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Title: Competition-mediated feedbacks in experimental multi-species epizootics

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Abstract

Competition structures ecological communities and alters host-pathogen interactions. In environmentally-transmitted pathogens, an infection-resistant competitor may influence infection dynamics in a susceptible species through the negative impacts of competition (e.g., by reducing host density or causing nutritional stress that increases susceptibility to infection) and/or the positive impacts of reducing transmission efficiency (e.g., by removing environmental pathogen stages). Thus, a non-susceptible competi-

8 tor may enhance, reduce, or have no net effect on susceptible host density and infection
9 prevalence. Here, we couple an epidemiological model with experimental epidemics to
10 test how resource competition with a non-susceptible competitor (*Daphnia pulicaria*)
11 influences fungal microparasite (*Metschnikowia bicuspidata*) infection dynamics in a
12 susceptible host species (*D. dentifera*). Our model and experiments suggest that com-
13 petitor density can mediate the direction and magnitude of the effect of competition
14 on infection dynamics, with a peak in infection prevalence occurring at intermediate
15 competitor densities. At low densities, the non-susceptible competitor *D. pulicaria* may
16 reduce infection prevalence in the susceptible host by removing fungal spores from the
17 environment through feeding. However, when competitor density is increased and re-
18 sources become limiting, *D. pulicaria* negatively impacts the susceptible host by in-
19 creasing susceptible host feeding rates, and therefore fungal spore intake, and further
20 by reducing susceptible host population size as it is driven towards competitive ex-
21 clusion. In conclusion, these results show that a tradeoff between the competitor as a
22 consumer of pathogen, which serves to reduce epidemic size, and as a modifier of sus-
23 ceptible host foraging ecology, which influences infection rates, may alternately enhance

24 or dampen the magnitude of local epidemics.

25 **Keywords**

26 Dilution effect, Diversity-disease, Competition, Environmentally transmitted pathogen,

27 Multi-host pathogen, Resource limitation

28 **Introduction**

29 Competition is a major structuring force of ecological communities, and the strength

30 of interactions between competitors can determine whether coexistence or competitive

31 exclusion occurs (Armstrong and McGehee, 1976; Wang et al., 2009). Further, environ-

32 mental conditions can influence the outcome of competition, as competitive outcomes

33 can be shaped by predators (Wollrab et al., 2013), temperature (Carmona-Catot et al.,

34 2013), resource availability (Riebesell, 1974), or natural enemies (Decaestecker et al.,

35 2015). Studies of competition typically focus on the long-term or equilibrium dynam-

36 ics of competing species (Wedin and Tilman, 1993), but many interesting competitive

37 interactions occur when populations are not at equilibrium, or are in the transient pe-

38 rioid before competitive coexistence or exclusion (Pickett, 1980). Epidemic pathogens
 39 or parasites are a prime example of this, as they disrupt host population dynamics, and
 40 can influence transient population dynamics and competitive outcomes. Pathogens may
 41 mediate the interactions between competing species by disproportionately affecting one
 42 of the competitors (Price et al., 1988), or by altering aspects of host life history such as
 43 development time or dispersal (Thomas et al., 2000). While the influence of parasites
 44 on competitive interactions has received ample attention (Park, 1948; Price et al., 1988;
 45 Preston and Johnson, 2010; Hatcher et al., 2012), there have been few studies into the
 46 role of competitors on parasite populations (Hall et al., 2009).
 47 Competitors may differ in competence (Kilpatrick et al., 2006) and susceptibility to
 48 pathogen infection (Hall et al., 2007), such that the addition of a competitor can re-
 49 duce, enhance, or have no net effect on infection dynamics in another competing host.
 50 For instance, a competitor that is a more competent host species could increase para-
 51 site population size, which would potentially elevate infection rates of the other com-
 52 peting host. However, if the competitor is a less competent host, or if the effect of com-
 53 petition reduces potential pathogen transmission events to the other competitor, infec-

tion risk may be reduced. The effect of competition therefore may offer a mechanistic explanation for some cases of the relationship between host diversity and disease risk (the “diversity-disease relationship”).

The diversity-disease relationship proposes that increases in host diversity may reduce (i.e. a dilution effect) or enhance (i.e. an amplification effect) infection risk in a focal host species (Ostfeld and Keesing, 2012; Orlofske et al., 2012). In theory, dilution effects may arise for many different reasons, but generally, the non-focal species are considered to be subject to “wasted” transmission events, so that pathogen fitness is reduced by infecting a less suitable host (Keesing et al., 2006). Studies of the dilution effect are typically phenomenological (Salkeld et al., 2013), and do not incorporate ecological interactions among species in the community. The inclusion of ecological interactions (e.g., competition) into studies of multi-host pathogen dynamics may inform a general theory for when we expect host diversity to reduce or enhance disease. Here, we use a combination of modeling and experiments to provide a link between diversity-disease relationships and parasite-mediated competition. Specifically, we investigate the impact of the addition of a non-susceptible superior competitor that consumes en-

70 vironmental pathogen on the infection and population dynamics of a susceptible host

71 species.

72 To do this, we use a model system comprised of two sympatric zooplankton competi-

73 tors, *Daphnia dentifera* and *Daphnia pulicaria*. These two species have been found to

74 co-occur in the north temperate lakes of the United States (Duffy et al., 2010). *Daph-*

75 *nia dentifera* is susceptible to infection by an environmentally-transmitted yeast pathogen

76 (*Metschnikowia bicuspidata*), and is also an inferior competitor to *D. pulicaria*, as *D.*

77 *pulicaria* has larger body size, reproductive rate (*unpublished data*), and foraging rate

78 (Gliwicz, 2004). Further, *D. pulicaria* has been found to outcompete *D. lumholtzi*, a

79 formidable invasive competitor (Engel and Tollrian, 2011). While *D. pulicaria* does not

80 become infected, it does consume pathogen spores during foraging, potentially reduc-

81 ing pathogen transmission to susceptible hosts (so-called “friendly competition” ; Hall

82 et al. (2009)). Reduced resources can nutritionally stress susceptible hosts, which can

83 result in enhanced pathogen transmission (Pulkkinen and Ebert, 2004) as a result of

84 increased filtering rate (Hall et al., 2007), providing a mechanistic link between host

85 foraging ecology and pathogen transmission. Therefore, the impact of competition on

86 infection dynamics will depend on the tradeoff between the role of the competitor as a
 87 consumer-of-pathogen and as a consumer-of-resources.

88 Previously, Hall et al. (2009) performed an experiment in which *D. pulicaria* were al-
 89 lowed to graze on pathogen spores, and then this media was exposed to susceptible
 90 *D. dentifera* to see if *D. pulicaria* grazing could reduce transmission by depleting en-
 91 vironmental pathogen spores. However, this study did not account for the role of the
 92 competitor (*D. pulicaria*) as a consumer. By reducing algal resources, the competitor
 93 may indirectly influence susceptible host foraging rate, which is intrinsically linked to
 94 pathogen transmission in this system. A theory for this complex of interactions was
 95 recently developed by Cáceres et al. (2014), who examined the equilibrium outcomes
 96 of competition between a susceptible and a non-susceptible competitor, finding com-
 97 petitive exclusion of the susceptible host species over long time scales when the non-
 98 susceptible species is a superior competitor. While previous studies of parasite-mediated
 99 competition have focused on directly transmitted pathogens, and superior competi-
 100 tors that are also susceptible to parasitism (Price et al., 1986, 1988), we focus on an
 101 environmentally-transmitted pathogen, and the interaction between a dominant com-

petitor that does not become infected and an inferior susceptible competitor. Theory predicts competitive exclusion of the inferior competitor in the long-term, although these species coexist in natural systems, most likely through niche partitioning or complex community interactions. We focus on the transient period where both species coexist, and examine infection dynamics as a result of competition for a limiting resource altering exposure to an environmentally transmitted pathogen. These transient dynamics are important, and ecologically relevant, given that the seasonal fluctuations in both zooplankton population sizes and infection dynamics may preclude zooplankton populations from achieving equilibrium dynamics (Hutchinson, 1961; Scheffer et al., 2003). Using a modified version of Cáceres et al. (2014) model that more closely matches our experimental system (see Supplementary Materials for a comparison of our model to Cáceres et al. (2014)), we extend this theory by examining the influence of competitor density on epidemic and population dynamics under non-equilibrium conditions. By examining the transient dynamics of our theoretical model, we generate several testable model predictions, and experimentally test these predictions using the same zooplankton-pathogen system examined by Hall et al. (2007). First, we predict that

the extent to which the competitor enhances or reduces infection prevalence in the susceptible host will depend on the initial density of the competitor and the availability of algal resources. When resources are limiting, we predict that the competitor will enhance infection prevalence in the susceptible host species by enhancing susceptible host foraging rate, and subsequent pathogen transmission. Second, we predict that susceptible host population size will decrease as a function of competitor density, since the susceptible host will be excluded more rapidly at higher densities of the superior competitor. This reduction in susceptible host population size may reduce infection prevalence if contact with pathogen (and therefore transmission) is reduced, or increase infection prevalence if susceptible host filtering rate is increased as a response to reduced resources. To test these hypotheses, we parameterized our epidemiological model, and compared model outputs with experimental epidemics. Experimental epidemics were initiated at three competitor densities, as increased competitor density serves to reduce resource availability through exploitative competition. We found that competition rarely benefited the susceptible host species, either enhancing infection prevalence at intermediate competitor densities, or competitively excluding the susceptible host at

high competitor densities. This work highlights the importance of competitive interactions in evaluating the direction of diversity-disease relationships.

Methods

Study system Clonal lines of two sympatric freshwater cladocerans were used in this study, *D. dentifera* (provided by M. Duffy) and *D. pulicaria* (originally isolated from Oneida Lake, New York, and provided by N. Hairston Jr.). *Metschnikowia bicuspidata* is a fungal pathogen that infects *D. dentifera*, but not *D. pulicaria*. Pathogen transmission can occur when the host ingests the pathogen, allowing the pathogen to pierce the gut wall and grow within the host. Parasite-induced mortality causes the release of a multitude of infectious spores (see Table 1), which are then filtered and ingested by other hosts. Recent studies have found essentially no genetic variation in the fungal pathogen, reducing the likelihood of genotype \times genotype interactions (Searle et al., 2015). However, genetic variation within natural *D. dentifera* populations could influence pathogen transmission dynamics through heterogeneity in resistance, or spore production per host (Auld et al., 2013; Carius et al., 2001). We acknowledge this as

an interesting avenue for further research. However, our focus is to elucidate patterns in infection dynamics due to competitor density. Therefore, we selected a single, well-studied *D. dentifera* clone with moderate susceptibility to infection (Dallas and Drake, 2014) for our experiments. Model sensitivity analyses (Supplementary Material) further suggest that our qualitative results are robust to variation in plausible ranges of host infection parameters.

Epidemiological model To examine the impact of a competitor on susceptible host infection dynamics, we used a two-host compartmental model, where the susceptible host species may be uninfected (S) or infected (I) by an environmentally transmitted fungal pathogen (with free-living spore population size P). This susceptible host species competes for resources (R) with a non-susceptible competitor (C). The model was formulated to correspond directly to the experimental treatments, allowing for the testing of model predictions with experimental data. The demographic and epidemic dynamics are described by the following system of differential equations:

$$\dot{S} = e_S f_S(R) R(S + I\phi) - \mu_S S - u f_S(R) S P \quad (1)$$

$$\dot{I} = u f_S(R) S P - \mu_I I \quad (2)$$

$$\dot{P} = \theta \mu_I I - \mu_P P - z_S f_S(R) (S + I) P - z_C f_C(R) C P \quad (3)$$

$$\dot{R} = \pi - \mu_R R - f_S(R) (S + I) R - f_C(R) C R \quad (4)$$

$$\dot{C} = e_C f_C(R) R C - \mu_C C \quad (5)$$

163 Susceptible (S) and competitor (C) populations grow proportionally to the rate at
 164 which individuals can acquire ($f_j(R)$, $j = S, C$) and assimilate (e_j , $j = S, C$) re-
 165 sources, die at rate μ_S (susceptible) or μ_C (competitor), and become infected at a rate
 166 determined by their filtering rate ($f_j(R)$, $j = S, C$) and a per spore infectivity pa-
 167 rameter (u). Filtering rates depend on the density of algal resources (R) (Hall et al.,
 168 2010; Cáceres et al., 2014) and filtering alters the rate of environmental pathogen and
 169 resource loss from the system, as well as the transmission of pathogen. Infected (I)

170 individuals still produce susceptible offspring, but at a rate reduced by ϕ . Infected
 171 individuals die at rate $\mu_I > \mu_S$ to account for pathogen-induced mortality (i.e. viru-
 172 lence). Upon death, hosts release a burst of pathogen spores (θ) to the environmental
 173 pathogen bank. Environmental pathogen (P) decays at a constant rate μ_P , and is also
 174 ingested by susceptible (S), infected (I), and competitor (C) individuals at rates deter-
 175 mined by their corresponding filtering rates ($f_j(R)$, $j = S, C$), and a parameter which
 176 determines the fraction of spores ingested that are rendered non-infectious after pas-
 177 sage through the host gut (z_j , $j = S, C$); this matches observations of spore survival
 178 after bluegill feeding (Duffy, 2009), and *Daphnia* hosts exposed to *Pasteuria ramosa*,
 179 a bacterial parasite (King et al., 2013). Resource (R) is introduced at a constant rate
 180 (π), and decays at a per capita rate μ_R plus additional decay as a function of host for-
 181 aging ($f_j(R)$, $j = S, C$).

182 While the exact relationship between algal resource concentration and *Daphnia* filter-
 183 ing rate is unclear, evidence suggests that clearance rate is negatively related to al-
 184 gal resource quantity, such that it is highest when algal resources are limiting (Porter
 185 et al., 1982; Hall et al., 2007, 2010; Sarnelle and Wilson, 2008). Therefore, we use a

186 type II functional response for filtering rates (Eq. 6 and 7). To establish the competi-
 187 tor as dominant, we increased the competitor's assimilation coefficient (e) and maxi-
 188 mum filtering rate (f_{C0}) relative to the susceptible host species, reflecting the biology of
 189 the system, as the competitor is a larger-bodied grazer with an elevated filtering rate,
 190 larger clutch sizes, and faster growth. The equations for host species and competitor
 191 filtering rates are provided below, where f_{S0} and f_{C0} are the maximum filtering rates
 192 at low resource availability for susceptible and competitor species respectively, and f_{S1}
 193 and f_{C1} determine how rapidly their foraging rates decline in response to increasing re-
 194 source availability.

$$f_S(R) = \frac{f_{S0}}{1 + f_{S1}R} = \frac{0.02}{1 + 4R} \quad (6)$$

$$f_C(R) = \frac{f_{C0}}{1 + f_{C1}R} = \frac{0.025}{1 + 4R} \quad (7)$$

195 The pathogen basic reproduction number (R_0) is a threshold quantity determining
 196 pathogen invasion. We provide it to highlight the effect of the opposing forces of spore
 197 removal through foraging (f_S and z_S), and spore creation through pathogen transmis-
 198 sion (u) and infected host death (θ). For our system, this can be expressed as

$$R_0 = \frac{\theta u f_S(R_S) S^*}{\mu_P + z_S f_S(R_S) S^*}$$

199 with the derivation outlined in the Supplementary Materials.
 200 Parameter definitions, units, and details of the parametrization are provided in Table
 201 1; parameter values were obtained largely from the published literature. To account
 202 for uncertainty in some parameter estimates, and to investigate the generality of the
 203 simulation results, we performed a sensitivity analysis (Supplementary Material). We
 204 solved this model numerically for a range of initial competitor densities (0 - 100 hosts
 205 L^{-1}). Simulations were initiated with 30 susceptible hosts, no infected hosts, and 10000
 206 pathogen spores. Simulated epidemics were run for 70 days, corresponding to condi-

207 tions in experimental epidemics. From epidemic simulations, we calculated mean infec-
 208 tion prevalence (i.e., average infection prevalence over 70 day time series), and mean
 209 susceptible host density as our response variables to changes in competitor density.
 210 **Experimental epidemics** To test our model predictions, we devised a mesocosm
 211 experiment where we manipulated competitor density as a means to modify resource
 212 availability, and therefore the effects of competition. Experimental populations were
 213 formed by dividing five gallon glass aquaria (16" x 8" x 10") in half, separating the two
 214 sides of the aquaria with partitions of 210 μ m Nitex mesh, and filling the tank with 6
 215 L of media; a combination of 2 L filtered pondwater (30 micron filter) and 4 L deion-
 216 ized water. Species were separated by this mesh partition, which allowed for the flow
 217 of resources and pathogen spores, but restricted movement of individuals, thereby iso-
 218 lating the effects of resource competition (i.e. exploitative competition) from any di-
 219 rect interaction (i.e. interference competition), and removing any confusion identifying
 220 *Daphnia* neonates to species.

221 Resource competition was produced by altering the density of *D. pulicaria* and restrict-
 222 ing algal resources. Every day, each half of experimental mesocosms was fed 1 mL of a

223 solution of 200 mg freeze-dried, pulverized *Spirulina* sp. suspended in 100 mL deion-
 224 ized water. We fed both partitions of the aquaria the same amount to ensure that re-
 225 sources were well-mixed between halves of each tank, and that the resource concen-
 226 tration throughout the aquaria was approximately 0.67 mg algal dry weight L⁻¹. Five
 227 mesocosms were formed for each of three initial *D. pulicaria* densities (0, 30, and 100
 228 individuals L⁻¹) for a total of fifteen aquaria. Populations of *D. dentifera* were estab-
 229 lished in each of the fifteen experimental aquaria at a density of 30 individuals L⁻¹ at
 230 the start of the experiment. Competitor densities were chosen based on our susceptible
 231 host density, where the 30 individual L⁻¹ treatment corresponds to both species start-
 232 ing at equal densities, and the 100 competitors L⁻¹ corresponding to a case where the
 233 competitor dominates the community. Both sides of the aquaria were inoculated with
 234 10 *Metschnikowia* spores mL⁻¹ one day after populations were established.

235 Mesocosms were sampled every 3-4 days until infection was no longer observed, which
 236 was after 70 days. We assessed infection prevalence and host density by stirring tanks
 237 and taking a 1 L water sample from each partition of each aquarium. Infection was as-
 238 sessed by visual inspection using a dissecting microscope (10× - 40×) under low light

239 and keeping hosts in a minimal amount of water to reduce host mortality. Hosts are
 240 translucent, and opaque pathogen clusters are present in host heart or gills approx-
 241 imately one week after pathogen transmission. Hosts were returned to their respec-
 242 tive aquarium. Sampling with replacement is ideal in this experiment, as spores that
 243 infected hosts liberate upon death drive subsequent infections in natural systems; re-
 244 moval of infected individuals would artificially reduce epidemic size or duration.

245 We analyzed the influence of competition on epidemic dynamics and host density. To
 246 examine epidemic dynamics in *D. dentifera* in response to competition with *D. puli-*
 247 *caria*, we calculated two quantities meant to capture aspects of epidemic size and du-
 248 ration: mean infection prevalence, and epidemic duration. Mean infection prevalence
 249 was quantified as the fraction of *D. dentifera* infected averaged over the total number
 250 of sampling points in which the susceptible host population persisted. Epidemic dura-
 251 tion was defined as the number of days epidemics had non-zero prevalence. These mea-
 252 sures were compared among initial competitor density treatments using Kruskal-Wallis
 253 tests. These tests addressed the influence of competitor density on infection dynamics
 254 and epidemic duration. Kruskal-Wallis tests were also used to investigate the relation-

ship between the time until *D. dentifera* population extinction and initial competitor density, which addressed the influence of competitor density on susceptible host demography and extinction dynamics. While it is possible that very small populations would not be detected in our 1 L sample, population extinction was noted only when a sample contained no hosts, and a visual inspection of the tank confirmed no living *D. dentifera* hosts.

Results

Comparison of model and experiments Equilibrium analysis of the model (see Supplementary Material) demonstrated that in the long term, *D. dentifera* would be excluded by *D. pulicaria*, and indeed our experimental populations went extinct within 70 days. Our epidemiological model revealed some outcomes that were not observed in our experimental epidemics. For instance, by examining numerous algal resource input values (Figure 1 and 2), we found that the theoretical hump-shaped relationship between initial competitor density and infection prevalence in the susceptible host species was not strongly influenced by resource availability. Consideration of the pathogen ba-

270 sic reproductive number in the absence of the competitor suggested that the addition of
 271 a competitor could enhance or reduce epidemic risk through antagonistic effects of in-
 272 creasing the filtering rate (and the chance of pathogen exposure) while simultaneously
 273 reducing the number of susceptible hosts and infectious propagules. Further exploration
 274 of the conditions where competition could reduce or enhance epidemic risk is outlined
 275 in the sensitivity analysis section of the Supplementary Materials. Overall, this effort
 276 suggested that the hump-shaped relationship between competitor density and infec-
 277 tion prevalence observed in both our experimental epidemics and epidemiological model
 278 is robust to a range of parameter values. The range of parameters in which competi-
 279 tor density strictly reduces infection prevalence is small, and corresponds to situations
 280 in which the competitor digests a much larger proportion of spores than the suscepti-
 281 ble host, or when susceptible hosts produce too few infectious spores to result in sus-
 282 tained transmission. Infection prevalence in the susceptible host species increased when
 283 competitors were first added to the system until a threshold was reached, and then de-
 284 clined. When resources were less limiting, competitors were able to reduce infection
 285 prevalence in the susceptible host more strongly, and mean susceptible host population

286 sizes were larger (Figure 2).

287 **Competitor density and susceptible host epidemic dynamics** Experi-
 288 tal epidemics were qualitatively similar to predictions derived from our epidemiolog-
 289 ical model (Figure 1), despite independent parameterization of the epidemiological
 290 model. The first testable prediction from our theoretical model was that prevalence
 291 has a hump-shaped relationship with the initial density of the competitor species. In
 292 our experimental epidemics, competitor density had a strong effect on susceptible host
 293 species infection dynamics (Figure 3 and 4). Mean infection prevalence (Kruskal-Wallis
 294 test; $\chi^2 = 6.74$, $df = 2$, $p = 0.034$) and epidemic duration (Kruskal-Wallis test; $\chi^2 =$
 295 6.31 , $df = 2$, $p = 0.043$) both increased at intermediate levels of competition (30 *Daph-*
 296 *nia* L^{-1} for both species). Further, it is interesting to note that at the early stages of
 297 epidemics, after *Daphnia* populations were exposed to free-living pathogen spores, in-
 298 fection prevalence increased monotonically with competitor density (Figure 3), sug-
 299 gesting that the competitor presence increased infection prevalence over very short
 300 timescales.

301 **Competitor density reduces susceptible host population size** Our model pre-
 302 dicts that mean susceptible host population size should decline with increasing initial
 303 competitor density (Figure 2). In our experiments, epidemics were smaller when *D.*
 304 *pulicaria* densities were at their highest (100 *D. pulicaria* L⁻¹), driven not by the re-
 305 moval of pathogen from the environment, but by the competitive exclusion of the in-
 306 ferior competitor (Figure 4b). The time until *D. dentifera* extinction was reduced by
 307 increasing *D. pulicaria* density, though not significantly (Kruskal-Wallis test; $\chi^2 = 4.92$,
 308 $df = 2$, $p = 0.085$). The resulting termination of epidemics with competitive exclusion
 309 is evident when examining the infection time series (Figure 3). However, experimen-
 310 tal epidemics also resulted in susceptible host extinction in the absence of competition,
 311 a phenomenon not predicted in our model. This is likely an experimental artifact; a
 312 result of deteriorating water quality, limited food supply, and enhanced mortality as a
 313 result of bi-weekly sampling.

314 Discussion

315 This study investigates support for the “friendly competition” concept (Hall et al.,
 316 2009) over ecologically relevant timescales, using a theoretical model and experiments.
 317 We found that when resources are limiting, competition with a superior competitor
 318 may be entirely “unfriendly” to susceptible host populations in two different ways:
 319 by increasing infection prevalence in susceptible hosts (through increased filtering in-
 320 duced by nutritional stress), and by reducing resources to levels below which suscep-
 321 tible host populations cannot maintain themselves. Prevalence increases through in-
 322 creased pathogen intake occurred at intermediate competitor densities; although preva-
 323 lence declined at high competitor densities (often interpreted as evidence for friendly
 324 competition), the net effect of competition was negative, reducing susceptible host pop-
 325 ulations towards competitive exclusion. Consistent with classic theory (Gause, 1934),
 326 our model predicts that eventual competitive exclusion of the inferior competitor is
 327 inevitable, barring niche partitioning or other coexistence mechanisms such as refu-
 328 gia. Overall, our findings suggest that the competitive effects of a dominant competitor
 329 are usually negative, and that any positive effect of the competitor removing pathogen

330 from the environment may be overwhelmed by the effect of reduced resource availabil-
 331 ity on inferior competitor feeding behavior and persistence.

332 Our model and experimental design make several simplifying assumptions that could
 333 influence competitor effects on infection dynamics in natural systems. For instance, our
 334 model allowed filtering rate, which is crucial to pathogen transmission, to vary only
 335 with resource quantity while other factors (e.g. pesticides; Fernández-Casalderrey et al.
 336 (1994)) may influence foraging ecology and therefore affect transmission independently
 337 of resource depletion by competitors. Within hosts, gut residence time may correspond
 338 to changes in the probability that a pathogen spore will pierce the gut wall and cause
 339 infection. This might explain the findings of previous studies in which both biotic and
 340 abiotic stressors decreased filtering rate and also increased pathogen infection success
 341 (Day and Kaushik, 1987; Fernández-Casalderrey et al., 1994; Coors and De Meester,
 342 2008; Coors et al., 2008; Jansen et al., 2011). Hosts experiencing stress, either through
 343 starvation (Pulkkinen and Ebert, 2004) or from the presence of secondary compounds
 344 from competitors, may experience higher transmission success due to the inability to
 345 resist pathogen infection (Lafferty and Holt, 2003); in this case, prevalence may con-

346 tinue to increase at higher competitor densities than predicted by our model. Finally,

347 *Daphnia* feeding selectivity (DeMott, 1982; Knisely and Geller, 1986), and spatial ag-

348 gregation of pathogen (given that pathogen spores settle quickly after host mortality)

349 may reduce spore encounter rates and host infection independently of resource and

350 competitor density.

351 Previous studies in the *Daphnia*-microparasite system have suggested that competition

352 with a non-susceptible host should reduce infection prevalence, resulting in so-called

353 “friendly competition” (Hall et al., 2009). Further, Civitello et al. (2013) argued that

354 increasing susceptible host density could inhibit disease spread as a result of pathogen

355 consumption and host foraging interference, suggesting another instance of foraging in-

356 fluencing infection prevalence. Lastly, Cáceres et al. (2014) examined the conditions

357 under which “friendly competition” could result in long term persistence of the infe-

358 rior competitor, using an epidemiological model nearly identical to ours. We arrived

359 at some conclusions also supported by Cáceres et al. (2014), including the fact that

360 two hosts competing for a limiting resource are unlikely to coexist indefinitely. How-

361 ever, our study also considers the transient dynamics before the susceptible host was

362 excluded. These transient dynamics are ecologically relevant (Hastings, 2004), both
 363 to zooplankton specifically and to studies of host-parasite interactions more generally
 364 (Dobson, 2004). Zooplankton populations are unlikely to have equilibrium population
 365 densities (McCauley and Murdoch, 1987) due to seasonal and stochastic changes in
 366 resource availability and environmental conditions, which influence host demographic
 367 rates. This means that models examining equilibrium conditions may not correspond
 368 to experimental data, making comparisons of models to experiments difficult (Hastings,
 369 2004). Our analysis suggests that friendly competition is unlikely to occur over shorter,
 370 biologically relevant timescales.

371 There are many ways that parasites can influence interactions between hosts (Hatcher
 372 et al., 2012). Many studies focus on how a parasite can handicap the superior com-
 373 petitor, leading to parasite-mediated coexistence (Freeland, 1983; Schall, 1992; Hatcher
 374 et al., 2006; Schmitz and Nudds, 1994). However, these studies typically do not con-
 375 sider how pathogen uptake is influenced by changes to foraging rates due to basal re-
 376 source availability. Our study suggests that competition-mediated foraging rates could
 377 increase prevalence in a pathogen-susceptible, superior resource competitor, reducing

378 its abundance relative to the inferior competitor below that expected when pathogen
 379 transmission is assumed to be independent of resource availability. Further theoretical
 380 and empirical work in this area is warranted.

381 Given its importance for transmission potential of zoonoses such as Lyme Disease (Os-
 382 tfeld and Keesing, 2012), there has been much recent interest in the role of host di-
 383 versity in either diluting or amplifying pathogen transmission (Civitello et al., 2015).

384 Many studies of diversity-disease relationships tend not to incorporate ecological inter-
 385 actions, most notably competition for basal resources. In simple systems where a host
 386 and non-host diluter acquire environmental pathogen stages, prevalence is predicted to
 387 decline monotonically with non-host density. Our results show that at least initially,
 388 intermediate diluter density maximizes epidemic size by increasing host acquisition of
 389 the pathogen. The situation in which the pathogen can become a food resource intro-
 390 duces complexity into the study of infectious disease in ecological communities, but is
 391 not specific to our study system. Many pathogens are environmentally transmitted,
 392 and are subject to incidental predation by hosts, and non-hosts alike (Thieltges et al.,
 393 2008; Parker et al., 2010). Therefore, our results suggest that evidence for dilution or

394 amplification may be influenced by the time scale of observation, resource availability,
 395 as well as the relative abundance or richness of lower-competency hosts.

396 Much like Strauss et al. (2015), this study attempts to unify two concepts in disease
 397 ecology by relating dilution theory to parasite-mediated competition. We highlight the
 398 importance of ecological context (resource availability) to competitive interactions be-
 399 tween hosts, and how this influences infection dynamics in the susceptible host through
 400 a mechanism related to host foraging ecology. Studies of diversity-disease relationships
 401 and parasite-mediated competition often do not incorporate the potentially strong ef-
 402 fect of environmental regulation, specifically with regards to resource availability. The
 403 incorporation of resource-mediated species interactions (direct and indirect) into stud-
 404 ies of diversity-disease relationships may yield a more mechanistic view of diversity-
 405 disease relationships and other areas of disease ecology.

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566 **Supplementary Materials**

567 Appendix A: Equilibrium analysis, non-dimensionalization, and sensitivity analysis of

568 epidemiological model

esa
preprint

Table 1: Parameters, definitions, and units used in our epidemiological model. We chose plausible estimates for parameters for which data were not readily available (denoted by an empty citation column). Our values of assimilation efficacy were estimated using information on *Daphnia* population growth rates (Civitello et al., 2013; Smith, 1963).

| Variable | Units | Definition | Value | Citation |
|------------------|------------------------|--|-----------------------|----------|
| e_S | – | Assimilation efficiency (Susceptible) | 26 | |
| e_C | – | Assimilation efficiency (Competitor) | 28 | |
| μ_S | day^{-1} | Death rate (Susceptible) | 0.10 | 1 |
| μ_I | day^{-1} | Death rate (Infected) | 0.15 | 2 |
| μ_C | day^{-1} | Death rate (Competitor) | 0.10 | |
| μ_P | day^{-1} | Death rate (Pathogen) | 0.25 | 3 |
| μ_R | day^{-1} | Death rate (Resource) | 0.25 | |
| z_S | – | Fraction spores digested (Susceptible) | 0.30 | |
| z_C | – | Fraction spores digested (Competitor) | 0.30 | |
| ϕ | – | Fecundity reduction by infection | 0.75 | 2 |
| f_{S0}, f_{S1} | $ml\ day^{-1}$ | Host filtering rate (Susceptible) | 0.020, 4 | 4, 5 |
| f_{C0}, f_{C1} | $ml\ day^{-1}$ | Host filtering rate (Competitor) | 0.025, 4 | 4, 5 |
| u | – | Per spore infectivity | 2.03×10^{-4} | 4 |
| θ | # | Mean spore load per infected host | 2×10^4 | 3 |
| π | $mg\ L^{-1}\ day^{-1}$ | Resource supply rate | 0.005 – 4 | 6 |

1: Stich and Maier (2007); 2: Duffy and Hall (2008); 3: Dallas and Drake (2014);

4: Hall et al. (2010); 5: DeMott (1982); 6: Tessier and Woodruff (2002)

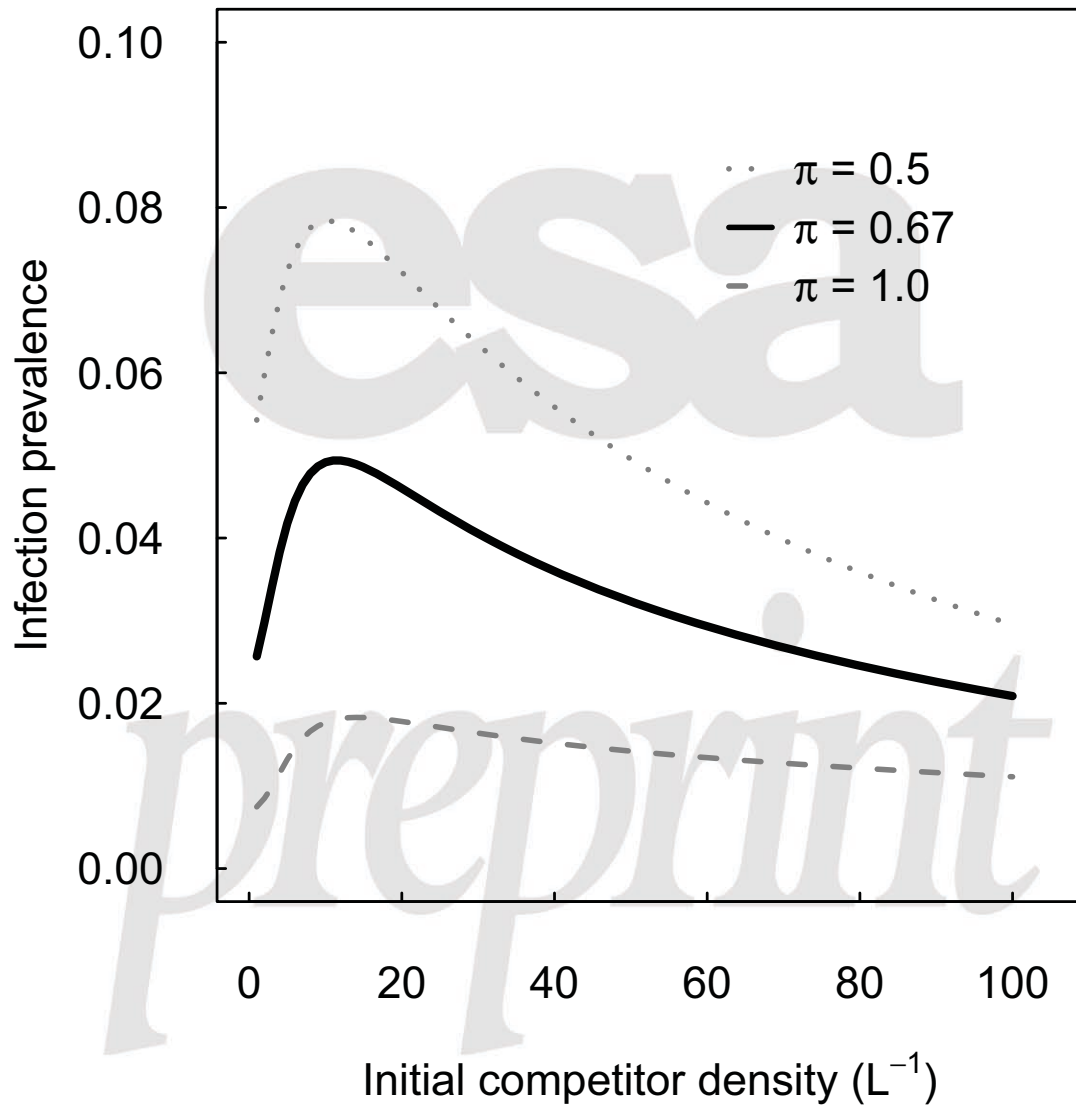
574 **Figures**

Figure 1: Non-monotonic relationship between mean infection prevalence and initial competitor density (x -axis) for three potential algal resource input levels (π ; y -axis). Enhanced resource input rates reduce infection prevalence by decreasing filtering rates (and hence pathogen exposure). The reduction in infection prevalence at larger initial competitor densities is a result of reduced susceptible host population sizes, and not a positive effect of the competitor removing environmental pathogen.

Figure 2: Mean population size of the susceptible host species ($S + I$) as a function of initial competitor density for three different algal resource input levels. $\pi = 0.5$ and $\pi = 1$ are the lower and upper dashed lines, respectively, and the solid black line corresponds to $\pi = 0.67$, the rate of algal resource supply in our experimental epidemics.

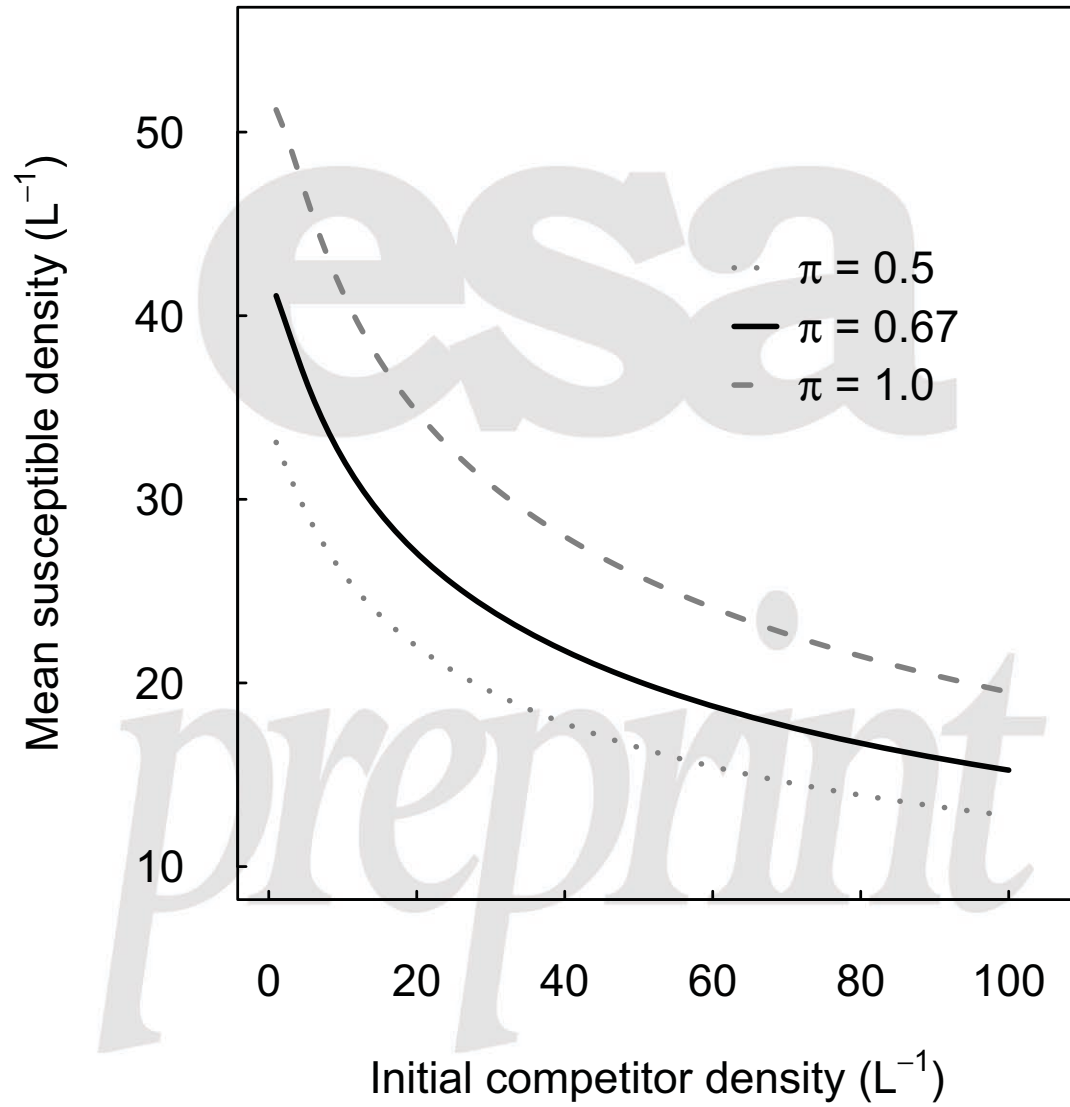
Figure 3: Mean infection prevalence (with standard error bars) for the epidemic time series. The inset barplot compares the infection prevalence at day 10, which corresponds to the first wave of infection, as the pathogen typically takes between 7 and 12 days to be readily identifiable, suggesting that competition initially increased infection prevalence proportional to *D. pulicaria* density.

Figure 4: Experimental epidemics at three competitor densities revealed that intermediate levels of competition significantly increased mean infection prevalence (*a*) and epidemic duration (*c*). There was no difference in epidemic measures between no competitor and high *D. pulicaria* competitor density treatments, driven by truncated epidemics at high competitor densities as a result of the competitive exclusion of the susceptible host (*D. dentifera*; *b*).



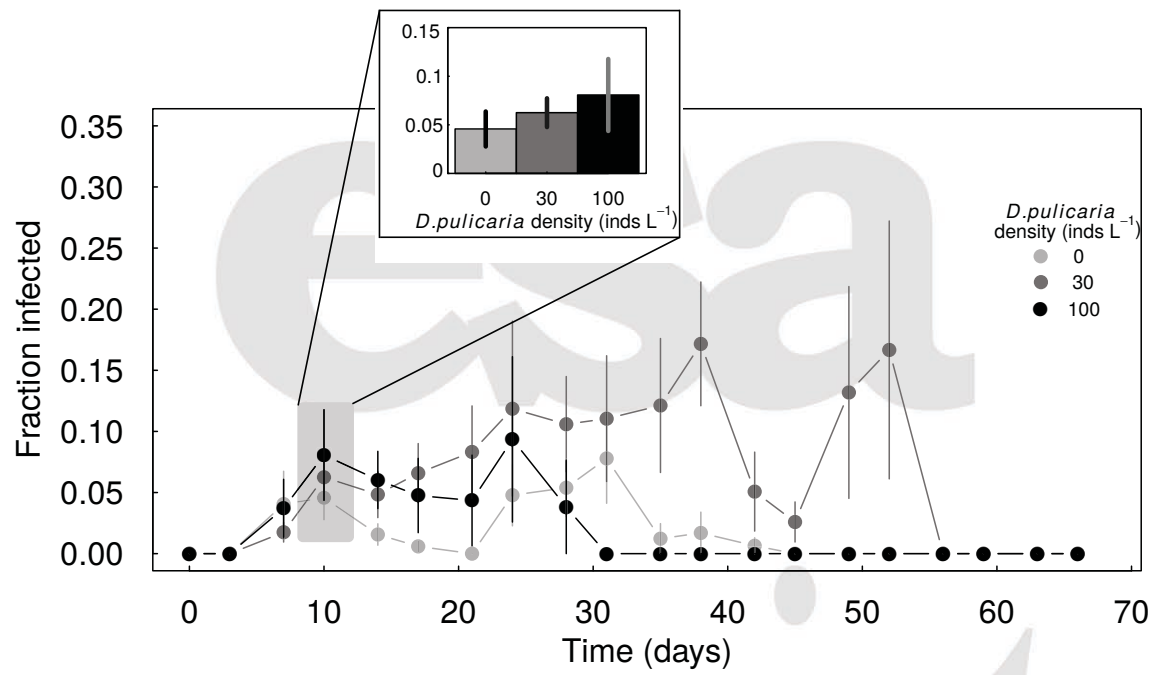
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576 Figure 1



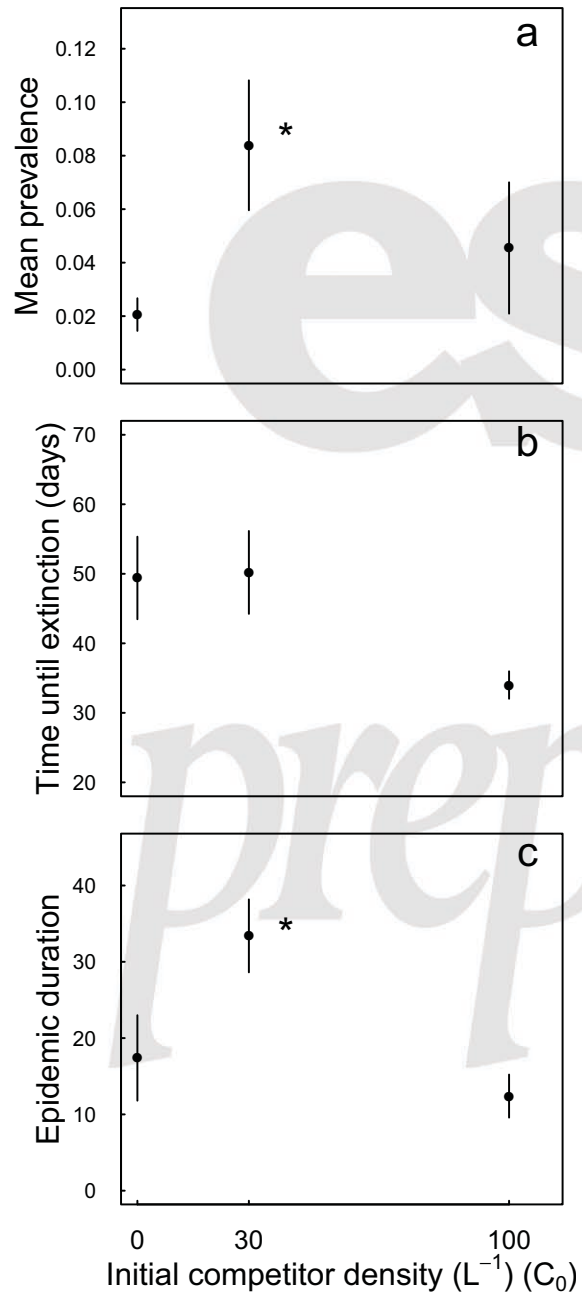
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578 Figure 2



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580 Figure 3



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582 Figure 4