fossil biodiversity, the PBDB (Alroy et al., 2008) and Sepkoski's compendium (Sepkoski, 1992) of fossil marine invertebrates (results from Sepkoski's compendium are presented in supplement section ??). We filter PBDB data to include only well preserved marine invertebrates following previously published collection inclusion criteria (Alroy et al., 2008; Alroy, 2010). We account for detection bias in the PBDB using an extension of the "three timer" correction (Alroy et al., 2008). "Three timer" correction accounts for the rate of failure to observe a genus, estimated by the number of times a gap occurs in its occurrence history. We extend this correction by also employing a new publication bias correction to help eliminate bias from preferential publication of novel taxa (see methods section). Results obtained from this correction strategy are similar to other published methods (Fig. ??). Fluctuations within a clade are computed as the difference in standing diversity between two time intervals, or equivalently the number of originations minus extinctions in one interval.

Phanerozoic biodiversity can be deconstructed and grouped into clades, or sub-systems, in several different ways. Lacking a full phylogenetic hypothesis for all marine invertebrates we use taxonomic classifications to identify potential sub-systems. Taxa ideally represent groups of organisms that descend from a common ancestor and share similar ecologically and evolutionary relevant morphological traits (Mayr, 1965; Erwin, 2007). For Phanerozoic marine invertebrates, Holman (Holman, 1989) has shown that variance in diversity dynamics is less between taxa belonging to the same order than taxa in different orders, indicating that the taxonomic level of orders is a likely candidate for sub-system delineation. To evaluate the optimal taxonomic level for sub-system designation, we test our superstatistical theory using taxonomic levels from order through class to phylum. Additionally we compare our results to randomized taxonomies to confirm that observed patterns are not an artifact of arbitrary classification but instead represent real biologically relevant differences between clades.

#### 1.2 Methods

#### Paleobiology Database data download and filtering

Data were downloaded from the Paleobiology Database (PBDB; www.pbdb.org) on 28 May 2013. Collections were filtered using the same approach as Alroy (Alroy *et al.*, 2008) to

insure that only well preserved marine invertebrate occurrences were used in subsequent analysis resulting in 221202 genus occurrences. These were further filtered to exclude those occurrences without order-level taxonomy and those collections with age estimate resolutions outside the 10my default bins of the PBDB resulting in 189516 occurrences left for analysis. To avoid basic sampling concerns we excluded the earliest Cambrian and the Cenozoic.

To focus attention on the variance of fluctuations we center each clade's fluctuation distribution. Because "equilibrium" in the statistical mechanical sense means a system undergoes coherent, concerted responses to perturbation the mean trend line is of less interest than deviations from it. We also note that most clades are already close to centered and so centering has little influence on clade-specific fluctuation distributions.

#### Three-timer and publication bias correction

We use a new and flexible method, described below, to correct for known sampling biases in publication-based specimen databases (Alroy et al., 2008; Alroy, 2010). We were motivated to use this method because rarefaction has been shown to under-perform compared to the more data-intensive shareholder quorum subsampling (SQS) method (Alroy, 2010). However, subsampling cannot be applied to small orders (i.e. the majority) because SQS becomes increasingly unreliable as sample size decreases (Alroy, 2010). We therefore develop a simple method based on first correcting for detection bias using the "three timer" correction (Alroy et al., 2008) in which the rate of failure to observe a genus is estimated by the number of times a gap occurs in the occurrence history of each genus. To eliminate further bias due to preferential publication of novel taxa we divide observed number of genera per order per time period by the expected number of genera given publications in that time period. The expected number is calculated by regressing the log-transformed number of genera on log-transformed number of publications. There is a weak trend toward higher diversity with more publications (Fig. ??) meaning that the most important correction comes from the three timer correction.

Our new method effectively re-scales each genus occurrence from 0/1 (absent/present), to a weighted number continuously ranging between 0 and 1. This method achieves similar results to more computationally intensive sub-sampling procedures (Alroy et al., 2008; Alroy, 2010). We directly compare our predicted time series of global genus diversity with results derived from SQS (Alroy, 2010) and the raw data (Fig. ??). Our method shows minor differences with the SQS prediction, However, these discrepancies do not have impact the distribution of fluctuations (Fig. ??) and super-statistical analysis on uncorrected PBDB data (see section ??) produces a similar result to the analysis on corrected PBDB data presented in the main text.

#### Numerical methods

To fit our super-statistical prediction we use the method of least squares instead and maximum likelihood. When building the prediction for P(x) by calculating order-level Gaussian

distributions and integrating over them, we use least squares to fit the variance term to each order. We do so because orders potentially show asymmetries in their distribution of fluctuations. Least squares is more flexible in fitting such distributions compared to maximum likelihood which will always estimate the empirical variance as the best-fitting parameters.

We also estimate P(x) directly from the raw data using maximum likelihood to compare the fit of our super-statistical prediction and that of a simple Gaussian distribution using AIC. To calculate a likelihood-based confidence interval on our prediction we bootstrapped the data, subsampling fluctuations with replacement from all orders combined.

#### 1.3 Results and Discussion

At the order level in both the sampling-corrected PBDB (Fig. 1.1) and Sepkoski's compendium (supplement section ??), fluctuations in genus diversity are well described by a Gaussian distribution (Fig. 1.1). Gaussian fluctuations would result from a homogeneous origination-extinction process under the condition of independence between orders. Independence could result from neutral-like processes (Hubbell, 2001), where the dynamics of one taxon are unaffected by those of another, or from dampening mechanisms that stabilize complex networks of interacting taxa (Brose et al., 2005). This is in direct contrast to the instability hypothesis underlying the self-organized criticality theory of paleo-biodiversity (Bak & Sneppen, 1993; Solé et al., 1997).

We estimate the distribution of  $\beta_k$ 's simply as the maximum likelihood distribution describing the set of inverse variances for all orders. In both the PBDB and Sepkoski's compendium, Phanerozoic marine invertebrate orders clearly follow a Gamma distribution in their  $\beta_k$  values (Fig. 1.1). While multiple processes could lead to a Gamma stationary distribution (e.g. (Cox et al., 1985)), one interesting possibility is a mean-reversion process (Cox et al., 1985). Mean reversion could be a consequence of niche conservatism if  $\beta_k$  values are associated with a clade's physiological and life history traits, themselves evolving via Ornstein-Uhlenbeck-like exploration of an adaptive landscape (Cox et al., 1985; Butler & King, 2004).

Using the observation of within order statistical equilibrium and Gamma-distributed  $\beta_k$  parameters we can calculate, without further adjusting free parameters, the distributions of order-level fluctuations for the entire marine Phanerozoic, P(x), as

$$P(x) = \int_0^\infty p_k(x \mid \beta) f(\beta) d\beta \tag{1.1}$$

where  $p_k(x \mid \beta)$  is the distribution of fluctuations within an order and  $f(\beta)$  is the stationary distribution of inverse variance in the magnitude of order-level fluctuations in diversity. This leads to a non-Gaussian, fat-tailed prediction for P(x) which matches both the PBDB and Sepkoski data closely (Fig. 1.2 and supplement section ??).

To quantitatively evaluate how well the super-statistical prediction matches the data we constructed a 95% confidence envelope from bootstrapped maximum likelihood estimates for

P(x). Observed fluctuations fall within this 95% confidence envelope (Fig. 1.2), indicating that the data do not reject the super-statistical prediction. For further comparison, we fit a Gaussian distribution to the observed fluctuations, which corresponds to the equilibrium hypothesis that all orders conform to the same statistics. Using Akaike Information Criterion (AIC) we find that observed fluctuations are considerably better explained by the super-statistical prediction than by the Gaussian hypothesis ( $\Delta$ AIC = 11285.18). Thus, as expected under the super-statistical hypothesis, the fat tailed distribution of fluctuations arise from the superposition of independent normal statistics for fluctuations within orders.

Computing the distribution of fluctuations using classes instead of orders leads to a substantially poorer fit to the observed data (Fig. ??). We quantify this shift with the Kolmogorov-Smirnov statistic, which changes from 0.041 in order to 0.062 in classes (Fig. 1.3). However, if super-statistical theory explains the data, this worsening fit with increasing taxonomic scale is expected as the different classes are not well defined subsystems in their fluctuation dynamics. Instead, classes aggregate increasingly disparate groups of organisms, and thus effectively mix their associated Gaussian fluctuations, meaning that one statistic should no longer be sufficient to describe class-level dynamics. Our analysis indicates that orders are evolutionarily coherent and independent entities, with all subsumed taxa sharing key ecological and evolutionary attributes allowing them to diversify concertedly and independently from other orders. Both the good fit at the order level and worsening fit at higher taxonomic levels is confirmed in Sepkoski's compendium, which also allows analysis of phylum-level patterns (Fig. ??).

To further test the evolutionary coherence of orders we conducted a permutation experiment in which genera were randomly reassigned to orders while maintaining the number of genera in each order. For each permutation, we calculated the super-statistical prediction and the Kolmogorov-Smirnov statistic. The permutation simulates a null model in which common evolutionary history is stripped away (genera are placed in random orders) but the total number of observed genera per order is held constant. Repeating this permutation 500 times yields a null distribution of Kolmogorov-Smirnov statistics that is far separated from the observed value (Fig. 1.3) suggesting the good fit at the order level is not merely a statistical artifact of classification but carries important biological information.

#### 1.4 Conclusion

Our analysis makes no assumption that orders should correspond to super-statistical subsystems, but identifies them as the appropriate level for marine invertebrates. Holman (Holman, 1989) has also shown that orders are "evolutionarily coherent" in that subtaxa within orders share common diversification dynamics. As we show, orders differ only in the variances of their diversity fluctuations (Fig. 1.1).

Our study is the first to demonstrate that complex patterns in the sequence of origination and extinction events in the fossil record are the result of a simple underlying process analogous to the statistical mechanisms by which complexity emerges in large physical (Beck, 2004) and social systems (Fuentes et al., 2009). We do so by identifying the biological scale at which clades conform to equilibrial dynamics, which could result from the process of niche conservatism. We then show that punctuated shifts to different equilibria between clades, a consequence of punctuated exploration of niche space by newly evolving clades, leads to a characteristically non-equilibrial distribution of diversity fluctuations when the marine Phanerozoic fauna is viewed as a whole macro-system.

Our work highlights the importance of both niche conservatism and punctuated adaptive radiation in producing the statistical behavior of the Phanerozoic; our theory thus provides new motivation for identifying the eco-evolutionary causes of innovations between lineages and how those innovations are eventually conserved within lineages. Armed with an understanding of the statistical behavior of diversification we can go on to examine mechanisms underlying additional patterns in the mean trend of biodiversity through the Phanerozoic. In particular, clades have been shown to wax and wane systematically through time (Liow & Stenseth, 2007; Quental & Marshall, 2013), a pattern that we cannot explain with superstatistics alone.

Super-statistics could also be applied to other areas of evolution and macroecology. For example new phylogenetic models already consider heterogeneous rates of diversification (e.g. Rabosky, 2006). The super-statistics of clades in adaptive landscapes could provide a means to build efficient models that jointly predict morphological change and diversification. This framework could also provide a new paradigm in modeling the distributions of diversity, abundance and resource use in non-neutral communities. Non-neutral models in ecology are criticized for their over-parameterization (Rosindell et al., 2011), yet a persistent counter argument to neutral theory (Hubbell, 2001) is the unrealistic assumption of ecological equivalency (Chave, 2004) and poor prediction of real dynamics (Ricklefs, 2006). If ecosystems are viewed as the super-position of many individualistically evolving clades, each exploiting the environment differently and thus obeying a different set of non-equivalent statistics, then diversity dynamics could be parsimoniously predicted while incorporating real biological information on ecological differences between taxa.

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Figure 1.1: The distributions of within-order fluctuations in genus diversity shown for the trajectories of three exemplar orders (A) and shown as an empirical cumulative density aggregated across all orders (B). To display all orders simultaneously we simply collapse their fluctuation distributions by dividing by their standard deviations. If orders conform to the Gaussian hypothesis their scaled fluctuations should fall along the cumulative density line of a normal N(0, 1) distribution, as shown in (B). In (C) the distribution of inverse variances  $\beta_k$  across all orders matches very closely to a Gamma distribution (black line); exemplar orders are again highlighted.

#### **Figures**

Figure 1.2: Distribution of fluctuations in genus diversity within orders of marine invertebrates in the Paleobiology Database (Alroy et al., 2008) after bias correction. The distribution is fat-tailed as compared to the maximum likelihood estimate of the normal distribution (blue line). At the order level the empirical distribution of fluctuations are well described by our super-statistical approach, both when computed from integrating over the distribution of observed variances (red line) and when fit via maximum likelihood (95% confidence interval; red shading).

Figure 1.3: Distribution of Kolmogorov-Smirnov (KS) statistics from randomly permuting genera within orders (gray shading represents 95% confidence interval). Solid black line is observed KS statistic at the order level, while the dashed black line shows the observed KS statistic at the class level.

## Chapter 2

## Community assembly on isolated islands: macroecology meets evolution

#### 2.1 Introduction

Current biodiversity is a product of speciation, extinction and dispersal, contingent on the ecological interactions of organisms with their biotic and abiotic environment. The evolutionary history leading to the assembly of any given ecological community must in some way shape current ecological assemblages. However, because the processes of evolution and ecology occur on different temporal and spatial scales, disentangling the relative influence of local ecological mechanisms from historical evolutionary processes on patterns of community structure remains a central challenge (?).

The evolutionary processes of speciation and extinction are classically viewed as constraints on regional species pools, occurring in a manner largely removed from local ecology (Hubbell, 2001; Cavender-Bares et al., 2009; ?). Conversely, ecological mechanisms tend to be viewed as packing standing diversity into local communities through consumption, competition, facilitation and, more recently, neutral ecological drift (Hubbell, 2001; Tilman, 2004; Bascompte & Jordano, 2007; Borer et al., 2014). While recent theoretical advances have provided greater insight into ecological drift (Hubbell, 2001; Rosindell & Phillimore, 2011), niche partitioning (Tilman, 2004), competition, predation (Borer et al., 2014) and species interaction networks (Williams & Martinez, 2000; Brose et al., 2006), these insights typically do not contain realistic evolutionary assumptions (Ricklefs, 2006) or ignore them entirely.

Insights into the genetic, biogeographic and selective mechanisms leading to diversification have also emerged based on inference from current patterns of species, genetic or phylogenetic diversity (e.g. ??). However, it is not possible to use current static patterns to infer the temporal dynamics of either the evolutionary mechanisms or their ecological consequences, nor can we understand what constitutes meaningful change in a system without a baseline for comparison. Here we show how testing idealized ecological theories—such as the

unified neutral theory (Hubbell, 2001) or the maximum entropy theory of ecology (Harte, 2011)—on archipelagos composed of islands formed in a discrete geological sequence can help identify the shifting balance and feedback between fast-acting, local ecological mechanisms, and longterm, large-scale evolutionary processes in determining ecological community structure. Islands having different ages of formation, along with discrete volcanoes within islands, provide the opportunity to study diversification of species and the assembly of communities in different stages. Ecological theory provides an idealized null baseline against which to compare observed patterns.

#### Hotspot oceanic archipelagos as model systems

Hotspot oceanic islands are opportune model systems for studying the interplay of local ecological mechanisms and the evolutionary drivers of biodiversity patterns. Due to their sequential formation as the tectonic plate moves over a volcanic hotspot, such island systems offer a range of spatial and temporal scales over which to analyze the outcomes of ecological and evolutionary processes (?). While many archipelagos around the world share these biotic and geological properties, the Hawaiian archipelago provides a particularly useful system for study because its linear geological chronology (Price & Clague, 2002), ecosystem developmental trajectories (?) and phylogeographic patterns of biodiversity are each well characterized (?). Moreover, studies of species diversity across the islands have revealed patterns that are non-uniform across the island chronosequence with marked differences among lineages (e.g. Gruner, 2007; Gillespie, 2009) that can be used to test for biologically meaningful differences among lineages that might drive their disparate diversification patterns.

#### Development of genetic structure

High levels of dispersal and associated gene flow among localities limit the extent to which populations can diverge genetically. However, when gene flow is low, distinct populations in different localities are free to diverge through local selective pressures and drift, which can lead to diversification (?) Thus, the magnitude of genetic connectivity among populations provides a measure of the relative importance of dispersal-driven assembly (dictated by processes removed from the local setting) in contrast to assembly by local (in situ) diversification in determining community composition. Using the chronosequence of the Hawaiian archipelago, we can analyze populations from multiple sets of taxa across trophic guilds occurring in geological contexts from young to old. We predict that dispersal-driven (ecological) processes will dominate in community assembly in young habitats, with the importance of in situ (evolutionary) processes increasing with habitat age. If evolutionary processes are not important, we predict that communities should reach a statistical steady state through ecological processes alone (Harte, 2011). If, as we expect, evolutionary processes become increasingly important in community assembly over time, we would expect to find associated deviations from an ecological null model of community assembly, provided by idealized ecological theory. Differences in population structure among taxa or trophic groups could indicate whether sufficient time has passed along the chronosequence for the group of interest to experience significant evolutionary pressures.

#### Macroecological metrics and idealized ecological theory

By their nature, unified theories of biodiversity (e.g. Hubbell, 2001; Harte, 2011) provide a simplified view of ecology, but deviations from theory can provide insights into which particular ecological patterns require additional biological mechanisms for their explanation (Harte, 2011). The maximum entropy theory of ecology (METE; Harte, 2011) in particular provides predictions of species abundance distributions, species are a relationships and metabolic rate and network linkage distributions for idealized ecological communities in which the behavior of a system is governed by a simple set of state variables. The principle of maximum information entropy (MaxEnt), from which the METE is derived, is an established inference procedure that has yielded accurate predictions of diverse patterns in fields as varied as thermodynamics (Jaynes, 1957), economics (?), forensics (?), imaging technologies (?) and, more recently, ecology (e.g. ??Harte, 2011). MaxEnt works by seeking the least-biased prediction of a distribution of interest (e.g. the distribution molecular velocities in the case of thermodynamics or of species abundances in the case of ecology) while constraining that prediction to be consistent with state variables describing the macroscopic attributes of the system (e.g. temperature or the total number of species and individuals). These are the most ignorant possible predictions about the system. Thus, studying the unique ecological conditions and evolutionary histories of real-world systems that deviate from the conditions predicted from maximizing information entropy can provide insights into the processes driving ecological systems away from the statistical steady state (Harte, 2011).

Ecological networks are complex systems forming hierarchical structures to which the principle of MaxEnt has recently been applied (Williams, 2010; Harte, 2011) and are a prime study focus because networks of interacting species embody both the ecology of trophic links and evolutionary processes such as co-evolution (Bascompte & Jordano, 2007; Donatti et al., 2011; Nuismer et al., 2013; ?). Thus they present an opportune starting place to study ecological and evolutionary feedbacks. The distribution of linkages in ecological networks can test whether plantanimal interaction networks assemble neutrally or through deterministic processes such as co-evolution of traits involved in foraging (Vázquez et al., 2005). Analysis of other network metrics such as modularity (the degree to which species interact in semi-autonomous modules) and nestedness (the degree of asymmetry in interaction between specialists and generalists) can further illuminate the underlying eco-evolutionary processes driving patterns of species interactions (Bascompte & Jordano, 2007; Donatti et al., 2011; Nuismer et al., 2013). In nested networks, species with fewer interactions (i.e. more specialized species) will interact with a subset of the species with which generalists interact. In this way interaction nestedness is mathematically equivalent to island nestedness (in which islands that are less species rich are subsets of islands that are more species rich). However, we only consider network nestedness here.

To gain insights into community assembly as it happens, we propose an integrative framework that harnesses advances in both evolutionary and ecological theory, placed in the context of age-structured archipelagos. Mechanistically simplified ecological theories such as the METE (Harte, 2011) can be used as powerful null models; deviations from theoretical expectations can flag biological phenomena that warrant further study. Here we demonstrate how community-level data from age-structured island systems, combined with population genetic and phylogenetic data, can test the extent to which the evolutionary histories behind such communities drive their deviation from theoretical expectations. We provide an initial test of this concept using a synthesis of published data on arthropod lineages in the Hawaiian islands. We provide metrics of ecological and evolutionary dynamics across communities from settings that range in geological age from 500 years to 5 Ma. We estimate taxon-specific timelines for the development of population genetic structure for both herbivores and predators and couple these results with macroecological measures of community structure, using predictions from statistical steady-state and ecological network theory to provide insights into changes in community structure over the extended timeframe provided by the island chronosequence.

#### 2.2 Methods

## Dispersal-driven processes to in situ differentiation across the island chronosequence

To evaluate the balance between regional immigration and the potential for local differentiation, we measured how molecular variation is partitioned among populations within species across locations of known substrate age on the islands of Hawaii and Maui (Fig. 1). We compiled published [DNA sequences, amplified fragment length polymorphism (AFLPs) and allozymes] and new data sets for a diversity of native Hawaiian arthropod groups that represent a spectrum of trophic levels (Table1). New sequences were included for sap-feeding Hemiptera group Nesosydne planthoppers [COI; data generated following the protocols in Goodman et al. (2012); GenBank accession numbers: KT023113KT023179] and Trioza psyllids [COI, cytB; data generated following protocols in Percy (2003); GenBank accession numbers: KR108061KR108144]. Samples were from the focal sites described below for the ecological analysis, as well as from other locations across Hawaii and Maui. These data were used to provide an estimate of how arthropod populations have accumulated genetic population structure within the focal sites of different geological age.

We used analysis of molecular variance (AMOVA) to examine how genetic variation is partitioned at two scales of population structure: among sites within volcanoes and among volcanoes on both the island of Hawaii and the islands of the Maui Nui complex (Maui, Molokai, Lanai). All analyses of allozyme and DNA sequence data were performed in Arlequin v.3.5 (Excoffier & Lischer, 2010) using the AMOVA procedure to compute  $F_{ST}$ , a measure of genetic variance, or, where possible,  $\Phi_{ST}$ , an  $F_{ST}$  analogue that incorporates genetic sequence

information. The *Laupala* AFLP data were analyzed using tfpga v.1.3 (Miller, 1997), using the same hierarchical approach of comparing within and among volcanoes as described above. To provide a temporal framework for the population differentiation analysis we assembled divergence-dating information from the literature for as many of the taxa as possible.

To explicitly test the association between landscape age and the potential for in situ genetic divergence we analyzed how within-site  $F_{ST}$  varies with the geological age of volcanoes on the islands of Hawaii and Maui Nui. For each volcano we calculated  $F_{ST}$  or  $\Phi_{ST}$  (Excoffier & Lischer, 2010) for each taxon among sites within volcanoes. This analysis assumes that volcano age parallels habitat age, allowing more or less time for the presence of the populations.

#### Ecological metrics across the island chronosequence

To investigate how ecological patterns change as communities age, we selected four focal sites across the chronosequence and island ages (two on the island of Hawaii, one on Maui and one on Kauai; Fig. 1) of approximately 12 km² (each was defined as a point with a 2 km radius buffer). Focal sites were selected to have similar forest composition (dominated by Metrosideros polymorpha; Myrtaceae), elevation (1100–1400 m) and rainfall (mean annual precipitation 2000–3000 mm). We then constructed bipartite interaction networks between native herbivorous Hemiptera species and native plants at each of the study sites. Bipartite networks describe the topology of ecological interactions between two guilds of organisms (e.g. herbivores and their plant hosts). Quantitative information on the relative importance of interaction links can be incorporated into network analyses (Vazquez et al., 2009). However, currently available data are restricted to binary networks: those that describe the potential for interaction between any two species but not the relative frequency of that interaction to each species.

We compiled species lists of all native herbivorous Hemiptera for each focal site from published species accounts (see Table S1 in the Supporting Information for a full list). Species accounts and other published sources were used to determine the presence, probable presence, or probable absence of each species at each of our four focal sites. A documented presence was defined as a known specimen collected at the focal site; a probable presence was defined as a species whose abiotic tolerances and known geographic range overlap with a focal site but no known specimen exists confirming its presence. Probable absence was assumed when the criteria for presence or for probable presence are not met. Two sets of species lists for each focal site were compiled: a conservative data set composed of only documented presence occurrences and a less conservative data set that also included probable presences.

Host plants for each species of Hemiptera were determined from published species accounts. Data on host plant use at each specific site were not available so we assumed that if a known host plant were present at a site it would eventually be used. Host plant occurrence in the focal sites was determined using distribution models for 1158 species of Hawaiian plants (Price, 2012). Each focal site was spatially joined in a geographic information system with all coincident plant distribution models that fell within its boundaries. Two sets of

resulting focal site-specific networks were constructed: one using the conservative data set of Hemiptera species presences and the other using the less conservative data set.

We hypothesized that potentially complex evolutionary feedbacks contributing to community assembly should result in departures from the predicted ecological statistical steady state. We used the METE (Williams, 2010; Harte, 2011) to compute the statistical steady state for the distribution of the number of host plants used by each Hemiptera species (hereafter referred to as degree distribution). To evaluate how well the METE predicts the data we simulated METE-conforming communities having the same number of species and links as observed. We then calculated the log-likelihood of each simulated data set and compared the resultant distribution of log-likelihoods under the hypothesis that the METE is true with the observed log-likelihood. This comparison is identical in approach to a z-score test using a Monte Carlo simulation to estimate the sampling distribution of log-likelihoods. R scripts (v.3.1.1; ?) used for METE estimation and Monte Carlo methods are available in Appendix 1. To investigate how speciation may in part drive network patterns and deviations from those predicted by idealized ecological theory, we analyzed the number of links assigned to each Hemiptera species (the degree distribution) separately for single-island endemics (those species found on only one island and thus probably derived from in situ diversification) versus multi-island endemics (those species found on multiple islands). Although multiple processes can lead to a species being a single-island endemic (?), such taxa provide a proxy for how much speciation occurs within islands. To compare species degree distributions between single-island endemics and multi-island endemics across sites of different ages we conducted a generalized linear model with binomial error, treating site identity as a categorical predictor. Binomial errors effectively account for network size due to the bounded support of the binomial distribution.

To understand how other network properties change with age of the ecosystem substrate, we calculated two widely used descriptive network metrics across sites—nestedness and modularity. Nestedness describes the degree of asymmetry of species interactions connecting specialists and generalists (Bascompte & Jordano, 2007; Ulrich et al., 2009). We calculated nestedness using the NODF metric (Almeida-Neto et al., 2008) as implemented in the R package vegan (Oksanen et al., 2013) and modularity using a variety of algorithms implemented in the R package igraph (Csardi & Nepusz, 2006). These metrics are not directly comparable across networks of different size and connectance (Ulrich et al., 2009), so for each metric in each network we calculate z-scores using a null model that randomizes network structure while maintaining certain aggregate network properties (Ulrich et al., 2009). These z-scores are calculated as the difference between the observed network metric minus the mean of the null model divided by the null model standard deviation, or  $(x_{obs}\bar{x}_{sim})/SD_{sim}$ . Because z-scores can be highly sensitive to the choice of null model (Ulrich et al., 2009) we implemented both a probabilistic null model (Bascompte & Jordano, 2007) and a null model that strictly constrains the degree distributions of plants and herbivores (Ulrich et al., 2009). The probabilistic null uses the frequency of interactions as the probability that a randomized link gets assigned to that cell in the interaction matrix (Bascompte & Jordano, 2007); thus the probabilistic null constrains row and column sums in probability but not absolutely.

#### 2.3 Results

## Dispersal-driven processes to in situ differentiation across the island chronosequence

The AMOVA revealed significant genetic population structure from the smallest to the largest spatial scales examined, all within a very recent timeframe. For mitochondrial loci, statistically significant molecular variation partitioned among sites within volcanoes ranged from 0.037 to 0.92 and among volcanoes from 0 to 0.30. Corresponding variation at multilocus nuclear loci among sites within volcanoes ranged from 0.21 to 0.58 and among volcanoes from 0.04 to 0.34. Taxa in the lower trophic levels (herbivorous sap-feeding Hemiptera: planthoppers and psyllids) had as much or more molecular variation partitioned at the among-site, within-volcano level than the among-volcano level, while the predatory spiders were less structured at localities within volcanoes compared with among them (Table 1). The analysis of genetic population structure across the chronosequence of localities revealed a similar pattern. The herbivores show high genetic population structure among localities even on young volcanoes (Fig. 2). By contrast, predatory spiders exhibited little genetic population structure within sites on the same volcano; this was higher among volcanoes, with values increasing with age across the chronosequence.

The observed levels of genetic divergence have evolved rapidly in many cases. For example, for species from the island of Hawaii for which phylogenetic data provide divergence times, estimates of dates of species divergence range from 0.5–4 Ma, with additional within-species genetic divergence having developed subsequently (Table 1). That some of these estimates are older than the known age of the Big Island suggests that genetic divergence pre-dates their colonization to Hawaii, or alternatively that estimates include sampling error. For the one species where population genetic data were used to estimate divergence times between populations, herbivorous Nesosydne planthoppers, it was determined that populations diverged as little as 2600 years ago (Goodman et al., 2012, Table 1).

#### Ecological metrics across the island chronosequence

The degree distribution of Hemiptera species varied across the chronosequence with both the youngest and oldest sites deviating most from the statistical steady-state maximum entropy predictions (Fig. 3). In the intermediate-aged site of Kohala, deviations are not significantly different from the predictions of maximum entropy. The generalized linear model revealed significant differences between the degree distributions of single-island endemics (species whose distributions are restricted to only one island) versus archipelagic endemics that are found across multiple islands (Fig. 3). Single-island endemics show significantly lower degree distributions overall (i.e. more specialization) compared with more generalist species found across multiple islands. Furthermore, single-island endemics use more host plant species on the intermediate-aged Maui site. The slightly younger Kohala shows increased generalization for both single-island endemics and archipelago endemics. However, when considering the

degree distribution defined by trophic links to plant genera instead of plant species, the pattern of increased generalization holds for Kohala, but endemics on Maui no longer show a difference in their degree distributions from other island endemics. This change in pattern suggests that increased generality of Maui endemics may be driven by increased plant species diversity within genera on that island.

Network nestedness decreased with habitat age while modularity increased (Fig. 4). This trend was recovered in networks constructed from both more and less stringent geographic criteria (Fig. S3). Choice of null model changed the magnitude of modularity and the sign of nestedness z-scores; however, the relative pattern of decreasing nestedness and increasing modularity remained across the different null models used to standardize network metrics (Fig. S2). The patterns were also robust to sampling intensity, as demonstrated by a rarefaction analysis (Fig. S4).

#### 2.4 Discussion

## Development of genetic population structure at different trophic levels

The analysis of available genetic data presented here indicates that divergence is occurring within the islands at small spatial scales and over short time periods (Table 1, Fig. 2). Furthermore, the scale of population structure varies with trophic position, with structure developing in sap-feeding herbivore lineages at smaller scales (and hence shorter timeframes in the context of the chronosequence) compared with detritivorous crickets and predatory spiders (Table 1, Fig. 2). Structure within species may allow populations to take independent evolutionary trajectories, especially when aided by other evolutionary processes acting differentially across species geographic ranges. A variety of factors have been associated with the genetic divergence of populations and species in the lineages described here, including combinations of genetic drift associated with geographic isolation (Percy, 2003; Mendelson & Shaw, 2005; O'Grady et al., 2011; Goodman et al., 2012), adaptation associated with competition, predation and mutualism (Gillespie, 2004; Roderick & Percy, 2008; ?) and sexual signaling (Mendelson & Shaw, 2005; Percy & Kennedy, 2006; Magnacca et al., 2008; ?).

The Nesosydne planthoppers provide evidence that some period of geographic isolation preceded the divergence of sexual signals (Goodman et al., 2012; ?). Shifts in plant host use are also associated with diversification in this group (Roderick & Percy, 2008). In a phylogenetic study of a radiation of sap-feeding Nesophrosyne (Cicadellidae) leafhoppers, species divergence was associated with host plant specialization between 1 and 5 Ma, but only with geography on the younger island (Bennett & O'Grady, 2013). Our network analysis shows that specialization and modularity are more pronounced on Maui than on Hawaii (Figs 3 and 4), consistent with the phylogenetic results from Nesophrosyne. Available dating analyses of other arthropod taxa indicate that population genetic structure can develop in much less than 1 Myr (Table 1), and suggest that landscape fragmentation processes (e.g.

lava flows) may dominate the earliest stages of diversification across taxa in the Hawaiian islands. Other taxa at low trophic levels, such as the herbivorous *Trioza psyllids*, detritivorous *Laupala* crickets and fungivorous Drosophila, show similar signals of geographic isolation combined with ecological and sexual processes driving genetic divergence and diversification across sites as young as those on Hawaii (Percy, 2003; Mendelson & Shaw, 2005; Percy & Kennedy, 2006; Magnacca *et al.*, 2008; O'Grady *et al.*, 2011). By contrast, spiders, which are predatory, develop genetic discontinuities at larger spatial and temporal scales with a strong signature of increasing structure with age of the chronosequence (Roderick *et al.*, 2012, Table 1). Further work is needed to assess the generality of this pattern of slower genetic differentiation in predators compared with herbivores.

#### Macroecological metrics: network structure and steady state

Across the Hawaiian archipelago, nestedness appears to decrease generally with site age, and is highest on the geologically youngest volcano, Kilauea. High nestedness on Kilauea may arise with high immigration of new species with high probabilities to eat or be eaten by the generalist species already present at the site (Bascompte & Jordano, 2007). However, despite high nestedness on Kilauea, and thus the potential for neutral colonization-driven assembly, this site did not conform to the statistical steady-state predication of the METE. The observed deviations from the METE at Kilauea appear to be largely driven by a surplus of singleton links (Fig. 3), which may reflect a state of incomplete assembly, possibly by lower species richness of the plant and herbivore biotas. Conversely, at Kohala, at intermediate age (150 ka), observations were not significantly different from the METE predictions. We posit that the reason why theoretical predictions fit Kohala so well is that the site has had sufficient time to undergo ecological succession and thus arrive at a statistical steady state, but is still too young to be affected by ecological specialization and rapid in situ diversification associated with host plants on older islands.

Interestingly, the communities on the older Maui and Kauai sites show strong deviations from the METE expectations (Fig. 4). The METE is agnostic about which mechanisms determine the values of the state variables that lead to its macroecological predictions (Harte, 2011). It does not account for the evolutionary history of biological systems. Thus, one possible explanation for the strong deviations from the METE expectations, compared with observations at our intermediate-aged site (Kohala), is that while the ages of Maui and Kauai are sufficient for evolutionary assembly driven by specialization and diversification on host plants, the older age of these islands may have led to range contractions and possibly extinction of plant species on the oldest island of Kauai (Whittaker et al., 2008).

Our results show decreased nestedness and increased modularity on Maui and Kauai. Co-evolution between interacting species should lead to greater modularity (Donatti *et al.*, 2011; Nuismer *et al.*, 2013). However, the influence of certain network properties, such as nestedness, on stability is still unknown, and so theoretical predictions of how network properties should change over evolutionary time, generally, are lacking. Theoretical and empirical studies have suggested that nestedness may or may not promote stability (Allesina

& Tang, 2012; ?). Furthermore, almost all studies of food webs have focused primarily on single or short ecological time spans of network development that do not span as much evolutionary time as is included here (e.g. Albrecht et al., 2010). Food webs are dynamic emergent entities, with broad topological characteristics that may change dramatically over time (e.g. Yeakel et al., 2013). To our knowledge, our study represents the first to evaluate network topology over larger temporal scales, and we argue that age-structured landscapes such as the Hawaiian archipelago are promising for resolving long-standing debates on the causes and consequences of network properties such as nestedness.

We found that single-island endemics were always more specialized than multiple-island endemics. Although dietary breadth has been positively associated with geographic range size (Lewinsohn et al., 2005), the direction of causality is unclear (Slatyer et al., 2013): while dietary breadth may allow some species to colonize other islands, it may also be driven by adaptation to exploit locally abundant hosts across a large range. Nevertheless, both scenarios are consistent with the hypothesis that in situ formation of single-island endemics may be the product of co-evolution and specialization. At the Kohala site, which showed the best fit to maximum entropy theory, single-island endemic and multiple-island endemic species alike showed increased generalization (i.e. a higher degree, or more links; Fig 3), while at the youngest site of Kilauea, specialist single-island endemics may be limited by low plant diversity and thus appear more specialized (Fig. 3). Conversely at the oldest site on Kauai, where plant diversity is high (Kitayama & Mueller-Dombois, 1995), single-island endemics are again associated with decreased degree and thus genuine specialization (Fig. 3). On Maui, single-island endemics show statistically significant increases in generalization, but this pattern disappears when analyzing the data at the resolution of plant genera, thus suggesting that Hemiptera species endemic to Maui may benefit from the diversification of plant species within genera.

#### Future research

The data and analyses presented here describing insect and plant communities across a chronosequence of habitats in Hawaii generate testable hypotheses concerning the relative importance of ecological and evolutionary processes in community assembly. Our work to date suggests the overarching hypothesis that ecological processes dominate community assembly in younger environments, with evolutionary processes becoming increasingly important as communities age. We can also make predictions about the sequence of community assembly based on proposed mechanisms.

In younger communities we predict characteristics of ecological assembly, with species resembling random samples through immigration from regional source pools. Thus, metrics describing these communities will approach expectations of an ecological statistical steady state. An exception will be communities that are still undergoing the initial stages of primary succession, which will change rapidly through time and represent nonrandom samples of source pools. We also predict that these communities will exhibit a nested network structure, assuming new species will eat or be eaten by the generalist species already present in

the community, as suggested by previous work on nestedness (bascompte2007) and by our finding that widespread species tend to be generalists (Fig. 4). Following the same logic, in older communities we expect to see characteristics of evolutionary assembly, dominated by processes such as adaptive exploration of niche space, giving way to speciation. Thus, we predict increasing specialization and modularity with time (Bascompte & Jordano, 2007; Donatti et al., 2011; Nuismer et al., 2013) as reflected by age across the chronosequence.

#### Ecological data: assembly of species into communities

In order to build a more rigorous understanding of the assembly process in both younger and older communities, fine-grained sampling of all macroscopic arthropod taxa is needed from a large number of sites across the island chronosequence. This will allow an assessment of changes in overall species composition and diversity across all players in the time-calibrated landscape (Gruner, 2007). Such data will allow us to test entire arthropod communities for deviations from METE predictions of statistical steady state (Harte, 2011) across substrates of different ages. For example, predators, whose assemblages are likely to be more dominated by immigration and ecological assembly (Fig. 2), may never show strong deviations from METE predictions, whereas herbivores could show increasing deviation with age in agreement with the network results of this paper (Fig. 3).

#### Evolutionary data: diversification within species

The current study demonstrates that taxa from different trophic guilds differ in the scale at which differentiation occurs and highlights the importance of fragmentation of the landscape in facilitating differentiation. Future work will be aimed at gathering data for additional focal taxa within this system, spanning different trophic levels. We will use these data to understand taxonomic and functional differences in the rate of differentiation, to assess the roles of genetic fusion and fission and the spatial scale over which they are important in fostering diversification (Gillespie & Roderick, 2014), and to detail the relative rates of speciation and extinction across the island chronosequence.

#### 2.5 Conclusions

We have shown how a chronosequence can be used to understand biodiversity dynamics across an ecological evolutionary continuum. Focusing on entire communities of arthropods in the Hawaiian islands allows us to incorporate predictions from idealized ecological theories to understand eco-evolutionary feedbacks and generate predictions about how entire communities develop over an extended time. Such an approach may prove fruitful for investigating the separate and interactive roles of ecological and evolutionary drivers of community assembly using age-structured systems as a simplified natural experiment, as exemplified by oceanic archipelagos.

We have demonstrated how taxa in the lower trophic levels developed genetic structure even in the youngest habitats of the observed chronosequence and at smaller spatial scales (Table 1, Fig. 2). Thus, lower trophic levels are affected by in situ processes of diversification very early in the chronosequence, compared with higher trophic levels, though in situ processes become more important over time in the latter. Network nestedness decreased while modularity increased with age (Fig. 4), again indicating a possible shift from assembly driven by ex situ immigration early on to one based on in situ diversification, such as in co-diversification of insect herbivores with host plants (Bascompte & Jordano, 2007; Donatti et al., 2011). That single-island endemics (probably the product of in situ diversification) show more specialization at older sites than more broadly distributed species (those taxa more likely to be initial colonists; Fig. 3) also supports this hypothesis.

This study provides a framework for using chronologically arranged oceanic island systems to examine the interplay between evolutionary and ecological processes in shaping biodiversity. Our initial results provide a clear hypothesis that ecological processes dominate community assembly in younger environments, with evolutionary processes becoming more important as communities age. We demonstrate how this approach can provide insights into the development of communities over ecological evolutionary time, and the dynamic feedbacks involved in assembly.

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Figure 2.1: Map of substrate age (millions of years, My) of the islands of Kauai, Maui and Hawaii. Colours correspond to substrate age from young (light) to old (dark). Focal sites are shown as black circles (on Hawaii, Kohala is in the north, Kilauea in the south) while sampling sites for genetic data are represented by grey circles..

#### **Figures**

Figure 2.2: Population genetic structure ( $\Phi_{ST}$  for all taxa except Laupala for which we used  $F_{ST}$ ) among sites within volcanoes with volcano age for insects and spiders. Calculations were based on mitochondrial DNA only (see Table 1 for details). The plant-feeding groups, specifically the sap-feeding Hemiptera, show higher genetic structure among sites on young volcanoes relative to older volcanoes, whereas detritivores (crickets), fungivores (Drosophila) and in particular predators (spiders) show little structure on young volcanoes. For spiders, substantial structure develops only later in the chronosequence, for example on Maui at approximately 1 Ma. Numbers refer to different species: 1, Nesosydne chambersi; 2, Nesosydne raillardiae; 3, Nesosydne bridwelli; 4, Trioza HB; 5, Trioza HC; 6, Drosophila sproati; 7, Laupala cerasina; 8, Tetragnatha anuenue; 9, Tetragnatha brevignatha; 10, Tetragnatha quasimodo; 11, Theridion grallator.

Figure 2.3: Patterns in degree distributions across sites, comparing archipelago-wide endemics (cosmopolitans) with single-island endemic (Endemics) taxa. The top panels show that networks deviate most from the predictions of the maximum entropy theory of ecology on the youngest and oldest sites. Inset figures show the distribution of simulated mean squared errors; if the vertical/red line falls within the grey region (95% confidence interval) the data are not significantly different from the predictions of maximum entropy theory. All sites except Kohala deviate from the predications. The bottom panel shows the number of links for endemics versus cosmopolitans. Endemics show lower linkage overall, but significantly increase on the intermediate-aged site Maui (highlighted with dotted box). Kohala shows increased linkage overall (highlighted with a solid box).

#### **Tables**

Table 2.1: Results of the analyses of molecular variance (AMOVA) that partitions molecular genetic variation among volcanoes and among sites within volcanoes for arthropod lineages found within the study sites on the island of Hawaii. Where estimates of divergence through molecular dating are available for the taxa, they are presented to show the timeframe within which this genetic structure has developed.

## Chapter 3

meteR: An R package for testing the Maximum Entropy Theory of Ecology

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## Conclusion

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## Bibliography

Albrecht, M., Riesen, M. & Schmid, B. (2010) Oikos, 119, 1610–1624.

Allesina, S. & Tang, S. (2012) Nature, 483, 205–208.

Almeida-Neto, M., Gumarães, P., Gumarães, P., Loyola, R. & Ulrich, W. (2008) *Oikos*, **117**, 1227–1239.

Alroy, J. (2010) Science, **329**, 1191–1194.

Alroy, J., Aberhan, M., Bottjer, D.J., Foote, M., Fürsich, F.T., Harries, P.J., Hendy, A.J.W., Holland, S.M., Ivany, L.C., Kiessling, W., Kosnik, M.A., Marshall, C.R., McGowan, A.J., Miller, A.I., Olszewski, T.D., Patzkowsky, M.E., Peters, S.E., Villier, L., Wagner, P.J., Bonuso, N., Borkow, P.S., Brenneis, B., Clapham, M.E., Fall, L.M., Ferguson, C.A., Hanson, V.L., Krug, A.Z., Layou, K.M., Leckey, E.H., Nürnberg, S., Powers, C.M., Sessa, J.A., Simpson, C., Tomašových, A. & Visaggi, C.C. (2008) Science, 321, 97–100.

Bak, P. & Sneppen, K. (1993) Phys. Rev. Lett., 71, 4083–4086.

Bascompte, J. & Jordano, P. (2007) Annu. Rev. Ecol. Evol. Syst., 38, 567–93.

Beck, C. (2004) *Physica D: Nonlinear Phenomena*, **193**, 195–207.

Beck, C. & Cohen, E. (2003) Physica A: Statistical Mechanics and its Applications, 322, 267–275.

Bennett, G.M. & O'Grady, P.M. (2013) Journal of Biogeography, 40, 1512–1523.

Benton, M. (1995) Science, 268, 52–58.

Borer, E.T., Seabloom, E.W., Gruner, D.S., Harpole, W.S., Hillebrand, H., Lind, E.M., Adler, P.B., Alberti, J., Anderson, T.M., Bakker, J.D. et al. (2014) Nature.

Brose, U., Berlow, E.L. & Martinez, N.D. (2005) *Ecology Letters*, 8, 1317–1325.

Brose, U., Williams, R.J. & Martinez, N.D. (2006) *Ecology Letters*, **9**, 1228–1236.

Butler, M.A. & King, A.A. (2004) The American Naturalist, 164, 683–695.

Cavender-Bares, J., Kozak, K.H., Fine, P.V. & Kembel, S.W. (2009) *Ecology letters*, 12, 693–715.

- Chave, J. (2004) Ecology letters, 7, 241–253.
- Cox, J.C., Ingersoll Jr, J.E. & Ross, S.A. (1985) Econometrica: Journal of the Econometric Society, 385–407.
- Csardi, G. & Nepusz, T. (2006) InterJournal, Complex Systems, 1695.
- Donatti, C.I., Guimarães, P.R., Galetti, M., Pizo, M.A., Marquitti, F. & Dirzo, R. (2011) Ecol. Lett., 14, 773–781.
- Eldredge, N. & Gould, S.J. (1972) Models in paleobiology, 82, 115.
- Erwin, D.H. (1998) Trends in Ecology and Evolution, 13, 344–349.
- Erwin, D.H. (2007) *Palaeontology*, **50**, 57–73.
- Erwin, D.H. (2012) Journal of Experimental Zoology Part B: Molecular and Developmental Evolution, 318, 460–465.
- Excoffier, L. & Lischer, H.E. (2010) Molecular ecology resources, 10, 564–567.
- Foote, M., Crampton, J.S., Beu, A.G. & Cooper, R.A. (2008) Paleobiology, 34, 421–433.
- Fuentes, M.A., Gerig, A. & Vicente, J. (2009) PLoS ONE, 4, e8243.
- Gavrilets, S. & Vose, A. (2005) Proceedings of the National academy of Sciences of the United States of America, 102, 18040–18045.
- Gilinsky, N.L. (1994) *Paleobiology*, 445–458.
- Gillespie, R. (2004) Science, **303**, 356–359.
- Gillespie, R. (2009) Adaptive radiation. S. Levin, S. Carpenter, C. Godfray, A. Kinzig, M. Loreau, J. Losos, B. Walker & D. Wilcove, eds., *Guide to Ecology*, Princeton University Press, Princeton, NJ, pp. 143–152.
- Gillespie, R. & Roderick, G. (2014) *Nature*, **509**, 207–298.
- Goodman, K., Welter, S. & Roderick, G. (2012) Evolution, 66, 27982814.
- Gruner, D.S. (2007) Biological Journal of the Linnean Society, 90, 551–570.
- Harnik, P.G. (2011) Proceedings of the National Academy of Sciences, 108, 13594–13599.
- Harte, J. (2011) The Maximum Entropy Theory of Ecology. Oxford University Press.

- Holman, E.W. (1989) *Paleobiology*, 357–363.
- Hopkins, M.J., Simpson, C. & Kiessling, W. (2014) Ecology letters, 17, 314–323.
- Hubbell, S.P. (2001) The unified neutral theory of biodiversity and biogeography (MPB-32), vol. 32. Princeton University Press.
- Jablonski, D. (2008) Annual Review of Ecology, Evolution, and Systematics, 39, 501–524.
- Kirchner, J.W. & Weil, A. (1998) *Nature*, **395**, 337–338.
- Kitayama, K. & Mueller-Dombois, D. (1995) Vegetatio, 120, 1–20.
- Lewinsohn, T.M., Novotny, V. & Basset, Y. (2005) Annual Review of Ecology, Evolution, and Systematics, 597–620.
- Liow, L.H. & Stenseth, N.C. (2007) Proceedings of the Royal Society B: Biological Sciences, 274, 2745–2752.
- Madin, J.S., Alroy, J., Aberhan, M., Fürsich, F.T., Kiessling, W., Kosnik, M.A. & Wagner, P.J. (2006) *Science*, **312**, 897–900.
- Magnacca, K.N., Foote, D. & OGrady, P.M. (2008) Zootaxa, 1728, 1–58.
- Mayr, E. (1965) Systematic Zoology, 14, 73–97.
- Mendelson, T.C. & Shaw, K.L. (2005) Nature, 433, 375–376.
- Newman, C., Cohen, J. & Kipnis, C. (1985) *Nature*, **315**, 400–401.
- Newman, M.E.J. & Roberts, B.W. (1995) Proceedings of the Royal Society of London B, **260**, 31–37.
- Nuismer, S.L., Jordano, P. & Bascompte, J. (2013) Evolution, 67, 338–354.
- O'Grady, P.M., Lapoint, R.T., Bonacum, J., Lasola, J., Owen, E., Wu, Y. & DeSalle, R. (2011) Molecular phylogenetics and evolution, 58, 244–256.
- Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H. & Wagner, H. (2013) vegan: Community Ecology Package.
- Percy, D. (2003) Evolution, 57, 2540–2556.
- Percy, D. & Kennedy, M. (2006) Invertebrate Systematics, 20, 431–445.
- Price, J. & Clague, D. (2002) Proc R Soc Lond B Biol Sci, 269, 2429–2435.

Price, J.P. (2012) Mapping plant species ranges in the Hawaiian Islands: developing a methodology and associated GIS layers. US Department of the Interior, US Geological Survey.

- Price, T.D., Hooper, D.M., Buchanan, C.D., Johansson, U.S., Tietze, D.T., Alström, P., Olsson, U., Ghosh-Harihar, M., Ishtiaq, F., Gupta, S.K. et al. (2014) Nature, **509**, 222–225.
- Quental, T.B. & Marshall, C.R. (2013) Science.
- Rabosky, D.L. (2006) Evolutionary bioinformatics online, 2, 247.
- Rabosky, D.L. (2009) Ecology Letters, 12, 735–743.
- Raup, D.M., Sepkoski Jr, J.J. et al. (1982) Science, 215, 1501–1503.
- Ricklefs, R.E. (2006) Ecology, 87, 1424–1431.
- Roderick, G., Croucher, P., Vandergast, A. & Gillespie, R. (2012) Evolutionary Biology, 32, 192–206.
- Roderick, G.K. & Percy, D. (2008) Host plant use, diversification, and coevolution: Insights from remote oceanic islands. K. Tilmon, ed., *Specialization, Speciation, and Radiation.* Evolutionary Biology of Herbivorous Insects, University of California Press, Berkeley, pp. 151–161.
- Rosindell, J., Hubbell, S.P. & Etienne, R.S. (2011) Trends in ecology & evolution, 26, 340–348.
- Rosindell, J. & Phillimore, A.B. (2011) Ecology Letters, 14, 552–560.
- Roy, K., Hunt, G., Jablonski, D., Krug, A.Z. & Valentine, J.W. (2009) Proceedings of the Royal Society B: Biological Sciences, 276, 1485–1493.
- Sepkoski, J.J. (1984) *Paleobiology*, **10**, 246–267.
- Sepkoski, J.J. (1992) A compendium of fossil marine animal families. Milwaukee Public Museum, Milwaukee, WI.
- Simpson, G. (1953) The Major Features of Evolution, 313–337.
- Slatyer, R.A., Hirst, M. & Sexton, J.P. (2013) Ecology letters, 16, 1104–1114.
- Solé, R.V., Manrubia, S.C., Benton, M. & Bak, P. (1997) Nature, 388, 764–767.
- Tilman, D. (2004) Proceedings of the National Academy of Sciences of the United States of America, 101, 10854–10861.

Ulrich, W., Almeida-Neto, M. & Gotelli, N.J. (2009) Oikos, 118, 3–17.

Vázquez, D., Poulin, R., Krasnov, B. & Shenbrot, G. (2005) Journal of Animal Ecology, 74, 946–955.

Vermeij, G.J. (1987) Evolution and Escalation. Princeton University Press, Princeton, N.J.

Vrba, E.S. (1983) Science, **221**, 387–389.

Williams, R.J. (2010) Theoretical Ecology, 3, 45–52.

Williams, R.J. & Martinez, N.D. (2000) Nature, 404, 180–183.

Yeakel, J.D., Guimarães, P.R., Bocherens, H. & Koch, P.L. (2013) Proceedings of the Royal Society of London B: Biological Sciences, 280, 20130239.