Model linearity breeds contempt: using Bayesian non-linear models to uncover broad macroecological patterns

Bernat Bramon Mora 1,* and Jake M. Alexander 1

$_{\scriptscriptstyle 1}$ Abstract

Species' distribution models have emerged as one of the most influential methodological advances in ecology and biogeography of the last decades. Useful to understand how populations of species will change along environmental gradients, they have become ecologists' compass to predict the effects of global climate change. That said, uncovering the mechanisms shaping species' realized niches has been one of the main driving forces behind the development of these models. That is, recent efforts have been often focused on understanding which biotic and abiotic factors are good predictors of species' niches—with an increasing effort placed in improving the predictive power of the statistical models. However, we still lack a general understanding of the shape of species' distributions, and much less is known about how these distributions compare to each other across gradients. Here, we use a set of Bayesian non-linear models to uncover the shape of species' realized niches. These models account for all prior knowledge we have regarding their shape, including expert knowledge 13 on species' environmental preferences and physiology. With this approach, we are able to 14 uncover the true shape of empirical species' distributions. Moreover, they allow us to tackle long-standing hypothesis regarding general biogeographical patterns. In particular, we found conclusive evidence of the relationship between several properties of distributions, including 17 the link between species' range size and elevation and their skewness along gradients. Finally, we are able to shed light on the extent to which some aspects of the shape of observed realized niches—such as kurtosis and skewness of the distributions—could be intrinsic properties of species. Overall, our approach offers a useful statistical framework to understand

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the shape of species' distributions, and our results provide an unprecedented perspective of the way systems of many species are distributed along environmental gradients.

24 Introduction

One of the central goals of ecology is to understand the ways species are distributed across space and time (ref). Over the last two decades, ecologists have developed multiple distribution models to try to untangle the factors that play a role in defining such distributions (Guisan & Zimmermann, 2000). Many ecological textbooks (Krebs, 1972) assume the shape of species distributions to be 29 unimodal and symmetric, but some have warned that empirical distributions can take many different forms (Austin, 1987, 2002) That said, a lot can be learned about species' geographic distributions from taking a 32 phenomenological approach, focusing instead on the description of basic properties of their realized niches. For example, the study of species' range sizes along environmental gradients can reveal general biodiversity patterns that are crucial from a conservation and management perspective (Stevens, 1992). Differences in species' responses to the environment could shed light on how climatic processes and historical contingencies have shaped their distributions (Rohde, 1992; ?). Other properties, such as the skewness of species' distributions, can also reveal general underlying processes regarding species' physiological tolerance to different environmental conditions (Kaufman, 1995). More generally, understanding the shape of species' realized niches and the extend to which these vary across species is a crucial issue in ecology and biogeography (ref); however, we do not have an effective way to parsimoniously 42 compare the realized niches of many species. Indeed, there is no general agreement on the shape of species' distributions (ref). 44 Many ecological textbooks (Krebs, 1972) assume the shape of species distributions to be unimodal and symmetric, but some have warned that empirical distributions can take many different forms (Austin, 1987, 2002). In practice, distribution frameworks often use logistic regressions with a linear relationship between covariates (but see? and?). This is useful because it simplifies the optimization process, but it comes with some statistical

shortcomings. First and foremost, such response curve and the linear relationship between

covariates often comes with a set of implicit mathematical constrains that might not be biologically justified. From a purely statistical perspective, if all that we are willing to assume is that species occupy finite geographic ranges—i.e. their probability distributions have finite variance—the most conservative statistical approach is to model these as a Gaussian distributions (Frank, 2009). Other factors might then condition species distributions to showcase heavy-tails or a skewed shapes, revealing interesting ecological processes shaping biodiversity 56 patterns (Austin, 1976; Minchin, 1987). The starting point, nevertheless, should be the one 57 that makes the fewest assumptions (i.e. the maximum entropy distribution; Frank 2009), and every new shape will imply a hypotheses on how communities are distributed (D'Amen et al., 2017). Second, the aforementioned structural constrains also limit our ability to include any 60 prior information to our parameter estimates. Observations on species' geographic varia-61 tion and optimal climatic conditions have long been documented, with extensive databases 62 compiled by botanists and field ecologists documenting basic knowledge on species' realized 63 niches (e.g. Landolt et al. 2010). That said, this information is rarely accounted for in most modelling approaches, potentially because there is not a straightforward way to feed this information into the parameters of a linear model (Scherrer & Guisan 2019; but see ter Braak & Looman 1986; Ovaskainen et al. 2017). Finally, and perhaps most importantly, a direct biological interpretation of parameter estimates in linear models becomes increasingly diffi-68 cult as one moves from unimodal and symmetric distributions (ter Braak & Looman, 1986; 69 Jamil & ter Braak, 2013) to skewed distributions (Huisman et al., 1993), making the tests of hypothesis on global biodiversity patterns particularly challenging. For example, Huisman et al. (1993) proposed several non-linear models to characterize several features of species' response curves; however, species' environmental indicator values, range size or distribution 73 skewness are difficult to understand altogether following these model structures.

Here, we instead rethink traditional modelling approaches and develop a conceptually simple—and yet statistical and computationally complex—statistical framework to revisit some classic hypothesis in ecology and biogeoraphy. In particular, we develop a Bayesian hierarchical model that accounts for all prior information that we have regarding the distribution of alpine plant species along an elevation gradient in the Swiss Alps, including expert knowledge on species environmental indicator values, range sizes, and plant physiology. We start by considering species' response curves as Gaussian distributed, and then we adapt our

model to allow for skewed and long-tailed distributions. Using this statistical framework, we are able to compare the basic properties of the realized niches of multiple species, testing for the existence of general biogeographical patterns.

One of the central goals of ecology is to understand the ways species are distributed across 85 space and time (ref). Over the last two decades, ecologists have developed multiple distri-86 bution models to try to untangle the factors that play a role in defining such distributions 87 (Guisan & Zimmermann, 2000). These models estimate species' realized niches using several covariates, including environmental variables (Guisan & Thuiller, 2005), species ecological traits' (Pollock et al., 2012) and phylogenetic relations (Ives & Helmus, 2011). More recently, some of the focus have shifted towards approaches that estimate and account for biotic factors, such as competitive or facilitative relationships between species (Ovaskainen et al., 2017). The idea is that by untangling the ways in which such biotic and abiotic factors shape species' distributions, we can gain a mechanistic understanding on how ecological communities are established and change over time. However, while these factors can increase the predictive performance of some of the models (Norberg et al., 2019), the interpretation of the corresponding parameter estimates has been recently questioned (Harris, 2016; Thur-97 man et al., 2019; Poggiato et al., 2021). This was best illustrated by Blanchet et al. (2020), 98 who used basic statistical arguments to highlight the artefactual nature of the link between co-occurrence and species' ecological interactions drawn by some distribution models.

The value of gaining a mechanistic understanding of species' distributions is unquestionable 101 (ref), with several studies highlighting the importance of factors such as biotic interactions 102 and dispersal ability in setting species' range limits (Wisz et al., 2013; Pollock et al., 2014; 103 Neuschulz et al., 2018). That said, a lot can be learned about species' geographic distri-104 butions from taking a phenomenological approach, focusing instead on the description of 105 basic properties of their realized niches. For example, the study of species' range sizes along 106 environmental gradients can reveal general biodiversity patterns that are crucial from a con-107 servation and management perspective (Stevens, 1992). Differences in species' responses 108 to the environment could shed light on how climatic processes and historical contingencies have shaped their distributions (Rohde, 1992; ?). Other properties, such as the skewness of 110 species' distributions, can also reveal general underlying processes regarding species' phys-111 iological tolerance to different environmental conditions (Kaufman, 1995). More generally, understanding the shape of species' realized niches and the extend to which these vary across
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Many ecological textbooks (Krebs, 1972) assume the shape of species distributions to 117 be unimodal and symmetric, but some have warned that empirical distributions can take 118 many different forms (Austin, 1987, 2002). In practice, distribution frameworks often use 119 logistic regressions with a linear relationship between covariates (but see? and?). This is useful because it simplifies the optimization process, but it comes with some statistical 121 shortcomings. First and foremost, such response curve and the linear relationship between 122 covariates often comes with a set of implicit mathematical constrains that might not be bio-123 logically justified. From a purely statistical perspective, if all that we are willing to assume 124 is that species occupy finite geographic ranges—i.e. their probability distributions have finite variance—the most conservative statistical approach is to model these as a Gaussian distributions (Frank, 2009). Other factors might then condition species distributions to showcase 127 heavy-tails or a skewed shapes, revealing interesting ecological processes shaping biodiversity 128 patterns (Austin, 1976; Minchin, 1987). The starting point, nevertheless, should be the one 129 that makes the fewest assumptions (i.e. the maximum entropy distribution; Frank 2009), and every new shape will imply a hypotheses on how communities are distributed (D'Amen et al., 131 2017). Second, the aforementioned structural constrains also limit our ability to include any 132 prior information to our parameter estimates. Observations on species' geographic varia-133 tion and optimal climatic conditions have long been documented, with extensive databases 134 compiled by botanists and field ecologists documenting basic knowledge on species' realized 135 niches (e.g. Landolt et al. 2010). That said, this information is rarely accounted for in most 136 modelling approaches, potentially because there is not a straightforward way to feed this in-137 formation into the parameters of a linear model (Scherrer & Guisan 2019; but see ter Braak 138 & Looman 1986; Ovaskainen et al. 2017). Finally, and perhaps most importantly, a direct 139 biological interpretation of parameter estimates in linear models becomes increasingly difficult as one moves from unimodal and symmetric distributions (ter Braak & Looman, 1986; Jamil & ter Braak, 2013) to skewed distributions (Huisman et al., 1993), making the tests of hypothesis on global biodiversity patterns particularly challenging. For example, Huisman et al. (1993) proposed several non-linear models to characterize several features of species' response curves; however, species' environmental indicator values, range size or distribution skewness are difficult to understand altogether following these model structures.

The field of ecology has quickly moved towards mechanistic and process-based approaches 147 to understand species' distributions (Warton et al., 2015). This has resulted in a plethora 148 of models accounting for several biotic and abiotic factors into the predictions of species 149 co-occurrence. Here, we instead rethink traditional modelling approaches and develop a 150 conceptually simple—and yet statistical and computationally complex—statistical framework to revisit some classic hypothesis in ecology and biogeoraphy. In particular, we develop 152 a Bayesian hierarchical model that accounts for all prior information that we have regard-153 ing the distribution of alpine plant species along an elevation gradient in the Swiss Alps, 154 including expert knowledge on species environmental indicator values, range sizes, and plant 155 physiology. We start by considering species' response curves as Gaussian distributed, and then we adapt our model to allow for skewed and long-tailed distributions. Using this sta-157 tistical framework, we are able to compare the basic properties of the realized niches of 158 multiple species, testing for the existence of general biogeographical patterns. First, we test 159 for the Rapopor's rule, which predicts a positive relationship between range size and eleva-160 tion (Stevens, 1992). While this pattern has been largely studied for multiple systems and 161 across gradients (McCain & Knight, 2013); contrasting evidence suggests this rule not to be 162 pervasive across species (Ribas & Schoereder, 2006; Bhattarai & Vetaas, 2006; McCain & 163 Knight, 2013). Our results not only allow us to properly test the existence of this geographi-164 cal pattern, but they also showcase variation in how different types of species, such as native 165 or neophytes, might respond to an environmental gradient. Second, we study whether or not 166 species' distributions show steeper declines towards stressful conditions, testing the so-called 167 abiotic stress limitation hypothesis (ref). Normand et al. (2009) tested this for vegetation 168 data using Huisman et al.'s statistical models for several independent species, finding no clear 169 support for such a hypothesis. Our results are able to shed light on this geographical pattern 170 as well as to highlight the degree to which different species will showcase different levels of 171 decline towards stressful conditions. Specifically, we are able to link plant physiological traits to the skewness of their distributions. Overall, we use models that are solely constrained 173 by the empirical information that we truly have regarding our system, relaxing as much as possible the structural constrains of the statistical framework. Using these models, we are able uncover the approximate shape of empirical plant distributions and answer fundamental questions regarding the way systems of many species are distributed along environmental gradients.

179 Methods

180 Empirical data

We studied the distribution of alpine plant communities along an elevation gradient. To do
so, we combined two different datasets: i) one describing the co-occurrence of species across
multiple open grasslands in the Swiss Alps, and ii) an extensive floristic database containing
environmental and physiological traits for all vegetation across Switzerland (Landolt *et al.*,
2010).

$_{186}$ Distribution data

We used data describing the distribution of 798 species across 912 sites covering most of the 187 mountain region of the Western Alps in the Canton de Vaud (Switzerland; Scherrer & Guisan 188 2019). Each of these sites is a 8×8 m plot placed somewhere along an elevation range from 189 375 m to 3210 m. In all sites, presence/absence data as well as Braun-Blanquet abundance-190 dominance classes were recorded for all species. Additionally, we used meteorological data 191 provided by Scherrer & Guisan (2019), containing multiple variables characterizing the cli-192 mate in each site at high spatial resolution (25 m). This dataset was compiled based on 193 30 years (1961–1990) of records from national weather stations. Since most of the data is highly correlated, we calculated the main axes of variation of the following scaled variables: 195 daily minimum, maximum and average temperature; sum of growing degree-days above 5°C; 196 mean temperature of wettest quarter; annual precipitation, precipitation seasonality, and 197 precipitation of driest quarter (see Supplementary Methods; Supplementary Fig. 1). 198

99 Floristic data

To complement the aforementioned distribution data, we used a floristic database of most 200 vegetation across Switzerland. This database was build based on expert knowledge and 201 field experience of botanists and ecologists, and contains information regarding species' 202 environmental preferences and physiological traits. Species' environmental preferences in 203 this database can be used to inform distribution models—e.g. as an informative prior in a 204 Bayesian framework. These are characterized following the ecological indicator values devel-205 oped by Landolt et al. (2010), providing both an estimate of the average conditions in which 206 a species can be found as well as a broad description of their range of variation. These values 207 are provided for a range of 10 climatic variables, including temperature, continentality, light 208 conditions, as well as moisture, acidity and nutrient content of the soil (see a full list and 209 description of the ecological indicators in the Supplementary Table 1; Landolt et al. 2010). 210 On the other hand, the information regarding species' physiological traits represents gen-211 eral descriptions of species' growth and life strategies—examples include their growth forms, 212 nature of the storage organs, dispersal ability and pollinator agents. In total, we identify 213 more than 120 binary traits that characterize the physiology of species (see a full list and 214 description of the ecological indicators in the Supplementary Table 1; Landolt et al. 2010). 215 Finally, and in addition to species' environmental preferences and physiological traits, the 216 floristic data also contains information on species types (e.g. identifying those species that 217 are neophytes) and change tendency (e.g. indicating species that have shown decline or in-218 crease in their populations over the recent decades). We describe this information in more 219 detail in the Supplementary Table 1. 220

221 Baseline model

There is a long list of model structures well suited to characterize species' distributions (see Norberg et al. 2019). As a baseline model, however, we were interested in a hierarchical model that does not make any assumptions regarding the shape of the distributions, and yet explicitly incorporates all information that we have regarding plant's environmental preferences. More specifically, we wanted to account for the climatic indicator values and range of variation registered in the floristic database for all plants in our dataset. These two

values provide basic information regarding plant's optimal environmental conditions and width of their distributions.

230 Response curve

To choose an appropriate response curve, we first need to agree on what we truly know 231 about the system. Given the prior information that we have about the system, we know 232 that species occupy specific geographic ranges; therefore, we know that their distributions 233 have finite variance. While we could also assume that many other factors might influence species' presence in a given site—e.g. the biotic interactions among specie in the site—we do 235 not necessarily have an a priori expectation of how exactly these factors will influence the 236 shape of species' distributions. Therefore, for this baseline model, if all that we are willing to 237 assume about species' realized niches is that these have finite variance, the most conservative 238 assumption and the safest bet—i.e. the one with the largest entropy—is that they follow a 239 Gaussian distribution (Fig 1a). That is, given the presence/absence or abundance y_{ij} of any species i in any given site j, and an environmental variable x_j , we can define species' 241 responses to the environment as

$$y_{ij} \sim F(p_{ij})$$

$$\log(p_{ij}) = -\alpha_i - \gamma_i (x_j - \beta_i)^2,$$
(1)

where F is the likelihood function, and α_i , β_i^k , and γ_i describe amplitude of the probability p_{ij} , species' average climatic suitability and range of variation along the an environmental gradient, respectively. Notice that F characterizes a Binomial distribution when considering binary data, and it characterizes an ordered categorical likelihood function when we consider Braun-Blanquet abundance-dominance classes as response variables (see the full description of both models in the Supplementary Methods). For the sake of simplicity, we use only one environmental variable to characterize species' probability distribution. That said, this model can easily be generalized to account for multiple predictors (see Supplementary Methods).

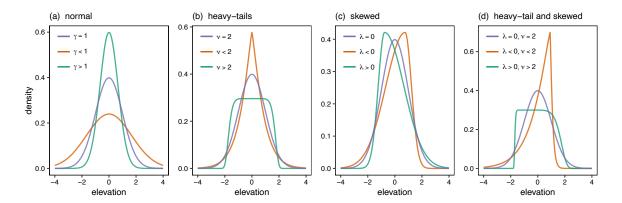


Figure 1: Different response curves. Panel (a) shows the probability density function characterized by Eq. (1) for different values of γ , when $\mu=0$. Panel (b) shows the probability density function characterized by Eq. (5) for different values of ν , when $\gamma=1$ and $\mu=0$. Panel (c) shows the probability density function characterized by Eq. (6) for different values of λ , when $\gamma=1$ and $\mu=0$. Panel (d) shows the probability density function characterized by Eq. (7) for different values of λ and ν , when $\gamma=1$ and $\mu=0$. Notice that for each case, we chose α values that normalize the probability distributions.

251 Model priors

The model structure described above allows us to explicitly incorporate all prior knowledge that we have regarding species' distributions contained in the floristic database. To do so, we define the prior distributions for the parameters in model (1) as:

$$\beta_{i} \sim \text{MVNormal}\left(\hat{\beta}, \Sigma^{\beta}\right)$$

$$\log(\gamma_{i}) \sim \text{MVNormal}\left(\hat{\gamma}, \Sigma^{\gamma}\right)$$

$$\log(\alpha_{i}) \sim \text{Normal}\left(\hat{\alpha}, \sigma_{\alpha}\right)$$

$$\hat{\beta}, \hat{\gamma}, \hat{\alpha} \sim \text{Normal}\left(0, 1\right)$$

$$\sigma_{\alpha} \sim \text{Exponential}\left(1\right)$$
(2)

where parameters γ_i and β_i are expressed as multivariate normal distributions—i.e. Gaussian processes—such that Σ^{β} and Σ^{γ} are variance-covariance matrices describing species' similarity in terms of their average climatic suitability and range of variation along the different environmental gradients, respectively. We define these variance-covariance matrices as follows:

$$\Sigma_{ij} = \eta \exp\left(-\rho D_{ij}^{2}\right) + \delta_{ij}\sigma,\tag{3}$$

where Σ_{ij} characterizes the covariance between any pair of species i and j, and δ_{ij} is the Kronecker delta. Notice that such a covariance structure declines exponentially with the 263 square of a distance matrix D_{ij} , which characterize differences between species computed 262 using our prior information. In the floristic database, this information is represented by the set of ordinal traits specified for the different species. While there are many different ways to turn ordinal data into distance matrices, we choose to use a mixed-membership stochastic 265 block model because it allows us to deal with cases of missing data (see Supplementary 266 Methods for extended details; Godoy-Lorite et al. 2016). In each covariance matrix, the hyperparameter ρ determines the rate of decline of the covariance between any two species, and η defines its maximum value. The hyperparameter σ describes the additional covariance 269 between the different observations for any given species. For all these hyperparameters, we 270 choose weekly informative priors such that $\sigma, \eta \sim \text{Exponential}(1)$ and $\rho \sim \text{Exponential}(0.5)$.

$_{272}$ Alternative variance-covariance structures

The model structure defined above allows us to test how different sources of information characterize each of the different parameters. Specifically, we can do this by modifying Eq. (3). For example, imagine that we have multiple matrices D^k characterizing species' differences along different axis of variation—e.g. two matrices characterizing physiological and environmental traits. One can modify Eq. (3) for a particular parameter—e.g. β_i —such that

$$\Sigma_{ij} = \eta \exp\left(-\sum_{k} \rho_k D_{ij}^{k^2}\right) + \delta_{ij}\sigma, \tag{4}$$

where now ρ_k are separate relevance hyperparameters for each distance matrix in the total variance of β_i .

Sampling the posterior

We generated the posterior samples for the Bayesian models with the Hamiltonian Monte
Carlo algorithm implementation provided by the R package 'rstan' to (Stan Developent
Team, 2021). Sampling models like the ones described above can be computationally very
intensive. This is especially true when using ordered categorical likelihood functions (see
Stan Development Team 2021). Therefore, we focus on those species for which we have at

least 20 occurrences when modelling binary data. When using ordinal data, we limit our study to a random subset of a 100 species of those with at least 20 occurrences to facilitate the sampling of the models.

To test the performance of the model as well as our choice of prior distributions, we modelled simulated data and compared the sampled posterior distributions to the data-generating parameters (see Supplementary Methods; Supplementary Fig. 2). Notice that using the link function in Eq. (1) could cause problems when sampling the model, and some adjustments need to be made when specifying the model (see Supplementary Methods and the Code Availability section).

$_{\scriptscriptstyle{296}}$ Modifying the baseline model

We proposed a baseline model that is naive regarding how the data is distributed, and yet accounts for all prior information that we have about the system. Now, we want to modify this model to test the extent to which empirical species' distributions showcase different properties, while preserving both the interpretation of the parameter estimates and prior information. To propose new species' response curves, we follow three criteria: (i) the probability distribution must have a defined variance and mean, (ii) the Gaussian shape must be a special case of the probability distribution, and (iii) there must be a re-parametrization of the model that allows us to keep the same prior information and interpretable parameters.

305 Heavy-tail response curve

Distributions with heavy-tails are very common across fields, as they can capture processes such as seasonality (e.g. in communications patterns; Malmgren et al. 2008) or some stochastic events (e.g. in the spread of infectious diseases; Wong & Collins 2020). Indeed, heavy-tail distributions are pervasive in ecology; for example, species' dispersal patterns have been shown to have heavy-tails due to natural variability among individuals (Petrovskii et al., 2009). Therefore, one might expect these properties to also emerge in alpine communities, where seasonality and dispersal patterns are crucial factors determining species' distributions. To accommodate this feature into our baseline model, we could consider a response curve that follows a generalized error distribution, since the Gaussian shape is a special case

of it and contains parameter that regulates the level of kurtosis (Fig 1b). In particular, we can adapt Eq. (1) to present this non-linear form as follows:

$$\log(p_{ij}) = -\alpha_i - \gamma_i' |x_j - \beta_i|^{\nu_i}, \tag{5}$$

where $\gamma_i' = g(\gamma_i, \nu_i)$, and ν_i is a parameter that describes the kurtosis of the distribution, which we define as $\nu_i \in (1, \infty)$. Following this, we choose an adaptive prior for this set of new parameter such that $\log(\nu_i - 1) \sim \text{Normal}(\hat{\nu}, \sigma_{\nu})$, where $\hat{\nu} \sim \text{Normal}(0, 1)$ and $\sigma_{\nu} \sim \text{Exponential}(2)$. Given the relationship between γ_i' and γ_i , we can re-parametrize the model and follow Eq. (2) to define the prior distributions (see Supplementary Table 2; Code Availability section; Nadarajah 2005). Notice that the Gaussian distribution will naturally emerge when $\nu_i = 2$.

fil the selection criteria described above. For example, the non-standardized Student's t-distributions is an interesting distribution because, as opposed to the generalized error distribution, it allows for heavy tails without generating a cusp at the center (see Fig 1b). However, we avoided using the non-standardized Student's t-distributions because it does not allow for tails that are lighter than normal (e.g. $\nu_i > 2$ in Eq. 5; Fig 1b), and the sampling of the model can be somewhat more challenging (ref).

331 Skewed response curve

When species experience abiotic or biotic pressures that increase or decrease along an en-332 vironmental gradient, one might expect their distributions to be skewed in one direction. 333 Likewise, this same skewed shape can also emerge as a result of species' asymmetric envi-334 ronmental tolerance. One way to accommodate this feature to our models is by considering 335 a skewed normal distribution (Supplementary Figure XX). As for the two cases described 336 above, the Gaussian is a special case of this distribution, and it contains a parameter that 337 controls for the level and direction of 'skewness' (Fig 1c). Importantly, this distribution 338 presents normal-like tails; therefore, the added skewness does not make additional assumptions regarding how species are distributed along the gradient. To test for the existence of this feature, we modified Eq. (1) as

$$\log(p_{ij}) = -\alpha_i - \gamma_i' \left(\frac{x_j - \beta_i'}{1 + \lambda_i \operatorname{sgn}(x_j - \beta_i')} \right)^2, \tag{6}$$

where $\gamma_i' = q_1(\gamma_i, \nu_i, \lambda_i)$, $\beta_i' = q_2(\gamma_i, \beta_i, \nu_i, \lambda_i)$, and λ_i is a parameter that describes the skewness of the distribution such that $\lambda_i \in (-1, 1)$. The function $\operatorname{sgn}(x)$ characterizes the sign function. We chose λ_i to have an adaptive prior such that $\operatorname{logit}\left(\frac{\lambda_i+1}{2}\right) \sim \operatorname{Normal}\left(\hat{\lambda}, \sigma_{\lambda}\right)$, where $\hat{\lambda} \sim \operatorname{Normal}(0, 1)$ and $\sigma_{\lambda} \sim \operatorname{Exponential}(1)$. Notice that this model can be reparametrized following q_1 and q_2 , allowing us to set the rest of the prior distributions as described for the baseline model (see Supplementary Table 2; Code Availability section). In this case, the Gaussian distribution is a special case of Eq. (6) when $\lambda_i = 0$ (Ashour & Abdel-hameed, 2010).

350 Heavy-tail and skewed response curve

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Finally, one could consider a response curve with both kurtosis and skewness. A convenient way to achieve this is by using a response curve that follows a skewed generalized error distribution (also see XX). This is a combination of the two distributions described above, containing two parameters that control for both the level and direction of kurtosis and skewness (Fig 1d). The skewed generalized error distribution can be considered by modifying the species' response curve in Eq. (1) as

$$\log(p_{ij}) = -\alpha_i - \left(\frac{\gamma_i' |x_j - \beta_i'|}{1 + \lambda_i \operatorname{sgn}(x_j - \beta_i')}\right)^{\nu_i}, \tag{7}$$

where $\gamma_i' = f_1(\gamma_i, \nu_i, \lambda_i)$, $\beta_i' = f_2(\gamma_i, \beta_i, \nu_i, \lambda_i)$, ν_i and λ_i are parameters that describe the kurtosis and skewness of the distribution, respectively. We define ν_i , λ_i and their prior distributions as in Eq. 5 and 6, respectively. Again, we can re-parametrize the model following f_1 and f_2 , and set the rest of the prior distributions as in the baseline model (see Supplementary Table 2; Code Availability section). Notice that the generalized error distribution (Eq. 5) and the skew normal distribution (Eq. 6) are special cases of Eq. (7) when $\lambda_i = 0$ and $\nu_i = 2$, respectively.

One way to test the extend to which species' physiological traits inform the kurtosis or

skewness of the distribution of any species i is by characterizing ν_i or λ_i as Gaussian processes.

That is, we can consider the prior distribution for these parameters as multivariate normal
with a variance covariance matrix Σ . As described before, this variance-covariance matrix
can be estimated following Eq. (3), and its structure can shed light on how informative the
physiological traits are. Likewise, we could instead use Eq. (4) and test the effectiveness of
other prior information in explaining the skewness of species' distributions.

1 Results

We studied the distribution data to characterize species' realized niches. To do so, we selected the main axis of variation of all environmental variables (Supplementary Fig. XX). Then, using the presence and absence of species across sites as the response variable, we 374 sampled the posterior distributions of the baseline model, accounting for the information in 375 the floristic database regarding species' indicator values and range of variation. Studying the relationship between the mean and variance of the distributions, we found clear evidence of the Rapopor's rule. Specifically, we found that species' range sizes decline with eleva-378 tion and temperature (i.e. β_i and γ_i in the baseline model are negatively correlated; Fig. 2; 379 Supplementary Fig. XX). The relationship was also found when using instead ordinal data 380 (Supplementary Fig. XX), but this pattern was not present along the second axis of varia-381 tion of the environmental variables (Supplementary Fig. XX). The comparison between the other parameter estimates revealed additional, somewhat more expected, relationships. In particular, we found the amplitude of distributions to be positively and negatively correlated 384 with the mean and the variance along the elevation gradient, respectively (Supplementary 385 Fig. XX). 386

Maintaining the symmetry of species' distributions, we then allowed the kurtosis—or shape of the tails—of these to vary in different ways. To do so, we changed the response curve of our Bayesian model to follow a generalized error distribution (Eq. 5). A comparison of the WAIC values showed this non-linear regression to outperform the baseline model (Supplementary Fig. XX). Studying the resulting posterior distributions, we found the parameter controlling for the kurtosis to be centred around $\nu_i \sim 2$, which corresponds to a distribution with longer tails than a Gaussian (Fig. 3). However, such parameter estimates displayed a lot of variation

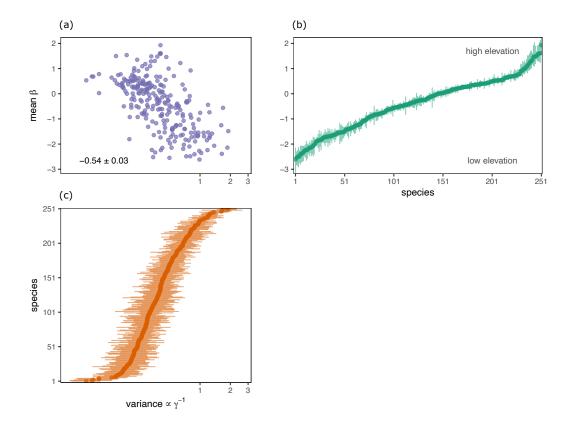


Figure 2: Relationship between mean and variance of species' distributions. Posterior distributions for parameters β_i and γ_i from Eq. (1) across species, and the relationship between them. Panel (a) describes the relationship between range size and elevation. Every dot represents the relationship between the mean values for the β_i and γ_i estimates of the different species. The value in the top-right corner of the plot displays the Pearson's correlation between these parameters calculated across samples of the posterior distributions. Panel (b) describes the β_i posterior distribution estimated for all species. Panel (c) describes the γ_i posterior distribution estimated for all species. In (b) and (c), the points represent the mean of the posterior distributions, and the corresponding lines characterize the 89% confidence intervals.

across species, which might indicate that the shape of the tails of the distribution is speciesspecific (Supplementary Fig. XX). Similarly, using Eq. (6), we next studied the skewness of species distributions. Based on the estimates for the WAIC values, this model outperformed the previous models (Supplementary Fig. XX), which sheds light on the naturally skewed nature of species' distributions. Perhaps most importantly, studying the mean value of the skewness across species (i.e. parameter $\hat{\lambda}$), we found that species' distributions generally present steeper declines towards higher elevations (Fig. 3).

Finally, we used a model that allowed for heavy-tails and skewed response curves (Eq. 7)
to understand the relationship between the kurtosis and skewness of distributions. This
model outperformed the rest, presenting Akaike weights close to 1 when comparing models

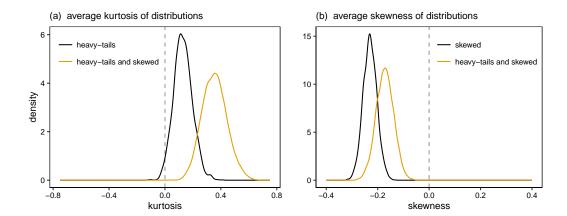


Figure 3: Average kurtosis and skewness of species' distributions. Calculated using the posterior distributions of parameters $\hat{\nu}$ and $\hat{\lambda}$ from the models (see Supplementary Table XX), the two panels describe the average (a) kurtosis and (b) skewness of distributions, respectively. Panel (a) displays the results obtained by using a response curve that follows a generalized error distribution. Panel (b) displays the results obtained by using a response curve that follows a skewed normal distribution. In both case, the red dotted line indicates the conditions by which species are normally distributed along the environmental axis.

(Supplementary Fig. XX). Comparing the posterior distribution for parameters ν_i and λ_i , we revealed the general constrains on the distribution of alpine plants studied. That is, we 405 uncovered the relationship between the mean, variance, amplitude, kurtosis and skewness of 406 empirical distribution of plants along an elevational gradient (Figs. XX and YY). Moreover, 407 different types of species seem to present characteristically different distributions (Fig. XX). 408 For example, invasive species... Notice the nature of these results does not depend on the presence or absence of a species at the edge of the sampling area (bootstrapping... Sup-410 plementary Fig. XX). Likewise, the posterior distributions obtained using binary data were 411 strikingly similar to those obtained using categorical data. This is important because... is 412 model identified similar relationships between parameters as the baseline model (Supplemen-413 tary Fig. XX), and measured similar average levels of kurtosis and skewness than the the other two models (Fig. 3). 415

416 Discussion

418

Structure of the discussion section:

1. Summary of results.

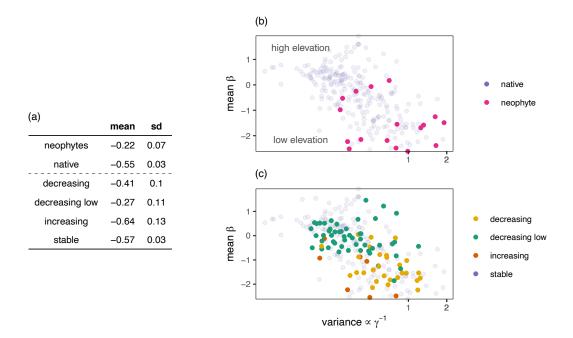


Figure 4: Universality of the relationship between mean and variance of species' distributions. Comparison between how different types of species are mapped in Fig. 2a. Panel (a) describes the correlation coefficient between β_i and γ_i for each type of species. Panel (b) shows the differences between neophytes and native species in the way these are distributed along the environmental gradient. Panel (c) shows the same differences for species that have decreased, decreased in low elevations, increase and remain stable over the last decades (see Supplementary Table 1 for further details). I might move this to the Supplementary Information.

- 2. Proper test of Rapopor's hypothesis. Different species follow different biogeographical patterns.
- 3. Proper test of skewed towards high altitude. Is species' physiology informative to explain the pattern?
- 423 4. What is the true shape of species' distributions? These display heavy-tail and skewed distributions.
- 5. Future directions. Missing bimodal curves. Using this information to understand where jSDMs estimate interactions between species. Further test of the ability of traits to predict those parameter estimates.
- First, we test for the Rapopor's rule, which predicts a positive relationship between range size and elevation (Stevens, 1992). While this pattern has been largely studied for multiple systems and across gradients (McCain & Knight, 2013); contrasting evidence suggests this

rule not to be pervasive across species (Ribas & Schoereder, 2006; Bhattarai & Vetaas, 431 2006; McCain & Knight, 2013). Our results not only allow us to properly test the existence 432 of this geographical pattern, but they also showcase variation in how different types of 433 species, such as native or neophytes, might respond to an environmental gradient. Second, we 434 study whether or not species' distributions show steeper declines towards stressful conditions, 435 testing the so-called abiotic stress limitation hypothesis (ref). Normand et al. (2009) tested 436 this for vegetation data using Huisman et al.'s statistical models for several independent 437 species, finding no clear support for such a hypothesis. Our results are able to shed light 438 on this geographical pattern as well as to highlight the degree to which different species will showcase different levels of decline towards stressful conditions. Specifically, we are able to link plant physiological traits to the skewness of their distributions.

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