

Overview of Ozone (O_3) Air Quality in the United States

Updated: June 29, 2023

1. Introduction

The overall purpose of this document is to maintain an up-to-date graphical summary of air quality information that supports the review of the National Ambient Air Quality Standards (NAAQS) for ozone (O_3). In previous reviews of the O_3 NAAQS, this type of information has generally been included in atmospheric sections of the Integrated Science Assessment (ISA) and Policy Assessment (PA) for O_3 and related photochemical oxidants. This stand-alone document will either replace or complement the air quality emissions and monitoring data in the atmospheric sections of future O_3 NAAQS review supporting documents and will be updated at regular intervals as new data becomes available.

The content of past NAAQS documents' atmospheric sections has included major sections on emissions and concentration trends utilizing maps and data from the Environmental Protection Agency's (EPA's) National Emissions Inventory (NEI) and the EPA's Air Quality System (AQS) database. In past NAAQS reviews, this often involved adaptation of figures and tables prepared for other reports or development of new figures and tables using data analysis and mapping software. Additionally, the release of updated emission inventories and ambient air monitoring data may not coincide with the schedule for the development of NAAQS review supporting documents. As a result, data access and resources can limit the availability of the most recent information for inclusion in NAAQS review supporting documents.

This stand-alone document allows the content to be updated as soon as new data becomes available, rather than relying on information that is available at the time of development of the NAAQS review supporting documents. It also ensures that the public will have access to a consistent set of maps and figures for each NAAQS pollutant that are updated on a routine basis, rather than separated by several years because of the disparate schedules of the various NAAQS reviews for each pollutant. Moreover, a stand-alone document can be expanded to include new air quality analyses as they are completed, rather than following the timeline for the public release of the NAAQS review supporting documents. Finally, this document takes advantage of a more flexible digital format for the routinely prepared maps and trends figures with an end product that more strongly emphasizes visual presentation of data and reduces the amount of text, while also creating a more interactive presentation of the information through the use of external links.

This document follows an organizational structure similar to that of the atmospheric sections of past O_3 NAAQS review supporting documents. The subsequent sections are as follows: 2. Atmospheric Chemistry; 3. Sources and Emissions of O_3 Precursors; 4. Ambient Air Monitoring Requirements and Monitoring Networks; 5. Data Handling Conventions and Computations for Determining Whether the Standards are Met; and 6. O_3 Concentrations Measured at Ambient Air Monitoring Sites Across the U.S. These sections are broad enough in scope to communicate relevant information about O_3 air quality, including scientific advances, but specific enough that the information needed to develop NAAQS review supporting documents can be quickly and readily retrieved.

2. Atmospheric Chemistry

O_3 is one of a group of photochemical oxidants formed in the troposphere¹ by photochemical reactions of precursor gases in the presence of sunlight and is generally not directly emitted from specific sources.² Tropospheric O_3 and other oxidants, such as peroxyacetyl nitrate (PAN) and hydrogen peroxide, form in polluted areas by atmospheric reactions involving two main classes of precursor pollutants: volatile organic compounds (VOCs) and nitrogen oxides (NO_x). This occurs especially during the summer, as a result of the photolysis of primary pollutants such as nitrogen dioxide (NO_2). The reaction is disrupted by the presence of VOCs, the radical that results from methane (CH_4) oxidation; or a reaction between carbon monoxide (CO) and the hydroxyl radical (OH) in the atmosphere. Thus, the substances NO_x , VOC, CH_4 and CO are considered to be the primary precursors of tropospheric O_3 . The formation of O_3 , other oxidants, and oxidation products from these precursors is a complex, nonlinear function of many factors including (1) the intensity and spectral distribution

¹ O_3 also occurs in the stratosphere, where it serves the beneficial role of absorbing the sun's harmful ultraviolet radiation and preventing the majority of this radiation from reaching the Earth's surface.

²The only other appreciable source of O_3 to the troposphere is transport from the stratosphere.

of sunlight; (2) atmospheric mixing; (3) concentrations of precursors in the ambient air and the rates of chemical reactions of these precursors; and (4) processing on cloud and aerosol particles.

O₃ is present not only in polluted urban atmospheres, but throughout the troposphere, even in remote areas of the globe. The same basic processes involving photochemical reactions of NO_x, VOCs, and CO contribute to O₃ formation throughout the troposphere. These processes also lead to the formation of other photochemical products, such as PAN, HNO₃, and H₂SO₄, and to other gaseous compounds, such as HCHO and other carbonyl compounds, as well as a number of particulate compounds.

Prior to 1979, the indicator for the NAAQS for photochemical oxidants was total photochemical oxidants. Early ambient air monitoring indicated similarities between O₃ measurements and the photochemical oxidant measurements, as well as reduced precision and accuracy of the latter. To address these issues, the EPA established O₃ as the indicator for the NAAQS for photochemical oxidants in 1979, and it is currently the only photochemical oxidant other than NO₂ that is routinely monitored in a national ambient air monitoring network.

Rather than varying directly with emissions of its precursors, O₃ changes in a nonlinear fashion with the concentrations of its precursors. Emissions of NO_x lead to both the formation and destruction of O₃, depending on the local quantities of NO_x, VOCs, radicals, and sunlight. O₃ chemistry is often described in terms of which precursors most directly impact formation rates. A NO_x-limited regime indicates that O₃ concentrations will decrease in response to decreases in ambient NO_x concentrations and vice-versa. These conditions tend to occur when NO_x concentrations are generally low compared to VOC concentrations and during warm, sunny conditions when NO_x photochemistry is relatively fast. NO_x-limited conditions are more common during daylight hours, in the summertime, in suburban and rural areas, and in portions of the country with high biogenic VOC emissions like the Southeast. In contrast, NO_x-saturated conditions (also referred to as VOC-limited or radical-limited) indicate that O₃ will increase as a result of NO_x reductions but will decrease as a result of VOC reductions. NO_x-saturated conditions occur at times and locations with lower levels of available sunlight, resulting in slower photochemical formation of O₃, and when NO_x concentrations are in excess compared to VOC concentrations. NO_x-saturated conditions are more common during nighttime hours, in the wintertime, and in densely populated urban areas or industrial plumes. These varied relationships between precursor emissions and O₃ chemistry result in localized areas in which O₃ concentrations are suppressed compared to surrounding areas, but which contain NO₂ that contributes to subsequent O₃ formation further downwind. Consequently, O₃ response to reductions in NO_x emissions is complex and may include decreases in O₃ concentrations at some times and locations and increases in O₃ concentrations at other times and locations. Over the past decade, there have been substantial decreases in NO_x emissions in the U.S. and many locations have transitioned from NO_x-saturated to NO_x-limited during times of year that are conducive to O₃ formation (generally summer). As these NO_x emissions reductions have occurred, lower O₃ concentrations have generally increased while the higher O₃ concentrations have generally decreased, resulting in a compressed O₃ distribution, relative to historical conditions.

As mentioned above, the formation of O₃ from precursor emissions is also affected by meteorological parameters such as the intensity of sunlight and atmospheric mixing. Major episodes of high O₃ concentrations in the eastern U.S. are often associated with slow-moving high-pressure systems which can persist for several days. High pressure systems during the warmer seasons are associated with the sinking of air, resulting in warm, generally cloudless skies, with light winds. The sinking of air results in the development of stable conditions near the surface which inhibit or reduce the vertical mixing of O₃ precursors, concentrating them near the surface. Photochemical activity involving these precursors is enhanced because of higher temperatures and the availability of sunlight during the warmer seasons. In the eastern U.S., concentrations of O₃ and other photochemical oxidants are determined by meteorological and chemical processes extending typically over areas of several hundred thousand square kilometers. Therefore, O₃ episodes are often regarded as regional in nature, although more localized episodes often occur in some areas, largely the result of local pollution sources during summer. In addition, in some parts of the U.S. (e.g., Los Angeles, CA), mountain barriers limit O₃ dispersion and result in a higher frequency and duration of days with elevated O₃ concentrations.

More recently, high O₃ concentrations of up to 150 parts per billion (ppb) have been measured during the wintertime in two western U.S. mountain basins. Wintertime mountain basin O₃ episodes occur on cold winter days with low wind speeds, clear skies, substantial snow cover, extremely shallow boundary layers driven by strong temperature inversions, and substantial precursor emissions activity from the oil and gas sector. The results of recent modeling studies suggest that photolysis of VOCs provides the source of reactive chemical species (radicals) needed to initiate the chemistry driving these wintertime O₃ episodes. This mechanism is markedly different from the chemistry driving summertime O₃ formation, which is initiated with the photolysis of NO₂ followed by the formation of the OH radicals.

O₃ concentrations in a region are affected both by local formation and by transport of O₃ and its precursors from upwind areas. O₃ transport occurs on many spatial scales including local transport within urban areas, regional transport over large regions of the U.S., and long-range transport which may also include international transport. In addition, O₃ can be transferred into the troposphere from the stratosphere, which is rich in naturally occurring O₃, through stratosphere-troposphere exchange (STE). These intrusions usually occur behind cold fronts, bringing stratospheric air with them and

typically affect O₃ concentrations in higher elevation areas (e.g., above 1500 m) more than areas at lower elevations.

3. Sources and Emissions of O₃ Precursors

Sources of emissions of O₃ precursor compounds can be divided into anthropogenic and natural source categories, with natural sources further divided into emissions from biological processes of living organisms (e.g., plants, microbes, and animals) and emissions from chemical or physical processes (e.g., biomass burning, lightning, and geogenic sources). Anthropogenic emissions associated with combustion processes, including mobile sources and power plants, accounted for the majority of U.S. NO_X and CO emissions in 2020 (Figure 1). Anthropogenic sources are also important for VOC emissions, though in some locations and times of the year (e.g., southeastern states during summer) the majority of VOC emissions come from vegetation.³ CH₄ is mostly emitted from anthropogenic sources, including production and consumption of fossil fuels for energy, agriculture and waste disposal systems. In practice, the distinction between natural and anthropogenic sources is often unclear, as human activities directly or indirectly affect emissions from what would have been considered natural sources during the preindustrial era. Thus, precursor emissions from plants, animals, and wildfires could be considered either natural or anthropogenic, depending on whether emissions result from agricultural practices, forest management practices, lightning strikes, or other types of events. Additional challenges are presented because much O₃ results from reactions between anthropogenic and natural precursors.

The National Emissions Inventory (NEI) is a comprehensive and detailed estimate of air emissions of criteria pollutants, precursors to criteria pollutants, and hazardous air pollutants from air emissions sources. The NEI is released every three years based primarily upon data provided by State, Local, and Tribal air agencies for sources in their jurisdictions and supplemented by data developed by the EPA. The NEI is built using the EPA's Emissions Inventory System (EIS) first to collect the data from State, Local, and Tribal air agencies and then to blend that data with other data sources.

Accuracy in an emissions inventory reflects the extent to which the inventory represents the actual emissions that occurred. Anthropogenic emissions of air pollutants result from a variety of sources such as power plants, industrial sources, motor vehicles and agriculture. The emissions from any individual source typically vary in both time and space. For the thousands of sources that make up the NEI, there is uncertainty in one or both of these factors. For some sources, such as power plants, direct emission measurements enable the emission factors derived from them to be more certain than sources without such direct measurements. However, it is not practically possible to directly monitor each of the emission sources individually and, therefore, emission inventories necessarily contain assumptions, interpolation and extrapolation from a limited set of sample data.

³It should be noted that the definition of VOCs used in this section does not include CH₄ because it is excluded from the EPA's regulatory definition of VOCs in [40 CFR 51.100\(s\)](#).

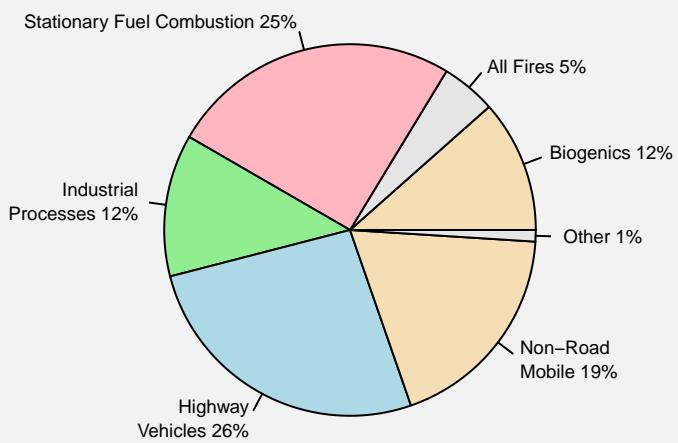
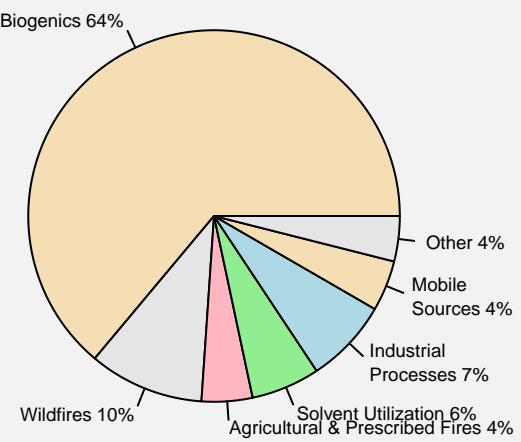
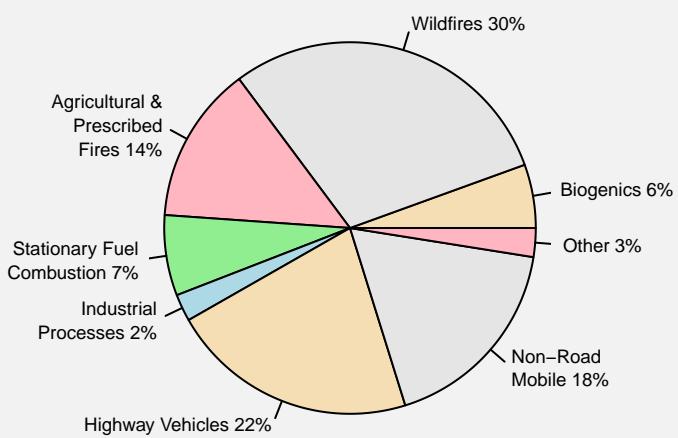
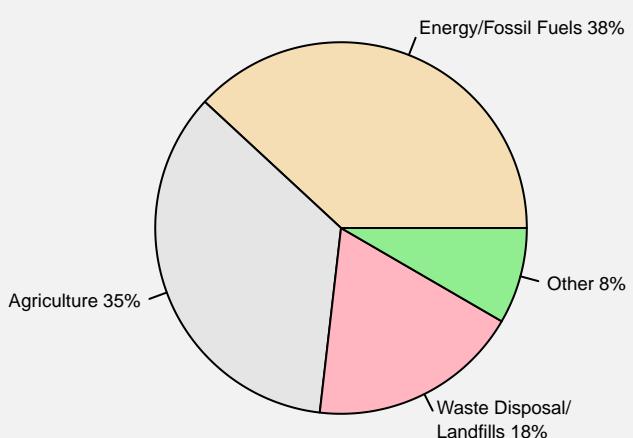
A) NO_x (8,916 kTon/yr)**B) VOCs (46,189 kTon/yr)****C) CO (66,153 kTon/yr)****D) CH₄ (34,983 kTon/yr)**

Figure 1. U.S. O₃ precursor emissions by sector: A) NO_x; B) CO; C) VOCs; D) CH₄. **Source:** [2020 NEI](#) for panels A)-C), [Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2021](#) for panel D).

Figure 2, Figure 3 and Figure 4 show county-level estimates of U.S. emissions densities (in tons/year/mi²) for CO, NO_x, and VOCs, respectively. In general, CO and NO_x emissions tend to be highest in urban areas which typically have the most anthropogenic sources of these precursors, however, CO emissions may be higher in some rural areas due to fires, and similarly NO_x emissions may be higher in some rural areas due to sources such as electricity generation, oil and gas extraction, and traffic along major highways. While there are some significant anthropogenic sources of VOC emissions in urban areas, in rural areas the vast majority of VOC emissions come from plants and trees (biogenics), particularly in the southeastern U.S. In other areas of the U.S., such as the Great Plains region and parts of the inter-mountain west, areas with higher levels of VOC emissions are largely due to oil and gas extraction.

It should be noted that O₃ levels in a given area are impacted by both local emissions of precursors that form O₃ in the area as well as remote emissions of precursors that form O₃ which is then transported into the area. Biogenic VOC emissions that lead to O₃ formation may vary greatly depending on the type and amount of vegetation, which is generally much lower in urban areas than in rural areas. However, biogenic VOC emissions that are upwind of an urban area can have a significant impact on urban O₃ levels. Thus, while the county-level maps shown in Figure 2, Figure 3 and Figure 4 illustrate the variability in precursor emissions in the U.S., it is not sufficient to look only at the patterns in local emissions when considering the impact on O₃ concentrations.

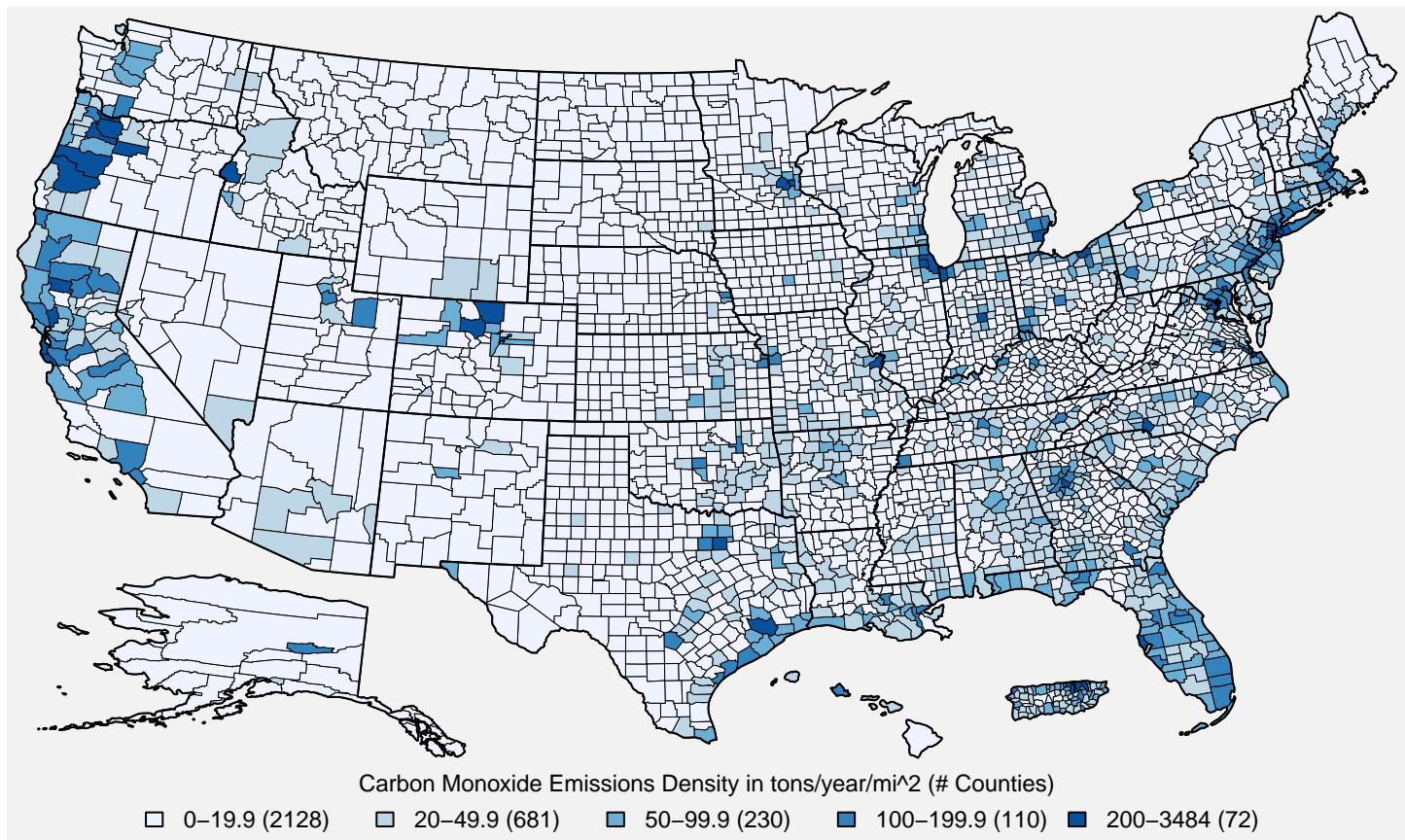


Figure 2. U.S. county-level CO emissions density estimates in tons/year/mi². **Source:** 2020 NEI

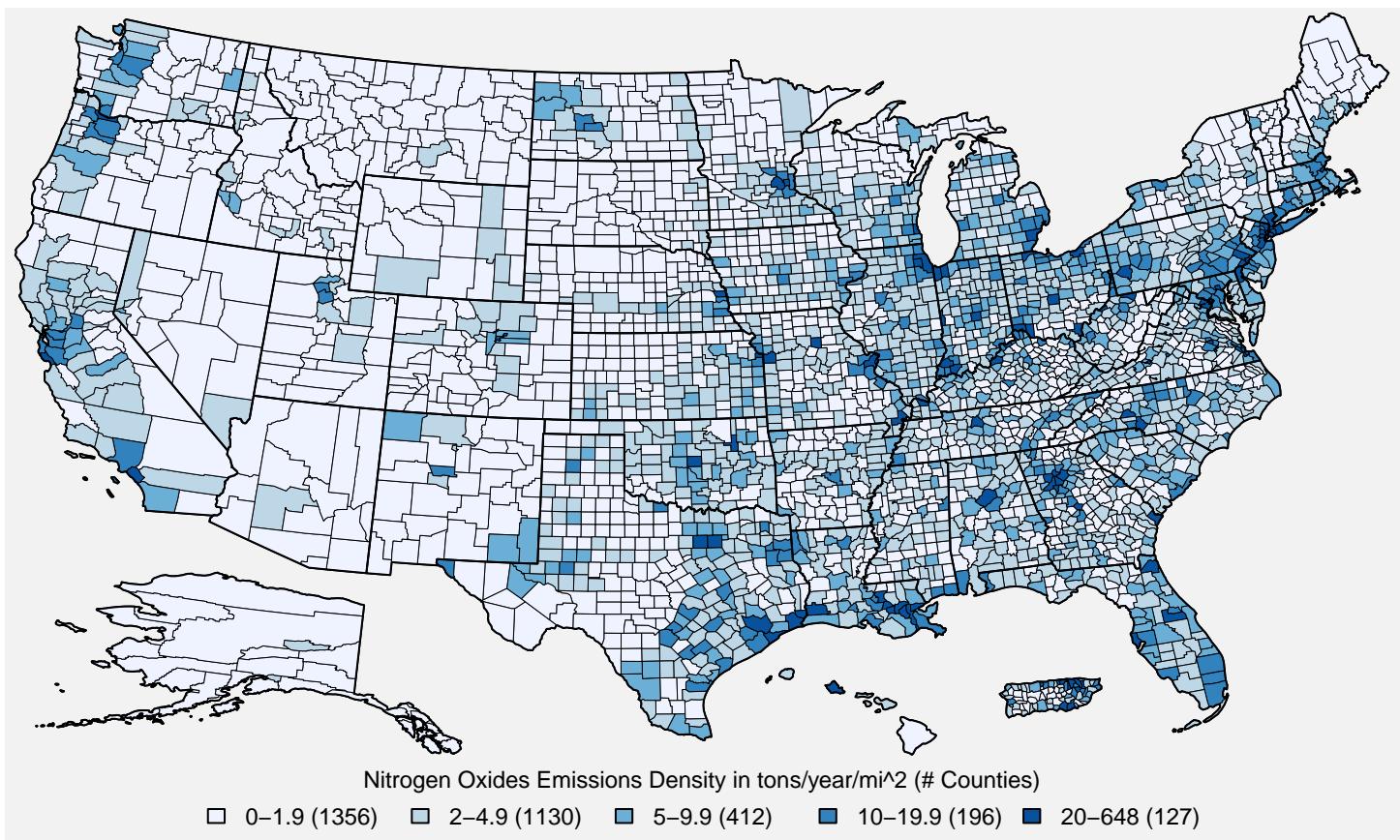


Figure 3. U.S. county-level NO_x emissions density estimates in tons/year/mi². **Source:** [2020 NEI](#)

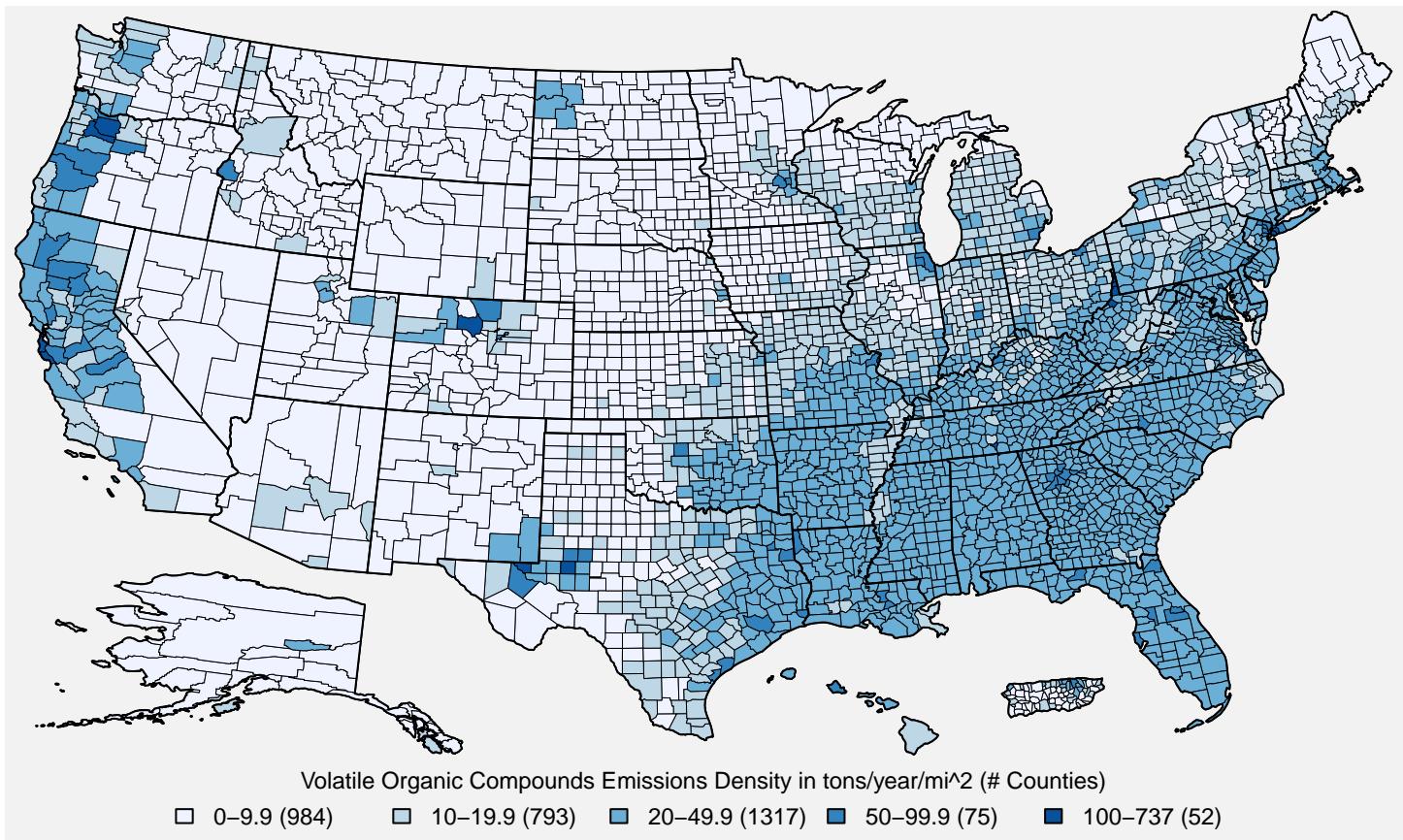


Figure 4. U.S. county-level VOC emissions density estimates in tons/year/mi². **Source:** [2020 NEI](#)

Figure 5 below shows the national trend in U.S. anthropogenic NO_x, VOC, CO, and CH₄ emissions from 2002 to 2022.⁴ Anthropogenic emissions of NO_x from mobile and stationary sources have declined appreciably since 2002, largely as a result of multiple regulatory programs implemented over the past two decades, including the NO_x SIP Call, the Cross-State Air Pollution Rule (CSAPR), and the Tier 3 Light-duty Vehicle Emissions and Fuel Standards. Similarly, EPA regulatory programs have contributed to a substantial reduction in mobile source CO emissions over the past two decades. The EPA's regulatory programs have also led to reductions in VOC emissions from mobile sources, however, industrial sources, which are the largest anthropogenic source of VOCs, have stayed relatively constant during this time period. Anthropogenic CH₄ emissions have decreased only slightly since 2002, as reductions in the energy sector have largely been offset by increasing emissions in the agricultural sector.

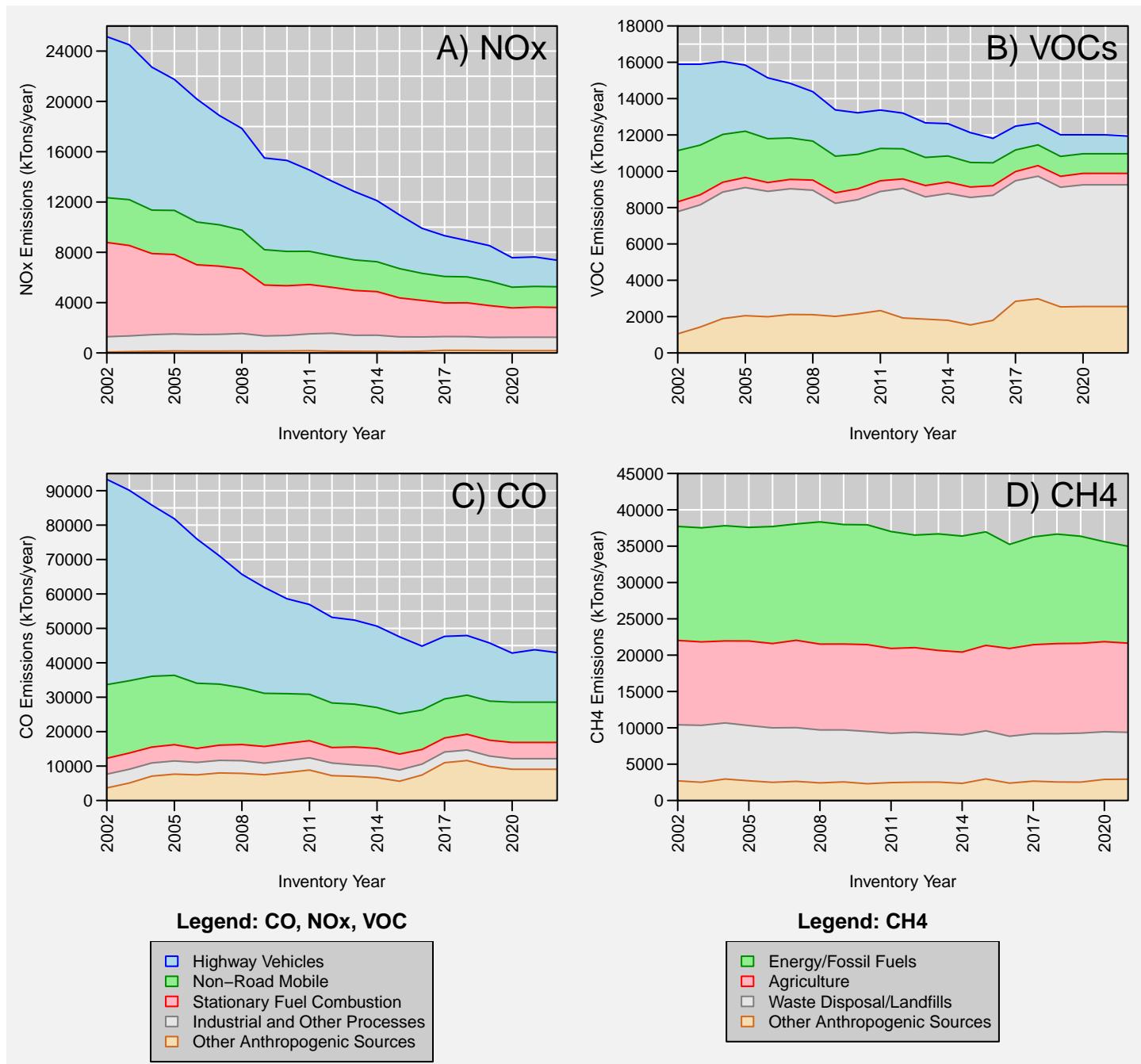


Figure 5. U.S. anthropogenic O₃ precursor emissions trends for: A) NO_x; B) CO; C) VOCs; and D) CH₄. **Source:** EPA's [Air Pollutant Emissions Trends Data](#) for panels A)-C), [Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2021](#) for panel D).

⁴NO_x, VOC, and CO data for Figure 5 come from the EPA's [Air Pollutant Emissions Trends Data](#). Note that emissions for some sectors are interpolated between inventory years, and the emissions for some sectors are held constant beyond the most recent inventory year (for details, see the "Development of Data" table in the [national emissions trends data file](#)). For the purposes of this document, wildfires are considered to be natural emissions and thus are not included in Figure 2.

4. Ambient Air Monitoring Requirements and Monitoring Networks

Ambient O₃ concentrations are measured by monitoring networks operated by State, Local and Tribal air agencies, which are typically funded in part by the EPA. The EPA provides monitoring requirements for O₃ and other pollutants in [40 CFR Part 58](#). Nearly all of the air monitoring stations that currently report data to the EPA use ultraviolet Federal Equivalent Methods (FEMs). The Federal Reference Method (FRM) was revised in 2015 to include a new chemiluminescence by nitric oxide (NO-CL) method. The previous ethylene (ET-CL) method, while still included in the CFR as an acceptable method, is no longer used due to lack of availability and safety concerns with ethylene.

In 2022, there were over 1,300 monitors reporting O₃ concentrations to the EPA. Figure 6 shows the locations of such monitoring sites that reported data to the EPA at any time during the 2020-2022 period. The main network of monitors providing ambient data for use in implementation activities related to the NAAQS is the State and Local Air Monitoring Stations (SLAMS) network, which comprises over 80% of all O₃ monitoring sites. The requirements for the SLAMS network depend on the population and most recent O₃ design values⁵ in the area. The minimum number of O₃ monitors required in a metropolitan statistical area (MSA) ranges from zero monitors for areas with a population less than 350,000 and no recent history of an O₃ design value greater than 85 percent of the level of the standard, to four monitors for areas with a population greater than 10 million and an O₃ design value greater than 85 percent of the standard level.⁶ At least one site in each MSA must be designed to record the maximum concentration for that particular metropolitan area.

Two important subsets of SLAMS sites separately make up the [National Core \(NCore\)](#) multi-pollutant monitoring network and the [Photochemical Assessment Monitoring Stations \(PAMS\)](#) network. Each State is required to have at least one NCore station, and O₃ monitors at NCore sites are required to operate year-round. At each NCore site located in a MSA with a population of 1 million or more (based on the most recent census), a PAMS network site is required.⁷ At a minimum, monitoring sites in the PAMS network are required to measure certain O₃ precursors during the months of June, July and August, although some precursor monitoring may be required for longer periods of time to improve the usefulness of data collected during an area's O₃ season. In addition to reporting O₃ concentrations, the NCore and PAMS networks provide data on O₃ precursor chemicals. The NCore sites feature co-located measurements of chemical species such as nitrogen oxide and total reactive nitrogen, along with various meteorological measurements. The additional data collected at the PAMS sites include measurements of NO_X, and a target set of VOCs. The enhanced monitoring at sites in these two networks informs our understanding of local O₃ formation.

While the SLAMS network has a largely urban and population-based focus, there are monitoring sites in other networks that can be used to track compliance with the NAAQS in rural areas. For example, the [Clean Air Status and Trends Network \(CASTNET\)](#) sites are located in rural areas. There were 83 CASTNET sites operating in 2022, with most of the sites in the eastern U.S. being operated by the EPA, and most of the sites in the western U.S. being operated by the National Park Service (NPS). Finally, there are also a number of Special Purpose Monitors (SPMs), which are not required but are often operated by air agencies for short periods of time (i.e., less than 3 years) to collect data for human health and welfare studies, as well as other types of monitoring sites, including monitors operated by tribes and industrial sources. The SPMs are typically not used to assess compliance with the NAAQS.⁸

⁵A design value is a statistic that summarizes the air quality data for a given area in terms of the indicator, averaging time, and form of the standard. Design values can be compared to the level of the standard and are typically used to designate areas as meeting or not meeting the standard and assess progress towards meeting the NAAQS.

⁶The SLAMS minimum monitoring requirements to meet the O₃ design criteria are specified in [Appendix D to 40 CFR Part 58](#).

⁷The requirements for PAMS, which were most recently updated in 2015, is fully described in section 5 of [Appendix D to 40 CFR Part 58](#).

⁸SPMs that use federal reference or equivalent methods, meet all applicable requirements in [40 CFR Part 58](#), and operate continuously for at least 3 years may be used to assess compliance with the NAAQS.

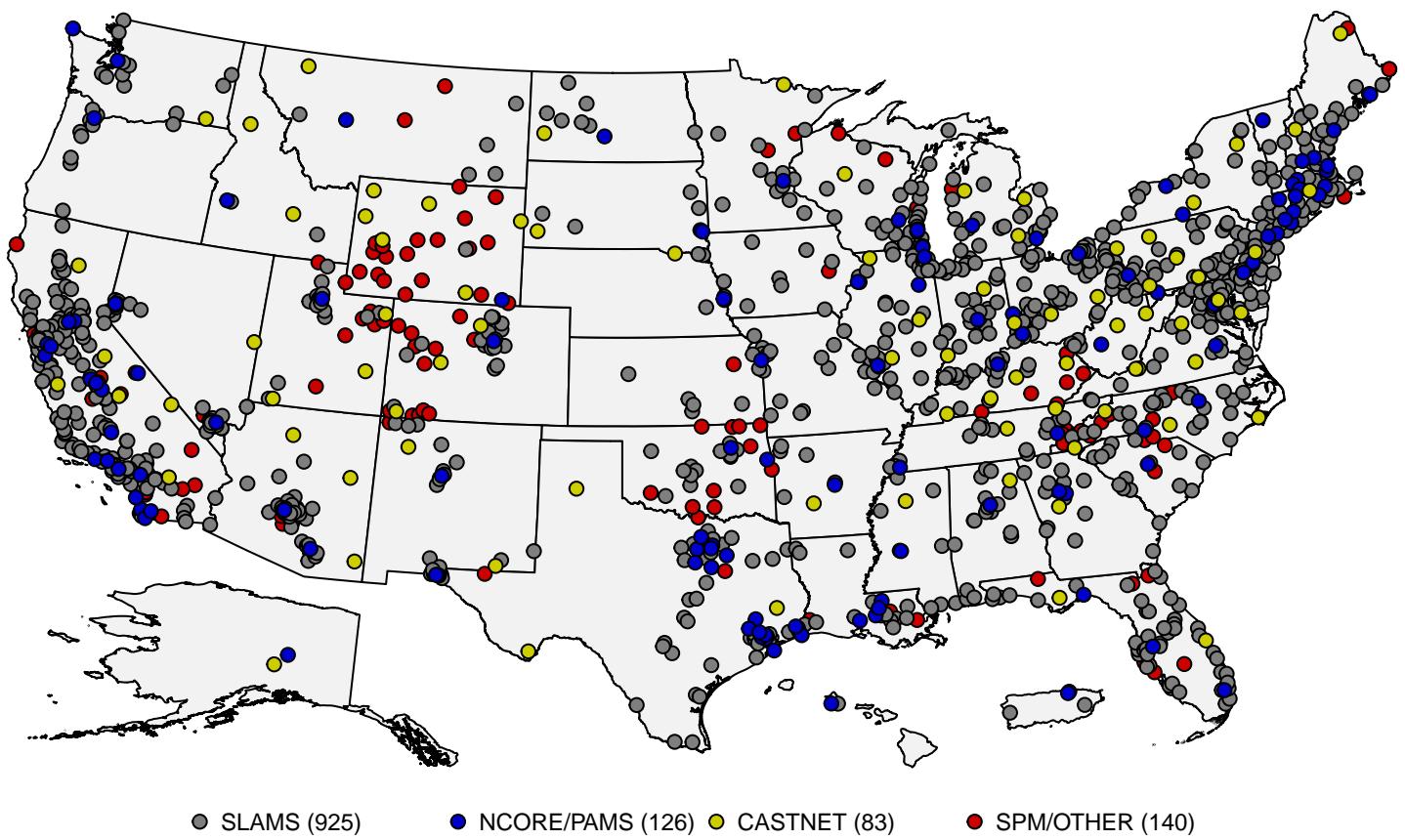


Figure 6: Map of U.S. O₃ monitoring sites reporting data to the EPA during the 2020-2022 period. **Source:** [AQS](#).

Since the highest O₃ concentrations tend to be associated with a particular season for various locations, the EPA requires O₃ monitoring only during specific O₃ monitoring seasons, as shown in Figure 7. The O₃ monitoring seasons vary by State and range in length from five months (May to September in Oregon and Washington) to all twelve months (in 11 states), with the most common season being March to October (in 27 states).⁹

To provide an assessment of data quality, monitoring agencies must perform quality assurance (QA) checks at least once every two weeks to derive estimates of precision and bias for O₃ and the other gaseous criteria pollutant measurements using calibration gas. For O₃ monitors, the data quality goal for precision and bias is 7 percent. Ambient air quality data and associated QA data are reported to the EPA via the [Air Quality System \(AQS\)](#).¹⁰ Data are reported quarterly and must be submitted to AQS within 90 days after the end of each calendar quarter (i.e. Jan/Feb/Mar, Apr/May/Jun, Jul/Aug/Sep, Oct/Nov/Dec). Additionally, each monitoring agency is required to certify all FRM/FEM data that is submitted to AQS annually, taking into consideration any QA findings, and a data certification letter must be sent to the EPA Regional Administrator by May 1st of the following year.

⁹The required O₃ monitoring seasons for each State are listed in Table D-3 of [Appendix D to 40 CFR Part 58](#).

¹⁰Quality assurance requirements for monitors used in evaluations of the NAAQS are provided in [Appendix A to 40 CFR Part 58](#). Annual summary reports of precision and bias can be obtained for each monitoring site at the EPA's [Air Data website](#).

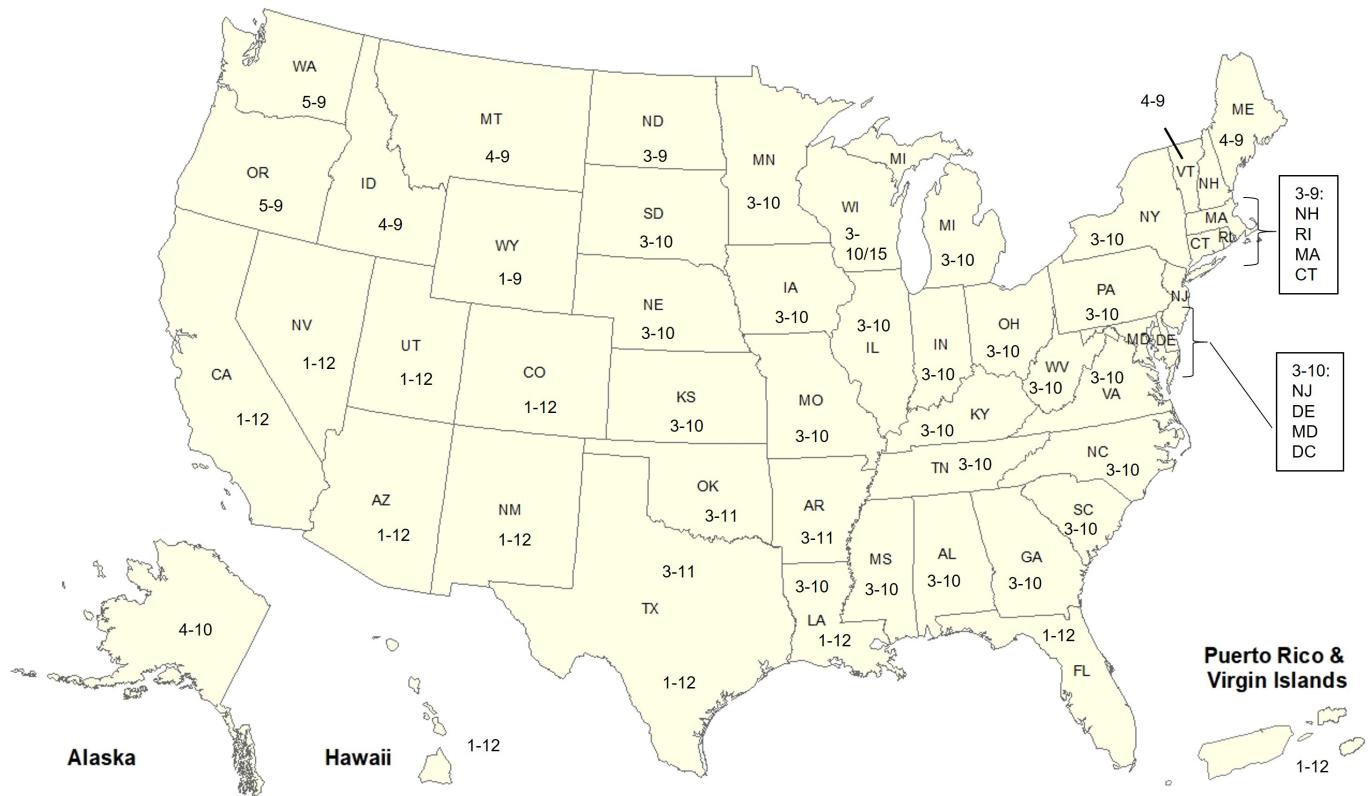


Figure 7: Current O₃ monitoring seasons in the U.S. Numbers in each State indicate the months of the year the State is required to monitor for O₃ (e.g., 3-10 means O₃ monitoring is required from March through October).

5. Data Handling Conventions and Computations for Determining Whether the Standards are Met

To assess whether a monitoring site or geographic area (usually a county or urban area) meets or exceeds a NAAQS, the monitoring data are analyzed consistent with the established regulatory requirements for the handling of monitoring data for the purposes of deriving a design value. A design value summarizes ambient air concentrations for an area in terms of the indicator, averaging time and form for a given standard such that its comparison to the level of the standard indicates whether the area meets or exceeds the standard. The procedures for calculating design values for the current O₃ NAAQS (established in 2015) are detailed in [Appendix U to 40 CFR Part 50](#) and are summarized below.

Hourly average O₃ concentrations at the monitoring sites used for assessing whether an area meets or exceeds the NAAQS are required to be reported in parts per million (ppm) to the third decimal place, with additional digits truncated, consistent with the typical measurement precision associated with most O₃ monitoring instruments. Monitored hourly O₃ concentrations flagged by the States as having been affected by an exceptional event, having been the subject of a demonstration submitted by the State, and having received concurrence from the appropriate EPA Regional Office, are excluded from design value calculations consistent with [40 CFR 50.14](#).¹¹ The hourly concentrations are used to compute moving 8-hour averages, which are stored in the first hour of each 8-hour period (e.g., the 8-hour average for the 7:00 AM to 3:00 PM period is stored in the 7:00 AM hour), and digits to the right of the third decimal place are truncated. Each 8-hour average is considered valid if six or more hourly concentrations are available for the 8-hour period.

Next, the daily maximum 8-hour average (MDA8) concentration for each day is identified as the highest of the 17 consecutive, valid 8-hour average concentrations beginning at 7:00 AM and ending at 11:00 PM (which includes hourly O₃ concentrations from the subsequent day). MDA8 values are considered valid if at least 13 valid 8-hour averages are available for the day, or if the MDA8 value is greater than the level of the NAAQS. Finally, the O₃ design value is calculated as the

¹¹A variety of resources and guidance documents related to identification and consideration of exceptional events in design value calculations are available at [<https://www.epa.gov/air-quality-analysis/final-2016-exceptional-events-rule-supporting-guidance-documents-updated-faqs>].

3-year average of the annual 4th highest MDA8 value.¹² An O₃ design value less than or equal to the level of the NAAQS is considered to be valid if valid MDA8 values are available for at least 90% of the days in the O₃ monitoring season (as defined for each State and shown in Figure 7) on average over the 3 years, with a minimum of 75% data completeness in any individual year. Design values greater than the level of the NAAQS are always considered to be valid. The current level of the O₃ NAAQS is 70 ppb.

An O₃ monitoring site meets the NAAQS if it has a valid design value less than or equal to the level of the standard, and it exceeds the NAAQS if it has a design value greater than the level of the standard. A geographic area meets the NAAQS if all ambient air monitoring sites in the area have valid design values meeting the standard. Conversely, if one or more monitoring sites has a design value exceeding the standard, then the area exceeds the NAAQS.

6. O₃ Concentrations Measured at Ambient Air Monitoring Sites Across the U.S.

Table 1 below presents summary statistics based on three daily O₃ metrics commonly used in health studies, the daily maximum 1-hour (MDA1) metric, the daily maximum 8-hour (MDA8) metric, and the daily 24-hour average (DA24) metric. These statistics are presented for year-round and each season (winter=Dec/Jan/Feb, spring=Mar/Apr/May, summer=Jun/Jul/Aug, autumn=Sep/Oct/Nov) for monitors that have complete year-round data (defined as having at least 75% data completeness based on a year-round O₃ monitoring season) in AQS for 2020-2022. Table 2 presents the same summary statistics as Table 1 for each NOAA Climate Region¹³ based on monitoring sites with at least 75% data completeness during the May - September period (i.e., the months where every State is required to monitor for O₃). Finally, Table 3 the same set of summary statistics based on the May-September period for the three daily O₃ metrics for four subsets of sites: urban NCORE and PAMS sites versus rural CASTNET sites, both in the eastern U.S. versus the in western U.S.¹⁴

Table 1 shows that the distribution of O₃ concentrations observed in the U.S. in recent years is nearly symmetric: the mean and median concentrations are within 2 ppb of each other for all combinations of metric and season. The highest median and mean O₃ concentrations generally occur in the spring and summer months, while the highest peak O₃ concentrations generally occur in the summer, and sometimes in the late spring or early fall when sunlight and temperature conditions are most favorable to O₃ production. Winter O₃ concentrations are generally lower, except for sporadic episodes occurring in mountain basins in the western U.S. under the right conditions as noted in section 2. Fewer low O₃ concentrations are observed in the spring, when background O₃ concentrations tend to be highest.

Table 1. National distribution of O₃ concentrations in ppb from the year-round dataset for 2020-2022.¹⁵ Source: AQS.

metric	season	N.sites	N.obs	mean	SD	min	p1	p5	p10	p25	p50	p75	p90	p95	p98	p99	max	max.site
MDA1	all	764	808,052	45	14	-3	18	25	29	36	44	52	62	68	78	85	275	060570005
MDA1	winter	709	187,504	37	9	-3	14	22	26	32	38	43	48	51	54	57	275	060570005
MDA1	spring	754	204,480	48	10	0	24	32	36	42	48	54	60	65	72	77	146	080677003
MDA1	summer	745	201,918	50	16	0	19	25	30	39	50	60	71	78	88	95	179	300750001
MDA1	autumn	734	195,184	43	14	0	18	24	28	34	42	50	60	68	78	86	185	060371103
MDA8	all	764	805,491	40	12	-3	14	21	25	32	40	48	56	61	68	73	139	060710005
MDA8	winter	707	186,593	33	9	-3	10	18	22	28	34	40	44	47	50	52	93	060570005
MDA8	spring	753	203,630	44	10	0	21	28	32	38	44	50	56	59	64	68	116	060714003
MDA8	summer	744	201,029	45	15	0	16	22	25	34	45	55	63	68	76	81	139	060710005
MDA8	autumn	730	193,810	38	12	0	14	20	24	30	37	45	54	60	67	73	138	060370016
DA24	all	764	808,052	30	10	-4	8	14	17	23	30	38	44	48	52	56	102	060570005
DA24	winter	709	187,504	26	9	-4	5	10	13	19	26	33	38	41	44	46	62	560350099
DA24	spring	754	204,480	35	8	-2	16	21	24	30	35	41	45	48	52	55	81	060714003
DA24	summer	745	201,918	33	11	0	11	15	18	24	33	41	48	52	57	60	102	060570005
DA24	autumn	734	195,184	28	10	-1	8	13	16	21	27	34	40	44	49	53	100	490037001

N.sites = number of sites; N.obs = number of observations; SD = standard deviation; min = minimum; p1, p5, p10, p25, p50, p90, p95, p98, p99 = 1st, 5th, 10th, 25th, 50th, 90th, 95th, 98th, 99th percentiles; max = maximum; max.site = AQS ID number for the monitoring site corresponding to the observation in the max column. winter = December/January/February; spring = March/April/May; summer = June/July/August; autumn = September/October/November.

¹²Design values are reported in ppm to the third decimal place, with additional digits truncated. This truncation step also applies to the initially calculated 8-hour average concentrations.

¹³For Table 2, monitoring sites in Alaska were assigned to the Northwest Region and monitoring sites in Hawaii were assigned to the West region.

¹⁴For Table 3, the eastern U.S. is defined as east of 100 degrees West longitude and the western U.S. is defined as west of 100 degrees West longitude.

¹⁵Negative concentration values may appear in AQS datasets down to the negative of the lower detection limit (LDL) to allow for normal instrument variability at very low concentrations. Data that exceed the negative of the LDL is typically indicative of a malfunction or another issue that affects the data defensibility.

Table 2 shows that the Southwest region typically measures the highest mean and median O₃ concentrations during the May-September period, while the highest peak O₃ concentrations are typically measured in California, in the West region. The distributions in the five eastern and central U.S. regions are all fairly similar, with the lowest concentrations occurring in the Southeast region. According to Table 3, the median and mean May-September MDA1 and MDA8 O₃ concentrations are roughly 10 ppb higher in the western U.S. than in the eastern U.S., with even greater differences seen in the upper tail of the distribution. In the eastern U.S., MDA1 and MDA8 O₃ concentrations in the lower half of the distribution tend to be similar across rural and urban areas, while concentrations in the upper half of the distribution tend to be higher in urban areas. In the Western U.S., MDA1 and MDA8 O₃ concentrations in the lower half of the distribution tend to be higher in rural areas, while concentrations in the upper half of the distribution tend to be higher in urban areas. This pattern reflects the compression of the O₃ distribution with lower levels of precursor emissions as noted in Section 2. Meanwhile, the DA24 metric tends to be higher in rural areas than in urban areas across the full distribution, owing to the tendency for more nighttime O₃ titration to occur in areas with higher precursor emissions.

Table 2. National distribution of O₃ concentrations in ppb from the May-September dataset for 2020-2022.¹⁵ **Source:** AQS.

metric	region	N.sites	N.obs	mean	SD	min	p1	p5	p10	p25	p50	p75	p90	p95	p98	p99	max	max.site
MDA1	all	1,141	513,816	49	15	-4	20	27	31	39	48	57	66	73	83	90	190	040217030
MDA1	Central	178	80,745	48	11	0	25	31	34	40	47	55	63	67	73	78	115	295100085
MDA1	East North Central	87	38,952	46	12	3	21	28	31	38	46	54	62	68	74	79	118	260050003
MDA1	Northeast	178	79,990	46	13	0	20	27	31	37	45	53	62	67	75	81	144	230112001
MDA1	Northwest	27	11,928	43	13	0	19	25	28	35	42	51	60	66	73	78	110	410050004
MDA1	South	134	59,944	45	15	-4	18	23	26	35	45	54	64	70	78	84	135	201730010
MDA1	Southeast	191	86,462	42	12	0	19	23	26	33	42	50	57	61	67	70	134	120550003
MDA1	Southwest	120	54,055	59	11	2	35	43	47	52	58	65	74	79	86	91	190	040217030
MDA1	West	178	80,163	56	19	0	19	28	33	43	55	67	80	89	100	109	185	060371103
MDA1	West North Central	48	21,577	50	10	4	26	34	38	44	50	56	62	66	71	74	179	300750001
MDA8	all	1,141	512,175	43	13	0	17	23	27	34	43	52	60	65	72	77	139	060710005
MDA8	Central	178	80,497	43	11	0	21	27	30	36	43	50	57	61	66	70	100	295100085
MDA8	East North Central	87	38,901	42	11	1	18	24	28	34	41	49	57	62	67	71	104	550590019
MDA8	Northeast	178	79,564	41	11	0	17	23	27	33	40	48	55	60	65	69	99	090019003
MDA8	Northwest	27	11,874	39	11	0	15	21	25	31	38	46	54	58	64	68	96	530330023
MDA8	South	134	59,742	40	13	0	15	20	23	30	40	49	57	62	67	71	114	484390075
MDA8	Southeast	191	86,189	37	11	0	16	20	22	28	37	45	52	55	60	62	93	131210055
MDA8	Southwest	120	53,937	54	9	2	31	39	43	48	54	60	66	70	75	79	108	040131004
MDA8	West	178	79,957	49	16	0	17	25	30	39	49	59	69	76	84	90	139	060710005
MDA8	West North Central	48	21,514	46	10	3	22	30	34	40	46	52	58	61	66	69	89	560019991
DA24	all	1,141	513,816	32	10	-4	11	16	19	24	32	39	46	50	55	58	102	060570005
DA24	Central	178	80,745	31	8	0	14	18	21	25	31	36	42	45	49	52	76	170317002
DA24	East North Central	87	38,952	32	9	3	14	18	21	26	32	38	44	48	52	56	79	260050003
DA24	Northeast	178	79,990	30	9	0	12	17	19	24	30	36	41	45	49	52	73	440090008
DA24	Northwest	27	11,928	29	9	0	10	15	18	22	28	35	41	44	49	51	65	160230101
DA24	South	134	59,944	28	10	-4	10	14	16	21	28	35	42	45	49	52	72	484390075
DA24	Southeast	191	86,462	26	9	0	10	13	15	19	25	32	38	42	46	49	65	371990004
DA24	Southwest	120	54,055	42	8	2	22	28	32	37	42	47	52	55	58	61	78	080590011
DA24	West	178	80,163	36	11	0	13	19	22	28	36	44	51	56	62	66	102	060570005
DA24	West North Central	48	21,577	37	9	2	16	23	26	32	37	43	48	52	56	58	76	560019991

N.sites = number of sites; N.obs = number of observations; SD = standard deviation; min = minimum; p1, p5, p10, p25, p50, p90, p95, p98, p99 = 1st, 5th, 10th, 25th, 50th, 90th, 95th, 98th, 99th percentiles; max = maximum; max.site = AQS ID number for the monitoring site corresponding to the observation in the max column. Central = Illinois, Indiana, Kentucky, Missouri, Ohio, Tennessee, West Virginia; East North Central = Iowa, Minnesota, Michigan, Wisconsin; Northeast = Connecticut, Delaware, Maine, Maryland, Massachusetts, New Hampshire, New Jersey, New York, Pennsylvania, Rhode Island, Vermont; Northwest = Alaska, Idaho, Oregon, Washington; South = Arkansas, Kansas, Louisiana, Mississippi, Oklahoma, Texas; Southeast = Alabama, Florida, Georgia, North Carolina, South Carolina, Virginia; Southwest = Arizona, Colorado, New Mexico, Utah; West = California, Hawaii, Nevada; West North Central = Montana, Nebraska, North Dakota, South Dakota, Wyoming.

Table 3. Distribution of O₃ concentrations in ppb from the May-September dataset by eastern vs. western U.S. and urban vs. rural sites for 2020-2022.¹⁵ **Source:** AQS.

metric	region	site.type	N.sites	N.obs	mean	SD	min	p1	p5	p10	p25	p50	p75	p90	p95	p98	p99	max	max.site
MDA1	All Sites	All Sites	1,141	513,816	49	15	-4	20	27	31	39	48	57	66	73	83	90	190	040217030
MDA1	Eastern U.S.	Urban	65	29,244	46	14	0	19	25	29	36	46	55	64	71	79	85	130	482011039
MDA1	Eastern U.S.	Rural	42	18,937	44	11	1	20	26	30	37	44	51	58	62	68	72	98	240339991
MDA1	Western U.S.	Urban	34	15,057	55	17	0	19	28	34	44	54	65	77	85	94	100	185	060371103
MDA1	Western U.S.	Rural	32	14,516	55	11	14	27	36	41	48	54	60	68	74	83	88	130	061070009
MDA8	All Sites	All Sites	1,141	512,175	43	13	0	17	23	27	34	43	52	60	65	72	77	139	060710005
MDA8	Eastern U.S.	Urban	65	29,157	41	13	0	16	21	24	32	41	49	57	62	68	72	102	484391002
MDA8	Eastern U.S.	Rural	42	18,839	40	11	1	17	22	26	32	40	47	53	58	62	65	84	171199991
MDA8	Western U.S.	Urban	34	15,003	49	14	0	16	25	30	39	49	58	67	72	79	83	118	060371103
MDA8	Western U.S.	Rural	32	14,469	51	11	12	24	32	37	45	51	57	63	68	75	79	114	061070009
DA24	All Sites	All Sites	1,141	513,816	32	10	-4	11	16	19	24	32	39	46	50	55	58	102	060570005
DA24	Eastern U.S.	Urban	65	29,244	29	9	0	11	15	17	22	29	35	41	45	48	51	71	170971007
DA24	Eastern U.S.	Rural	42	18,937	30	9	0	11	15	18	24	30	37	43	46	50	53	66	361099991
DA24	Western U.S.	Urban	34	15,057	35	10	0	12	18	22	28	35	41	47	50	54	57	75	060731006
DA24	Western U.S.	Rural	32	14,516	44	10	6	18	26	30	38	44	50	56	59	64	67	91	061070009

N.sites = number of sites; N.obs = number of observations; SD = standard deviation; min = minimum; p1, p5, p10, p25, p50, p90, p95, p98, p99 = 1st, 5th, 10th, 25th, 50th, 90th, 95th, 98th, 99th percentiles; max = maximum; max.site = AQS ID number for the monitoring site corresponding to the observation in the max column. Region is defined such that the eastern U.S. includes sites east of 100 degrees west longitude and the western U.S. includes sites west of 100 degrees west longitude; site.type = Urban includes Urban NCore and PAMS sites; Site.type = Rural includes CASTNET sites.

Figure 8 below shows a map of the O₃ design values at U.S. ambient air monitoring sites based on data from the 2020-2022 period. From the figure it is apparent that many monitoring sites have design values exceeding the current NAAQS, and that most of these sites are located in or near urban areas. The highest design values occurred in California, Texas, along the shoreline of Lake Michigan, and near large urban areas in the northeastern and western U.S. There are also high design values associated with wintertime O₃ in the Uinta Basin in Utah. The lowest design values occurred in the north central region of the U.S., rural parts of New England and the southeastern U.S., and along the Pacific Ocean, including Alaska and Hawaii.

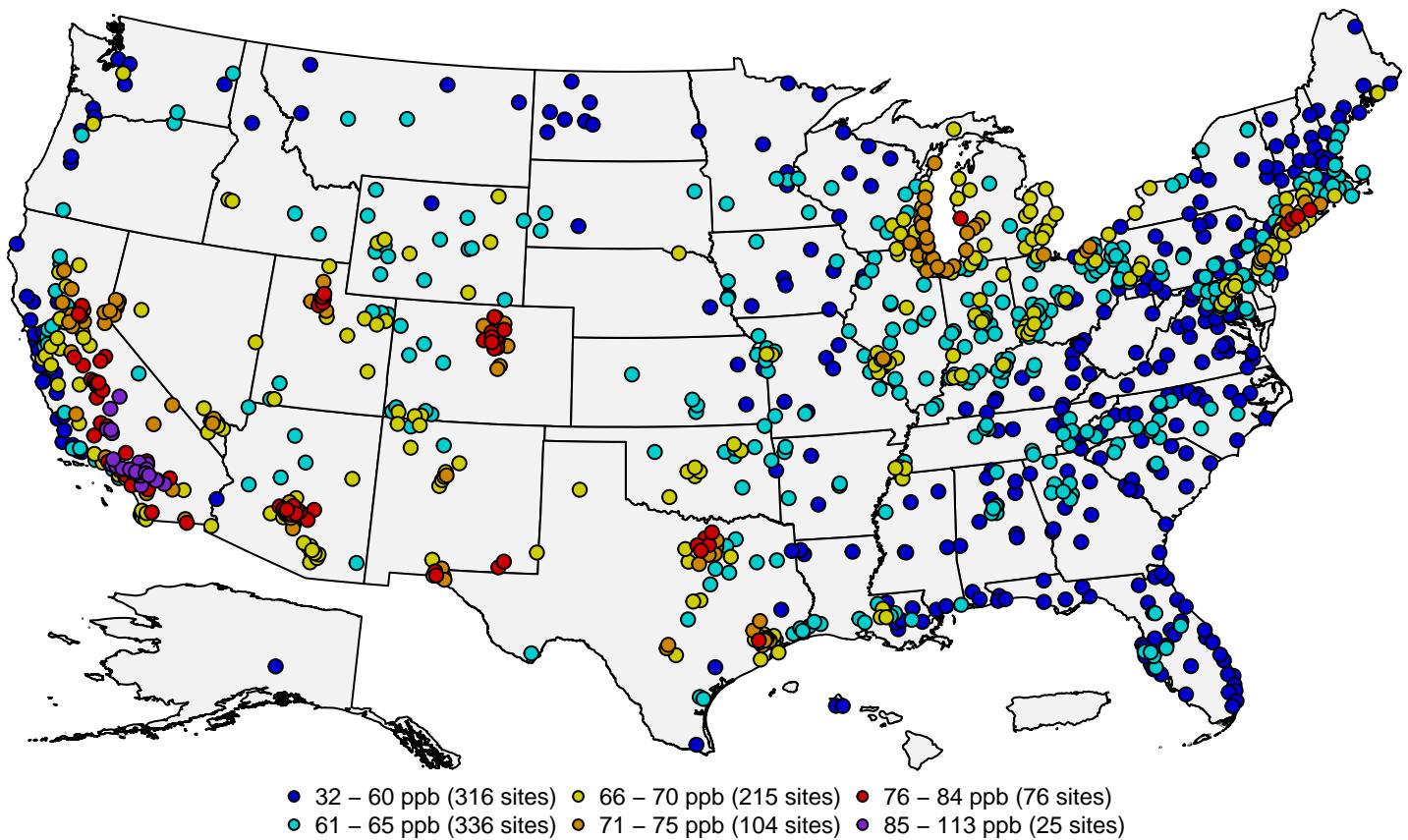


Figure 8: O₃ design values in ppb for the 2020-2022 period. **Source:** AQS.

Figure 9 below shows a map of the site-level trends in the O₃ design values at U.S. monitoring sites having valid design values in at least 16 of the 21 3-year periods from 2000 through 2022. The trends were computed using the Thiel-Sen estimator, and tests for significance (p -value < 0.05) were computed using the Mann-Kendall test. From this figure it is apparent that design values have decreased significantly over most of the eastern U.S. during this period. These decreases are in part due to EPA programs such as the Clean Air Interstate Rule (CAIR) and the Cross-State Air Pollution Rule (CSAPR) with the goal of achieving broad, regional reductions in summertime NO_x emissions, as well as mobile emission reductions from federal motor vehicle emissions and fuel standards and local controls resulting from implementation of the existing O₃ standards. Other areas of the country have also experienced decreases in design values, most notably in California and near urban areas in the intermountain west.

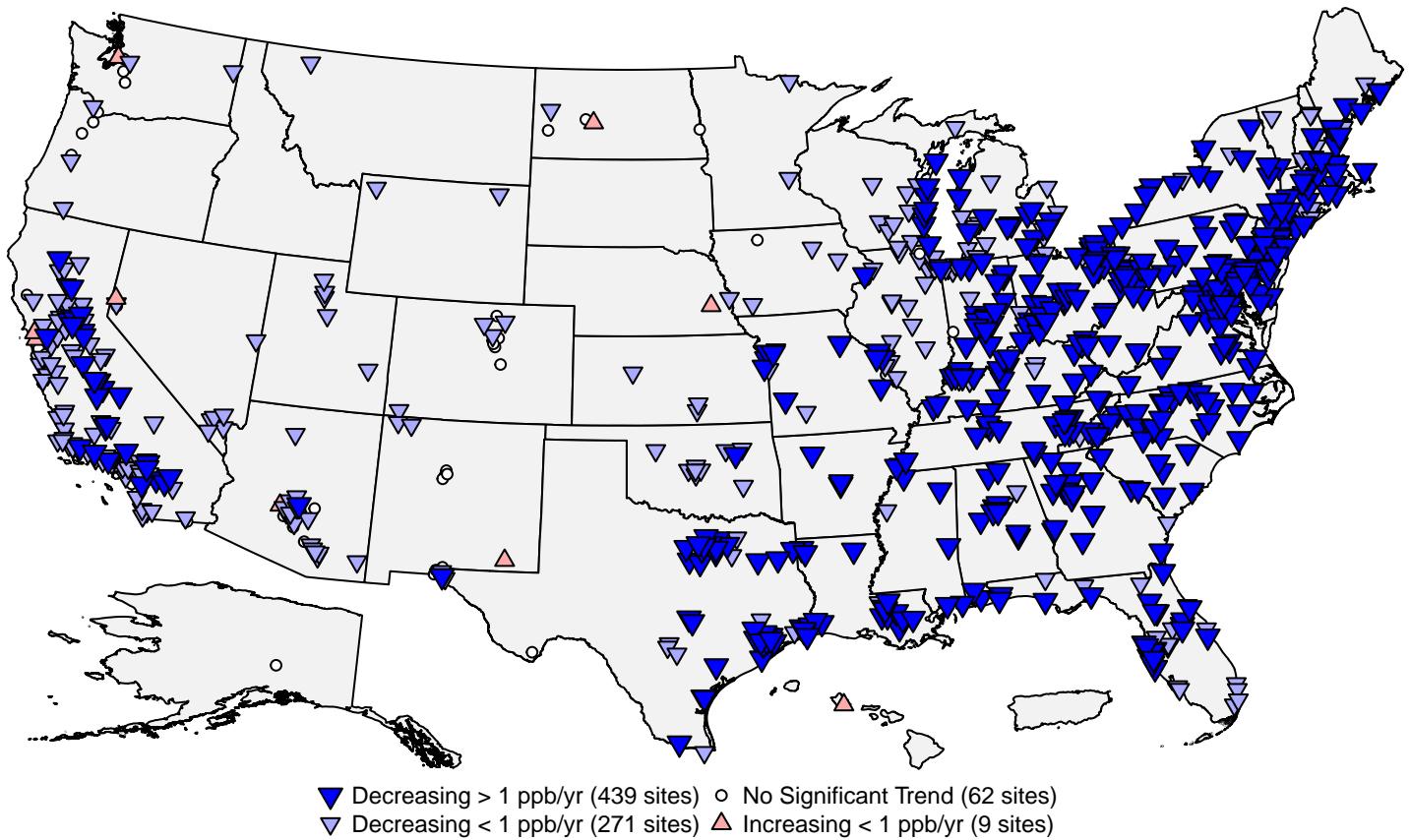


Figure 9: Site-level trends in O₃ design values based on data from 2000 through 2022. **Source:** AQS, trends computed using R statistical software.

Figure 10 below shows the national trend in the design values based on the 781 monitoring sites shown in Figure 9. The U.S. median design value decreased by 26% from 2000 (86 ppb) to 2022 (64 ppb). Additional information from the published literature has examined trends in MDA8 concentrations across the distribution of high and low O₃ days. O₃ metrics impacted by high hourly O₃ concentrations, such as the design value, decreased at most U.S. sites during the time periods studied. Concurrently, metrics that are impacted by averaging longer time periods of hourly O₃ measurements were more varied with fewer sites exhibiting decreases and most other sites exhibiting no trend.

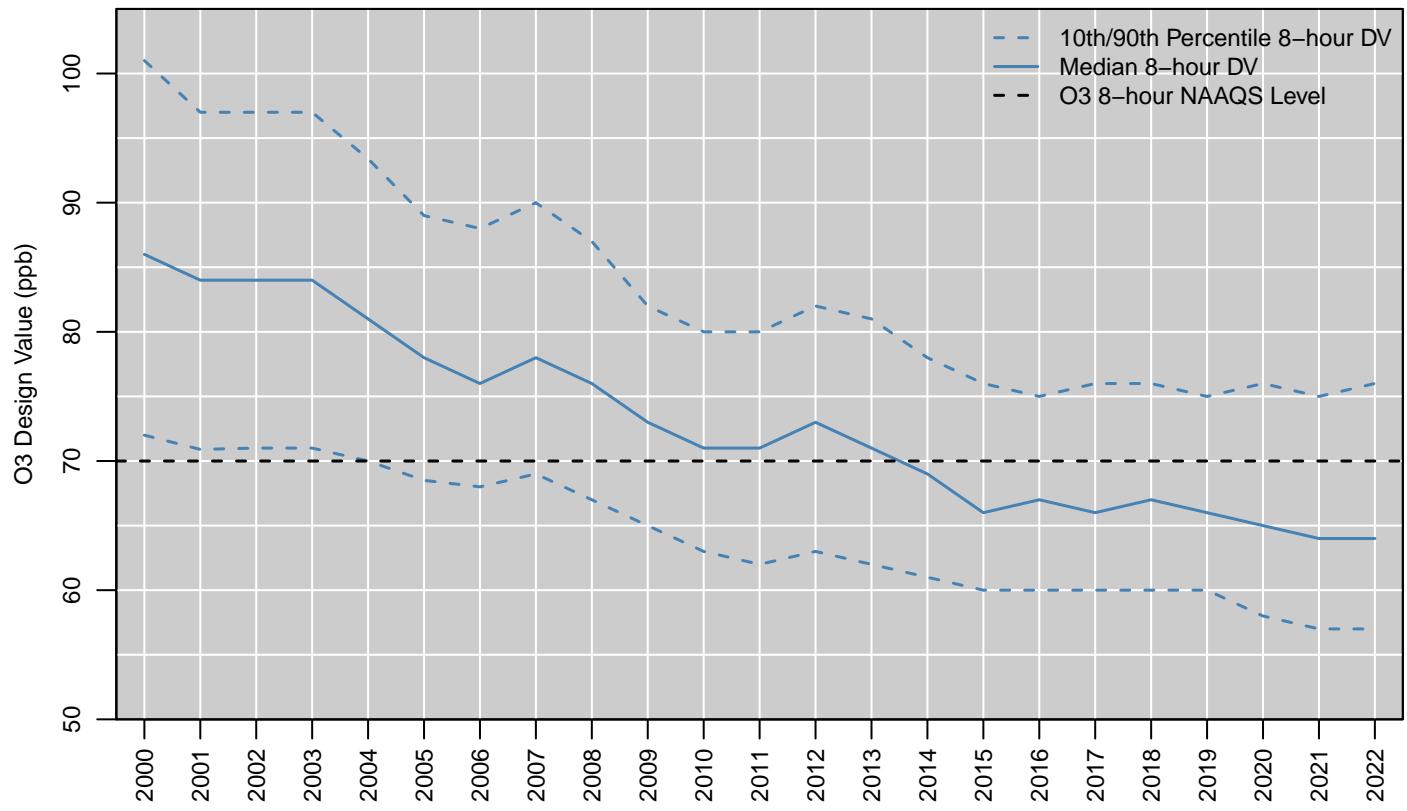


Figure 10: National trend in O₃ design values in ppb, 2000 to 2022. **Source:** [AQS](#).

Figure 11 below shows the national distribution of the annual 4th highest MDA8 O₃ concentrations reported in each year from 1980 to 2022.¹⁶ The red line shows the number of sites included in the boxplot for each year. The median annual 4th highest MDA8 O₃ concentration decreased by 31%, from 93 ppb in 1980 to 64 ppb in 2022. While the magnitude of the highest O₃ concentrations declined between 1980 and the early 2000s, the center of the distribution stayed relatively constant, with over 75% of sites measuring annual 4th highest MDA8 O₃ concentrations above the current NAAQS level. Since 2003, regional control programs such as the NO_x SIP Call and CSAPR have contributed to reduced concentrations over the eastern U.S., so that roughly 75% of all sites have been meeting the NAAQS each year since 2013. The size of the O₃ monitoring network increased from 1980 through the early 2000s and has stayed relatively constant at around 1,200 sites since then.

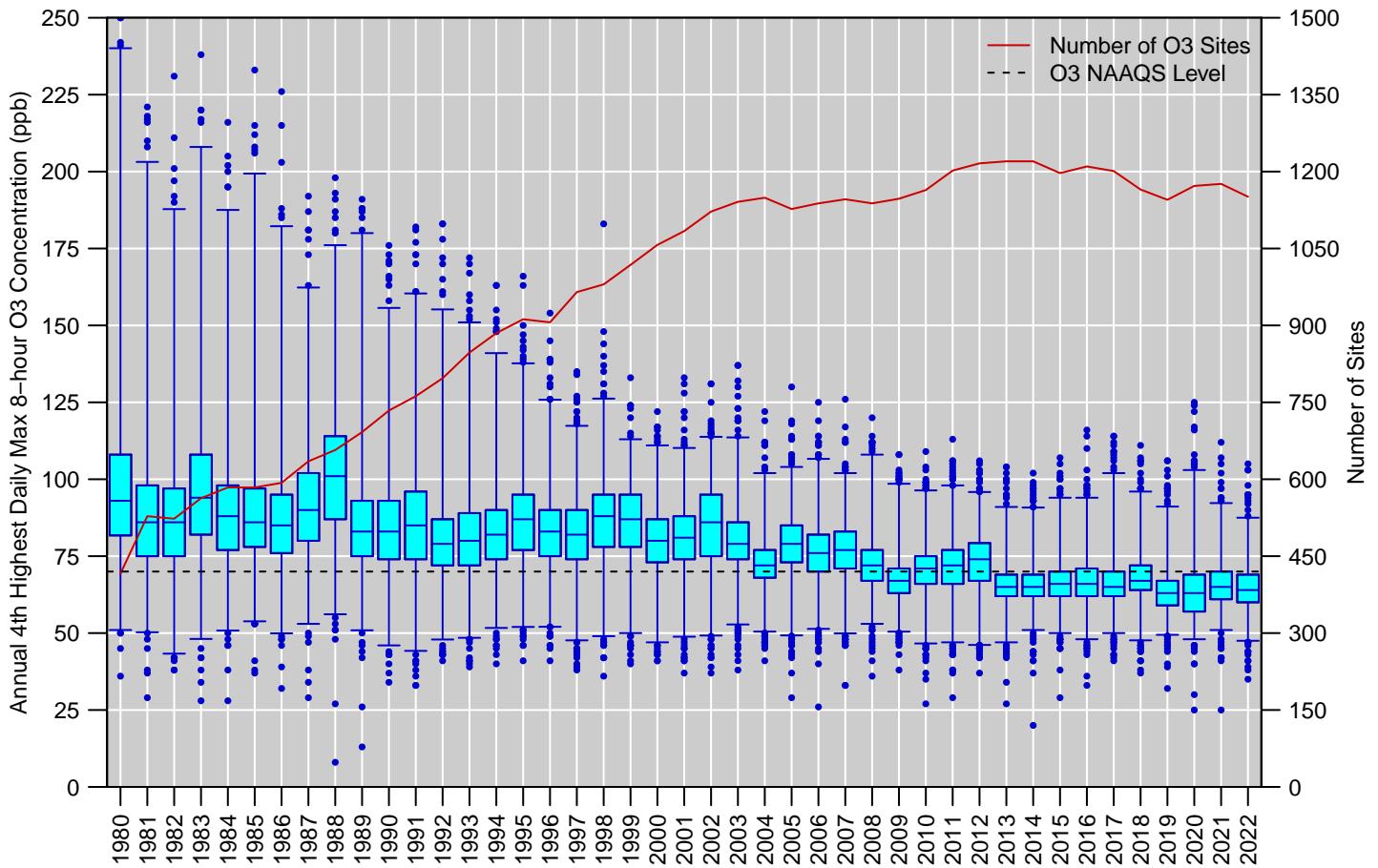


Figure 11: Distribution of annual 4th highest MDA8 O₃ concentrations measured at U.S. monitoring sites, 1980 to 2022. Boxes represent the median and interquartile range, whiskers extend to the 1st and 99th percentiles, and values outside this range are shown as circles. The red line shows the number of O₃ monitoring sites reporting data to EPA in each year. Source: [AQS](#).

¹⁶For this analysis, the annual 4th highest MDA8 O₃ concentrations were retrieved from AQS for all U.S. sites for years that had at least 75% annual data completeness during the O₃ monitoring season.

References

U.S. EPA. Integrated Science Assessment for Ozone and Related Photochemical Oxidants (Final Report, April 2020). U.S. Environmental Protection Agency, Washington, DC, EPA/600/R-20/012, 2020.

U.S. EPA. Policy Assessment for the Review of the Ozone NAAQS (Final Report, May 2020). U.S. Environmental Protection Agency, Research Triangle Park, NC, EPA-452/R-20-001, 2020.

Additional Resources

- [Ground-level Ozone Pollution](#)
- [Ozone \(\$O_3\$ \) Air Quality Standards](#)
- [National Emissions Inventory \(NEI\)](#)
- [Ambient Monitoring Technology Information Center \(AMTIC\)](#)
- [Air Quality Design Values](#)
- [National Air Quality: Status and Trends of Key Air Pollutants](#)
- [Air Data: Air Quality Data Collected at Outdoor Monitors Across the U.S.](#)