

Foundations of ecological and evolutionary change

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Abstract

Biological evolution is realised through the same mechanisms of birth and death that underlie change in population density. The deep interdependence between ecology and evolution is well-established, but much theory in each discipline has been developed in isolation. While recent work has accomplished eco-evolutionary synthesis, a gap remains between the logical foundations of ecology and evolution. We bridge this gap with a new equation that defines a summed value for a characteristic across individuals in a population, from which the fundamental equations of population ecology and evolutionary biology (the Price equation) are derived. We thereby unify the fundamental equations of population ecology and biological evolution under a general framework. Our unification further demonstrates the equivalence between mean population growth rate and evolutionary fitness and links this change to ecosystem function. Finally, we outline how our proposed framework can be used to unify social evolution and density-dependent population growth.

Key words: Ecology, Evolution, Eco-Evolutionary Theory, Fundamental Theorem, Price Equation, Population Growth

Introduction

Theoretical unification of phenomena is a powerful tool for scientific advancement. Such unification has been a major goal in scientific research throughout history ([Kitcher 1993](#)), and its value is perhaps most evident in reconciling unconnected models and revealing new and unexpected empirical predictions. In evolutionary

biology, the Price equation (Box 1) provides a unifying framework for evolutionary theory by exhaustively
 24 and exactly describing evolutionary change for any closed evolving population (Price 1970; Luque 2017;
 Lehtonen et al. 2020). The Price equation is therefore fundamental, in the sense that it binds together
 26 all of evolutionary theory by formally defining what evolutionary change is and is not (Price 1970; Rice
 2004; Gardner 2008; Frank 2017; Luque 2017; Luque and Baravalle 2021). Using this formal definition,
 28 the scope of, and relationships among, sub-disciplines within evolutionary theory can be clarified. For
 example, fundamental equations of both population and quantitative genetics can be derived from the Price
 30 equation (Queller 2017). This provides conceptual clarity by demonstrating the logical consistency of different
 theoretical frameworks within evolutionary biology. Our aim here is to propose an equation that extends this
 32 conceptual clarity to include population change and thereby provide a formal and exact definition of joint
 ecological and evolutionary change.

34 In biological populations, ecological change is caused by the same processes of individual birth and death
 that cause changes in allele frequencies and phenotypes (Turchin 2001; Connor and Hartl 2004; Barfield et al.
 36 2011). As with evolution, a fundamental equation can exhaustively and exactly define population change.
 Unlike the Price equation, this fundamental equation is perhaps self-evident; population change is simply the
 38 addition of births and subtraction of deaths from current population size (N_t) to recover the new population
 size (N_{t+1} ; Box 2). By definition, this relationship of $N_{t+1} = N_t + Births - Deaths$ applies to any closed
 40 population. Turchin (2001) argues that general principles are needed to establish a logical foundation for
 population ecology, and this simple birth and death model and the consequences that logically follow from it
 42 (e.g., exponential population growth) is fundamental to population ecology. Any unifying definition of joint
 ecological and evolutionary change must be able to derive both the Price equation and this birth and death
 44 model.

The interdependence of ecological and evolutionary processes has long been recognised (e.g., Darwin 1859;
 46 Fisher 1958; Pelletier et al. 2009), but the rise of eco-evolutionary models, which incorporate both, is relatively
 recent following a widespread recognition that ecology and evolution can happen on similar timescales (Govaert
 48 et al. 2019; Yamamichi et al. 2023). Currently, a universally recognised formal definition of eco-evolutionary
 change is lacking, with some theoreticians broadly interpreting “eco-evolutionary dynamics” to allow for a

separation of ecological and evolutionary timescales (Lion et al. 2023) and others advocating for a more narrow interpretation in which no such separation is permitted (Bassar et al. 2021). Bassar et al. (2021) identify two types of eco-evolutionary models that follow from these interpretations. The first type uses separate equations to model population change versus evolutionary change, thereby allowing for any number of ecological, evolutionary, or environmental feedbacks (e.g., Lion 2018; Patel et al. 2018; Lion et al. 2023). The second type models population demographics as functions of quantitative traits, with ecological and evolutionary change following from demographic processes and trait distributions operating on the same timescale (e.g., Barfield et al. 2011; Simmonds et al. 2020; Jaggi et al. 2024). Both model types can be very general, but like all predictive models, they rely on simplifying assumptions for tractability (Levins 1966; Luque 2017). These simplifying assumptions are often grounded in the Price equation to demonstrate accuracy and logical consistency when modelling evolutionary change (e.g., Coulson and Tuljapurkar 2008; Barfield et al. 2011; Rees and Ellner 2016; Lion 2018). For example, Barfield et al. (2011) link their model back to Price (1970), which they consider to be a “universal law of evolution”, to place their conclusions concerning stage structured evolution in the broader context of evolutionary theory. The role of fundamental equations is therefore important for unifying theory (Luque 2017), and we believe that a fundamental equation of eco-evolutionary change is needed.

We present an equation from which the fundamental equations of ecology and evolutionary biology can be derived. Derivation follows by adding assumptions that are specific to population ecology or evolution in the same way that key equations of population genetics (e.g., average excess) or quantitative genetics (e.g., Breeder’s equation) can be derived from the Price equation (Queller 2017). We propose our equation as a formal definition of eco-evolutionary change.

Box 1: The Price equation is an abstract formula to represent evolutionary change. Formulated originally in the early 1970s by George Price (Price 1970, 1972), it postulates some basic properties that all evolutionary systems must satisfy: change over time, ancestor and descendant relations, and a character or phenotype (Rice 2004). Using simple algebraic language, the Price equation

represents evolutionary change with the predominant notation,

$$\bar{w}\Delta\bar{z} = \text{Cov}(w, z) + \text{E}(w\Delta z).$$

In the above equation, $\Delta\bar{z}$ is the change in the average character value z over a time step of arbitrary length, w is an individual's absolute fitness, and \bar{w} average population fitness. On the right-hand side of the equation, the first term is the covariance between a character value z and fitness w , which reflects \bar{z} change attributable to differential survival and reproduction. The second term is the expected value of Δz , which reflects the extent to which offspring deviate from parents in z (Rice 2004; Okasha 2006; Frank 2012). A more specific version of the covariance term was already known within the quantitative and population genetics tradition (Robertson 1966), usually representing the action of natural selection. The Price equation adds an expectation term and abstracts away from any specific mechanisms of replication or reproduction, or mechanisms of inheritance. Its definitional nature and lack of substantive biological assumptions has been portrayed both as a strength (Rice 2004; Frank 2012; Luque 2017; Baravalle and Luque 2022), and its greatest weakness. The abstract nature of the Price equation places it at the top of the hierarchy of fundamental theorems of evolution from which the rest (Robertson's theorem, Fisher's fundamental theorem, breeder's equation, Hamilton's rule, adaptive dynamics, etc.) can be easily derived (Lehtonen 2016, 2018; Queller 2017). This abstractness is also key to developing a more general view of evolution (Rice 2020; Luque and Baravalle 2021; Edelaar et al. 2023). In contrast, some researchers consider the Price equation just a triviality (even tautological), and useless without further modelling assumptions (van Veelen 2005; van Veelen et al. 2012). The debate remains open (van Veelen 2020; Baravalle et al. 2024).

Box 2: The number of individuals in any closed population (N) at any given time ($t + 1$) is determined by the existing number of individuals (N_t), plus the number of births (*Births*) minus

the number of deaths (*Deaths*),

$$N_{t+1} = N_t + Births - Deaths.$$

This equation is necessarily true for any closed population. Despite its simplicity, it is a general equation for defining population change and a starting point for understanding population ecology. Turchin (2001) notes that a consequence of this fundamental equation is the tendency for populations to grow exponentially (technically geometrically in the above case where time is discrete). This inherent underlying tendency towards exponential growth persists even as the complexities of real populations, such as structure, stochasticity, or density-dependent effects are added to population models (Turchin 2001). Given the assumption that all individuals in the population are identical, a per capita rate of birth, $Births = bN_t$ and death, $Deaths = dN_t$ can be defined. Rearranging and defining $\lambda = 1 + b - d$ gives, $N_{t+1} = N_t\lambda$. Here λ is the finite rate of increase (Gotelli 2001), and note that because $0 \leq d \leq 1$, $\lambda \geq 0$. Verbally, the change in size of any closed population equals its existing size times its finite rate of increase.

A foundation for biological evolution and population ecology

To fully unify biological evolution and population ecology, we must reconcile the the Price equation (Box 1) with the general equation for population change (Box 2). The Price equation is critical for partitioning different components of biological change (Price 1970; Frank 1997; Gardner 2008; Luque 2017; Queller 2017; Lehtonen 2018). It has also been highly useful for integrating evolutionary theory across disciplines (Fox 2006; Brantingham 2007; MacCallum et al. 2012; Frank 2015; Godsoe et al. 2021; Ulrich et al. 2024). These properties would seem to make it an intuitive starting point for a logical foundation of ecology and evolution, perhaps through some kind of mathematical equivalence (Page and Nowak 2002) or addition of terms (Collins and Gardner 2009), or through the use of its recursive structure (Kerr and Godfrey-Smith 2009; Frank 2012). But despite its flexibility, the Price equation still relies on relative frequencies, which must by definition sum

to one (Frank 2015). This is because the Price equation describes the average change in a population; the frequency of entities is scaled thereby conserving total probability (Frank 2015, 2016). But to recover the fundamental principle of exponential population growth (Turchin 2001), this scaling must be avoided in a fundamental equation of ecology and evolution.

We therefore begin with the most fundamental axioms underlying the ecology and evolution of living systems (Rice 2004; Rice and Papadopoulos 2009). In such systems, diversity is discontinuous and can be defined in terms of discrete entities (Dobzhansky 1970). Our framework is general enough that entities can be anything discrete, but we will focus on each entity i as an individual organism. Individuals give rise to new individuals through birth such that β_i is the number of births attributable to i . Individuals are removed from the population through death such that δ_i is an indicator variable that takes a value of 1 (death of i) or 0 (persistence of i). All individuals are defined by some characteristic z_i , and Δz_i defines any change in z_i from one time step t to the next $t + 1$. The total number of individuals in the population is N . From this foundation, we can define Ω to be a summed characteristic across N entities,

$$\Omega = \sum_{i=1}^N (\beta_i - \delta_i + 1) (z_i + \Delta z_i). \quad (1)$$

The foundation of eco-evolutionary change is an interaction between the demographic processes of birth and death ($\beta_i - \delta_i + 1$) and some characteristic of an individual ($z_i + \Delta z_i$). The value of Ω is a summation of individual characteristics, which takes the same units as z . From eqn 1, we can derive the most fundamental equations of population ecology (Box 2) and evolutionary biology (Box 1) through an appropriate interpretation of z . Under many interpretations of z , we can also interpret Ω as a metric of ecosystem function.

Population ecology

To recover the general equation for population ecology (Box 2), we define z_i as the identity of i belonging to the population. In other words, we set $z_i = 1$ to simply indicate that i is a member of the population. In this case, z is count, which takes the unit 1 [note that ‘individual’ is a label, not a unit; Newell and Tiesinga

(2019)]. Further, we assume that individuals do not change species by setting $\Delta z_i = 0$. In this case,

$$\Omega = \sum_{i=1}^N (\beta_i - \delta_i + 1).$$

146 We can now interpret Ω as the population size at $t + 1$, N_{t+1} and note that $N\beta$, $N\delta$, and N are the total births, total deaths, and size of the population at t , respectively,

$$N_{t+1} = N_t + \text{Births} - \text{Deaths}. \quad (2)$$

148 If we assume that individuals are identical, then we can drop the subscript i such that $\beta_i = \beta$ and $\delta_i = \delta$. Summing from 1 to N , we can rewrite the above,

$$\Omega = N(1 + \beta - \delta).$$

150 If we define $\lambda = 1 + \beta - \delta$ (Box 2), then we can rewrite,

$$N_{t+1} = N_t \lambda. \quad (3)$$

We therefore recover the general equation for population ecology (eqn 2) and the fundamental property of
152 exponential growth in populations (Turchin 2001) (eqn 3).

Evolutionary biology

154 Recovering the Price equation requires a few more steps. We start by defining individual fitness,

$$w_i = \beta_i - \delta_i + 1. \quad (4)$$

In this definition, the longevity of the individual matters. An individual that survives from t to $t + 1$ has a

higher fitness than one that dies, even if both have the same reproductive output. With this definition of fitness (eqn 4), we substitute,

$$\Omega = \sum_{i=1}^N (w_i z_i + w_i \Delta z_i). \quad (5)$$

We can break eqn 5 down further and multiply each side by $1/N$,

$$\frac{1}{N}\Omega = \frac{1}{N} \sum_{i=1}^N (w_i z_i) + \frac{1}{N} \sum_{i=1}^N (w_i \Delta z_i). \quad (6)$$

We can rewrite the terms on the right-hand side of eqn 6 as expected values and remove the subscripts,

$$\frac{1}{N}\Omega = E(wz) + E(w\Delta z). \quad (7)$$

Now we must consider the total conservation of probability (Frank 2015, 2016). In eqn 7, Ω is the total sum trait values (z_i) across the entire population at $t + 1$ divided by the number of individuals (N) in the population at t . But the size of the population can change from t to $t + 1$. To recover mean trait change for the Price equation (and therefore conserve total probability), we need to account for this change in population size. The left-hand side of eqn 7 describes contributions to the sum trait value from the new population at $t + 1$. But we cannot treat Ω/N as the mean of z at $t + 1$ (\bar{z}') because we need to weigh N by the mean fitness of the population at t to account for any change in population size from t to $t + 1$.

For example, if mean fitness at t is 2 (i.e., $\bar{w} = 2$), then half as many individuals will have contributed to Ω in $t + 1$ than would have if $\bar{w} = 1$ (i.e., there are N individuals at t and $2N$ individuals at $t + 1$). We therefore need to multiply the mean trait value \bar{z}' (at $t + 1$) by the mean fitness \bar{w} (at t) to recover the mean contribution of the N individuals at t to the total Ω . Consequently,

$$\Omega = N\bar{w}\bar{z}' \quad (8)$$

Equation 8 conserves the total probability and now clarifies Ω as a summed trait value, which has the same
 172 units as z and equals expected population growth at t times mean trait value at $t + 1$. This is consistent with
 the population ecology derivation from the previous section where $z_i = 1$ by definition, and $\Omega = N_{t+1}$ (note
 174 $\lambda = \bar{w}$). We can therefore rewrite eqn 7,

$$\bar{w}\bar{z}' = E(wz) + E(w\Delta z). \quad (9)$$

We can rearrange eqn 9 to derive the Price equation by expressing covariance as, $\text{Cov}(X, Y) = E(XY) -$
 176 $E(X)E(Y)$, and therefore $E(XY) = \text{Cov}(X, Y) + E(X)E(Y)$. Substituting into eqn 9,

$$\bar{w}\bar{z}' = \text{Cov}(w, z) + \bar{w}\bar{z} + E(w\Delta z).$$

From here,

$$\bar{w}(\bar{z}' - \bar{z}) = \text{Cov}(w, z) + E(w\Delta z).$$

178 Since $\Delta\bar{z} = (\bar{z}' - \bar{z})$,

$$\bar{w}\Delta\bar{z} = \text{Cov}(w, z) + E(w\Delta z). \quad (10)$$

From eqn 1, which describes fundamental birth and death processes in a population, we can derive both the
 180 most fundamental model of population ecology (eqn 2; Box 2) and the fundamental equation of evolution
 (eqn 10; Box 1).

182 Ecosystem function

In some cases, we might also consider Ω to be the total contribution of a population to ecosystem function.
 184 For example, z might be interpreted as trait that underlies a function of interest such as carbon capture or

photosynthetic rate. When that the relevant characteristic (z_i) is additive across individuals (e.g., the total amount of carbon capture in a population is the sum of carbon captured across individuals), Ω could be an appropriate metric of ecosystem function. This will not always be the case. The sum of some measurable traits such as leaf pH (Cornelissen et al. 2011) will not be ecologically meaningful when summed across individuals. And in other cases, the relationship between Ω and some ecosystem function of interest is likely to be subadditive. For example, we could interpret z_i as the frequency at which a pollinator i visits plants (which might change from t to $t + 1$ for any number of reasons, including ontogenic or environmental, as reflected in Δz_i). While Ω would then be the total frequency of pollinator visits in a population, pollinator saturation effects would likely have a logarithmic relationship between Ω and total seed set (e.g., Mu et al. 2018). While we have argued that ecological and evolutionary change can both be derived from the same fundamental relationship between birth, death, and individual characteristics, the same is not true for ecosystem function. Nevertheless, Box 3 provides a toy example of four plants with different fitnesses and carbon captures. We use this toy example to demonstrate a concrete example of our proposed framework.

Discussion

A classical sign of scientific progress is the ability to connect disparate theories and models to show how empirical and theoretical models are logical (mathematical) consequences of more fundamental ones (Nagel 1961). Conceptual unification therefore has a critical role in advancing scientific theory (Morrison 2000). Showing how apparently disparate phenomena follow from the same shared principles reveals what is most fundamental about the natural world and provides a foundation for further scientific investigation and the construction of more coherent and predictive models. Our eqn 1 provides a foundation for the conceptual unification of the fundamental equations of population ecology and evolution.

Rather than making simplifying assumptions, as is the approach for most ecological and evolutionary models, we focus on fundamental axioms that are universal to closed biological systems: discrete individuals, birth, death, and change over time. We define an abstract sum (Ω), to which all individuals within the population contribute. From the simple assumptions of population identity ($z_i = 1$) and invariability ($\Delta z_i = 0$),

we recover the most general equation of population ecology (Box 2) and principle of exponential growth ($N_{t+1} = N_t\lambda$). By defining individual fitness (w_i) and applying the total conservation of probability to individual frequencies (Frank 2015, 2016), we recover the most fundamental equation of evolution (Box 1). We therefore deliver a general framework to unify fundamental equations of ecological and evolutionary change.

The Price equation provides a complete and exact description of evolution in any closed evolving system (Price 1970; Frank 2012). It is derived by rearranging the mathematical notation defining changes in the frequencies and characteristics of any type of entity (Price 1970; Gardner 2008; Luque 2017) (e.g., individuals, alleles). This derivation partitions total characteristic change into different components, making it possible to isolate evolutionary mechanisms (e.g., selection) and levels of biological organisation (e.g., group, individual) (Frank 1995, 2012; Kerr and Godfrey-Smith 2009; Luque 2017; Okasha and Otsuka 2020). Because of its abstract nature and lack of any system-specific assumptions, the Price equation makes no predictions about what will happen in any particular system (Gardner 2020). Its role is not to predict, but to formally and completely define and separate components of evolutionary change. The same is true of the general equation for population change (eqns 2 and 3), at least as we have used it here where it serves to define what population growth means in ecology. This equation formally and completely describes population change in terms of births and deaths. In eqn 1, we therefore have a fundamental equation from which we can derive complete ecological and evolutionary change in any closed biological population. We anticipate that this will be useful for eco-evolutionary theory in the same way that the Price equation is useful for evolutionary theory: facilitating specific model development and identifying new conceptual insights, unresolved errors, and sources of model disagreements.

Our unification demonstrates the equivalence between the finite rate of increase λ (Box 2) (Gotelli 2001) and population mean evolutionary fitness \bar{w} (Box 1). The population growth equation $N_{t+1} = N_t\lambda$ could therefore be rewritten as $N_{t+1} = N_t\bar{w}$. We do not claim to be the first to notice this equivalence between population growth and mean population fitness. Indeed, the relationship between population growth rate and evolutionary fitness has been proposed and applied many times before (Fisher 1930; Charlesworth 1980; Lande 1982; Roff 2008; Lion 2018). For example, Lande (1982) explicitly concludes mean absolute fitness per unit time is $\bar{w} = e^r$, where $r = \ln(\lambda)$. We show this from first principles and clarify the relationship between

fitness and population growth. Over an arbitrary length of time, fitness is properly defined as $w_i = \beta_i - \delta_i + 1$.

Over an individual's lifetime (which, by definition, includes death), fitness is therefore β_i . Interestingly, the rate of change in ecology and evolution are reflected in the first and second statistical moments of fitness, respectively. Population growth rate reflects mean fitness \bar{w} , while the rate of evolutionary change reflects the variance in fitness $\text{Var}(w)$ (i.e., Fisher's fundamental theorem) (Frank 1997; Rice 2004; Queller 2017).

Our unification may also help explain, at least partially, some of the success of classical population genetic models. For decades, population genetics (and to some extent quantitative genetics) has been accused of being a reductionistic view of evolution, reducing everything to changes in allele frequencies and abstracting away from individuals and their environments (the ecological interactions) (MacColl 2011). This has been a line of argumentation by some defenders of the so-called Extended Synthesis (Pigliucci 2009), especially in relation to niche construction (Odling-Smee et al. 2003). Famously, Mayr (1959) characterized population genetics as a simple input and output of genes, analogous to "the adding of certain beans to a beanbag and the withdrawing of others" (also called "beanbag genetics").

Historical critics of population genetics could not articulate a clear explanation for why it works so well despite all of its idealisations and simplifications. From the Price equation, we are able to recover classical population and quantitative genetic models (Queller 2017) and develop new ones (Rice 2004, 2020; Luque 2017; Lion 2018). Our eqn 1 contains ecology at its core, and we show how the Price equation logically follows from it after accounting for absolute population growth (eqn 8). We therefore conclude that population and quantitative genetic equations contain ecology (no matter how hidden), and the ecological nature of evolution is implicit in population and quantitative genetic models.

We have focused on the dynamics of a closed population, and in doing so leave ecological and evolutionary change attributable to migration for future work. In population ecology, immigration and emigration can be incorporated by adding a term for each to the right-hand side of the equation in Box 2 (Gotelli 2001). In evolution, because the Price equation relies on mapping ancestor-descendant relationships, accounting for migration is more challenging. Kerr and Godfrey-Smith (2009) demonstrate how the Price equation can be extended to allow for arbitrary links between ancestors and descendants, thereby extending the Price

equation to allow for immigration and emigration. Frank (2012) presents a simplified version of Kerr and
 264 Godfrey-Smith (2009) that allows some fraction of descendants to be unconnected to ancestors. In both
 ecology and evolution, accounting for migration is done through the use of additional terms on the right-hand
 266 side of the equations.

Our fundamental equation is complete and exact. It therefore implicitly includes any effects of density
 268 dependence on population growth (see Box 2), or any social effects on evolutionary change (see Box 1).

We can make both of these effects explicit with the same partition in the summation on the right-hand
 270 side of eqn 1. To do this, we multiply the expression $(\beta_i - \delta_i + 1)(z_i + \Delta z_i)$ by $\left(1 - \sum_{j=1}^N a_{ij}(z_j + \Delta z_j)\right)$,
 where j indexes the same individuals as i , and a_{ij} defines the effect j has on the contribution of i to Ω . In
 272 evolutionary terms, a_{ij} defines how j modulates the fitness of i (w_i). In ecological terms, a_{ij} defines how
 j modulates the contribution of i to total population growth. If we define α as the summed effect of all
 274 individuals on a focal i (i.e., $\alpha = \sum_{j=1}^N a_{ij}$), then we can partition ecological change into density-independent
 and density-dependent effects and derive a discrete time logistic growth equation. We save these derivations
 276 for a future investigations focused on evolutionary and community ecology.

We have shown that we can derive the fundamental equations of population ecology and biological evolution
 278 from a single unifying equation. Lastly, we propose our eqn 1 as a potential starting point for defining
 ecosystem function and further conceptual unification between ecology, evolution, and ecosystem function.
 280 The Price equation has previously been used to investigate ecosystem function (Loreau and Hector 2001; Fox
 2006), but not with any attempt towards conceptual unification with evolutionary biology. For example, Fox
 282 (2006) applied the abstract properties of the Price equation to partition total change in ecosystem function
 into separate components attributable to species richness, species composition, and context dependent effects.
 284 This approach provides a framework for comparing the effects of biodiversity on ecosystem function in
 empirical systems (Fox 2006; Winfree et al. 2015; Mateo-Tomás et al. 2017). Instead, our eqn 1 defines
 286 Ω as total ecosystem function contributed by a focal population (e.g., biomass production, decomposition,
 photosynthetic rate). It is therefore possible to investigate ecological, evolutionary, and ecosystem function
 288 change from the same shared framework.

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Author Contributions

Both authors came up with the question idea. Following many discussions between the authors, ABD proposed the initial equation with subsequent exploration and development from both authors. Both authors contributed to the writing.

Competing Interests

The authors declare no competing interests.

Data Availability

This work does not include any data.

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- 428