

## Bacterial community structure in fumigated soil



Cristina A. Domínguez-Mendoza, Juan M. Bello-López, Yendi E. Navarro-Noya, Arit S. de León-Lorenzana, Laura Delgado-Balbuena, Selene Gómez-Acata, Victor M. Ruíz-Valdiviezo, Daniel A. Ramírez-Villanueva, Marco Luna-Guido, Luc Dendooven\*

Laboratory of Soil Ecology, ABACUS, Cinvestav, Avenida Instituto Politécnico Nacional 2508, Z.C. 07360 México DF, Mexico

### ARTICLE INFO

#### Article history:

Received 8 November 2013

Received in revised form

27 January 2014

Accepted 17 February 2014

Available online 3 March 2014

#### Keywords:

C and N mineralization

Fumigated and unfumigated arable soil

*Bacillus*

*Cohnella*

*Micromonospora*

*Paenibacillus*

### ABSTRACT

Soil microbial biomass has been determined since the mid 1970's by the chloroform fumigation incubation technique as proposed by Jenkinson and Powlson (1976). The microbial biomass C can be determined by subtracting the CO<sub>2</sub> emitted from an unfumigated soil (mineralization of soil organic matter) from that emitted from a chloroform fumigated inoculated soil (mineralization of soil organic matter and killed soil microorganisms) and dividing the difference by a proportionality factor ( $k_c = 0.45$ ). The question remained which microorganisms recolonized a fumigated soil. An arable soil was fumigated for one day with ethanol-free chloroform or left unfumigated and incubated aerobically after removal of the chloroform for 10 days. The bacterial population structures were determined in the fumigated and unfumigated soil after 0, 1, 5 and 10 days by means of 454 pyrosequencing of the 16S rRNA gene. Fumigating the arable soil reduced significantly the relative abundance of phylotypes belonging to different groups, but increased the relative abundance of only four genera belonging to two phyla (*Actinobacteria* and *Firmicutes*) and two orders (*Actinomycetales* and *Bacillales*). The relative abundance of phylotypes belonging to the *Micromonospora* (*Micromonosporaceae*) increased significantly from 0.2% in the unfumigated soil to 6.7% in the fumigated soil and that of *Bacillus* (*Bacillaceae*) from 3.6% to 40.8%, *Cohnella* (*Paenibacillaceae*) from undetectable amounts to 0.6% and *Paenibacillus* (*Paenibacillaceae*) from 0.3% to 4.2%. The relative percentage of phylotypes belonging to the *Acidobacteria*, *Bacteroidetes*, *Chloroflexi*, *Gemmatimonadetes* and *Proteobacteria* ( $\alpha$ - $\beta$ -,  $\delta$ - and  $\gamma$ -*Proteobacteria*) were significantly lower in the fumigated than in the unfumigated soil and in most of them the relative abundance of different bacterial orders (i.e. Gp3, Gp4, Gp6, *Sphingobacteriales*, *Gemmatimonadales*, *Rhodospirillales*, *Burkholderiales*, *Xanthomonadales*) was reduced strongly ( $P < 0.001$ ). It was found that the relative abundance of a wide range of bacteria was reduced shortly after fumigating an arable soil, but only a limited group of bacteria increased in a fumigated arable soil indicating a capacity to metabolize the killed soil microorganisms or recolonize a fumigated soil.

© 2014 Elsevier Ltd. All rights reserved.

### 1. Introduction

Soil microorganisms are involved in most of the processes that occur in soil. However, it has always been difficult to have an accurate estimation of the total soil microbial population (Kirchman, 2012). In a first attempt to measure the soil microbial population, microorganisms were stained and counted. Later soil bacteria were enumerated by culturing techniques in nutrient rich or specific media. It was hoped that the determination of the amount of

microorganisms in a soil would allow to estimate the magnitude of certain processes, such as C and N mineralization (Fierer et al., 2009). As such, it would be possible to predict the organic matter dynamics and the amount of mineral N that would be available for growing crops (Abera et al., 2012). However, it became clear that these measurements did not often relate well with microbial activity and soil processes.

In 1976, Jenkinson and Powlson wrote a seminal paper describing a method to measure the soil microbial biomass carbon. They proposed to fumigate a soil and measure the amount of CO<sub>2</sub> emitted within a given period, i.e. 10 days. They assumed that the killed biomass would be mineralized and a certain part of the biomass C would be emitted as CO<sub>2</sub> that could be determined easily.

\* Corresponding author. Tel.: +52 55 5747 3319; fax: +52 55 5747 3313.  
E-mail address: [dendooven@me.com](mailto:dendooven@me.com) (L. Dendooven).

At the same time, the unfumigated soil would serve as control and allow to measure the soil organic matter that apart from the killed biomass would be mineralized in the fumigated soil. As such, the microbial biomass could be calculated as the CO<sub>2</sub> emitted from the fumigated soil minus the CO<sub>2</sub> emitted from the unfumigated soil divided by a factor  $K_C$  considering the amount of biomass C mineralized to CO<sub>2</sub>. The  $K_C$  value reported by Jenkinson and Powlson (1976) was 0.45.

It has always been questioned which microorganisms would recolonize the fumigated soil and mineralize the killed soil microbial population. Kemmit et al. (2008) using phospholipid fatty acids analysis showed that fumigation induced changes in the biomass structure. The PLFA technique has given a first insight into which microorganisms colonize a fumigated soil, but the identification of the bacterial groups remains limited (Kemmit et al., 2008; Rousk et al., 2009; Dungait et al., 2013). New molecular and metagenomic techniques do allow to identify the microorganisms in soil in great detail now (Fierer et al., 2009). An arable soil was fumigated with ethanol-free chloroform for 1 day, inoculated with 0.01 g unfumigated soil as in the original technique and incubated together with unfumigated soil for 10 days. At the onset of the incubation and after 1, 5 and 10 days the emission of CO<sub>2</sub> was measured and the microbial population monitored. Our objective was to determine the bacterial community structure in both a fumigated and unfumigated soil by means of 454 pyrosequencing of the 16S rRNA gene.

## 2. Materials and methods

### 2.1. Soil sampling

The sampling site is located in Otumba. Details of the sampling site can be found in Aguilar-Chávez et al. (2012). The soil was classified as Typic Fragiudepts. The site was cultivated mainly with maize for >20 y and the cultivated crops were not irrigated. Soil was collected by augering randomly 20 times the 0–15 cm top-layer of three plots of approximately 400 m<sup>2</sup> spatially separated. The soil samples from each plot ( $n = 20$ ) were pooled so that three soil samples were obtained ( $n = 3$ ) and characterized. The sandy loam soil had an organic C content of 6.7 g C kg<sup>-1</sup> soil and total N content of 0.73 g N kg<sup>-1</sup> soil. The water holding capacity (WHC) was 665 g kg<sup>-1</sup> dry soil, pH 7.7 and electrolytic conductivity (EC) 1.23 dS m<sup>-1</sup>. Details of techniques used to characterize the soil can be found in Ruíz-Valdiviezo et al. (2010).

### 2.2. Fumigation incubation

Eight sub-samples of 25 g fresh weight soil of each plot ( $n = 3$ ) were added separately to 120 ml glass flasks. Half of the flasks were placed in a desiccator and fumigated with ethanol-free chloroform (Mueller et al., 1992) for one day (i.e. the fumigated soil samples) as described by Jenkinson and Powlson (1976), while the other half were placed separately in 1 l glass jars containing a vessel with 20 ml 1 M NaOH to trap the CO<sub>2</sub> evolved and a vessel with distilled water to avoid desiccation of the soil samples (i.e. the unfumigated soil samples). The glass jars were closed airtight and incubated in the dark at 25 ± 2 °C for 10 days. After one day of fumigation, the desiccator was aired for 1 h and the headspace was vacuum evacuated until all chloroform was removed. The fumigated soil samples were inoculated with 0.01 g unfumigated soil, mixed, placed separately in a 1 l glass jar and incubated as described for the unfumigated soil samples.

After 0, 1, 5 and 10 days, a jar with fumigated and unfumigated soil from each plot ( $n = 3$ ) was selected at random, opened, the vessel with 1 M NaOH stoppered and the soil removed from the

glass flask for extraction of DNA. The 1 M NaOH was titrated with 0.1 M HCl to determine the CO<sub>2</sub> trapped (Jenkinson and Powlson, 1976). The soil microbial biomass was calculated as the [(CO<sub>2</sub> emitted from the fumigated soil – CO<sub>2</sub> emitted from the unfumigated soil)/0.45] (Jenkinson and Powlson, 1976).

### 2.3. DNA extraction and PCR amplification of bacterial 16S rRNA genes

Each soil sample was washed first with 0.15 M sodium pyrophosphate and 0.15 M pH 8 phosphate buffer to eliminate the soil organic material (Ceja-Navarro et al., 2010). The DNA was extracted from the washed soil as described by Ceja-Navarro et al. (2010) and consisted in a chemical and thermal shock for cell lysis. DNA was extracted four times from 0.5 g soil (a total 2 g soil) and pooled. A total of 2 g soil was extracted for DNA per plot so overall 6 g soil was extracted. The precipitation of the proteins and purification of the DNA are described in Ceja-Navarro et al. (2010).

The V1–V6 region of the 16S rRNA bacterial genes was amplified with 10-pb barcoded primers 8-F (5'-AGA GTT TGA TCI TGG CTC A-3') and 949-R (5'-CCG TCW ATT KCT TTG AGT T-3') and containing the A and B 454 FLX adapters (Navarro-Noya et al., 2013). The PCR reactions were done as previously described by Navarro-Noya et al. (2013). The product of five reactions of each metagenomic DNA sample was pooled to minimize PCR bias and constituted a single library (Acinas et al., 2004). All the pyrosequencing libraries were purified using the DNA Clean & Concentrator purification kit as recommended by the manufacturer (Zymo Research, Irvine, CA, USA), and quantified using the PicoGreen® dsDNA assay (Invitrogen, Carlsbad, Ca, USA) and the NanoDrop™ 3300 Fluorometer (Thermo Scientific NanoDrop). Sequencing was done by Macrogen Inc. (DNA Sequencing Service, Seoul, Korea) by using a Roche 454 GS-FLX Plus System pyrosequencer (Roche, Mannheim, Germany). The 24 pyrosequencing-derived 16S rRNA gene sequence datasets were submitted to the NCBI Sequence Read Archive (SRA) under accession number SRA108903.

### 2.4. Analysis of pyrosequencing data

The QIIME version 1.5.0 software pipeline was used to analyze the pyrosequencing data (Caporaso et al., 2010b). Firstly, the poor quality reads were eliminated from the data sets, i.e. quality score <25, containing homopolymers >6, length <400 nt, and containing errors in primers and barcodes. Operational taxonomic units (OTU) were determined at 97% similarity level with UCLUST algorithm (Edgar, 2010). Chimeras were detected and removed from the data sets using the Chimera Slayer (Haas et al., 2011). Sequence alignments were done against the Greengenes core set and using representative sequences of each OTU using PyNAST, and filtered at a threshold of 75% (Caporaso et al., 2010a).

### 2.5. Phylogenetic and statistical analysis

The taxonomic distribution estimates at different levels was done using the taxonomy assignment at a confidence threshold of 80% by the naïve Bayesian rRNA classifier from the Ribosomal Data Project (<http://rdp.cme.msu.edu/classifier/classifier.jsp>) (Wang et al., 2007).

The beta diversity analysis was done using phylogenetic information (UniFrac analysis) with Fast UniFrac (Hamady et al., 2010) using a phylogenetic tree computed with FastTree (Price et al., 2009) and a rarefied 'biom' table at 250 reads as inputs. Also, abundance of the bacterial groups at different taxonomic levels (phylum, order, genus) was separately explored with a PCA using PROC FACTOR (SAS Institute, 1989). Significant difference between

the relative abundance of the different bacterial groups as a result of fumigation was determined by analysis of variance (ANOVA) and based on the least significant difference using the general linear model procedure (PROC GLM, SAS Institute, 1989).

### 3. Results

#### 3.1. Microbial biomass: fumigation incubation technique

The emission of CO<sub>2</sub> was larger from the fumigated soil than from the unfumigated soil and that from the first day onwards (Fig. S1). Afterward, however, emissions of CO<sub>2</sub> were similar in the fumigated and unfumigated soil. The microbial biomass C was 116.4 mg C kg<sup>-1</sup> soil or 1.74% of the organic soil C content.

#### 3.2. The bacterial community structure in the arable soil

Phylotypes belonged to 13 different phyla in the unfumigated arable soil (Fig. 1). The relative abundance of phylotypes belonging to the Proteobacteria was the highest (41.7%), followed by Acidobacteria (14.6%) and the Bacteroidetes (14.6%). The relative

abundance of phylotypes belonging to the Firmicutes was (5.8%) and three phyla could be described as rare (<1%). Changes in the relative abundance of the different groups on the phylum, class and order level in the unfumigated soil were generally small over time (No data shown).

Fumigating the arable soil reduced significantly the relative abundance of phylotypes belonging to different groups, but increased the relative abundance of only a small proportion of the soil bacteria (Table 1, Fig. 2a and b). The relative abundance of phylotypes belonging to only two phyla (Actinobacteria and Firmicutes), two orders (Actinomycetales and Bacillales) and four genera significantly increased after fumigation. The relative abundance of phylotypes belonging to the *Micromonospora* (Actinomycetales, Actinobacteria) increased significantly from 0.2% in the unfumigated soil to 6.7% in the fumigated soil and that of *Bacillus* from 3.6% to 40.8%, that of *Cohnella* from undetectable amounts to 0.6% and *Paenibacillus* from 0.3% to 4.2% (Fig. 3).

Fumigating the soil strongly reduced the relative abundance of a wide range of phylotypes belonging to different bacterial groups; at the level of genus approximately 30 different groups were affected. The relative percentage of phylotypes belonging to the phyla

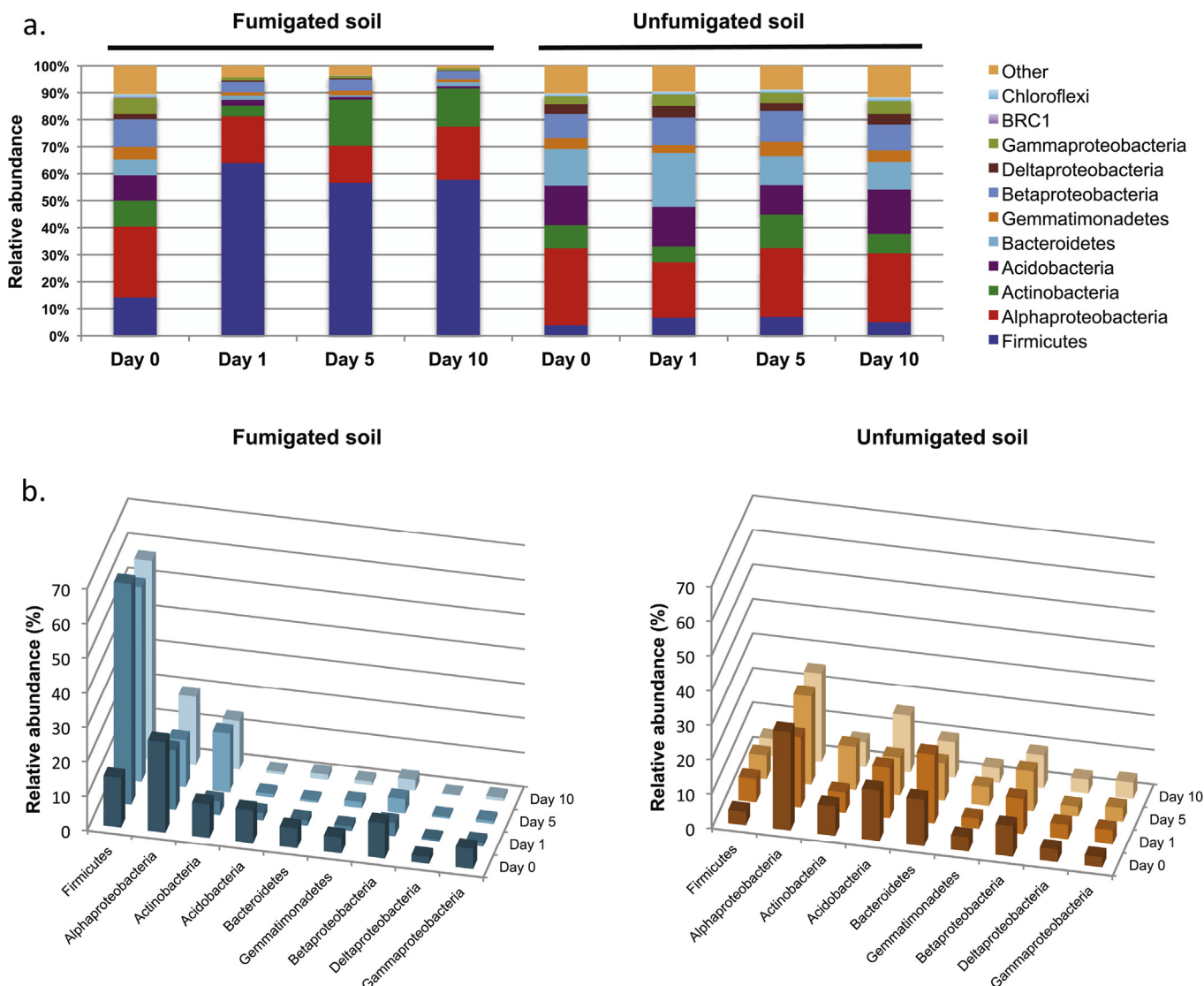


Fig. 1. Relative abundance of the different phyla and Proteobacteria classes found in the fumigated and unfumigated soil incubated aerobically for 10 days.

**Table 1**

The relative abundance of different bacterial groups (level of phylum, class, order and genus) in an unfumigated or fumigated arable soil incubated aerobically for 10 days. Bacterial groups were affected significantly by fumigation at the  $P < 0.05$  level.

Phylum/Class	The relative abundance of the bacterial group was larger in unfumigated soil than in the fumigated soil.
Acidobacteria	Acidobacteria <sup>***</sup> , Gp16 <sup>*</sup> , Gp3 <sup>***</sup> , Gp4 <sup>***</sup> , Gp6 <sup>***</sup> , Gp7
Actinobacteria	<i>Marmoricola</i> (Actinomycetales; Nocardioidaceae)
Bacteroidetes	Bacteroidetes <sup>***</sup> , Sphingobacteria <sup>***</sup> , Sphingobacteriales <sup>***</sup> , <i>Flavisolibacter</i> <sup>**</sup> (Chitinophagaceae; Sphingobacteriales), <i>Adhaeribacter</i> <sup>***</sup> , <i>Sporocytophaga</i> <sup>*</sup> (Cytophagaceae; Sphingobacteriales)
Chloroflexi	Chloroflexi <sup>***</sup> , Anaerolineae, Chloroflexi, Thermomicrobia <sup>*</sup> , Anaerolineales, Sphaerobacteriales, <i>Sphaerobacter</i> (Sphaerobacteriales; Sphaerobacteraceae)
Gemmatimonadetes	Gemmatimonadetes <sup>**</sup> , Gemmatimonadales <sup>**</sup> , <i>Gemmatimonas</i> <sup>**</sup>
Proteobacteria	Proteobacteria <sup>***</sup>
α-Proteobacteria	α-Proteobacteria, Caulobacteriales, Rhodospirillales <sup>**</sup> , <i>Skermanella</i> <sup>**</sup> (Rhodospirillaceae; Rhodospirillales), <i>Porphyrobacter</i> (Erythrobacteraceae; Sphingomonadales)
β-Proteobacteria	β-Proteobacteria <sup>***</sup> , Burkholderiales <sup>**</sup> Nitrosomonadales, <i>Azohydromonas</i> (Alcaligenaceae; Burkholderiales)
δ-Proteobacteria	<i>Methylobium</i> <sup>*</sup> (Burkholderiales_incertae_sedis; Burkholderiales), <i>Variovorax</i> <sup>*</sup> (Comamonadaceae; Burkholderiales), <i>Herbaspirillum</i> (Oxalobacteraceae; Burkholderiales), <i>Nitrospira</i> (Nitrosomonadaceae; Nitrosomonadales)
γ-Proteobacteria	δ-Proteobacteria <sup>***</sup> , Myxococcales
	γ-Proteobacteria <sup>***</sup> , Xanthomonadales <sup>***</sup> , <i>Steroidobacter</i> (Sinobacteraceae; Xanthomonadales), <i>Lysobacter</i> <sup>*</sup> (Xanthomonadaceae; Xanthomonadales)
Actinobacteria	The relative abundance of the bacterial group was larger in fumigated soil than in the unfumigated soil.
Firmicutes	<i>Micromonospora</i> <sup>**</sup> (Micromonosporaceae; Actinomycetales)
	Firmicutes <sup>***</sup> , Bacillaceae <sup>***</sup> , Bacillales <sup>***</sup> , <i>Bacillus</i> <sup>***</sup> (Bacillaceae; Bacillales), <i>Paenibacillus</i> <sup>**</sup> and <i>Cohnella</i> (Paenibacillaceae; Bacillales)

\* Significant at the  $P < 0.01$ , \*\* Significant at the  $P < 0.001$ , \*\*\* Significant at the  $P < 0.0001$ .

Acidobacteria, Bacteroidetes, Chloroflexi, Gemmatimonadetes and Proteobacteria were significantly lower in the fumigated than in the unfumigated soil and in most of them the relative abundance of different bacterial groups were reduced (Table 1, Fig. 1). The relative abundance of phylotypes of the Acidobacteria showed the largest drop from an average of 14.6% in the unfumigated soil to 1.4% in the fumigated soil at day 10 and within the Acidobacteria, the largest decrease in relative abundance was found in the class Gp4 (unfumigated soil, 7.3%; fumigated soil, 0.04%).

### 3.3. Changes in the bacterial community structure

Bacterial community structure in the fumigated and unfumigated soil was compared using taxonomic and phylogenetic information. PCA analysis considering the relative abundance at different taxonomic levels (phyla, orders or genera) clearly separated the fumigated soil from the unfumigated soil. For instance, considering the relative abundance of the different orders, the fumigated soil samples were found in the lower left quadrant (a negative PC1 and PC2) or upper left quadrant (one fumigated soil sample at day 0), while the unfumigated soil samples were found in the lower or upper right quadrant together with the fumigated soil samples at day 0 (a positive PC1) (Fig. 4). The fumigated soil samples were characterized by a higher relative abundance of Actinomycetales and Bacillales, while the unfumigated soil samples by a higher relative abundance of a different orders, e.g. Gp4, Gp6, Gemmatimonadales, Burkholderiales, Rhodospirillales, Sphingobacteriales and Xanthomonadales. The UniFrac analysis also separated clearly the communities of the fumigated from the unfumigated soils (Fig. S2).

## 4. Discussion

### 4.1. Microbial biomass: fumigation incubation technique

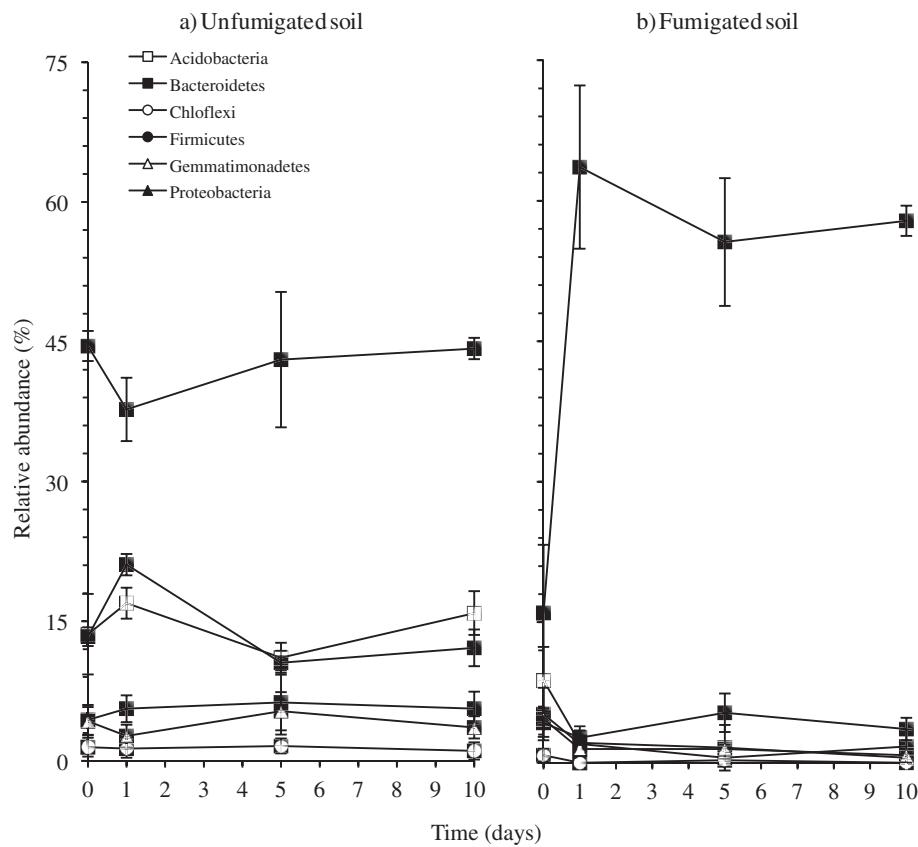
The emission of CO<sub>2</sub> in the fumigated and unfumigated arable soil followed a pattern described already by Jenkinson and Powelson (1976). The difference in emitted CO<sub>2</sub> between the fumigated and unfumigated soil was evident from day 1 onwards, but afterward the emission of CO<sub>2</sub> was similar in the fumigated and unfumigated soil (Kemmit et al., 2008).

### 4.2. The bacterial community structure in the arable soil

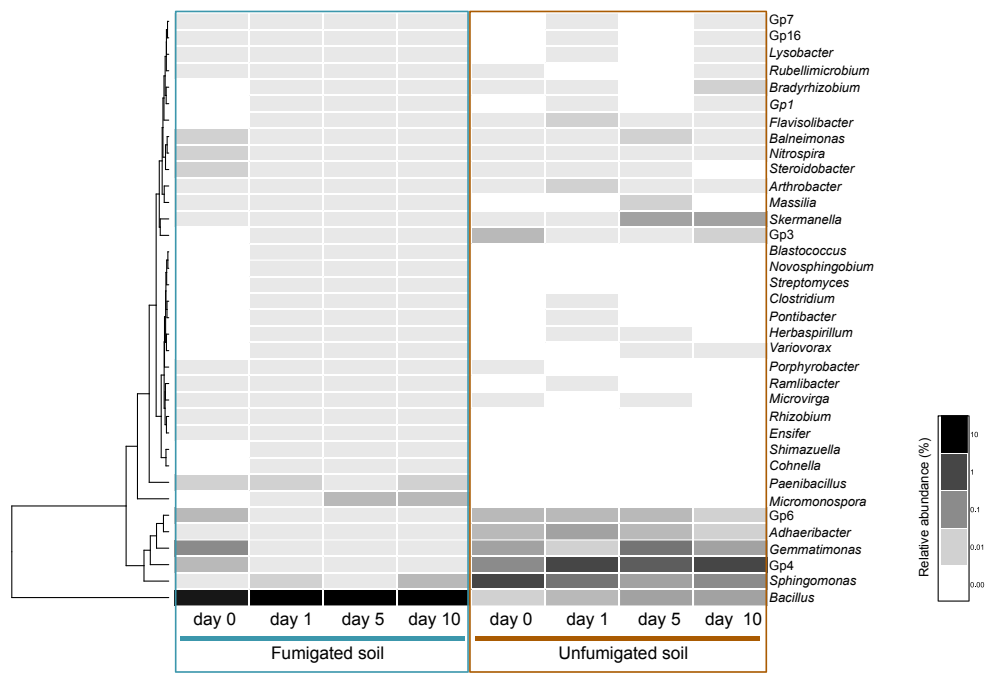
Fumigating the arable soil, inoculating it with a small soil sample and incubating it aerobically for 10 days, increased the relative abundance of a small number of bacterial groups. The relative abundance of only four genera belonging to only two phyla (Actinobacteria, Firmicutes) and two orders (Actinomycetales, Bacillales), known to be involved in degradation of organic material and composting (Partanen et al., 2010; Martins et al., 2013), increased in the incubated soil after fumigation, i.e. phylotypes belonging to genera *Micromonospora* (Micromonosporaceae), *Bacillus* (Bacillaceae), *Paenibacillus* and *Cohnella* (Paenibacillaceae). Consequently, phylotypes belonging to these genera are early colonizers of fumigated soil and involved in the early stages of the mineralization of decomposable organic material derived from the killed soil microorganisms. Strains belonging to the genera *Bacillus*, *Paenibacillus*, and *Micromonospora* are metabolic versatile, while less is known about strains belonging to *Cohnella*. Although the fumigated soil was inoculated with unfumigated soil and mixed thoroughly, the capacity of phylotypes belonging to the Bacillales (Kolter et al., 1993; Cano and Borucki, 1995; García-Fraile et al., 2008) and *Micromonospora* to form spores (Hardisson and Suarez, 1979) will have favored them in recolonizing the fumigated soil.

Strains belonging to the genus *Bacillus* are extremely versatile. They are well known degraders of organic residue, such as soybean and maize (España et al., 2011), show cellulolytic activity (Cucurachi et al., 2013), have been found during various phases of composting (He et al., 2013) and are known to produce proteases, chitinases and cellulases (Pathma and Sakthivel, 2013). A *Bacillus* sp. strain LD003 showed growth on high-molecular and low-molecular weight lignin fractions and utilization of lignin-associated aromatic monomers and degradation of ligninolytic indicator dyes, e.g. Azure B, Methylene Blue and Toluidene Blue O (Bandounas et al., 2011).

Several members of the genus *Paenibacillus* produces cellulases (Pathma and Sakthivel, 2013 [http://apps.webofknowledge.com/OneClickSearch.do?product=WOS&search\\_mode=OneClickSearch&excludeEventConfig=ExcludelfFromFullRecPage&colName=WOS&SID=3BxDSAavvo3aH9QEXnWj&field=AU&value=Sakthivel,%20N](http://apps.webofknowledge.com/OneClickSearch.do?product=WOS&search_mode=OneClickSearch&excludeEventConfig=ExcludelfFromFullRecPage&colName=WOS&SID=3BxDSAavvo3aH9QEXnWj&field=AU&value=Sakthivel,%20N)), phosphatases (Kim et al., 2009), xylanases (Ko et al., 2010a), enzymes which can specifically decompose xylans the main constituent of hemicelluloses, have shown cellulolytic and



**Fig. 2.** Relative abundance of the bacterial phyla in a) unfumigated soil and b) soil that was fumigated with ethanol free chloroform, inoculated with 0.01 g unfumigated soil incubated soil aerobically for 10 days. Bars are  $\pm$  one STD of three soil samples ( $n = 3$ ).



**Fig. 3.** Heat map of the most important genera and Acidobacteria groups (Gp) in the fumigated and unfumigated soil incubated aerobically for 10 days. Values are the mean of three soil samples ( $n = 3$ ).



hemicellulolytic activity (Hatayama et al., 2006; Ko et al., 2010b; Eida et al., 2012). Additionally, *Paenibacillus* can degrade kraft lignin (Chandra et al., 2007), they consisted 5.1% of the phylotypes in a lignin degradation bacterial consortium (Wang et al., 2013) and identification of genes for lignocellulose processing and metabolic reconstructions suggested *Paenibacillus* a key group in decomposition of lignocellulosic material (D'haeseleer et al., 2013). By genome mining, seven chitinase and four 1,3-beta glucanase encoding genes were identified in the *Paenibacillus vortex* genome (Sirota-Madi et al., 2010); these enzymes are involved in degradation of polysaccharide components of the fungal cell wall (El-Katatny et al., 2001), while the genome of a *Paenibacillus lautus* strain is enriched in GH18 chitinases, GH28 polygalacturonases, GH88 unsaturated glucuronyl hydrolases, GH105 unsaturated rhamnogalacturonyl hydrolases and pectate lyase (PL) family enzymes (Meade et al., 2012).

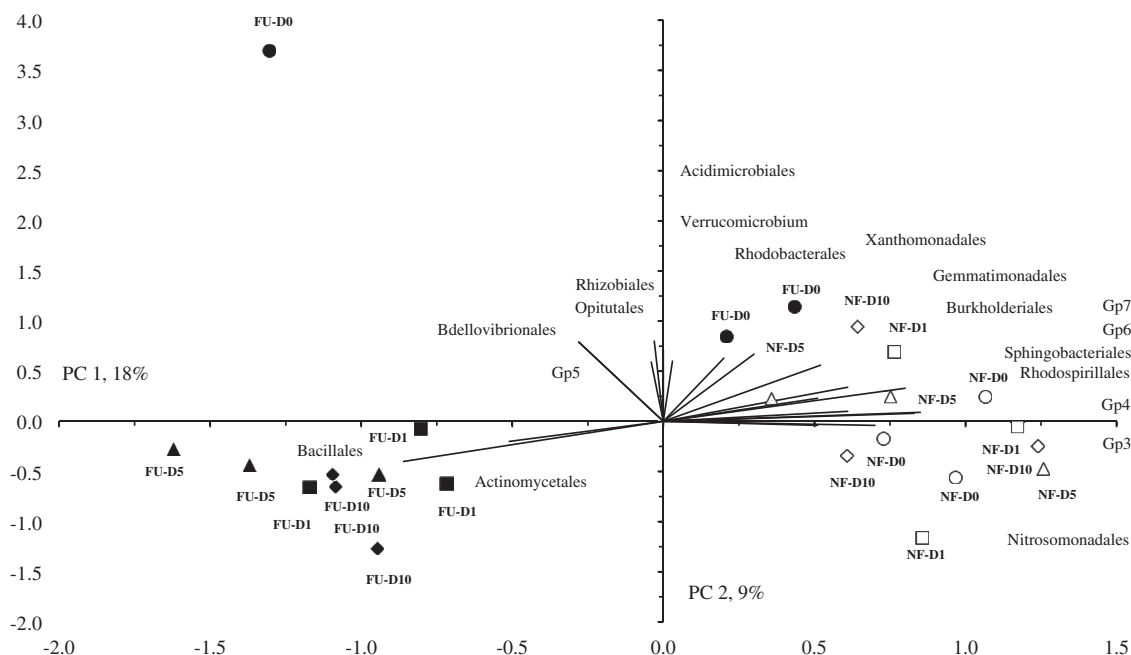
*Micromonospora* has been described as one of the most important genus of the soil actinobacterias (Genilloud, 2012). However, it is frequently reported as endophyte and nodulating bacteria or in aquatic ecosystems (de Menezes et al., 2012). *Micromonospora*s are well known for their ability to degrade complex recalcitrant materials, such as cellulose and chitin (Jendrossek et al., 1997; Wohl and McArthur, 2001; Talia et al., 2012) and produce acid and alkaline phosphatases (El-Tarabily et al., 2008).

Strains belonging to the genus *Cohnella* were found during various phases of composting (He et al., 2013). *Cohnella* strains isolated from coffee residue compost contributed to an efficient cellulolytic and hemicellulolytic processes during composting, and they consisted 2.2% in a lignin degradation bacterial consortium mentioned earlier (Wang et al., 2013).

Although the abundance of phylotypes belonging to the genus *Micromonospora* increased >50-times in the fumigated soil compared to the unfumigated soil, the relative abundance of other phylotypes belonging to Actinobacteria decreased significantly, i.e. *Marmoricola* ( $P < 0.05$ ), while the relative abundance of 44 other groups belonging to the Actinobacteria were not affected by the

fumigation process. It is thus clear that not all phylotypes belonging to the same group respond in the same way to fumigation. For instance, the relative abundance of phylotypes belonging to the genus *Marmoricola* was so low in the fumigated soil that they were not detectable, while they consisted of 0.3% of all sequences in the unfumigated soil.

Strains belonging to the phyla Bacteroidetes and Proteobacteria are generally described as responding easily to labile carbon sources (Fierer et al., 2007). Fumigating the soil had a large effect on phylotypes belonging to the Bacteroidetes. Bacteroidetes were the second most abundant phylum in the arable soil (14.6%), but their relative abundance dropped to 1.4% in the fumigated soil. Strains belonging to the  $\alpha$ - and  $\beta$ -Proteobacteria are well known to respond to easily decomposable organic material and are generally considered to be r-strategist, or weedy fast-growing microbiota whose populations fluctuate opportunistically (Fierer et al., 2007). In this study, however, they did not colonize the fumigated arable soil or metabolize the liberated microbial organic material as the relative abundance of phylotypes belonging to the Proteobacteria, the most abundant phylum in the unfumigated arable soil (41.7%), dropped to 22.6% in the fumigated soil. The fumigation reduced each of the different classes of Proteobacteria, except for the  $\alpha$ -Proteobacteria. The relative abundance of the  $\beta$ -Proteobacteria dropped to 3.8% in the fumigated soil compared to 10.4% in the unfumigated soil. The relative abundance of the Burkholderiales ( $\beta$ -Proteobacteria) increases normally in response to labile substrates (Goldfarb et al., 2011), but they were not favored by the liberation of organic material in the fumigated soil as their relative abundance dropped from 8.9% in the unfumigated soil to 3.3% in the fumigated soil. Mature compost was found to be preferred by  $\delta$ -Proteobacteria, so the decrease in their relative abundance from 3.0% in the unfumigated soil to 0.5% in the fumigated soil can be expected.  $\gamma$ -Proteobacteria, however, were found in the earlier stages of composting (de Gannes et al., 2013), but in this study they were not involved in metabolizing cell material or



**Fig. 4.** Principal component analysis considering the percentages of the different orders in the fumigated and unfumigated soil incubated aerobically for 10 days. Unfumigated soil incubated for 0 (○), 1 (□), 5 (△) or 10 days (◇) and fumigated soil incubated for 0 (●), 1 (■), 5 (▲) or 10 days (◆).

colonizing the fumigated soil as their relative abundance dropped from 4.0% in the unfumigated soil to 0.9% in the fumigated soil.

Although the above mentioned Proteobacteria and Bacteroidetes are favored generally by the application of easily decomposable material, their relative abundance did not increase after fumigating the soil. It might be that they lack the necessary enzymes to degrade the killed microorganisms, e.g. chitinase and 1,3-beta glucanase to degrade polysaccharide components of the fungal cell, but most likely they are not the first degraders of these components released after fumigation of the microbial biomass or the available easily decomposable soil organic material; they grow slower than phylotypes belonging to the genera *Micromonospora*, *Bacillus*, *Paenibacillus* and *Cohnella* under the given conditions. It has to be remembered that the increase in the relative abundance of the Firmicutes occurred within a day, so the augmentation of the phylotypes belonging to the Proteobacteria, Bacteroidetes and other groups was slower and they will only later on, i.e. after 10 days, replace the fast growing Firmicutes. The graph showing the changes in relative abundance of the different groups affected significantly by fumigation already indicates this as the relative abundance of the Proteobacteria starts to increase at day 10 compared to day 5.

## 5. Conclusion

It was found that fumigating an arable soil, reduced the relative abundance of phylotypes belonging to different bacterial groups, but increased the relative abundance of only four genera. The relative abundance of phylotypes belonging to the *Micromonospora* (Actinomycetales, Actinobacteria) increased significantly from 0.2% in the unfumigated soil to 6.7% in the fumigated soil and that of *Bacillus* from 3.6% to 40.8%, *Cohnella* from undetectable amounts to 0.6% and *Paenibacillus* from 0.3% to 4.2%. The genera *Bacillus*, *Paenibacillus* and *Micromonospora* are metabolic versatile with different phylotypes producing proteases, phosphatases, chitinases, cellulases and showing lignolytic activity. This gives them the metabolic capacity to degrade the killed microorganisms and recolonize the fumigated soil. Less is known about the metabolic characteristics of the genus *Cohnella*, although strains belonging to this genus have shown cellulolytic and hemicellulolytic activity, and degrade lignin.

The relative abundance of phylotypes belonging to the Acidobacteria, Bacteroidetes, Chloroflexi, Gemmatimonadetes and Proteobacteria ( $\alpha$ -,  $\beta$ -,  $\delta$ - and  $\gamma$ -Proteobacteria) were significantly lower in the fumigated than in the unfumigated soil and in most of them the relative abundance of different bacterial groups (i.e. Gp3, Gp4, Gp6, Sphingobacteriales, Gemmatimonadales, Rhodospirillales, Burkholderiales, Xanthomonadales) was reduced strongly. It might be that they lack the necessary enzyme to degrade the organic material released after lysis of the microbial cells, but most likely their growth is slower than that of phylotypes belonging to the genera *Bacillus*, *Paenibacillus*, *Micromonospora* and *Cohnella* in the arable soil under the given conditions.

## Acknowledgments

This research was funded by Cinvestav and 'Apoyo Especial para Fortalecimiento de Doctorado PNPC 2013' from 'Consejo Nacional de Ciencia y Tecnología' (CONACYT, Mexico). The authors thank ABACUS-CONACYT. Y.E.N.-N. received a postdoctoral grant from CONACYT and ABACUS and J.M. B.-L. a postdoctoral grant from CONACYT. C.A. D.-M., A.S.dL.-L., L.D.-B., S. G.-A., V.M. R.-V. and D.A. R.-V. received grant-aided support from CONACYT.

## Appendix A. Supplementary data

Supplementary data related to this article can be found at <http://dx.doi.org/10.1016/j.soilbio.2014.02.012>.

## References

- Abera, G., Wolde-meskel, E., Bakken, L.R., 2012. Carbon and nitrogen mineralization dynamics in different soils of the tropics amended with legume residues and contrasting soil moisture contents. *Biology and Fertility of Soils* 48, 51–66.
- Acinas, S.G., Klepac-Ceraj, V., Hunt, D.E., Pharino, C., Ceraj, I., Distel, D.L., Polz, M.F., 2004. Fine-scale phylogenetic architecture of a complex bacterial community. *Nature* 430, 551–553.
- Aguilar-Chávez, A., Díaz-Rojas, M., Cárdenas-Aquino, M.R., Dendooven, L., Luna-Guido, M., 2012. Greenhouse gas emissions from a wastewater sludge-amended soil cultivated with wheat (*Triticum* spp. L.) as affected by different application rates of charcoal. *Soil Biology and Biochemistry* 52, 90–95.
- Bandounas, L., Wierckx, N.J.P., de Winde, J.H., Ruijsenaars, H.J., 2011. Isolation and characterization of novel bacterial strains exhibiting ligninolytic potential. *BMC Biotechnology* 11, 94.
- Cano, R.J., Borucki, M.K., 1995. Revival and identification of bacterial spores in 25- to 40-million-year-old dominican amber. *Science* 268, 1060–1064.
- Caporaso, J.G., Bittinger, K., Bushman, F.D., DeSantis, T.Z., Andersen, G.L., Knight, R., 2010a. PyNAST: a flexible tool for aligning sequences to a template alignment. *Bioinformatics* 26, 266e267.
- Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K., Fierer, N., Peña, A.G., Goodrich, J.K., Gordon, J.I., Huttley, G.A., Kelley, S.T., Knights, D., Koenig, J.E., Ley, R.E., Lozupone, C.A., McDonald, D., Muegge, B.D., Pirrung, M., Reeder, J., Sevinsky, J.R., Turnbaugh, P.J., Walters, W.A., Widmann, J., Yatsunenko, T., Zaneveld, J., Knight, R., 2010b. QIIME allows analysis of high-throughput community sequencing data. *Nature Methods* 7, 335–336.
- Ceja-Navarro, J.A., Rivera-Orduña, F.N., Patiño-Zúñiga, L., Vila-Sanjurjo, A., Crossa, J., Govaerts, B., Dendooven, L., 2010. Phylogenetic and multivariate analyses to determine the effects of different tillage and residue management practices on soil bacterial communities. *Applied and Environmental Microbiology* 76, 3685–3691.
- Chandra, R., Raj, A., Purohit, H.J., Kapley, A., 2007. Characterization and optimization of three potential aerobic bacterial strains for kraft lignin degradation from pulp paper waste. *Chemosphere* 67, 839–846.
- Cucurachi, M., Busconi, M., Marudelli, M., Soffritti, G., Fogher, C., 2013. Direct amplification of new cellulase genes from woodland soil purified DNA. *Molecular Biology Reports* 40, 4317–4325.
- de Gannes, V., Eudoxie, G., Hickey, W.J., 2013. Prokaryotic successions and diversity in composts as revealed by 454-pyrosequencing. *Bioresource Technology* 133, 573–580.
- de Menezes, A.B., McDonald, J.E., Allison, H.E., McCarthy, A.J., 2012. Importance of *Micromonospora* spp. as colonizers of cellulose in freshwater lakes as demonstrated by quantitative reverse transcriptase PCR of 16S rRNA. *Applied and Environmental Microbiology* 78, 3495–3499.
- D'haeseleer, P., Gladden, J.M., Allgaier, M., Chain, P.S.G., Tringe, S.G., Malfatti, S.A., Aldrich, J.T., Nicora, C.D., Robinson, E.W., Paša-Tolić, L., Hugenholtz, P., Simmons, B.A., Singer, S.W., 2013. Proteogenomic analysis of a thermophilic bacterial consortium adapted to deconstruct switchgrass. *PLoS One* 8, e68465.
- Dungait, J.A.J., Kemmitt, S.J., Michallon, L., Guo, S., Wen, Q., Brookes, P.C., 2013. The variable response of soil microorganisms to trace concentrations of low molecular weight organic substrates of increasing complexity. *Soil Biology & Biochemistry* 64, 57–64.
- Edgar, R.C., 2010. Search and clustering orders of magnitude faster than BLAST. *Bioinformatics* 26, 2460–2461.
- Eida, M.F., Nagaoka, T., Kouno, K., 2012. Isolation and characterization of cellulose-decomposing bacteria inhabiting sawdust and coffee residue composts. *Microbes and Environments* 27, 226–233.
- El-Katatny, M.H., Gudeli, M., Robra, K.H., Elnaghy, M.A., Gubitz, G.M., 2001. Characterization of a chitinase and an endo-beta-1,3-glucanase from *Trichoderma harzianum* Rifai T24 involved in control of the phytopathogen *Sclerotium rolfsii*. *Applied Microbiology and Biotechnology* 56, 137–143.
- El-Tarabily, K.A., Nassar, A.H., Sivasithamparan, K., 2008. Promotion of growth of bean (*Phaseolus vulgaris* L.) in a calcareous soil by a phosphate-solubilizing, rhizosphere-competent isolate of *Micromonospora endolithica*. *Applied Soil Ecology* 39, 161–171.
- España, M., Rasche, F., Kandeler, E., Brune, T., Rodriguez, B., Bending, G.D., Cadisch, G., 2011. Identification of active bacteria involved in decomposition of complex maize and soybean residues in a tropical vertisol using <sup>15</sup>N-DNA stable isotope probing. *Pedobiologia* 54, 187–193.
- Fierer, N., Bradford, M.A., Jackson, R.B., 2007. Toward an ecological classification of soil bacteria. *Ecology* 88, 1354–1364.
- Fierer, N., Strickland, M.S., Liptzin, D., Bradford, M.A., Cleveland, C.C., 2009. Global patterns in belowground communities. *Ecology Letters* 12, 1238–1249.
- García-Fraile, P., Velázquez, E., Mateos, P.F., Martínez-Molina, E., Rivas, R., 2008. *Cohnella phaseoli* sp. nov., isolated from root nodules of *Phaseolus coccineus* in Spain, and emended description of the genus *Cohnella*. *International Journal of Systematic and Evolutionary Microbiology* 58, 1855–1859.

- Genilloud, O., 2012. Genus I. Micromonospora. In: Goodfellow, M., Kämpfer, P., Busse, H.-J., Trujillo, M.E., Suzuki, K.I., Ludwig, W., Whitman, W.B. (Eds.), *Bergey's Manual of Systematic Bacteriology: the Actinobacteria*, second ed. Bergey's Manual Trust, Springer, Athens, GA, pp. 1039–1057.
- Goldfarb, K.C., Karaoz, U., Hanson, C.A., Santee, C.A., Bradford, M.A., Treseder, K.K., Wallenstein, M.D., Brodie, E.L., 2011. Differential growth responses of soil bacterial taxa to carbon substrates of varying chemical recalcitrance. *Frontiers in Microbiology* 2, 94.
- Haas, B.J., Gevers, D., Earl, A.M., Feldgarden, M., Ward, D.V., Giannoukos, G., Ciulla, D., Tabbaa, D., Highlander, S.K., Sodergren, E., Methé, B., DeSantis, T.Z., The Human Microbiome Consortium, Petrosino, J.F., Knight, R., Birren, B.W., 2011. Chimeric 16S rRNA sequence formation and detection in Sanger and 454-pyrosequenced PCR amplicons. *Genome Research* 21, 494e504.
- Hamady, M., Lozupone, C., Knight, R., 2010. Fast UniFrac: facilitating high-throughput phylogenetic analyses of microbial communities including analysis of pyrosequencing and PhyloChip data. *The ISME Journal* 4, 17–27.
- Hardisson, C., Suarez, J.E., 1979. Fine structure of spore formation and germination in *Micromonospora chacea*. *Journal of General Microbiology* 110, 233–237.
- Hatayama, K., Shoun, H., Ueda, Y., Nakamura, A., 2006. *Tuberibacillus calidus* gen. nov., sp. nov., isolated from a compost pile and reclassification of *Bacillus naganensis* Tomimura et al. 1990 as *Pullulanibacillus naganensis* gen. nov., comb. nov. and *Bacillus laevolacticus* Andersch et al. 1994 as *Sporolactobacillus laevolacticus* comb. nov. *International Journal of Systematic and Evolutionary Microbiology* 56, 2545–2551.
- He, Y.M., Xie, K.Z., Xu, P.Z., Huang, X., Gu, W.J., Zhang, F.B., Tang, S.H., 2013. Evolution of microbial community diversity and enzymatic activity during composting. *Research in Microbiology* 164, 189–198.
- Jendrossek, D., Tomasi, G., Kroppenstedt, R.M., 1997. Bacterial degradation of natural rubber: a privilege of actinomycetes? *FEMS Microbiology Letters* 150, 179–188.
- Jenkinson, D.S., Powelson, D.S., 1976. The effects of biocidal treatments on metabolism in soil. V. A method for measuring soil biomass. *Soil Biology and Biochemistry* 8, 209–213.
- Kemmit, S.J., Lanyon, C.V., Waite, I.S., Wen, Q., Addiscott, T.M., Bird, N.R.A., O'Donnell, A.G., Brookes, P.C., 2008. Mineralization of native soil organic matter is not regulated by the size, activity or composition of the soil microbial biomass—a new perspective. *Soil Biology & Biochemistry* 40, 61–73.
- Kim, B.C., Lee, K.H., Kim, M.N., Kim, E.M., Rhee, M.S., Kwon, O.Y., Shin, K.S., 2009. *Paenibacillus pinihi* sp. nov., a cellulolytic bacterium isolated from the rhizosphere of *Pinus densiflora*. *Journal of Microbiology* 47, 530–535.
- Kirchman, D.L., 2012. *Processes in Microbial Ecology*, first ed. Oxford University Press Inc., New York.
- Ko, C.H., Lin, Z.P., Tu, J., Tsai, C.H., Liu, C.C., Chen, H.T., Wang, T.P., 2010a. Xylanase production by *Paenibacillus campinasensis* BL11 and its pretreatment of hardwood kraft pulp bleaching. *International Biodeterioration and Biodegradation* 64, 13–19.
- Ko, C.H., Tsai, C.H., Lin, P.H., Chang, K.C., Tu, J., Wang, Y.N., Yang, C.Y., 2010b. Characterization and pulp refining activity of a *Paenibacillus campinasensis* cellulase expressed in *Escherichia coli*. *Bioresource Technology* 101, 7882–7888.
- Kolter, R., Siegele, D.A., Tormo, A., 1993. The stationary phase of the bacterial life cycle. *Annual Review of Microbiology* 47, 855–874.
- Martins, L.F., Antunes, L.P., Pascon, R.C., de Oliveira, J.C.F., Digiampietri, L.A., Barbosa, D., Peixoto, B.M., Vallim, M.A., Viana-Niero, C., Ostroski, E.H., Telles, G.P., Dias, Z., da Cruz, J.B., Juliano, L., Verjovski-Almeida, S., da Silva, A.M., Setubal, J.C., 2013. Metagenomic analysis of a tropical composting operation at the São Paulo Zoo Park reveals diversity of biomass degradation functions and organisms. *PLoS One* 8, e61928.
- Mead, D.A., Lucas, S., Copeland, A., Lapidus, A., Cheng, J.F., Bruce, D.C., Goodwin, L.A., Pitluck, S., Chertkov, O., Zhang, X., Detter, J.C., Han, C.S., Tapia, R., Land, M., Hauser, L.J., Chang, Y.J., Kyrpides, K.C., Ivanova, N.N., Ovchinnikova, G., Woyke, T., Brumm, C., Hochstein, R., Schoenfeld, T., Brumm, P., 2012. Complete genome sequence of *Paenibacillus* strain Y4.12MC10, a novel *Paenibacillus lautus* strain isolated from obsidian hot spring in Yellowstone National Park. *Standards in Genomic Sciences* 6, 381–400.
- Mueller, T., Joergensen, R.G., Meyer, B., 1992. Estimation of soil microbial biomass-C in the presence of living roots by fumigation extraction. *Soil Biology and Biochemistry* 24, 179–181.
- Navarro-Noya, Y.E., Gómez-Acata, S., Montoya-Ciriaco, N., Rojas-Valdez, A., Suárez-Arriaga, M.C., Valenzuela-Encinas, C., Jiménez-Bueno, N., Verhulst, N., Govaerts, B., Dendooven, L., 2013. Relative impacts of tillage, residue management and crop-rotation on soil bacterial communities in a semi-arid agroecosystem. *Soil Biology and Biochemistry* 65, 86–95.
- Partanen, P., Hultman, J., Paulin, L., Auvinen, P., Romantschuk, M., 2010. Bacterial diversity at different stages of the composting process. *BMC Microbiology* 10, 94.
- Pathma, J., Sakthivel, N., 2013. Molecular and functional characterization of bacteria isolated from straw and goat manure based vermicompost. *Applied Soil Ecology* 70, 33–47.
- Price, M.N., Dehal, P.S., Arkin, A.P., 2009. FastTree: computing large minimum evolution trees with profiles instead of a distance matrix. *Molecular Biology and Evolution* 26, 1641–1650.
- Rousk, J., Brookes, P.C., Bååth, E., 2009. Contrasting soil pH effects on fungal and bacterial growth suggests functional redundancy in carbon mineralization. *Applied and Environmental Microbiology* 75, 1589–1596.
- Ruiz-Valdiviezo, V.M., Luna-Guido, M., Galzy, A., Gutiérrez-Miceli, F.A., Dendooven, L., 2010. Greenhouse gas emissions and C and N mineralization in soils of Chiapas (México) amended with leaves of piñón (*Jatropha curcas* L.). *Applied Soil Ecology* 46, 17–25.
- SAS Institute, 1989. *Statistic Guide for Personal Computers*, Version 6.04. SAS Institute, Cary.
- Sirota-Madi, A., Olender, T., Helman, Y., Ingham, C., Brainis, I., Roth, D., Hagi, E., Brodsky, L., Leshkowitz, D., Galatenko, V., Nikolaev, V., Mugasimangalam, R.C., Bransburg-Zabary, S., Gutnick, D.L., Lancet, D., Ben-Jacob, E., 2010. Genome sequence of the pattern forming *Paenibacillus vortex* bacterium reveals potential for thriving in complex environments. *BMC Genomics* 11, 710.
- Talia, P., Sede, S.M., Campos, E., Rorig, M., Principi, D., Tosto, D., Hopp, H.E., Grasso, D., Cataldi, A., 2012. Biodiversity characterization of cellulolytic bacteria present on native Chaco soil by comparison of ribosomal RNA genes. *Research in Microbiology* 163, 221–232.
- Wang, Q., Garrity, G.M., Tiedje, J.M., Cole, J.R., 2007. Native bayesian classifier for rapid assignment of rRNA sequences into the new bacterial taxonomy. *Applied and Environmental Microbiology* 73, 5261–5267.
- Wang, Y.X., Liu, Q., Yan, L., Gao, Y.M., Wang, Y.J., Wang, W.D., 2013. A novel lignin degradation bacterial consortium for efficient pulping. *Bioresource Technology* 139, 113–119.
- Wohl, D.L., McArthur, J.V., 2001. Aquatic actinomycete-fungal interactions and their effects on organic matter decomposition: a microcosm study. *Microbial Ecology* 42, 446–457.