SIMULTANEOUSLY ESTIMATING FOOD WEB COMPLEXITY AND STRUCTURE WITH UNCERTAINTY

A Preprint

Anubhav Gupta *

Department of Evolutionary Biology and Environmental Studies

University of Zurich

8057 Zurich, Switzerland

anubhav.gupta@ieu.uzh.ch

Owen L. Petchey

Department of Evolutionary Biology and Environmental Studies

University of Zurich

8057 Zurich, Switzerland

owen.petchey@ieu.uzh.ch

May 6, 2021

Abstract

- 1) Food web models explain and predict the trophic interactions in a food web, and they can infer missing interactions among the organisms. The allometric diet breadth model (ADBM) is a food web model based on the foraging theory. In the ADBM the foraging parameters are allometrically scaled to body sizes of predators and prey. In Petchey et al. (2008), the parameterisation of the ADBM had two limitations: (a) the model parameters were point estimates, and (b) food web connectance was not estimated.
- 2) The novelty of our current approach is: (a) we consider multiple predictions from the ADBM by parameterising it with approximate Bayesian computation, to estimate parameter distributions and not point estimates. (b) Connectance emerges from the parameterisation, by measuring model fit using the true skill statistic, which takes into account prediction of both the presences and absences of links.

10

11

^{*}Corresponding author

- 3) We fit the ADBM using approximate Bayesian computation to 16 observed food webs from a wide variety of ecosystems. Connectance was consistently overestimated in the new parameterisation method. In some of the food webs, considerable variation in estimated parameter distributions occurred, and resulted in considerable variation (i.e. uncertainty) in predicted food web structure.
 - 4) We conclude that the observed food web data is likely missing some trophic links that do actually occur, and that the ADBM likely predicts some links that do not exist. The latter could be addressed by accounting in the ADBM for additional traits other than body size. Further work could also address the significance of uncertainty in parameter estimates for predicted food web responses to environmental change.
- 2 **Keywords** connectance \cdot ABC \cdot ADBM \cdot food web \cdot true skill statistic \cdot uncertainty

23 1 Introduction

12

13

14

15

16

17

18

19

20

21

Knowledge about the trophic interactions among the organisms in a community is crucial for understanding the structure and dynamics of ecological communities and for predicting their response to environmental change (Dunne, Williams, and Martinez 2002; Tylianakis and Binzer 2014; O'Connor et al. 2009; Bergamino, Lercari, and Defeo 2011; Krause et al. 2003; Lurgi, López, and Montoya 2012; Morris, Sinclair, and Burwell 2015). The network of trophic interactions is often referred to as a food web. The food web structure can provide answers to key ecological questions: which species are more vulnerable to environmental changes such as temperature (Petchey et al. 1999); how robust a food web is to extinctions (Dunne, Williams, and Martinez 2002); and how a food web reacts if the top predators are removed (Knight et al. 2005)?

Trophic interactions information from multiple sources can be used to infer a food web, e.g. gut contents 32 (Peralta-Maraver, López-Rodríguez, and de Figueroa 2017), stable isotope composition of tissues (Layman et 33 al. 2007), and experimentation (Warren 1989). Sometimes the methods used to infer the interactions may 34 lead to uncertainty in the constructed food web. E.g. In gut content analysis of some fish predators, there can be tissues which are not identifiable and cannot be assigned with certainty to a specific prey item (Baker, Buckland, and Sheaves 2014). With stable isotope ratios of tissues, uncertainty may be due to factors such as variability in the isotopic fractionation values across multiple combinations of diets and tissues/species, unquantified temporal or spatial variation in prey isotopic values, and variation caused by routing of particular 39 dietary nutrients into particular tissues (Crawford, Mcdonald, and Bearhop 2008). Furthermore, complete 40 recording of all interactions usually requires a large sampling effort even at small spatial and temporal scales (Hobson, Piatt, and Pitocchelli 1994). Food web structure is very difficult to record at larger spatial and temporal scales without losing resolution (spatial, temporal, and taxonomic) (Gravel et al. 2013; Martinez 1991; Jordán and Osváth 2009). Less than complete sampling of interactions can result in no interaction being observed between a pair of individuals that in fact do interact, which results in missing links in a food web. Due to under-sampling, food webs can be poorly understood, which may hinder further advances in the field (Martinez et al. 1999).

When interactions are difficult to observe, and hence well-documented food webs are not available, models which predict species interactions may provide a solution (Tamaddoni-Nezhad et al. 2013; Gravel et al. 2013; Petchey et al. 2008; Allesina, Alonso, and Pascual 2008; Cohen, Newman, and Steele 1985). A food web model can be used to predict missing information about species interactions. For example, Petchey et al. (2008) showed how a model of species interactions (and therefore food web structure) could be parameterised from data on the known presence and absence of trophic interactions. The model and its parameter values encode the rules for occurrence or absence of species interactions to predict food web structure. Observed data may be used to select and parameterise the model. Tamaddoni-Nezhad et al. (2013) used large agricultural datasets, logic-based machine learning and text mining to assign interactions between nodes to automatically construct food webs. Gravel et al. (2013), inspired by the niche model of food web structure developed a method that used the statistical relationship between predator and prey body size to infer the food web.

Food web models are also useful for ecological forecasting. Lindegren et al. (2010) used a stochastic food web model driven by regional climate scenarios to produce quantitative forecasts of cod dynamics in the twenty-first century. Hattab et al. (2016) forecasted the potential impacts of climate change on the local food web structure of the highly threatened Gulf of Gabes ecosystem, located in the south of the Mediterranean Sea. Hence, food web models have an important role in filling gaps in knowledge about species interactions, including predicting future changes in food web structure.

The allometric diet breadth model (ADBM) was the first model able to predict food web complexity and structure (Beckerman, Petchey, and Warren 2006; Petchey et al. 2008). The ADBM uses foraging theory, specifically the contingency model (MacArthur and Pianka 1966), to predict which set of the available prey species would be consumed by a predator. This set is the prey species that maximises the energy intake rate of the predator. The model requires the foraging related traits of species, such as energy content of a potential prey item, the rate of space clearance (also known as attack rate), the density of prey items, and handling time (the amount of time required to handle food items). The model is termed "allometric" because each of these quantities is derived from the body size of the prey and predator using several allometric relationships. The ADBM has also been used to investigate the effect of temperature on an observed food web structure (O'Gorman et al. 2019).

The ADBM had variable success in explaining the structure of 15 different food webs, with the proportion of links correctly predicted ranging from 5 % to 65 % (Table 1). The ADBM correctly predicted 65% of the presence of links in the Coachella valley food web. The poorest prediction of presence of links was for the Grasslands food web with only 7% of the presence of links correctly predicted. When trophic interactions

were more strongly dependent on size, the model correctly predicted a greater proportion of links. Indeed, constructing a food web based only on body size (i.e. ignoring taxonomy) resulted in almost twice the number of correctly predicted links, i.e. 83%, in contrast to taxonomy (Woodward et al. 2010).

gg

Although Petchey et al. (2008) demonstrated that foraging theory could predict food web structure, their implementation of the ADBM included at least two limitations. The parameterisation method provided estimates of the parameters with no uncertainty: a single set of parameter values that maximised the explanatory power was selected. In other words, the parameterisation method led to point estimates of the parameters that predicted a single food web structure (because the ADBM is purely deterministic). Moreover, the best predicted food web was not exactly the same as the observed one. In a sense then, the parameterisation method used in Petchey et al. (2008) was akin to estimating the intercept and slope of a regression line, but not any uncertainty in those parameters. Given that uncertainty is an essential dimension in ecological models, and in predictions about the future state of ecological communities (Petchey et al. 2015; Carpenter 2016), this is an important limitation.

The second limitation was in the estimation of the connectance of the food web, which is the number of realised trophic links divided by the total number of possible trophic links. Although the ADBM can in principle predict connectance, Petchey et al. (2008) prevented the model from doing so. They set the value of relevant parameters in the model to instead ensure the predicted connectance was equal to the observed connectance. The ADBM was not therefore used to simultaneously predict complexity and structure of food webs. Moreover, fixing predicted connectance to be equal to observed connectance does not account for the possibility that the observed connectance was imperfectly measured. Indeed, if low effort was used to observe the trophic links in a community, the observed connectance are likely to be lower than if all trophic links were observed. Connectance is an important driver for the stability and dynamics of a food web (May 1972) and most of the structural properties of food webs co-vary with connectance (Dunne, Williams, and Martinez 2002; Poisot and Gravel 2014), thus this limitation must be addressed.

In this article we report on how we address these limitations. We removed the first limitation by applying approximate Bayesian computation (ABC). The approach originated in population genetics and has been used in a wide range of research fields: systems biology (Toni et al. 2009), ecology (Jabot and Chave 2009), epidemiology (Shriner et al. 2006) and ecological networks (Ibanez 2012; Poisot and Stouffer 2016). One of the advantages of ABC is that it does not require a likelihood function. As ADBM is a complex deterministic model where the likelihood can not be explicitly evaluated, ABC is a good choice of parameterisation.

We addressed the second limitation by allowing estimation of number of links as well as arrangement of links. To accomplish this, we measured model fit by using the true skill statistic, which takes into account both the number of presences and absences of links correctly predicted. High values of the true skill statistic

occurs when both the predicted arrangement of links and the predicted number of links are close to the

observed arrangement and number of links, respectively.

Table 1: Information about the food webs predicted using the ADBM.

Common food web name	Predation matrix source	Body size source	General ecosystem	Number of species	Connectance	Body size range (approximate)	Proportion of presence of links correct	Type of interactions
Benguela Pelagic	(Yodzis 1998)	(Yodzis 1998)	Marine	29	0.23	$10^{-8} \text{ to } 10^{6}$	0.57	Predation
Broadstone Stream (taxonomic aggregation)	(Woodward and Hildrew 2001; Woodward et al. 2005)	(Brose et al. 2005)	Freshwater	29	0.19	$10^{-6} \text{ to } 10^{-2}$	0.62	Predation
Broom	Memmott et al. 2000)	(Brose et al. 2005)	Terrestrial	68	0.02	$10^{-6} \text{ to } 10^{0}$	0.08	Herbivory, Parasitism, Predation, Pathogenic
Capinteria	(Lafferty et al. 2006)	(Reide, unpublished)	Marine (Salt Marsh)	72	0.05	$10^{-14} \text{ to } 10^5$	0.16	Predator-parasite, Parasite-parasite
Caricaie Lakes	(Cattin et al. 2004)	(Brose et al. 2005)	Freshwater	158	0.05	$10^{-5} \text{ to } 10^{5}$	0.13	Predation, Parasitism
Coachella Valley	(Polis 1991)	(Reide, unpublished)	Terrestrial (Desert)	26	0.34	$10^{-8} \text{ to } 10^4$	0.65	Herbivory, Predation
EcoWEB41	(Cohen 1989)	(Jonsson 1998)	Marine	19	0.14	10^{-11} to 10^6	0.47	Predation
EcoWEB60	(Cohen 1989)	(Jonsson 1998)	Terrestrial	33	0.06	10^{-5} to 10^{6}	0.24	Predation, Parasitism, Herbivory
Grasslands	(Dawah et al. 1995)	(Brose et al. 2005)	Terrestrial	65	0.03	10^{-3} to 10^{-2}	0.07	Herbivory, Parasitism
Mill Stream	(Ledger, Edwards, Woodward unpublished)	(Brose et al. 2005)	Freshwater	80	0.06	10^{-6} to 10^{-1}	0.37	Herbivory, Predation
Sierra Lakes	(Harper-Smith et al. 2005)	(Brose et al. 2005)	Freshwater	33	0.16	$10^{-4} \text{ to } 10^{0}$	0.60	Predation
Skipwith Pond	(Warren 1989)	(Brose et al. 2005)	Freshwater	71	0.07	10^{-4} to 10^{-1}	0.14	Predation
Small Reef	(Opitz 1996)	(Reide unpublished)	Marine (Reef)	50	0.22	10^{-11} to 10^5	0.41	Predation, Herbivory
Tuesday Lake	(Jonsson et al. 2005)	(Brose et al. 2005)	Freshwater	73	0.08	10^{-11} to 10^3	0.46	Predation
Ythan	(Emmerson and Raffaelli 2004)	(Emmerson and Raffaelli 2004)	Marine (Estuarine)	88	0.05	10^{-12} to 10^0	0.22	Predation
Broadstone Stream (size aggregation)	(Woodward et al. 2010)	(Woodward et al. 2010)	Freshwater	29	0.24	$10^{-7} \text{ to } 10^2$	0.83	Predation

115 2 Materials and Methods

131

In the upcoming sections, we present a detailed account of the application of ABC to parameterise the ADBM, the description of the ADBM and of the food web data we used. We explain the rejection Monte Carlo ABC method in the main text, and Markov chain Monte Carlo ABC and sequential Monte Carlo ABC methods in the Supplementary information (hereafter SI) Section S1 (hereafter SI-S1). We computed an accuracy measure known as true skill statistic to assess the ADBM's predictions and calculated different food web properties to compare these predictions across food webs.

2.1 Allometric Diet Breadth Model (ADBM)

The allometric diet breadth model (ADBM) is based on optimal foraging theory, specifically the contingency foraging model (MacArthur and Pianka 1966). The ADBM predicts the set of prey species a consumer should feed upon to maximise its rate of energy intake (Petchey et al. 2008) (hereafter referred as PBRW study). The species in this set are assumed to have the trophic link with the predator. To make these predictions, the model assumes that a foraging predator is in one of two exclusive states: searching for prey or handling a prey item. The model requires four variables for each potential predator-prey interaction:

- The energy content of the resources E_i (only prey *i* specific) (energy).
- The handling times H_{ij} , which is the time not spent searching caused by consuming a prey item (prey i and predator j specific) (time).
- The space clearance rates A_{ij} (also known as the attack rate; prey i and predator j specific)

 (area or volume per time).
- The prey densities N_i (only prey i specific) (individuals per area or volume).

The term "Allometric" in the ADBM refers to the use of four allometric relationships, one for each of these four variables, including the body size of the predator M_j and prey M_i (Table 2). With these four allometric relationships, and the body size of each of the species in a community, we can predict the four variables (energy, handling time, space clearance rate, and prey density), and then use the contingency foraging model to predict diets.

Traits (Unit)	Allometric function	Parameters	Comments
Energy (Joules)	$E_i = eM_i$	e	Arbitrary. No effect on structure
Abundance (individual/ m^2 or individual/ m^3)	$N_i = nM_i^{n_i}$	n	Connectance affected by the product nah*
		n_i	Assumed value of $\frac{-3}{4}$ based on empirical data
Space Clearance Rate $(m^2/s \text{ or } m^3/s)$	$A_{ij} = aM_i^{a_i}M_j^{a_j}$	a	Connectance affected by the product nah^* ;
			Estimated using ABC
		a_i	Estimated using ABC
		a_j	Estimated using ABC
Handling time (s)	$H_{ij} = \frac{h}{b - \frac{M_i}{M_j}}$ if $\frac{M_i}{M_j} < b$	h	Connectance affected by the product nah^*
	$H_{ij} = \infty \text{ if } \frac{M_i}{M_i} \ge b$	b	Estimated using ABC

Table 2: Traits with their allometric function and corresponding parameters in ADBM.

Each of the four allometric equations has parameters: a constant and/or at least one exponent (Table 2). It is the value of some of these parameters that can be estimated to have the predicted food web structure match (as closely as possible) the structure of an observed food web. This is akin to choosing values of slope and intercept of a linear regression that maximises the fit of the regression line to the observed data.

Because some of the allometric constants and exponents are known, and because others are redundant with respect to each other (see Table 2 for details), we estimate only the following parameters: a, a_i , a_j and b in the model (Table 2).

2.2 Observed food web data

The observed food webs that we fit the ADBM to belong to marine, freshwater and terrestrial ecosystems (Table 1). The observed connectance of these food webs is from 0.02 to 0.34 and there are 19 to 158 species. The food webs contain primary producers, herbivores, carnivores, parasites, and parasitoids. They also contain various types of feeding interactions, including predation, herbivory, bacterivory, parasitism, pathogenic, and parasitoid.

The goodness of fit of the ADBM's predictions depends on the types of interactions in the food webs in the PBRW study. Because some of the interactions are more size structured than other interactions. Predacious and aquatic herbivore interactions were predicted better than parasitoid and herbivory ones (PBRW study). Moreover, the ADBM's predictions improve when the species (taxonomic) are replaced by size classes (ignoring taxonomy) (Woodward et al. 2010).

All food webs with one exception (Broadstone Stream) was available only at the species level, i.e. with information about interactions between species and the body size of species. We use the

term "species" in this study to indicate a "node" in a food web in which nodes are connected by trophic interactions, and nodes are a collection of individuals that share links. These species/nodes are not always taxonomic species, but can be broader taxonomic ranks.

In contrast, the Broadstone Stream food web data contained interactions between individuals and the individual body sizes. Thus, the Broadstone Stream food web can be constructed by aggregating by either taxonomy or size (Woodward et al. 2010). The ADBM can predict the food web irrespective of the aggregation method, and aggregation by size, ignoring taxonomy led to higher explanatory power i.e. the match between observed and predicted food web structure was higher. With aggregation by size, 83% of the links were correctly predicted than with taxonomic aggregation where 40% of links were correctly predicted (Woodward et al. 2010).

2.3 Parameter estimation: Approximate Bayesian Computation

We used approximate Bayesian computation (ABC) to identify sets of parameter values that resulted 174 in predicted food webs that were close in structure to the observed food web. ABC is an approach 175 that does not require a likelihood function. Instead, there is a distance function that measures the 176 distance between a model's prediction and the observed data. The approximation of the likelihood 177 depends on the ABC method used, which is further discussed below and SI. The model parameter 178 values are sampled from a prior distribution. The accepted parameter values form an approximate 179 posterior distribution for the model parameter. We implemented three ABC methods to parameterise 180 the ADBM: namely rejection Monte Carlo (Fig. 1), Markov chain Monte Carlo, and sequential 181 Monte Carlo. The three methods produced very similar results (SI Figs S33-S34) and we therefore 182 only include the simplest (rejection) in this main text. 183

184 2.3.1 Prior distribution

166

167

168

169 170

171

172

The prior distributions for a_i and a_j were chosen to be uniform distributions. The range of distribution was from -1.5 to 1.5 and 0 to 3 for a_i and a_j respectively, informed by the estimates in Rall et al. (2012). However, we chose a prior range specific to food webs for the parameter b because body size varies greatly among the species in the observed food webs. For example: in the Benguela Pelagic food web, the body sizes of species range from the order of 10^{-8} gm to 10^{5} gm. Hence, the range of prey-predator ratio was from the order of 10^{-14} to 10^{14} . To take this into

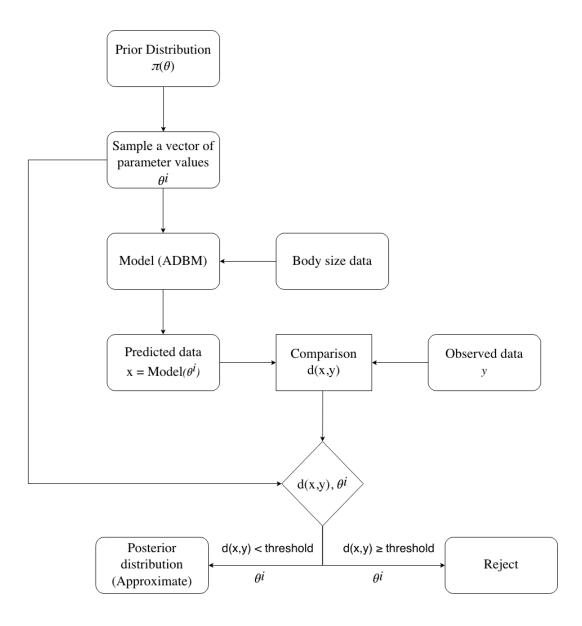


Figure 1: Flowchart of rejection approximate Bayesian computation method implemented to parameterise the ADBM.

account, we took the prior of $log_{10}(b)$ from a uniform distribution ranging from -15 to 15. In the case of parameter a, we chose the prior of $log_{10}(a)$ to be a uniform distribution. However, the prior range varied between food webs. For example, the prior range for Benguela Pelagic was chosen to be -12 to 10. The method for choosing the specific range of the prior distribution of the a parameter is detailed in SI-S4.

196 2.3.2 Comparison of observed and predicted

The difference between the model's prediction and the observed data (e.g. the sum of squared residuals is such a distance in linear regression) is quantified by a distance measure. The distance is lower when there is a closer match between the model's prediction and the observation. A perfect match would result in zero distance.

The magnitude of the distance is used for the acceptance or rejection of a set of parameter values. An accepted set of parameter values contributes to the posterior distribution, rejected ones do not. This makes the distance measure one of the important features of ABC. A threshold distance is chosen, and if the distance for a particular set of parameter values is less than the threshold, then that set of parameter values contributes to the posterior distribution. When the distance is greater than the threshold, the parameter values do not contribute to the posterior. Hence, the magnitude of the distance threshold determines the proportion of a model's parameters that are accepted. A higher threshold causes a high proportion of acceptances but less accuracy with the acceptance of some parameter sets that result in predictions quite unlike the observed data. Below, we first describe and justify our choice of distance measure, and then our choice of threshold.

In the PBRW study the measure of distance was equivalent to 1 - a/(a + c), where a is the number of observed links that were predicted (the number of true positives) and c is the number of observed links that were not predicted (the number of false negatives). A distance of 0 indicates that all observed links were correctly predicted. One way for the ADBM to achieve this is to predict that every species has a trophic link with every other species including itself – a fully connected food web with connectance of 1. The PBRW study prevented this by constraining the number of predicted links to be equal to the number of observed links, i.e. the model connectance was fixed to be the same as the observed connectance. In this study, we relaxed this constraint, with the

number of links as well as the arrangement of links being estimated. The first step was to choose an appropriate distance measure.

The distance measure used in this study is 1 minus the true skill statistic: distance = 1 - TSS.

This distance ranges from 0 to 2.

TSS is defined as:

$$TSS = \frac{ad - bc}{(a+c)(b+d)}$$

where a is the number of observed links that are predicted by the model (true positives), d is the number of observed absences of links that are correctly predicted (true negatives), b is the number of false positives, and c is the number of false negatives.

The TSS ranges from -1 to 1, where +1 indicates a perfect prediction. A TSS value of zero or less indicates a performance no better than random.

The inclusion of true and false negatives in the distance measure means that the best theoretically possible prediction (smallest distance) is a unique prediction, and specifically the one in which the predicted presence and absence of links matches exactly with the observed presence and absence of links.

Food web dynamics and stability are strongly dependent on connectance (May 1972), we therefore set the distance threshold (for acceptance) such that the model had a reasonable chance of predicting the observed value of connectance.

To do this, we examined how the predicted connectance varied with the distance threshold. An example of this relationship is given in Fig. 2 for the Benguela Pelagic food web. We chose the minimum threshold value that gave a range of predicted connectance containing the observed connectance.

Furthermore, it is useful to note that in Fig. 2 there are no connectance values below a distance threshold value of less than 0.5 because for this particular food web there were no sets of parameter values that achieved a better model fit than is indicated by 1 - TSS = 0.5. I.e. it is impossible for the ADBM to make better predictions than this. One reason for this is that the ADBM, when body

size is the only trait, can only predict contiguous diets in trait space, whereas the observed data contains gaps in the diet.

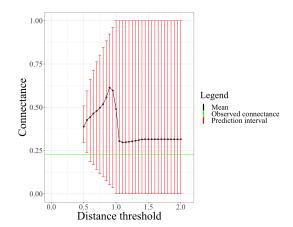


Figure 2: The prediction interval of the predicted connectance increases with increasing distance threshold for the Benguela Pelagic food web. The green line and black line represent the observed connectance and mean of predicted connectance respectively.

246 2.3.3 The Rejection ABC method

In the rejection ABC method, a set of parameter values are sampled from the prior distributions. This set of parameter values is either accepted, and thereby added to the posterior distribution of the parameter values, or it is rejected (based on if the distance 1 - TSS is less than or greater than the threshold distance, as mentioned above). This process is repeated until there are enough acceptances to give stable (approximate) posterior distributions. In addition, we used a kernel function that assigns weight to each set of parameter values, where the weight is inversely proportional to the distance (1 - TSS).

In the upcoming section, we further detail the rejection ABC method.

Properties:

254

255

256

257

258

259

- A prior distribution $\pi(\theta)$: π is the uniform distribution for parameters $\theta = (a, a_i, a_j, b)$
- A model prediction $model(\theta)$: ADBM(θ). This is a predicted food web, x_i , given by a particular set of parameter values θ_i . Hence, $x_i = ADBM(\theta_i)$
- A summary statistic s(x): x is the predation matrix predicted by the ADBM.

A kernel function
$$K(u)$$
 : epanechnikov $K(u)=\frac{3}{4}(1-\frac{u}{tol})$ if $u\leq tol$ = 0 otherwise

- where tol is the distance threshold
- A distance function $d(x_i, y)$: $d(x_i, y) = 1 TSS(x_i, y)$
- An observed food web y, in the form of a predation matrix containing zeros and ones.
- Sampling:
- for $i = 1 \dots n = 1000$
- Draw a set of parameter values θ_i from the prior distribution $\pi(\theta)$.
- Compute the model result $x_i = model(\theta_i)$
- Compute $s(x_i)$ and $d(s(x_i), s(y))$
- Accept or reject the parameter set probabilistically:
- Assign a probability p_i to θ_i as per the kernel K; p = K(d), where d is the distance evaluated in the previous step.
- Compute $\alpha \sim U(0,1)$
- If $p_i \leq \alpha$, then accept θ_i and i = i + 1
- Output:
- An approximate joint posterior distribution using the accepted $\theta_1, \ldots, \theta_n$.

276 2.4 Assessment of model fit

- Accuracy is how close the model prediction is to the observation. The ADBM's prediction is a predation matrix that consists of the presence and absence of links thus comparing how close the prediction is to the observation is not straightforward as comparing two numerical values. We defined the accuracy of the ADBM using true skill statistics to take into account the true and false predictions of both the presence and absence of links, which is defined above.
- We examined how closely structural properties of the predicted food web matched those of the observed food webs. We evaluated properties such as proportion of basal species, proportion of

intermediate species, proportion of top species, proportion of herbivores, mean omnivory, clustering coefficient, standard deviation of generality, standard deviation of vulnerability, diet similarity, mean path length and nestedness. We did not compute mean trophic level and maximum trophic level because their computation did not converge in the R *cheddar* package (Hudson et al. 2013) for all the food webs.

We investigated the performance of the ADBM parameterised with the ABC by computing standardised error of the food web properties, where the standardised error is the absolute raw error (the difference between observed and predicted value) divided by the maximum absolute raw error for that property. We did not calculate the standardised error for mean omnivory and mean path length because it had some NA values and infinite values for all the food webs respectively.

3 Results

As an example of the model outcomes, we first present the results for the Benguela food web (e.g. predicted food web structure, variation in predicted food web structure, and posterior parameter distributions). We chose this food web as it was well explained using the method of Petchey et al (2008) (hereafter referred as PBRW study). The results of the other food webs are included in the SI Figs S1-S32. We then compare model outcomes across all empirical food webs between the PBRW study and our current work. We compare the true skill statistic of the two approaches and compare some food web properties, such as proportions of basal, intermediate, and top species.

The true skill statistic (TSS) of the predicted Benguela Pelagic food web varied between 0.4 and 0.52. This variation in the TSS is represented in terms of predation matrices displayed in Fig. 3(a), which overlays 1000 independent predation matrices created from the posterior parameter distributions. In all the 1000 independent predation matrices, the predicted links are mostly present in the upper triangular portion of the matrix where most of the observed links are also present. Links in the upper right triangle of the predation matrix are for predators feeding on prey smaller than themselves.

In the 1000 predicted predation matrices, there predators are sometimes smaller than their predicted prey, the links in the lower left triangle of the predation matrix. This is also portrayed in the marginal distribution of $log_{10}(b)$ in Fig. 4(d), as it includes values greater than b=2

 $log_{10}(b) = 0.3$). This is relevant as values of b = 2 make the most profitable prey item equal in size to the predator size. Lower values of b make the most profitable prey item smaller than the size of the predator.

There were around 250 potential links in the lower left triangle of the predation matrix that were never predicted in any of the 1000 predicted predation matrix (Fig. 3(b)). This strongly suggests that the predator-prey size ratio of these links is so small (i.e. very large prey, very small predator) that the links cannot occur, given that the preponderance of observed links are predators consuming prey smaller than themselves.

The marginal posterior of parameter b in the Benguela Pelagic food web was more constrained than the marginal posterior distribution of the other three allometric parameters (Fig. 4) as the posterior range was the narrowest.

The mean true skill statistic using the ABC approach was higher than the point estimates from the PBRW study (Fig. 5(a)) across all food webs except one. Our present approach led to estimates of connectance greater than the values of connectance of the PBRW study, which were fixed to equal the observed values of connectance.

We did not find a consistent relationship between the parameters estimated using the current approach and those estimated in the PBRW study (Fig. 5(c-f)), except for in the case of parameter b. The mean using the ABC approach was always higher than the estimates from the PBRW study (Fig. 5(f)) and the 95% credible interval of the posterior of b includes the estimate from the PBRW study.

The marginal posterior of parameter b was more constrained than the other three allometric parameters, i.e. the posterior range was the narrowest (SI Figs S17-S32). In most of the food webs, the parameter b had a unimodal distribution (SI Figs S17-S32). EcoWEB60 and Grasslands had a bimodal distribution and Sierra Lakes had three modes.

The structural food web properties proportion of intermediate species, mean omnivory, clustering coefficient, sd of generality, sd of vulnerability, diet similarity and nestedness estimated from the current ABC approach were generally higher than the PBRW study (SI Fig. S36(b, e-j)). The properties proportion of basal species, proportion of top species, and proportion of herbivores were generally lower (SI Fig. S36(a, c, d)).

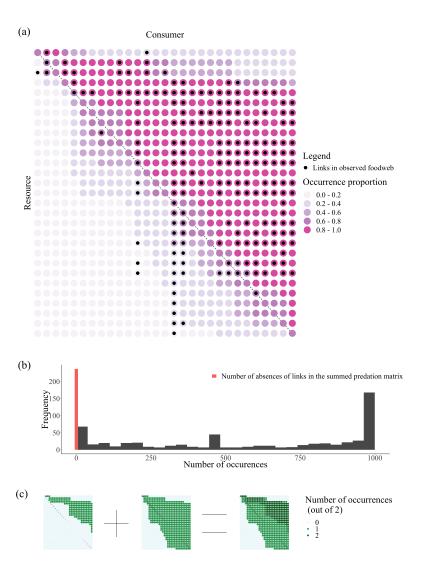


Figure 3: (a) Observed and predicted predation matrices for Benguela Pelagic food web. Body size increases from left to right and top to bottom along the predation matrix. Black circles show where there is an observed trophic link. The intensity of the pink circles shows the proportion of 1000 predicted food webs that had a trophic link between the corresponding species. This type of overlay is shown for two examples predicted in panel (c). (b) The histogram of the number of times a link was predicted across 1000 independently predicted food webs. There were 841 species pairs in this food web. About 150 of these were predicted to have a trophic link in all 1000 predicted predation matrices. The red bar shows the number of pairs of species for which a trophic link was never predicted. (c) Two predicted predation matrices for Benguela Pelagic food web corresponding to the minimum and the maximum value of estimated b, and their sum.

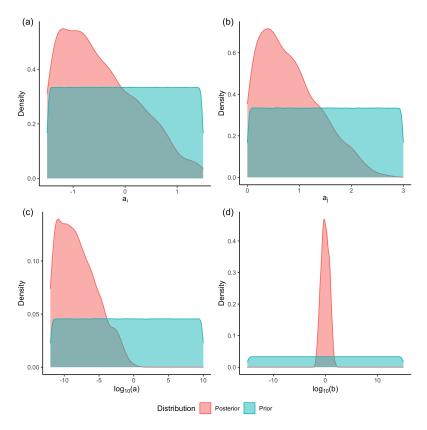


Figure 4: Marginal prior and marginal posterior distribution of the ADBM parameters for the Benguela Pelagic food web estimated using rejection ABC.

The real values of the proportion of intermediate species, mean omnivory, clustering coefficient, sd of generality, sd of vulnerability and nestedness, were mostly within the lower range of the predicted 95% interval. The proportion of basal species, proportion of top species, proportion of herbivores were underestimated in comparison to the real values for most of the food webs.

The ADBM, when parameterised with the ABC, generally better predicted the structural food web properties, such as proportion of basal species when the true skill statistics was higher (Fig. 6(a)) across the 16 food webs. However, the ABC parameterised ADBM less accurately predicted food web properties on average than in the PBRW study (Fig. 6(b)).

Within each food web, we found various relationships between the standardised error and true skill statistic (SI Figs S37 and S38). E.g. For Skipwith Pond food web (SI Fig. S37(l)), high values

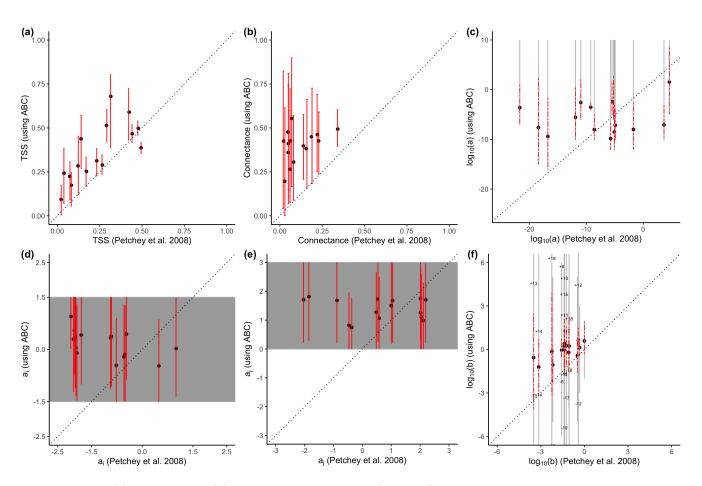


Figure 5: TSS (a), connectance (b) and ADBM parameters (c, d, e, f) computed using the ABC method compared with the corresponding point estimates from Petchey et al (2008). The red lines are the 95% credible/prediction intervals and the black filled circles represent the corresponding means. The grey region represents the intervals of the prior distributions for a_i and a_j . The grey lines represent the prior range of the parameters a and b in the log_{10} scale. The prior range for the parameter b extends above and below the y-axis limits for some food webs and so the values of the limits are shown on the plot. The dashed black lines are the 1:1 relationships for reference.

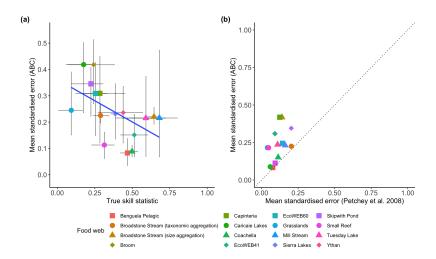


Figure 6: (a) The mean standardised error of the food web properties predicted from the ADBM parameterised using rejection ABC plotted against the mean true skill statistic for each food webs. The vertical and horizontal bars correspond to 95% prediction intervals of the standardised error and true skill statistic respectively. Solid blue line is linear regression through the means (t = -2.44, df = 14, P = 0.028). (b) The mean standardised error computed from the ABC method plotted against the mean standardised error from Petchey et al. (2008). The dashed line is the 1:1 relationship for reference.

of TSS were associated with high error, whereas the opposite was true for other food webs, such as Broadstone Stream (SI Fig. S37(b, p)). The other food webs showed more complex relationships.

353 4 Discussion

The ABC parameterisation method employed here improves on the basic parameterisation methods applied in Petchey et al. (2008) (PBRW). The ABC method provides uncertainty in parameter estimates, and thereby a range of predicted food webs (Fig. 5(c-f)). It also allowed us to estimate parameters that were fixed by the PBRW study, and thereby also predicts connectance (Fig. 5(b)). Including uncertainty and predicting connectance are significant advances in ADBM. They allow predictions in changes of food web structure caused by environmental changes that include uncertainty in the predicted food web structure and including uncertainty in such predictions is critical (Petchey et al. 2015; Cressie et al. 2009; Lindegren et al. 2010). A future development will be to partition the contribution of different sources of uncertainty such as incomplete sampling and model deficiencies to make improvements in the model with the aim of reducing uncertainty.

Future research should investigate the functional and dynamical significance of the uncertainty in the predicted food web structure. Below we discuss some of the results of our study, and expand on these opportunities and priorities for future research.

In all cases, the predicted connectance was greater than the observed connectance (Fig. 5(b)). Why did this occur? Firstly, it is important to recognise that the ADBM (when using only body size as a trait) can only predict diets that are contiguous with respect to the size of prey. I.e. it cannot predict that a predator will consume prey of size 1 and 3, and not prey of size 2. Such patterns can however be predicted if a trait other than size and which is not perfectly correlated with size, influences foraging parameters (Petchey et al. 2008). Secondly, it is important to note that the observed diets were not contiguous when prey are ordered by their size. The estimation process will result in a greater number of predicted links than observed given these features, and the model attempts to maximise the coincidence of predicted and observed link presence and absence (i.e. the true skill statistic).

These findings raise the question as to whether the model or the observed data is incorrect. We expect that the observed data does not contain some links that would occur in reality. This can be possible due to low sampling effort causing some links that do occur to be not observed. In this case, the model may correctly predict a link that was not yet observed as the data was incorrect. More intensive and more complete sampling of links in food webs has been recognised as important, due to the potential that a low sampling effort will influence the perceived food web structure (Martinez et al. 1999).

We expect there are cases where the model incorrectly predicts a feeding link despite no possibility that such a link would occur in reality. This may be the case when a trait other than, or in addition to, prey size is influential. For example, a particular prey species may have a defensive trait that means it takes longer to consume it than an undefended prey of the same size. Incorporating traits other than body size in the ADBM would allow for discontiguous diets along the size axis. Furthermore, the ADBM's current form is a biology-only model; it does not include an observation process, although this could be included. The model would then be able to predict the absence of a link due to incomplete observations.

It would be interesting to take a very well sampled food web, and test if the ABC parameter estimation applied to a subset of the observed links in a simulated poorly sampled food web predicts the connectance of a well sampled food web. Such an outcome would indicate the potential to compensate for under-sampling with an appropriate food web model and estimation procedure.

The ABC parameterisation resulted in a lower prediction accuracy of structural features of the food webs (Fig. 6 (b)) due to the overestimation of connectance. This was confirmed by principal component analysis of variation in the food web structural properties which revealed a first PC axis representing on average 62% of the overall variance, and this first axis was highly correlated with connectance, with an average Spearman correlation of 0.87.

Our parameterisation approach was to maximise the true skill statistic (the coincidence of predicted and observed link presences, and the coincidence of predicted and observed link absences). The TSS assigns equal importance to the collection of presence and absence of observed links with the weight of an observed single presence or absence link being dependent on the connectance of the food web. If the connectance is less than 0.5, the TSS assigns more weight to a presence of link than to an absence of a link and vice versa.

Because the connectance of the observed food webs is less than 0.5 (Table 1), the TSS assigned more weight to a single presence of link than to a single absence of link. This result is as expected, as the chance that a recorded link is a correct is likely to be greater than the chance that a recorded absence is correct. This is because the observation of a single feeding interaction is sufficient to record the presence of a link. However, this is not true for the absence of links: one observation of a predator not consuming a prey does not mean that it never do so. Nevertheless, if we observe no interaction between two species during the sampling period, we conclude that there is an absence of link.

To improve our estimation procedure we could quantify the uncertainty in the recorded absence of links and include this uncertainty in the parameterisation method. Weight/importance could be assigned to true positives, true negatives, false positives and false negatives calculated from empirical studies which may be specific to that food web. Alternatively, an observation process could be added to the model, such that the biological part of the model can predict that a feeding link is possible, but then the observation process in the model leads to that link not being predicted.

In the PBRW study, the parameter b played a major role in maintaining the maximum predictive power of the ADBM. Indeed, they found that estimating b only, and not estimating either a_i or a_j slightly decreased model performance, and that estimating only b and a_j did not decrease model performance relative to when all three parameters were estimated.

We found that the posterior distribution of the parameter b was the most constrained of all the parameters (Fig. 4). Parameter b defines the range of prey body size which has a finite handling time, and the prey size with the highest energetic profitability. As the parameter b relates to the prey-predator body size ratio, the constrained posterior of b (Fig. 4(d)) indicates the importance of the ratio of body size of prey and predator in determining the food web structure with the ADBM.

The marginal posterior of parameter a was right-skewed (Fig. 4(c)). This may be because the ABC parameterisation overestimates the connectance, which means that lower values of a are preferred over higher values of a (a lower value of a leads to a lower space clearance/attack rate, and a lower space clearance rate results in a higher connectance).

Information about who eats who can be collected from multiple sources, such as gut contents of organisms, stable isotope composition of tissues, and experimentation (Peralta-Maraver, López-Rodríguez, and de Figueroa 2017; Layman et al. 2007; Warren 1989). Moreover, experimentation provides independent estimates of allometric foraging parameters, such as b, a_i , and a_j (Rall et al. 2012). Diverse data could be used to parameterise the ADBM's predictions to test how uncertainty in the different datasets influences the ADBM's predictions using ABC. Appropriate summary statistics in the ABC method could be used to address such challenges. We could use, as an example, the approximate trophic position inferred from stable isotope ratio data from an individual tissue and gut content data of a predator simultaneously to parameterise the ADBM. The trophic position and the gut content information would be the summary statistics in this example. A further question that could be addressed in future studies is how the quantity of data affects the ADBM's predictions. The outcome of such a study could help food web researchers decide on how much data from a specific source is needed to predict the food web structure, and help further optimise the deployment of limited sampling resources.

When only partial food web data is available (Patonai and Jordán 2017), the summary statistics in ABC can be used to infer these food web structures from the ADBM. It would be possible to use

gut content data of only some of the species in a food web to parameterise the ADBM and predict the
food web structure. Summary statistics opens up a broad spectrum of possibilities in parameterising
food web models. There are multiple empirical and theoretical studies on a range of different food
web properties of food webs across different ecosystems (Williams and Martinez 2000; Goldwasser
and Roughgarden 1993; Martinez 1991). These can conceivably be used in parameterising food web
models using ABC to constrain the model predictions.

456 5 Acknowledgements

This work was supported by the University Research Priority Program Global Change and Biodiversity (Grant number: U-704-04-11) of the University of Zurich. We thank the Petchey group members for their valuable suggestions in the manuscript. We thank Debra Zuppinger-Dingley for proofreading the manuscript.

461 6 Author contributions

Anubhav Gupta: Conceptualization (equal), formal analysis (lead), methodology (lead), software (lead), writing – original draft preparation (lead), writing – review and editing (equal). Owen L. Petchey: Conceptualization (equal), funding acquisition (lead), methodology (supporting), resources (lead), writing – original draft preparation (supporting), writing – review and editing (equal).

466 References

- Allesina, Stefano, David Alonso, and Mercedes Pascual. 2008. "A General Model for Food Web Structure." *Science* 320 (5876). American Association for the Advancement of Science: 658–61. https://doi.org/10.1126/science.1156269.
- Baker, Ronald, Amanda Buckland, and Marcus Sheaves. 2014. "Fish Gut Content Analysis:
 Robust Measures of Diet Composition." Fish and Fisheries 15 (1): 170–77. https://doi.org/10.
 1111/faf.12026.

- Beckerman, A. P., O. L. Petchey, and P. H. Warren. 2006. "Foraging Biology Predicts Food
- Web Complexity." Proceedings of the National Academy of Sciences of the United States of America
- 475 103: 13745-9.
- Bergamino, Leandro, Diego Lercari, and Omar Defeo. 2011. "Food Web Structure of Sandy
- Beaches: Temporal and Spatial Variation Using Stable Isotope Analysis." Estuarine, Coastal and
- 478 Shelf Science 91 (4): 536-43. https://doi.org/10.1016/j.ecss.2010.12.007.
- Brose, Ulrich, Lara Cushing, Eric L. Berlow, Tomas Jonsson, Carolin Banasek-Richter, Louis-
- 480 Felix Bersier, Julia L. Blanchard, et al. 2005. "Body Sizes of Consumers and Their Resources."
- 481 Ecology 86 (9): 2545-5. https://doi.org/10.1890/05-0379.
- 482 Carpenter, Stephen R. 2016. "Ecological Futures: Building an Ecology
- 483 of the Long Now." *Ecology*, October, 2069–83. https://doi.org/10.1890/0012-
- $9658 (2002) 083 [2069 : EFBAEO] \\ 2.0.CO; \\ 2@10.1002/(ISSN) \\ 1939-9170. \\ MacArthur Award. \\$
- Cattin, Marie-France, Louis-Félix Bersier, Carolin Banašek-Richter, Richard Baltensperger,
- and Jean-Pierre Gabriel. 2004. "Phylogenetic Constraints and Adaptation Explain Food-Web
- 487 Structure." Nature 427 (6977, 6977). Nature Publishing Group: 835-39. https://doi.org/10.
- 488 1038/nature02327.
- Cohen, Joel E. 1989. "Just Proportions in Food Webs." Nature 341 (6238, 6238). Nature
- 490 Publishing Group: 104-5. https://doi.org/10.1038/341104b0.
- Cohen, Joel E., C. M. Newman, and John Hyslop Steele. 1985. "A Stochastic Theory of
- 492 Community Food Webs I. Models and Aggregated Data." Proceedings of the Royal Society of London.
- 493 Series B. Biological Sciences 224 (1237). Royal Society: 421-48. https://doi.org/10.1098/rspb.
- 494 1985.0042.
- Crawford, Kerry, Robbie A. Mcdonald, and Stuart Bearhop. 2008. "Applications of Stable
- 496 Isotope Techniques to the Ecology of Mammals." Mammal Review 38 (1): 87-107. https://doi.
- org/10.1111/j.1365-2907.2008.00120.x.
- 498 Cressie, Noel, Catherine A. Calder, James S. Clark, Jay M. Ver Hoef, and Christopher K.
- 499 Wikle. 2009. "Accounting for Uncertainty in Ecological Analysis: The Strengths and Limitations of

- Hierarchical Statistical Modeling." Ecological Applications 19 (3): 553-70. https://doi.org/10.
 1890/07-0744.1.
- Dawah, Hassan Ali, Bradford A. Hawkins, and Michael F. Claridge. 1995. "Structure of the Parasitoid Communities of Grass-Feeding Chalcid Wasps." *The Journal of Animal Ecology* 64 (6): 708. https://doi.org/10.2307/5850.
- Dunne, Jennifer A., Richard J. Williams, and Neo D. Martinez. 2002. "Network Structure and Biodiversity Loss in Food Webs: Robustness Increases with Connectance." *Ecology Letters* 5 (4): 558–67.
- Emmerson, Mark C., and Dave Raffaelli. 2004. "Predator-Prey Body Size, Interaction
 Strength and the Stability of a Real Food Web." Journal of Animal Ecology 73 (3): 399–409.
 https://doi.org/10.1111/j.0021-8790.2004.00818.x.
- Goldwasser, Lloyd, and Jonathan Roughgarden. 1993. "Construction and Analysis of a Large Caribbean Food Web: Ecological Archives E074-001." Ecology 74 (4): 1216–33. https://doi.org/10.2307/1940492.
- Gravel, Dominique, Timothée Poisot, Camille Albouy, Laure Velez, and David Mouillot. 2013. "Inferring Food Web Structure from Predator-Prey Body Size Relationships." Edited by Robert Freckleton. *Methods in Ecology and Evolution* 4 (11): 1083–90. https://doi.org/10.1111/ 2041-210X.12103.
- Harper-Smith, Sarah, Eric L. Berlow, Roland A. Knapp, Richard J. Williams, and Neo D. Martinez. 2005. "COMMUNICATING ECOLOGY THROUGH FOOD WEBS: VISUALIZING AND QUANTIFYING THE EFFECTS OF STOCKING ALPINE LAKES WITH TROUT." In Dynamic Food Webs, 407–23. Elsevier. https://doi.org/10.1016/B978-012088458-2/50038-2.
- Hattab, Tarek, Fabien Leprieur, Frida Ben Rais Lasram, Dominique Gravel, François Le Loc'h, and Camille Albouy. 2016. "Forecasting Fine-Scale Changes in the Food-Web Structure of Coastal Marine Communities Under Climate Change." *Ecography* 39 (12): 1227–37. https://doi.org/10.1111/ecog.01937.

- Hobson, Keith A., John F. Piatt, and Jay Pitocchelli. 1994. "Using Stable Isotopes to Determine Seabird Trophic Relationships." *Journal of Animal Ecology* 63 (4). [Wiley, British Ecological Society]: 786–98. https://doi.org/10.2307/5256.
- Hudson, Lawrence N., Rob Emerson, Gareth B. Jenkins, Katrin Layer, Mark E. Ledger, Doris E. Pichler, Murray S. A. Thompson, Eoin J. O'Gorman, Guy Woodward, and Daniel C. Reuman. 2013. "Cheddar: Analysis and Visualisation of Ecological Communities in R." *Methods in Ecology* and Evolution 4 (1): 99–104. https://doi.org/10.1111/2041-210X.12005.
- Ibanez, Sébastien. 2012. "Optimizing Size Thresholds in a Plant-Pollinator Interaction
 Web: Towards a Mechanistic Understanding of Ecological Networks." *Oecologia* 170 (1): 233–42.
 https://doi.org/10.1007/s00442-012-2290-3.
- Jabot, Franck, and Jérôme Chave. 2009. "Inferring the Parameters of the Neutral Theory of Biodiversity Using Phylogenetic Information and Implications for Tropical Forests." *Ecology Letters* 12 (3): 239–48. https://doi.org/10.1111/j.1461-0248.2008.01280.x.
- Jonsson, Tomas. 1998. "Food Webs and the Distribution of Body Sizes." PhD.
- Jonsson, Tomas, Joel E. Cohen, and Stephen R. Carpenter. 2005. "Food Webs, Body Size, and Species Abundance in Ecological Community Description." In *Advances in Ecological Research*, 36:1–84. Elsevier. https://doi.org/10.1016/S0065-2504(05)36001-6.
- Jordán, Ferenc, and Györgyi Osváth. 2009. "The Sensitivity of Food Web Topology to
 Temporal Data Aggregation." *Ecological Modelling* 220 (22): 3141-6. https://doi.org/10.1016/
 j.ecolmodel.2009.05.002.
- Knight, Tiffany M., Michael W. McCoy, Jonathan M. Chase, Krista A. McCoy, and Robert D. Holt. 2005. "Trophic Cascades Across Ecosystems." *Nature* 437 (7060): 880–83. https://doi.org/10.1038/nature03962.
- Krause, Ann E., Kenneth A. Frank, Doran M. Mason, Robert E. Ulanowicz, and William W. Taylor. 2003. "Compartments Revealed in Food-Web Structure." *Nature* 426 (6964): 282–85. https://doi.org/10.1038/nature02115.

- Lafferty, K. D., A. P. Dobson, and A. M. Kuris. 2006. "Parasites Dominate Food Web Links."
- Proceedings of the National Academy of Sciences 103 (30): 11211-6. https://doi.org/10.1073/
- pnas.0604755103.
- Layman, Craig A., D. Albrey Arrington, Carmen G. Montaña, and David M. Post. 2007. "Can
- 556 Stable Isotope Ratios Provide for Community-Wide Measures of Trophic Structure?" Ecology 88
- 557 (1): 42-48. https://doi.org/10.1890/0012-9658(2007)88[42:CSIRPF]2.0.CO;2.
- Lindegren, Martin, Christian Möllmann, Anders Nielsen, Keith Brander, Brian R. MacKenzie,
- 559 and Nils Chr. Stenseth. 2010. "Ecological Forecasting Under Climate Change: The Case of
- Baltic Cod." Proceedings of the Royal Society B: Biological Sciences 277 (1691): 2121-30. https:
- 561 //doi.org/10.1098/rspb.2010.0353.
- Lurgi, Miguel, Bernat C. López, and José M. Montoya. 2012. "Climate Change Impacts on Body
- 563 Size and Food Web Structure on Mountain Ecosystems." Philosophical Transactions of the Royal
- 564 Society B: Biological Sciences 367 (1605): 3050-7. https://doi.org/10.1098/rstb.2012.0239.
- MacArthur, Robert H., and Eric R. Pianka. 1966. "On Optimal Use of a Patchy Environment."
- 566 The American Naturalist 100 (916): 603-9.
- Martinez, Neo D. 1991. "Artifacts or Attributes? Effects of Resolution on the Little Rock
- 568 Lake Food Web." Ecological Monographs 61 (4): 367-92. https://doi.org/10.2307/2937047.
- Martinez, Neo D., Bradford A. Hawkins, Hassan Ali Dawah, and Brian P. Feifarek. 1999.
- *Effects of Sampling Effort on Characterization of Food-Web Structure." Ecology 80 (3): 1044-55.
- 571 https://doi.org/10.1890/0012-9658(1999)080[1044:EOSEOC]2.0.CO:2.
- May, Robert M. 1972. "Will a Large Complex System Be Stable?" Nature 238 (5364): 413.
- 573 https://doi.org/10.1038/238413a0.
- Memmott, J., N.D. Martinez, and J.E. Cohen. 2000. "Predators, Parasitoids and Pathogens:
- 575 Species Richness, Trophic Generality and Body Sizes in a Natural Food Web." Journal of Animal
- 576 Ecology 69 (1): 1-15. https://doi.org/10.1046/j.1365-2656.2000.00367.x.
- Morris, Rebecca J., Frazer H. Sinclair, and Chris J. Burwell. 2015. "Food Web Structure
- 578 Changes with Elevation but Not Rainforest Stratum." Ecography 38 (8): 792-802. https://doi.
- org/10.1111/ecog.01078.

- O'Connor, Mary I., Michael F. Piehler, Dina M. Leech, Andrea Anton, and John F. Bruno. 2009.

 "Warming and Resource Availability Shift Food Web Structure and Metabolism." Edited by Michel

 Loreau. *PLoS Biology* 7 (8): e1000178. https://doi.org/10.1371/journal.pbio.1000178.
- O'Gorman, Eoin J., Owen L. Petchey, Katy J. Faulkner, Bruno Gallo, Timothy A. C. Gordon, Joana Neto-Cerejeira, Jón S. Ólafsson, Doris E. Pichler, Murray S. A. Thompson, and Guy Woodward. 2019. "A Simple Model Predicts How Warming Simplifies Wild Food Webs." *Nature Climate Change* 9 (8): 611–16. https://doi.org/10.1038/s41558-019-0513-x.
- Opitz, Silvia. 1996. "Quantitative Models of Trophic Interactions in Caribbean Coral Reefs."

 Iclarm.
- Patonai, Katalin, and Ferenc Jordán. 2017. "Aggregation of Incomplete Food Web Data 590 May Help to Suggest Sampling Strategies." *Ecological Modelling* 352 (May): 77–89. https: 591 //doi.org/10.1016/j.ecolmodel.2017.02.024.
- Peralta-Maraver, I., M. J. López-Rodríguez, and J. M. Tierno de Figueroa. 2017. "Structure, Dynamics and Stability of a Mediterranean River Food Web." *Marine and Freshwater Research* 68 (3). CSIRO PUBLISHING: 484–95. https://doi.org/10.1071/MF15154.
- Petchey, Owen L, Andrew P Beckerman, Jens O Riede, and Philip H Warren. 2008. "Size, Foraging, and Food Web Structure." *Proceedings of the National Academy of Sciences* 105 (11). National Acad Sciences: 4191–6.
- Petchey, Owen L., P. Timon McPhearson, Timothy M. Casey, and Peter J. Morin. 1999. "Environmental Warming Alters Food-Web Structure and Ecosystem Function." *Nature* 402 (6757): 600 69–72. https://doi.org/10.1038/47023.
- Petchey, Owen L., Mikael Pontarp, Thomas M. Massie, Sonia Kéfi, Arpat Ozgul, Maja Weilenmann, Gian Marco Palamara, et al. 2015. "The Ecological Forecast Horizon, and Examples of Its Uses and Determinants." *Ecology Letters* 18 (7): 597–611. https://doi.org/10.1111/ele.
- Poisot, Timothée, and Dominique Gravel. 2014. "When Is an Ecological Network Complex?

 Connectance Drives Degree Distribution and Emerging Network Properties." *PeerJ* 2 (February).

 PeerJ Inc.: e251. https://doi.org/10.7717/peerj.251.

- Poisot, Timothée, and Daniel B. Stouffer. 2016. "How Ecological Networks Evolve." Preprint.
- $_{\rm 609}$ Ecology. https://doi.org/10.1101/071993.
- Polis, Gary A. 1991. "Complex Trophic Interactions in Deserts: An Empirical Critique of Food-Web Theory." *The American Naturalist* 138 (1): 123–55.
- Rall, B. C., U. Brose, M. Hartvig, G. Kalinkat, F. Schwarzmuller, O. Vucic-Pestic, and O. L.
- Petchey. 2012. "Universal Temperature and Body-Mass Scaling of Feeding Rates." Philosophical
- 614 Transactions of the Royal Society B: Biological Sciences 367 (1605): 2923-34. https://doi.org/
- 615 10.1098/rstb.2012.0242.
- Shriner, Daniel, Yi Liu, David C. Nickle, and James I. Mullins. 2006. "Evolution of Intrahost
- 617 Hiv 1 Genetic Diversity During Chronic Infection." Evolution 60 (6): 1165-76. https://doi.org/
- 618 10.1111/j.0014-3820.2006.tb01195.x.
- Tamaddoni-Nezhad, Alireza, Ghazal Afroozi Milani, Alan Raybould, Stephen Muggleton,
- 620 and David A. Bohan. 2013. "Construction and Validation of Food Webs Using Logic-Based
- 621 Machine Learning and Text Mining." In Advances in Ecological Research, 49:225–89. Elsevier.
- 622 https://doi.org/10.1016/B978-0-12-420002-9.00004-4.
- Toni, Tina, David Welch, Natalja Strelkowa, Andreas Ipsen, and Michael P.H. Stumpf. 2009.
- 624 "Approximate Bayesian Computation Scheme for Parameter Inference and Model Selection in
- 625 Dynamical Systems." Journal of the Royal Society Interface 6 (31): 187-202. https://doi.org/10.
- 626 1098/rsif.2008.0172.
- Tylianakis, Jason M., and Amrei Binzer. 2014. "Effects of Global Environmental Changes
- on ParasitoidHost Food Webs and Biological Control." Biological Control 75 (August): 77–86.
- 629 https://doi.org/10.1016/j.biocontrol.2013.10.003.
- Warren, Philip H. 1989. "Spatial and Temporal Variation in the Structure of a Freshwater
- 631 Food Web." Oikos 55 (3): 299. https://doi.org/10.2307/3565588.
- Williams, Richard J., and Neo D. Martinez. 2000. "Simple Rules Yield Complex Food Webs."
- 633 Nature 404 (6774, 6774). Nature Publishing Group: 180-83. https://doi.org/10.1038/35004572.
- Woodward, Guy, Julia Blanchard, Rasmus B. Lauridsen, Francois K. Edwards, J. Iwan Jones,
- David Figueroa, Philip H. Warren, and Owen L. Petchey. 2010. "Chapter 6 Individual-Based Food

- Webs: Species Identity, Body Size and Sampling Effects." In Advances in Ecological Research, edited
 by Guy Woodward, 43:211–66. Integrative Ecology: From Molecules to Ecosystems. Academic
- $_{\rm 638}$ Press. https://doi.org/10.1016/B978-0-12-385005-8.00006-X.
- Woodward, Guy, and Alan G. Hildrew. 2001. "Invasion of a Stream Food Web by a New Top
- 640 Predator." Journal of Animal Ecology 70 (2): 273-88. https://doi.org/10.1111/j.1365-2656.
- 641 2001.00497.x.
- Woodward, Guy, Dougie C. Speirs, and Alan G. Hildrew. 2005. "Quantification and Resolution
- of a Complex, Size-Structured Food Web." In Advances in Ecological Research, 36:85–135. Elsevier.
- 644 https://doi.org/10.1016/S0065-2504(05)36002-8.
- Yodzis, Peter. 1998. "Local Trophodynamics and the Interaction of Marine Mammals
- and Fisheries in the Benguela Ecosystem." Journal of Animal Ecology 67 (4): 635–58. https:
- 647 //doi.org/10.1046/j.1365-2656.1998.00224.x.