

PERSPECTIVE OPEN



"Reactive Mineral Sink" drives soil organic matter dynamics and stabilization

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Reactive primary and secondary minerals play a critical role in the transformation and stabilization of organic matter (OM) in soil, a critical aspect that has been largely overlooked in existing literature. In this regard, we propose a new model known as the "reactive mineral sink" (RMS) to illustrate three primary mechanisms through which these minerals drive the bioprocessing, transformation, transport and stabilization of OM in soil. Firstly, from a biological perspective, reactive minerals influence enzymatic and microbial OM processing through binding enzymatic proteins or influencing the structure of microbial communities. Secondly, from a chemical standpoint, these minerals have the capacity to adsorb OM and/or coprecipitate with it, leading to a more diverse distribution of OM in the soil. This distribution, in turn, triggers OM transformation through chemical catalysis and redox reactions. Thirdly, on a physical level, reactive minerals have a substantial impact on soil architecture, aggregate dynamics, porosity development, and hydrological processes. These physical changes then affect the transport, reprocessing and stabilization of OM. The RMS model provides a conceptual framework that underscores the fundamental role of reactive minerals in driving the dynamics of OM and carbon (C) sequestration in natural soil. Furthermore, it promotes the restoration of soil biogeochemical processes and ecological resilience. We advocate for the implementation of strategies based on the RMS model to enhance the sequestration of organic C in soils for the purposes of rejuvenating soil health and mitigating CO₂ emission.

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INTRODUCTION

Soil organic matter (OM) comprises a wide array of organic compounds within the soil that undergo progressive decomposition¹. It includes organic material from plant and animal sources, their decomposition products, as well as biomass and metabolic products of soil microorganisms. The formation and stabilization of OM are critical to establishing a resilient soil structure and ensuring optimal biogeochemical functionality2. In the context of mitigating global climate change, soil OM serves as the predominant carbon (C) reservoir within terrestrial ecosystem, and its long-term stability is governed by its physical and chemical properties^{1,3}. In the conventional view, soil OM consisted primarily of recalcitrant lignin-rich compounds derived from plant biomass¹. However, it later became evident that microbial-derived OM, formed within the soil profile itself, also contributes significantly to the stable pool of soil OM. To emphasize the role of soil microbes in the decomposition of OM and the creation of new forms of microbial-derived OM, the concept of the "microbial carbon pump" (MCP)⁴ was introduced. More recently, the protection of OM by soil minerals gained recognition as a crucial mechanism for the storage of OM over the long term⁵. Consequently, numerous studies have focused on the adsorption of OM compounds to a variety of soil minerals, ranging from ferric oxyhydroxides to clay minerals^{6,7}. The critical roles of minerals in OM persistence and accumulation in soils were further introduced in the "soil mineral carbon pump" model proposed by Xiao et al.8.

Nevertheless, critical knowledge gaps still need to be addressed concerning the types of minerals involved in stabilizing OM, as well as biological, chemical and physical roles of soil minerals in OM processing, transformation and stabilization. Indeed, as suggested by Kleber et al.9, the adsorption of OM onto mineral

surfaces merely represents the initial phase of organo-mineral interactions. Subsequently, a series of additional reactions take place, encompassing mineral-catalyzed OM transformations, the generation of reactive oxygen species (ROS) induced by minerals, and the oxidation of OM. Previous studies^{10–12} have highlighted mineral-mediated biological processes, such as mineral-enzymemediated OM transformation¹⁰. Reactive minerals in soil can also shape soil microbial communities and activities^{11,12}, thus influencing microbial OM processing. Furthermore, minerals act as cementing agents, crucial in binding and preserving soil aggregate structure, thereby governing the storage and stabilization of OM^{13,14}. From these, we propose a greater focus on the pivotal roles of reactive minerals in regulating OM dynamics, since minerals are enduring substrates within soil.

To facilitate a clear and comprehensive discussion, we propose a conceptual framework, namely the "reactive mineral sink" (RMS) to illustrate plausible pathways of soil reactive minerals in OM biogeochemistry. The RMS can be defined as a matrix composed of a variety of minerals characterized by a high density of surface charges and/or significant thermodynamic potential. These minerals encompass both less stable primary minerals (e.g., biotite, amphibole) and relatively stable secondary minerals, such as smectite, montmorillonite, vermiculite, kaolinite, as well as amorphous minerals like Fe(III)/Mn(IV)/Al(III) oxyhydroxides (as listed in Table 1).

We evaluate mineral stability using the "Gibbs free energy of formation of minerals". This parameter governs the phase equilibria of mineral systems and serves as the basis for their geochemical behavior in soil, including processes such as mineral weathering and transformation 15. Minerals with higher Gibbs free energy of formation, such as primary minerals, typically exhibit

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Mineral class	Key minerals	Characteristics
Phyllosilicates	2:1 layer type (present in Cambisols, Cryosols, Luvisols, Chernozems, Acrisol, Lixisol, Vertisols etc): Mica Examples: Biotite- K(Mg, Fe) ₃ AlSi ₃ O ₁₀ (F, OH) ₂ Muscovite- (KF) ₂ (Al ₂ O ₃) ₃ (SiO ₂) ₆ Illite- (K,H ₃ O)(Al,Mg,Fe) ₂ (Si,Al) ₄ O ₁₀ [(OH) ₂ ,(H ₂ O)]) Phlogopite- KMg ₃ (AlSi ₃ O ₁₀)(F,OH) ₂ Annite- KFe ₃ AlSi ₃ O ₁₀ (OH) ₂ Smectite Examples: Montmorillonite- (Na,Ca) _{0,33} (Al,Mg) ₂ (Si ₄ O ₁₀)(OH) ₂ ·nH ₂ O Nontronite- (CaO _{0,5} ,Na) _{0,3} Fe ³⁺ ₂ (Si,Al) ₄ O ₁₀ (OH) ₂ ·nH ₂ O) Vermiculite- (Mg,Fe ²⁺ ,Fe ³⁺) ₃ [(Al,Si) ₄ O ₁₀](OH) ₂ ·4H ₂ O; Chlorite- (Mg,Fe) ₃ (Si,Al) ₄ O ₁₀ (OH) ₂ ·(Mg,Fe) ₃ (OH) ₆	Negative surface charge, low to high surface area (1–800m ² g ⁻¹), high cation exchange capacity (10–200 cmol kg ⁻¹), swell, relatively low stability.
	1:1 layer type (present in Acrisols, Lixisols, Ferralsols, Nitisols, Vertisols etc): Examples: Kaolinite/Halloysite- $Al_2Si_2O_5(OH)_4$	Negative surface charge, moderate surface area (6–60 m² g⁻¹), low cation exchange capacity (0–10 cmol kg⁻¹), platy, relatively high stability.
	Modulated phyllosilicates (present in Andosols): Examples: Imogolite- Al ₂ SiO ₃ (OH) ₄ Allophane- Al ₂ O ₃ ·(SiO ₂) _{1.3-2} ·(2.5-3)H ₂ O	Positive surface charge, high surface area (600–900 m 2 g $^{-1}$), pH dependent cation exchange capacity.
Inosilicates	Amphibole minerals, Pyroxene minerals (Present in Podzols, Ferralsols, Andosols etc): Examples: Richterite-(Na,K) ₂ (Mg,Mn,Ca) ₆ Si ₈ O ₂₂ (OH) ₂ Hedenbergite-CaFeSi ₂ O ₆	Low surface area (< 1 $\mathrm{m^2~g^{\text{-1}}}$), low stability, susceptible to oxidation and weathering.
Metal oxides	Ferric oxide and oxyhydroxide (present in Podzols, Ferralsols, Acrisol, Lixisol, Nitisol, Andosols): Examples: Hematite-Fe $_2$ O $_3$ Goethite- α -FeO(OH) Magnetite-Fe 2 +Fe 3 + $_2$ O $_4$ Lepidocrocite- γ -FeO(OH) Maghemite- γ -Fe $_2$ O $_3$ Ferrihydrite- (Fe 3 + $_2$ O $_3$ *0.5H $_2$ O	Positive surface charge, moderate to high surface area (20–800 m ² g ⁻¹), pH dependent cation exchange capacity.
	Al oxide and oxyhydroxide (present in Podzols, Ferralsols, Acrisol, Lixisol, Nitisol, Andosols): Examples: Boehmite-γ-AlO(OH) Gibbsite-Al(OH) ₃ Amorphous Al oxides	Positive surface charge, moderate to high surface area (10–600 ${\rm m}^2~{\rm g}^{-1}$), pH dependent cation exchange capacity.
	Mn oxide (present in various soils): Examples: Birnessite-MnO ₂ ·nH ₂ O Todorokite-(Na,Ca,K,Ba,Sr) _{1-x} (Mn,Mq,Al) ₆ O ₁₂ ·3–4H ₂ O	Negative surface charge, high surface area (>1000 m 2 g $^{-1}$).

These mineral characteristics were adapted from Beckingham et al.¹⁸, Kleber et al.⁹, Singh et al.¹⁹ and Kögel-Knabner et al.²⁰. This summary highlights the mos abundant reactive minerals in various soils; specific soils may contain other unlisted minerals.

lower stability and are more prone to weathering, leading to the formation of secondary minerals^{16,17}.

The RMS is expected to systematically influence the character and stabilization of OM through multiple pathways. It may impact OM composition by regulating microbial OM processing, resynthesis, and chemical reactions, ultimately leading to the creation of persistent OM. Additionally, the RMS drives OM diversity and its stability within minerals and/or soil aggregates by influencing the physical transport of OM and interactions between OM and minerals. The functional significance of RMS is paramount in augmenting the storage of soil OM, bolstering biogeochemical processes, promoting soil health, and fostering the development of soil resilience and sustainable productivity.

RMS DRIVES OM DYNAMICS IN SOILS

Reactive minerals in soils play a central role within the soil ecosystem. They function as a dynamic hub, driving the bioprocessing, transformation, transport, and stabilization of OM via an interplay of biological, chemical, and physical processes, as illustrated in Fig. 1.

Biological processes

The RMS influences soil OM biodegradation and resynthesis by driving biological processes in the following ways:

(1) Reactive minerals exert their influence on extracellular enzyme, commonly referred to as exoenzymes²¹, by binding to enzyme proteins. This interaction subsequently affects the

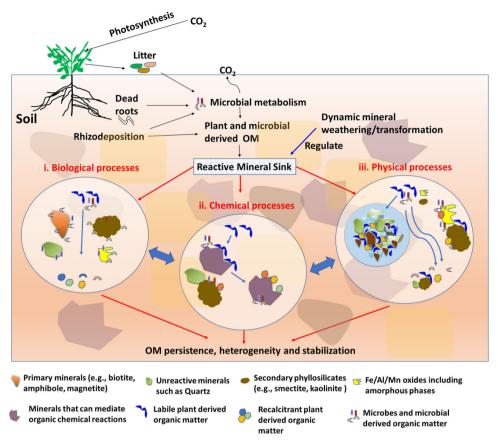


Fig. 1 Diagram showing the "Reactive Mineral Sink" (RMS)-driven OM processing, resynthesis, transformation, transport and stabilization in soils. (i) Biological processes: The reactive minerals can directly bind exoenzymes and influence enzymatic OM decomposition and transformation. The mineral matrix can also create micron-sized niches that promote different microbial communities that, in turn, can alter OM decomposition, reprocessing and resynthesis. (ii) Chemical processes: The reactive mineral matrix causes heterogeneous OM distribution and molecular transformation because different mineral surfaces have different chemical reactivities that will impact OM adsorption, chemical reactions and subsequent organo-mineral association. (iii) Physical processes: Mineral coagulation and dispersion influence aggregate formation and turnover, leading to OM protection and/or liberation. The mineral matrix can also influence hydrological processes that regulate the transport of OM, microbes, colloids and solutes through the matrix, resulting in OM relocation, reprocessing and stabilization.

extracellular pathway of OM decomposition and modification¹⁰, resulting in alterations to OM chemistry and its stability. Exoenzymes, being proteins rich in nitrogen-containing moieties, are well known for their strong tendency to bind with a variety of minerals²². When associated with minerals, functional groups within enzymes may undergo modifications^{10,22}, which can impact their catalytic capacity at transforming OM. Furthermore, certain reactive minerals, including Fe(II) minerals, can trigger the production of ROS²³, which in turn, influence exoenzyme activities and thereby regulate their role in OM transformation.

(2) Reactive minerals create microenvironments and pore spaces through compartmentalization, providing localized habitats with diverse water, nutrient, and oxygen conditions for microorganisms. The assemblage of minerals and the formation of aggregates contribute to the generation of these pore spaces. Pore spaces within aggregates and interstitial spaces between aggregates (typically measuring tens of micrometers in diameter) are essential microhabitats for soil microorganisms^{24,25}. This spatial heterogeneity can lead to diversity within soil microbial communities and exerts control over OM decomposition and reprocessing via microbial catabolism/anabolism, often referred to as the "microbial carbon pump"⁴. Pores within the size range of 30–90 µm have been identified as playing significant roles in the protection and/or mineralization of OM²⁵. The mineral phases,

composition, and characteristics directly impact porosity structure and pore sizes, which, in turn, regulates microbial OM processes.

(3) Reactive minerals have the capacity to alter microbial community structure by preferentially favouring the development of specific microbial communities in the "mineralosphere" 12. Minerals play a significant role in shaping microbial communities and their activities. They provide microbial habitats on mineral surfaces, essential nutrients, energy sources (e.g., during Fe redox reactions), and can also introduce potentially harmful elements like heavy metals and ROS^{11,23}. Changes in microbial community structure and functions have consequences for the pathways for microbial OM decomposition and regeneration⁴. For instance, microorganisms in the mineralosphere can utilize low-molecularweight organic acids and monosaccharides during mineral weathering²⁶, potentially leading to increased production and accumulation of microbial metabolites and/or necromass⁴. Different types of minerals have distinct effects on microbial OM processes and molecular composition. For example, a study demonstrated that kaolinite and montmorillonite had differing impacts on microbial communities, resulting in differences in the molecular composition of OM²⁷.

Reactive minerals can influence microbial activities and metabolisms, transforming OM. For instance, minerals rich in Fe(II) can stimulate the activities of Fe(II) oxidizing bacteria, especially heterotrophic/mixotrophic Fe(II) oxidizers such as



nitrate-reducing Fe(II) oxidizers, which utilize OM as an energy source²⁸. Fe(II)-rich reactive minerals can also induce the production of ROS, selectively inhibiting specific microorganisms by causing oxidative damage to lipids and proteins^{29,30}. In addition, minerals rich in Fe(III) can facilitate Fe(III)-reducing bacteria that metabolize by oxidizing simple organic compounds and discarding electrons onto the Fe(III)-bearing minerals, while producing dissolved Fe(II) and bicarbonate^{31,32}.

Soils in the early stages of pedogenesis, including Cambisols and post mine tailing soils³³, have experienced limited soil development and retain a substantial amount of less stable primary minerals. These minerals undergo weathering driven by both biological and non-biological factors, resulting in significant mineral changes and subsequent mineral-mediated bioprocessing of OM. Notably, when minerals undergo congruent dissolution, all constituent elements are dissolved³⁴. These elements released from the minerals can provide essential nutrients for soil microbes, thus facilitating microbial growth and the decomposition/reprocessing of OM³⁴.

Furthermore, the mineralogical composition can influence the dynamics of OM through the regulation of plant root activities. For instance, Liang et al. (2023)³⁵ demonstrated that minerals and roots interactively influence the stabilization and loss of OM. Under conditions where low-reactivity minerals (e.g., kaolinite) exist, roots enhanced OM mineralization and reduced the association of OM with minerals. Conversely, when highly reactive minerals (e.g., ferric oxyhydroxides) are present, root exudates promoted greater association of OM with minerals. Moreover, a recent study³⁶ revealed that minerals, such as kaolinite and ferrihydrite, resulted in significant differences in the organic molecular composition, particularly lipids, in rhizosphere soil.

Chemical processes

The RMS exerts its influence on the distribution patterns and molecular properties of OM through adsorption and mineralmediated chemical reactions. OM compounds can be adsorbed onto mineral surfaces through mechanisms such as ligand exchange, hydrogen bonding, cation bridging, and hydrophobic interactions³⁷. The heterogeneous distribution of OM arises from the presence of mixed minerals with varying affinities and capacities for the adsorption and stabilization of OM. Minerals that have a higher specific surface area and a greater density of surface charges can stabilize a greater number of OM compounds compared to minerals with lower surface area and reactivity. These properties lead to enrichment of OM in specific minerals within the soil^{9,37}. For example, it has been shown that OM compounds in soil aggregates are primarily found in a heterogeneous pattern, often associated with Al/Fe rich phyllosilicates rather than quartz at the submicrometer scale³⁸. In paddy soil, an increase in ferric oxyhydroxide content from 13.7 to 55.8 g kg^{-1} Fe resulted in an enhanced stabilization of OM, affirming the pivotal role of ferric oxyhydroxides in preserving soil OM. The impact of soil minerals on OM stabilization is contingent on the properties of the minerals and the characteristics of OM³⁹. In a recent study⁴⁰, it was found that hematite-rich tailings exhibited a greater capacity to stabilize OM during early pedogenesis in comparison to biotiterich tailings⁴⁰. Additionally, vermiculites, in contrast to kaolinite and illite, demonstrated a propensity for promoting the preservation of plant litter-derived carbohydrates through organo-mineral associations. This phenomenon is likely attributed to the higher specific surface area and stronger affinity for carbohydrates that vermiculites possess relative to kaolinite and illite⁴¹.

In addition to adsorbing OM, reactive minerals or their derived ions can also coprecipitate with organic molecules, giving rise to organo-mineral associations. The size of these associations can vary, spanning from nano- to micro-scales, which results from interactions between inorganic oligomers and organic

compounds. For example, soluble ions derived from biotite minerals, such as Fe, Si, and Al, have been observed to coprecipitate with organic ligands, forming loosely bounded organomineral associations⁴². These organo-mineral associations, including coprecipitates, play a vital role in protecting OM from mineralization in soils. For example, the coprecipitation of ferrihydrite with OM has been shown to effectively inhibit OM mineralization in anoxic soils^{43,44}. This protective effect is attributed to the capacity of organo-mineral coprecipitates to function as essential cementing agents or nuclei for formation of soil aggregates. This process acts as a shield, safeguarding OM from microbial mineralization⁴⁵.

Minerals can facilitate the transformation of OM through multiple processes, including electrolytic and/or hydrolytic breakdown of macromolecules, heterogeneous oxidation, and direct oxidation mediated by transition elements (e.g., Fe or Mn)⁹. Some minerals, such as Mn(IV) oxides, sulfides, and magnetite may induce hydrolytic breakdown of large macromolecular organic molecules into small molecules⁴⁶. Some nanosized mineral particles exhibit intrinsic enzyme-like activities akin to those derived from biological sources, effectively catalyzing the transformation of OM47. For instance, a recent study revealed that magnetite nanoparticles possess an intrinsic enzyme-like activity that is similar to that of natural peroxidases, facilitating the oxidation of OM47. These nanominerals with enzyme-like characteristics are called "nanozymes" ⁴⁸. In another study, it was observed that ferrihydrite and/or birnessite displayed efficient catalytic activity in promoting Maillard reactions. This catalysis led to the geopolymerization of small and simple molecules, resulting in the formation of complex and more recalcitrant compounds⁴⁹

Redox sensitive minerals, such as biotite, magnetite, Fe(II) sulfides, and Mn/Ti oxides, can efficiently catalyze the transformation of OM into form new molecules⁹. In floodplain soils, it was discovered that the primary driving force behind the oxidation and preservation of OM was reactive Fe(II), surpassing microbial activity⁵⁰. Ferrous iron-rich minerals have the ability to induce the production of ROS, leading to the oxidation of OM⁵¹. ROS, including hydroxyl radical (HO*) and hydrogen peroxide (H₂O₂), can enhance the mineralization and oxidative transformation of OM. This, in turn, leads to an increase in the biochemical stability of OM and the formation of organo-mineral complexes that can contribute to the long-term stabilization of OM^{9,52}. The interactions between OM and ROS can further lead to the oxidation of complex OM, resulting in the production of low molecular weight organic compounds characterized by an abundance of carboxyl groups. These compounds readily associate with minerals, contributing to organo-mineral associations^{52,53}. Additionally, certain fungi-mediated redox reactions have been observed to facilitate both the decomposition of OM and the oxidative preservation of mineral-stabilized OM on fungal surfaces⁵⁴.

ROS is generated through reactions between water and O_2 , primarily occurring at defect sites on minerals containing Fe(II). Furthermore, ROS production can occur as a result of photochemical electron transfer, especially involving minerals like Fe(II)bearing phyllosilicates (e.g., biotite, chlorite) and Fe(III) oxides (e.g., hematite)^{55,56}. It is well-documented that HO* can be created through Fenton oxidation, which involves a catalytic reaction chain incorporating H₂O₂, Fe(II), and ferric oxyhydroxides⁵². ROS stimulate OM mineralization and promote long-term C stabilization by enhancing the resilience of residual OM and encouraging the formation of organo-mineral complexes⁵². Studies^{33,57} have also found that mineral-associated OM was enriched in oxidized compounds, such as carboxyl groups, in soils with different mineralogical compositions. Moreover, electron transfer processes occurring during mineral-mineral interactions in soil can lead to the oxidation or reduction of OM compounds. For instance, electron transfer from Fe(II)-bearing phyllosilicates, such as nontronite, to ferric oxyhydroxides, like ferrihydrite, may induce

the reduction of OM compounds during their association with these minerals⁵⁸.

The weathering of poorly stable primary minerals in soils typically results in the development of surface-reactive secondary minerals and the accumulation of poorly crystalline minerals. These secondary minerals often possess high specific surface areas and variable charge characteristics, rendering them suitable for the binding and stabilization of OM³⁷. For example, our research⁵⁹ found that the mineral weathering processes in early soils enriched with biotite could generate secondary ferrihydrite and/ or Fe(III)-Si rich amorphous minerals with elevated affinity for OM. This leads to shifts in the chemical diversity and composition of dissolved OM⁵⁹. Additionally, mineral weathering processes result in the release of cations like Ca²⁺, Mq²⁺ and Fe³⁺ into the soil solution. These cations can then form complexes with organic compounds³⁷. Because of the complexity and dynamic changes in mineralogical composition and properties, as well as the molecular profiles of soil OM, it is crucial to acknowledge that the strength of organo-mineral associations is influenced by the combined effects of minerals within the matrix, rather than examining individual isolated mineral phases. This consideration should be a central focus in future studies.

Physical processes

The reactive mineral matrix in soil can encompass a variety of minerals that can aggregate together with OM to create soil aggregates. Within these aggregates, OM becomes occluded and thus safeguarded, resulting in low bioaccessibility¹³. Key factors in this process include secondary phyllosilicate minerals, such as smectite, and amorphous minerals, such as short-range ordered Fe/Si minerals, which function as mineral cements that maintain aggregate stability and protect OM in the forms of both occluded particulate OM and mineral-associated OM within the aggregates¹³. The assembly of minerals and the formation of soil aggregates can further produce pore spaces and provide localized niches both inside and outside the aggregates with varying environmental conditions (oxygen, water, nutrient) for microbial colonization²⁴ and microbial processing of OM. In addition to protecting OM in soil aggregates, certain minerals like clay, calcite, and metal oxides have been reported as having the capacity to encase OM, such as humic substances, sugars, and amino acids, within their developing crystalline formations⁶⁰. In a similar manner, authigenic minerals precipitated by bacteria can completely encase the cells, thus enhancing their preservation potential^{61,62}.

It is well-recognized that soil aggregates themselves cannot represent the entire soil structure underpinning its functionality. As proposed by Vogel et al. (2022)⁶³, soil architecture resulting from soil aggregation, which includes the spatial arrangement of pore networks as well as various minerals and organic components, is critical to soil biogeochemical functions. Properties of soil reactive minerals, such as surface charge, surface area, particle size, crystallinity and covalent reactivity, contribute to the soil architecture and hierarchical pore networks. These factors ultimately regulate hydrological processes⁶⁴, including the transport of OM, colloids and microbes together with water flow⁶⁵. Biological and chemical weathering processes could bring about changes to mineral composition that influence soil hydrological processes and the transport of OM. As a result, some OM can be vertically transported from topsoil into subsoil⁶⁶. These latter processes create feedback loops influencing the biological and chemical processes described above for the transformation and stabilization of OM under fluctuating environmental conditions. For example, in grassland soil (Eutric Fluvisol), there was a decrease in low-molecular weight compounds, while larger microbial-derived molecules increased in the subsoil compared to topsoil⁶⁶.

MULTI-PROCESS INTERACTIONS

RMS-mediated biological, chemical, and physical processes occur simultaneously in soil and collectively determine the composition. distribution, and stabilization of OM in soil. For instance, mineral assemblage and soil aggregate formation (physical processes) can both occlude OM and provide microporosity within aggregates. These, in turn, exhibit variable environmental conditions (i.e., redox, water, salts, nutrient etc) which give rise to the development of specific microbial communities⁶⁷ and associated microbial OM processing (biological processes). Mineral phases within aggregates can be shielded from interacting with OM compounds in the pore spaces between aggregates, thereby reducing OM transformation processes (chemical processes). Direct reactions between redox sensitive minerals and OM can also lead to alterations in mineral composition and surface properties⁶⁸, causing changes to mineral matrix-driven physical and biological processes. Furthermore, reactive minerals can influence pore spaces amongst aggregates of different sizes, leading to a steep gradient of oxygen and nutrient availability. This, in turn, influences the soil food web at micrometer scale and activities of microbial communities within aerobic and anaerobic niches, resulting in diverse microbial OM processing²⁴ (biological processes).

Some OM can be vertically transported from the topsoil into the subsoil via water flow (physical processes). Subsoil properties, such as mineralogical composition, oxygen availability, temperature, OM content, and microbial abundance and diversity, differ from those of the topsoil. These distinctions collectively influence microbial-mediated OM processes (biological processes) in the two soil layers⁶⁶. However, due to the presence of reducing conditions, the presence of transition elements in the subsoil, such as Fe(II), can actively mediate chemical reactions for OM oxidation⁵⁰ (chemical processes). This underscores the critical role of the mineral matrix in driving chemical changes in OM within the subsoil, highlighting the significance of these processes in soil OM dynamics.

IMPLICATIONS OF RMS DRIVEN SOIL OM DYNAMICS

The RMS model offers a comprehensive framework for understanding the dynamics of OM mediated by a changing reactive mineral matrix and suggests sustainable strategies for fostering soil ecological restoration, enhancing soil resilience, and facilitating carbon sequestration. The quantity and quality of soil OM underpin soil physical structure, its hydrological conditions, chemical stability, as well as various biogeochemical processes such as nutrient cycling, microbial community and functions⁶⁹. Consequently, the dynamics of OM driven by RMS are critical for preserving and restoring the physical structure and biogeochemical functionality of degraded soils resulting from poor land management and mining impacts^{33,70}.

It is important to emphasize that the dynamics of OM driven by RMS plays a crucial role in regulating and improving soil health. According to Lehmann et al. 71, soil health is defined as "the capacity of soil to function as a vital living ecosystem that sustains plants, animals and humans and connects agriculture and soil science to policy, stakeholder needs and sustainable supply-chain management." When conducting a comprehensive assessment of soil health, a range of physical, chemical, and biological indicators are used to evaluate the overall condition of the soil^{71,72}. Generally, physical indicators include soil aggregate stability, water content, and bulk density; chemical indicators involve factors such as pH, extractable N, P, K, and other essential elements such as Fe, Mg, Mn and Zn; while biological indicators include soil organic matter fractions, labile carbon content, soil protein content, and respiration. Therefore, a comprehensive soil health index should be constructed by considering the



significance and interplay of these physical, chemical, and biological indicators, all of which collectively contribute to the assessment and management of soil health⁷¹.

Reactive minerals and their interactions with OM can contribute to soil health through various pathways: (1) They can impact the forms and partitioning of OM via organo-mineral interactions, leading to changes in OM fractions (such as labile or persistent OM, mineral unprotected or mineral-associated OM). These changes serve as key indicators of soil health^{71,72}. (2) These minerals play a significant role in enhancing soil aggregate stability, which is an important indicator of soil health⁷². Minerals and/or organo-mineral association are essential components and cementing agents for soil aggregation. (3) They have the potential to influence the availability of essential nutrients like N, P, K, Ca, Mg and Fe, thereby regulating plant growth—one of the key ecosystem services considered in comprehensive assessments of soil health⁷¹. (4) Additionally, they participate in retaining pollutants, such as heavy metals and organic toxins, by stabilizing them with minerals. This retention helps reduce the leaching of pollutants from soil into water sources. This enhances water guality, another crucial ecosystem service assessed in soil health⁷¹

Furthermore, RMS provides a new framework for estimating and enhancing soil organic carbon (OC) sequestration in soils with different mineralogy. This framework holds significant in augmenting soil OM stocks as part of climate change mitigation efforts. The existing C models (e.g., RothC, Socrates, SOMM, NCSOIL, and DNDC) emphasize the biochemical persistence of OM compounds⁷³, often overlooking the intrinsic influence of soil mineral composition, mineral phase and energy dynamics in regulating OM chemistry and stabilization⁹. While some models, such as CENTURY and DAISY, consider the role of clay minerals in OM stabilization through adsorption processes, they do not account for the comprehensive range of processes mediated by reactive minerals, as described in the RMS model. For instance, in the RMS model, it is demonstrated that "labile" plant derived OM can transform into "persistent" OM after undergoing adsorption and mineral-driven transformations⁷⁴. Consequently, for accurate estimates of terrestrial soil C sequestration, modeling should consider the parameters of the reactive mineral matrix within a soil, including mineral composition, abundance, and phases, as well as surface reactivity characteristics such as charge and redox

RMS-BASED STRATEGY FOR ENHANCING SOIL OC SEQUESTRATION

The multiple functionalities of reactive minerals in soil OM dynamics give rise to their roles in mitigating and managing soil OC sequestration and stabilization. Here, soil OC refers specifically to the carbon component of soil OM. It is a critical component in understanding soil's role in carbon sequestration for climate change mitigation, as well as soil conservation practices. The knowledge of RMS driven OM dynamics provides a fundamental basis for the development of RMS based strategies towards increasing the capacity of OC sequestration and storage in terrestrial soil.

Soil OC fractions

Soil OC is sequestered through a combination of physical and chemical protection within particulate OM (POM) and mineral-associated OM (MAOM)⁷⁵. POM is primarilty composed of OM derived from plant litter and partially decomposed plant tissues, whereas MAOM is predominantly composed of OM originating from microbial biomass and their metabolites intimately bound to minerals²². MAOM, due to its mineral protection, is more persistent than POM⁷⁵. POM can be progressively decomposed by microbes, giving rise to a broader range of OM, including

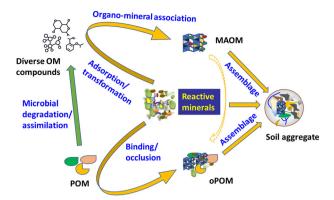


Fig. 2 Diagram showing the importance of reactive minerals in soil organic matter distribution and stabilization, as well as aggregate formation in soils. Note: POM: particulate organic matter; MAOM: mineral associated organic matter; oPOM: occluded POM within soil aggregates.

microbial metabolites and/or necromass. As soil aggregates form and undergo turnover, POM can be trapped, transforming into occluded POM (oPOM), thereby protecting OM and reducing its susceptibility to biological degradation. The interactions between reactive minerals and OM suggest that changes in reactive minerals and their dynamics play a pivotal role in regulating OC sequestration by facilitating the formation of MAOM and oPOM (Fig. 2). The reactive minerals can also bind organic compounds to form MAOM (chemical processes)⁷⁶, acting as a nucleus for this sequestration. Moreover, reactive minerals, along with the newly formed MAOM with active surfaces, can physically bind free POM to form oPOM within soil aggregates (physical processes)⁷⁷ (Fig. 2). Therefore, the composition and properties of reactive minerals should be given increased attention to gain a comprehensive understanding of the forms and status of OM stabilization in soil.

The RMS-based approach builds upon the most current insights and rationale regarding the various functions of reactive minerals in OM dynamics. It emphasizes coordination of both direct mineral-OM interactions and indirect interactions involving minerals, microorganisms, and plants, all within the context of relevant biological and abiological environmental factors (Fig. 3). The RMS strategy promotes the pivotal role of reactive minerals as an integrative and optimized solution for the long-term seguestration of OC in soil. The dynamics of reactive mineral transformation is closely linked to the soil's weathering history and age, influenced by a range of abiotic conditions (e.g., climate, initial primary mineralogy) and biotic factors (e.g., microbial and plant growth). Therefore, it is necessary to consider the soil's mineralogy in both young and mature soil systems when formulating strategies aimed at enhancing and mitigating the sequestration of OC in soils (Fig. 3).

Soil age and mineral dynamics

In young soils, like Cambisols, Cryosols and Podzols, early pedogenesis is in progress. These soils still contain abundant primary minerals that are actively undergoing weathering. Due to limited microbial and plant growth during the initial stages, these soils have relatively low levels of OM originating from biological sources (Fig. 3). Over the course of pedogenetic processes, the soil's mineralogical composition, geochemical properties, and physical structure undergo dynamic changes, closely intertwined with the succession of plant and microbial communities. In contrast, mature soils, such as Acrisols, Lixisols, Ferralsols and Nitisols, have a relatively stable mineralogy. These soils have experienced extensive weathering processes and major pedogenic phases. Before erosive loss, mature soils are expected to

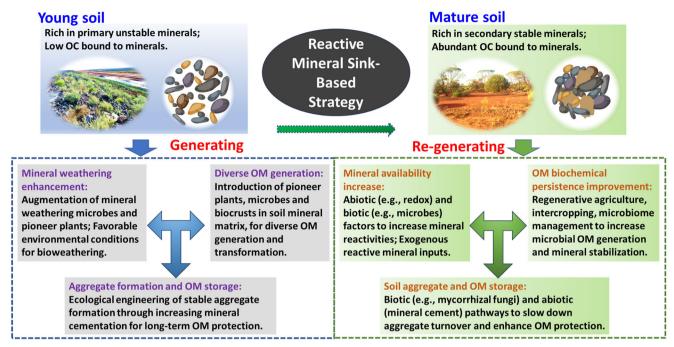


Fig. 3 Diagram showing the "Reactive Mineral Sink" (RMS)-based strategy for increasing soil organic carbon sequestration in young soil and mature soil.

contain an abundance of weathered primary minerals and secondary minerals. They exhibit relatively stable geochemical properties, physical structure, and host well-established plant and microbial communities. The RMS-based strategy is considered to be closely related to the primary mineral-dominated young soils and secondary mineral-dominated mature soils which drive the dynamics of soil OM.

Young soil

In young soils during early pedogenesis, an abundance of primary minerals still exists. These soils have only developed limited populations of biota, including microorganisms and lower plants, resulting in inherently low OC content. Therefore, enhancing the abundance and functionality of the biota is critical to initiating and accelerating pedogenetic processes and OC sequestration and storage. Biotas are powerful drivers in mineral weathering. Based on this premise, strategies aimed at restoring or rejuvenating the capacity of young soils to stabilize OM should take into account the interlinked processes illustrated in Fig. 3. These include: (1) Stimulating bio-weathering of primary minerals to generate secondary minerals with low crystallinity, thereby increasing the potential for mineral-OM reaction; (2) Promoting diverse plant and microbial successions to foster the generation of complex and long-lasting OM; and (3) Implementing eco-engineering approaches that facilitate the formation of aggregates containing reactive minerals where OM can be speciated and occluded or stored.

Enhancing mineral weathering. Enhancing mineral weathering in young soil or soils predominantly composed of primary minerals is considered as a fundamental requirement for increasing the soil's capacity to sequester OC. In these substrates, the less stable primary minerals can be readily weathered, giving rise to secondary minerals which are essential constituents of the reactive mineral pool. These secondary minerals exhibit significantly increased surface area, leading to the formation of MAOM and the stabilization of OC. As mineral weathering and secondary mineral formation progress, these minerals can also drive chemical

transformation of OM molecule profiles⁹. In a recent study conducted at the continental scale⁷⁸, a strong link was made between primary mineral weathering rate and the geographic distribution of poorly crystalline minerals. This study revealed that soil OC was effectively stabilized in regions abundant in poorly crystalline minerals⁷⁸, e.g., those mineral products resulting from the weathering of primary minerals. This work highlighted the importance of secondary minerals which are integral to the RMS in the sequestration and stabilization of soil OC. To stimulate mineral weathering and secondary mineral formation, and increase the reactivity of mineral substrates in the soil, microorganism proficient in mineral weathering²⁶ and pioneering, as well as resilient plants with robust root system⁷⁹, can be employed. These strategies lead to an increase in the quantity of OC adsorbed. transformed, and stabilized in MAOM. Ultimately, maintaining a certain rate of mineral weathering and transformation is critical to replenishing and maintaining the size and functionality of the RMS in the soil. This process can be enhanced by activating biota to sustain mineral weathering processes.

Diverse and persistent OM generation and inputs. In addition to mineral weathering, increasing inputs of a wide range of OM is critical to sustaining OC stabilization in young soils. This may be facilitated by promoting the colonization of native microbes and plants that are well-suited and adaptive to the physical and chemical conditions present in the young soil dominated with primary minerals. In harsh environments, biocrust, biofilm and mosses could establish a presence in young soils, contributing to an increase OM inputs⁸⁰. Native pioneer plant species help to mobilize and recycle mineral nutrients inherent to the primary mineral. Their biomass litters enhance the development and growth of the microbial communities and biocrusts, which, in turn, leads to increased microbial-derived OM in soil. This includes microbial biomass, amino acids, sugars, chitin, lipids, and N-rich compounds such as proteins⁸¹. Notably, proteins are known to have a higher affinity for minerals compared to OM with lower N content^{22,40}. It is worth highlighting that the interplay between root and microbial activities and the types of minerals present in



the soil significantly influence OM generation, stabilization and loss³⁵. Thus, it is imperative to systematically consider plantmicrobe-mineral interactions when striving to enhance OC sequestration in young soils.

Stable aggregate formation. The processes of mineral weathering and the sustainable input of OM are fundamental to the formation of soil aggregates. These aggregates occlude POM and MAOM, enabling long-term storage of OC. The generation of secondary minerals is a critical factor in the durability of these cemented aggregate, which contributes to the stabilization of OC⁴⁵. In young soils rich in poorly stable primary minerals, ecological engineering processes, such as the introduction of native plant species and the addition of plant litter as a source of OM, have been used to accelerate mineral weathering, foster organo-mineral associations, and promote aggregate formation^{33,45}. These strategies have led to an improvement in the stabilization of OC through both occluded POM and MAOM⁸². Eco-engineering strategies can be devised to enhance RMS, and consequently, boost the sequestration and storage of OC in developing young soils.

Mature soil

Many types of mature or aged soils have experienced prolonged weathering processes, resulting in a high degree of crystallinity and relatively low reactivity. This diminished reactivity is often evident through metrics like the ratio of amorphous to crystalline Fe-minerals⁸³. Regeneration of reactive minerals in mature soil represents an important strategy for restoring and rejuvenating their capacity to sequester OC. "Regeneration" in this context means to reactivate mineral transformation to enhance mineral reactivity which is required for mineral-microbial-organic interactions. This can be effectively achieved by adding reactive minerals, such as through eco-engineering practices, or through adding alluvium or colluvium, in addition to the involvement of soil biota and native plants (Fig. 3). Briefly, the strategies include: (1) Increasing mineral transformation and encouraging the neoformation of reactive minerals and/or the introduction of exogenous reactive mineral amendments to increase reaction sites for organo-mineral interactions through biological and abiological regulation. (2) Improving microbial assimilation of OM and increasing the biochemical persistence of OM by implementing efficient revegetation and microbiome management in soil. These actions should consider the existing mineral composition and properties. (3) Slowing down soil aggregate turnover and enhancing the occlusion and storage of POM and MAOM within aggregates using biological and abiological treatments. This strategy is particularly relevant for the restoration or rehabilitation of degraded soils, especially those where mineral reactivity and OC sequestration capacity have largely been diminished due to factors such as long-term intensive cropping, grazing, or mismanagement.

Activating mineral dynamics. Although minerals and MAOM in mature soil are relatively stable, they remain subject to dynamic change, such as dispersion, mineral transformation, and new organo-mineral association. These occur in response to seasonal wet and dry cycles as well as bioweathering processes driven by microbes and roots^{84–86}. For instance, during wet seasons, shifts in soil redox conditions may influence the transformation of Fe and Mn minerals⁸⁷, triggering alterations in the dynamics of OM chemistry and stabilization⁹. In mature soil that are enriched in Fe(III) minerals, heterotrophic bacteria such as *Geobacter* sp. stimulate the reduction of Fe(III), leading to the transformation of mineral phases and the initiation of new organo-mineral interactions⁸⁸. Furthermore, according to the conceptual theory of organo-mineral association proposed by Kleber et al. (2007)²², increasing amounts of OM can be stabilized in the outer layer of

the existing MAOM, forming MAOM clusters. These clusters consist of numerous MAOM components assembled through organo-organic and/or organo-mineral interactions⁸⁹. This phenomenon enhances the mature soil's capacity for sequestering OC⁹⁰.

In the context of soil quality management within agriculture systems, replenishing reactive minerals into structurally degraded soils may be an effective strategy to enhance organo-mineral interactions and promoting the stabilization of OC⁹¹. In fact, multiple studies have shown that the addition of clay minerals (e.g., smectite, kaolinite, illite) and/or oxyhydroxides like ferrihydrite and allophane increased the long-term sequestration of OC in soil¹⁹. Therefore, it is proposed that soil amendments enriched with such minerals should be used as soil remedies to enhance the sequestration of OC in mature soil, particularly those that have been degraded or highly weathered. These mineral fertilizers not only provide essential nutrients (such as P, S, K, Ca, Mg) for the development of plants and microorganisms but they also provide more reactive binding sites for developing organo-mineral associations, thereby facilitating the long-term sequestration of OC. In summary, even mature soils, such as forest soils, still have the potential to substantially increase the sequestration of OC by encouraging the regeneration of reactive minerals and/or by introducing exogenous reactive mineral amendments (Fig. 3).

Improving OM biochemical persistence. Apart from RMS abundance in mature soil, the biochemical persistence of OM is important to long-term OC stabilization. This persistence can be achieved by increasing the presence of native plants and the microbiome in soil. For instance, mycorrhizal fungi working together with fine roots have the potential to increase overall input of OM and its chemical complexity^{92,93}. Within agricultural soil systems, introducing natural conservation into cropping and pasture landscapes can improve bioturbation, OM diversity and enhance its stability⁹⁴. The adoption of regenerative agriculture practices could also increase OC accumulation by bolstering the enrichment of both POM and MAOM⁹⁵. Additionally, implementing crop rotations that include legumes and transitioning land use from cropping to pasture could increase OC sequestration⁷³.

Among biota-derived OM sources, those derived from microorganisms are considered to be more persistent and prone to stabilization⁴. As a result, enriching microbial-derived OM formation is one of the important strategies to increase OC stabilization and storage in mature soil. For instance, well-managed and longterm grassland practices have been shown to increase the presence of microbial necromass in MAOM fractions when compared to annual cropping systems⁹⁶. Likewise, multi-species intercropping has been found to increase microbial-derived OM accumulation in macroaggregates in contrast to mono-species cropping system⁹⁷. The generation of microbial-derived OM is closely linked to the concept of carbon use efficiency (CUE) within the soil. Management of the soil microbiome can be facilitated by the presence of a reactive mineral matrix^{27,98}, which can influence microbial communities and temporary microenvironmental variations like temperature, moisture, and oxygen. These factors, in turn, improve CUE, microbial OM production and the development of organo-microbial-mineral assemblages – factors that are crucial to long-term OC sequestration.

Enhancing OC storage in soil aggregates. In mature soils, it is important to protect and stabilize OC through enhancing soil aggregate stability and functional turnover⁷⁷. For instance, biomass such as mycorrhizal fungi^{99,100}, and/or mineral cements such as reactive minerals with high specific surface area⁴⁵, can slow down the turnover of aggregates. This, in turn, enhances the protection of OC by occluding POM and MAOM with the aggregates.

Overall, the RMS-based strategy emphasizes the dynamics of reactive minerals and mineral-microbe/plant-OM interactions.

These interactions play a pivotal role in influencing the biochemical persistence and stabilization of OM.

ADVANCED TECHNIQUES FOR UNRAVELING RMS-DRIVEN OM DYNAMICS

Understanding the behavior of soil OM and its relationships with minerals presents a challenge when using conventional analytical approaches, primarily because of the intricate and heterogeneous nature of soil OM pool¹. Previous research often relied on chemical extraction techniques to characterize humic substrates within soil, but these methods are limited in their ability to provide high-resolution and intrinsic insights into organo-mineral interactions¹⁰¹. Instead, advanced mass spectrometry (MS) and high resolution micro-spectroscopic analysis methods are required to examine the interfaces between microbes-minerals-OM. Such techniques will unravel the relationships among OM transformation and stabilization and mineral reactivity in soil OC storage and cycling^{101,102}.

Ultrahigh-resolution analytical techniques, such as Electrospray ionization ultrahigh-resolution Fourier transform cyclotron resonance mass spectrometry (ESI-FT-ICR-MS)¹⁰³ and Orbitrap-MS¹⁰⁴, are highly effective in deciphering mineral-mediated OM reactions by capturing the products of newly formed organic molecules at the molecular level. These methodologies provide valuable insights into OM chemodiversity and composition¹⁰³. High resolution MS can also reveal the fractionation of organic molecules resulting from the preferential adsorption of specific molecules by minerals¹⁰⁵. Using FT-ICR-MS, we have recently uncovered thousands of organic molecules and observed molecular shifts associated with the weathering of Fe-bearing minerals in ore tailings⁵⁹. However, it is important to note that high-resolution MS techniques are not suitable for characterizing the forms of OM interfacing with minerals at the micro-and nanoscale.

To complement the FT-ICR-MS method, advanced techniques such as synchrotron-based scanning transmission X-ray microscopy (STXM) coupled to near edge X-ray absorption fine structure spectroscopy (NEXAFS) and nanoscale secondary ion mass spectrometry (NanoSIMS) can be used to resolve OC forms and their relationship with minerals at micro- and nano-scale 106. A range of imaging techniques, including high resolution microcomputed tomography (CT)¹⁰⁷, synchrotron-based X-ray fluorescence microscopy (XFM)^{45,108}, and micro-Fourier transform infrared spectrometry (FTIR) mapping¹⁰⁹, can be utilized for *in situ* characterization of soil architecture^{40,110}. These methods enable the study of organic and mineral distribution in relation to the pore network, providing valuable insights into soil structure and functions. Information obtained through such in situ high resolution micro-spectroscopic analyses offers direct evidence of the spatial distribution and chemical properties of reactive minerals and their association with OM. This aids in the characterization of soil architecture and fine-scale structure. Furthermore, such spatial imaging data, collected over a period, can track the dynamic movement and transformation of minerals and OM in response to various treatments in the soil. This is especially critical for gaining an in-depth understanding of OM dynamics and soil structure development driven by reactive minerals. However, it is recognized that micro-spectroscopic methods can only capture signals of OC distribution and/or overall OC forms. They do not provide information about the quantity and molecular composition of OM profile. Therefore, a combination of a suite of high-resolution MS and microspectroscopic methodologies, together with macroanalyses (bulk, physical, chemical fractionation)¹¹¹, is in our opinion the best way forward for unraveling the processes and mechanisms of RMSdriven OM dynamics and stabilization in soil.

CONCLUDING REMARKS

The RMS model advocates the previously overlooked yet crucial role of reactive minerals in governing the dynamics, longevity, and stabilization of OM in soil. This novel perspective emphasizes the significance of RMS in preserving and revitalizing soil OM storage and biogeochemical functionality through a combination of biological, physical, and chemical processes. The RMS-based strategy takes into consideration the interplay between reactive minerals and OM derived from plants and microorganisms. This holistic approach provides a new pathway for further research on enhancing the biochemical persistence and stabilization of OM, distinguishing itself from other approaches that have primarily emphasized either MAOM or POM¹¹².

In the RMS-based strategy, we advocate for a coordinated approach that considers various factors related to management and treatment, including mineral availability, OM inputs and transformations, and the stability of the soil aggregates. Moreover, RMS highlights the critical importance of soil mineralogy in estimating soil OC sequestration and storage as a means of mitigating climate change – an area that has largely been overlooked in conventional C modeling⁷³. Further studies by soil OC researchers will be essential to enrich the much-needed data and validate the RMS model. A world-wide initiative involving sampling and investigation can compile a comprehensive database on OM dynamics across diverse soil types, particularly in relation to the abundance and composition of reactive minerals.

Given the intricate spatial and functional complexity of soil mineral matrix and its ever-changing nature due to mineral weathering, it is imperative to employ a suite of cutting-edge methodologies, including high-resolution mass spectrometry and various micro-spectroscopic techniques, to unravel the dynamics and stabilization mechanisms driven by RMS under a variety of environmental conditions. The insights gained from the RMS model form the fundamental basis for advancing soil health, promoting sustainable food and fiber production, and reinforcing the resilience of terrestrial ecosystems.

DATA AVAILABILITY

All the information mentioned in this manuscript is available in the main text.

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AUTHOR CONTRIBUTIONS

S.W.: Conceptualization, Methodology, Discussion, Writing-original draft, review & editing. K.O.K.: Discussion, Writing-review & editing. B.C.: Discussion, Writing-review & editing. L.H.: Conceptualization, Discussion, Funding acquisition, Writing-review & editing.

COMPETING INTERESTS

The authors declare no competing interests.

ADDITIONAL INFORMATION

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