

Title: Are super-predators also super-scary? A systematic review and meta-analysis of animal responses to human interactions

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Abstract

Human induced rapid environmental change has been recognized as a global threat to natural systems and the organisms that inhabit them. Human hunters and fishers interact with animals in natural spaces as predators, and are more effective in capturing prey and reducing populations than natural predators overall. On the other hand, seemingly benign interactions with humans, such as tourism, may also be perceived as threatening and have negative impacts on animal populations. Here, we provide a synthesis of the current knowledge of the combined effects of lethal and non-lethal human interactions on the behavioural responses of animals in natural spaces. We conducted a systematic review and meta-analysis of literature published over the past three decades, and built a dataset to determine the relative effect of lethal and non-lethal human interactions on foraging, vigilance, and movement behaviors of wild animals. Lethal human interactions included hunting, fishing, and retaliatory killing in response to conflict, and non-lethal human interactions were distinguished as active (snorkeling, scuba diving, pedestrians, hiking, and tourists) or passive (roads, human settlements, sonar, and boat traffic). We also considered how the trophic level of the species and body size could influence the effect of human interactions on animal behavior. Our findings show that lethal human interactions had a significant effect on the behaviors of animals, causing animals that are actively hunted by humans to increase vigilance, reduce foraging, and alter movements. Both active and passive non-lethal interactions had weaker and non-significant effects on altering these behaviours of animals. Overall, our meta-analysis shows that the impact of non-lethal human activities in eliciting fear-driven behavioural changes in prey does not seem to have broad empirical support. It also suggests that the intensity of effect of the “human super-predator” may depend on the trophic level of affected species, or the history of the human-animal interactions in that landscape, In

43 order to understand the nuances of these effects, systematic studies across more species and
44 geographic regions are needed.

45

46 **Keywords:** HIREC, risk – disturbance hypothesis, non – consumptive effects, predator – prey
47 interactions, fear

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I. Introduction

Humans are a pervasive species that interact with and change their environment in a multitude of ways. Over the past two centuries, human development has led to the alteration of global ecosystems at a rate that has not been witnessed in the history of the planet. From climate change to the rampant extraction of biological resources, the effects of human-induced rapid environmental change (HIREC) are large in scale and magnitude (Estes et al. 2011; Sih 2013). Given their unique ecology, humans not only occupy every eco-region of the world, but have multiple functional roles that can shift trophic ecology in intensive and extensive ways. Animals therefore have no choice but to respond to human activities in the natural world.

Throughout most of their evolutionary history, humans have been hunter–gatherers (Treves and Naughton-Treves 1999). As predators, however, humans are extraordinarily deadly, as they out-compete most natural predators in the rate and efficacy of killing their prey (Darimont et al. 2009; 2015). Through social organisation and tool use, humans can target multiple trophic levels at high efficiency (Treves and Naughton-Treves 1999). Thus, humans have been referred to as generalist super-predators (Clinchy et al. 2016), taking a disproportionately large number of animals from ecosystems. In fact, the rate of mortality of prey species from human predators, such as hunters and fishers, are much greater in magnitude than the sum of kill by all natural predators in both terrestrial and marine ecosystems (Darimont et al. 2015). Thus, humans create truly risky conditions for both wild prey and predators in natural systems (Oriol-Cotterill et al. 2015).

In addition to hunting and fishing, humans also interact with wild animals through non–lethal activities (Boyle and Samson 1985). Non-lethal interactions of humans with animals can be

further distinguished as active or passive. Active non-lethal interactions include tourism, walking in parks, hiking in nature reserves, mountaineering and off-road bicycling, and have been shown to negatively influence animals even-though the humans in their habitats are not actively killing them (e.g. Loehr et al. 2005; Newey 2007; Brown et al. 2012; Montero-Quintana et al. 2020). Passive non-lethal interactions are even more pervasive, as they include alterations to environments. Roads and settlements, in particular, have the greatest potential to influence the behaviour of animals (e.g. Mehlhoop et al. 2002; Cappa et al. 2017, etc). Given the number of bird collisions on glass (Loss, Will, and Marra 2015) and high occurrence of roadkill globally (Benson, Mahoney, and Patterson 2015), animals are facing mortality from human interactions that they may not perceive as risky.

Other non-lethal activities such as air and sea travel may also affect animal behavior (e.g. Van der Kolk et al. 2024; see Harris et al. 2019 for review). Although compelling evidence suggests that non-lethal human activities can be extremely strong in changing animal behaviours (Boyle and Samson 1985; selected examples: (Davis et al. 1997; Christopher M. Papouchis et al. 2001; Nevin and Gilbert 2005; Westekemper et al. 2018; Ladle et al. 2019), many other studies show that animals habituate quickly to human encounters (see Smith et al. 2024). Thus, it remains unclear whether the non-lethal interactions of humans on animals are as strong as the lethal interactions in eliciting fear and changing behaviours.

Given that humans have both lethal and non-lethal interactions with animals in the wild, the individual and combined effects of these on animal behaviors can potentially be very high. In a recent narrative review, Smith et al. (2024) outlined a framework describing the pathways by which lethal and non-lethal human interactions can change phenotypic (behavioral and

physiological) responses of animals, with consequence for population growth and abundance. They find that the perceived risk of humans can induce phenotypic changes in wild animals, but as expected, the evidence is mixed (Smith et al 2024). What is still missing is a quantitative assessment of the relative magnitude of effect for different types of human activities, which would illustrate when and why animals respond to human-induced risk.

In this paper, we use a meta-analysis to synthesize the current evidence for human-induced behavioural changes in wild animals. We separated studies based on whether the interactions were lethal (hunting, fishing), active and nonlethal (ecotourism, hiking) and passive and nonlethal (roads and human settlements). One of the clearest and fastest responses that prey use to minimize risk is by shifting behaviour. Decades of research has shown that natural predators change their foraging and vigilance patterns, as well as the movement and habitat use patterns of prey (Brown and Kotler 2004; Sih 1984). Thus, we focused on changes in foraging behaviours (bite rates and foraging time), vigilance (vigilance time and proportion) and movement (displacement, home range size and rate) of animals due to human interactions. We predict that in areas where humans hunt, animals would strongly invest in anti-predator behavioural strategies, spending more time being vigilant, less time in foraging, and increasing their movement rate or displacement (Figure 1). We also expect animals to perceive non-lethal activities of humans as threatening but to a lesser degree; thus, we expect a lower intensity of anti predator responses to active and passive non-lethal interactions of humans (Figure 1). We also tested the potential influence of body size on the magnitude of behavioural effects based on the hypothesis that larger animals would react to a greater extent to human disturbance as they are disproportionately targeted by the lethal actions of humans.

From a systematic review of published studies, we derived effect sizes that describe the magnitude of behavioural change across these categories of human interactions. We synthesise our meta-analysis to determine whether non-lethal effects of humans that are active or passive influence wild animal behaviours to the same extent as lethal human effects. Finally, we highlight where gaps in the literature have yielded skewed conclusions about human impacts on the behaviour of animals in natural spaces and suggest avenues for future research in this field.

II. Methods

To determine the strength of behavioural change (i.e. effect sizes) caused by human activities, we conducted a systematic literature survey that included studies where foraging, vigilance or movement behaviours were measured in animals in response to lethal human interactions (hunting, including poaching and trapping, and fishing), nonlethal but active interactions (walking in parks, hiking in nature reserves, off-road bicycling and animal watching), and nonlethal but passive interactions (roads and human settlements).

1. Search and literature database

We conducted a scoping search using the web of science database and their advanced search function. Our initial search term was “(risk OR fear) AND (human OR anthropogenic) AND (behavio*) AND ALL=(predate*)”. The search terms were applied across all fields. The initial search returned 2319 results on 12th January 2023. We downloaded the data and conducted a literature mapping analysis to identify additional search terms and updated our search string to “(human OR anthropo*) AND (risk OR fear OR NCE OR trait mediated effect OR

nonconsumptive* effect) AND (behavio*) AND (predat*)”. Using our updated search string, we conducted searches on the SCOPUS and WOS database as well as grey literature databases such as OpenThesis on 14th February 2023. We then used Rayyan, a literature management and screening software to deduplicate our database (Ouzzani et al. 2016; Foo et al. 2021). After deduplication we were left with 7562 abstracts. This review was not registered.

2. Initial screening

We piloted our initial screening protocol on 100 randomly selected papers using Rayyan (Ouzzani et al., 2016). We decided inclusion and exclusion criteria by evaluating titles, abstracts, and key words to determine the population studied, the exposure or intervention to the population, the control if any and the outcome (see Appendix A.1 for detailed inclusion criteria). The initial screening was conducted by three independent researchers. The three researchers discussed conflicting decisions on inclusion or exclusion of abstracts until a consensus was reached. We applied the finalised screening protocol to all abstracts in our database and were left with 436 papers to further evaluate.

3. Full text screening and data collection

After conducting initial screening, the screened database was downloaded from Rayyan (Ouzzani et al., 2016). We then cleaned the database and extracted PICO data in a workable format using R. Papers chosen for inclusion during our initial screening were full-text screened. We determined our protocol based on variables in our initial screening, the methods used, study design and reporting (Appendix A.2) (Foo et al., 2021). We collected data on the study species

identity and trophic level, location of the study, treatment, controls, responses, mean effects, variance measures and sample size. We piloted the protocol on 39 papers. Where results were not clearly reported we either used published data to calculate mean effects and variance measures or used a plot digitiser to extract data from figures. Studies in which data extraction was not possible were excluded from our meta-analysis. We included 71 studies at this stage.

4. Backward and forward search

We identified key studies and reviews during our screening for backward and forward search. Backward search identified papers cited by these studies and reviews, and forward search identified papers that cite these key papers. We added the results to our existing database and deduplicated it to create the final paper database for the review (Foo et al., 2021). After screening the additional abstracts, we added an additional 40 studies to our final selection. Of these, 14 were added after full-text screening. We have included a PRISMA (EcoEvo) checklist and flowchart in Appendix A and E (Page et al. 2021; O’Dea et al. 2021).

5. Types of studies

There were several types of study approaches that measured animal responses to human activities. The majority of studies were observational or natural experiments. This included spatial or temporal contrast. Spatial contrasts included designs in which behaviors were compared across protected and unprotected areas for lethal interactions ($n = 8$), areas with and without tourist for active interactions ($n = 6$) and areas close and far away from roads or settlements for passive interactions ($n = 4$). Studies also employed auditory stimuli ($n = 5$) for example road

noise or human speech in areas with hunting pressure or high traffic to test for the effect of lethal interactions and passive interactions respectively. Other experimental studies employed direct pedestrian disturbance to simulate active interactions ($n = 3$) and lethal interactions in areas with hunting ($n = 3$). Four studies used temporal contrast in human activities such as hunting closures and tourist seasons to determine the effects of active ($n = 1$) and lethal ($n = 3$) interactions. Finally, ten studies did not implement or utilize a contrast for their comparison and instead determined correlations between human activities for active interactions and distance from road for passive interaction to study their effects on behavior.

We have focused on studies that deal with foraging, vigilance and movement behaviors. Studies dealing with foraging and vigilance behaviors used focal individual sampling and scan sampling. The measured behaviors were time spent vigilant, time spent foraging, frequency of vigilance behavior in a group, rate of individual vigilance behavior, and foraging (bite) rates. We included movement studies that measured movement rates (speed), displacement (distance traveled), and home range size using GPS tags, collars and transmitters. Home range size was included as it is function of displacement and speed in a given sampling period.

6. *Meta-analysis*

We used studies that reported all necessary metrics for comparison including treatment means, measures of uncertainty, and sample sizes for further analysis (Mikolajewicz and Komarova 2019, Appendix A.2). We did not include papers that only reported test statistics in our meta-analysis. We gathered secondary data including trophic level, dietary guild and body mass, on the species from each study from mammalbase (Lintulaakso 2021), animaltraits database

(Herberstein et al. 2022), handbook of the birds of the world (Del Hoyo et al. 1992) and fishbase (Froese and Pauly 2010).

We divided the studies into categories based on their study designs as experimental or observational. We converted proportional and percentage data (means and measures of uncertainty) to absolute values (number of individuals or time in seconds), and we converted all measures of uncertainty to variance before analysis. We calculated standardised mean difference for studies that reported an explicit treatment and control (negative or positive), as

$$d = \frac{x_{\text{control}} - x_{\text{treatment}}}{s_{\text{pooled}}}, \text{ where } x_i \text{ is the mean of treatment and control outcomes respectively and}$$

$$s_{\text{pooled}} = \sqrt{\frac{s_{\text{control}}^2(n_{\text{control}} - 1) + s_{\text{treatment}}^2(n_{\text{treatment}} - 1)}{n_{\text{treatment}} + n_{\text{control}} - 2}}, \text{ here } s_i^2 \text{ is the sample variance and } n_i \text{ is the}$$

sample size. All the correlational studies we included reported unstandardized regression

coefficients (b), we thus calculated the standardised mean differences from b as $d = \frac{b}{s_{\text{pooled}}}$, where

$$s_{\text{pooled}} = \sqrt{\frac{s_y^2(N - 1) - b^2\left(\frac{n_{\text{treatment}}n_{\text{control}}}{n_{\text{treatment}} + n_{\text{control}}}\right)}{N - 2}} \text{ here } N \text{ is the total sample size (Lipsey and Wilson}$$

2001). The model coefficients of studies which used a treatment:control contrast were multiplied by -1 to maintain consistency across studies. We assigned each study a unique identifier (study ID); when we extracted multiple data points from studies, we assigned a unique identifier to each datapoint (data ID).

We first fit an intercept only multilevel meta-analytic (MLMA) model with species, study and data id as random effects to determine the degree of heterogeneity across studies after accounting for known sources of variation as follows:

$$y_{ijk} = \mu + u_{ijk} + s_i + d_j + \eta_{ij}$$

where, y_{ij} is the effect from j^{th} data point of the i^{th} study, μ is the overall mean across studies, u_{ij} of the i^{th} estimate withing the j^{th} study, s_i is the random effect due to the i^{th} study, d_j is the random effect due to the j^{th} data point and η_{ij} is the error of the j^{th} estimate from the i^{th} study. The random effect framework assumes that each study has its own true effect size which is derived from a population of true effect sizes. Since this meta-analysis involves a comparison across taxa in different habitats and geographies, the random effect framework is appropriate (Nakagawa et al. 2022; Noble et al. 2022). We also fit a phylogenetic model with relatedness as random effect.

$$y_{ijk} = \mu + u_{ij} + s_i + d_j + p_k + \eta_{ij}$$

where, p_k is random effect due to the k^{th} species. Phylogeny did not capture significant variation across all outcomes (Chamberlain et al. 2012; Cinar, et al. 2022). We thus used a model with only study identity, and data identity as random effects. We used the Q statistic to determine left over variation. We used both I^2 and H^2 to determine heterogeneity across studies (Harrer et al. 2021). We used the final MLMA to determine whether there was an overall significant effect across studies. Finally, we used the MLMA to perform a multiple regression to determine the effect of type of human interactions and animal size on the observed effects.

7. Sensitivity analysis

Our analysis draws from a global cross taxonomic dataset and, thus, there are multiple possible sources of non-independence. Bias in terms of the taxa studies and geographic location of study meant that certain populations were represented multiple times in our data set. Phylogenetic relatedness may also lead to non-independence across studies (Noble et al. 2017). We account for potential non-independence by implementing the MLMA with a random effect. We evaluated publication bias visually in our dataset using a funnel plot and used regression test for funnel plot asymmetry (Nakagawa et al. 2022, Appendix C). We also used a multilevel meta-regression on sample variance with study identity and data identity as random effect to determine publication bias. As significant results have a tendency to be published first, we plotted the effect size against the year of publication to determine any time-lag effects. To determine the robustness of our analysis, we used a leave-one-out method where we dropped each study from our data set in sequence and ran the MLMA to determine if any single study had a disproportionate effect on the summary effect size (Nakagawa et al. 2022; Noble et al. 2017) (Appendix D). All data preparation and analysis were conducted in R – 4.0.1 (R Core Team 2024). Data and code are available at <https://www.github.com/cheesesnakes/superpredator-or-superscary>.

III. Results

After screening, we included 85 studies in our systematic review, the majority of which were published in the past two decades (Appendix A5). Of these, 44 studies reported enough information to be included in the meta-analysis, covering 38 species and families. These studies were widely distributed across geographies and habitats ranging from 78°N to 43°S and 169°E to 123°W (Figure 2). Across outcomes, 24 studies reported measures of foraging behaviours, which included bite rate and time spent foraging, 29 studies focused on vigilance behaviours, which

included time spent vigilant and vigilance rate, and 9 studies dealt with movement behaviour, which involved home range size, displacement and rate of movement (Figure 3). The majority of these papers (n = 36) studied primary consumers, such as *Cervus elaphus* and *Capreolus capreolus* (Figure 4, Appendix A.6). Only 7 studies included secondary and tertiary consumers (Figure 4), thus statistical comparisons across trophic levels and functional guilds were omitted.

The body size of the animal did not significantly influence the effect of human activities on foraging ($\beta = -4.139 \times 10^{-6} \pm 4.352 \times 10^{-5}$, $t_{34} = -0.09$, $p = 0.924$), movement ($\beta = 0.006 \pm 0.006$, $T_5 = 0.948$, $p = 0.288$), or vigilance behaviour ($\beta = -4.139 \times 10^{-6} \pm 4.347 \times 10^{-5}$, $t_{46} = -0.0517$, $p = 0.961$). Overall, animals responded similarly to human disturbance regardless of their size.

1. Changes in behaviours across types of human interactions

The overall effect of human interactions across studies on foraging behaviours was significantly negative (SMD = -0.411, 95% CI = -0.893 – 0.071, $T_{22} = -1.73$, $p = 0.092$) and on vigilance was significantly positive (SMD = 1.31, 95% CI = 0.376 – 2.237, $T_{27} = 2.83$, $p = 0.007$). However, across studies, the effect of human interactions on movement behaviour across studies was not significant (SMD = 0.012, 95% CI = -0.694 – 0.718, $T_6 = -0.04$, $p = 0.971$, Appendix C.2 – C.4, Figure 4). Upon fitting the MLMA model, there was still significant heterogeneity among studies that was not accounted for by random effects (Appendix C).

Foraging behaviour of animals was negatively affected by lethal human interactions (SMD = -1.65, 95% CI = -3.56 - 0.27), but not active non-lethal interactions (SMD = 0.071, 95% CI = -

0.38 – 0.53) or passive non-lethal interactions (SMD = -0.688, 95% CI = -1.83 – 0.46). Five data points showed a significant negative effect of active interactions on foraging behaviour and 6 data points showed a significant positive effect. Similarly, 5 data points showed a significant negative effect of passive non-lethal interactions on foraging and 2 data points showed a significant positive effect of passive interactions on foraging.

Across studies, lethal interactions had a significant positive effect on vigilance behaviours of study animals (SMD = 3.192, 95% CI = 0.85 – 5.53) but active non-lethal interactions did not have a significant effect (SMD = 0.57, 95% CI = -0.46 – 1.61) and nor did passive non-lethal interactions (SMD = 0.44, 95% CI = -1.45 – 2.34, Figure 4). There was a significant positive effect of active interactions on vigilance in 13 data points and significant positive effect of passive interactions in 5 data points (Figure 4). There was no significant effect of human interactions on movement.

Overall, when we consider all studies in this analysis, foraging behaviours were significantly more suppressed by lethal human activities than by active non-lethal interactions ($\beta = -1.141 \pm 0.529$, $t_{34} = -2.155$, $p = 0.038$, figure 5). The magnitude of behavioural effects on vigilance behaviour due to lethal human interactions was significantly higher than that of active non-lethal interactions ($\beta = 2.359 \pm 0.983$, $t_{46} = 2.398$, $p = 0.02$, figure 5). There was no significant difference in the magnitude of interactions between active and passive non-lethal human interactions. There were no significant differences in the effect of human interactions on movement behaviours across the three types of interactions (Appendix D, figure 5).

IV. Discussion

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350 Humans occupy multiple niches across almost every ecosystem. As predators, humans are more
 351 deadly than most other predators in both marine and terrestrial systems because they can target
 352 multiple trophic levels simultaneously and can extract animals at a faster rate than natural
 353 predators (Darimont et al. 2015). Humans also occupy natural spaces for seemingly benign
 354 activities such as resource extractions, tourism, hiking, and so on. Thus, humans can be predators
 355 in the classic sense, but their non-lethal activities, such as their presence in the environment or
 356 modification of the environment, can also be perceived as threatening. The mere presence of
 357 humans has been shown to alter animal behaviour, a phenomenon considered under the risk
 358 disturbance hypothesis (Frid and Dill, 2002). Understanding how animals perceive and respond
 359 to humans in their environments is vital for successful management of animal populations as well
 360 as the management of ecosystems they inhabit. The review and meta-analysis presented here
 361 covers a wide range of studies both geographically and taxonomically over the past three decades
 362 (Figure 2, Appendix A). We find that the magnitude of effects on movement, foraging and
 363 vigilance behaviours of wild animals depends significantly on the type of human interaction with
 364 lethal human-wildlife interactions creating the greatest and only overall significant effects on
 365 behaviours compared to both active and passive non-lethal interactions (Figure 5).

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367 *1. Lethal interactions: hunting, fishing, etc.*

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369 Humans have hunted other for a majority of their evolutionary history (Treves and Naughton-
 370 Treves 1999). In this context, hunting includes fishing, and trapping that result in removal of prey
 371 individuals through direct human intervention. Darimont et al.'s (2015) analysis revealed that
 372 humans have killed a larger biomass of prey species than most other predators in their respective

ecosystems. The efficiency and scale at which humans can kill prey adds to their perceived fierceness, a predator trait (Wirsing et al. 2021). Thus, prey should respond to human hunters very strongly (Clinchy et al. 2016; Crawford et al. 2022). Clinchy et al. (2016) showed that in areas where badgers are hunted, badgers respond to playbacks of human voices by significantly reducing their foraging activity. Our meta-analysis showed that, as expected, lethal interactions had a significant and adverse effect on both foraging and vigilance behavior of animals across taxa and geographical locations. In line with the foraging – vigilance trade off (Brown and Kotler 2004), our metanalysis shows that animals significantly reduced foraging and significantly increased vigilance in response to human hunters. We find an overall larger magnitude of effect on vigilance rather than foraging (Figure 5), but this is most likely because fewer studies have reported the effects of lethal interactions on foraging behavior (Figure 3).

The behavioral response to lethal interactions was largest for large game species such as deer (*Dama dama*), elk (*Cervus elaphus*) and moose (*Alces alces*), whereas the response of terrestrial predators, such as puma (*Puma concolor*), hyena (*Crocuta crocuta*) and badgers (*Meles meles*), was limited (Suraci et al. 2019; Clinchy et al. 2016; Pangle and Holekamp 2010). Across prey species, the type of response was varied. Pecorella et al. (2016) showed that fallow deer (*Dama dama*) were more vigilant and foraged less in areas where culling was allowed than where culling was not allowed in Maremma Regional Park (Italy). Large ungulates such as Sable antelope (*Hippotragus niger*) and greater Kudu (*Tragelaphus stepsiceros*) also showed large increases in vigilance in the areas around Hwange National Park (Crosmarby et al. 2012). Fallow deer showed a proportional decrease in foraging as vigilance increased (Pecorella et al. 2016), however, this trade-off is not always proportional. Red deer (*Cervus elaphus*) in the Scottish Highlands showed a disproportionate increase in vigilance behaviors relative to the decrease in foraging behavior in

response to lethal interactions (Jayakody et al. 2008). Conversely, Moose (*Alces alces*) showed a larger decrease in foraging behavior than increase in vigilance (Bhardwaj et al. 2022). Notably, roe deer (*Capreolus capreolus*) showed large changes in movement behavior as opposed to foraging and vigilance (Benhaïem et al. 2008; Picardi et al. 2019).

2. Active non-lethal disturbance

According to the risk disturbance hypothesis, animals are expected to perceive non-lethal human disturbance similarly to the threat of predation (Frid and Dill 2002). For instance, bird species have been found to reduce foraging and increase vigilance in response to pedestrians in urban parks of Madrid (Fernández-Juricic et al. 2001). These behavioural changes are akin to how birds would respond to a natural predator, suggesting that human presence is perceived as a threat. Thus, animals may respond to human disturbances in the same way that they do to mitigate predation risk from natural predators. Multiple studies have sought to test the risk disturbance hypothesis across contexts. Price (2008), in a review, showed that human disturbance in the form of pedestrians alters the flight, foraging and vigilance behaviours of multiple bird species. Stansell et al. (2022) showed that dark-eyed juncos (*Junco hyemalis*) are more shy in urban areas than in their wild habitat.

However, the relative magnitude of perceived risk from lethal humans, and non-lethal humans may greatly differ (Mols et al. 2022). Our analysis shows variation across studies in terms of the magnitude and direction of observed effects for both types of non-lethal interaction with humans, but the effects of active interactions were small for most studies. In addition, lethal interactions with humans elicited the greatest magnitude of effects across outcomes. We hypothesize that this

could be attributed to the divergent ways in which animals perceive benign humans in their environment (Figure 1, Smith et al. 2021).

In some cases, the presence of humans may benefit some species by altering interactions between them; female grizzly bears with cubs have been observed using areas with recreational human activity to avoid depredation by aggressive males (Ladle et al. 2019). In particular, they may influence the interactions between predators and prey (Moller 2008; Shannon et al. 2014; Proudman et al. 2021). For example, humans may deter predators by their mere presence and create safe havens for prey who are ambivalent to human presence to forage. This is known as the human shield hypothesis (sensu Berger 2007). However, lack of data across trophic levels makes it difficult to test this hypothesis (Reimers, Lund, and Ergon 2011; Ward et al. 1997; Uchida and Blumstein 2021).

In addition, a large number of studies in our analysis showed no significant effect of active non-lethal interactions on foraging ($n = 10$) and vigilance ($n = 9$) behavior, i.e., the effect sizes were close to 0. We hypothesise that this subdued response to humans may be attributed to habituation as a function of frequency of interaction. Many studies have pointed to substantial habituation to tourist interactions, such as with mammals in the African Savannah (Knight 2009), killer whale and Hector's dolphin interactions with whale watchers (Bejder et al. 1999, Williams et al. 2002) and penguin interactions with researchers (Shelton et al. 2004). More specifically, Uchida and Blumstein (2021) showed that yellow bellied marmots (*Marmota flaviventer*) in highly disturbed forests grew habituated to humans after repeated exposure and reduced their flight initiation distances; however, marmots in groups that were habituated gained less body mass over the duration of the study. On the other hand, mammal population densities are on the whole higher

near human settlements, indicating that the benefits of associated resources may outweigh the negative effects of exposure (Tucker et al, 2021). Overall, there is a need to examine both the immediate effects of non-lethal human disturbance as well as the long-term positive or negative fitness consequences of such exposure.

3. *Passive non-lethal interactions*

Humans are the most elaborate ecosystem engineer on the planet, completely altering environments to make it more suitable for themselves. Predation risk is highly dependent on whether animals perceive these environments as safe or threatening (Wirsing et al. 2021). Van der Kolk et al. (2024) found that various shore birds altered their flight responses based on the frequency of airplanes over their nesting areas, with intermittent disturbance having the greatest negative effect. Similarly, multiple studies have highlighted the negative effects of ship sonar on cetacean behavior (see Harris et al. 2018 for a meta-analysis and review).

By altering habitats, human development alters the distribution of risk in the landscape. For example, roads can clear away thickets or other cover that ambush predators use thus making it "safer" for some prey species. Thus, one would expect prey species to congregate around roads and invest more in foraging. We see evidence for this in guancos (*Lama guanicoe*) and mice (*Peromyscus leucopus*) (Cappa et al. 2017; Grignolio et al. 2011). Mehlhoop et al. (2002) showed that moose (*Rangifer tarandus*) in Norwegian forests avoid human infrastructure in areas where hunting is permitted. On the other hand, Cappa, Giannoni, and Borghi (2017) showed that guancos (*Lama guanicoe*) in large groups were less vigilant and foraged closer to roads in an

Argentine wildlife reserve and thus may perceive roads as beneficial. Widespread habitat alteration has also led to temporal shifts in activity of multiple taxa globally (Gaynor et al. 2018). Conflict with predator species can also make human settlements safe spaces for consumers to forage (eg. wild boar, Podgorski et al. 2013). Conversely, development may also open up habitats making it more difficult for prey individuals to escape predation (Smith, et al. 2016). Human development may also pose a direct threat to animals, especially in the case of conflict with livestock or roadkill (Loss et al. 2015, van der Kolk et al. 2024, Harris et al. 2018). Across mammal species, Tucker et al (2021) find an increase in population densities near human settlements, and speculate that more resources and lower competition/predation may lead to this pattern. However, they do caution that while some species may get benefits, many others be negatively affected and decline. Similarly, our meta-analysis found considerable variation across studies on the effect of passive interactions on foraging, vigilance and movement behaviors (Figure 4). Thus, the response of animals to passive interactions with roads and human settlements may be highly context and species dependent.

4. Context dependence of the behavioural effects human interactions on animals

Animals can make mistakes in how they respond behaviourally to human interactions (Smith et al. 2021), by overreacting or underreacting. Animals can overreact to non-lethal interactions leading to high energy investment and even chronic stress responses (e.g. (Blumstein 2003; Nunes et al. 2019; Westekemper et al. 2018; review: Larson et al. 2016). On the other hand, misunderstanding lethal interactions can lead to the death of the individual. Thus, there are a number of contingencies that may apply to the expression of risk avoidance and mitigation

behaviours. For example, the openness of the habitat, predator traits such as habitat domain, hunting mode and fierceness, and prey body state are some important variables to consider (Schmitz et al. 2001, Wirsing et al. 2021). While all of these also apply to the human predator, there may be some additional considerations for a predictive framework (Smith et al. 2024).

i. Trophic level

Human hunters can augment existing predation risk from natural predators, create new predation risk or both depending on which animals they target (Supplementary Table B2). Hunters tend to disproportionately target predatory and larger species for their higher social and economic value. For example, marine apex predators such as sharks, barracuda and tuna are highly valuable in fish markets compared to consumer species such as sardines (Pauly et al. 1998; Jackson et al. 2001). Similarly, in many developed countries large herbivores such as deer are hunted for sport or culling (Pecorella et al. 2016). Thus, interactions with humans may be perceived differently across trophic levels.

When humans target top predators, they create new risk that these animals have not previously experienced. Predators may or may not be equipped to respond to new threats leading to many extirpations (e.g. Wolves in Yellowstone (Laundré, Hernández, and Altendorf et al. 2001), bears in Alaska (Merkle et al. 2013), puma in California and Florida (Nickel et al. 2021)). Thus, predators must respond as naive prey when targeted by humans. The severity and direction of the response may be proportional to the extent and duration of persecution by humans. Predators may be overly cautious around human hunters leading to greater missed opportunities. Alternatively, predators may not be cautious enough around humans leading to greater direct mortality (Smith

et al. 2021). The direct removal of predator individuals and their behavioural responses to the risk of the human super-predator may have beneficial effects for consumers effectively subtracting predation risk for these animals (e.g. mesopredator release and population explosions of large herbivores, see Terbourg et al. 2015).

The majority of studies we found in our search focused on lower trophic levels and not predators (Appendix A). This may be because it is particularly difficult to observe foraging and vigilance behaviours in predators (but see Smith et al. (2017) for examples of human disturbance effects on *Puma concolor* foraging behaviour). We were unable to test our hypotheses regarding predator behavioural responses to humans in this meta – analysis, however, we highlight this as an important topic for future research. Incorporating other behaviours such as movement and habitat selection may allow us to gain further insights into how predators themselves respond to the threat of predation.

Humans may also compete with top predators by targeting consumers (eg. deer hunters in California). This may have a multiplicative effect on predation risk perceived by consumer individuals. Consumers may have fewer safe habitats to forage and thus may experience a higher energy cost of anti-predator strategies. There may also be mutualisms between predators and humans creating areas of higher predation success (corralling) or even subsidizing predator energy costs (cattle depredation). In a majority of studies, humans were not the only predator present. For example, humans compete with puma in the Santa Cruz Mountains of California, with wolves in parts of north America, and with bears in the arctic. Thus, hypotheses about multiplicative effects of human predation risk and competition with top predators may be testable in the field.

541

542 In some cases, humans may target multiple trophic levels simultaneously. Trawl fishing, for
543 example, is particularly non-selective. It is difficult to determine the effects of such ubiquitous
544 predation risk. Suraci et al. (2019) tested effects of lethal human interactions across multiple
545 trophic levels in the Santa Cruz Mountains and found evidence for a behavioural cascade. To our
546 knowledge, no other study has simultaneously tested behaviours across multiple trophic levels.

547

548 ii. Body mass

549

550 As with predators, larger animals may also be disproportionately targeted for economic reasons
551 and as trophies. Larger and healthy animals are rarely targeted by predators and thus have little
552 reason to invest in anti-predator behaviour. Thus, we expected the magnitude of effects to
553 increase with the body mass of the study animal. This, however, was not the case in our analysis
554 as the change in effect size was negligible across outcomes.

555

556 iii. History

557

558 The disproportionate targeting of apex predators has led to the local extirpation of many species
559 in the global north. Though conservation efforts have led to the creation of no-take and no-
560 hunting zones in the past few decades, a history of hunting may still affect how animals perceive
561 humans in their environment (Clinchy et al. 2016). For example, playback experiments of
562 extirpated predators indicates that fear may persist across generations (Clinchy et al. 2016).
563 However, the degree of response by animals may depend on a number of factors, starting from
564 their own responses (whether innate or learned), nature of hunting, duration of hunting cessation,

which species were historically targeted and the nature of current human interactions. Repeated exposure and learned tolerance may also play a role in response attenuation (Uchida and Blumstein, 2021). Thus, it may be worthwhile to investigate whether the animal response over time to the cessation of human activity, both lethal and non-lethal.

5. Using fear of humans for management

The fear of human hunters can also be valuable tool for conservation and management. Miller and Schmitz (2019) proposed a framework for reducing the conflict between large carnivores (*Panthera tigris*) and cattle herders in central India. They showed that by utilizing tigers' intrinsic fear of humans in the central Indian landscape and common human signs such as traps, fences and scents, conflict can be significantly reduced. Blackwell et al. (2016) proposed using fear induced by periodic playbacks of human noises to deter animals from roads and reduce the toll of road kill. Gaynor et al. (2021) proposed an applied ecology of fear framework for using non-consumptive behavioural effects of the human super predator for conservation and management. If indeed non-lethal cues work best when reinforced by actual lethal actions, a combination of human cues and periodic and regulated hunting may offer a viable model for the management of wild populations and the reduction of conflict.

V. Conclusions

The effects of natural animal predators on prey have been studied and debated for many decades (Lotka 1920; Sih 1984; Dill, Heithaus, and Walters 2003; Steneck 2012; S.L. Lima et al. 2021). It

is clear that through their consumption of prey, natural predators create a dangerous environment to which prey respond (Lima and Dill 1990; Laundré 2010). Thus, a wide range of anti-predator responses have evolved and are induced in animals to minimize the probability of being killed (Lima and Dill 1990). Recent evidence suggests that the fear of predators itself even without direct threat of being killed changes prey responses (i.e. non-consumptive effects of predators), with potential consequences to fitness (Schmitz, Beckerman, and O'Brien 1997; Heithaus et al. 2008; Hawlena and Schmitz 2010; Clinchy, Sheriff, and Zanette 2013; Heithaus et al. 2008; Sheriff et al. 2020; Michael Clinchy, Sheriff, and Zanette 2013).

We have presented an extensive synthesis of the available literature on the effect of human interactions on animal behaviour. Our general conclusions and suggestions for future research are listed below:

1. While lethal interactions with humans have a profound effect on animal behaviour, there is mixed evidence of such effects from non-lethal interactions.
2. There is considerable variation in the response of animals to non-lethal interactions that has yet to be disentangled. Factors such as frequency of interactions, history of hunting, and trophic level may attenuate the response of animals to human disturbance.
3. The effects of human interactions on animal fitness and movement would be an important consideration for the management of wild populations.
4. Future research may benefit from contrasting humans as lethal predators with natural predators in the same space and across trophic levels, given the potential for compounding effects on animals and ecosystems.

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References

- Atwood, TB., Madin, EMP, Harborne, AR, Hammill, E., Luiz, OJ., Ollivier, QR., Roelfsema, CM., Macreadie, PI., and Lovelock, CE.. 2018. “Predators Shape Sedimentary Organic Carbon Storage in a Coral Reef Ecosystem.” *Frontiers in Ecology and Evolution* 6 (August):110. <https://doi.org/10.3389/fevo.2018.00110>.
- Bejder, L., S. M. Dawson, and J. A. Harraway. 1999. “Responses by Hector’s Dolphins to Boats and Swimmers in Porpoise Bay, New Zealand.” *Marine Mammal Science* 15 (3): 738–50. <https://doi.org/10.1111/j.1748-7692.1999.tb00840.x>.
- Benhaiem, S., Delon, M. , Lourtet, B., Cargnelutti, B., Aulagnier, S., Hewison, A.J.M., Morellet, N., and Verheyden, H.. 2008. “Hunting Increases Vigilance Levels in Roe Deer and Modifies Feeding Site Selection.” *Animal Behaviour* 76 (3): 611–18. <https://doi.org/10.1016/j.anbehav.2008.03.012>.
- Benson, JF, Mahoney, PJ, and Patterson, BR. 2015. “Spatiotemporal Variation in Selection of Roads Influences Mortality Risk for Canids in an Unprotected Landscape.” *OIKOS* 124 (12): 1664–73. <https://doi.org/10.1111/oik.01883>.
- Berger, J. 2007. “Fear, Human Shields and the Redistribution of Prey and Predators in Protected Areas.” *BIOLOGY LETTERS* 3 (6): 620–23. <https://doi.org/10.1098/rsbl.2007.0415>.
- Bhardwaj, M., Lodnert, D., Olsson, M., Winsvold, A., Eilertsen, S.M., Kjellander, P., and Seiler, A.. 2022. “Inducing Fear Using Acoustic Stimuli—A Behavioral Experiment on Moose (Alces Alces) in Sweden.” *Ecology and Evolution* 12 (11). <https://doi.org/10.1002/ece3.9492>.
- Blackwell, BF, DeVault, TL , Fernandez-Juricic, E, Gese, EM, Gilbert-Norton, L, and Breck, SW. 2016. “No Single Solution: Application of Behavioural Principles in Mitigating Human-Wildlife Conflict.” *Animal Behaviour* 120:245–54.

<https://doi.org/10.1016/j.anbehav.2016.07.013>.

Blumstein, DT. 2003. “Flight-Initiation Distance in Birds Is Dependent on Intruder Starting Distance.” *Journal of Wildlife Management* 67 (4): 852–57.

<https://doi.org/10.2307/3802692>.

Boyle, S.A., and Samson, F.B. 1985. “Effects of Nonconsumptive Recreation on Wildlife: A Review.” *Wildlife Society Bulletin (1973-2006)* 13 (2): 110–16.

Brown, CL, Hardy, A.R., Barber, J.R., Fristrup, K.M., Crooks, K.R., and Angeloni, L.M. 2012. “The Effect of Human Activities and Their Associated Noise on Ungulate Behavior.” *PLOS One* 7 (7). <https://doi.org/10.1371/journal.pone.0040505>.

Brown, J.S., and Kotler, B.P. 2004. “Hazardous Duty Pay and the Foraging Cost of Predation: Foraging Cost of Predation.” *Ecology Letters* 7 (10): 999–1014.

<https://doi.org/10.1111/j.1461-0248.2004.00661.x>.

Cappa, FM, SM Giannoni, and CE Borghi. 2017. “Effects of Roads on the Behaviour of the Largest South American Artiodactyl (Lama Guanicoe) in an Argentine Reserve.” *ANIMAL BEHAVIOUR* 131:131–36. <https://doi.org/10.1016/j.anbehav.2017.07.020>.

Chamberlain, A.A., Hovick, S.M., Dibble, C.J., Rasmussen, N.L., Van Allen, B.G., Maitner, B.S., Ahern, J.R., et al. 2012. “Does Phylogeny Matter? Assessing the Impact of Phylogenetic Information in Ecological Meta-analysis.” *Ecology Letters* 15 (6): 627–36. <https://doi.org/10.1111/j.1461-0248.2012.01776.x>.

Cinar, O., Nakagawa, S., and Viechtbauer, W. 2022. “Phylogenetic Multilevel Meta-analysis: A Simulation Study on the Importance of Modelling the Phylogeny.” *Methods in Ecology and Evolution* 13 (2): 383–95. <https://doi.org/10.1111/2041-210X.13760>.

Clinchy, M., Sheriff, M.J., and Zanette, L.Y. 2013. “Predator-Induced Stress and the Ecology of Fear.” *Functional Ecology* 27 (1): 56–65. <https://doi.org/10.1111/1365-2435.12007>.

- 664 Clinchy, M., Zanette, L.Y., Roberts, D., Suraci, J.P., Buesching, C.D., Newman, C., and
665 Macdonald, D.W. 2016. “Fear of the Human ‘Super Predator’ Far Exceeds the Fear of
666 Large Carnivores in a Model Mesocarnivore.” *Behavioral Ecology* 27 (6): 1826–32.
- 667 Crawford, D.A., Conner, L.M., Clinchy, M., Zanette, L.Y., and Cherry, M.J. 2022. “Prey Tells,
668 Large Herbivores Fear the Human ‘Super Predator.’” *Oecologia* 198 (1): 91–98.
669 <https://doi.org/10.1007/s00442-021-05080-w>.
- 670 Crosmar, W.G., Makumbe, P., Cote, S.D., and Fritz, H. 2012. “Vulnerability to Predation and
671 Water Constraints Limit Behavioural Adjustments of Ungulates in Response to Hunting
672 Risk.” *Animal Behaviour* 83 (6): 1367–76. <https://doi.org/10.1016/j.anbehav.2012.03.004>.
- 673 Darimont, C.T., Carlson, S.M., Kinnison, M.T., Paquet, P.C., Reimchen, T.E., and Wilmer, C.C.
674 2009. “Human Predators Outpace Other Agents of Trait Change in the Wild.” *Proceedings*
675 *of the National Academy of Sciences* 106 (3): 952–54.
676 <https://doi.org/10.1073/pnas.0809235106>.
- 677 Darimont, C.T., Fox, C.H., Bryan, H.M., and Reimchen, T.E. 2015. “The Unique Ecology of
678 Human Predators.” *Science* 349 (6250): 858–60. <https://doi.org/10.1126/science.aac4249>.
- 679 Darimont, C.T., Cooke, R., Bourbonnais, M.L., Bryan, H.M., Carlson, S.M., Estes, J.A., Galetti,
680 M., et al. 2023. “Humanity’s Diverse Predatory Niche and Its Ecological Consequences.”
681 *Communications Biology* 6 (1): 609. <https://doi.org/10.1038/s42003-023-04940-w>.
- 682 Davis, D., Banks, S., Birtles, A., Valentine, P., and Cuthill, M. 1997. “Whale Sharks in Ningaloo
683 Marine Park: Managing Tourism in an Australian Marine Protected Area.” *Tourism*
684 *Management* 18 (5): 259–71. [https://doi.org/10.1016/S0261-5177\(97\)00015-0](https://doi.org/10.1016/S0261-5177(97)00015-0).
- 685 Del Hoyo, J., Elliott, A., and Sargatal, J. 1992. *Handbook of the Birds of the World*. Vol. 1. Lynx

686 Edicions: Barcelona.

687 http://www.aviornis.nl/uploads/media/Parelhoenders_in_Handbook_of_the_Birds_of_the_World

688 [_01.pdf](#).

689 Dill, L.M., Heithaus, M.R., and Walters, C.J. 2003. “Behaviorally Mediated Indirect Interactions

690 in Marine Communities and Their Conservation Implications.” *Ecology* 84 (5): 1151–57.

691 [https://doi.org/10.1890/0012-9658\(2003\)084\[1151:BMIIM\]2.0.CO;2](https://doi.org/10.1890/0012-9658(2003)084[1151:BMIIM]2.0.CO;2).

692 Estes, J.A., Brashares, J.S., and Power, M.E. 2013. “Predicting and Detecting Reciprocity

693 between Indirect Ecological Interactions and Evolution.” *The American Naturalist* 181

694 (Suppl 1): S76–99. <https://doi.org/10.1086/668120>.

695 Fernández-Juricic, E., Clavijo, M., Jiménez, M.D., Asensio, E., and Lucas, E. 2001. “Bird

696 Tolerance to Human Disturbance in Urban Parks of Madrid (Spain): Management

697 Implications.” *In Ecology of Urban Environments*.

698 https://doi.org/10.1007/978-1-4615-1531-9_12.

699 Foo, Y.Z., O’Dea, R.E., Koricheva, J., Nakagawa, S., and Lagisz, M. 2021. “A Practical Guide to

700 Question Formation, Systematic Searching and Study Screening for Literature Reviews in

701 Ecology and Evolution.” *Methods in Ecology and Evolution* 12 (9): 1705–20.

702 <https://doi.org/10.1111/2041-210X.13654>.

703 Frid, A., and Dill, L.M. 2002. “Human-Caused Disturbance Stimuli as a Form of Predation Risk.”

704 *Conservation Ecology* 6 (1). <https://doi.org/10.5751/ES-00404-060111>.

705 Gaynor, K.M., Cherry, M.J., Gilbert, S.L., Kohl, M.T., Larson, C.L., Newsome, T.M., Prugh,

706 T.R., Suraci, J.P., Young, J.K., and Smith, J.A.. 2021. “An Applied Ecology of Fear

707 Framework: Linking Theory to Conservation Practice.” *Animal Conservation* 24 (3): 308–

- 708 21. <https://doi.org/10.1111/acv.12629>.
- 709 Harrer, M., Cuijpers P., Furukawa, T.A., and Ebert, D.D.. 2021. *Doing Meta-Analysis With R: A*
710 *Hands-On Guide*. 1st ed. Boca Raton, FL and London: Chapman & Hall/CRC Press.
711 [https://www.routledge.com/Doing-Meta-Analysis-with-R-A-Hands-On-Guide/Harrer-](https://www.routledge.com/Doing-Meta-Analysis-with-R-A-Hands-On-Guide/Harrer-Cuijpers-Furukawa-Ebert/p/book/9780367610074)
712 [Cuijpers-Furukawa-Ebert/p/book/9780367610074](https://www.routledge.com/Doing-Meta-Analysis-with-R-A-Hands-On-Guide/Harrer-Cuijpers-Furukawa-Ebert/p/book/9780367610074).
- 713 Harris, C.M., Thomas, L., Falcone, E.A., Hildebrand, J., Houser, D., Kvadsheim, P.M., Lam,
714 F.P.A., et al. 2018. “Marine Mammals and Sonar: Dose-Response Studies, the Risk-
715 Disturbance Hypothesis and the Role of Exposure Context.” *JOURNAL OF APPLIED*
716 *ECOLOGY* 55 (1): 396–404. <https://doi.org/10.1111/1365-2664.12955>.
- 717 Hawlena, D., and Schmitz, O.J.. 2010. “Herbivore Physiological Response to Predation Risk and
718 Implications for Ecosystem Nutrient Dynamics.” *Proceedings of the National Academy*
719 *of Sciences* 107 (35): 15503–7. <https://doi.org/10.1073/pnas.1009300107>.
- 720 Heithaus, M.R., Wirsing, A.J., Thomson, J.A., and Burkholder, D.A.. 2008. “A Review of Lethal
721 and Non-Lethal Effects of Predators on Adult Marine Turtles.” *Journal of Experimental*
722 *Marine Biology and Ecology* 356 (1–2): 43–51.
723 <https://doi.org/10.1016/j.jembe.2007.12.013>.
- 724 Herberstein, M.E., McLean, E.L., Wolff J.O., Khan M.K., Smith K., Allen A.P., et al. 2022.
725 “AnimalTraits - a Curated Animal Trait Database for Body Mass, Metabolic Rate
726 and Brain Size.” *Scientific Data* 9 (1): 265. <https://doi.org/10.1038/s41597-022-01364-9>.
- 727 Jeremy, J.B.C., Kirby, M., Berger W.H., Bjorndal, K.A., Botsford, L.W., Bourque, B.J., Bradbury,
728 R.H., et al. 2001. “Historical Overfishing and the Recent Collapse of Coastal
729 Ecosystems.” *Science* 293 (5530): 629–37. <https://doi.org/10.1126/science.1059199>.

- 730 Jayakody, S., Sibbald, A.M., Gordon, I.J., and Lambin, X.. 2008. “Red Deer *Cervus Elephus*
731 Vigilance Behaviour Differs with Habitat and Type of Human Disturbance.” *WILDLIFE*
732 *BIOLOGY* 14 (1): 81–91. [https://doi.org/10.2981/0909-](https://doi.org/10.2981/0909-6396(2008)14[81:RDCEVB]2.0.CO;2)
733 6396(2008)14[81:RDCEVB]2.0.CO;2.
- 734 Knight, J. 2009. “Making Wildlife Viewable: Habituation and Attraction.” *Society & Animals*
735 17 (2): 167–84. <https://doi.org/10.1163/156853009X418091>.
- 736 Ladle, A., Avgar, T., Wheatley, M., Stenhouse, M.B., Nielsen, S.E., and Boyce, M.S.. 2019.
737 “Grizzly Bear Response to Spatio-Temporal Variability in Human Recreational Activity.”
738 *JOURNAL OF APPLIED ECOLOGY* 56 (2): 375–86.
739 <https://doi.org/10.1111/13652664.13277>.
- 740 Courtney L., Reed, S.E., Merenlender, A.M., and Crooks, K.R.. 2016. “Effects of Recreation on
741 Animals Revealed as Widespread through a Global Systematic Review.” *PLOS ONE* 11
742 (12): e0167259. <https://doi.org/10.1371/journal.pone.0167259>.
- 743 Laundré, J.W. 2010. “Behavioral Response Races, Predator–Prey Shell Games, Ecology of Fear,
744 and Patch Use of Pumas and Their Ungulate Prey.” *Ecology* 91 (10): 2995–3007.
745 <https://doi.org/10.1890/08-2345.1>.
- 746 Laundré, J.W., Hernández, L., and Altendorf, K.B.. 2001. “Wolves, Elk, and Bison:
747 Reestablishing the ‘Landscape of Fear’ in Yellowstone National Park, U.S.A.” *Canadian*
748 *Journal of Zoology* 79 (8): 1401–9. <https://doi.org/10.1139/z01-094>.
- 749 Lima, S.L., Gámez, S., Arringdale, N., and Harris, N.C. 2021. “Vigilance Response of a Key
750 Prey Species to Anthropogenic and Natural Threats in Detroit.” *Frontiers in Ecology and*
751 *Evolution* 9. <https://doi.org/10.3389/fevo.2021.570734>.

- 752 Lima, S.L., and Dill, L.M. 1990. “Behavioral Decisions Made under the Risk of Predation: A
753 Review and Prospectus.” *Canadian Journal of Zoology* 68 (4): 619–40.
754 <https://doi.org/10.1139/z90-092>.
- 755 Lintulaakso, K. 2021. “MammalBase — Database of Traits, Measurements and Diets of the
756 Species in Class Mammalia.” Zenodo. <https://doi.org/10.5281/ZENODO.7462864>.
- 757 Lipsey, M.W., and Wilson, D.B. 2001. *Practical Meta-Analysis*. Applied Social Research
758 Methods Series, v. 49. Thousand Oaks, Calif: Sage Publications.
- 759 Loehr, J., Kovanen, M., Carey, J., Hogmander, H., Jurasz, C., Karkkainen, S., Suhonen, J., and
760 Ylonen, H.. 2005. “Gender- and Age-Class-Specific Reactions to Human Disturbance in a
761 Sexually Dimorphic Ungulate.” *CANADIAN JOURNAL OF ZOOLOGY* 83 (12): 1602–7.
762 <https://doi.org/10.1139/Z05-162>.
- 763 Loss, S.R., Will, T., and Marra, P.P. 2015. “Direct Mortality of Birds from Anthropogenic
764 Causes.” *Annual Review of Ecology, Evolution, and Systematics* 46 (Volume 46, 2015):
765 99–120. <https://doi.org/10.1146/annurev-ecolsys-112414-054133>.
- 766 Lotka, A.J. 1920. “Analytical Note on Certain Rhythmic Relations in Organic Systems.”
767 *Proceedings of the National Academy of Sciences* 6 (7): 410–15.
768 <https://doi.org/10.1073/pnas.6.7.410>.
- 769 Madin, E.M.P., Gaines, S.D., Madin, J.S., and Warner, R.R. 2010. “Fishing Indirectly Structures
770 Macroalgal Assemblages by Altering Herbivore Behavior.” *The American Naturalist* 176
771 (6): 785–801. <https://doi.org/10.1086/657039>.
- 772 Mehlhoop, A.C., Van Moorter, B., Rolandsen, C.M., Hagen,, D. Granhus, A., Eriksen, R.,
773 Ringsby, T.H., and Solberg, E.J. 2022. “Moose in Our Neighborhood: Does Perceived

774 Hunting Risk Have Cascading Effects on Tree Performance in Vicinity of Roads and
775 Houses?” *ECOLOGY AND EVOLUTION* 12 (4). <https://doi.org/10.1002/ece3.8795>.

776 Merkle, J.A., Robinson, H.S., Krausman, P.R., and Alaback, P. 2013. “Food Availability and
777 Foraging near Human Developments by Black Bears.” *JOURNAL OF MAMMALOGY* 94
778 (2): 378–85. <https://doi.org/10.1644/12-MAMM-A-002.1>.

779 Mikolajewicz, N., and Komarova, S.V. 2019. “Meta-Analytic Methodology for Basic Research: A
780 Practical Guide.” *Frontiers in Physiology* 10 (March):203.
781 <https://doi.org/10.3389/fphys.2019.00203>.

782 Miller, JRB, and Schmitz, OJ. 2019. “Landscape of Fear and Human-Predator Coexistence:
783 Applying Spatial Predator-Prey Interaction Theory to Understand and Reduce Carnivore-
784 Livestock Conflict.” *BIOLOGICAL CONSERVATION* 236:464–73.
785 <https://doi.org/10.1016/j.biocon.2019.06.009>.

786 Moller, AP. 2008. “Flight Distance of Urban Birds, Predation, and Selection for Urban Life.”
787 *BEHAVIORAL ECOLOGY AND SOCIOBIOLOGY* 63 (1): 63–75.
788 <https://doi.org/10.1007/s00265-008-0636-y>.

789 Moller, AP. 2012. “Urban Areas as Refuges from Predators and Flight Distance of Prey.”
790 *BEHAVIORAL ECOLOGY* 23 (5): 1030–35. <https://doi.org/10.1093/beheco/ars067>.

791 Moller, AP, Grim T., Ibanez-Alamo JD, Marko, G, and Tryjanowski, P. 2013. “Change in Flight
792 Initiation Distance between Urban and Rural Habitats Following a Cold Winter.”
793 *BEHAVIORAL ECOLOGY* 24 (5): 1211–17. <https://doi.org/10.1093/beheco/art054>.

794 Mols, B, Lambers, E, Cromsigt, JPM, Kuijper, DPJ, and Smit, C. 2022. “Recreation and
795 Hunting Differentially Affect Deer Behaviour and Sapling Performance.” *OIKOS* 2022

796 (1). <https://doi.org/10.1111/oik.08448>.

797 Montero-Quintana, AN, Vazquez-Haikin, JA, Merklings, T, Blanchard, P, and Osorio-Beristain, M.

798 2020. “Ecotourism Impacts on the Behaviour of Whale Sharks: An Experimental

799 Approach.” *ORYX* 54 (2): 270–75. <https://doi.org/10.1017/S0030605318000017>.

800 Nakagawa, S, Lagisz, M, Jennions, MD, Koricheva, J, Noble, DWA, Parker, TH, Sánchez-Tójar,

801 A, Yang, Y, and O’Dea, RE. 2022. “Methods for Testing Publication Bias in Ecological

802 and Evolutionary Meta-Analyses.” *Methods in Ecology and Evolution* 13 (1): 4–21.

803 <https://doi.org/10.1111/2041-210X.13724>.

804 Nevin, OT, and Gilbert, BK. 2005. “Measuring the Cost of Risk Avoidance in Brown Bears:

805 Further Evidence of Positive Impacts of Ecotourism.” *BIOLOGICAL CONSERVATION*

806 123 (4): 453–60. <https://doi.org/10.1016/j.biocon.2005.01.007>.

807 Newey, P. 2007. “Foraging Behaviour of the Common Myna (*Acridotheres Tristis*) in Relation to

808 Vigilance and Group Size.” *EMU* 107 (4): 315–20. <https://doi.org/10.1071/MU06054>.

809 Nickel, BA, Suraci, JP, Nisi, AC, and Wilmers, CC. 2021. “Energetics and Fear of Humans

810 Constrain the Spatial Ecology of Pumas.” *Proceedings of the National Academy of*

811 *Science of the United States of America* 118 (5).

812 Noble, D.W. A., Pottier, P, Lagisz, M, Burke, S, Drobniak, SM, O’Dea, RE, and Nakagawa, S.

813 2022. “Meta-Analytic Approaches and Effect Sizes to Account for ‘Nuisance

814 Heterogeneity’ in Comparative Physiology.” *Journal of Experimental Biology* 225

815 (Suppl_1): jeb243225. <https://doi.org/10.1242/jeb.243225>.

816 Noble, D.W.A., Lagisz, M, O’dea, RE, and Nakagawa, S. 2017. “Nonindependence and

817 Sensitivity Analyses in Ecological and Evolutionary Meta-analyses.” *Molecular Ecology*

818 26 (9): 2410–25. <https://doi.org/10.1111/mec.14031>.

819 Nunes, JACC, Blumstein, DT, Giglio, VJ, Barros, F, and Quimbayo, JP. 2019. “Reef Fish
820 Antipredator Behavior in Remote Islands Does Not Reflect Patterns Seen in Coastal
821 Areas.” *Ethology, Ecology and Evolution* 31 (6): 557–67.
822 <https://doi.org/10.1080/03949370.2019.1636141>.

823 O’Dea, R.E., Lagisz, M, Jennions, MD, Koricheva, J, Noble, DWA, Parker, TH, Gurevitch, J, et
824 al. 2021. “Preferred Reporting Items for Systematic Reviews and Meta-analyses in
825 Ecology and Evolutionary Biology: A PRISMA Extension.” *Biological Reviews* 96 (5):
826 1695–1722. <https://doi.org/10.1111/brv.12721>.

827 Ordiz, A, Stoen, OG, Saebo, S, Sahlen, V, Pedersen, BE, Kindberg, J, and Swenson, JE. 2013.
828 “Lasting Behavioural Responses of Brown Bears to Experimental Encounters with
829 Humans.” *Journal of Applied Ecology* 50 (2): 306–14.
830 <https://doi.org/10.1111/1365-2664.12047>.

831 Oriol-Cotterill, A, Valeix, M, Frank, LG, Riginos, C, and Macdonald, DW. 2015. “Landscapes of
832 Coexistence for Terrestrial Carnivores: The Ecological Consequences of Being
833 Downgraded from Ultimate to Penultimate Predator by Humans.” *OIKOS* 124 (10): 1263–
834 73. <https://doi.org/10.1111/oik.02224>.

835 Ouzzani, M, Hammady, H, Fedorowicz, Z, and Elmagarmid, A. 2016. “Rayyan—a Web and
836 Mobile App for Systematic Reviews.” *Systematic Reviews* 5 (1): 210.
837 <https://doi.org/10.1186/s13643-016-0384-4>.

838 Page, MJ., McKenzie, JE, Bossuyt, PM, Boutron, I, Hoffmann, TC, Mulrow, CD, Shamseer, L, et
839 al. 2021. “The PRISMA 2020 Statement: An Updated Guideline for Reporting Systematic

- 840 Reviews.” *BMJ* 372 (March):n71. <https://doi.org/10.1136/bmj.n71>.
- 841 Pangle, WM, and Holekamp, KE. 2010. “Lethal and Nonlethal Anthropogenic Effects on Spotted
842 Hyenas in the Masai Mara National Reserve.” *Journal of Mammology* 91 (1): 154–64.
843 <https://doi.org/10.1644/08-mamm-a-359r.1>.
- 844 Papouchis, C.M., Singer, F.J., and Sloan, W.B. 2001. “Responses of Desert Bighorn Sheep to
845 Increased Human Recreation.” *Journal of Wildlife Management*.
846 <https://doi.org/10.2307/3803110>.
- 847 Pauly, D, Christensen, V, Dalsgaard, J, Froese, R, and Torres, F. 1998. “Fishing Down Marine
848 Food Webs.” *Science* 279 (5352): 860–63. <https://doi.org/10.1126/science.279.5352.860>.
- 849 Pecorella, I, Ferretti, F, Sforzi, A, and Macchi, E. 2016. “Effects of Culling on Vigilance
850 Behaviour and Endogenous Stress Response of Female Fallow Deer.” *Wildlife Research*
851 43 (3): 189–96. <https://doi.org/10.1071/WR15118>.
- 852 Picardi, S, Basille, M, Peters, W, Ponciano, JM, Boitani, L, and Cagnacci F. 2019. “Movement
853 Responses of Roe Deer to Hunting Risk.” *Journal of Wildlife Management* 83 (1): 43–51.
854 <https://doi.org/10.1002/jwmg.21576>.
- 855 Podgorski, T, Bas, G, Jedrzejewska, B, Sonnichsen, L, Snieszko, S, Jedrzejewski, S, and Okarma,
856 H. 2013. “Spatiotemporal Behavioral Plasticity of Wild Boar (*Sus Scrofa*) under
857 Contrasting Conditions of Human Pressure: Primeval Forest and Metropolitan Area.”
858 *Journal of Mammology* 94 (1): 109–19. <https://doi.org/10.1644/12-MAMM-A-038.1>.
- 859 Price, M. 2008. “The Impact of Human Disturbance on Birds: A Selective Review.” *Australian*
860 *Zoologist* 34:163–96. <https://doi.org/10.7882/fs.2008.023>.
- 861 Proudman, NJ, Churski, M, Bubnicki, JW, Nilsson, JA, and Kuijper, DPJ. 2021. “Red Deer

862 Allocate Vigilance Differently in Response to Spatio-Temporal Patterns of Risk from
863 Human Hunters and Wolves.” *Wildlife Research* 48 (2): 163–74.
864 <https://doi.org/10.1071/WR20059>.

865 R Core Team. 2024. “R: A Language and Environment for Statistical Computing.” Manual.
866 Vienna, Austria: R Foundation for Statistical Computing. <https://www.R-project.org/>.

867 Rainer, F. and Pauly, D. 2010. “FishBase.” Fisheries Centre, University of British Columbia Los
868 Baños, Philippines.

869 Reimers, E, Lund, S, and Ergon, T. 2011. “Vigilance and Fright Behaviour in the Insular Svalbard
870 Reindeer (*Rangifer Tarandus Platyrrhynchus*).” *Canadian Journal of Zoology* 89 (8): 753–
871 64. <https://doi.org/10.1139/Z11-040>.

872 Sbragaglia, V, Morroni, L, Bramanti, L, Weitzmann, B, Arlinghaus, R, and Azzurro, E. 2018.
873 “Spearfishing Modulates Flight Initiation Distance of Fishes: The Effects of Protection,
874 Individual Size, and Bearing a Speargun.” *ICES Journal of Marine Sciences* 75 (5):
875 1779–89. <https://doi.org/10.1093/icesjms/fsy059>.

876 Schmitz, OJ, Beckerman, AP, and O’Brien, KM. 1997. “Behaviorally Mediated Trophic
877 Cascades: Effects of Predation Risk on Trophic Interactions” *Ecology* 78 (6).

878 Shannon, G, Angeloni, LM, Wittemyer, G, Fristrup, GM, and Crooks, KR. 2014. “Road Traffic
879 Noise Modifies Behaviour of a Keystone Species.” *Animal Research* 94:135–41.
880 <https://doi.org/10.1016/j.anbehav.2014.06.004>.

881 Shelton, E. J., J. E. S. Higham, and P. Seddon. "Habituation." *penguin research and ecotourism:*
882 *some thoughts from left field. New Zealand Journal of Zoology* 31, no. 1 (2004): 119.

883 Sheriff, MJ., Peacor, SD, Hawlena, D, and Thaker, M. 2020. “Non-consumptive Predator Effects

884 on Prey Population Size: A Dearth of Evidence.” Edited by Jean-Michel Gaillard. *Journal*
885 *of Animal Ecology* 89 (6): 1302–16. <https://doi.org/10.1111/1365-2656.13213>.

886 Sih, A. 1984. “The Behavioral Response Race Between Predator and Prey.” *The American*
887 *Naturalist* 123 (1): 143–50. <https://doi.org/10.1086/284193>.

888 Smith, J.A., Wang Y, and Wilmers, CC. 2016. “Spatial Characteristics of Residential
889 Development Shift Large Carnivore Prey Habits.” *Journal of Wildlife Management* 80 (6):
890 1040–48. <https://doi.org/10.1002/jwmg.21098>.

891 Smith, JA., Gaynor, KM, and Suraci, JP. 2021. “Mismatch Between Risk and Response May
892 Amplify Lethal and Non-Lethal Effects of Humans on Wild Animal Populations.”
893 *Frontiers in Ecology and Evolution* 9 (March):604973.
894 <https://doi.org/10.3389/fevo.2021.604973>.

895 Smith, JA., McDaniels, ME, Peacor, SD, Bolas, EC, Cherry, MJ, Dorn, NJ, Feldman, OK, et al.
896 2024. “Population and Community Consequences of Perceived Risk from Humans in
897 Wildlife.” *Ecology Letters* 27 (6): e14456. <https://doi.org/10.1111/ele.14456>.

898 Stansell, HM, Blumstein, DT, Yeh PJ, and Nonacs P. 2022. “Individual Variation in Tolerance of
899 Human Activity by Urban Dark-Eyed Juncos (*Junco Hyemalis*).” *Wilson Journal of*
900 *Ornithology* 134 (1): 43–51. <https://doi.org/10.1676/21-00001>.

901 Steneck, RS. 2012. “Apex Predators and Trophic Cascades in Large Marine Ecosystems:
902 Learning from Serendipity.” *Proceedings of the National Academy of Sciences* 109 (21):
903 7953–54. <https://doi.org/10.1073/pnas.1205591109>.

904 Suraci, JP, Clinchy, M, Zanette, LY, and Wilmers, CC. 2019. “Fear of Humans as Apex Predators
905 Has Landscape-Scale Impacts from Mountain Lions to Mice.” *Ecology Letters* 22

906 (10): 1578–86. <https://doi.org/10.1111/ele.13344>.

907 Suraci, JP, Smith, JA, Clinchy, M, Zanette, LY, and Wilmers, CC. 2019. “Humans, but Not Their
908 Dogs, Displace Pumas from Their Kills: An Experimental Approach.” *Scientific Reports*
909 9. <https://doi.org/10.1038/s41598-019-48742-9>.

910 Tran, DSC, Langel, KA, Thomas, MJ, and Blumstein, DT. 2016. “Spearfishing-Induced
911 Behavioral Changes of an Unharvested Species inside and Outside a Marine Protected
912 Area.” *Current Zoology* 62 (1): 39–44. <https://doi.org/10.1093/cz/zov006>.

913 Treves, A, and Naughton-Treves, L. 1999. “Risk and Opportunity for Humans Coexisting with
914 Large Carnivores.” *Journal of Human Evolution* 36 (3): 275–82.
915 <https://doi.org/10.1006/jhev.1998.0268>.

916 Tucker, M. A., Santini, L., Carbone, C., & Mueller, T. (2021). Mammal population densities at a
917 global scale are higher in human-modified areas. *Ecography*, 44(1), 1-13.

918 Uchida, K, and Blumstein, DT. 2021. “Habituation or Sensitization? Long-Term Responses of
919 Yellow-Bellied Marmots to Human Disturbance.” *Behavioral Ecology* 32 (4): 668–
920 78. <https://doi.org/10.1093/beheco/arab016>.

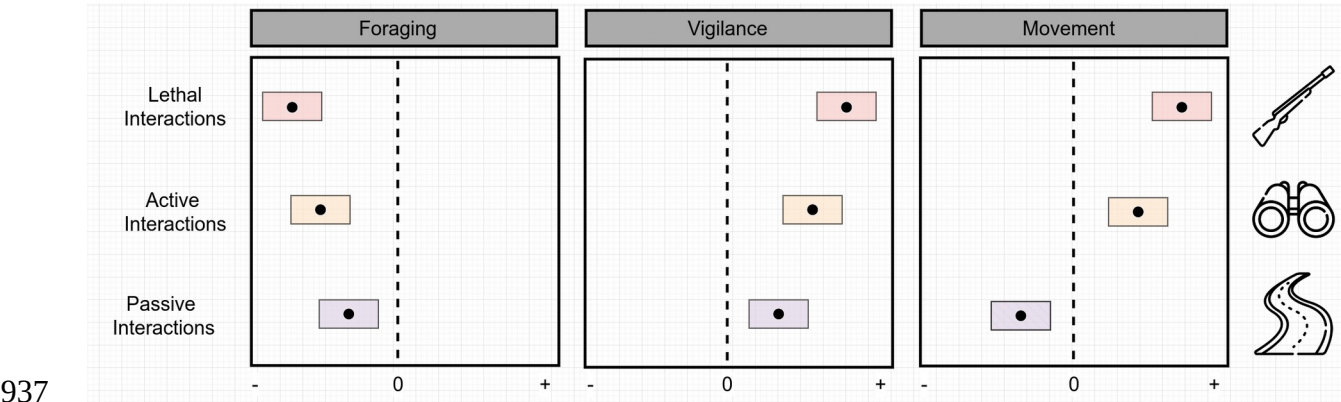
921 van der Kolk, H., Allen, M., Bruno J., and van de Pol, M. 2024. “Frequency-Dependent
922 Tolerance to Aircraft Disturbance Drastically Alters Predicted Impact on Shorebirds.”
923 *Ecology Letters* 27 (6): e14452. <https://doi.org/10.1111/ele.14452>.

924 Ward, C, and Low, B.S. 1997. “Predictors of Vigilance for American Crows Foraging in an Urban
925 Environment.” *The Wilson Journal of Ornithology*.

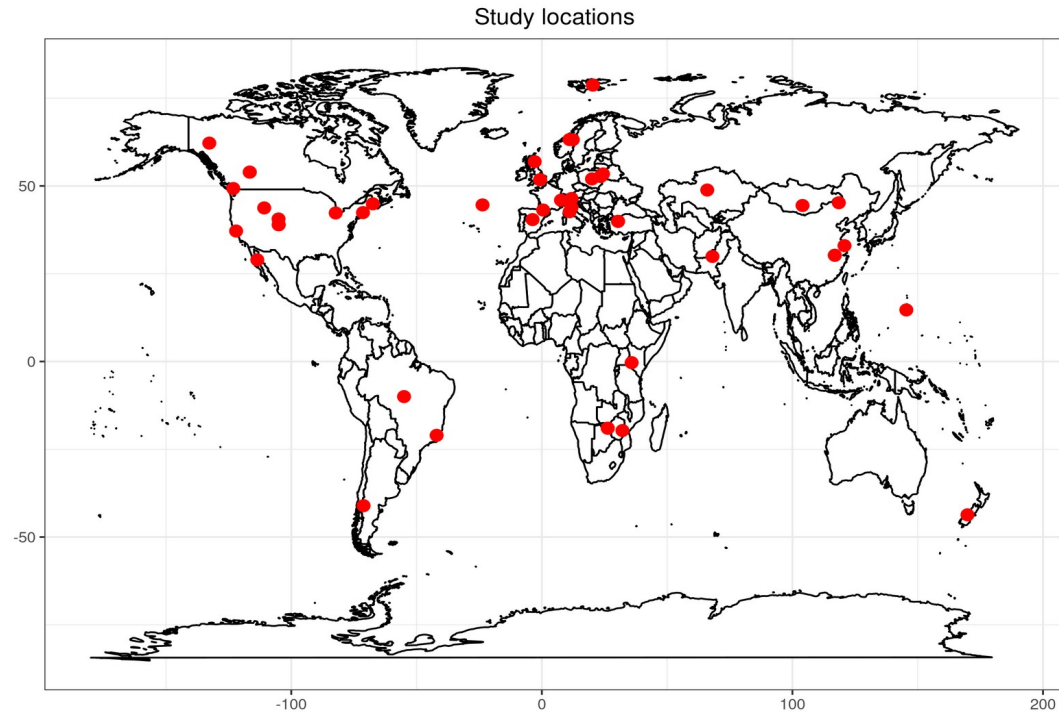
926 Westekemper K., Reinecke, H., Signer, J., Meißner, M., et al. 2018. “Stay on Trails - Effects of
927 Human Recreation on the Spatiotemporal Behavior of Red Deer *Cervus Elaphus* in a

- 928 German National Park.” *Wildlife Biology*. <https://doi.org/10.2981/wlb.00403>.
- 929 Williams, Rob, Andrew W. Trites, and David E. Bain. 2002. “Behavioural Responses of Killer
- 930 Whales (*Orcinus Orca*) to Whale-Watching Boats: Opportunistic Observations and
- 931 Experimental Approaches.” *Journal of Zoology* 256 (2): 255–70.
- 932 <https://doi.org/10.1017/S0952836902000298>.
- 933 Wirsing, AJ., Heithaus, MR, Brown, JS, Kotler, BP, and Schmitz, OJ. 2021. “The Context
- 934 Dependence of Non-consumptive Predator Effects.” Edited by Jonathan Chase. *Ecology*
- 935 *Letters* 24 (1): 113–29. <https://doi.org/10.1111/ele.13614>.

936 **Figures and tables**



938 **Figure 1:** The different scenarios in which humans interact with animals in wild spaces and the
939 hypothesised consequences on animal behavior. We expect a gradient of response with the
940 strongest behavioral changes resulting from lethal interactions with humans, followed by active
941 interactions and passive interactions. The scenarios are: a) A hunter that targets top predators
942 (top) b) A group of ecotourists watching animals in their natural habitat (middle) c) A road cutting
943 across a natural habitat (bottom).



945 **Figure 2:** Geographical distribution of studies included in the systematic review.

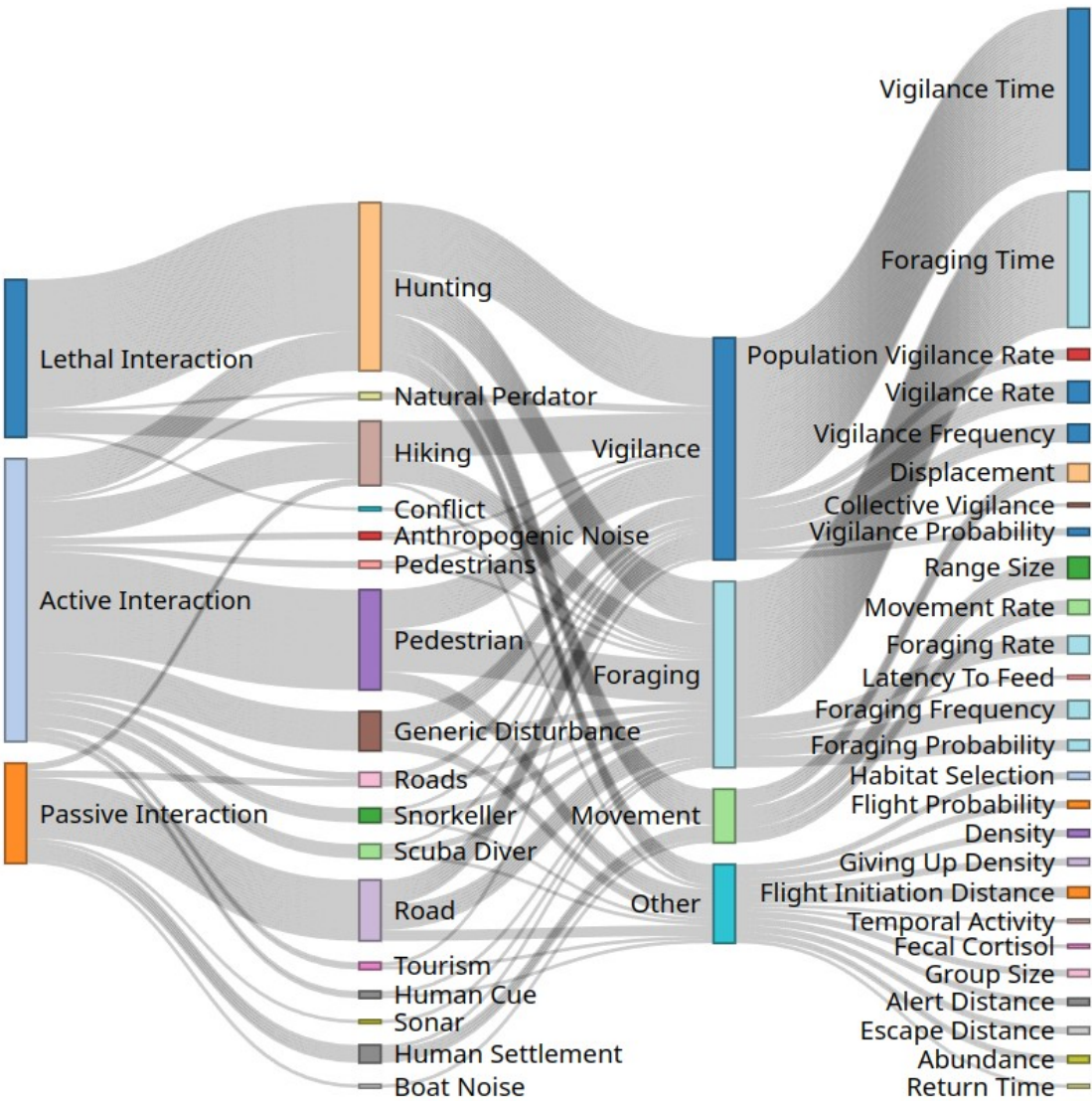


Figure 3: Distribution of included studies across type of interaction, treatment, behavior and measured outcome (in order).

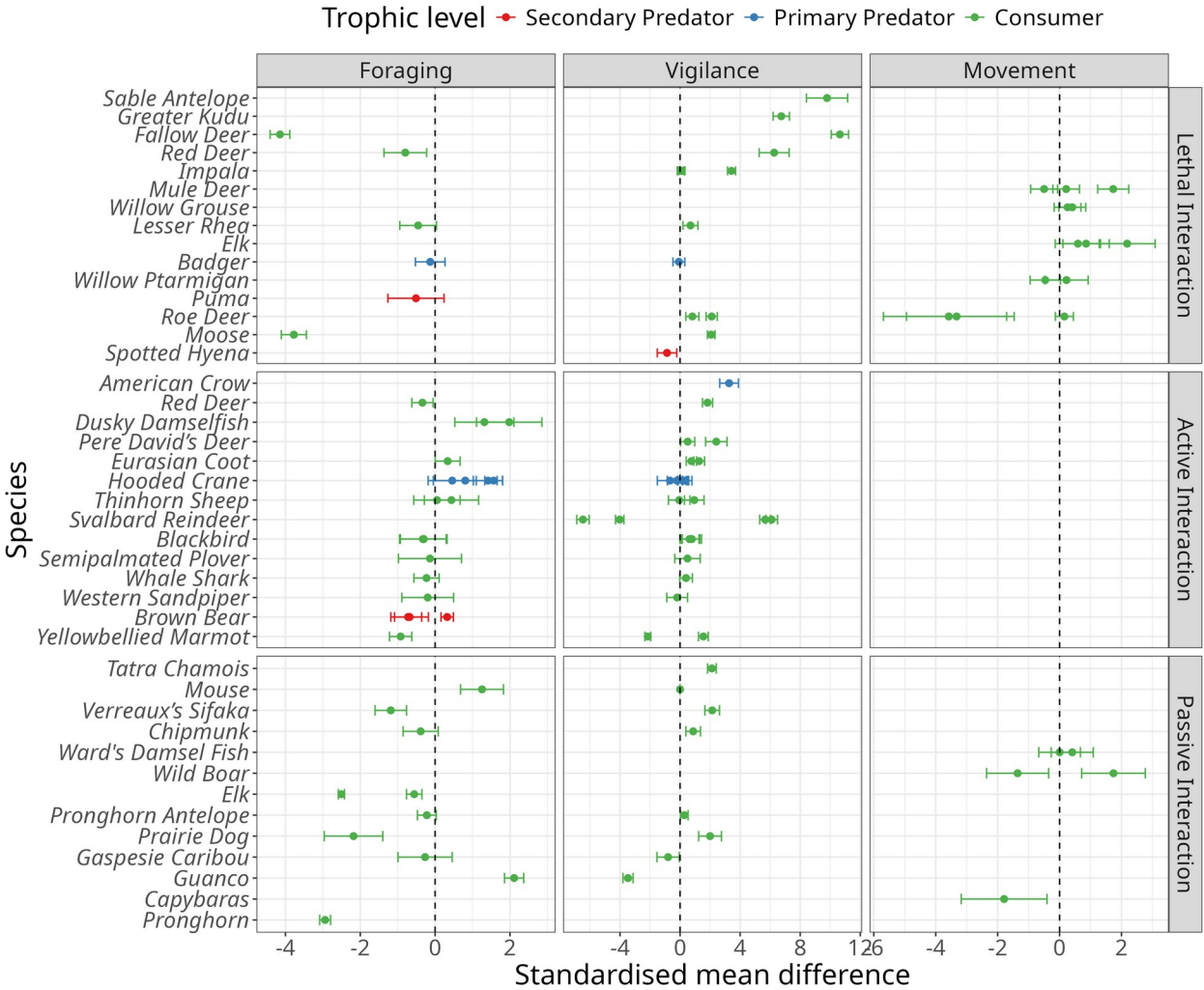


Figure 4: The effect of human disturbance (lethal, active disturbance, passive disturbance) across taxa and trophic levels. Negative values on the x-axis indicate a reduction in that behaviour, whereas positive values indicate an increase in that behaviour, relative to control conditions where no human activity is present.

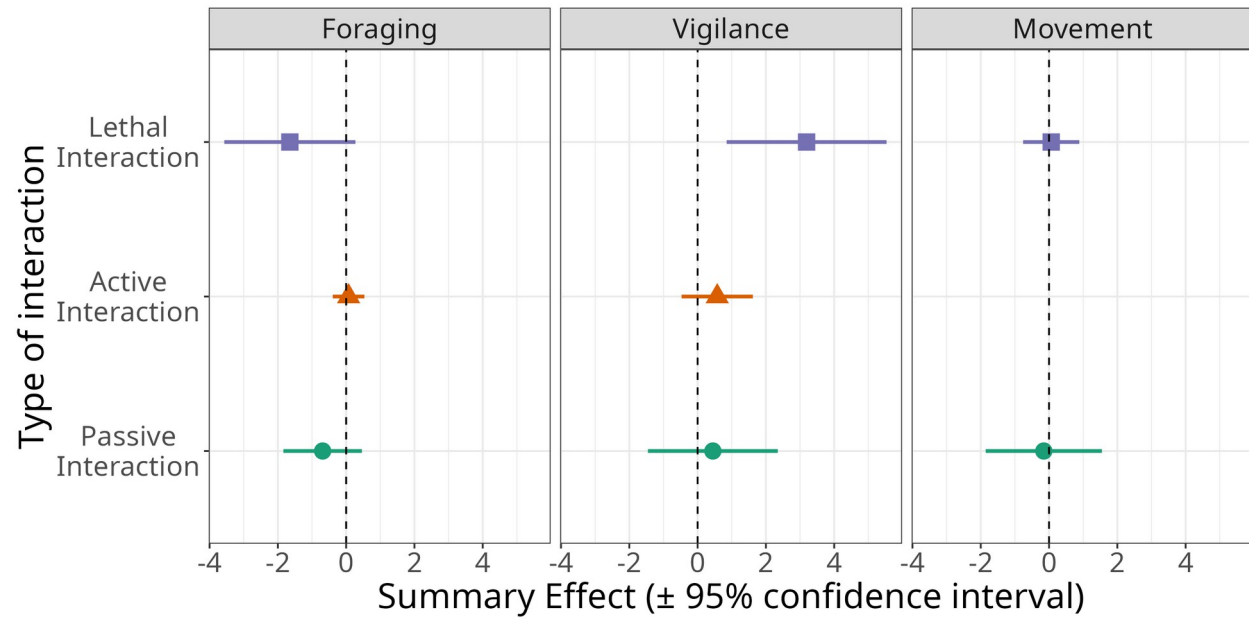


Figure 5: Summary effect of human interactions on foraging, movement, and vigilance behaviour across studies included in this meta-analysis. For movement, active disturbance was excluded due to low sample size.