

1 **Title page**

2 **Article title:** The Effect of Phylogenetic Uncertainty and Imputation on EDGE Scores

3 **Running head:** The Effect of Phylogenetic Uncertainty and Imputation on EDGE Scores

4 **Authors:** K. Bodie Weedop¹, Arne Ø. Mooers², Caroline M. Tucker³, and William D. Pearse^{1*}

5 ¹ Department of Biology & Ecology Center, Utah State University, 5305 Old Main Hill, Logan UT, 84322

6 ² Department of Biological Sciences, Simon Fraser University, Burnaby, British Columbia, Canada

7 ³ Department of Biology, University of North Carolina–Chapel Hill

8 ^{*}To whom correspondence should be addressed: `will.pearse@usu.edu`

9 **Word-count:** XXX (abstract, main text, acknowledgments, and references)

Abstract

Faced with the challenge of saving as much diversity as possible given financial and time constraints, conservation biologists are increasingly prioritizing species on the basis of their overall contribution to evolutionary diversity. Metrics such as EDGE (Evolutionary Distinct and Globally Endangered) have been used to set such evolutionarily-based conservation priorities for a number of taxa, such as mammals, birds, corals, amphibians, and sharks. Each application of EDGE has required some form of correction to account for species whose position within the tree of life are unknown. Perhaps the most advanced of these corrections is phylogenetic imputation, but to date there has been no systematic assessment of both the sensitivity of EDGE scores to a phylogeny missing species, and the impact of using imputation to correct for this. Here we perform such a systematic assessment, simulating incomplete phylogenies from which some species are missing, then imputating the position of those species, and measuring the impact each of these processes has on accurately estimating species' true EDGE scores. We find that EDGE ranking for species on a tree is remarkably robust to missing species from that tree, but that phylogenetic imputation for missing species, while unbiased, is not accurate in reconstructing species' true evolutionary distinctiveness. On the basis of these results, we provide clear guidance for EDGE scoring in the face of phylogenetic uncertainty.

Keywords: conservation prioritization, evolutionary distinctiveness, EDGE, phylogenetic imputation.

Introduction

Evidence from the fossil record and present-day studies argue we are in the midst of, or entering, a sixth mass extinction (Barnosky et al. 2011; Ceballos et al. 2015), such that more populations than ever are declining and species face heightened danger of extinction (Wake & Vredenburg 2008; Thomas et al. 2004). Habitat destruction (Brooks et al. 2002), invasive species (Molnar et al. 2008), climate change (Pounds et al. 2006), and disease (Lips et al. 2006) are some of the leading causes of species declines globally. Conservation biologists seek to reduce these detrimental effects on species populations, but in reality they have limited resources with which to do so. This challenge, termed the “Noah’s Ark problem” (Weitzman 1998), has driven conservation biologists to identify different ways by which to prioritize, or triage, their resource allocation (Bottrill et al. 2008).

Conservation triage, like all sound decision-making, requires an index with which to quantify the relative urgency or importance for conservation among a set of options. This allows scientists and policy-makers to use data to quantify need and inform conservation decision-making and management activities. One triage strategy uses the EDGE metric to identify and prioritize species that are Evolutionarily Distinct and Globally Endangered (Isaac et al. 2007). Evolutionary Distinctiveness (ED) measures the relative contributions made by each species within a particular clade to phylogenetic diversity, assigning each branch length equally to all the subtending species (Redding n.d.; Isaac et al. 2007). Global Endangerment (GE), assigns numerical values to each of the World Conservation Union (IUCN) Red List Categories. As species become increasingly threatened and are placed into categories of increasing concern (*e.g.* from Vulnerable to Endangered), the GE numerical value increases. A species’ EDGE score is an aggregate value intended to equally reflect a species’ evolutionary distinctiveness and conservation status (even if it does not always in practice; see Pearse et al. 2015).

Usage of the EDGE metric has expanded greatly. Though originally used to prioritize global mammals, EDGE scores are now available now for a variety of taxonomic groups, including amphibians (Isaac et al. 2012), birds (Jetz et al. 2014), corals (Curnick et al. 2015), squamate reptiles (Tonini et al. 2016), sharks (Stein et al. 2018), and all tetrapods (Gumbs et al. 2018). Related metrics are also now available, each subtly emphasizing different things, such as the expected contribution of each species to future phylogenetic diversity (HEDGE, I-HEDGE; Steel et al. 2007; Jensen et al. 2016), our uncertainty over a species’ future (EDAM; Pearse et al. 2015), and the complementarity of a set of species (Faith et al. 2003). The development and expansion of EDGE-like metrics mirrors progress in other areas of conservation biology, and as such the likelihood of success in conservation (Wilson et al. 2007; McBride et al. 2007), relative cost of certain

interventions (Naidoo et al. 2006), and complementarity of interventions (Pressey et al. 1993; Myers et al. 2000) can now be considered in its calculation. The EDGE index was developed explicitly with the intention of informing conservation triage, and is now the basis of the global EDGE of Existence Program (<http://www.edgeofexistence.org/>). The successful application of EDGE in this program highlights the potential for phylogenetic conservation prioritization metrics to provide actionable insights in the face of uncertainty about species' attributes. Nonetheless, almost every application of an EDGE-type approach must address uncertainty resulting from missing data.

Missing data can affect EDGE scores in several ways. First, the IUCN identifies some species as Data Deficient (IUCN 2001; IUCN 2008), which affects the GE component of a species' EDGE score. Fortunately, the IUCN provides guidance for using any available contextual data to assign some threat status to such species. A number of studies illustrate how to assign threat categories to Data Deficient species, which in turn should reduce the uncertainty in GE (Good et al. 2006; Butchart & Bird 2010; Morais et al. 2013; Dulvy et al. 2014). The issue of missing phylogenetic data is arguably more complicated because not only does the focal species have no ED score, but its absence from the phylogeny may affect the ED scores of related species. Species of conservation concern are almost by definition rare, and frequently lack sufficient DNA (or even morphological) data to be placed with certainty on a phylogeny. In most cases, taxonomic information rather than sequence data alone has been used to place species in the tree of life when constructing EDGE lists (see Isaac et al. 2007; Collen et al. 2011; Isaac et al. 2012; Jetz et al. 2014; Curnick et al. 2015; Stein et al. 2018; Gumbs et al. 2018; Forest et al. 2018). Yet, to our knowledge, there is no systematic study of the effect of imputation on species' EDGE scores, despite this practice having received attention in other areas of comparative biology (Kuhn et al. 2011; Thomas et al. 2013; Rabosky 2015). Thus we do not know the fidelity of EDGE scores between imputed and non-imputed phylogenies or the magnitude of error that phylogenetic imputation introduces. As interest in using EDGE-type measures and phylogenies for conservation triage grows, the need for consensus on how to resolve cases of phylogenetic uncertainty becomes increasingly urgent.

Here we attempt to quantify the effect of missing species on EDGE rankings and assess the degree to which subsequent imputation affects the accuracy of EDGE scores. We do so by simulating phylogenies and missing species that are distributed either at random, or with bias, across those phylogenies. By contrasting the ED scores of the species before and after the loss of other species from the phylogenies, we measure the impact of missing species on ED scores. We then assess the extent to which phylogenetic imputation can accurately estimate the EDGE scores of missing species in simulated data. We also examine the extent to which such imputation affects the scores of species for which we have data. In so doing, we hope to provide clear

guidance as to the applicability of phylogenetic imputation as a solution for species missing phylogenetic data. From our results, we argue that species' ED values are remarkably robust to missing species, and that phylogenetic imputation does not reliably reconstruct the true ranking of those missing species.

Methods

Here we use a simulation approach to test the effect of having missing species on a phylogeny (through species removal from simulated phylogenies) and then imputing species for species' ED (Evolutionary Distinctiveness) scores. We focus exclusively on the ED-component of the EDGE metric, since uncertainty in species GE scores has already been addressed by the IUCN's proposal to assign Data Deficient species scores (IUCN 2001; IUCN 2008). Because EDGE is the product of both ED and GE components, even perfectly accurate GE values could be associated with imperfect EDGE scores if the ED scores were inaccurate.

All trees (both starting and imputed) were simulated under a pure-birth Yule model using `sim.bdtree` in `geiger` (setting parameters `b=1` and `d=0`; Pennell et al. 2014). This model was chosen because it is the simplest model possible: speciation rates are constant across the entire tree of life and there is no extinction. We acknowledge that more complex and/or biologically realistic models of diversification could potentially improve the performance of imputation. However, we suggest that imputation under a simple model that is identical to that used to simulate the data is a low, and fair, benchmark for a method to meet. We used `caper::ed.calc` to calculate ED values (Orme et al. 2013). All simulations and analyses were performed using R (version 3.4.0; R Core Team 2017), and we performed 100 replicate simulations of each parameter combination. All our analysis code is available online (<https://github.com/bweedop/edgeSims>) and in the supplementary materials.

The impact of missing species on EDGE scores

Our first set of simulations assess the impact of missing species data on the ED scores of remaining species, considering data missing either in a random or phylogenetically-biased fashion. We simulated phylogenies of different sizes (number of species: 64, 128, 256, ..., 2048, 4096) and then removed constant fractions of tips from the tree (0%, 1%, 2%, ..., 19%, ..., 99%). To simulate species missing at random throughout the phylogeny, we used `sample()` to select the relevant fraction of species (rounded to the nearest whole number) without replacement. To remove species in a phylogenetically-biased manner, we used Felsenstein (2005)'s

threshold model. We simulated a trait under a constant rate Brownian-motion model ($\sigma=0.5$, starting root value = 1) using `sim.char` in `geiger` (Pennell et al. 2014). Species were then removed from the tree if their simulated trait was in the upper quantile matching the fraction of species to be removed. For example, if 10% of species were to be removed from the tree, the species with the highest 10% of values would be removed. This results in closely related species being removed more often than expected by chance.

To quantify the effect of these manipulations, we calculated the ED values of species that are not removed from a tree both before and after removal. We then correlated these ED scores: if missing species do not affect ED values of the remaining species, we would expect a strong, positive correlation between the ED scores of the remaining species calculated before and after species were removed from the phylogeny. We emphasize that species removed from the phylogeny are omitted from this comparison. We outline our approach in figure 1.

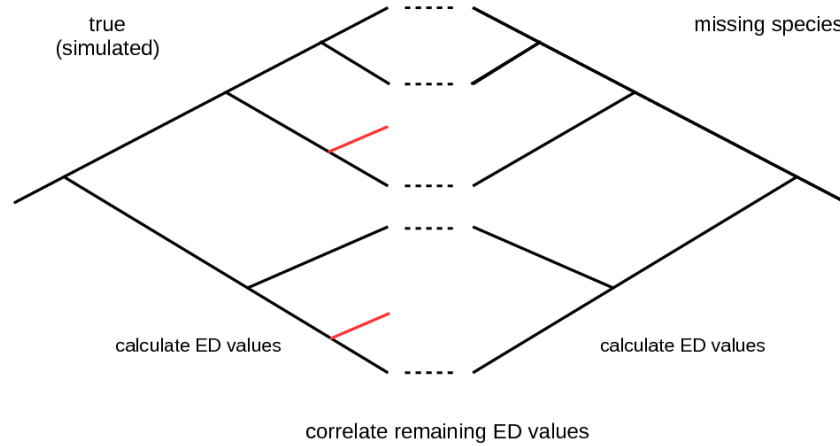


Figure 1: Conceptual overview of the missing-species simulations in this study. The simulated tree on the left is the true tree prior to removal of missing species. On the right is the same tree after missing species have been removed. Species that are removed are shown in red. To compare the ED values of the remaining species, we correlate their ED values before (left) and after (right) removal of the missing species. Dashed lines can be seen for the species which would have ED scores compared.

The impact of phylogenetic imputation on EDGE scores

Our second set of simulations tested the impact of imputation on ED scores within an imputed clade. We used relatively small clades (5, 6, 7, ..., 30, 31, 32 species) from phylogenies of different sizes [128 (2^7), 147 ($2^{7.2}$), 168 ($2^{7.4}$), ... , 776 ($2^{9.6}$), 891 ($2^{9.8}$), 1024 (2^{10}) species]. We first randomly selected a clade to be removed from the ‘true’ tree and then simulated a new phylogeny of the same size as the removed clade. This newly simulated clade was generated under the same pure-birth model as the original phylogeny. We then placed the newly simulated clade in the full phylogeny, in the same location as the removed clade. If a newly simulated clade was so old that it was not possible to graft it into place, we discarded that clade and simulated another. In an empirical study the model of evolution under which the phylogeny had evolved would have to be estimated, which is an additional source of error not considered here. We simulated each combination of clade and total phylogeny sizes 100 times. An overview of our approach is given in figure 2.

To assess whether clades, once imputed, had similar ED scores to their true values, we correlated the imputed ED scores with the true ED scores. We also calculated the sum of the absolute change in ranked ED for each species, which is particularly relevant for EDGE-listing as conservation actions are often focused around the top 100, 200, etc., species. We modeled both of these metrics (the change in ranking and the correlation) as a function of a number of potential explanatory variables. Specifically, we included in our models: the estimated speciation rate of the original phylogeny (using `ape::yule`; Paradis et al. 2004), the sum of all phylogenetic branch-lengths in the original phylogeny (Faith’s PD; Faith 1992), the sum of all phylogenetic branch-lengths in the original focal clade (Faith’s PD; Faith 1992), the value of γ in the original phylogeny (using `phytools::gammatest`; Pybus & Harvey 2000; Revell 2012), Colless’ index of the original phylogeny (using `ape::treeShape::as.treeShape`; Colless 1982; Bortolussi et al. 2009), the kurtosis of species’ ED values in the original phylogeny (using `moments::kurtosis`; Komsta & Novomestky 2015), the skew of species’ ED values in the original phylogeny (using `moments::skew`; Komsta & Novomestky 2015), the total number of species in the original phylogeny, the total number of species within the imputed clade, and the depth (age) of the imputed clade in the phylogeny. Although the expectations of many of these explanatory variables are known for Yule trees, in each simulation they are expected to vary somewhat by chance.

Recently, there has been interest in assigning missing species the mean ED score within a clade (see Gumbs et al. 2018). To test the efficacy of such methods, we assigned the average ED of the selected clade to each of its’ species and calculated (as above) the mean change in absolute ranking under this scheme. Note that we could not correlate ED scores (as we do above), since such a correlation would require variation in species’

158 scores and under this approach a single score (the mean ED) is assigned to all imputed species.

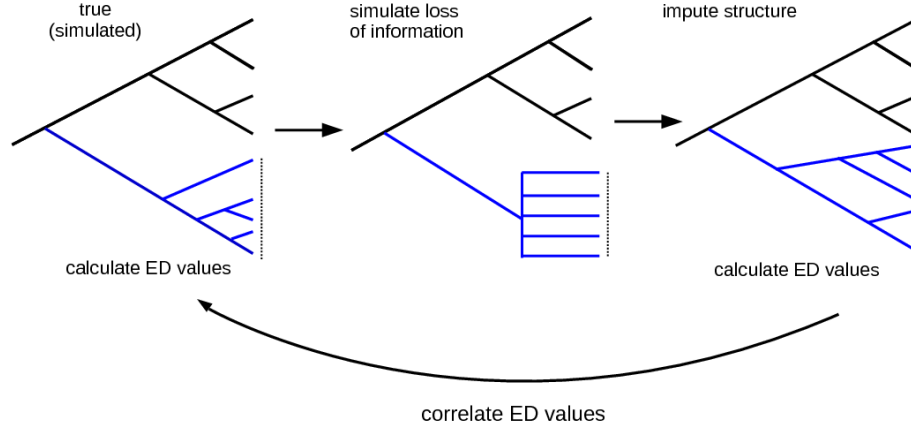


Figure 2: Conceptual overview of the imputation simulations conducted in this study. The simulated tree on the left is the ‘true tree’. We selected a clade to treat as ‘missing’ (highlighted with a dashed line and in blue) by replacing it with a polytomy (middle panel), and then imputed the ‘missing’ species following the procedure described in the Methods to produce the imputed clade in the right panel. To compare true and imputed ED values within the imputed clade, we correlated ED values calculated for the true clade (left) with those for the imputed clade (right).

159 Results

160 With increasing numbers of missing species from a phylogeny, remaining species’ ED scores become less ac-
 161 curate (table 1; figure 3). ED values were less robust to phylogenetically-biased missing species, as compared
 162 to those that are randomly missing from the tree. The effect of missing species is not necessarily severe,
 163 however; if 20% of species are missing, the average correlation coefficient between true and estimated ED is
 164 0.88 and 0.94 for phylogenetically-biased and random missing species, respectively.

165 When clades were imputed on the tree, we found a weak (if any) average positive correlation between the
 166 imputed ED and true ED values for species within the imputed clades (0.1974 in a statistical model with
 167 an r^2 of 0.5%; figure 4, table 2). We also found no explanatory variables that explained significant variation
 168 in this relationship (table 2; see also Supplementary Materials). However, we did find evidence that, when
 169 imputing larger clades, the variation in the correlation between true and imputed ED scores decreases,

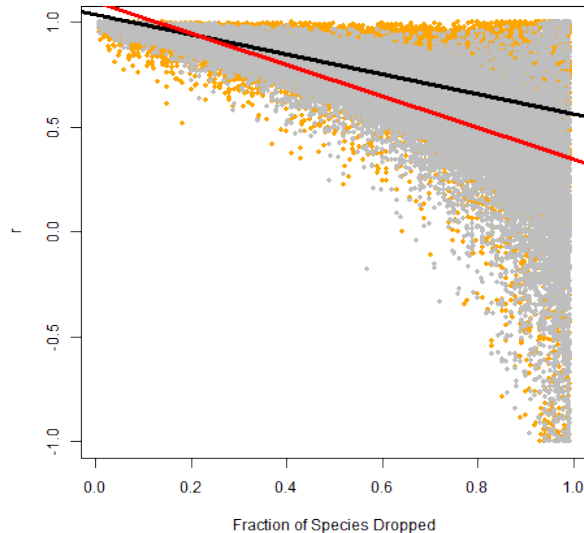


Figure 3: The effect of missing data on the calculation of the remaining species' ED values. The correlation coefficient of species' ED values in full (simulated) phylogenies, comparing values before and after the random loss of (other) species from the tree. The color of data points denote whether the species were removed from the phylogeny completely at random (orange) or in a phylogenetically biased fashion (see text; grey). Lines show regressions for random (red) or phylogenetically biased (black) species loss; see table 1 for model coefficients. This plot shows that the accuracy of estimation of ED values is inversely proportional to the number of species missing from the phylogeny, and that phylogenetically-biased species loss has a greater impact on accuracy.

although it remains centered around 0 (see Supplementary Materials). When considering rankings rather than raw scores, we similarly found that imputation can introduce error into the estimation of species' ED values (figure 5 and table 3). This ranking error increased with the size of the imputed clade and phylogeny (table 3), and can affect ranking error within the top 100 and 250 species (see Supplementary Materials). To give an example of the magnitude of the effect, within a phylogeny of 1024 species, the members of an imputed clade of 30 species are, on average, ± 315 rankings from their true rankings. Importantly, we found similar effects in ranking error when using the average ED value of clade for a missing species (see Supplementary Materials).

	Estimate	Std. Error	t value	Pr(> t)
Reference—Phylogenetically biased				
Intercept	1.0315	0.0013	821.39	<0.0001
Fraction of species removed	-0.4696	0.0020	-233.16	<0.0001
Number of species overall	2.500×10^{-6}	2.984×10^{-7}	8.38	<0.0001
Contrast—Random				
Intercept	0.0630	0.0018	35.47	<0.0001
Fraction of species removed	-0.2774	0.0028	-97.45	<0.0001
Number of species overall	5.013×10^{-6}	4.219×10^{-7}	11.88	<0.0001

Table 1: Statistical model of the effect of missing data on the calculation of the remaining species’ ED values. Results of a multiple regression fit to the data shown in figure 3, regressing the correlation coefficient of (remaining) species’ ED scores before and after other species were removed from the phylogeny ($F_{139696,5} = 40,350$, $r^2 = 0.5908$, $p < 0.0001$). The first three rows refer to the overall intercept, effect of the fraction of species removed from the phylogeny, and the overall size of the phylogeny when species were removed in a phylogenetically biased fashion. The last three rows are contrasts, reporting the difference (contrast) of each parameter when species were removed at random from the phylogeny, whether random loss of species has a statistically different effect. The correlation of ED scores appears affected by an interaction between the number of species removed from the tree and whether those species were removed at random or in a phylogenetically-biased fashion. The overall size of the phylogeny has little discernible effect, and its statistical significance is likely driven by the large number of simulations we performed (139,700).

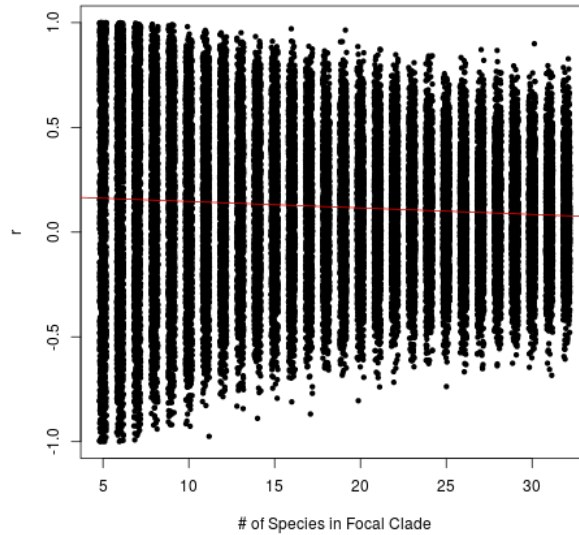


Figure 4: The correlation between species’ imputed and true ED scores plotted as a function of the number of species imputed (focal clade size). Each data point represents the correlation between ED values within the focal clades where imputation has occurred, comparing species’ true ED values with their imputed ED values. This plot, and the statistical analysis of it in table 2, show limited support for an association between true and imputed ED values.

	Estimate	Std. Error	t value	Pr(> t)
Intercept	0.1974	0.0501	3.94	0.0001
Size of Focal Clade	-0.0036	0.0005	-7.60	< 0.0001
Size of Phylogeny	0.0001	0.0001	0.60	0.5497
PD	-0.0001	0.0001	-0.64	0.5241
Estimated speciation rate	-0.0199	0.0493	-0.40	0.6865
Colless' Index	-0.0000	0.0000	-0.08	0.9380
Skew	0.0022	0.0083	0.27	0.7885
Kurtosis	-0.0001	0.0008	-0.16	0.8736
Depth of Imputed Clade	0.0006	0.0005	1.27	0.2045

Table 2: Statistical model of the potential drivers of the correlation between imputed and true ED values. Results of a multiple regression fitted to the data shown in figure 4, showing a relatively poor correlation between imputed and true ED scores ($F_{44791,8} = 29.1$, $r^2 = 0.005$, $p < 0.0001$). Given the extremely low predictive power of this statistical model we are reticent to make strong claims about drivers of the correlation between imputed and observed ED.

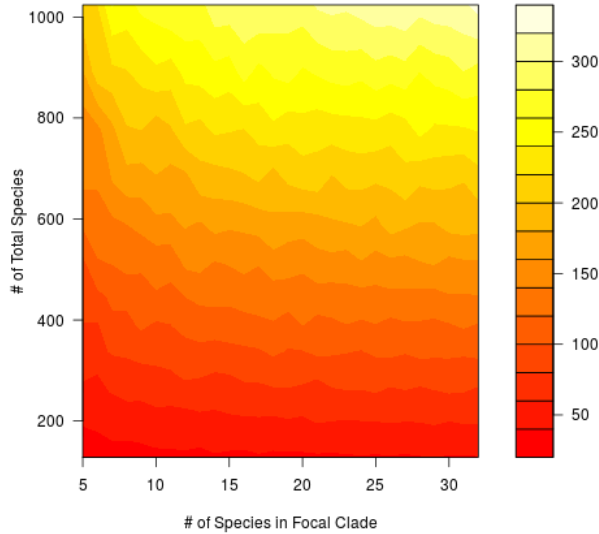


Figure 5: Mean ranking error of imputed species. An interpolated heat-map of the mean ranking error of imputed species as a function of the total number of species in the phylogeny (vertical axis) and number of species in the focal (imputed) clade (horizontal axis). Table 3 gives statistical support for the trend of increased error in larger phylogenies and imputed clades.

	Estimate	Std. Error	t value	Pr(> t)
Intercept	-1.6344	0.0332	-49.29	0.0001
Size of focal (imputed) clade	0.0900	0.0010	91.22	<0.0001
Size of phylogeny	0.5179	0.0013	383.99	<0.0001

Table 3: Statistical mode of the effect of clade and phylogeny size on ranking error. Model of the raw data underlying figure 5, regressing the ranking error of imputed species against the number of species in the imputed clade and the entire phylogeny ($F_{47997,2} = 77890$, $r^2 = 0.7644$, $p < 0.0001$). As can be seen in figure 5, the average ranking error is positively correlated with the size of the clade being imputed and the entire phylogeny. Square-root transformations have been applied to both ranking error and size of phylogeny.

Discussion

Phylogenies are increasingly playing a role in conservation prioritization, decision-making, and policy (Vézquez & Gittleman 1998). A major obstacle to a more widespread adoption of phylogenetic prioritization methods such as EDGE is phylogenetic uncertainty (Collen 2015). There is a tension between the need to make decisions to preserve biodiversity—including evolutionary history—now, and the reality that we do not have complete information about the phylogenetic placement of many species of conservation concern. Isaac & Pearse (in press), in a review of EDGE-like metrics, framed this as an urgent wish not to “*fiddle while Rome burns*” tempered by a fear of “*saving the wrong city*”. The intention of our study is to provide concrete information about the impact of such phylogenetic uncertainty on conservation prioritization. To address this uncertainty, we answered two key questions: (1) the extent to which species that are missing from the tree of life impact the ED scores of species for which we do have data, and (2) the extent to which phylogenetic imputation can accurately estimate ED scores for taxa with no phylogenetic data. For the first question (1), we found that while missing species do impact the ED scores of other species, the effects are not always severe, particularly if species are missing at random from the tree of life. For the second question (2), we found limited evidence that phylogenetic imputation can reconstruct species’ ED scores and rankings.

Our results are derived solely from simulations under a simple model of diversification—the Yule model. We do acknowledge that, in reality, lineages evolve in more complex ways than are captured by such a simple model. However, such complexities are unlikely to make imputation easier, and as they would have to be modelled and quantified in empirical studies, they would likely make imputation more difficult and uncertain. We suggest that focusing on the simplest model of diversification makes our results more generalizable. Further, we focus here solely on the results from a single imputation in each simulation, despite, empirically, biologists reporting average ED scores calculated across pseudo-posterior distributions of many imputed phylogenies (Kuhn et al. 2011). Thus our results show that the variation within these pseudo-posterior distributions is likely very large. It is well-known that such imputation methods are not biased (indeed, this was originally shown by Kuhn et al. 2011): here we emphasize that the uncertainty they introduce is sufficiently large that they may be less informative than previously been thought.

ED scores are relatively robust to missing species

Missing species and poor phylogenetic resolution have been identified as causes of uncertainty when calculating ED (Isaac et al. 2007), but we were unable to find a quantitative assessment of how missing species might

affect ED values of species for which data is available. Empirically in corals and gymnosperms, incomplete phylogenies produced similar results as later, more complete trees (Curnick et al. 2015; Forest et al. 2018). Our results support this finding. Indeed, our analysis suggests that, on average (and we emphasize that there is a good amount of variation about that average; see figure 3), a phylogeny missing 20% of species at random will still have ED scores that are strongly correlated ($\rho = 0.94$) with the true ED scores.

We did find that missing species are more problematic when those species are non-randomly distributed across the phylogeny. Our simulations do not examine extreme phylogenetic patterning, such as if an entire clade were missing. This is notable because clades that are geographically restricted to difficult-to-reach regions are both difficult to sequence and not uncommon (as is seen with 27 coral species in the Indian Ocean; Arrigoni et al. 2012). We also do not attempt to comprehensively simulate all of the different ways in which species could be missing from a phylogeny. We simply demonstrate that, compared to a scenario in which species are missing at random, phylogenetically patterned missing species can have a greater effect on the ED scores of species for which we have data.

Imputation does not reconstruct the ED value of species with great precision

Our results show that neither imputation (figures 4 and 5), nor clade-averages of ED (see Supplementary Materials), accurately recover the true ED values or the true ED rank of missing species. Thus we argue that, even though imputation allows missing species to be incorporated into EDGE lists, their associated EDGE scores may not accurately reflect their true scores. We acknowledge these are averages and may change depending on particular phylogeny, but we can find no statistically significant predictors of that variation.

While we did not assess clades with fewer than five species (we do not consider correlations or averages to be reliable with so few data-points), we cannot think why smaller clades would necessarily be more reliable (and this would be a reversal of the trend in figure 4). Indeed, in the smallest possible clade (two species), imputation is essentially sampling a terminal branch length from an exponential distribution (Kuhn et al. 2011); such a process should still lead to a great degree of uncertainty. It is, perhaps, unsurprising that imputed ED values do not correlate with their true values (see figure 4), but we were surprised at the degree of ranking error. Indeed, larger phylogenies showed *greater* ranking error; we naïvely would have expected the opposite. We would have expected our upper bound on the age of the imputed clade, which we would have expected to be relatively younger in larger phylogenies, to have somewhat controlled the range of the

ranks of our imputed species. ED is known to be driven mostly by terminal branch length (Redding et al. 2008; Isaac et al. 2007; Steel et al. 2007); our results therefore emphasize this.

Imputation is not the only way to incorporate missing species into EDGE-like frameworks (see Collen et al. 2011; Gumbs et al. 2018), but it is likely the most common. 3,330 of the birds (~30%; Jetz et al. 2014), 250 of the mammals (~5.6%; Collen et al. 2011), and 610 of the sharks (~49%; Stein et al. 2018) in recent EDGE lists were imputed. It is well-known that phylogenetic imputation can cause biases in other statistical methods, such as the estimation of evolutionary phylogenetic signal (Rabosky 2015). We emphasize that we are not suggesting that imputation *biases* ED scores: we are, instead, suggesting that it is less precise than has previously been acknowledged.

Guidelines for the use of imputation

The impact of imputation on EDGE scores is almost certainly less than its impact on ED scores, because EDGE scores are a product of both ED and IUCN status ('GE'). However, the goal of EDGE-like measures is to incorporate phylogeny, and if imputed EDGE scores are driven by their GE component because of uncertainty introduced by imputation, this essentially creates another metric of IUCN status.

Our results suggest that incomplete phylogenies can be used to estimate ED scores with remarkably high degrees of accuracy. Instead of using imputation to account for the relatively minor impact of missing species, we suggest that conservation biologists concerned without accounting for phylogenetic uncertainty should focus on the species for which they have data. While we have not explored this uncertainty here, evolutionary biologists commonly work with distributions of trees generated from genetic data (reviewed in Huelsenbeck et al. 2001; Bollback 2005), since the precise topology and dating of a phylogeny is almost always uncertain. This uncertainty has, indeed, already been shown to affect EDGE scores and rankings (Pearse et al. 2015). If biologists are concerned about the impact of missing species on known species' ED(GE) scores we see no harm in being precautionary and using imputation. It is important, however, to focus on the major sources of potential error first, and so we would encourage biologists to incorporate uncertainty in species with phylogenetic data as a first priority.

Our results suggest that prioritizing species whose phylogenetic structure has been imputed should be done with extreme care, if it at all. In the case that an imputed species is imputed to be below a threshold set for conservation (most EDGE studies focus on the 'top 100' species or something similar), then the path forward is clear: that species should not have conservation funds allocated to it at this time. The case where

a species, on average, passes a threshold is more complex, but the theory underlying imputation can give some guidance. Imputed distributions of trees essentially represent Bayesian posterior distributions (Kuhn et al. 2011), and so the 95% posterior densities of these distributions' ED values represent a range within which we can be 95% certain the true ED scores lie (if the model assumptions are met). Thus we suggest that conservation action should only be initiated for a species if there is a 95% (or 80%, or whatever confidence is deemed appropriate) probability that it is above that threshold. For example, a species whose ranking is estimated to have a 20% probability of being between the 30th and 100th highest-ranked species could not, with confidence, be called a top-100 species. Our results suggest that, on average, very few imputed species will meet such a criterion. Regardless, the calculations of such probabilities is trivial with the data users of imputation have in-hand already.

Ultimately, we are currently fighting a losing battle to preserve the tree of life. Our results are good news: they suggest that we can start right away using the phylogenies we already have in-hand. The effect of missing species is mild enough that we often do not need costly and time-consuming imputation, and imputation rarely gives us sufficiently precise estimates of species' ED scores anyway. We suggest that, given we do not have the resources to save everything, we should consider focusing our efforts on those species whose ED scores we can know with greater certainty: those for which we have data.

Acknowledgments

We are grateful to E. Simpson, M. Sneddon, J. Stachewicz, J. Rosindell, and XXX anonymous reviewers for providing constructive feedback on this manuscript.

References

- Arrigoni, R., Stefani, F., Pichon, M., Galli, P., & Benzoni, F. (2012). Molecular phylogeny of the robust clade (Faviidae, Mussidae, Merulinidae, and Pectiniidae): an Indian Ocean perspective. *Molecular Phylogenetics and Evolution* 65.1, 183–193.
- Barnosky, A. D., Matzke, N., Tomiya, S., Wogan, G. O., Swartz, B., Quental, T. B., Marshall, C., McGuire, J. L., Lindsey, E. L., Maguire, K. C., et al. (2011). Has the Earth’s sixth mass extinction already arrived? *Nature* 471.7336, 51–57.
- Bollback, J. (2005). “Posterior Mapping and Posterior Predictive Distributions”. *Statistical Methods in Molecular Evolution*. Ed. by R. Nielsen. Springer New York. Chap. 16, pp. 439–462.
- Bortolussi, N., Durand, E., Blum, M., & Blum, M. M. (2009). Package ‘apTreeshape’.
- Bottrill, M. C., Joseph, L. N., Carwardine, J., Bode, M., Cook, C., Game, E. T., Grantham, H., Kark, S., Linke, S., McDonald-Madden, E., et al. (2008). Is conservation triage just smart decision making? *Trends in Ecology & Evolution* 23.12, 649–654.
- Brooks, T. M., Mittermeier, R. A., Mittermeier, C. G., Da Fonseca, G. A., Rylands, A. B., Konstant, W. R., Flick, P., Pilgrim, J., Oldfield, S., Magin, G., et al. (2002). Habitat loss and extinction in the hotspots of biodiversity. *Conservation biology* 16.4, 909–923.
- Butchart, S. H. & Bird, J. P. (2010). Data deficient birds on the IUCN Red List: what don’t we know and why does it matter? *Biological Conservation* 143.1, 239–247.
- Ceballos, G., Ehrlich, P. R., Barnosky, A. D., García, A., Pringle, R. M., & Palmer, T. M. (2015). Accelerated modern human-induced species losses: Entering the sixth mass extinction. *Science advances* 1.5, e1400253.
- Collen, B. (2015). Conservation prioritization in the context of uncertainty. *Animal Conservation* 18.4, 315–317.
- Collen, B., Turvey, S. T., Waterman, C., Meredith, H. M., Kuhn, T. S., Baillie, J. E., & Isaac, N. J. (2011). Investing in evolutionary history: implementing a phylogenetic approach for mammal conservation. *Philosophical Transactions of the Royal Society of London B: Biological Sciences* 366.1578, 2611–2622.
- Colless, D. H. (1982). Review of phylogenetics: the theory and practice of phylogenetic systematics. *Systematic Zoology* 31.1, 100–104.
- Curnick, D., Head, C., Huang, D., Crabbe, M. J. C., Gollock, M., Hoeksema, B., Johnson, K., Jones, R., Koldewey, H., Obura, D., et al. (2015). Setting evolutionary-based conservation priorities for a phylogenetically data-poor taxonomic group (Scleractinia). *Animal Conservation* 18.4, 303–312.

- Dulvy, N. K., Fowler, S. L., Musick, J. A., Cavanagh, R. D., Kyne, P. M., Harrison, L. R., Carlson, J. K., Davidson, L. N., Fordham, S. V., Francis, M. P., et al. (2014). Extinction risk and conservation of the world’s sharks and rays. *Elife* 3.
- Faith, D. P. (1992). Conservation evaluation and phylogenetic diversity. *Biological conservation* 61.1, 1–10.
- Faith, D., Carter, G., Cassis, G., Ferrier, S., & Wilkie, L. (2003). Complementarity, biodiversity viability analysis, and policy-based algorithms for conservation. *Environmental Science & Policy* 6.3, 311–328.
- Felsenstein, J. (2005). Using the quantitative genetic threshold model for inferences between and within species. *Philosophical Transactions of the Royal Society of London B: Biological Sciences* 360.1459, 1427–1434.
- Forest, F., Moat, J., Baloch, E., Brummitt, N. A., Bachman, S. P., Ickert-Bond, S., Hollingsworth, P. M., Liston, A., Little, D. P., Mathews, S., et al. (2018). Gymnosperms on the EDGE. *Scientific reports* 8.1, 6053.
- Good, T. C., Zihra, M. L., & Kremen, C. (2006). Addressing data deficiency in classifying extinction risk: a case study of a radiation of Bignoniaceae from Madagascar. *Conservation Biology* 20.4, 1099–1110.
- Gumbs, R., Gray, C. L., Wearn, O. R., & Owen, N. R. (2018). Tetrapods on the EDGE: Overcoming data limitations to identify phylogenetic conservation priorities. *PloS One* 13.4, e0194680.
- Huelsenbeck, J. P., Ronquist, F., Nielsen, R., & Bollback, J. P. (2001). Bayesian inference of phylogeny and its impact on evolutionary biology. *science* 294.5550, 2310–2314.
- Isaac, N. J. B. & Pearse, W. D. (in press). “The Use of EDGE (Evolutionary Distinct Globally Endangered) and EDGE-Like Metrics to Evaluate Taxa for Conservation”. *Phylogenetic Diversity*. Ed. by R. Scherson & D. Faith. Springer International Publishing AG.
- Isaac, N. J., Redding, D. W., Meredith, H. M., & Safi, K. (2012). Phylogenetically-informed priorities for amphibian conservation. *PLoS one* 7.8, e43912.
- Isaac, N. J., Turvey, S. T., Collen, B., Waterman, C., & Baillie, J. E. (2007). Mammals on the EDGE: conservation priorities based on threat and phylogeny. *PloS one* 2.3, e296.
- IUCN (2001). *IUCN Red List categories and criteria*. IUCN.
- (2008). Guidelines for Using the IUCN Red List Categories and Criteria, Version 7.0. *Prepared by the Standards and Petitions Working Group of the IUCN SSC Biodiversity Assessments Sub-Committee in August 2008*.
- Jensen, E. L., Mooers, A. Ø., Cacccone, A., & Russello, M. A. (2016). I-HEDGE: determining the optimum complementary sets of taxa for conservation using evolutionary isolation. *PeerJ* 4, e2350.
- Jetz, W., Thomas, G. H., Joy, J. B., Redding, D. W., Hartmann, K., & Mooers, A. O. (2014). Global distribution and conservation of evolutionary distinctness in birds. *Current Biology* 24.9, 919–930.

- Komsta, L. & Novomestky, F. (2015). Moments, cumulants, skewness, kurtosis and related tests. *R package version 0.14*.
- Kuhn, T. S., Mooers, A. Ø., & Thomas, G. H. (2011). A simple polytomy resolver for dated phylogenies. *Methods in Ecology and Evolution* 2.5, 427–436.
- Lips, K. R., Brem, F., Brenes, R., Reeve, J. D., Alford, R. A., Voyles, J., Carey, C., Livo, L., Pessier, A. P., & Collins, J. P. (2006). Emerging infectious disease and the loss of biodiversity in a Neotropical amphibian community. *Proceedings of the national academy of sciences of the United States of America* 103.9, 3165–3170.
- McBride, M. F., Wilson, K. A., Bode, M., & Possingham, H. P. (2007). Incorporating the effects of socioeconomic uncertainty into priority setting for conservation investment. *Conservation Biology* 21.6, 1463–1474.
- Molnar, J. L., Gamboa, R. L., Revenga, C., & Spalding, M. D. (2008). Assessing the global threat of invasive species to marine biodiversity. *Frontiers in Ecology and the Environment* 6.9, 485–492.
- Morais, A. R., Siqueira, M. N., Lemes, P., Maciel, N. M., De Marco Jr, P., & Brito, D. (2013). Unraveling the conservation status of Data Deficient species. *Biological conservation* 166, 98–102.
- Myers, N., Mittermeier, R. A., Mittermeier, C. G., Da Fonseca, G. A., & Kent, J. (2000). Biodiversity hotspots for conservation priorities. *Nature* 403.6772, 853–858.
- Naidoo, R., Balmford, A., Ferraro, P. J., Polasky, S., Ricketts, T. H., & Rouget, M. (2006). Integrating economic costs into conservation planning. *Trends in ecology & evolution* 21.12, 681–687.
- Orme, D., Freckleton, R., Thomas, G., Petzoldt, T., Fritz, S., Isaac, N., & Pearse, W. D. (2013). *caper: Comparative Analyses of Phylogenetics and Evolution in R*. R package version 0.5.2.
- Paradis, E., Claude, J., & Strimmer, K. (2004). APE: analyses of phylogenetics and evolution in R language. *Bioinformatics* 20, 289–290.
- Pearse, W. D., Chase, M. W., Crawley, M. J., Dolphin, K., Fay, M. F., Joseph, J. A., Powney, G., Preston, C. D., Rapacciolo, G., Roy, D. B., et al. (2015). Beyond the EDGE with EDAM: prioritising British plant species according to evolutionary distinctiveness, and accuracy and magnitude of decline. *PloS one* 10.5, e0126524.
- Pennell, M. W., Eastman, J. M., Slater, G. J., Brown, J. W., Uyeda, J. C., FitzJohn, R. G., Alfaro, M. E., & Harmon, L. J. (2014). geiger v2. 0: an expanded suite of methods for fitting macroevolutionary models to phylogenetic trees. *Bioinformatics* 30.15, 2216–2218.
- Pounds, J. A., Bustamante, M. R., Coloma, L. A., Consuegra, J. A., Fogden, M. P., Foster, P. N., La Marca, E., Masters, K. L., Merino-Viteri, A., Puschendorf, R., et al. (2006). Widespread amphibian extinctions from epidemic disease driven by global warming. *Nature* 439.7073, 161–167.

- Pressey, R., Humphries, C., Margules, C. R., Vane-Wright, R., & Williams, P. (1993). Beyond opportunism: key principles for systematic reserve selection. *Trends in ecology & evolution* 8.4, 124–128.
- Pybus, O. G. & Harvey, P. H. (2000). Testing macro-evolutionary models using incomplete molecular phylogenies. *Proceedings of the Royal Society of London B: Biological Sciences* 267.1459, 2267–2272.
- R Core Team (2017). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing. Vienna, Austria.
- Rabosky, D. L. (2015). No substitute for real data: a cautionary note on the use of phylogenies from birth–death polytomy resolvers for downstream comparative analyses. *Evolution* 69.12, 3207–3216.
- Redding, D. (n.d.). Incorporating genetic distinctness and reserve occupancy into a conservation prioritisation approach 2003. *Master’s thesis, University of East Anglia, Norwich* [OpenURL](#).
- Redding, D. W., Hartmann, K., Mimoto, A., Bokal, D., DeVos, M., & Mooers, A. Ø. (2008). Evolutionarily distinctive species often capture more phylogenetic diversity than expected. *Journal of theoretical biology* 251.4, 606–615.
- Revell, L. J. (2012). phytools: An R package for phylogenetic comparative biology (and other things). *Methods in Ecology and Evolution* 3, 217–223.
- Steel, M., Mimoto, A., & Mooers, A. Ø. (2007). Hedging our bets: the expected contribution of species to future phylogenetic diversity. *Evolutionary bioinformatics online* 3, 237.
- Stein, R. W., Mull, C. G., Kuhn, T. S., Aschliman, N. C., Davidson, L. N., Joy, J. B., Smith, G. J., Dulvy, N. K., & Mooers, A. O. (2018). Global priorities for conserving the evolutionary history of sharks, rays and chimaeras. *Nature Ecology & Evolution*, 1.
- Thomas, C. D., Cameron, A., Green, R. E., Bakkenes, M., Beaumont, L. J., Collingham, Y. C., Erasmus, B. F., De Siqueira, M. F., Grainger, A., Hannah, L., et al. (2004). Extinction risk from climate change. *Nature* 427.6970, 145–148.
- Thomas, G. H., Hartmann, K., Jetz, W., Joy, J. B., Mimoto, A., & Mooers, A. O. (2013). PASTIS: an R package to facilitate phylogenetic assembly with soft taxonomic inferences. *Methods in Ecology and Evolution* 4.11, 1011–1017.
- Tonini, J. F. R., Beard, K. H., Ferreira, R. B., Jetz, W., & Pyron, R. A. (2016). Fully-sampled phylogenies of squamates reveal evolutionary patterns in threat status. *Biological Conservation* 204, 23–31.
- Vézquez, D. P. & Gittleman, J. L. (1998). Biodiversity conservation: does phylogeny matter? *Current Biology* 8.11, R379–R381.
- Wake, D. B. & Vredenburg, V. T. (2008). Are we in the midst of the sixth mass extinction? A view from the world of amphibians. *Proceedings of the National Academy of Sciences* 105.Supplement 1, 11466–11473.
- Weitzman, M. L. (1998). The Noah’s ark problem. *Econometrica*, 1279–1298.

413 Wilson, K. A., Underwood, E. C., Morrison, S. A., Klausmeyer, K. R., Murdoch, W. W., Reyers, B., Wardell-
414 Johnson, G., Marquet, P. A., Rundel, P. W., McBride, M. F., et al. (2007). Conserving biodiversity
415 efficiently: what to do, where, and when. *PLOS biology* 5.9, e223.

Supplementary material and analyses

Article title: The Effect of Phylogenetic Uncertainty and Imputation on EDGE Scores

Running head: The Effect of Phylogenetic Uncertainty and Imputation on EDGE Scores

Authors: K. Bodie Weedop¹, Arne Ø. Mooers², Caroline M. Tucker³, and William D. Pearse¹

¹ Department of Biology & Ecology Center, Utah State University, 5305 Old Main Hill, Logan UT, 84322

² Department of Biological Sciences, Simon Fraser University, Burnaby, British Columbia, Canada

³ Department of Biology, University of North Carolina–Chapel Hill

*To whom correspondence should be addressed: will.pearse@usu.edu

A. Effect of Measures of the True, Full Phylogenies

This section contains analyses of a number of metrics from the original phylogenies and their effect on the correlation coefficient of ED values described in body of the manuscript.

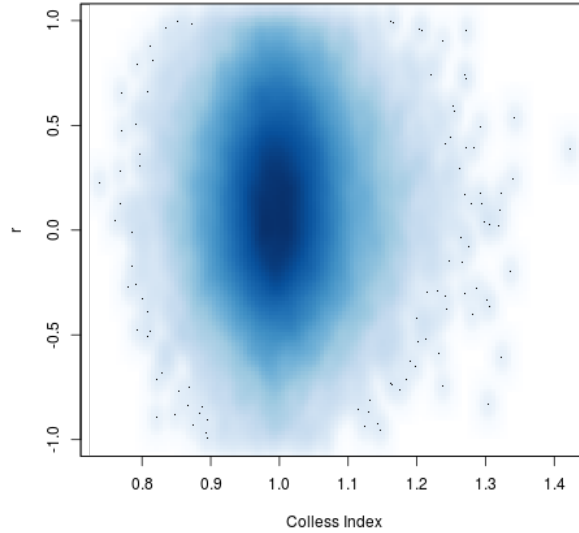


Figure 1A: **Effect of the True Colless Index of Full Phylogeny.** Smoothed color density plot of the effect of colless' index (prior to imputation) on the correlation coefficient of ED values.

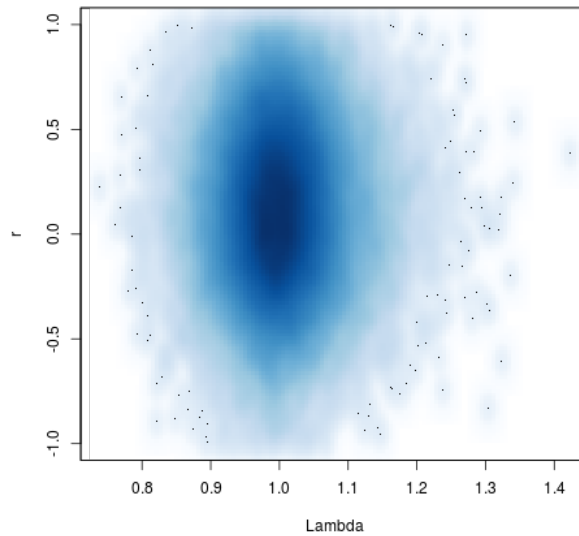


Figure 2A: **Effect of the True Lambda of Full Phylogeny.** Smoothed color density plot of estimated speciation rate (λ ; prior to imputation) and its effect on the correlation coefficient of ED values.

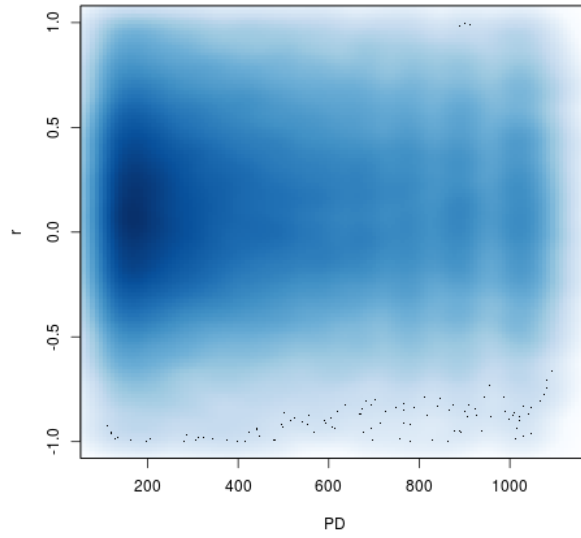


Figure 3A: **Effect of True PD of Full Phylogeny.** Smoothed color density plot of the total phylogenetic diversity (prior to imputation) and its effect the correlation coefficient of ED values.

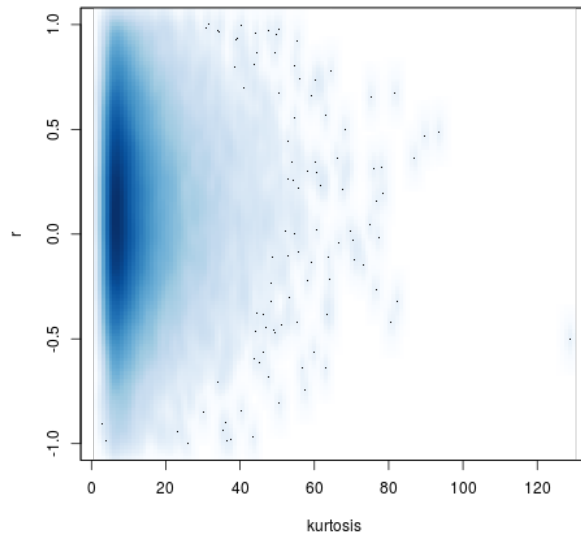


Figure 4A: **Effect of the True Kurtosis of Full Phylogeny.** Smoothed color density plot of kurtosis of ED values (prior to imputation) and its effect the correlation coefficient of ED values.

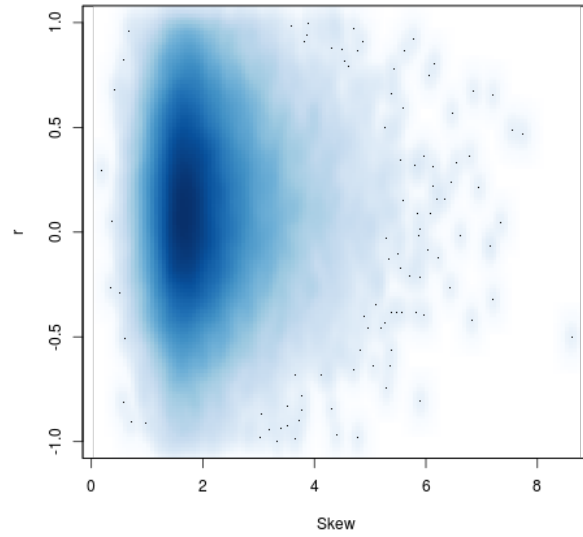


Figure 5A: **Effect of the True Skew of Full Phylogeny.** Smoothed color density plot of the skew of ED values (prior to imputation) and its effect the correlation coefficient of ED values.

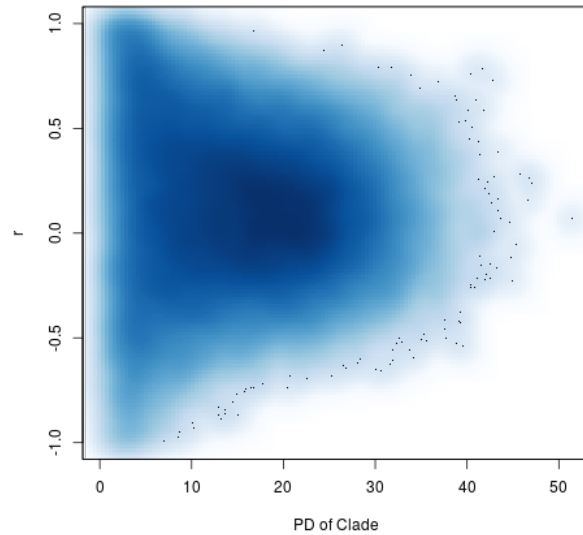


Figure 6A: **Effect of the True PD of The Selected Clade.** Smoothed color density plot of the total phylogenetic diversity of the focal clade (prior to imputation) and its effect the correlation coefficient of ED values.

427 **B. Error Rate in Top Rankings**

428 This section contains plots of mean error rates in ED rankings of the top species overall.

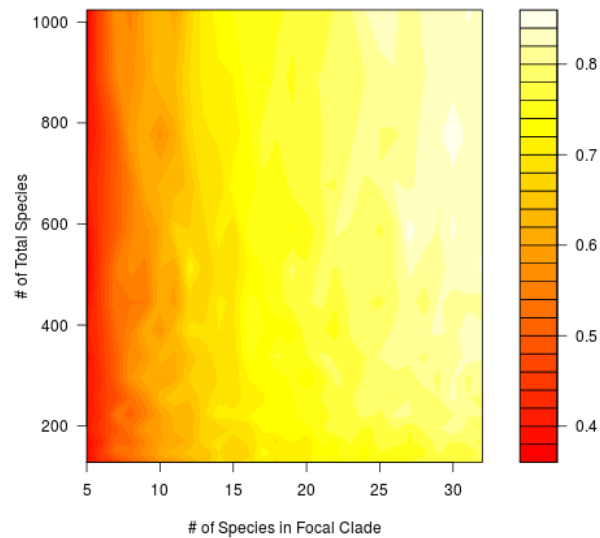


Figure 1B: **Mean error rate in the ranking of top 50 species.** Heat map plot demonstrating the average error rate in ED ranking of the top 50 species. The gradient on the right gives the average error rate and the corresponding color.

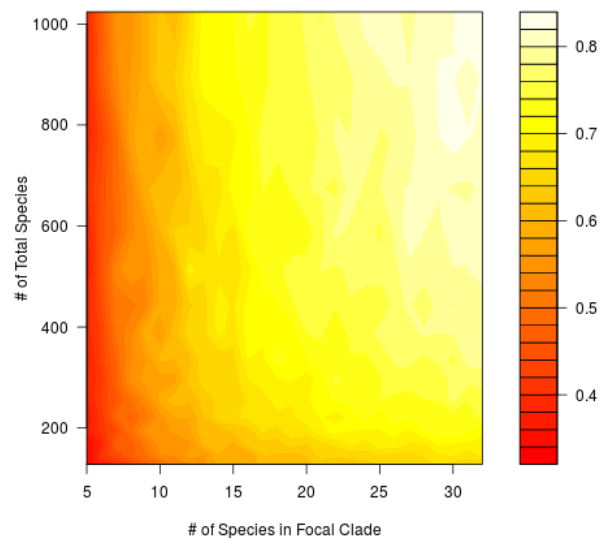


Figure 2B: **Mean error rate in the ranking of top 100 species.** Heat map plot demonstrating the average error rate in ED ranking of the top 100 species. The gradient on the right gives the average error rate and the corresponding color.

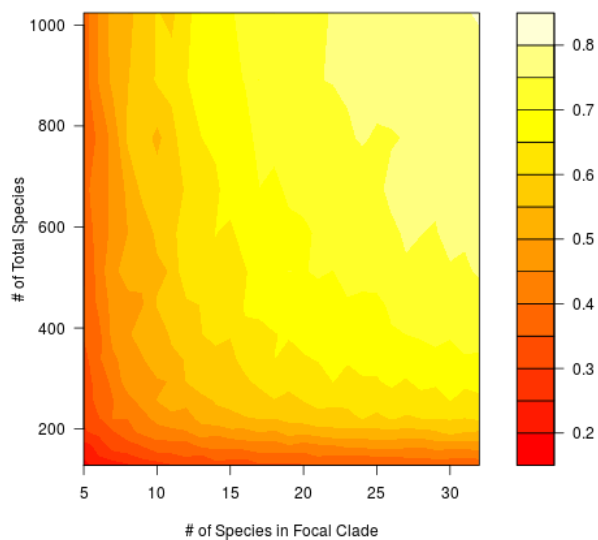


Figure 3B: **Mean error rate in the ranking of top 200 species.** Heat map plot demonstrating the average error rate in ED ranking of the top 50 species. The gradient on the right gives the average error rate and the corresponding color.

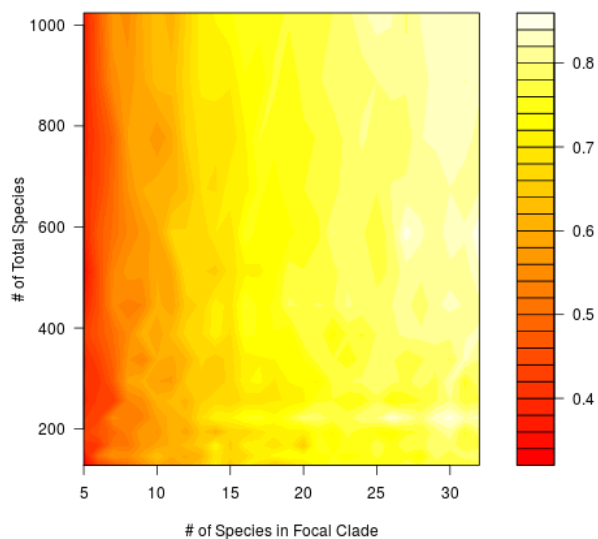


Figure 4B: **Mean error rate in the ranking of top 5% of species.** Heat map plot demonstrating the average error rate in ED ranking of the top 5% of all species in the phylogeny. The gradient on the right gives the average error rate and the corresponding color.

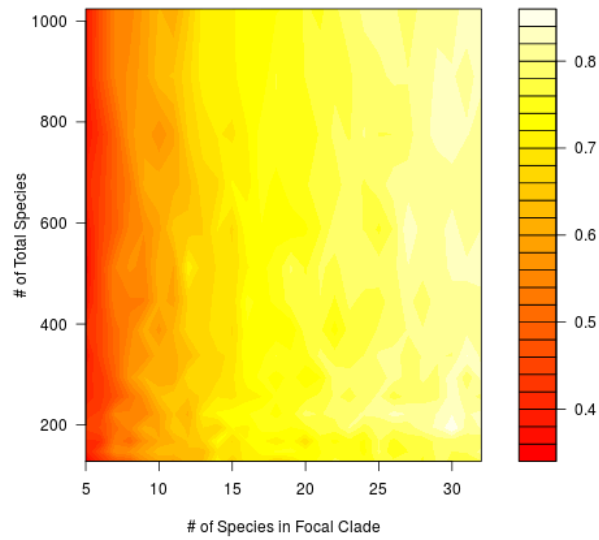


Figure 5B: **Mean error rate in the ranking of top 10% of species.** Heat map plot demonstrating the average error rate in ED ranking of the top 10% of all species in the phylogeny. The gradient on the right gives the average error rate and the corresponding color.

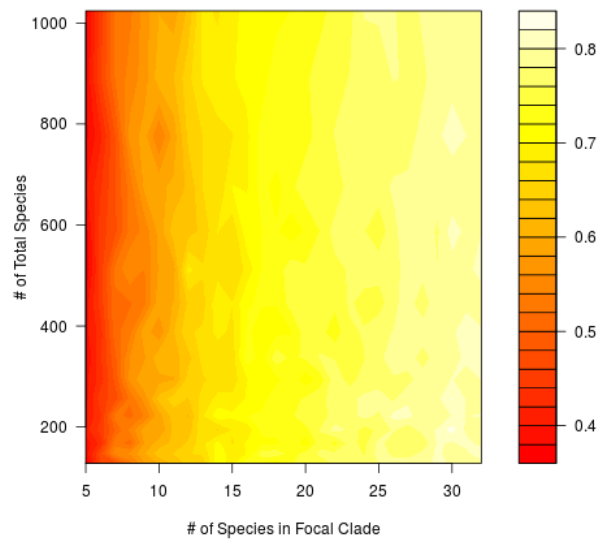


Figure 6B: **Mean error rate in the ranking of top 20% of species.**Heat map plot demonstrating the average error rate in ED ranking of the top 20% of all species in the phylogeny. The gradient on the right gives the average error rate and the corresponding color.

C. Ranking Error When Using Average ED Value

This section contains a plot of the mean ranking error when species in the focal clades were assigned the average ED value of the clade.

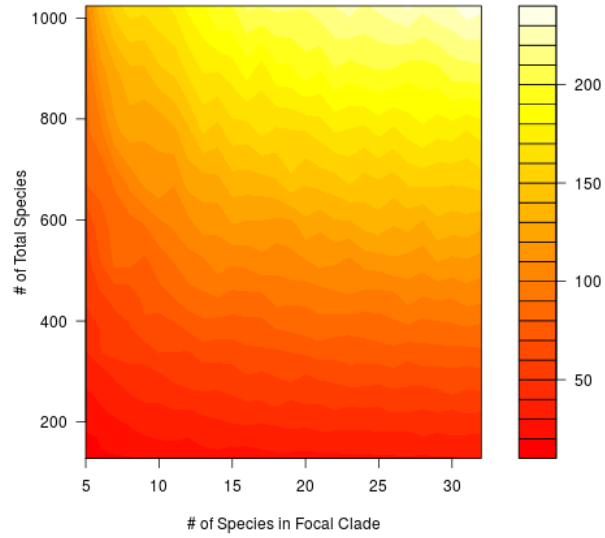


Figure 1C: **Mean Ranking Error of Species Assigned Average ED.** Heat map plot demonstrating the average ranking error when species in the focal clade were assigned the average ED value of the clade. The gradient on the right demonstrates average number of positions within the full ranking that focal clade species shifted from their true rank.

D. Ranking Error of Non-imputed Species

This section contains a plot of visualizing the mean ranking error of of non-imputed species when species in the focal clade were imputed.

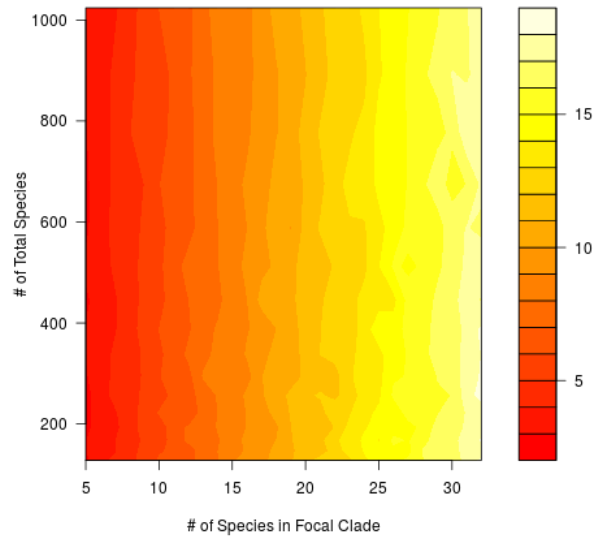


Figure 1D: **Ranking Error of Non-imputed Species.** Heat map plot demonstrating the average ranking error of non-imputed species when species in the focal clade were imputed. The gradient on the right demonstrates average number of positions within the full ranking that focal clade species shifted from their true rank.