# Suitability prevalence area index in late quaternary explains genetic diversity in Tassel eared Squirrels

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# Abstract

The current distributions of species do not always correspond to their historical distributions over evolutionarily significant periods. This is because environmental conditions are not static over time; species tend to distribute where conditions are most favorable.

It follows from this relationship between environment and species distribution that population size also varies with time. This variation in population size is related to the effective population size  $N_e$  (Karlin 1968), so that an index reflecting changes in environmental conditions in geography might be expected to be related to  $N_e$  and thus to indicators of population structure such as the fixation index  $F_{st}$ . Thus, it is possible to relate patterns of changes in the distribution of species to the genetic structure of their populations using a statistical model that explains this relationship.

With this approach we can predict the geographic pattern of population structure from environmental information. This approach is strongly driven by advances in currently available climate simulations (Leonardi et al. 2023; Krapp et al. 2021), as well as next generation sequencing data, and supported by both ecological niche (Thorup et al. 2021; Nogués-Bravo 2009) and population genetics theories (Lira-Noriega and Manthey 2014).

In this work we propose a method to find the Suitability Prevalence Area (SPA) as an index with a double purpose: 1) to find endemic areas to delimit the historical distribution of the species and 2) to explain the patterns of genetic diversity.

To obtain the SPA, we performed a historical reconstruction of the geographical range back to 120 000 BC at 2 000 year intervals and recorded the environmental suitability at each site.

Subsequently, we delimited historical endemic areas to locations where the prevalence of suitability remained at 90% during this period. From the fixation index calculated with respect to populations in the historical endemic areas, a statistical model of the fixation index as a function of SPA was performed. With this statistical model, the fixation index values were projected to the current distribution to obtain a map with the geographic pattern of this index.

As a case study we consider squirrels (*Sciurus aberti*), a species currently distributed in disjunct patches from the southern Rocky Mountains in the United States to the northern Sierra Madre Occidental in Mexico for which it is possible to find reliable information on both the genetic structure of its populations and its current distribution (Bono et al. 2018; Burgin et al. 2018).

Our results reveal that suitability prevalence corresponds to the fixation index of S. aberti populations with respect to a source population. Populations closer to the historical endemic area present a higher genetic diversity and a lower  $F_{st}$  value. Finally, this study allows us to add a biogeographic explanation to the results obtained with population genetic methods and to generate maps of this structure as tools to support conservation with a perspective that integrates both population genetics and historical patterns of species distribution.

# Introduction

Relating the genetic structure of the populations of a species to its geographic distribution is a complicated challenge from the point of view of biogeography. The task is made easier if evolutionary history is incorporated on both sides of the relationship. On the one hand, giving depth and an evolutionary perspective to the geographic distribution of a taxon relies primarily on ecological niche theory, and is driven by the massive amount of information currently available from both digitized biological collections and citizen science (through which records of taxa occurrence are through which records of taxa occurrence are available) as well as from environmental information available for geographic information systems and climate models. information systems and accurate climatological models of both the present and the past.

This change in reproductive success is driven by a set of factors, however we propose in this work a conceptual framework where we consider only environmental factors at a macroecological scale, inspired by the concept of Grinnellian niches (those composed by sets of points of non-interactive variables in multidimensional spaces), seeking as a principle the necessary conditions for fitness and not the response of taxa. points of non-interactive variables in multidimensional spaces), seeking as a principle the conditions necessary for fitness and not the response of taxa.

That is, we simplify our model (as is done in the physical sciences) to link through fitness the changes in biological populations and their structure, observed through population genetics, with patterns of spatial distribution over time. Thus, the goal is to link biogeographic patterns and evolutionary processes of populations in a single conceptual framework.

It is from this Grinellian niche where we start, assuming that taxa in general have to be distributed in the geography where environmental conditions are more favorable, noting that the distribution of taxa is not static, depending on dynamic environmental conditions and therefore the current known distribution of a taxon does not always correspond to its historical distribution over time, since the current climate has not remained so for periods of time on scales relative to the processes of population genetics. (i. e. tens or hundreds of thousands of years). Even for some species these dynamics seem to induce ecological processes such as seasonal migrations in birds, or hibernation in mammals and frogs. But in general, environmental changes are reflected in geography as range expansions and contractions over time. That is, processes of population dynamics, driven by environmental fitness, in our approach.

This change in population dynamics is implied by the genetic structure of populations through the concept of effective population size  $N_e$  (Wright 1931; Karlin 1968). So if we set a similar time scale for these changes, we will be able to observe the fitness relationship (through population dynamics) in both genetic structure and distribution range changes over time. That is, we need an index that reflects environmental changes in the geographic distribution of species and changes in the genetic structure of the population. In this case, we can consider the fixation index  $F_{st}$  as an indicator of the difference between populations that come from a source population and that have undergone range shifts that we assume are driven by environmental changes. Under these assumptions we expect that it is possible to relate changes in the distribution of taxa to the genetic structure of their populations using statistical models to explain the relationship.

## Theoretical basis

Environment, fitness and suitability For a given region  ${\bf G}$  in te geography, at any point in time, we have a set of existing environmental conditions  ${\bf E}({\bf G},t)=\left\{\vec{e}_i;{\bf G},t\right\}$ . We define  $\vec{e}_i=(e_1,e_2,...,e_v)_i$  as a vector of environmental variables, that is, a point in the space of v environmental variables  ${\bf E}({\bf G},t)$ . Then  $R_0(\vec{e}_i)$  is a function that relates fitness to environment, that is, for a species with non-overlapping generations, the population's net reproductive rate  $R_0$  is a measure of fitness that depends on environmental conditions. If we take the set of places in environmental space where these values are greater than a threshold, we can define a volume in multidimensional space such that  $R(\vec{e})$  is always positive and greater than this threshold, that is  ${\bf N}_F=\{\vec{e}|R_0(\vec{e})\geq k\}$ . i. e. the set of environmental conditions for which values of fitness are capable of sustaining populations. We call  ${\bf N}_F\subseteq {\bf E}({\bf G},t)$ . This is a way of expressing the fundamental niche as a function of environmental conditions. So we can correspong each point of geography to environmental space, with different environmental conditions and different growth rates. However all sites within the region

defined by  $\mathbf{N}_F$  will have viable populations and positive rates depending on time because we assume that the environment is dynamic and each geographic point is static, so for each site in geographic space this growth rate depends on the environment and therefore is different at each time:  $R_0 = R_0(\vec{e}(t))$ . Thus locally the growth rate will be related to the local population size as long as it is within the boundary defined by  $\mathbf{N}_F$ .

This equation was proposed by Soberon and Peterson Soberón and Peterson (2020) to relate demographics to a function of fitness and environment. In that study they tested the validity with monthly variables and empirical values on a short time scale. relating life history parameters and the environmental suitability of niche models.

This set can be seen as a convex volume in environmental space and we relate this growth rate to an environmental suitability function bounded between 0 and 1, for example, for a unimodal multivariate pdf, it would be equivalent to the distance from the maximum of the function. That is, the fitness-related environmental suitability can be any function such that it is 0 if  $R_0(\vec{e}_i) > 1$  and 1 when  $R_0$  is evaluated at the expected value of  $\vec{e}(t)$  for all values of the set  $\mathbf{N}_F$  where  $R_0(\vec{e}_i) > 1$ . In a practical way, instead of  $\mathbf{N}_F$ , we approximated  $\mathbf{N}_F^*(\mathbf{G},\mathbf{t}) = \mathbf{E} \cap \mathbf{N}_F$  (the existing niche) from known presence points in the geography  $\mathbf{G}$  obtained from emprical observations stored in database, and map to  $\mathbf{E}$  in order to numerically fitting a convex volume of the environmental space that enclose the known information.

The objective of our approach is not to know the function of  $R_0$  but rather to relate this local variation and the function  $\mathbf{N}_F$  with a function  $S_i(\vec{e}_i)$  of environmental suitability that can be estimated from known methods of applied mathematics and machine learning (for example the case of Maxent, Phillips, Anderson, and Schapire (2006)).

Population size locally is a function of environmental suitability, where if suitability is 0 the population is 0 and if suitability is 1 the population has a maximum growth rate  $R_i(\vec{e}_i)$ 

$$N_i \propto S_i(\vec{e}_i)$$

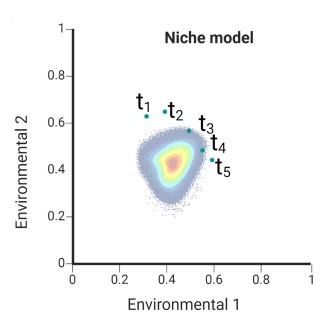


Figure 1: Hypothetical niche in environmental space. An arbitrary point moves within the environmental space entering and exiting the volume (niche) where the intrinsic growth rate is positive.

Once we have an approximation of this volume, we consider that each point in the geography is static and has a trajectory in the environmental space and for each trajectory the fitness function changes and therefore the suitability function changes but not in a homogeneous way, so each site presents different displacement rates in environmental spaces and as a consequence, different demographic rates. So we can establish that

for each site i in geographic space  $\mathbf{G}$ , there is a population size  $N_i$  that changes with respect to a fitness function dependent on time-varying environmental conditions:

$$N_i(\vec{e}_i(t)) \propto S_i(\vec{e}_i(t))$$

Genetic variation and effective population size On the other hand, loss of heterozygosity may be related to the amount of genetic variation present in the absence of mutation and selection, so theoretically, one would expect a strong correlation between effective population size and heterozygosity on the basis of population genetics. We start from the assumption that in the absence of gene flow between subpopulations, the rate of fixation by isolation depends on the effective population size  $N_e$  when subpopulations diverge from a common ancestor in t generations. Wright (1943):

$$F_{ct}(t) = 1 - e^{-kt/N_e}$$

where we assume, according to Nei (Nei 1986), that genetic drift within each subpopulation causes the average heterozygosity between populations  $(H_s)$  to approach zero  $(F_{st} = 1 - H_s/H_t)$ , i.e., for t = 0  $F_{st} = 0$  and for very large t the index is 1. that is, for t = 0  $F_{st} = 0$  and for very large t the index is 1. We have included a constant t without loss of generalization.

Using  $e(x) = \sum_{k=0}^{\infty} \frac{x^k}{k!} = 1 + x + \frac{x^2}{2} + \dots$  and leaving to first order for  $N_e$  large enough

$$F_{st} \approx \frac{kt}{N_c} + C$$

Moreover, the effective population can be estimated as the harmonic mean of the population in a demographic time series (Karlin 1968) for  $\tau$  generations  $\frac{1}{N_e} \approx \frac{1}{\tau} \left( \frac{1}{N_1} + \frac{1}{N_2} + \ldots + \frac{1}{N_{\tau}} \right)$  for each i site in  $\mathbf{G}$ . We substitute the harmonic mean into the above equation, considering that although the number of generations is not exactly equal to the time of the fixation index, they are expected to be very close:

$$F_i^{st} \approx \left(kt\right) \frac{1}{\tau} \sum_{j=1}^{\tau} \frac{1}{N_{ij}}$$

Relationship between the fixation index and environmental suitability. We can infer locally for each population  $N_i$  in the geography expect to be proportional to the environmental suitability for each time considered, For simplification in the index we select a  $N_i$  population in the geography and infer that locally, this population is proportional to the environmental suitability in each time considered, i. e.  $N_j \propto S_j(\vec{e}_j)$  (each index j is for the generation time), as we describe above, so we can assume an harmonic mean of suitabilities  $\frac{1}{\tau} \sum_{j=1}^{\tau} \frac{1}{S_j} = \frac{1}{S}$  and substitue in the previous equation:

$$F_{st} \approx \left(kt\right) \frac{1}{\tau} \sum_{j=1}^{\tau} \frac{1}{S_{j}(\vec{e_{j}})}$$

We call this harmonic mean of environmental suitability Suitability prevalence index (S) given between 0 and 1, since it reflects the rate of population growth as a function of environment over a period of time. Thus we have a way to check that the loss of heterozygosity (through the fixation index  $F_{st}$ ) is inversely proportional to the suitability prevalence for an i site in the geography:

$$F_i^{st} = \frac{\beta_0}{S_i} + \beta_1$$

If we consider all sites in the available geography, we call this harmonic Suitability Prevalence Area  $S = \{S_i | S_i > k\}$  since we expect that areas where the  $F_{st}$  is lower are in areas where the prevalence over time

is higher, i.e. environmental conditions remain suitable for a significant period of time and therefore no effects of genetic drift of populations have occurred. Conversely, it is expected that in populations with a high fixation index value, fitness will not have prevailed over the same time interval even though conditions currently exist and populations are present.

#### Methods

We summarize our 5 steps predictive approach in the figure 2.



Figure 2: Realized niche in environmental space. A point dynamically moves along different time slices

We propose a method to find the Suitability Prevalence Area (SPA) as an index with a dual purpose: 1) to find endemic areas to delimit the historical distribution of species and 2) to explain patterns of genetic diversity.

The method for estimating SPA consists of a series of 4 steps:

- 1. Model the potential distribution of a taxon with present-day environmental conditions and obtain an environmental suitability map.
- 2. Project the potential distribution model to various environmental scenarios in the past with constant time steps and obtain the environmental suitability for each time scenario.
- 3. Characterize past environments and retain projections only for sites with analogous environmental conditions to the present.
- 4. Calculate the harmonic mean of the suitability for all the sites all time slices.

Subsequently, one application of the SPA is to explain genetic diversity (the fixation index) as a function of the SPA and project it to the geography. For this it is necessary to perform a series of steps, which consist of:

- Obtain georeferenced points of genetically structured populations with fixation index values.
- Extract values from the SPA map with these points.
- A linear model of the fixation index value with respect to the SPA is performed.
- With the linear model all SPA values of suitability are interpolated.
- A map is generated with all the values interpolated by the model.

To test our model we use the Aberti squirrel (Sciurus aberti) as a case study because:

- Since it is a medium-sized mammal (0.5 kg) its identification is easy.
- The subspecies are geographically isolated, since they are found in pine forests surrounded by desert. Therefore we can consider the assumption of isolation by distance (Wright 1943).
- There are sufficient occurrence records for the entire range of the known distribution.
- It has low speciation rate (0.25) (Upham, Esselstyn, and Jetz 2019), therefore, we assume that its fundamental niche is stable over time.
- There is fixation index information (Bono et al. 2018) for different populations of the subspecies, which is useful for using SPA as a way to explain genetic differentiation between populations.

### Unique occurrence records

We have downloaded geographic information on Sciurus aberti from the open access platform Global Biodiversity Information Facility (GBIF). Subsequently, we eliminated duplicate records, those without precise coordinates and those with uncorroborated collection information. Once the unique occurrence records were obtained, we filtered the geographic space in order to avoid overpredictions in the spatial distribution models caused by the agglomeration of occurrence points. The filtering consisted of superimposing the records in the geographic space, then we made 5 latitudinal slices of 5 degrees of separation between each one, from latitude 20 degrees south to latitude 45 degrees north, taking as criteria the known distribution of the species and the ecoregions of North America, level 1 (downloaded from the US Environmental Protection Agency, 2010). Once the geographic space was segmented, we left the same number of records for each band (30 records per band) obtaining a total of 150 unique records of presence.

#### **Environmental characterization**

The environmental information to generate the models, from the present and the reconstruction to the past, was obtained from Pastclim 1.2, an R statistical software package designed to download and manipulate paleoclimatic datasets. In our particular case we chose the Beyer at al. Beyer, Krapp, and Manica (2020) set as it has environmental information available up to 120 000 BE, significantly evolutionary time, at 2000 year time intervals based on Global Circulation Models HadCM3 (Singarayer and Valdes 2010) and HadAM3H (Valdes et al. 2017). The environmental dataset has 17 bias-corrected bioclimatic variables, reduced spatial scale and spatial resolution of 0.5°square cells.

# Accessibility region (M).

Species distribution models according to the BAM diagram (Soberon and Peterson 2005) states that the geographic range occupied by a species ( $\mathbf{G}_o$ ) is the region of appropriate assemblages in terms of abiotic conditions. The region  $\mathbf{M}$  represents in geography areas in which the species has access due to its colonization capabilities and the structure of geographic barriers within a specific period of time (Soberón and Nakamura 2009).

In this study we selected a set of North American ecoregions for modeling to delimit our Accessibility region. We took into account ecoregions that overlap with the known distribution of the species and also took into account neighboring ecoregions to increase the potential accessibility in projection scenarios to the past. Thus, regions 5, 6, 9, 10, 12, 13 of the level I ecoregions (https://www.epa.gov/) were selected.

#### Niche models

Current distribution Estimating niche boundaries for species occurrence is called ecological niche modeling (ENM), and when the emphasis is on geographic distribution, it is known as species distribution modeling (SDM) (Guisan and Thuiller 2005; Peterson and Soberón 2012; Saupe et al. 2012). We model the current distribution of S. aberti using the MAXENT 3.4.4 algorithm (Phillips, Anderson, and Schapire 2006), which is a correlative model based on the maximum entropy principle used to estimate species distributions. The output of Maxent is a relative occurrence rate (ROR) interpreted as a probability of habitat suitability given observed environmental conditions by presence-only points (Merow, Smith, and Silander Jr 2013).

The parameterization was performed with the default values of the program with the exception of the "extrapolate" and "clamping" options to avoid artificial extrapolations in the extreme values of the climatic variables used in the models. To calibrate the models we used 70% of the records and the remaining 30% to validate them.

In order to give statistical certainty and to consider uncertainty, as well as to reduce overfitting and to have a suitable model to extrapolate to past scenarios, 10 replications were performed for each model using a cross-validation with the training data. In the end, the average model of these replications was considered as the result.

We evaluated the statistical significance of the resulting models for the present using the partial ROC test

(Peterson, Papeş, and Soberón 2008), which is a modification of the ROC (receiver Operation Characteristic) test. The results of this test are proportions (ratios) of the area under the curve of the model with respect to a null model, product of repetitions that allow to statistically evaluate the areas under the curve (AUC) in relation to that expected by chance (Peterson, Papeş, and Soberón 2008), where a value derived by chance would be 0 and an acceptable value, according to the proportion of minimum omission errors tolerated in the model, would be greater than 1. A different set of test data was used on the average model of the replicates and in this way the reduction of overfitting was guaranteed.

**Reconstruction to the past** The average model obtained from Maxent was extrapolated to environmental conditions into the past. Environmental layers were taken from pastclim (Leonardi et al. 2023) in 2000 year time slices to cover a total of 120 000 years. For each scenario into the past, 10 replicates were cross-validated and averaged.

Analogy of environmental conditions To be certain that areas in the past are climatically analogous to the present we implemented the mobility-orient parity (MOP) test. This method generates a value from 0 to 1 according to the environmental analogy of past conditions to the present to eliminate geographic areas in the projections to the past outside of climates analogous to the present (Owens et al. 2013).

# Suitability Prevalence Area (SPA)

Finally, to generate the SPA, both the average projection of the present and all the average projections of the past are considered. The output files are read in raster format.

A harmonic mean is performed as follows: For each layer the inverse of each cell value is obtained. Then the inverse values are summed by cells and divided by the number of scenarios. Finally, the inverse of each cell is obtained. It is important to consider that the Maxent outputs may contain very small values close to zero in many cells, for this reason we added a small amount of 0.001 to all the values in all the maps so that when obtaining the harmonic mean we could have an inverse value for values close to zero without altering the biological sense of environmental suitability.

## Endemic historic Area

To calculate the historic endemic area, the SPA result is taken and a threshold is applied to generate a binary map according to the **S** equation. This threshold is the proportion of time that pixel has had ideal conditions. For our study, we consider the value of 90% since for a period of 120 000 years it is a coverage of 100 000 years, which is a significantly evolutionary time to fix genetic patterns at the population level.

#### Fixation index as a functino of SPA

To test our hypothesis of genetic diversity as a function of environmental suitability prevalence, from a set of genetically structured populations, a matrix of fixation indices between populations is considered. For this work we considerer the data from @Bono et al. (2018). It consideres 10 populations along the known distribution and provides values of  $F_{st}$  between this populations in matrix form. The matrix was transformed into a table as pairs between populations with index values between each pair. The values of the population with itself, i.e. 0, were included. Subsequently, georeferences were obtained for each population considered in the fixation index matrix. From this geo-referencing the SPA value for each population was extracted. Finally, the SPA value was aligned in the table of fixation index pairs. In total the table has 4 columns: source population, target population, fixation index and SPA of the target population. This table was separated by source population groups to generate a set of linear models between SPA values and fixation indexes  $F_{st}$ . An analysis was performed for each source population with the hypothesis that the correlation between SPA and fixation index exists and is non-zero if a dispersal process has occurred from the historical endemic area of the taxon.

## Fixation index projected in geography

The linear model that presents the highest value of  $R^2$  is considered suitable to make a geographic projection of the fixation index and observe genetic diversity in geographic space. To generate this map, the SPA values are taken and a table of coordinates and SPA values is generated for each cell of the raster. The SPA values are then extrapolated to obtain fixation index values from the linear model obtained previously. This column is added to the coordinate table and a raster is generated with the projected  $F_{st}$  values. Finally, a cutoff of this raster is performed using a map of the distribution of the species at present with a cutoff threshold of 90% in binary format (where 1 =presence and 0 = absence), since the records of the presence localities were obtained from bases that might contain some errors in the data (Phillips, Anderson, and Schapire 2006).

## Results

# Current and past distribution

We show the species distribution model for *Sciurus aberti* along the time line of the projected area per time slice. We show 5 notable moments in time line. and 5 notable points

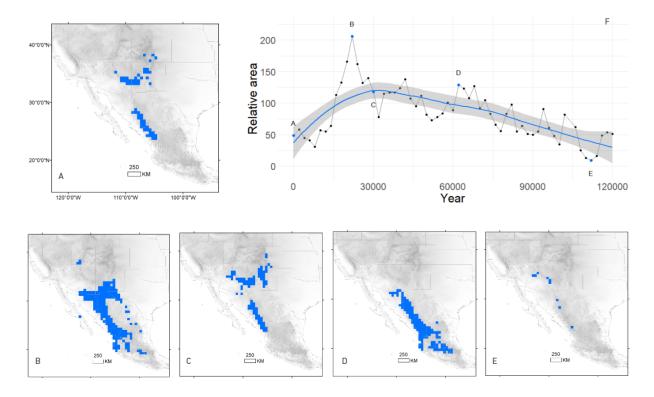


Figure 3: A) Species distribution model for *Sciurus aberti* present distribution. B) Relative projected area for SDM in pixels according a threshold of 0.9 suitability for each past time-sliced environmental conditions. We can observe a maximum area corresponding to the last interglacial era and an inverted U-shape in the trend of the area along the time considered. Also It can be noted that the actual distribution has similar area to the coditions 120 000 years ago. B) and D) corresponds to distribution peaks along the time scale and C) and E) corresponds to minimum

## SPA and Endemic Historic Area

We get the suitability presence area (figure 2) delmited by a 0.5 and 0.9 threshold (Figure 4). The endemic region where the enviornmental conditions 90 % del tiempo is the Sierra Madre Occidental, in the limits of Sonora, Chihuahua, Sinaloa and Durango. We correspond this Area to the *Sciurus aberti barbieri* subspecies.

Also is notable that at 0.5 threshold, other regions seems to have stable conditions. We observe the area close to Coronado National Forest as well as Gila National Forest and Mescalero Reservation as suitable areas along time in the United States and some Areas of the Volcanic Belt in Mexico (Puebla, north of Guerrero State and Jalisco.

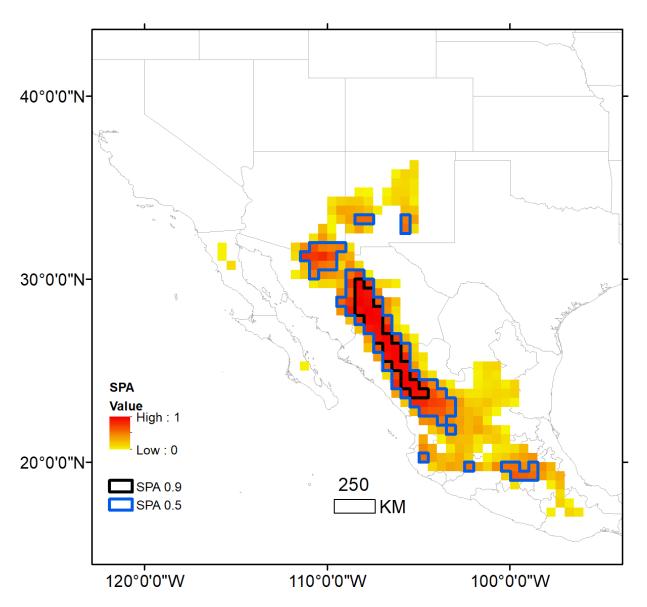


Figure 4: Suitability prevalence area. Here we observe the harmonic mean of suitability for each pixel along the distribucion from 0 to 1. The bold lines shows the area when the harmonic mean es above 0.5 threshold (blud line) and 0.9 threshold (black line)

## Fst as a function of SPA and Fixation index projected in geography

For all the linear regressions performed (Table 1), the only one that had statistical significance is the one for the Sciurus aberti barberi population with an R-squared value of 0.794.

Table 1: R squared from linear models (fixation index as a function of suitability prevalence area). The only R squared with statistical significance is barberi

population	r.squared	p.value
S. a. aberti Carson-SFW	0.000	0.962
S. a. aberti Coconino-Gila	0.131	0.304
S. a. aberti MT-Zuni	0.002	0.894
S. a. aberti San Juan	0.000	0.952
S. a. barberi	0.794	0.001
S. a. chuscensis E	0.002	0.892
S. a. chuscensis W	0.018	0.713
S. a. ferreus Carson E	0.040	0.581
S. a. ferreus Pike	0.000	0.967
S. a. kaibabensis	0.005	0.841

It can be noted that S. a. aberti Coconino-Gila had an R-squared value greater than 0 (0.131) but it is not statistically significant. Furthermore, in our approach we imputed the value of the origin with an Fst of zero. To show that the model hypothesis still holds, we performed a regression for barberi without the point of origin imposed (Figure 5b), in addition to eliminating the S. aberti chuscensis populations due to their low uncertainty in georeference, obtaining an R-squared of 0.72 (p value = 0.016).

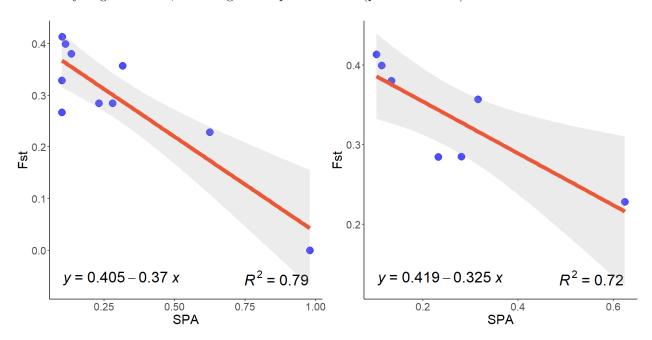


Figure 5: Linear regression for S a. barberi population A) SPA in all populations B) withouth barberi and chusensis populations. We can observe that in both cases the fixation index is explained with the SPA and the hypothesis holds without the imputation of the origin of the population (this means to add a 0  $F_{st}$  value for barberi population)

From the model we extrapolated the values from SPA along the geography to new  $F_{st}$  values to get a geographical projection of fixation index from S.~a.~barberi population. (Figure 6)

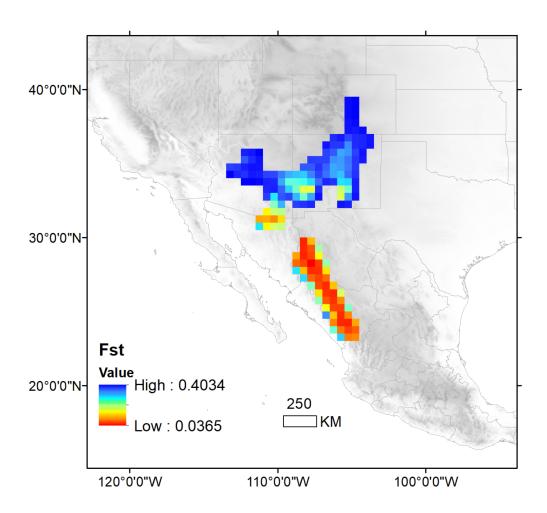


Figure 6: Projection of the  ${\cal F}_{st}$  model to the geography in the actual distribution of S aberti

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