

Doctoral Dissertation

Tracking atmospheric chemical components in accordance with the Sustainable Development Goals (SDGs)

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Tracking atmospheric chemical components in accordance with the Sustainable Development Goals (SDGs)*

Phan Anh

Abstract

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Keywords:

π , astronomy, mathematics, computer, algorithm

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Nara Institute of Science and Technology, November 16, 2023.

Contents

List of Figures	v
List of Tables	vii
1 Introduction	1
1.1 Background	1
2 Background	2
2.1 Air pollution	2
2.2 Greenhouse gas	2
AIR POLLUTION INDUCED BY INTERVENTION EVENTS	3
3 Ukraine's case study	4
3.1 Introduction	4
3.2 Data	7
3.2.1 Selection of analysis periods	7
3.2.2 TROPOMI NO ₂ from Sentinel 5P	8
3.2.3 Meteorological and surface NO ₂ data	11
3.2.4 Fire spots database and Ukraine crisis hub	12
3.2.5 Population data	13
3.3 Business-as-usual (BAU) modelling	13
3.4 COVID-19 induced NO ₂ changes	17
3.4.1 Lockdown and pre-lockdown meteorological patterns	18
3.4.2 NO ₂ changes in populous Ukrainian cities	19
3.5 NO ₂ changes induced by the armed conflict	23
3.5.1 S5P NO ₂ level changes in conflict hotspots	24
3.5.2 Changes of S5P NO ₂ levels in other affected areas	26

3.6 Conclusion	32
4 Japan's case study	35
4.1 Introduction	35
4.2 Data	39
4.2.1 Study area	39
4.2.2 Ground observation	39
4.2.3 ERA5 reanalysis dataset	41
4.2.4 Sentinel 5P TROPOMI	41
4.2.5 Biogeochemical modelled CH ₄ budget	41
4.3 Method	42
4.3.1 Business-as-usual (BAU) modelling	42
4.3.2 Experiments design	45
4.4 Results	45
4.4.1 NO ₂ level changes	45
4.4.2 O ₃ level changes	48
4.4.3 CH ₄ level changes	53
4.5 Discussion	56
4.5.1 Variations in spatial resolution of multisource data	56
4.5.2 Limitations	57
4.6 Conclusion	58
GREENHOUSE GAS ESTIMATION AND MONITORING	60
5 Plant functional types mapping	61
5.1 Introduction	61
5.2 Data	62
5.2.1 Study area	62
5.2.2 Data collection	62
5.3 Methodology	64
5.4 Experimental results	67
5.5 Conclusion	72
6 Global upscaled of carbon fluxes	74

7 CO2 monitoring platform	75
7.1 Introduction	75
7.2 Method	75
7.2.1 Data collection	75
7.2.2 API development	75
7.3 Result and discussion	75
7.3.1 Result	75
7.3.2 Discussion	75
8 Conclusion	76
Bibliography	78

List of Figures

3.1	Monthly map of S5P NO ₂ in Ukraine	9
3.2	Monthly number of qualitfied S5P NO ₂ observations in Ukraine	10
3.3	Feature importance estimated using LightGBM split method.	14
3.4	Model performance evaluation	15
3.5	Examples of OBS and BAU data in 2022	16
3.6	Meteorological variations during pre-lockdown and lockdown	19
3.7	S5P NO ₂ level changes for most populous cities	21
3.8	Analyzed conflict hotspots using satellite and ACLED data	25
3.9	NO ₂ level changes estimates for conflict events	27
3.10	OBS and BAU S5P NO ₂ trends (2020-2022) in populous cities	28
3.11	OBS and BAU S5P NO ₂ trends (2020-2022) for selected CPPs	31
4.1	Mobility changes for 6 prefectures in Japan	37
4.2	Study area	40
4.3	Model performance evaluation	43
4.4	NO ₂ reduction trends in 2020	47
4.5	NO ₂ , O ₃ , SR, T2M, FNR, and HCHO variations in 2020	50
4.6	2020 O ₃ mean trends (4 MAs)	51
4.7	NO ₂ , CO and CH ₄ , and HCHO variations in 2020	54
4.8	2020 CO and CH ₄ mean trends (4 MAs)	55
5.1	Study area and annotated data	63
5.2	Study area and annotated data	65
5.3	Forest map in Ena City, Japan.	69
5.4	Study area and annotated data	70
5.5	Inferred tree age map in Ena City	71

5.6 Inferred tree species map in Ena City	72
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List of Tables

3.1	Model performance evaluation	14
3.2	The hyperparameters used to develop the LightGBM model	17
3.3	S5P NO ₂ level changes estimates for most populous cities	22
3.4	Comparison of NO ₂ changes during 2020 and 2022 lockdown period	29
3.5	S5P NO ₂ variations in populous cities from February to July 2022	30
4.1	Model performance evaluation	44
4.2	short	46
4.3	short	49
4.4	short	56
5.1	Samples, weights for cross-entropy loss training	66
5.2	The experimental results of UNET and our model.	68

1 Introduction

1.1 Background

2 Background

2.1 Air pollution

2.2 Greenhouse gas

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AIR POLLUTION INDUCED BY INTERVENTION EVENTS

3 Ukraine's case study

3.1 Introduction

Nitrogen dioxide (NO_2) is a key air pollutant that can have harmful effects on human health. An increase in nitrogen oxide ($\text{NOx} = \text{NO} + \text{NO}_2$) concentrations contributes to global warming through a chemical reaction that leads to the formation of ozone (O_3), a short-lived climate pollutant with a potent warming effect (Stocker et al., 2013). The lifetime of NO_2 is strongly influenced by photochemical reactions and meteorological parameters (Barré et al., 2021) and varies seasonally (Dragomir et al., 2015; Kendrick et al., 2015). During winter, photochemical reaction activity is reduced, resulting in a longer lifetime of the NO_2 . Additionally, seasonal variations in NO_2 concentration are controlled by dispersion processes which are significantly affected by changes in boundary layer height (BLH), wind speed and direction patterns due to temperature inversions in summer and winter (Barré et al., 2021; Kendrick et al., 2015). NO_2 concentration levels have been widely used to evaluate decreases in emissions associated with intervention events such as the COVID-19 pandemic lockdown and impacts on the air quality due to the short lifetime of NO_2 in the atmosphere (Barré et al., 2021; Cooper et al., 2022). In Europe, anthropogenic NOx emissions are mainly attributed to combustion processes in transportation, as well as energy production and distribution.

In Ukraine, coal-fired power plants (CPPs) dominantly account for 80% of total SO_2 and 25% of total NOx emissions, and some have been identified as the highest-emitting CPPs in the region and in the world (Lauri and Rosa, 2021). Since the pandemic started in March 2020, and now with the ongoing armed conflict with Russia, Ukraine has faced a series of threats to the economy, human security and the environment, as well as geopolitical tensions (Pereira et al.,

2022). During the pandemic response starting in 2020, many national and local lockdown restrictions were issued to prevent the spread of the virus, causing a sharp decrease in gross domestic product growth rate, as well as industrial and energy production (Danylyshyn, 2020). In 2021, Ukraine's economy started to recover from the pandemic but the recovery was eventually upended by an armed conflict with Russia that started on February 24, 2022. The conflict has been causing a multi-pronged crisis not only in Ukraine but also in Europe, with increased prices and exacerbated inflation among the many impacts. Many facilities and extensive areas of housing and other infrastructure, including some CPPs, have been reported destroyed or damaged in Ukraine. These impacts have consequently triggered an unprecedented refugee crisis in Ukraine, clogging border crossings between Ukraine and bordering European countries (Júlia et al., 2022). The many socio-economic changes that have occurred during the pandemic and the conflict could be expected to contribute to major variability in air quality in Ukraine, including NO₂ pollution levels, during the 2020–2022 period.

A report by the United Nations Development Programme (UNDP) (Dumitru et al., 2020), estimated the impacts of the pandemic lockdown on NO₂ levels in Ukraine by using Sentinel 5P (S5P) NO₂ column concentrations and Copernicus Atmosphere Monitoring Service (CAMS) surface NO₂ data (Marécal et al., 2015). However, meteorological variables were not acknowledged, although ignoring weather factors could strongly affect final estimates of changes in pollution concentration levels induced by the lockdown (Schiermeier, 2020). A more recent study (Zalakeviciute et al., 2022) utilized direct satellite observation from 2019 and early 2020 as business-as-usual data to evaluate the impact of the Russia-Ukraine conflict in 2022 on air quality, but again, without acknowledging weather effects. These two studies utilized estimates of year-to-year differences. However, such estimates can easily be affected and dominated by changes in meteorological parameters rather than emission sources (Grange et al., 2021; Shi et al., 2021). Therefore, a more sophisticated method is needed to measure the impacts of intervention events through better quantification of actual air quality.

In order to normalize the meteorological effects to accurately and reliably quantify the impact of intervention events, the use of machine learning is increasingly being adopted, but mostly applied for ground-based measurements following the

original idea proposed by (Grange et al., 2018) and (Grange and Carslaw, 2019). The objective of this approach is to construct a business-as-usual (BAU) model for predicting air pollution levels independently of the impacts of any intervention events. This is achieved by integrating meteorological, spatial, and temporal features into the model during the BAU period to accurately represent air pollution levels. An intervention event, in this context, refers to an occurrence that has caused changes in air quality. Recently, (Barré et al., 2021) have introduced their weather normalization approach to improve estimates of lockdown impacts not only on NO₂ levels from ground-based observations and CAMS simulations, but also in satellite measurements from S5P. The original method in (Grange et al., 2018; Grange and Carslaw, 2019) has been altered in order to work with satellite retrieval column NO₂ concentration levels from S5P by adopting a new feature, the forecast surface NO₂ level from CAMS data. Alternatively, gradient boosting machines (GBMs) (Friedman, 2001) have been also utilized instead of random forests (Grange et al., 2018) to develop weather-normalization models under the BAU conditions. (Barré et al., 2021) reported an overall reduction (ranging from 23% to 32%) in major European cities using the three datasets. Their study showed an average difference of 14% between satellite-based and ground-based estimates, and 11% between simulations from the CAMS regional ensemble of air quality models and ground-based estimates. These findings suggest that estimates of the impacts of the lockdown on NO₂ levels can vary depending on the source of the data.

This chapter aims to investigate the actual satellite-derived column NO₂ pollution levels induced by pandemic lockdown restrictions and the armed conflict with Russia, which have been two major changes in human activities in Ukraine since 2019. In order to do so, we developed a weather-normalization model under BAU scenarios for S5P column NO₂ levels to decouple the meteorological effects from the intervention effects. The BAU simulation NO₂ levels are then used to quantify changes in S5P column NO₂ concentrations during the lockdown and the armed conflict. We describe the data used in the study in section 3.2 and the methodology in section 3.3. The results and discussion on NO₂ level changes are summarized in section 3.4 for the lockdown, and section 3.5 for the armed conflict. Finally, we conclude the results of the study in section 3.6.

3.2 Data

3.2.1 Selection of analysis periods

In this study, we consider the three years 2019, 2020, and 2022 for our analysis. We assumed that in 2019, before the lockdown in 2020 and the armed conflict with Russia in 2022, there were no other significant factors impacting socio-economic activities. Hence, we used 2019 NO₂ pollution levels as the reference data for development of the BAU model.

Ukraine reported its first active case of COVID-19 on March 3, 2020, and began closing its borders to foreign citizens from March 15 onwards. Around the same time, the country also witnessed its first COVID-19 related death. On April 6, the government introduced a strict lockdown, imposing significant restrictions on movement and requiring the public to wear masks in public spaces. This lockdown was eventually extended until June, although certain restrictions were already lifted starting from May 11. For the lockdown component of our study, we focused on two specific periods: the pre-lockdown period, which ran from March 1 to 15, 2020, and the strict lockdown period, spanning from April 6 to May 10, 2020. The decision to count the pre-lockdown period from March 1 was based on the lack of qualified S5P data available for analysis before March, as indicated in 3.2. In 2021, even though COVID-19 vaccines had been developed and distributed to citizens of Ukraine (vaccinations started on February 24, 2021), many local lockdowns and restrictions continued to be issued to cope with growing numbers of daily COVID-19 active cases, while trying to keep socio-economic activities on track for recovery.

The Russia-Ukraine conflict began on February 24, 2022. We employed data for the period February 1 to July 31 each year from 2019 to 2022 for NO₂ variability analysis. This time frame covers the pre-lockdown and lockdown periods in 2020 and extends beyond the first five months (February 24 to July 31) of the armed conflict in 2022.

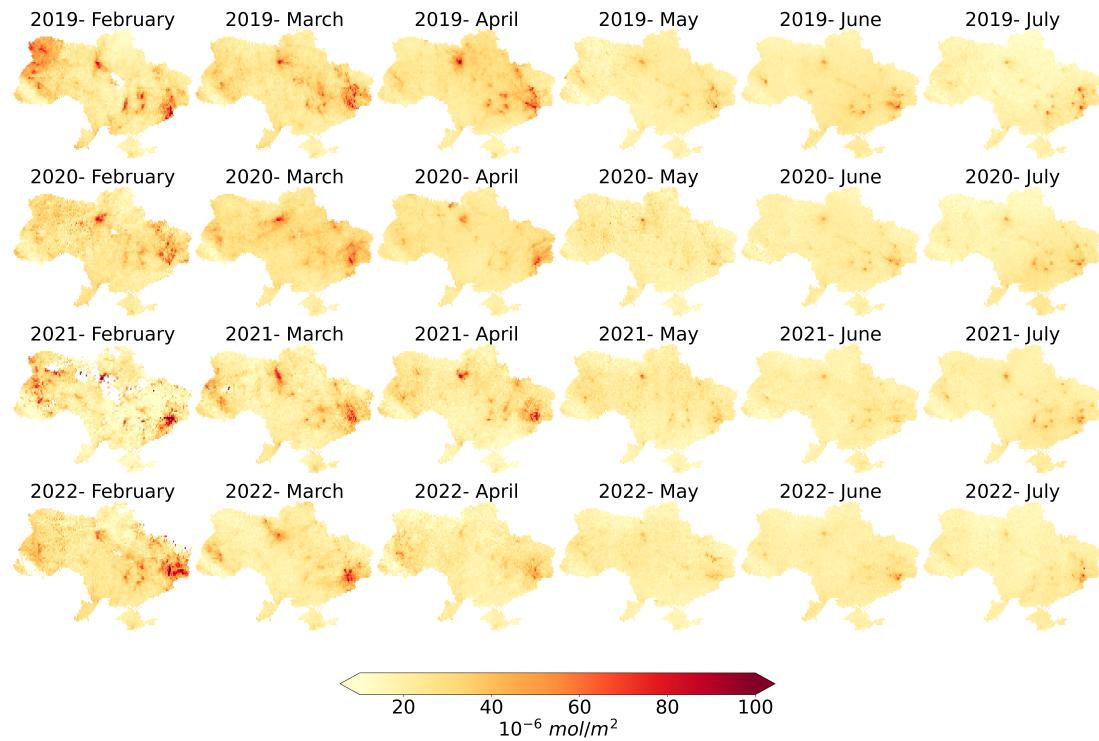
3.2.2 TROPOMI NO₂ from Sentinel 5P

Most previous studies assessing the impacts of intervention involved ground observations in their analysis. However, reliable ground measurement data was only available in Kyiv (capital of Ukraine) as other sites had been damaged or destroyed in the armed conflict and taken out of service (Savenets, 2021). Thus, open satellite data is considered the most efficient way to monitor air quality for all parts of Ukrainian territory (Shelestov et al., 2021).

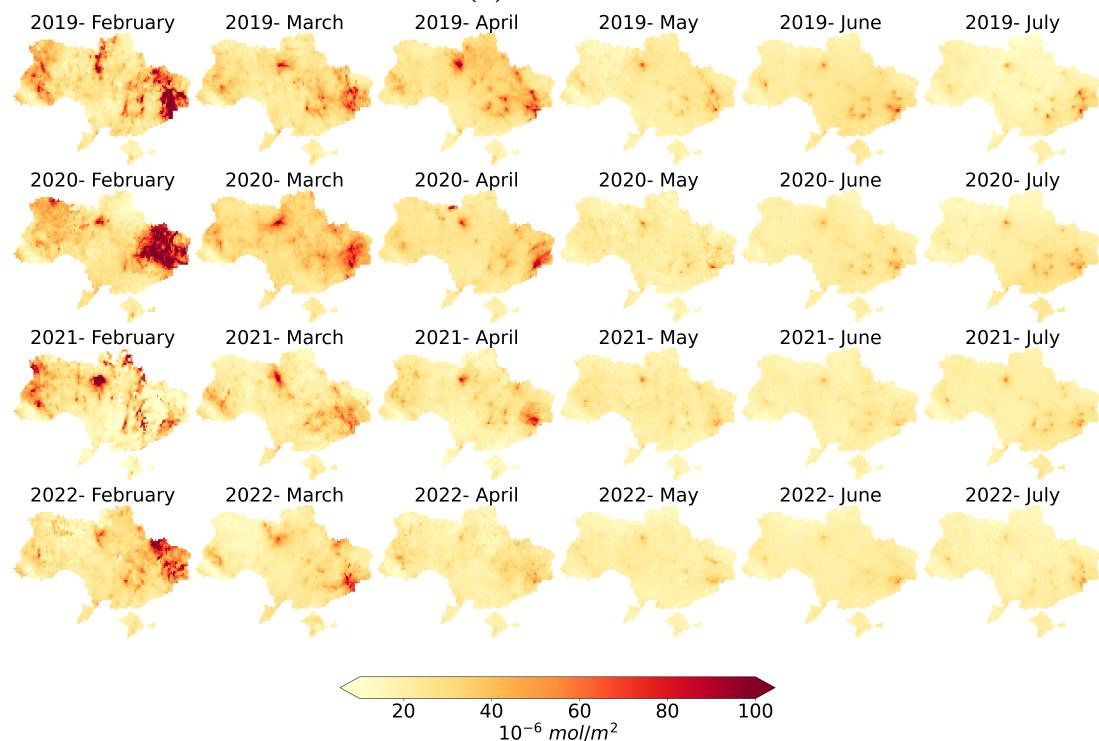
The S5P data has been distributed from 2018 to the present with two available options. The first is original data (ORG) processed with either of two versions of processor, v1.x (5/2018–6/2021) or v2.x (7/2021 onwards). The second is reprocessed datasets (RPRO) with the processor (v2.x) for the full mission. According to (Van Geffen et al., 2022), the S5P NO₂ v2.2 data has larger vertical column density (VCDs) than v1.x data, ranging from 10% to 40%, mostly found at mid and high latitudes in winter. Therefore, bias between S5P v1.x and v2.x could lead to overestimation and underestimation when comparing air pollution data in 2022 versus 2019, thereby affecting evaluations of the conflict's impacts on S5P NO₂ levels.

In this study, we conducted experiments using two versions of S5P NO₂ data. The first dataset is ORG data which was collected through level 3 (L3) offline processing (OFF) of the S5P product available on Google Earth Engine (Gorelick et al., 2017). This dataset comprises processed data from different processor versions for each year from 2019 to 2022 (v1.3.1 in 2019, v1.3.2 in 2020, and v2.3.1 in 2022). The second dataset, denoted as the RPRO product, employs processor version v2.4.0 for the full mission duration. This dataset was acquired from the Sentinel-5P Pre-Operations Data Hub (s5phub.copernicus.eu) using the Sentinelsat API.

Regarding the RPRO data, we began by downloading the level 2 (L2) dataset. In order to generate the L3 NO₂ dataset, each operational L2 product underwent mosaicking and filtering of low-quality pixels, which involved removing items with quality assurance (QA) values less than 75% for the “tropospheric_NO₂_column_number_density” band. The harpconvert tool was utilized to perform the conversion from L2 to L3 product. Subsequently, both datasets were linearly interpolated to a spatial resolution of 0.1×0.1 degree. At

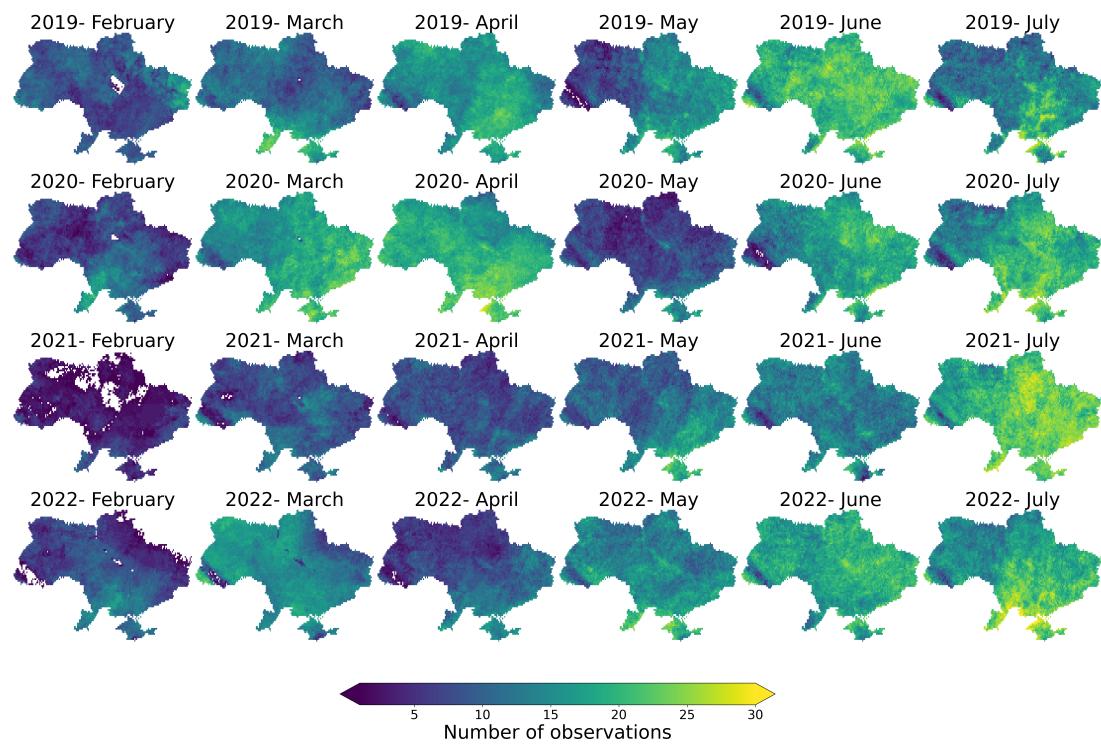


(a) ORG data

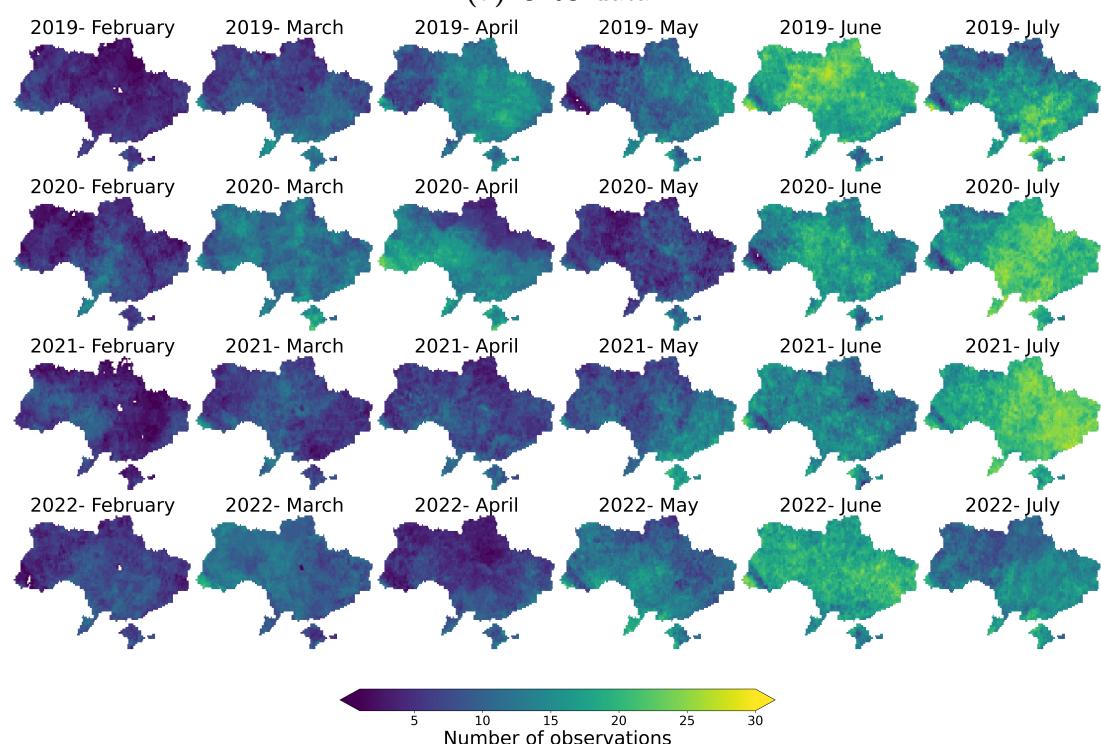


(b) RPRO data

Figure 3.1. Monthly (February to July) average map of TROPOMI S5P NO₂ tropospheric columns for Ukraine from 2019 to 2022



(a) ORG data



(b) RPRO data

Figure 3.2. Monthly (from February to July) number of TROPOMI S5P NO₂ tropospheric columns observations for Ukraine from 2019 to 2022

the time of the experiment, the RPRO data was only accessible until July 2022.

Plots presented in Figure 3.1 display the average monthly TROPOMI NO₂ tropospheric column over Ukraine from 2019 to 2022 (February to July) using the ORG data (Figure 3.1a) and RPRO data (Figure 3.1b), respectively. In 2020, a reduction of 4.8% (ORG data) and 8.3% (RPRO data) in mean NO₂ levels over the Ukrainian territory was observed from April to May, compared to levels recorded in 2019. In 2022, a reduction of 2.4% (ORG data) and 2.9% (RPRO data) was seen from March to July, compared to levels recorded in 2021. Additionally, during the same period, a reduction of 10.3% (ORG data) and 15% (RPRO data) was observed, compared to the NO₂ levels recorded in 2019. We observed that the reduction in NO₂ levels was more significant in the RPRO data compared to the ORG data, both during the lockdown in 2020 and the first five months (March–July) of the conflict in 2022 in Ukraine.

We summarize the number of qualified observations available for each month from 2019 to 2022 (February to July) in Ukraine using the ORG data (Figure 3.2a) and RPRO data (Figure 3.2b). The quantification of seasonal NO₂ levels can be challenging, particularly during the selected months in winter (February) and spring (March, April) of 2021 and 2022, due to the limited availability of qualified observations. This is further complicated when attempting to estimate changes before and after intervention events such as the lockdown and the armed conflict in Ukraine, as the before period falls within the winter months when observations are scarce.

3.2.3 Meteorological and surface NO₂ data

In this study, the meteorological and surface NO₂ data are utilized as the predictors for the estimation of NO₂ under BAU conditions as suggested by (Barré et al., 2021). The meteorological data is ERA5 reanalysis data which is collected from the Climate Data Store of the Copernicus Climate Change Service (Hersbach et al., 2018). We use the following weather variables: 10 m wind speed (u and v component, m/s) and direction (degrees), 2m air temperature (K), 2m dewpoint temperature (K), relative humidity (%), geopotential (m²/s²), and BLH (m). All the variables are downloaded at the original resolution of 0.25×0.25 degree and then linearly interpolated to 0.1×0.1 degree (about $10\text{km} \times 10\text{km}$) resolution. The

utilized surface NO₂ data is collected from CAMS European air quality forecast and reanalyses and forecast (Marécal et al., 2015) by using the Atmosphere Data Store of the CAMS (<https://ads.atmosphere.copernicus.eu/>). Since the forecast data is a 3-year rolling archive from the present, we utilized the analysis data for 2019. The surface NO₂ forecast data served as the predictors under the BAU scenario for 2020 to 2022. As forecast predictions do not involve an assimilation process (Barré et al., 2021), we expect no effect of the pandemic lockdown, and the impact of the armed conflict related events on air pollution was included in the surface NO₂ pollution level. Both forecast and analysis data are available at the resolution of 0.1×0.1 degree. We calculated the mean values based on data from 13:00 and 14:00 hours local time to represent the surface NO₂ and meteorology value at the time the satellite S5P overpassed Ukraine.

3.2.4 Fire spots database and Ukraine crisis hub

In order to draw a detailed picture of the battle spots, we utilized data from Fire Information for Resource Management System (FIRMS) provided by National Aeronautics and Space Administration (NASA) and Ukraine Crisis Hub data from the Armed Conflict Location and Event Data Project (ACLED) (Raleigh et al., 2010). The NASA FIRMS portal provides active fire data at three-hour intervals based on satellite observations from products of the Moderate Resolution Imaging Spectroradiometer (MODIS) and Visible Infrared Imaging Radiometer Suite (VIIRS). For the study, data from the VIIRS product was employed to access the active fire spots due to its superior fire detection capabilities compared to the MODIS products (Csiszar et al., 2014; Schroeder et al., 2014).

Detailed data on conflict hotspot locations are extracted from the Ukraine Crisis Hub which is distributed by ACLED (Raleigh et al., 2010). Information regarding the conflict events is updated weekly and disaggregated to event type with time and location (latitude and longitude) in Ukraine and the Black Sea region available from 2018 until the present. As a result of the conflict, we expect to see and identify corresponding patterns between locations of active fire spots and the locations of conflict events.

3.2.5 Population data

As NO₂ pollution levels are closely related to human socio-economic activities and frequently high in populous urban areas, we downloaded 2020 population data for Ukraine from the WorldPop Global Project (www.worldpop.org), available annually at the spatial resolution of 100m×100m as one of the features for the BAU NO₂ model. The population data was collected, clipped to the Ukrainian territory, and linearly interpolated to 0.1×0.1 degree (about 10km×10km).

3.3 Business-as-usual (BAU) modelling

When considering changes induced by the pandemic lockdown and the armed conflict, especially for before-after analysis, an important factor is the meteorology variations. In this study, we use a suggested list of predictors by (Barré et al., 2021), which consists of meteorological, spatial, and temporal features, population counts from WorldPop Global Project, and surface NO₂ pollution levels from CAMS European analysis data for 2019 and forecast data for 2020 to 2022 for BAU model development. The spatial and temporal features contain latitude, longitude, Julian date (number of the day from January 1), and day of the week, respectively. However, unlike the study cited (Barré et al., 2021), for machine learning model selection, instead of GBM we utilized LightGBM (Ke et al., 2017), which is a gradient boosting decision tree, to build the BAU model. During the training process, other than in studies that used the grid search with an n-fold cross-validation approach to tune the model's hyperparameters (Barré et al., 2021; Petetin et al., 2020), we employed the Fast Library for Automated Machine Learning (FLAML) (Wang et al., 2021), which is a new lightweight library for quickly determining the accurate model, to find the optimum hyperparameters for the LightGBM model in our case.

In order to assess the performance of the BAU simulation model, we randomly selected and used 80% of the data for the training set and 20% for the validation set. We used the following metrics: mean bias (MB), normalized mean bias (nMB), root mean square error (RMSE), normalized root mean square error (nRMSE) and Pearson correlation coefficient (R). As shown in the detailed results presented in Table 3.1, the model achieved high R on the validation set

Table 3.1. The performance of the BAU model on the validation set described using the following metrics: mean bias (MB), normalized mean bias (nMB), root mean square error (RMSE), normalized root mean square error (nRMSE) and Pearson correlation coefficient (R). N represents the number of points in both the training set and validation set, where each point is associated with unique latitude and longitude values. There are no duplicate points shared between the training and validation sets.

	MB	nMB	RMSE	nRMSE	R	n
Performance with S5P data version 1.x—ORG data						
Training set	3.68×10^{-5}	1.53×10^{-4}	7.80	7.40	0.87	5022
Validation set	0.03	0.10	9.53	10.98	0.80	1269
Performance with S5P data version 2.4—RPRO data						
Training set	2.67×10^{-4}	1.04×10^{-3}	6.97	5.12	0.91	5051
Validation set	0.07	0.26	8.47	7.75	0.86	1242

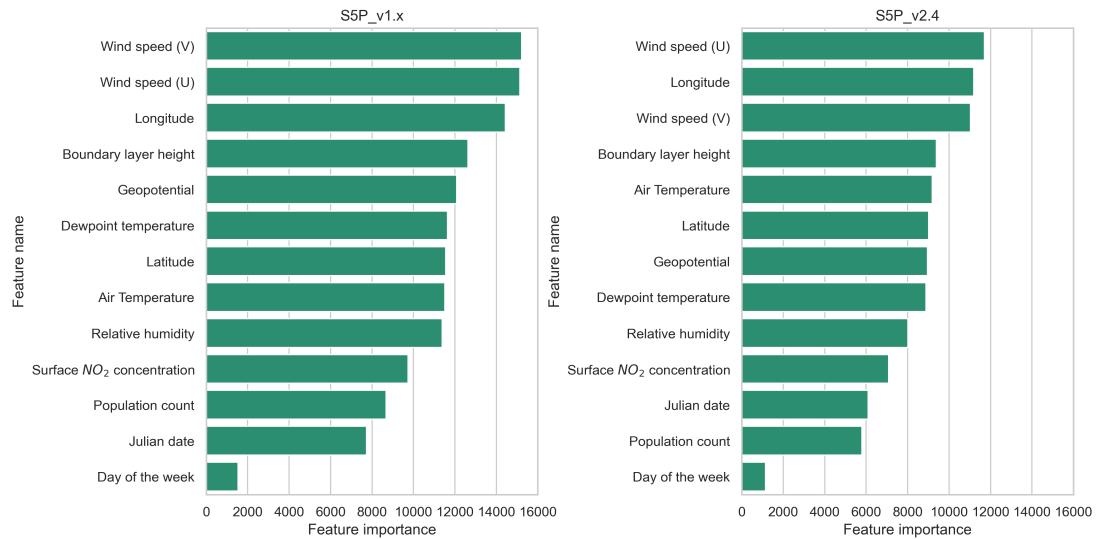


Figure 3.3. Feature importance estimated using LightGBM split method.

(0.8 for ORG data, 0.86 for RPRO data), with low MB and RMSE indicating that the column NO₂ levels are well represented by the input features. Based on the feature importance measure as shown in Figure 3.3, we found that the most

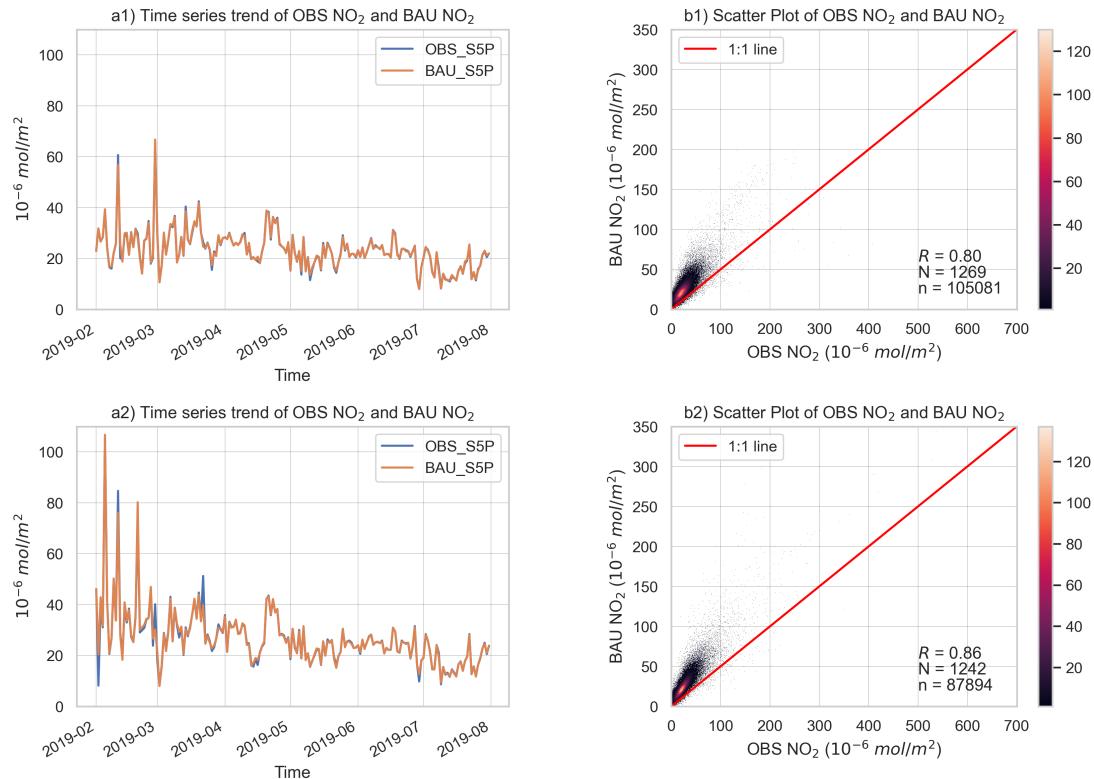
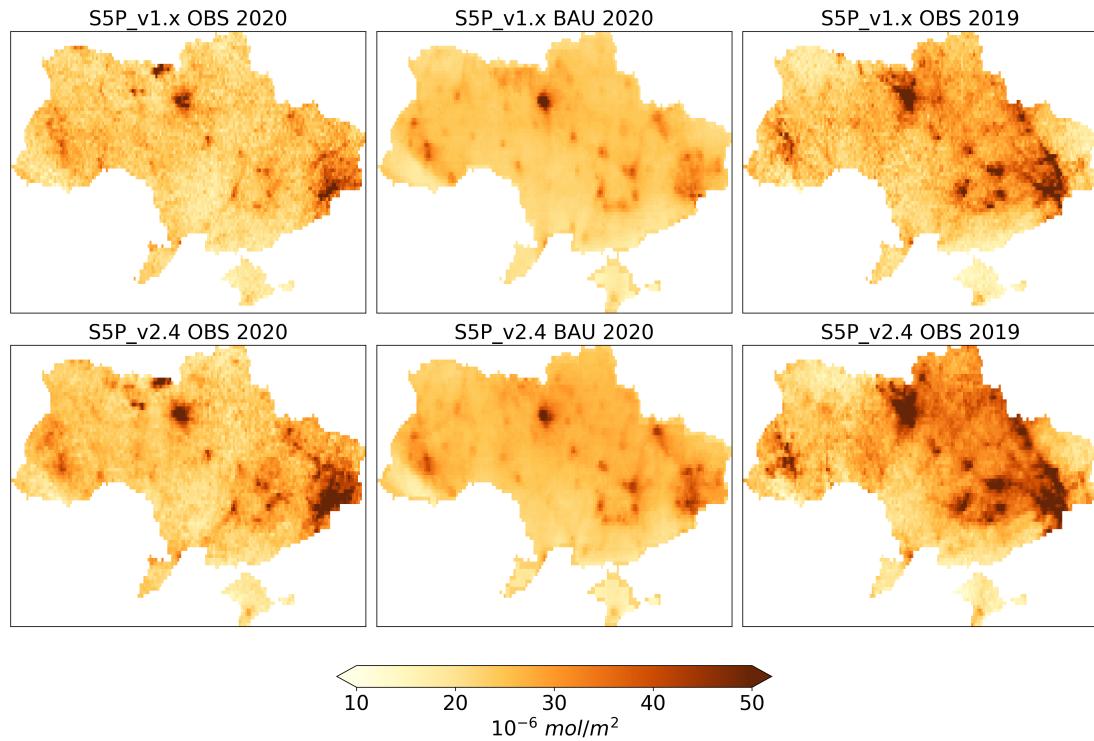
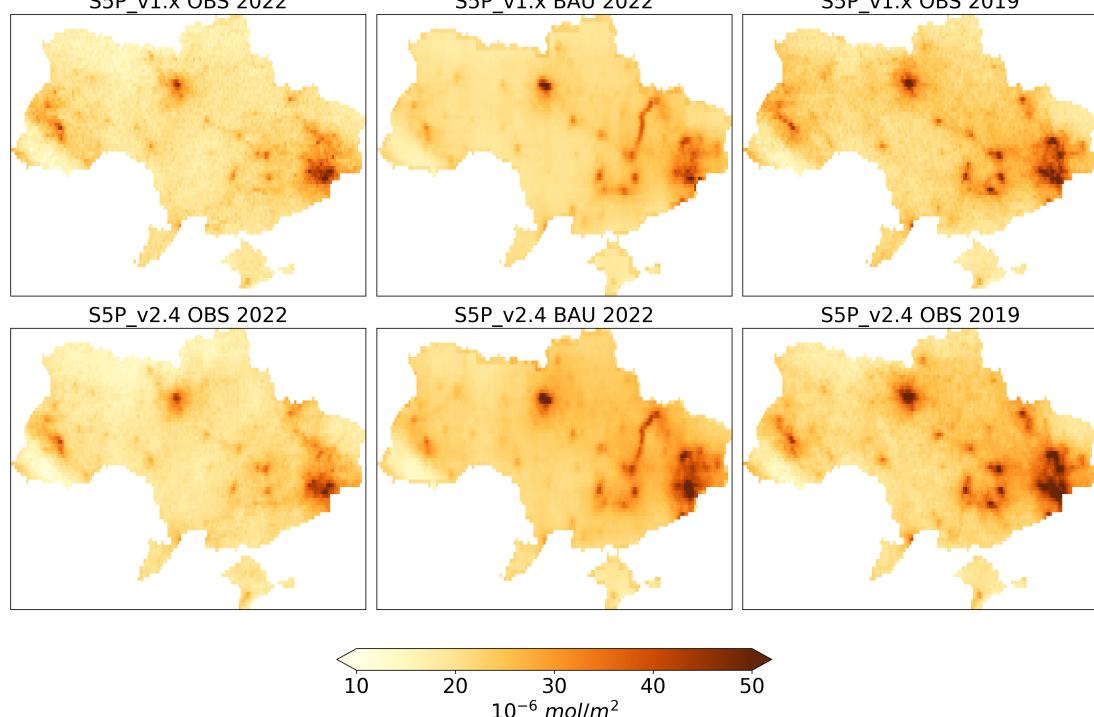


Figure 3.4. The timeseries trend lines (a1) and (a2) and scatter plots (b1) and (b2) depict the OBS_{NO_2} and BAU_{NO_2} on the validation set in 2019. Sub-figures (a1) and (b1) correspond to the S5P version 1.x data, while sub-figures (a2) and (b2) represent the S5P version 2.4 data. In the scatter plot, we showed the 1:1 line, Pearson correlation coefficient (R), N represents the number of points in both the training set and validation set, where each point is associated with unique latitude and longitude values. At each point, we used the available daily data from February 1 to July 31, 2019, to make the training and validation set with total number samples is denoted as n . There are no duplicate points and samples shared between the training and validation sets.

important predictors are wind speed and direction, and BLH, which is also consistent with our hypothesis about the impact of the meteorological parameters on column NO_2 levels mentioned above. In Figure 3.4, we present the performance of the BAU model on the validation set using trend lines and scatter plots to



(a) The OBS, BAU data in 2020 (April 6 to May 10) with reference data in 2019



(b) The OBS, BAU data in 2022 (February 24 to July 31) with reference data in 2019

Figure 3.5. The OBS (1st column), BAU (2nd column) data from April 6 to May 10, 2020 (a) and from February 24 to July 31, 2022 (b) with the corresponding reference data in 2019 (3rd column)

compare the predictions with the actual ground truth data. Furthermore, Figure 3.5 displays the OBS data, BAU model's predictions during the lockdown period in 2020, and more than five months of the conflict (February 24–July 31) in 2022. This data is accompanied by the reference NO₂ levels from 2019 which were utilized to train the BAU for corresponding periods. The hyperparameters used to develop the LightGBM model are listed in Table 3.2 for S5P data version 1.x and version 2.4.

Table 3.2. The hyperparameters used to develop the LightGBM model with S5P data version 1.x and version 2.4. We used FLAML library (Wang et al., 2021) for tuning these following parameters: shrinkage rate (learning_rate), minimal number of data in one leaf (min_data_in_leaf), minimal sum hessian in one leaf (min_sum_hessian_in_leaf), number of boosting iterations (num_iterations), max number of leaves in one tree (num_leaves).

Parameter	S5P v1.x	S5P v2.4
learning_rate	0.30775042929674906	0.3858774543125185
min_data_in_leaf	11	5
min_sum_hessian_in_leaf	0.001	0.001
num_iterations	907	3451
num_leaves	8604	4342

The main shortcoming of this method is the lack of qualified reference data to develop the weather normalization model under BAU conditions, as the S5P TROPOMI data has been only available since mid-2018. Only one year of training data in 2019 is considered relatively small, thus resulting in large errors in BAU simulations in winter months as during this time, limited qualified S5P observations are available and NO₂ pollution levels are quite unpredictable due to the inconsistency in heating activities and NO₂ intake from Poland.

3.4 COVID-19 induced NO₂ changes

The purpose of this section is to examine the effect of the lockdown on changes in NO₂ column levels in populous urban areas, namely the nine cities Kyiv, Kharkiv,

Odessa, Dnipro, Donetsk, Zaporizhzhia, Lviv, Kryvyi Rih, and Mykolaiv (listed in declining order of population). To begin, we analyse the meteorological patterns during the pre-lockdown and lockdown periods and discuss how these might influence the NO₂ levels, apart from the impacts of the lockdown measures. Next, we utilize two methods to estimate changes in NO₂ levels. The first method, known as the year-to-year approach suggested by (Barré et al., 2021), involves calculating the median value of the actual S5P observation data in 2020 and subtracting the observation data from 2019. The second method, OBS-BAU, utilizes the median value of the actual observation data (OBS) in 2020 and subtracts the simulated NO₂ levels that represent the BAU scenario, which are predicted by the S5P tropospheric NO₂ column levels without any lockdown measures. The BAU simulations are based on the representation of meteorological, spatial, and temporal parameters.

3.4.1 Lockdown and pre-lockdown meteorological patterns

Figure 3.6a and 3.6b display the probability density functions of the BLH, and Figure 3.6c and 3.6d display wind speed and direction during the pre-lockdown and lockdown periods of 2019 and 2020 based on data from the nine selected cities. In 2020, the BLH exhibited a similar distribution to that of 2019 during the pre-lockdown period, but with lower values. This decrease in BLH would have resulted in an increase in NO₂ levels in 2020 compared to 2019, as the reduced BLH restricts the dispersion of NO₂ emissions, leading to an increase in NO₂ concentration levels (see Figure 3.6a).

Conversely, during the lockdown period (see Figure 3.6b), we observed higher values of BLH in 2020 compared to 2019. This increase in BLH could have contributed to the dispersion of NO₂ concentration, resulting in a reduction of NO₂ levels during the lockdown in 2020. This phenomenon, in addition to the effects of the lockdown restrictions, may have also contributed to minimizing the NO₂ levels over major cities in Ukraine. Therefore, it is essential to consider the impacts of meteorological variables on NO₂ level variability analysis.

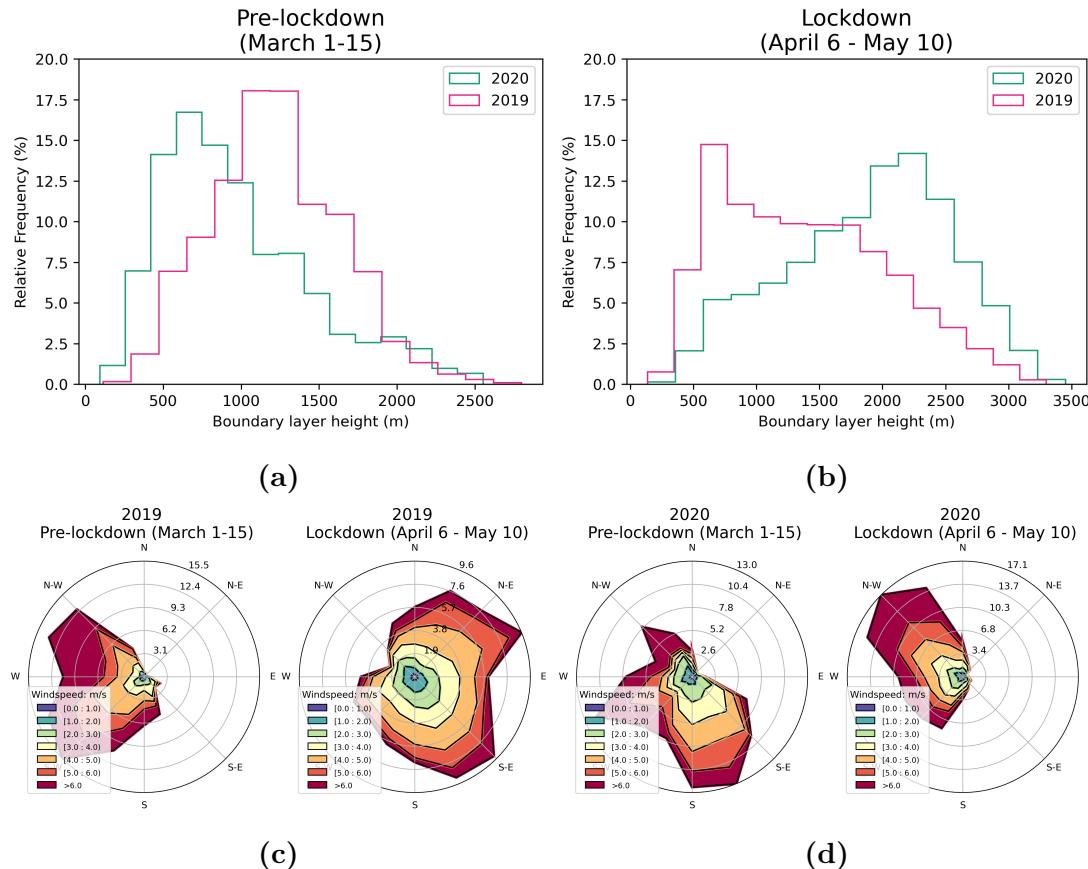


Figure 3.6. Probability density functions of BLH during (a) the pre-lockdown (March 1–15) and (b) the lockdown period (April 6–May 10) between 2019 and 2020 based on data from the nine most populous cities of Ukraine. Wind rose plots for wind speed and direction for pre-lockdown (March 1–15) and lockdown (April 6–May 10) periods in (c) 2019 and (d) 2020 based on data from the nine most populous cities of Ukraine

3.4.2 NO₂ changes in populous Ukrainian cities

In Figures (3.7a, 3.7b), and Table 3.3, we present the result of the year-to-year approach. We assumed that there would be a minimal change in NO₂ pollution levels during the pre-lockdown period, but a significant reduction during the lockdown when comparing the same time frame in 2019 and 2020 due to the implemented lockdown measures and social distancing practices. In Figure 3.7,

two different methods, namely the OBS-BAU and year-to-year approaches, were used for the analysis. The circle size in the figures corresponds to the population of each city. For each sub-figure (a) and (b), the first row (a1, a2, b1, b2) contains two plots showing the results based on the ORG data (S5P v1.x), while the second row (a3, a4, b3, b4) includes two plots presenting the results based on the RPRO data (S5P v2.4). The left column plots (a1, a3, b1, b3) of Figures (3.7a, 3.7b) display the year-to-year estimates, while the right column plots (a2, a4, b2, b4) display the OBS-BAU estimates. Figure 3.7a illustrates that the prevailing trend in the nine selected cities during the pre-lockdown period showed an increase, with an average of 5.2% (ORG data) and 13.9% (RPRO data) in NO₂ levels, while during the lockdown period (Figure 3.7b), a general reduction was observed in most cities with an average of 15.6% (ORG data) and 11.1% (RPRO data). This confirms that the lockdown measures reduced the NO₂ column concentrations in major urban areas of Ukraine, as we anticipated. It is worth noting that the year-to-year approach using the original satellite observations has been widely used in many studies and online resources. However, as mentioned in (Barré et al., 2021; Grange et al., 2021), it is heavily influenced by meteorological variables such as wind speed and direction, and BLH (Wallace and Kanaroglou, 2009).

In order to quantify the true improvement in air quality with respect to column NO₂ levels due to the lockdown restrictions, we calculated the difference between the actual observation data and the simulated data under BAU conditions with the meteorological effects decoupled. Like the year-to-year approach, we anticipate a slight variation between the OBS NO₂ levels and the BAU NO₂ levels during the pre-lockdown period. Furthermore, we expect to observe an overall reduction in the OBS data compared to the BAU data, or at least, a lesser increase during the lockdown when compared to the pre-lockdown levels, due to the impact of the lockdown measures. Figure 3.7((a2, a4) and (b2, b4)) shows the OBS-BAU estimates for pre-lockdown and lockdown in 2020. During the pre-lockdown (Figure 3.7(a2, a4)), we observed an average increase of 3.7% (ORG data) and 12.5% (RPRO data), which is smaller than the year-to-year estimate. However, during the lockdown period (Figure 3.7(b2, b4)) a smaller increase trend was observed, with an average of 0.5% (ORG data) and 10.2% (RPRO data). This indicates that while the OBS NO₂ levels in 2020 were higher than those predicted under the

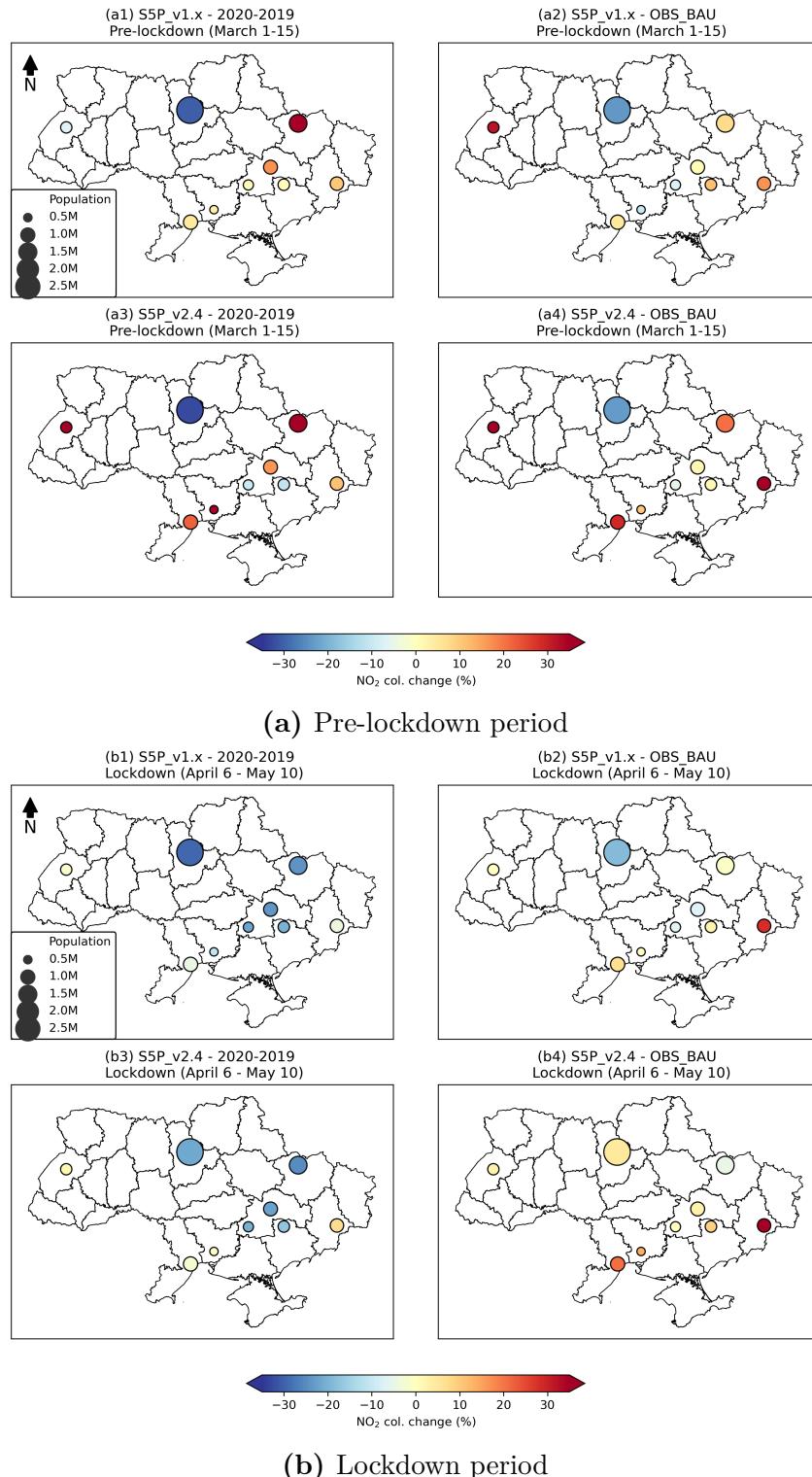


Figure 3.7. Estimates of S5P NO₂ column changes for the nine most populous cities in Ukraine during the (a) pre-lockdown and (b) lockdown periods.

Table 3.3. The OBS-BAU and year-to-year (2020–2019) estimates (in percentage) during pre-lockdown and lockdown periods in the nine most populous cities in Ukraine. The values are represented as mean, while standard deviation is not presented here due to lack of space.

City	Pre-lockdown (March 1 – 15)				Lockdown (April 6 – May 10)			
	OBS-BAU		2020–2019		OBS-BAU		2020–2019	
	ORG	RPRO	ORG	RPRO	ORG	RPRO	ORG	RPRO
Kyiv	−23.7	−23.1	−30.6	−32.8	−18.8	4.9	−29.4	−21.4
Kharkiv	7.6	20.8	47.9	49.1	−0.9	−4.9	−24.1	−24.9
Odessa	5.1	29.0	4.8	22.4	6.9	21.0	−4.4	−1.9
Dnipro	1.3	1.5	17.0	16.7	−6.6	2.8	−23.9	−22.3
Donetsk	16.8	41.9	10.3	11.2	28.2	42.0	−4.0	7.2
Zaporizhzhia	11.5	1.9	0.6	−11.1	2.5	9.1	−20.1	−17.2
Lviv	32.2	35.7	−7.3	37.7	0.0	3.0	−1.2	1.4
Kryvyi Rih	−7.3	−5.3	1.2	−9.8	−6.4	0.1	−21.9	−20.5
Mykolaiv	−10.2	10.1	3.3	41.5	−0.6	13.8	−11.1	−0.4
Mean	3.7	12.5	5.2	13.9	0.5	10.2	−15.6	−11.1

BAU scenario during the lockdown period, the measures implemented during the lockdown effectively curbed the increase in NO₂ column concentrations in major urban areas of Ukraine when compared to the pre-lockdown levels, aligning with our initial expectations. By using the OBS-BAU estimate based on the ORG data, the most significant reduction was observed in Kyiv (18.8%), with Dnipro and Kryvyi Rih experiencing smaller reductions of 6.6% and 6.4%, respectively. However, when using RPRO data, a reduction was only seen in Kharkiv (4.9%).

In comparison with the year-to-year approach with respect to the pre-lockdown (see Table 3.3), the OBS-BAU estimates (3.7% for ORG data, 12.5% for RPRO data) show a smaller change than in year-to-year estimates (5.2% for ORG data, 13.9% for RPRO data). We consider the OBS-BAU estimate to be more reasonable as mentioned above, and the lower values in BLH in 2020 could result in higher year-to-year estimates during the pre-lockdown period between 2020 and 2019. Therefore, we anticipate a lower estimate, which is a smaller increase, after the weather effects are decoupled. Similar findings are seen during the lockdown

for OBS-BAU and year-to-year estimates. The contribution from the lower BLH in 2019 could overestimate the reduction of NO₂ concentrations by 15.6% (ORG data) and 11.1% (RPRO data) in the year-to-year lockdown estimates. By normalizing the weather effects, a lower reduction in the increase is anticipated and estimated from the OBS-BAU approach (0.5% for ORG data, 10.2% for RPRO data). Additionally, the year-to-year approaches mostly present a larger standard deviation than the OBS-BAU approach, which could be attributed to local biases caused by meteorological variabilities (Barré et al., 2021). Using weather-normalization techniques, we observed that much of the reduction in NO₂ levels between 2020 and 2019 can be attributed to weather variability. This suggests that stricter measures may need to be considered in the future to achieve significant NO₂ reductions in densely populated areas of Ukraine.

3.5 NO₂ changes induced by the armed conflict

In the previous section, we discussed the influence of meteorological factors on the concentration of NO₂ and how using OBS-BAU estimates can mitigate overestimation or underestimation in the year-to-year approach. In this section, we shift our focus solely to the OBS-BAU estimates to explore the impacts of the armed conflict on NO₂ column concentration. The year-to-year estimates are displayed together for the purpose of comparison.

During the lockdown, one might reasonably assume that pollution levels were likely to decrease as the result of an anticipated reduction in socio-economic activities in major urban areas. However, trends in NO₂ levels during the conflict are likely to be unpredictable in the chaos of armed conflict actions and regionally attributed to various type of emissions at multiple locations, especially at the beginning of the conflict. On one hand, the NO₂ levels should be expected to decline as anthropogenic emissions would be expected to decline due to minimized activities in transportation, industry and other socio-economic activities. On the other hand, surges in conflict activities – such as attacks with missiles, artillery shelling, bomb and mine explosions, etc., as well as the constant usage of military vehicles and the transportation of civilian populations from conflict zones in such a short time – could result in a rise in air pollution levels. Therefore, we extend our

study beyond the most populous cities and include other territories in Ukraine affected by the conflict. To accomplish this, we begin by locating the conflict hotspots where military actions and battles took place, and then analyse the changes in NO₂ concentrations in the hotspots, which are highly contested zones. We estimated the changes in pollution levels from individual conflict points, and the results are presented in Section 5.1. In Section 5.2, we analyse the impacts of the conflict on NO₂ levels in other affected regions, such as major cities with populations exceeding 0.5 million, and the areas surrounding CPPs.

3.5.1 S5P NO₂ level changes in conflict hotspots

Satellite-captured fire spots and statistics in conflict hotspots

To understand the distribution of conflict hotspots, we utilized both the satellite-capture fire data from the NASA FIRMS portal, and in particular, the locations of battles provided by ACLED (Raleigh et al., 2010). First, we inspect the fire data from the VIIRS fire product for two consecutive years (2021 and 2022), searching for patterns representing the appearance of conflict hotspots. Then, we visually compare the pattern of fire spots captured by satellite with the reported battle locations.

Figure 3.8 displays the satellite-captured fire spots for 2021 (1st column), 2022 (2nd column) and locations of conflict hotspots (3rd column). We only show the similar patterns captured from the monthly NASA FIRMS product and the reported conflict hotspot locations from Ukraine Crisis Hub, to avoid the overwhelming plots of 12 months. We observe that from February 24 until the end of March, the distribution of the detected fire spots forms no certain pattern and is scattered over the Ukrainian territory. From April to July 2022 the fire pattern starts to form and gradually be identifiable as similar to the conflict spots in the eastern part of the Ukrainian territory, while no special pattern is found in the 2021 figures for the corresponding periods. It is notable that the eastern region comprising of five oblasts (typically translated as regions or provinces, namely Dnipropetrovsk, Donetsk, Kharkiv, Luhansk, and Zaporizhzhia) has been at the frontline of the armed conflict and subject to intense conflict hotspots since the conflict began. Given our understanding that the ongoing armed conflict is the

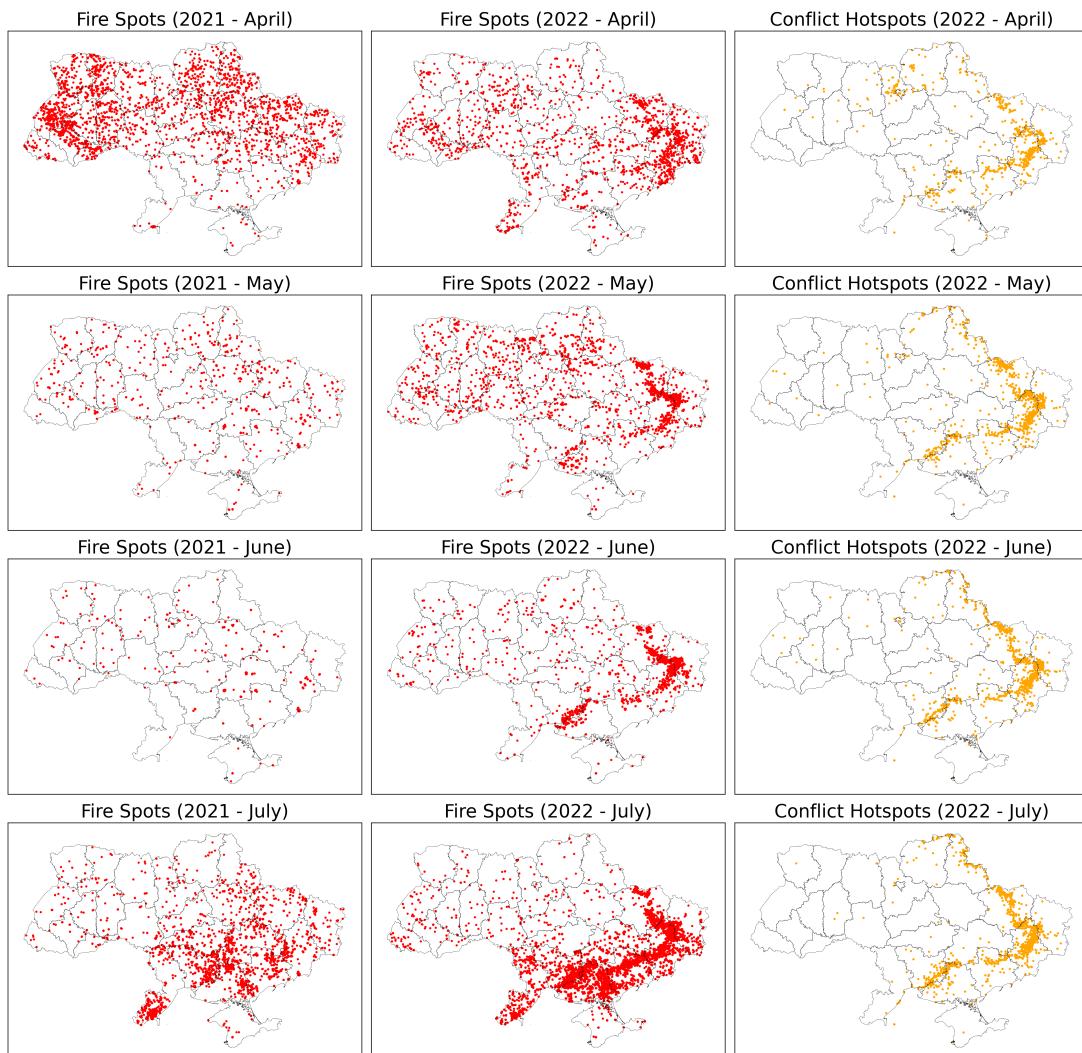


Figure 3.8. Satellite-captured fire spots for 2021 (1st column), 2022 (2nd column) and conflict hotspots (3rd column) in April, May, June and July. The patterns of conflict hotspots are clearly recognizable in the satellite-capture fire product from NASA FIRMS

source of explosions and smoke, it is reasonable to assume that the conflict has resulted in a significant increase in air pollution (Pereira et al., 2022), particularly in the areas directly affected by the conflict events that are detectable via VIIRS satellite products, so we would expect that S5P observations have the capability to show the resulting impacts on both overall air quality and concentrations of

NO₂ in the affected areas.

Changes of S5P NO₂ column levels

Until March 2023, as reported by (Nichita and Ana, 2023), nearly 40,000 events related to the conflict were recorded across the Ukrainian territory by the ACLED project (Raleigh et al., 2010). The five oblasts Dnipropetrovsk, Donetsk, Kharkiv, Luhansk and Zaporizhzhia have been on the frontline of the Russia-Ukraine armed conflict since February 24, 2022. In these areas, shelling, artillery, and missile attacks accounted for 71% of conflict events recorded between February 24 and July 31, 2022 (Nichita and Ana, 2023). In order to evaluate the impacts of conflict events at the smallest level, we quantify changes in NO₂ column levels directly at the reported event location using OBS-BAU and year-to-year estimates for the corresponding pixel from S5P data, which is equivalent a 10 km²-area containing the event location (Figure 3.9).

The OBS-BAU estimates based on ORG data indicate an average increase of 0.3%, while the year-to-year estimates show a more substantial increase of 13.2%. However, when using RPRO data, we observed an 11% reduction in the OBS-BAU estimate and a 1.35% increase in the year-to-year estimate. Although there is a high level of uncertainty in estimating changes at the event location-pixel level, and the inconsistent timing between the reported conflict related events and S5P overpass may lead to an underestimation of changes in air pollution levels, the information gathered can still be useful in identifying changes in the NO₂ columns associated with conflict related event locations in the five oblasts.

3.5.2 Changes of S5P NO₂ levels in other affected areas

Most populous cities of Ukraine

In the nine most populous cities in Ukraine, both the lockdown and the conflict have led to a reduction in daily anthropogenic activities. Although this reduction was expected to lower the NO₂ levels, as discussed in Section 4, the lockdown measures did not result in a significant reduction in NO₂ column levels in 2020. To quantify the changes caused by the conflict and compare them with the effects of the lockdown measures, we analysed the OBS-BAU estimate for the most

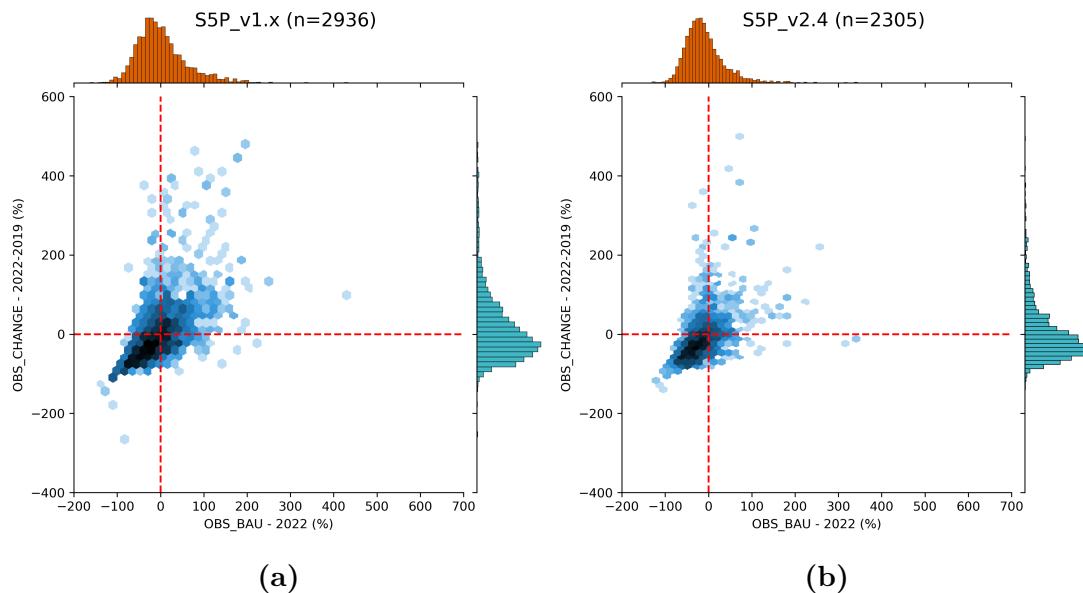


Figure 3.9. OBS-BAU and year-to-year estimates for the individual conflict events including air/drone strikes, armed clashes, remote explosive/landmine occurrences, shelling/artillery/missile attacks, and other forms of attacks that occurred between February 24 and July 31, 2022, for five frontline oblasts, Dnipropetrovsk, Donetsk, Kharkiv, Luhansk and Zaporizhzhia. The number of data points is denoted by (n).

populous cities in Ukraine during the strict lockdown period from April 6 to May 10 in 2020 and 2022 (Table 3.4). To avoid overwhelming plots, Figure 3.10 displays the NO₂ column trend lines for OBS data and BAU predictions from February to July in 2020 and 2022 for five cities (Kyiv, Kharkiv, Dnipro, Zaporizhzhia, and Kryvyi Rih) only.

Table 3.4 presents the OBS-BAU estimates corresponding to the strict lockdown period (April 6 to May 10) in 2020 and 2022 for the nine most populous cities in Ukraine. Our findings indicate that the conflict has caused more significant reductions in NO₂ levels, compared to the lockdown measures. While minor reductions to increases were observed during the 2020 lockdown, a consistent and continuous reduction has been noticed in most cities, during the same lockdown period (April 6 to May 10) in 2022. The average reduction across all the cities of interest, as shown in Table 3.4, is about 12.1% (based on ORG data) and 18.1%

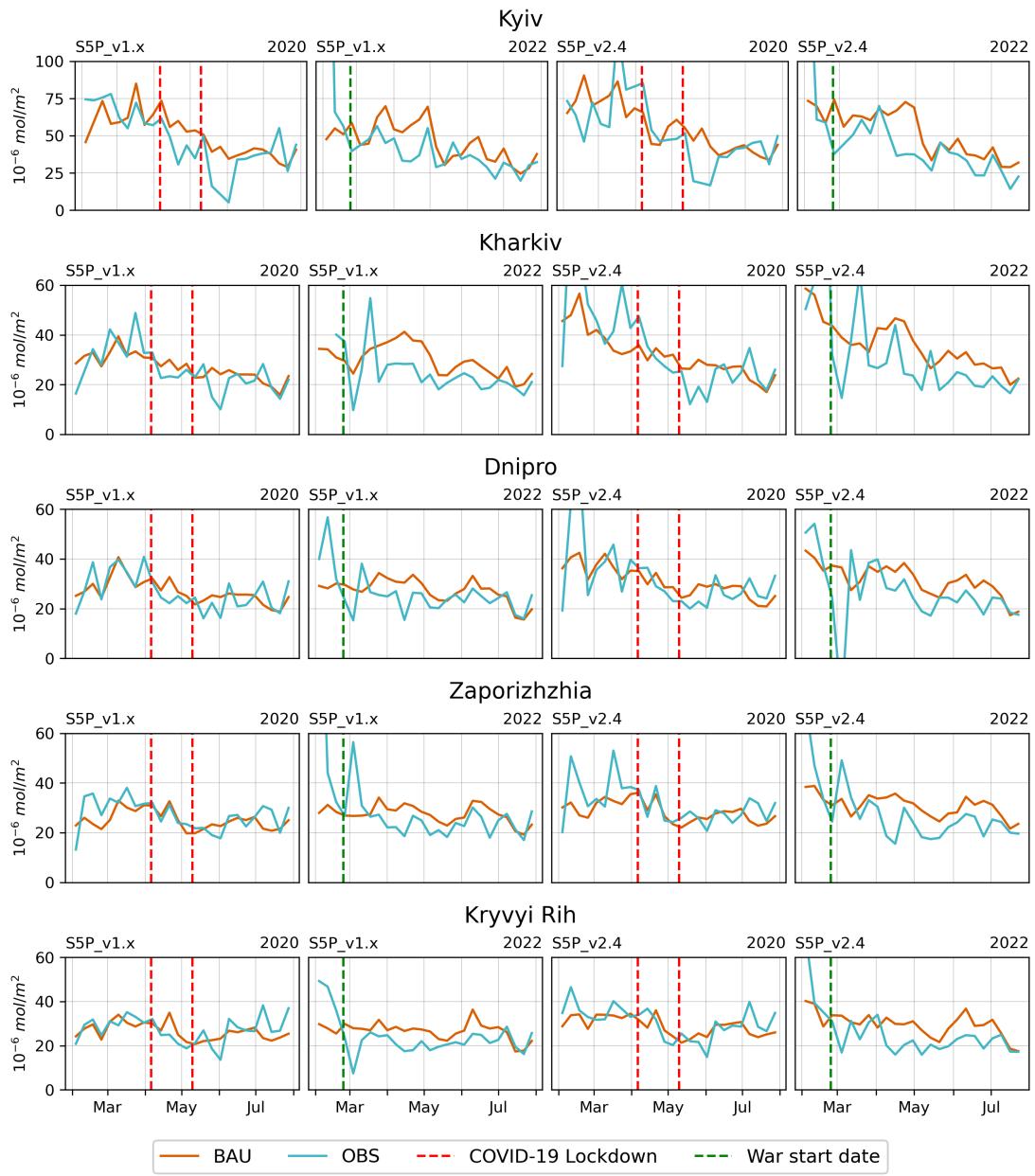


Figure 3.10. The trend lines of OBS and BAU S5P NO₂ column levels from February to July in 2020 and 2022 for five cities in Ukraine. Each row displays plots for a different city. The first and second column plots represent the ORG data (S5P version 1.x), while the third and last column plots show the RPRO data (S5P version 2.4). The first and third column plots pertain to 2020, while the second and last column plots pertain to 2022.

Table 3.4. The OBS-BAU estimate (in percentage) of ORG data and RPRO data for the strict lockdown period (April 6 to May 10) in 2020 and in 2022 for the nine most populous cities in Ukraine. The values are represented as mean (with standard deviation in parentheses). The mean and standard deviation in the last row were calculated across the nine cities.

City	2020 (April 6 –May 10)		2022 (April 6 –May 10)	
	ORG	RPRO	ORG	RPRO
Kyiv	-18.8 (6.5)	4.9 (17.4)	-29.3 (9.5)	-34.6 (7.6)
Kharkiv	-0.9 (10.3)	-4.9 (15.9)	-24.9 (17.9)	-29.7 (20.8)
Odessa	6.9 (12.4)	21.0 (16.4)	-7.6 (14.3)	-14.5 (9.7)
Dnipro	-6.6 (9.2)	2.8 (10.9)	-17.4 (10.0)	-19.5 (8.6)
Donetsk	28.2 (35.2)	42.0 (29.8)	3.5 (19.9)	3.2 (18.7)
Zaporizhzhia	2.5 (9.1)	9.1 (12.7)	-12.6 (13.7)	-18.4 (11.6)
Lviv	0.0 (10.9)	3.0 (8.5)	14.9 (17.9)	-3.3 (9.9)
Kryvyi Rih	-6.4 (8.7)	0.1 (9.9)	-20.8 (9.8)	-27.7 (8.1)
Mykolaiv	-0.6 (9.8)	13.8 (17.6)	-14.6 (10.1)	-18.0 (6.8)
Mean	0.5 (11.9)	10.2 (13.3)	-12.1 (13.2)	-18.1 (11.5)

(based on RPRO data). The largest reduction was observed in Kyiv, while the increase occurred in Lviv (14.9% based on ORG data) and in Donetsk (3.5% based on ORG data, 3.2% based on RPRO data).

In more than the first five months after the conflict began until the end of July 2022, an overall reduction is observed across the nine cities (see Table 3.5) with an average of 3.1% (ORG data) and 7% (RPRO data). The largest reductions in NO₂ levels were observed in Kyiv, with an average of 14.9% (ORG data) and 27.6% (RPRO data). Conversely, Donetsk and Lviv experienced increases in NO₂ levels, with both ORG and RPRO data, while in Mykolaiv only RPRO data showed the increases. The rise in Donetsk can be attributed to it being where major armed conflicts occurred during this period.

Table 3.5. Average OBS-BAU and year-to-year estimate (in percentage) of ORG data and RPRO data from February 24 to July 31, 2022, for the nine most populous cities in Ukraine. The values are represented as mean (with standard deviation in parentheses). The mean and standard deviation in the last row were calculated across the nine cities.

City	ORG		RPRO	
	OBS-BAU	year-to-year	OBS-BAU	year-to-year
Kyiv	-14.9 (17.3)	-30.5 (14.7)	-27.6 (12.1)	-37.3 (11.3)
Kharkiv	-3.2 (28.5)	20.7 (39.8)	-3.0 (33.3)	2.4 (23.6)
Odessa	-6.8 (15.4)	-13.6 (16.4)	-5.4 (13.0)	4.5 (61.0)
Dnipro	-12.4 (16.6)	-15.0 (21.3)	-17.6 (13.8)	-17.0 (20.5)
Donetsk	19.4 (26.6)	4.2 (21.5)	17.0 (22.8)	-9.4 (15.8)
Zaporizhzhia	-10.5 (16.4)	-15.7 (27.3)	-13.7 (14.4)	-19.1 (18.6)
Lviv	20.8 (21.9)	-9.0 (24.1)	2.2 (16.8)	-9.8 (17.3)
Kryvyi Rih	-15.5 (15.7)	-22.4 (21.7)	-17.4 (15.0)	-26.2 (42.7)
Mykolaiv	-4.8 (13.1)	-7.8 (23.5)	2.1 (14.8)	12.9 (21.9)
Mean	-3.1 (13)	-9.9 (14.1)	-7 (12.7)	-11 (15)

Coal power plants

Besides anthropogenic activities in major cities, the contribution of CPPs to NO₂ concentration levels is considered to be significant in Ukraine (Lauri and Rosa, 2021). The Zaporizhzhia CPP is one of the largest emitters among CPPs in Ukraine, emitting 21,830 tonnes of NOx in 2019. Many power plants have been targeted in the conflict, and their damage or destruction has resulted in power blackouts affecting millions of people.

According to Draft Ukraine Recovery Plan, Materials of the “Energy Security” Working Group covering the period to the end of June 2022, significant damage has been reported at the Zaporizhzhia, Luhansk, and Sievierodonetsk power stations, as well as other CPPs. This damage could be expected to affect NO₂ levels in the areas surrounding the damaged power plants. To investigate such changes, we also compare trends in the NO₂ column levels between OBS data and BAU simulations for 2020 and 2022, utilizing both ORG and RPRO data as presented

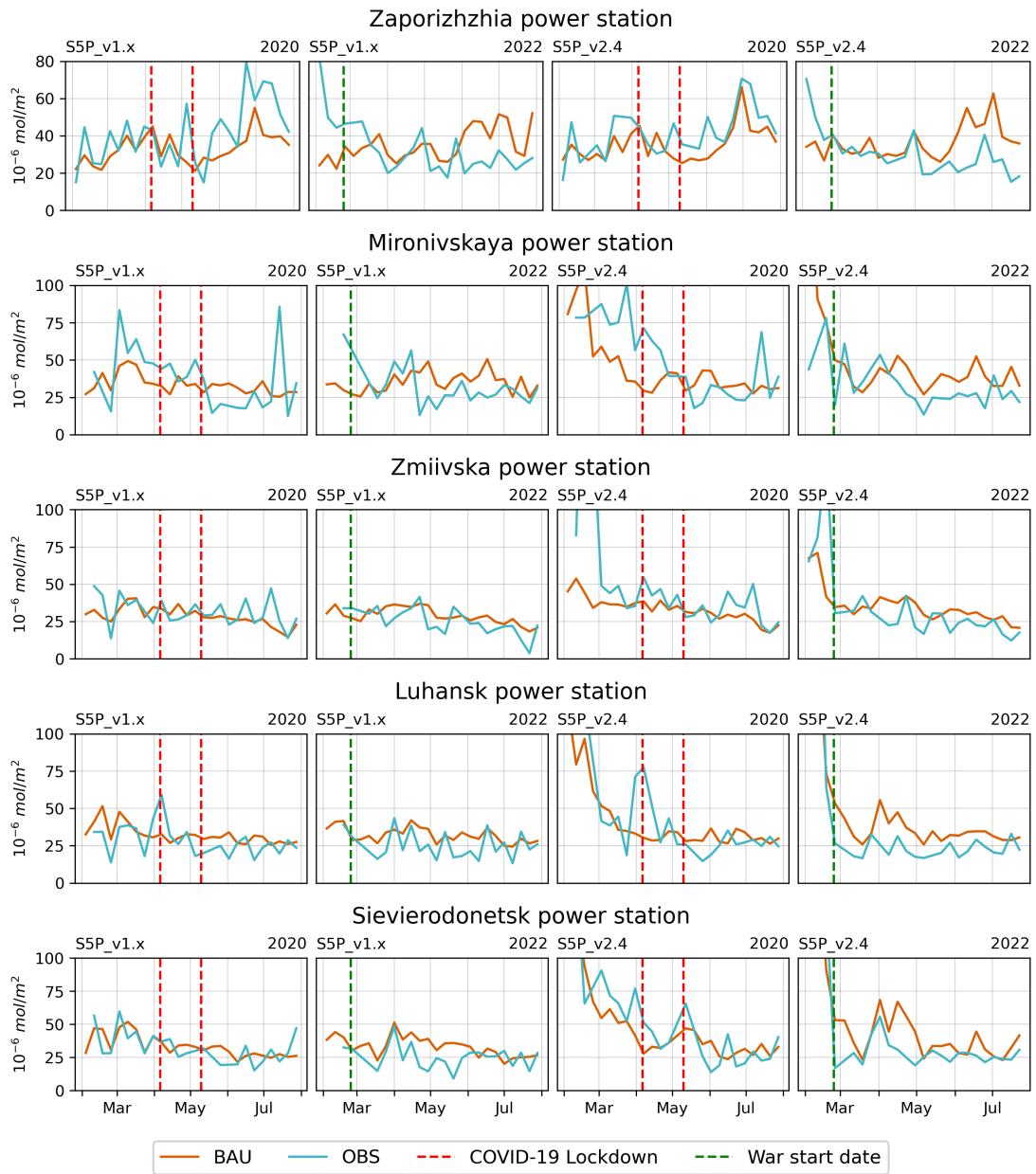


Figure 3.11. The trend lines for the OBS and BAU S5P NO₂ column levels from February to July in 2020 and 2022 are presented for selected CPPs. Each row displays plots for a different CPP. The first and second column plots represent ORG data (S5P version 1.x), while the third and last column plots show RPRO data (S5P version 2.4). The first and third column plots pertain to 2020, while the second and last column plots pertain to 2022.

in Figure 3.11. Examining an area of 10km² around each CPP, we find that, similar to previous discussions on lockdown effects, little changes are observed around most CPPs during the pandemic lockdown in 2020. However, a clear reduction is evident between the time when the conflict began and July 2022 at the Zaporizhzhia, Mironivskaya, Zmiivska, Luhansk, and Sievierodonetsk power stations. At areas surrounding other power stations, no noticeable reduction is observed.

3.6 Conclusion

In this study, we performed a comprehensive assessment of variations in the S5P column NO₂ levels in Ukraine during the COVID-19 pandemic lockdown in 2020 and the armed conflict with Russia in 2022. For this purpose, we utilized two S5P products, namely, original and reprocessing data. We first developed a weather normalization model under business-as-usual conditions, using meteorological parameters from ERA5 reanalysis, ensembled surface forecasts, and analysis NO₂ data from 11 CAMS models, along with other spatial and temporal features. Next, we applied the BAU prediction to estimate the change in NO₂ levels during the lockdown period in 2020 for the nine most populous cities in Ukraine (Kyiv, Kharkiv, Odessa, Dnipro, Donetsk, Zaporizhzhia, Lviv, Kryvyi Rih, and Mykolaiv). We extended the analysis using BAU predictions to estimate the impact of the armed conflict from February 24 to July 31, 2022, in conflict hotspot locations, the nine most populous cities, and areas surrounding selected CPPs (Zaporizhzhia, Mironivskaya, Zmiivska, Luhansk, and Sievierodonetsk) in Ukraine.

The main outcomes of the study can be summarized as follows:

- In 2020, meteorological parameters also heavily influenced the NO₂ tropospheric column levels, contributing to decreases in levels during the lockdown period.
- After normalizing the meteorological parameters, we found that the lockdown did not lead to lower NO₂ levels than the BAU prediction in 2020, although it did manage to mitigate the increase in NO₂ compared to the

pre-lockdown period. Our study indicates that stricter measures may need to be considered in the future to achieve a significant reduction in NO₂ levels in densely populated areas of Ukraine.

- We observed that satellite-capture fire data from the VIIRS product can capture the spatial patterns of the conflict related events on the ground. From this product, conflict location patterns are clearly represented during the April–July 2022 period.
- Upon examining changes in NO₂ levels at conflict hotspots at the location-pixel level, we observed changes ranging from an 11% reduction to a slight increase of 0.3% when comparing the OBS to BAU predictions using RPRO and ORG data, respectively.
- During the strict lockdown period from April 6 to May 10, 2022, the reduction in NO₂ levels in the nine most populous cities was more significant compared to 2020. Across most cities, an average reduction of 12.1% (ORG data) and 18.1% (RPRO data) was observed. However, it is worth noting that Lviv and Donetsk showed an increase in NO₂ levels during this period.
- From February 24 to July 31, 2022, the nine most populous cities in Ukraine experienced an overall reduction of 3.1% (ORG data) and 7% (RPRO data) in NO₂ levels. The most significant reduction was observed in Kyiv, with an average decrease of 14.9% (ORG data) and 27.6% (RPRO data). However, in contrast, NO₂ levels increased in Lviv, Donetsk and Mykolaiv during this period.
- The conflict has resulted in damage to several CPPs, which are considered as major sources of NO₂ emissions in the country. Our analysis indicates a clear reduction in NO₂ levels in the areas closely surrounding Zaporizhzhia, Mironivskaya, Zmiivska, Luhansk, Sievierodonetsk CPPs.
- By utilizing the OBS-BAU estimate for both ORG data and RPRO data to analyse NO₂ variations during the 2022 conflict, we found that discrepancies resulting from changes in the processor during the S5P lifetime in ORG data might lead to a slight underestimation of NO₂ reductions. Specifically, we

observed a smaller decrease using ORG data (3.1%) than with RPRO data (7%) in the most populous cities of Ukraine.

The consideration of meteorological effects is crucial in regulating pollution levels. Neglecting these effects could introduce errors in quantifying actual air quality changes attributed to an intervention event. For future studies assessing the impacts of conflict in Ukraine on air quality, it will be essential to account for meteorological variability to achieve genuine and quantitative estimates.

NO₂ is a significant precursor to tropospheric O₃ and also affects the lifetime of methane (CH₄) (Akimoto and Tanimoto, 2022). Additionally, it has the potential to serve as an indicator for monitoring CO₂ emissions (Miyazaki and Bowman, 2023). In future studies, it would be valuable to explore how changes in NO₂ levels during conflict could impact O₃ and CH₄ concentrations in Ukraine as both are important short-lived climate pollutants that contribute to positive radiative forcing, thereby exacerbating global warming.

4 Japan's case study

4.1 Introduction

Nitrogen dioxide (NO_2) is an important air pollutant that raises significant concerns due to its negative effects on human health (Hamra et al., 2015). Additionally, it serves as a crucial precursor to tropospheric ozone (O_3), along with volatile organic compounds (VOCs) (Akimoto and Tanimoto, 2022). Nitrogen oxides ($\text{NOx} = \text{NO} + \text{NO}_2$), carbon monoxide (CO) and non-methane volatile organic compounds (NMVOCs) have an influence on the methane (CH_4) lifetime by affecting the atmospheric mixing ratio of hydroxyl radicals (OH) (Akimoto and Tanimoto, 2022), which act as a primary sink for CH_4 (Turner et al., 2019). Both O_3 and CH_4 are short-lived climate pollutants (SLCPs) that contribute to positive radiative forcing, thereby intensifying global warming (Akimoto and Tanimoto, 2022). Moreover, owing to its short lifetime in the atmosphere and significant signal compared to carbon dioxide (CO_2), NO_2 possesses the potential to serve as an indicator for monitoring localized fossil fuel CO_2 emissions (Miyazaki and Bowman, 2023).

In 2020, the implementation of COVID-19 social distancing policies in multiple countries led to a significant decrease in human activities worldwide (de Palma et al., 2022). While the general anticipation was for a reduction in NO_2 emissions in many cities due to the decline in anthropogenic activities (Bauwens et al., 2020; Barré et al., 2021; Cooper et al., 2022), the response of O_3 and CH_4 has been unexpected.

Increased levels of O_3 have been observed in northern Europe, China, and South Africa as a consequence of the COVID-19 lockdown, according to sensitivity simulations conducted using the MIROC-CHASER global chemical transport model (Miyazaki et al., 2021). This rise in O_3 can be attributed to the general

reduction in NOx, which enhances O3 production by reducing NO titration in areas with high levels of NOx pollution or VOC-limited areas (Akimoto and Tanimoto, 2022). Furthermore, meteorological effects have played a significant role in the changes observed in O3 levels between 2020 and the reference year (Ordóñez et al., 2020; Liu et al., 2021). Despite accounting for the influence of weather conditions, significant variations in O3 level estimates have been reported across studies, particularly in European countries (Ordóñez et al., 2020; Grange et al., 2021), and China (Liu et al., 2021; Shi et al., 2021). The presence of sunlight is essential for the O3 generation in response to the decrease in NOx during the lockdown period. As a result, the lack of sunny conditions in specific urban areas at the time of the atmospheric response to NO2 reduction may have led to differing time delays before observable changes in O3 levels occurred (Grange et al., 2021) (Grange et al. 2021).

In 2020, during the COVID-19 pandemic, global CH4 emissions experienced a significant growth rate, which was contrary to the expected decrease in anthropogenic CH4 emissions due to the implementation of lockdown measures (Peng et al., 2022). In 2020, anthropogenic CH4 emissions only slightly decreased compared to 2019, while wetland emissions rose sharply. This increase in wetland emissions was likely influenced by unusually warm and wet weather in the Northern Hemisphere (Peng et al., 2022), which could be connected to the impact of climate change (Zhang et al., 2023). Apart from the variation in CH4 emission itself, it was found that the decrease in hydroxyl radical (OH) concentration due to changes in air pollutants like NOx, CO, and NMVOCs during the COVID-19 pandemic mainly accounted for approximately half ($53 \pm 10\%$) of the observed global CH4 level growth in 2020 (Peng et al., 2022). A similar finding regarding the effect of NOx, CO, and NMVOCs emission changes on the 2020 methane levels is reported by (Stevenson et al., 2022). However, other studies using Greenhouse gases Observing SATellite (GOSAT) observations indicated that most of observed increase in atmospheric CH4 during 2020 and 2021 can be attributed to increased CH4 emission itself (Qu et al., 2022; Feng et al., 2023). Although CH4 has a long estimated lifetime of 8-10 years and has mostly been discussed at the global level, it is important to note that policies and approaches to address CH4 emissions may vary locally.

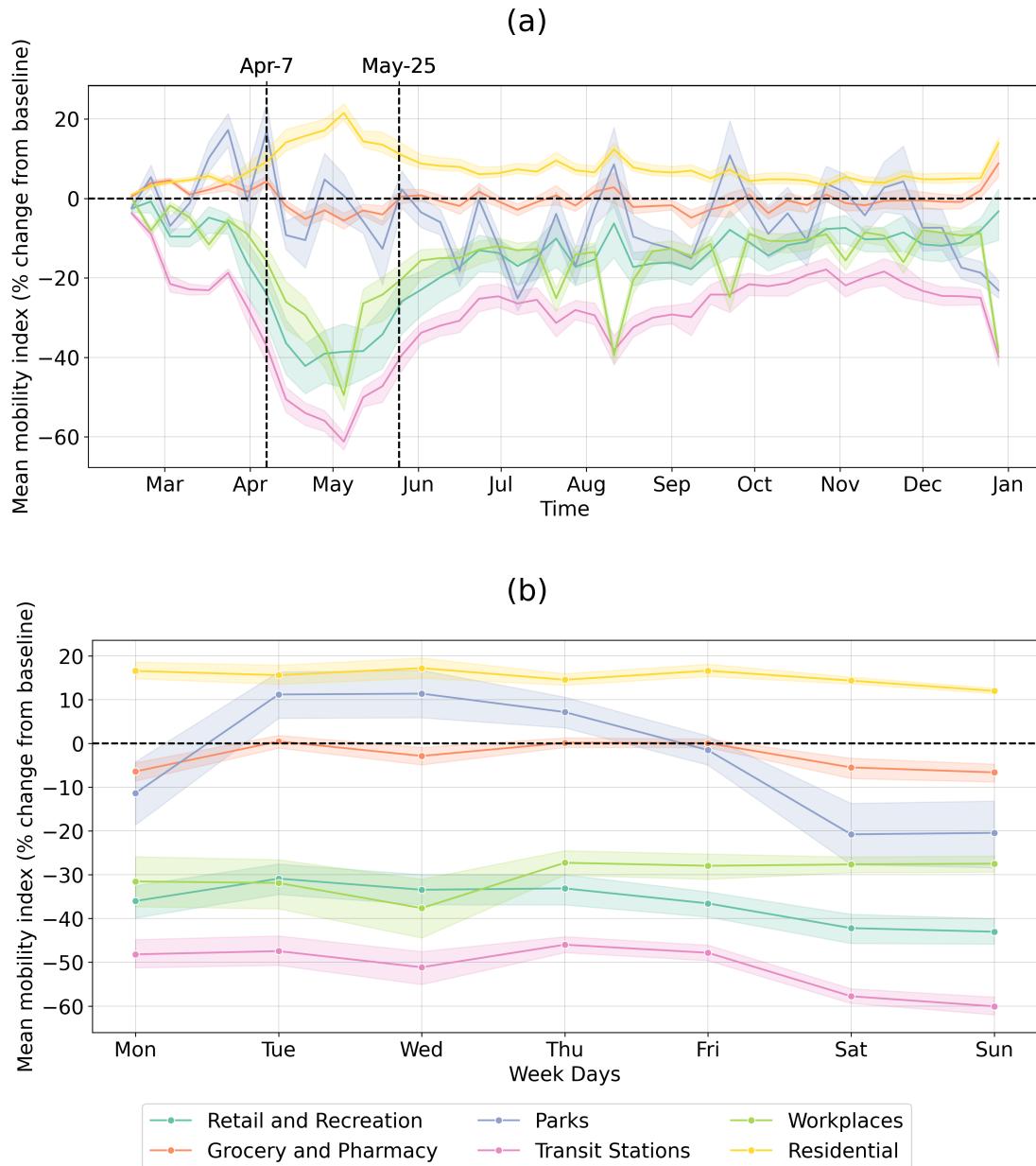


Figure 4.1. Mobility changes for 6 prefectures in Japan (Aichi, Fukuoka, Tokyo, Osaka, Kyoto, and Hyogo) in 2020 based on Google's mobility indices for time-series (a) and days of the week (b)

In 2020, Japan also experienced the impact of the COVID-19 pandemic, and in response to prevent the virus's spread, a state of emergency was declared from April 7 to May 25. This measure resulted in the suspension of various economic activities and imposed restrictions on people's mobility. As a consequence, there was a significant decline (Figure 1) in a unique weekend movement trend (Damiani et al., 2022).

Although the primary aim of the lockdown was not specifically to address air pollution and greenhouse gas emissions, the implementation of these measures offers valuable insights for atmospheric modelling. It provides practical knowledge and first-hand experience to develop more efficient strategies for mitigating air pollution and reducing greenhouse gas emissions in the future (Grange et al., 2021). It is important to note that the changes in air pollutants during this period varied across regions and were strongly influenced by meteorological conditions. Performing a regional analysis of these changes can provide evidence to support the formulation of appropriate regional policies in the future. In this study, our objective is to evaluate the impact of changes in anthropogenic activities during the COVID-19 pandemic (from April 7 to December 31) on NO₂, O₃, CO and CH₄ in metropolitan areas (MAs) of Japan in 2020, which have not been thoroughly investigated in previous studies.

In the first phase (Section 4.2), we gathered data from ground observations, satellite sources, and biogeochemical model simulations. Subsequently, we constructed a weather normalization model under business-as-usual (BAU) conditions utilizing machine learning techniques, incorporating meteorological, spatial, and temporal predictors (Section 4.3). We investigated variations in air pollution levels by analysing the BAU predictions alongside additional data in Section 4.4. Lastly, we provided discussions in Section 4.5, while in Section 4.6, we present our study's findings, conclusions, and recommendations for future policy considerations.

4.2 Data

4.2.1 Study area

Prior research primarily focused on assessing the impact of pandemic lockdown measures on air quality within the Greater Tokyo Area, being the most densely populated metropolitan area globally (Damiani et al., 2022; Zoran et al., 2023). Nevertheless, there's a notable absence of similar analyses for other MAs. Our study covers 14 MAs in Japan, extending from Sapporo in the north to Kagoshima in the south, as depicted in Figure 2. We focus on these metropolitan areas due to their housing of Japan's highly populated and vibrant cities, which are intricately connected with human activities and air pollution in Japan.

4.2.2 Ground observation

To acquire air quality data, we gathered ground observations for NO₂, O₃, CO, and CH₄ from the air quality monitoring data archive published by the National Institute for Environmental Studies (NIES). These observations spanned a ten-year period from 2010 to 2020 and were collected from 1,180 stations for NO₂, 835 stations for O₃, 383 stations for CH₄, and 237 stations for CO. The study utilized two types of stations: roadside air monitoring stations (RsAMS), which are placed in areas prone to air pollution from vehicle exhaust caused by traffic congestion, like intersections, roads, and near road edges, and ambient air monitoring stations (AAMS), which are established to assess air pollution in general living spaces such as residential areas. These station types have been categorized by NIES, and the data can be readily acquired from the original downloadable dataset.

Apart from air quality data, we incorporated ground observations of meteorological data from Japan Meteorological Agency (JMA) as input features for the BAU models used in the study. Specifically, we obtained daily records from 52 weather stations located within the same 14 MAs. At each weather station, we gathered temperature, wind direction and speed, local atmospheric pressure, and relative humidity, as suggested by (Grange et al., 2021). The corresponding meteorological parameters were extracted from the nearest weather observation site for each air quality station.

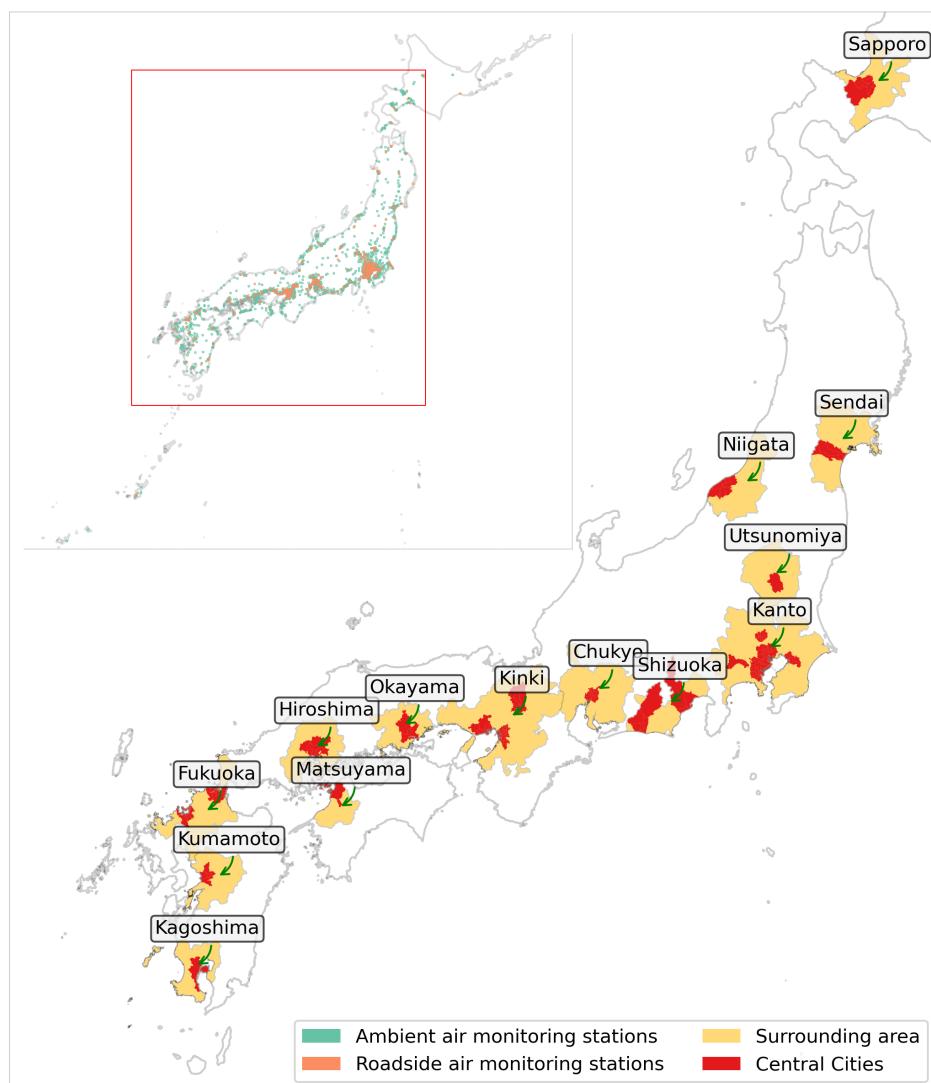


Figure 4.2. The locations of 14 metropolitans' areas and the distribution of ground observations for air quality monitoring in Japan

4.2.3 ERA5 reanalysis dataset

Alongside the weather data collected from the ground stations in the NIES database, for the features of the BAU models, we incorporated additional daily data pertaining to boundary layer height, total cloud cover, downward solar radiation (SR), and total precipitation, as recommended by (Shi et al., 2021). This supplementary information was sourced from the ERA5 reanalysis dataset (ERA5 hourly data on single levels from 1940 to the present) obtained from the Climate Data Store of the Copernicus Climate Change Service. Additionally, the ERA5 2m temperature variable (T2M) and SR will be utilized to assess the variation of sunny conditions during both the lockdown and post-lockdown periods within the study area. The original ERA5 data possesses a spatial resolution of $0.25^\circ \times 0.25^\circ$.

4.2.4 Sentinel 5P TROPOMI

In this study, we utilized the Sentinel 5P (S5P) Tropospheric Monitoring Instrument (TROPOMI) data to evaluate the tropospheric formaldehyde-to-NO₂ ratio (FNR) specifically for the year 2020. This ratio serves as a key indicator for the sensitivity of tropospheric ozone production. The tropospheric NO₂ and formaldehyde (HCHO – as a proxy for NMVOCs) data was obtained from the S5P L3 product “OFFL/L3_NO2” (based on processor version 1.2.x and 1.3.x) and “OFFL/L3_HCHO” (based on processor version 1.1.x) collections from Google Earth Engine, respectively. To generate the comprehensive L3 S5P product, each operational level (L2) product underwent preprocessing and mosaicking using the harpconvert tool. The low-quality pixels were filtered out in L3 NO₂ product by excluding those with AQ (Air Quality) values below 75% for the band “tropospheric_NO2_column_number_density”. The resulting data, ready for download, is available with a spatial resolution of about $1 \times 1 \text{ km}^2$.

4.2.5 Biogeochemical modelled CH₄ budget

In our assessment of CH₄ emission variations, with a specific focus on emissions from natural sources such as wetlands, we utilized CH₄ budget data obtained from the Vegetation Integrative Simulator for Trace gases (VISIT) (Ito et al.,

2019). VISIT is a biogeochemical model that takes into account historical land use and climatic conditions to estimate CH₄ emissions (Ito et al., 2019). The CH₄ budgets generated by the VISIT model are now available and accessible through the Global Environmental Database provided by NIES, Japan (Ito et al., 2019). We utilized the global data versions “Ver.2021.1_CH4Wetl_Cao” (Ito, 2021a), and “Ver.2021.1_CH4Wetl_WH” (Ito, 2021b), which incorporate Cao scheme (Cao et al., 1996), and Walter and Heimann scheme (WH scheme) (Walter and Heimann, 2000), to estimate CH₄ emission for each MA, which offers CH₄ emission information at a spatial resolution of 0.5° × 0.5°.

4.3 Method

4.3.1 Business-as-usual (BAU) modelling

To accurately quantify the actual change in the levels of the four pollutants, we developed a weather normalization model under BAU conditions using machine learning. This model was specifically designed to simulate pollutant levels without the influence of COVID-19 restriction measures, using meteorological, spatial, and temporal features as inputs. The meteorological predictors utilized in our model include ground observation data such as temperature, wind direction and speed, local atmospheric pressure, and relative humidity. Additionally, we incorporated data from the ERA5 reanalysis dataset, which comprises boundary layer height, total cloud cover, downward solar radiation, and total precipitation. Temporal predictors included the Julian date (the number of days since January 1) and the day of the week. Furthermore, latitude and longitude coordinates of each station were utilized as spatial predictors. To develop the weather normalization models for each pollutant at both AAMS and RsAMS, we utilized data from the years 2016 to 2019, which offers a comprehensive timeframe to account for the diverse air pollution concentration fluctuations experienced across various meteorological conditions. Extending the period, such as from 2010 to 2019, would not accurately represent recent air quality trends due to the impact of past air pollution reduction policies. Conversely, a shorter timeframe, such as the pre-lockdown period months would not adequately capture the full range of

meteorological variations. Overall, four separate weather normalization models were developed for each pollutant (NO_2 , O_3 , CO , and CH_4), taking into account the specific station type (RsAMS and AAMS).

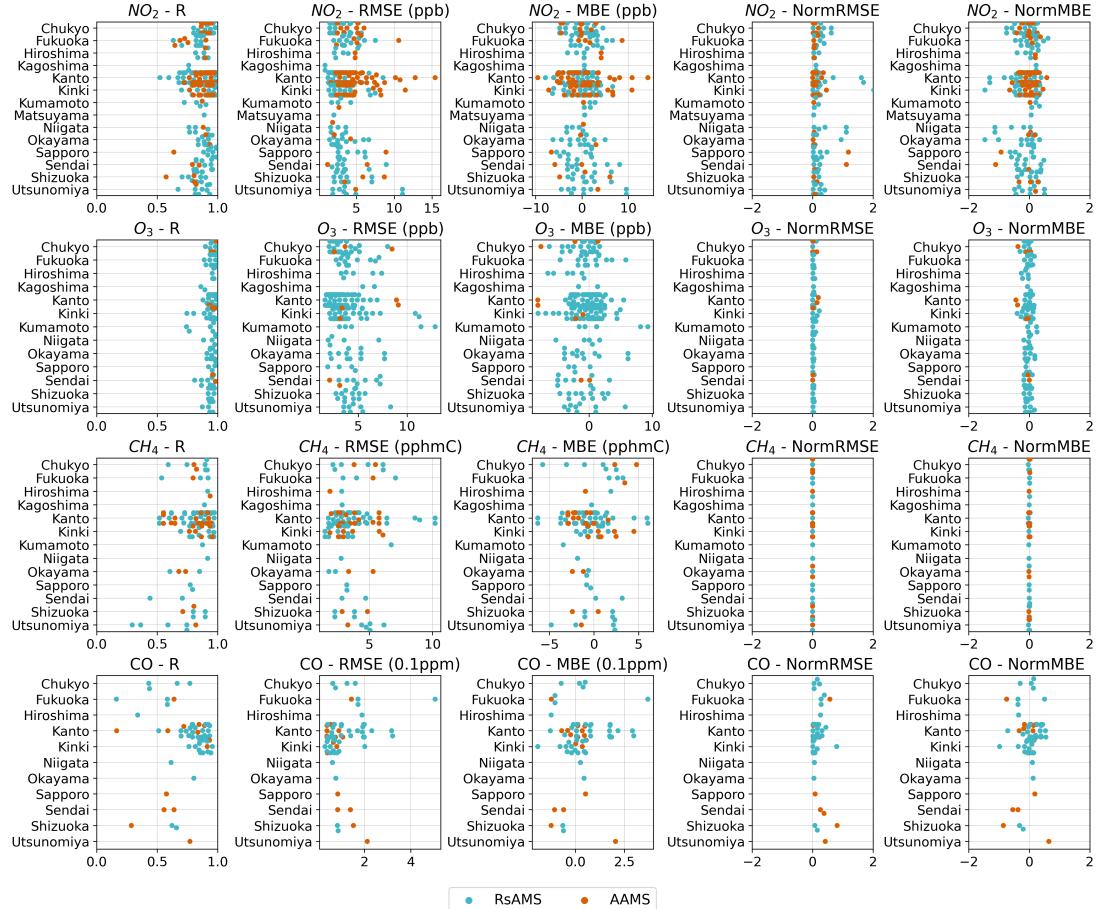


Figure 4.3. The details score of each station on the test set. For each station on the test set, we calculated the following scores and display it in this figure: Pearson correlation coefficient (R), root mean square error (RMSE), normalized root mean square error (NormRMSE) and mean bias error (MBE), normalized mean bias error (NormMBE)

We employed the LightGBM machine learning model (Ke et al., 2017), a gradient boosting decision tree algorithm, to construct the BAU model using the aforementioned predictors. To fine-tune the model's hyperparameters, we utilized Fast and Lightweight AutoML Library (FLAML) (Wang et al., 2021), a

Table 4.1. The performance of BAU model on the test set (30% station data) with the following metrics: Pearson correlation coefficient (R), root mean square error (RMSE), normalized root mean square error (NormRMSE) and mean bias error (MBE), normalized mean bias error (NormMBE). For the normalized MBE and RMSE, we normalize values for each station and then compute the mean

Pollutants	Station type	R	RMSE	NormRMSE	MBE	NormMBE
NO ₂	AAMS	0.89	3.13	0.15	-0.12	-0.07
	RsAMS	0.88	4.84	0.10	0.30	-0.03
O ₃	AAMS	0.96	3.75	0.02	-0.37	-0.02
	RsAMS	0.96	4.92	0.06	-3.18	-0.16
CO	AAMS	0.73	0.84	0.17	0.00	-0.07
	RsAMS	0.77	1.23	0.13	0.39	0.04
CH ₄	AAMS	0.82	3.75	0.00	-0.29	0.00
	RsAMS	0.80	3.82	0.00	-0.26	0.00

lightweight library specifically designed for accurately identifying optimal hyperparameters for models. During the training process, we utilized 70% of the station data within each metropolitan area (MA), while the remaining 30% was reserved for validating the model's performance. Both the training and test data sets were randomly selected for each MA, ensuring unbiased representation across the dataset.

In order to evaluate the performance of the BAU model we utilized the following metrics mean bias error (MBE), normalized mean bias error (NormMBE), root mean square error (RMSE), normalized root mean square error (NormRMSE) and Pearson correlation coefficient (R) as suggested by (Grange et al., 2021). The detailed results are presented in Figure 3 for each pollutant and station, average scores are shown in Table 1. In general, the model demonstrated strong performance with high R values (mostly R > 0.8) and low MBE and RMSE scores when applied to the test set for NO₂, O₃, and CH₄. Regarding CO, the model achieved a satisfactory R value (R > 0.73).

4.3.2 Experiments design

Our aim is to assess the alterations in NO₂ levels within 14 MAs during both the lockdown and post-lockdown periods in 2020. We also intend to explore how changes in NO₂ may influence the shifts in O₃ and CH₄ levels in each of these timeframes. Notably, we were encouraged to undertake this investigation by an observation of an unusual O₃ response to NO₂ reduction in the Greater Tokyo Area (Damiani et al., 2022), prompting me to study the response of O₃ and CH₄ in all 14 MAs across Japan.

We conducted three experiments to assess the impact of NO₂ changes on O₃ and CH₄ levels. In the first experiment, we focused solely on quantifying the change in NO₂ levels using the time series observations and "OBS-BAU" estimate which involved subtracting the BAU prediction from the observed data (OBS). In the second experiment, we expanded the analysis to include O₃, incorporating additional variables from the ERA5 (temperature – T2M and SR) and S5P datasets (FNR and HCHO). The last experiment included CH₄, incorporating the "OBS-BAU" estimate for CH₄ and NO₂, as well as the "OBS-BAU" estimate for CO and simulated CH₄ emissions from wetlands using the VISIT model.

For the experiments, we selected April 7 to May 25 as the lockdown period, August 1–31 as the post-lockdown period for O₃ analysis, and June 1 to December 31 for CH₄ analysis. We selected these timeframes to better understand how the four air pollutants changed in response to the unforeseen COVID-19 lockdown measures and the period after the lockdown.

4.4 Results

4.4.1 NO₂ level changes

We initially examined the monthly trend of observed NO₂ concentration levels across 1,180 stations in the 14 MAs from 2010 to 2019, and we compared these trends with the NO₂ levels observed during the lockdown in 2020 as depicted in Figure 4a. The results indicate that the actual reduction in NO₂ levels during the lockdown in 2020 is lower than the trend observed during 2010-2019, specifically 2.7 ppb for RsAMS and 2.2 ppb for AAMS. This implies that the NO₂ levels

observed during the lockdown were equivalent to those in 2023 for RsAMS and 2025 for AAMS, based on the trend observed during 2010-2019.

Prior studies have indicated the importance of considering meteorological factors when evaluating the effects of intervention measures (Ordóñez et al., 2020; Grange et al., 2021; Shi et al., 2021). In order to accurately assess the impact of the lockdown while isolating the effects of weather conditions, we computed the "OBS-BAU" estimates for all MAs as depicted in Figure 4b. Additionally, Figure 4c presents the complete time series of NO₂ levels in 2020 (OBS), the expected levels without the lockdown (BAU), and the average data from 2016-2019 for four MAs (Kanto, Kinki, Chukyo, Fukuoka). We only show the figures for four MAs to avoid overwhelming complexity and to provide a more manageable representation of the figures.

Table 4.2. OBS-BAU estimates for NO₂ during the lockdown (April 7 to May 25) and post-lockdown (August 1 to 31). For timeseries estimate, we considered all days of the week. However, when considering weekday, we only included Monday to Friday, while for weekends, we only accounted for Sunday and Saturday. The values are represented as mean (standard deviation)

Station type	Lockdown (April 7 –May 25)			Post-lockdown (August 1–31)		
	Timeseries (%)	Weekday (%)	Weekend (%)	Timeseries (%)	Weekday (%)	Weekend (%)
AAMS	-14.5 (12.1)	-12.9 (14.3)	-18.4 (8.6)	-10.2 (7.3)	-6.8 (7.8)	-17.2 (8.3)
RsAMS	-19.1 (13.5)	-18.0 (14.2)	-21.9 (13.9)	-18.1 (11.2)	-13.6 (12.3)	-27.4 (10.0)

Overall, NO₂ levels exhibited a decline across most MAs. The decline in emissions was particularly significant in RsAMS compared to AAMS in most MAs, with an average reduction of 19.1% and 14.5% respectively. However, these reductions were smaller compared to those observed in European cities (Barré et al., 2021; Grange et al., 2021). Additionally, we observed that the reduction in NO₂ levels during weekends was more significant than on weekdays, primarily due to a substantial decrease in mobility during weekends compared to weekdays (refer to Figure 1b). During the lockdown the average reduction in NO₂ levels for AAMS was 12.9% on weekdays and 18.4% on weekends. As for RsAMS, the average reduction stood at 18% on weekdays and 21.9% on weekends. For most MAs,

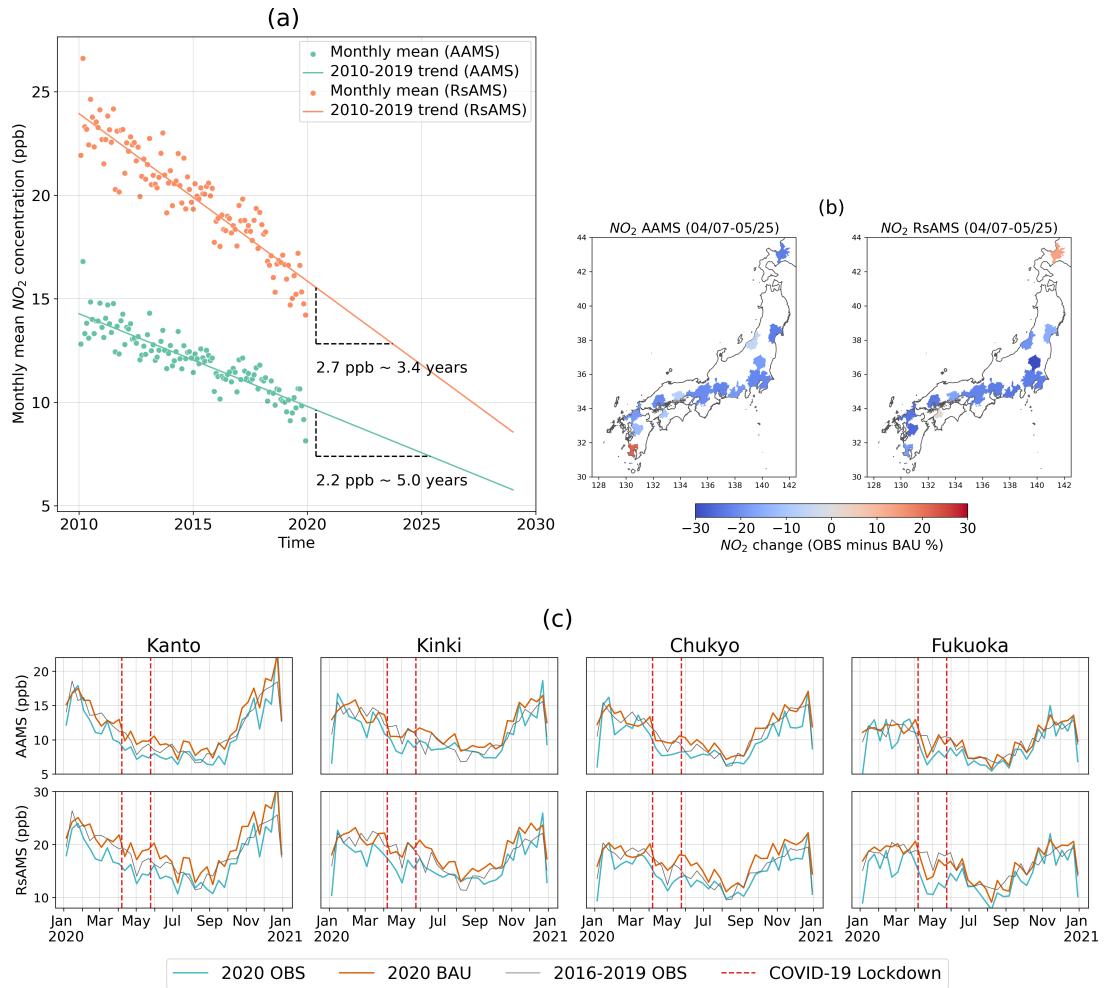


Figure 4.4. (a) Mean ground observation trend with the reduction in NO_2 due to the lockdown in 2020 for AAMS and RsAMS. (b) Map visualization of the “OBS-BAU” estimate for NO_2 during the lockdown period. (c) The 7-day rolling mean of 2020 observation (OBS), BAU prediction (BAU), and mean level of NO_2 from 2016 to 2019 for 4 MAs

even the lockdown has been lifted in the end of May 2020, the NO_2 level still continue to decline until the end of December 2020. This ongoing decrease may be attributed to the sustained reduction in mobility trends from the period of the lockdown through the end of 2020 (as illustrated in Figure 1a). These findings

are summarized in Table 2 and Table 3.

4.4.2 O₃ level changes

In this experiment, we investigated various parameters to gain a better understanding of the changes in O₃ in response to the reduction of NO₂ caused by COVID-19 social distancing policies. Alongside the "OBS-BAU" estimates, we examined standardized anomalies of T2M and SR between 2020 and 2016-2019 period, S5P FNR in 2020, and changes in S5P HCHO between 2020 and 2019. These parameters were analyzed for two distinct periods: the lockdown period and the post-lockdown (August 1 –31), 2020.

Changes during the lockdown period

During the lockdown period (April 7 to May 25), we observed a slight change in O₃ levels across most MAs (Figure 5 second row and Figure 6). On average, there was a reduction of 2.3% in AAMS and 0.6% in RsAMS, as indicated in Table 2. Although the overall trend showed a decrease, we did find instances of increased O₃ levels in certain MAs, particularly in RsAMS such as Kanto (1.6%), Kinki (2.2 %), and Fukuoka (3.5 %), as depicted in Figure 5 (second row). Moreover, we have observed the existence of an "ozone weekend effect" in the changes of O₃ levels, indicating higher increase in O₃ mixing ratios during weekends in comparison to weekdays (Akimoto and Tanimoto 2022). This effect was observed in the "OBS-BAU" estimates for RsAMS in Fukuoka (increased 8.8% - weekends, 1.3% - weekdays) and Kinki (increased 4.9% - weekends, 1.2% - weekdays). The observed slight decrease in O₃ levels across most MAs in Japan contrasts with the trends observed in many other major cities worldwide (Shi et al., 2021; Grange et al., 2021), where significant increases in O₃ levels have been observed. For instance, after accounting for weather effects, notable increases have been reported in Beijing (28.9 %), Wuhan (44.5 %), Milan (66.8 %) Rome (55.8 %), New York (17.4 %), Los Angeles (14.8 %), and Delhi (26.2 %) by (Shi et al., 2021).

To explore this variation further, we analyzed the disparity in T2M and SR between the corresponding period of 2020 and the reference period 2016-2019 as

shown in Figure 5 (3rd row). We observed small positive SR anomalies in the southeast region of Japan and negative SR anomalies in the northeast region. Additionally, across the entire country, negative T2M anomalies were observed. The presence of negative T2M anomalies and fluctuating SR levels suggests that the prevailing weather conditions during this period impeded the production of O₃.

Table 4.3. OBS-BAU estimates for O₃ during the lockdown (April 7 to May 25) and post-lockdown (August 1 to 31). For timeseries estimate, we considered all days of the week. However, when considering weekday, we only included Monday to Friday, while for weekends, we only accounted for Sunday and Saturday. The values are represented as mean (standard deviation)

Station type	Lockdown (April 7 –May 25)			Post-lockdown (August 1–31)		
	Timeseries	Weekday	Weekend	Timeseries	Weekday	Weekend
	(%)	(%)	(%)	(%)	(%)	(%)
AAMS	-2.3 (2.7)	-2.7 (3.2)	-1.2 (2.7)	2.2 (15.6)	3.2 (15.3)	0.0 (18.8)
RsAMS	-0.6 (2.7)	-1.4 (2.7)	1.4 (3.7)	8.9 (10.7)	8.9 (12.3)	8.6 (12.7)

Changes during the August, 2020

In August 2020, the NO₂ levels continued to decline in all MAs, albeit at a slower rate compared to the lockdown period, as shown in Table 2. However, during this period, we observed a more noticeable increase in O₃ levels across most MAs compared to the lockdown. On average, there was a 8.9% increase for RsAMS and a 2.2% increase for AAMS. Notably, the increase in O₃ levels during weekends was more significant than on weekdays in Niigata, Okayama, Kinki and Sendai. Specifically, For AAMS of Niigata, O₃ levels experienced a 9.4% increase on weekends and a 5.8% increase on weekdays. In RsAMS of Okayama, O₃ levels saw a 13% increase on weekends, exceeding the 10.6% increase observed on weekdays. Similarly, in AAMS in the Kinki region, O₃ levels exhibited a weekend increase of 19.8%, surpassing the 17.4% increase observed on weekdays. In Sendai, the increase during weekends was even more pronounced, with a 15.6% increase for AAMS and a 22% increase for RsAMS, whereas on weekdays the increase

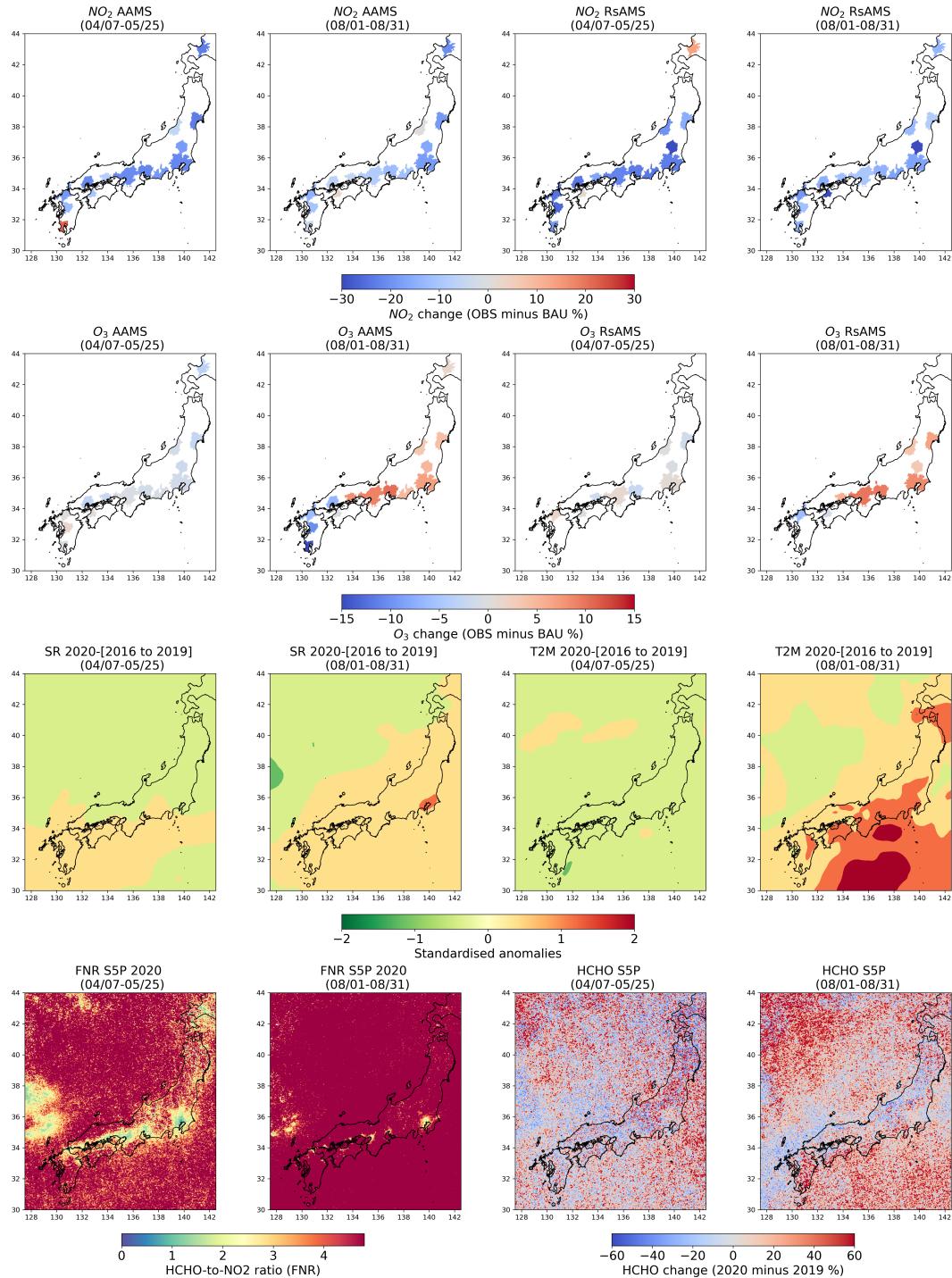


Figure 4.5. The 1st and 3rd columns show the plots for the lockdown (April 7 to May 25). The 2nd and last columns show the plots for August 1 – 31. The 1st row: The OBS-BAU estimates of NO_2 for AAMS and RsAMS. The 2nd row: The OBS-BAU estimates of O_3 for AAMS and RsAMS. The 3rd row: The standardised anomalies of downward solar radiation (SR) and temperature (T2M) from ERA5 dataset. The last row: The formamide-to- NO_2 (FNR) ratio in 2020 and the HCHO change between 2020 and 2019 from Sentinel 5P data.

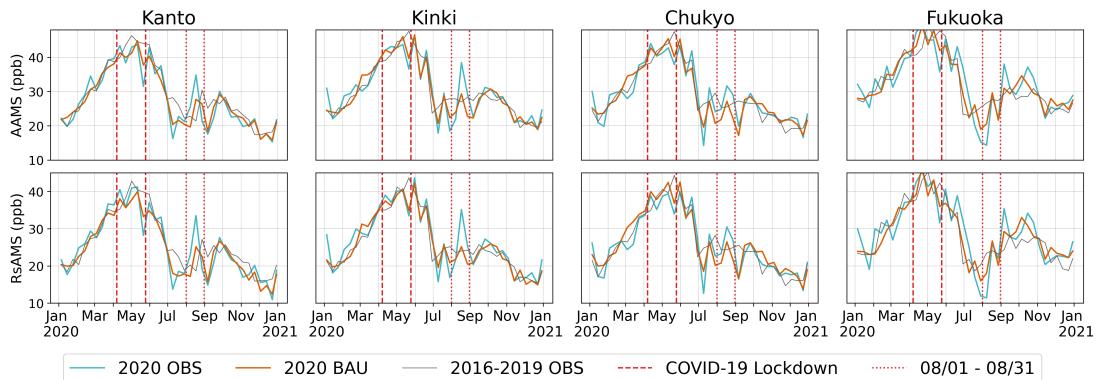


Figure 4.6. The 7-day rolling mean of 2020 observation (OBS), BAU prediction (BAU), and mean level of O₃ from 2016 to 2019 for 4 MAs (Kanto, Kinki, Chukyo, and Fukuoka)

was 5.1% for AAMS and 9.8% for RsAMS. This observation could be attributed to the greater reduction in movement during weekends compared to weekdays in these MAs as shown in Figure 1b.

In order to investigate the differences in O₃ levels between August and the lockdown period, we examined the standard anomalies of SR and T2M in August 2020, comparing them to the 2016-2019 period. Our analysis revealed positive anomalies in both SR and T2M across all MAs, as shown in Figure 5 (3rd row). These favorable weather conditions, combined with the reduced levels of NO₂, likely facilitated increased O₃ production.

Although there was an overall trend of increasing O₃ levels during this period, we did observe a reduction in O₃ levels in five MAs which is located in the southern region: Hiroshima (AAMS: 13.7%), Matsuyama (AAMS: 1%, RsAMS: 3%), Fukuoka (AAMS: 12.5%, RsAMS: 12.3%), Kumamoto (AAMS: 20.7%), and Kagoshima (AAMS: 29.9%). To understand the decrease in O₃ levels observed in these five MAs, we utilized the S5P FNR for 2020, as well as the changes in HCHO as a proxy for NMVOCs between 2020 and 2019. The FNR is commonly used to assess the sensitivity of near-surface O₃ levels (Martin et al., 2004). As suggested by (Duncan et al., 2010), when the FNR is below 1, the O₃ production regime is considered VOC-limited, and when it exceeds 2, it is considered NO_x-limited. When the FNR values fall within the range of 1–2, O₃ is expected to be

in the transition regime (Duncan et al., 2010). However, it has been observed that the FNR can vary by region (Jin et al., 2020; Irie et al., 2021; Souri et al., 2023; Ren et al., 2022), and the assumption that it lies within the 1–2 range may not hold true at the global level (Schroeder et al., 2017). Hence, it might be essential to calculate this ratio on a regional scale (Damiani et al., 2022; Schroeder et al., 2017). Despite the FNR showing high variability in the region, it still provides information about the trend of O₃ production regimes in our study.

Figure 5 (last row) presents the FNR across all MAs indicating a shift in the O₃ production regime from VOC-limited during the initial lockdown to NO_x-limited in August. This transition is evident as the FNR changes from $0 < \text{FNR} < 2$ during the lockdown to $\text{FNR} > 4$ in August. During the VOC-limited regime, a decrease in NO_x typically leads to an increase in O₃ levels (Duncan et al., 2010). However, in the NO_x-limited regime, a reduction in NO_x can also result in a decrease in O₃ levels (Duncan et al., 2010). In Figure 5 (last row), we can observe that the NO_x-limited regime dominates the five MAs of Hiroshima, Matsuyama, Fukuoka, Kumamoto, and Kagoshima. Despite NO₂ levels continuing to decline during this period, the HCHO levels exhibited a more significant increase in these MAs compared to the lockdown period. Hence, this could explain the reduction in O₃ levels observed in these five southern MAs.

We elucidated the difference in O₃ levels between major MAs in Japan and other large urban areas worldwide by examining meteorological changes (T2M, SR), and variations in O₃ precursors levels by utilizing S5P FNR derived from S5P NO₂ and HCHO measurements. The difference can be attributed to the absence of sunny conditions during the lockdown period. However, in August, when sunny conditions became more prevalent, we observed an increase in O₃ levels in response to the sustained reduction in NO₂ levels across most MAs, which are likely VOC-limited areas. Based on the analysis of S5P data, it appears that the southern metropolitan areas (MAs) exhibited a predominant NO_x-limited trend during August 2020, potentially due to the increased presence of biogenic VOCs (BVOCs). However, the monitoring of BVOCs emissions remains challenging due to limited observations (Tani and MOCHIZUKI, 2021; Ito and Ichii, 2021). Therefore, it is also important to pay attention to those NO_x-limited areas, as future reductions in anthropogenic NMVOCs may have minimal effectiveness in

reducing O₃ levels (Akimoto and Tanimoto 2022).

4.4.3 CH₄ level changes

In this experiment, we analyze the "OBS-BAU" estimates for NO₂, CO, and CH₄, and incorporate the VISIT model's simulated CH₄ emissions from wetlands to investigate the changes in CH₄ levels during the 2020 lockdown and post-lockdown period. Our focus is on understanding the relationship between the reduction in NO₂ and its potential impact on OH (hydroxyl radicals), as well as the contrasting effect of CO. The decrease in NO₂ levels is expected to result in a reduction in OH, while reductions in CO can increase OH levels and shorten the lifetime of CH₄ (Akimoto and Tanimoto, 2022).

During the lockdown period, we observed a marginal rise in CH₄ levels across most MAs (Figure 7 third row and Figure 8b), with an average increase of 0.6% for AAMS and 0.8% for RsAMS (Table 3). While NO₂ levels decreased in most MAs (Figure 7 first row), the trend for CO varied (Figure 7 second row and Figure 8a). AAMS showed an average decrease of 10.9% in CO levels, while RsAMS saw a slightly smaller reduction 8.8%. Notably, CO levels significantly increased in RsAMS of Kagoshima (60.6%), while slight increases were observed in Kanto AAMS, and in Matsuyama for both RsAMS and AAMS. It is worth noting that although the increases in CO levels in Kagoshima were significant, this region have among the lowest natural CH₄ emissions in Japan as Figure 7 (last row), which explains the slight increase in CH₄ observed in this MA.

During the post-lockdown period from June to December 2020, NO₂ levels continued to decrease, showing an average reduction of 12.8% for AAMS and 18.3% for RsAMS (Table 3) which is smaller than during the lockdown period. In contrast, CO levels started to recover as the COVID-19 lockdown was lifted, with a smaller reduction of 5.7% for AAMS and 5.5% for RsAMS. Notably, significant increases in CO levels were still evident at RsAMS in Kagoshima (62.2%). In Fukuoka we also observed a steady rise of CO levels in both RsAMS (13%) and AAMS (11.5%). In response to these changes in NO₂ and CO, we observed a greater increase in CH₄ levels during this period, with a rise of 1.3% for AAMS and 1.1% for RsAMS.

In general, we saw a slight increase in CH₄ levels both during the lockdown

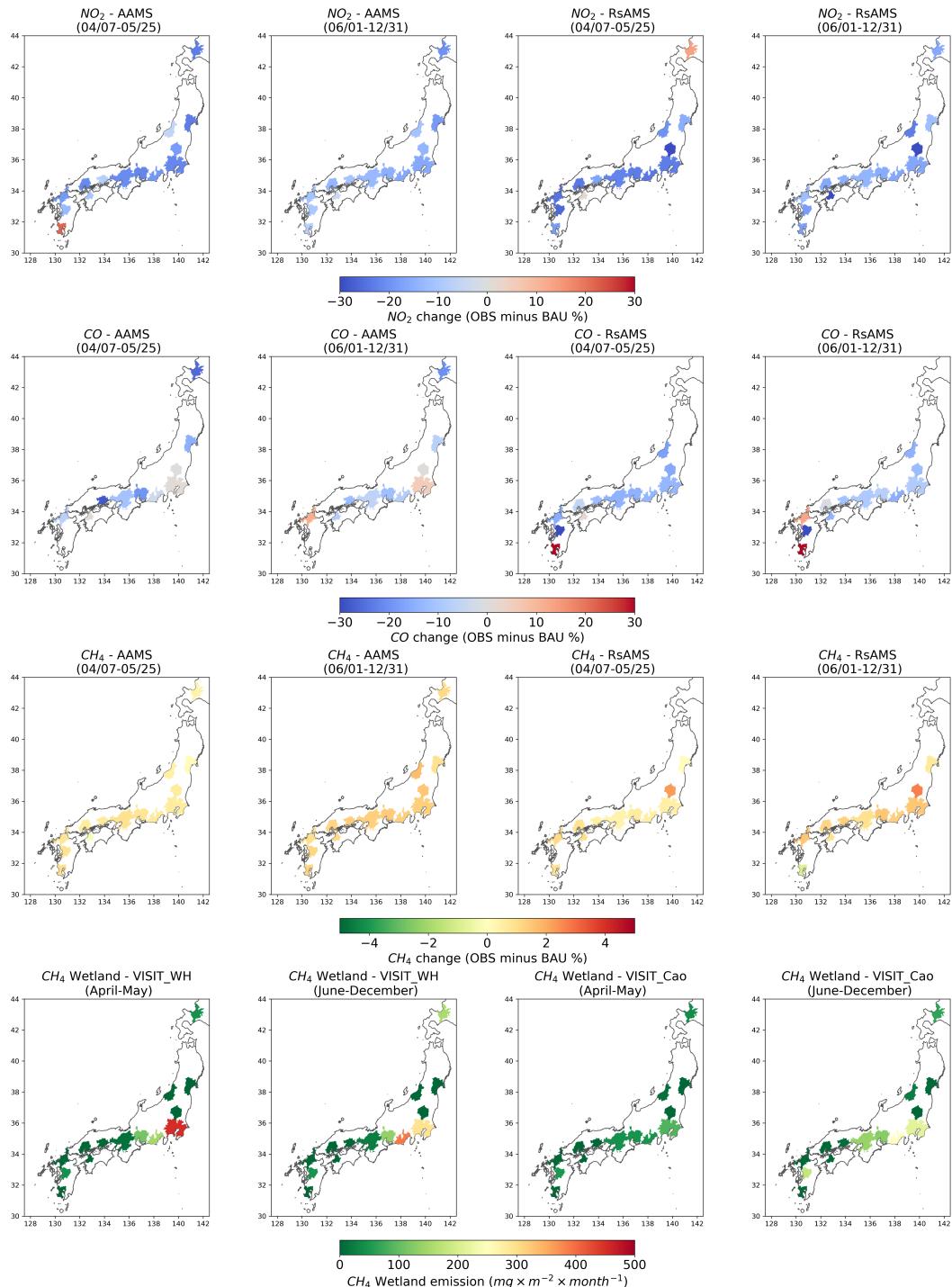


Figure 4.7. The 1st and 3rd columns show the plots for the lockdown (April to May). The 2nd and last columns show the plots for the post-lockdown (June to December). The 1st row: The “OBS-BAU” estimates of NO₂ for AAMS and RsAMS. The 2nd row: The “OBS-BAU” estimates of CO for AAMS and RsAMS. The 3rd row: The “OBS-BAU” estimate of CH₄ for AAMS and RsAMS. The last row: The CH₄ emission from wetland based on the simulation of VISIT model with Walter and Heimann scheme and Cao scheme.

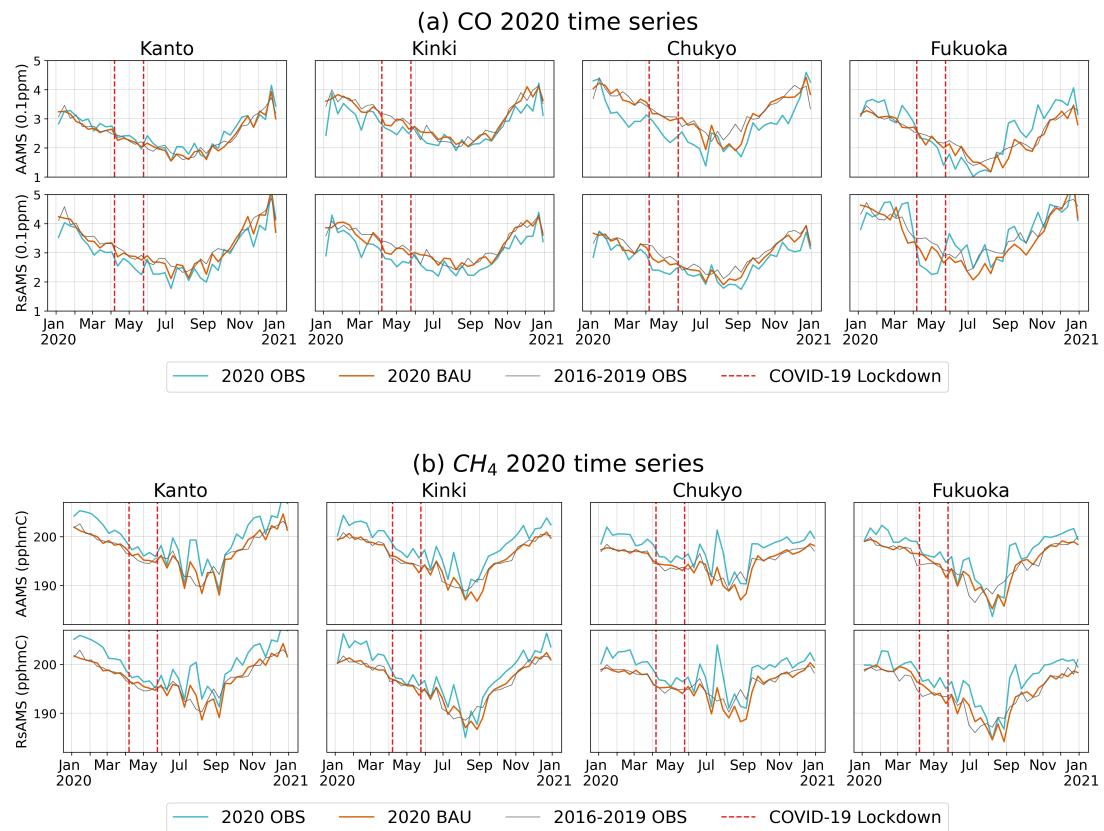


Figure 4.8. The 7-day rolling mean of 2020 observation (OBS), BAU prediction (BAU), and mean level of CO (a) and CH₄ (b) from 2016 to 2019 for 4 MAs (Kanto, Kinki, Chukyo, and Fukuoka)

and the post-lockdown periods, based on the "OBS-BAU" estimates. However, a more pronounced increase in CH₄ was observed during the post-lockdown phase in AAMS when compared to RsAMS, which can be attributed to the more substantial recovery of CO levels in AAMS relative to the lockdown period. Although it has been reported that global CH₄ growth in 2020 is primarily attributed to the atmospheric sink resulting from lower anthropogenic NO_x emissions (Stevenson et al., 2022; Peng et al., 2022), our findings regarding the contribution of NO_x reduction to the CH₄ growth in Japan in 2020 align with a previous study (Akimoto and Tanimoto, 2022; Qu et al., 2022; Feng et al., 2023), indicating that the impact of NO_x and CO change on the increase in CH₄ growth in Japan during

Table 4.4. OBS-BAU estimates for NO₂ and CO and CH₄ during the lockdown (April 7 to May 25) and the post-lockdown (June 1 to December 31). For CH₄ analysis we only consider timeseries estimate which include all days of the week. The values are represented as mean (standard deviation)

Pollutant	Station type	(April 7 – May 25)	(June 1 – December 31)
		(%)	(%)
NO ₂	AAMS	-14.5 (12.1)	-12.8 (4.3)
	RsAMS	-19.1 (13.5)	-18.3 (6.4)
CO	AAMS	-10.9 (11.0)	-5.7 (9.4)
	RsAMS	-8.8 (24.6)	-5.5 (25.2)
CH ₄	AAMS	0.6 (0.3)	1.3 (0.2)
	RsAMS	0.8 (0.6)	1.1 (0.9)

the lockdown and post-lockdown period is not as significant as the direct CH₄ emission itself.

4.5 Discussion

4.5.1 Variations in spatial resolution of multisource data

Since we utilized multisource data for the analysis, we acknowledge that variations in spatial resolution among input data can influence the consistency and reliability of data analysis. In certain situations, the need for interpolation to achieve a uniform grid may arise, particularly when generating inputs for a Convolutional Neural Network (CNN). This interpolation process inadvertently introduces uncertainty into the results. However, in this study, we refrained from any data interpolation and used it at its provided original resolution. The multisource data was employed for two primary objectives: weather-normalization model development and visual examination purpose.

For weather-normalization model development, we used ERA5 data and ground station data to construct the weather-normalization model. Certain variables, such as total cloud cover and boundary layer height, are exclusively available from ERA5. The ERA5 data we employed has a resolution of $0.25^\circ \times 0.25^\circ$, meaning

that some stations might share identical ERA5 records. This can influence the model development, even though, ideally, local ERA5 values for each station should be distinct, albeit not significantly deviating from the $0.25^\circ \times 0.25^\circ$ spatial resolution value. To mitigate this effect on the model development, we have integrated spatial context values (latitude and longitude) and station types as additional inputs. Since these features are distinct for each station, we anticipate that they can help minimize the impact of the coarse spatial resolution from ERA5 on the model.

To visually inspect the sensitivity of tropospheric O₃ production utilizing S5P HCHO and NO₂, as well as CH₄ emission estimates from wetland, we rely exclusively on original data with consistent spatial resolution. It's important to note that our primary focus is to visually inspect the prevailing trends at the MA level, which has a spatial resolution coarser than that of any input data we utilized. Therefore, we believe that the dominant trends at the MA level remain unaffected by these spatial disparities in this particular MA-level context.

4.5.2 Limitations

In this research, we utilized the S5P FNR to examine the sensitivity of O₃ production. Although HCHO could be an alternative indicator for NMVOCs presence, the significant uncertainty in the FNR threshold from previous studies, along with the lack of NMVOCs observations and reliable satellite HCHO and NO₂ data, poses challenges in understanding O₃ level variations during and after the lock-down period. This issue is particularly crucial and warrants in-depth exploration in future studies.

Additionally, it's important to mention that the study did not include an analysis of long-range air pollution transportation from China to western MAs of Japan following the Chinese economic recovery from the pandemic (Itahashi et al., 2022). This aspect was beyond the scope of the current research but should be considered in future investigations.

4.6 Conclusion

This study presents an air quality analysis that examines the changes in four air pollutants, namely NO₂, O₃, CO, and CH₄, during the COVID-19 pandemic in 14 MAs of Japan from April 7 to December 31 in 2020. Firstly, we developed a machine learning BAU model that incorporates meteorological, spatial, and temporal features to account for weather variability in air quality time series. Next, we utilized the BAU model predictions and observation data to estimate the actual reduction (OBS-BAU estimate) in NO₂ levels. We then integrated temperature and solar radiation anomalies from ERA5 reanalysis data and S5P TROPOMI data (FNR and HCHO) along with the “OBS-BAU” estimate to investigate the unique response of O₃ to the NO₂ reduction during the lockdown and post-lockdown period (August 1 – 31, 2020). Finally, we evaluated the impact of NO₂ and CO changes on the CH₄ levels using a combination of “OBS-BAU” estimate and wetland CH₄ emission simulations from the VISIT model. The main findings of the study can be summarized as follows:

Based on ground observations of NO₂, the reduction of NO₂ during the lockdown period in 2020 corresponds to a decrease equivalent to 3.4 years and 5 years of the 2010-2019 trend of NO₂ for roadside and ambient air monitoring stations respectively. After normalizing the meteorological effects by BAU predictions, the NO₂ reduction was 14.5% for AAMS and 19.1% for RsAMS. The decrease in NO₂ levels is more pronounced during the weekend than on weekdays.

By analyzing ground observations of NO₂ and O₃, along with BAU simulations and meteorological data from ERA5, as well as FNR and HCHO data from S5P TROPOMI, we found that the reduction in NO₂ levels during the lockdown did not immediately result in an increase in O₃. Instead, we observed that the increase in O₃ occurred after the lockdown, specifically in August when sunny conditions were reinforced. This finding is significant for Japan, as it has not been previously reported in other studies.

Furthermore, when analyzing the ground observations of NO₂, CO, and CH₄ alongside BAU simulations and model-simulated CH₄ emissions from wetlands, we found that the changes in NO₂ and CO contributed marginally to the variations in CH₄ levels, ranging from 0.6% to 1.3%, across the study areas. This finding aligns with previous studies (Akimoto and Tanimoto, 2022; Qu et al.,

2022; Feng et al., 2023), but also differs from others where the reduction in atmospheric sink has been reported as a major contributor to increased CH₄ levels (Stevenson et al., 2022; Peng et al., 2022).

Based on the findings of this study, we recommend simultaneous reduction of air pollutants and anthropogenic VOCs as well as biogenic VOCs to mitigate the adverse effects on O₃ and CH₄. These pollutants are significant SLCPs that can have detrimental impacts on future climate mitigation efforts. Therefore, it is crucial to address both air pollutants and VOCs emissions to effectively mitigate these adverse effects in the future policies.

GREENHOUSE GAS ESTIMATION, FORECASTING AND MONITORING

5 Plant functional types mapping

We proposed a combined machine learning approach with a deep convolutional neural network (CNN) to monitor forest utilization toward Sustainable Development Goals (SDGs) for data-scarce regions. First, we employed the Random Forest (RF) classifier using Google Earth Engine (GEE) for forest mapping. Then, we designed a deep CNN architecture that works for tree species/age mapping from coarse and polygonal ground-truth data. The proposed network has U-shape and comprises 3D Atrous Convolutions. The model was optimized by a weighted cross-entropy loss function. We trained the model with times-series Sentinel 1, 2, and Digital Elevation Model (DEM) data with sparse annotations. Our proposed models achieved 94.5% overall accuracy (OA) for forest mapping, 77.80% (OA) for tree species, and 81.74% (OA) for tree age classification, respectively in Ena city, Japan. The outcome of our study indicates the potential of remote sensing and machine learning in monitoring forest development, conservation, and utilization toward SDGs from coarse ground-truth data. Our source code for the implementation is available at: https://github.com/anhp95/forest_attr_segment

5.1 Introduction

Forest plays an important role in achieving SDG15 and mitigating the climate change at the global level. With the help of remote sensing and state-of-the-art machine learning algorithms, mapping the forest area with tree species/age could contribute to monitoring the forest-related SDG issues such as SDG indicators 15.1.1, 15.2.1, 15.4.2.

Forest mapping is a well-known task in land-cover/ land-use classification problems. However, tree species/age map is obtained at the cost of much higher complexity. In the previous studies in tree species/age classification, either high-

resolution input (Schiefer et al., 2020; La Rosa et al., 2021) or ground-truth data (point-level) is commonly used. These materials are considered to be expensive, time-consuming to collect, and rarely available in specific regions (e.g., developing regions). Therefore, in this chapter, we present a methodology to help monitor forest areas, tree species/age from coarse annotations and free remote sensing data. Firstly, we classify forest area using RF classifier. Subsequently, we designed a deep CNN architecture for tree species/age segmentation. We demonstrated that our proposed approach is considered to be useful for data-scarce regions.

The paper outline is described as follows. Section 5.2 presents the study area and data used in the paper. Section 5.3 provides the overall methodology and the experimental results in the study area are introduced in Section 5.4. Finally, the conclusion of the paper and the future works is remarked in Section 5.5.

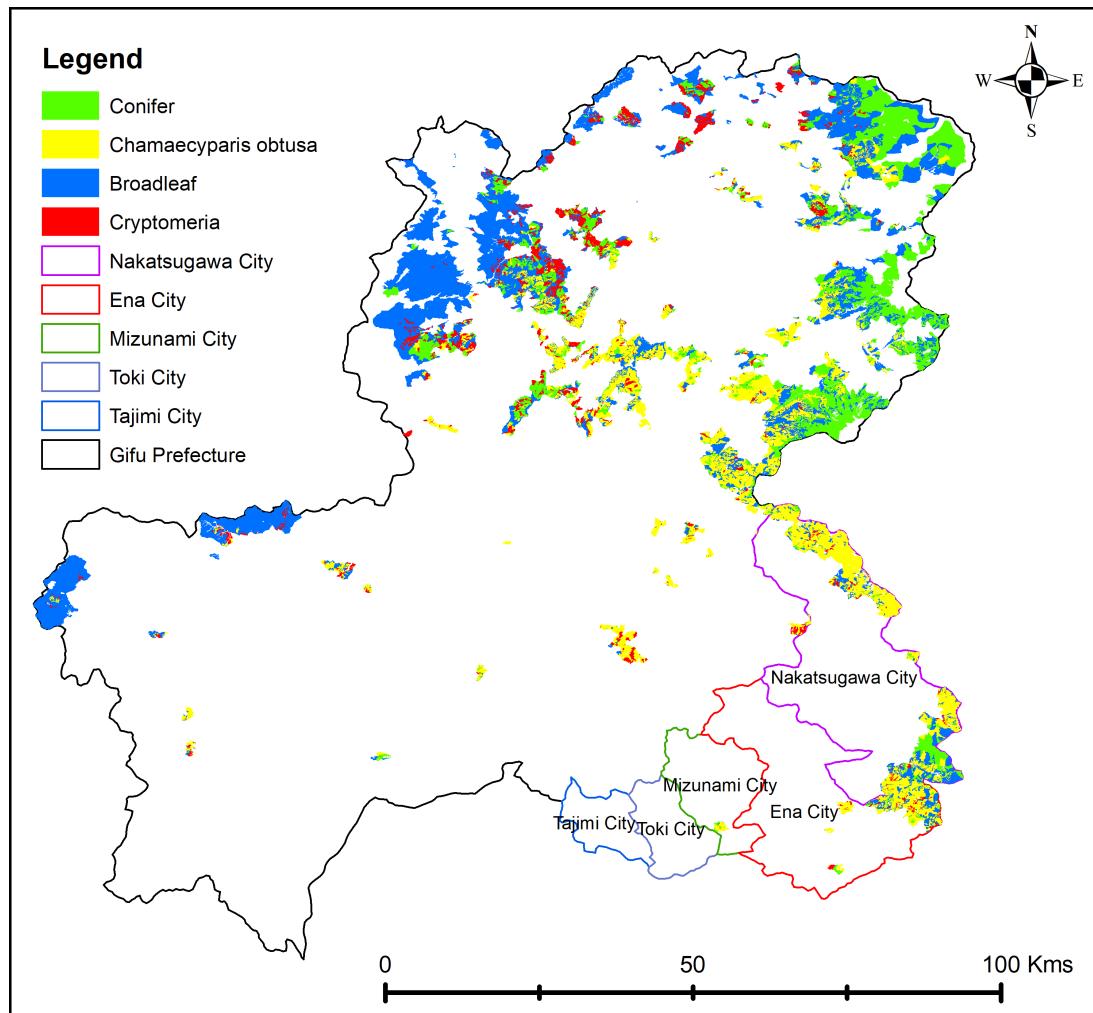
5.2 Data

5.2.1 Study area

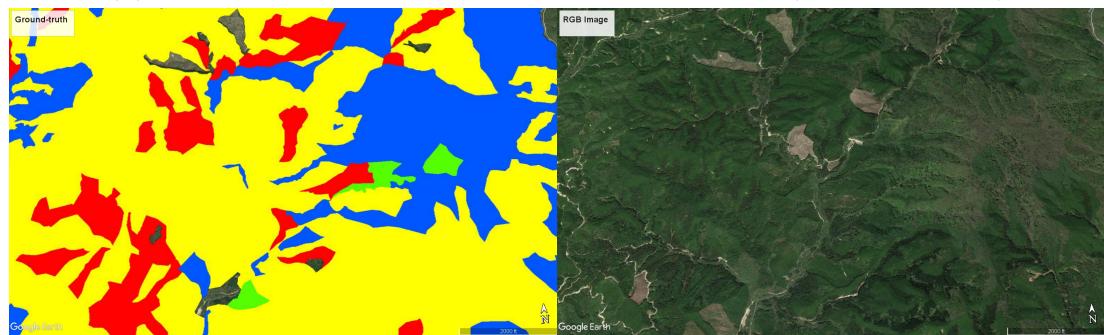
The study area is Ena city, located in the southeastern part of Gifu prefecture, middle of Japan. The total area of Ena city is approximately 504 km² with an elevation of 282 meters. The annual temperature of the city ranges from 2 °C to around 26.4 °C. According to the statistics of the local government, 60% of the forest in Ena is artificial forest which is dominated by *Chamaecyparis obtusa*. Here, the forest plays a vital role in timber production, water-related disaster prevention and CO₂ sequestration.

5.2.2 Data collection

For forest mapping, we randomly selected a training set (forest/non-forest: 750/250 points) and a validation set (forest/non-forest: 300/100 points) based on the land-use map in 2016 provided by National Land Information Portal. For tree species/age mapping, we collected the labeled data from the same resource. The ground-truth data is available as coarse polygons each covering a mixed-species zone that was annotated by the most dominant tree species of that area. Only na-



(a) Annotations of national forest in Gifu prefecture (black boundary)



(b) Example of annotated area

(c) The corresponding RGB image

Figure 5.1. (a) The designated study area outlined in red is Ena city, (b) demonstrates coarse annotations as an illustrative example, and (c) showcases the corresponding RGB image sourced from Google Earth.

tional forest data is available in this database. Due to a small portion of national forest data in Ena city (11%), we utilized the data collected in 2018 from Gifu prefecture to train the model. The remote sensing resource we applied includes Sentinel 1A, Sentinel 2 L1C, and DEM with a spatial resolution of 10m, 10m, and 30m, respectively. The dataset contains 11 spectral channels: Red, Green, Blue, Red Edge, Near-infrared, Short-wave infrared, and Normalized difference vegetation index of Sentinel 2, and VV, VH bands of Sentinel 1A, respectively. The DEM data is NASA Shuttle Radar Topography Mission digital elevation model. The details of the acquisition time of Sentinel 1,2 are described in Section 4.1 and 4.2 for tree species/age segmentation and forest mapping, respectively.

5.3 Methodology

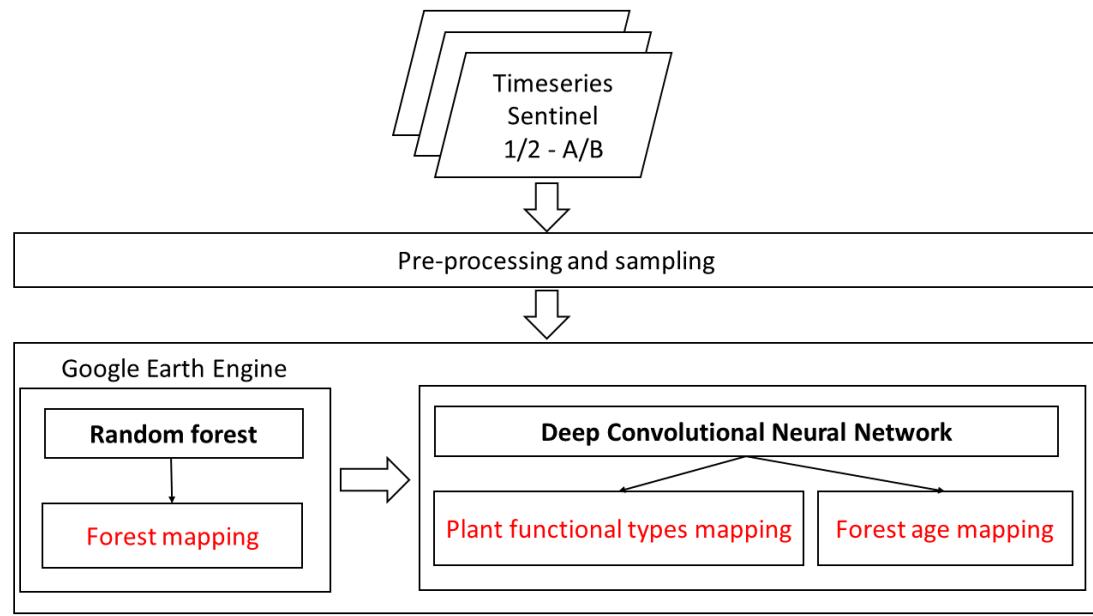
The proposed overall workflow is shown in Fig. 1. First, the Sentinel 1 data was directly downloaded from GEE as each pixel is the backscatter coefficient derived from a chain of preprocessing steps including applying orbit file, GRD border noise removal, thermal noise removal, radiometric calibration, and terrain correction. Sentinel 2 data was mosaicked and monthly average to remove clouds and missing values. The training and validation set for forest mapping; tree species/age segmentation were sampled from the satellite images. In the following sections, we illustrate how the obtained training and validation sets are utilized to train the machine learning models.

Forest mapping

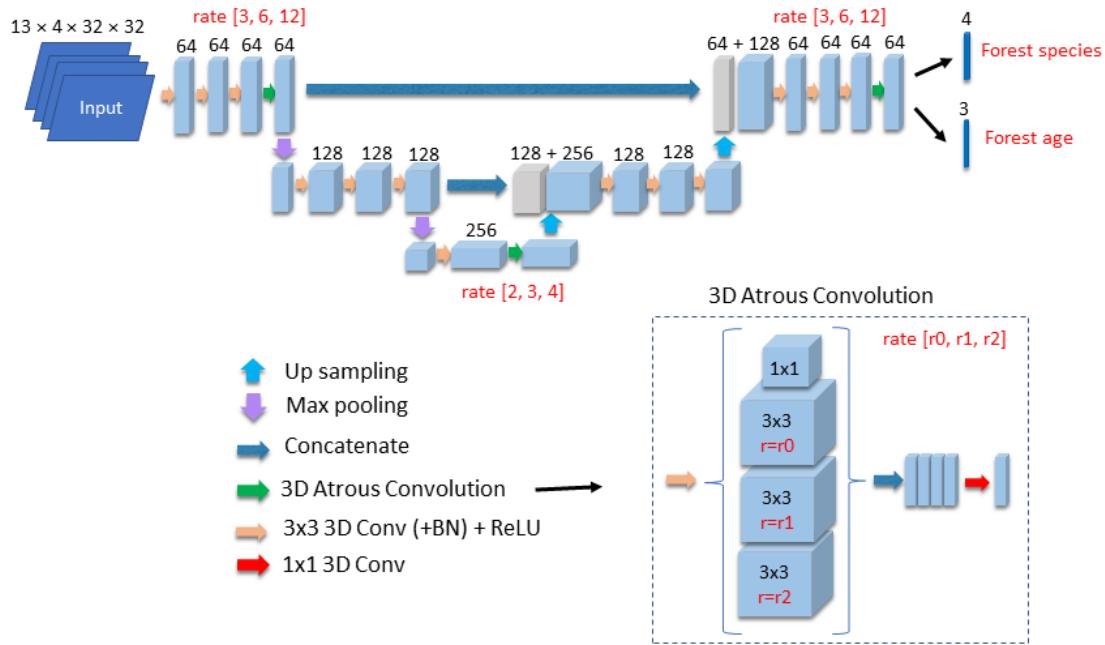
In order to quickly map the forest, we deployed the RF model which is a popular ensemble machine learning classifier for land-cover/land-use classification (Gislason et al., 2006). In addition, RF has been proven to be effective in land-cover mapping from low-resolution ground-truth data (Robinson et al., 2021).

Tree species/age mapping

Although RF is able to perform well with low-resolution labeled data, it is not sufficient to achieve a better result than our proposed deep learning model (Ta-



(a) Annotations of national forest in Gifu prefecture (black boundary)



(b) Annotations of national forest in Gifu prefecture (black boundary)

Figure 5.2. (a) The designated study area outlined in red is Ena city, (b) demonstrates coarse annotations as an illustrative example, and (c) showcases the corresponding RGB image sourced from Google Earth.

ble. 2). The proposed network was designed based on the UNET architecture (Ronneberger et al., 2015) and is shallower than the original version (see Fig. 2). We applied 3D Atrous Convolution (3DAConv) to the model as atrous convolution is proven to be effective in semantic segmentation from coarse annotations (Chen et al., 2017b). Atrous convolution was introduced in DeepLab architecture (Chen et al., 2017a) as a term for convolution with upsampled filters. The model backbone is built from encoder and decoder paths. The encoder path consists of three layers. The first one has three 3D convolutions (3DConv) followed by a 3DAConv. The second layer only contains three 3DConvs. The last layer consists of one 3DConv followed by a 3DAConv. A $2 \times 2 \times 2$ max pooling layer with strides of two follows each encoder layer. Each 3DConv is followed by a rectified linear unit (ReLU), before each ReLU is a batch normalization (BN). We avoid doubling the number of channels right before the max pooling as introduced in 3D UNET (Çiçek et al., 2016). In the decoder path, the up-convolution (ConvTranspose3D) was applied to upsample the feature map. A 3DAConv was added at the end of decoder path. The output dimensions will be reduced to the number of labels by a $1 \times 1 \times 1$ 3DConv from the last 3DAConv. The number of labels in our case is 4 for tree species: Broadleaf, Conifer, Cryptomeria, Chamaecyparis obtusa; 3 for tree age: young age (≤ 20 years), mature age (21-50 years), harvesting age (≥ 50 years).

Table 5.1. Training and validation samples and the corresponding weights for cross-entropy loss function.

Class	Training set	Validation set	Weight
Tree species (number of input images)			
Broadleaf	5017	264	0.153
Conifer	3048	160	0.252
Chamaecyparis obtusa	3191	168	0.241
Cryptomeria	768	40	1
Tree age (number of input images)			
Harvesting age	4000	205	0.05
Mature age	2095	110	0.1
Young age	186	10	1

Experiment design and settings

For tree species/age mapping, we designed the following experiment to evaluate the performance of the proposed network with the RF, 2D, 3D UNET, and our model. The time-series satellite data of Sentinel 1, 2 in 2018 was organized into three periods: January-April (P1), May-August (P2), October-December (P3). For each period, the satellite image was mosaicked and composited. We first investigate the effect of seasonal changes on the performance of tree species/age mapping from satellite data employing RF and 2D UNET. Due to the input shape constraint, we only can examine the performance of 3D UNET and the proposed model with the data of the entire year. In order to train the data with our network or 3D UNET, the input shape must be $13 \times 4 \times 32 \times 32$. In doing so, the DEM band was stacked to the Sentinel 1/2 in P1, P2, P3 (each $13 \times 32 \times 32$) to make the input trainable to the network.

To evaluate the performance of the segmentation models, we used overall accuracy (OA) score on the validation set. Finally, the results map obtained by each model were generated for visual examination.

We implemented the deep learning model in Pytorch and trained the model on NDVIA GeForce RTX 3080 Ti GPU. The model was trained with 100 epochs and was optimized by Adam optimizer with the initial learning rate 10^{-5} . The learning rate is divided by 2 after every 10 epochs.

For forest mapping, the implementation is directly supported in the GEE API which effortlessly enhances the computing performance for mapping stages.

5.4 Experimental results

For forest mapping, only Sentinel 2 in June 2018 with 10m-resampled DEM data derived from GEE were used to train the RF classifier as by our observation, June data has least slope effect in regions with high elevations. The model achieved 94.5% OA for forest/non-forest classification. The forest map produced by the model is shown in Fig. 3.

As observed from Table. 2, RF outperformed 2D UNET in all the tests. Both RF and 2D UNET experiments had a certain outcome that employed only P2 period resulting in the lowest OA in both species and age segmentation. The

Table 5.2. The experimental results of UNET and our model.

Model	Time-series period	Highest OA (%)	
		Species	Age
RF	P2	67.41	73.94
	P1 + P2	71.68	78.66
	P1 + P2 + P3	71.65	78.68
2D UNET	P2	59.81	65.67
	P1 + P2	67.25	75.4
	P1 + P2 + P3	65.02	74.55
3D UNET	P1 + P2 + P3	76.91	80.53
Our model	P1 + P2 + P3	77.80	81.74

OA is significantly improved when a longer time-series scheme is adopted from P2 to P1 + P2. Adding P3 to the training set did not enhance the performance of RF and 2D UNET compared to P1 + P2. That means, in terms of utilizing time-series data for forest species/age segmentation in the study area, it would be preferable to use the data collected from January to August period with an ensemble learning model from multiple decision trees like RF or a 2D UNET-based CNN architecture.

Despite a minor effect of P3 data to the performance of RF and 2D UNET, with 3D CNN scheme in 3D UNET and our suggested model, incorporating P3 data has essentially increased the OA of tree species/age discrimination. The performance of our model and 2D/3D UNET in 100 epochs is presented in Table. 2, Fig. 4, and Fig. 5. The OA has importantly improved from 71.68% to 76.91% with 3D UNET, 77.80% with our model for tree species segmentation, and from 78.66% to 80.53% with 3D UNET, 81.74% with our model for tree age segmentation.

The OA profiles in Table. 2, Fig. 4, and Fig. 5 show that our model outperformed RF, 2D UNET and 3D UNET by approximately 6.12%, 10.55%, 0.89% OA for tree species and 3.03%, 6.31, 1.18% OA for tree age segmentation, respectively.

Fig. 6 is the tree age map produced by our model. The major harvesting-age

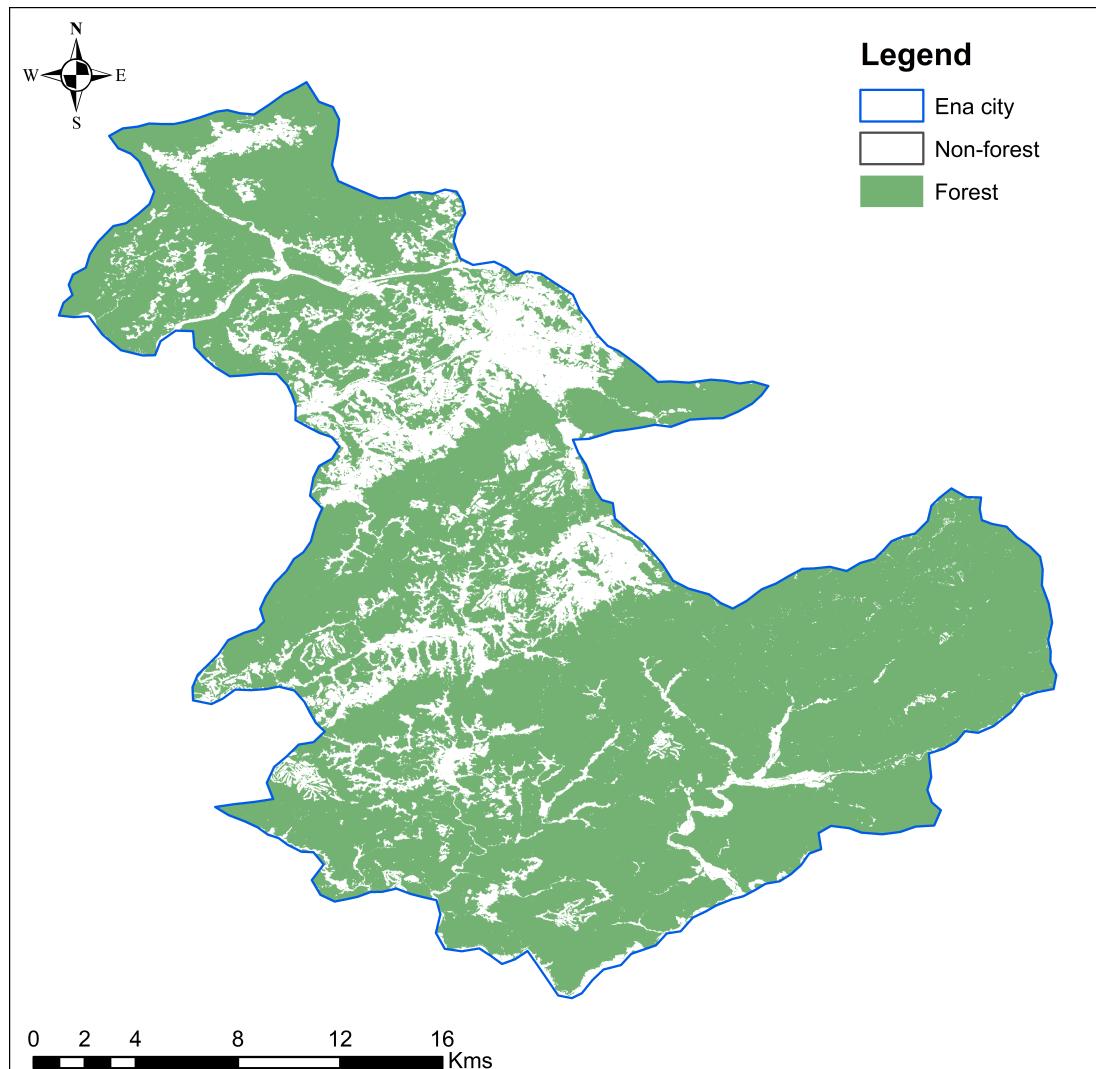
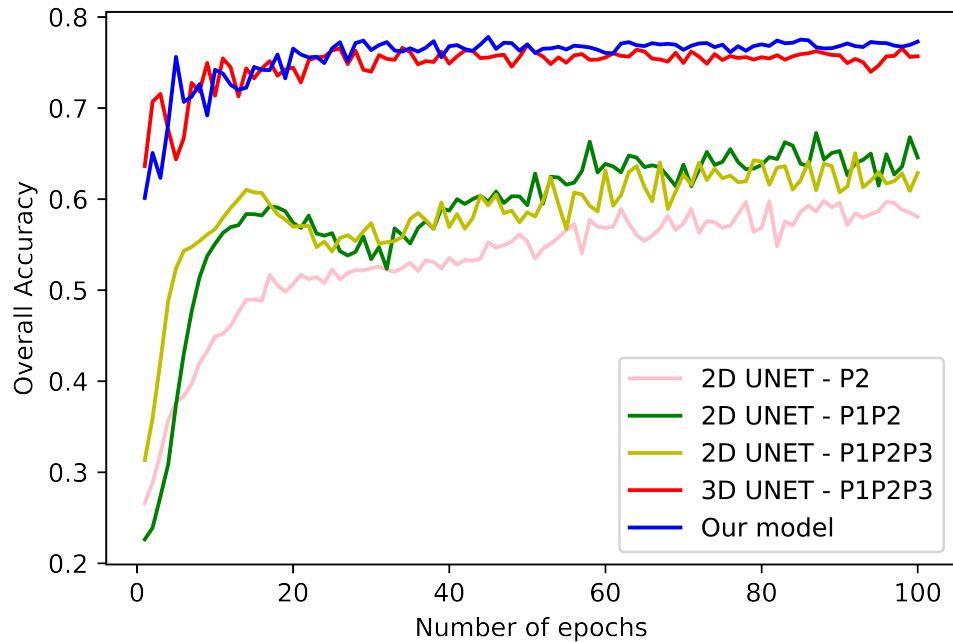


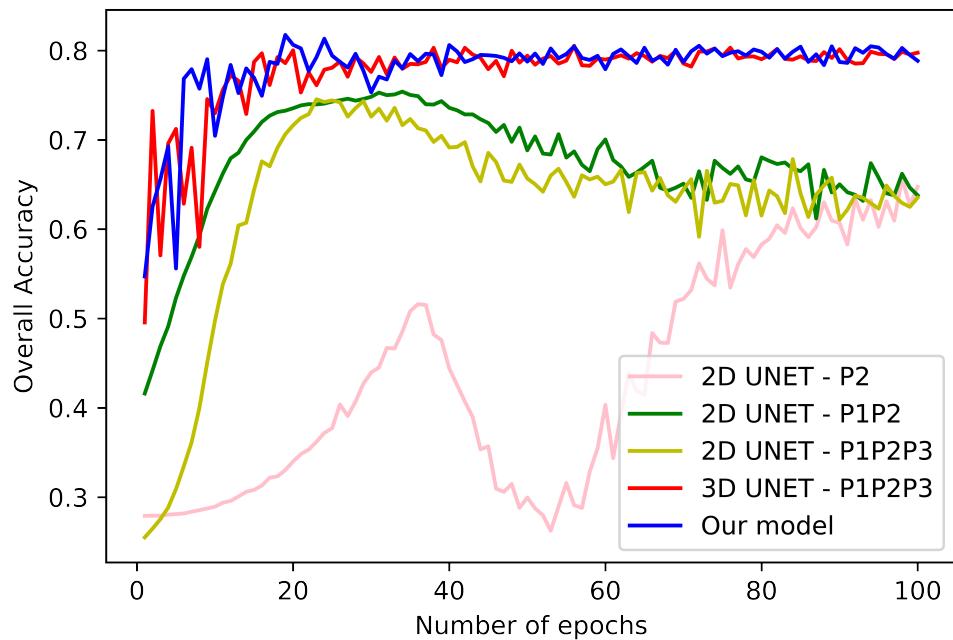
Figure 5.3. Forest map in Ena City, Japan.

area was inferred to be mostly located in the Northern, Southern, and central parts of the city. The major mature-age forest is distributed throughout the region while small areas of young-age forest are observed scattering over the city from the west to the south.

The inferred tree species map is illustrated in Fig. 7. *Chamaecyparis obtusa* is the most dominant tree and is distributed all over the region. *Cryptomeria* is predominantly occupied in the central Southeast and Northwest part of the study area. The majority of Broadleaf are mostly allocated in the Northeast,



(a) Tree species segmentation.



(b) Tree age segmentation.

Figure 5.4. OA profile of tre species (a) and tree age (b) segmentation

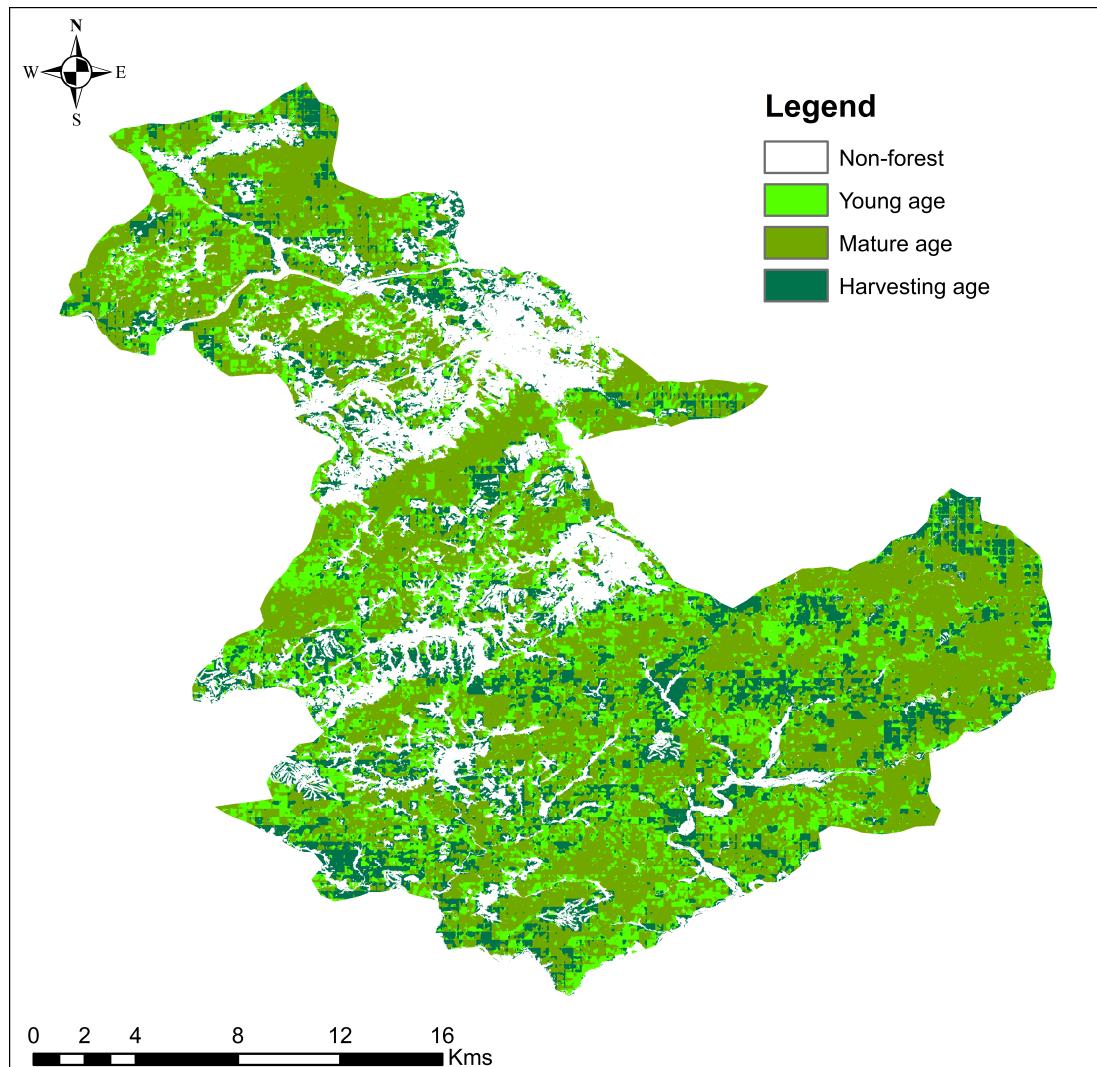


Figure 5.5. Inferred tree age map in Ena City, Japan – 2018.

Southern, and Northwest parts of the region. The detected Conifer is scattered in the area in the Northern and Southern parts, a minor contribution comes from the Northwest of the city

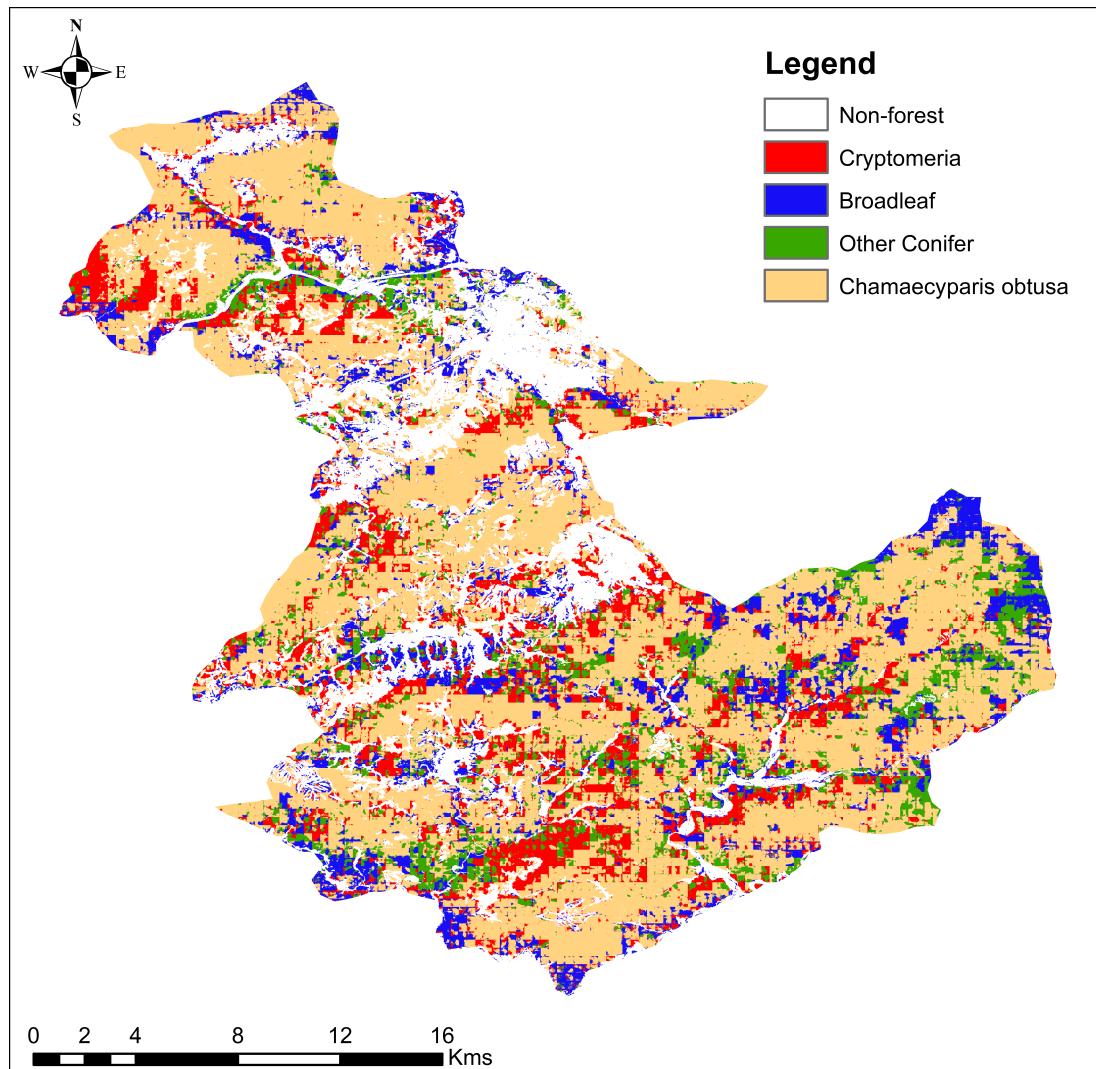


Figure 5.6. Inferred tree species map in Ena City, Japan – 2018.

5.5 Conclusion

In this study, by utilizing remote sensing, RF classifier, and deep learning, the approach for forest-related SDG issues monitoring in data-scarce regions has been proposed. We examined the approach in Ena City, Japan and achieved promising results in forest mapping, and tree species/age mapping. Our proposed model outperforms the RF, 2D/3D UNET in tree species/age segmentation with coarse-polygonal ground-truth data. The outcome of this study could be served as an

input for further steps to produce high-resolution land cover map for the data-scarce regions. In the future, we will investigate the postprocessing method to improve the map quality from coarse annotations.

6 Global upscaled of carbon fluxes

7 CO2 monitoring platform

7.1 Introduction

7.2 Method

7.2.1 Data collection

7.2.2 API development

7.3 Result and discussion

7.3.1 Result

7.3.2 Discussion

8 Conclusion

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Acknowledgements

Thank you, and thank you.

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