

Saint Matthew Island Blue King Crab Stock Assessment 2016

D’Arcy Webber

Jie Zheng

James Ianelli

September 2016

Executive Summary

1. **Stock:** Blue king crab, *Paralithodes platypus*, Saint Matthew Island (SMBKC), Alaska.
2. **Catches:** Peak historical harvest was 4288 tonnes (9.454 million pounds) in 1983/84¹. The fishery was closed for 10 years after the stock was declared overfished in 1999. Fishing resumed in 2009/10 with a fishery-reported retained catch of 209 tonnes (0.461 million pounds), less than half the 529.3 tonne (1.167 million pound) TAC. Following three more years of modest harvests supported by a fishery catch per unit effort (CPUE) of around 10 crab per pot lift, the fishery was again closed in 2013/14 due to declining trawl-survey estimates of abundance and concerns about the health of the stock. The directed fishery resumed again in 2014/15 with a TAC of 300 tonnes (0.655 million pounds), but the fishery performance was relatively poor with a retained catch of 140 tonnes (0.309 million pounds).
3. **Stock biomass:** Following a period of low numbers after the stock was declared overfished in 1999, trawl-survey indices of SMBKC stock abundance and biomass generally increased in subsequent years, with survey estimated mature male biomass reaching 9516 tonnes (20.98 million pounds; CV = 0.55) in 2011, the second highest in the 39-year time series used in this assessment. Survey mature male biomass then declined to 5652 tonnes (12.46 million pounds; CV = 0.33) in 2012 and to 2202 tonnes (4.459 million pounds; CV = 0.22) in 2013 before going back up to 5472 tonnes (12.06 million pounds; CV = 0.44) in 2014 and 5134 tonnes (11.32 million pounds; CV = 0.76) in 2015.
4. **Recruitment:** Because little information about the abundance of small crab is available for this stock, recruitment has been assessed in terms of the number of male crab within the 90-104 mm carapace length (CL) size class in each year. The 2013 trawl-survey area-swept estimate of 0.335 million male SMBKC in this size class marked a three-year decline and was the lowest since 2005. That decline did not continue as the 2014 survey estimate is 0.723 million. The survey recruitment is 0.992 million in 2015, but the majority of this survey estimate is from one tow with a great deal of uncertainty.
5. **Management performance:** In recent assessments, estimated total male catch has been determined as the sum of fishery-reported retained catch, estimated male discard mortality in the directed fishery, and estimated male bycatch mortality in the groundfish fisheries, as these have been the only sources of non-negligible fishing mortality to consider. The stock was above the minimum stock-size threshold (MSST) in 2014/15 and is hence not overfished. Overfishing did not occur in 2014/15 (Table 1).
6. **Basis for the OFL:** Estimated mature-male biomass (MMB) on 15 February is used as the measure of biomass for this Tier 4 stock, with males measuring 105 mm CL or more considered mature. The B_{MSY} proxy is obtained by averaging estimated MMB over a specific reference time period, and current CPT/SSC guidance recommends using the full assessment time frame as the default reference period (Table 2).

¹1983/84 refers to a fishing year that extends from 1 July 1983 to 30 June 1984.

Table 1: Status and catch specifications (1000 tonnes) (scenario 1).

| Year | MSST | Biomass ($MMB_{\text{mat}}_{\text{ing}}$) | TAC | Retained catch | Total male catch | OFL | ABC |
|---------|------|--|------|-------------------|---------------------|------|------|
| 2011/12 | 1.50 | 5.03 | 1.15 | 0.85 | 0.95 | 1.70 | 1.54 |
| 2012/13 | 1.80 | 2.85 | 0.74 | 0.73 | 0.82 | 1.02 | 0.92 |
| 2013/14 | 1.50 | 3.01 | 0.00 | 0.00 | 0.00 | 0.56 | 0.45 |
| 2014/15 | 1.86 | 2.48 | 0.30 | 0.14 | 0.15 | 0.43 | 0.34 |
| 2015/16 | | 2.45 | | | | 0.28 | 0.22 |
| 2016/17 | | | | | | | |

Table 2: Basis for the OFL (1000 tonnes) (scenario 1).

| Year | Tier | B_{MSY} | Biomass ($MMB_{\text{mat}}_{\text{ing}}$) | B/B_{MSY} | F_{OFL} | γ | Basis for B_{MSY} | Natural mortality |
|---------|------|-----------|--|-------------|-----------|----------|---------------------|----------------------|
| 2011/12 | 4a | 3.11 | 7.17 | 2.31 | 0.18 | 1 | 1989-2010 | 0.18 |
| 2012/13 | 4a | 3.56 | 5.63 | 1.56 | 0.18 | 1 | 1978-2012 | 0.18 |
| 2013/14 | 4b | 3.06 | 3.01 | 0.98 | 0.18 | 1 | 1978-2013 | 0.18 |
| 2014/15 | 4b | 3.28 | 2.71 | 0.82 | 0.14 | 1 | 1978-2014 | 0.18 |
| 2015/16 | 4b | 3.71 | 2.45 | 0.66 | 0.11 | 1 | 1978-2015 | 0.18 |
| 2016/17 | 4b | | | | | 1 | 1978-2016 | 0.18 |

A. Summary of Major Changes

Changes in Management of the Fishery

There are no new changes in management of the fishery.

Changes to the Input Data

All of the time series used in this assessment have been updated to include the most recent fishery and survey results. This assessment makes use of an updated full trawl-survey time series supplied by R. Foy in August 2015, updated groundfish bycatch estimates based on 1999-2014 NMFS AKRO data also supplied by R. Foy, and the ADF&G pot survey data in 2016.

Changes in Assessment Methodology

This assessment is done using Gmacs. The model is based upon the 3-stage length-based assessment model first presented in May 2011 by Bill Gaeuman and accepted by the CPT in May 2012. There are several differences between the Gmacs assessment and the previous model. One of the major differences being that natural and fishing mortality are continuous within 5 discrete seasons. Season length in Gmacs is controlled by changing the proportion of natural mortality that is applied during each season.

Changes in Assessment Results

Changes in assessment results depend on model scenario. The Gmacs match model scenario attempts to match the 2015 assessment by specifying the same (or similar) dynamics and the same (fixed) parameter values. However, a different Gmacs scenario (Gmacs sele) provides a much better match to the 2015 model assessment.

B. Responses to SSC and CPT Comments

CPT and SSC Comments on Assessments in General

Comment: *Regarding general code development, the CPT had the following requests:*

1. 1-year projection for calculating Tier 3 or 4 OFLs
2. specify catchability as a fixed or estimated parameter or use the analytic calculation for the MLE
3. specify priors (e.g., gamma) using mean and variance/standard deviation for all parameters to ease specifying priors
4. include an option to calculate dynamic B_{MSY}
5. add the ability to “jitter” initial parameter values
6. add the ability to conduct retrospective analyses
7. add ability to estimate bycatch fishing mortality rates when observer data are missing but effort data is available
8. allow different phases for “rec_ini”, “rec_dev” estimation

Response:

1. Done
2. Done
3. XXX
4. XXX
5. XXX
6. XXX
7. XXX
8. Done

Comment: *Andre Punt pointed out the need to use a fixed-iteration Newton’s method to calculate OFL, not bisection, to keep the calculation differentiable so that OFL can be reported as an sdreport variable.*

Response: This has been done and the F_{OFL} and OFL have both been reported as an sdreport variables in this document.

CPT and SSC Comments Specific to the SMBKC Stock Assessment

Comment: *the CPT requests that some evaluation should also be included in the September report to the CPT which compares against the previous assessment model corrected for the error.*

Response: The error in the 2015 was fixed and this model was run again. Comparisons between the Gmacs models and the 2015 model are presented throughout this document.

Comment: *The SSC and CPT requested the following models for review at the spring 2016 meeting:*

1. Base: try to match 2015 model but prevent dome shaped selectivity
2. Base + add CV for both surveys
3. Above + Francis re-weighting
4. Above + remove M spike

Response: Models 1, 3, and 4 are all included and evaluated in this document as the **Gmacs base**, **Gmacs Francis**, and **Gmacs M** scenarios. Model 2 was not included in this document for two reasons. Firstly, if doing Francis iterative re-weighting then additional CV should not be added as well. Secondly, the SSC recommended against the model runs with additional CV (see the comment from the SSC below).

Comment: *The SSC is not convinced that the model runs with extra CV are very informative. The inclusion of extra CV seems to be rather arbitrary based on the numbers of points that fall within confidence intervals estimated from trawl surveys. The SSC recommends coming up with some alternative way to consider extra variability, which could be informed by simulation testing.*

Response: All model runs that estimate additional CV were dropped from this document. Instead we provide three model runs that use the Francis iterative re-weighting method to weight the length-frequency data relative to the abundance indices. These runs are the **Gmacs Francis**, **Gmacs M**, and **Gmacs force** scenarios.

Comment: *The descriptions of seasons in the model is confusing and currently reads as if M differs among seasons. More justification is needed on how seasons are defined and how they were selected, as well as clarification on M during these seasons.*

Response: This description has been updated and justification provided in Appendix A.

Comment: *During the presentation to the SSC, uncertainty was expressed about the origins of the growth transition matrix, but page 7 of the report indicates that the matrix was derived by Otto and Cummiskey (1990). As this matrix is critical to the model, the origin and integrity of the growth transition matrix should be carefully explained in the assessment for fall 2016. In some other models, the transition matrix can be estimated. If there are doubts about the veracity of the transition matrix, perhaps this can be explored in the modeling framework.*

Response: The report is correct, the growth matrix was derived by Otto and Cummiskey (1990) and used in this assessment.

Comment: *The selectivities were constrained so that they do not exceed 1.0, but the tables of log-transformed parameter estimates do not indicate that this upper bound was approached. This should be clarified.*

Response:

Comment: *It would be helpful to include a table of NMFS trawl survey CPUE by crab stage, just as was provided for the ADF&G pot survey (Table 1).*

Response: This table has been added.

Comment: *Page 10 refers to a table of observed and estimated sample size, but no such table was provided.*

Response: This table has been added.

Comment: *As with the 2015 model, GMACS consistently overestimates trawl survey estimates of male biomass in the last decade, whereas GMACS tends to underestimate the last couple of pot survey estimates (Figure 9, 12). This is also reflected in patterns in residuals, and the proportions of stage-3 crab tend to be overestimated in recent years (Figure 14). These patterns should be discussed in the assessment.*

Response:

Comment: *The report contains very little description and interpretation of results. Moreover, not all figures are cited in the document. The document should highlight the major features of the results and offer some explanation, as well.*

Response:

Comment: *A brief explanation was provided about the future outlook (page 12) that indicated a declining stock. However, stock trends shown in Figure 24 generally suggest population growth since 1993. Closer examination of Tables 9-11 suggest that trends depend somewhat on model run and life stage. Statements about future outlook should be qualified and refer to figures and tables and explain any differences in outcomes.*

Response:

Comment: *The SSC discussed the possibility that these patterns could be indicative of spatial patterns in stock distribution. The trawl survey covers a much larger geographic distribution than the pot survey (Figure 4). Crab distribution may vary with sex (females tend to be found close to shore) and life stage. Thus, the trawl and pot surveys may sample the crab stock differentially. Moreover, the geographic distributions of these stages may vary with stock density and temperature. It could be informative to conduct some spatial analyses, which could include: (1) estimation of survey catchability as a function of temperature, (2) a stock assessment model run that includes pot surveys and only those trawl stations that fall within the pot survey distribution*

as a comparison the runs that include the full trawl survey data, and (3) analysis of the spatial distribution of surveyed crabs by stage at high and low biomass and during warm and cold years.

Response:

C. Introduction

Scientific Name

The blue king crab is a lithodid crab, *Paralithodes platypus* (Brant 1850).

Distribution

Blue king crab are sporadically distributed throughout the North Pacific Ocean from Hokkaido, Japan, to southeastern Alaska (Figure 1). In the eastern Bering Sea small populations are distributed around St. Matthew Island, the Pribilof Islands, St. Lawrence Island, and Nunivak Island. Isolated populations also exist in some other cold water areas of the Gulf of Alaska (NPFMC 1998). The St. Matthew Island Section for blue king crab is within Area Q2 (Figure 2), which is the Northern District of the Bering Sea king crab registration area and includes the waters north of Cape Newenham (58°39' N. lat.) and south of Cape Romanzof (61°49' N. lat.).

Stock Structure

The Alaska Department of Fish and Game (ADF&G) Gene Conservation Laboratory division has detected regional population differences between blue king crab collected from St. Matthew Island and the Pribilof Islands². NMFS tag-return data from studies on blue king crab in the Pribilof Islands and St. Matthew Island support the idea that legal-sized males do not migrate between the two areas (Otto and Cummiskey 1990). St. Matthew Island blue king crab tend to be smaller than their Pribilof conspecifics, and the two stocks are managed separately.

Life History

Like the red king crab, *Paralithodes camtschaticus*, the blue king crab is considered a shallow water species by comparison with other lithodids such as golden king crab, *Lithodes aequispinus*, and the scarlet king crab, *Lithodes couesi* (Donaldson and Byersdorfer 2005). Adult male blue king crab are found at an average depth of 70 m (NPFMC 1998). The reproductive cycle appears to be annual for the first two reproductive cycles and biennial thereafter (cf. Jensen and Armstrong 1989) and mature crab seasonally migrate inshore where they molt and mate. Unlike red king crab, juvenile blue king crab do not form pods, but instead rely on cryptic coloration for protection from predators and require suitable habitat such as cobble and shell hash. Somerton and MacIntosh (1983) estimated SMBKC male size at sexual maturity to be 77 mm carapace length (CL). Paul et al. (1991) found that spermatophores were present in the vas deferens of 50% of the St. Matthew Island blue king crab males examined with sizes of 40-49 mm CL and in 100% of the males at least 100 mm CL. Spermatophore diameter also increased with increasing CL with an asymptote at ~ 100 mm CL. They noted, however, that although spermatophore presence indicates physiological sexual maturity, it may not be an indicator of functional sexual maturity. For purposes of management of the St. Matthew Island blue king crab fishery, the State of Alaska uses 105 mm CL to define the lower size bound of functionally mature males (Pengilly and Schmidt 1995). Otto and Cummiskey (1990) report an average growth increment of 14.1 mm CL for adult SMBKC males.

²NOAA grant Bering Sea Crab Research II, NA16FN2621, 1997.



Figure 1: Distribution of blue king crab (*Paralithodes platypus*) in the Gulf of Alaska, Bering Sea, and Aleutian Islands waters (shown in blue).



Figure 2: King crab Registration Area Q (Bering Sea).

Management History

The SMBKC fishery developed subsequent to baseline ecological studies associated with oil exploration (Otto 1990). Ten U.S. vessels harvested 545 tonnes (1.202 million pounds) in 1977, and harvests peaked in 1983 when 164 vessels landed 4288 tonnes (9.454 million pounds) (Fitch et al. 2012; Table 3).

Table 3: The 1978/79 to 2014/15 directed St. Matthew Island blue king crab pot fishery. The Guideline Harvest Level (GHL) and Total Allowable Catch (TAC) are in millions of pounds. Harvest includes deadloss. Catch per unit effort (CPUE) in this table is simply the harvest number / pot lifts. The average weight is the harvest weight / harvest number in pounds. The average CL is the average of retained crab in mm from dockside sampling of delivered crab. Source: Fitch et al 2012; ADF&G Dutch Harbor staff, pers. comm.

| Year | Dates | GHL/TAC | Harvest | | Pot lifts | CPUE | avg wt | avg CL |
|-------------------|---------------|---------|----------------|-----------|-----------|------|--------|--------|
| | | | Crab | Pounds | | | | |
| 1978/79 | 07/15 - 09/03 | | 436,126 | 1,984,251 | 43,754 | 10 | 4.5 | 132.2 |
| 1979/80 | 07/15 - 08/24 | | 52,966 | 210,819 | 9,877 | 5 | 4.0 | 128.8 |
| 1980/81 | 07/15 - 09/03 | | CONFIDENTIAL | | | | | |
| 1981/82 | 07/15 - 08/21 | | 1,045,619 | 4,627,761 | 58,550 | 18 | 4.4 | NA |
| 1982/83 | 08/01 - 08/16 | | 1,935,886 | 8,844,789 | 165,618 | 12 | 4.6 | 135.1 |
| 1983/84 | 08/20 - 09/06 | 8.0 | 1,931,990 | 9,454,323 | 133,944 | 14 | 4.9 | 137.2 |
| 1984/85 | 09/01 - 09/08 | 2.0-4.0 | 841,017 | 3,764,592 | 73,320 | 11 | 4.5 | 135.5 |
| 1985/86 | 09/01 - 09/06 | 0.9-1.9 | 436,021 | 2,175,087 | 46,988 | 9 | 5.0 | 139.0 |
| 1986/87 | 09/01 - 09/06 | 0.2-0.5 | 219,548 | 1,003,162 | 22,073 | 10 | 4.6 | 134.3 |
| 1987/88 | 09/01 - 09/05 | 0.6-1.3 | 227,447 | 1,039,779 | 28,230 | 8 | 4.6 | 134.1 |
| 1988/89 | 09/01 - 09/05 | 0.7-1.5 | 280,401 | 1,236,462 | 21,678 | 13 | 4.4 | 133.3 |
| 1989/90 | 09/01 - 09/04 | 1.7 | 247,641 | 1,166,258 | 30,803 | 8 | 4.7 | 134.6 |
| 1990/91 | 09/01 - 09/07 | 1.9 | 391,405 | 1,725,349 | 26,264 | 15 | 4.4 | 134.3 |
| 1991/92 | 09/16 - 09/20 | 3.2 | 726,519 | 3,372,066 | 37,104 | 20 | 4.6 | 134.1 |
| 1992/93 | 09/04 - 09/07 | 3.1 | 545,222 | 2,475,916 | 56,630 | 10 | 4.5 | 134.1 |
| 1993/94 | 09/15 - 09/21 | 4.4 | 630,353 | 3,003,089 | 58,647 | 11 | 4.8 | 135.4 |
| 1994/95 | 09/15 - 09/22 | 3.0 | 827,015 | 3,764,262 | 60,860 | 14 | 4.9 | 133.3 |
| 1995/96 | 09/15 - 09/20 | 2.4 | 666,905 | 3,166,093 | 48,560 | 14 | 4.7 | 135.0 |
| 1996/97 | 09/15 - 09/23 | 4.3 | 660,665 | 3,078,959 | 91,085 | 7 | 4.7 | 134.6 |
| 1997/98 | 09/15 - 09/22 | 5.0 | 939,822 | 4,649,660 | 81,117 | 12 | 4.9 | 139.5 |
| 1998/99 | 09/15 - 09/26 | 4.0 | 635,370 | 2,968,573 | 91,826 | 7 | 4.7 | 135.8 |
| 1999/00 - 2008/09 | | | FISHERY CLOSED | | | | | |
| 2009/10 | 10/15 - 02/01 | 1.17 | 103,376 | 460,859 | 10,697 | 10 | 4.5 | 134.9 |
| 2010/11 | 10/15 - 02/01 | 1.60 | 298,669 | 1,263,982 | 29,344 | 10 | 4.2 | 129.3 |
| 2011/12 | 10/15 - 02/01 | 2.54 | 437,862 | 1,881,322 | 48,554 | 9 | 4.3 | 130.0 |
| 2012/13 | 10/15 - 02/01 | 1.63 | 379,386 | 1,616,054 | 37,065 | 10 | 4.3 | 129.8 |
| 2013/14 | | | FISHERY CLOSED | | | | | |
| 2014/15 | 10/15 - 02/05 | 0.66 | 69,109 | 308,582 | 10,133 | 7 | 4.5 | 132.3 |
| 2015/16 | | | | | | | | |

The fishing seasons were generally short, often lasting only a few days. The fishery was declared overfished and closed in 1999 when the stock biomass estimate was below the minimum stock-size threshold (MSST) of 4990 tonnes (11.0 million pounds) as defined by the Fishery Management Plan for the Bering Sea/Aleutian Islands King and Tanner crabs (NPFMC 1999). Zheng and Kruse (2002) hypothesized a high level of SMBKC natural mortality from 1998 to 1999 as an explanation for the low catch per unit effort (CPUE) in the 1998/99 commercial fishery and the low numbers across all male crab size groups caught in the annual NMFS eastern Bering Sea trawl survey from 1999 to 2005 (Table 7). In November 2000, Amendment 15 to the FMP for Bering Sea/Aleutian Islands king and Tanner crabs was approved to implement a rebuilding plan for the SMBKC stock (NPFMC 2000). The rebuilding plan included a regulatory harvest strategy (5 AAC 34.917), area closures, and gear modifications. In addition, commercial crab fisheries near St. Matthew Island were

scheduled in fall and early winter to reduce the potential for bycatch mortality of vulnerable molting and mating crab.

NMFS declared the stock rebuilt on 21 September 2009, and the fishery was reopened after a 10-year closure on 15 October 2009 with a TAC of 529 tonnes (1.167 million pounds), closing again by regulation on 1 February 2010. Seven participating vessels landed a catch of 209 tonnes (460,859 pounds) with a reported effort of 10,697 pot lifts and an estimated CPUE of 9.9 retained individual crab per pot lift. The fishery remained open the next three years with modest harvests and similar CPUE, but large declines in the NMFS trawl-survey estimate of stock abundance raised concerns about the health of the stock, prompting ADF&G to close the fishery again for the 2013/14 season. Due to an abundance above thresholds, the fishery was reopened for the 2014/15 season with a low TAC of 297 tonnes (0.655 million pounds) and in 2015/16 the TAC was further reduced to 186 tonnes (0.411 million pounds).

Though historical observer data are limited due to very limited sampling, bycatch of female and sublegal male crab from the directed blue king crab fishery off St. Matthew Island was relatively high historically, with estimated total bycatch in terms of number of crab captured sometimes more than twice as high as the catch of legal crab (Moore et al. 2000; ADF&G Crab Observer Database). Pot-lift sampling by ADF&G crab observers (Gaeuman 2013; ADF&G Crab Observer Database) indicates similar bycatch rates of discarded male crab since the reopening of the fishery (Table 4), with total male discard mortality in the 2012/13 directed fishery estimated at about 12% (88 tonnes or 0.193 million pounds) of the reported retained catch weight, assuming 20% handling mortality.

Table 4: Observed proportion of crab by size class during the ADF&G crab observer pot-lift sampling. Source: ADF&G Crab Observer Database.

| Year | Total pot lifts | Pot lifts sampled | Number of crab (90 mm+ CL) | Stage 1 | Stage 2 | Stage 3 |
|-------------------|-----------------|-------------------|----------------------------|---------|---------|---------|
| 1990/91 | 26,264 | 10 | 150 | 0.113 | 0.393 | 0.493 |
| 1991/92 | 37,104 | 125 | 3,393 | 0.133 | 0.177 | 0.690 |
| 1992/93 | 56,630 | 71 | 1,606 | 0.191 | 0.268 | 0.542 |
| 1993/94 | 58,647 | 84 | 2,241 | 0.281 | 0.210 | 0.510 |
| 1994/95 | 60,860 | 203 | 4,735 | 0.294 | 0.271 | 0.434 |
| 1995/96 | 48,560 | 47 | 663 | 0.148 | 0.212 | 0.640 |
| 1996/97 | 91,085 | 96 | 489 | 0.160 | 0.223 | 0.618 |
| 1997/98 | 81,117 | 133 | 3,195 | 0.182 | 0.205 | 0.613 |
| 1998/99 | 91,826 | 135 | 1,322 | 0.193 | 0.216 | 0.591 |
| 1999/00 - 2008/09 | | | FISHERY CLOSED | | | |
| 2009/10 | 10,484 | 989 | 19,802 | 0.141 | 0.324 | 0.535 |
| 2010/11 | 29,356 | 2,419 | 45,466 | 0.131 | 0.315 | 0.553 |
| 2011/12 | 48,554 | 3,359 | 58,666 | 0.131 | 0.305 | 0.564 |
| 2012/13 | 37,065 | 2,841 | 57,298 | 0.141 | 0.318 | 0.541 |
| 2013/14 | | | FISHERY CLOSED | | | |
| 2014/15 | 10,133 | 895 | 9,906 | 0.094 | 0.228 | 0.679 |
| 2015/16 | | | | | | |

On the other hand, these same data suggest a significant reduction in the bycatch of females, which may be attributable to the later timing of the contemporary fishery and the more offshore distribution of fishery effort since reopening in 2009/10³. Some bycatch of discarded blue king crab has also been observed historically in the eastern Bering Sea snow crab fishery, but in recent years it has generally been negligible, and observers recorded no bycatch of blue king crab in sampled pot lifts during 2013/14. The St. Matthew Island golden king crab fishery, the third commercial crab fishery to have taken place in the area, typically occurred in areas with depths exceeding blue king crab distribution. NMFS observer data suggest that variable but mostly limited SMBKC bycatch has also occurred in the eastern Bering Sea groundfish fisheries (Table 5).

³D. Pengilly, ADF&G, pers. comm.

Table 5: Groundfish SMBKC male bycatch biomass (tonnes) estimates. Trawl includes pelagic trawl and non-pelagic trawl types. Source: J. Zheng, ADF&G, and author estimates based on data from R. Foy, NMFS. AKRO estimates used after 2008/09.

| Year | Trawl bycatch | Fixed gear bycatch |
|------|---------------|--------------------|
| 1991 | 3.538 | 0.045 |
| 1992 | 1.996 | 2.268 |
| 1993 | 1.542 | |
| 1994 | 0.318 | 0.091 |
| 1995 | 0.635 | 0.136 |
| 1996 | | 0.045 |
| 1997 | | 0.181 |
| 1998 | | 0.907 |
| 1999 | | 1.361 |
| 2001 | | 0.862 |
| 2002 | 0.726 | 0.408 |
| 2003 | 0.998 | 1.134 |
| 2004 | 0.091 | 0.635 |
| 2005 | | 0.590 |
| 2006 | 2.812 | 1.451 |
| 2007 | 0.045 | 69.717 |
| 2008 | 0.272 | 6.622 |
| 2009 | 0.635 | 7.530 |
| 2010 | 0.363 | 9.571 |
| 2011 | 0.181 | 0.590 |
| 2012 | | 0.590 |
| 2013 | 0.181 | 0.272 |
| 2014 | | 0.272 |
| 2015 | | 0.635 |

D. Data

Summary of New Information

Data used in this assessment have been updated to include the most recently available fishery and survey numbers. In addition, this assessment makes use of an updated trawl-survey time series provided by R. Foy in August 2015, as well as updated 1993-2014 groundfish bycatch estimates based on AKRO data also supplied by R. Foy. The data used in each of the new models is shown in Figure 3.

Major Data Sources

Major data sources used in this assessment include annual directed-fishery retained-catch statistics from fish tickets (1978/79-1998/99, 2009/10-2012/13, and 2014/15-2015/16; Table 3); results from the annual NMFS eastern Bering Sea trawl survey (1978-2016; Table 7); results from the triennial ADF&G SMBKC pot survey (every third year during 1995-2013), the 2015 pot survey, and the 2016 pot survey (Table 6); size-frequency information from ADF&G crab-observer pot-lift sampling (1990/91-1998/99, 2009/10-2012/13, and 2014/15-2015/16; Table 4); and NMFS groundfish-observer bycatch biomass estimates (1992/93-2015/16; Table 5).

Figure 4 maps stations from which SMBKC trawl-survey and pot-survey data were obtained. Further information concerning the NMFS trawl survey as it relates to commercial crab species is available in Daly et al. (2014); see Gish et al. (2012) for a description of ADF&G SMBKC pot-survey methods. It should be noted that the two surveys cover different geographic regions and that each has in some years encountered

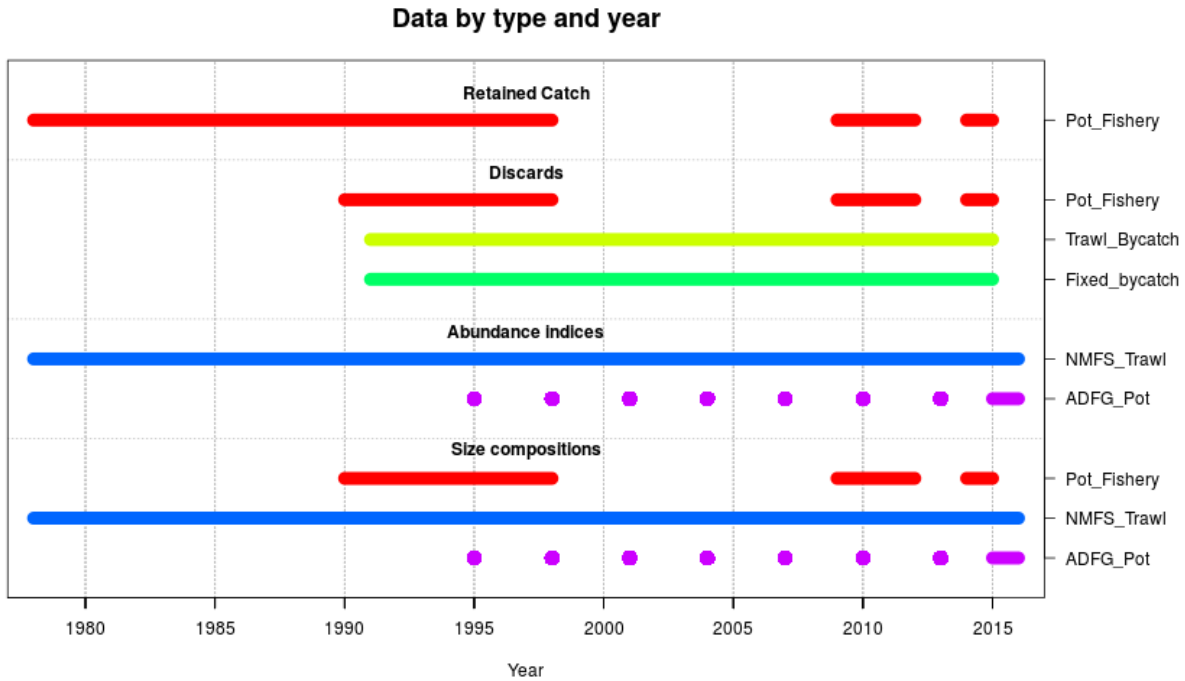


Figure 3: Data extent for the SMBKC assessment.

proportionally large numbers of male blue king crab in areas where the other is not represented (Figure 5). Crab-observer sampling protocols are detailed in the crab-observer training manual (ADF&G 2013). Groundfish SMBKC bycatch data come from NMFS Bering Sea reporting areas 521 and 524 (Figure 6). Note that for this assessment the newly available NMFS groundfish observer data reported by ADF&G statistical area was not used.

Other Data Sources

Recent model configurations developed for SMBKC makes use of a growth transition matrix based on Otto and Cummiskey (1990), the same growth transition matrix is used in this assessment. Other relevant data sources, including assumed population and fishery parameters, are presented in Appendix A, which also provides a detailed description of the model configuration used for this assessment.

Excluded Data Sources

Groundfish bycatch size-frequency data are available for selected years. These data were used in model-based assessments prior to 2011. However, they have since been excluded because these data tend to be severely limited: for example, 2012/13 data include a total of just 4 90 mm+ CL male blue king crab from reporting areas 521 and 524.



Figure 4: Trawl and pot-survey stations used in the SMBKC stock assessment.

Table 6: Size-class and total CPUE (90+ mm CL) with estimated CV and total number of captured crab (90+ mm CL) from the 96 common stations surveyed during the seven triennial ADF&G SMBKC pot surveys and the 2015 and 2016 surveys. Source: D. Pengilly and R. Gish, ADF&G.

| | Stage-1 | Stage-2 | Stage-3 | | | |
|------|-------------|--------------|-----------|------------|------|-----------------|
| Year | (90-104 mm) | (105-119 mm) | (120+ mm) | Total CPUE | CV | Number of crabs |
| 1995 | 1.919 | 3.198 | 6.922 | 12.042 | 0.13 | 4624 |
| 1998 | 0.964 | 2.763 | 8.804 | 12.531 | 0.06 | 4812 |
| 2001 | 1.266 | 1.737 | 5.487 | 8.477 | 0.08 | 3255 |
| 2004 | 0.112 | 0.414 | 1.141 | 1.667 | 0.15 | 640 |
| 2007 | 1.086 | 2.721 | 4.836 | 8.643 | 0.09 | 3319 |
| 2010 | 1.326 | 3.276 | 5.607 | 10.209 | 0.13 | 3920 |
| 2013 | 0.878 | 1.398 | 3.367 | 5.643 | 0.19 | 2167 |
| 2015 | 0.198 | 0.682 | 1.924 | 2.805 | 0.18 | 1077 |
| 2016 | | | | | | |

Table 7: NMFS EBS trawl-survey area-swept estimates of male crab abundance (10^6 crab) and of mature male biomass (10^6 lbs). Total number of captured male crab ≥ 90 mm CL is also given. Source: R. Foy, NMFS. The "+" refer to plus group.

| Year | Abundance | | | | | Biomass | | Number of crabs |
|------|------------------------|-------------------------|----------------------|-------|-------|----------------------|-------|-----------------|
| | Stage-1 (90-104 mm) | Stage-2 (105-119 mm) | Stage-3 (120+ mm) | Total | CV | Total (90+ mm CL) | CV | |
| 1978 | 2.213 | 1.991 | 1.521 | 5.726 | 0.411 | 15.064 | 0.394 | 157 |
| 1979 | 3.061 | 2.281 | 1.808 | 7.150 | 0.472 | 17.615 | 0.463 | 178 |
| 1980 | 2.856 | 2.563 | 2.541 | 7.959 | 0.572 | 22.017 | 0.507 | 185 |
| 1981 | 0.483 | 1.213 | 2.263 | 3.960 | 0.368 | 14.443 | 0.402 | 140 |
| 1982 | 1.669 | 2.431 | 5.884 | 9.984 | 0.401 | 35.763 | 0.344 | 271 |
| 1983 | 1.061 | 1.651 | 3.345 | 6.057 | 0.332 | 21.240 | 0.298 | 231 |
| 1984 | 0.435 | 0.497 | 1.452 | 2.383 | 0.175 | 8.976 | 0.179 | 105 |
| 1985 | 0.379 | 0.376 | 1.117 | 1.872 | 0.216 | 6.858 | 0.210 | 93 |
| 1986 | 0.203 | 0.447 | 0.374 | 1.025 | 0.428 | 3.124 | 0.388 | 46 |
| 1987 | 0.325 | 0.631 | 0.715 | 1.671 | 0.302 | 5.024 | 0.291 | 71 |
| 1988 | 0.410 | 0.816 | 0.957 | 2.183 | 0.285 | 6.963 | 0.252 | 81 |
| 1989 | 2.169 | 1.154 | 1.786 | 5.109 | 0.314 | 13.974 | 0.271 | 208 |
| 1990 | 1.053 | 1.031 | 2.338 | 4.422 | 0.302 | 14.837 | 0.274 | 170 |
| 1991 | 1.147 | 1.665 | 2.233 | 5.046 | 0.259 | 15.318 | 0.248 | 197 |
| 1992 | 1.074 | 1.382 | 2.291 | 4.746 | 0.206 | 15.638 | 0.201 | 220 |
| 1993 | 1.521 | 1.828 | 3.276 | 6.626 | 0.185 | 21.051 | 0.169 | 324 |
| 1994 | 0.883 | 1.298 | 2.257 | 4.438 | 0.187 | 14.416 | 0.176 | 211 |
| 1995 | 1.025 | 1.188 | 1.741 | 3.953 | 0.187 | 12.574 | 0.178 | 178 |
| 1996 | 1.238 | 1.891 | 3.064 | 6.193 | 0.263 | 20.746 | 0.241 | 285 |
| 1997 | 1.165 | 2.228 | 3.789 | 7.182 | 0.367 | 24.084 | 0.337 | 296 |
| 1998 | 0.660 | 1.661 | 2.849 | 5.170 | 0.373 | 17.586 | 0.355 | 243 |
| 1998 | 0.223 | 0.222 | 0.558 | 1.003 | 0.192 | 3.515 | 0.182 | 52 |
| 2000 | 0.282 | 0.285 | 0.740 | 1.307 | 0.303 | 4.623 | 0.310 | 61 |
| 2001 | 0.419 | 0.502 | 0.938 | 1.859 | 0.243 | 6.242 | 0.245 | 91 |
| 2002 | 0.111 | 0.230 | 0.640 | 0.981 | 0.311 | 3.820 | 0.320 | 38 |
| 2003 | 0.449 | 0.280 | 0.465 | 1.194 | 0.399 | 3.454 | 0.336 | 65 |
| 2004 | 0.247 | 0.184 | 0.562 | 0.993 | 0.369 | 3.360 | 0.305 | 48 |
| 2005 | 0.319 | 0.310 | 0.501 | 1.130 | 0.403 | 3.620 | 0.371 | 42 |
| 2006 | 0.917 | 0.642 | 1.240 | 2.798 | 0.339 | 8.585 | 0.334 | 126 |
| 2007 | 2.518 | 2.020 | 1.193 | 5.730 | 0.420 | 14.266 | 0.385 | 250 |
| 2008 | 1.352 | 0.801 | 1.457 | 3.609 | 0.289 | 10.261 | 0.284 | 167 |
| 2009 | 1.573 | 2.161 | 1.410 | 5.144 | 0.263 | 13.892 | 0.256 | 251 |
| 2010 | 3.937 | 3.253 | 2.458 | 9.648 | 0.544 | 24.539 | 0.466 | 388 |
| 2011 | 1.800 | 3.255 | 3.207 | 8.263 | 0.587 | 24.099 | 0.558 | 318 |
| 2012 | 0.705 | 1.970 | 1.808 | 4.483 | 0.361 | 13.669 | 0.339 | 193 |
| 2013 | 0.335 | 0.452 | 0.807 | 1.593 | 0.215 | 5.043 | 0.217 | 74 |
| 2014 | 0.723 | 1.627 | 1.809 | 4.160 | 0.503 | 13.292 | 0.449 | 181 |
| 2015 | 0.992 | 1.269 | 1.979 | 4.240 | 0.774 | 12.958 | 0.770 | 153 |
| 2016 | | | | | | | | |

E. Analytic Approach

History of Modeling Approaches for this Stock

A four-stage catch-survey-analysis (CSA) assessment model was used before 2011 to estimate abundance and biomass and prescribe fishery quotas for the SMBKC stock (2010 SAFE; Zheng et al. 1997). The four-stage CSA is similar to a full length-based analysis, the major difference being coarser length groups, which are more suited to a small stock with consistently low survey catches. In this approach, the abundance of male crab with a CL of 90 mm or above is modeled in terms of four crab stages: stage 1: 90-104 mm CL; stage 2: 105-119 mm CL; stage 3: newshell 120-133 mm CL; and stage 4: oldshell ≥ 120 mm CL and newshell ≥ 134 mm CL. Motivation for these stage definitions comes from the fact that for management of the SMBKC stock, male crab measuring at least 105 mm CL are considered mature, whereas 120 mm CL is considered a proxy for the legal size of 5.5 in carapace width, including spines. Additional motivation for these stage definitions comes from an estimated average growth increment of about 14 mm per molt for SMBKC (Otto and Cummiskey 1990).

Concerns about the pre-2011 assessment model led to the CPT and SSC recommendations that included development of an alternative model with provisional assessment based on survey biomass or some other index of abundance. An alternative 3-stage model was proposed to the CPT in May 2011 but was requested to proceed with a survey-based approach for the Fall 2011 assessment. In May 2012 the CPT approved a slightly revised and better documented version of the alternative model for assessment.

The 2015 SMBKC stock assessment model, first used in Fall 2012, was a variant of the previous four-stage SMBKC CSA model and similar in complexity to that described by Collie et al. (2005). Like the earlier model, it considered only male crab at least 90 mm in CL, but it combined stages 3 and 4 of the earlier model resulting in just three stages (male size classes) determined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) 120 mm+ (i.e., 120 mm and above). This consolidation was driven by concern about the accuracy and consistency of shell-condition information, which had been used in distinguishing stages 3 and 4 of the earlier model.

Assessment Methodology

The 2016 SMBKC assessment model makes use of the modeling framework Gmacs. The aim when developing this model was to first provide a fit to the data that best matched the 2015 SMBKC stock assessment model. A detailed description of the Gmacs model and its implementation is presented in Appendix A.

Model Selection and Evaluation

Five different Gmacs model scenarios were considered, in this document results from these models and the 2015 model are compared. The models include:

1. **2015 Model:** the 2015 provided by Jie. Note that an error was found in the 2015 model code⁴. This error was fixed before making comparisons. Fixing this error caused the NMFS trawl survey selectivity to exceed 1 for stage-2 crab.
2. **Gmacs match:** tries to match as closely as possible with the 2015 Model by fixing the stage-1 and stage-2 selectivity parameters and the catchability coefficient (q) for the ADF&G pot survey at those values estimated in the 2015 model (and allows the NMFS trawl survey selectivity to exceed 1 for stage-2 crab). The parameters that are estimated in this model include the average recruitment (\bar{R}), the recruitment deviations (δ_y^R), the initial numbers in each stage (n_0), the natural mortality deviation

⁴The error in the 2015 model code was in the population dynamics function where the growth transition matrix is applied to the numbers at length to calculate the numbers during the following time-step, specifically ' $N(t+1,3)=TM(2,3)*NN(2)+NN(3);$ ' which should be ' $N(t+1,3)=TM(1,3)*NN(1)+TM(2,3)*NN(2)+NN(3);$ '.

1998 (δ_{1998}^M), and the fishing mortalities for the directed pot fishery, the trawl bycatch fishery, and the fixed bycatch fishery ($\bar{F}^{\text{df}}, \bar{F}^{\text{tb}}, \bar{F}^{\text{fb}}, \delta_{t,y}^{\text{df}}, \delta_{t,y}^{\text{tb}}, \delta_{t,y}^{\text{fb}}$).

3. **Gmacs base:** directed pot, NMFS trawl survey and ADF&G pot survey selectivities are estimated for stage-1 and stage-2 crab (and fixed at 1 for stage-3 crab). These selectivities are bounded so that they cannot be greater than 1.
4. **Gmacs Francis:** uses the Francis iterative re-weighting method (Francis 2011) as well as estimating the directed pot, NMFS trawl survey and ADF&G pot survey selectivities for stage-1 and stage-2 crab. These selectivities are bounded so that they cannot be greater than 1.
5. **Gmacs M:** natural mortality (M) is fixed at 0.18 yr^{-1} during all years as well as using the Francis iterative re-weighting method and estimating the directed pot, NMFS trawl survey and ADF&G pot survey selectivities for stage-1 and stage-2 crab. These selectivities are bounded so that they cannot be greater than 1.
6. **Gmacs force:** natural mortality (M) is fixed at 0.18 yr^{-1} during all years and the directed pot, NMFS trawl survey and ADF&G pot survey selectivities for stage-1 and stage-2 crab are estimated. These selectivities are bounded so that they cannot be greater than 1. The Francis iterative re-weighting method is also used but the NMFS trawl survey is up-weighted by $\lambda^{\text{NMFS}} = 1.5$ and the ADF&G pot survey is up-weighted by $\lambda^{\text{ADFG}} = 2$.

Table 8 outlines the major features of each of the models.

Table 8: Description of the five different Gmacs scenarios.

| Scenario | Selectivity estimated | Use Francis weighting | Estimate M_{1998} |
|---------------|-----------------------|-----------------------|---------------------|
| Gmacs match | No | No | Yes |
| Gmacs base | Yes | No | Yes |
| Gmacs Francis | Yes | Yes | Yes |
| Gmacs M | Yes | Yes | No |
| Gmacs force | Yes | Yes | No |

Results

Results for all Gmacs scenarios are provided here with comparisons to the 2015 model. We recommend scenario XXX to be used for the overfishing determination in 2016, based on the fit to the data and the plausibility of parameter estimates.

a. Effective sample sizes and weighting factors.

Observed and estimated effective sample sizes are compared in Table 11. Effective sample sizes are also shown on size-frequency plots (Figures 14, 15, and 16). Data weighting factors, SDNRs, and MARs are presented in Table 18.

b. Tables of estimates.

Model parameter estimates for each of the Gmacs scenarios are summarized in Tables 12, 13, 14, 15, and 16. These parameter estimates are compared in Table 17. Negative log-likelihood values and management measures for each of the Gmacs scenarios are compared in Tables 19 and 20.

c. Graphs of estimates.

Estimated (and fixed) selectivities are compared in Figure 7.

The various model fits to total male (> 89 mm CL) trawl survey biomass are compared in Figures 8 and 9. The fits to pot survey CPUE are compared in Figures 10 and 11. Standardized residuals of total male trawl survey biomass and pot survey CPUE are plotted in Figures 12 and 13.

Fits to stage compositions for trawl survey, pot survey, and commercial observer data are shown in Figures 14, 15, and 16 for the all scenarios. Bubble plots of stage composition residuals for trawl survey, pot survey, and commercial observer data are shown for the **Gmacs base**, **Gmacs Francis**, **Gmacs M**, and **Gmacs force** scenarios in Figures 17, 18, 19, and 20, respectively.

Fits to retained catch numbers and bycatch biomass are shown for all Gmacs scenarios in Figure 21.

Estimated recruitment is compared in Figure 22. Estimated abundances by stage and mature male biomasses for all scenarios (including the 2015 model) are shown in Figures 26 and 23. Estimated natural mortality each year (M_t) is presented in Figure 27.

d. Graphic evaluation of the fit to the data.

There is little difference between model estimated survey biomass in the gmacs scenarios when compared with the 2015 model (Figures 8 and 10). Looking at the model fits to the NMFS trawl survey biomass (Figure 8), the **Gmacs match** scenario is the most similar to the 2015 model, and the **Gmacs base** model is very similar as well. In all scenarios, Gmacs produces a better fit during the mid-late 1980s. However, since about 2010 Gmacs estimates a slightly lower survey biomass than the 2015 model in an attempt to better fit the ADF&G pot survey CPUE (Figure 10). The two Gmacs scenarios that do not attempt to estimate natural mortality in 1998/99 (**Gmacs M** and **Gmacs force**) predict lower survey biomass from 1992 to 1998 than the other scenarios and the 2015 model. These same two runs also predict a lower survey biomass in recent years (since about 2010). While these two models may result in slightly worse fits to the data, they do not risk over-fitting the data in the same way the other scenarios do. As expected the model that upweights the NMFS survey biomass and ADF&G pot survey CPUE (**Gmacs force**) provides a better fit to the survey biomass during the mid-late 1980s and a much better fit to the pot survey CPUE in the most recent two years (Figures 8, 9, 10, and 11). Keep in mind that this scenario was only included for exploratory purposes and forcing these weights resulted in worse SDNR and MAR values for the two abundance indices.

LFs

Estimated recruitment to the model is variable over time (Figure 22). Estimated recruitment during recent years is generally low in all scenarios. Estimated mature male biomass on 15 February also fluctuates strongly over time. The high biomass estimates in recent years for the Gmacs CV scenario is quite different to the other scenarios (Figure 23).

e. Retrospective and historic analyses.

Gmacs retrospective analyses under development.

f. Uncertainty and sensitivity analyses.

Estimated standard deviations of parameters and selected management measures for the five Gmacs scenarios are summarized in Tables 12, 13, 14, 15, and 16. Probabilities for mature male biomass and OFL in 2016 are illustrated in Section F.

g. Comparison of alternative model scenarios.

Discussion to come.

F. Calculation of the OFL and ABC

The overfishing level (OFL) is the fishery-related mortality biomass associated with fishing mortality F_{OFL} . The SMBKC stock is currently managed as Tier 4 (2013 SAFE), and only a Tier 4 analysis is presented here. Thus given stock estimates or suitable proxy values of B_{MSY} and F_{MSY} , along with two additional parameters α and β , F_{OFL} is determined by the control rule

$$F_{OFL} = \begin{cases} F_{MSY}, & \text{when } B/B_{MSY} > 1 \\ F_{MSY} \frac{(B/B_{MSY} - \alpha)}{(1 - \alpha)}, & \text{when } \beta < B/B_{MSY} \leq 1 \end{cases} \quad (1)$$

$F_{OFL} < F_{MSY}$ with directed fishery $F = 0$ when $B/B_{MSY} \leq \beta$

where B is quantified as mature-male biomass (MMB) at mating with time of mating assigned a nominal date of 15 February. Note that as B itself is a function of the fishing mortality F_{OFL} (therefore numerical approximation of F_{OFL} is required). As implemented for this assessment, all calculations proceed according to the model equations given in Appendix A. F_{OFL} is taken to be full-selection fishing mortality in the directed pot fishery and groundfish trawl and fixed-gear fishing mortalities set at their model geometric mean values over years for which there are data-based estimates of bycatch-mortality biomass.

The currently recommended Tier 4 convention is to use the full assessment period, currently 1978-2016, to define a B_{MSY} proxy in terms of average estimated MMB and to set $\gamma = 1.0$ with assumed stock natural mortality $M = 0.18 \text{ yr}^{-1}$ in setting the F_{MSY} proxy value γM . The parameters α and β are assigned their default values $\alpha = 0.10$ and $\beta = 0.25$. The F_{OFL} , OFL, and MMB in 2016 for all scenarios are summarized in Table XXX. ABC is 80% of the OFL.

OFL, ABC, retained catch and bycatches for 2016 are summarized for all scenarios in Table 9.

Table 9: Summary of the OFL, ABC, retained catch and bycatches for the five different Gmacs scenarios.

| Scenario | OFL | ABC | Retained catch | Pot male bycatch | Groundfish bycatch |
|---------------|-----|-----|----------------|------------------|--------------------|
| Gmacs match | | | | | |
| Gmacs base | | | | | |
| Gmacs Francis | | | | | |
| Gmacs M | | | | | |
| Gmacs force | | | | | |

G. Rebuilding Analysis

This stock is not currently subject to a rebuilding plan.

H. Data Gaps and Research Priorities

1. Growth increments and molting probabilities as a function of size.
2. Trawl survey catchability and selectivities.
3. Temporal changes in spatial distributions near the island.
4. Natural mortality.

I. Projections and Future Outlook

With the decline of estimated population biomass during recent years, outlook for this stock is not promising. If the decline continues, the stock will fall to depleted status soon.

J. Acknowledgements

We thank the Crab Plan Team, Doug Pengilly for reviewing the earlier draft of this manuscript. Some materials in the report are from the SAFE report prepared by Bill Gaeuman in 2014. We thank Andre Punt for his input into the Gmacs model and for finding the error in the old SMBKC model code.

K. References

- Alaska Department of Fish and Game (ADF&G). 2013. Crab observer training and deployment manual. Alaska Department of Fish and Game Shellfish Observer Program, Dutch Harbor. Unpublished.
- Collie, J.S., A.K. Delong, and G.H. Kruse. 2005. Three-stage catch-survey analysis applied to blue king crabs. Pages 683-714 [In] Fisheries assessment and management in data-limited situations. University of Alaska Fairbanks, Alaska Sea Grant Report 05-02, Fairbanks.
- Daly, B., R. Foy, and C. Armistead. 2014. The 2013 eastern Bering Sea continental shelf bottom trawl survey: results for commercial crab species. NOAA Technical Memorandum, NMFS-AFSC.
- Donaldson, W.E., and S.C. Byersdorfer. 2005. Biological field techniques for lithodid crabs. University of Alaska Fairbanks, Alaska Sea Grant Report 05-03, Fairbanks.
- Fitch, H., M. Deiman, J. Shaishnikoff, and K. Herring. 2012. Annual management report for the commercial and subsistence shellfish fisheries of the Bering Sea, 2010/11. Pages 75-1776 [In] Fitch, H., M. Schwenzfeier, B. Baechler, T. Hartill, M. Salmon, M. Deiman, E.
- Evans, E. Henry, L. Wald, J. Shaishnikoff, K. Herring, and J. Wilson. 2012. Annual management report for the commercial and subsistence shellfish fisheries of the Aleutian Islands, Bering Sea and the Westward Region's Shellfish Observer Program, 2010/11. Alaska Department of Fish and Game, Fishery Management Report No. 12-22, Anchorage.
- Fournier, D.A., H.J. Skaug, J. Ancheta, J. Ianelli, A. Magnusson, M.N. Maunder, A. Nielsen, and J. Sibert. 2012. AD Model Builder: using automatic differentiation for statistical inference of highly parameterized complex nonlinear models. *Optim. Methods Softw.* 27:233-249.
- Francis, R.I.C.C. 2011. Data weighting in statistical fisheries stock assessment models. *Can. J. Fish. Aquat. Sci.* 68: 1124-1138.
- Gaeuman, W.B. 2013. Summary of the 2012/13 mandatory crab observer program database for the Bering Sea/Aleutian Islands commercial crab fisheries. Alaska Department of Fish and Game, Fishery Data Series No. 13-54, Anchorage.
- Gish, R.K., V.A. Vanek, and D. Pengilly. 2012. Results of the 2010 triennial St. Matthew Island blue king crab pot survey and 2010/11 tagging study. Alaska Department of Fish and Game, Fishery Management Report No. 12-24, Anchorage.
- Jensen, G.C. and D.A. Armstrong. 1989. Biennial reproductive cycle of blue king crab, *Paralithodes platypus*, at the Pribilof Islands, Alaska and comparison to a congener, *P. camtschatica*. *Can. J. Fish. Aquat. Sci.* 46: 932-940.

Moore, H., L.C. Byrne, and D. Connolly. 2000. Alaska Department of Fish and Game summary of the 1998 mandatory shellfish observer program database. Alaska Dept. Fish and Game, Commercial Fisheries Division, Reg. Inf. Rep. 4J00-21, Kodiak.

North Pacific Fishery Management Council (NPFMC). 1998. Fishery Management Plan for Bering Sea/Aleutian Islands king and Tanner crabs. North Pacific Fishery Management Council, Anchorage.

North Pacific Fishery Management Council (NPFMC). 1999. Environmental assessment/regulatory impact review/initial regulatory flexibility analysis for Amendment 11 to the Fishery Management Plan for Bering Sea/Aleutian Islands king and Tanner crabs. North Pacific Fishery Management Council, Anchorage.

North Pacific Fishery Management Council (NPFMC). 2000. Environmental assessment/regulatory impact review/initial regulatory flexibility analysis for proposed Amendment 15 to the Fishery Management Plan for king and Tanner crab fisheries in the Bering Sea/Aleutian Islands and regulatory amendment to the Fishery Management Plan for the groundfish fishery of the Bering Sea and Aleutian Islands area: A rebuilding plan for the St. Matthew blue king crab stock. North Pacific Fishery Management Council, Anchorage. Draft report.

North Pacific Fishery Management Council (NPFMC). 2007. Public Review Draft: Environmental assessment for proposed Amendment 24 to the Fishery Management Plan for Bering Sea and Aleutian Islands king and Tanner crabs to revise overfishing definitions. 14 November 2007. North Pacific Fishery Management Council, Anchorage.

Otto, R.S. 1990. An overview of eastern Bering Sea king and Tanner crab fisheries. Pages 9-26 [In] Proceedings of the international symposium on king and Tanner crabs. University of Alaska Fairbanks, Alaska Sea Grant Program Report 90-4, Fairbanks.

Otto, R.S., and P.A. Cummiskey. 1990. Growth of adult male blue king crab (*Paralithodes platypus*). Pages 245-258 [In] Proceedings of the international symposium on king and Tanner crabs. University of Alaska Fairbanks, Alaska Sea Grant Report 90-4, Fairbanks.

Paul, J.M., A. J. Paul, R.S. Otto, and R.A. MacIntosh. 1991. Spermatophore presence in relation to carapace length for eastern Bering Sea blue king crab (*Paralithodes platypus*, Brandt, 1850) and red king crab (*P. Camtschaticus*, Tilesius, 1815). *J. Shellfish Res.* 10: 157-163.

Pengilly, D. and D. Schmidt. 1995. Harvest Strategy for Kodiak and Bristol Bay Red king Crab and St. Matthew Island and Pribilof Blue King Crab. Alaska Department of Fish and Game, Commercial Fisheries Management and Development Division, Special Publication Number 7, Juneau.

Schirripa, M.J., C.P. Goodyear, and R.M. Methot. 2009. Testing different methods of incorporating climate data into the assessment of US West Coast sablefish. *ICES Journal of Marine Science*, 66: 1605–1613.

Somerton, D.A., and R.A. MacIntosh. 1983. The size at sexual maturity of blue king crab, *Paralithodes platypus*, in Alaska. *Fishery Bulletin* 81: 621-828.

Wilderbuer, T., D. G. Nichol, and J. Ianelli. 2013. Assessment of the yellowfin sole stock in the Bering Sea and Aleutian Islands. Pages 619-708 in 2013 North Pacific Groundfish Stock Assessment and Fishery Evaluation Reports for 2014. North Pacific Fishery Management Council, Anchorage.

Zheng, J. 2005. A review of natural mortality estimation for crab stocks: data-limited for every stock? Pages 595-612 [In] *Fisheries Assessment and Management in Data-Limited Situations*. University of Alaska Fairbanks, Alaska Sea Grant Program Report 05-02, Fairbanks.

Zheng, J., and G.H. Kruse. 2002. Assessment and management of crab stocks under uncertainty of massive die-offs and rapid changes in survey catchability. Pages 367-384 [In] A.J. Paul, E.G. Dawe, R. Elnor, G.S. Jamieson, G.H. Kruse, R.S. Otto, B. Sainte-Marie, T.C. Shirley, and D. Woodby (eds.). *Crabs in Cold Water Regions: Biology, Management, and Economics*. University of Alaska Fairbanks, Alaska Sea Grant Report 02-01, Fairbanks.

Zheng, J., M.C. Murphy, and G.H. Kruse. 1997. Application of catch-survey analysis to blue king crab stocks near Pribilof and St. Matthew Islands. *Alaska Fish. Res. Bull.* 4:62-74.

Table 10: Mean weight (kg) by stage in used in all of the models (provided as a vector of weights at length each year to Gmacs).

| Year | Stage-1 | Stage-2 | Stage-3 |
|------|---------|---------|---------|
| 1978 | 0.7 | 1.2 | 1.9 |
| 1979 | 0.7 | 1.2 | 1.7 |
| 1980 | 0.7 | 1.2 | 1.9 |
| 1981 | 0.7 | 1.2 | 1.9 |
| 1982 | 0.7 | 1.2 | 1.9 |
| 1983 | 0.7 | 1.2 | 2.1 |
| 1984 | 0.7 | 1.2 | 1.9 |
| 1985 | 0.7 | 1.2 | 2.1 |
| 1986 | 0.7 | 1.2 | 1.9 |
| 1987 | 0.7 | 1.2 | 1.9 |
| 1988 | 0.7 | 1.2 | 1.9 |
| 1989 | 0.7 | 1.2 | 2.0 |
| 1990 | 0.7 | 1.2 | 1.9 |
| 1991 | 0.7 | 1.2 | 2.0 |
| 1992 | 0.7 | 1.2 | 1.9 |
| 1993 | 0.7 | 1.2 | 2.0 |
| 1994 | 0.7 | 1.2 | 1.9 |
| 1995 | 0.7 | 1.2 | 2.0 |
| 1996 | 0.7 | 1.2 | 2.0 |
| 1997 | 0.7 | 1.2 | 2.1 |
| 1998 | 0.7 | 1.2 | 2.0 |
| 1999 | 0.7 | 1.2 | 1.9 |
| 2000 | 0.7 | 1.2 | 1.9 |
| 2001 | 0.7 | 1.2 | 1.9 |
| 2002 | 0.7 | 1.2 | 1.9 |
| 2003 | 0.7 | 1.2 | 1.9 |
| 2004 | 0.7 | 1.2 | 1.9 |
| 2005 | 0.7 | 1.2 | 1.9 |
| 2006 | 0.7 | 1.2 | 1.9 |
| 2007 | 0.7 | 1.2 | 1.9 |
| 2008 | 0.7 | 1.2 | 1.9 |
| 2009 | 0.7 | 1.2 | 1.9 |
| 2010 | 0.7 | 1.2 | 1.8 |
| 2011 | 0.7 | 1.2 | 1.8 |
| 2012 | 0.7 | 1.2 | 1.8 |
| 2013 | 0.7 | 1.2 | 1.9 |
| 2014 | 0.7 | 1.2 | 1.9 |
| 2015 | 0.7 | 1.2 | 1.9 |
| 2016 | 0.7 | 1.2 | 1.9 |

Table 11: Observed and effective sample sizes for observer data from the directed pot fishery, the NMFS trawl survey, and the ADF&G pot survey.

| Year | Observed sample sizes | | | Effective sample sizes | | |
|------|-----------------------|------------|-----------|------------------------|------------|-----------|
| | Observer pot | NMFS trawl | ADF&G pot | Observer pot | NMFS trawl | ADF&G pot |
| 1978 | | 157 | | | 50 | |
| 1979 | | 178 | | | 50 | |
| 1980 | | 185 | | | 50 | |
| 1981 | | 140 | | | 50 | |
| 1982 | | 271 | | | 50 | |
| 1983 | | 231 | | | 50 | |
| 1984 | | 105 | | | 50 | |
| 1985 | | 93 | | | 46.5 | |
| 1986 | | 46 | | | 23 | |
| 1987 | | 71 | | | 35.5 | |
| 1988 | | 81 | | | 40.5 | |
| 1989 | | 208 | | | 50 | |
| 1990 | 150 | 170 | | 15 | 50 | |
| 1991 | 3393 | 197 | | 25 | 50 | |
| 1992 | 1606 | 220 | | 25 | 50 | |
| 1993 | 2241 | 324 | | 25 | 50 | |
| 1994 | 4735 | 211 | | 25 | 50 | |
| 1995 | 663 | 178 | 4624 | 25 | 50 | 100 |
| 1996 | 489 | 285 | | 25 | 50 | |
| 1997 | 3195 | 296 | | 25 | 50 | |
| 1998 | 1323 | 243 | 4812 | 25 | 50 | 100 |
| 1999 | | 52 | | | 26 | |
| 2000 | | 61 | | | 30.5 | |
| 2001 | | 91 | 3255 | | 45.5 | 100 |
| 2002 | | 38 | | | 19 | |
| 2003 | | 65 | | | 32.5 | |
| 2004 | | 48 | 640 | | 24 | 100 |
| 2005 | | 42 | | | 21 | |
| 2006 | | 126 | | | 50 | |
| 2007 | | 250 | 3319 | | 50 | 100 |
| 2008 | | 167 | | | 50 | |
| 2009 | 19802 | 251 | | 50 | 50 | |
| 2010 | 45466 | 388 | 3920 | 50 | 50 | 100 |
| 2011 | 58667 | 318 | | 50 | 50 | |
| 2012 | 57282 | 193 | | 50 | 50 | |
| 2013 | | 74 | 2167 | | 37 | 100 |
| 2014 | 9906 | 181 | | 50 | 50 | |
| 2015 | | 153 | 1077 | | 50 | 100 |

Table 12: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the **Gmacs match** model.

| Parameter | Estimate | SD |
|--|----------|---------|
| Natural mortality deviation in 1998/99 δ_{1998}^M | 1.667 | 0.116 |
| $\log(\bar{R})$ | 13.360 | 0.048 |
| $\log(N_1)$ | 14.894 | 0.169 |
| $\log(N_2)$ | 14.477 | 0.194 |
| $\log(N_3)$ | 14.285 | 0.200 |
| $\log(\bar{F}_{\text{pot}})$ | -1.519 | 0.045 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.228 | 0.068 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.130 | 0.068 |
| F_{OFL} | 0.088 | 0.009 |
| OFL | 826.120 | 152.500 |

Table 13: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the **Gmacs base** model.

| Parameter | Estimate | SD |
|--|----------|---------|
| Natural mortality deviation in 1998/99 δ_{1998}^M | 1.659 | 0.128 |
| $\log(\bar{R})$ | 13.373 | 0.059 |
| $\log(N_1)$ | 14.861 | 0.171 |
| $\log(N_2)$ | 14.509 | 0.197 |
| $\log(N_3)$ | 14.213 | 0.210 |
| ADF&G pot survey catchability ($q \times 1000$) | 3.834 | 0.293 |
| $\log(\bar{F}_{\text{pot}})$ | -1.517 | 0.054 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.258 | 0.082 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.160 | 0.082 |
| log Stage-1 directed pot selectivity 1978-2008 | -0.709 | 0.175 |
| log Stage-2 directed pot selectivity 1978-2008 | -0.399 | 0.126 |
| log Stage-1 directed pot selectivity 2009-2016 | -0.611 | 0.164 |
| log Stage-2 directed pot selectivity 2009-2016 | -0.000 | 0.000 |
| log Stage-1 NMFS trawl selectivity | -0.222 | 0.066 |
| log Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| log Stage-1 ADF&G pot selectivity | -0.750 | 0.136 |
| log Stage-2 ADF&G pot selectivity | -0.062 | 0.083 |
| F_{OFL} | 0.088 | 0.011 |
| OFL | 826.770 | 191.960 |

Table 14: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the **Gmacs Francis** model.

| Parameter | Estimate | SD |
|--|----------|---------|
| Natural mortality deviation in 1998/99 δ_{1998}^M | 1.668 | 0.136 |
| $\log(\bar{R})$ | 13.376 | 0.059 |
| $\log(N_1)$ | 14.836 | 0.205 |
| $\log(N_2)$ | 14.538 | 0.227 |
| $\log(N_3)$ | 14.231 | 0.236 |
| ADF&G pot survey catchability ($q \times 1000$) | 3.767 | 0.274 |
| $\log(\bar{F}_{\text{pot}})$ | -1.488 | 0.057 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.253 | 0.082 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.155 | 0.082 |
| log Stage-1 directed pot selectivity 1978-2008 | -0.621 | 0.183 |
| log Stage-2 directed pot selectivity 1978-2008 | -0.417 | 0.149 |
| log Stage-1 directed pot selectivity 2009-2016 | -0.498 | 0.174 |
| log Stage-2 directed pot selectivity 2009-2016 | -0.000 | 0.000 |
| log Stage-1 NMFS trawl selectivity | -0.150 | 0.062 |
| log Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| log Stage-1 ADF&G pot selectivity | -0.792 | 0.136 |
| log Stage-2 ADF&G pot selectivity | -0.000 | 0.000 |
| F_{OFL} | 0.089 | 0.011 |
| OFL | 833.790 | 193.920 |

Table 15: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the **Gmacs M** model.

| Parameter | Estimate | SD |
|---|----------|---------|
| $\log(\bar{R})$ | 13.227 | 0.054 |
| $\log(N_1)$ | 14.836 | 0.207 |
| $\log(N_2)$ | 14.601 | 0.224 |
| $\log(N_3)$ | 14.275 | 0.236 |
| ADF&G pot survey catchability ($q \times 1000$) | 4.522 | 0.298 |
| $\log(\bar{F}_{\text{pot}})$ | -1.421 | 0.056 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.158 | 0.080 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.060 | 0.080 |
| log Stage-1 directed pot selectivity 1978-2008 | -0.509 | 0.183 |
| log Stage-2 directed pot selectivity 1978-2008 | -0.394 | 0.150 |
| log Stage-1 directed pot selectivity 2009-2016 | -0.500 | 0.175 |
| log Stage-2 directed pot selectivity 2009-2016 | -0.000 | 0.000 |
| log Stage-1 NMFS trawl selectivity | -0.071 | 0.061 |
| log Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| log Stage-1 ADF&G pot selectivity | -0.747 | 0.136 |
| log Stage-2 ADF&G pot selectivity | -0.000 | 0.000 |
| F_{OFL} | 0.073 | 0.010 |
| OFL | 564.370 | 132.530 |

Table 16: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the **Gmacs force** model.

| Parameter | Estimate | SD |
|---|----------|--------|
| $\log(\bar{R})$ | 13.096 | 0.049 |
| $\log(N_1)$ | 14.788 | 0.206 |
| $\log(N_2)$ | 14.588 | 0.218 |
| $\log(N_3)$ | 14.248 | 0.228 |
| ADF&G pot survey catchability ($q \times 1000$) | 4.006 | 0.193 |
| $\log(\bar{F}_{\text{pot}})$ | -1.337 | 0.044 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.175 | 0.070 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.076 | 0.070 |
| log Stage-1 directed pot selectivity 1978-2008 | -0.635 | 0.178 |
| log Stage-2 directed pot selectivity 1978-2008 | -0.502 | 0.147 |
| log Stage-1 directed pot selectivity 2009-2016 | -0.228 | 0.169 |
| log Stage-2 directed pot selectivity 2009-2016 | -0.000 | 0.000 |
| log Stage-1 NMFS trawl selectivity | -0.029 | 0.059 |
| log Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| log Stage-1 ADF&G pot selectivity | -0.283 | 0.175 |
| log Stage-2 ADF&G pot selectivity | -0.000 | 0.000 |
| F_{OFL} | 0.057 | 0.005 |
| OFL | 355.850 | 50.454 |

Table 17: Comparisons of model parameter estimates for the five Gmacs model scenarios.

| Parameter | Base | Force | Francis | M | Match |
|--|---------|---------|---------|---------|---------|
| ADF&G pot survey catchability (q) | 0.004 | 0.004 | 0.004 | 0.005 | -1.519 |
| $\log(\bar{F}_{\text{fixed bycatch}})$ | -9.160 | -9.076 | -9.155 | -9.060 | - |
| $\log(\bar{F}_{\text{pot}})$ | -1.517 | -1.337 | -1.488 | -1.421 | -12.228 |
| $\log(\bar{F}_{\text{trawl bycatch}})$ | -12.258 | -12.175 | -12.253 | -12.158 | -9.130 |
| $\log(\bar{R})$ | 13.373 | 13.096 | 13.376 | 13.227 | 13.360 |
| $\log(N_1)$ | 14.861 | 14.788 | 14.836 | 14.836 | 14.894 |
| $\log(N_2)$ | 14.509 | 14.588 | 14.538 | 14.601 | 14.477 |
| $\log(N_3)$ | 14.213 | 14.248 | 14.231 | 14.275 | 14.285 |
| log Stage-1 ADF&G pot selectivity | -0.750 | -0.283 | -0.792 | -0.747 | - |
| log Stage-1 directed pot selectivity 1978-2008 | -0.709 | -0.635 | -0.621 | -0.509 | - |
| log Stage-1 directed pot selectivity 2009-2015 | -0.611 | -0.228 | -0.498 | -0.500 | - |
| log Stage-1 NMFS trawl selectivity | -0.222 | -0.029 | -0.150 | -0.071 | - |
| log Stage-2 ADF&G pot selectivity | -0.062 | -0.000 | -0.000 | -0.000 | - |
| log Stage-2 directed pot selectivity 1978-2008 | -0.399 | -0.502 | -0.417 | -0.394 | - |
| log Stage-2 directed pot selectivity 2009-2015 | -0.000 | -0.000 | -0.000 | -0.000 | - |
| log Stage-2 NMFS trawl selectivity | -0.000 | -0.000 | -0.000 | -0.000 | - |
| Natural mortality (M) deviation in 1998/99 | 1.659 | - | 1.668 | - | 1.667 |

Table 18: Comparisons of data weights, Francis LF weights, SDNR values, and MAR values for the five Gmacs model scenarios.

| Component | Match | Base | Francis | M | Force |
|---|-------|------|---------|------|-------|
| NMFS trawl survey weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.50 |
| ADF&G pot survey weight | 1.00 | 1.00 | 1.00 | 1.00 | 2.00 |
| Directed pot LF weight | 1.00 | 1.00 | 1.72 | 1.61 | 1.40 |
| NMFS trawl survey LF weight | 1.00 | 1.00 | 0.50 | 0.51 | 0.26 |
| ADF&G pot survey LF weight | 1.00 | 1.00 | 1.70 | 1.21 | 0.38 |
| Francis weight for directed pot LF | 1.71 | 1.71 | 1.72 | 1.61 | 1.40 |
| Francis weight for NMFS trawl survey LF | 0.49 | 0.47 | 0.50 | 0.51 | 0.26 |
| Francis weight for ADF&G pot survey LF | 1.96 | 2.24 | 1.70 | 1.21 | 0.38 |
| SDNR NMFS trawl survey | 1.44 | 1.41 | 1.36 | 1.54 | 2.26 |
| SDNR ADF&G pot survey | 3.97 | 3.83 | 3.77 | 3.79 | 5.94 |
| SDNR directed pot LF | 0.68 | 0.65 | 0.67 | 0.68 | 0.80 |
| SDNR NMFS trawl survey LF | 1.25 | 1.32 | 1.31 | 1.35 | 1.79 |
| SDNR ADF&G pot survey LF | 0.79 | 0.83 | 0.96 | 1.00 | 1.67 |
| MAR NMFS trawl survey | 1.06 | 1.15 | 1.14 | 1.25 | 1.69 |
| MAR ADF&G pot survey | 3.03 | 2.94 | 2.74 | 3.41 | 4.70 |
| MAR directed pot LF | 0.47 | 0.43 | 0.55 | 0.49 | 0.60 |
| MAR NMFS trawl survey LF | 0.55 | 0.53 | 0.68 | 0.70 | 1.01 |
| MAR ADF&G pot survey LF | 0.56 | 0.43 | 0.57 | 0.62 | 0.92 |

Table 19: Comparisons of negative log-likelihood values for the five Gmacs model scenarios.

| Component | Match | Base | Francis | M | Force |
|-------------------------------|--------|--------|---------|--------|--------|
| Pot Retained Catch | -69.05 | -69.21 | -69.25 | -69.07 | -67.37 |
| Pot Discarded Catch | 6.58 | 6.21 | 6.33 | 5.71 | 8.22 |
| Trawl bycatch Discarded Catch | -6.88 | -6.88 | -6.88 | -6.88 | -6.88 |
| Fixed bycatch Discarded Catch | -6.85 | -6.86 | -6.86 | -6.87 | -6.86 |
| NMFS Trawl Survey | -6.12 | -7.93 | -10.40 | 1.53 | 41.59 |
| ADF&G Pot Survey CPUE | 57.08 | 52.01 | 49.68 | 52.73 | 145.40 |
| Directed Pot LF | -12.03 | -12.84 | 11.42 | 11.66 | 14.56 |
| NMFS Trawl LF | 20.24 | 26.65 | 54.72 | 58.11 | 97.49 |
| ADF&G Pot LF | -6.70 | -6.00 | 1.50 | 2.01 | 14.92 |
| Recruitment deviations | 58.13 | 58.05 | 57.67 | 58.67 | 63.07 |
| F penalty | 14.49 | 14.49 | 14.49 | 14.49 | 14.49 |
| M penalty | 6.47 | 6.47 | 6.47 | 0.00 | 0.00 |
| Prior | 13.72 | 13.71 | 13.71 | 13.71 | 13.71 |
| Total | 69.09 | 67.87 | 122.60 | 135.81 | 332.34 |
| Total estimated parameters | 282.00 | 291.00 | 291.00 | 289.00 | 289.00 |

Table 20: Comparisons of management measures for the five Gmacs model scenarios. Biomass and OFL are in tonnes.

| Component | Gmacs match | Gmacs base | Gmacs Francis | Gmacs M | Gmacs force |
|--------------|-------------|------------|---------------|----------|-------------|
| MMB_{2016} | 2225.552 | 2229.252 | 2205.083 | 1803.586 | 1447.430 |
| B_{MSY} | 3681.508 | 3692.217 | 3614.103 | 3462.258 | 3335.423 |
| F_{OFL} | 0.088 | 0.088 | 0.089 | 0.073 | 0.058 |
| OFL_{2016} | 826.123 | 826.770 | 833.788 | 564.369 | 355.853 |
| ABC_{2016} | 660.898 | 661.416 | 667.030 | 451.495 | 284.682 |

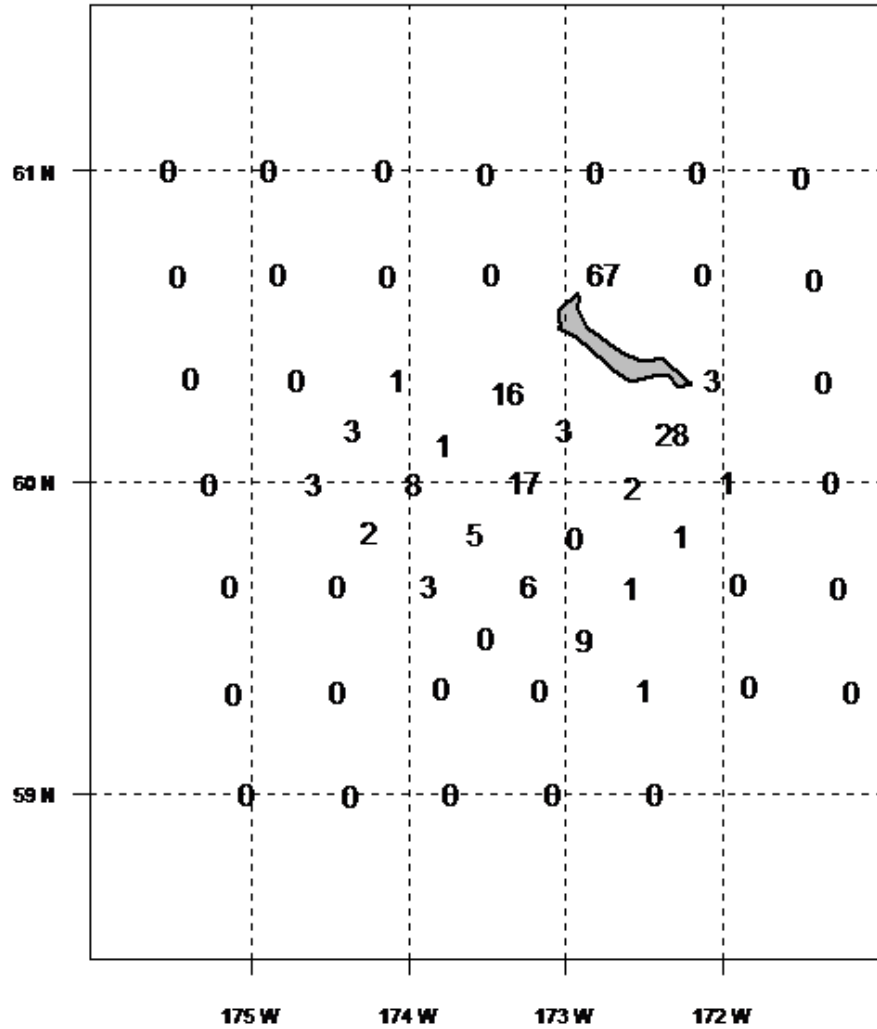


Figure 5: Catches of 181 male blue king crab measuring at least 90 mm CL from the 2014 NMFS trawl-survey at the 56 stations used to assess the SMBKC stock. Note that the area north of St. Matthew Island, which includes the large catch of 67 crab at station R-24, is not represented in the ADF&G pot-survey data used in the assessment.

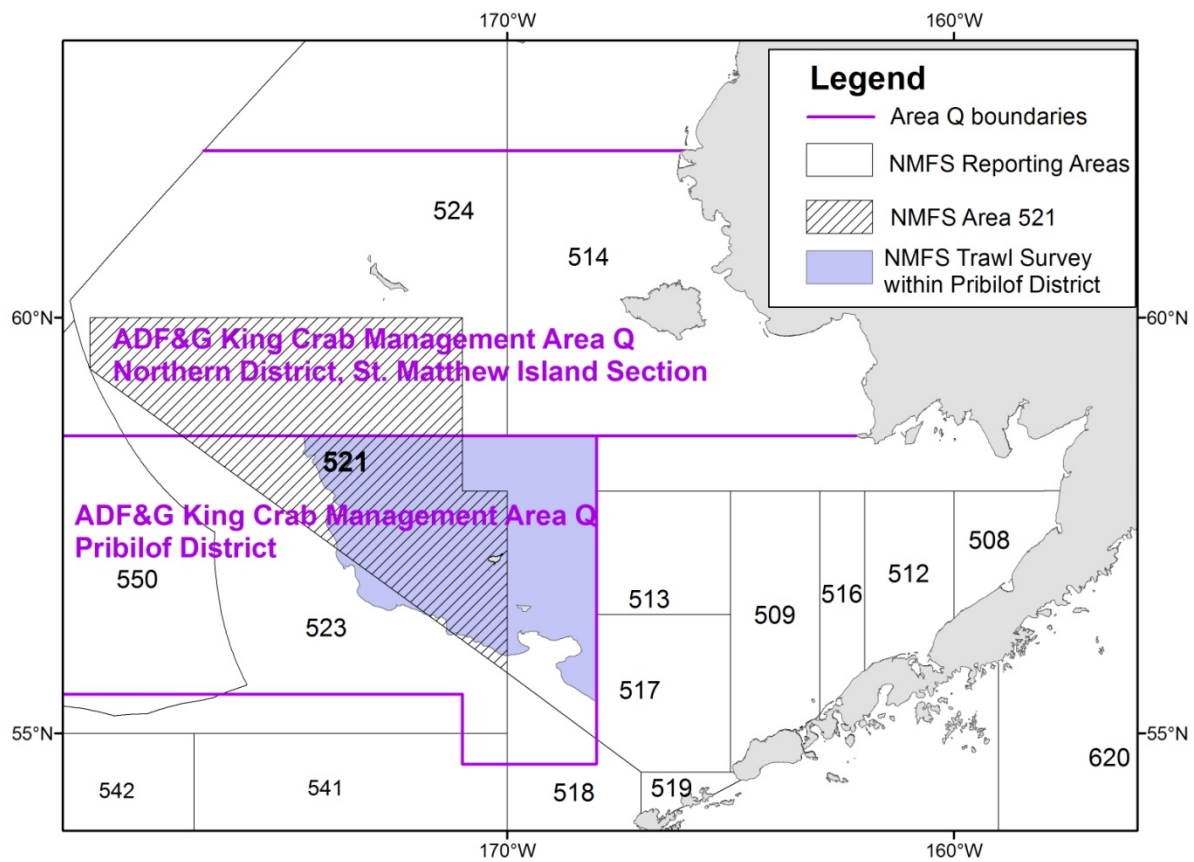


Figure 6: NFMS Bering Sea reporting areas. Estimates of SMBKC bycatch in the groundfish fisheries are based on NMFS observer data from reporting areas 524 and 521.

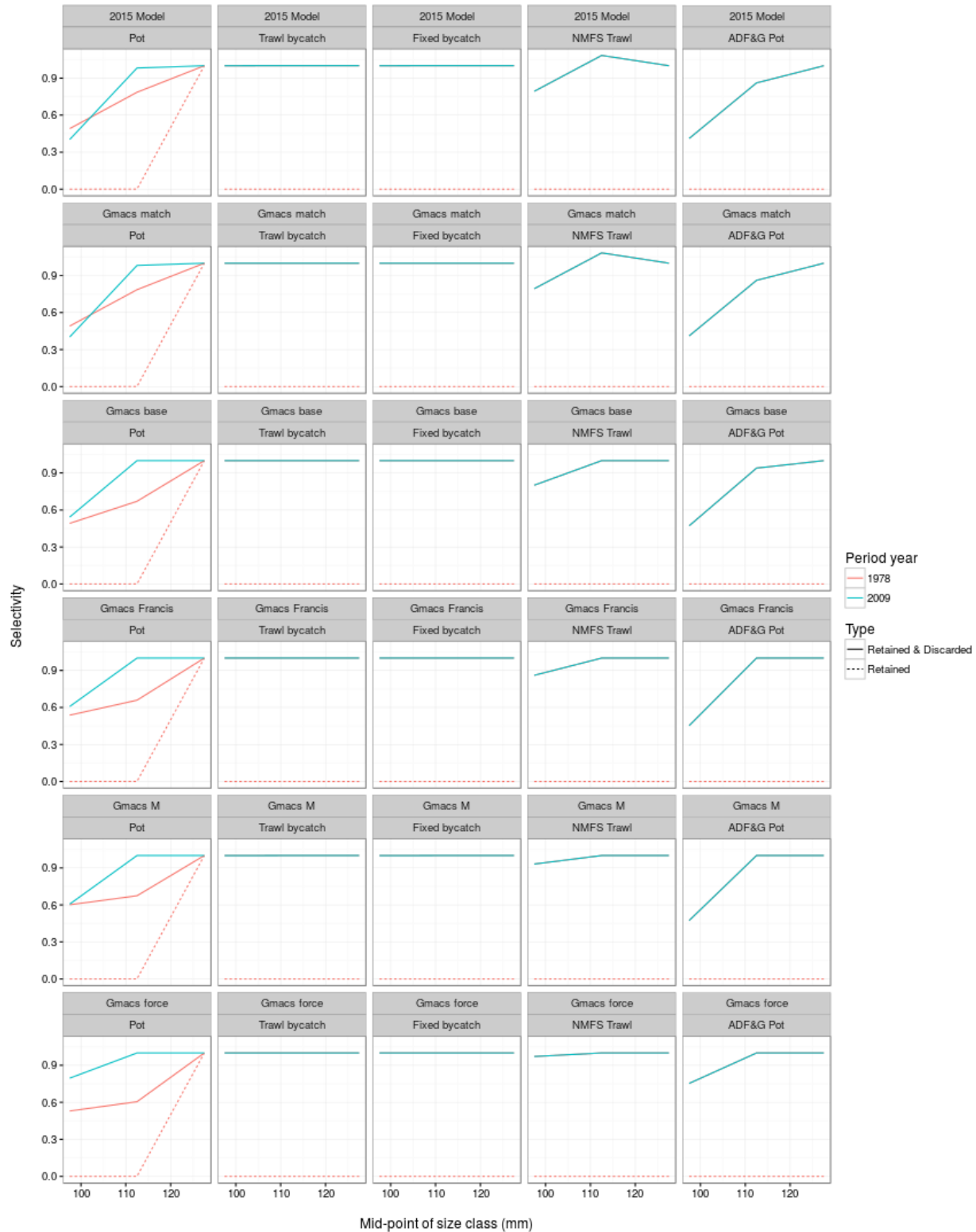


Figure 7: Comparisons of the estimated (and fixed to match the 2015 model selectivities in the Gmacs base scenario) stage-1 and stage-2 selectivities for each of the different model scenarios (the stage-3 selectivities are all fixed at 1). Estimated selectivities are shown for the directed pot fishery, the trawl bycatch fishery, the fixed bycatch fishery, the NMFS trawl survey, and the ADF&G pot survey. Two selectivity periods are estimated in the directed pot fishery, from 1978-2008 and 2009-2016.

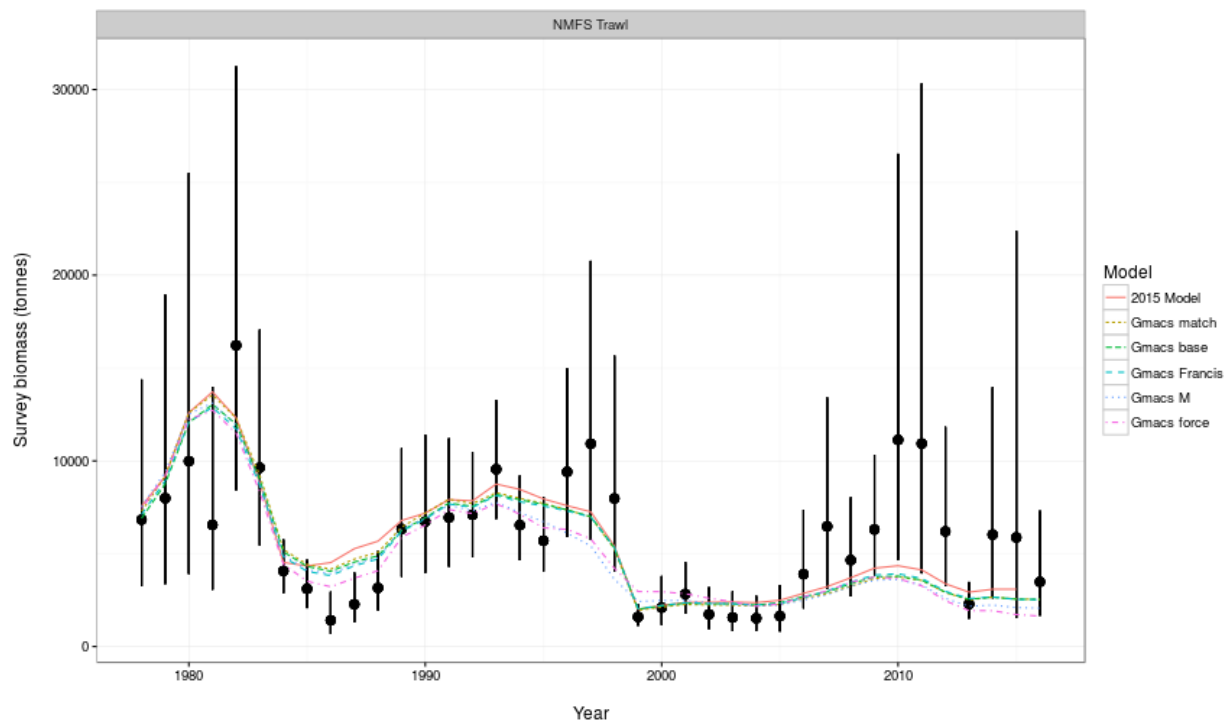


Figure 8: Comparisons of area-swept estimates of total male survey biomass (tonnes) and model predictions for the 2015 model and each of the Gmacs model scenarios. The error bars are plus and minus 2 standard deviations.

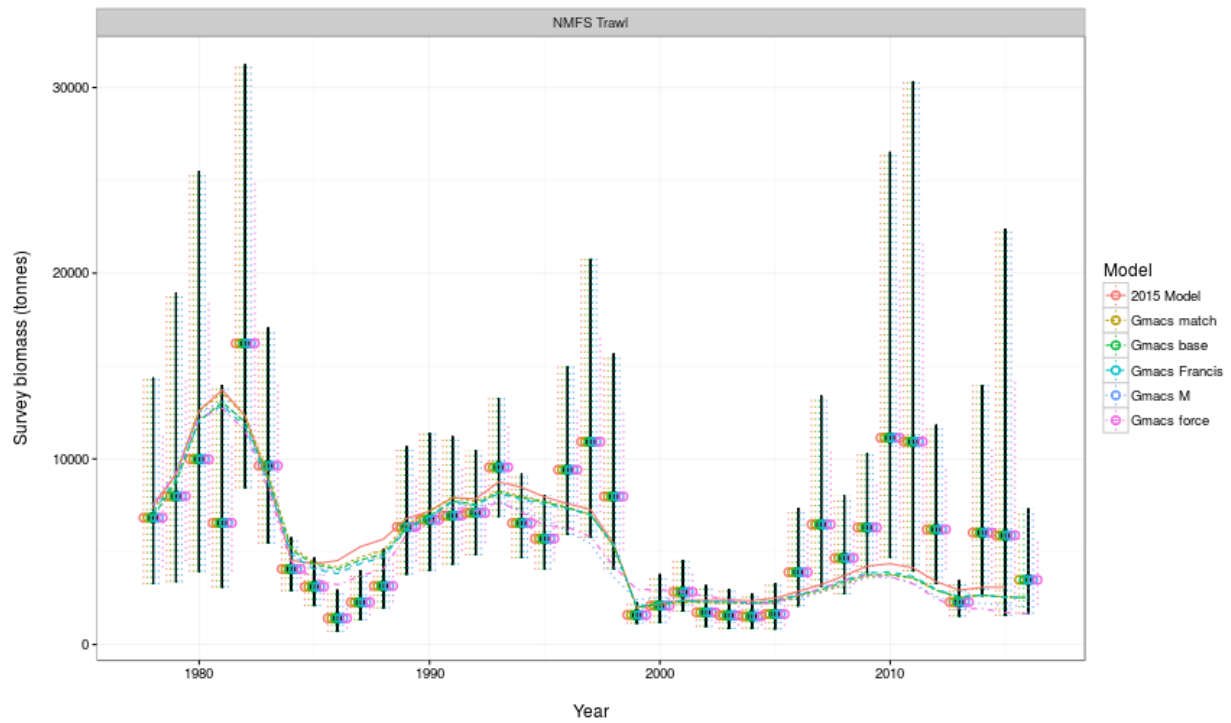


Figure 9: Comparisons of area-swept estimates of total male survey biomass (tonnes) and model predictions for the 2015 model and each of the Gmacs model scenarios. The solid black error bars are plus and minus 2 standard deviations derived using the original survey CVs. The dotted error bars are plus and minus 2 standard deviations but represent the weighted survey CVs.

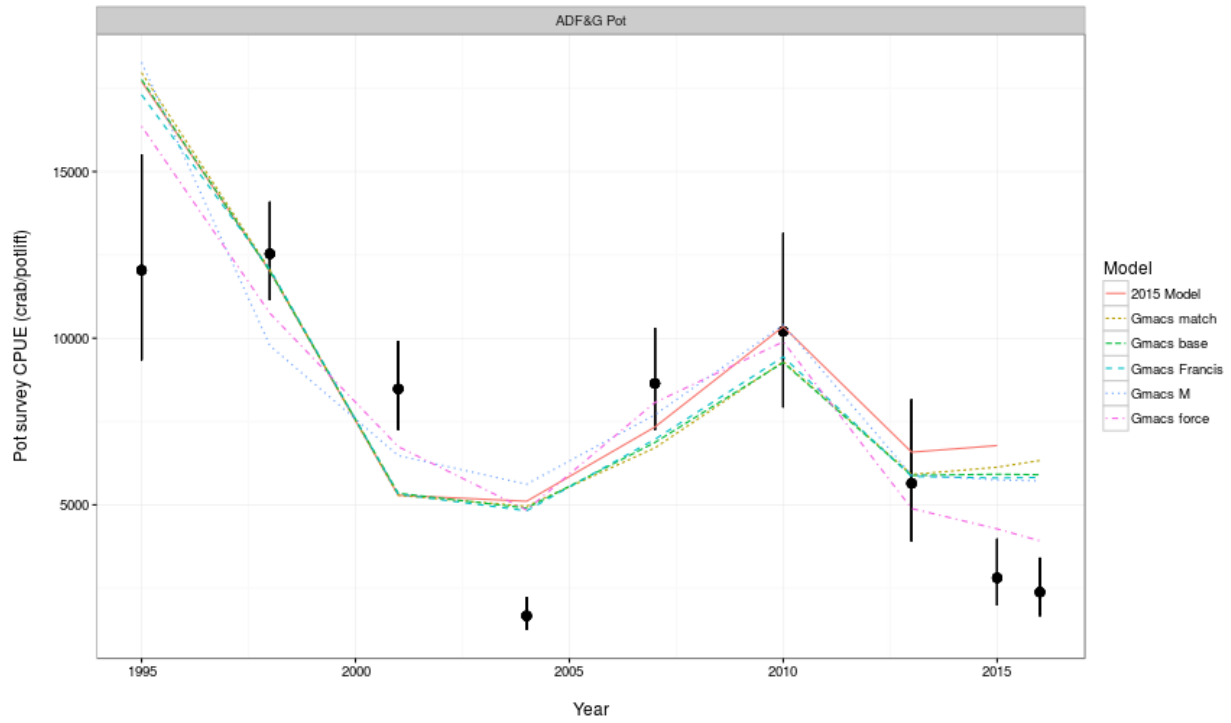


Figure 10: Comparisons of total male pot survey CPUEs and model predictions for the 2015 model and each of the Gmacs model scenarios. The error bars are plus and minus 2 standard deviations.

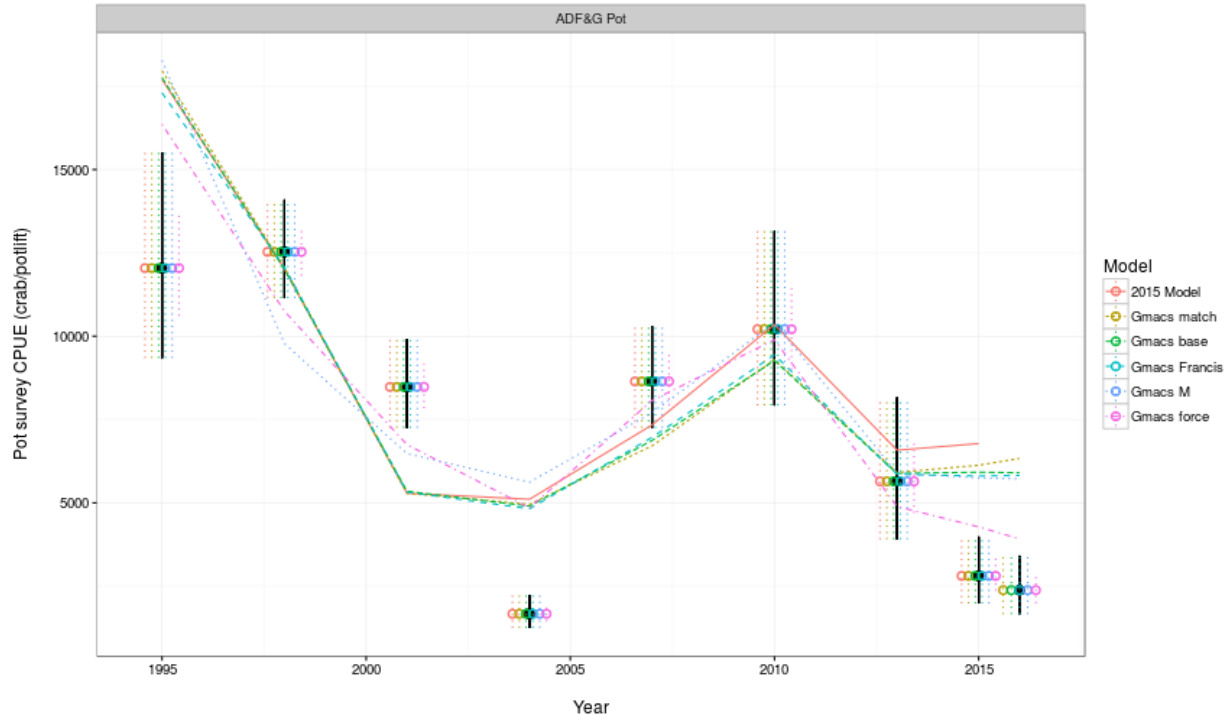


Figure 11: Comparisons of total male pot survey CPUEs and model predictions for the 2015 model and each of the Gmacs model scenarios. The solid black error bars are plus and minus 2 standard deviations derived using the original survey CVs. The dotted error bars are plus and minus 2 standard deviations but represent the weighted survey CVs.

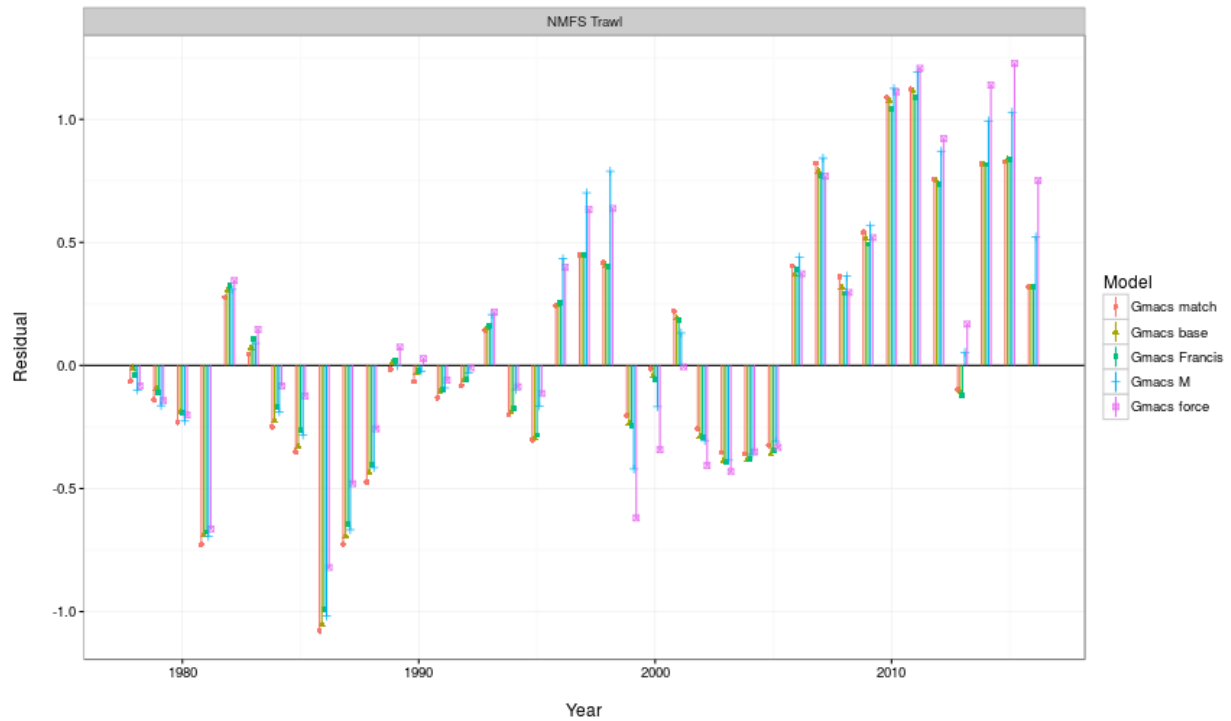


Figure 12: Standardized residuals for area-swept estimates of total male survey biomass for each of the Gmacs model scenarios.

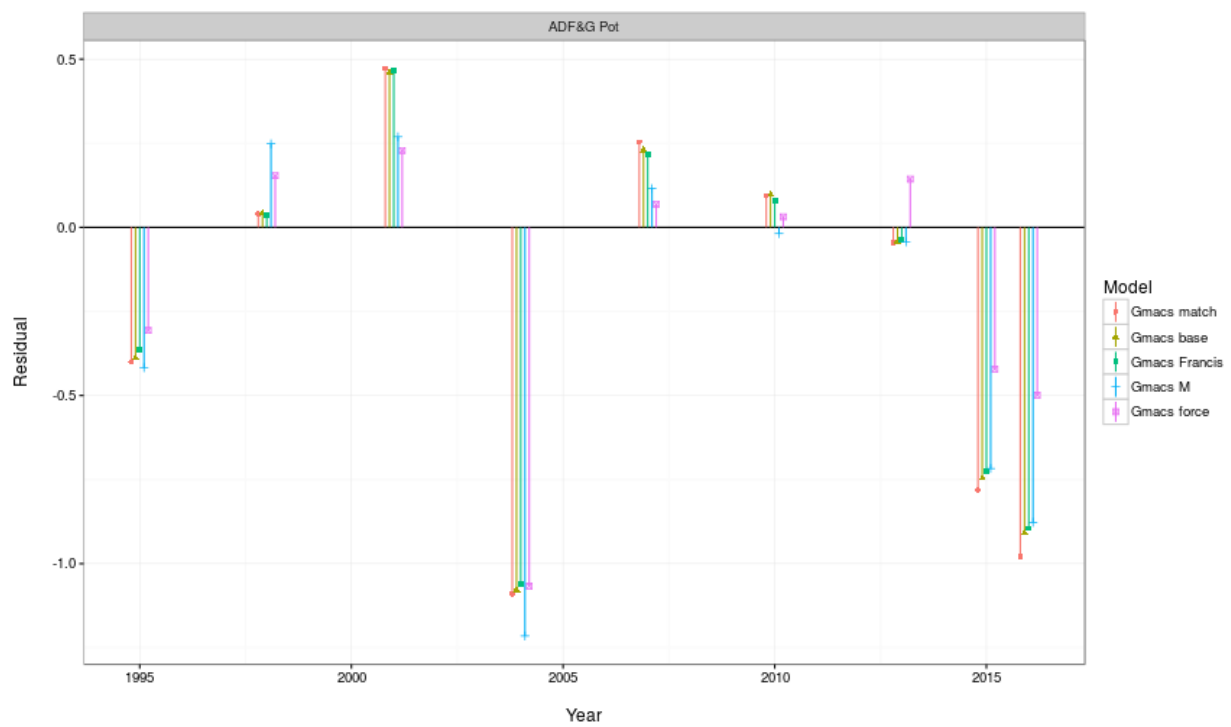


Figure 13: Standardized residuals for total male pot survey CPUEs for each of the Gmacs model scenarios.

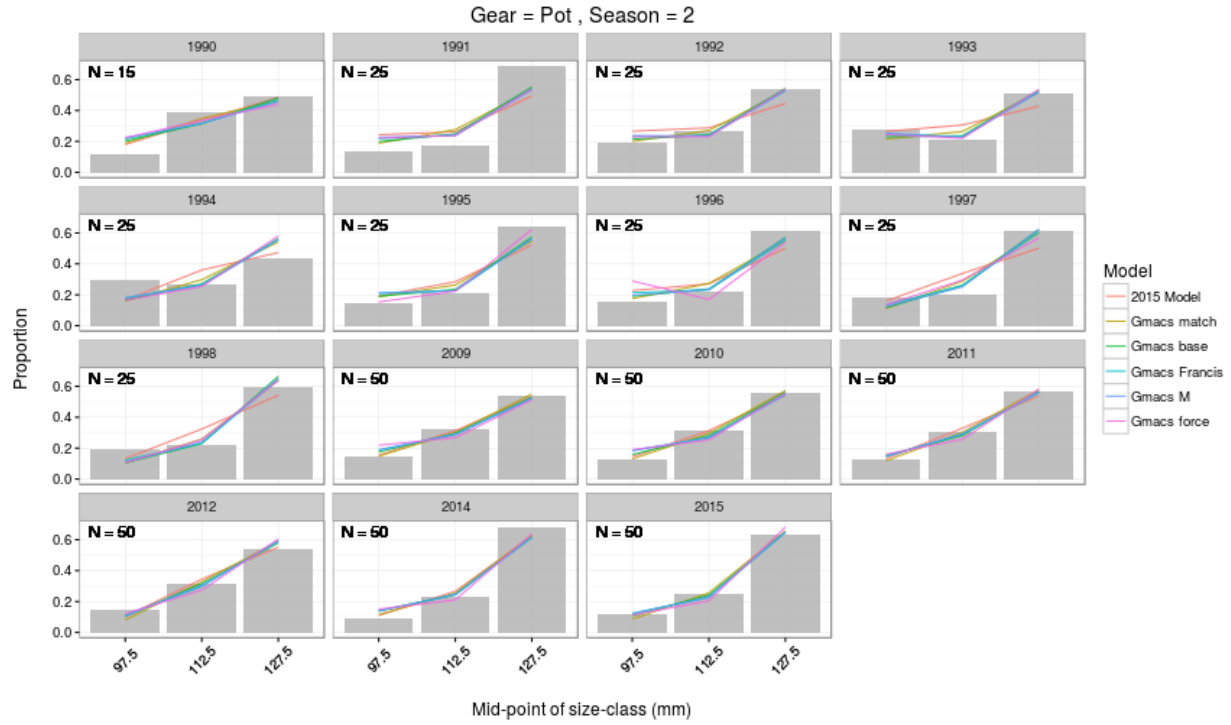


Figure 14: Observed and model estimated size-frequencies of SMBKC by year retained in the directed pot fishery for the 2015 model and each of the Gmacs model scenarios. Note that there is no model estimated size-frequency for the 2015 model during the 2015 year.

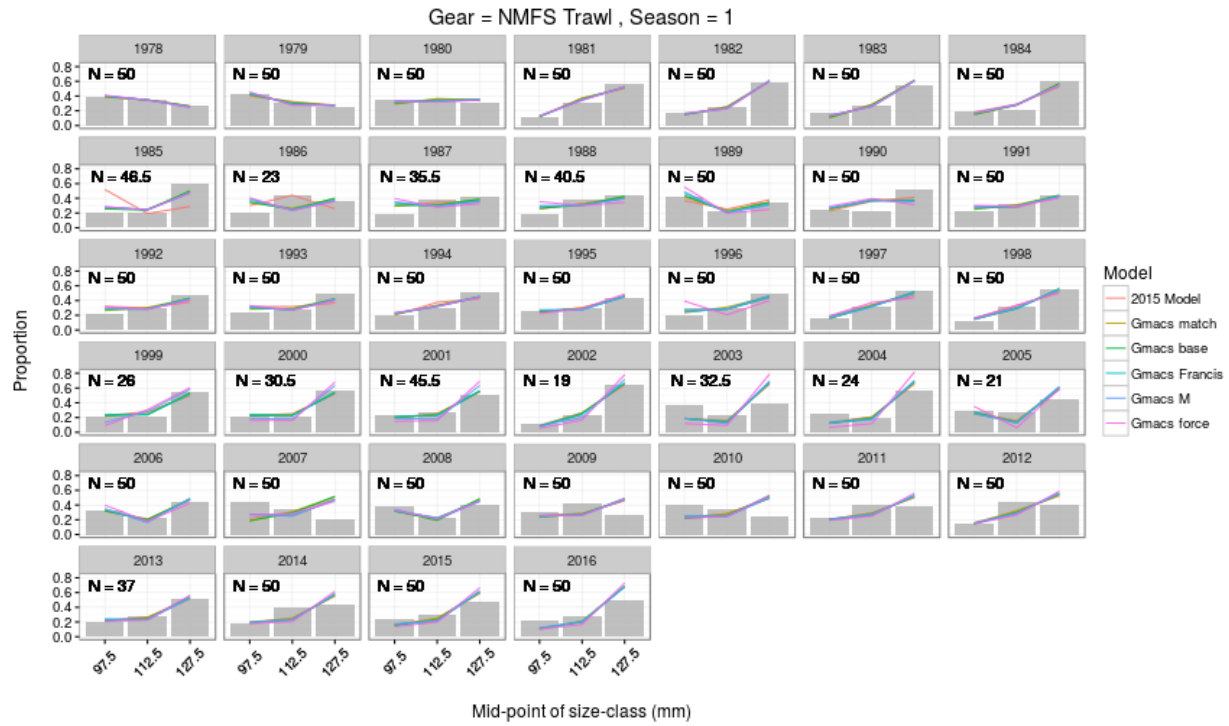


Figure 15: Observed and model estimated size-frequencies of discarded male SMBKC by year in the NMFS trawl survey for the 2015 model and each of the Gmacs model scenarios. Note that there is no model estimated size-frequency for the 2015 model during the 2016 year.

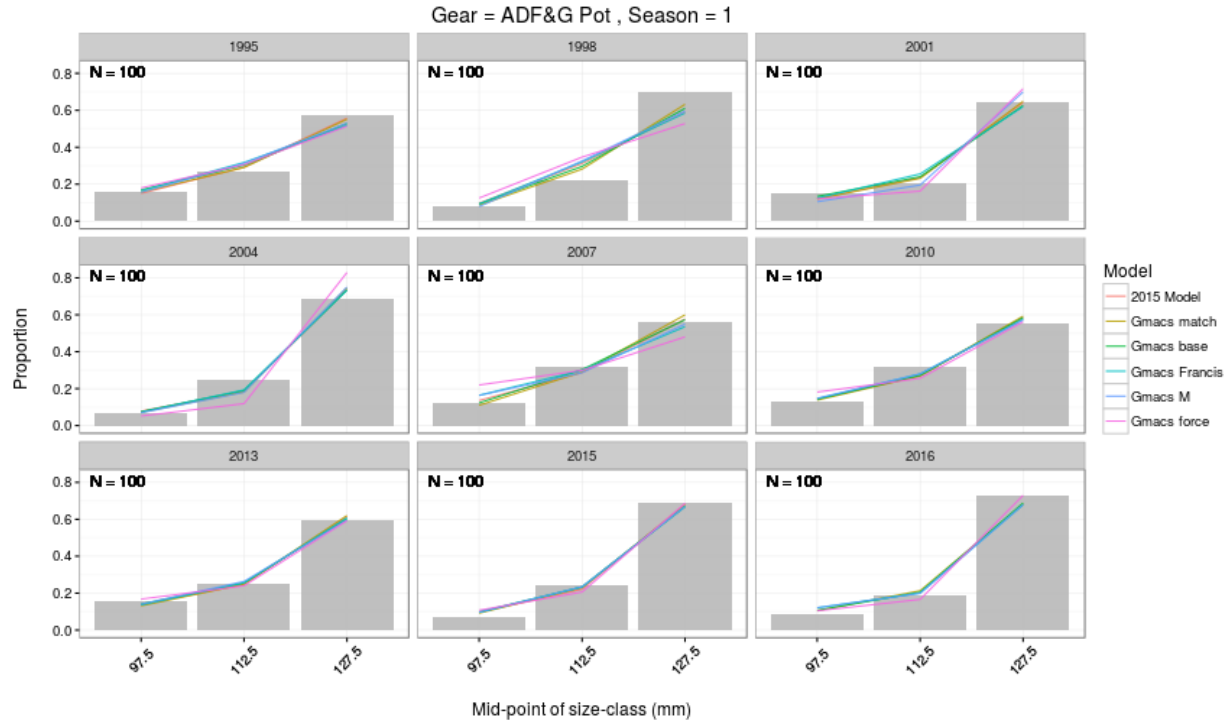


Figure 16: Observed and model estimated size-frequencies of discarded SMBKC by year in the ADF&G pot survey for the 2015 model and each of the Gmacs model scenarios. Note that there is no model estimated size-frequency for the 2015 model during the 2016 year.

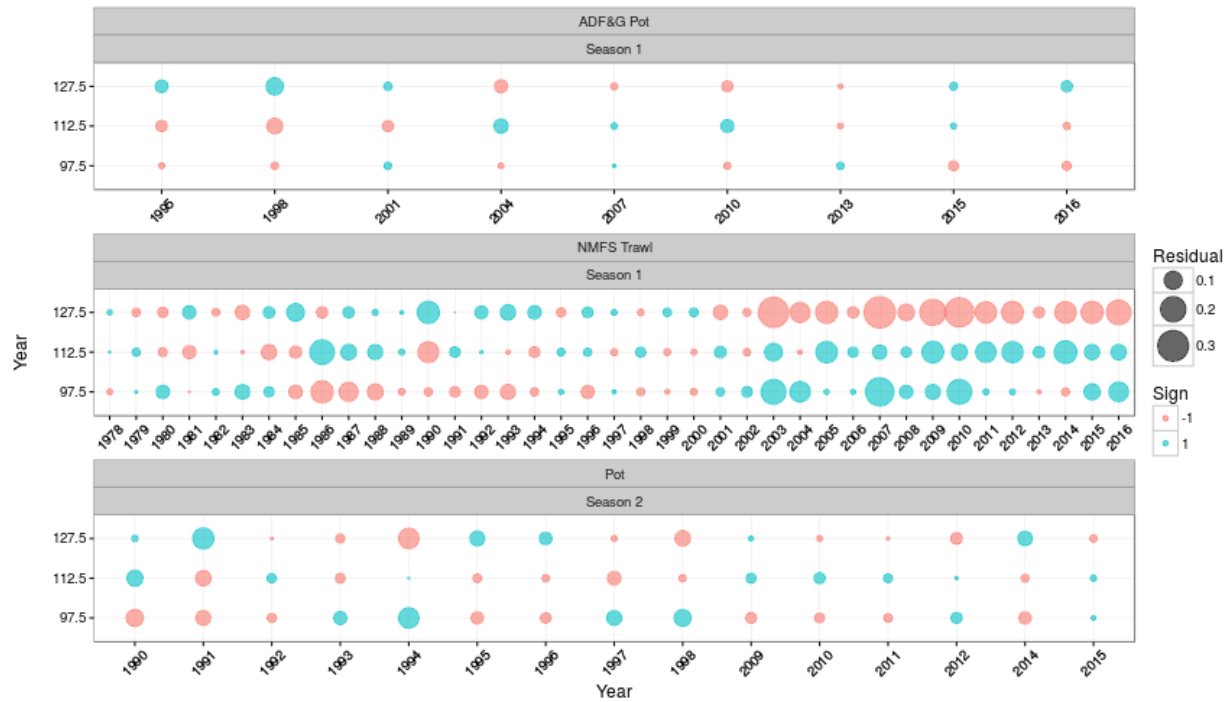


Figure 17: Bubble plots of residuals by stage and year for the directed pot fishery size composition data for St. Mathew Island blue king crab (SMBKC) in the **Gmacs base** model.

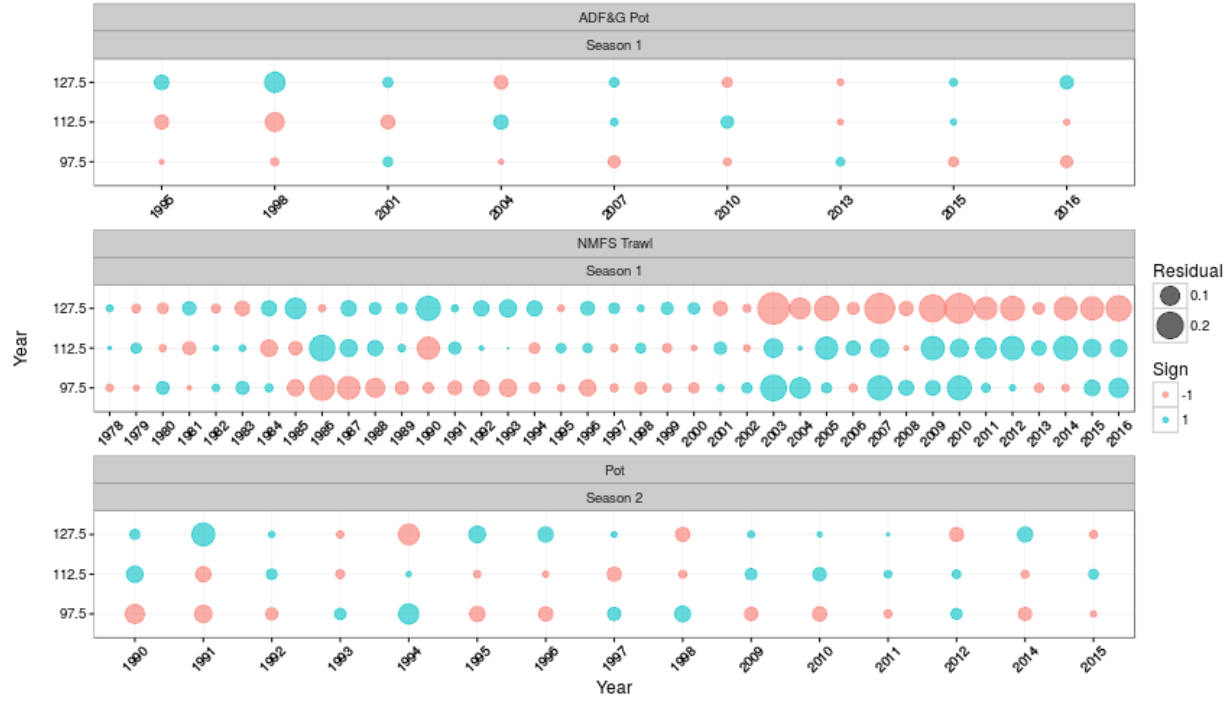


Figure 18: Bubble plots of residuals by stage and year for the NMFS trawl survey size composition data for St. Mathew Island blue king crab (SMBKC) in the **Gmacs Francis** model.

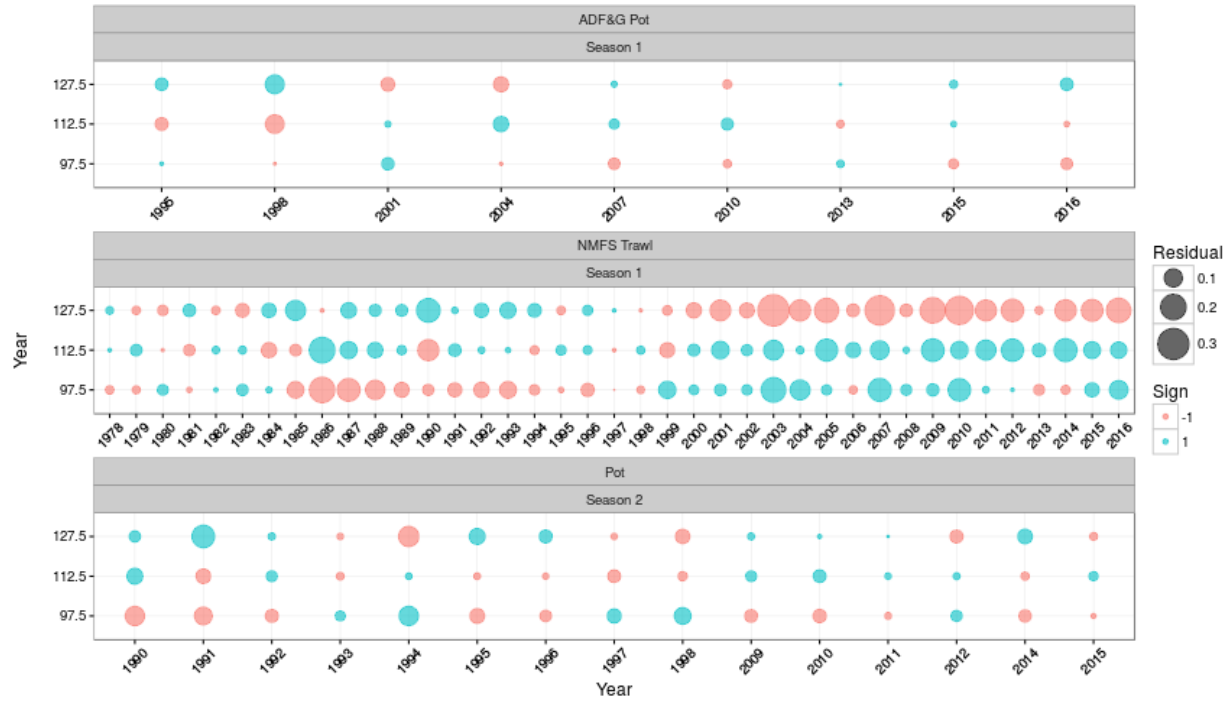


Figure 19: Bubble plots of residuals by stage and year for the NMFS trawl survey size composition data for St. Mathew Island blue king crab (SMBKC) in the **Gmacs M** model.



Figure 20: Bubble plots of residuals by stage and year for the ADF&G pot survey size composition data for St. Mathew Island blue king crab (SMBKC) in the **Gmacs force** model.

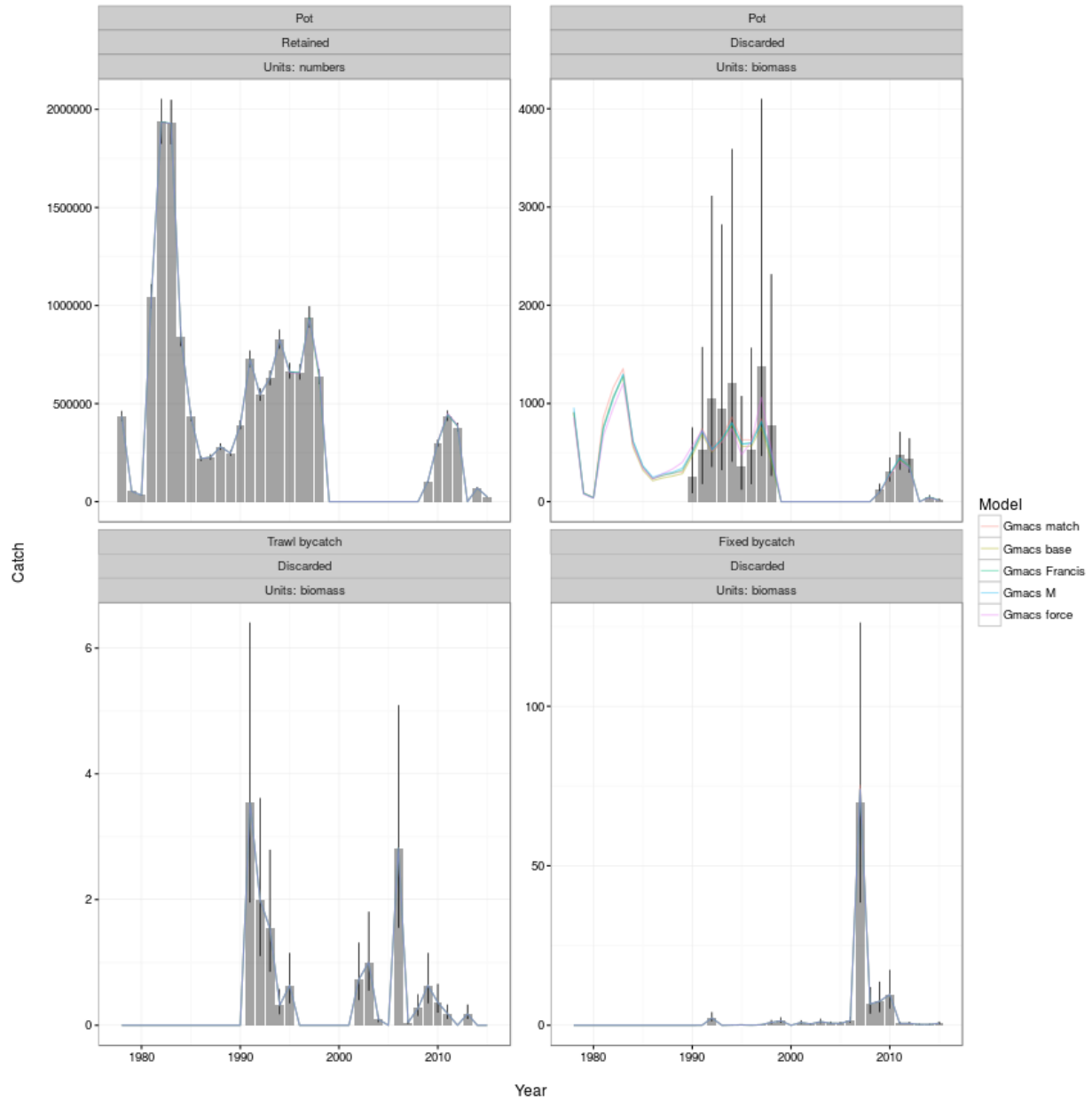


Figure 21: Comparison of observed and model predicted retained catch and bycatches in each of the Gmacs models. Note that difference in units between each of the panels, some panels are expressed in numbers of crab, some as biomass (tonnes).

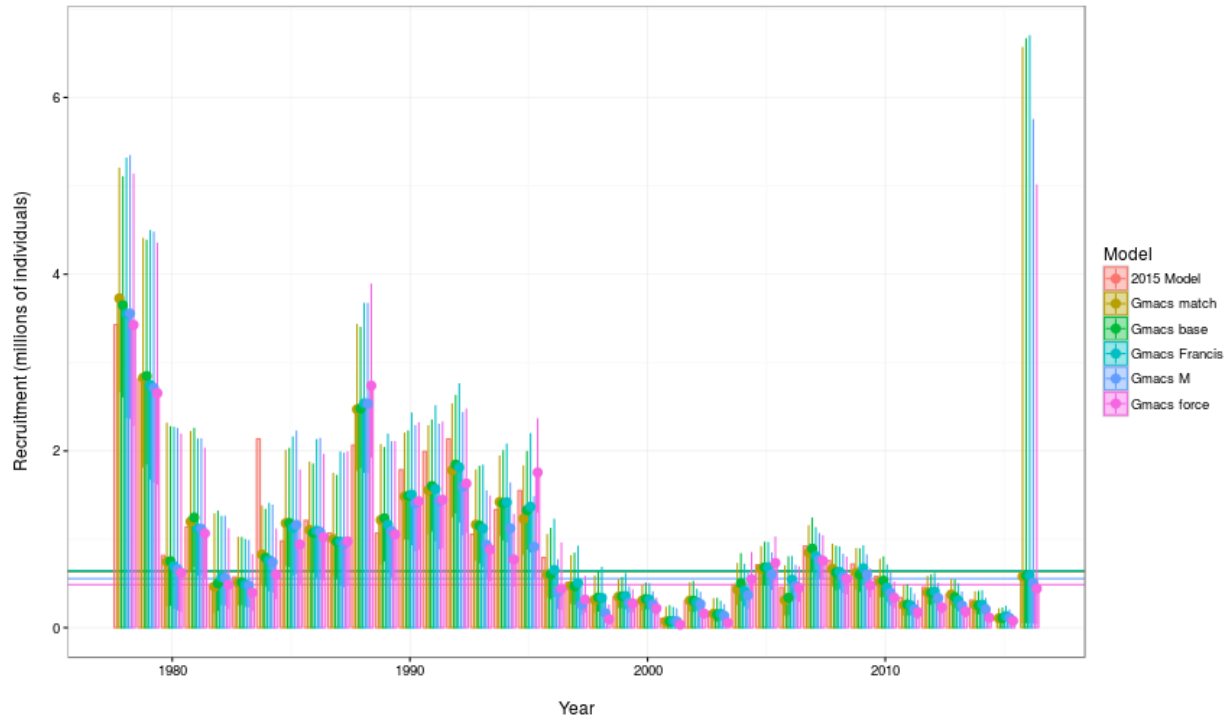


Figure 22: Comparisons of estimated recruitment time series during 1979-2016 in each of the scenarios. The solid horizontal lines in the background represent the estimate of \bar{R} in each model scenario.

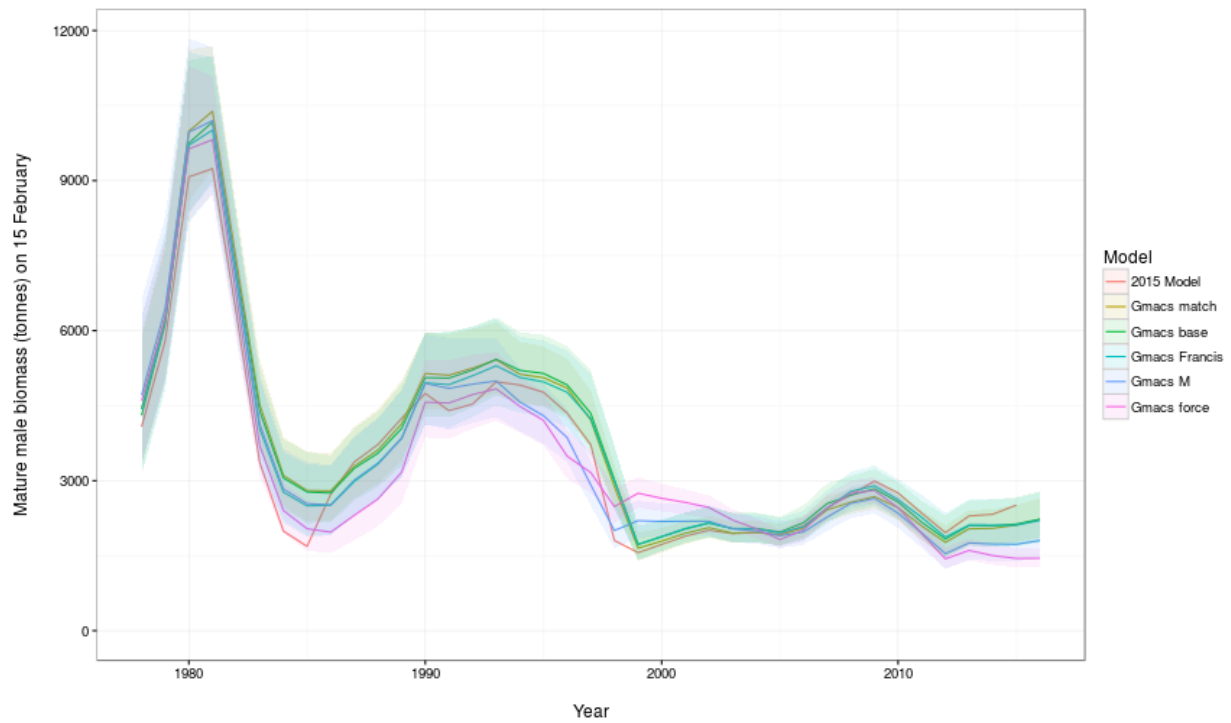


Figure 23: Comparisons of estimated mature male biomass (MMB) time series on 15 February during 1978-2016 for each of the model scenarios.

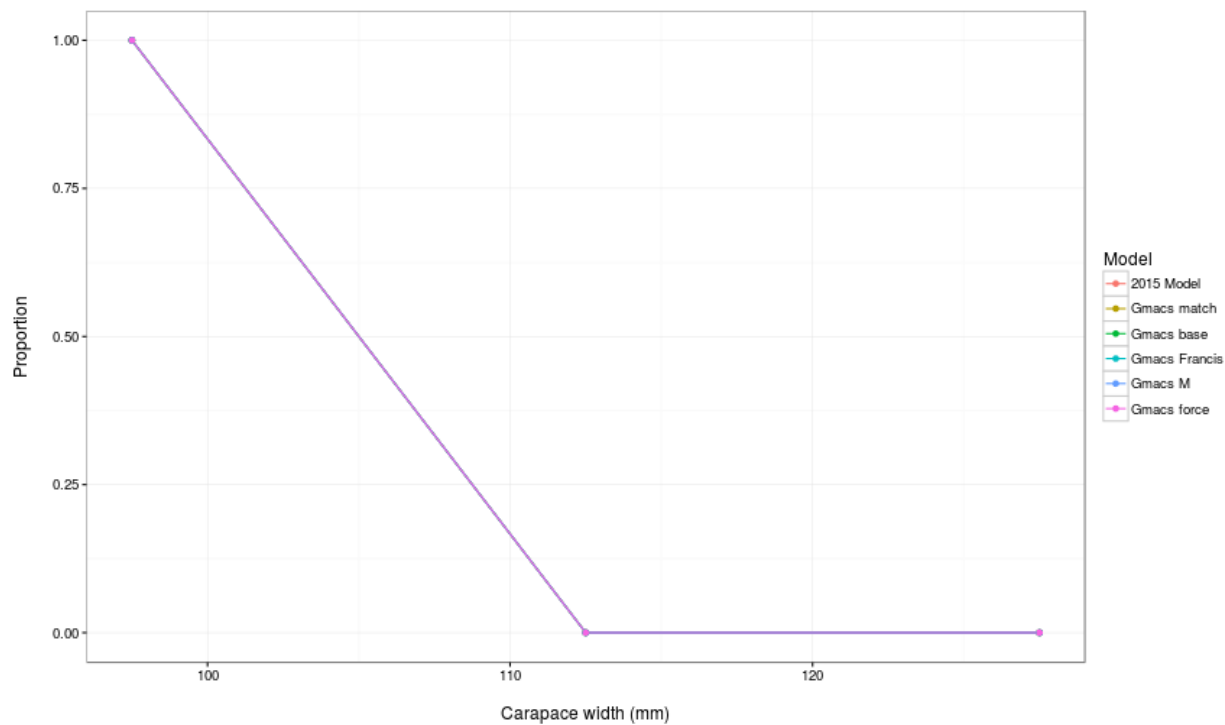


Figure 24: Distribution of carapace width (mm) at recruitment.

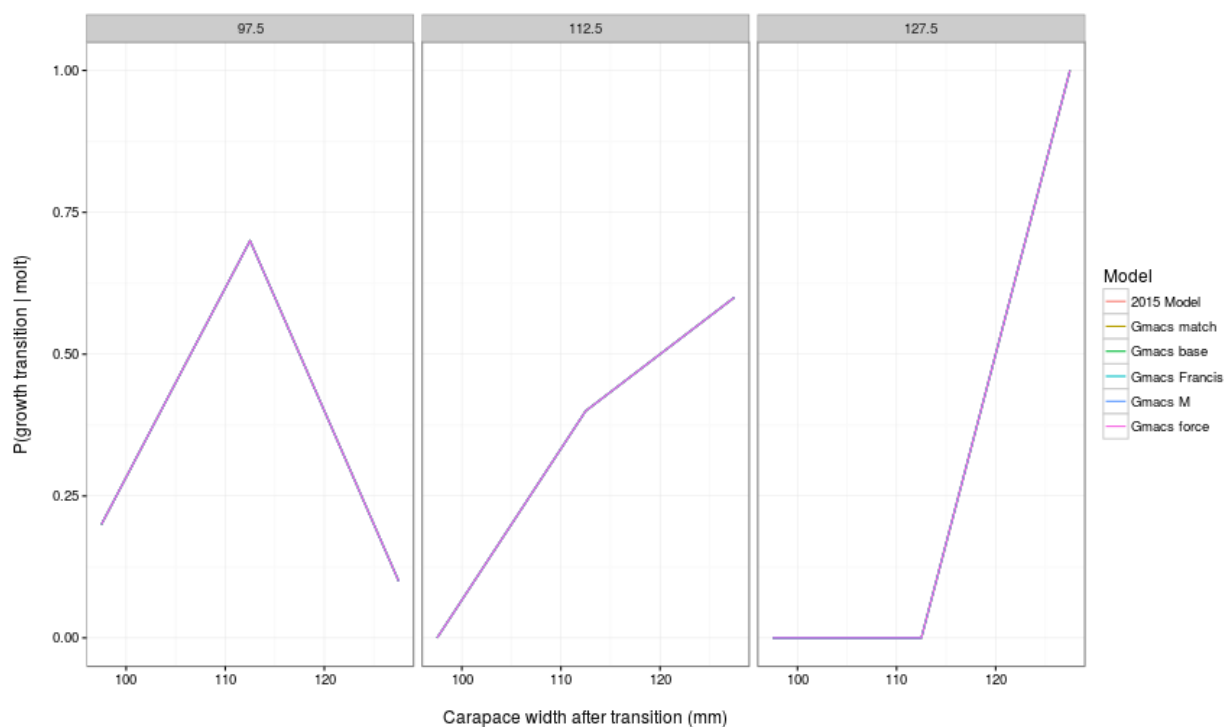


Figure 25: Probability of size transition by stage (i.e. the combination of the growth matrix and molting probabilities). Each of the panels represent the stage before a transition. The x-axes represent the stage after a transition. The size transition matrix was provided as an input directly to Gmacs (as it was during the 2015 SMBKC assessment).

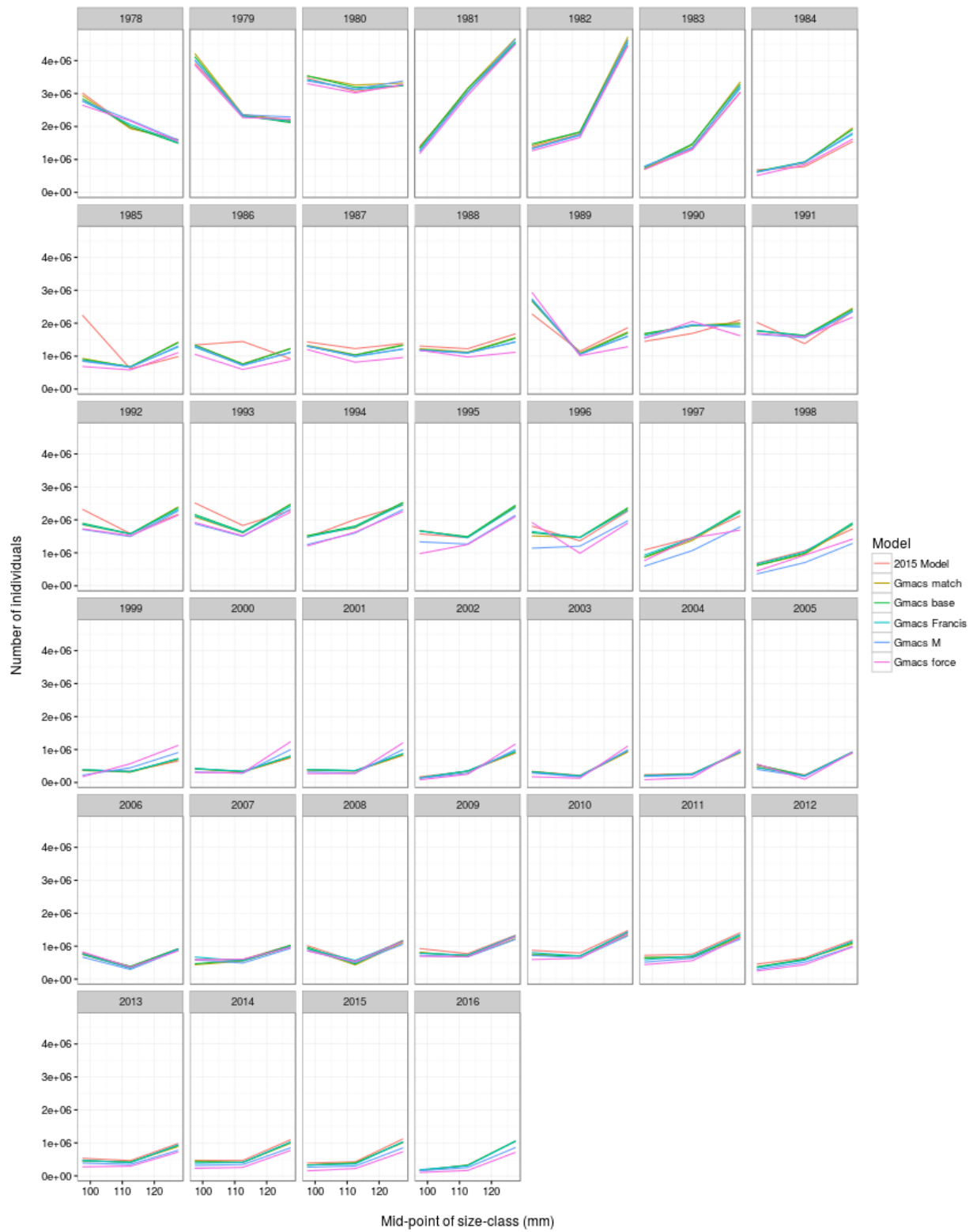


Figure 26: Numbers by stage each year (at the beginning of the model year, i.e. 1 July) in each of the models including the 2015 model.

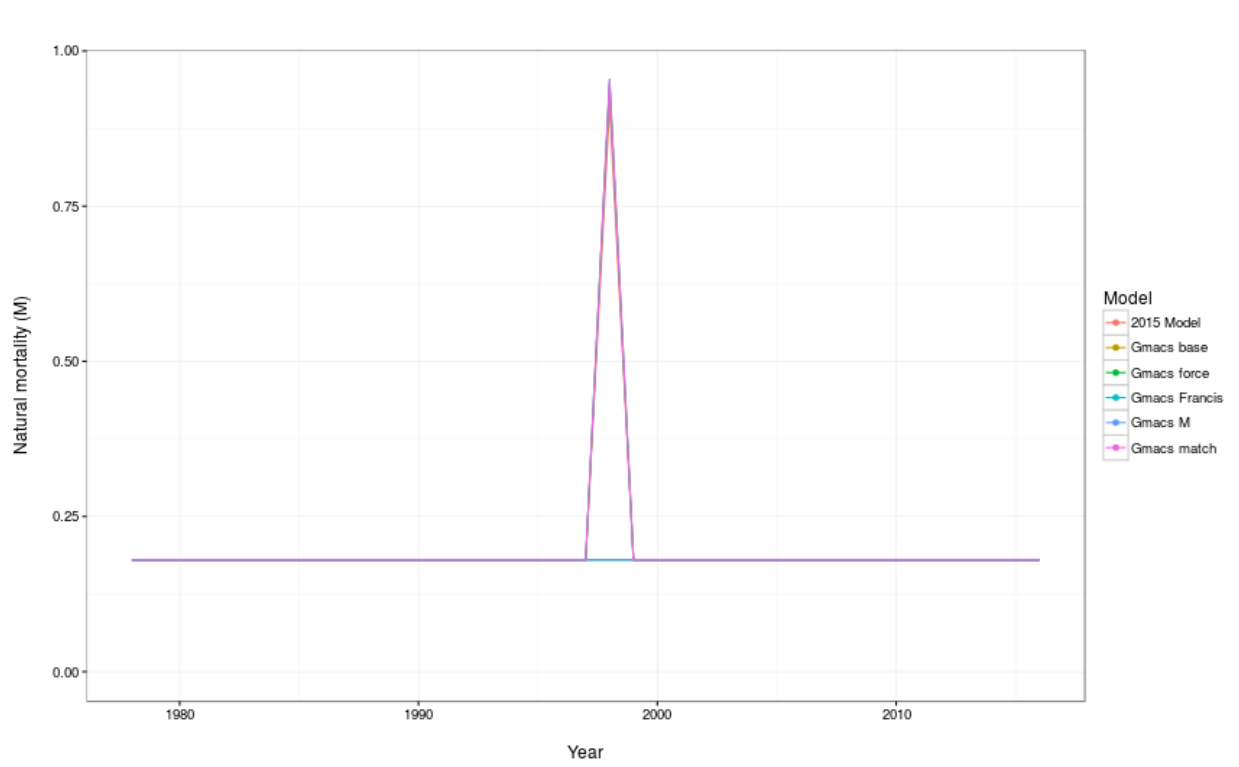


Figure 27: Time-varying natural mortality (M_t). Estimated pulse period occurs in 1998/99 (i.e. M_{1998}).

Appendix A: SMBKC Model Description

1. Introduction

The Gmacs model has been specified to account only for male crab at least 90 mm in carapace length (CL). These are partitioned into three stages (size-classes) determined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) 120+ mm. For management of the St. Matthew Island blue king crab (SMBKC) fishery, 120 mm CL is used as the proxy value for the legal measurement of 5.5 mm in carapace width (CW), whereas 105 mm CL is the management proxy for mature-male size (5 AAC 34.917 (d)). Accordingly, within the model only stage-3 crab are retained in the directed fishery, and stage-2 and stage-3 crab together comprise the collection of mature males. Some justification for the 105 mm value is presented in Pengilly and Schmidt (1995), who used it in developing the current regulatory SMBKC harvest strategy. The term “recruit” here designates recruits to the model, i.e., annual new stage-1 crab, rather than recruits to the fishery. The following description of model structure reflects the Gmacs base model configuration.

2. Model Population Dynamics

Within the model, the beginning of the crab year is assumed contemporaneous with the NMFS trawl survey, nominally assigned a date of 1 July. Although the timing of the fishery is different each year, MMB is measured 15 February, which is the reference date for calculation of federal management biomass quantities. To accommodate this, each model year is split into 5 seasons (t) and a proportion of the natural mortality (τ_t) is applied in each of these seasons where $\sum_{t=1}^{t=5} \tau_t = 1$. Each model year consists of the following processes:

1. Season 1

- Beginning of the SMBKC fishing year (1 July)
- $\tau_1 = 0$
- Surveys

2. Season 2

- τ_2 ranges from 0.05 to 0.44 depending on the time of year the fishery begins each year (i.e. a higher value indicates the fishery begins later in the year; see Table 3)

3. Season 3

- $\tau_3 = 0$
- Fishing mortality applied

4. Season 4

- $\tau_4 = 0.63 - \sum_{i=1}^{i=4} \tau_i$
- Calculate MMB (15 February)

5. Season 5

- $\tau_5 = 0.37$
- Growth and molting
- Recruitment (all to stage-1)

The proportion of natural mortality (τ_t) applied during each season in the model is provided in Table 21. The beginning of the year (1 July) to the date that MMB is measured (15 February) is 63% of the year. Therefore 63% of the natural mortality must be applied before the MMB is calculated. Because the timing of the fishery is different each year τ_2 is different each year and thus τ_4 differs each year.

With boldface lower-case letters indicating vector quantities we designate the vector of stage abundances during season t and year y as

$$\mathbf{n}_{t,y} = n_{l,t,y} = [n_{1,t,y}, n_{2,t,y}, n_{3,t,y}]^\top. \quad (2)$$

Using boldface upper-case letters to indicate a matrix, we describe the size transition matrix \mathbf{G} as

$$\mathbf{G} = \begin{bmatrix} 1 - \pi_{12} - \pi_{13} & \pi_{12} & \pi_{13} \\ 0 & 1 - \pi_{23} & \pi_{23} \\ 0 & 0 & 1 \end{bmatrix}, \quad (3)$$

with π_{jk} equal to the proportion of stage- j crab that molt and grow into stage- k within a season or year. Similarly, the survival matrix $\mathbf{S}_{t,y}$ during season t and year y is

$$\mathbf{S}_{t,y} = \begin{bmatrix} 1 - e^{-Z_{1,t,y}} & 0 & 0 \\ 0 & 1 - e^{-Z_{2,t,y}} & 0 \\ 0 & 0 & 1 - e^{-Z_{3,t,y}} \end{bmatrix}, \quad (4)$$

where $Z_{l,t,y}$ represents the combination of natural mortality $M_{t,y}$ and fishing mortality $F_{t,y}$ during season t and year y

$$\mathbf{Z}_{t,y} = Z_{l,t,y} = M_{t,y} + F_{t,y}. \quad (5)$$

The number of new crab, or recruits, of each stage entering the model each season t and year y is represented as the vector $\mathbf{r}_{t,y}$. The SMBKC formulation of Gmacs specifies recruitment to stage-1 only during season $t = 5$, thus

$$\mathbf{r}_{t,y} = [\bar{R}, 0, 0]^\top \quad \text{for } t = 5, \quad (6)$$

where \bar{R} is the average annual recruitment. The basic population dynamics underlying Gmacs can thus be described as

$$\begin{aligned} \mathbf{n}_{t+1,y} &= \mathbf{S}_{t,y} \mathbf{n}_{t,y}, & \text{if } t < 5 \\ \mathbf{n}_{t,y+1} &= \mathbf{G} \mathbf{S}_{t,y} \mathbf{n}_{t,y} + \mathbf{r}_{t,y}, & \text{if } t = 5 \end{aligned} \quad (7)$$

The natural mortality

$$M_{t,y} = \bar{M} \tau_t + \delta_y^M \quad \text{where } \delta_y^M \sim \mathcal{N}(0, \sigma_M^2) \quad (8)$$

Fishing mortality by year y and season t is denoted $F_{t,y}$ and calculated as

$$F_{t,y} = F_{t,y}^{\text{df}} + F_{t,y}^{\text{tb}} + F_{t,y}^{\text{fb}} \quad (9)$$

where $F_{t,y}^{\text{df}}$ is the fishing mortality associated with the directed fishery, $F_{t,y}^{\text{tb}}$ is the fishing mortality associated with the trawl bycatch fishery, $F_{t,y}^{\text{fb}}$ is the fishing mortality associated with the fixed bycatch fishery. Each of these are derived as

$$\begin{aligned}
F_{t,y}^{exdf} &= \bar{F}^{exdf} + \delta^{exdf}_{t,y}, \\
F_{t,y}^{extb} &= \bar{F}^{extb} + \delta^{extb}_{t,y}, \\
F_{t,y}^{exfb} &= \bar{F}^{exfb} + \delta^{exfb}_{t,y},
\end{aligned} \tag{10}$$

where $\delta^{exdf}_{t,y}$, $\delta^{extb}_{t,y}$, and $\delta^{exfb}_{t,y}$ are the fishing mortality deviations for each of the fisheries, each season t during each year y , \bar{F}^{exdf} , \bar{F}^{extb} , and \bar{F}^{exfb} are the average fishing mortalities for each fishery. Also add selectivity, retention, discard mortality in here. Also catch, CPUE.

3. Model Data

Data inputs used in model estimation are listed in Table 22.

4. Model Parameters

Table 23 lists fixed (externally determined) parameters used in model computations. In all scenarios, the stage-transition matrix is

$$\mathbf{G} = \begin{bmatrix} 0.2 & 0.7 & 0.1 \\ 0 & 0.4 & 0.6 \\ 0 & 0 & 1 \end{bmatrix} \tag{11}$$

which is the combination of the growth matrix and molting probabilities.

Estimated parameters are listed in Table 24 and include an estimated parameter for natural mortality (M) in 1998/99 assuming an anomalous mortality event in that year, as hypothesized by Zheng and Kruse (2002), with natural mortality otherwise fixed at 0.18 yr^{-1} .

$$\theta = \{\bar{R}, \mathbf{n}_0, q_{\text{pot}}, cv, \delta_{1998}^M, s_{1,l=1}^{\text{pot}}, s_{1,l=2}^{\text{pot}}, s_{2,l=1}^{\text{pot}}, s_{2,l=2}^{\text{pot}}, s_{l=1}^{\text{NMFS}}, s_{l=2}^{\text{NMFS}}, s_{l=1}^{\text{ADFG}}, s_{l=2}^{\text{ADFG}}\}$$

Also Fs

In any year with no directed fishery, and hence zero retained catch, F_t^{df} is set to zero rather than model estimated. Similarly, for years in which no groundfish bycatch data are available, F_t^{gf} and F_t^{gt} are imputed to be the geometric means of the estimates from years for which there are data.

Both surveys are assigned a nominal date of 1 July, the start of the crab year.

5. Model Objective Function and Weighting Scheme

The objective function consists of a sum of eight “negative log-likelihood” terms characterizing the hypothesized error structure of the principal data inputs with respect to their true, i.e., model-predicted, values and four “penalty” terms associated with year-to-year variation in model recruit abundance and fishing mortality in the directed fishery and groundfish trawl and fixed-gear fisheries (Table 25). See Table 6, where upper and lower case letters designate model-predicted and data-computed quantities, respectively, and boldface letters again indicate vector quantities. Sample sizes n_t (observed number of male SMBKC $\leq 90 \text{ mm CL}$) and estimated coefficients of variation \hat{cv}_t were used to develop appropriate variances for stage-proportion and abundance-index components. The weights λ_j appearing in the objective function component expressions in Table 6 play the role of “tuning” parameters in the modeling procedure.

$$\sigma_i = \frac{1}{\lambda} \sqrt{\log(1.0 + c_i^2)} \quad (12)$$

$$\delta_i = \frac{\log(obs_i/pred_i)}{\sigma_i} + 0.5\sigma_i \quad (13)$$

The standard deviation of the normalized (or standardized) residuals (SDNR) is calculated as

$$SDNR = \sqrt{\frac{1}{n} \sum_{i=1}^n (\delta_i - \bar{\delta})^2} \quad (14)$$

For an abundance data set to be well fitted, the SDNR should not be much greater than 1 (a value much less than 1, which means that the data set is fitted better than was expected, is not a cause for concern). What is meant by “much greater than 1” depends on m (the number of years in the data set). Francis (2011) suggests upper limits of 1.54, 1.37, and 1.26 for $m = 5, 10$, and 20 , respectively. Although an SDNR not much greater than 1 is a necessary condition for a good fit, it is not sufficient. It is important to plot the observed and expected abundances to ensure that the fit is good.

MAR, Francis weighting (Francis 2011).

6. Estimation

The model was implemented using the software AD Model Builder (Fournier et al. 2012), with parameter estimation by minimization of the model objective function using automatic differentiation. Parameter estimates and standard deviations provided in this document are AD Model Builder reported values assuming maximum likelihood theory asymptotics.

Table 21: Proportion of the natural mortality (τ_t) that is applied during each season (t) in the model.

| Year | Season 1 | Season 2 | Season 3 | Season 4 | Season 5 |
|------|----------|----------|----------|----------|----------|
| 1978 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1979 | 0.00 | 0.06 | 0.00 | 0.57 | 0.37 |
| 1980 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1981 | 0.00 | 0.05 | 0.00 | 0.58 | 0.37 |
| 1982 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1983 | 0.00 | 0.12 | 0.00 | 0.51 | 0.37 |
| 1984 | 0.00 | 0.10 | 0.00 | 0.53 | 0.37 |
| 1985 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1986 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1987 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1988 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1989 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1990 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1991 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1992 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1993 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1994 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1995 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1996 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1997 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1998 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1999 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2000 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2001 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2002 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2003 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2004 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2005 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2006 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2007 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2008 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2009 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2010 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2011 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2012 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2013 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2014 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2015 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2016 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |

Table 22: Data inputs used in model estimation.

| Data | Years | Source |
|---|--|---|
| Directed pot-fishery retained-catch number (not biomass) | 1978/79 - 1998/99 2009/10 - 2015/16 | Fish tickets (fishery closed 1999/00 - 2008/09) |
| Groundfish trawl bycatch biomass | 1992/93 - 2015/16 | NMFS groundfish observer program |
| Groundfish fixed-gear bycatch biomass | 1992/93 - 2015/16 | NMFS groundfish observer program |
| NMFS trawl-survey biomass index (area-swept estimate) and CV | 1978-2016 | NMFS EBS trawl survey |
| ADF&G pot-survey abundance index (CPUE) and CV | Triennial 1995-2016 | ADF&G SMBKC pot survey |
| NMFS trawl-survey stage proportions and total number of measured crab | 1978-2016 | NMFS EBS trawl survey |
| ADF&G pot-survey stage proportions and total number of measured crab | Triennial 1995-2016 | ADF&G SMBKC pot survey |
| Directed pot-fishery stage proportions and total number of measured crab | 1990/91 - 1998/99 2009/10 - 2015/16 | ADF&G crab observer program (fishery closed 1999/00 - 2008/09) |

Table 23: Fixed model parameters for all scenarios.

| Parameter | Symbol | Value | Source/rationale |
|---|--------------|------------------------|---|
| Trawl-survey catchability | q | 1.0 | Default |
| Natural mortality | M | 0.18 yr^{-1} | NPFMC (2007) |
| Size transition matrix | \mathbf{G} | Equation 11 | Otto and Cummiskey (1990) |
| Stage-1 and stage-2 mean weights | w_1, w_2 | 0.7, 1.2 kg | Length-weight equation (B. Foy, NMFS) |
| Stage-3 mean weight | $w_{3,y}$ | Depends on year | applied to stage midpoints Fishery reported average retained weight from fish tickets, or its average, and mean weights of legal males |
| Recruitment SD | σ_R | 1.2 | High value |
| Natural mortality SD | σ_M | 10.0 | High value (basically free parameter) |
| Directed fishery handling mortality | | 0.2 | 2010 Crab SAFE |
| Groundfish trawl handling mortality | | 0.8 | 2010 Crab SAFE |
| Groundfish fixed-gear handling mortality | | 0.5 | 2010 Crab SAFE |

Table 24: The lower bound (LB), upper bound (UB), initial value, prior, and estimation phase for each estimated model parameter.

| Parameter | LB | Initial value | UB | Prior | Phase |
|---|----|---------------|----|---|-------|
| Average recruitment $\log(\bar{R})$ | -7 | 10.0 | 20 | Uniform(-7,20) | 1 |
| Stage-1 initial numbers $\log(N_1)$ | 5 | 14.5 | 20 | Uniform(5,20) | 1 |
| Stage-2 initial numbers $\log(N_2)$ | 5 | 14.0 | 20 | Uniform(5,20) | 1 |
| Stage-3 initial numbers $\log(N_3)$ | 5 | 13.5 | 20 | Uniform(5,20) | 1 |
| ADF&G pot survey catchability q | 0 | 4.0 | 5 | Uniform(0,5) | 1 |
| Stage-1 directed fishery selectivity 1978-2008 | 0 | 0.4 | 1 | Uniform(0,1) | 3 |
| Stage-2 directed fishery selectivity 1978-2008 | 0 | 0.7 | 1 | Uniform(0,1) | 3 |
| Stage-1 directed fishery selectivity 2009-2015 | 0 | 0.4 | 1 | Uniform(0,1) | 3 |
| Stage-2 directed fishery selectivity 2009-2015 | 0 | 0.7 | 1 | Uniform(0,1) | 3 |
| Stage-1 NMFS trawl survey selectivity | 0 | 0.4 | 1 | Uniform(0,1) | 4 |
| Stage-2 NMFS trawl survey selectivity | 0 | 0.7 | 1 | Uniform(0,1) | 4 |
| Stage-1 ADF&G pot survey selectivity | 0 | 0.4 | 1 | Uniform(0,1) | 4 |
| Stage-2 ADF&G pot survey selectivity | 0 | 0.7 | 1 | Uniform(0,1) | 4 |
| Natural mortality deviation during 1998 δ_{1998}^M | -3 | 0.0 | 3 | Normal($\delta_{1997}^M, \sigma_M^2$) | 4 |
| Recruitment deviations δ_y^R | -7 | 0.0 | 7 | Normal(0, σ_R^2) | 3 |

Table 25: Log-likelihood and penalty components of base-model objective function. The λ_k are weights, described in text; the neff t are effective sample sizes, also described in text. All summations are with respect to years over each data series.

| Component | Distribution | Form |
|------------------------------|--------------|--|
| Legal retained-catch biomass | Lognormal | $-0.5 \sum (\log(c_t/C_t)^2 / \log(1 + cv_c^2))$ |
| Dis. Pot bycatch biomass | Lognormal | $-0.5 \sum (\log(c_t/C_t)^2 / \log(1 + cv_c^2))$ |