



Environmental indicators combined with risk analysis to evaluate potential wildfire incidence on the Dadu Plateau in Taiwan

Chao-Yuan Lin¹ · Pei-Ying Shieh¹ · Shao-Wei Wu¹ · Po-Cheng Wang¹ · Yung-Chau Chen¹

Received: 15 November 2021 / Accepted: 22 February 2022 / Published online: 19 March 2022
© The Author(s) 2022

Abstract

Wildfire is a common disaster in the world, and it has a considerable impact on the safety of residents and ecological disturbance. Periodic wildfires are an urgent problem to be solved. This research uses big data from relevant departments to extract environmental indicators that affect wildfires, including satellite images, meteorological observations, and field surveys and establishes a risk model for the Spatio-temporal distribution of wildfires based on risk analysis. Previous studies using Differenced Normalized Burn Ratio (*dNBR*) to assess fire severity and distinguish wildfire ruins did not deal with the impact of atmospheric humidity on *dNBR* values. In this study, an adjustable fire threshold was developed to enable *dNBR* to improve the accuracy of identifying wildfire locations. Regarding the temporal distribution of wildfire risks, environmental vulnerability cannot specifically reflect the frequency of actual wildfires. If the hazard degree is introduced to calculate the wildfire risk, the coefficient of determination can be increased from 0.49 to 0.79. The verification of the village boundary zone depicts that the risk analysis can effectively show the temporal and spatial distribution of wildfire hotspots. On this basis, a village-level wildfire disaster prevention strategy can be formulated.

Keywords Remote sensing · Principal component analysis · Fire hazard · Fire vulnerability · Normalized burn ratio

1 Introduction

Global wildfire research mainly focuses on forests, tropical forests, and grasslands in North America and Northern Eurasia, Mediterranean climate zones in Europe and Australia, and Southeast Asian monsoon regions (Bowman et al. 2009; Ying et al. 2021). Taiwan is located in the East Asian monsoon region, and it is also a place where wildfires occur frequently. In addition to being affected by climate, changes in land use caused by human

✉ Chao-Yuan Lin
yuanlin@dragon.nchu.edu.tw

¹ Department of Soil and Water Conservation, National Chung Hsing University, Taichung, Taiwan, ROC

activities (Chiu et al. 2012) are another key factor causing frequent wildfires. Over the past decade, the average seasonal length of wildfires has increased by 18% due to climate impacts (Jolly et al. 2015), leading to an increase in the frequency and severity of wildfires, a trend that has worsened with the increase in man-made development. With infrastructure construction and rapid economic development, special attention has been paid to the control and management of potential wildfire risks in the region, especially in hillside areas. Wildfires will burn vegetation, expose the soil, change the physiochemical properties of the soil, and accelerate erosion and degradation (Moody et al. 2013; Shakesby and Doerr 2006). Furthermore, wildfire could contribute to the formation of water-repellent layers on the surface (Bowker et al. 2004; Ravi et al. 2009; Shakesby and Doerr 2006). This phenomenon, in addition to reducing soil infiltration, also increases surface runoff (Neary and Gottfried 2002; Pausas et al. 2008) and peak flow (Costa 1987; Hasan et al. 2020), which reduce water storage capacity and can easily lead to flooding in downstream urban areas of Dadu Plateau during heavy rains (Cho 2017). Slope stability of burned site can also lead to forest degradation and collapse (Benda and Dunne 1997; Jackson and Roering 2009), posing a significant threat to downstream urban infrastructure and livelihoods. Wildfire is an essential factor that interferes with the ecosystem, and fire intensity and frequency affect the structure/function of the ecosystem (Sankey et al. 2013; Westerling et al. 2006). Frequent wildfires reduce forest area and hinder plant succession. Ecologically, migratory birds may also be affected by reduced species abundance and forest ecological habitat loss (Tsai 2005). In addition, fires release carbon and increase the burden of global greenhouse gases, leading to intensified global warming. Extreme weather makes wildfires more frequent and destructive (Westerling et al. 2006; Turco et al. 2018).

An average of 57 wildfires occurs in Taiwan's forests each year, which is the most frequent area of Taiwan's wildfires. In comparison, there are more than 400 in Dadu Plateau (Chuang 2015). This is related to the land development dominated by local human activities and plant characteristics. The forest form of Dadu Plateau is dominated by Taiwan acacia (*Acacia confusa* Merr.), and the grassland is mainly distributed by Guinea grass (*Panicum maximum* Jacq.) (Lin and Yang 2001; Tsai 2005). Many studies have shown that wildfires are associated with meteorological factors such as rainfall (Hsiao 2003), relative humidity (Lai 2003; Lin et al. 2005), temperature (Hsiao 2003; Chiu et al. 2004), and wind speed (Chiu et al. 2004). The growth of Guinea grass in the Dadu Plateau will quickly wither and become a source of fire, which will burn quickly after October due to plummeted rainfall and the effects (the formation of burning wind on the leeward side) of the northeast monsoon (Chuang 2015; Lin and Yang 2001; Tsai 2005). In addition, most fires are caused by farming and animal husbandry, as well as the common sacrificial customs in traditional Chinese festivals, such as burning paper and burning incense (Lin et al. 2008; Chuang 2015).

Wildfire is a coupled phenomenon, involving multiple factors, such as environmental humidity conditions, fuel flammability (fuel moisture content), natural and man-made ignitions, weather, and other comprehensive factors (Gudmundsson et al. 2014; Pyne and Andrews 1984). The vital information of wildfires can be extracted through satellite remote sensing data. It has the characteristics of real-time and large area and has been widely used in fire modeling at different scales around the world (Tian et al. 2013; Abatzoglou et al. 2018; Chen et al. 2020). Therefore, understanding the changes in land cover, environmental climate conditions and past historical fire data will help wildfire risk modeling and wildfire management and prevention. Due to the many factors involved, most wildfire models are built under the risk concept framework, using machine learning and artificial intelligence statistical methods to aggregate complex variables into a comprehensive index (Oliveira

et al. 2021a, b). However, there are still research gaps in the spatial and temporal distribution of potential wildfires. This study used Guinea grass samples and satellite imagery, meteorology and land-use data to detect changes in biomass and tissue moisture content on a monthly basis to determine the associated indicators affecting the occurrence of wildfires. Finally, a wildfire risk model at the village level was established for reference by local fire stations and residents for wildfire control.

2 Materials and methods

2.1 Study area

Dadu Plateau is located in central Taiwan (120°30' E–120°40' E, 24°20'N–24°05'N). The altitude of the Plateau is about 200~300 m (Fig. 1-left). The terrain is uplifted by the thrust of the fault, forming a terraced terrain with steep west and gentle east. Weather and climate are mainly affected by the East Asian monsoon belt. Based on the data of the Wuchi Meteorological Station, the average annual temperature is about 23°C, the coldest monthly mean temperature is 15.9°C, and the hottest monthly average temperature is 29.2°C (Central weather bureau, 2006 to 2015). The average monthly rainfall is about 47.5 mm and 182.5 mm for the dry duration (October to the following April) and wet duration (May to September), respectively. The land use is mainly forest, grassland, agricultural land, and building (Fig. 1-right), which is affected by the local soil properties. The soil is laterite with the properties of acidity, lack of organic matter, high viscosity/plasticity and well-drained and is not suitable for cultivation due to low productivity and lack of irrigation facilities (Chiu et al. 2012). The grassland is dominated by Guinea grass, which is a fire-adaptive species with tiller ability and can survive well on acidic and barren soils (Lin 2000). Taiwan acacia with symbiotic rhizobia in its root systems (Lee 2013) is the main

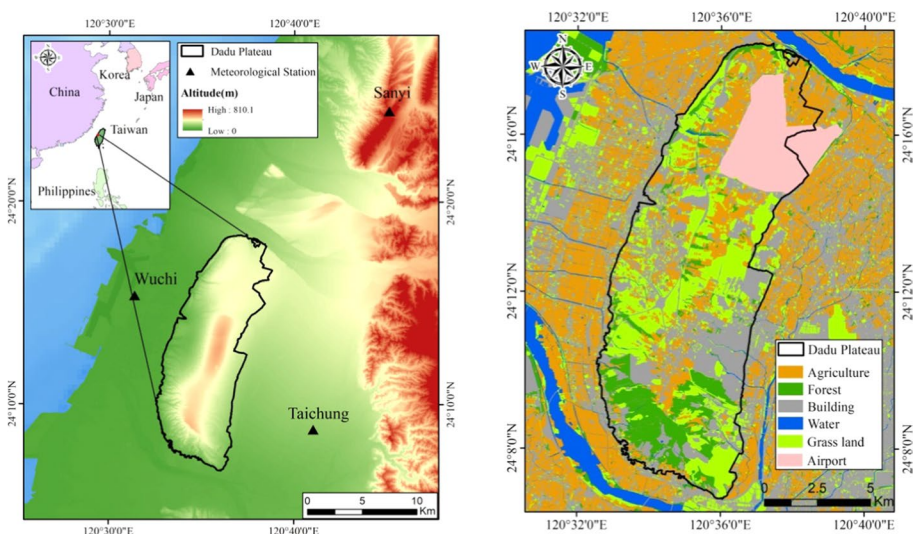


Fig. 1 The location and elevation distribution of Dadu Plateau (left) and land use type (right) extracted from DEM and map data of the National Land Surveying and Mapping Center (<https://maps.nlsc.gov.tw/>)

woody plant at the afforestation areas. According to fire reports, 83% of the fire ruins can be found on the grasslands, and the clear ecotones formed by wildfires are usually spotted at the adjacent grassland.

2.2 Study materials

This study collects and compiles big data from various agencies and considers the limitations of fire reporting materials. The spatial and temporal distribution of wildfire risk was analyzed using data from 2006 to 2015. Meteorological data are extracted from the annual weather report of the central weather bureau (<https://e-service.cwb.gov.tw/HistoryDataQuery/>) to calculate the monthly Evenoff's moisture coefficient (EMC) and average wind speed. Satellite images Landsat 4–5 TM and Landsat 8 OLI/TIRS are derived from the center of the US Geological Survey website (<http://earthexplorer.USGS.gov/>) to extract relevant environmental indices (NDVI, NDWI, and NBR) for establishing a spatio-temporal distribution of wildfire risk model. In addition, biomass and tissue moisture content are derived from the monthly data of the on-site survey from 2014–2015, and fire case provided by the Fire bureau is used for model validation.

3 Methods

3.1 Wildfire model based on risk concept

Wildfire risk is often a combination of potential environmental vulnerability factors and external hazard disturbances, such as persistent drought, excessive accumulation of flammable fuels, environmental humidity, man-made activity and development. In the theory of natural disaster risk assessment, wildfire risk assessment is widely and prominent. The IPCC risk model (IPCC 2014) used to define wildfire risk as the product of wildfire hazard and environmental vulnerability is shown in Eq. (1).

$$\text{Risk} = \text{Hazard} \times \text{Vulnerability} \quad (1)$$

The risk is the probability of wildfire occurrence, the hazard is the probability of ignition, and vulnerability is the flammability of fuels. In this study, the risk of wildfires will be discussed based on temporal and spatial distribution and extracted the relevant environmental indicators according to the vulnerability and hazard. The establishment of the temporal wildfire risk model is to evaluate monthly wildfire risk; the spatial wildfire risk model aims to draw a wildfire risk map of the Dadu Plateau during high wildfire risk.

3.2 Temporal distribution of wildfire vulnerability indicators

3.2.1 Wildfire vulnerability indicators

The research first takes environmental factors from weather data, remote sensing data, and field surveys as environmental vulnerability, uses principal component analysis (PCA) to combine various environmental vulnerability factors into a comprehensive index, and uses monthly wildfire vulnerability to assess wildfire risk. EMC is one of the drought indices based on cumulative rainfall, temperature, and relative humidity and can be considered a

composite index of atmospheric moisture content (Li, 1990; Yen and Wu, 2004). Based on accumulated rainfall, average temperature and mean relative humidity at nearby weather stations (Wuchi, Taichung, Sanyi) from 2006 to 2015 (Fig. 1-left), EMC of each station is calculated using monthly data and then, uses Thiessen's polygon approach to interpolate the research site's EMC. Thiessen's polygon approach is a common method of estimating meteorological factors in sample areas using weather stations, which calculate the weight of weather factors based on their control areas to understand EMC's monthly change. The EMC formula can be expressed as the following formula (2).

$$EMC = \frac{R}{E_0} \quad E_0 = 0.0018(25 + T)^2(100 - r) \quad (2)$$

T is the average temperature ($^{\circ}\text{C}$); r is the mean relative humidity (%); R is accumulated rainfall (mm). Monthly mean wind speed of the site is also estimated from the stations by Thiessen's polygon method.

Vegetation such as forests, grasslands, and farmland in the Dadu Plateau can be considered the main fuel for wildfires. Normalized Difference Vegetation Index can be extracted from the red and near-infrared light of satellite images (Rouse et al. 1973). In order to facilitate large-scale vegetation monitoring, the monthly NDVI is used to understand the trend of NDVI changes. NDVI changes are mainly influenced by environmental stresses such as soil moisture content, nutrient supply, tissue disease, and farming behavior (Fiorucci and Gaetani, 2007). Over the years, many authors have proposed using NDVI changes to assess the likelihood of wildfires in vegetation (Lasaponara 2005) and using NDVI as an index of wildfire vulnerability (Fiorucci and Gaetani, 2007).

Moisture content of plant tissue is a key factor affecting wildfire occurrence and development. Normalized Difference Water Index (NDWI) is defined as the reflectance difference of near-infrared (NIR) and short-wave-infrared light (SWIR) divided by the sum of the two, the more moisture content in the plant tissue the less probability of ignition and could be applied to reflect the variations of plant's moisture content. NDWI has a higher sensitivity compared to NDVI for vegetation moisture content (Karamihalaki et al. 2016).

According to fire report statistics, wildfires mainly occur in the dry season and usually occur in grasslands. This may be due to the rapid withering and fuel accumulation during the dry season. Guinea grass has become a major fire-adapting species on the local grasslands under long-term periodic wildfires. In addition to using the overall NDVI and NDWI as environmental vulnerability factors, monthly changes of biomass and/or moisture content of the Guinea tissue were investigated to be as factors of fire vulnerability. According to the quantitative plant survey (Lin 2013), random plots (1 m \times 1 m, three replications) of grassland located at the western Cheng-kung-ling were selected for regular monthly sampling from 2014 to 2015. The biomass and/or moisture content were calculated according to the manual of herbal treatment (Shyu et al. 2003). The Guinea grass of the sample plot was cut off to measure its fresh weight on an electronic scale and dried in the oven for 48 h at 85 degrees Celsius, and the resulting dry weight was biomass. The moisture content percentage is (fresh weight—dry weight)/dry weight \times 100%. Finally, the monthly NDVI, NDWI, biomass, and moisture content are used as wildfire vulnerability indicators.

3.2.2 Statistical analysis

The environmental vulnerability consists of a variety of index variables, which are difficult to explain due to the complexity of the data. Therefore, PCA is used to reduce the

dimensionality of seven environmental vulnerability variables. PCA is a common machine learning method widely used to process big data today (Holland 2019; Jolliffe and Cadima 2016). By transforming a large set of variables into a smaller set of variables, which still contains most of the information in the large set, the principal components with eigenvalue greater than 1 are selected to calculate the comprehensive vulnerability index for wildfire occurrence.

Before PCA, the value range of each factor must be standardized to obtain its influence weight. Due to the positive and negative relationship between impact factors and fire occurrence, biomass, NDVI and wind speed factors that have a positive contribution to fire vulnerability can be standardized as normalized biomass (nBM), normalized NDVI (nNDVI), and normalized Wind speed (nWS) are expressed by the following Eq. (3).

$$y = \frac{x_i - x_{\min}}{x_{\max} - x_{\min}} \quad (3)$$

where y is the standardized value, x_i is the original value of index i , and x_{\max} and x_{\min} are the maximum and minimum values of the index, respectively.

Factors that have a negative impact on fire vulnerability, such as EMC, NDWI, and moisture content, can be inverted into iEMC, iNDWI, and iMC; then, normalized to niEMC, niNDWI, and niMC through the above equation.

3.3 Spatial distribution of wildfire risk

Select the period most prone to wildfires, and use the environmental indicators of wildfire vulnerability and hazards to draw a wildfire risk spatial distribution map.

3.3.1 Spatial distribution of wildfire vulnerability

Wildfire vulnerability represents the ability of the environment to resist the fire that may occur from external forces. Areas with high wildfire vulnerability are accessible to ignite and burn after contact with a fire source. Among the factors that affect the temporal distribution of wildfire vulnerability, only NDVI, and NDWI extracted from satellite images are used to map the spatial distribution in a specific time interval environment. Due to insufficient spatial resolution of meteorological data, NDVI and NDWI already reflect the information of biomass and tissue moisture content, such temporal distribution modeling variables are excluded from the spatial distribution modeling. Generally speaking, areas with vigorous vegetation growth have greater potential for fuel accumulation, which can promote combustion or the spread of wildfires; fuel moisture also affects the burning rate. Therefore, NDVI and NDWI are used to express fuel quantity and moisture index, respectively.

Fuel properties also affect the occurrence of wildfires, so the type of land cover is closely related to the fire event. Brown and Davis (1973) pointed out that grass is a kind of light fuel, which is characterized by the rapid loss of water from plant tissues and air supply flux, making it easy to catch fire and spread, but burning does not usually last for a long time. However, the heavy fuel of wood is not easy to burn because the tissue moisture is not easy to lose, but after being ignited, it burns more and lasts longer.

Landsat images are classified according to maximum likelihood classification (Lillesand et al. 2000) and supervised classification. The images provided by Google Earth are used for ground truth training. Google Earth satellite images are mainly high-resolution images

of the DigitalGlobe series of satellites, which can be used to interpret the spectral statistical data of each training sample as a basis for image data classification. After using supervised classification for image classification, ground truth data need to be compared with each other to determine whether the classification result is accurate and whether it meets the requirements. Accuracy evaluation uses classification results to compare ground truth data, calculates the number of random samples using the formula proposed by Yamane in 1967, and uses the overall accuracy of the error matrix and Kappa coefficient as the accuracy evaluation criteria.

The dry season images most prone to wildfires are used to describe land use in agriculture, grassland, forest, and bare land. The vegetation types of agriculture and grassland are easy to burn and can be classified as light fuels. Forests can be classified as heavy fuel, and bare land is classified as non-fuel. Standardized fuel characteristic values can be divided into three levels (1, 0.5, 0), and the spatial distribution of fuel properties can be distributed according to the type of land cover.

3.4 Temporal/ Spatial distribution of wildfire hazard

Hazard can be defined as how external forces affect natural disasters. Lightning and human activities may cause wildfires, but they are challenging to observe and quantify. Under periodic wildfire conditions, the burning frequency of a place can be considered as one of the fire hazard indexes. Historic wildfire sites can be considered as fire spots and are most likely to catch fire in the future. Therefore, the extraction of wildfire locations in the past is a fire hazard index for assessing the fire risk of a specific location.

Normalized Burn Ratio (*NBR*) proposed by Caselles and López García in (1991) was applied to extract the burned sites from satellite images. The near infrared (*NIR*) with a wavelength of 0.76–0.90 μm and the short-wave infrared (*SWIR*₂) with a wavelength of 2.08–2.35 μm in the Landsat satellite sensor are used to describe possible burning locations. The value of *NBR* ranges from –1 to 1. The higher the value, the denser the vegetation. Conversely, the lower the value, the sparser the vegetation coverage. *NBR* can be expressed as Eq. (4).

$$NBR = \frac{NIR - SWIR_2}{NIR + SWIR_2} \quad (4)$$

In order to avoid misjudging areas with low *NBR* values as burning locations, (Key and Benson (2005) proposed *dNBR* to quantify environmental changes in fire disturbances. *dNBR* can be expressed as Eq. (5).

$$dNBR = NBR_{\text{prefire}} - NBR_{\text{postfire}} \quad (5)$$

By subtracting two *NBR* images, the changes in grid *dNBR* values in the images before and after the fire can be understood to reflect the severity of the combustion. Key and Benson (2005) used *dNBR* values in conjunction with wildfire ground surveys to propose fire severity classifications (Table 1). These thresholds are usually used to determine the severity of combustion and effectively extract the location of combustion (Epting et al. 2005; Hall et al. 2008; Soverel et al. 2011).

However, when using fire severity to classify the fire range, it should be noted that the initial threshold of *dNBR* is not a fixed value. Since the initial threshold of *dNBR* is affected by atmospheric humidity, if a fixed initial threshold of *dNBR* is used to extract

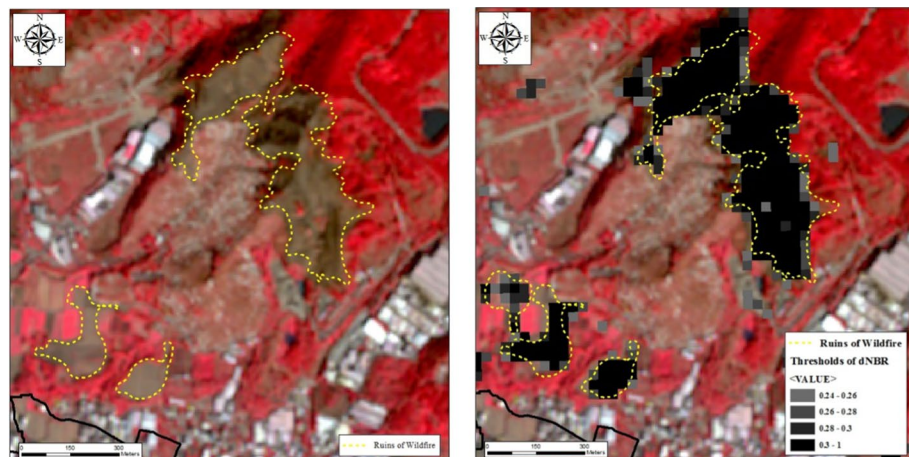
Table 1 Classification of burn severity

Burn severity	<i>dNBR</i>
Unburned	– 0.100~0.099
Low	0.100~0.269
Low-medium	0.270~0.439
Medium-high	0.440~0.659
High	0.660~1.300

Key and Benson (2005)

the fire range, the image must be limited to the same period (Fang and Yang 2014). In order to effectively and automatically use *dNBR* to identify the fire scene, it is necessary to consider the influence of atmospheric humidity changes on the telemetry images taken in the two stages. In this study, the actual fire scene image was used as the basic image, and the *dNBR* was adjusted to the boundary of the fire site based on the pseudo-color image (Fig. 2-left) to obtain the initial threshold of the *dNBR* (Fig. 2-right). In addition, we use the difference (Overall *dNBR*) of the average *NBR* of all grids of the two images in the study area as an indicator of the influence of changes in atmospheric humidity during the two periods. Use the initial threshold of *dNBR* obtained from each wildfire event as the y variable and the overall *dNBR* as the x variable to find the correlation between the two variables. Then, the initial threshold of *dNBR* can be dynamically evaluated through regression analysis.

Based on the characteristics of *dNBR*, this research studies the hazards of spatio-temporal distribution in risk assessment. The temporal distribution is based on the total area of fire sites in each period, and the spatial distribution analysis includes not only the frequency of fires in each region during a fixed period, but also the severity of wildfires.

**Fig. 2** Fire ruins explained from false color images based on Landsat image data (left). Fire remains extracted from different initial thresholds of *dNBR* (right)

3.4.1 Wildfire risk model verification

(1) Burned sites extraction

The changes in NDVI before and after wildfires were used to extract burned locations. Using K-means analysis to define high and low NDVI from the images before and after the fire, then the vegetation changes can be divided into four categories to identify the burning location. Therefore, the NDVI change level remains the same or higher and is classified as a non-fire area, and the NDVI change belongs to a lower category, which can be regarded as a burning site that can be used to verify wildfire risk models.

(2) Verification of village units

Spatial wildfire risk modeling is based on the grid scale of the image as the analysis unit. Once a wildfire occurs, it will spread rapidly and will not stop until there is not enough fuel to burn, so it is easy to break through the grid unit. Firebreak can effectively reduce and control the spread of fire (Ascoli et al. 2018; Fernández et al. 2019). Firebreaks comprise both natural and artificial barriers. Common natural barriers include cliffs, streams, valleys and ridge lines (Whitman et al. 2018). Artificial fire belts include green firebreaks (Cui et al. 2019) and roads (Laschi et al. 2019). Therefore, it is more appropriate to introduce fire belts as a concept for spatial wildfire risk assessment and verification. This study uses the village boundary as the division unit. The criteria for dividing villages are based on 1. Valley or ridgeline. 2. The center line of roads, alleys and river ditch. 3. Permanent gates, ponds, bridges and buildings that can be used as boundaries, all have the function of fire protection belts. A total of 62 villages (including airports) on the Dadu Plateau were used as the verification basis for the spatial wildfire risk assessment. The modeling results of the village districts can be more visually applied in fire management.

4 Results

4.1 Temporal distribution of wildfire vulnerability

4.1.1 EMC and mean wind speed

According to the statistical analysis of meteorological data from 2006 to 2015, the highest average temperature occurred in July (29.17 °C) and the lowest occurred in January (16.13 °C); the lowest average relative humidity (72.8%) occurred in October and the highest (81%) occurs in February; there is relatively high precipitation in June to August and relatively low rainfall during October to February. The EMC of each month could be obtained from the calculation of the above three meteorological factors and shows relatively high value during May to August. A high EMC value implies greater moisture content in the air. The mean wind speed for each month is relatively high during October to February due to the prevailed northeast monsoon. The monthly normalized linear reversed EMC (niEMC) and the normalized wind speed (nWS) are listed in Table 2.

Table 2 Meteorological data, EMC, NDWI, and NDVI

Items	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
Mean Temp.(°C)	16.13	17.10	19.10	22.82	26.06	27.95	29.17	28.89	28.05	25.24	22.12	17.83
Mean RH(%)	78.74	81.03	76.49	76.83	77.29	78.31	74.99	76.70	75.36	72.84	75.25	74.07
Accumulated rainfall (mm)	26.48	58.64	79.38	125.86	202.78	316.88	252.91	336.98	146.67	37.78	55.83	36.01
EMC	0.40	0.15	0.96	1.31	1.90	2.89	1.91	2.76	1.17	0.30	0.56	0.42
nEMC	0.90	1.00	0.70	0.57	0.36	0.00	0.35	0.04	0.62	0.94	0.84	0.90
Mean wind speed,(m/s)	4.97	4.39	3.84	3.12	3.01	3.06	3.17	2.82	3.23	4.53	4.14	4.58
nWS	1.00	0.73	0.48	0.14	0.09	0.11	0.16	0.00	0.19	0.80	0.61	0.82
NDWI	-0.05	-0.11	-0.04	0.02	0.33	0.35	0.31	0.24	0.25	0.18	0.04	-0.01
nNDWI	0.12	0.00	0.14	0.27	0.94	1.00	0.91	0.75	0.77	0.61	0.33	0.20
niNDWI	0.88	1.00	0.86	0.73	0.06	0.00	0.09	0.25	0.23	0.39	0.67	0.80
NDVI	0.35	0.29	0.34	0.49	0.76	0.72	0.72	0.71	0.71	0.54	0.45	0.39
nNDVI	0.13	0.00	0.11	0.42	1.00	0.92	0.92	0.91	0.91	0.54	0.34	0.21

*The data are derived from the average of the data of each month from 2006 to 2015 and then, normalized

4.1.2 NDWI and NDVI

Satellite images are used to extract monthly averages of NDVI and NDWI (Table 2). Figure 3 shows a significant increase in nNDWI and nNDVI from April to May due to abundant rainfall during the rainy season and adequate water use for plant photosynthesis. At the beginning of the dry season (September to October), nNDWI is on a downward trend due to reduced rainfall, while nNDVI is sluggish. NDWI is more sensitive to changes in atmospheric moisture than NDVI. Overall, the monthly changes in nNDWI and nNDVI are fairly consistent.

4.1.3 Monthly biomass and moisture content of Guinea grassland

A sampling of Guinea grass began in November 2014 and lasted for one year. Unfortunately, the sample sites were destroyed from May to October 2015 due to fire incident. Biomass accumulates with the growth of the plant, and the biomass is supplemented by the accumulation of NDVI throughout the year; NDWI is more sensitive to changes in atmospheric humidity than NDVI, and the moisture content of the plant is supplemented by NDWI. Multiple imputation method was used to supplement the missing data, and regression analysis by statistical software SPSS. The imputation regression equations of biomass and tissue moisture content are shown in Eq. (6) and (7).

$$Y_1 = 1.002 + 0.068X_1 (R^2 = 0.88, P = 0.006) \quad (6)$$

$$Y_2 = 56.009 + 89.089X_2 (R^2 = 0.60, P = 0.07) \quad (7)$$

Y_1 is biomass, X_1 is nNDVI accumulation; Y_2 is moisture content, X_2 is NDWI. The biomass of grass should increase annually with the growth of grass, but the results show that in the early dry season, the biomass is the highest. During the drought season, plant growth slows down or even decreases slightly due to soil water shortage and feeding by pests or animals. Overall, there is little change in biomass over the year. The results showed no

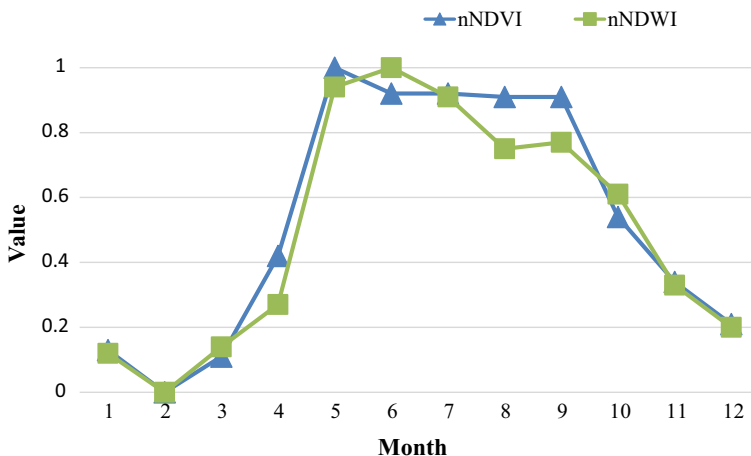


Fig. 3 Monthly changes of nNDVI and nNDWI

significant difference in the monthly biomass of the Guinea grasslands, which means that the Guinea grasslands are abundant in fuel all the year round. However, changes in plant moisture content showed a significant difference in the monthly change (Table 3).

4.1.4 Statistical analysis result

The principal component analysis was performed using the niEMC, nWS, niNDWI, nNDVI, nBM, and niMC factors described above for variable dimension reduction. After the KMO and Bartlett's Test, there is a common medium factor between the variables ($KMO > 0.7$, $P < 0.01$) for principal component analysis. The eigenvalue and variance explained percentages for components 1 and 2 are 4.39, 73.12, and 1.15, 19.17, respectively (Table 4). Because their eigenvalue is greater than 1, two components are retained, and the cumulative variance explained by the two components is 92.29%. This means that the interpretation is good, and most of the original variables can be interpreted.

We hide the absolute value of the factor less than 0.5 and use the Varimax method to rotate the original factor matrix. The rotating factor matrix shows that component 1 consists of five factors (NDWI, NDVI, MC, EMC, and WS) and component 2 contains only BM factors (Table 5). Using the rotation factor matrix to calculate the eigenvalue and variance explained percentage, components 1 and 2 show 4.20, 82% and 0.92, 18%, respectively, which can be considered as the effect of each component on the vulnerability of wildfires. The larger the coefficient in the factor matrix, the greater the weight it occupies in the component. Therefore, the factor load in the component multiplied by the percentage of variance explained can be calculated as a weight for the effect of each factor on the vulnerability of wildfires.

4.1.5 Monthly wildfire nVulnerability

Monthly wildfire vulnerability is calculated based on the weight of the factors. Analysis of potential wildfires shows that December is the most vulnerable, indicating a period of dryness, fuel-rich and good wildfire conditions. Overall, there is a high degree of vulnerability from October to the following April ($nVulnerability > 0.5$), and low vulnerability from May to September.

According to the fire department's 2006–2015 records, the most significant number was about 200 in March and the second-highest number of fires in October (Fig. 4). The degree of vulnerability from November to the following February was slightly different from the record of fire cases, while the trend of March to October was more consistent with the record of fire cases. It can be seen that the month of high fire vulnerability does not reflect the actual occurrence of fire records.

Correlation of monthly wildfire nVulnerability and fire case is expressed as a positive linear relationship, but overall can only explain 49% ($p < 0.05$). Only considering vulnerability factors alone is not sufficient to reflect wildfire risk. Fire nVulnerability during the rainy season (May–September) and dry season (October–April) was analyzed (Fig. 5). The monthly wildfire nVulnerability and fire case records show highly positive correlation ($R^2 = 0.77$, $p = 0.05$) during the rainy season, while the dry season is negatively correlated ($R^2 = 0.22$, $p = 0.28$). Wildfire nVulnerability during the dry season can only explain the 22% wildfire risk, so it is necessary to add the hazard to explore the risk of wildfires, especially during the dry season.

Table 3 Monthly biomass and moisture content of the Guinea grassland

Items	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
Biomass (kg/m ²)	1.04	1.07	1.04	1.08	1.15*	1.20*	1.25*	1.30*	1.35*	1.38*	1.53	1.34
nBM	0.00	0.06	0.00	0.08	0.23	0.33	0.43	0.53	0.63	0.70	1.00	0.61
Moisture content (%)	53.94	48.66	46.79	55.39	85.41*	87.19*	83.63*	77.39*	78.28*	72.05*	64.43	53.48
nMC	0.82	0.95	1.00	0.79	0.04	0.00	0.09	0.24	0.22	0.37	0.56	0.83

* Data imputation based on regression equations of biomass and tissue moisture content

Table 4 Eigenvalue and variance explained of principal component analysis

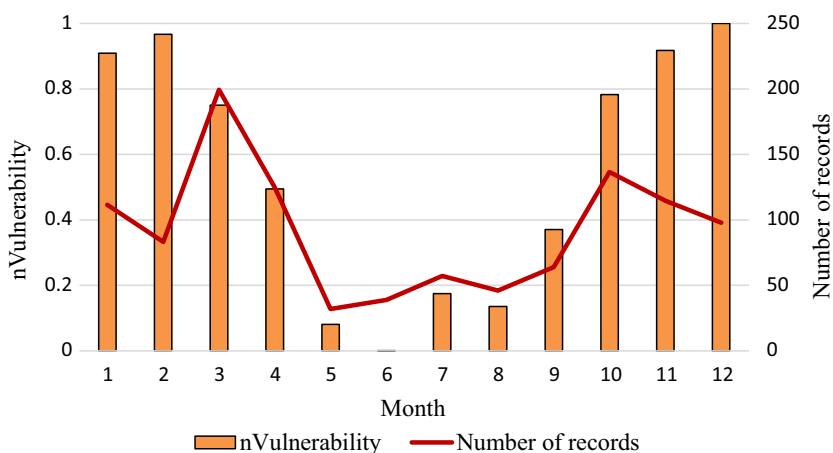
Component	Eigenvalue	Variance explained(%)	Cumulative variance explained(%)
1	4.39	73.12	73.12
2	1.15	19.17	92.29
3	0.32	5.24	97.52
4	0.13	2.09	99.62
5	0.01	0.24	99.85
6	0.01	0.15	100.00

Bold font indicates retained components

Table 5 Post-rotation factor matrix

Factor	Post-rotation factor matrix		Score factor matrix		Weight (%)
	Component 1	Component 2	Component 1	Component 2	
niNDWI	0.92		0.20		16.4
nNDVI	– 0.94		0.21		17.3
niMC	0.87		0.18		14.9
nBM		0.96		1	18.0
niEMC	0.94		0.21		17.1
nWS	0.91		0.20		16.2
Eigenvalue	4.20	0.92	–	–	–

Bold indicates that the absolute value of the post-rotation factor is greater than 0.5

**Fig. 4** Monthly changes in wildfire nVulnerability and fire case records

4.1.6 The initial threshold of *dNBR*

The initial threshold of *dNBR* can be obtained from the false color image of the burned sites by the adjustment of the grid *dNBR* value. The difference between the average values

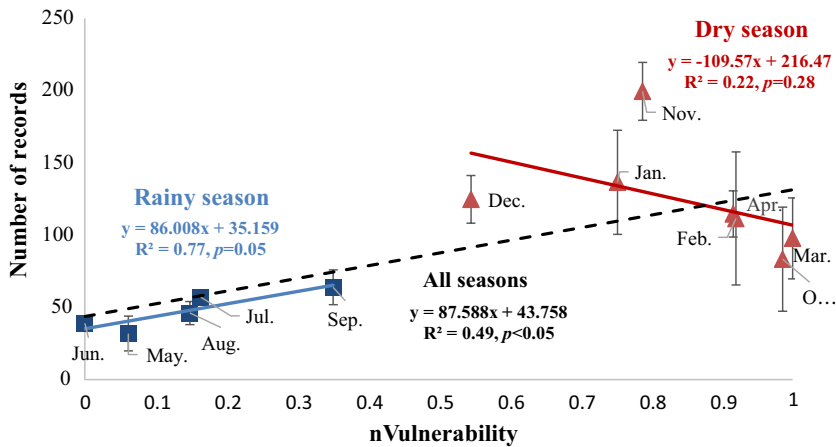


Fig. 5 The correlation between wildfire vulnerability and fire cases in each month from 2006 to 2015, with an average monthly explanation rate of 49%; if divided into dry and rainy seasons, the explanation rate for the dry season is only 22%

of the *NBR* of the two phases of remotely measured image samples (Overall *dNBR*) is used as the degree of influence of the atmospheric humidity change between the two phases of images. In each wildfire event, the initial threshold of *dNBR* obtained by adjusting the grid *dNBR* value was used as the y variable, and overall *dNBR* was used as the x variable for regression analysis. The correlation of the overall *dNBR* and initial threshold of *dNBR* shows a high correlation ($R^2 = 0.78$, $p < 0.001$) (Fig. 6). Based on the change in atmospheric humidity between the two periods, the initial threshold of *dNBR* can be obtained through the application of regression, which is beneficial to extract the burned area accurately.

The overall *dNBR* > 0 indicates that atmospheric humidity of pre-fire is wetter than that of post-fire, and the initial threshold increases as the *dNBR* is larger (the threshold

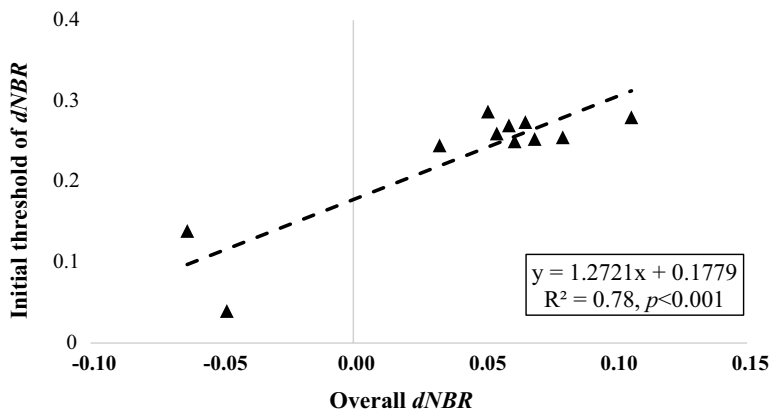


Fig. 6 The correlation between the overall *dNBR* and the initial threshold of *dNBR*. 11 sets of *dNBR* (22 telemetry images) are used to evaluate the influence of atmospheric humidity on the initial threshold of *dNBR*

converges to between 0.25 and 0.30). The overall $dNBR < 0$ indicates atmospheric humidity of pre-fire is drier than that of post-fire, and the initial threshold can be converged below 0.15.

The monthly wildfire area for the 2006–2015 dry seasons (October–April) is calculated based on the $dNBR$ threshold method described above, and the average wildfire sits per period is standardized as a monthly risk index. Satellite image extraction is not possible for month-to-month analysis because it is affected by satellite trajectories and clouds. However, we have a long analysis year and provide a sufficient number of samples. The total area of wildfires is extracted from the interval between the selected images, during which the average area of the monthly wildfires is used as the area of the wildfires that occur each month during the dry season. The area ratio is calculated (compared to the maximum value of the average combustion area) and standardize the monthly hazard estimates (Table 6 and 7) to derive the risk value of the monthly wildfire using the risk model (vulnerability \times hazard). The risk model is mainly used to correct the vulnerability of the dry season, while the level of risk during the rainy season is not considered, so the risk of the rainy season will not work.

After correcting the monthly risk value, the fire risk model is obtained, and the result shows that the overall correlation has increased (Fig. 7) to a very significant level, showing a positive correlation trend. Hazard has been added so that the temporal distribution of the wildfire risk model can be verified.

4.2 Spatial distribution of wildfire risk

4.2.1 Wildfire vulnerability

According to the temporal distribution of wildfire risk, the peak of wildfires is at the beginning and end of the dry season. As a result, recent spatial environmental vulnerability information (2013–2015) has been used to simulate the spatial distribution of future wildfire hot spots. The spatial risk model is built with data of early dry season of 2013–2015, and maps of wildfire risk are delineated for reference in wildfire prevention. Therefore, NDVI and iNDWI were extracted from satellite images taken over three years in different years at the beginning of the dry season and standardized as factors reflecting biomass and plant moisture content.

Table 6 Monthly hazards calculated from $dNBR$ in dry seasons

Month	Number of samples	Total area of burn site (m ²)	Average burn area (m ²)	Weight*	Normalized weight	Hazard**
10	5	4,978,050	995,610	0.88	0.67	1.67
11	5	3,659,250	731,850	0.65	0.00	1.00
12	5	3,901,350	780,270	0.69	0.12	1.12
1	3	2,212,350	737,450	0.66	0.01	1.01
2	3	1,496,700	748,350	0.67	0.04	1.04
3	2	1,984,500	992,250	0.88	0.66	1.66
4	1	1,125,000	1,125,000	1.00	1.00	2.00

*Area ratio, compared with the maximum value of average burn area

**Normalized weight plus 1 as the hazard to avoid subsequent risk analysis to dilute vulnerability

Table 7 Monthly risks calculated from hazard multiplying vulnerability

Month	1	2	3	4	5	6	7	8	9	10	11	12
nVulnerability	0.92	0.99	0.79	0.55	0.06	0.00	0.16	0.15	0.35	0.75	0.92	1.00
Hazard	1.01	1.04	1.66	2	1*	1*	1*	1*	1*	1.67	1	1.12
Risk	0.93	1.03	1.31	1.10	0.06	0.00	0.16	0.15	0.35	1.26	0.92	1.12
nRisk*	0.71	0.78	1.00	0.84	0.05	0.00	0.12	0.11	0.27	0.96	0.70	0.86
Number of report	111.50	83.30	199.44	124.8	31.89	39.00	57.20	46.10	63.90	136.50	114.60	97.78

*The minimum hazard of the dry season is used as hazard of the rainy season

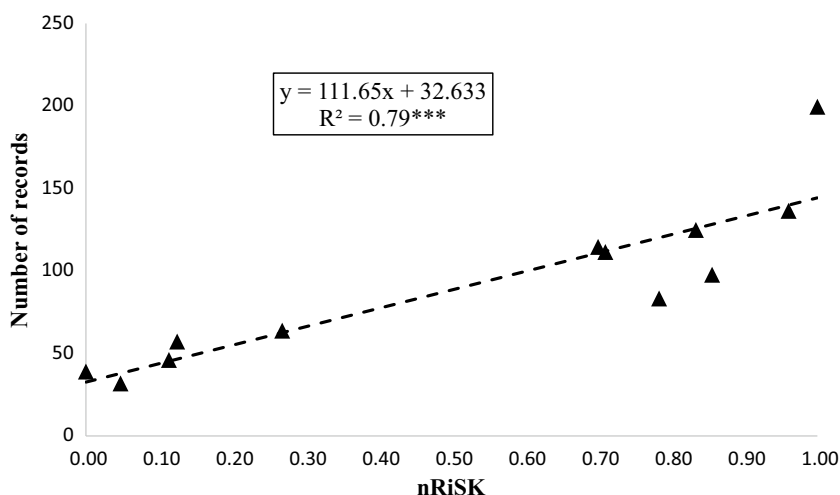


Fig. 7 Correlation of risk and number of records

Satellite image taken on October 16 and manual selection of training samples derived from Google Earth image of Nov. 23 (Fig. 8a), coupled with Maximum Likelihood Classification, were used to estimate land cover types (Fig. 8b) of the most wildfire vulnerability in 2013, and the classification of fuel property can be derived as Fig. 8c. In order to evaluate the accuracy of land cover classification, the results of classification and real land cover were evaluated. A total of 400 samples of land cover type were randomly to calculate accuracy analysis. The overall accuracy of image processing is 78%, and the Kappa coefficient is 0.69, indicating that the classification results are acceptable.

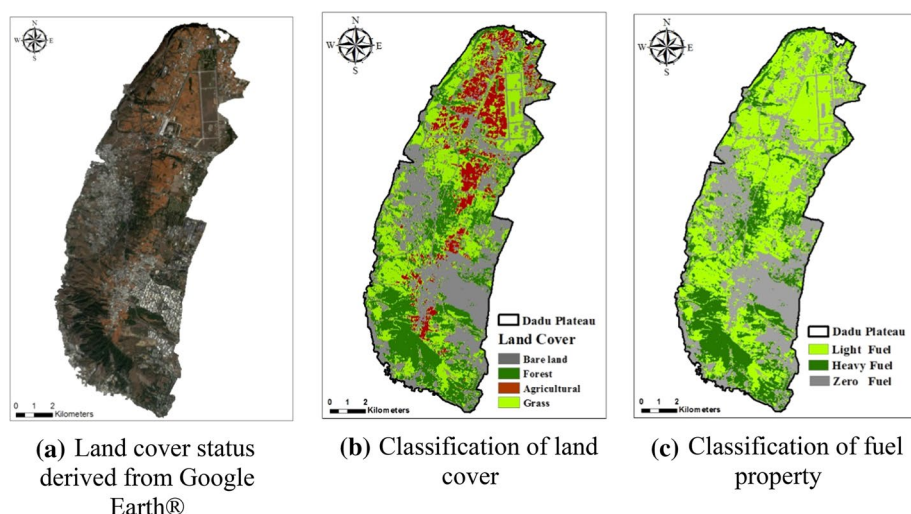


Fig. 8 Maps of Google Earth®, classification of land cover and fuel property

The spatial distribution of fuel properties is extracted from the land cover map. We add a standardized map of iNDWI, NDVI and fuel properties to get a spatial distribution of wildfire vulnerability (Fig. 9), which shows us that the higher the value representing the region's rich fuel, low moisture content and light fuel characteristics, the more likely it is for those areas to ignite when exposed to fire sources. In contrast, low-vulnerability areas are less likely to be ignited.

4.2.2 Wildfire hazard

Satellite images taken at the beginning of the high-risk period are used as pre-fire images, and post-fire images are taken at the end of the high-risk period. The spatial distribution of burned sites during this period was extracted using the initial threshold of *dNBR*. Figure 10 shows that the initial threshold of *dNBR* varies in different years. The results of the *dNBR* reflect low number of burned sites during 2014–2015 due to dry season with sufficient rainfall. Five *dNBR* maps were overlapped and standardized to obtain a wildfire hazard index. A higher hazard index indicates that the region is more likely to come into contact with fire sources and that wildfires occur more frequently, which also means that the area is more affected by external forces (fire sources).

4.2.3 Wildfire risk and model verification

The spatial distribution of wildfire risk during the dry season in the Dadu Plateau can be obtained through the product of hazard and vulnerability (Fig. 11a). The NDVI change matrix is used to determine the combustion areas of the dry season (2008–2009, 2010–2011, and 2013–2014) for spatial model validation (Fig. 11b). Using the village boundary map as a partition unit, the risk level of each village is calculated as an independent variable, and the wildfire frequency of the partition unit extracted from NDVI is used as a dependent variable. The results show that the risk of wildfires is linear positively correlated (the determination coefficient increased to 0.78, $p < 0.001$) with the frequency of wildfires (Fig. 12), which depicts that the spatial wildfire risk model established in this study has high practical value.

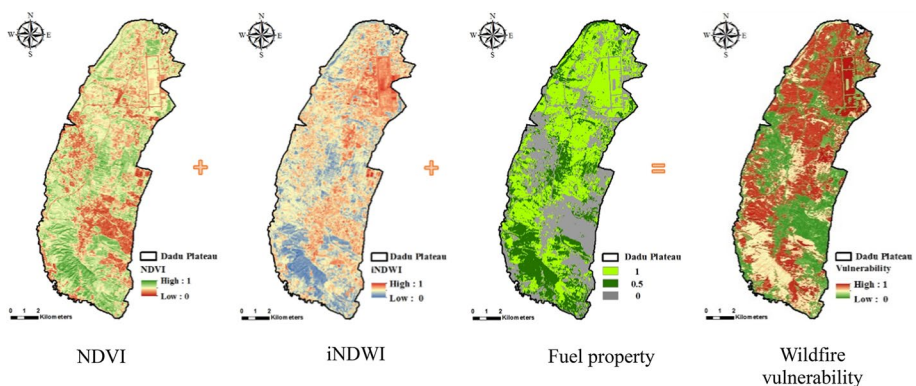


Fig. 9 Spatial distribution of wildfire vulnerability calculated by NDVI, iNDWI, and fuel property

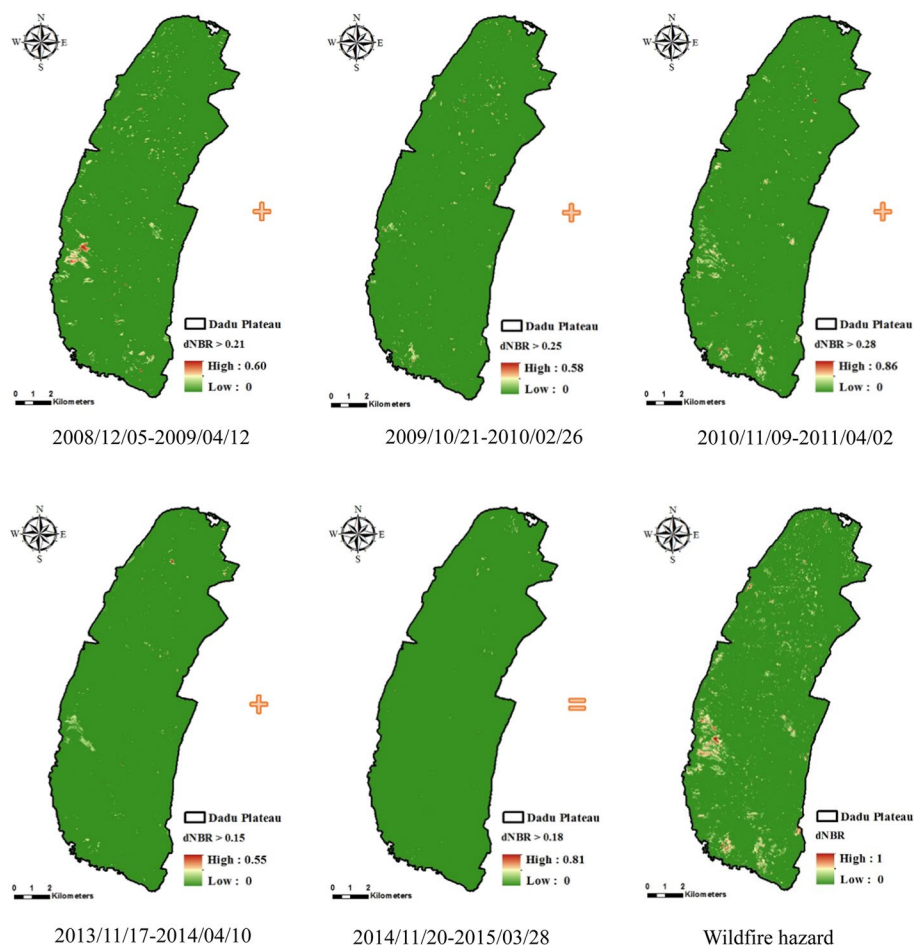


Fig. 10 Wildfire hazard index calculated by analyzing the overlay of the annual burning sites

5 Discussions

Disaster risk is defined as the probability of an event occurring under certain circumstances. However, many studies use hazard, danger, or environmental vulnerability as one or more indicators of wildfire risk (Oliveira et al. 2021a, b; Parente and Pereira 2016; Silva et al. 2020), which may cause confusion to follow-up researchers. The Canadian Forest Fire Risk Assessment System (CFFDRS) and the US National Fire Risk Assessment System (NFDRS) both determine the wildfire hazard levels based on fuel moisture and weather conditions and regulate artificial fire activities to reduce the risk of wildfires. This shows that the coupling relationship between the dry flammable environment and the fire source can be regarded as a wildfire risk. The study uses wildfire vulnerability to predict the hot spots of wildfire disasters. Through field sampling observation and statistical analysis, the Dadu Plateau grassland is rich in biomass throughout the year, and the moisture content of plant tissue changes with the seasons, which affects the occurrence of wildfires. Therefore, the environmental vulnerability analysis can only explain 49% of wildfires, and

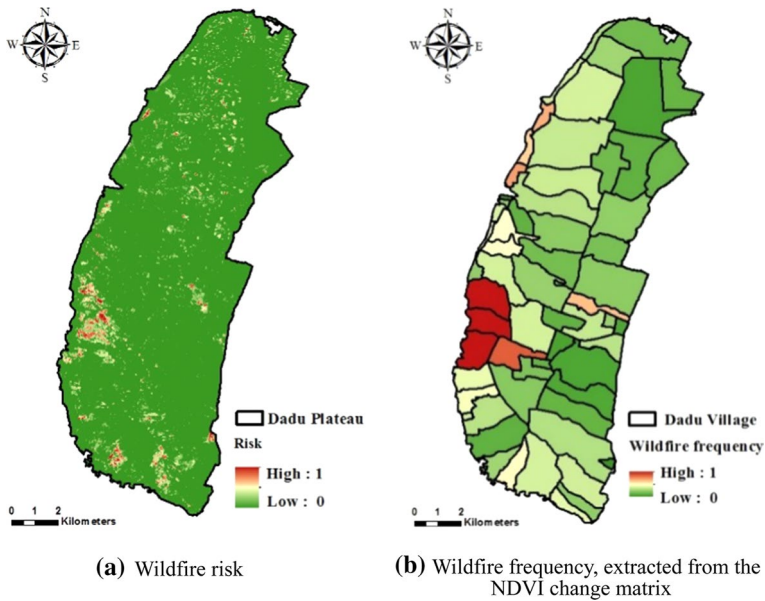


Fig. 11 Spatial distribution of wildfire risk and frequency

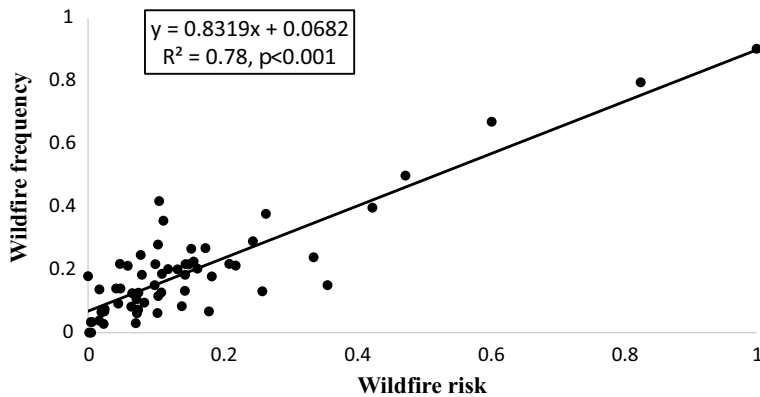


Fig. 12 The correlation of wildfire risk and wildfire frequency

the explanation rate in the dry season is even lower. Since wildfire risk is defined as the hazard and vulnerability of the environment to fire sources, adding hazard analysis helps to explain the factors that affect the occurrence of wildfires clearly and is of great significance to fire rescue.

Taiwan is located in the Asian monsoon climate zone. Summer is the rainy season, and the temperature is lower in autumn and winter in the dry season. This is different from the nature of wildfires that occur at the high temperature during the dry season in the western USA (Westerling et al. 2003, 2006; Williams et al. 2019). In the study area, moisture content is more important for wildfire frequency. The monthly vulnerability and fire frequency

indicate that drought and abundant fuel are suitable environments for wildfires from October to April. However, environmental vulnerability only accounts for 49% of monthly fire records. According to the statistics of forest fires in various countries, the vast majority of forest fires are caused by man-made fires (Soares and Batista, 2007), which is similar to the cause of wildfires on the Dadu Plateau (Chuang 2015). By analyzing the high-risk March (2013 to 2015) fire records (Fig. 4), the relationship between wildfires and land use in the study area was discussed (Fig. 13). The wildfires around the cemetery accounted for 68% of the total. The main reason is that March in Taiwan is the Qingming Festival for ancestors and tomb sweeping. Activities such as weeding, worshiping, and smoking in the cemetery inadvertently increase the chances of fire and fuel contact, leading to frequent wildfires. However, it is not the environment most prone to wildfires. This is mainly because human activities increase the risk of wildfires, and it has been proven that only considering environmental vulnerability does not reflect the actual situation of wildfires. This is similar to other Asian island countries such as Japan and Vietnam. The former is due to religious festivals (A Summery Festival of the Dead) (Zorn et al. 2001), while the latter is caused by behaviors such as smoking bees in order to collect honey from bees (Le et al. 2014).

Human factors are important for wildfire risk models, but the current Canadian Forest Fire Hazard Assessment System (CFFDRS) and the US National Fire Hazard Assessment System (NFDRS) do not explicitly include human factors (Silva et al. 2020). This may be related to the huge uncertainty and unpredictability of wildfires caused by man-made fires. Population density, distance from roads, location, and number of historical fires are often used as human factors (González et al. 2018; Silva et al. 2020). In countries where wildfires are frequent, historical wildfire events are often used to assess potential wildfire risks (González et al. 2018).

The fire site can be considered as a potential wildfire location (hazard factor) due to its high repeatability risk, and the vulnerability factor can be combined to assess the wildfire risk. Since historical report records cannot determine the location of the fire, this study extracts the *dNBR* of the telemetry image to describe the location and burning degree of the wildfire and conducts fire hazard analysis. However, the use of *dNBR* to extract wildfire sites is severely affected by atmospheric humidity. Compared with forests, herbs are more affected by seasonal humidity changes (Birch et al. 2015; Key and Benson 2005; Miller and Thode 2007). Therefore, the traditional *dNBR* extraction method needs to extract images from the same month every other year to reduce the interference of environmental factors. Using satellite images taken under different

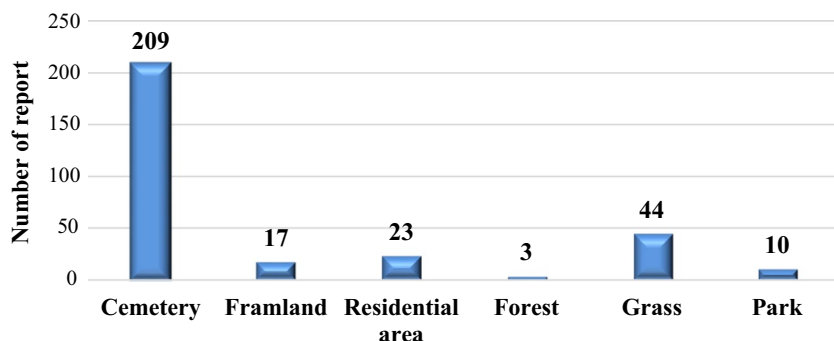


Fig. 13 Statistics of wildfire reports by different land use types from 2013 to March 2015

atmospheric humidity conditions, the relationship between the initial threshold of $dNBR$ (grid $dNBR$) and the overall $dNBR$ can be established to effectively extract wildfire locations. This research proposes an innovative method that uses the correlation between the overall $dNBR$ and the initial threshold of $dNBR$ to eliminate the defects of the traditional fixed threshold. This method can effectively extract wildfire relics in a short time, which is beneficial to the study of periodic wildfire areas.

As important metropolitan areas surround the study area, the long-term periodic burning of the Dadu Plateau has severely reduced the forest area. After a wildfire, guinea grass may grow quickly and hinder the succession of other plants (Fig. 14). Ecologically, migratory birds are affected by the decline in species richness and loss of forest ecological habitat (Tsai 2005). Large numbers of grassland are subject to periodic wildfire, which may reduce water storage capacity of the study area and easily lead to flooding in the lower urban area during heavy rains (Cho 2017).

In the case of periodic wildfires, fuel is not easy to accumulate, and wildfires are generally not too large and can be extinguished quickly; however, society must bear the damage of fires, and the local ecological and natural landscape system is gradually changing. Considering the spatial wildfire risk management, it is recommended to introduce the concept of fire belts in densely populated areas and use village boundaries as zoning units for wildfire control. The village boundary map is used as the zoning unit, the risk level of each village is calculated as a separate variable, and the wildfire frequency of the zoning unit extracted from NDVI is used as the dependent variable. The results show that the spatial wildfire risk model established in this study has high practical value. The risk value of wildfires in each village is divided into five levels (Fig. 15); it can be graded according to the risk value of wildfires in each village, which can provide reference and application for related departments in zoning management and disaster relief manpower layout.

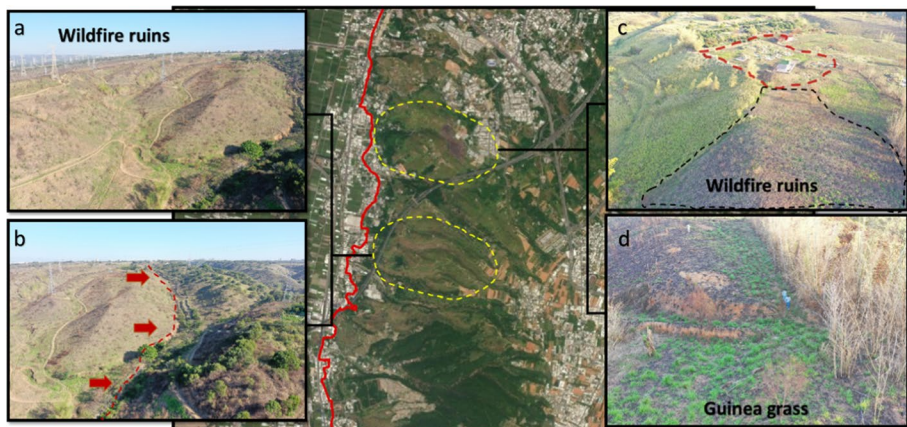
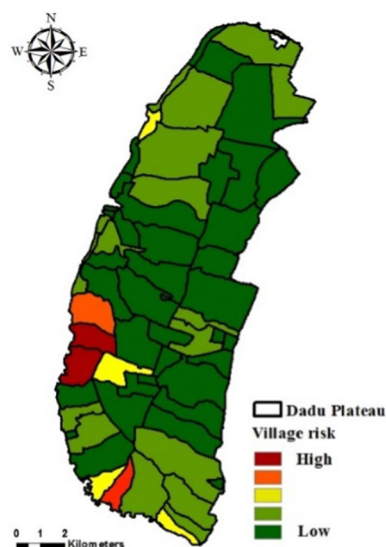


Fig. 14 Scene photographs of areas with high fire risk (a: mosaic landscape, b: obvious interlaced zone, c: delayed plant succession, d: wildfire-adapted vegetation germinates one week later)

Fig. 15 Risk levels of wildfires in each village



6 Conclusion

As fire behavior involves complex environmental issues, this study uses environmental indicators combined with risk analysis to explore the temporal and spatial distribution of fire hotspots. In areas with periodic wildfires, if only the vulnerability of wildfires is considered, the frequency of wildfires cannot be explained; this shows the importance of wildfire hazards. Therefore, in addition to environmental vulnerability, man-made fire hazards play an important role in the management of periodic fire areas. In the field of wildfire prevention and management, attention should be paid to the cooperative relationship between stakeholders. The establishment of village-level fire risk maps can enable stakeholders to improve their fire drill skills, guide them to form a consensus on wildfire risk and management, implement actual fire safety actions, and gradually improve and reduce the damage caused by periodic fires to the environmental landscape.

Acknowledgements This research was supported by Grants from the Ministry of Science and Technology of Taiwan R.O.C. (104-2313-B-005-030 and 105-2621-M-005-002-MY3).

Funding Chao-Yuan Lin Phd is funded by Ministry of Science and Technology, Taiwan (grant no: 104-2313-B-005-030, 105-2621-M-005-002-MY3).

Data availability Meteorological data are extracted from the annual weather report of the central weather bureau (<https://e-service.cwb.gov.tw/HistoryDataQuery/>). Satellite images are derived from the center of the US Geological Survey website (<http://earthexplorer.USGS.gov/>). Fire case provided by the Fire bureau (<https://www.fire.taichung.gov.tw/home.asp>).

Declarations

Conflicts of interest The authors declare no conflicts of interest.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons

licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by/4.0/>.

References

- Abatzoglou JT, Williams AP, Boschetti L et al (2018) Global patterns of interannual climate–fire relationships. *Glob Chang Biol* 24:5164–5175. <https://doi.org/10.1111/GCB.14405>
- Ascoli D, Russo L, Giannino F et al (2018) Firebreak and fuelbreak. *Encycl Wildfires Wildland-Urban Interface Fires*. https://doi.org/10.1007/978-3-319-51727-8_70-1
- Benda L, Dunne T (1997) Stochastic forcing of sediment supply to channel networks from landsliding and debris flow. *Water Resour Res* 33:2849–2863. <https://doi.org/10.1029/97WR02388>
- Birch DS, Morgan P, Kolden CA et al (2015) Vegetation, topography and daily weather influenced burn severity in central Idaho and western Montana forests. *Ecosphere*. <https://doi.org/10.1890/ES14-00213.1>
- Bowker MA, Belnap J, Rosentreter R, Graham B (2004) Wildfire-resistant biological soil crusts and fire-induced loss of soil stability in Palouse prairies, USA. *Appl Soil Ecol* 26:41–52. <https://doi.org/10.1016/J.APSOIL.2003.10.005>
- Bowman DMJS, Balch JK, Artaxo P et al (2009) Fire in the Earth system. *Science*. 324:481–484. <https://doi.org/10.1126/SCIENCE.1163886>
- Brown AA, Davis KP (1973) Forest fire: control and use, 2d edn. McGraw-Hill Book Company, New York
- Caselles V, López García MJ (1991) Mapping burns and natural reforestation using thematic mapper data. *Geocarto Int* 6:31–37. <https://doi.org/10.1080/10106049109354290>
- Chen A, Tang R, Mao J et al (2020) Spatiotemporal dynamics of ecosystem fires and biomass burning-induced carbon emissions in China over the past two decades. *Geogr Sustain* 1:47–58. <https://doi.org/10.1016/J.GEOSUS.2020.03.002>
- Chiu CR, Tseng JC, Huang WD, Yang CM (2004) Grey relational analysis of the effect of climate factors on the satellite remote sensing brightness index (BRI) of Guinea grass in Mt. Dadu Area Crop Environ Bioinforma 1:207–214. <https://doi.org/10.30061/CEB.200409.0006>
- Chiu CR, Hsueh IC, Liu YA, Lai YJ (2012) Trend analysis of land-use and land-cover changes in Dadu Hill, Taichung. *Urban Plan* 39:25–50. <https://doi.org/10.6128/CP.39.1.25>
- Cho CC (2017) Water conservation hotspots selection at the Dadu Plateau by using index of land cover changes. *J Soil Water Conserv* 49:1963–1978
- Chuang YL (2015) Application of fire risk analysis for wildfire management at Dadu terrace. National Chung Hsing University
- Costa JE (1987) A comparison of the largest rainfall-runoff floods in the United States with those of the People's Republic of China and the world. *J Hydrol* 96:101–115. [https://doi.org/10.1016/0022-1694\(87\)90146-6](https://doi.org/10.1016/0022-1694(87)90146-6)
- Cui X, Alam MA, Perry GL et al (2019) Green firebreaks as a management tool for wildfires: lessons from China. *J Environ Manage* 233:329–336. <https://doi.org/10.1016/J.JENVMAN.2018.12.043>
- Epting J, Verbyla D, Sorbel B (2005) Evaluation of remotely sensed indices for assessing burn severity in interior Alaska using Landsat TM and ETM+. *Remote Sens Environ* 96:328–339. <https://doi.org/10.1016/J.RSE.2005.03.002>
- Fang L, Yang J (2014) Atmospheric effects on the performance and threshold extrapolation of multi-temporal Landsat derived dNBR for burn severity assessment. *Int J Appl Earth Obs Geoinf* 33:10–20. <https://doi.org/10.1016/J.JAG.2014.04.017>
- Fernández P, Rodríguez A, Gutiérrez D et al (2019) Firebreaks as a barrier to movement: the case of a butterfly in a Mediterranean landscape. *J Insect Conserv* 235(23):843–856. <https://doi.org/10.1007/S10841-019-00175-5>
- Fiorucci P, Gaetani F (2007) Dynamic fire danger mapping from satellite imagery and meteorological forecast data. *Earth Interact* 11:1–17. <https://doi.org/10.1175/EI199.1>
- González C, Castillo M, García-Chevesich P, Barrios J (2018) Dempster-Shafer theory of evidence: a new approach to spatially model wildfire risk potential in central Chile. *Sci Total Environ* 613–614:1024–1030. <https://doi.org/10.1016/j.scitotenv.2017.09.105>

- Gudmundsson L, Rego FC, Rocha M, Seneviratne SI (2014) Predicting above normal wildfire activity in southern Europe as a function of meteorological drought. *Environ Res Lett* 9:084008. <https://doi.org/10.1088/1748-9326/9/8/084008>
- Hall RJ, Freeburn JT, DeGroot WJ et al (2008) Remote sensing of burn severity: experience from western Canada boreal fires. *Int J Wildl Fire* 17:476–489. <https://doi.org/10.1071/WF08013>
- Hasan MM, Burian SJ, Barber ME (2020) Determining the impacts of wildfires on peak flood flows in high mountain watersheds. *Int J Environ Impacts Manag Mitig Recover* 3:339–351. <https://doi.org/10.2495/EI-V3-N4-339-351>
- Holland SM (2019) principal components analysis (PCA). Department of Geology, University of Georgia, Athens, GA 30602–2501, 2455 Teller Road, Thousand Oaks, California 91320
- Hsiao CW (2003) The Study on Estimating temporal-spatial distribution of Forest Fire Danger Rating in Taiwan. National Taiwan University
- Intergovernmental Panel on Climate Change (2014) Fifth Assessment Report(AR5). Intergov Panel Clim Chang
- Jackson M, Roering JJ (2009) Post-fire geomorphic response in steep, forested landscapes: oregon coast range, USA. *Quat Sci Rev* 28:1131–1146. <https://doi.org/10.1016/j.quascirev.2008.05.003>
- Jolliffe IT, Cadima J (2016) Principal component analysis: a review and recent developments. *Philos Trans R Soc A Math Phys Eng Sci*. <https://doi.org/10.1098/RSTA.2015.0202>
- Jolly WM, Cochrane MA, Freeborn PH et al (2015) Climate-induced variations in global wildfire danger from 1979 to 2013. *Nat Commun*. <https://doi.org/10.1038/ncomms8537>
- Karamihalaki M, Stagakis S, Sykioti O et al (2016) Monitoring drought effects on mediterranean conifer forests using spot-vegetation NDVI and NDWI timeseries. *Eur Sp Agency, Spec Publ* 740:188–192
- Key CH, Benson NC (2005) Landscape assessment: Ground measure of severity; the Composite Burn Index, and remote sensing of severity, the Normalized Burn Index. General Technical Report RMRS. USDA Forest Service, Rocky Mountain Research Station, Ogden, UT, pp 1–51
- Lai YJ (2003) The study on estimating temporal-spatial distribution of solar irradiance in watershed. National Taiwan University
- Lasaponara R (2005) Inter-comparison of AVHRR-based fire susceptibility indicators for the Mediterranean ecosystems of southern Italy. *Int J Remote Sens* 26:853–870. <https://doi.org/10.1080/014316042000274131>
- Laschi A, Foderi C, Fabiano F, Neri F (2019) Forest road planning, construction and maintenance to improve forest fire fighting: a review. *Croat J for Eng* 40:207–219
- Le TH, Nguyen TNT, Lasko K et al (2014) Vegetation fires and air pollution in Vietnam. *Environ Pollut* 195:267–275. <https://doi.org/10.1016/j.envpol.2014.07.023>
- Lee CY (2013) Changes in soil nutrients, arbuscular mycorrhizal fungi and invasive plant growth under repeated fire at dadu tableland. National Taiwan University, Central Taiwan
- Li B (1990) Natural resources and environment of Ordos plateau in Inner Mongolia (In Chinese). Sciences Press, Beijing
- Lillesand T, Kiefer RW, Chipman JW (2000) Remote sensing and image interpretation. Wiley, New York
- Lin SH (2000) Character and application of soil-conserving grasses. *Weed Sci Bull* 21:51–58. [https://doi.org/10.6274/WSSROC-2000-021\(1\)-051](https://doi.org/10.6274/WSSROC-2000-021(1)-051)
- Lin CC, Chiu CR, Chen MY et al (2005) modeling fire danger prediction of Dadu mountain area. *Q J Chinese for* 38:83–94. <https://doi.org/10.30064/QJCF.200503.0006>
- Lin MD, Rau JY, Tseng HH et al (2008) Characterizing PAH emission concentrations in ambient air during a large-scale joss paper open-burning event. *J Hazard Mater* 156:223–229. <https://doi.org/10.1016/J.JHAZMAT.2007.12.015>
- Lin JK, Yang KC (2001) Wildfire on the Dadu Terrace. In: *Environ. Inf. Cent*. <https://e-info.org.tw/node/12089>. Accessed 14Sep2021
- Lin SH (2013) Vegetation engineering of slopland. Wu-Nan Book Inc.
- Miller JD, Thode AE (2007) Quantifying burn severity in a heterogeneous landscape with a relative version of the delta Normalized Burn Ratio (RdNBR). *Sci Total Environ* 109:80. <https://doi.org/10.1016/j.rse.2006.12.006>
- Moody JA, Shakesby RA, Robichaud PR et al (2013) Current research issues related to post-wildfire runoff and erosion processes. *Earth-Science Rev* 122:10–37. <https://doi.org/10.1016/j.earscirev.2013.03.004>
- Neary DG, Gottfried G (2002) Fires and floods: post-fire watershed responses. Millpress Science Publishers, Rotterdam, The Netherlands
- Oliveira S, Gonçalves A, Zêzere JL (2021a) Reassessing wildfire susceptibility and hazard for mainland Portugal. *Sci Total Environ* 762:143121. <https://doi.org/10.1016/J.SCITOTENV.2020.143121>

- Oliveira S, Rocha J, Sá A (2021b) Wildfire risk modeling. *Curr Opin Environ Sci Heal* 23:100274. <https://doi.org/10.1016/J.COESH.2021.100274>
- Parente J, Pereira MG (2016) Structural fire risk: the case of Portugal. *Sci Total Environ* 573:883–893. <https://doi.org/10.1016/j.scitotenv.2016.08.164>
- Pausas JG, Llovet J, Rodrigo A, Vallejo R (2008) Are wildfires a disaster in the Mediterranean basin? A review. *Int J Wildl Fire* 17:713–723. <https://doi.org/10.1071/WF07151>
- Pyne SJ, Andrews PL (1984) Introduction to wildland fire: fire management in the United States. Wiley, New York
- Ravi S, D'Odorico P, Wang L et al (2009) Post-fire resource redistribution in desert grasslands: a possible negative feedback on land degradation. *Ecosystems* 12:434–444. <https://doi.org/10.1007/s10021-009-9233-9>
- Rouse JW, Hass RH, Schel IJA, Deering DW (1973) Monitoring vegetation systems in the Great Plains with ERTS
- Sankey JB, Wallace CSA, Ravi S (2013) Phenology-based, remote sensing of post-burn disturbance windows in rangelands. *Ecol Indic* 30:35–44. <https://doi.org/10.1016/j.ecolind.2013.02.004>
- Shakesby RA, Doerr SH (2006) Wildfire as a hydrological and geomorphological agent. *Earth-Science Rev* 74:269–307. <https://doi.org/10.1016/j.earscirev.2005.10.006>
- Shyu TH, Lee YH, Tsou TP, Chiang MY (2003) Comp Responses Nine Weeds Ozone Fumigation 24:75–87
- Silva IDB, Valle ME, Barros LC, Meyer JFCA (2020) A wildfire warning system applied to the state of Acre in the Brazilian Amazon. *Appl Soft Comput J*. <https://doi.org/10.1016/J.ASOC.2020.106075>
- Soares RV, Batista AC (2007) Incêndios florestais controle, efeitos e uso do fogo. Universidade Federal do Paraná
- Soverel NO, Coops NC, Perrakis DDB et al (2011) The transferability of a dNBR-derived model to predict burn severity across 10 wildland fires in western Canada. *Int J Wildl Fire* 20:518–531. <https://doi.org/10.1071/WF10081>
- Tian X, Zhao F, Shu L, Wang M (2013) Distribution characteristics and the influence factors of forest fires in China. *For Ecol Manage* 310:460–467. <https://doi.org/10.1016/J.FORECO.2013.08.025>
- Tsai JH (2005) The relationship between the distribution of plant community and environment factor in the belt plot in Jhukeng-beikeng of Dadu Terrace, Taichung: a four-years study. Providence University
- Turco M, Rosa-Cánovas JJ, Bedia J et al (2018) (2018) Exacerbated fires in Mediterranean Europe due to anthropogenic warming projected with non-stationary climate-fire models. *Nat Commun* 9(1):1–9. <https://doi.org/10.1038/s41467-018-06358-z>
- Westerling AL, Gershunov A, Brown TJ et al (2003) Climate and wildfire in the western United States. *Bull Am Meteorol Soc* 84:595–604. <https://doi.org/10.1175/BAMS-84-5-595>
- Westerling AL, Hidalgo HG, Cayan DR, Swetnam TW (2006) Warming and earlier spring increase Western US forest wildfire activity. *Science* 313:940–943. <https://doi.org/10.1126/science.1128834>
- Whitman E, Parisien M, Thompson DK, Flannigan MD (2018) Topoedaphic and forest controls on post-fire vegetation assemblies are modified by fire history and burn severity in the northwestern canadian boreal forest. *For*, 9:151 9:151. <https://doi.org/10.3390/F9030151>
- Williams AP, Abatzoglou JT, Gershunov A et al (2019) Observed impacts of anthropogenic climate change on wildfire in California. *Earth's Futur* 7:892–910. <https://doi.org/10.1029/2019EF001210>
- Yamane T (1967) Statistics: an introductory analysis, 2nd edn. Harper and Row, New York
- Yen TM, Wu YW (2004) Intergovernmental panel on climate change. *Q J for Res* 26:47–60
- Ying L, Cheng H, Shen Z et al (2021) Relative humidity and agricultural activities dominate wildfire ignitions in Yunnan, Southwest China: Patterns, thresholds, and implications. *Agric for Meteorol* 307:108540. <https://doi.org/10.1016/J.AGRFORMET.2021.108540>
- Zorn T, Nakayama K, Hashimoto O (2001) Global forest fire assessment 1990–2000 - FRA WP 55