

Foundations of ecological and evolutionary change

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Abstract

Biological evolution is realised through the same mechanisms of birth and death that underlie change in population density. The deep interdependence between ecology and evolution is well-established, and recent models focus on integrating eco-evolutionary dynamics to demonstrate how ecological and evolutionary processes interact and feed back upon each other. Nevertheless, a gap remains between the logical foundations of ecology and evolution. Both population ecology and evolution have fundamental equations that define how the size of a population (ecology) and the average characteristic within a population (evolution) change, respectively, over time. These fundamental equations are a complete and exact description of change for any closed population, but how they are formally linked remains unclear. We link the fundamental equations of population ecology and evolution with an equation that sums how individual characteristics interact with individual fitness in a population. From this equation, we derive the fundamental equations of population ecology and evolutionary biology (the Price equation). We thereby identify an overlooked bridge between ecology and biological evolution. Our unification formally recovers the equivalence between mean population growth rate and evolutionary fitness and links this change to ecosystem function. We outline how our framework can be used to further develop eco-evolutionary theory.

Key words: Ecology, Evolution, Eco-Evolutionary Theory, Fundamental Theorem, Price Equation, Population Growth

Introduction

Theoretical unification is a powerful tool for scientific advancement. Such unification has been a major goal in scientific research throughout history (Smocovitis 1992; Kitcher 1993), and its value is perhaps most evident in reconciling unconnected models and revealing new and unexpected empirical predictions. In evolutionary biology, the Price equation (Box 1) provides a unifying framework for evolutionary theory by exhaustively and exactly describing evolutionary change for any closed evolving population (Price 1970; Luque 2017; Lehtonen et al. 2020). The Price equation is therefore fundamental, in the sense that it binds together all of evolutionary theory by formally defining what evolutionary change is and is not (Price 1970; Rice 2004; Gardner 2008; Frank 2017; Luque 2017; Luque and Baravalle 2021). Using this formal definition, the scope of, and relationships among, sub-disciplines within evolutionary theory can be clarified. For example, fundamental equations of both population and quantitative genetics can be derived from the Price equation (Queller 2017). This provides conceptual clarity by demonstrating the logical consistency of different theoretical frameworks within evolutionary biology. Our aim here is to propose an equation that extends this conceptual clarity to include population size change and thereby provide a formal and exact definition for joint ecological and evolutionary change.

In biological populations, ecological change is caused by the same processes of individual birth and death that cause evolutionary changes in allele frequencies and phenotypes (Turchin 2001; Connor and Hartl 2004; Barfield et al. 2011). As with evolution, a fundamental equation can exhaustively and exactly define population change. Unlike the Price equation, this fundamental equation is perhaps self-evident. Population change is simply the addition of individuals minus the removal of individuals from current population size (N_t), which recovers the new population size (N_{t+1} ; Box 2). By definition, this relationship of $N_{t+1} = N_t + \text{Births} - \text{Deaths}$ applies to any closed population. Turchin (2001) argues that general principles are needed to establish a logical foundation for population ecology, and this simple birth and death model and the consequences that logically follow from it (e.g., exponential population growth) are fundamental to population ecology. Any unifying definition of joint ecological and evolutionary change must be able to derive both the Price equation and this birth and death model.

The union of ecological and evolutionary processes has long been recognised (e.g., [Darwin 1859](#); [Fisher 1958](#); [Pelletier et al. 2009](#)), but the rise of eco-evolutionary models, which incorporate both, is relatively recent following a widespread recognition that ecology and evolution can happen on similar timescales ([Govaert et al. 2019](#); [Yamamichi et al. 2023](#)). Currently, a universally recognised formal definition of eco-evolutionary change is lacking, with some theoreticians broadly interpreting “eco-evolutionary dynamics” to allow for a separation of ecological and evolutionary timescales ([Lion et al. 2023](#)) and others advocating for a more narrow interpretation in which no such separation is permitted ([Bassar et al. 2021](#)). [Bassar et al. \(2021\)](#) identify two types of eco-evolutionary models that follow from these interpretations. The first type uses separate equations to model population size change versus evolutionary change, thereby allowing for any number of ecological, evolutionary, or environmental feedbacks (e.g., [Lion 2018](#); [Patel et al. 2018](#); [Lion et al. 2023](#)). The second type models population demographics as functions of quantitative traits with ecological and evolutionary change following from demographic processes and trait distributions operating on the same timescale (e.g., [Barfield et al. 2011](#); [Simmonds et al. 2020](#); [Jaggi et al. 2024](#)). Both model types can be very general, but like all predictive models, they rely on simplifying assumptions for tractability ([Levins 1966](#); [Luque 2017](#)). These simplifying assumptions are often grounded in the Price equation to demonstrate accuracy and logical consistency when modelling evolutionary change (e.g., [Coulson and Tuljapurkar 2008](#); [Barfield et al. 2011](#); [Rees and Ellner 2016](#); [Lion 2018](#)). For example, [Barfield et al. \(2011\)](#) link their model back to [Price \(1970\)](#), which they consider to be a “universal law of evolution”, to place their conclusions concerning stage structured evolution in the broader context of evolutionary theory. The role of fundamental equations is therefore important for unifying theory ([Luque and Baravalle 2021](#)), and we believe that a fundamental equation of eco-evolutionary change has been curiously overlooked.

We present an equation from which the fundamental equations of ecology and evolutionary biology can be derived. Derivation follows by adding assumptions that are specific to population ecology or evolution in the same way that key equations of population genetics or quantitative genetics can be derived from the Price equation by restricting the domain of interest (e.g., to allele frequencies in the case of population genetics, or to continuous phenotypes in the case of quantitative genetics, [Queller 2017](#)). We propose our equation as a formal definition of eco-evolutionary change.

Box 1: The Price equation is an abstract formula to represent evolutionary change. Formulated originally in the early 1970s by George Price (Price 1970, 1972), it postulates some basic properties that all evolutionary systems must satisfy: change over time, ancestor and descendant relations, and a character or phenotype (Rice 2004). Using simple algebraic language, the Price equation represents evolutionary change with the predominant notation,

$$\bar{w}\Delta\bar{z} = \text{Cov}(w, z) + \text{E}(w\Delta z).$$

In the above equation, $\Delta\bar{z}$ is the change in the average character value z over a time step of arbitrary length, w is an individual's fitness, and \bar{w} average population fitness. On the right-hand side of the equation, the first term is the covariance between a character value z and fitness w , which reflects \bar{z} change attributable to differential survival and reproduction. The second term is the expected value of Δz , which reflects the extent to which offspring deviate from parents in z (Rice 2004; Okasha 2006; Frank 2012). A more specific version of the covariance term was already known within the quantitative and population genetics tradition (Robertson 1966), usually representing the action of natural selection. The Price equation adds an expectation term and abstracts away from any specific mechanisms of replication or reproduction, or mechanisms of inheritance. Its definitional nature and lack of substantive biological assumptions has been portrayed both as a strength (Rice 2004; Frank 2012; Luque 2017; Baravalle and Luque 2022), and its greatest weakness. The abstract nature of the Price equation places it at the top of the hierarchy of fundamental theorems of evolution from which the rest (Robertson's theorem, Fisher's fundamental theorem, breeder's equation, Hamilton's rule, adaptive dynamics, etc.) can be derived (Lehtonen 2016, 2018, 2020; Queller 2017). This abstractness is also key to developing a more general view of evolution (Rice 2020; Luque and Baravalle 2021; Edelaar et al. 2023). In contrast, some researchers consider the Price equation just a triviality (even tautological), and useless without further modelling assumptions (van Veelen 2005; van Veelen et al. 2012). The

debate remains open (van Veelen 2020; Baravalle et al. 2025).

Box 2: The number of individuals in any closed population (N) at any given time ($t + 1$) is determined by the number of individuals at t (N_t), plus the number of births ($Births$) minus the number of deaths ($Deaths$),

$$N_{t+1} = N_t + Births - Deaths.$$

This equation is necessarily true for any closed population. Despite its simplicity, it is a general equation for defining population change and a starting point for understanding population ecology. Turchin (2001) notes that a consequence of this fundamental equation is the tendency for populations to grow exponentially (technically geometrically in the above case where time is discrete). This inherent underlying tendency towards exponential growth persists even as the complexities of real populations, such as structure, stochasticity, or density-dependent effects are added to population models (Turchin 2001). Given the assumption that all individuals in the population are identical, a per capita rate of birth, $Births = bN_t$, and death, $Deaths = dN_t$, can be defined. Rearranging and defining $\lambda = 1 + b - d$ gives, $N_{t+1} = N_t\lambda$. Here λ is the finite rate of increase (Gotelli 2001), and note that because $0 \leq d \leq 1$, $\lambda \geq 0$. Verbally, the change in size of any closed population equals its existing size times its finite rate of increase.

A foundation for biological evolution and population ecology

To unify biological evolution and population ecology, we must reconcile the Price equation (Box 1) with the general equation for population change (Box 2). The Price equation is critical for partitioning different components of biological change (Price 1970; Frank 1997; Gardner 2008; Luque 2017; Queller 2017; Lehtonen

2018, 2020). It has also been highly useful for integrating evolutionary theory across disciplines (Fox 2006; Brantingham 2007; MacCallum et al. 2012; Frank 2015; Godsoe et al. 2021; Ulrich et al. 2024). These properties would seem to make it an intuitive starting point for a logical foundation of ecology and evolution, perhaps through some kind of mathematical equivalence (Page and Nowak 2002) or addition of terms (Collins and Gardner 2009), or through the use of its recursive structure (Kerr and Godfrey-Smith 2009; Frank 2012). But despite its flexibility, the Price equation still relies on relative frequencies, which must by definition sum to one (Frank 2015). This is because the Price equation describes the average change in a population; the frequency of entities is scaled thereby conserving total probability (Frank 2015, 2016). But to recover the fundamental principle of exponential population growth (Turchin 2001), this scaling must be avoided.

We therefore begin with the most fundamental axioms underlying the ecology and evolution of living systems (Rice 2004; Rice and Papadopoulos 2009). In such systems, diversity is discontinuous, in the sense that living systems are composed of discrete entities including individual organisms and groups of organisms (Dobzhansky 1970). Our framework is general enough that entities can be anything discrete, but we will focus on each entity i as an individual organism. Individuals give rise to new individuals through birth such that β_i is the number of births attributable to i . Individuals are removed from the population through death such that δ_i is an indicator variable that takes a value of 1 (death of i) or 0 (persistence of i). All individuals are defined by some characteristic z_i , and Δz_i defines any change in z_i from one time step t to the next $t + 1$. The total number of individuals in the population is N . From this foundation, we can define Ω to be a summed characteristic across N entities,

$$\Omega = \sum_{i=1}^N (\beta_i - \delta_i + 1) (z_i + \Delta z_i). \quad (1)$$

This foundation of eco-evolutionary change is a statistical interaction between the demographic processes of birth and death ($\beta_i - \delta_i + 1$) and some characteristic of an individual ($z_i + \Delta z_i$). The value of Ω is a summation of individual characteristics and takes the same units as z . From eqn 1, we can derive the most fundamental equations of population ecology and evolutionary biology through an appropriate interpretation of z . Under more limited interpretations of z , we can also interpret Ω as a metric of ecosystem function.

Population ecology

To recover the general equation for population ecology (Box 2), we define z_i as the identity of i belonging to the population. In other words, we set $z_i = 1$ to simply indicate that i is one member of the population. In this restricted case, z is count, which takes the unit 1 (note that ‘individual’ is a label, not a unit, see [Newell and Tiesinga 2019](#)). Further, we assume that individuals do not change species by setting $\Delta z_i = 0$. In this case,

$$\Omega = \sum_{i=1}^N (\beta_i - \delta_i + 1).$$

We can now interpret Ω as the population size at $t + 1$, N_{t+1} . Summing up β_i , δ_i , and current individuals tallies up the total number of individuals in the next time step,

$$N_{t+1} = N_t + \text{Births} - \text{Deaths}. \quad (2)$$

If we assume that individuals are identical, then we can drop the subscript i such that $\beta_i = \beta$ and $\delta_i = \delta$ and note that $N\beta$, $N\delta$, and N are the total births, total deaths, and size of the population at t , respectively,

Summing from 1 to N , we can rewrite the above,

$$\Omega = N (1 + \beta - \delta).$$

If we define $\lambda = 1 + \beta - \delta$ (Box 2), then we can rewrite,

$$N_{t+1} = N_t \lambda. \quad (3)$$

We therefore recover the general equation for population ecology (eqn 2) and the fundamental tendency for populations to grow (or decline) exponentially (eqn 3, [Turchin 2001](#)).

Evolutionary biology

Recovering the Price equation requires a few more steps. We start by defining individual fitness,

$$w_i = \beta_i - \delta_i + 1. \quad (4)$$

In this definition, the longevity of the individual matters. By definition, an individual that survives from t to $t + 1$ has a higher fitness than one that dies, even if both have the same reproductive output. With this definition of fitness (eqn 4), we substitute,

$$\Omega = \sum_{i=1}^N (w_i z_i + w_i \Delta z_i). \quad (5)$$

We can break eqn 5 down further and multiply each side by $1/N$,

$$\frac{1}{N}\Omega = \frac{1}{N} \sum_{i=1}^N (w_i z_i) + \frac{1}{N} \sum_{i=1}^N (w_i \Delta z_i). \quad (6)$$

We can rewrite the terms on the right-hand side of eqn 6 as expected values and remove the subscripts,

$$\frac{1}{N}\Omega = E(wz) + E(w\Delta z). \quad (7)$$

Now we must consider the total conservation of probability (Frank 2015, 2016). In eqn 7, Ω is the total sum trait values (z_i) across the entire population at $t + 1$ divided by the number of individuals (N) in the population at t . But the size of the population can change from t to $t + 1$. To recover mean trait change for the Price equation (and therefore conserve total probability), we need to account for this change in population size. The left-hand side of eqn 7 describes contributions to the sum trait value from the new population at $t + 1$. But we cannot treat Ω/N as the mean of z at $t + 1$ (\bar{z}') because we need to weight N by the mean fitness of the population at t to account for any change in population size from t to $t + 1$.

For example, if mean fitness at t is 2 (i.e., $\bar{w} = 2$), then half as many individuals will have contributed to

Ω in $t + 1$ than would have if $\bar{w} = 1$ (i.e., there are N individuals at t and $2N$ individuals at $t + 1$). We
 176 therefore need to multiply the mean trait value \bar{z}' (at $t + 1$) by the mean fitness \bar{w} (at t) to recover the mean
 contribution of the N individuals at t to the total Ω (this has been applied in the context of population
 178 genetics by [Ewens 2014](#)). Consequently,

$$\Omega = N\bar{w}\bar{z}' \quad (8)$$

Equation 8 conserves the total probability and now clarifies Ω as a summed trait value, which has the same
 180 units as z and equals expected population growth at t times mean trait value at $t + 1$. This is consistent with
 the population ecology derivation from the previous section where $z_i = 1$ by definition, and $\Omega = N_{t+1}$ (note
 182 $\lambda = \bar{w}$). We can therefore rewrite eqn 7,

$$\bar{w}\bar{z}' = \text{E}(wz) + \text{E}(w\Delta z). \quad (9)$$

We can rearrange eqn 9 to derive the Price equation by expressing covariance as $\text{Cov}(X, Y) = \text{E}(XY) -$
 184 $\text{E}(X)\text{E}(Y)$, and therefore $\text{E}(XY) = \text{Cov}(X, Y) + \text{E}(X)\text{E}(Y)$. Substituting into eqn 9,

$$\bar{w}\bar{z}' = \text{Cov}(w, z) + \bar{w}\bar{z} + \text{E}(w\Delta z).$$

From here,

$$\bar{w}(\bar{z}' - \bar{z}) = \text{Cov}(w, z) + \text{E}(w\Delta z).$$

186 Since $\Delta\bar{z} = (\bar{z}' - \bar{z})$,

$$\bar{w}\Delta\bar{z} = \text{Cov}(w, z) + \text{E}(w\Delta z). \quad (10)$$

Equation 10 is the Price equation. From eqn 1, which describes fundamental birth and death processes in a population, we can therefore derive both the most fundamental model of population ecology (eqn 2; Box 2) and the fundamental equation of evolution (eqn 10; Box 1).

Ecosystem function

In some cases, Ω could also be interpreted as the total contribution of a population to ecosystem function. This is restricted to cases in which z is a characteristic defining an absolute quantity measured at the whole organism level such as biomass, seed production, carbon capture, flower visits, or nutrient consumption. In such cases, the sum across individuals gives a meaningful total quantity for the population (e.g., the total biomass or seeds produced in the population). When z is instead defined by relative organism-level measurements such wing loading, nutrient ratio, or diet composition, or when z is measured at a level of biological organisation below the organism (e.g., average cell volume or leaf surface area), Ω does not have a clear population-level interpretation. In such cases, eqn 1 still defines eco-evolutionary change; the interpretation of Ω by itself is just not as interesting, biologically.

Box 3 provides an instructive example of three plants with different fitnesses and carbon captures. We use this example to demonstrate a concrete example of our proposed framework.

Box 3: As an instructive example of our framework, consider a population of $N_t = 3$ annual plants in which individual total fruit mass (kg) is measured, and change is observed over a year. For all plants, $\delta_i = 1$ (because annual plants die between t and $t + 1$), and let plant fecundities be $\beta_1 = 1$, $\beta_2 = 1$, and $\beta_3 = 2$. Applying eqn 1 to population change such that $\Omega = N_{t+1}$, $z_i = 1$ and $\Delta z_i = 0$, $N_{t+1} = (1 - 1 + 1)(1 + 0) + (1 - 1 + 1)(1 + 0) + (2 - 1 + 1)(1 + 0) = 4$ (note $\bar{w} = 4/3$, so $N_t \bar{w} = 3(4/3) = 4$). Focusing next on the characteristic of total fruit mass, let $z_1 = 0.8$ kg, $z_2 = 1.0$ kg, and $z_3 = 1.5$ kg. Also let $\Delta z_i = 0.1$ for all plants to reflect a change in soil environment from t to $t + 1$. In this case, population fruit yield at $t + 1$ is $\Omega = (1 - 1 + 1)(0.8 + 0.1) + (1 - 1 + 1)(1.0 + 0.1) + (2 - 1 + 1)(1.5 + 0.1) = 5.2$ kg. At t , mean

fruit yield per plant was $(0.8 + 1.0 + 1.5)/3 = 1.1$ kg, but at $t + 1$, mean fruit yield per plant is $[(0.8 + 0.1) + (1.0 + 0.1) + 2(1.5 + 0.1)]/4 = 1.3$ kg. Note that $\text{Cov}(w, z) = (1/N) \sum_i^N (z_i - \bar{z})(w_i - \bar{w}) = (1/3) [(0.8 - 1.1)(1 - 4/3) + (1 - 1.1)(1 - 4/3) + (1.5 - 1.1)(2 - 4/3)] = 2/15$ and $E(w\Delta z) = 4/3 \times 0.1 = 2/15$, so applying the Price equation, $\bar{w}\Delta\bar{z} = 2/15 + 2/15 = 4/15$. Since $\bar{w} = 4/3$, multiplying both sides of the equation by $3/4$ returns $\Delta z = 0.2$, which is the mean difference in fruit yield between $t + 1$ and t . The framework expressed in eqn 1 thereby links population change, evolutionary change, and ecosystem function.

Discussion

An important aspect of scientific progress is the ability to connect disparate theories and models to show how specific empirical and theoretical models are logical (mathematical) consequences of more fundamental ones (Nagel 1961; Morrison 2000). Rather than making simplifying assumptions, as is the approach for specific ecological and evolutionary models, we focus on fundamental axioms that are universal to closed biological systems: discrete individuals, birth, death, and change over time. We define an abstract sum (Ω), to which all individuals within the population contribute. From the basic axiom that each individual is one member of a specific population ($z_i = 1$ and $\Delta z_i = 0$), we recover the most general equation of population ecology (Box 2) and the principle of exponential growth ($N_{t+1} = N_t\lambda$). By defining individual fitness (w_i) and applying the total conservation of probability to individual frequencies (Frank 2015, 2016), we recover the most fundamental equation of evolution, the Price equation (Box 1). Our eqn 1 thereby provides a foundation for defining eco-evolutionary change in any population.

The Price equation provides a complete and exact description of evolution in any closed evolving system (Price 1970; Frank 2012). It is derived by rearranging the mathematical notation defining changes in the frequencies and characteristics of any type of entity (e.g., individuals, alleles, Price 1970; Gardner 2008; Luque 2017). This derivation partitions total characteristic change into different components, making it possible to isolate evolutionary mechanisms (e.g., selection) and levels of biological organisation (e.g., group,

individual) (Frank 1995, 2012; Kerr and Godfrey-Smith 2009; Luque 2017; Okasha and Otsuka 2020). Because
 238 of its abstract nature and lack of any system-specific assumptions, the Price equation is not dynamically
 sufficient and makes no predictions about what will happen in any particular system (Gardner 2020). Its role
 240 is not to predict, but to formally and completely define and separate components of evolutionary change
 (Baravalle et al. 2025). The same is true of the general equation for population change (eqn 2), at least
 242 as we have used it here where it serves to define what population growth means in ecology. This equation
 formally and completely describes population change in terms of births and deaths. In eqn 1, we therefore
 244 have a fundamental equation from which we can derive complete ecological and evolutionary change in any
 closed biological population. Like all fundamental equations, our equation is necessarily abstract and not
 246 dynamically sufficient. We believe that it will be useful for eco-evolutionary theory in a similar way that
 the Price equation is useful for evolutionary theory: potentially facilitating specific model development and
 248 identifying new conceptual insights, unresolved errors, and sources of model disagreements (see below and
 Supporting Information S1).

250 Our unification recovers the equivalence between the finite rate of increase λ (Box 2) and population mean
 evolutionary fitness \bar{w} (Box 1). The population growth equation $N_{t+1} = N_t \lambda$ can always be rewritten as
 252 $N_{t+1} = N_t \bar{w}$. This specific equivalence has been proposed before (e.g., Lande 1976), as has the broader
 relationship between population growth rate and evolutionary fitness (Fisher 1930; Charlesworth 1980; Lande
 254 1982; Roff 2008; Lion 2018). We show this from first principles and clarify the relationship between fitness
 and population growth. Over an arbitrary length of time, fitness is properly defined as $w_i = \beta_i - \delta_i + 1$.
 256 Over an individual's lifetime (which, by definition, includes death), fitness is therefore β_i . Rates of change
 in ecology and evolution are reflected in the first and second statistical moments of fitness, respectively.
 258 Population growth rate reflects mean fitness \bar{w} , while the rate of evolutionary change reflects the variance in
 fitness $\text{Var}(w)/\bar{w}$ (i.e., Fisher's fundamental theorem, Frank 1997; Rice 2004; Queller 2017).

260 Our unification may also help explain, at least partially, some of the success of classical population genetic
 models. For decades, population genetics (and to some extent quantitative genetics) has been accused of
 262 being a reductionist view of evolution, reducing everything to changes in allele frequencies and abstracting
 away from individuals and their environments (the ecological interactions, MacColl 2011). This has been a

line of argumentation by some defenders of the so-called Extended Synthesis (Pigliucci 2009), especially in relation to niche construction (Odling-Smee et al. 2003). Famously, Mayr (1959) characterised population genetics as a simple input and output of genes, analogous to “the adding of certain beans to a beanbag and the withdrawing of others” (also called “beanbag genetics”). Historical critics of population genetics could not articulate a clear explanation for why it works so well despite all of its idealisations and simplifications. From the Price equation, we are able to recover classical population and quantitative genetic models (Queller 2017) and develop new ones (Rice 2004, 2020; Luque 2017; Lion 2018). Our eqn 1 contains ecology at its core, and we show how the Price equation logically follows from it after accounting for absolute population growth (eqn 8). We therefore conclude that population and quantitative genetic equations contain ecology (no matter how hidden), and the ecological nature of evolution is implicit in population and quantitative genetic models. We have focused on the dynamics of a closed population, and in doing so leave ecological and evolutionary change attributable to migration for future work. In population ecology, immigration and emigration can be incorporated by adding a term for each to the right-hand side of the equation in Box 2 (Gotelli 2001). In evolution, because the Price equation relies on mapping ancestor-descendant relationships, accounting for migration is more challenging. Kerr and Godfrey-Smith (2009) demonstrate how the Price equation can be extended to allow for arbitrary links between ancestors and descendants, thereby extending the Price equation to allow for immigration and emigration. Frank (2012) presents a simplified version of Kerr and Godfrey-Smith (2009) that allows some fraction of descendants to be unconnected to ancestors. In both ecology and evolution, accounting for migration is done through the use of additional terms on the right-hand side of the equations.

We believe our fundamental equation to be complete and exact for any closed population. It therefore implicitly includes any effects of density dependence on population growth (see Box 2), or any social effects on evolutionary change (see Box 1). Both of these effects can be made explicit by specifying how other individuals in a population affect birth and death of a focal individual. We demonstrate this by deriving more specific models of density dependent population growth and multi-level selection in Supporting Information S1. In Supporting Information S1, we also suggest how new theoretical insights might be gained by integrating density-dependent growth and social interactions into our framework.

We have shown that we can derive the fundamental equations of population ecology and biological evolution from a single unifying equation. Lastly, we propose our eqn 1 as a potential starting point for defining ecosystem function and further conceptual unification between ecology, evolution, and ecosystem function. The Price equation has previously been used to investigate ecosystem function (Loreau and Hector 2001; Fox 2006), but not with any attempt towards conceptual unification with evolutionary biology. For example, Fox (2006) applied the abstract properties of the Price equation to partition total change in ecosystem function into separate components attributable to species richness, species composition, and context dependent effects. This approach provides a framework for comparing the effects of biodiversity on ecosystem function in empirical systems (Fox 2006; Winfree et al. 2015; Mateo-Tomás et al. 2017). Instead, our eqn 1 defines Ω as total ecosystem function contributed by a focal population. It is therefore possible to investigate ecological, evolutionary, and ecosystem function change from the same shared framework.

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Author Contributions

314 Both authors came up with the question idea. Following many discussions between the authors, ABD
proposed the initial equation with subsequent exploration and development from both authors. Both authors
316 contributed to the writing.

Competing Interests

318 The authors declare no competing interests.

Data Availability

320 This work does not include any data.

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