

A. GULF OF MAINE ATLANTIC COD (*GADUS MORHUA*) STOCK ASSESSMENT FOR 2012, UPDATED THROUGH 2011

Executive Summary

TOR 1. Estimate catch from all sources including landings and discards. Characterize the uncertainty in these sources of data and take into account the recommendations and subsequent work from the March 2012 MRIP workshop. Evaluate available information on discard mortality and, if appropriate, update mortality rates applied to discard components of the catch.

Since 1964, catch of Gulf of Maine Atlantic cod has ranged from 3,242 mt to 22,272 mt. Recent catches over the past five years have ranged from approximately 5,500 mt to 8,400 mt. Catch estimates prior to 1981 do not include commercial discards or estimates of recreational removals. Given the smaller mesh sizes and lower minimum retention sizes that existed pre-1977, commercial discards could have been substantial, particularly given the presence of several strong year classes in the 1970s. Since 1982, commercial landings have been the largest source of fishery removals, comprising 40-90% of the total catch. Commercial discards constituted a large proportion of the catch between 1998 and 2003 when trip limits ranged from 30-500 lb/day (13.6 – 226.8 kg/day). Since 2006 commercial discards have accounted for <10% of the total catch and <3% of the catch since 2010. Major uncertainties in the commercial catch include the mis-allocation of commercial landings stemming from industry mis-reporting of statistical area and uncertainty in the discard estimation method. The uncertainty with respect to mis-reporting is estimated to be small (5%). In recent years precision of the estimated discards has been high with coefficients of variation (CV) <20%. Beginning with the SAW 53 assessment, the Gulf of Maine cod assessment has included hindcasted commercial discard back to 1982; however, the uncertainty on these estimates is unknown.

A notable contraction of both the commercial trawl and gillnet fleets has been observed since the mid-1990's. Generally, the fishery has become highly concentrated in the western Gulf of Maine, exhibiting similar trends as those observed in the resource as a whole as evidenced from fishery-independent surveys. Between 2006 and 2010, there was an intense aggregation of the commercial fishery within a small geographic area of approximately 260 km² (<0.5% of the total Gulf of Maine surface area). By 2010, this area (known as ten minute square '427044') was responsible for >45% of the total commercial landings. There are several likely causes for this concentration in the fishery including concentration of the cod resource as well as regulatory changes. These factors are described in more depth under TOR2.

There is a large recreational fishery in the Gulf of Maine that, over the last decade, has accounted for approximately 20-31% of the total catch. Previous assessments have used data collected under the Marine Recreational Fisheries Statistical Survey (MRFSS). Beginning with this current assessment, MRFSS data have been re-estimated using revised methodologies consistent with the new Marine Recreational Information Program (MRIP) which has replaced the MRFSS program. The revised MRIP recreational catch estimates are approximately 25% lower than the MRFSS estimates pre-2003 and range from 4% higher to 50% lower between 2004 and 2011.

With increases in the recreational minimum retention size, the discard component has become an increasingly important component of recreational catch with discards more than

two times greater than the recreational landings in terms of numbers of fish between 2006 and 2011. This assessment includes revised estimates of the survival of discarded fish, with only 30% of recreationally released fish estimated to die. The true percentage of recreational discards suffering mortality remains a key source of uncertainty in the estimate of recreational removals. The uncertainty associated with the estimates of total recreational catch is on the order of 10-25% in terms of percent standard error (PSE). An additional source of uncertainty is the age composition of recreational discards prior to 2005. Beginning with the SAW 53 assessment, the recreational discard length frequency distributions were hindcasted to 1981 in an effort to incorporate recreational discards into an age-based assessment.

As noted previously, the current assessment incorporates revised estimates of the mortality of fish discarded in both the commercial and recreational fishery. The previous assessment of this stock (SAW 53, NEFSC 2012) assumed 100% mortality of all discarded fish. The revised estimates are a product of a Discard Mortality Working Group (DMWG) convened in July 2012 to evaluate the available scientific information on the survival of cod on a gear-by-gear basis. The working group consisted of scientific experts with experience in field estimation of discard survival and stock assessments as well as both recreational and commercial fishermen and other industry representative. The revised mortality estimates developed by the DMWG ranged from 20-80% depending on gear type. The impacts of the revised mortality rates on the total estimates of fishery removals are most pronounced in the recreational fishery where 30% of discarded cod are estimated to die. In the commercial fishery, where discards are dominated by otter trawl and sink gillnet gear, the revised discards mortality estimates had a much smaller impact since 75% and 80% of the fish discarded by otter trawl and sink gillnet gear are estimated to die.

TOR 2. Present the survey data and calibration information being used in the assessment (e.g., indices of abundance, recruitment, state surveys, age-length data, etc.). Consider model-based (e.g. GLM) as well as design-based analyses of the survey data in developing trends in relative abundance. Investigate the utility of commercial or recreational LPUE as a measure of relative abundance. Characterize the uncertainty and any bias in these sources of data.

The Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl surveys began in 1968 and 1963 respectively, providing a long time series of fishery independent indices. Age-specific indices for Gulf of Maine cod began in 1970. All previous Gulf of Maine cod assessments have used only the offshore survey strata. The aggregate indices of abundance (numbers) and biomass have generally declined since time series highs in the 1960/1970s. Current indices are at, or near, all time lows. The number of stations and strata where cod have been observed in the Gulf of Maine has generally decreased over time as the resource has become increasingly concentrated in the western Gulf of Maine. It appears that two related, but separate, processes may be underway with respect to the concentration of the resource. Over the longer term, there has been a loss of cod from the eastern and central Gulf of Maine with an apparent concentration of cod in the western area. In addition to this, since 2006 there has been a further aggregation of cod within the western Gulf of Maine into highly localized areas which are hypothesized to be driven by prey availability. While it is difficult to prove definitely that these processes are responsible for the observed distribution changes, the evidence is suggestive.

The impacts of including the inshore survey strata in the NEFSC survey indices were examined by the 55th Stock Assessment Workshop Working Group (SAW 55 WG). The overall trends in both the aggregate and age specific indices of older fish was not markedly different with the inclusion of the inshore strata, and more importantly, there was inconsistent sampling of the inshore survey strata throughout the time series which impedes the construction of consistent and stable survey indices. For this reason, and because the inshore areas that were sampled by the NEFSC survey are largely covered by the Massachusetts Department of Marine Fisheries (MADMF) bottom trawl survey, the SAW 55 WG concluded that the status quo should be maintained with the inshore strata excluded from NEFSC indices. The NEFSC survey vessel was replaced in spring 2009 resulting in changes to the survey protocol. Calibration experiments to estimate differences in catchability between the two survey series were conducted and peer-reviewed. Length based calibration models were used to express the 2009-2012 NEFSC indices in units equivalent to the longer time series. Preliminary attempts to estimate length-based survey calibration factors internally within a Statistical Catch at Age (SCAA) assessment model were conducted and reviewed by the SAW 55 WG and found to be very similar to the externally estimated calibration factors used in previous assessment. While the SAW 55 WG generally supported the internally estimated approach as a longer-term research recommendation, given the high level of agreement between the internally and externally estimated calibration coefficients the SAW 55 WG supported continued use of the existing calibration coefficients.

The SAW 55 WG also considered model-based estimates of the NEFSC survey indices as opposed to the design-based estimators that have been employed in past assessments. The model-based estimates were based on a generalized linear model (GLM) that attempted to standardize for multiple factors including stratum, time of day and depth. Overall there was a high degree of agreement between the GLM-based estimates and the design-based estimates; however, the variability about the GLM-based estimates was considerably higher than the design-based estimates. The SAW 55 WG was concerned about the incorporation of GLM-based smoothed indices into the assessment model, which then effectively applies an additional smooth as it fits the survey index. Given the similarity of the indices, the increased variability in the GLM-based indices, and concerns over the use of smoothed series in assessment models, the SAW 55 WG concluded that the existing design-based indices be used as inputs to the assessment model.

The MADMF bottom trawl survey began in 1978, with two surveys (spring and fall) conducted annually. Age-specific indices are available beginning in 1982. The MADMF fall survey catches very few older fish and there is poor cohort tracking within the survey. For this reason, the MADMF fall survey is not used in the Gulf of Maine cod assessment. MADMF spring biomass index is currently at a time series lows and the abundance (numbers) index is the third lowest observed. Similar to what has been observed in the NEFSC survey, the number of stations and stratum in which cod have been observed has declined over time.

The SAW 55 WG spent considerable time evaluating catch per unit effort indices and their utility as indices of abundance within the Gulf of Maine cod stock assessment. A number of analyses were undertaken to describe Gulf of Maine cod distributional changes, which particularly since 2006, appear to have been driven by fine-scale spatial processes of prey (primarily sand lance). A number of surveys indicate that the Stellwagen Bank area appears to be a foraging ‘hot spot’ for cod feeding on sand lance. Additionally, the VTR, observer

and VMS information from the commercial fishery indicates that fishing effort since the mid-2000s has become concentrated in this area. Over the longer term, there have been a number of regulatory changes (e.g. seasonal closures, trip limits, etc) which call into question the utility of commercial LPUE as an index of GOM cod biomass. Based on these concerns, the SAW 55 WG recommended that the commercial LPUE index not be used in the SAW 55 assessment model. This recommendation is consistent with the findings of the recent NEFSC sponsored LPUE workshop. Given concerns comparable to those of the commercial fishery, the SAW 55 WG recommended that the recreational LPUE index also not be included in the GOM cod assessment model. It should be noted that sensitivity runs were conducted which incorporated LPUE indices and these model results are similar to those of the base model (described in Appendix A.6).

The SAW 55 WG also evaluated data from the Maine – New Hampshire (ME/NH) inshore groundfish survey which began in the fall of 2000. Because of lack of age-specific information and the short time series of the survey, the survey was not included in the assessment models. Progress has been made on the implementation and analysis of the data collected since the start of the survey; specifically, spring and fall 2005 and spring 2011 ageing has been completed and spring 2006 is in progress (S. Sherman, ME DMR, pers. comm.). The SAW 55 WG recommended that the complete ageing of the entire time series of collected otoliths be considered a high priority.

TOR 3. Summarize the findings of recent workshops on stock structure of cod of the Northeastern US and Atlantic Canada. Summarize the findings of recent workshops on stock structure of cod of the Northeastern US and Atlantic Canada.

A work plan on the topic of Atlantic cod stock structure in the Northeast United States/Scotian Shelf region was recommended by the New England Fishery Management Council's Scientific and Statistical Committee (SSC). The work plan laid out a three-phase process for re-evaluating, and possibly revising, the spatial basis for assessment and management of Atlantic cod. The first phase was to review data (genetic, life history, tagging, etc.) in order to evaluate the “null hypothesis” of the status quo management units.

The NEFSC sponsored a public workshop on cod stock structure, held June 12-14, 2012, facilitated by the Gulf of Maine Research Institute to address Phase I. Invited participants from the fishing and scientific communities presented on a range of topics with opportunities for discussion. The full workshop report is available at <http://www.gmri.org/minis/index.asp?ID=52&p=149>.

Many of the workshop participants felt that there was compelling evidence that the current management units need to be revised. The Workshop did not reach any conclusions on what the most appropriate management units might be. This will require further data analysis and modeling in order to complete Phase I of the SSC recommended process. The workshop report also identifies gaps in the data and analyses and recommended action to address them.

The Workshop did not explicitly address and propose the next steps in the process. The Steering Committee recommended that an inclusive, but focused, Working Group meeting be held involving a small group of Canadian and US scientists to consider the results of the Workshop. This Working Group should be provided the short-term data and analyses identified as missing by the Workshop. Using that information, as well as the conclusions

from the Workshop, the Working Group should determine the most appropriate representations of biological stock structure to complete Phase I of the process. The results from this Working Group meeting should be evaluated through an independent peer-review process.

Since the phased review process of cod stock structure that was recommended by the SSC has not been completed, no changes to stock structure were incorporated into this assessment.

TOR 4. Investigate the evidence for natural mortality rates which are time- and/or age-specific. If appropriate, integrate these into the stock assessment (TOR 5).

Previous assessments of Gulf of Maine cod have assumed a constant, age-invariant rate of instantaneous natural mortality (M) of 0.2. The SAW 55 WG evaluated the sufficiency of this assumption through life history analyses of natural mortality. From the meta-analysis of life history-based estimates, the evidence available with respect to Gulf of Maine cod life history parameters suggests that an assumption of $M = 0.2$ is reasonable. It should be noted that maximum age as high as 16 has been observed in the commercial fishery as recently as 2009 which suggests comparable natural mortalities relative to earlier in the time series. Also, examinations of maturity-at-age and condition factor over time show no evidence of strong trends both of which can relate to changes in natural mortality.

The method of Lorenzen (1996) was used to provide an aged-based estimate of M . This method, which is based upon the relationship between body weight and M across a wide range of species, was used in SAW 54 to provide age-based estimates of M for Southern New England – Mid Atlantic Bight yellowtail flounder. The peer review panel of SAW 54 (O’Boyle et al. 2012) considered that applying an inter-species relationship to infer within-species dynamics was an over-interpretation of the method. While M no doubt may be age-specific, the pattern estimated from the Lorenzen method may not be appropriate.

Two working papers considered the predator field of cod in the Gulf of Maine area (Link 2012, Waring 2012). Link (2012) noted that directed piscivory of cod by other fish was not common, with fewer than 200 cod observed in over 550,000 stomachs examined. Similarly, the evidence for cannibalism is weak with only 20 cod found in over 20,000 stomachs. Studies to date suggest that M due to fish predation is likely low and is focused on juvenile and smaller size groups (Smith and Link 2010). Waring (2012) considered marine mammals as a potential source of elevated M in the Gulf of Maine area. Four species of seals (Harbor, Grey, Harp and Hooded) are found in New England with Harbor and Grey seals being the most numerous. The Harbor seal population, which was about 38,000 individuals in 2001, has been growing at an annual rate 6.6%. The Grey seal herd has increased from tens of animals in the early 1980s to thousands of animals in the late 2000s. Firm estimates on the size of the current herds are not available. Notwithstanding this, the food habit research suggests that cod mortality due to seals is low. Additionally, while seals are known to prey on cod, they are generalist feeders and the importance of cod in the diet of Gulf of Maine grey seals is unknown. There is limited information that suggests that cod represent only a minor component of harbor seal diet along the Maine coast (Wood 2001).

An analysis of tagging data collected during 2003 – 2006 to jointly estimate natural and fishing mortality was undertaken during GARM III (Miller and Tallack 2007). This analysis was updated for SAW 55 (Miller 2012). Contrary to the earlier work, this analysis was not

length-based. Estimates of M ranged 0.4 – 0.7 for the Gulf of Maine. It also provided evidence of significant cod movements between GOM and GB and area 4X on the order of 4.1 to 29.7%. While M was relatively high compared to current estimates, F was comparatively low, prompting discussion on whether or not it was representative of the fishery due to local effects. The results were sensitive to the assumptions on the return rate of high-reward tags. High-reward return rates on the order of 50% were associated with Gulf of Maine cod M estimates of 0.3, with M increasing as the high-reward tag rate increased. Model preference (based on log-likelihood function) was for assumptions of near-100% on reporting rates of the high-reward tags. Estimates of fishing mortality, F , were inversely related to the M response with F declining with higher assumptions of high-reward tags reporting rates. Across the full range of high-reward tag reporting rates total mortality (Z) was estimated at approximately 1.0.

Concerns were raised with the tagging conducted in the Cape Cod area, which represented over 50% of the date in the database. The tagging had been conducted employing a wide range of expertise with mostly small cod tagged. This in combination with the warm water in the area may have resulted in higher tag induced mortality than assumed in the model. There were additional concerns with the assumed tag reporting rate (100%) for high reward tags. There is evidence to suggest differential reporting rates among some sectors of the commercial fishery, most notably the reporting rate by gillnet vessels was five times lower than that of trawl vessels (Tallack 2006). It is unknown if these same reporting trends also apply to the high-reward tags. There was also discussion on the age groups of cod represented by the study. GOM cod of 50 cm are approximately 2.5 – 3 years old, implying that the estimates of M are for ages 2.5 – 3 plus with it weighted towards the younger ages.

The SAW 55 WG discussed how best to use these estimates of M . It was hesitant to conclude that M was in the range of 0.6 – 0.7 and to recommend that these estimates be directly included in the assessment models. Rather, the tagging analysis is another form of modeling that should be considered. The SAW 55 WG discussed the availability of historical tagging to which the current estimates could be compared. It was reported that tagging work conducted in the Gulf of Maine area during the 1970s and 1980s suggested M estimates in the range of 0.2 – 0.3 whereas tagging in the 1990s was suggestive of M similar to the more recent results. These observations are based upon unpublished work that could not be corroborated at the meeting. Much of the historical work (e.g. Hunt et al. 1999) had been focused on cod movements and did not provide estimates of natural, fishing or total mortality. Further, concerns were raised that there was no obvious mechanism (e.g. predation) that could explain a recent increase in M . While counter arguments were raised that no mechanism has been identified for the current M estimate of 0.2, it should be noted that this estimate is supported by life history parameters. The SAW 55 WG recommended profiling natural mortality across both the historical and more recent periods of the assessment to inform the discussion as to whether or not there has been a long-term change in M . The SAW 55 WG agreed that an option with an M change should be considered as an alternate to a base assessment model which would assume no change in M (i.e. $M = 0.2$).

TOR 5. Estimate annual fishing mortality, recruitment and stock biomass (both total and spawning stock) for the time series, and estimate their uncertainty. Consider feasibility of survey catchability estimates, the starting year for the assessment, estimation of the stock recruitment curve, inclusion of multiple fleets, and whether to use domed or flat selectivity-at-age for the NEFSC surveys. Provide a summary of steps in the model building process. Include a historical retrospective analysis to allow a

comparison with previous assessment results. Review the performance of historical projections with respect to stock size, catch recruitment and fishing mortality.

There were several changes to the input data and the impacts of these changes on the existing SAW 53 model have been documented in this report. The primary changes to the data inputs were the revised recreational catch estimates, updated assumptions about the mortality of discarded fish and minor updates to the MADMF spring survey indices-at-age. The data updates resulted in only a -54 mt difference in the 2010 estimate of spawning stock biomass (< 1% difference). The data updates did result in moderate differences in the terminal estimate of fishing mortality due to the revisions to recreational catch and discard mortality assumptions. The combined effect of these revisions adjusted the SAW 53 estimate of 2010 age 5 fishing mortality downward from 1.14 to 0.67. Revising the discard mortality assumption increased the retrospective patterning associated with spawning stock biomass, fishing mortality and age 1 recruitment. This increase in the retrospective pattern may suggest that the revised discard mortality estimates underestimate the true mortality.

In addition to the data updates, there have been changes to the model formulation of the ASAP model. The most notable change is the move from two to three fishery selectivity blocks and the assumption of flat-topped selectivity in the fishery compared to the SAW 53 model which allowed fishery selectivity to be freely estimated. There was also a minor change in the functional form used to estimate selectivity for the MADMF spring survey. Whereas the SAW 53 model used a double logistic function to fit age 1-9 indices, the approach used in the revised ASAP model utilizes a non-parametric approach with the selectivity at ages 1-6 estimated independently.

The SAW 55 WG selected four different models for review by the SARC 55 Panel. Between the four models, there were two issues in terms of the science that arose in the SAW 55 WG that resulted in significant differences in interpretation of the Gulf of Maine cod assessment, different assessment results, and consequently led to lack of consensus. The first issue involved the use of data prior to 1982 in conducting the assessment and in determining the stock recruitment relationship based on an assessment using this data. The second issue involved whether or not natural mortality was changing in the Gulf of Maine system. Two SCAA models were put forward which evaluated the performance of models using pre-1982 with internal stock recruit relationship to assumptions of a) constant natural mortality; and, b) natural mortality ramping (linearly) from a constant 0.2 in the years pre-1989 to a constant natural mortality of 0.4 from 2003 onward. Similarly, two ASAP models were put forward which used only data from 1982 onward but explored the two natural mortality scenarios.

With respect to the first issue, the SARC 55 Panel expressed a number of concerns regarding the use of pre-1982 data and the fitting of a stock-recruit function, but ultimately the Panel discounted the results and eliminated the SCAA approach from further consideration. Given all of the information provided to the Panel, there remained considerable uncertainty in the estimates of M . The evidence for and against constant and ramped natural mortality was equivocal. As with the Working Group, the Panel was unable to reach a decision on which natural mortality values or time varying scenarios best characterized this system. The SARC 55 Panel recognized that one of the motivations for examining how, or if, changes in natural mortality had occurred was driven by an effort to reduce the retrospective pattern present in the $M = 0.2$ model. However, the Panel concluded that “*...finding that including a changing M provides a better fit, is generally not sufficient to justify using such a model modification without other ecologically directed information to back it up*” (SARC 55 Panel Summary

Report, 2012). Noting the lack of conclusive evidence to support a change in M they determined that it was unclear as to whether a change in natural mortality was influencing the retrospective pattern or some other factor. For example, a Delphi method had been applied prior to the working group meetings to find alternative values of discard mortality rates for different gears. The retrospective pattern was worse with the lower discard mortality rates, implying that the ramp M approach could be partially aliasing unaccounted fishing mortality.

Given that there was no clear way forward for providing a single model for guiding management advice, the SARC 55 Panel put forward (accepted) both the ASAP $M = 0.2$ and M -ramp models. The consequences associated with using or disregarding either approach are outlined under TOR 8.

The assessment results for the two ASAP models accepted by the SARC Panel are as follows:

$M = 0.2$ ASAP model

The ASAP $M = 0.2$ (SAW55_3BLOCK_BASE) assessment model indicates that total SSB has ranged from 6,268 mt to 22,036 mt during the assessment time period, with current SSB in 2011 estimated at 9,903 mt (90% posterior probability 7,644 – 13,503 mt). The base model estimates SSB in 2010 at 11,141 mt which is 6% lower than the SAW 53 estimate of 11,868 mt. Total January 1 biomass in 2011 is estimated at 14,728 mt (90% posterior probability 11,890 – 19,149 mt) and F's at the end of the time series are estimated between 0.75 and 1.00 with the 2011 fully recruited, $F_{full} = 0.86$ (90% posterior probability 0.53 – 1.05).

Recruitment over the past decade has been poor despite modest increases in SSB. Age-1 recruitment has not exceeded 10 million fish in the last two decades and has been below 7 hundred thousand fish over the last decade. The five highest recruitment events in the time series were spawned during a six year period from 1982 to 1987 where the SSB was near the highest observed in the time series, averaging over 14,000 mt annually. The current population structure is comprised primarily of fish that have not yet fully recruited to the fishery (fish age 1-3), with >80% of the population age 4 and younger.

Retrospective analysis for the 2004-2011 terminal years indicates retrospective error in both F and SSB with the tendency for the model to underestimate F and overestimate SSB. The 5-year Mohn's rho value for SSB and F were 0.40 and -0.27 respectively. While the retrospective pattern is larger than that observed in the SAW53 model, the directionality in the terminal year has shifted such that spawning stock biomass tended to be underestimated and fishing mortality overestimate. It appeared that the retrospective pattern was transient with a one year peel showing no bias. Both the SAW 55 WG and SARC 55 Panel agreed that no adjustment be made for retrospective pattern given that the retrospective pattern is small, it may be transient in nature and that SAW 53 made no retrospective adjustment.

M -ramp ASAP model

The ASAP M -ramp (SAW55_3BLOCK_BASE_M_SPLIT) assessment model indicates that total SSB has ranged from 7,930 mt to 21,531 mt during the assessment time period, with current SSB in 2011 estimated at 10,221 mt (90% posterior probability 7,943 – 13,676 mt). Total January 1 biomass in 2011 is estimated at 16,312 mt (90% posterior probability 13,173 – 20,771 mt) and F's at the end of the time series are estimated between 0.60 and 0.90 with the 2011 fully recruited, $F_{full} = 0.90$ (90% posterior probability 0.57 – 1.09).

Recruitment over the past decade has been poor to moderate despite modest increases in SSB. Age-1 recruitment has been below ten thousand fish since 2008. The current population structure is comprised primarily of fish that have not yet fully recruited to the fishery (fish age 1-3), with >80% of the population age 4 and younger.

Retrospective analysis for the 2004-2011 terminal years indicates retrospective error in both F and SSB with the tendency for the model to underestimate F and overestimate SSB. The 5-year Mohn's rho value for SSB and F were -0.01 and 0.06 respectively. The retrospective error is considerably reduced relative to the $M = 0.2$ (SAW55_3BLOCK_BASE) model. Both the SAW 55 WG and SARC 55 Panel agreed that no retrospective adjustment should be conducted for the purposes of stock status determination or short-term projections.

TOR 6. State the existing stock status definitions for “overfished” and “overfishing”. Then update or redefine biological reference points (BRPs; point estimates or proxies for B_{MSY} , $B_{THRESHOLD}$, F_{MSY} , and MSY) and provide estimates of their uncertainty. Consider alternative parametric models of the stock recruitment relationship. If analytic model-based estimates are unavailable, consider recommending alternative measurable proxies for BRPs. Comment on the appropriateness of existing BRPs and any “new” (i.e., updated, redefined, or alternative) BRPs.

The existing MSY reference points based on a spawning potential ratio (SPR) of 40% were established at SAW 53 (NEFSC 2012). The overfishing definition is $F_{MSYproxy} = F_{40\%} = 0.20$. A stock is considered to be overfished if spawning biomass is less than half of SSB_{MSY} . The existing overfished definition is $\frac{1}{2} SSB_{40\%} = 0.5 \cdot 61,218 \text{ mt} = 30,609 \text{ mt}$. New reference points are warranted given the changes in fishery selectivity and fishery weights-at-age due to the revisions in recreational catch estimates and discard mortality assumptions. Additionally the M -ramp assumption has considerable impacts on recruitment estimates which will impact the estimation of SSB_{MSY} and MSY.

Analytic model-based reference points are not estimable because of insufficient contrast in the ASAP base model time series of estimated SSB and recruitment (1982-2011). As no standard stock-recruitment relationship could be found, the use of proxy reference points for this stock was necessary. A yield per recruit (YPR) analysis was performed using a 3-year average of weights-at-age (2009-2011) which was consistent with the approach used in SAW 53 and supported by recent observed trends. The remaining YPR inputs were time invariant (maturity-at-age) or were constant in the most recent time block of the assessment model (selectivity, natural mortality). The SARC 55 Panel concluded that for long-term projections (i.e., the establishment of reference points) natural mortality should be assumed equal to 0.2, because the longer-term historical evidence seems to indicate that $M=0.2$ is more plausible than the more recent 0.4 assumed under the M -ramp model. Given the SARC 55 Panel's conclusions regarding natural mortality, there are only minor differences in the selectivity vectors between the $M = 0.2$ and M -ramp YPR inputs; all other inputs are identical. YPR inputs are summarized in Table A.93 for both the $M= 0.2$ and M -ramp models.

The basis for the existing reference points was derived at GARM III (NEFSC 2008), and is based on $F_{40\%}$. The SARC 55 Panel recommended to maintain the $F_{40\%}$ basis for reference points for both the $M = 0.2$ and M -ramp models but noted that “*We do not suggest that $F_{40\%}$ is necessarily the best proxy to use, rather there has yet to be compelling reasons to abandon*

it" (SARC 55 Panel Summary Report, 2012).

To arrive at estimates for SSB_{MSY} and a corresponding MSY, long term projections were run sampling from the empirical distribution of recruitment estimates from the preferred ASAP model. The recruitment vector included years 1982-2009; recruitment in 2010 and 2011 were not included due to their greater variance. The projection model samples from a cumulative density function derived from estimated age-1 recruitment. However, the revised model adjusts projected recruitment when SSB falls below some specified spawning biomass threshold based on a linear function that declines to zero at zero spawning stock biomass. Consistent with the SAW 53 assessment, the 'hinge' was set at the lowest observed SSB in the time series. For the $M = 0.2$ scenario, this was 6,300 mt and 7,900 mt for the M-ramp scenario. To approximate the distribution of the SSB and MSY distributions, the long term projections were made from 1000 estimates of numbers at age in 2011, which were estimated by performing MCMC simulation of the ASAP models (described above under TOR 5). The 2011 age 1 estimates were based on sampling from the empirical distribution of recruitment estimates from only the ten year period 2000-2009. All projections were conducted with the AGEPRO software (Age Structured Projection Model v4.1).

For the ASAP, 1982 start, $M = 0.2$ scenario, the resulting reference points and their 90% confidence intervals corresponding to $F_{MSYproxy} = F_{40\%}$ (0.18) are SSB_{MSY} = 54,743 mt (40,207 – 73,354 mt) and MSY = 9,399 mt (6,806 – 13,153 mt).

For the ASAP, 1982 start, M -ramp scenario, the resulting reference points and their 90% confidence intervals corresponding to $F_{MSYproxy} = F_{40\%}$ (0.18) are SSB_{MSY} = 80,200 mt (64,081 – 99,972 mt) and MSY = 13,786 mt (10,900 – 17,329 mt).

TOR 7. Evaluate stock status with respect to the existing model (from the most recent accepted peer reviewed assessment) and with respect to a new model developed for this peer review. In both cases, evaluate whether the stock is rebuilt.

- a. When working with the existing model, update it with new data and evaluate stock status (overfished and overfishing) with respect to the existing BRP estimates.

The updated SAW 53 model (SAW55_BASE) estimates 2011 SSB at 11,874 mt. This is less than the existing overfished threshold of 30,609 mt; therefore, the stock is overfished. The updated estimate of fully recruited fishing mortality (F_{full}) in 2011 is 0.59. This is greater than the overfishing limit of 0.20, and therefore, overfishing is occurring.

- b. Then use the newly proposed model and evaluate stock status with respect to "new" BRPs (from Cod TOR-6).

For the ASAP, 1982 start, $M = 0.2$ scenario, the revised reference points are $F_{MSYproxy} = F_{40\%} = 0.18$ and SSB_{MSY} = 54,743 mt (0.5 x SSB_{MSY} = 27,372 mt). The model estimates 2011 SSB at 9,903 mt. This is less than the overfished threshold of 27,372 mt; therefore, the stock is overfished. The estimate of 2011 fully recruited fishing mortality (F_{full}) is 0.86. This is greater than the overfishing limit of 0.18, and therefore, overfishing is occurring.

For the ASAP, 1982 start, M -ramp scenario, the revised reference points are $F_{MSYproxy} = F_{40\%} = 0.18$ and $SSB_{MSY} = 80,200$ mt ($0.5 \times SSB_{MSY} = 40,100$ mt). The model estimates 2011 SSB at 10,221 mt. This is less than the overfished threshold of 40,100 mt; therefore, the stock is overfished. The estimate of 2011 fully recruited fishing mortality (F_{ful}) is 0.90. This is greater than the overfishing limit of 0.18, and therefore, overfishing is occurring.

Under both the $M = 0.2$ and M -ramp scenarios the stock is assessed to be overfished and overfishing is occurring. It is notable that this stock has experienced a long history of overfishing relative to current reference points.

TOR 8. Develop and apply analytical approaches to conduct single and multi-year stock projections to compute the pdf (probability density function) of the OFL (overfishing level) and candidate ABCs (Acceptable Biological Catch; see Appendix to the SAW TORs).

- a. **Provide numerical annual projections (3-5 years). Each projection should estimate and report annual probabilities of exceeding threshold BRPs for F , and probabilities of falling below threshold BRPs for biomass. Use a sensitivity analysis approach in which a range of assumptions about the most important uncertainties in the assessment are considered (e.g., terminal year abundance, variability in recruitment).**

Short term projections of future stock status were conducted based on the current assessment results without accounting for retrospective bias. This rationale was identical to that of stock status determination. Numbers-at-age in 2012 were derived from 1000 different vectors of numbers-at-age produced from the MCMC chain with 2011 age 1 estimates based on sampling from the empirical distribution of recruitment estimates from only the ten year period 2000-2009. Biological inputs were identical to those used for reference point determination. Short term projections have used an assumed catch in 2012 of 3,767 mt. This estimate is based on the current commercial and recreational catches as well as the expected catch over the remainder of the year which has been extrapolated using the harvest trajectories from the past two years (NEFMC PDT, T. Nies pers. comm.).

Recruitment was sampled from a cumulative density function (CDF) of estimated age 1 recruitment from 1982 to 2009. The same AGEPRO model used for reference point determination was used to conduct short-term projections (i.e., model adjusts projected recruitment based on a linear function that declines to zero at zero SSB when SSB falls below some ‘hinge’ SSB-level corresponding to the lowest SSB observed in the time series). For the $M = 0.2$ scenario, the ‘hinge’ SSB value was set at 6,300 mt and 7,900 mt for the M -ramp scenario. All projections were run under the assumption of 75% F_{MSY} ($0.18 \cdot 0.75 = 0.135$).

A consequence analysis was conducted to evaluate the sensitivity of management advice to the assumptions about M (i.e. $M = 0.2$ or M -ramp). For the M -ramp scenario the projections were provided assuming that: a) M remained at 0.4; or, b) that M returns to 0.2 in the projection period.

Under 75% F_{MSY} exploitation, the stock is projected to rebuild under the $M = 0.2$ and

M -ramp ($M = 0.2$) scenarios by 2022. The stock cannot rebuild under the M -ramp ($M = 0.4$) scenario since the reference points are based on an assumption of M returning to 0.2 in the long-term. It is important to note that the SARC Panel was not willing to conclude that M would remain at 0.4 in perpetuity and so did not provide reference points for the M -ramp model under a long-term assumption of $M = 0.4$. A full discussion of the three scenarios evaluated is provided under TOR 8b.

b. Comment on which projections seem most realistic. Consider the major uncertainties in the assessment as well as sensitivity of the projections to various assumptions.

The risks associated with management actions taken during 2013 – 2015 were examined by undertaking stock projections under the competing assumptions for the state of nature. For example, if the true state of nature is that natural mortality has remained unchanged at 0.2 and that stock productivity is best reflected by the 1982 – present dataset (SPR, $M = 0.2$ model), then the consequences of management actions by setting projected catch according to 75% F_{MSY} based on the two alternative states of nature were examined (M -ramp scenario with $M = 0.2$ in short-term and M -ramp scenario with $M = 0.4$ in the short term). In all cases, the 2012 catch was provided by the NEFMC Groundfish Plan Development Team. Projections were only conducted until 2015. There may be longer term consequences which might be revealed through a more extensive analysis. This is beyond the current terms of reference.

The three states of nature considered were:

- $M = 0.2$: stock dynamics and assessment based on 1982 – present dataset with M remaining at 0.2 for the projection period.
- M -ramp: stock dynamics and assessment based on 1982 – present dataset with M returning to 0.2 in the projection period.
- M -ramp: stock dynamics and assessment based on 1982 – present dataset with M remaining at 0.4 for the projection period.

When management actions are correctly based upon a particular state of nature, a modest (5,300 – 13,000 mt) increase in SSB is projected between 2013 and 2015 for all three scenarios explored. The M -ramp ($M = 0.2$) scenario has the greatest rebuilding potential whereas the M -ramp ($M = 0.4$) has the lowest rebuilding potential. Fully recruited fishing mortality declines from 0.86 ($M = 0.2$) or 0.90 (M -ramp) to 0.14 (all scenarios). Catch declines from 6,830 mt in 2011 to 1,313 - 2,582 mt in 2015 depending on the scenario with the M -ramp ($M = 0.4$) scenario resulting in the lowest yield and the M -ramp ($M = 0.2$) having the highest yield. The $M = 0.2$ scenario is an intermediate case. If the management actions are correctly based upon the ‘true’ state of nature all scenarios indicate that the stock will be in an overfished state as of 2013.

The SARC 55 Panel concluded that the $M = 0.2$ projections and the M -ramp projections with M remaining at 0.4 in the short-term were equally realistic. Like the SAW 55 WG, the SARC 55 Panel could not decide which option was more plausible. The Panel concluded that if M is currently 0.2 [0.4] then it seemed more reasonable to assume that in the short-term M would remain at 0.2 [0.4]. Note that for long-term

projections that Review Panel decided that M should be 0.2 under all scenarios, because the longer-term historical evidence seems to indicate that $M=0.2$ is more plausible.

The consequences of mis-specifying natural mortality (e.g., $M = 0.2$ is true state of nature and manage under M -ramp, $M = 0.4$) will not impact status determination in 2013; under all consequence analyses considered the stock will be in an overfished state in 2013. Considering only the $M = 0.2$ and M -ramp ($M = 0.4$) scenarios, the consequence of mis-specifying natural mortality will result in at most 717 mt of an over-/under-harvest of fishery yield in 2015. While the magnitude is small in terms of historical catch, this amounts to 55% of over-harvest (M -ramp is true state of nature and manage under $M = 0.2$) or a 35% under-harvest ($M = 0.2$ is true state of nature and manage under M -ramp, $M = 0.4$). Assuming an M -ramp ($M = 0.4$) when M is actually equal to 0.2 results in a lower than ‘planned’ fishing mortality and catch and higher than ‘planned’ SSB. When M is assumed to be 0.2 but an M -ramp ($M = 0.4$) is correct, fishing mortality and thus catch would be considerably higher than ‘planned’ with the result that in 2013 the stock would be experiencing overfishing.

c. Describe this stock’s vulnerability (see “Appendix to the SAW TORs”) to becoming overfished, and how this could affect the choice of ABC.

The Gulf of Maine cod stock is currently undergoing processes that have not been incorporated into the analytical formulations. Nevertheless, they should be considered when setting the ABC.

Since the mid-1990s, as observed in the NEFSC bottom trawl surveys and consistent with the trends in the fishery, the distribution of cod has become increasingly concentrated in the western part of the Gulf, with a gradual loss of cod from the coastal and central Gulf. Since the mid-2000s, the stock has become particularly concentrated in a small region of the western Gulf, an area which appears to be a forage ‘hotspot’ due to the presence of sand lance, a prey of cod. This biases CPUE as an indicator of the abundance of the stock as a whole.

There is uncertainty associated with natural mortality rates. Natural mortality of cod may be increasing through consumption by other fishes and marine mammals as these populations increase; however, evidence of this is lacking in the food habits data and among life history parameters. On the other hand, tagging studies suggest natural mortality levels higher than 0.2 during 2003 – 2006 time period. The tagging studies, combined with the reduced assessment model retrospective patterns were the basis of the M -ramp model. However, the states of nature as reflected in the natural mortality rates included in the models are uncertain. For example, a Delphi method had been applied prior to the working group meetings to find alternative values of discard mortality rates for different gears. The retrospective pattern was worse with the lower discard mortality rates, implying that the ramp M approach could be partially aliasing unaccounted fishing mortality.

It may be that at low population sizes, cod experience mortality from a number of unidentified sources. High mortality, both fishing and natural will lead to a truncated age structure, implying that spawning success is increasing dependent upon younger

individuals. Murawski et al. (2001) suggest that reproduction by older females is more successful than by young females. There are a number of other factors that are known to negatively influence cod spawning success at low population sizes (Rowe et al., 2004).

If weak recruitment and low reproductive rates of Gulf of Maine cod continue, productivity and rebuilding of the stock will be less than projected. Over the last five years recruitment estimates have declined to a low level in both the $M = 0.2$ and M -ramp assessment models. Recent survey indices of recruitment indicate continued poor recruitment. Additionally, the NEFSC 2011 fall and 2012 spring survey abundance indices were the 4th lowest and the lowest in their respective time series. The MADMF 2012 spring survey biomass index was the lowest in its times series. The 2012 spring survey observations were not incorporated into the assessment formulations, implying that projections may be optimistic.

The current assessment provides a range of views of current stock status, all of which indicate that the resource is in an overfished state and has experienced a long history of overfishing. Concerns for stock status may also be apparent in the fishery. Cumulative commercial and recreational catches to date in 2012 are projected to be less than 60% of the total allocated quota (based on projected catch provided by NEFMC PDT, T. Nies pers. comm.). While this is suggestive of an overall difficulty on the part of the commercial fishery to locate Gulf of Maine cod it is not definitive given other possible explanations such as sector quota restrictions on other co-occurring species. However, observations from the recreational fishery which is not subject to the same catch share system as the commercial fishery has also reported difficulty locating Gulf of Maine cod.

TOR 9. Review, evaluate and report on the status of the SARC and Working Group research recommendations listed in recent SARC reviewed assessments and review panel reports. Identify new research recommendations.

The SAW 55 WG reviewed the status of previous research recommendations and proposed new ones to address issues raised during the three SAW 55 WG meetings. There was a single research recommendation carried forward from GARM III which has been addressed in this report. Of the nine research recommendations brought forward from SAW 53, six have been either partially or fully addressed. The remaining research recommendations from SAW 53 include estimation of cod bycatch in both the nearshore and offshore lobster fishery, ageing of the backlog of otoliths collected from the Maine – New Hampshire inshore groundfish survey and the re-evaluation of Atlantic cod stock structure in the northeast region. There is currently work in progress to address all three of these research recommendations, but progress was not sufficient to inform this assessment.

The SAW 55 WG proposed eight new research recommendations which primarily focus on improving estimates of natural mortality and the survival of post-capture fish as well as advances in assessment methods. All new research recommendations proposed by the SAW 55 WG have been assigned relative priorities (high, medium, low) as appropriate. Many of these recommendations were felt to be common to both the Gulf of Maine and Georges Bank Atlantic cod stocks and are labeled as ‘general’. The SARC 55 Panel also contributed seven additional research recommendations which are included in this section.

SAW 55 Terms of Reference for Gulf of Maine Atlantic cod (*Gadus morhua*)

1. Estimate catch from all sources including landings and discards. Characterize the uncertainty in these sources of data and take into account the recommendations and subsequent work from the March 2012 MRIP workshop. Evaluate available information on discard mortality and, if appropriate, update mortality rates applied to discard components of the catch.
2. Present the survey data and calibration information being used in the assessment (e.g., indices of abundance, recruitment, state surveys, age-length data, etc.). Consider model-based (e.g. GLM) as well as design-based analyses of the survey data in developing trends in relative abundance. Investigate the utility of commercial or recreational LPUE as a measure of relative abundance. Characterize the uncertainty and any bias in these sources of data.
3. Summarize the findings of recent workshops on stock structure of cod of the Northeastern US and Atlantic Canada.
4. Investigate the evidence for natural mortality rates which are time- and/or age-specific. If appropriate, integrate these into the stock assessment (TOR 5).
5. Estimate annual fishing mortality, recruitment and stock biomass (both total and spawning stock) for the time series, and estimate their uncertainty. Consider feasibility of survey catchability estimates, the starting year for the assessment, estimation of the stock recruitment curve, inclusion of multiple fleets, and whether to use domed or flat selectivity-at-age for the NEFSC surveys. Provide a summary of steps in the model building process. Include a historical retrospective analysis to allow a comparison with previous assessment results. Review the performance of historical projections with respect to stock size, catch recruitment and fishing mortality.
6. State the existing stock status definitions for “overfished” and “overfishing”. Then update or redefine biological reference points (BRPs; point estimates or proxies for B_{MSY} , $B_{THRESHOLD}$, F_{MSY} , and MSY) and provide estimates of their uncertainty. Consider alternative parametric models of the stock recruitment relationship. If analytic model-based estimates are unavailable, consider recommending alternative measurable proxies for BRPs. Comment on the appropriateness of existing BRPs and any “new” (i.e., updated, redefined, or alternative) BRPs.
7. Evaluate stock status with respect to the existing model (from the most recent accepted peer reviewed assessment) and with respect to a new model developed for this peer review. In both cases, evaluate whether the stock is rebuilt.
 - a. When working with the existing model, update it with new data and evaluate stock status (overfished and overfishing) with respect to the existing BRP estimates.
 - b. Then use the newly proposed model and evaluate stock status with respect to “new” BRPs (from Cod TOR-6).
8. Develop and apply analytical approaches to conduct single and multi-year stock projections to compute the pdf (probability density function) of the OFL (overfishing level) and candidate ABCs (Acceptable Biological Catch; see Appendix to the SAW TORs).

- a. Provide numerical annual projections (3-5 years). Each projection should estimate and report annual probabilities of exceeding threshold BRPs for F, and probabilities of falling below threshold BRPs for biomass. Use a sensitivity analysis approach in which a range of assumptions about the most important uncertainties in the assessment are considered (e.g., terminal year abundance, variability in recruitment).
 - b. Comment on which projections seem most realistic. Consider the major uncertainties in the assessment as well as sensitivity of the projections to various assumptions.
 - c. Describe this stock's vulnerability (see "Appendix to the SAW TORs") to becoming overfished, and how this could affect the choice of ABC.
9. Review, evaluate and report on the status of the SARC and Working Group research recommendations listed in recent SARC reviewed assessments and review panel reports. Identify new research recommendations.

Introduction

The 55th Stock Assessment Workshop Working Group (SAW 55 WG) prepared the assessment. The working group held three meetings during 27 August – 2 November 2012. The meeting dates and locations are listed below. Working group participation varied by meeting. A complete list of working group participants can be found in Appendix A.1.

- SAW 55 Data Working Group Meeting
 - August 27-31, 2012
 - Northeast Fisheries Science Center (NEFSC), Woods Hole, MA
- SAW 55 Models Working Group Meeting
 - October 15-19, 2012
 - Northeast Fisheries Science Center (NEFSC), Woods Hole, MA
- SAW 55 Models and Biological Reference Points Working Group Meeting
 - October 30-November 2, 2012
 - Northeast Fisheries Science Center (NEFSC), Woods Hole, MA

Assessment history

The initial analytical assessment of the Gulf of Maine stock was conducted using a virtual population analysis (VPA) model by Serchuk and Wigley (1986) and presented at the 7th Northeast Fisheries Science Center (NEFSC) Stock Assessment Workshop (SAW) in 1988 (NEFSC 1989). Subsequently, the stock was reviewed again at SAW 12, 15, 19, and 24 (NEFSC 1991, 1993, 1995, 1997, 1998; Mayo 1995, 1998, Mayo et al. 1993, 1998, 2002). Additionally, interim assessments were reviewed outside of the SAW framework by the Northern Demersal Working Group in July 1999 (NEFSC 2000) and again in August 2000 (NEFSC 2001a).

Amendment 4 (1991) to the Multispecies Fisheries Management Plan implemented $F_{20\%}$ as an overfishing mortality threshold for Gulf of Maine cod. Estimates of $F_{20\%}$ and F_{max} are shown below (*note $F_{20\%}$ was not reported in the SAW 7 documents):

| Stock assessment workshop | Year | $F_{20\%}$ | F_{max} | Model type | Notes |
|---------------------------|------|------------|-----------|------------|--------------------------|
| SAW 7 | 1988 | | 0.27 | VPA | Commercial landings only |
| SAW 12 | 1991 | 0.40 | 0.27 | VPA | Commercial landings only |
| SAW 15 | 1993 | 0.36 | 0.25 | VPA | Commercial landings only |
| SAW 19 | 1995 | 0.35 | 0.27 | VPA | Commercial landings only |
| SAW 24 | 1997 | 0.37 | 0.29 | VPA | Commercial landings only |
| SAW 27 | 1998 | 0.39 | 0.29 | VPA | Commercial landings only |

The 1996 re-authorization of Magnuson-Stevens Conservation and Management Act required the redefinition of overfishing and overfished with respect to the rate of fishing mortality associated with producing maximum sustainable yield. SAW 27 provided estimates of F_{MSY} and B_{MSY} based on the ASPIC surplus production model with survey catchability coefficients conditioned on biomass estimates from the SAW 27 VPA. These estimates were mean age 1⁺ $B_{MSY} = 33,000$ mt (total biomass) and age 1⁺ total biomass weighted $F_{MSY} = 0.31$. This method was used in the Report of the Overfishing Definition Review Panel (Applegate et al. 1998) and the corresponding reference points were adopted in Amendment 9 to the multispecies

FMP. The biomass threshold was set at $\frac{1}{4} B_{MSY}$ (8,300 t).

In the last eleven years, the Gulf of Maine cod stock has undergone five peer-reviewed assessments: SAW 33 (NEFSC 2001), the Groundfish Assessment Review Meeting (GARM, NEFSC 2002), GARM II (NEFSC 2005), GARM III (NEFSC 2008) and SAW 53 (NEFSC 2012a). Summaries of these assessments and the resulting stock status are provided in Tables A.1 and A.2. All of these assessments, with the exception of SAW 53 were conducted using the ADAPT VPA model. All assessments began the assessment time series in 1982 which corresponds to the availability of age data from the commercial fishery. The data inputs from SAW 33 through GARM II were nearly identical, with GARM I and II representing updates to the SAW 33 model inputs. Commercial discards were accounted for by increasing the total landings by 500 mt increments; the size of the increase was determined based on the estimated discards. This method assumes that the discarded fraction of the catch is of the same size composition as the landed catch. In the existence of trip limits, this assumption may be appropriate, but when discarding is occurring primarily as a result of minimum retention sizes, such a method may incorrectly characterize the age composition of the catch. The SAW 53 assessment included direct estimates of commercial discards since 1989 as well as including hindcasted estimates for the years 1982-1988. Recreational landings were included in these assessments prior to SAW 53, but recreational discards were not. Beginning with the SAW 53, estimates of recreational discards were included in the fishery removals. Additionally, prior to SAW 53 catch and stock weights-at-age were estimated solely from the landed fraction of the catch. When discards due to minimum sizes restrictions contribute a sizeable fraction of overall removals, this method has the potential to overestimate stock biomass. The weights-at-age used as inputs to SAW 53 provided a more realistic characterization of fish weights relative to the approach taken in previous assessments.

SAW 33 included catch through 2000 and survey indices through 2001 (spring only). SAW 33 re-evaluated reference points using an age based production model with a Beverton-Holt stock recruit relationship (NEFSC 2001b). Reference points were estimated as total stock age 1^+ total biomass $B_{MSY}=90,300$ mt, $SSB_{MSY}=78,000$ mt, and $F_{MSY}=0.23$. The SAW 33 assessment concluded that Gulf of Maine cod were not over fished, but overfishing was occurring. It is noteworthy that the stock status determination applied at SAW 33 was different than the current basis. For SAW 33 the overfished definition was based on $\frac{1}{4} B_{MSY}$ criteria (Applegate et al. 1998) unlike the $\frac{1}{2} SSB_{MSY}$ that was later adopted by the Working Group on Re-estimation of Biological Reference Points for New England Groundfish (NEFSC 2002b). The 2001 total stock biomass was estimated at 24,000 mt (18,000 mt SSB); just over 25% of B_{MSY} . Fishing mortality (F) was estimated at 0.73 which was over three times higher than F_{MSY} .

The Working Group on Re-evaluation of Biological Reference points for New England Groundfish (NEFSC 2002a) further revised Gulf of Maine cod reference points; SSB_{MSY} was revised to 82,800 mt based on change in the period used to derive mean stock weights. F remained unchanged. Amendment 13 (2004) to the Multispecies FMP adopted the Working Group's revised reference points ($SSB_{MSY}=82,800$ mt, $F_{MSY}=0.23$). The biomass threshold was revised to $\frac{1}{2} SSB_{MSY}$ (41,400 t). GARM I updated the data inputs by one year (through 2001) using the same VPA formulation as SAW 33. Spawning stock biomass in 2001 was estimated at 22,040 mt, approximately 25% of SSB_{MSY} . F was estimated at 0.47, two times greater than F_{MSY} . As of 2002 Gulf of Maine cod were overfished and overfishing was occurring. GARM II was a three year update (through 2004) to the GARM I assessment. Biological reference points remained unchanged from GARM I. Spawning stock biomass had

declined to 18,800 mt in 2004 and F had increased to 0.63. The stock complex was still overfished and overfishing was occurring. The GARM II assessment exhibited a retrospective pattern in both F and SSB, with a tendency for F to be underestimated and SSB to be overestimated in the most recent three years.

The 2008 GARM III assessment represented a benchmark assessment update. Major changes from the previous assessments include a more thorough consideration of commercial discards and updates to the biological reference points. Unlike previous assessments where landings-at-age were increased in fixed amounts, the GARM III method applied an estimated discard ratio to the landings-at-age. While this method better characterizes the true trends in discards, it still makes the assumption that the age composition of the discards is identical to the landed fraction. It should be noted that the ratio increase in landings-at-age was only applied from 1999 to 2007. Prior to 1999, commercial discards were not accounted for. As in previous assessments, catch and stock weights-at-age were estimated solely from the landed fraction of the catch and recreational discards were not included in the catch estimates. Biological reference points were based on the non-parametric yield and SSB per recruit analysis with $F_{40\%}$ used as a proxy for F_{MSY} . The reference points were estimated as follows: $F_{MSY} = 0.237$ and $SSB_{MSY} = 58,248$ mt. Terminal year estimates of F were 0.46 and SSB was estimated to have increased to 33,877 mt. The stock was perceived to no longer be overfished, but overfishing was still occurring. The large increase in SSB was contingent on the relative strength of the 2003, and to a greater degree, the 2005 year classes. The 2005 year class was estimated at 23.9 million fish (age 1) which represented the second largest observed year class in the assessment time period. The 2005 year class was only age 2 in 2007 and had yet to enter the fishery; the 2007 estimates of partial recruitment indicated that the vulnerability of this year class to the fishery was at less than 1%. The entire strength of the 2005 year class was primarily derived from the NEFSC spring and MADMF fall survey indices.

The SAW 53 assessment was also a benchmark assessment. For SAW 53 a new assessment model (ASAP) was developed that incorporated updated estimates of the length-weight equation, maturity at age, and weights at age. Additional changes from the GARM III assessment include the inclusion of direct estimates of commercial discards and the inclusion of recreational discards as mentioned previously. Given the major changes in data that have occurred in the most recent update, the SAW 53 assessment is not entirely comparable with previous assessments. Much of the scale differences between the SAW 53 assessment and previous assessments were the result of changes to the underlying data (e.g., weights-at-age) and not as a result of the assessment or choice of model. Biological reference points were based on the non-parametric yield and SSB per recruit analysis with $F_{40\%}$ used as a proxy for F_{MSY} . The reference points were estimated as follows: $F_{MSY} = 0.20$ and $SSB_{MSY} = 61,218$ mt. Terminal year estimates of F were 1.14 and SSB was estimated at 11,868 mt. The stock was perceived to be overfished, and overfishing was occurring. The major change in estimated biomass from the GARM III assessment resulted primarily from the revised estimates of weights at age, and more importantly, from revised estimates of the strength of the 2003 and 2005 year class. With three years of additional survey and fishery data it appears that the GARM III overestimated the strength of these two year classes and subsequently the spawning stock biomass in 2007. The overestimation by the GARM III assessment was partly the result of too much confidence put on highly uncertain survey observations near the end of the time series. The ASAP model developed for SAW 53 allows direct incorporation of survey and fishery precision estimates and is more robust to uncertain data inputs.

Fisheries Management

Gulf of Maine Atlantic cod have been managed under two different management authorities in recent history. Prior to 1977 the 5Y component (statistical areas 511-515) of the stock was managed under an international treaty through the International Commission for the Northwest Atlantic Fisheries (ICNAF). Fisheries management was primarily controlled through annual total allowable catches (TACs) and minimum mesh sizes (Serchuk et al. 1994). The TACs remained constant at 10,000 mt between 1973 and 1975 followed by reductions to 8,000 mt in 1976 and then to 5,000 mt in 1977. The Magnuson Fishery Conservation and Management Act (MFMCA) was passed in 1977 and subsequently the management authority of the Gulf of Maine cod stock, as well as all other New England groundfish stocks, shifted to the New England Fishery Management Council (NEFMC).

The use of TACs continued under the NEFMC authority through 1982, with TACs dispersed among quarters and vessel tonnage classes. The early quota period was accompanied by poor catch monitoring and reported black markets for quota managed species and may have contributed to increased uncertainty over catches. The system adopted in the mid-80's had numerous exceptions and special programs to mesh and minimum size requirements that make it difficult to draw conclusions about how regulations influenced fishery selectivity. In 1982, the "Interim" Groundfish Fisheries Management Plan (FMP) was implemented which replaced the quota system (TAC) with input controls such as mesh sizes and minimum retention sizes (Table A. 3). The "Interim" FMP was replaced by the initial Groundfish FMP in 1985 which largely carried forward the existing measures from the interim FMP.

Amendment 4 to the FMP required the use of a Nordmore grate in the northern shrimp fishery as well as placing a prohibition on the retention of groundfish bycatch. Beginning with Amendment 5 (1994), there was a concerted attempt to reduce fishing effort through a days-at-sea (DAS) reduction schedule. Additionally, Amendment 5 brought about mandatory vessel reporting in the way of the Vessel Trip Reports (VTRs). Effort controls were increased under Amendment 7 through further acceleration of the DAS reduction schedule, and the addition of seasonal and year round closures in the Gulf of Maine. Between 1997 and 1999 trips limits on Gulf of Maine cod were reduced from 1000 lbs/day to 30 lbs/day. Amendment 13, implemented in May 2004, placed additional restrictions on DAS usage while allowing for the use of regular B DAS to target healthy stocks. Additionally, Amendment 13 implemented mandatory electronic reporting for all primary federally permitted seafood dealers. In 2006, Framework 42 established reference point thresholds for the 18 groundfish stocks reviewed at GARM II as well as formalized rebuilding plans for all overfished stocks ($< \frac{1}{2} SSB_{MSY}$), such as Gulf of Maine cod. Through 2010 a series of additional framework actions and interim rules placed additional restrictions on DAS usage and seasonal closures on the recreational fishery.

The effort controls first adopted in 1994 were frequently changed, making it difficult to isolate the effects of individual regulations. The use of often-changing trip limits led to increased discard rates and may have contributed to high-grading. Seasonal (rolling) and year-round closures may have limited fishery access to larger spawning fish, and strict DAS limits focused effort on easily caught nearshore cod and led to the increased use of sink gillnet gear.

In 2010, the groundfish fishery experienced a major management change with the passage of Amendment 16. Amendment 16, with the introduction of annual catch limits (ACLs),

represented a return to the use of hard TACs. Additionally, 17 new groundfish sectors were approved and those vessels not members of a groundfish sector were subject to additional cuts in DAS and restrictive trip limits. Vessels fishing under the sector management were exempt from DAS restrictions and instead, each sector was given a share of the total commercial groundfish sub-ACL. How the catch was divided up amongst sector vessels or how catch was allocated throughout the year was left to the sole discretion of the sector. One of the requirements of Amendment 16 was an increase in the overall level of observer coverage. This was accomplished using observers trained through the existing Northeast Fisheries Observer Program (NEFOP) as well as a new class of observers termed At-Sea Monitors (ASMs). The data collection protocols for ASMs were restricted to catch estimation and the collection of limited biological information (e.g., lengths). The recent shift to a catch share system in 2010 appears to have dramatically reduced discards but it is too soon to fully understand the overall impacts of the sector management system.

Since the passage of Amendment 16, two framework modifications have been made to the FMP with direct impacts on the management of Gulf of Maine cod. Framework 45 implemented in May 2011 created the Whaleback closure area in the western Gulf of Maine designed to protect spawning aggregations of cod that occur in this area. This is a seasonal closure area extending from April 1 to June 30 and applies to both the recreational and commercial fisheries. Framework 47 was implemented in May 2012 and reduced the minimum retention size in the recreational fishery to 19 inches as well as reducing the recreational bag limit to 9 fish per angler. It should be noted that Framework 47 took effect outside of the time frame being used for the current assessment which only considers data through December 31st 2011.

Biology

Stock structure

Atlantic cod (*Gadus morhua*) is a demersal gadoid species whose range in United States (US) waters extends from Cape Hatteras north to the Canadian border. Globally, Atlantic cod occur on both sides of the North Atlantic Ocean, extending southward in the eastern Atlantic to the Bay of Biscay. Within the United States Exclusive Economic Zone (EEZ) there are two recognized stocks of cod: Gulf of Maine and Georges Bank. The existing Gulf of Maine of Maine stock complex extends from the northern tip of Cape Cod east to the US/Canadian border and north to the coast of Maine (Fig. A.1).

Recent reviews of historical and contemporary tagging studies (O'Brien et al. 2005, Tallack 2007, Loehrke and Cadrian 2007) suggest that there is movement of fish between the Gulf of Maine and Georges Bank stocks with the degree of mixing < 30% (Hunt et al. 1999, Tallack 2009, Miller 2012). The SAW 55 WG reviewed some preliminary analyses evaluating possible impacts of stock mixing on assessment results (Chen and Cao 2012). Overall, the results indicated that the lack of consideration of inter-stock mixing had little impact on the GOM cod assessment results. The importance of the quality of the catch information was highlighted. The WG expressed several concerns and possible areas of improvement in the analysis. While the study is a work in progress with many assumptions and issues to be resolved, it highlighted the value of undertaking modeling to explore complex spatial processes influencing cod in the Gulf of Maine.

Several meta-analyses of the life history parameters of Atlantic cod in the region have been

conducted over the last four decades that generally support the current stock boundaries. These investigations have highlighted differences in both the growth and maturity rates between the Gulf of Maine and Georges Bank stocks (Pentilla and Gifford 1976, Begg et al. 1999). These differences are highlighted in the current assessment as well in addition to differences that have been noted in the trends of condition factor between stocks. There are recognized localized metapopulations within the Gulf of Maine (e.g., Ames 2004, Kovach et al. 2010, Siceloff and Howell 2012), between which, the degree of mixing is unknown. Additionally, there is recent work showing possible genetic connections between the Gulf of Maine and Nantucket Shoals (current considered part of the Georges Bank stock; Kovach et al. 2010). Investigations into the relative importance of these connections as well as a re-investigation of all stock structure information are under way and summarized under TOR 3. The existing stock structure boundaries constitute the best available science for the current assessment.

Length-weight relationship

The SAW 53 assessment used seasonal survey-based length-weight (LW) equations as the basis for converting catch weights to numbers-at-age (Equations 1-3, Fig. A.2). Since 1992, the NEFSC bottom trawl surveys have used digital scales to record individual fish lengths. Using these data, updated survey-based length weight equations were compared to the existing length weight equation. Both seasonal (spring/fall) and annual updates were evaluated. The use of a time-invariant LW equation is only appropriate if the LW relationship has remained stable over time. Information presented in the SAW 53 assessment report (NEFSC 2012a) showed the temporal stability of LW relationships over time. Additionally, seasonal condition factors have remained relatively stable over time (Fig. A.3), providing additional support to the use of a time-invariant LW equation.

$$(1) \quad W = 0.000004714L^{3.1741} \text{ (Spring)}$$

$$(2) \quad W = 0.000006178L^{3.1322} \text{ (Fall)}$$

$$(3) \quad W = 0.000005132L^{3.1625} \text{ (Annual)}$$

There are divergent opinions as to whether it is more appropriate to use a landings-based length-weight equation versus a survey-based length-weight equation to convert catch weights to numbers-at-age. Advocates for a landings-based derivation argue that since the fishery may catch larger (heavier) fish at length, there is the possibility that a survey-based length weight equation may be biased low, particularly at greater lengths. A survey-based approach may be preferred when a large portion of the catch is comprised of discards (or some other fraction not sampled such as recreational landings) or when the catch weights-at-age are also used to estimate stock weights due to sparse sampling of older ages in the surveys (missing or highly variable estimates of weights-at-age). In the case of Gulf of Maine Atlantic cod, the arguments for a survey-based LW relationship are valid (large fraction of catches not from commercial landings and use of catch weights to estimate stock weights). Currently in the Northeast Region, fishery surveys are the only source of individual length-weight sampling.

The suitability of applying a survey-based LW equation to commercial landings was evaluated by applying the seasonal LW relationships in equations 1 and 2 to the observed length frequency distributions of commercial biological samples collected between 1982 and 2011. The estimated weights were then compared to the recorded sample weight and the

distributions of differences were examined for the presence of bias. Examinations across years showed no evidence of strong temporal trends and across all market categories the interquartile ranges of the differences overlapped the equality line in the majority of years for both the ‘scrod’ and ‘market’ landings market categories (Fig. A.4). There was some indication that the estimated weights were greater than the recorded weights for the ‘large’ market category which could suggest that the survey LW relationships estimate heavier fish at length relative to the true relationship within the commercial landings. Interestingly, using the arguments made against the use of survey-based LW presented above, this is opposite of the expectation.

Since cod are typically landed in gutted form, a more likely explanation for the discrepancies noted in the ‘large’ market category is that the current conversion factor for converting gutted cod to its live weight equivalent is incorrectly specified. There has been an ongoing data collection effort by the NEFSC’s Cooperative Research Program to collecting information to support a re-evaluation of the established conversion factors for a variety of groundfish species. Preliminary analyses of the data collected to date suggest that a more appropriate conversion factor should be 1.20 compared to the established 1.17 that has been used to date (Table A.4). While the differences are quite small, applying the preliminary 1.20 estimate to the above analysis resolves much of the apparent bias in the large market category (Fig. A.5). The preliminary work is suggestive that a revisions may be needed to the gutted-live weight conversion factor; however the work is still preliminary and additional data are needed before the work can be finalized. It should be stressed that the small changes in the conversion factor that are suggested by the preliminary work will have negligible impact on the assessment results.

Growth and maturity

Atlantic cod in the Gulf of Maine and Georges Bank reach a maximum size around 130 cm (≈ 25 kg). Cod in the Gulf of Maine tend to grow slower than on Georges Bank (Fig. A.6 - 7). Generally, the differences in growth parameters lend support to the treatment of Gulf of Maine and Georges Bank as separate stocks. These results are consistent with that of previous research on the topic (Penttila and Gifford 1976, Begg et al. 1999). The Gulf of Maine cod von Bertalanffy growth parameters were re-estimated using NEFSC survey data from 1970 to 2012 (Equation 4). A summary of the number of ages included in the analysis are presented in Table A.5. Given the sparseness of the sampling of older ages, the L_∞ may be poorly estimated.

$$(4) \quad L = 150.93 \cdot (1 - e^{-0.11(t-0.13)}) \text{ (Annual)}$$

To examine whether there have been large scale changes in the growth rates over time as well as evidence of cohort-specific growth rates associated with year class strength, von Bertalanffy growth curves were fit to the 1960-2005 year classes. Since K and L_∞ tend to be correlated and given the limited observations when fitting growth curves on a cohort basis, L_∞ was fixed at the time series estimate of 150.93 cm. Plots of K by cohort show oscillations about the mean cohort K value of 0.12 with K values of the most recent cohorts tending to fall below the mean (Fig. A.8). Comparison of cohort K values to age 1 estimated recruitment from the SAW 53 assessment shows no discernible relationship between K and cohort strength (Fig. A.9), suggesting a lack of density dependent growth within the range of recruitment events that have been observed.

Examination of monthly trends in the mean length of Gulf of Maine cod landed in the commercial fishery suggests that the majority of somatic growth occurs between March and December, with little growth occurring January through February (Fig. A.10). Examination of mean survey weights-at-age suggests that fish size-at-age has oscillated about the long-term mean with no strong increasing or decreasing trends (Fig.'s A.11-12).

A logistic regression method (O'Brien et al. 1993) was used to fit maturity-at-age from the NEFSC spring survey data from 1970 to 2012. The number of maturity samples taken per year ranges from 23 to 229 (Table A.6). The SAW 55 WG examined the trends in annual age-at-50% maturity (A_{50} ; Fig. A.13), and determined that the estimated A_{50} (age at which 50% of fish are mature) varied about the time series average, but without any persistent trends. Based on the lack of persistent trends, the SAW 55 WG supported the use of a time-invariant maturity ogive to characterize the maturity schedule of GOM cod (Fig. A.14). The input to the stock assessment model is based on the female maturity ogive presented in Table A.7. The time series $A_{50\%}$ for male cod was 2.85 and 2.66 for females. The approach is identical to that used for the SAW 53 assessment and the estimated values are similar, with the only changes resulting from incorporation of an additional year of survey data.

The SAW 55 WG also evaluated trends in length-based maturity (Fig. A. 15). Similar to the age-based maturity examinations, the annual L_{50} estimates lacked persistent trends, though they were generally more variable than the patterns present in the A_{50} . The WG discussed possible reasons for this and concluded that this was likely due to the high variability in the length distributions observed in any given year (Fig. A.16). The time series average L_{50} was estimated at 40.8 cm for females and 42.9 cm for males. Length-based maturity ogives are presented in Figure A.17.

In relation to spawning time, genetic and growth research presented to the WG (Kovach et al. 2010, Dean et al. 2012) indicated that cod in the western Gulf of Maine are comprised of northern spring (May/June) and southern winter (December/January) spawning components, with homing of each group to their respective and distinct spawning grounds. However, there was no information available to evaluate the relative contribution of each of the spawning components. This raised the issue as to what date to use for peak spawning. The previous assumed peak was April 1st. The WG recommended that rather than defining the assessment spawning period as a 'peak' spawning date it was more appropriate to base the date used for the estimation of SSB on the mid-point between the spawning periods, which is approximately 1st April (SAW 55 WG 2012a).

Response to the Terms of Reference

TOR A.1. Estimate catch from all sources including landings and discards

Overview

In the recent period (1982 to present) total catch has ranged from 21.0 thousand metric tons (mt) to 3.1 thousand mt (Table A.8, Fig. A18). Commercial landings are the predominant source of fishery removals averaging 75% of the total catch between 1982 and 2011. The current levels of commercial landings are on similar scales with historical (pre-1982) estimates (Table A.9 and Fig. A.19). There is an indication that a higher level of commercial

landing were supported pre-1932, however these landings were estimated from a proration of combined Georges Bank and Gulf of Maine landings and have a higher degree of uncertainty than the landings estimates beginning in 1932 (Serchuk and Wigley 1992). Some research has suggested that landings of Gulf of Maine cod during the mid-1800's could have been as high as 60,000 mt (Alexander et al. 2009, Fig. A.19). Georges Bank landings have typically exceeded Gulf of Maine landings throughout much of the time series, though Gulf of Maine landings have generally exceeded those of Georges Bank over the last decade (Fig. A.20).

Prior to 1999, landings were the primary component of the total catch, constituting 63-93% of the total catch. Since 1999, landings have made up only about 46-73% of the total catch (Table A.8, Fig. A.18). There are three primary reasons for this shift: (1) significant restrictions on commercial landings leading to (2) an increase in commercial discards, and (3) increased contribution from the recreational fishery.

Beginning in 1999, commercial discards became a significant component of the catch, accounting for greater than 30% of the overall catch (Fig. A.18). Notable increases in commercial discards were primarily the result of restrictive trip limits between 1998 and 2000 (Table A.3). Trip limits were gradually relaxed from 2000 through 2004 resulting in an overall decrease in the contribution of commercial discards to the overall catch.

Recreational landings peaked in 1987, but generally, recreational landings prior to 1999 constituted approximately 13% of the overall catch, whereas they accounted for, on average, about 25% from 1999 through 2011. Recreational discards became an increasingly important component of the overall Gulf of Maine cod catch as the minimum retention size of cod was progressively increased from 15 in. in 1982 to the current size limit of 24 in., which has been in effect since 2006. The recreational minimum size was decreased to 19 in. in 2012, but that is beyond the time series used in the current assessment.

Commercial landings

In 1982, the United Nations Convention on the Law of the Sea (UNCLOS) defined a countries exclusive economic zone (EEZ) as a zone extending up to 200 nautical miles from a nation's coast. The EEZ defines the region where each country has sovereign rights to marine resources including fisheries. The geographic proximity of the US and Canada in the Gulf of Maine and Georges Bank Regions resulted in an overlap of each nation's EEZ. Given the importance of these areas with respect to resource extraction (among other reasons), the US and Canada both submitted cases to the International Court of Justice at The Hague, Netherlands seeking clarification. The Court issued a final ruling on October 12, 1984 formally delineating the US and Canadian EEZ. Hereafter, this demarcation line became informally known as the "Hague Line".

Within the Gulf of Maine, the US EEZ splits statistical areas 464, 465 and 467 (Fig. A.1). Prior to Hague line implementation, landings of cod in US ports from these statistical areas could have been either from the Gulf of Maine or Scotian Shelf stocks. Current management of Gulf of Maine cod includes catch from these areas against the fisheries ACLs. Previous assessments have not included these catches. While landings from these statistical areas have been low since 1985, accounting for less than two percent of the total Gulf of Maine landings (Fig. A.21), these landings are included in the current assessment to maintain consistency with the existing ACL monitoring programs. Based on the recommendations of the SAW 53

and 55 WGs, no attempt was made to adjust landings prior to 1985.

Since 1964, when modern catch statistics began, commercial landings of Gulf of Maine cod have ranged from 1.4 thousand mt to nearly 18 thousand mt (Tables A.8 and A.9). Landings statistics for area 5 (Gulf of Maine and part of Georges Bank stocks) exist back to 1893 (e.g., Mayo et al. 2009, Fig. A.19) and there are isolated estimates of commercial landings in the mid-1800's (Alexander et al. 2009). The methods used to apportion landings to individual stock complex prior to 1964 are not well documented and generally, these stock landings are considered less certain. Estimates of historical Gulf of Maine cod landings are of similar magnitude as landings between 1964 and 2010, though there was a period of sustained high landings above 8,000 mt prior to 1930. Total species landings are derived from the weighout reports of commercial seafood dealers and these data are generally considered a census of total landings. While un-reported landings are possible, no estimates exist to evaluate their magnitude. A secondary data source is required to apportion dealer landings to statistical area (stock) and assign basic information on fishing effort (e.g., gear, mesh, tow duration). Prior to 1994, the partitioning of stocks from total cod landings was accomplished, in part, through a port-interview process conducted by port agents working for the National Marine Fisheries Service (NMFS). The percent of Gulf of Maine cod landings attributed to interviewed trips was generally less than 40% (Fig. A.22). When trips were not interviewed, the port-agent would attribute area and fishing effort characteristics to the landings using their knowledge of the fishery and or information obtained during the interview process about vessels operating in the vicinity of the interviewed captain.

With the requirement of vessel-reported VTRs starting in 1994, the port interview process stopped and the area and effort information was inferred directly from the VTRs. Currently, a standardized procedure is used to assign area and effort from VTRs to dealer-reported landings from 1994 onward (Wigley et al. 2008). The product from this process is stored the NEFSC allocation (AA) database tables. Landings are matched to VTRs in a hierachal manner, with landings matched at the top tier (level A, direct matching) having a higher confidence in the area and fishing effort attribution than those matched at the lower tiers. The matching rates have improved over time with approximately 80% of Gulf of Maine cod landings being matched at the highest level since 2004 (Figs. A.22 and 23). Interestingly, there is a seasonal component to the matching success, with generally poor matching success around the month of May (Fig. A.24). This phenomenon has not been fully explained, but does coincide with the start of the groundfish fishing year and annual renewal of vessel permits during which reporting compliance checks have historically been conducted. The overall precision associated with the allocation process, in terms of a CV is estimated at less than 0.1 (Table A.10).

An additional area of uncertainty with stock landings stems from the mis-reporting and/or under reporting of statistical areas on VTRs. Federal regulations require that a separate VTR logbook sheet be filled out for each statistical area or gear/mesh fished. Vessels fishing in multiple statistical areas frequently under-report the number of statistical areas fished (Palmer and Wigley 2007, 2009 and 2012). Following the SAW 53 assessment of Gulf of Maine cod there was also public concern that vessels were fishing in the vicinity of Stellwagen Bank in the western Gulf of Maine, but hauling back their catches across the Gulf of Maine/Georges Bank stock boundary and reporting the catch as Georges Bank cod. Based on comparisons of VTR reports with the observer and vessel monitoring system (VMS) data, the impacts of this misreporting on Gulf of Maine landings estimates are estimated to be small. Since 2006, the magnitude of this error is $\leq 2\%$ (Table A.11). While VTR mis-reporting remains problematic,

it is not likely to be a large source of error with respect to the quantification of Gulf of Maine cod landings.

For some species, there may be a component of the catch that does not get reported by seafood dealers. In the case of Gulf of Maine cod, fish retained by the crew for home consumption are the most significant component of commercial landings that would not be reported by seafood dealers. Estimates of home consumption can be derived from VTRs, but these estimates are likely underestimates of total home consumption landings due to incomplete reporting. From 1994 to 2011, home consumption landings are estimated at \leq 0.3% of total commercial landings (Table A.12). Even if these represent underestimates, it is unlikely that home consumption landings represent a significant source of fishery removals. Because of the low magnitude, home consumption estimates are not included in estimates of commercial landings.

The commercial fishery is primarily conducted by vessels fishing trawl and gillnet gear with gillnet gear having become progressively more important over time (Fig. A.25). Current landings by trawl and gillnet gear are about equal and account for nearly 95% of the total landings. Landings by longline and handline (jig) are minor. There is a seasonal component to fleet activity in the Gulf of Maine with gillnet landings decreasing during the spring months (March through June) when parts of the western Gulf of Maine are inaccessible due to rolling closures. Larger trawl vessels which have the capacity to fish further off shore, to the east of the rolling closures, dominate the landings during the spring months (Fig. A.26).

The ports of Gloucester and Portland have historically been the primary offload ports of Gulf of Maine cod (Fig. A.27). Portland landings have declined over the last twenty years and Gloucester now accounts for over 60% of total commercial landings. The rolling closures in the western Gulf of Maine affects port landing patterns in a manner similar to their impact on the gear trends. Landings in Gloucester drop off during the months of April and May when the nearshore waters in the western Gulf of Maine are closed to groundfishing (Fig. A.28). During these months, cod are primarily landed in ports along the Maine coast. The rolling closures cycle clockwise around the western Gulf of Maine, and by June, when the rolling closures are off the coast of Maine, Gloucester again becomes the dominant port for Gulf of Maine cod landings.

The patterns for landings by statistical area are similar to the port trends. Over the last twenty years, landings have become increasingly concentrated in statistical area 514 (Fig. A.29), which is the statistical area in closest proximity to Gloucester. Landings from statistical areas to the north and east have declined. Currently, statistical area 514 accounts for >70% of total stock landings. The rolling closures have had impacted the statistical area landing patterns in manner much like the port and gear trends (Fig. A.30).

The waters of the northeast U.S. are divided into ten minutes squares which are rectangular areas ten minutes of latitude by ten minutes of longitude. Each ten minute square covers approximately 260 km², representing approximately 0.5% of the total 52,462 km² surface area of the Gulf of Maine cod management area. Annual Lorenz curves were estimated for both the commercial trawl and gillnet fishery based on the cumulative catch by ten minute square (following methods outlined in Wigley 1996). From the Lorenz curve an annual Gini index, or concentration index, can be estimated using Equation (5).

$$(5) \quad G = A/(A+B)$$

Where G is the Gini index, A is the area between 1:1 equality line and B is the area under the Lorenz curve.

Annual Gini indices were developed for both the commercial trawl and gillnet fishery based on the cumulative catch by ten minute square. The gillnet Gini index was relatively flat between 1994 and 2002 but then increased steadily through 2010, with a slight decline in 2011. Comparatively, the trawl Gini index has increased steadily over the VTR time series, though the index has been relatively flat since 2008 (Fig. A.31). The concentration in the commercial fleet is characterized by a directional shift in the catch-weighted center (centroid) of fishing activity to the southwest. The current center of fish of fishing activity is located in the western Gulf of Maine in the vicinity of 42.6° N x 70.3° W (Fig. A.32).

There has been a decline of approximately 40% in the number of ten minute squares contributing to the annual Gulf of Maine cod landings (Fig. A.33). However because approximately 60% of the landings prior to 1994 were not from interviewed trips and the spatial information from non-interviewed trips is only precise to the quarter degree square level, the true decline is likely greater. Comparison of ten minute square landings patterns from the mid 1990's to the late 2000's showed two noticeable patterns: (1) cod were being caught in fewer ten minute squares, particularly along coastal Maine, and (2) in the 1990's landings were evenly distributed across the Gulf of Maine, where as in the late 2000's landings were dominated by only a few ten minute squares in the western Gulf of Maine (Fig. A.34). The increases in the contribution of landings from a relatively small area of the Gulf of Maine are consistent with the concentration indicated by the Gini indices.

The top five most important ten minute squares throughout the VTR time series (1994-2011) were identified based on their average annual contribution to the total Gulf of Maine cod landings. The use of these ten minute squares over time in terms of total annual landings was investigated to determine if particular regions within the Gulf of Maine appeared to exhibit unique properties in terms of cod removals. These trends can be thought of in terms of a utilization index for small areas (i.e., does the fishery tend to differentially utilize certain areas and are there persistent trends?).

The top five ten minute squares with respect to annual contribution to Gulf of Maine cod landings are all located to the west of the Western Gulf of Maine Closure Area (Fig. A.35). Three of these ten minute squares (427034, 427044, 427054) correspond with Stellwagen Bank, a prominent bathymetric feature in the western Gulf of Maine. These five ten minute squares account for 10 to 65% of the total Gulf of Maine cod commercial landings in any year, with the contribution generally increasing over time (Fig. A.36). Examination of the annual trends of each of the ten minute squares shows that one ten minute square, 427044, is the predominant ten minute square, accounting for > 45% of the total commercial landings in 2010. In terms of total landings contribution, the 427044 square is unlike any other region in the Gulf of Maine. The second most important ten minute square only contributed 10% to the total landings in a given year (427034) and interestingly is located directly to the west of 427044.

As previously mentioned, a shortcoming of VTR data is that they are self reported and the catch amounts and location are subject mis-reporting. To verify that the apparent trends were not byproducts of VTR misreporting, utilization indices were created using observer and VMS data (using methods outlined in Palmer and Wigley 2009). The utilization trends for

ten minute square 427044 estimated from VMS and observer data exhibit similar trends to those from the VTR data for both the trawl and gillnet fleet (Fig. A.37). The observer data indicates slightly more utilization of the area during the 2000 period relative to the VTR data. Caution should be taken to not over interpret the earlier trends in the observer data due to the low number of observer trips in the early part of the time series. Overall, three sources provide evidence of a large increase in the utilization of ten minute square 427044 beginning around 2006 and persisting through 2010. The three data sources provide slightly different perceptions about when the increase began, but all three data sources suggest that by 2011, the level of utilization had dropped off.

The increased utilization of 427044 occurred not only in terms of fraction of annual landings, but also in terms of the number of trips and vessels. Between 1994 and 2010 there was an increase in the number of vessels and fishing trips into this ten minute square while outside of 427044 there was an overall decline in the number of vessels and trips landing Gulf of Maine cod (Fig. A.38). One hypothesis for the utilization trends of 427044 is increased sand lance abundance on Stellwagen Bank from 2006 through 2010 (Richardson et al. 2012).

Landings of Gulf of Maine cod have been dominated by ton class 2 (5-50 tons) and 3 (51-150 tons) vessels. Prior to 1994 ton class 4 (151-500 tons) contributed between 10-25% of the total commercial landings (Fig. A.39). Partly as results of the trip limits that were introduced in the late 1990's it became unprofitable for the larger vessels to target cod, and over the past decade Gulf of Maine cod has been predominately targeted by the smaller ton class 2 day boat vessels. With the implementation of groundfish sectors in May 2010 and the removal of trip limits there has been an increase in the relative landings by ton class 3 and 4 vessels, but they have not returned to the same levels observed in the 1980s through early 1990s. The rolling closures affect the seasonal patterns of ton class landings similar to the patterns observed for gear type and area (Fig. A.40).

Commercial landings of Gulf of Maine cod are classified by four primary market categories: scrod, market, large and unclassified. Other market categories exist such as snapper, whale and steaker, but these are considered variants of the scrod (snapper) and large (whale and steaker) market categories. Market sized fish typically dominate annual landings with scrod sized fish having become less common over time, possibly in response to increasing minimum retention sizes (Fig. A.41). Over the past six years, market cod have accounted for approximately 70% of the total landings (Fig. A.42).

The temporal landing patterns of Gulf of Maine cod has been relatively consistent over the past six years with the exception of 2010 (Fig. A.43). From 2006 through 2009, the fishery was most active from May through March, with very little landings occurring during the months of March and April. Presumably, the low landings during these months were the result of a combination of limited availability of DAS and rolling closures. However, a large increase was observed in April 2010, likely in response to the major changes brought about by Amendment 16. It is possible that vessels that were entering sectors in May 2010 sought to fully utilize any remaining DAS as its currency would be useless under a sector-based system. Seasonal landing patterns in 2011 were similar to those observed from 2006-2009.

Commercial landings: biosampling

Biological sampling (length and age) of Gulf of Maine cod prior to 1982 was poor to non-

existent (Table A.13). The sufficiency of biological sampling has always limited age-based assessments of Gulf of Maine cod to the period from 1982 onward. Prior to 1982 it was not uncommon for sampling to be absent across entire market categories, or even for an entire year. From 1982 to 1995 sampling was relatively constant at around approximately 30 to 60 samples per year. When sampling dropped off, it was typically sampling of the smaller (scrod) and larger (large) market categories that suffered. Beginning in 1996 there was a notable increase in overall sampling. The years 1998 to 2000 were exceptions to this trend and were marked by years of low landings, including the lowest level of commercial landings (i.e., 1,407 mt in 1999).

Since 1982 length sampling of the commercial landings has varied from 28.1 to 517.9 mt per 100 lengths (Table A.14). Sampling intensities less than 200 mt per 100 lengths has traditionally been considered an unofficial NAFO/ICNAF standard. Sampling intensity has generally increased over time and has exceeded the standard since 1996. Prior to 1982 length sampling was poor with sampling intensities exceeding 1000 mt per 100 lengths sampled. The sampling density (number of lengths per sample) has ranged from 3 to 345 lengths per sample with an average of 79 lengths per sample (Table A.13). In the earlier periods, while sampling intensity was lower than the current period, the density was generally higher. Part of the trend in declining sampling densities has come about from a relaxation of the requirement to collect the full number of desired lengths per sample. In the past, samplers would frequently not sample unless they could collect a full sample (typically 100 lengths, but has varied by market category over time). Given that age sampling is conducted at the same time as length sampling (but lower density), it is not surprising that the sampling of age structures (otoliths) has followed similar trends as lengths. From 1995 onward the metric tons per 100 ages have been less than 1000 mt with sampling in the last five years on the order of 100 mt per 100 ages (Table A.15).

Prior to SAW 53, Gulf of Maine cod assessments have estimated numbers-at-age by aggregating lengths into 3 cm bins. A complete update of the catch-at-age was conducted for SAW 53 and in doing so, an attempt was made to use 1 cm intervals. This required a greater degree of age imputation to manually fill in gaps in the age length key (ALK). The majority of market/time blocks required no imputation and for those that did, generally the percentage of landings requiring imputation was less than 5% (Table A.16). ALK imputation was primarily restricted to the older ages; given the small numbers of the population in these ages combined with the plus group handling of older ages, the impacts of this imputation are likely negligible.

When estimating the number of fish landed-at-age, every attempt was made to maintain the market category/quarter sampling design. However, when the availability of lengths for a particular market/quarter block was low, either a semiannual or annual time block was used. A criterion of 100 lengths per block was applied to the commercial landings for use as an objective basis to decide when it was appropriate to bin across quarters. In situations where an annual time block was required, the annual LW relationship (Equation 3) was used to convert landings to numbers-at-age. Otherwise, the appropriate seasonal LW equation was applied (Equations 1 and 3). A summary of the amount of binning that was required is presented in Table A.14. Total numbers-at-age are presented in Table A.17. The bootstrapped generated CVs on the landings-at-age estimates are shown in Table A.18. CVs are generally less than 30% for those ages that make up the majority of the landings (Ages 3-6). Prior to 1984, the calculation of bootstrap CVs were not possible due to the inability to identify individual sampling events. There is considerable uncertainty in the estimates of landings-at-

age among some of the older ages, particularly beyond age 9 where the average CV begins to exceed 40%. Overall, younger ages have become less prevalent in the commercial landings with increases in the minimum retention size (Fig. A.44). Older fish were less common in the landings back in the late 1990's, likely due to a truncated population age structure. Estimates of weights-at-age from landings in the commercial fishery are presented in Table A.19.

Commercial discards

Gulf of Maine Atlantic cod are primarily discarded in the commercial fishery for three reasons: (1) fish are below the minimum retention size (too small), (2) fish are of poor quality, and (3) high grading of smaller or poor quality fish in situations where a limited amount of fish can be landed (e.g., under trip limits). Discarding of smaller/poor quality fish became increasingly important from 1999 onward when the trip limits became more restrictive. However, the primary reported reason for fish discards has been because the fish were too small (Fig. A.45). With increases to the commercial minimum retention sizes in 2002, discarding due to undersized fish accounts for approximately 70% of total fish discards. This finding is in contrast to the conclusions of the GARM III assessment that “*...presumed that cod of all sizes and ages are discarded without prejudice.*” The GARM III conclusion was based on an examination of the years 1998 to 2000 when trip limits were most restrictive; however, this conclusion does not hold for other periods. This distinction is important to consider when determining how best to estimate the discards-at-age. Given that the majority of discards are of fish that are below minimum retention size, the method used in GARM III to account for discards in the catch-at-age was is not appropriate for the full time series and would lead to an underestimation in the fishing mortality on younger fish and an overestimation in older fish.

Direct sampling of the commercial fishery for discards has been conducted by fisheries observers since 1989. Of the Gulf of Maine cod that were observed to have been discarded by fishery observers, the following gear types account for greater than 99% of the total observed discards: benthic longline, small mesh (<5.5") otter trawl, large mesh ($\geq 5.5"$) otter trawl, shrimp trawl, and large mesh (5.5"-7.99") and extra large mesh ($\geq 8.0"$) sink gillnet gear (Table A.20).

While handline gear does not constitute a large fraction of observed discards, this is partly because this gear type is not frequently observed owing to the small size of these vessels and regulatory exemptions from observer coverage for some handline permit categories. Regardless, it is known that discarding by this gear does occur and it is accounted for in the in-season groundfish monitoring programs. The SAW 53 assessment attempted to estimate discards for this gear type, but the SAW 53 WG concluded that the proportion of observed trips for handline was too low and the imprecision of the discard estimates was too high to give confidence in the derived estimates (Table A.21). This decision was supported by the SAW 55 WG.

The in-season groundfish catch monitoring program makes a distinction between otter trawl gear types, specifically between standard otter trawls and the two modified otter trawl gear types, the Ruhle trawl and the haddock separator trawl. An examination of dealer data and observer data was conducted to determine if the data would support such a distinction when estimating discards for the stock assessment. The data indicate that there are more trips observed that use these modified gear types than report the gear types on the VTR (Table

A.22). This suggests that these gear types are not being accurately reported in the VTR data and no distinction can be made between the modified gear types and the standard otter trawl. However, given that the use of these gear types did not begin until 2009 and the frequency of use is low, this should have negligible impacts on discard estimates.

The total number of observed trawl, gillnet and longline trips ranged from a low of 62 in 1997 to a current high of 2,850 trips (Table A.23). The large increase in the number of observed trips in 2010 was due to the additional contribution of ASMs that were required for the groundfish fishery under Amendment 16. ASM coverage averaged approximately 25% of total groundfish trips whereas regular observer coverage (NEFOP) averaged about 7% (M. Palmer, NEFSC, unpublished data). A comparison of the estimated discard rates between ASM and NEFOP observers (Wigley et al. 2012) showed no statistical difference for the majority of gears and quarters examined. Generally, the Gulf of Maine cod ASM discard rates were statistically indistinguishable from the NEFOP discard rates as evidenced by the fact that the 95% confidence intervals of the difference between estimates include zero (Figs. A.46 and A.47). A comparison of the length frequency distributions showed only small differences in the longline and extra-large mesh gillnet distributions with the large mesh otter trawl and gillnet being nearly identical (Figs. A.48 and A.49). As in the SAW 53 assessment, no distinction has been made between data collected by ASM and NEFOP observers with respect to discard estimation.

The SAW 53 assessment evaluated several different temporal stratification schemes with respect to their impact on total discards and relative precision. Quarterly, semi-annual and annual stratifications were explored. All achieved nearly identical results with respect to total discards, with the annual stratification having slightly lower CVs, though generally all CVs were below the informal target of 30%. Given the lack of sensitivity to choice of temporal stratification, a decision was made to use a semi-annual stratification owing to its ease of use from an operational perspective when estimating discards-at-age. The current assessment has retained this approach.

Final estimates of discards ranged from under 100 mt in 1998 to a high of 2,198 mt in 1990 (Table A.24). While there are exceptions, large-mesh otter trawl is the major source of cod discards. Shrimp trawl discards were an important component of cod discards in the early years, but the required use of a Nordmore grate for the Gulf of Maine shrimp fishery beginning in 1992 was highly effective at reducing cod discards. The resulting CVs on the discard estimates are variable on a gear-specific basis. At the aggregate level, CVs of total discards are typically less than 30% and below 20% over the last four years (Table A.25). As a means of evaluating the accuracy of the discard estimation procedure, a check was conducted to attempt to estimate total landings using the same methodology used to estimate discards. Instead of estimating a d_{cod}/k_{all} ratio, a k_{cod}/k_{all} ratio is estimated. When compared to the total cod landings, the results show close agreement with respect to scale and trends lending support to the accuracy of the discard estimation procedure (Fig. A.50).

Discard mortality

The SAW53 assessment assumed 100% mortality of all discarded fish. A working group was convened in July 2012 to evaluate the available scientific information on the survival of cod on a gear-by-gear basis (summarized in Palmer 2012a). The working group consisted of scientific experts with experience in field estimation of discard survival and stock

assessments as well as both recreational and commercial fishermen and other industry representative. Using a modified Delphi approach the WG developed revised mortality estimates of cod (NEFSC 2012b). The revised estimates ranged from 20-80% depending on gear (Table A.26, Fig. A.51) which are generally consistent with the literature reviewed in Palmer (2012a). However, an accurate quantification of true discard mortality is difficult. Sole reliance on the results of the available literature is likely to bias the discard estimates low both because of the largely unaccounted impacts of long-term post-release mortality as well as unobserved escapement mortality. The discards incorporated into stock assessments only account for the observed discards brought on deck. There is some additional and unquantified mortality associated with fish that escape the capture process. While this fraction is likely small, it is an additional component that if not considered will result in negatively biased estimates of discard mortality (ICES 2005). Additional research is needed to better quantify both the true mortality of fish discarded at sea as well as quantify the magnitude of unobserved mortality in the Gulf of Maine cod fisheries.

Given that the otter trawl and gillnet gear are the dominant gear types with respect to commercial Gulf of Maine cod discards, the revised mortality estimates had only a moderate impact on the total estimates of commercial discards (Fig. A.52).

Commercial discards: biosampling

Observers collect length and age information from the discarded fraction of the catch (as well as on the retained catch); however, only length samples are currently available. ALKs were created using both commercial landings and NEFSC survey ALK corresponding to the appropriate season (spring/fall). Length sampling extends back to 1989 and has generally been quite good with sampling intensities for most years less than 100 mt of discards per 100 lengths (Table A.27). The length distributions by gear are shown in Figure A.53 on an aggregate basis and by year in Figure A.54. Increases in the minimum fish size as well as the impacts of trip limits leading to the discarding of larger sized fish are evident in the time series plots. Generally, shrimp trawl captures the smallest fish with the sink gillnet gear having a much broader distribution of lengths including a large proportion of lengths in excess of the minimum size. The reasoning for the expanded length distribution in the gillnet fishery is largely due to the prevalence of poor quality discards in this fishery (e.g., damage due to seals, dogfish or sand fleas that occurs during the gear soak).

When estimating discards at length, attempts were made to maintain the separate semi-annual estimates so that the most appropriate seasonal LW equation could be applied. For some years and gear types this was not possible owing to limited sampling. A criterion of 50 lengths per block was applied to the commercial landings to provide an objective basis to decide when it was appropriate to bin across semesters and or gear types. Binning across gear types was only done between the two gillnet gears owing to the similarities of their length frequency distributions.

Commercial discard hindcasting: pre-1989

Direct observations of discards by fishery observers only exist from 1989 to present. The model formulations used in past assessments have started in 1982 owing to the availability of information on the age composition of commercial landings. Prior to the SAW 53 assessment, no attempt was made to hindcast discards back to 1982. The SAW 53 assessment applied a survey filter method described in Palmer et al. (2008) and previously applied to

groundfish stocks in the Northeast Region (e.g., Mayo et al. 1992, O'Brien and Esteves 2001) to extend discard estimates back to 1982. Discards were only hindcasted for the three primary discard gear types during this period: large mesh otter trawl, shrimp trawl and large mesh sink gillnet. This same approach has been used in the current assessment.

The survey filter method requires information on survey numbers at length (N_i), estimates of gear selectivity at length (m_i), a scaling factor (q) and an estimate of total fishery effort (f). Assuming these are available, discard-at-length can be estimated using the following equations:

If:

$$(6.a) \quad C_i/f = q \cdot (N_i \cdot m_i), \text{ then}$$

$$(6.b) \quad C_i = (q \cdot f) \cdot (N_i \cdot m_i) \text{ as above.}$$

If :

$$(6.c) \quad K_i = C_i \cdot s_i, \text{ and}$$

$$(6.d) \quad D_i = C_i \cdot (1-s_i), \text{ then}$$

$$(6.e) \quad D_i = (q \cdot f) \cdot (N_i \cdot m_i) \cdot (1-s_i), \text{ and}$$

$$(6.f) \quad D_i/f = q \cdot [N_i \cdot m_i \cdot (1-s_i)]$$

where:

C_i is the catch retained by a given commercial mesh at length i ,
 N_i is the abundance of fish in the survey at length i ,
 m_i is the proportion of the available population retained by a given mesh at length i ,
 s_i is the proportion of the retained catch kept at length i ,
 K_i is the kept portion of the catch at length i , and
 D_i is the discarded portion of the catch at length i .
 f is some estimate of total fishing effort.

If it is assumed that the fish discarded pre-1989 were all less than the minimum size, the above equation can be simplified by setting s_i to 0. This assumption is likely valid for large mesh otter trawl and shrimp trawl, but may not hold for large mesh sink gillnet gear (Fig. A.55). The impacts of this assumption on the estimation of proportion at age is evaluated later. Using a set of years when management was similar to the hindcast years, gear selectivity at length (m_i), and the appropriate scaling factor (q) can be estimated and the accuracy of the overall method can be evaluated. The years 1989 to 1993 were used for method development and evaluation of trawl and gillnet gear and the years 1989 to 1991 for shrimp trawl due to major changes in the shrimp trawl discard patterns that occurred in 1992 (i.e., Nordmore grate).

Using Pope's (1975) 'alternate tow' approach, the ratios of observed proportion-at-length discarded from the fishery to the proportion-at-length present in the survey are generated (e.g., Fig. A.56). Equation 7 (Wileman et al. 1996) is then fit to the aggregate ratios (across all years) to generate selectivity ogives (Fig. A.57). The fits to the shrimp trawl were poor, and given the small size distribution of cod discarded in the shrimp trawl fishery, an assumption was made that the selectivity of the shrimp trawl was identical to that of the NEFSC bottom trawl survey. The mesh sizes of the shrimp fishery during this period (1.75"/4.45 cm) were not all together dissimilar from those of the survey gear (11.5 cm codend with a 1.27 cm liner). Comparison of the proportions at length between the survey-filter method and the direct observations recorded by observers shows reasonably close

agreement in the length distributions across years for large mesh otter trawl and shrimp trawl gears (Figs. A.58 and A.59). There was less agreement among the length frequency distributions for sink gillnet gear, with only two of the five years showing close agreement (Fig. A.60). Conversion of the number-at-length to numbers-at-age using a combined spring and fall NEFSC survey ALK showed even closer agreement between the survey-filter approach and the direct estimates (Fig. A.61 – A.63). This suggests that while the assumptions of the survey filter method may not accurately reflect the length distribution of gillnet discards, the overall impacts on the age distribution are mitigated.

$$(7) \quad r(l) = \left[\frac{\exp(a+bl)}{1+\exp(a+bl)} \right]$$

By regressing the ratio of observed discards-at-length to the total fishing effort (K_{all} was used similar to the contemporary discard estimates) on the ratio of selectivity-adjusted survey numbers-at-length, the gear-specific scaling factor (q) can be estimated as the slope of the regression line (Equation 6.f, Fig. A.64). In performing these regressions, it was noted that the relationship of the two ratios was different in 1990 relative to other years. It's possible that this reflects some effects of the 1987 year class moving into the fishery; the 1987 year class was the largest year class observed during the SAW 53 assessment time series (NEFSC 2012a).

Total discards estimated using the survey filter approach reflected the relative trends and scales from the direct estimates (Table A.28). The large mesh gillnet estimates were underestimated relative to the direct estimates, possibly due to the assumption of smaller fish in the survey filter method. In 1990 the survey filter underestimated across all gear types, possibly due to poor fit of q in that year as described above.

The SAW 53 WG considered an alternative metric to the survey-filter hindcast: use of an average of the d_{cod}/k_{all} ratio from years 1989-1993 and raise it by the annual K_{all} in years 1982-1988. The SAW 53 WG discussed whether the average d_{cod}/k_{all} ratio could be biased from including the 1990 value in the estimate, which may have been much higher owing to the anomalously large 1987 year class. As an intermediate approach, the WG recommended a third calculation of hindcasted discards using the average d_{cod}/k_{all} ratio for years 1989 to 1993, *excluding 1990* (Fig. A.65). The SAW 53 WG discussed the appropriateness of hindcasting, and whether assuming that discards are zero is better than making assumptions to derive estimated amounts. Ultimately, the SAW 53 WG concluded that the true discards are likely between zero and the d_{cod}/k_{all} ratio estimates that included the 1990 value (which provides a likely upper bound). The final approach applied the average d_{cod}/k_{all} ratio for years 1989 to 1993, *excluding 1990* as the basis for the amount of hindcasted annual discards with the proportion at age determined using the survey filter method. These estimates have been retained in the current assessment. Commercial discards-at-age and weights-at-age are presented in Tables A.29 and A.30 respectively. Figure A.66 shows the impact of the revised discard mortality estimates on estimates of total discards in terms of numbers. Bubble plots of commercial discards-at-age over time are shown in Fig. A.67.

Recreational landings

There is a large recreational fishery for cod in the Gulf of Maine that, over the last decade, has accounted for approximately 20-31% of the total catch. Previous assessments have used data collected under the Marine Recreational Fisheries Statistical Survey (MRFSS).

Beginning with this current assessment MRFSS data have been re-estimated using revised methodologies consistent with the new Marine Recreational Information Program (MRIP) which has replaced the MRFSS program (NMFS 2012). Since the existing data were collected under the MRFSS program, they will be referred to as MRFSS data. The conversion of MRFSS data to MRIP estimates is described later. The MRFSS data collection program began in 1979, though estimates of recreationally caught cod are not available until 1981.

Recreational catch data are divided into three components: directly observed landings (A), unobserved landings (B1), and unobserved discards (B2). Recreational catch is partitioned into Gulf of Maine and Georges Bank stocks using annual site register lists; catches attributed to intercept/interview sites in Maine and New Hampshire as well as Massachusetts landings from Essex, Suffolk, and Plymouth counties are allocated to the Gulf of Maine stock.

Landings from Barnstable County (Massachusetts) are split such that intercept sites bordering Cape Cod Bay are allocated to the Gulf of Maine stock and those on the east and south side of Cape Cod are allocated to the Georges Bank stock.

While MRFSS/MRIP is the source for official recreational catch estimates, VTRs provide a useful source for understanding some of the finer spatial and temporal trends that cannot be easily determined from the MRFSS data. They also help inform the validity of the MRFSS sampling scheme and treatment of data. VTR data are only available for the federally permitted party (head boats) and charter modes. Early in the time series party vessels were the predominate source of VTR-reported recreational catch, though charter boat landings have increased over the last five years (Fig. A.68). VTR data do not cover the private recreational fleet or party/charter vessels operating only within state waters. Federally permitted recreational vessels only represent from 14 to 69% of the total recreational harvest in a given year (Table A.31), thus VTR-based estimates will underestimate the total recreational landings (Fig. A.69). The MRFSS program did not sample the New England region in Wave 1 (January/February); however, VTR data suggest that historically, very low recreational activity occurs in these months (Table A.32). Since May 1, 2006 the recreational fishery has been prohibited from possessing cod in the Gulf of Maine between November 1st and March 31st. This prohibition was extended to April 15th in 2009. MRFSS-based estimates of total catch by sampling wave show highly variable temporal patterns, but are generally consistent with VTR data, with waves 2-5 having the highest proportion of total annual catch (Table A.33). Based on the VTRs, there are virtually no landings of Gulf of Maine cod in ports south of Massachusetts (Table A.34). This finding supports the existing allocation scheme based on the site register lists that is used to assign MRFSS recreational catch to the Gulf of Maine and Georges Bank stock components.

Unlike the commercial trawl fishery the recreational fishery has always been relatively concentrated with Gini indices ranging from 0.81-0.92 (Fig. A.70). There have been no large scale changes in the center of recreational effort over time (Fig. A.71). The majority of VTR-reported recreational landings come almost exclusively from statistical areas 513 – 515 (Table A.35), with most recreational activity located to the west of 70° W (Fig. A.72). The recreational fleet does not utilize 427044 to the same extent as the commercial fleet (Fig. A.73). While the charter boat fleet does have two notable periods of high utilization of this

area (1998-2000 and 2007-2010) the relative use is much less than that of the commercial fleet (Fig. A.74).

The MRFSS data collection program is a numbers-based survey and conversion of MRFSS estimates to removals in terms of total biomass can be accomplished in several ways. Total weight estimates typically provided by the MRFSS program convert numbers to weight using the average sampling weights by state and semester. In the earlier time periods, sampling was poor such that average MRFSS weights did not exist for all cells. This can lead to an underestimation of removals in terms of average weight (Method 1). Imputing the missing cells using the averages from other cells within the same year addresses the issue of missing cells (Method 2). The quality of the MRFSS weight sampling is unknown, though it is generally perceived that the quality of the length information is more reliable. Length sampling of recreational landings has improved over time, though the sampling intensity is not as good as that of the commercial fishery (Table A.36). An alternative method is to use the annual length frequency distributions (Figs. A.75 and A.76) to generate numbers at length and then apply the annual LW equation to estimate total removals in terms of weight (Method 3). Because the majority of recreational catch occurs mid-way between the spring and fall NEFSC surveys, it was not appropriate to partition out catch into spring and fall components. Methods 2 and 3 achieve similar results in terms of total landings, Method 1 tends to underestimate total removals early in the time series when sampling was sparse (NEFSC 2012a). Consistent with SAW 53, method 3 has been used to report out total recreational catch in terms of biomass, though these estimates are not used in the stock assessment model.

The numbers-based estimates of recreational landings were converted to numbers-at-age using ALKs borrowed from the NEFSC survey which include age information collected from the inshore strata. The inclusion of the inshore strata provided a better spatial overlap with the recreational fishery compared to the use of just the offshore strata (Fig. A.72). Recreational landings-at-age show similar trends with respect to the impacts of increasing minimum retention sizes (Fig. A.77). Like the commercial landings, older ages are absent from the recreational landings throughout much of the 1990s.

Recreational discards

With increases in the minimum recreational retention sizes, the contribution of recreational discards to total recreational catch has been increasing over time (Fig. A.18). Prior to the SAW 53 assessment, recreational discards were reported, but they were not included in the catch-at-age used in the assessment models. The primary reason for the exclusion was that historically, there had been no length sampling of recreational discards, and thus no information to convert the total recreational discard estimates (B2 catch) to estimates of discards-at-age. The largest fraction of discards is attributed to the party/charter mode in areas that are greater than 3 miles from shore and the private/rental mode, which has seen an increasing trend in the fraction taken more than 3 miles from shore (Table A.37). Beginning in 2005 direct sampling of cod discards from party boats began in the Gulf of Maine (i9 sampling; Table A.38). Sampling intensities have averaged approximately 200 mt of discards per 100 lengths sampled which is slightly higher relative to the length sampling of recreational landings during the same period.

Because of the increasing importance of recreational discards over time, the SAW 53 assessment attempted a hindcast of recreational discards using the available length frequency information and a variant of the survey filter method used to hindcast commercial discards.

Unlike commercial discards, estimates on the magnitude of recreational discards in terms of total numbers were already available from the MRFSS data. The survey filter method was needed only to construct the length frequency distribution of the recreational discard catch back in time. Similar to commercial discards, the assumption was made that all discarding was done due to minimum retention sizes. This assumption appears to be valid for the recreational fishery, with almost no discarding of legal-sized fish occurring in the 2005 – 2010 period (Fig. A.78). Using the alternate-tow approach used for commercial discards, a gear selectivity ogive was constructed (Fig. A.79). Comparing the survey-filter length frequency distributions to the observed length frequency distributions showed close agreement (Fig. A.80). Applying the survey filter method back to 1981 (start of the length sampling of recreational landings) yielded the length distributions shown in Fig. A.81. The same NEFSC survey ALKs applied to the recreational landing was used for the recreational discards resulting in the discard-at-age patterns shown in Figure A.82. As with commercial discards, the SAW 53 assessment assumed 100% mortality of all recreationally discarded fish. The revised estimate of 30% mortality was applied the recreational discards for the current assessment.

Conversion of MRFSS data using MRIP methodologies

In 2012 NMFS released revised estimate of recreational catch extending back to 2004. The revised estimates were based on the application of the MRIP sampling design to the existing MRFSS data. For Gulf of Maine cod, the revised MRIP estimates ranged from 48.4-98.1% of the MRFSS landings estimates and 52.5-101% of the MRFSS discard estimates (Table A.39). A working group convened in March 2012 recommended applying a ratio estimator to MRFSS data collected pre-2004 to convert the old data into scales consistent with the revised MRIP estimates. The WG concluded that the ratio estimator be based on the “ratio of means” (across all comparison years included) rather than based on the “mean of ratios” for individual years (NMFS 2012). Consistent with the recommendations of the WG, that approach has been employed in the current assessment yielding a ratio estimator of 0.742 for AB1 catch and 0.756 for B2 catch (Table A.39).

Total recreational catch has been re-estimated since SAW 53 due to minor updates to the MRFSS data and to accommodate the MRIP re-estimation. The minor updates to the MRFSS data resulted in differences generally < 1%, but some larger differences were present in the more recent year, most notably in the estimates of B2 catch (Table A.40). Conversion of the MRFSS estimates to MRIP-based estimates resulted in differences ranging from 1-50% for AB1 catch and -4-44% for B2 catch (Table A.41).

A summary of recreational catch from 1981 to 2010 is presented in Table A.42. Recreational catch has ranged between 0.3 and 4.1 thousand mt. Because of the method used to apportion MRFSS/MRIP cod estimates to stock areas, there are no direct estimates of precision available for recreational catches; however, the published estimates of percent standard error (PSE) provide some gauge as to the relative precision of the recreational catch estimates (Table A.43). Overall the general precision of these estimates is about equal to the commercial discards. Total cumulative recreational landings-at-age and landing weights-at-age are presented in Tables A.44 and A.45. Recreational discards-at-age and discard weights-at-age are presented in Table A.46 and A.47.

Total catch-at-age and mean weight-at-age

Estimates of total catch-at-age were determined by summing the numbers-at-age across all of the catch components: commercial landings, commercial discards, recreational landings and recreational discards (Table A.48). The age structure of fishery catch was truncated in the early 1990s relative to that observed in the 1980s. The truncation persisted through 2000 with age 9 and older fish beginning to reappear in the fishery in greater numbers beginning in 2001. These older age classes persisted through 2007 but have become less common in the fishery catches over the most recent four years. Mean catch weights-at-age were estimated by using a numbers weighted average of the individual catch component's mean weights-at-age (Table A.49). There is evidence of declines in the mean weights-at-age for fish older than age 5 over the last decade (Fig. A.83).

Estimation of January 1/spawning stock weights

Sampling of older age fish in the trawl surveys has historically been low, and use of survey-based weights-at-age to estimate January 1 and spawning stock weights for use as model inputs would require extensive imputation. For this reason, catch weights-at-age were used to estimate January 1 and spawning stock weights. Prior to estimation of stock/spawning stock weights, minor imputation of the catch weights at-age were required to fill in gaps in the older age classes (primarily in the age 9+ group). An examination of possible approaches (e.g., moving averages or time series averages) showed that imputation using a 5-year centered moving average would be most appropriate.

January 1 and spawning stock weights were estimated from catch weights using a method described in Rivard (1980, 1982). March 1 is the assumed spawning event in the base model. Given that there is little somatic growth between January 1 and the assumed start of the major spawning period (April 1; Fig. A.10), spawning stock weights were set equal to January 1 weights-at-age. The Rivard method adjusts the catch mean weights-at-age, which are generally presumed to represent mid-year weights, back to January 1. Mean weights at the beginning of the year for a given age class are calculated as the geometric mean of the weight in the same year and of the same cohort in the previous year. No adjustments are made for the plus group calculation. Calculations for the initial and final years and ages are described in Rivard (1980, 1982). January 1/spawning stock weights are shown in Table A.50.

Brooks et al. (2012) evaluated the sufficiency of applying Rivard-adjusted catch weights as a proxy for January 1/spawning stock weights. The analyses found the Rivard-adjusted age-specific catch weights to have similar trends and scale compared to NEFSC spring survey weights but had far less variability than survey weights and were not subject to the large number of missing ages and years of observation.

TOR A.2. Present the survey data and calibration information being used in the assessment; investigate the utility of commercial or recreational LPUE as a measure of relative abundance

There are three primary fishery independent surveys that operate bi-annually in the Gulf of Maine: the NEFSC bottom trawl survey, Massachusetts Department of Marine Fisheries (MADMF) bottom trawl survey and the Maine-New Hampshire (ME/NH) inshore groundfish survey. All three surveys operate in both the spring and fall with the seasonal timing differing

slightly between surveys. The NEFSC survey occurs the earliest of the three spring surveys with MADMF and ME/NH having similar timing (Fig. A.84). Conversely, the MADMF survey occurs first in the fall with the NEFSC and ME/NH survey having similar timing.

NEFSC bottom trawl survey

The NEFSC spring and fall bottom trawl surveys began in 1968 and 1963 respectively, providing the longest regional time series of fishery independent information. All previous Gulf of Maine cod assessments used only the offshore survey strata (Fig. A.85). The current approach to generating NEFSC indices ignores the inshore strata (Figs. A.86 and A.87) because a) historically they are not consistently sampled (Figs. A.88 and A.89); and b) the Massachusetts Department of Marine Fisheries (MADMF) survey covers the inshore areas and this survey has traditionally been included in the Gulf of Maine cod assessments. The impacts of including the inshore survey strata in the NEFSC survey indices were examined by both the SAW 53 and SAW 55 WGs. The overall trend in the aggregate abundance (numbers) and biomass indices were similar between the offshore-only indices and the combined inshore-offshore indices (Fig. A.90). There were several years in which the spring survey indices were noticeably higher due to inclusion of the inshore survey strata, but the general trends were similar suggesting that inclusion of the inshore variability increased the between year variability of the survey. The observed increases were primarily due to increases in age 0-2 fish with minimal impact on the age-specific indices of older age classes (Figs. A.91 and A.92). Due to the inconsistent sampling and minimal impact on the index trends, the SAW 55 WG supported the conclusions of the SAW 53 WG to exclude the inshore survey strata from the NEFSC Gulf of Maine cod survey index.

A frequent criticism of the NEFSC bottom trawl survey is that it does not cover the same areas where the commercial and recreational fisheries catch cod, and thus ‘misses’ much of the cod that exists in the Gulf of Maine. A comparison of the NEFSC spring and fall survey catches to commercial (total observed cod catches by ten minute square) and recreational activity (total number of trips catching cod by ten minute square) show close agreement between the location of survey and fishery catches (Fig. A.93).

The NEFSC bottom trawl survey has utilized three different vessels and three different door configurations throughout the time series of the survey (Table A.51). In an effort to maintain a consistent survey time series, survey indices are converted to ‘Albatross IV/Polyvalent door’ equivalents using several different conversion factors (Table A.52). The largest change in the survey time series occurred in 2009 when the *RV Albatross IV* was decommissioned and replaced by the *FSV Henry B. Bigelow*. This resulted in changes not only to the vessel and doors, but also to the overall trawl gear as well as the survey protocols (summarized in Table A.53). Calibration experiments to estimate survey differences were conducted in the spring and fall of 2008 (Brown 2009). The results of those experiments were peer reviewed by a panel of external (non-NMFS) experts and summarized in Miller et al. (2010). These results provide annual calibration coefficients both in terms of abundance (numbers) and biomass (weight). Further work by Brooks et al. (2010) developed length-specific abundance calibration coefficients for Atlantic cod. This method uses a segmented regression model where a constant conversion factor is applied to fish ≤ 20 cm and ≥ 54 cm, and a constantly decreasing linear regression is fit to fish between 20 and 54 cm (Fig. A.94). A comparison of the converted and unconverted spring and fall survey indices is presented in Figure A.95. It should be noted that while considerable focus has been placed on the Albatross/Bigelow

calibration, the effects of door calibration are generally larger than those of the *Albatross/Bigelow* calibration. Attempts to estimate *Albatross/Bigelow* calibration coefficients directly within an assessment model yielded similar coefficients as those estimated by Brooks et al. (2010), thus leading the SAW 55 WG to support the continued use of the existing *Albatross/Bigelow* calibrations coefficients (see Appendix A.5 for a description of the estimation of *Albatross/Bigelow* calibration coefficients within a statistical catch-at-age model).

During the SAW 53 fishing industry meeting (August 16, 2011 in Gloucester, MA), industry expressed concern with the 24-hour operation of the survey. There was a sense that there were differences in the relative catchability of cod between daytime and nighttime hours. These observations are supported in the scientific literature (e.g., Beamish 1966), though the nature of off bottom movements is highly variable. An analysis was pursued as to whether there were appreciable differences in survey catchability between daytime and nighttime tows. The results showed that generally catchability was slightly higher in the daytime tows. However, the trends between day and night tows were similar, and in most years the day/night survey indices fell within the 80% CI of the aggregate index (Fig. A.96). Because of the similarity in the trends it is appropriate to use both day and night tows to calculate indices for the assessment. Splitting by day and night would result in reduced tows and lost strata (Table A.54), which would increase the likelihood that survey indices could be influenced by a single large tow in any year.

Aggregate survey indices over time are presented in Table A.55 and the corresponding CVs are presented in Table A.56. It is worth noting that some of the highest survey indices are associated with relatively high CVs/confidence intervals. This is an important consideration in determining how to interpret survey indices; i.e., do increases in survey indices represent true increases in the relative size of the resource, or are the indices being driven by a few influential tows that are not indicative of the resource abundance/biomass? Generally, survey indices were higher in the earlier time periods, reaching lows in the mid-1990s. During the early to mid-2000s there was a slight increase in survey indices relative to the mid-1990, but subsequently survey indices have declined and are at, or near, time series lows (Figs. A.97 and A.98). There is reasonably good agreement between the intra-season survey indices (spring numbers vs. biomass) and inter-season indices (e.g., spring biomass vs. fall biomass), but poor agreement between inter-season and inter-index comparisons (e.g., spring biomass vs. fall numbers; Fig. A.99).

Indices-at-age for both the spring and fall surveys are presented in Tables A.57-A.60 and Figures A.100 and A.101. It should be noted that age information for the spring and fall survey does not begin until 1970. Similar to the trends observed in the commercial and recreational fisheries, there were few older fish present in the survey catch-at-age throughout most of the 1990s. Within the spring survey there is strong cohort tracking out to age 6 (Fig. A.102) and out to age 9 in the fall survey (Fig. A.103).

Plots of the spring and fall survey catches (number/tow) show a general decline in the overall abundance and spatial extent of the resource from the 1970s through the 1990s (Fig. A.104). There is an increase in the 2000-2010 period, but the increase appears to be restricted to the western Gulf of Maine. Moderate survey catches occurred along the coast of Maine in the 1970s, but these have not been observed in the past twenty years. To further address the aspect of spatial aggregation, a time series of Gini indices were calculated following the techniques outlined in Wigley (1996). These results support the patterns shown in distribution

plots and suggest an overall concentration of the resource over the last twenty years (Fig. A.105). The number of stations and strata where cod have been observed in the Gulf of Maine has generally decreased over time as the resource has become increasingly concentrated in the western Gulf of Maine (Figs. A.106 and A.107). Not surprisingly, the largest declines have been observed in those strata (01380-01400) off the coast of eastern Maine. These patterns are similar to the spatial aggregation that has occurred in the commercial fishery.

NEFSC model-based survey indices

The SAW 55 WG considered a generalized linear model (GLM) of the survey data, in which the factors considered included cruise (proxy for year), stratum, temperature, depth and time of day (Terceiro 2012). This model highlighted the highly contagious and over dispersed nature of the data, which called for use of a negative binomial distribution (one of many explored) in the fitting of the model. The best fit to the data was achieved with a model using cruise, stratum and time of day as factors. Overall, the temporal trends estimated by the model were similar to those of the design-based estimators described above.

The WG considered that use of the GLM estimates in the assessment model would result in an underestimation of the variability in the survey indices as the GLM is effectively acting as a smoothing function of each time series. The WG therefore recommended that the design-based survey indices be used in the assessment models. However, it noted that the CVs from the GLM could be compared to those generated during the stage two iterative re-weighting process as the latter incorporate both observation and process error, similar to what the GLM produces.

MADMF bottom trawl survey

The MADMF has conducted research bottom trawl surveys during the spring and fall since 1978, though age information is not available until 1982. A complete description of the MADMF trawl survey is provided in King et al. (2010). The survey strata included in the MADMF survey primarily includes the nearshore habitat within Massachusetts state waters in the southwestern Gulf of Maine (Fig. A.108). Because the MADMF surveys are conducted in relatively shallow waters and are limited in their spatial extent, they do not provide an index of the total stock resource, but may provide some information on the younger age classes inhabiting the nearshore environment (i.e., a recruitment index). Additionally, given the limited spatial extent, the MADMF survey may be more susceptible to resource availability due to timing of onshore/offshore seasonal movements (i.e., process error).

The abundance indices of these surveys exhibit the same overall trends as the NEFSC surveys, with the spring index currently at an all-time low (Table A.61 and Fig. A.109). The corresponding CVs for the aggregate indices are presented in Table A.62. Fall abundance indices are near time series lows, but the biomass index is currently above average (Fig. A.110). There is moderate agreement between the intra-season survey indices (spring numbers vs. biomass), but poor agreement among other index comparisons (Fig. A.111). Similar to what has been observed in the NEFSC survey, the number of stations and stratum in which cod have been observed has declined since highs early in the time series (Figs. A.112 and A.113).

In constructing the proportions at age in Massachusetts inshore survey, it was noted that a

number of length groups in the ALK were missing age information. While there was a modest (20 days) difference in the timing of the MADMF and the NEFSC spring survey, an attempt was made to augment the ALK of the inshore survey using aging data collected during the sampling of the inshore strata of the NEFSC survey consistent with the approach used for the SAW 53 assessment (NEFSC 2012a). The number of otoliths sampled in both surveys was about the same. After analysis conducted during the SAW 55 WG meeting, it was agreed that such augmentation was not necessary, with the ALK before and after this treatment being very similar. It was therefore recommended that the aging data in the Massachusetts inshore survey not be augmented with the NEFSC ageing data. MADMF indices at-age are presented in Tables A.63-66 and Figs. A.114 and 115.

The SAW 55 WG considered diagnostics of the Massachusetts spring and fall surveys, specifically how well the abundance of year-classes was being tracked by each survey. In general, year-class tracking in the spring survey was reasonable between ages 1-6 (Fig. A.116) but only reasonable between ages 0-1 in the fall survey (Fig. A.117). The WG discussed reasons why this might be the case, including seasonal movements of cod between the inshore and offshore. Based upon this analysis, the WG recommended that the MADMF spring, but not fall, survey time series be used in the SAW 55 assessment model of GOM cod consistent with the SAW 53 assessment.

ME/NH inshore groundfish trawl survey

The ME/NH inshore groundfish trawl survey has not been included in previous assessments, though previous assessment reviews have encouraged a thorough examination of the information available from this survey (e.g., NEFSC 2002b, NEFSC 2012a). The ME/NH survey began in fall 2000 and has been conducted in the spring and fall annually in the nearshore waters of the Gulf of Maine (Fig. A.118; Sherman et al. 2005). The ten year time series of abundance and biomass indices do not exhibit strong interannual fluctuations (Fig. A.118). Overall, there is moderate agreement between seasonal abundance and biomass indices, but poor agreement between spring and fall similar to the patterns observed in the MADMF survey (Fig. A.120). The SAW 55 WG discussed the possibility that seasonal north/south movements of cod along the Maine coast may be partly responsible for the lack of cohesion between the spring and fall survey, though no definitive information was available to evaluate these hypotheses. Similar to the NEFSC and MADMF surveys there has been a general decline in the percent of positive tows over the last decade in both the spring and fall surveys (Fig. A.121), though the fall survey has exhibited small increases the number of positive tows over the last three years.

The spatial distribution of catches seems consistent with the patterns observed in the NEFSC surveys with the highest catches occurring in the southwestern Gulf of Maine off the coasts of New Hampshire (Fig. A.122). There are some indications of high catches along the eastern Maine coast and could be indicative of spawning aggregations. The length frequency distributions suggests that the survey captures primarily age 0 through 2 fish (<40 cm; Fig. A.123). The size frequencies seem to suggest that ME/NH captures the same age classes observed in MADMF survey.

The biggest impediment to inclusion of this survey is the absence of age information. Progress has been made on the implementation and analysis of the data collected since the start of the survey; specifically, spring and fall 2005 and spring 2011 ageing has been completed and spring 2006 is in progress (S. Sherman, ME DMR, pers. comm.). The WG

recommended that the complete ageing of the entire time series of collected otoliths be considered a high priority. The WG concluded that while this survey may be valuable in the longer term, it is both too short and lacking the aging data to be used in the SAW 55 assessment.

Following up on the research recommendations of the SAW 53 WG, the SAW 55 WG evaluated available reproductive information to determine whether any of the fish sampled in this survey were mature and whether there was evidence to suggest the presence of spawning aggregations along the Maine coast. Since 2004 over 100 maturity samples have been taken annually in the ME/NH survey (Table A.67). Trends in the length-at-50% (L_{50}) maturity were evaluated which did not indicate large shifts over the short times series (Fig. A.124), but did show that the fish captured in the ME/NH tend to mature at a smaller size relative to those captured in the NEFSC survey. The L_{50} for cod captured in the ME/NH survey was 31.8 cm and 32.5 cm for females and males respectively (Fig. A.125) compared to 40.8 and 42.9 cm in the NEFSC spring survey. The cause of the discrepancy in the maturity schedule is not known, but similar patterns have been observed in other species such as winter flounder (S. Sherman, ME DMR, pers. comm.). Plots of fish greater than 25 cm show the possibility of spawning aggregations at both the southern and northern extents of the survey (Fig. A.126). Examination of maturity samples by region indicate a higher proportion of mature fish in the northern regions (Table A.68). It's not clear whether these patterns are confounded by north/south differences in the maturity schedule (i.e., fish at the southern extents mature at a larger size).

Inter-survey comparisons

Comparisons of inter-survey indices show moderate levels of agreement between NEFSC and MADMF surveys within seasons, but generally poor agreement across seasons (Figs. A.127 and A.128). Neither the NEFSC and MADMF surveys showed good agreement with the ME/NH survey, but this may be partly related to the short time series of ME/NH survey and general lack of contrast.

Catch per unit effort (CPUE) indices

Trends in commercial landings per unit effort (LPUE) had been used in Gulf of Maine cod stock assessments prior to SAW 53. The 1982-1993 age composition of the landings corresponding to the effort of an otter trawl sub-fleet (summarized in Mayo et al. 1994) had been used to calculate LPUE-at-age indices for ages 2 through 6 (Mayo et al. 2009). The index was never extended beyond 1994 due to major changes occurring in the Gulf of Maine groundfish fishery (Table A.3) including regulatory measure to reduce fishing effort, closed areas, changes in mesh size and trip limits in addition to a switch in the fisheries-dependent data collection system from a landings interview/intercet program to a self reported logbook program. All of these issues affect the comparability of LPUEs estimated from 1994 onward with the earlier time series. Additionally, these same issues would make standardization of a contemporary catch per unit effort (CPUE) index difficult. The SAW 53 WG examined model sensitivity runs to assess the utility of including the Mayo et al. (2009) LPUE index. Model results were insensitive to the index and the SAW 53 WG concluded to remove the index from the SAW 53 assessment.

The apparent disconnect between the increasing catch per unit effort (CPUE) reported by groundfish fishermen and the comparatively limited rebuilding suggested in SAW 53 assessment (NEFSC 2012a) received notable attention following the release of the final assessment results. To address the criticism the NEFSC convened a CPUE WG in August 2012 to review and evaluate the information available on both commercial and recreational catch per unit effort (NEFSC 2012c). The CPUE WG concluded that ideally, LPUE indices should be formally considered and vetted as inputs into the assessment model. If a LPUE index is determined to be a poor index of fish abundance, while it may not be formally included as a model input, the index should be described in the assessment report and explanations put forward describing why the information in the LPUE index may be inconsistent with other assessment tuning indices.

The SAW 55 WG considered a number of analyses in an attempt to develop representative indices of GOM cod fishable biomass based on commercial and recreational LPUE. One analysis updated the LPUE index used by Mayo et al. (1994) through 2011 (Palmer 2012b). This index used year, depth, tonnage class, quarter and statistical unit area as factors in a GLM assuming lognormal error (Fig. A.129, Table A.69). Trends produced by the analysis tracked spawning biomass (SSB) as estimated during the SAW 53 relatively well up until 2006 after which time LPUE increased much faster than SSB. The reasons for this divergence were discussed at length by the WG. A hypothesis considered by the WG is that sand lance abundance, which is a forage species of cod, became abundant in a small region of the western GOM (near Stellwagen Bank) between 2006 and 2010 (Richardson et al. 2012). It was theorized that this resulted in the aggregation of cod in the area and thus elevated commercial catch rates. The incidence of occurrence of sand lance in cod stomachs collected during the spring and fall NEFSC BTS surveys has increased since 2006. These surveys indicate that the Stellwagen Bank area appears to be a forage ‘hot spot’ for cod feeding on sand lance. The VTR, observer and VMS information from the commercial fishery indicates that fishing effort since the mid-2000s has become highly concentrated in this area as documented previously in this report (Palmer 2012b). A large abundance of cod in a region easily exploitable by the day boat fleet was likely responsible for the increase in CPUE reported by the fishing industry between 2006 and 2010. Interestingly, the two large NEFSC spring survey tows that were identified to have contributed to the large increase in estimated biomass in the GARM III assessment (NEFSC 2012a) both occurred on Stellwagen Bank - one in 2007 and the other in 2008. The same processes that led to the overestimation of biomass in the GARM III assessment were also responsible for the increases in CPUE reported by the fishing industry.

The SAW 55 WG discussed at length the processes that may be influencing the cod distribution in the Gulf of Maine. It appears that two related but separate processes may be underway. Over the longer term, there has been a loss of cod from the central and coastal areas of the Gulf with an apparent concentration of cod in the western area. Additionally, since 2006, there has been further aggregation of cod within the western Gulf into forage hot spots, hypothesized to be driven by sand lance. While it is difficult to prove definitely that these processes are responsible for the observed distribution changes, the evidence is suggestive. Notwithstanding the causes of the observed patterns, cod appear now to be aggregated in a small area of the Gulf, which suggests that the catchability (relationship between LPUE and biomass) has changed over the LPUE time series and has likely increased more recently. Over the longer term, there have a number of regulatory changes (e.g. seasonal closures, trip limits, etc) which call into question the utility of commercial LPUE as an index of GOM cod biomass. Similar issues with commercial catch rate indices have been observed

elsewhere (e.g. Harley et al., 2001). Based on these concerns, the WG recommended that the commercial LPUE index not be used in the SAW 55 assessment model. This recommendation is consistent with the findings of the recent NEFSC-sponsored LPUE workshop.

An LPUE index was also developed for the recreational fishery (Wood 2012). A GLM using year, month, area, permit and fishing category as factors was applied to the 1994 – 2011 recreational landings data (Fig. A.129, Table A.70). A number of error structures were explored with a lognormal model ultimately chosen. Contrary to the commercial fishery, recreational fishing has consistently occurred within a restricted region of the western Gulf. As with the commercial fishery, recreational fishing has been impacted by a series of regulations (e.g. seasonal closures, bag limits, etc). The analysis only included landings and was not able to include the release information which has become an increasing component of the catch. Further, the GLM analysis was only able to include party-charter boats. Overall, given concerns comparable to those of the commercial fishery, the WG recommended that the recreational LPUE index not be included in the GOM cod assessment model.

TOR A.3. Summarize the findings of recent workshops on stock structure of cod

A work plan on the topic of Atlantic cod stock structure in the Northeast United States/Scotian Shelf region was recommended by the New England Fishery Management Council's Scientific and Statistical Committee (SSC). The work plan laid out a three-phase process for re-evaluating, and possibly revising, the spatial basis for assessment and management of Atlantic cod. The first phase was to review data (genetic, life history, tagging, etc.) in order to evaluate the “null hypothesis” of the status quo management units.

The NEFSC sponsored a public workshop on cod stock structure, held June 12-14, 2012, facilitated by the Gulf of Maine Research Institute to address Phase I. Invited participants from the fishing and scientific communities presented on a range of topics with opportunities for discussion. The full workshop report is available at
<http://www.gmri.org/minis/index.asp?ID=52&p=149>.

Many of the workshop participants felt that there was compelling evidence that the current management units need to be revised. The Workshop did not reach any conclusions on what the most appropriate management units might be. This will require further data analysis and modeling in order to complete Phase I of the SSC recommended process. The workshop report also identifies gaps in the data and analyses and recommended actions to address them.

The Workshop did not explicitly address and propose the next steps in the process. The Steering Committee recommended that an inclusive but focused Working Group meeting be held involving a small group of Canadian and US scientists to consider the results of the Workshop. This Working Group should be provided the short-term data and analyses identified as missing by the Workshop. Using that information, as well as the conclusions from the Workshop, the Working Group should determine the most appropriate representations of biological stock structure to complete Phase I of the process. The results from this Working Group meeting should be evaluated through an independent peer-review process.

Since the phased review process of cod stock structure that was recommended by the SSC has not been completed, no changes to stock structure were incorporated into this assessment.

TOR A.4. Investigate the evidence for natural mortality rates

Previous assessments of Gulf of Maine cod have assumed a constant, age-invariant rate of instantaneous natural mortality (M) of 0.2 (NEFSC 2012a, NEFSC 2008, Mayo et al. 2009). The SAW 55 WG evaluated the sufficiency of this assumption through life history analyses of natural mortality. Hoenig (1983) demonstrated that natural mortality can be estimated as a function of the maximum observed age (t_{max}) in a population (*ibid*; Equation 8). Depending on whether the maximum age observed from the surveys ($t_{max} = 17$) or the maximum age observed in the fishery ($t_{max} = 16$) is used, this approach yields estimates of $M = 0.25$ or 0.26 . This approach was further refined by Hewitt and Hoenig (2005; Equation 9), and through the revised approach yields similar results of $M = 0.24$ or 0.26 . Because the Gulf of Maine cod stock has been heavily exploited for most of its recent history (post-1970; Figure A.19), and age samples are only available from the 1970s, M values in the range of 0.246 to 0.281 estimated from maximum age likely overestimate the true M .

An alternative approach relies on the gonadosomatic index (GSI) which used the ratio of gonad weight to somatic weight (Gunderson 1997). The general premise is that M is positively correlated with reproductive effort (*ibid*; Equation 10), more specifically, female reproductive effort. Estimates of GSI were not readily available for Gulf of Maine cod; however using a GSI value of 0.117 reported for the adjacent Georges Bank cod (McIntyre and Hutchings 2003) yields an M estimate of 0.209. Pauly (1980) first showed that M is proportional to the von Bertalanffy growth parameter, K . Using a variant of the relationship (Jensen 1996; Equation 11) and an estimate of $g=1.598$ (Gunderson et al. 2003) provides estimates of $M = 0.165$ or 0.201 depending on whether the K value is taken from the growth parameters estimated from the fall or spring surveys respectively. The lack of observed change in condition, as evidenced by a constant LW equation, does not support a hypothesis for a shift in life history parameters.

$$(8) \quad \ln(Z) = a + b * \ln(t_{max})$$

$$(9) \quad M = 4.22/t_{max}$$

$$(10) \quad M = 1.79 * GSI$$

$$(11) \quad M = gK$$

where:

Z is total mortality,

$a = 1.46$,

$b = -1.01$,

t_{max} is the maximum observed age in a population,

M is natural mortality,

GSI is the gonadosomatic index,

$g = 1.598$ (after Gunderson et al. 2003),

K is the von Bertalanffy growth parameter

From this, the meta-analysis of life history-based estimates the evidence available with respect to Gulf of Maine cod life history parameters suggests that an assumption of $M = 0.2$ is reasonable. It should be noted that maximum age as high as 16 has been observed in the

commercial fishery as recent as 2009 which suggests comparable natural mortalities relative to earlier in the time series (Table A.19). Also examinations of maturity-at-age and condition factor over time show no evidence of strong trends both of which can relate to changes in natural mortality.

The method of Lorenzen (1996) was used to provide an aged-based estimate of M (Fig. A.130, Table A.71). This method, which is based upon the relationship between body weight and M across a wide range of species, was used in SAW 54 to provide age-based estimates of M for Southern New England – Mid Atlantic Bight yellowtail flounder (Equation 12). The peer review panel of SAW 54 (O’Boyle et al. 2012) concluded that the application of an inter-species relationship to infer within-species dynamics was an over-interpretation of the method. While M no doubt may be age-specific, the pattern estimated from the Lorenzen method may not be appropriate. Recent work performed by Deroba and Shueller (<https://afs.confex.com/afs/2012/webprogram/Paper10183.html>) indicated that using constant or age varying mortality would have similar impacts on the assessment. The SAW 55 WG thus concluded that the parsimonious approach is for the SAW 55 assessment models to use a single M for all ages.

$$(12) \quad M_w = M_u W^b$$

where:

M_w = natural mortality associated with fish of weight, W ,
 M_u = natural mortality at unit weight, (3.69, consistent with Lorenzen ocean ecosystem constant)
 W = weight (g),
 b = allometric scaling factor (-0.305, consistent with Lorenzen ocean ecosystem constant)

Two working papers considered the predator field of cod in the Gulf of Maine area (Link 2012, Waring 2012). Link (2012) noted that directed piscivory of cod by other fish was not common, with well less than 200 cod in over 550,000 stomachs observed. Similarly, the evidence for cannibalism is weak with only 20 cod found in over 20,000 stomachs. Studies to date suggest that M due to fish predation is likely low and is focused on juvenile and smaller size groups (Smith and Link 2010). Waring (2012) considered marine mammals as a potential source of elevated M in the Gulf of Maine area. Four species of seals (Harbor, Grey, Harp and Hooded) are found in New England with Harbor and Grey seals being the most numerous. The Harbor seal population, which was about 38,000 individuals in 2001, has been growing at an annual rate 6.6%. The Grey seal herd has increased from tens of animals in the early 1980s to thousands of animals in the late 2000s. Firm estimates on the size of the current herds are not available. Notwithstanding this, the food habit research suggests that cod mortality due to seals is low. Additionally, while seals are known to prey on cod, they are generalist feeders and the importance of cod in the diet of Gulf of Maine grey seals is unknown. There is limited information that suggests that cod represent only a minor component of harbor seal diet along the Maine coast (Wood 2001).

An analysis of tagging data collected during 2003 – 2006 to jointly estimate natural and fishing mortality was undertaken during GARM III (Miller and Tallack 2007). This analysis was updated for SAW 55 (Miller 2012). Contrary to the earlier work, this analysis was not length-based. Estimates of M ranged from 0.4 to 0.7 for the Gulf of Maine with Gulf of Maine M estimates tending to be lower than Georges Bank estimates. It also provided evidence of significant cod movements between GOM and GB and area 4X on the order of 4.1 to 29.7%. While M was relatively high compared to current estimates, F was

comparatively low, prompting discussion on whether or not it was representative of the fishery due to local effects. The results were highly sensitive to the assumed return rate of high-reward tags. High-reward return rates on the order of 50% were associated with Gulf of Maine cod M estimates of 0.3, with M increasing as the high-reward tag rate increased. Model preference (based on log-likelihood function) was for assumptions of near-100% on reporting rates of the high-reward tags. Estimates of fishing mortality, F , were inversely related to the M response with F declining with higher assumptions of high-reward tags reporting rates. Across all ranges total mortality (Z) was estimated at approximately 1.0.

Concerns were raised with the tagging conducted in the Cape Cod area, which represented over 50% of the data in the database. The tagging had been conducted employing a wide range of expertise with mostly small cod tagged. This in combination with the warm water in the area may have resulted in higher tag induced mortality than assumed in the model. There were additional concerns with the assumed tag reporting rate (100%) for high reward tags. There is evidence to suggest differential reporting rates among some sectors of the commercial fishery, most notably the reporting rate by gillnet vessels was five times lower than that of trawl vessels (Tallack 2006). It is unknown if these same reporting trends also apply to the high-reward tags. There was also discussion on the age groups of cod represented by the study. GOM cod of 50 cm of about 2.5 – 3 years old, implying that the estimates of M are for ages 2.5 – 3 plus with it weighted towards the younger ages.

The SAW 55 WG discussed how best to use these estimates of M . It was hesitant to conclude that M was in the range of 0.6 – 0.7 and to recommend that these estimates be directly included in the assessment models. Rather, the tagging analysis is another form of modeling that should be considered. The WG discussed the availability of historical tagging to which the current estimates could be compared. It was reported that tagging work conducted in the Gulf of Maine area during the 1970s and 1980s suggested M estimates in the order of 0.2 – 0.3 whereas tagging in the 1990s was suggestive of M similar to the more recent results. These observations are based upon unpublished work that could not be corroborated at the meeting. Much of the historical work (e.g. Hunt et al. 1999) had been focused on cod movements and did not provide estimates of natural, fishing or total mortality. Further, concerns were raised that there was no obvious mechanism (e.g. predation) that could explain a recent increase in M , although it was countered that no mechanism has been identified for the current M estimate of 0.2, though this estimate is supported by life history parameters. The SAW 55 WG recommended profiling natural mortality across both the historical and more recent periods of the assessment to inform the discussion as to whether or not there has been a long-term change in M . The WG agreed that an option (M -ramp) with an M change should be considered as an alternate to a base model which would assume no change in M (i.e. $M = 0.2$).

Catch-curve analyses

Catch curves were constructed for the aggregate fishery catches (commercial and recreational landings and discards; Fig. A.131) and for the NEFSC spring (Fig. A.132) and fall surveys (Fig. A.133) based on the methods of Robson and Chapman (1961). Catch curves were conducted on a cohort basis rather than an annual basis which removed the confounding effects of differential year class strength on the interpretation of catch curve results. Linear regressions were fit to the log transformed catches of ages 4-8. While ages 4-8 may not

precisely match the fully recruited age classes in both the catch and the survey it offers a compromise between full selection and having sufficient ages to fit a reliable regression (i.e., few fish beyond age 8 are regularly observed across the survey time series). The slope of the regressions provides an indication of cohort Z which is useful when interpreting the implied total mortality of both tagging models (e.g., Miller 2012) and assessment models. The analyses suggest time series Z estimates on the order of 1.0 with the survey estimates being considerably more variable than the catch-based analyses (Fig. A.134). The catch-based Z estimates indicate total mortality around 1.0 beginning with the 1979 cohort and increasing above 1.5 with the 1988 and 1989 cohorts before dropping to time series lows near 0.6 with the 1994-1996 cohorts. Current Z estimates are estimated at approximately 1.0 for the 2004 cohort which is consistent with the total mortality suggested by the Miller (2012) tagging analysis.

Catch curves can also be useful for making general inferences on the selectivity of both fisheries and surveys. While selectivities can be estimated from the fitting of stock assessment models, it is useful to have model-independent estimates of selectivity that can be used to validate model-based estimates and/or provide some *a priori* understanding of selectivity. A method described in Restrepo et al. (2007) uses the residuals from the log-transformed linear catch curve analysis to infer relative selectivity-at-age. Selectivities are estimated using the ratio of observed to predicted catch proportions and then rescaling the residuals from each curve so that the maximum positive residual equals 1. This analysis was conducted on the catch curves from the total catch and NEFSC spring and fall surveys. The distribution of selectivities-at-age from all cohorts was examined to evaluate the time series distributions of selectivity at age. While this approach masks any changes that may be occurring in the selectivity across time, it is useful for gaining a general understanding of catch and survey selectivities and evaluating whether there is strong evidence for the presence of domed-selectivity (i.e., lower selectivity at older ages). Examination of the residual patterns from total catch (Fig. A.135) and the NEFSC spring (Fig. A.136) and fall surveys (Fig. A.137) do not provide evidence for domed selectivity. The selectivity distributions on the younger ages are consistent with the model-based selectivity from the SAW 53 assessment (NEFSC 2012a) with age at 50% ($A_{50\%}$) selectivity roughly between ages 3 to 3.5 for the total catch, ages 3 to 4 for the NEFSC spring survey and ages 2 to 3 for the NEFSC fall survey.

Additionally, a comparison of proportion of fish age 5 and older caught in the NEFSC surveys relative to the fishery shows a higher ratio of old fish caught by the NEFSC surveys (Tables A.72 and A.73). This in itself does not confirm the presence of flat top survey selectivity in the survey, but does indicate that the surveys may have higher selectivity on the older ages relative to the fishery.

There have been discussions during previous assessment meetings and working group meetings that adult cod may be unavailable to the NEFSC surveys due to the presence of fixed gear (primarily lobster pots) in the inshore areas. However, the ME/NH survey actively works with the lobster industry to have gear removed in advance of the survey and as noted before, this survey is not capturing large cod (Fig. A.123). Decreased selectivity in the fishery may be plausible, particularly if large cod are exploiting closed areas unavailable to the fishery (either permanent or seasonal). However, the SAW 53 WG examined the Massachusetts cod industry-based survey (Hoffman et al. 2006) which sampled in closed areas. The length frequencies from this survey did not indicate the presence of larger cod in the rolling closure areas relative to those captured in the fishery or surveys. Additionally, an

analysis of cod tagging data conducted by Hart and Miller (2008) concluded that there was no evidence that larger/older Atlantic cod are subjected to lower fishing mortality in the Gulf of Maine than smaller cod.

TOR A.5. Estimate annual fishing mortality, recruitment and stock biomass

Summary of the SAW 53 assessment model

The SAW 53 Gulf of Maine cod assessment applied the statistical catch-at-age model, ASAP (Age Structured Assessment Program v2.0.20, Legault and Restrepo 1998, Legault 2008), which can be obtained from the NOAA Fisheries Toolbox (<http://nft.nefsc.noaa.gov/>). This represented a change from previous stock assessments which historically had been assessed using VPA models. The reasons for selecting the ASAP model included the ability to explore alternative model formulations to counter/lend support to VPA results, additional flexibility to explore starting condition assumptions (e.g., extending the time series beyond 1982), ability to estimate a stock-recruit relationship internal to the model, and the ability to explicitly handle data uncertainty, particularly given the lessons learned from the update of the VPA model with respect to uncertainty in the survey data.

ASAP is an age-structured model that uses forward computations assuming separability of fishing mortality into year and age components to estimate population sizes given observed catches, catch-at-age, and indices of abundance. Discards can be treated explicitly. The separability assumption is partially relaxed by allowing for fleet-specific computations and by allowing the selectivity-at-age to change in blocks of years. Weights are input for different components of the objective function which allows for configurations ranging from relatively simple age-structured production models to fully parameterized statistical catch-at-age models. The objective function is the sum of the negative log-likelihood of the fit to various model components. Catch-at-age and survey age composition are modeled assuming a multinomial distribution, while most other model components are assumed to have lognormal error. Specifically, lognormal error is assumed for: total catch in weight by fleet, survey indices, stock recruit relationship, and annual deviations in fishing mortality. Recruitment deviations are also assumed to follow a lognormal distribution, with annual deviations estimated as a bounded vector to force them to sum to zero (this centers the predictions on the expected stock recruit relationship). For more technical details, the reader is referred to the technical manual (Legault 2012a).

The SAW 53 assessment covered the years 1982 to 2010. The choice of the 1982 start year (as opposed to early start years) was chosen because this is the period which has the highest data density. Data are available on the quantity and size composition of the landings and discards, both commercial and recreational. Several survey indices are available (NEFSC and MADMF), each with aggregate indices of abundance and biomass, along with data on age/size composition. Biological information such as growth, maturity and length / weight relationships are also available. Extending the time series before 1982 results in not only loss of information quality, but also introduces additional uncertainty into the assessment. Prior to 1982 there was no information on commercial discards and there was no information on recreational catch prior to 1981. Extending the assessment back in time requires tenuous assumptions about unrecorded historical catches. Any hindcasting of both unrecorded catches and assumptions on the selectivity of the fishery back in time are confounded by the

extensive regulatory changes back in time (Table A.3). Alternative start years were examined in the SAW 53 assessment and had negligible impact on the terminal year assessment result (NEFSC 2012a).

An age 9⁺ group was utilized in the assessment due to the difficulties encountered when attempting to estimate older ages in the population. Additionally there was evidence of truncation in the population age structure over the most recent three years and the difficulties in precisely estimating fishery selectivities of the older ages in preliminary developmental ASAP runs. The mid-point of the spawning period was assumed to be April 1 (25% through fishing year). Recruitment is modeled as deviations from the geometric mean (steepness fixed at 1.0). During the SAW 53 assessment, unsuccessful attempts were made to fit a Beverton-Holt function internally within the model because of insufficient contrast in the ASAP base model time series of estimated SSB and recruitment (1982-2010).

The model included two different fishery selectivity blocks with the first extending from 1982 to 1990 and the second from 1991 to 2010. The choice of selectivity blocks was informed on known periods of major change in the fishery with respect to mesh size, minimum retention size and changes in the regulatory reporting system. Different split years ranging from 1989 to 1994 were explored that encompassed these major changes. Sensitivity runs indicated that the 1990/1991 split had the lowest objective function and offered improved fit to the age composition in the way of reduced residual patterning. For the fishery, selectivity-at-age is freely estimated within each block for 8 out of 9 ages, with one age class fixed at full selectivity in each block. In block 1, age 5 was assumed to be fully selected, while in block 2 age 6 was assumed to be fully selected. This decision was informed on the basis of smaller mesh sizes and minimum retention sizes during the years included in block 1.

Each of the two NEFSC surveys included a single time invariant selectivity vector with selectivity-at-age being freely estimated from ages 1 to 5 and fixed at age 6 and older. The choice of the flat-topped selectivity pattern for the NEFSC survey indices was informed in part by the VPA results from SAW 53 results, which suggested increasing catchability with age, and the likelihood calculated in ASAP for domed versus flat-topped scenarios. Additionally, comparison of proportion of fish age 5 and older caught in the NEFSC surveys relative to the fishery shows a higher ratio of old fish caught by the NEFSC surveys (NEFSC 2012a). This in itself does not confirm the presence of flat top survey selectivity, but does support a conclusion of higher selectivity-at-age in the survey relative to the fishery. The MADMF spring survey was fit using a double logistic function to account for the sharp declines in selectivity-at-age observed in the VPA results. The descending slope of the double logistic function experienced boundary problems in preliminary runs and was subsequently fixed at 10 in the base model.

The effective sample size (ESS) estimated for both the fishery and survey catch-at-age data (which are treated as multinomial) was compared to the input effective sample size in an iterative fashion until the effective sample size specified more or less matched the mean model estimated value, or until no further improvement in trying to match the estimated value could be made. Additionally, following Francis (2011), minor adjustment in the effective sample sizes were informed by the overall fit between the predicted and observed mean age of the catch. The final ESS for the fishery was set to 75, the two NEFSC surveys set to 30 and the MADMF spring set to 15. The CVs on the surveys were initially set equal to the bootstrapped CVs presented in Tables A.47 and A.52). The bootstrapped CVs characterize the sampling error, but additional process error may be present in the survey indices that are

not reflected in the bootstrapped CVs. Subsequent examination of the model fits to the survey indices resulted in adjustments to the survey CVs by adding the following constants to each of the survey CV vectors to account for additional process error: 0.2 (NEFSC spring), 0.1 (NEFSC fall), 0.3 (MADMF spring). It should be noted that these minor adjustments offered slight improvements to the statistical fit of the SAW 53 model but had little impact on the model results (NEFSC 2012a).

An annual CV of 0.05 was assumed for the fishery catch. This was a trade-off in forcing an exact fit to the catch (as in a VPA-like formulation) versus accounting for some of the uncertainty in catch owing to the uncertainty in stock allocation, discard estimation and hindcasting procedure. Commercial landings in the assessment time period are assumed to be very precise. There is a limited amount of error introduced in the allocation procedure and through VTR misreporting, but generally, these uncertainties are low. CVs on commercial discards are in the range of 0.11 – 0.38 and recreational catch PSEs are in the vicinity of 20%. Given the overall uncertainties, the assumption of a constant catch CV=0.05 was not unreasonable. Model sensitivities to alternate CV assumptions were also explored during SAW 53, but overall, the model results are robust to alternate estimates of catch precision.

Update of the SAW 53 assessment model using revised data inputs

The general approach used to build the bridge from the SAW 53 ASAP model to an updated SAW 55 model was as follows (run numbers correspond to the run summaries presented in Tables A.74 and A.75):

- **Run 1:** Base model from SAW 53 (**SAW53_BASE**).
- **Run 2:** Update the recreational catch to account for changes from MRFSS to MRIP methodology. Requires updates to the total catch-at-age and catch weights-at-age matrices (**SAW55_B1**).
- **Run 3:** Update commercial and recreational discards to account for the revised assumptions of discard mortality. Requires updates to the total catch-at-age and catch weights-at-age matrices (**SAW55_B2**).
- **Run 4:** Update the stock weights-at-age matrix to account for the revisions in recreational catch numbers resulting from the changes from MRFSS to MRIP methodology (**SAW55_B3**).
- **Run 5:** Update the maturity ogive to account for the minor changes in maturity resulting from the inclusion of an additional year of maturity observations (**SAW55_B4**).
- **Run 6:** Update the MADMF spring survey to account for changes in the indices-at-age resulting from the use of only the MADMF ALK. Also, update the timing of the survey from April to May to account for a misspecification in the SAW 53 model (**SAW55_B5**).
- **Run 7:** Add 2011 data (**SAW55_B6**)
- **Run 8:** Update ASAP software to version 3.0.8 (Legault 2012a). ***This model represents the new SAW 55 reference model (SAW55_BASE)***.
- **Run 9:** Run SAW55_BASE under the assumption of 100% discard mortality to evaluate the impacts of the alternate discard mortality assumption on the SAW 55 reference model (**SAW55_BASE_100MORT**).

The results from the bridge building exercise are presented in Table A.75.

Overall, the impacts of the data updates were minimal on the 1982-2010 model formulation with a total 54 mt difference in the 2010 spawning stock biomass between the SAW55_B5 and SAW53_BASE models (Fig. A.138). There were moderate differences in the terminal estimate of age 5 fishing mortality driven by the effects of the revisions to recreational catch and discard mortality assumptions (Fig. A.139). Both of these revisions revised the SAW 53 estimate of 2010 age 5 fishing mortality downward from 1.14 to 0.67. The lower assumed catches attributed to revisions to the recreational catch and discard mortality assumptions resulted in minor negative re-scaling of age 1 recruitment estimates (Fig. A.140). Adding an additional year of data to the assessment model (2011) resulted in a 7% increase in 2010 spawning stock biomass and 7% decrease in 2010 age 5 fishing mortality relative to the SAW 53 results. The ASAP software change had no impact on the assessment model results. There were small changes in the estimated fishery and survey selectivities associated with the data updates (Table A.76). The selectivities were primarily affected by the changes in the discard mortality assumptions which shifted the selectivity curves to the right.

Compared to the impacts of the data update process on the assessment results, there was a larger impact on the observed retrospective patterns by discard mortality estimates. The SAW 53 assessment assumed 100% mortality of all discarded fish. Discard mortality has been revised in the SAW 55 assessment based on the recommendations of the Discard Mortality WG (NEFSC 2012b, Table A.26). Revising the discard mortality assumption increased the retrospective patterning associated with spawning stock biomass, age 5 fishing mortality and age 1 recruitment (Fig. A.141). To confirm that the discard mortality assumptions had a similar effect on the revised SAW55_BASE model, a variant of the SAW55 reference model was run using an assumption of 100% discard mortality (SAW55_BASE_100MORT). Introducing 100% discard mortality back into the SAW 55 model reduced the retrospective patterns to levels below those observed in the SAW 53 assessment (Figs. A.141 and A.142). Based on the minor retrospective patterns observed in the SAW 53 model, the SARC 53 Panel recommended that stock status determination should not be based on retrospective adjusted estimates of SSB and F (NEFSC 2012a). There are a number of potential sources of retrospective patterns, including missing catch (Legault 2009) which would be the expected effect if the true discard mortality was closer to the 100% assumption used in SAW 53 as opposed the revised estimates that are being used in SAW 55. While it is difficult to identify the exact cause of a retrospective pattern, the change in discard mortality assumptions from SAW 53 to SAW 55 does introduce additional retrospective patterning which negatively impacts the reliability of the model. Given the previously noted concerns with the revised discard mortality assumptions, further work is needed to revisit these assumptions and conduct field studies to better quantify discard mortality.

One interesting aspect of the SAW 55 retrospective pattern is that while the magnitude of the previous retrospective biases has increased, the additional year of data has caused the sign of the retrospective bias to switch such that the 2011 model underestimates spawning stock biomass relative to the 2010 model and fishing mortality is overestimated. These retrospective patterns will be further discussed as they relate to the final ASAP model(s).

Further refinement of the SAW 55 reference model

In developing the final ASAP model for SAW 55, over one hundred different model

configurations were explored. The nature of the sensitivity runs fell into two different categories: 1) determining whether an alternate model formulation offered improved fit to the data; and 2) evaluating the sensitivity of the model with respect to a range of assumptions. These investigations explored the model's sensitivity to the following:

Model fit explorations

- Survey calibration coefficients
- Use of survey numbers vs. biomass indices
- Survey catchability
- Multiple fleet definitions
- Inclusion of catch-per-unit-effort indices
- Plus group assumption (age 9⁺ vs. 11⁺)
- Survey selectivity assumptions (dome vs. flat topped)

Evaluating the sensitivity of the model with respect to a range of assumptions

- Inclusion/exclusion of survey indices
- Assessment starting points (e.g., 1964, 1970 vs. 1982)
- Catch precision assumptions
- Stock structure considerations

With a few exceptions, the distributions of the results from these sensitivity runs were similar to the SAW 55 reference case (SAW55_BASE) indicating that the model results are robust to a wide range of alternate assumptions and configurations (Fig. A.143). The major sensitivities runs that were explored are described in detail in Appendix A.6. Only the primary sensitivity runs that describe the transition from the SAW55_BASE model to the final SAW 55 model(s) are described with the main body of this report.

Placement of selectivity blocks in a two-block model

The SAW 53 model included two fishery selectivity blocks with a split between 1990/1991. Examination of residual patterns in the fits to catch-at-age from the SAW55_BASE model indicated problems with the model fits to the catch-at-age both in the early (pre-1990) and late (post-2004) time periods (Fig. A.144). Alternate model configurations were explored within the two-block model that attempted to reduce the residual patterning by adjusting the years in which the split occurs between the selectivity blocks from 1986/1987 to 1992/1993 in two year increments. There was a ten point improvement in the objective function associated with the 1986/1987 and 1988/1989 splits (Table A.77). Both of these earlier splits reduced the residual patterning in the fits to the catch-at-age in the earlier period but offered no improvement in the residual patterns occurring post-2004. With the earlier selectivity block splits, there was minor degradation in the precision of the selectivity-at-age estimates (Table A.77). The CVs on ages 1-3 increased from 17 to 44% and 24% for age 9⁺. While the percent change in the selectivity CVs was large, the CVs were still relatively small for ages 1-3. The increase in CVs is likely a result of having fewer observations within the earlier time blocks with which to estimate selectivity. Overall this is a small tradeoff given the overall improvement in objective function and improved fits to the catch-at-age. Additionally, while this analysis was instructive in informing placement for the first selectivity block (between 1987 and 1989) it does not address the residual patterning in the latter part of the time series which is perhaps better addressed in a three-block model. An exploration of three-block models will be conducted later in this report.

Fitting of the MADMF spring survey selectivity-at-age

An area of concern with the SAW 53 assessment model were boundary solutions on the selectivity parameter estimates for the MADMF spring survey. The survey selectivity is estimated using a double logistic function; in the SAW 53 assessment the ascending slope parameter was fixed at 10.0 to avoid boundary problems with this parameter, but other boundary problems existed for the $A_{50\%}$ and $A_{50\%}$ descending parameters (Table A.76). These problems persist in the SAW55_BASE model. In an effort to address these concerns, attempts were made to fit the MADMF spring survey using a non-parametric approach with each age having an independent selectivity parameter. Informed by the double logistic fit, selectivity at age 1 was fixed at 1.0 and selectivity at all other ages was freely estimated. The first modeling approach (SAW55_BASE_FIXED_MADMF_AGE1_9) fitted all ages 1-9. The estimated selectivity curve was similar to that double logistic fit of the SAW55_BASE model (Figure A.145). There was a high degree of imprecision with the selectivity estimates beyond age 6. This is consistent with the finding that the year-class tracking in the spring survey was reasonable between ages 1-6 (Fig. A.116). Based on these results a second attempt was made to restrict the model to fitting only ages 1-6 in the MADMF spring survey (SAW55_BASE_FIXED_MADMF_AGE1_6). The estimated selectivity curve from the age 1-6 fit is nearly identical in to the age 1-9 fit between ages 1 and 6 both with respect to the estimated selectivity at age and CVs. Additionally, there is no perceptible change in the residual patterns observed in the survey fits to age (Fig. A.146). Fitting the model to only MADMF spring survey ages 1-6 resulted in improved fits to the catch at age compositions (Table A.78). These investigations indicate that further Gulf of Maine Atlantic cod assessment models should restrict the ages used when fitting the MADMF spring survey to ages 1-6.

Development of a three selectivity block model

Based on the results of the two-block examinations as well as the fits to the MADMF spring survey, attempts were made to fit a three-block model with the MADMF spring survey fit non-parametrically to ages 1-6. The two-block model examinations showed support for a split between the first and second survey blocks somewhere between 1987/1988 and 1988/1989. Additionally, the catch-at-age residuals suggested that a third selectivity block between 2004/2005 may address some of the observed residual patterning. There were no major regulatory changes in specifically in 2004 or 2005 that would give *a priori* expectation for a change in selectivity at this precise cut-off; however, there was a major change in the reporting system used for commercial landings with a switch from a paper weighout system to mandatory electronic self-reporting for all federally permitted dealers. Additionally, in 2006 there were increases in the recreational minimum retention size, seasonal recreational closures and the implementation of 2:1 DAS accounting in the commercial fishery (Table A.3).

Several different attempts were made to fit a three-block model including:

- **SAW55_3BLOCK:** a simple implementation of the SAW55_BASE non-parametric selectivity with the addition of a third-selectivity block starting in 2005.
- **SAW55_3BLOCK_DL:** Identical blocking to that used in the SAW55_3BLOCK model, but a utilizing a double logistic function to estimate fishery selectivity.

- **SAW55_3BLOCK_SL**: Identical blocking to that used in the SAW55_3BLOCK model, but utilizing a single logistic function to estimate fishery selectivity (flat-topped).

The SAW55_3BLOCK model offered improved model fit relative to the two-block SAW55_BASE in the way of improved objective function while having minimal impact on the assessment results (Table A.79). The SAW55_3BLOCK model resolved the residual patterns that were evident late in the time series in the previous two-block explorations (Fig. A.147). One of the problems with the two-block model were some of the boundary problems observed in fitting the fleet selectivities as well as the high CVs on the selectivity estimates on older ages. These same problems existed in the SAW55_3BLOCK model; specifically, boundaries were hit at age 4 and age 8 in the first selectivity block and age 5 in the second selectivity block and CVs exceeded 0.3 in all ages ≥ 8 when boundary problems were not encountered (Table A.80). Fitting the fleet selectivity using a double logistic function attempted to address these issues but proved problematic due to the difficulty in estimating the downward portion of the double logistic function. The $A_{50\%}$ and downward slope parameter estimates either hit boundary solutions or had excessively high CVs (Table A.80, Fig. A.148). The problems encountered in both the parametric and double logistic selectivity fits suggest general problems with estimating the downward component of the dome-shaped selectivity. Overall, there did not appear to be sufficient information within the data with which to estimate dome-shaped selectivity for the fleet. Given these results, an attempt was made to estimate fleet selectivity used a single logistic fit. The SAW55_3BLOCK_SL estimated similar selectivities for the younger ages and the parameter estimates were all well estimated with CVs < 0.1 (Table A.80). There were minor differences in the catch-at-age residuals between the SAW55_3BLOCK and SAW55_3BLOCK_SL models (Fig. A.147). Additionally, the SAW55_3BLOCK model did not offer an improved model fit relative to the single logistic formulation (10 objective point difference and 18 parameter difference). The SAW55_3BLOCK_SL model resolved the diagnostic problems present in the SAW55_3BLOCK model and offered a more parsimonious model formulation with negligible difference in the assessment results (no change in 2011 F, <1% change in 2011 SSB).

While there was some evidence of higher selectivity at older ages relative to the fishery (Tables A.72 and A.73), the examination of the catch curve residuals (Figs. A.135) did not provide compelling evidence for a dome-shaped fleet selectivity. The use of single logistic form to estimate fleet selectivity does not negate that there may be minor doming of the fleet selectivity, but the weight of evidence combined with the model fit diagnostics indicate that the evidence is weak and the assumption of a dome has a negligible influence on the assessment results. Additionally, a working paper considered by the SAW 55 WG, Legault (2012b), examined the effects of different error assumptions on model estimated selectivities and concluded that “[t]his argues for greater reliance on external information for the existence of domes. Or as a corollary, more forcing of flat tops in the selectivity functions unless strong external evidence is available to support the presence of domed selectivity.”

The next steps in the formulation of the final three block model were to incorporate an earlier split between the first and second blocks and incorporate the fitting of the age 1-6 MADMF spring survey indices at age using a non-parametric approach. Based on the observed lack of improvement in the objective function seen between the 1986/1987 and 1988/1989 split models a 1988/1989 split was applied to ensure that there were sufficient data within the first block with which to precisely estimate selectivity. The move from the 1990/1991 split

(SAW55_3BLOCK_BASE_SL) to the 1988/1989 split addressed the residual patterning observed in the catch-at-age during the early part of the time series (Fig. A.147) as well as offering a 12 point improvement in the objective function.

The penultimate step in the development of a three-block model was the modification in fitting the MADMF spring survey selectivity (SAW55_3BLOCK_SL_MADMF_1_6). Changing the MADMF spring survey selectivity from a double logistic to a non-parametric selectivity-at-age fit to ages 1-6 had only minor impacts on the assessment results (1.5% decrease in SSB, and 3% increase in age 5 F and no perceptible changes in fishery selectivity patterns; Fig. A.148). More importantly, this change addressed the diagnostics issues with the fitting of the MADMF spring survey that were previously mentioned. Examination of the root mean square error (RMSE) fits to the aggregate survey indices did not provide any indication that the further adjustments were needed to the survey CVs. All SAW 55 model explorations incorporated the same CV adjustments used in the SAW53_BASE model to account for additional process error: 0.2 (NEFSC spring), 0.1 (NEFSC fall), 0.3 (MADMF spring).

The final steps in the development of the three-block model were to a) modify the penalty function applied to the recruitment deviations and b) to adjust the input effective sample sizes (ESS). In all previous models there was a penalty function, lambda, applied to the recruitment deviations. Since the existing model does not fit a stock recruit relationship the SAW 55 WG consensus was that the model should place less constraint on recruitment estimates. Through an iterative approach a final agreed approach set the lambda value at 0.2 with CVs set at 0.5. This approach addressed the WG concerns and provided some constraints at the end of the time series where there is little information to inform recent recruitment. The ESS adjustments were based on the application of ESS multipliers consistent with Method 1.8 of Francis (2011). The multipliers are computed such that the stage 2 input effective sample sizes are equal to the current input effective sample sizes times the multiplier. Thus, a value of 1 leaves the input sample size unchanged, while values greater than 1 increase the input sample size and values less than 1 decrease the input sample size. Francis (2011) recommends only applying these multipliers once after all other model formulations have been determined. The new input ESS values are the result of applying these stage 2 multipliers to the original input ESS (rounded to the nearest integer). The ESS adjustments applied following this approach are as follows (multipliers are in brackets):

- Fleet catch: $75 \cdot (1.064) = 80$
- NEFSC spring survey: $30 \cdot (0.516) = 15$
- NEFSC fall survey: $30 \cdot (0.494) = 15$
- MADMF spring survey: $15 \cdot (0.588) = 9$

The net effect of these was moderate with respect to the terminal estimates of spawning stock biomass and age 5 fishing mortality (Table A.79). The 2011 SSB estimate decreased by 16% and the age 5 F increased by 22%. The final base model is referred to as the SAW55_3BLOCK_BASE.

Natural mortality

As noted earlier, the SAW 55 WG spent considerable time discussing natural mortality (SAW 55 WG 2012a, 2012b, 2012c). There was conflicting evidence for both the scale and trends in natural mortality with the tagging information providing the only evidence for changes in M .

Meta-analyses that were considered as well as food habits information provided no compelling evidence for changes in M over time. To address the conflicts in information, the WG recommended profiling the models over a wide range of M values. The profiles were conducted for three separate time blocks: 1982 – 2002, 2003 – 2011 and 1982 – 2011. The first two time blocks correspond to the period before/during the contemporary tagging study analyzed in Miller (2012). Profiling was conducted on the SAW55_3BLOCK_BASE model across a wide range of M values. Profiling over the entire 1982-2011 time series showed support for M between 0.3 and 0.5. When profiling was conducted on the restricted time blocks an M of between 0.1 and 0.2 was preferred for the 1982-2002 period whereas profiling conducted on 2003-2011 period suggested an M between 0.1 and 0.6 (Fig. A.149). These profiles were consistent with the tagging evidence for M being greater than 0.2 in the 2000s and a change in M over the longer term. Interestingly, when profiling was conducted over the full 1982-2011 time period on a variation of the base model under an assumption of 100% discard mortality there was model preference for M in the range of 0.2 to 0.4. Discard mortality assumptions have implications for model-based inferences of natural mortality.

It should be noted that ASAP profiling exercises conducted using modified time blocks (1982-2004, 2005-2011) which showed clear support for an M between 0.1 and 0.2 in the early time period and an M between 0.4-0.6 in the later time period. These profile results are not shown because they are not entirely consistent with the tagging period examined (2003-2006). However, they do illustrate that the assessment model-based evidence for a higher M in the more recent time period is sensitive to the time blocks examined. This highlights the low discriminatory power of the models to estimate M . Despite the low-discriminatory power of the models, the SAW 55 WG did agree to explore an M -ramp model with M during the 1982 – 88 period set equal to 0.2, during 2003 – 2011 at 0.4, with a linear ramping up of M during 1989 – 2002 between 0.2 and 0.4 (Fig. A.150).

Sensitivity runs of the final three-block model

The SAW55_3BLOCK_BASE model presented above constitutes the preferred ASAP model for the SAW 55 Gulf of Maine Atlantic cod assessment. The SAW 55 WG explored several sensitivities of this model with respect to different assumptions of discard mortality (revised discard mortality vs. 100% discard mortality), natural mortality ($M = 0.2$ vs. M -ramp) and fishery selectivity (flat-topped vs. dome). A factorial comparison of the various sensitivity assumptions was conducted to fully evaluate model sensitivity. The examined models are displayed in Table A.81 and the sensitivities are described below. Plots of the model estimated fishery selectivities are provided in Figs. A.151 (flat-topped) and A.152 (domed). The time series of spawning stock biomass, age 5 fishing mortality and age 1 recruitment are provided in Fig. A.153 (flat-topped) A.154 (domed). Retrospective plots are provided in Figs. A.155 (flat-topped) and A.156 (domed). Model diagnostics are provided in Tables A.82 (flat-topped) and A.83 (domed).

While the SAW 55 WG had previously agreed to move forward with the use of the alternate discard mortality rates, concern were raised due to the degradation of model performance when the alternate discard mortalities were incorporated into both the SAW53_BASE and SAW 55_BASE models, specifically the increase in retrospective patterning (Figs. A.141-142). The incorporation of the 100% discard mortality has only small effects on the fishery selectivity estimates, primarily in the way of causing a slight shift towards smaller fish in all selectivity blocks (Figs. A.151-152). The 100% discard mortality assumption causes a slight

positive re-scaling of both age 1 recruitment and spawning stock biomass with minimal effects on the 2011 estimates. It does however result in an increase in the fishing mortality estimates in 2010 and 2011 under all scenarios (Tables A.82-83, Figs. A.153-154). As observed in previous models, there was a reduction in the SSB and F retrospective patterning on the order of 30-40% (Tables A.82-83) when 100% discard mortality was used compared to the alternate discard mortality rates. The WG noted that assuming 100% mortality of discards (as done by SAW 53) moderately improved model fits and reduced the retrospective pattern and was more consistent with tagging studies in which carefully handled cod can experience high (e.g. 50%) mortality within two days of being released (Miller 2012). Notwithstanding this, the WG agreed to use the estimates from the Discard Mortality WG (NEFSC 2012b) for status determination and projections but to show the impact of the 100% discard mortality estimates on the 2011 spawning stock biomass (SSB) and fishing mortality (F) estimates without bringing these through to reference points and projections.

Earlier formulations of the SAW55 ASAP model indicated no statistical basis to choose a dome over a flat-top and stock trends were the same. It was noted that during GARM III, the principle was adopted that a flat-top should be assumed unless there was evidence for a dome (NEFSC 2008). Tagging analyses considered at that time indicated that flat-top relationships were to be expected (Hart and Miller 2008). The WG discussed other processes which could explain a dome or a flat top (e.g. gear mix) but there were no specific explanations for a dome. In response, it was noted that the SCAA models favored domes although over-parameterization could be an issue (Legault 2012b). The SCAA models were rerun with the flat-top selectivities from the ASAP models to see how this assumption is influencing the difference between the two formulations. These runs confirmed that use of a flat-topped fishery selectivity was not consequential to the difference and thus the WG agreed that further formulations would use flat top fishery selectivity relationships. It should be stressed that for the fishery selectivity curves that were estimated for the Gulf of Maine cod in this assessment, the choice of a flat-topped or domed shape has negligible impact on the assessment results (Fig. A.157).

The influence of the M -ramp (M_SPLIT) had almost no impact on fishery selectivity estimates, but resulted in positive re-scaling of age 1 recruitment and spawning stock biomass and negative re-scaling of fishing mortality from about 1991 onward. Interestingly, there were only small impacts on the 2011 terminal estimates (Tables A.82-83, Figs. A.153-154). Under an assumption of M ramping to 0.4 in the later period of the time series, the removals attributed to natural mortality exceed fishery removals from 1998 to 2010 (Fig. 158). There was considerable improvement in the retrospective patterns both in the flat-topped (Fig. A.155) and domed (A.156) ASAP formulations. There was an 8 point improvement in the objective function under both the flat-topped and domed assumptions (Table A.82-83). Support for an M -ramp rests primarily on its ability to reduce the retrospective pattern, although the retrospective patterning in the Gulf of Maine stock is not as severe as that of the Georges Bank cod stock assessment. It should be noted that the Miller (2012) tagging analysis supported a much higher M (0.6) than used in the M -ramp model (0.4). A sensitivity run of the ASAP M -ramp model (flat-topped) was conducted using an M of 0.6 during the recent period. Compared to the model using an M ramped up to 0.6, the fit with M ramped up to 0.4 improved fit by 22 log-likelihood points. These analyses indicated that while estimation of current spawning stock biomass (SSB) was generally comparable between models with different M options, the bigger issue is the impact of these options on reference point and thus stock status determination. The WG agreed to pursue two M options ($M = 0.2$ and M -ramp) with respect to their potential impact on reference points and short-term

projections (SAW 55 WG 2012c). **The SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) ASAP models were forwarded to the SARC 55 Panel for consideration.**

Recommendations of the SAW 55 WG

The SAW 55 WG could not reach consensus on which model should serve as the basis of current stock status determination and management advice, but noted that “*...lack of consensus should not be interpreted as implying equal support for the models...*”

Consequently, both the $M = 0.2$ and M -ramp ASAP models were put forward for consideration by the SARC 55 Panel along with the list of support for and against both modeling approaches which is outlined below:

$M = 0.2$ approach

The features that lend support to the assumption that M has remained constant throughout the time series are those features which do not support the M ramp assumption, which is discussed below. The main feature against the assumption of constant M is the presence of a retrospective pattern. However, there is some evidence to suggest that this may be transitory and becoming less of an issue (SAW 55 WG, 2012c). It was for this reason that no adjustment for the retrospective pattern has been made to any of the models.

M -ramp approach

One of the main features supporting the assumption of a recent change in natural mortality is that it employs an $M = 0.4$ which is generally consistent with the results of the 2003 – 2006 GMRI tagging data and associated analyses (if one assumes a 50% reporting rate of high reward tags). The tagging analysis indicated that M could be as high as 0.6. Tag reporting rates would have to be very low in order to be consistent with an M of 0.2.

Another line of support for this assumption is the model fits. The value of the objective function for the M -ramp model was lower (by 8-10 log-likelihood points depending on the specific formulation) than that of the $M = 0.2$ model. Further, compared to the $M = 0.2$ model, assuming that M had changed more recently reduces the retrospective pattern.

The final observation supporting a recently elevated M in Gulf of Maine Cod is evidence of increasing M in the adjacent NAFO Div. 4X Cod stock, based on both tagging analyses and assessment model fits.

A number of features don't lend support to a recently increasing M . There is no evidence for increased predation, either by fish or pinnipeds, in the diet compositional data collected by the NEFSC. Regarding the GMRI tagging analyses, if reporting rates of high reward tags were less than 50%, natural mortality would be less than 0.4. It is unfortunate that there are little or no historical tagging studies to which the results of the GMRI study could be compared. Besides using different assumptions, these earlier studies did not formally incorporate parameters to estimate movement. For

these reasons, the tagging studies which suggested higher (than 0.2) M in 4X may not apply to Gulf of Maine Cod (SAW 55 WG 2012a).

Regarding model fits, the likelihood profile of M for the 2003 – 2011 period was relatively flat, with estimates between 0.1 and 0.6 potentially possible. Exploratory runs indicated that M profiling was sensitive to which years to include in the recent period of high M . A change of two years would result in a more informative profile (favoring higher M).

The final lines of evidence against a recently elevated M relate to the life history information. Compared to adjacent stocks, there have been little or no long-term changes in maturity at age, fish condition and growth. Meta-analyses of life history parameters suggest an M of 0.2 with no trend over time. For example, fish as old as age 16 have been observed in the population within the past five years, seemingly inconsistent with a two-fold increase in natural mortality.

Diagnostics and results of the $M = 0.2$ ASAP model (SAW55_3BLOCK_BASE)

Model fits to the fishery catches were good, with no strong patterning of residuals over time and generally good agreement between modeled and observed catches (Fig. A.159). An ESS of 80 on the fishery catch-at-age appeared reasonable (Fig. A.160) though the application of the Francis (2011) stage 2 multipliers results in slightly lower ESS than would have been achieved using the iterative mean approach used in SAW 53 (NEFSC 2012a). The input ESS did achieve reasonable fits to the observed catch-at-age (Fig. A.161.a-c) with no large residual runs or obvious year class effects apparent in the residual patterning (Fig. A.162). The Francis approach focuses on the model fits to the observed mean catch-at-age which are generally good (Fig. A.163). Overall, the fits to the mean age in the SAW55_3BLOCK_BASE model are improved over those from SAW 53 (0.96 vs. the SAW 53 values of 1.28). Fishery selectivities were flat-topped as described in depth in previous sections (Fig. A.164). The trends in selectivity, with decreasing selectivity on the younger ages through time is consistent with management measures that have gradually increased mesh sizes and minimum retention sizes. The fishery selectivity parameters are well estimated with CVs ≤ 0.10 on all parameters (Table A.84).

Fits to the NEFSC spring survey index exhibited no strong residual patterning (Fig. A.165). It is notable that the ASAP model did not fit the 2007 and 2008 index values well, with the model fits being influenced by the high CVs in these years. These two index values were the subject of considerable discussions during SAW 53 and are partly responsible for the large discrepancies between the GARM III and SAW 53 assessment results (NEFSC 2012a). The input ESS value of 15 was generally supported by the modeled estimates (Fig. A.166), though as noted with the fishery ESS values, they appear to be lower than those that would have resulted from the iterative mean approach used in SAW 53. There is a decent fit of observed to predicted age compositions (Fig. A.167). There was no strong residual patterning to the index age composition fits, although there are some small transient year class effects in the early to mid-1990s. Fits to the mean age were comparable to the fishery mean ages (Fig. A.168, RMSE=1.02) lending additional support to the input ESS.

Models fits to the NEFSC fall survey were generally better than the spring fits, with stronger coherence between the observed index and modeled estimate and less residual patterning

(Fig. A.169). ESS values of 15 are generally lower than the modeled estimates; additionally there is some suggestion of decreased ESS more recently in the time series (Fig. A.170). The fit to the age composition was generally good, with very little patterning to the survey indices age composition residuals (Fig. A.171). There does appear to be a small increase in the residuals in the more recent years which is likely related to the trends observed in the model estimated ESS. The overall fit to the mean catch-at-age is reasonable, though there is some indication of reduced fit in the most recent period (Fig. A.172) as suggested by the comparison of the input ESS to the modeled ESS values and the residual patterns in the fits to the indices-at-age.

Similar to the fits to the NEFSC surveys, the fit to the MADMF spring survey is reasonably good with the model tracking the observed index values moderately well, with no strong residual patterning (Fig. A.173). The input ESS appears generally reasonable (Fig. A.174). The MADMF spring age compositions were not fit as well as the NEFSC surveys, with the magnitude of residuals being somewhat larger for this survey relative to the others, particularly at the younger ages (Fig. A.175). However, no long runs of residuals (either positive or negative) are observed and there are no indications of year class effects. Estimated mean ages were fairly close to the observed mean ages, with a RMSE of 1.06 (Fig. A.176).

The NEFSC fall survey exhibits higher selectivity at younger ages relative to the spring survey (Table A.84, Fig. A.177). Survey catchabilities (q) are presented in Figure A.178. The q CVs were less than 20%. The NEFSC spring survey $q=0.92$ which would appear to suggest that the NEFSC spring is close to 100% efficient. Considering the calibration coefficients applied to the Bigelow survey years, this would suggest greater than 100% efficiency over the last two years. This is not necessarily a valid assumption and caution needs to be taken when interpreting the area-swept converted values of q . A full exploration of the survey q estimates is provided in Appendix A.6 along with model independent estimates of total stock biomass which support the general scale of biomass estimated by the BASE model. Long-term retrospective analyses were conducted to evaluate the patterns of survey catchability changes over time (NEFSC spring and fall and MADMF spring). While the NEFSC spring (Fig. A.179) survey q exhibited a sharp increase during the 1990s, q has remained relatively stable for the NEFSC fall (Fig. A.180) and MADMF spring (Fig. A.181) surveys. The WG discussed potential causes for this pattern, though no concrete hypotheses were put forward. One hypothesis concerned the intense aggregation of cod that was observed in the vicinity of Stellwagen Bank between 2006 and 2010; however a q retrospective on just the 1982-2002 time series suggests that the shift in q was independent of this process (Fig. A.182).

The SAW55_3BLOCK_BASE assessment model indicates that total SSB has ranged from 6,268 mt to 22,036 mt during the assessment time period, with current SSB in 2011 estimated at 9,903 mt (Table A.85, Fig. A.183). The base model estimates SSB in 2010 at 11,141 mt which is 6% lower than the SAW 53 estimate of 11,868 mt. Total January 1 biomass in 2011 is estimated at 14,728 mt (Table A.85, Fig. A.184) and F's at the end of the time series are estimated between 0.75 and 0.98 (Fig. A.183) with the 2011 fully recruited, $F_{full} = 0.86$ (Table A.86). Fishing mortalities-at-age are presented in Table A.87. The low fishing mortality on ages 1 through 3 is notable given that the maturity $A_{50\%}$ is between ages 2 and 3. The current fishery selectivity allows one to two spawning events, on average prior to entering the fishery. These patterns partly explain the persistence of the population in the presence of the high Fs over the past decade. The coefficients of variation on SSB and F have generally been less than 0.1 except at the end of the time series where CVs increased to at or

near 0.2 (Fig. A.185).

Recruitment over the past decade has been poor despite modest increases in SSB (Fig. A.186 and A.187). Age-1 recruitment has not exceeded 10 million fish in the last two decades and has been below 7 hundred thousand fish over the last decade (Table A.88). While there is an absence of a well defined stock-recruit relationship there is some indication of a relationship (Fig. A.188). The five highest recruitment events in the time series were spawned during a six year period from 1982 to 1987 where the SSB was near the highest observed in the time series, averaging over 14,000 mt annually. The current population structure is comprised primarily of fish that have not yet fully recruited to the fishery (fish age 1-3), with >80% of the population age 4 and younger (Table A.88 and Fig. A.189).

MCMC simulation was performed to obtain posterior distributions of the SSB, total B, F_{full} and F_{5-7} time series. Two MCMC chains of initial length of ten thousand were simulated with every thousandth value saved. The trace of each chain's saved draws suggests good mixing (Fig. A.190 and A.191). The lagged autocorrelations showed decreasing correlation with increased lag with correlations ≤ 0.1 beyond lag 1 (Fig. A.192 and A.193). From the MCMC distributions, 90% posterior probability intervals (PI) were calculated to provide a measure of uncertainty for the model point estimates. Time series plots of the SSB and F_{full} 90% PIs as well as plots of the posterior probability distributions for SSB_{2011} and $F_{full(2011)}$ are shown in Figures A.194 through A.197. ASAP point estimates and the 90% PIs are reported in Table A.91.

Retrospective analyses for the 2004-2011 terminal years indicates retrospective error in both F and SSB with the tendency for the model to underestimate F and overestimate SSB (Fig. A.155). The 5-year Mohn's rho value for SSB and F were 0.40 and -0.27 respectively (Table A.82). While the retrospective pattern is larger than that observed in the SAW53_BASE model, the directionality in the terminal year has shifted such that spawning stock biomass tended to be underestimated and fishing mortality overestimate. The SAW 55 WG discussed criteria to judge when to adjust for a retrospective pattern (SAW 55 WG 2012c). It was mentioned that there are no firm guidelines on when to (or not) adjust for a retrospective pattern. There was however SAW 55 WG agreement to always adjust for a consistent retrospective pattern and to do this on the numbers at age.

The ASAP model presented retrospective patterns based upon five year peels. It appeared that the retrospective pattern was transient with a one year peel showing little bias. The SAW 55 WG could not agree on general criteria to adjust for the retrospective pattern, noting that this is a broader issue than the Gulf of Maine cod assessment. The group agreed that further formulations should not adjust for the retrospective pattern given that the retrospective pattern is small, it may be transient in nature and that SAW 53 made no retrospective adjustment. This decision was supported by the SARC 55 Panel (i.e., no retrospective adjustment should be conducted for the purposes of stock status determination or short-term projections).

Diagnostics and results of the M-ramp ASAP model (SAW55_3BLOCK_BASE_M_SPLIT)

Model fits to the fishery catches were good, with no strong patterning of residuals over time and generally good agreement between modeled and observed catches (Fig. A.198). An ESS of 80 on the fishery catch-at-age appeared reasonable (Fig. A.199) though the application of

the Francis (2011) stage 2 multipliers results in slightly lower ESS than would have been achieved using the iterative mean approach used in SAW 53 (NEFSC 2012a). The input ESS did achieve reasonable fits to the observed catch-at-age (Fig. A.200.a-c) with no large residual runs or obvious year class effects apparent in the residual patterning (Fig. A.201). As noted previously, the Francis approach focuses on the model fits to the observed mean catch-at-age which were generally good (Fig. A.202). As with the $M = 0.2$ model, fishery selectivities were flat-topped as described in depth in previous sections (Fig. A.203). Similar to the $M = 0.2$ model, selectivity on the younger ages decreased over time blocks consistent with management measures that have gradually increased mesh sizes and minimum retention sizes. The fishery selectivity parameters are well estimated with CVs ≤ 0.10 on all parameters (Table A.84).

Fits to the NEFSC spring survey index exhibited no strong residual patterning (Fig. A.204). It is notable that the ASAP model did not fit the 2007 and 2008 index values well, with the model fits being influenced by the high CVs in these years. The input ESS value of 15 were generally supported by the modeled estimates (Fig. A.205), though as noted with the fishery ESS values, they appear to be lower than those that would have resulted from the iterative mean approached used in SAW 53. There is a decent fit of observed to predicted age compositions (Fig. A.206). There was no strong residual patterning to the index age composition fits, although there are some small transient year class effects in the early to mid-1990s. Fits to the mean age were comparable to the fishery mean ages (Fig. A.207, RMSE=1.02) lending additional support to the input ESS.

Models fits to the NEFSC fall survey were generally better than the spring fits, with stronger coherence between the observed index and modeled estimate and less residual patterning (Fig. A.208). ESS values of 15 are generally lower than the modeled estimates; additionally there is some suggestion of decreased ESS more recently in the time series (Fig. A.209). The fit to the age composition was generally good, with very little patterning to the survey indices age composition residuals (Fig. A.210). There's a small increase in the residuals in the more recent years which is likely related to the trends observed in the model estimated ESS. The overall fit to the mean catch-at-age is reasonable, though there is some indication of reduced fit in the most recent period (Fig. A.211) as suggested by the comparison of the input ESS to the modeled ESS values and the residual patterns in the fits to the indices-at-age.

Similar to the fits to the NEFSC surveys, the fit to the MADMF spring survey is reasonably good with the model tracking the observed index values moderately well, with no strong residual patterning (Fig. A.212). The input ESS appears generally reasonable (Fig. A.213). The MADMF spring age compositions were not fit as well as the NEFSC surveys, with the magnitude of residuals being somewhat larger for this survey relative to the others, particularly at the younger ages (Fig. A.214). However, no long runs of residuals (either positive or negative) are observed and there are no indications of year class effects. Estimated mean ages were fairly close to the observed mean ages, with a RMSE of 1.06 (Fig. A.215).

The NEFSC fall survey exhibits higher selectivity at younger ages relative to the spring survey (Table A.84, Fig. A.216). Survey catchabilities (q) are presented in Figure A.217. The q CVs were less than 20%.

The SAW55_3BLOCK_BASE_M_SPLIT assessment model indicates that total SSB has ranged from 7,930 mt to 21,531 mt during the assessment time period, with current SSB in 2011 estimated at 10,221 mt (Table A.85, Fig. A.218). Total January 1 biomass in 2011 is

estimated at 16,312 mt (Table A.85, Fig. A.219) and F's at the end of the time series are estimated between 0.60 and 0.90 (Fig. A.218) with the 2011 fully recruited, $F_{full} = 0.90$ (Table A.86). Fishing mortalities-at-age are presented in Table A.89. The low fishing mortality on ages 1 through 3 is notable given that the maturity $A_{50\%}$ is between ages 2 and 3. The current fishery selectivity allows one to two spawning events on average prior to entering the fishery. These patterns partly explain the persistence of the population in the presence of the high Fs over the past decade. The coefficients of variation on SSB and F have generally been less 0.1 except at the end of the time series where CVs increased to at or near 0.2 (Fig. A.220).

Recruitment over the past decade has been poor to moderate despite modest increases in SSB (Fig. A.221). There is no well defined stock-recruit with very little relationship between age 1 recruitment and spawning stock biomass (Fig. A.222). Age-1 recruitment has been below ten thousand fish since 2008 (Table A.90 and Fig. A.223). The current population structure is comprised primarily of fish that have not yet fully recruited to the fishery (fish age 1-3), with >80% of the population age 4 and younger (Table A.90 and Fig. A.224).

Identical to the $M = 0.2$ model, MCMC simulation was performed to obtain posterior distributions of the SSB, total B, F_{full} and F_{5-7} time series. Two MCMC chains of initial length of ten thousand were simulated with every thousandth value saved. The trace of each chain's saved draws suggests good mixing (Fig. A.225 and A.226). The lagged autocorrelations showed decreasing correlation with increased lag with correlations ≤ 0.1 beyond lag 1 (Fig. A.227 and A.228). From the MCMC distributions, 90% PIs were calculated to provide a measure of uncertainty for the model point estimates. Time series plots of the SSB and F_{full} 90% PIs as well as plots of the posterior probability distributions for SSB_{2011} and $F_{full(2011)}$ are shown in Figures A.229 through A.232. ASAP point estimates and the 90% PIs are reported in Table A.91.

Retrospective analysis for the 2004-2011 terminal years indicates retrospective error in both F and SSB with the tendency for the model to underestimate F and overestimate SSB (Fig. A.155). The 5-year Mohn's rho value for SSB and F were -0.01 and 0.06 respectively (Table A.82). This retrospective is considerably reduced relative to the SAW55_3BLOCK_BASE ($M = 0.2$) model. Both the SAW 55 WG and SARC 55 Panel agreed that no retrospective adjustment should be conducted for the purposes of stock status determination or short-term projections.

Conclusions of the SARC 55 Panel

The SARC 55 Panel recognized that one of the motivations for examining how, or if, changes in natural mortality had occurred was driven by an effort to reduce the retrospective pattern present in the $M = 0.2$ model. Given all of the information provided to the Panel, there remained considerable uncertainty in the estimates of M . The evidence for and against constant and ramped natural mortality was equivocal. As with the Working Group, the Panel was unable to reach a decision on which natural mortality values or time varying scenarios best characterized this system.

With respect to the improved diagnostics of the M -ramp model, the Panel concluded that “*...finding that including a changing M provides a better fit, is generally not sufficient to justify using such a model modification without other ecologically directed information to*

back it up" (SARC 55 2012). Noting the lack of conclusive evidence to support a change in M they determined that it was unclear as to whether a change in natural mortality was influencing the retrospective pattern or some other factor. For example, a Delphi method had been applied prior to the working group meetings to find alternative values of discard mortality rates for different gears. The retrospective pattern was worse with the lower discard mortality rates, implying that the ramp M approach could be partially aliasing unaccounted fishing mortality.

Given that there was no clear way forward for providing a single model for guiding management advice, the SARC 55 Panel put forward (accepted) both the ASAP $M = 0.2$ and M -ramp models. The consequences associated with using or disregarding either approach are outlined under TOR 8.

Other models considered by the SARC 55 Panel, but not accepted:

Historical (1932) ASAP model with Beverton-Holt stock recruit relationship

While the SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models constituted the accepted models, the SAW 55 WG felt it was worthwhile to develop candidate ASAP models that both a) utilized the historical information back to 1932; and b) fit a Beverton-Holt (BH) stock recruit (SR) function internally within the model. In this respect they provided a more similar comparison to the SCAA candidate models (though a BH model was also developed for the SCAA model). The current version of ASAP does not allow the fitting of a Ricker SR relationship. Extending the assessment back in time is necessary to establish sufficient contrast in the SR relationship such that a SR function can be estimated. Such an approach, by necessity, requires that the assessment incorporate data of lower informational quality, but also data with higher uncertainty. **It is important to note that the ASAP BH models were not presented to the SARC 55 Panel as preferred ASAP models, rather they were prepared to provide ASAP equivalents to the SCAA models (described in next section).** The SARC 55 Panel did not accept the SCAA modeling approach for the reasons outlined in the next section. Many of the SARC 55 Panel's objections to the SCAA modeling approach would also apply to the ASAP BH models (e.g., incorporation of highly uncertain historical data, influence of uncertain recruitment on steepness parameters, volatility of MSY reference points to SR functional form).

To adjust historical catches to account for un-recorded commercial discards and recreational catch the average ratio of these catches to the commercial landings between 1982 and 1988 (0.32) was applied as an adjustment factor. While the constant ratio approach was the best that could be developed by the SAW 55 WG, it is based on several critical assumptions, namely that the commercial discards were constantly proportional to commercial landings prior to 1982. This assumption may not be valid, particularly when historical landings of non-targeted fisheries such as the northern shrimp fishery are considered. There is evidence that landings, and presumably effort, of northern shrimp were greater back during the late 1960s to mid-1970s (Fig. A.233). This fishery was responsible for large amounts of Gulf of Maine cod discards, particularly during the mid-1980s (Fig. A.24) prior to implementation of the Nordmore grate. Given the fisheries tendency to catch small cod (Fig. A.53), the discard patterns of this fishery would be subject not only to relative effort, but also the year-class strength of Gulf of Maine cod. These observations suggest that the discards of Gulf of Maine

cod could have been much greater relative to landings during the late-1960s to mid-1970s. While no better direct estimates are available for the historical catches, these types of issues should be considered when determining the reliability of historical catch estimates used in stock assessments. Given the high uncertainty in the historical catches, the SAW 55 WG agreed to apply a CV of 0.4 between 1932 and 1963 and 0.2 between 1964 and 1982. Time series averages of catch weights and stock weights were used for the period prior to 1982. The assumptions used for historical catches were identical between ASAP and SCAA runs.

For the ASAP 1932 BH runs, both $M = 0.2$ and M -ramp models were developed. In each model initial guesses for numbers-at-age, fishing mortality and steepness were set. The model was free to estimate fishing mortality and steepness with no imposed penalty function/prior, though the initial numbers at age were fixed. Preliminary investigations of the ASAP 1932 Beverton-Holt model explored alternate starting points, with model convergence and results robust to these alternate starting points. The final model applied an initial $F = 0.2$ in 1932. A summary of model diagnostics are presented in Table A.92. An interesting finding from the 1932 BH runs is that there is no model preference for the M -ramp, in fact imposing the M -ramp on the 1932 BH model results in a loss of 17 objective points, though most of these differences were due to differences in the recruitment deviations. Additionally, unlike the 1982 ASAP model, the retrospective pattern is worse under the M -ramp assumption (Fig. A.234). Since the support for the M -ramp in the 1982 ASAP formulation rested in part on its ability to reduce the retrospective pattern, these results call into question the justification for an M -ramp model.

The spawning stock biomass, fishing mortality and age 1 recruitment estimate time series were nearly identical to their 1982 equivalents for the years in which overlap occurred (1982-2011; Fig. A.235). The imposed M -ramp did affect the historical time series due to the effects on the estimated SR relationship (Fig. A.236). Steepness was estimated at 0.90 in the $M = 0.2$ model and 0.82 in the M -ramp model. Overall, the corresponding reference points appeared well-estimated with CVs < 0.15 (Table A.92). Profiling over various values of steepness shows that for the $M = 0.2$ run, there is equal evidence for a steepness between about 0.85 and 0.95 (Fig. A.237) between which there is no considerable change with respect to reference points or 2011 estimates of SSB or F (Fig. A.238). Consequently, across the range of likely steepness values, there is little impact on stock status determination (Fig. A.239). The converse is not true for the M -ramp model where there is equal evidence for steepness between 0.7 and 0.9 with large implications on the estimate of F_{MSY} (ranges from approximately 0.5 to 1.1). While B_{MSY} and the 2011 SSB estimate were not highly sensitive to the steepness estimate, given that the 2011 SSB estimate (8,442 mt) was close to the B_{MSY} estimate of 7,713 mt, steepness values in the range of 0.7 and 0.9 can lead to very different perceptions about stock status (Fig. A.239).

These results are important to consider in both in the context of current stock biomass and fishing mortality and in the justification of an M -ramp. The 2011 estimates of the 1932 BH ASAP models were both below those of the 1982 models, but both generally exhibited the same time series trends and scale. The implementation of an M -ramp into the 1932 BH model degraded the model performance, particularly with respect to the retrospective patterns which was a justification for its consideration in the 1982 model.

SCAA model

Statistical catch-at-age (SCAA) assessment models were also considered by the SAW 55 WG, the details of which are provided in Appendices A.2-A.5. The primary differences between the ASAP and SCAA model formulations are the choice of starting points with the SCAA model starting in 1932 and the fitting of an internal Ricker stock-recruit relationship. The treatment of the historical input data in the SCAA model was identical to that described in the ASAP 1932 BH models detailed in Appendix A.6. There were other minor differences between the models but the WG concluded that each model series estimated similar spawning stock biomass across the range of the time series (Fig. A.240), but did note that the terminal 2011 estimates exhibited differences in scale with the SCAA model tending to estimate high biomass at the end of the time series. The WG discussed the source(s) of this difference and identified it as the weightings given to recent stock – recruitment data, with the SCAA model applying greater shrinkage to the SR relationship in the more recent years.

Both a constant $M = 0.2$ and an M -ramp model SCAA were developed and brought forward for consideration by the SARC 55 Panel. **A full description of the comparison of the ASAP and SCAA modeling approaches put forward by the SAW 55 WG is provided in Appendix A.7.** Ultimately, the SARC 55 Panel did not accept the SCAA approach. Below is the justification provided by the SARC 55 Panel (SARC 55 2012):

“While using information in the earlier part of the time series to help define a stock-recruitment relationship is laudable, it can be tricky. A number of concerns were raised and discussed regarding the use of the pre-1982 data (which was not of the same detail and quality as the post-1982 series) and the results from fitting the stock-recruitment curves to these data. Any one concern, by itself, might not have been enough to preclude the use of these methods in the assessment, but together these concerns led the Review Panel to discount the results and consequently the approach was eliminated from further consideration. These concerns can be examined from the point of view of the two parametric stock recruitment models (Ricker and Beverton-Holt) and then from the point of view of the data. These concerns are outlined below:

- *The FMSY reference point derived from the Ricker model based on the longer data series was sometimes higher than total mortality derived from surveys suggesting that FMSY estimated in this way is higher than would make sense as the stock decreased at these mortality levels. The Review Panel acknowledges that the criterion for determining survey total mortality integrates selectivity as well, but believes the above argument still holds.*
- *Although the Ricker model fit the longer data series better than other models (neither the Ricker or Beverton-Holt could be reasonably fit without including some other information, as that derived from the longer data series or some other external piece of prior information), the fit was clearly influenced by low recruitments in earlier years associated with high spawning stock biomass (SSB). The Review Panel could not decide if this was a period with low recruitment productivity driven by external forces or if it was a low recruitment period because of high SSB. If the low productivity had been estimated at two or more periods of high SSB then the Review Panel would have had more consideration into the Ricker model. There was also no evidence of density dependent effects on recruitment rate such as cannibalism.*

- *The Beverton-Holt stock-recruitment model was similarly rejected because these low recruitment points also inflated the steepness parameter to values beyond what seemed reasonable.*
- *Including the earlier catch series was necessary to fit a stock recruit relationship, however, because of the above arguments and concerns about the quality and the less detailed information available in earlier part of the data series, the Review Panel concluded that these relationships were too unreliable to provide MSY reference points for characterizing assessment advice and so all model formulations (either ASAP and SCAA) that included a stock recruitment relationship were not considered further.*
- *Regarding the low recruitment values of the 1960s, it looked like there were other avenues that could be pursued to help validate whether or not they should be included in determining stock recruitment model fits and associated reference point calculations. For example, examining evidence of ecosystem drivers would help determine if these recruitments were more likely to be evidence of density dependence or alternatively an environmental regime shift or a change in predation by other species. A general concern about the quality of the data in the earlier part of the series provides further motivation for examining the credibility of these influential points.*
- *As no standard stock-recruitment relationship could be found, the use of proxy reference points for this stock was supported.*
- *One other important related issue should be noted when using the Ricker or the Beverton-Holt relationships for data like these. The two models result in very different SSB_{MSY} and F_{MSY} reference points although the resulting recruitment levels at these points may be close to indistinguishable. Basing overfishing thresholds on such a volatile criterion may not be the best approach for establishing stable and sustainable management actions for stocks with this type of recruitment history.”*

Historical assessment retrospective

A comparison between the results of the current SAW 55 assessment (both $M = 0.2$ and M -ramp model) and the five previous assessment (SAW 33, GARM I, GARM II, GARM III, and SAW 53) is provided in Figure A.241. This historical “retrospective” examination of past model performance illustrates the general tendency of updated models to achieve higher estimates of F and lower estimates of SSB, total biomass and overall stock size over the last decade. These patterns are in addition to the intra-model retrospective patterns that are present in the existing ASAP model as well as past VPA models. Given the major changes in data that have occurred in both the SAW 55 as well as the SAW 53 benchmark assessments, the current assessment is not entirely comparable with previous assessments. Much of the scale differences between the current assessment and previous assessments are driven by changes to the underlying data (e.g., recreational catch estimates, discard mortality assumptions, weights-at-age) and not as a result of the assessment or choice of model.

TOR A.6. State the existing stock status definitions for “overfished” and “overfishing”; update or redefine biological reference points.

The existing MSY reference points based on a spawning potential ratio (SPR) of 40% were established at SAW 53 (NEFSC 2012). The overfishing definition is $F_{MSYproxy} = F_{40\%} = 0.20$. A stock is considered to be overfished if spawning biomass is less than half of SSB_{MSY} . The existing overfished definition is $\frac{1}{2} SSB_{40\%} = 0.5 \cdot 61,218 \text{ mt} = 30,609 \text{ mt}$. New reference points are warranted given the changes in fishery selectivity and fishery weights-at-age due to the revisions in recreational catch estimates and discard mortality assumptions. Additionally the M -ramp assumption has considerable impacts on recruitment estimates which will impact the estimation of SSB_{MSY} and MSY.

As noted under TOR 5, the ASAP model has the capability to estimate a Beverton-Holt stock recruit function within the model; however, model runs attempting to fit a Beverton-Holt function were unsuccessful when 1982 is used as a starting year. Analytic model-based reference points are not estimable because of insufficient contrast in the ASAP base model time series of estimated SSB and recruitment (1982-2011). *As no standard stock-recruitment relationship could be found, the use of proxy reference points for this stock was necessary.* A yield per recruit (YPR) analysis was performed using a 3-year average of weights-at-age (2009-2011) which was consistent with the approach used in SAW 53 and supported by recent observed trends. The remaining YPR inputs were time invariant (maturity-at-age) or were constant in the most recent time block of the assessment model (selectivity, natural mortality). The SARC 55 Panel concluded that for long-term projections (i.e., the establishment of reference points) natural mortality should be assumed equal to 0.2, because the longer-term historical evidence seems to indicate that $M=0.2$ is more plausible than the more recent 0.4 assumed under the M -ramp model. Given the SARC 55 Panel’s conclusions regarding natural mortality, there are only minor differences in the selectivity vectors between the $M = 0.2$ and M -ramp YPR inputs; all other inputs are identical. YPR inputs are summarized in Table A.93 for both the $M = 0.2$ and M -ramp models.

The basis for the existing reference points was derived at GARM III (NEFSC 2008), and is based on $F_{40\%}$. This decision was based on an assumed natural mortality of $M = 0.2$. The decision to use $F_{40\%}$ as a proxy was endorsed by the independent reviewers at GARM III meeting, stating that “*If recruitment and spawning stock biomass derived from the assessment are not informative about a relationship, the panel recommended use of $F_{40\%MSY}$ as a proxy for F_{MSY} (NEFSC 2002) and SSB_{MSY} proxy computed using a stochastic projection approach, also referred to as the “nonparametric approach”*” (NEFSC 2008, p979). Additional analyses by the SAW 55 WG evaluated various proxies for F_{MSY} by comparing estimated SSB and recruitment ratios (SSB/R) with expected spawning biomass per recruit (SPR) over a range of fishing mortalities ($F=20\%$ to $F80\%$ in 5% increments) to investigate the potential for replacement under equilibrium assumptions (i.e. constant harvest rate and biology over the lifespan). The SAW WG considered an analysis of replacement lines under recent productivity (approximately last 10 years) and concluded that for the $M = 0.2$ option, $F_{40\%}$ (0.18) was still appropriate (Fig. A.242). It should be noted that subsequent to the SAW/SARC 55, work was presented at SAW 56 WG that invalidates the replacement line approach for determining an appropriate spawning potential ratio and suggested that $F_{40\%}$ be maintained for fish with typical groundfish life histories (Legault and Brooks 2013). The

SARC 55 Panel recommended to maintain the $F_{40\%}$ basis for reference points for both the $M = 0.2$ and M -ramp models but noted that “... $F_{40\%}$ is necessarily the best proxy to use, rather there has yet to be compelling reasons to abandon it” (SARC 55 Panel Summary Report, 2012).

To arrive at estimates for SSB_{MSY} and a corresponding MSY, long term projections were run sampling from the empirical distribution of recruitment estimates from the preferred ASAP model. The recruitment vector included years 1982-2009; recruitment in 2010 and 2011 were not included due to their greater variance. The projection model samples from a cumulative density function derived from estimated age-1 recruitment. However, the revised model adjusts projected recruitment when SSB falls below some specified spawning biomass threshold based on a linear function that declines to zero at zero spawning stock biomass. Consistent with the SAW 53 assessment, the ‘hinge’ was set at the lowest observed SSB in the time series. For the $M = 0.2$ scenario, this was 6,300 mt and 7,900 mt for the M-ramp scenario. To approximate the distribution of the SSB and MSY distributions, the long term projections were made from 1000 estimates of numbers at age in 2011, which were estimated by performing MCMC simulation of the ASAP models (described above under TOR 5). The 2011 age 1 estimates were based on sampling from the empirical distribution of recruitment estimates from only the ten year period 2000-2009. All projections were conducted with the AGEPRO software (Age Structured Projection Model v4.1).

For the ASAP, 1982 start, $M = 0.2$ scenario, the resulting reference points and their 90% confidence intervals corresponding to $F_{MSYproxy} = F_{40\%}$ (0.18) are $SSB_{MSY} = 54,743$ mt (40,207 – 73,354 mt) and $MSY = 9,399$ mt (6,806 – 13,153 mt).

For the ASAP, 1982 start, M -ramp scenario, the resulting reference points and their 90% confidence intervals corresponding to $F_{MSYproxy} = F_{40\%}$ (0.18) are $SSB_{MSY} = 80,200$ mt (64,081 – 99,972 mt) and $MSY = 13,786$ mt (10,900 – 17,329 mt).

A detailed summary of these reference points is also provided in Table A.94.

TOR A.7. Evaluate stock status with respect to the existing model

TOR A.7.a. When working with the existing model, update it with new data and evaluate stock status (overfished and overfishing) with respect to the existing BRP estimates.

The updated SAW 53 model (SAW55_BASE) estimates 2011 SSB at 11,874 mt. This is less than the existing overfished threshold of 30,609 mt; therefore, the stock is overfished. The updated estimate of fully recruited fishing mortality (F_{full}) in 2011 is 0.59. This is greater than the overfishing limit of 0.20, and therefore, overfishing is occurring.

TOR A.7.b. Then use the newly proposed model and evaluate stock status with respect to “new” BRPs (from Cod TOR-6).

For the ASAP, 1982 start, $M = 0.2$ scenario, the revised reference points are $F_{MSYproxy} = F_{40\%} = 0.18$ and $SSB_{MSY} = 54,743$ mt (0.5 x $SSB_{MSY} = 27,372$ mt). The model estimates 2011 SSB at 9,903 mt. This is less than the overfished threshold of 27,372 mt; therefore, the stock is overfished. The estimate of 2011 fully recruited fishing mortality (F_{full}) is 0.86. This is

greater than the overfishing limit of 0.18, and therefore, overfishing is occurring.

For the ASAP, 1982 start, M -ramp scenario, the revised reference points are $F_{MSYproxy} = F_{40\%} = 0.18$ and $SSB_{MSY} = 80,200$ mt ($0.5 \times SSB_{MSY} = 40,100$ mt). The model estimates 2011 SSB at 10,221 mt. This is less than the overfished threshold of 40,100 mt; therefore, the stock is overfished. The estimate of 2011 fully recruited fishing mortality (F_{full}) is 0.90. This is greater than the overfishing limit of 0.18, and therefore, overfishing is occurring.

Under both the $M = 0.2$ and M -ramp scenarios the stock is assessed to be overfished and overfishing is occurring. It is notable that this stock has experienced a long history of overfishing relative to current reference points (Fig. A.243).

TOR A.8. Develop and apply analytical approaches to conduct single and multi-year stock projections

TOR A.8.a. Provide numerical annual projections

Short term projections of future stock status were conducted based on the current assessment results without accounting for retrospective bias. This rationale was identical to that of stock status determination. Numbers-at-age in 2012 were derived from 1000 different vectors of numbers-at-age produced from the MCMC chain with 2011 age 1 estimates based on sampling from the empirical distribution of recruitment estimates from only the ten year period 2000-2009. Biological inputs were identical to those used for reference point determination. Short term projections have used an assumed catch in 2012 of 3,767 mt. This estimate is based on the current commercial and recreational catches as well as the expected catch over the remainder of the year which has been extrapolated using the harvest trajectories from the past two years (NEFMC PDT, T. Nies pers. comm.).

Recruitment was sampled from a cumulative density function (CDF) of estimated age 1 recruitment from 1982 to 2009. The same AGEPRO model used for reference point determination was used to conduct short-term projections (i.e., model adjusts projected recruitment based on a linear function that declines to zero at zero SSB when SSB falls below some ‘hinge’ SSB-level corresponding to the lowest SSB observed in the time series). For the $M = 0.2$ scenario, the ‘hinge’ SSB value was set at 6,300 mt and 7,900 mt for the M -ramp scenario. All projections were run under the assumption of 75% F_{MSY} ($0.18 \cdot 0.75 = 0.135$).

A consequence analysis was conducted to evaluate the sensitivity of management advice to the assumptions about M (i.e. $M = 0.2$ or M -ramp). For the M -ramp scenario the projections were provided assuming that: a) M remained at 0.4; or, b) that M returns to 0.2 in the projection period.

Projection results are summarized in terms of median SSB and fishery catch (yield) under all three scenarios outlined above in Table A.95. Under 75% F_{MSY} exploitation, the stock is projected to rebuild under the $M = 0.2$ and M -ramp ($M = 0.2$) scenarios by 2022. The stock cannot rebuild under the M -ramp ($M = 0.4$) scenario since the reference points are based on an assumption of M returning to 0.2 in the long-term. It is important to note that the SARC 55 Panel was not willing to conclude that M would remain at 0.4 in perpetuity and so did not provide reference points for the M -ramp model under a long-term assumption of $M = 0.4$. A full discussion of the three scenarios evaluated is provided under TOR 8b.

TOR A.8.b. Comment on which projections seem most realistic assumptions

Consequence Analysis

The risks associated with management actions taken during 2013 – 2015 were examined by undertaking stock projections under the competing assumptions for the state of nature. For example, if the true state of nature is that natural mortality has remained unchanged at 0.2 and that stock productivity is best reflected by the 1982 – present dataset (SPR, $M = 0.2$ model), then the consequences of management actions by setting projected catch according to 75% F_{MSY} based on the two alternative states of nature were examined (M -ramp scenario with $M = 0.2$ in short-term and M -ramp scenario with $M = 0.4$ in the short term). In all cases, the 2012 catch was provided by the NEFMC Groundfish Plan Development Team. Projections were only conducted until 2015. There may be longer term consequences which might be revealed through a more extensive analysis. This is beyond the current terms of reference.

The column headers in Table A.96 and Figure A.244 represent the ‘true’ states of nature considered, these being:

- $M = 0.2$: stock dynamics and assessment based on 1982 – present dataset with M remaining at 0.2 for the projection period.
- M -ramp: stock dynamics and assessment based on 1982 – present dataset with M returning to 0.2 in the projection period.
- M -ramp: stock dynamics and assessment based on 1982 – present dataset with M remaining at 0.4 for the projection period.

The row headers in Table A.96 indicate the basis of the management action during the projected period (2013 – 2015). For example, the row header ‘ASAP, 1982 start, $M = 0.2$ ’ indicates that catch was projected assuming that the stock conditions and reference points were as per these dynamics. All projections were conducted at 75% F_{MSY} , based on the assumed state of nature and thus which establishes the catch in each cell. This is the ‘planned’ catch. The cells of the table indicate the SSB and fully recruited fishing mortality (F_{full}) which are a consequence of applying the catch based on the assumed state of nature to the SSB of the ‘true’ state of nature. The diagonal rows represent the situation in which the management actions based upon the assumed state of nature are in fact correct.

The consequence analysis is summarized in Figure A.244. As with Table A.96, the column headers indicate one of the ‘true’ states of nature. The row headers indicate whether or not catch, SSB or F_{full} is being displayed along the row. The content of each cell summarizes the consequences (reflected by the medians of the distributions in question) of assuming one state of nature when another is true. The black line in each cell indicates the catch, SSB and F_{full} for the ‘true’ state of nature. The coloured lines (for the projected period only) indicate the catch, SSB and F_{full} which result when the 75% F_{MSY} estimated catch is incorrectly based upon an alternate state of nature. The dashed lines in each figure are the B_{MSY} , F_{MSY} and MSY for the ‘true’ states of nature.

When management actions are correctly based upon a particular state of nature (the diagonals of Table A.96), a modest increase in SSB is projected between 2013 and 2015 for all three scenarios explored. The M -ramp ($M = 0.2$) scenario has the greatest rebuilding potential

whereas the M-ramp ($M = 0.4$) has the lowest rebuilding potential. Fully recruited fishing mortality declines from 0.86 ($M = 0.2$) or 0.90 (M -ramp) to 0.14 (all scenarios). Catch declines from 6,830 mt in 2011 to 1,313 - 2,582 mt in 2015 depending on the scenario with the M -ramp ($M = 0.4$) scenario resulting in the lowest yield and the M -ramp ($M = 0.2$) having the highest yield. The $M = 0.2$ scenario is an intermediate case. If the management actions are correctly based upon the ‘true’ state of nature all scenarios indicate that the stock will be in an overfished state as of 2013 (Table A.96).

The SARC 55 Panel concluded that the $M = 0.2$ projections and the M -ramp projections with M remaining at 0.4 in the short-term were equally realistic. Like the SAW 55 WG, the SARC 55 Panel could not decide which option was more plausible. The Panel concluded that if M is currently 0.2 [0.4] then it seemed more reasonable to assume that in the short-term M would remain at 0.2 [0.4]. Note that for long-term projections that Review Panel decided that M should be 0.2 under all scenarios, because the longer-term historical evidence seems to indicate that $M=0.2$ is more plausible.

The consequences of mis-specifying natural mortality (e.g., $M = 0.2$ is true state of nature and manage under M -ramp, $M = 0.4$) will not impact status determination in 2013; under all consequence analyses considered the stock will be in an overfished state in 2013. Considering only the $M = 0.2$ and M -ramp ($M = 0.4$) scenarios, the consequence of mis-specifying natural mortality will result in at most 717 mt of an over-/under-harvest of fishery yield in 2015. While the magnitude is small in terms of historical catch, this amounts to 55% of over-harvest (M -ramp is true state of nature and manage under $M = 0.2$) or a 35% under-harvest ($M = 0.2$ is true state of nature and manage under M -ramp, $M = 0.4$). Assuming an M -ramp ($M = 0.4$) when M is actually equal to 0.2 results in a lower than ‘planned’ fishing mortality and catch and higher than ‘planned’ SSB. When M is assumed to be 0.2 but an M -ramp ($M = 0.4$) is correct, fishing mortality and thus catch would be considerably higher than ‘planned’ with the result that in 2013 the stock would be experiencing overfishing (Table A.97).

TOR A.8.c. Describe this stock’s vulnerability

The Gulf of Maine cod stock is currently undergoing processes that have not been incorporated into the analytical formulations. Nevertheless, they should be considered when setting the ABC.

Since the mid-1990s, as observed in the NEFSC bottom trawl surveys and consistent with the trends in the fishery, the distribution of cod has become increasingly concentrated in the western part of the Gulf, with a gradual loss of cod from the coastal and central Gulf. Since the mid-2000s, the stock has become particularly concentrated in a small region of the western Gulf, an area which appears to be a forage ‘hotspot’ due to the presence of sand lance, a prey of cod. This biases CPUE as an indicator of the abundance of the stock as a whole.

There is uncertainty associated with natural mortality rates. Natural mortality of cod may be increasing through consumption by other fishes and marine mammals as these populations increase; however, evidence of this is lacking in the food habits data and among life history parameters. On the other hand, tagging studies suggest natural mortality levels higher than 0.2 during 2003 – 2006 time period. The tagging studies, combined with the reduced assessment model retrospective patterns were the basis of the M -ramp model. However, the

states of nature as reflected in the natural mortality rates included in the models are uncertain. For example, a Delphi method had been applied prior to the working group meetings to find alternative values of discard mortality rates for different gears. The retrospective pattern was worse with the lower discard mortality rates, implying that the ramp M approach could be partially aliasing unaccounted fishing mortality.

It may be that at low population sizes, cod experience mortality from a number of unidentified sources. High mortality, both fishing and natural will lead to a truncated age structure, implying that spawning success is increasingly dependent upon younger individuals. Murawski et al. (2001) suggest that reproduction by older females is more successful than by young females. There are a number of other factors that are known to negatively influence cod spawning success at low population sizes (Rowe et al., 2004).

If weak recruitment and low reproductive rates of Gulf of Maine cod continue, productivity and rebuilding of the stock will be less than projected. Over the last five years recruitment estimates have declined to a low level in both the $M = 0.2$ and M -ramp assessment models. Recent survey indices of recruitment indicate continued poor recruitment. Additionally, the NEFSC 2011 fall and 2012 spring survey abundance indices were the 4th lowest and the lowest in their respective time series. The MADMF 2012 spring survey biomass index was the lowest in its times series. The 2012 spring survey observations were not incorporated into the assessment formulations, implying that projections may be optimistic.

The current assessment provides a range of views of current stock status, all of which indicate that the resource is in an overfished state and has experienced a long history of overfishing. Concerns for stock status may also be apparent in the fishery. Cumulative commercial and recreational catches to date in 2012 are projected to be less than 60% of the total allocated quota (based on projected catch provided by NEFMC PDT, T. Nies pers. comm.). While this is suggestive of an overall difficulty on the part of industry to locate Gulf of Maine cod it is not definitive given other possible explanations such as sector quota restrictions on other co-occurring species. However, observations from the recreational fishery which is not subject to the same catch share system as the commercial fishery has also reported difficulty locating Gulf of Maine cod.

TOR A.9. Review, evaluate and report on the status of the SARC and Working Group research recommendations

The SAW 55 WG reviewed the status of previous research recommendations and proposed new ones to address issues raised during the three WG meetings. For all new research recommendations proposed by the SAW 55 WG, the WG has indicated relative priorities (high, medium, low) as appropriate. Many of these recommendations were felt to be common to both the Gulf of Maine and Georges Bank Atlantic cod stocks. These are indicated as ‘General’ below. The SARC 55 Panel also contributed several additional research recommendations which are included in this section.

GARM III

- The Panel recommended that historical data be used to hindcast recruitments as far back in time as possible for use in the estimation of reference points and projections.

- *Analyses to explore the use of the historical information were undertaken by the WG with the sensitivity of reference points examined.*

SAW 53

- Examine historical and contemporary estimates of cod catch in the lobster fishery. Preliminary discussions with Maine DMR suggest that the lobster bycatch may be relatively small proportional to other fishery removals.
 - *There is ongoing work through a collaboration between the University of Maine and the Maine Department of Marine Resources to estimate Atlantic cod bycatch in the Maine lobster fishery. Work is still in progress and no information was available for evaluation during SAW 55 (Y. Chen, University of Maine Orono, pers. comm.).*
 - *Observer coverage of both nearshore and offshore lobster vessels has been allocated by the Northeast Fishery Observer Program for period April 2012 – March 2013 with the specific objective of obtaining information on fishery bycatch.*
 - *The WG recommended that this research recommendation be carried forward.*
- The SAW 53 data WG had recommended that consideration be given to inclusion of the inshore strata data when switching to the *FRV* Bigelow survey time series. Sampling in these strata during both spring and fall surveys has been inconsistent or non-existent, dependent upon the stratum.
 - *The analysis presented to the SAW 55 WG indicated that inclusion of these inshore strata had minimal influence on the trends in both survey indices. It was thus recommended that these inshore strata be excluded from the SAW 55 analyses.*
 - *When it is judged that the Bigelow time series is long enough to include as a separate series, reconsideration needs to be given to adding these strata back into the survey index since there has been consistent sampling of these survey strata since the change in survey vessels in 2009.*
- Further pursue the incorporation of the Maine/New Hampshire Inshore Trawl (MENH) Survey in future assessments. The unavailability of age information and short time series have precluded this survey from being used in past assessments. While age structures are currently collected from this survey, they have not been aged.
 - *Progress has been made on the implementation and analysis of the data collected since the start of the ME/NH survey in 2000/2001; specifically, spring and fall 2005 and spring 2011 ageing has been completed and spring 2006 is in progress (S. Sherman, ME DMR, pers. comm.). Continued progress towards ageing the entire time series of collected otoliths should be considered a high priority.*
- The SARC 53 Data Working Group suggested exploration of the maturity information collected by the ME/NH survey to examine agreement with the NEFSC maturity ogives.

- *Maine DMR (S. Sherman, ME DMR, pers. comm.) provided the maturity info for the ME/NH inshore groundfish survey. These data were analyzed and presented at the SAW 55 data meeting and summarized in this report.*
- Examine the reproductive information collected from the ME/NH survey for the early years (e.g., where Downeast Maine stations were sampled to evaluate whether any of the fish were mature and if it could possibly suggest the presence of a spawning aggregation.
 - *ME DMR (S. Sherman ME DMR pers. comm.) provided maps of cod ≥ 25 cm broken down in to two time blocks (2001- 2006, 2007-2011). Additionally, maturity data were examined in terms of proportion mature by region. These data were presented at the SAW 55 data meeting and summarized in this report.*
- Examine the impacts of excluding the Commercial LPUE index from the assessment. The Commercial LPUE index exists for the year 1982 – 1993 and is no longer updated. Regulations implemented since 1994 (e.g., trip limits, area closures) limit the utility of a LPUE index that extends beyond these years. Initial modeling to explore this recommendation indicated no impact to the updated VPA and negligible impact to the ASAP base model if the Commercial LPUE index is excluded. The NDMBRPWG therefore decided to drop the Commercial LPUE index from this, and all future assessments of Gulf of Maine cod.
 - *This recommendation was included in TOR 2 of SAW 55. A number of surveys indicate that the Stellwagen Bank area appears to be a forage ‘hot spot’ for cod feeding on sand lance. As well, the VTR, observer and VMS information from the commercial fishery indicates that fishing effort since the mid-2000s has become concentrated in this area. Over the longer term, there have been a number of regulatory changes (e.g. seasonal closures, trip limits, etc) which call into question the utility of commercial LPUE as an index of GOM cod biomass. Based on these concerns, the WG recommended that the commercial LPUE index not be used in the SAW 55 assessment model. This recommendation is consistent with the findings of the recent NEFSC sponsored LPUE workshop. Given concerns comparable to those of the commercial fishery, the WG recommended that the recreational LPUE index also not be included in the GOM cod assessment model.*
- Stock definition should be re-assessed. The SARC 53 panel recommended that efforts be undertaken to reassess the stock definition for Gulf of Maine Atlantic cod. Cod is a very population-rich species, and matching the scale of the assessment to the spatial scale of the population dynamics is important to achieve reliable, accurate assessments. Several lines of evidence support this recommendation: 1) the assessment under review presents compelling evidence of a change in the distribution of cod within the current stock area. The SARC 53 panel was not able to determine whether this is solely a demographic response, but comments made during the SARC indicate that it may also relate to a reduction in the diversity of spawning times and locations; 2) there is compelling historical and contemporary evidence from natural history information and tagging studies of movements across stock boundaries that compromises the integrity of existing stock definitions, and 3) There is a wealth of historical and more recent genetic information of local stock structure and local

adaption in cod and in fish populations general at finer spatial scales than previously admitted.

- *As indicated under TOR 3, a separate process has been initiated to address this recommendation. The SAW 55 WG reported on the findings of a recent workshop on stock structure which was an element of this initiative.*
- The level, schedule and variability of natural mortality should be evaluated. Currently, the level of fishing mortality, F , estimated in Gulf of Maine Atlantic cod is substantially higher than the estimated rate of natural mortality, M . However, as managers begin to regulate harvests more effectively, F will decline and approach M . Under such circumstances the accuracy of the assumed M becomes more important. Accordingly, the SARC 53 panel recommended that efforts be increased to evaluate size-specific, age-specific and inter-annual variation in M be expanded.
 - *This was considered and reported on under TOR 4 of SAW 55. The SAW 55 WG considered analyses which provide evidence for M greater (up to 0.6) than the currently assumed value of 0.2 during 2003 – 2006. These and other analyses were the basis of the WG's decision to consider an M change model option.*
- Study of the behavior of fishers in response to changes in the distribution of the stock and to changes in management. There was clear evidence presented in the assessment and at SARC 53 of changes in the distribution of cod within the stock area. The SARC 53 panel recommended that research and analyses be conducted to: 1) understand and characterize changes in the distribution of the stock, 2) understand and characterize changes in the distribution of fishing effort and to evaluate the impacts of such changes on the pattern and biological characteristics of removals from the stock and 3) evaluate the potential for changes in the distribution of effort to be associated with changes in the distribution of vulnerability of different components of the stock to fishing mortality.
 - *As reported under TOR 2, a number of analyses were undertaken to describe GOM cod distributional changes, which particularly since 2006, appear to have been driven by prey (sand lance) spatial processes. The associated changes in commercial and recreational effort distribution, as well as regulatory changes over the longer term, imply that LPUE is no longer a representative index of abundance and led to the WG's decision to exclude these time series from the base models.*

SAW 55 WG (new recommendations)

- The tagging analysis of Miller (2012, WP 31) provided evidence of natural mortality greater than 0.4 during the 2003 – 2006 period. Historical tagging data were reported to exist, but there was no comparable analysis to which this could be compared. It would be useful to reconsider historical tagging data using modern analytical methods similar to that in Miller (2012) to allow comparisons of the estimates of natural and fishing mortality. (High)
- Improved estimates of discard mortality/survival (i.e. post capture mortality) are needed, particularly in the recreational fishery. Studies which incorporate electronic

tags and acoustic arrays would allow confirmation of the currently used estimates. (Low)

- Studies to provide information on the natural mortality of cod and inferred temporal trends are needed. Specifically, predator population estimates (i.e. pinnipeds) specific to Gulf of Maine/Georges Bank and focused stomach collection and analyses of fish and other predators would assist in evaluating whether or not natural mortality may have changed. (High)
- The SAW 55 WG noted that there may be advantages to inclusion of the tagging analysis formally within the stock assessment model. This would allow consideration of the factors affecting tagging estimates of F and M, including age/size based processes. This would be a longer-term project given the complexity of integrating the two analyses. (High, General)
- The SAW 55 WG discussed at length the appropriateness of the methods used to weight the proportions at age data within the ASAP and SCAA models. The current ASAP error assumption (multinomial) assumes that the standardized variance on the proportions at age is constant. Analyses were presented to the WG that indicated that the variance on the proportions at age was not constant and that in order to properly account for this in the model fitting process, it was necessary to employ an age-dependent weighting, as the adjusted log-normal does. While use of the multinomial would not produce biased estimates, it would likely result in the variance being over-estimated. Further, the AIC criterion would not be valid in model selection, although it was countered that the ASAP uses a penalized likelihood. This issue could not be fully resolved by the WG and further work is required to explore the appropriate weighting of the proportions at age data. (Medium, General)
- The SAW 55 WG considered an approach that incorporated the *Bigelow/Albatross* calibration coefficients within an SCAA assessment model. This allowed re-estimation of the coefficients as data on year-classes was updated. While the effect in this assessment was small, the approach may have merit and consideration should be given for the incorporation into the ASAP software. (Medium, General)
- Explore the utility of applying a random errors approach to the internal fitting of stock – recruitment relationships. This would require extensive software changes to ASAP code. (Medium, General)
- Simulations (conditioned on data) of the internal estimation of stock - recruitment functions be used to explore potential bias in the fitting of these relationships. (Medium, General)

SARC 55 Panel (new recommendations)

- Provide analysis on changes in the location and quality of preferred environment and habitats for cod and potential implications on M (adult and juvenile) and spawning potential.

- Telemetry tagging may provide a more direct way to measure natural mortality, particularly if there are local cod populations with high site fidelity.
- Consider other assessment models that include ‘smoothing’ approaches (e.g. penalized random walks) to deal with changes in fishery selectivity and natural mortality.
- Consider accounting for residual patterns and retrospective patterns using process errors. A rationale for this is that process errors can be projected into the future to potentially better account for the model/process uncertainty (indicated by residual and retrospective patterns) in projections and MSY reference points. The current approach of retrospective correcting for process error does not seem sufficient particularly in long-term projections for rebuilding analyses and reference point calculations. Uncertainty in calibrations to standardize survey time series for changes in vessels and fishing gear (i.e. doors) was not accounted for in the stock size indices. This may be a useful area for future research, although hopefully the time-series will soon be long enough that direct calibration will not be required.
- A GLM approach could be used to combine NEFSC and MADMF survey indices into two more complete indices for the Spring and Fall. The NEFSC surveys have better coverage in offshore strata, and the MADMF surveys had better coverage in inshore strata. Combining surveys would result in better coverage of the whole stock and hopefully better stock size indices.
- As part of the model building exercise, consider summarizing the information about mortality rates and trends in stock size using a survey-only assessment model such as SURBA. This could replace catch-curve estimation of Z’s. It can also be used to explore conflict (or lack thereof) between surveys and catches.
- When stock-recruit data are uncertain but the time-series is long, consider constraining R_{max} to be some reasonable value (e.g. maximum of historic assessment values) and derive MSY reference points using the constrained stock-recruit curve. There are nonparametric approaches that could be used to address sensitivity of MSY reference points to simple parametric assumptions about stock-recruitment relationships

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Tables

Table A.1. Summary of model inputs and formulations used to assess the Gulf of Maine Atlantic cod stock over the last eleven years.
 Notes: ¹1999-2000 commercial landings raised to account for commercial discards, ²1999-2001 commercial landings raised to account for commercial discards, ³Not known with certainty that MADMF time series included the spring 2002 survey, ⁴1999-2004 commercial landings were raised to account for commercial discards.

| Year | Meeting | Model | Starting year | Catch data series | | | | Survey series | | | Plus group |
|------|----------|-------|---------------|------------------------|---------------------|-----------------------|-----------------------|---------------|------------------------|-----------------|------------|
| | | | | Commercial landings | Commercial discards | Recreational landings | Recreational discards | NEFSC | MADMF | Commercial LPUE | |
| 2001 | SAW 33 | VPA | 1982 | 1982-2000 ¹ | | 1982-2000 | | 1982-2000 | 1982-2000 | 1982-1993 | 7+ |
| 2002 | GARM I | VPA | 1982 | 1982-2001 ² | | 1982-2001 | | 1982-2002 | 1982-2002 ³ | 1982-1993 | 7+ |
| 2005 | GARM II | VPA | 1982 | 1982-2004 ⁴ | | 1982-2004 | | 1982-2005 | 1982-2005 | 1982-1993 | 7+ |
| 2008 | GARM III | VPA | 1982 | 1982-2007 | 1999-2007 | 1982-2007 | | 1982-2008 | 1982-2008 | 1982-1993 | 11+ |
| 2011 | SARC 53 | ASAP | 1982 | 1982-2010 | 1982-2010 | 1982-2010 | 1982-2010 | 1982-2010 | 1982-2010 | | 9+ |

Table A.2. Summary of the results of the Gulf of Maine Atlantic cod assessments over the last eleven years and the resulting stock status determinations based on the existing biological reference points at the time of the assessment. Notes: ¹SR (BH) = Beverton-Holt stock recruitment; ²Stock status was determined using a different basis in 2001 (total biomass, 25% of BMSY; Applegate et al. 1998); ³YPR = Yield per recruit, based on 5-year averages of weights-at-age, maturity-at-age and selectivity-at-age, $F_{MSY}=F_{40\%}$. ⁴YPR = Yield per recruit, based on 3-year averages of weights-at-age, $F_{MSY}=F_{40\%}$.

| Year | Meeting | SSB (mt) _{terminal} | $F_{terminal}$ | F note | Reference point basis | SSB _{msy} (mt) | F _{msy} | MSY (mt) | Stock status |
|------|----------|------------------------------|----------------|--------------|-----------------------|------------------------------|------------------|----------|---|
| 2001 | SAW 33 | 13,100 (B=24,400) | 0.73 | F_{avg4-5} | SR (BH) ¹ | 8,000 ($B_{MSY}=90,300$ mt) | 0.230 | N/A | Not overfished, overfishing is occurring ² |
| 2002 | GARM I | 22,040 | 0.47 | F_{avg4-5} | SR (BH) ¹ | 82,830 | 0.225 | 16,600 | Overfished, overfishing is occurring |
| 2005 | GARM II | 18,800 | 0.63 | F_{avg4-5} | SR (BH) ¹ | 82,830 | 0.225 | 16,600 | Overfished, overfishing is occurring |
| 2008 | GARM III | 33,877 | 0.46 | F_{avg5-7} | YPR ³ | 58,248 | 0.237 | 10,014 | Not overfished, overfishing is occurring |
| 2011 | SARC 53 | 11,868 | 1.14 | F_{mult} | YPR ⁴ | 61,218 | 0.200 | 10,392 | Overfished, overfishing is occurring |

Table A.3. Summary of major regulatory actions that have affected the Gulf of Maine Atlantic cod fishery since 1973. For a more detailed summary of recent regulatory actions see Nies (2011).

| Date | Regulatory action | Cod end minimum mesh size (in) | Minimum fish size (in) | | Commercial trip limits | Recreational trip limits | Closures | Differential DAS Counting |
|----------|--|--------------------------------|------------------------|--------------|---|---|---|---|
| | | | Commercial | Recreational | | | | |
| 01/01/73 | | 4.5 | ? | ? | | | | |
| 01/01/77 | Groundfish FMP | 5.125 | 16 | 16 | | | | |
| 01/01/82 | | | 17 | 15 | | | | |
| 01/01/83 | | 5.5 | | | | | | |
| 01/01/89 | | | 19 | 19 | | | | |
| 04/01/92 | Shrimp trawl fishery: Nordmore grate regulation, groundfish bycatch prohibited | | | | | | | |
| 05/01/94 | Amendment 5 | 6.0 | | | | | | DAS monitor w/ reduction schedule, mandatory reporting |
| 05/01/96 | Amendment 7 | | | 20 | | | | Accelerated DAS reduction |
| 05/01/97 | Framework 20 | | | | 1000 lbs day for first 4 days, then 1500 lbs/day; no overall cap but RA had authority to reduce limit | | | |
| 05/01/98 | Framework 25 | | | | 700 lbs/day; no overall cap but RA had authority to reduce limit | | WGOM (Jeffreys Ledge, Stellwagen Bank) | |
| 06/25/98 | | | | | 400 lbs/day; no overall cap but RA had authority to reduce limit | | | |
| 02/01/99 | Framework 26 | | | | | | Additional month-block closures for February to April | |
| 05/01/99 | Framework 27 | 6.5 square/6.0 diamond | | | 200 lbs/day; no overall cap | | | |
| 05/28/99 | | | | | 30 lbs/day | | | |
| 08/03/99 | Interim rule | | | | 100 lbs/day, 500 lbs max per trip; modifications to running clock | | | |
| 01/05/00 | Framework 31 | | | | 400 lbs/day, 4000 lb/trip | | Additional month-block closures for February | |
| 06/01/00 | Framework 33 | 6.5 square/6.5 diamond | | | | | | |
| 11/01/00 | | | | | | | One month closure of Cashes Ledge | |
| 05/01/02 | Interim rule | | 22 | 23 | 500 lb/day, 4000 lb/trip | 10 cod/person | Additional month-block closures for May - June 2003; Cashes Ledge Closed year round | 20% reduction in DAS |
| 06/01/02 | Revised interim rule | | | | | | | |
| 08/01/02 | Emergency rule | | 19 | | | | | |
| 05/01/04 | Amendment 13 | | | | 800 lb/day, 4000 lb/trip | | WGOM, Cashes Ledge and rolling closures continued | Further reduction in DAS |
| 05/01/06 | Emergency rule | | | | 600 lb/day, 4000 lb/trip | | | |
| 11/22/06 | FW 42 | | | 24 | 800 lb/day, 4000 lb/trip | Possession prohibited November to March 31st | | DAS counted 2:1 in inshore GOM |
| 05/01/09 | Interim rule | | | | | Possession prohibited November to April 15 | | |
| 05/01/10 | Amendment 16 | | | | None for sector vessels, varies in-season for common pool, handgear A and B vessels (50 lb/trip - 800 lb/day, 4000 lb/trip) | 10 cod/person, Possession prohibited November to April 15 | Some changes to rolling closures for sector vessels | DAS counted in 24-hour blocks; no differential DAS counting except as AMs |
| 05/01/11 | Framework 45 | | | | | | Whaleback closure April 1 - June 30 (commercial and recreational) | |
| 05/01/12 | Framework 47 | | | 19 | | 9 cod/person, Possession prohibited November to April 15 | | |

Table A.4. Preliminary estimates of updated Atlantic cod gutted-to-live weight conversion factors summarized by quarter, sex and maturity stage. Raw data were provided by the Northeast Fisheries Science Center's Cooperative Research Program (C. Sarro pers. comm.).

| Class variable | | N | Mean | Standard Deviation | Coefficient of Variation | Median | 5th Percentile | 95th Percentile |
|-----------------------|------------------|----------|-------------|---------------------------|---------------------------------|---------------|-----------------------|------------------------|
| All | | 422 | 1.200 | 0.091 | 0.076 | 1.190 | 1.077 | 1.381 |
| Quarter | 1 | 123 | 1.170 | 0.076 | 0.065 | 1.154 | 1.077 | 1.308 |
| | 2 | 179 | 1.191 | 0.099 | 0.083 | 1.172 | 1.070 | 1.381 |
| | 3 | 120 | 1.244 | 0.077 | 0.062 | 1.235 | 1.144 | 1.387 |
| | 4 | 0 | | | | | | |
| Sex | Male | 98 | 1.172 | 0.085 | 0.073 | 1.157 | 1.070 | 1.323 |
| | Female | 84 | 1.214 | 0.110 | 0.091 | 1.189 | 1.077 | 1.415 |
| Maturity | Immature | 5 | 1.125 | 0.086 | 0.076 | 1.125 | 1.050 | 1.263 |
| | Developing | 35 | 1.209 | 0.072 | 0.059 | 1.200 | 1.107 | 1.333 |
| | Ripe | 19 | 1.241 | 0.098 | 0.079 | 1.222 | 1.103 | 1.488 |
| | Ripe and running | 11 | 1.193 | 0.068 | 0.057 | 1.174 | 1.111 | 1.333 |
| | Spent | 7 | 1.119 | 0.033 | 0.030 | 1.133 | 1.077 | 1.160 |
| | Resting | 85 | 1.154 | 0.063 | 0.054 | 1.143 | 1.077 | 1.267 |
| | Unknown | 1 | 1.222 | | | 1.222 | 1.222 | 1.222 |

Table A.5. Summary of the number of Atlantic cod otoliths sampled from Northeast Fisheries Science Center (NEFSC) bottom trawl surveys from 1970 to 2012 by stock and age. Otoliths that have not been aged are not included in this summary.

| Age | Georges Bank | Gulf of Maine |
|------------|---------------------|----------------------|
| 0 | 372 | 188 |
| 1 | 2353 | 1378 |
| 2 | 4112 | 2544 |
| 3 | 3919 | 2832 |
| 4 | 2813 | 2462 |
| 5 | 1556 | 1420 |
| 6 | 781 | 782 |
| 7 | 369 | 378 |
| 8 | 190 | 171 |
| 9 | 79 | 116 |
| 10 | 54 | 59 |
| 11 | 28 | 33 |
| 12 | 14 | 30 |
| 13 | 4 | 16 |
| 14 | 6 | 18 |
| 15 | 3 | 3 |
| 16 | 1 | 3 |
| 17 | | 1 |
| 18 | 1 | |

Table A.6. Number of Atlantic cod maturity samples taken from the Northeast Fisheries Science Center (NEFSC) spring survey from 1970 to 2012 by year.

| Year | Unknown | Male | Female | Total |
|-------------|----------------|-------------|---------------|--------------|
| 1970 | 6 | 41 | 51 | 98 |
| 1971 | | 23 | 40 | 63 |
| 1972 | 2 | 31 | 50 | 83 |
| 1974 | 1 | 35 | 66 | 102 |
| 1975 | 4 | 42 | 75 | 121 |
| 1976 | 3 | 75 | 71 | 149 |
| 1977 | | 70 | 88 | 158 |
| 1978 | | 37 | 64 | 101 |
| 1979 | 13 | 96 | 119 | 228 |
| 1980 | | 35 | 56 | 91 |
| 1981 | 5 | 112 | 106 | 223 |
| 1982 | 4 | 74 | 91 | 169 |
| 1983 | 2 | 77 | 66 | 145 |
| 1984 | 1 | 40 | 65 | 106 |
| 1985 | | 47 | 81 | 128 |
| 1986 | 1 | 44 | 56 | 101 |
| 1987 | 2 | 77 | 46 | 125 |
| 1988 | 32 | 64 | 59 | 155 |
| 1989 | 3 | 68 | 74 | 145 |
| 1990 | 1 | 56 | 57 | 114 |
| 1991 | 1 | 62 | 70 | 133 |
| 1992 | 1 | 51 | 61 | 113 |
| 1993 | | 45 | 63 | 108 |
| 1994 | 1 | 61 | 45 | 107 |
| 1995 | | 39 | 36 | 75 |
| 1996 | | 58 | 60 | 118 |
| 1997 | | 60 | 63 | 123 |
| 1998 | | 73 | 55 | 128 |
| 1999 | 5 | 80 | 71 | 156 |
| 2000 | 9 | 78 | 70 | 157 |
| 2001 | 1 | 46 | 79 | 126 |
| 2002 | 3 | 121 | 135 | 259 |
| 2003 | | 156 | 121 | 277 |
| 2004 | 2 | 23 | 40 | 65 |
| 2005 | | 52 | 52 | 104 |
| 2006 | 7 | 63 | 59 | 129 |
| 2007 | | 85 | 127 | 212 |
| 2008 | 1 | 60 | 79 | 140 |
| 2009 | 6 | 148 | 229 | 383 |
| 2010 | | 118 | 130 | 248 |
| 2011 | | 46 | 58 | 104 |
| 2012 | 8 | 152 | 177 | 337 |

Table A.7. Gulf of Maine Atlantic cod female maturity ogive. The time series average incorporated data collected the Northeast Fisheries Science Center (NEFSC) spring survey between 1970 and 2012.

| Age (years) | Proportion Mature | Lower 95% CI | Upper 95% CI |
|-------------|-------------------|--------------|--------------|
| 0 | 0.03 | 0.02 | 0.03 |
| 1 | 0.09 | 0.08 | 0.11 |
| 2 | 0.29 | 0.26 | 0.31 |
| 3 | 0.61 | 0.59 | 0.64 |
| 4 | 0.86 | 0.84 | 0.88 |
| 5 | 0.96 | 0.95 | 0.97 |
| 6 | 0.99 | 0.99 | 0.99 |
| 7 | 1.00 | 1.00 | 1.00 |
| 8 | 1.00 | 1.00 | 1.00 |
| 9 | 1.00 | 1.00 | 1.00 |
| 10 | 1.00 | 1.00 | 1.00 |
| 11 | 1.00 | 1.00 | 1.00 |

Table A.8. Estimates of Gulf of Maine of Atlantic cod catch (mt) by fleet (commercial, recreational) and disposition (landed, discarded) from 1982 to 2011.

| Year | Recreational discards (mt) | Recreational landings (mt) | Commercial discards (mt) | Commercial landings (mt) | Total catch (mt) |
|------|----------------------------|----------------------------|--------------------------|--------------------------|------------------|
| 1982 | 8.1 | 2816.7 | 805.4 | 13465.9 | 17096.2 |
| 1983 | 17.6 | 1772.8 | 829.1 | 13867.4 | 16486.8 |
| 1984 | 16.6 | 1266.8 | 858.9 | 10725.3 | 12867.6 |
| 1985 | 16.9 | 2765.9 | 962.9 | 10645.3 | 14390.9 |
| 1986 | 10.1 | 1928.4 | 964.2 | 9669.6 | 12572.4 |
| 1987 | 48.0 | 3547.2 | 884.0 | 7526.2 | 12005.4 |
| 1988 | 13.5 | 1688.5 | 682.9 | 7948.2 | 10333.2 |
| 1989 | 76.1 | 1957.2 | 786.9 | 10550.7 | 13370.8 |
| 1990 | 66.7 | 2246.7 | 1560.6 | 15439.7 | 19313.7 |
| 1991 | 68.0 | 2287.2 | 663.9 | 17959.0 | 20978.1 |
| 1992 | 35.5 | 623.6 | 668.6 | 11019.4 | 12347.0 |
| 1993 | 101.9 | 1011.9 | 479.8 | 8366.7 | 9960.4 |
| 1994 | 100.6 | 721.7 | 207.5 | 8030.2 | 9060.1 |
| 1995 | 96.2 | 627.2 | 235.4 | 6606.8 | 7565.6 |
| 1996 | 81.0 | 498.6 | 157.2 | 7019.8 | 7756.7 |
| 1997 | 58.8 | 236.3 | 87.1 | 5432.1 | 5814.3 |
| 1998 | 72.2 | 353.1 | 78.5 | 4074.3 | 4578.1 |
| 1999 | 71.7 | 577.2 | 1021.9 | 1407.4 | 3078.1 |
| 2000 | 137.6 | 967.1 | 946.1 | 3771.8 | 5822.7 |
| 2001 | 227.5 | 1967.6 | 1545.4 | 4314.4 | 8054.9 |
| 2002 | 286.9 | 1254.8 | 1329.1 | 3638.3 | 6509.1 |
| 2003 | 282.4 | 1607.7 | 741.0 | 3865.6 | 6496.7 |
| 2004 | 201.4 | 1150.9 | 631.1 | 3782.3 | 5765.7 |
| 2005 | 267.3 | 1346.9 | 269.5 | 3557.6 | 5441.3 |
| 2006 | 194.0 | 702.3 | 342.3 | 3029.4 | 4268.0 |
| 2007 | 316.9 | 1042.2 | 178.4 | 3989.8 | 5527.3 |
| 2008 | 315.4 | 1267.2 | 349.2 | 5443.5 | 7375.2 |
| 2009 | 292.4 | 1357.1 | 752.3 | 5952.9 | 8354.7 |
| 2010 | 384.5 | 1758.2 | 170.8 | 5356.4 | 7669.9 |
| 2011 | 334.2 | 1799.1 | 98.8 | 4597.9 | 6830.0 |

Table A.9. Historical estimates of Gulf of Maine of Atlantic cod catch (mt) by fleet (commercial, recreational) and disposition (landed, discarded) from 1932 to 1981. *Estimates of both United States (US) and foreign fleet commercial landings are shown. No estimates of recreational catch are available prior to 1981 and no estimates of commercial discards are available pre-1982.*

| Year | US recreational discards (mt) | US recreational catch (mt) | US commercial discards (mt) | US commercial landings (mt) | Foreign landings (mt) | Total catch (mt) |
|------|-------------------------------|----------------------------|-----------------------------|-----------------------------|-----------------------|------------------|
| 1932 | | | | 5,858.0 | | 5,858.0 |
| 1933 | | | | 7,025.0 | | 7,025.0 |
| 1934 | | | | 11,619.0 | | 11,619.0 |
| 1935 | | | | 9,679.0 | | 9,679.0 |
| 1936 | | | | 7,442.0 | | 7,442.0 |
| 1937 | | | | 7,432.0 | | 7,432.0 |
| 1938 | | | | 7,547.0 | | 7,547.0 |
| 1939 | | | | 5,504.0 | | 5,504.0 |
| 1940 | | | | 5,836.0 | | 5,836.0 |
| 1941 | | | | 6,124.0 | | 6,124.0 |
| 1942 | | | | 6,679.0 | | 6,679.0 |
| 1943 | | | | 9,397.0 | | 9,397.0 |
| 1944 | | | | 10,516.0 | | 10,516.0 |
| 1945 | | | | 14,532.0 | | 14,532.0 |
| 1946 | | | | 9,248.0 | | 9,248.0 |
| 1947 | | | | 6,916.0 | | 6,916.0 |
| 1948 | | | | 7,462.0 | | 7,462.0 |
| 1949 | | | | 7,033.0 | | 7,033.0 |
| 1950 | | | | 5,062.0 | | 5,062.0 |
| 1951 | | | | 3,567.0 | | 3,567.0 |
| 1952 | | | | 3,011.0 | | 3,011.0 |
| 1953 | | | | 3,121.0 | | 3,121.0 |
| 1954 | | | | 3,411.0 | | 3,411.0 |
| 1955 | | | | 3,171.0 | | 3,171.0 |
| 1956 | | | | 2,693.0 | | 2,693.0 |
| 1957 | | | | 2,562.0 | | 2,562.0 |
| 1958 | | | | 4,670.0 | | 4,670.0 |
| 1959 | | | | 3,795.0 | | 3,795.0 |
| 1960 | | | | 3,448.0 | | 3,448.0 |
| 1961 | | | | 3,216.0 | | 3,216.0 |
| 1962 | | | | 2,989.0 | | 2,989.0 |
| 1963 | | | | 2,595.0 | | 2,595.0 |
| 1964 | | | | 3,217.4 | 25.0 | 3,242.4 |
| 1965 | | | | 3,611.5 | 148.0 | 3,759.5 |
| 1966 | | | | 3,841.1 | 384.0 | 4,225.1 |
| 1967 | | | | 5,526.6 | 297.0 | 5,823.6 |
| 1968 | | | | 6,076.0 | 61.0 | 6,137.0 |
| 1969 | | | | 7,828.4 | 327.0 | 8,155.4 |
| 1970 | | | | 7,511.7 | 449.0 | 7,960.7 |
| 1971 | | | | 7,192.5 | 282.0 | 7,474.5 |
| 1972 | | | | 6,786.1 | 141.0 | 6,927.1 |
| 1973 | | | | 6,061.1 | 77.0 | 6,138.1 |
| 1974 | | | | 7,425.4 | 125.0 | 7,550.4 |
| 1975 | | | | 8,676.1 | 112.0 | 8,788.1 |
| 1976 | | | | 9,877.7 | 16.0 | 9,893.7 |
| 1977 | | | | 11,992.8 | | 11,992.8 |
| 1978 | | | | 11,890.1 | | 11,890.1 |
| 1979 | | | | 10,972.3 | | 10,972.3 |
| 1980 | | | | 12,514.9 | | 12,514.9 |
| 1981 | 18.8 | 4,111.5 | | 12,381.6 | | 16,512.0 |

Table A.10. Coefficients of variation (CV) associated with the landings allocation procedure (AA tables, Wigley et al. 2008) for Gulf of Maine Atlantic cod commercial landings.

| Year | CV |
|------|-------|
| 1994 | 0.003 |
| 1995 | 0.012 |
| 1996 | 0.003 |
| 1997 | 0.003 |
| 1998 | 0.003 |
| 1999 | 0.007 |
| 2000 | 0.003 |
| 2001 | 0.002 |
| 2002 | 0.003 |
| 2003 | 0.002 |
| 2004 | 0.003 |
| 2005 | 0.002 |
| 2006 | 0.002 |
| 2007 | 0.001 |
| 2008 | 0.001 |
| 2009 | 0.001 |
| 2010 | 0.003 |
| 2011 | 0.002 |

Table A.11. Relative differences between VTR and VMS-based allocation of Gulf of Maine Atlantic cod by stock and year (from Palmer and Wigley 2012).

| Year | Stock | |
|------|-------|------|
| | GBK | GOM |
| 2004 | 0.7 | -1.9 |
| 2005 | 2.2 | -5.0 |
| 2006 | 0.2 | -0.2 |
| 2007 | 0.8 | -0.8 |
| 2008 | 0.6 | -0.4 |
| 2009 | 1.0 | -0.6 |
| 2010 | 2.0 | -1.0 |
| 2011 | 2.0 | -1.0 |

Table A.12. Estimates of total United States landings of Gulf of Maine Atlantic cod associated with ‘non-dealer’ transactions from 1994 to 2011. These estimates are obtained from information reported on Vessel Trip Reports (VTRs).

| Year | Home consumption (mt) | Bait (mt) | LUMF (mt) | Total non-dealer transactions (mt) | Total commercial landings (mt) | Percentage of total dealer landings (%) |
|------|-----------------------|-----------|-----------|------------------------------------|--------------------------------|---|
| 1994 | 0.9 | | | 0.9 | 8030.2 | 0.0 |
| 1995 | 3.5 | 0.1 | | 3.5 | 6606.8 | 0.1 |
| 1996 | 8.3 | 0.1 | | 8.4 | 7019.8 | 0.1 |
| 1997 | 3.2 | | | 3.2 | 5432.1 | 0.1 |
| 1998 | 3.3 | 0.0 | | 3.3 | 4074.3 | 0.1 |
| 1999 | 4.0 | 0.0 | | 4.1 | 1407.4 | 0.3 |
| 2000 | 5.3 | 0.0 | | 5.4 | 3771.8 | 0.1 |
| 2001 | 6.7 | 0.2 | | 6.9 | 4314.4 | 0.2 |
| 2002 | 6.6 | | | 6.6 | 3638.3 | 0.2 |
| 2003 | 6.3 | | | 6.3 | 3865.6 | 0.2 |
| 2004 | 4.0 | | | 4.0 | 3782.3 | 0.1 |
| 2005 | 3.1 | 0.0 | | 3.1 | 3557.6 | 0.1 |
| 2006 | 2.4 | | | 2.4 | 3029.4 | 0.1 |
| 2007 | 1.6 | 0.1 | | 1.7 | 3989.8 | 0.0 |
| 2008 | 2.0 | | | 2.0 | 5443.5 | 0.0 |
| 2009 | 1.2 | 0.0 | | 1.2 | 5952.9 | 0.0 |
| 2010 | 3.5 | 0.0 | 0.5 | 4.0 | 5356.4 | 0.1 |
| 2011 | 0.7 | 0.1 | 0.5 | 1.3 | 4597.9 | 0.0 |

Table A.13. Total number of Gulf of Maine Atlantic cod biological samples taken from commercial landings by market category and year between 1969 and 2011.

| Year | Scrod (0814) | | | | Market (0813) | | | | Large (0811) | | | | Total | Total lengths (excludes unclassified) | Sampling intensity (lengths/sample) | | | |
|------|--------------|----|----|----|---------------|----|----|----|--------------|----|-----|-----|-------|---|---|-------|--|--|
| | Quarter | | | | Quarter | | | | Quarter | | | | | | | | | |
| | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | | | | | | |
| 1969 | | | | | | | | | | | | | | | | | | |
| 1970 | | | | | 1 | | | | | | | | | 1 | 100 | 100.0 | | |
| 1971 | | | | | | | | | | | | | | | | | | |
| 1972 | | | | | | | | | | | | | | | | | | |
| 1973 | | | | | | | | | | | | | | | | | | |
| 1974 | 1 | | | | | | | | 1 | | | | | 2 | 203 | 101.5 | | |
| 1975 | | 1 | | 1 | | | | | | | | | | 2 | 248 | 124.0 | | |
| 1976 | | | | | | | | | | | | | | | | | | |
| 1977 | 1 | 1 | 3 | 3 | | 1 | 2 | 1 | | 1 | 1 | | | 14 | 2,525 | 180.4 | | |
| 1978 | 3 | 2 | 1 | | 2 | 2 | 2 | 1 | | | | | | 14 | 2,256 | 161.1 | | |
| 1979 | 1 | | 1 | 2 | | 1 | 2 | 1 | | | | | | 8 | 755 | 94.4 | | |
| 1980 | 3 | 1 | 1 | | | | | | | | | | | 5 | 364 | 72.8 | | |
| 1981 | 1 | 1 | 1 | 3 | | 1 | 3 | | | 1 | | | | 11 | 1,189 | 108.1 | | |
| 1982 | 2 | 3 | 3 | 2 | 2 | 2 | 3 | 1 | | 2 | 1 | 2 | | 23 | 3,848 | 167.3 | | |
| 1983 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 1 | 1 | 2 | 2 | 2 | | 29 | 5,241 | 180.7 | | |
| 1984 | 7 | 5 | 6 | 7 | 4 | 3 | 5 | 6 | 1 | 6 | 3 | 2 | | 55 | 3,925 | 71.4 | | |
| 1985 | 5 | 6 | 7 | 5 | 9 | 6 | 7 | 4 | 7 | 5 | 3 | 6 | | 70 | 5,284 | 75.5 | | |
| 1986 | 5 | 5 | 6 | 3 | 5 | 6 | 8 | 3 | 1 | 5 | 4 | 3 | | 54 | 4,069 | 75.4 | | |
| 1987 | 5 | 4 | 3 | 4 | 4 | 5 | 3 | 5 | 4 | 2 | 3 | 1 | | 43 | 3,188 | 74.1 | | |
| 1988 | 4 | 2 | 4 | 4 | 1 | 5 | 3 | 5 | 1 | 2 | 2 | | | 33 | 2,619 | 79.4 | | |
| 1989 | 3 | 3 | 4 | 3 | 4 | 2 | 5 | 4 | 2 | 1 | 1 | 1 | | 33 | 2,718 | 82.4 | | |
| 1990 | 3 | 7 | 3 | 5 | 4 | 7 | 4 | 3 | | 2 | 1 | | | 39 | 2,981 | 76.4 | | |
| 1991 | 2 | 10 | 4 | 4 | 5 | 11 | 12 | 3 | | 3 | 3 | 1 | | 58 | 4,676 | 80.6 | | |
| 1992 | 2 | 8 | 6 | 3 | 6 | 7 | 7 | 3 | 3 | 1 | 1 | 4 | | 51 | 4,086 | 80.1 | | |
| 1993 | 3 | 3 | 3 | 1 | 1 | 2 | 4 | 1 | 1 | 1 | 2 | 1 | | 23 | 1,686 | 73.3 | | |
| 1994 | | 2 | 2 | 4 | 1 | 6 | 3 | 5 | | 2 | 3 | 2 | | 30 | 2,658 | 88.6 | | |
| 1995 | 4 | 3 | 2 | 4 | 2 | 8 | 2 | 2 | | 3 | | 1 | | 31 | 2,557 | 82.5 | | |
| 1996 | 5 | 4 | 7 | 9 | 6 | 9 | 11 | 11 | 1 | 2 | 3 | 3 | | 71 | 6,486 | 91.4 | | |
| 1997 | 7 | 13 | 3 | 10 | 12 | 11 | 10 | 9 | 2 | 8 | 2 | 2 | | 89 | 7,559 | 84.9 | | |
| 1998 | 4 | 7 | | 3 | 9 | 9 | 9 | 5 | 1 | | 2 | 1 | | 50 | 4,536 | 90.7 | | |
| 1999 | 6 | | | | 3 | 1 | 1 | | 2 | | | | | 13 | 1,073 | 82.5 | | |
| 2000 | 13 | 6 | 5 | 7 | 16 | 14 | 5 | 9 | | | | | | 76 | 5,921 | 77.9 | | |
| 2001 | 4 | 4 | 4 | 7 | 4 | 10 | 8 | 16 | 2 | 15 | 18 | 20 | | 112 | 7,117 | 63.5 | | |
| 2002 | 3 | 2 | | 1 | 16 | 3 | 6 | 5 | 50 | 8 | 16 | 19 | | 129 | 5,263 | 40.8 | | |
| 2003 | 5 | 1 | 17 | 8 | 14 | 8 | 25 | 19 | 50 | 34 | 34 | 33 | | 248 | 11,479 | 46.3 | | |
| 2004 | 17 | 11 | 6 | 22 | 18 | 23 | 15 | 15 | 37 | 20 | 11 | 27 | | 222 | 11,210 | 50.5 | | |
| 2005 | 23 | 29 | 33 | 16 | 14 | 15 | 22 | 19 | 21 | 41 | 72 | 64 | | 369 | 10,163 | 27.5 | | |
| 2006 | 15 | 8 | 8 | 3 | 17 | 21 | 18 | 12 | 48 | 49 | 62 | 63 | | 324 | 10,770 | 33.2 | | |
| 2007 | 10 | 6 | 11 | 8 | 7 | 14 | 18 | 17 | 43 | 73 | 102 | 60 | | 371 | 10,623 | 28.6 | | |
| 2008 | 13 | 7 | 5 | 7 | 12 | 15 | 13 | 11 | 58 | 72 | 73 | 71 | | 357 | 10,922 | 30.6 | | |
| 2009 | 9 | | 2 | 14 | 10 | 17 | 20 | 37 | 61 | 97 | 114 | 135 | | 516 | 14,871 | 28.8 | | |
| 2010 | 4 | 2 | | 9 | 30 | 22 | 42 | 21 | 79 | 52 | 77 | 33 | | 371 | 17,451 | 47.0 | | |
| 2011 | 6 | 7 | 3 | 13 | 23 | 33 | 32 | 36 | 23 | 71 | 49 | 41 | | 337 | 18,682 | 55.4 | | |

Table A.14. Total numbers of Gulf of Maine Atlantic cod lengths sampled from commercial landings by market category and year between 1969 and 2011. Sampling intensity is expressed as metric tons landings per 100 lengths sampled (*200 metric tons per 100 lengths is an unofficial NAFO/ICNAF standard*). Cells shaded in grey indicate where lengths were aggregated semi-annually. Cells shaded orange indicate where lengths were aggregated annually. Aggregation occurred when length sampling was insufficient; a general criterion of 100 lengths/block was used to determine sufficiency.

| Year | Scrod (0814) | | | | Market (0813) | | | | Large (0811) | | | | Unclassified (0815) | | | | Total lengths | Landings (mt) | Metric tons/100 lengths | |
|------|--------------|-----|-----|------|---------------|------|------|------|--------------|------|------|------|---------------------|-----|-----|---|---------------|---------------|-------------------------|--------|
| | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | | | | |
| 1969 | | | | | | | | | | | | | | | | | 114 | 7,928 | 6867.0 | |
| 1970 | | | | | 100 | | | | | | | | | | | | 287 | 387 | 7,512 | 1941.0 |
| 1971 | | | | | | | | | | | | | | | | | | 0 | 7,193 | |
| 1972 | | | | | | | | | | | | | | | | | | 0 | 6,786 | |
| 1973 | | | | | | | | | | | | | | | | | | 0 | 6,061 | |
| 1974 | 102 | | | | | | | | 101 | | | | | | | | | 203 | 7,425 | 3657.8 |
| 1975 | | 186 | | 62 | | | | | | | | | | | | | | 248 | 8,676 | 3498.4 |
| 1976 | | | | | | | | | | | | | | | | | | 101 | 9,878 | 6291.5 |
| 1977 | 101 | 66 | 402 | 1012 | | 277 | 371 | 64 | | 80 | 152 | | | | | | | 2,525 | 11,993 | 475.0 |
| 1978 | 407 | 455 | 65 | | 370 | 304 | 500 | 100 | | | 55 | | | | | | | 2,256 | 11,890 | 527.0 |
| 1979 | 56 | | 58 | 116 | | 100 | 237 | 188 | | | | | | | | | | 755 | 10,972 | 1453.3 |
| 1980 | 213 | 100 | 51 | | | | | | | | | | | | | | | 212 | 576 | 12,515 |
| 1981 | 52 | 57 | 81 | 236 | | | | 82 | 471 | | 210 | | | | | | | 1,189 | 12,382 | 1041.3 |
| 1982 | 401 | 488 | 484 | 308 | 418 | 309 | 665 | 345 | | 208 | 64 | 158 | 97 | 102 | 122 | | 4,169 | 13,466 | 323.0 | |
| 1983 | 712 | 626 | 578 | 253 | 396 | 1021 | 583 | 200 | 56 | 205 | 514 | 97 | | | | | 5,294 | 13,867 | | |
| 1984 | 344 | 271 | 342 | 378 | 396 | 264 | 443 | 551 | 75 | 552 | 204 | 105 | 94 | | | | 4,019 | 10,725 | 266.9 | |
| 1985 | 263 | 352 | 449 | 241 | 837 | 565 | 677 | 351 | 542 | 341 | 263 | 403 | | | | | 5,284 | 10,645 | 201.5 | |
| 1986 | 229 | 264 | 319 | 160 | 520 | 608 | 834 | 329 | 75 | 279 | 269 | 183 | | | | | 4,069 | 9,670 | 237.6 | |
| 1987 | 281 | 232 | 165 | 271 | 344 | 490 | 351 | 399 | 157 | 150 | 258 | 90 | | | | | 3,188 | 7,526 | 236.1 | |
| 1988 | 298 | 99 | 215 | 249 | 59 | 539 | 291 | 481 | 59 | 194 | 135 | | | | | | 2,619 | 7,948 | 303.5 | |
| 1989 | 154 | 170 | 201 | 174 | 401 | 204 | 506 | 409 | 195 | 102 | 104 | 98 | | | | | 2,718 | 10,551 | 388.2 | |
| 1990 | 156 | 362 | 165 | 260 | 409 | 715 | 370 | 300 | | 136 | 108 | | | | | | 2,981 | 15,440 | 517.9 | |
| 1991 | 100 | 533 | 192 | 215 | 514 | 1034 | 1137 | 275 | | 302 | 273 | 101 | | | | | 4,676 | 17,959 | 384.1 | |
| 1992 | 118 | 443 | 320 | 180 | 633 | 725 | 592 | 263 | 297 | 142 | 75 | 298 | | | | | 4,086 | 11,019 | 269.7 | |
| 1993 | 159 | 173 | 174 | 55 | 97 | 173 | 393 | 106 | 65 | 87 | 141 | 63 | | | | | 1,753 | 8,367 | 477.3 | |
| 1994 | | 102 | 107 | 181 | 97 | 576 | 324 | 567 | | 184 | 322 | 198 | | | | | 2,658 | 8,030 | 302.1 | |
| 1995 | 211 | 196 | 107 | 249 | 170 | 807 | 215 | 224 | | 280 | | 98 | | | | | 2,557 | 6,607 | 258.4 | |
| 1996 | 278 | 275 | 491 | 691 | 596 | 961 | 1165 | 1178 | 68 | 200 | 303 | 280 | | | | | 6,486 | 7,020 | 108.2 | |
| 1997 | 520 | 848 | 188 | 751 | 1235 | 1071 | 991 | 880 | 190 | 539 | 201 | 145 | | | | | 7,559 | 5,432 | 71.9 | |
| 1998 | 295 | 383 | | 101 | 911 | 951 | 1103 | 436 | 99 | | 175 | 82 | | | | | 4,536 | 4,074 | 89.8 | |
| 1999 | 385 | | | | 311 | 108 | 58 | | 211 | | | | | | | | 1,073 | 1,407 | 131.2 | |
| 2000 | 694 | 304 | 294 | 426 | 1588 | 1167 | 409 | 924 | | | | 115 | | | | | 5,921 | 3,772 | 63.7 | |
| 2001 | 189 | 215 | 216 | 404 | 428 | 984 | 697 | 1548 | 172 | 474 | 892 | 898 | | | | | 7,117 | 4,314 | 60.6 | |
| 2002 | 106 | 80 | 39 | | 1365 | 260 | 411 | 395 | 1192 | 397 | 524 | 494 | | | | | 5,263 | 3,638 | 69.1 | |
| 2003 | 254 | 66 | 214 | 73 | 1121 | 705 | 1762 | 1402 | 1179 | 1432 | 1583 | 1688 | | | | | 11,479 | 3,866 | 33.7 | |
| 2004 | 361 | 299 | 233 | 73 | 1384 | 1887 | 1288 | 994 | 2049 | 1419 | 283 | 940 | 25 | | | | 11,235 | 3,782 | 33.7 | |
| 2005 | 73 | 193 | 324 | 506 | 919 | 1095 | 1384 | 1362 | 790 | 709 | 1330 | 1478 | | | | | 10,404 | 3,558 | 34.2 | |
| 2006 | 494 | 167 | 294 | 125 | 1291 | 1412 | 1075 | 753 | 1552 | 871 | 1348 | 1388 | | | | | 10,770 | 3,029 | 28.1 | |
| 2007 | 291 | 174 | 315 | 293 | 584 | 1188 | 1521 | 1488 | 654 | 811 | 1887 | 1417 | | | | | 10,689 | 3,990 | 37.3 | |
| 2008 | 536 | 251 | 203 | 85 | 969 | 1403 | 1196 | 927 | 712 | 1314 | 1753 | 1573 | | | | | 10,922 | 5,443 | 49.8 | |
| 2009 | 407 | | 62 | 141 | 800 | 1601 | 1791 | 2601 | 954 | 1656 | 2304 | 2554 | | | | | 14,871 | 5,953 | 40.0 | |
| 2010 | 150 | 53 | | 199 | 2679 | 1762 | 2788 | 1741 | 1428 | 2106 | 2561 | 1984 | | | | | 17,451 | 5,356 | 30.7 | |
| 2011 | 287 | 320 | 144 | 577 | 2005 | 2848 | 2674 | 3260 | 1141 | 2250 | 1884 | 1292 | | | | | 18,682 | 4,598 | 24.6 | |

Table A.15. Total numbers of Gulf of Maine Atlantic cod ages sampled from commercial landings by quarter between 1977 and 2011.

| Year | Quarter | | | | Landings (mt) | Metric tons/100 ages |
|------|---------|------|------|------|---------------|----------------------|
| | 1 | 2 | 3 | 4 | | |
| 1977 | 20 | 114 | 229 | 205 | 568 | 11992.8 |
| 1978 | 124 | 124 | 115 | 20 | 383 | 11890.1 |
| 1979 | 10 | 20 | 48 | 52 | 130 | 10972.3 |
| 1980 | 35 | 27 | 15 | | 77 | 12514.9 |
| 1981 | 12 | 15 | 67 | 170 | 264 | 12381.6 |
| 1982 | 194 | 237 | 251 | 183 | 865 | 13465.9 |
| 1983 | 277 | 513 | 400 | 158 | 1348 | 13867.4 |
| 1984 | 245 | 350 | 296 | 337 | 1228 | 10725.3 |
| 1985 | 446 | 377 | 397 | 323 | 1543 | 10645.3 |
| 1986 | 243 | 360 | 398 | 173 | 1174 | 9669.6 |
| 1987 | 252 | 229 | 226 | 228 | 935 | 7526.2 |
| 1988 | 131 | 223 | 187 | 196 | 737 | 7948.2 |
| 1989 | 206 | 129 | 203 | 165 | 703 | 10550.7 |
| 1990 | 140 | 302 | 171 | 150 | 763 | 15439.7 |
| 1991 | 126 | 447 | 385 | 152 | 1110 | 17959.0 |
| 1992 | 220 | 298 | 264 | 178 | 960 | 11019.4 |
| 1993 | 72 | 130 | 186 | 49 | 437 | 8366.7 |
| 1994 | 21 | 195 | 149 | 308 | 673 | 8030.2 |
| 1995 | 144 | 311 | 101 | 126 | 682 | 6606.8 |
| 1996 | 190 | 315 | 426 | 449 | 1380 | 7019.8 |
| 1997 | 395 | 632 | 331 | 285 | 1643 | 5432.1 |
| 1998 | 192 | 325 | 276 | 199 | 992 | 4074.3 |
| 1999 | 227 | 27 | 11 | | 265 | 1407.4 |
| 2000 | 639 | 481 | 205 | 396 | 1721 | 3771.8 |
| 2001 | 280 | 574 | 674 | 950 | 2478 | 4314.4 |
| 2002 | 1320 | 301 | 437 | 347 | 2405 | 3638.3 |
| 2003 | 1046 | 1111 | 1948 | 1525 | 5630 | 3865.6 |
| 2004 | 1880 | 1011 | 425 | 228 | 3544 | 3782.3 |
| 2005 | 494 | 644 | 1117 | 1287 | 3542 | 3557.6 |
| 2006 | 1109 | 806 | 1225 | 1197 | 4337 | 3029.4 |
| 2007 | 719 | 1020 | 1138 | 1030 | 3907 | 3989.8 |
| 2008 | 858 | 1225 | 1213 | 1173 | 4469 | 5443.5 |
| 2009 | 947 | 1407 | 1684 | 2222 | 6260 | 5952.9 |
| 2010 | 1335 | 1235 | 1856 | 1103 | 5529 | 5356.4 |
| 2011 | 735 | 1867 | 1555 | 1412 | 5569 | 4597.9 |

Table A.16. Percent of Gulf of Maine Atlantic cod length observations missing corresponding age information by market category and quarter. Cells shaded in grey indicate where lengths were aggregated semi-annually. Cells where the imputation percentage exceeded 5% are highlighted in bold italics. Cells where no imputation was required are null.

| Year | Scrod (0814) | | | | Market (0813) | | | | Large (0811) | | | |
|------|--------------|-----|-----|-----|---------------|---|-----|-------------|--------------|-------------|-------------|------------|
| | Quarter | | | | Quarter | | | | Quarter | | | |
| | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 |
| 1982 | 0.5 | 0.4 | 3.3 | 0.3 | | | | | 22.6 | | 12.2 | |
| 1983 | | 0.2 | | | | | | | 1.1 | | 0.2 | |
| 1984 | | | 0.3 | 0.3 | | | | | | | 0.5 | 1.0 |
| 1985 | | | | | | | | | 0.4 | | 1.1 | 0.2 |
| 1986 | | | | | | | | | 0.8 | | | 3.8 |
| 1987 | | | | | | | | | 0.6 | | 0.6 | |
| 1988 | 1.3 | | | | | | | 0.8 | | | 2.2 | |
| 1989 | | | | | | | | 18.8 | | 19.6 | | 2.5 |
| 1990 | | | | | | | | | | | | |
| 1991 | 44.0 | | 4.7 | 0.9 | 2.5 | | 4.3 | | 4.3 | | 11.4 | 33.7 |
| 1992 | 1.7 | | | | | | | | 5.4 | 12.0 | | 10.2 |
| 1993 | | | 0.4 | | 8.1 | | 0.5 | 12.3 | | 2.6 | | 9.3 |
| 1994 | | | 0.9 | | | | 0.3 | | | 19.6 | 7.5 | 1.0 |
| 1995 | 21.3 | | | | 1.2 | | | | 0.7 | | 28.6 | |
| 1996 | 3.6 | | | | 0.2 | | | | | | 1.0 | 5.0 |
| 1997 | | 0.7 | | 0.3 | 0.5 | | | | 14.7 | 0.4 | | 2.1 |
| 1998 | 5.7 | | | | | | | | 4.8 | | | |
| 1999 | | | | | | | | | | | 2.8 | |
| 2000 | 0.1 | | | | | | | | | | 2.6 | |
| 2001 | | | | | | | | | 1.7 | 0.6 | 0.2 | 0.1 |
| 2002 | | | | | | | | | 0.3 | 1.0 | | |
| 2003 | | | 1.0 | | | | | | 0.4 | 0.1 | 0.1 | 0.1 |
| 2004 | | 0.7 | | | | | | | 0.1 | | 1.1 | 0.7 |
| 2005 | | | | | | | | | 0.4 | 0.1 | 0.3 | 0.1 |
| 2006 | | | | | | | | | 0.2 | 0.9 | 0.3 | 0.5 |
| 2007 | | | | | | | | | 0.6 | 0.1 | 0.7 | 0.4 |
| 2008 | | | | | | | | 0.2 | | 0.1 | 0.2 | |
| 2009 | | | 0.5 | | | | | | | 0.2 | | |
| 2010 | | | | | | | | | | | 0.0 | 0.1 |
| 2011 | | | | | | | | | | | | |

Table A.17. Total Gulf of Maine Atlantic cod commercial landings-at-age (numbers) from 1982 to 2011.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|--------|-----------|-----------|-----------|-----------|---------|--------|--------|--------|--------|--------|-------|-------|-------|-------|-------|-------|
| 1982 | 27,609 | 1,335,509 | 1,634,173 | 1,116,072 | 619,571 | 51,241 | 69,146 | 59,375 | 43,415 | 32,683 | 6,285 | 898 | | | | | |
| 1983 | | 833,083 | 2,413,843 | 1,067,910 | 627,331 | 407,393 | 44,212 | 57,669 | 25,845 | 12,747 | 3,800 | 3,515 | 1,719 | 2,599 | | | |
| 1984 | 2,782 | 425,538 | 1,227,232 | 1,504,575 | 396,710 | 195,918 | 96,402 | 9,105 | 16,794 | 14,229 | 11,957 | 2,335 | 3,863 | 1,235 | | | |
| 1985 | | 387,614 | 1,440,985 | 1,002,193 | 615,000 | 123,315 | 73,198 | 32,430 | 3,962 | 10,619 | 2,438 | 4,573 | 1,583 | 470 | | | |
| 1986 | | 85,363 | 2,187,322 | 818,717 | 239,742 | 161,736 | 38,700 | 27,497 | 19,813 | 4,745 | 1,497 | 3,940 | 2,434 | 306 | | | |
| 1987 | 442 | 193,735 | 627,766 | 1,116,907 | 267,706 | 64,579 | 45,981 | 5,481 | 8,410 | 9,270 | 182 | 607 | | 2,129 | | | |
| 1988 | | 167,468 | 1,356,369 | 907,960 | 400,942 | 58,792 | 21,864 | 20,247 | 3,257 | 2,438 | 1,213 | | | 606 | | | |
| 1989 | | 322,130 | 1,486,592 | 1,354,890 | 451,857 | 70,570 | 58,876 | 7,931 | 2,238 | 9,000 | 3,945 | | 1,127 | 1,127 | | | |
| 1990 | | 210,618 | 3,403,626 | 2,227,578 | 452,797 | 151,887 | 25,246 | 24,675 | 7,680 | 16,034 | 11,764 | 2,353 | 3,597 | | | | |
| 1991 | | 198,915 | 609,915 | 4,543,525 | 904,421 | 138,556 | 42,961 | 25,983 | 7,877 | 4,698 | 2,571 | | | | | | |
| 1992 | | 302,552 | 527,720 | 432,280 | 1,969,905 | 213,021 | 77,420 | 5,837 | 4,488 | 1,042 | | | | | | | |
| 1993 | | 25,866 | 1,543,228 | 729,548 | 92,745 | 464,198 | 37,780 | 11,264 | | | | | | | | | |
| 1994 | | 29,014 | 1,055,313 | 1,170,244 | 240,940 | 63,586 | 69,917 | 28,114 | 6,108 | 384 | 1,008 | | | | | | |
| 1995 | | 183,724 | 938,703 | 1,056,404 | 207,195 | 28,494 | 6,521 | 17,992 | 580 | 2,228 | | | | | | | |
| 1996 | | 55,763 | 507,349 | 1,763,068 | 375,559 | 35,144 | 3,903 | 413 | 845 | | | | | | | | |
| 1997 | | 77,455 | 434,378 | 435,036 | 800,750 | 67,415 | 5,368 | 2,080 | 393 | 636 | | | | | | | |
| 1998 | | 87,919 | 391,916 | 544,744 | 139,369 | 187,088 | 27,507 | 4,853 | 1,495 | 762 | | | | | | | |
| 1999 | | 2,858 | 179,688 | 191,438 | 66,127 | 23,995 | 22,398 | 7,504 | 1,035 | | | | | | | | |
| 2000 | | 102,341 | 258,469 | 501,545 | 124,105 | 66,295 | 9,007 | 6,465 | | | | | | | | | |
| 2001 | | 43,737 | 471,763 | 326,442 | 206,475 | 65,902 | 38,490 | 5,509 | 8,803 | 1,006 | | | | | | | |
| 2002 | | 1,439 | 111,287 | 433,957 | 170,415 | 102,971 | 41,667 | 12,019 | 3,750 | 4,055 | 434 | 80 | | 40 | | | |
| 2003 | | 8,113 | 47,543 | 198,476 | 380,859 | 120,697 | 52,001 | 19,769 | 9,173 | 4,250 | 2,812 | 472 | | | | | |
| 2004 | | 492 | 142,749 | 130,172 | 220,142 | 170,502 | 52,305 | 26,442 | 13,941 | 6,789 | 1,414 | 620 | | | | | |
| 2005 | | 1,217 | 37,890 | 423,154 | 64,419 | 178,040 | 83,220 | 21,459 | 12,366 | 5,056 | 3,125 | 1,817 | 500 | | | | |
| 2006 | | 777 | 115,306 | 181,958 | 300,653 | 21,412 | 62,692 | 29,111 | 10,477 | 5,994 | 2,537 | 1,242 | 953 | 180 | | | |
| 2007 | | 5,209 | 95,694 | 629,852 | 99,105 | 178,429 | 5,952 | 15,582 | 7,698 | 3,753 | 1,468 | 1,323 | 1,174 | 126 | 345 | | |
| 2008 | | 4,142 | 283,069 | 465,757 | 600,316 | 53,944 | 82,494 | 2,490 | 6,652 | 3,224 | 986 | 473 | 367 | 234 | 104 | 21 | |
| 2009 | | 2,700 | 283,610 | 718,934 | 333,800 | 199,827 | 16,653 | 20,518 | 857 | 2,311 | 1,072 | 952 | 224 | 127 | 61 | 49 | |
| 2010 | | 1,683 | 121,449 | 578,192 | 463,641 | 114,076 | 59,845 | 8,069 | 2,947 | 446 | 476 | 162 | 112 | 17 | 28 | | |
| 2011 | | 534 | 97,964 | 296,737 | 396,070 | 256,786 | 26,149 | 29,090 | 4,906 | 1,177 | 196 | 538 | 68 | 178 | | | |

Table A.18. Coefficients of variation (CV) associated with the estimates of Gulf of Maine Atlantic cod commercial landings numbers-at-age from 1982 to 2011. CVs greater than 0.3 are shaded grey.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 | Age 11 | Age 12 | Age 13 | Age 14 | Age 15 | Age 16 |
|---------|--------|-------|-------|-------|-------|-------|-------|-------|-------|--------|--------|--------|--------|--------|--------|--------|
| 1984 | 0.7443 | 0.12 | 0.04 | 0.02 | 0.04 | 0.06 | 0.06 | 0.17 | 0.16 | 0.22 | 0.20 | 0.39 | 0.29 | 0.69 | | |
| 1985 | | 0.08 | 0.06 | 0.04 | 0.03 | 0.05 | 0.05 | 0.10 | 0.25 | 0.14 | 0.27 | 0.35 | 0.48 | 0.76 | | |
| 1986 | | 0.18 | 0.05 | 0.04 | 0.06 | 0.08 | 0.14 | 0.13 | 0.20 | 0.44 | 0.56 | 0.37 | 0.65 | 0.89 | | |
| 1987 | 1.3501 | 0.19 | 0.07 | 0.04 | 0.07 | 0.09 | 0.15 | 0.29 | 0.28 | 0.43 | 0.90 | 0.44 | | 0.68 | | |
| 1988 | | 0.29 | 0.06 | 0.05 | 0.06 | 0.09 | 0.15 | 0.24 | 0.48 | 0.81 | 0.81 | | | 1.32 | | |
| 1989 | | 0.38 | 0.08 | 0.09 | 0.07 | 0.14 | 0.24 | 0.33 | 0.56 | 0.23 | 0.34 | | 0.68 | 0.69 | | |
| 1990 | | 0.26 | 0.07 | 0.08 | 0.13 | 0.24 | 0.47 | 0.36 | 0.41 | 0.26 | 0.28 | 0.67 | 0.70 | | | |
| 1991 | | 0.23 | 0.15 | 0.04 | 0.11 | 0.12 | 0.23 | 0.31 | 0.27 | 1.02 | 0.64 | | | | | |
| 1992 | | 0.18 | 0.20 | 0.13 | 0.06 | 0.11 | 0.18 | 0.62 | 0.56 | 0.88 | | | | | | |
| 1993 | | 0.89 | 0.09 | 0.18 | 0.29 | 0.11 | 0.34 | 0.41 | | | | | | | | |
| 1994 | | 0.49 | 0.10 | 0.07 | 0.27 | 0.25 | 0.21 | 0.22 | 0.64 | 1.02 | 0.89 | | | | | |
| 1995 | | 0.25 | 0.12 | 0.09 | 0.10 | 0.35 | 0.23 | 0.21 | 1.05 | 0.61 | | | | | | |
| 1996 | | 0.27 | 0.10 | 0.04 | 0.14 | 0.20 | 0.28 | 0.95 | 0.69 | | | | | | | |
| 1997 | | 0.20 | 0.09 | 0.07 | 0.06 | 0.14 | 0.32 | 0.27 | 0.62 | 0.60 | | | | | | |
| 1998 | | 0.16 | 0.11 | 0.07 | 0.15 | 0.15 | 0.27 | 0.37 | 0.49 | 0.99 | | | | | | |
| 1999 | | | 0.19 | 0.12 | 0.31 | 0.36 | 0.23 | 0.17 | 0.58 | | | | | | | |
| 2000 | | 0.14 | 0.08 | 0.06 | 0.12 | 0.23 | 0.49 | 0.55 | | | | | | | | |
| 2001 | | 0.24 | 0.06 | 0.07 | 0.08 | 0.11 | 0.14 | 0.30 | 0.28 | 0.59 | | | | | | |
| 2002 | | 1.11 | 0.22 | 0.05 | 0.09 | 0.07 | 0.11 | 0.15 | 0.29 | 0.26 | 0.48 | 1.21 | | 1.38 | | |
| 2003 | | 0.35 | 0.17 | 0.05 | 0.03 | 0.06 | 0.07 | 0.10 | 0.17 | 0.19 | 0.23 | 0.46 | | | | |
| 2004 | | 1.38 | 0.11 | 0.07 | 0.07 | 0.06 | 0.09 | 0.13 | 0.21 | 0.23 | 0.49 | 0.75 | | | | |
| 2005 | | 0.66 | 0.15 | 0.05 | 0.08 | 0.09 | 0.08 | 0.12 | 0.12 | 0.15 | 0.21 | 0.26 | 0.42 | | | |
| 2006 | | 1.02 | 0.17 | 0.06 | 0.04 | 0.14 | 0.09 | 0.09 | 0.14 | 0.11 | 0.17 | 0.22 | 0.27 | 0.56 | | |
| 2007 | | 0.49 | 0.13 | 0.04 | 0.08 | 0.10 | 0.27 | 0.19 | 0.12 | 0.15 | 0.25 | 0.23 | 0.27 | 0.69 | 0.46 | |
| 2008 | | 0.72 | 0.10 | 0.05 | 0.05 | 0.13 | 0.08 | 0.39 | 0.16 | 0.17 | 0.29 | 0.38 | 0.44 | 0.56 | 0.80 | 1.43 |
| 2009 | | 0.52 | 0.10 | 0.05 | 0.09 | 0.07 | 0.18 | 0.12 | 0.25 | 0.17 | 0.26 | 0.26 | 0.40 | 0.59 | 0.90 | 1.01 |
| 2010 | | 0.50 | 0.12 | 0.04 | 0.04 | 0.08 | 0.10 | 0.13 | 0.16 | 0.38 | 0.34 | 0.66 | 0.67 | 1.38 | 1.42 | |
| 2011 | | 0.28 | 0.04 | 0.02 | 0.01 | 0.01 | 0.02 | 0.02 | 0.07 | 0.14 | 0.34 | 0.26 | 0.72 | 0.45 | | |
| Average | 1.05 | 0.43 | 0.11 | 0.06 | 0.10 | 0.13 | 0.19 | 0.27 | 0.35 | 0.43 | 0.42 | 0.46 | 0.50 | 0.82 | 0.90 | 1.22 |

Table A.19. Mean weights-at-age (kg) of commercially landed Gulf of Maine Atlantic cod from 1982 to 2011.

| Year | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 | Age 10 | Age 11 | Age 12 | Age 13 | Age 14 | Age 15 | Age 16 |
|---------|-------|-------|-------|-------|-------|-------|-------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 1982 | | 0.831 | 1.177 | 1.669 | 2.790 | 5.006 | 7.097 | 9.580 | 9.945 | 12.789 | 19.365 | 16.480 | 22.443 | | | | |
| 1983 | | | 1.172 | 1.621 | 2.428 | 3.812 | 6.058 | 5.982 | 10.480 | 11.548 | 11.138 | 18.890 | 12.669 | 24.552 | 22.224 | | |
| 1984 | | 0.569 | 1.179 | 1.656 | 2.679 | 3.568 | 5.563 | 8.541 | 10.290 | 13.711 | 14.485 | 14.318 | 15.430 | 17.886 | 19.285 | | |
| 1985 | | | 1.312 | 1.740 | 2.820 | 4.528 | 5.610 | 8.436 | 11.238 | 12.479 | 14.280 | 13.394 | 16.112 | 16.739 | 22.012 | | |
| 1986 | | | 1.392 | 1.819 | 2.905 | 4.691 | 6.272 | 7.994 | 9.826 | 13.592 | 13.496 | 15.888 | 15.808 | 20.232 | 16.834 | | |
| 1987 | | 0.998 | 1.369 | 1.719 | 3.252 | 4.805 | 6.912 | 9.318 | 10.769 | 14.810 | 16.101 | 13.418 | 8.066 | | 22.379 | | |
| 1988 | | | 1.293 | 1.943 | 2.448 | 5.282 | 5.315 | 6.374 | 9.951 | 10.434 | 17.787 | 9.857 | | | 21.886 | | |
| 1989 | | | 1.314 | 1.763 | 3.055 | 4.242 | 5.943 | 9.379 | 13.425 | 16.500 | 20.410 | 22.606 | | 27.911 | 27.896 | | |
| 1990 | | | 1.247 | 1.660 | 2.238 | 4.380 | 7.816 | 11.229 | 12.270 | 15.999 | 16.344 | 22.690 | 23.134 | 22.138 | | | |
| 1991 | | | 1.489 | 1.834 | 2.412 | 4.031 | 7.164 | 9.689 | 12.261 | 15.093 | 6.203 | 24.937 | | | | | |
| 1992 | | | 1.608 | 1.941 | 2.899 | 3.070 | 5.699 | 10.984 | 10.766 | 13.418 | 19.072 | | | | | | |
| 1993 | | | 1.356 | 1.930 | 2.350 | 4.595 | 5.802 | 9.649 | 13.673 | | | | | | | | |
| 1994 | | | 1.434 | 1.955 | 3.186 | 3.349 | 6.350 | 7.787 | 12.422 | 10.012 | 22.008 | 22.643 | | | | | |
| 1995 | | | 1.588 | 1.774 | 2.838 | 5.187 | 7.054 | 11.466 | 13.223 | 19.756 | 23.143 | | | | | | |
| 1996 | | | 1.746 | 2.258 | 2.337 | 3.532 | 7.523 | 11.759 | 14.795 | 16.331 | | | | | | | |
| 1997 | | | 1.846 | 2.291 | 3.093 | 3.162 | 4.829 | 9.027 | 12.177 | 15.625 | 17.749 | | | | | | |
| 1998 | | | 1.396 | 2.020 | 2.726 | 4.025 | 4.376 | 7.235 | 12.111 | 17.500 | 15.060 | | | | | | |
| 1999 | | | 1.545 | 1.741 | 2.539 | 3.390 | 5.049 | 7.563 | 10.220 | 12.279 | | | | | | | |
| 2000 | | | 1.736 | 2.608 | 3.635 | 4.678 | 6.158 | 5.600 | 8.939 | | | | | | | | |
| 2001 | | | 1.937 | 2.556 | 3.400 | 5.036 | 6.544 | 7.684 | 9.213 | 8.945 | 17.660 | | | | | | |
| 2002 | | | 1.326 | 2.706 | 3.378 | 4.269 | 6.300 | 7.072 | 8.965 | 10.167 | 10.786 | 15.353 | 17.249 | | 18.746 | | |
| 2003 | | | 1.871 | 2.475 | 3.279 | 4.321 | 5.544 | 7.584 | 8.892 | 10.909 | 12.121 | 13.709 | 14.362 | | | | |
| 2004 | | | 1.648 | 2.689 | 3.686 | 4.261 | 5.976 | 7.590 | 9.902 | 12.654 | 14.059 | 11.423 | 22.553 | | | | |
| 2005 | | | 1.926 | 2.274 | 3.118 | 4.584 | 4.793 | 6.447 | 8.066 | 11.054 | 13.942 | 14.901 | 15.362 | 19.605 | | | |
| 2006 | | | 2.671 | 2.540 | 3.437 | 3.877 | 4.905 | 5.673 | 7.605 | 9.709 | 12.724 | 16.000 | 15.761 | 20.480 | 20.326 | | |
| 2007 | | | 2.090 | 2.616 | 3.317 | 4.053 | 5.014 | 6.518 | 7.182 | 10.140 | 12.199 | 13.344 | 14.213 | 17.126 | 21.784 | 21.757 | |
| 2008 | | | 1.848 | 2.768 | 3.145 | 3.811 | 4.777 | 6.036 | 6.106 | 8.583 | 11.258 | 13.800 | 16.189 | 19.251 | 19.918 | 18.735 | |
| 2009 | | | 1.939 | 2.766 | 3.532 | 3.972 | 4.775 | 6.007 | 8.367 | 11.208 | 10.805 | 12.934 | 15.971 | 15.803 | 22.452 | 22.812 | |
| 2010 | | | 2.228 | 2.731 | 3.528 | 4.268 | 4.874 | 5.550 | 8.478 | 10.152 | 11.016 | 13.209 | 12.519 | 16.891 | 20.103 | 16.834 | |
| 2011 | | | 1.746 | 2.724 | 3.389 | 4.094 | 4.988 | 5.934 | 6.076 | 11.750 | 12.190 | 17.376 | 17.827 | 23.845 | 19.502 | | |
| Average | | 0.799 | 1.614 | 2.160 | 2.995 | 4.196 | 5.836 | 7.990 | 10.254 | 12.755 | 14.823 | 16.056 | 16.216 | 20.189 | 21.096 | 19.946 | 24.398 |

Table A.20. Fractions of the total Gulf of Maine Atlantic cod observed to have been discarded by the commercial fishery from 1989 to 2011, broken down by gear type. Gears contributing greater than 5% of the total observed discards in any year are shaded grey.

| Year | Total observed discards (mt) | Fraction of total observed discards | | | | | | | | | |
|------|------------------------------|-------------------------------------|----------|---------------------|----------------------|--------------|---------------------|--------------------------|----------------------------|-------|--|
| | | Longline | Handline | Otter trawl (mt) | | Shrimp trawl | Sink Gillnet (mt) | | | Other | |
| | | | | Small mesh (< 5.5") | Large mesh (>= 5.5") | | Small mesh (< 5.5") | Large mesh (5.5 - 7.99") | Extra large mesh (>= 8.0") | | |
| 1989 | 4.1 | 0.00 | 0.00 | 0.03 | 0.37 | 0.37 | 0.00 | 0.23 | 0.00 | 0.00 | |
| 1990 | 5.7 | 0.00 | 0.00 | 0.00 | 0.37 | 0.34 | 0.00 | 0.29 | 0.00 | 0.00 | |
| 1991 | 11.3 | 0.00 | 0.00 | 0.00 | 0.23 | 0.14 | 0.00 | 0.63 | 0.00 | 0.00 | |
| 1992 | 9.7 | 0.01 | 0.00 | 0.00 | 0.35 | 0.06 | 0.00 | 0.58 | 0.00 | 0.00 | |
| 1993 | 4.6 | 0.01 | 0.00 | 0.00 | 0.21 | 0.02 | 0.00 | 0.76 | 0.00 | 0.00 | |
| 1994 | 1.0 | 0.00 | 0.00 | 0.00 | 0.24 | 0.10 | 0.00 | 0.62 | 0.04 | 0.01 | |
| 1995 | 2.0 | 0.00 | 0.00 | 0.10 | 0.50 | 0.02 | 0.00 | 0.33 | 0.06 | 0.00 | |
| 1996 | 1.1 | 0.00 | 0.01 | 0.10 | 0.12 | 0.01 | 0.00 | 0.65 | 0.11 | 0.01 | |
| 1997 | 0.4 | 0.00 | 0.00 | 0.06 | 0.21 | 0.02 | 0.00 | 0.62 | 0.07 | 0.03 | |
| 1998 | 0.9 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.96 | 0.02 | 0.01 | |
| 1999 | 11.3 | 0.00 | 0.00 | 0.02 | 0.07 | 0.00 | 0.00 | 0.91 | 0.00 | 0.00 | |
| 2000 | 11.3 | 0.00 | 0.00 | 0.00 | 0.68 | 0.00 | 0.00 | 0.31 | 0.01 | 0.00 | |
| 2001 | 14.5 | 0.00 | 0.00 | 0.00 | 0.66 | 0.00 | 0.00 | 0.32 | 0.01 | 0.00 | |
| 2002 | 21.3 | 0.00 | 0.00 | 0.04 | 0.65 | 0.00 | 0.00 | 0.28 | 0.03 | 0.00 | |
| 2003 | 36.5 | 0.02 | 0.00 | 0.04 | 0.63 | 0.00 | 0.00 | 0.24 | 0.06 | 0.00 | |
| 2004 | 34.0 | 0.00 | 0.00 | 0.02 | 0.34 | 0.00 | 0.00 | 0.43 | 0.21 | 0.00 | |
| 2005 | 28.1 | 0.16 | 0.00 | 0.07 | 0.36 | 0.00 | 0.00 | 0.31 | 0.09 | 0.00 | |
| 2006 | 14.3 | 0.17 | 0.00 | 0.04 | 0.61 | 0.00 | 0.00 | 0.16 | 0.02 | 0.00 | |
| 2007 | 13.2 | 0.14 | 0.00 | 0.01 | 0.67 | 0.00 | 0.00 | 0.14 | 0.03 | 0.00 | |
| 2008 | 33.3 | 0.06 | 0.00 | 0.01 | 0.86 | 0.00 | 0.00 | 0.05 | 0.02 | 0.00 | |
| 2009 | 80.9 | 0.02 | 0.00 | 0.00 | 0.86 | 0.00 | 0.00 | 0.10 | 0.01 | 0.00 | |
| 2010 | 33.8 | 0.03 | 0.00 | 0.01 | 0.61 | 0.00 | 0.00 | 0.26 | 0.07 | 0.01 | |
| 2011 | 39.4 | 0.12 | 0.04 | 0.01 | 0.59 | 0.00 | 0.00 | 0.21 | 0.02 | 0.00 | |

Table A.21. Preliminary estimates of Gulf of Maine Atlantic cod commercial handline discards from SAW 53 (NEFSC 2012).

| Year | Trips | Handline (mt) | CV |
|------|-------|---------------|------|
| 1989 | | | |
| 1990 | | | |
| 1991 | | | |
| 1992 | 2 | 0.0 | |
| 1993 | | | |
| 1994 | 2 | | |
| 1995 | 1 | | |
| 1996 | 2 | | |
| 1997 | | | |
| 1998 | | | |
| 1999 | 1 | | |
| 2000 | | | |
| 2001 | | | |
| 2002 | | | |
| 2003 | 1 | | |
| 2004 | 3 | 0.0 | |
| 2005 | 4 | 34.4 | 0.69 |
| 2006 | 2 | 0.0 | |
| 2007 | 5 | 6.9 | 0.62 |
| 2008 | | | |
| 2009 | 3 | 75.9 | 0.49 |
| 2010 | 10 | 44.1 | 0.98 |
| 2011 | 30 | | |

Table A.22. Number of Ruhle and haddock separator trawl trips recorded in the commercial dealer data and in the at-sea observer data from the Gulf of Maine. Fractional trips in the dealer data are a function of the stock allocation process used to partition dealer-reported landings to stock area.

| Year | NEGEAR | Number of dealer trips | Number of observed trips |
|------|--------|---------------------------|-----------------------------|
| 2009 | 054 | 6.0 | |
| 2009 | 057 | | 2 |
| 2010 | 054 | 0.3 | |
| 2010 | 057 | 12.1 | 17 |
| 2011 | 054 | 1.0 | |
| 2011 | 057 | 5.9 | 31 |

Table A.23. Total number of Gulf of Maine trips (statistical areas 464, 465, 467, 511-515) observed from 1989 to 2011, summarized by gear type. *The 2010-11 numbers include trips observed by both at-sea monitors and observers.*

| Year | Longline | Otter trawl | | Shrimp trawl | Sink Gillnet | | Total |
|------|----------|------------------------|-------------------------|--------------|------------------------------|-------------------------------|-------|
| | | Small mesh (< 5.5") | Large mesh (>= 5.5") | | Large mesh (5.5" - 7.99") | Extra large mesh (>= 8.0") | |
| 1989 | | 23 | 44 | 40 | 84 | | 191 |
| 1990 | | 8 | 26 | 31 | 120 | | 185 |
| 1991 | 2 | 29 | 53 | 52 | 801 | | 937 |
| 1992 | 9 | 15 | 45 | 82 | 896 | | 1047 |
| 1993 | 2 | 6 | 17 | 81 | 560 | | 666 |
| 1994 | | | 9 | 77 | 82 | 7 | 175 |
| 1995 | | 30 | 29 | 73 | 62 | 14 | 208 |
| 1996 | | 40 | 19 | 35 | 39 | 10 | 143 |
| 1997 | | 3 | 7 | 16 | 31 | 5 | 62 |
| 1998 | | | 7 | | 78 | 6 | 91 |
| 1999 | | 11 | 25 | | 70 | 8 | 114 |
| 2000 | | | 122 | | 70 | 19 | 211 |
| 2001 | | 4 | 136 | 3 | 39 | 21 | 203 |
| 2002 | | 34 | 199 | | 62 | 25 | 320 |
| 2003 | 14 | 19 | 278 | 15 | 254 | 95 | 675 |
| 2004 | 8 | 68 | 321 | 12 | 587 | 340 | 1336 |
| 2005 | 58 | 69 | 534 | 17 | 505 | 251 | 1434 |
| 2006 | 36 | 24 | 209 | 20 | 109 | 35 | 433 |
| 2007 | 36 | 16 | 234 | 14 | 92 | 46 | 438 |
| 2008 | 20 | 12 | 260 | 19 | 130 | 49 | 490 |
| 2009 | 35 | 22 | 428 | 12 | 271 | 30 | 798 |
| 2010 | 52 | 30 | 685 | 15 | 1080 | 379 | 2241 |
| 2011 | 80 | 25 | 1098 | 1 | 1382 | 264 | 2850 |

Table A.24. Estimates of total Gulf of Maine Atlantic cod commercial discards (mt) by gear from 1982 to 2011 by gear. Estimates from 1989 to 2011 were estimated using an approach consistent with the Standardized Bycatch Report Methodology (Wigley et al., 2007). Estimates from 1982 to 1988 were hindcasted using an approach documented in this report. *Gear-specific estimates do not account for survival of discarded fish.*

| Year | Longline | Otter trawl | | Shrimp trawl | Sink Gillnet | | Total | Total w/ discard survival |
|------|----------|------------------------|-------------------------|--------------|------------------------------|-------------------------------|--------|---------------------------------|
| | | Small mesh (< 5.5") | Large mesh (>= 5.5") | | Large mesh (5.5" - 7.99") | Extra large mesh (>= 8.0") | | |
| 1982 | | | 882.9 | 144.0 | 108.3 | | 1135.2 | 805.4 |
| 1983 | | | 904.5 | 160.1 | 104.9 | | 1169.4 | 829.1 |
| 1984 | | | 861.4 | 228.6 | 120.0 | | 1209.9 | 858.9 |
| 1985 | | | 943.4 | 311.2 | 105.9 | | 1360.5 | 962.9 |
| 1986 | | | 853.5 | 380.6 | 125.5 | | 1359.5 | 964.2 |
| 1987 | | | 774.1 | 345.9 | 125.1 | | 1245.0 | 884.0 |
| 1988 | | | 612.0 | 216.7 | 128.5 | | 957.2 | 682.9 |
| 1989 | | 6.1 | 677.3 | 256.4 | 161.2 | | 1101.1 | 786.9 |
| 1990 | | 0.9 | 1567.6 | 410.7 | 219.0 | | 2198.2 | 1560.6 |
| 1991 | 0.3 | 0.8 | 621.1 | 205.2 | 106.0 | | 933.5 | 663.9 |
| 1992 | 8.0 | 0.0 | 778.7 | 48.9 | 108.2 | | 943.8 | 668.6 |
| 1993 | 281.7 | 0.0 | 370.8 | 6.3 | 153.6 | | 812.4 | 479.8 |
| 1994 | | | 163.8 | 7.5 | 105.1 | 4.3 | 280.8 | 207.5 |
| 1995 | | 8.3 | 152.5 | 4.0 | 129.7 | 20.3 | 314.9 | 235.4 |
| 1996 | | 3.3 | 25.1 | 3.0 | 145.2 | 23.7 | 200.4 | 157.2 |
| 1997 | | 16.6 | 27.9 | 4.7 | 59.1 | 6.8 | 115.0 | 87.1 |
| 1998 | | | 11.6 | | 82.4 | 5.5 | 99.5 | 78.5 |
| 1999 | | 11.6 | 826.5 | | 536.0 | 8.1 | 1382.1 | 1021.9 |
| 2000 | | | 789.0 | | 473.8 | 18.5 | 1281.3 | 946.1 |
| 2001 | | 0.2 | 873.0 | 0.0 | 1113.5 | 54.2 | 2040.9 | 1545.4 |
| 2002 | | 16.4 | 868.6 | | 828.6 | 58.4 | 1772.0 | 1329.1 |
| 2003 | 66.4 | 22.0 | 553.8 | 2.6 | 321.8 | 71.0 | 1037.6 | 741.0 |
| 2004 | 7.9 | 2.9 | 532.4 | 0.9 | 231.8 | 84.6 | 860.6 | 631.1 |
| 2005 | 123.9 | 3.8 | 166.0 | 1.1 | 109.5 | 26.7 | 431.0 | 269.5 |
| 2006 | 47.7 | 2.6 | 337.7 | 0.3 | 94.3 | 15.8 | 498.4 | 342.3 |
| 2007 | 67.3 | 2.0 | 102.6 | 0.9 | 83.6 | 19.3 | 275.7 | 178.4 |
| 2008 | 58.4 | 6.1 | 343.1 | 0.2 | 84.8 | 21.8 | 514.5 | 349.2 |
| 2009 | 19.1 | 2.1 | 719.9 | 0.1 | 263.2 | 37.4 | 1041.8 | 752.3 |
| 2010 | 11.6 | 6.3 | 159.6 | 0.3 | 52.6 | 10.6 | 241.1 | 170.8 |
| 2011 | 31.9 | 4.6 | 77.9 | 0.0 | 34.5 | 3.7 | 152.6 | 98.8 |

Table A.25. Coefficients of variation (CV) for the Gulf of Maine Atlantic cod commercial discard (mt) estimates from 1989 to 2011 by gear; CVs greater than 0.3 are shaded in grey. *CVs are not available for hindcasted discards (pre-1989).*

| Year | Longline | Otter trawl | | Shrimp trawl | Sink Gillnet | | Total |
|------|-------------|------------------------|-------------------------|--------------|------------------------------|-------------------------------|-------------|
| | | Small mesh (< 5.5") | Large mesh (>= 5.5") | | Large mesh (5.5" - 7.99") | Extra large mesh (>= 8.0") | |
| 1989 | | 0.67 | 0.34 | 0.25 | 0.29 | | 0.22 |
| 1990 | | 0.79 | 0.37 | 0.42 | 0.23 | | 0.28 |
| 1991 | 0.40 | 0.60 | 0.37 | 0.32 | 0.10 | | 0.26 |
| 1992 | 0.64 | 3.72 | 0.33 | 0.24 | 0.07 | | 0.27 |
| 1993 | 0.20 | | 0.44 | 0.13 | 0.09 | | 0.22 |
| 1994 | | | 0.63 | 0.15 | 0.32 | 0.75 | 0.38 |
| 1995 | | 0.24 | 0.59 | 0.24 | 0.26 | 0.45 | 0.31 |
| 1996 | | 2.84 | 0.91 | 0.34 | 0.30 | 0.28 | 0.25 |
| 1997 | | 0.25 | 0.44 | 0.41 | 0.42 | 0.85 | 0.25 |
| 1998 | | | 0.55 | | 0.28 | 0.95 | 0.25 |
| 1999 | | 0.62 | 0.56 | | 0.37 | 0.51 | 0.36 |
| 2000 | | | 0.28 | | 0.27 | 0.31 | 0.20 |
| 2001 | | 1.84 | 0.27 | | 0.52 | 0.58 | 0.31 |
| 2002 | | 0.55 | 0.34 | | 0.24 | 0.59 | 0.20 |
| 2003 | 0.30 | 0.72 | 0.29 | 0.42 | 0.14 | 0.28 | 0.16 |
| 2004 | 0.48 | 0.44 | 0.34 | 0.37 | 0.13 | 0.12 | 0.22 |
| 2005 | 0.24 | 0.27 | 0.19 | 0.38 | 0.13 | 0.12 | 0.11 |
| 2006 | 0.29 | 0.27 | 0.39 | 0.44 | 0.38 | 0.32 | 0.28 |
| 2007 | 0.17 | 0.43 | 0.22 | 0.70 | 0.29 | 0.31 | 0.13 |
| 2008 | 0.42 | 0.37 | 0.21 | 0.55 | 0.18 | 0.49 | 0.16 |
| 2009 | 0.17 | 0.28 | 0.14 | 0.64 | 0.19 | 0.49 | 0.11 |
| 2010 | 0.33 | 0.28 | 0.19 | 0.90 | 0.11 | 0.17 | 0.13 |
| 2011 | 0.18 | 0.41 | 0.09 | | 0.04 | 0.07 | 0.06 |

Table A.26. Median, 25th and 75th percentiles of the discard % death estimates by gear type developed by the Discard Mortality Working Group (expressed as percent dead; NEFSC 2012b). Median estimates were used to adjust discard estimates in the current assessment.

| Percentile | Commercial handline | Longline | Otter trawl | Recreational hook and line | Sink gillnet |
|--------------------------|------------------------|----------|-------------|-------------------------------|--------------|
| Median (50th percentile) | 20 | 33 | 75 | 30 | 80 |
| 25th percentile | 13 | 26 | 70 | 20 | 68 |
| 75th percentile | 25 | 39 | 80 | 35 | 86 |

Table A.27. Length sampling of Gulf of Maine Atlantic cod commercial discards from 1989 to 2011 by gear type and semester. Sampling intensity is expressed as metric tons landings per 100 lengths sampled (*200 metric tons per 100 lengths is an unofficial NAFO/ICNAF standard*). Colors denote specific gear/mesh sizes; in all years except 2003-2005, 2007/08 and 2010/11 the length frequency distributions from large mesh gillnet were applied to extra large mesh gillnet due to insufficient sampling. A general criterion of 50 lengths/block was used to determine sufficiency.

| Year | Longline | | Otter trawl - small mesh | | Otter trawl - large mesh | | Shrimp trawl | | Gillnet - large mesh | | Gillnet - extra large mesh | | Total | Total discards (mt) | mt/100 lengths | |
|------|----------|--------|--------------------------|--------|--------------------------|--------|--------------|--------|----------------------|--------|----------------------------|--------|-------|---------------------|----------------|-------|
| | Semi 1 | Semi 2 | Semi 1 | Semi 2 | Semi 1 | Semi 2 | Semi 1 | Semi 2 | Semi 1 | Semi 2 | Semi 1 | Semi 2 | | | | |
| 1989 | | | 125 | 14 | 542 | 1053 | 2011 | 77 | 104 | | | | | 3926 | 1101.1 | 28.0 |
| 1990 | | | ** | | 587 | 818 | 607 | 31 | 138 | 3 | | | | 2184 | 2198.2 | 100.6 |
| 1991 | * | | ** | | 706 | 124 | 397 | | 65 | 30 | | | | 1322 | 933.5 | 70.6 |
| 1992 | * | | ** | | 924 | 924 | 401 | 10 | 78 | 130 | | | | 2467 | 943.8 | 38.3 |
| 1993 | 48 | | ** | | 68 | 866 | 591 | | 90 | 223 | | | | 1886 | 812.4 | 43.1 |
| 1994 | | | ** | | 194 | | 563 | 40 | 274 | 112 | | 7 | | 1190 | 280.8 | 23.6 |
| 1995 | | | 69 | 225 | 473 | | 377 | 3 | 60 | 147 | 20 | 3 | | 1377 | 314.9 | 22.9 |
| 1996 | | | 52 | 19 | 15 | 73 | 44 | 21 | 109 | 31 | 16 | 20 | | 400 | 200.4 | 50.1 |
| 1997 | | | 7*** | | 104 | 1 | 17***** | | 34 | 11 | 1 | 2 | | 153 | 115.0 | 75.2 |
| 1998 | | | *** | | 5**** | | | | 43 | 40 | 9 | 3 | | 95 | 99.5 | 104.8 |
| 1999 | | | 6*** | | 220 | | | | 130 | 1156 | | 14 | | 1520 | 1382.1 | 90.9 |
| 2000 | | | *** | | 248 | 85 | | | 125 | 157 | 6 | 6 | | 627 | 1281.3 | 204.3 |
| 2001 | | | *** | | 61 | 647 | | | 223 | 144 | 3 | 4 | | 1082 | 2040.9 | 188.6 |
| 2002 | | | | | 192 | 104 | 1162 | | 412 | 845 | 1 | 39 | | 2755 | 1772.0 | 64.3 |
| 2003 | 718 | | 173 | 131 | 1109 | 234 | 192 | | 603 | 1352 | 38 | 205 | | 4755 | 1037.6 | 21.8 |
| 2004 | 197 | | 103 | 519 | 385 | 771 | 76 | | 1165 | 1524 | 27 | 536 | | 5303 | 860.6 | 16.2 |
| 2005 | 2283 | 147 | 180 | 183 | 986 | 2939 | 70 | | 190 | 663 | 47 | 104 | | 7792 | 431.0 | 5.5 |
| 2006 | 880 | 3 | 43 | 9 | 1899 | 339 | 96 | | 44 | 59 | 6 | 15 | | 3393 | 498.4 | 14.7 |
| 2007 | 817 | 327 | 1 | 62 | 1172 | 1103 | 12***** | | 91 | 310 | 53 | 164 | | 4100 | 275.7 | 6.7 |
| 2008 | 958 | | | 18 | 2316 | 1639 | 42***** | | 142 | 73 | 72 | 26 | | 5244 | 514.5 | 9.8 |
| 2009 | 552 | 187 | | 22 | 2219 | 1744 | 2***** | | 502 | 112 | 7 | 15 | | 5360 | 1041.8 | 19.4 |
| 2010 | 239 | 57 | 4 | 51 | 716 | 2672 | 5***** | | 289 | 903 | 30 | 94 | | 5055 | 241.1 | 4.8 |
| 2011 | 1322 | 107 | | 27 | 2522 | 3612 | | | 792 | 694 | 33 | 41 | | 9150 | 152.6 | 1.7 |

*Borrowed from 1993 LF
**Used 1989-1995 aggregate LF
***Used 1996-2002 aggregate LF
****Borrowed from 1997 LF
*****Used 1996-1997 aggregate LF
*****Used 2007 - 2010 aggregate LF

Table A.28. Comparison of the survey-filter discard estimates to direct observer-based discard estimates for large mesh otter trawl, shrimp trawl and large mesh gillnet between 1989 and 1993 for Gulf of Maine Atlantic cod.

| Year | Otter trawl, large mesh ($\geq 5.5"$) | | Shrimp trawl | | Sink gillnet, large mesh (5.5" - 7.99") | |
|------|---|-----------------------------|-----------------------|-----------------------------|---|-----------------------------|
| | Discard estimate (mt) | Survey-filter estimate (mt) | Discard estimate (mt) | Survey-filter estimate (mt) | Discard estimate (mt) | Survey-filter estimate (mt) |
| 1989 | 677.3 | 499.8 | 256.4 | 215.6 | 161.2 | 70.9 |
| 1990 | 1567.6 | 722.0 | 410.7 | 273.2 | 219.0 | 80.5 |
| 1991 | 621.1 | 917.3 | 205.2 | 243.8 | 106.0 | 71.4 |
| 1992 | 778.7 | 769.4 | | | 108.2 | 62.4 |
| 1993 | 370.8 | 572.6 | | | 153.6 | 73.1 |

Table A.29. Total Gulf of Maine Atlantic cod commercial discards-at-age (numbers) from 1982 to 2011. *These estimates include gear-specific assumptions of discard survival.*

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|---------|---------|-----------|-----------|---------|--------|--------|--------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1982 | 581 | 347,720 | 1,156,034 | 224,521 | 50,895 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 13,645 | 562,544 | 1,281,940 | 158,839 | 5,416 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 18,275 | 347,694 | 1,445,433 | 219,644 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 67,101 | 459,681 | 1,162,717 | 516,585 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 17,767 | 731,053 | 1,522,658 | 208,195 | 48,007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 100,702 | 252,248 | 1,375,956 | 406,263 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 3,446 | 405,259 | 1,149,396 | 275,330 | 23,306 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 43 | 157,339 | 733,450 | 415,475 | 51,442 | 5,129 | 1,380 | 502 | 109 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 61,442 | 539,508 | 1,619,321 | 185,562 | 1,188 | 216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 3,251 | 115,661 | 244,750 | 156,398 | 273,359 | 23,658 | 945 | 211 | 0 | 494 | 22 | 0 | | | | | |
| 1992 | 23,803 | 364,755 | 481,485 | 278,021 | 32,164 | 91,688 | 2,805 | 119 | 14 | 0 | 0 | | | | | | |
| 1993 | 26,570 | 100,225 | 345,799 | 212,563 | 62,392 | 47 | 682 | 187 | 0 | 0 | 0 | 0 | 0 | | | | |
| 1994 | 11,734 | 119,195 | 93,081 | 140,124 | 14,606 | 816 | 234 | 270 | 0 | 0 | 0 | 0 | | | | | |
| 1995 | 11,572 | 75,059 | 57,584 | 104,772 | 42,720 | 3,914 | 413 | 0 | 0 | 0 | 0 | | | | | | |
| 1996 | 22,067 | 31,719 | 22,411 | 24,451 | 38,147 | 6,928 | 657 | 102 | 78 | 542 | 0 | | | | | | |
| 1997 | 1,472 | 66,116 | 33,817 | 27,941 | 5,256 | 13,811 | 766 | 120 | 0 | 0 | 0 | | | | | | |
| 1998 | 699 | 2,565 | 36,073 | 20,996 | 13,651 | 1,615 | 1,536 | 82 | 0 | 0 | 0 | | | | | | |
| 1999 | 63 | 58,620 | 35,442 | 77,449 | 78,134 | 64,863 | 19,741 | 22,472 | 3,779 | 32 | 0 | 0 | 0 | 0 | | | |
| 2000 | 0 | 10,977 | 192,879 | 122,257 | 137,216 | 26,040 | 8,080 | 1,471 | 315 | | | | | | | | |
| 2001 | 0 | 584 | 166,381 | 181,295 | 117,448 | 89,585 | 23,098 | 9,463 | 1,433 | 1,304 | 0 | | | | | | |
| 2002 | 0 | 10,379 | 26,625 | 95,299 | 150,797 | 58,039 | 36,422 | 15,103 | 9,627 | 3,784 | 3,221 | 270 | 220 | 0 | 0 | 0 | |
| 2003 | 22,873 | 30,227 | 60,078 | 48,552 | 131,760 | 95,818 | 18,452 | 5,589 | 1,985 | 819 | 315 | 204 | 15 | | | | |
| 2004 | 187 | 130,674 | 71,594 | 234,041 | 42,241 | 41,615 | 19,027 | 4,267 | 1,900 | 569 | 231 | 88 | 11 | | | | |
| 2005 | 1,487 | 19,746 | 72,822 | 27,925 | 88,613 | 2,854 | 7,378 | 2,689 | 588 | 435 | 156 | 176 | 80 | 43 | | | |
| 2006 | 204 | 10,521 | 29,696 | 159,504 | 38,366 | 53,974 | 2,405 | 2,150 | 1,902 | 93 | 34 | 5 | 0 | 1 | 0 | | |
| 2007 | 407 | 10,720 | 49,447 | 57,421 | 49,909 | 4,291 | 2,782 | 49 | 53 | 6 | 0 | 2 | 0 | 0 | 0 | 0 | |
| 2008 | 305 | 7,598 | 58,021 | 104,763 | 59,668 | 40,918 | 1,629 | 1,361 | 75 | 17 | 27 | 26 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 81 | 5,791 | 52,840 | 167,603 | 143,740 | 56,239 | 26,856 | 734 | 1,259 | 13 | 33 | 7 | 0 | 8 | 0 | 0 | 0 |
| 2010 | 213 | 4,607 | 23,503 | 52,319 | 27,322 | 15,926 | 3,289 | 989 | 20 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 27 | 1,612 | 13,351 | 31,934 | 28,579 | 6,662 | 1,533 | 153 | 29 | 87 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table A.30. Mean weights-at-age (kg) of commercially discarded Gulf of Maine Atlantic cod from 1982 to 2011.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|-------|-------|-------|-------|-------|-------|--------|--------|--------|--------|--------|--------|--------|--------|-------|-------|-------|
| 1982 | 0.000 | 0.315 | 0.500 | 0.608 | 0.648 | | | | | | | | | | | | |
| 1983 | 0.024 | 0.218 | 0.509 | 0.649 | 0.752 | | | | | | | | | | | | |
| 1984 | 0.001 | 0.225 | 0.485 | 0.610 | | | | | | | | | | | | | |
| 1985 | 0.039 | 0.194 | 0.541 | 0.589 | | | | | | | | | | | | | |
| 1986 | 0.005 | 0.274 | 0.439 | 0.621 | 0.573 | | | | | | | | | | | | |
| 1987 | 0.004 | 0.143 | 0.492 | 0.559 | | | | | | | | | | | | | |
| 1988 | 0.003 | 0.121 | 0.442 | 0.554 | 0.615 | | | | | | | | | | | | |
| 1989 | 0.046 | 0.224 | 0.490 | 0.751 | 1.751 | 4.112 | 5.534 | 9.336 | 6.408 | | | | | | | | |
| 1990 | | 0.195 | 0.645 | 0.703 | 0.846 | 4.340 | 4.564 | | | | | | | | | | |
| 1991 | 0.014 | 0.238 | 0.859 | 0.917 | 0.993 | 1.401 | 6.746 | 8.389 | | 18.191 | 3.705 | | | | | | |
| 1992 | 0.023 | 0.053 | 0.680 | 0.773 | 1.082 | 1.154 | 1.614 | 5.239 | 2.425 | | | | | | | | |
| 1993 | 0.021 | 0.073 | 0.684 | 0.944 | 0.926 | 1.953 | 4.309 | 7.342 | | | | | | | | | |
| 1994 | 0.022 | 0.049 | 0.629 | 0.827 | 1.798 | 3.872 | 12.083 | 9.439 | | | | | | | | | |
| 1995 | 0.027 | 0.093 | 0.809 | 0.925 | 1.637 | 4.928 | 4.682 | | | | | | | | | | |
| 1996 | 0.033 | 0.067 | 0.676 | 1.126 | 1.840 | 3.752 | 6.768 | 11.559 | 12.656 | 17.406 | | | | | | | |
| 1997 | 0.017 | 0.058 | 0.590 | 0.928 | 1.984 | 1.785 | 4.381 | 8.657 | | | | | | | | | |
| 1998 | 0.007 | 0.200 | 0.603 | 1.093 | 1.686 | 3.316 | 3.287 | 3.285 | | | | | | | | | |
| 1999 | 0.052 | 0.201 | 0.595 | 1.940 | 3.353 | 4.626 | 6.586 | 6.605 | 9.634 | 12.279 | | | | | | | |
| 2000 | | 0.292 | 0.962 | 1.843 | 3.041 | 3.882 | 4.881 | 4.279 | 6.121 | | | | | | | | |
| 2001 | | 0.316 | 0.669 | 2.023 | 3.777 | 4.898 | 5.908 | 6.594 | 7.159 | 8.790 | | | | | | | |
| 2002 | | 0.203 | 0.923 | 1.415 | 2.987 | 4.222 | 6.258 | 7.030 | 9.453 | 12.322 | 10.912 | 10.519 | 14.222 | | | | |
| 2003 | 0.038 | 0.133 | 0.804 | 1.364 | 1.672 | 2.772 | 4.085 | 6.911 | 9.868 | 8.622 | 11.658 | 10.100 | 12.774 | | | | |
| 2004 | 0.025 | 0.106 | 0.455 | 1.128 | 1.879 | 2.800 | 4.834 | 6.755 | 8.763 | 11.588 | 11.820 | 10.579 | 11.694 | | | | |
| 2005 | 0.027 | 0.109 | 0.564 | 1.170 | 1.400 | 3.246 | 3.573 | 5.707 | 7.370 | 10.673 | 15.830 | 16.405 | 17.950 | 23.098 | | | |
| 2006 | 0.069 | 0.276 | 0.665 | 1.066 | 1.494 | 1.604 | 1.871 | 3.857 | 2.822 | 7.902 | 8.238 | 13.434 | | 13.434 | | | |
| 2007 | 0.024 | 0.227 | 0.658 | 1.063 | 1.394 | 1.710 | 2.171 | 4.447 | 5.197 | 6.529 | | 7.736 | | | | | |
| 2008 | 0.078 | 0.203 | 0.770 | 1.273 | 1.572 | 1.741 | 3.047 | 6.283 | 6.021 | 5.514 | 10.341 | 10.660 | | | | | |
| 2009 | 0.026 | 0.356 | 0.913 | 1.515 | 2.010 | 2.109 | 2.402 | 3.970 | 3.288 | 8.250 | 8.733 | 7.259 | | 10.510 | | | |
| 2010 | 0.023 | 0.251 | 1.047 | 1.251 | 1.743 | 1.912 | 1.962 | 2.184 | 4.322 | 8.210 | | | | | | | |
| 2011 | 0.122 | 0.361 | 0.875 | 1.181 | 1.303 | 1.473 | 1.592 | 1.669 | 2.623 | 16.409 | | | | | | | |

Table A.31. Proportion of recreationally harvested (type A, and B1 catch) Gulf of Maine Atlantic cod by mode and area as estimated by the Marine Recreational Information Program from 1981 to 2011. *The summary only includes catch from Maine, New Hampshire and Massachusetts. The 'Shore' category includes man-made and beach catch. Due to the proration step that is required to split Massachusetts landed fish between the Gulf of Maine and Georges Bank, these estimates are not directly translatable to the aggregate estimates of Gulf of Maine recreational catch; they are provided for informational purposes only.

| Year | Shore (beach/bank/structure) | | Party/charter | | | Private/rental | | |
|---------|------------------------------|-----------------|---------------|-----------------|-----------------|----------------|-----------------|-----------------|
| | Inland | Ocean ≤ 3 miles | Inland | Ocean ≤ 3 miles | Ocean ≥ 3 miles | Inland | Ocean ≤ 3 miles | Ocean ≥ 3 miles |
| 1981 | 0.00 | 0.00 | 0.04 | 0.06 | 0.54 | 0.03 | 0.28 | 0.05 |
| 1982 | 0.00 | 0.00 | 0.00 | 0.02 | 0.46 | 0.10 | 0.32 | 0.09 |
| 1983 | 0.00 | 0.01 | 0.01 | 0.02 | 0.35 | 0.01 | 0.40 | 0.20 |
| 1984 | 0.01 | 0.00 | 0.01 | 0.05 | 0.36 | 0.03 | 0.28 | 0.26 |
| 1985 | 0.01 | 0.00 | 0.00 | 0.07 | 0.27 | 0.13 | 0.26 | 0.27 |
| 1986 | 0.00 | 0.05 | 0.00 | 0.08 | 0.59 | 0.05 | 0.12 | 0.10 |
| 1987 | 0.00 | 0.00 | 0.00 | 0.19 | 0.53 | 0.01 | 0.14 | 0.14 |
| 1988 | 0.00 | 0.02 | 0.01 | 0.03 | 0.36 | 0.03 | 0.09 | 0.47 |
| 1989 | 0.00 | 0.00 | 0.05 | 0.05 | 0.37 | 0.22 | 0.08 | 0.22 |
| 1990 | 0.00 | 0.01 | 0.01 | 0.05 | 0.53 | 0.02 | 0.10 | 0.27 |
| 1991 | 0.00 | 0.00 | 0.00 | 0.00 | 0.34 | 0.05 | 0.10 | 0.51 |
| 1992 | 0.00 | 0.03 | 0.00 | 0.00 | 0.39 | 0.02 | 0.09 | 0.47 |
| 1993 | 0.00 | 0.00 | 0.00 | 0.01 | 0.66 | 0.03 | 0.10 | 0.19 |
| 1994 | 0.00 | 0.00 | 0.00 | 0.02 | 0.37 | 0.17 | 0.16 | 0.28 |
| 1995 | 0.00 | 0.00 | 0.00 | 0.04 | 0.69 | 0.04 | 0.05 | 0.17 |
| 1996 | 0.00 | 0.00 | 0.02 | 0.03 | 0.56 | 0.01 | 0.05 | 0.34 |
| 1997 | 0.00 | 0.00 | 0.01 | 0.09 | 0.65 | 0.02 | 0.04 | 0.17 |
| 1998 | 0.00 | 0.00 | 0.00 | 0.05 | 0.57 | 0.02 | 0.09 | 0.28 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.03 | 0.51 | 0.00 | 0.11 | 0.34 |
| 2000 | 0.00 | 0.00 | 0.01 | 0.01 | 0.51 | 0.04 | 0.16 | 0.28 |
| 2001 | 0.00 | 0.00 | 0.02 | 0.01 | 0.24 | 0.12 | 0.20 | 0.41 |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 | 0.17 | 0.03 | 0.23 | 0.57 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.00 | 0.26 | 0.00 | 0.11 | 0.63 |
| 2004 | 0.00 | 0.00 | 0.00 | 0.00 | 0.29 | 0.18 | 0.09 | 0.43 |
| 2005 | 0.00 | 0.00 | 0.00 | 0.00 | 0.39 | 0.03 | 0.14 | 0.43 |
| 2006 | 0.00 | 0.00 | 0.00 | 0.00 | 0.56 | 0.02 | 0.08 | 0.34 |
| 2007 | 0.00 | 0.00 | 0.00 | 0.02 | 0.42 | 0.17 | 0.02 | 0.37 |
| 2008 | 0.00 | 0.00 | 0.00 | 0.00 | 0.32 | 0.07 | 0.02 | 0.58 |
| 2009 | 0.00 | 0.00 | 0.02 | 0.00 | 0.47 | 0.04 | 0.01 | 0.47 |
| 2010 | 0.00 | 0.00 | 0.01 | 0.00 | 0.27 | 0.02 | 0.02 | 0.69 |
| 2011 | 0.00 | 0.00 | 0.22 | 0.03 | 0.31 | 0.04 | 0.02 | 0.37 |
| Average | 0.00 | 0.00 | 0.01 | 0.03 | 0.43 | 0.06 | 0.13 | 0.34 |

Table A.32. Proportion of recreationally landed Gulf of Maine Atlantic cod reported on Vessel Trip Reports (VTRs) by month from 1994 to 2011. Recreational vessels are prohibited from possessing Gulf of Maine Atlantic cod in the months shaded grey. Since May 1, 2006 recreational possession was prohibited from November 1st to March 31st. In 2009 the prohibition period was extended to November 1st to April 15th.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
|---------|------|------|------|------|------|------|------|------|------|------|------|------|
| 1994 | 0.00 | 0.00 | 0.00 | 0.01 | 0.02 | 0.17 | 0.15 | 0.22 | 0.13 | 0.16 | 0.03 | 0.11 |
| 1995 | 0.02 | 0.02 | 0.02 | 0.10 | 0.16 | 0.16 | 0.12 | 0.16 | 0.10 | 0.05 | 0.06 | 0.01 |
| 1996 | 0.00 | 0.00 | 0.02 | 0.14 | 0.22 | 0.18 | 0.14 | 0.15 | 0.09 | 0.05 | 0.00 | 0.00 |
| 1997 | 0.00 | 0.00 | 0.01 | 0.14 | 0.23 | 0.16 | 0.15 | 0.15 | 0.10 | 0.05 | 0.01 | 0.00 |
| 1998 | 0.00 | 0.00 | 0.01 | 0.15 | 0.21 | 0.19 | 0.17 | 0.12 | 0.10 | 0.04 | 0.01 | 0.00 |
| 1999 | 0.00 | 0.00 | 0.02 | 0.20 | 0.24 | 0.14 | 0.13 | 0.12 | 0.09 | 0.05 | 0.01 | 0.00 |
| 2000 | 0.00 | 0.01 | 0.03 | 0.18 | 0.22 | 0.15 | 0.13 | 0.12 | 0.11 | 0.05 | 0.01 | 0.00 |
| 2001 | 0.01 | 0.03 | 0.06 | 0.15 | 0.18 | 0.16 | 0.16 | 0.12 | 0.09 | 0.04 | 0.01 | 0.00 |
| 2002 | 0.01 | 0.02 | 0.05 | 0.25 | 0.19 | 0.14 | 0.14 | 0.10 | 0.07 | 0.02 | 0.01 | 0.00 |
| 2003 | 0.00 | 0.00 | 0.02 | 0.12 | 0.24 | 0.16 | 0.15 | 0.15 | 0.09 | 0.04 | 0.01 | 0.01 |
| 2004 | 0.00 | 0.01 | 0.01 | 0.14 | 0.27 | 0.17 | 0.13 | 0.12 | 0.09 | 0.04 | 0.02 | 0.00 |
| 2005 | 0.00 | 0.00 | 0.03 | 0.15 | 0.17 | 0.21 | 0.13 | 0.14 | 0.10 | 0.03 | 0.03 | 0.00 |
| 2006 | 0.01 | 0.02 | 0.09 | 0.19 | 0.18 | 0.18 | 0.13 | 0.09 | 0.08 | 0.03 | 0.00 | 0.00 |
| 2007 | 0.00 | 0.00 | 0.00 | 0.16 | 0.23 | 0.17 | 0.14 | 0.12 | 0.12 | 0.05 | 0.00 | 0.00 |
| 2008 | 0.00 | 0.00 | 0.00 | 0.20 | 0.26 | 0.17 | 0.13 | 0.11 | 0.08 | 0.06 | 0.00 | 0.00 |
| 2009 | 0.00 | 0.00 | 0.00 | 0.17 | 0.30 | 0.17 | 0.10 | 0.09 | 0.11 | 0.06 | 0.00 | 0.00 |
| 2010 | 0.00 | 0.00 | 0.00 | 0.14 | 0.25 | 0.23 | 0.12 | 0.13 | 0.08 | 0.04 | 0.00 | 0.00 |
| 2011 | 0.00 | 0.00 | 0.00 | 0.13 | 0.27 | 0.24 | 0.15 | 0.11 | 0.08 | 0.03 | 0.00 | 0.00 |
| Average | 0.00 | 0.01 | 0.02 | 0.15 | 0.21 | 0.18 | 0.14 | 0.13 | 0.09 | 0.05 | 0.01 | 0.01 |

Table A.33. Proportion of recreationally caught (type A, B1 and B2) Gulf of Maine Atlantic cod by sampling wave as estimated by the Marine Recreational Information Program between 1981 and 2011.

| Year | Wave | | | | |
|---------|------|------|------|------|------|
| | 2 | 3 | 4 | 5 | 6 |
| 1981 | 0.16 | 0.63 | 0.11 | 0.10 | 0.00 |
| 1982 | 0.33 | 0.29 | 0.22 | 0.16 | 0.01 |
| 1983 | 0.11 | 0.29 | 0.26 | 0.32 | 0.02 |
| 1984 | 0.08 | 0.40 | 0.39 | 0.12 | 0.01 |
| 1985 | 0.19 | 0.53 | 0.16 | 0.09 | 0.02 |
| 1986 | 0.22 | 0.13 | 0.21 | 0.26 | 0.18 |
| 1987 | 0.41 | 0.26 | 0.11 | 0.12 | 0.11 |
| 1988 | 0.04 | 0.41 | 0.12 | 0.41 | 0.02 |
| 1989 | 0.04 | 0.35 | 0.25 | 0.29 | 0.06 |
| 1990 | 0.11 | 0.46 | 0.15 | 0.25 | 0.03 |
| 1991 | 0.14 | 0.49 | 0.06 | 0.20 | 0.10 |
| 1992 | 0.26 | 0.24 | 0.19 | 0.29 | 0.03 |
| 1993 | 0.17 | 0.39 | 0.17 | 0.20 | 0.07 |
| 1994 | 0.05 | 0.31 | 0.20 | 0.14 | 0.31 |
| 1995 | 0.18 | 0.23 | 0.08 | 0.41 | 0.10 |
| 1996 | 0.12 | 0.32 | 0.19 | 0.21 | 0.15 |
| 1997 | 0.31 | 0.28 | 0.18 | 0.07 | 0.16 |
| 1998 | 0.30 | 0.26 | 0.23 | 0.06 | 0.16 |
| 1999 | 0.33 | 0.22 | 0.16 | 0.23 | 0.06 |
| 2000 | 0.22 | 0.37 | 0.16 | 0.20 | 0.04 |
| 2001 | 0.12 | 0.31 | 0.22 | 0.23 | 0.12 |
| 2002 | 0.17 | 0.28 | 0.19 | 0.17 | 0.19 |
| 2003 | 0.19 | 0.40 | 0.18 | 0.15 | 0.09 |
| 2004 | 0.06 | 0.39 | 0.13 | 0.27 | 0.14 |
| 2005 | 0.21 | 0.36 | 0.25 | 0.12 | 0.07 |
| 2006 | 0.20 | 0.32 | 0.31 | 0.15 | 0.03 |
| 2007 | 0.22 | 0.30 | 0.25 | 0.12 | 0.10 |
| 2008 | 0.07 | 0.54 | 0.27 | 0.12 | 0.00 |
| 2009 | 0.11 | 0.57 | 0.16 | 0.12 | 0.03 |
| 2010 | 0.13 | 0.45 | 0.20 | 0.22 | 0.00 |
| 2011 | 0.04 | 0.69 | 0.17 | 0.08 | 0.02 |
| Average | 0.17 | 0.37 | 0.19 | 0.19 | 0.08 |

Table A.34. Proportion of recreationally landed Gulf of Maine Atlantic cod reported on Vessel Trip Reports (VTRs) by state from 1994 to 2011.

| Year | CT | FL | MA | ME | NH | NJ | NK | NY | RI | VA |
|-------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| 1994 | 0.00 | 0.00 | 0.59 | 0.32 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1995 | 0.00 | 0.00 | 0.72 | 0.18 | 0.10 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1996 | 0.00 | 0.00 | 0.69 | 0.21 | 0.10 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1997 | 0.00 | 0.00 | 0.63 | 0.25 | 0.12 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1998 | 0.00 | 0.00 | 0.59 | 0.27 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1999 | 0.00 | 0.00 | 0.67 | 0.19 | 0.14 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| 2000 | 0.00 | 0.00 | 0.67 | 0.17 | 0.16 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2001 | 0.00 | 0.00 | 0.71 | 0.13 | 0.16 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2002 | 0.00 | 0.00 | 0.64 | 0.11 | 0.24 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2003 | 0.00 | 0.00 | 0.66 | 0.14 | 0.19 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2004 | 0.00 | 0.00 | 0.60 | 0.12 | 0.26 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 |
| 2005 | 0.00 | 0.00 | 0.56 | 0.10 | 0.33 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 |
| 2006 | 0.00 | 0.00 | 0.55 | 0.13 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2007 | 0.00 | 0.00 | 0.48 | 0.17 | 0.34 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| 2008 | 0.00 | 0.00 | 0.52 | 0.15 | 0.34 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2009 | 0.00 | 0.00 | 0.50 | 0.17 | 0.33 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2010 | 0.00 | 0.00 | 0.54 | 0.11 | 0.34 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2011 | 0.00 | 0.02 | 0.62 | 0.07 | 0.30 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Average | 0.00 | 0.00 | 0.61 | 0.17 | 0.22 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Table A.35. Proportion of recreationally landed Gulf of Maine Atlantic cod reported on Vessel Trip Reports (VTRs) by statistical area from 1994 to 2011.

| Year | 464 | 465 | 510 | 511 | 512 | 513 | 514 | 515 |
|-------------|------------|------------|------------|------------|------------|------------|------------|------------|
| 1994 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.29 | 0.43 | 0.26 |
| 1995 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.36 | 0.51 | 0.12 |
| 1996 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.38 | 0.59 | 0.03 |
| 1997 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.48 | 0.51 | 0.01 |
| 1998 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.49 | 0.50 | 0.01 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.39 | 0.58 | 0.02 |
| 2000 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.34 | 0.61 | 0.05 |
| 2001 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.31 | 0.66 | 0.03 |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.37 | 0.60 | 0.03 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.36 | 0.54 | 0.10 |
| 2004 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.33 | 0.62 | 0.04 |
| 2005 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.39 | 0.57 | 0.04 |
| 2006 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.42 | 0.54 | 0.05 |
| 2007 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.45 | 0.52 | 0.01 |
| 2008 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.44 | 0.54 | 0.02 |
| 2009 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.42 | 0.58 | 0.00 |
| 2010 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.45 | 0.50 | 0.06 |
| 2011 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.31 | 0.66 | 0.03 |
| Average | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.39 | 0.56 | 0.05 |

Table A.36. Length sampling intensity of recreationally harvested (type A, and B1) Gulf of Maine Atlantic cod by semester and year as estimated by the Marine Recreational Information Program from 1981 to 2011. Sampling intensity is expressed as metric tons of landings per 100 lengths sampled (*200 metric tons per 100 lengths is an unofficial NAFO/ICNAF standard*).

| Year | Semester | | Total | A,B1 estimated numbers (000s) | AB1 Landings (mt) | Lengths per 1000 fish | mt per 100 lengths |
|------|----------|-----|-------|-------------------------------|-------------------|-----------------------|--------------------|
| | 1 | 2 | | | | | |
| 1981 | 355 | 366 | 721 | 2011.2 | 4111.5 | 0.4 | 570.3 |
| 1982 | 320 | 276 | 596 | 1368.7 | 2816.7 | 0.4 | 472.6 |
| 1983 | 609 | 560 | 1169 | 937.1 | 1772.8 | 1.2 | 151.7 |
| 1984 | 394 | 391 | 785 | 678.9 | 1266.8 | 1.2 | 161.4 |
| 1985 | 272 | 155 | 427 | 1212.5 | 2765.9 | 0.4 | 647.7 |
| 1986 | 77 | 90 | 167 | 734.0 | 1928.4 | 0.2 | 1154.8 |
| 1987 | 167 | 367 | 534 | 1504.5 | 3547.2 | 0.4 | 664.3 |
| 1988 | 325 | 213 | 538 | 943.2 | 1688.5 | 0.6 | 313.9 |
| 1989 | 208 | 352 | 560 | 893.2 | 1957.2 | 0.6 | 349.5 |
| 1990 | 160 | 210 | 370 | 930.9 | 2246.7 | 0.4 | 607.2 |
| 1991 | 377 | 83 | 460 | 1023.1 | 2287.2 | 0.4 | 497.2 |
| 1992 | 710 | 268 | 978 | 238.4 | 623.6 | 4.1 | 63.8 |
| 1993 | 136 | 200 | 336 | 568.3 | 1011.9 | 0.6 | 301.2 |
| 1994 | 333 | 485 | 818 | 392.9 | 721.7 | 2.1 | 88.2 |
| 1995 | 663 | 434 | 1097 | 378.6 | 627.2 | 2.9 | 57.2 |
| 1996 | 585 | 515 | 1100 | 260.0 | 498.6 | 4.2 | 45.3 |
| 1997 | 190 | 392 | 582 | 105.0 | 236.3 | 5.5 | 40.6 |
| 1998 | 447 | 215 | 662 | 144.2 | 353.1 | 4.6 | 53.3 |
| 1999 | 111 | 117 | 228 | 184.7 | 577.2 | 1.2 | 253.1 |
| 2000 | 70 | 77 | 147 | 388.5 | 967.1 | 0.4 | 657.9 |
| 2001 | 124 | 121 | 245 | 755.6 | 1967.6 | 0.3 | 803.1 |
| 2002 | 181 | 196 | 377 | 409.1 | 1254.8 | 0.9 | 332.8 |
| 2003 | 361 | 322 | 683 | 454.9 | 1607.7 | 1.5 | 235.4 |
| 2004 | 422 | 473 | 895 | 379.4 | 1150.9 | 2.4 | 128.6 |
| 2005 | 391 | 382 | 773 | 446.9 | 1346.9 | 1.7 | 174.2 |
| 2006 | 681 | 155 | 836 | 188.7 | 702.3 | 4.4 | 84.0 |
| 2007 | 479 | 220 | 699 | 303.5 | 1042.2 | 2.3 | 149.1 |
| 2008 | 590 | 231 | 821 | 382.6 | 1267.2 | 2.1 | 154.3 |
| 2009 | 852 | 488 | 1340 | 386.9 | 1357.1 | 3.5 | 101.3 |
| 2010 | 621 | 508 | 1129 | 503.9 | 1758.2 | 2.2 | 155.7 |
| 2011 | 711 | 496 | 1207 | 516.0 | 1799.1 | 2.3 | 149.1 |

Table A.37. Percentage of recreationally discarded (type B2) Gulf of Maine Atlantic cod by mode and area as estimated by the Marine Recreational Information Program from 1981 to 2011.

**The summary only includes catch from Maine, New Hampshire and Massachusetts. The 'Shore' category includes man-made and beach catch. Due to the proration step that is required to split Massachusetts landed fish between the Gulf of Maine and Georges Bank, these estimates are not directly translatable to the aggregate estimates of Gulf of Maine recreational catch; they are provided for informational purposes only.*

| Year | Shore (beach/bank/structure) | | Party/charter | | | Private/rental | | |
|---------|------------------------------|-----------------|---------------|-----------------|-----------------|----------------|-----------------|-----------------|
| | Inland | Ocean ≤ 3 miles | Inland | Ocean ≤ 3 miles | Ocean ≥ 3 miles | Inland | Ocean ≤ 3 miles | Ocean ≥ 3 miles |
| 1981 | 0.00 | 0.00 | 0.00 | 0.00 | 0.16 | 0.11 | 0.63 | 0.10 |
| 1982 | 0.00 | 0.00 | 0.00 | 0.00 | 0.44 | 0.01 | 0.26 | 0.29 |
| 1983 | 0.00 | 0.04 | 0.00 | 0.00 | 0.15 | 0.10 | 0.54 | 0.17 |
| 1984 | 0.01 | 0.00 | 0.00 | 0.03 | 0.26 | 0.00 | 0.45 | 0.25 |
| 1985 | 0.01 | 0.00 | 0.00 | 0.23 | 0.35 | 0.02 | 0.03 | 0.36 |
| 1986 | 0.00 | 0.00 | 0.01 | 0.16 | 0.36 | 0.06 | 0.19 | 0.21 |
| 1987 | 0.00 | 0.00 | 0.00 | 0.29 | 0.47 | 0.02 | 0.08 | 0.14 |
| 1988 | 0.00 | 0.00 | 0.02 | 0.04 | 0.49 | 0.01 | 0.12 | 0.31 |
| 1989 | 0.00 | 0.00 | 0.04 | 0.06 | 0.37 | 0.14 | 0.08 | 0.30 |
| 1990 | 0.00 | 0.00 | 0.02 | 0.06 | 0.44 | 0.02 | 0.08 | 0.38 |
| 1991 | 0.00 | 0.00 | 0.00 | 0.00 | 0.35 | 0.04 | 0.09 | 0.51 |
| 1992 | 0.00 | 0.03 | 0.00 | 0.00 | 0.34 | 0.05 | 0.07 | 0.50 |
| 1993 | 0.00 | 0.00 | 0.00 | 0.01 | 0.65 | 0.04 | 0.14 | 0.16 |
| 1994 | 0.00 | 0.00 | 0.00 | 0.01 | 0.37 | 0.22 | 0.13 | 0.27 |
| 1995 | 0.00 | 0.00 | 0.00 | 0.05 | 0.68 | 0.04 | 0.06 | 0.17 |
| 1996 | 0.00 | 0.00 | 0.01 | 0.03 | 0.56 | 0.02 | 0.06 | 0.34 |
| 1997 | 0.00 | 0.00 | 0.03 | 0.10 | 0.56 | 0.04 | 0.06 | 0.22 |
| 1998 | 0.00 | 0.00 | 0.00 | 0.06 | 0.52 | 0.02 | 0.12 | 0.28 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.03 | 0.44 | 0.01 | 0.11 | 0.42 |
| 2000 | 0.00 | 0.00 | 0.01 | 0.01 | 0.48 | 0.03 | 0.18 | 0.29 |
| 2001 | 0.00 | 0.00 | 0.03 | 0.01 | 0.21 | 0.13 | 0.22 | 0.39 |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 | 0.14 | 0.03 | 0.25 | 0.58 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.00 | 0.23 | 0.00 | 0.12 | 0.65 |
| 2004 | 0.00 | 0.00 | 0.00 | 0.00 | 0.29 | 0.20 | 0.10 | 0.41 |
| 2005 | 0.00 | 0.00 | 0.00 | 0.00 | 0.39 | 0.04 | 0.13 | 0.44 |
| 2006 | 0.00 | 0.00 | 0.00 | 0.00 | 0.55 | 0.03 | 0.07 | 0.36 |
| 2007 | 0.00 | 0.00 | 0.00 | 0.00 | 0.41 | 0.14 | 0.02 | 0.44 |
| 2008 | 0.00 | 0.00 | 0.00 | 0.00 | 0.34 | 0.09 | 0.02 | 0.55 |
| 2009 | 0.00 | 0.00 | 0.01 | 0.00 | 0.46 | 0.05 | 0.00 | 0.47 |
| 2010 | 0.00 | 0.00 | 0.01 | 0.00 | 0.25 | 0.02 | 0.02 | 0.71 |
| 2011 | 0.00 | 0.00 | 0.23 | 0.03 | 0.28 | 0.04 | 0.02 | 0.41 |
| Average | 0.00 | 0.00 | 0.01 | 0.04 | 0.39 | 0.06 | 0.14 | 0.36 |

Table A.38. Length sampling intensity of recreationally discarded (type B2) Gulf of Maine Atlantic cod by semester and year as estimated by the Marine Recreational Information Program from 2005 to 2011. Length samples of recreationally discarded (19 samples) Atlantic cod were unavailable prior to 2005. Sampling intensity is expressed as metric tons landings per 100 lengths sampled (*200 metric tons per 100 lengths is an unofficial NAFO/ICNAF standard*).

| Year | Semester | | Total | B2 releases (000s) | B2 releases (mt) | Lengths per thousand fish | Metric tons per 100 lengths |
|------|----------|------|-------|--------------------------|------------------------|---------------------------------|-----------------------------------|
| | 1 | 2 | | | | | |
| 2005 | 577 | 624 | 1201 | 1038.1 | 891.0 | 1.2 | 208.1 |
| 2006 | 952 | 599 | 1551 | 708.4 | 646.7 | 2.2 | 162.9 |
| 2007 | 728 | 846 | 1574 | 964.4 | 1056.2 | 1.6 | 216.2 |
| 2008 | 1258 | 709 | 1967 | 952.1 | 1051.2 | 2.1 | 156.4 |
| 2009 | 765 | 889 | 1654 | 826.0 | 974.8 | 2.0 | 216.2 |
| 2010 | 715 | 1024 | 1739 | 1049.4 | 1281.6 | 1.7 | 243.2 |
| 2011 | 493 | 937 | 1430 | 892.4 | 1114.1 | 1.6 | 290.1 |

Table A.39. Annual ratios of Marine Recreational Fisheries Statistical Survey (MRFSS) and Marine Recreational Information Program (MRIP) catch estimates and aggregate time series ratios (ratio of means).

| Year | MRIP | | MRFSS | | Ratio | |
|------|----------------------------|------------------------|----------------------------|------------------------|---------------|--------------|
| | Landings, AB1 (numbers) | Releases, B2 (numbers) | Landings, AB1 (numbers) | Releases B2, (numbers) | Landings, AB1 | Releases, B2 |
| 2004 | 379,444 | 736,820 | 536,147 | 885,537 | 0.708 | 0.832 |
| 2005 | 446,894 | 1,038,133 | 590,390 | 1,356,379 | 0.757 | 0.765 |
| 2006 | 188,699 | 708,360 | 227,980 | 763,402 | 0.828 | 0.928 |
| 2007 | 303,540 | 964,427 | 309,786 | 1,180,096 | 0.980 | 0.817 |
| 2008 | 382,555 | 952,120 | 477,913 | 1,281,510 | 0.800 | 0.743 |
| 2009 | 386,913 | 826,019 | 478,765 | 1,130,115 | 0.808 | 0.731 |
| 2010 | 503,887 | 1,049,409 | 1,041,480 | 2,000,702 | 0.484 | 0.525 |
| 2011 | 516,049 | 892,438 | 526,101 | 882,038 | 0.981 | 1.012 |
| Sum | 3,107,981 | 7,167,726 | 4,188,561 | 9,479,780 | 0.742 | 0.756 |

Table A.40. Relative difference between SAW 53 recreational catch estimates (numbers) and the unadjusted updated Marine Recreational Fisheries Statistical Survey (MRFSS) estimates. Positive numbers indicate SAW 53 estimates were larger (e.g., 0.50 implies the updated estimates are 50% lower than the SAW 53 estimates).

| Relative difference between SARC 53 numbers and adjusted MRFSS numbers | | | | |
|--|--------|--------|-----------------------|--------------------|
| Year | Type A | Type B | Type AB1 (harvest) | Type B2 (releases) |
| 1981 | -0.04 | 0.00 | -0.02 | 0.00 |
| 1982 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1983 | -0.01 | 0.00 | 0.00 | 0.00 |
| 1984 | -0.01 | 0.00 | 0.00 | 0.00 |
| 1985 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1986 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1987 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1988 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1989 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1990 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1991 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1992 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1993 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1994 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1995 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1996 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1997 | -0.04 | 0.00 | -0.01 | 0.00 |
| 1998 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2000 | -0.01 | 0.00 | 0.00 | 0.00 |
| 2001 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2004 | -0.19 | 0.06 | -0.01 | 0.01 |
| 2005 | 0.03 | -0.01 | 0.00 | -0.08 |
| 2006 | -0.01 | 0.00 | 0.00 | -0.12 |
| 2007 | -0.03 | 0.00 | -0.01 | -0.15 |
| 2008 | -0.02 | 0.00 | 0.00 | -0.10 |
| 2009 | -0.01 | 0.00 | 0.00 | -0.07 |
| 2010 | -0.11 | -0.03 | -0.04 | -0.07 |

Table A.41. Relative difference between SAW 53 recreational catch estimates (numbers) and the updated Marine Recreational Information Program (MRIP) estimates. Positive numbers indicate SAW 53 estimates were larger (e.g., 0.50 implies the updated estimates are 50% lower than the SAW 53 estimates).

| Year | Type AB1 (harvest) | Type B2 (releases) |
|------|-----------------------|-----------------------|
| 1981 | 0.24 | 0.24 |
| 1982 | 0.26 | 0.24 |
| 1983 | 0.25 | 0.24 |
| 1984 | 0.25 | 0.24 |
| 1985 | 0.26 | 0.24 |
| 1986 | 0.26 | 0.24 |
| 1987 | 0.26 | 0.24 |
| 1988 | 0.26 | 0.24 |
| 1989 | 0.26 | 0.24 |
| 1990 | 0.26 | 0.24 |
| 1991 | 0.26 | 0.24 |
| 1992 | 0.26 | 0.24 |
| 1993 | 0.26 | 0.24 |
| 1994 | 0.26 | 0.24 |
| 1995 | 0.26 | 0.24 |
| 1996 | 0.26 | 0.24 |
| 1997 | 0.25 | 0.24 |
| 1998 | 0.26 | 0.24 |
| 1999 | 0.26 | 0.24 |
| 2000 | 0.26 | 0.24 |
| 2001 | 0.26 | 0.24 |
| 2002 | 0.26 | 0.24 |
| 2003 | 0.26 | 0.24 |
| 2004 | 0.29 | 0.18 |
| 2005 | 0.24 | 0.18 |
| 2006 | 0.17 | -0.04 |
| 2007 | 0.01 | 0.06 |
| 2008 | 0.20 | 0.18 |
| 2009 | 0.19 | 0.22 |
| 2010 | 0.50 | 0.44 |

Table A.42. Estimates of Gulf of Maine Atlantic cod recreational catch in numbers (000's) and weight (mt). Recreational releases are shown using both the 100% discard mortality (grey) and Discard WG revised 30% mortality assumptions.

| Year | Numbers (000s) | | | | Mass (mt) | | | | Released:harvest ratio |
|------|----------------|---|--|---|---------------|---|--|---|------------------------|
| | Harvest (AB1) | Released (B2) w/ 100% discard mortality | Released (B2) w/ 30% discard mortality | Total catch w/ 30% discard mortality | Harvest (AB1) | Released (B2) w/ 100% discard mortality | Released (B2) w/ 30% discard mortality | Total catch w/ 30% discard mortality | |
| 1981 | 2011.2 | 145.1 | 43.5 | 2054.7 | 4111.5 | 62.8 | 18.8 | 4130.4 | 0.07 |
| 1982 | 1368.7 | 71.6 | 21.5 | 1390.2 | 2816.7 | 27.2 | 8.1 | 2824.9 | 0.05 |
| 1983 | 937.1 | 174.2 | 52.2 | 989.4 | 1772.8 | 58.6 | 17.6 | 1790.4 | 0.19 |
| 1984 | 678.9 | 148.7 | 44.6 | 723.6 | 1266.8 | 55.3 | 16.6 | 1283.4 | 0.22 |
| 1985 | 1212.5 | 150.9 | 45.3 | 1257.8 | 2765.9 | 56.2 | 16.9 | 2782.7 | 0.12 |
| 1986 | 734.0 | 91.9 | 27.6 | 761.6 | 1928.4 | 33.7 | 10.1 | 1938.5 | 0.13 |
| 1987 | 1504.5 | 428.5 | 128.6 | 1633.1 | 3547.2 | 160.0 | 48.0 | 3595.2 | 0.28 |
| 1988 | 943.2 | 133.3 | 40.0 | 983.2 | 1688.5 | 45.2 | 13.5 | 1702.1 | 0.14 |
| 1989 | 893.2 | 432.6 | 129.8 | 1023.0 | 1957.2 | 253.6 | 76.1 | 2033.2 | 0.48 |
| 1990 | 930.9 | 357.6 | 107.3 | 1038.2 | 2246.7 | 222.4 | 66.7 | 2313.4 | 0.38 |
| 1991 | 1023.1 | 310.3 | 93.1 | 1116.1 | 2287.2 | 226.7 | 68.0 | 2355.3 | 0.30 |
| 1992 | 238.4 | 180.8 | 54.2 | 292.7 | 623.6 | 118.2 | 35.5 | 659.0 | 0.76 |
| 1993 | 568.3 | 568.0 | 170.4 | 738.7 | 1011.9 | 339.8 | 101.9 | 1113.9 | 1.00 |
| 1994 | 392.9 | 543.6 | 163.1 | 556.0 | 721.7 | 335.3 | 100.6 | 822.3 | 1.38 |
| 1995 | 378.6 | 516.2 | 154.8 | 533.5 | 627.2 | 320.5 | 96.2 | 723.4 | 1.36 |
| 1996 | 260.0 | 340.8 | 102.2 | 362.3 | 498.6 | 270.1 | 81.0 | 579.7 | 1.31 |
| 1997 | 105.0 | 227.0 | 68.1 | 173.1 | 236.3 | 195.9 | 58.8 | 295.1 | 2.16 |
| 1998 | 144.2 | 289.6 | 86.9 | 231.1 | 353.1 | 240.8 | 72.2 | 425.3 | 2.01 |
| 1999 | 184.7 | 359.7 | 107.9 | 292.6 | 577.2 | 238.8 | 71.7 | 648.8 | 1.95 |
| 2000 | 388.5 | 696.2 | 208.8 | 597.4 | 967.1 | 458.7 | 137.6 | 1104.7 | 1.79 |
| 2001 | 755.6 | 992.0 | 297.6 | 1053.2 | 1967.6 | 758.3 | 227.5 | 2195.0 | 1.31 |
| 2002 | 409.1 | 823.5 | 247.1 | 656.1 | 1254.8 | 956.2 | 286.9 | 1541.7 | 2.01 |
| 2003 | 454.9 | 837.8 | 251.3 | 706.3 | 1607.7 | 941.4 | 282.4 | 1890.1 | 1.84 |
| 2004 | 379.4 | 736.8 | 221.0 | 600.5 | 1150.9 | 671.2 | 201.4 | 1352.2 | 1.94 |
| 2005 | 446.9 | 1038.1 | 311.4 | 758.3 | 1346.9 | 891.0 | 267.3 | 1614.2 | 2.32 |
| 2006 | 188.7 | 708.4 | 212.5 | 401.2 | 702.3 | 646.7 | 194.0 | 896.3 | 3.75 |
| 2007 | 303.5 | 964.4 | 289.3 | 592.9 | 1042.2 | 1056.2 | 316.9 | 1359.1 | 3.18 |
| 2008 | 382.6 | 952.1 | 285.6 | 668.2 | 1267.2 | 1051.2 | 315.4 | 1582.6 | 2.49 |
| 2009 | 386.9 | 826.0 | 247.8 | 634.7 | 1357.1 | 974.8 | 292.4 | 1649.6 | 2.13 |
| 2010 | 503.9 | 1049.4 | 314.8 | 818.7 | 1758.2 | 1281.6 | 384.5 | 2142.6 | 2.08 |
| 2011 | 516.0 | 892.4 | 267.7 | 783.8 | 1799.1 | 1114.1 | 334.2 | 2133.3 | 1.73 |

Table A.43. Percent standard error (PSE) of Gulf of Maine Atlantic cod recreational catch estimates (A, B1 and B2) by state by the Marine Recreational Information Program between 1981 and 2011. *Note: due to the proration step that is required to split Massachusetts landed fish between the Gulf of Maine and Georges Bank, these estimates of PSE are not directly translatable to the aggregate estimates of Gulf of Maine recreational catch. The PSEs are provided for informational purposes only.

| Year | Maine | New Hampshire | Massachusetts |
|---------|-------|---------------|---------------|
| 1981 | 35.7 | 24.6 | 23.4 |
| 1982 | 22.0 | 47.1 | 39.1 |
| 1983 | 20.6 | 18.5 | 13.6 |
| 1984 | 16.7 | 14.7 | 13.9 |
| 1985 | 24.2 | 26.3 | 23.3 |
| 1986 | 18.4 | 24.0 | 22.6 |
| 1987 | 40.4 | 36.1 | 14.3 |
| 1988 | 75.4 | 25.6 | 10.6 |
| 1989 | 21.1 | 19.6 | 14.6 |
| 1990 | 29.8 | 24.9 | 11.2 |
| 1991 | 33.9 | 36.5 | 9.5 |
| 1992 | 43.3 | 31.1 | 13.5 |
| 1993 | 33.6 | 30.2 | 13.1 |
| 1994 | 32.2 | 31.3 | 9.2 |
| 1995 | 34.9 | 16.3 | 11.2 |
| 1996 | 38.6 | 20.2 | 13.2 |
| 1997 | 36.3 | 23.8 | 17.6 |
| 1998 | 47.0 | 17.9 | 17.4 |
| 1999 | 43.7 | 14.7 | 17.7 |
| 2000 | 21.9 | 12.6 | 14.5 |
| 2001 | 26.1 | 10.6 | 8.0 |
| 2002 | 20.3 | 11.9 | 9.1 |
| 2003 | 28.1 | 11.7 | 9.5 |
| 2004 | 40.5 | 15.5 | 19.7 |
| 2005 | 21.5 | 14.8 | 15.1 |
| 2006 | 16.6 | 10.9 | 13.9 |
| 2007 | 32.7 | 14.4 | 16.8 |
| 2008 | 23.0 | 14.9 | 17.7 |
| 2009 | 17.1 | 13.6 | 18.0 |
| 2010 | 20.0 | 12.1 | 17.6 |
| 2011 | 27.6 | 18.9 | 12.0 |
| Average | 30.4 | 20.8 | 15.5 |

Table A.44. Total Gulf of Maine Atlantic cod recreational landings-at-age (numbers) from 1982 to 2011.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|------|---------|---------|---------|---------|---------|---------|---------|--------|--------|--------|-------|-------|-------|-------|-------|-------|
| 1981 | | 159,922 | 623,992 | 622,091 | 426,564 | 69,951 | 42,612 | 7,392 | 29,365 | | 25,105 | | 4,184 | | | | |
| 1982 | 765 | 67,908 | 420,464 | 427,446 | 263,437 | 129,184 | 14,639 | 24,905 | 13,178 | 3,904 | 574 | | 2,296 | | | | |
| 1983 | | 14,924 | 315,694 | 339,632 | 128,267 | 76,679 | 45,287 | 5,810 | 4,873 | 1,777 | 1,390 | 802 | 2,004 | | | | |
| 1984 | | 11,741 | 224,928 | 226,199 | 139,013 | 40,743 | 23,707 | 9,247 | 390 | 420 | 350 | 627 | | 432 | 1,153 | | |
| 1985 | | 35,163 | 368,684 | 438,416 | 149,622 | 123,096 | 38,047 | 33,994 | 15,929 | 2,206 | 5,509 | 316 | 1,005 | 532 | | | |
| 1986 | | 21,723 | 120,551 | 351,802 | 124,583 | 39,540 | 40,989 | 9,316 | 10,691 | 6,281 | 3,579 | 865 | 3,202 | 865 | | | |
| 1987 | | 16,878 | 348,751 | 517,856 | 457,592 | 77,647 | 24,836 | 35,051 | 8,978 | 8,452 | 6,339 | 1,878 | 282 | | | | |
| 1988 | | 3,134 | 197,888 | 449,655 | 225,659 | 46,787 | 8,638 | 3,696 | 6,000 | | | 1,753 | | | | | |
| 1989 | | 3,619 | 116,660 | 436,314 | 242,898 | 64,122 | 15,197 | 10,911 | 1,329 | 2,127 | | | | | | | |
| 1990 | | 2,812 | 40,204 | 449,749 | 295,754 | 87,368 | 36,966 | 4,457 | 11,742 | 1,887 | | | | | | | |
| 1991 | | 3,614 | 35,323 | 152,702 | 701,569 | 106,170 | 11,169 | 12,368 | | | 143 | | | | | | |
| 1992 | | 2,101 | 21,451 | 43,626 | 35,194 | 123,077 | 10,143 | 2,642 | 193 | | | | | | | | |
| 1993 | | 1,913 | 42,807 | 343,796 | 133,450 | 10,536 | 32,237 | 3,594 | | | | | | | | | |
| 1994 | | 475 | 13,965 | 243,207 | 103,423 | 24,535 | 2,404 | 3,971 | 600 | 370 | | | | | | | |
| 1995 | | 25 | 35,494 | 187,086 | 144,820 | 9,965 | 1,024 | | 192 | | | | | | | | |
| 1996 | | | 11,977 | 64,661 | 162,532 | 19,752 | 850 | 34 | | 236 | | | | | | | |
| 1997 | | 78 | 5,075 | 31,836 | 21,300 | 42,823 | 3,631 | 35 | 192 | | | | | | | | |
| 1998 | 218 | | 9,310 | 52,886 | 52,992 | 11,547 | 15,851 | 1,107 | 315 | | | | | | | | |
| 1999 | | 552 | 5,301 | 53,525 | 61,018 | 39,039 | 9,650 | 14,515 | 1,105 | | | | | | | | |
| 2000 | | | 52,606 | 130,285 | 163,854 | 25,350 | 10,670 | 2,007 | 3,741 | | | | | | | | |
| 2001 | | | 42,329 | 386,498 | 214,243 | 84,322 | 17,177 | 9,279 | 1,320 | 464 | | | | | | | |
| 2002 | | | 310 | 57,771 | 233,715 | 73,361 | 23,839 | 9,622 | 6,047 | 785 | 1,454 | | 2,170 | | | | |
| 2003 | | | | 4,884 | 37,189 | 149,359 | 188,046 | 41,113 | 18,104 | 7,470 | 5,073 | 1,170 | 1,724 | 817 | | | |
| 2004 | | | | 97 | 98,544 | 72,720 | 129,126 | 58,696 | 11,806 | 4,675 | 1,764 | 1,182 | 224 | 609 | | | |
| 2005 | | | | | 3,181 | 47,690 | 280,723 | 19,902 | 57,931 | 23,160 | 6,401 | 4,575 | 1,601 | 830 | 649 | 251 | |
| 2006 | | | | | 167 | 29,903 | 47,416 | 78,493 | 5,155 | 14,283 | 7,461 | 2,864 | 1,753 | 636 | 344 | 184 | 41 |
| 2007 | | | | | 1,762 | 35,777 | 186,312 | 25,702 | 42,350 | 1,937 | 3,598 | 2,781 | 1,394 | 737 | 392 | 595 | 96 |
| 2008 | | | | | 3,945 | 93,103 | 123,240 | 101,819 | 27,956 | 26,590 | 1,476 | 2,097 | 2,330 | | | | |
| 2009 | | | | | 1,529 | 74,035 | 162,755 | 66,702 | 66,208 | 3,325 | 8,426 | 210 | 1,685 | 931 | 914 | 192 | |
| 2010 | | | | | 10,155 | 93,506 | 204,897 | 141,754 | 37,562 | 9,467 | 3,124 | 1,413 | 223 | | 1,785 | | |
| 2011 | | | | | 3,419 | 88,254 | 176,415 | 150,699 | 77,558 | 8,261 | 9,161 | 1,523 | 394 | 143 | 95 | 107 | 21 |

Table A.45. Mean weights-at-age (kg) of recreationally landed Gulf of Maine Atlantic cod from 1982 to 2011.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|-------|-------|-------|-------|-------|-------|-------|--------|--------|--------|--------|--------|--------|--------|--------|--------|-------|
| 1981 | | 0.341 | 0.995 | 1.524 | 2.915 | 4.715 | 5.645 | 5.863 | 8.359 | 12.339 | | 18.100 | | | | | |
| 1982 | 0.022 | 0.372 | 0.848 | 1.401 | 2.209 | 5.362 | 6.955 | 9.732 | 8.990 | 11.008 | 11.547 | | 21.416 | | | | |
| 1983 | | 0.378 | 0.791 | 1.398 | 2.401 | 3.772 | 6.032 | 6.745 | 8.393 | 9.627 | 15.175 | 19.306 | 19.182 | | | | |
| 1984 | | 0.372 | 0.775 | 1.365 | 2.668 | 4.005 | 5.349 | 6.559 | 6.583 | 8.955 | 11.743 | 13.474 | | 17.780 | 27.103 | | |
| 1985 | | 0.346 | 0.752 | 1.281 | 2.810 | 5.310 | 6.771 | 8.645 | 11.257 | 11.854 | 12.252 | 8.049 | 9.297 | 8.332 | | | |
| 1986 | | 0.375 | 0.668 | 1.589 | 2.770 | 5.308 | 7.418 | 8.584 | 11.185 | 11.839 | 14.266 | 14.560 | 22.376 | 14.560 | | | |
| 1987 | | 0.243 | 0.900 | 1.472 | 2.696 | 4.196 | 8.162 | 10.978 | 11.301 | 12.673 | 13.141 | 13.835 | 8.332 | | | | |
| 1988 | | 0.170 | 0.787 | 1.528 | 2.188 | 4.550 | 4.414 | 5.123 | 10.614 | | | 10.175 | | | | | |
| 1989 | | 0.539 | 0.989 | 1.500 | 2.700 | 4.579 | 6.191 | 8.715 | 7.616 | 17.137 | | | | | | | |
| 1990 | | 0.132 | 0.916 | 1.439 | 2.261 | 4.965 | 7.351 | 8.502 | 10.658 | 13.166 | | | | | | | |
| 1991 | | 0.180 | 1.088 | 1.499 | 2.025 | 3.388 | 6.933 | 13.033 | | | 3.838 | | | | | | |
| 1992 | | 0.106 | 1.360 | 1.715 | 2.541 | 2.923 | 4.437 | 9.324 | 2.516 | | | | | | | | |
| 1993 | | 0.184 | 0.805 | 1.566 | 1.827 | 2.890 | 3.791 | 11.707 | | | | | | | | | |
| 1994 | | 0.136 | 1.169 | 1.514 | 2.262 | 2.270 | 5.374 | 5.751 | 18.165 | 2.156 | | | | | | | |
| 1995 | | 0.509 | 1.432 | 1.514 | 1.769 | 3.381 | 2.479 | | 4.244 | | | | | | | | |
| 1996 | | | 1.483 | 1.809 | 1.863 | 2.502 | 9.632 | 8.622 | | 13.434 | | | | | | | |
| 1997 | | 0.307 | 1.626 | 1.924 | 2.389 | 2.396 | 2.964 | 6.038 | 11.932 | | | | | | | | |
| 1998 | 0.010 | | 1.600 | 2.071 | 2.435 | 3.491 | 3.179 | 4.591 | 12.220 | | | | | | | | |
| 1999 | | 0.290 | 1.296 | 1.943 | 2.951 | 3.687 | 5.490 | 5.561 | 7.637 | | | | | | | | |
| 2000 | | | 1.561 | 1.961 | 2.718 | 3.199 | 5.103 | 5.023 | 10.277 | | | | | | | | |
| 2001 | | | 1.709 | 2.199 | 2.659 | 3.732 | 5.019 | 6.259 | 10.560 | 5.813 | | | | | | | |
| 2002 | | | 1.275 | 2.135 | 2.581 | 3.048 | 5.265 | 6.429 | 7.919 | 8.984 | 10.569 | | 21.420 | | | | |
| 2003 | | | 1.954 | 2.237 | 2.525 | 3.225 | 4.822 | 8.064 | 9.802 | 11.167 | 11.115 | 15.401 | 21.534 | | | | |
| 2004 | | | 1.545 | 2.045 | 2.612 | 2.829 | 3.911 | 5.747 | 9.387 | 12.100 | 13.609 | 13.256 | 20.155 | | | | |
| 2005 | | | 1.510 | 1.968 | 2.374 | 3.566 | 3.904 | 6.089 | 7.852 | 9.766 | 13.574 | 14.627 | 16.347 | 17.544 | | | |
| 2006 | | | 2.321 | 2.270 | 2.969 | 3.301 | 4.683 | 5.470 | 8.339 | 10.105 | 12.466 | 15.021 | 15.090 | 18.390 | 17.774 | | |
| 2007 | | | 2.226 | 2.503 | 2.965 | 3.535 | 4.418 | 5.147 | 7.863 | 11.709 | 12.713 | 14.426 | 14.231 | 16.520 | 15.964 | 19.820 | |
| 2008 | | | 1.922 | 2.746 | 2.910 | 3.415 | 2.747 | 5.124 | 10.004 | 12.290 | 18.942 | | | | | | |
| 2009 | | | 2.197 | 2.506 | 3.066 | 3.518 | 4.444 | 6.371 | 8.034 | 9.777 | 10.005 | 12.269 | 18.736 | 19.782 | | | |
| 2010 | | | 2.563 | 2.728 | 3.151 | 3.771 | 4.115 | 7.441 | 9.409 | 9.584 | 9.850 | | 15.000 | | | | |
| 2011 | | | 1.798 | 2.474 | 3.032 | 3.707 | 4.577 | 5.274 | 5.624 | 12.022 | 16.019 | 18.353 | 14.407 | 19.306 | 13.835 | | |

Table A.46. Total Gulf of Maine Atlantic cod recreational discards-at-age (numbers) from 1982 to 2011. *These estimates include assumptions of 30% discard survival.*

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|------|--------|---------|---------|---------|--------|--------|-------|------|------|-------|-------|-------|-------|-------|-------|-------|
| 1981 | | 13,575 | 24,578 | 5,363 | | | | | | | | | | | | | |
| 1982 | | 5,612 | 14,535 | 1,052 | 278 | | | | | | | | | | | | |
| 1983 | | 20,028 | 31,320 | 901 | | | | | | | | | | | | | |
| 1984 | | 8,107 | 33,657 | 2,856 | | | | | | | | | | | | | |
| 1985 | | 10,816 | 25,312 | 9,151 | | | | | | | | | | | | | |
| 1986 | | 7,925 | 18,474 | 492 | 675 | | | | | | | | | | | | |
| 1987 | | 12,226 | 99,875 | 16,449 | | | | | | | | | | | | | |
| 1988 | | 6,688 | 28,038 | 5,279 | | | | | | | | | | | | | |
| 1989 | | 5,478 | 74,963 | 46,707 | 2,626 | | | | | | | | | | | | |
| 1990 | | 1,273 | 22,214 | 75,071 | 8,729 | | | | | | | | | | | | |
| 1991 | | 2,352 | 20,600 | 23,716 | 42,819 | 3,603 | | | | | | | | | | | |
| 1992 | | 3,446 | 24,659 | 18,197 | 2,446 | 5,287 | 198 | | | | | | | | | | |
| 1993 | | 3,791 | 97,835 | 49,454 | 19,319 | | | | | | | | | | | | |
| 1994 | | 4,326 | 65,863 | 86,959 | 5,930 | | | | | | | | | | | | |
| 1995 | | 3,848 | 42,660 | 91,272 | 16,491 | 579 | | | | | | | | | | | |
| 1996 | | 5,817 | 21,418 | 31,232 | 40,139 | 3,642 | | | | | | | | | | | |
| 1997 | | 2,950 | 21,137 | 25,402 | 6,176 | 11,777 | 660 | | | | | | | | | | |
| 1998 | | 3,376 | 37,760 | 26,503 | 17,554 | 289 | 1,398 | | | | | | | | | | |
| 1999 | | 14,776 | 47,252 | 37,178 | 6,006 | 2,315 | 313 | 84 | | | | | | | | | |
| 2000 | | 13,781 | 137,217 | 45,526 | 11,069 | 1,145 | 112 | | | | | | | | | | |
| 2001 | | | 141,504 | 124,214 | 26,316 | 5,148 | 423 | | | | | | | | | | |
| 2002 | | 6,452 | 13,217 | 110,592 | 94,169 | 21,982 | 244 | | 394 | | | | | | | | |
| 2003 | | 14,672 | 52,512 | 34,528 | 102,484 | 41,375 | 5,760 | | | | | | | | | | |
| 2004 | | 18,746 | 33,734 | 134,010 | 14,587 | 16,564 | 3,407 | | | | | | | | | | |
| 2005 | | 3,799 | 102,844 | 46,076 | 153,325 | 2,048 | 3,247 | 79 | 9 | 14 | | | | | | | |
| 2006 | 27 | 8,728 | 28,442 | 121,853 | 22,392 | 28,622 | 1,369 | 530 | 542 | 5 | | | | | | | |
| 2007 | 23 | 1,451 | 52,053 | 110,524 | 110,351 | 8,306 | 6,602 | 9 | 11 | | | | | | | | |
| 2008 | 110 | 4,558 | 64,400 | 117,489 | 58,727 | 37,397 | 2,826 | 131 | | | | | | | | | |
| 2009 | 18 | 4,860 | 44,423 | 97,205 | 67,844 | 21,111 | 11,863 | 184 | 303 | | | | | | | | |
| 2010 | | 3,552 | 48,239 | 127,212 | 78,138 | 46,935 | 9,364 | 1,382 | | | | | | | | | |
| 2011 | | 626 | 7,071 | 43,222 | 104,012 | 87,852 | 20,425 | 4,033 | 363 | 128 | | | | | | | |

Table A.47. Mean weights-at-age (kg) of recreationally discarded Gulf of Maine Atlantic cod from 1982 to 2011.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1981 | 0.367 | 0.456 | 0.492 | | | | | | | | | | | | | | |
| 1982 | 0.307 | 0.400 | 0.450 | 0.509 | | | | | | | | | | | | | |
| 1983 | 0.260 | 0.386 | 0.326 | | | | | | | | | | | | | | |
| 1984 | 0.288 | 0.387 | 0.436 | | | | | | | | | | | | | | |
| 1985 | 0.272 | 0.395 | 0.426 | | | | | | | | | | | | | | |
| 1986 | 0.319 | 0.380 | 0.429 | 0.499 | | | | | | | | | | | | | |
| 1987 | 0.221 | 0.393 | 0.371 | | | | | | | | | | | | | | |
| 1988 | 0.185 | 0.357 | 0.438 | | | | | | | | | | | | | | |
| 1989 | 0.395 | 0.524 | 0.692 | 0.867 | | | | | | | | | | | | | |
| 1990 | 0.231 | 0.528 | 0.637 | 0.786 | | | | | | | | | | | | | |
| 1991 | 0.234 | 0.536 | 0.776 | 0.819 | 0.818 | | | | | | | | | | | | |
| 1992 | 0.217 | 0.590 | 0.724 | 0.837 | 0.902 | 0.868 | | | | | | | | | | | |
| 1993 | 0.252 | 0.487 | 0.769 | 0.794 | | | | | | | | | | | | | |
| 1994 | 0.283 | 0.470 | 0.740 | 0.683 | | | | | | | | | | | | | |
| 1995 | 0.302 | 0.520 | 0.635 | 0.870 | 0.931 | | | | | | | | | | | | |
| 1996 | 0.277 | 0.655 | 0.827 | 0.902 | 0.918 | | | | | | | | | | | | |
| 1997 | 0.196 | 0.685 | 0.915 | 1.095 | 1.092 | 1.294 | | | | | | | | | | | |
| 1998 | 0.203 | 0.630 | 1.007 | 1.072 | 1.211 | 1.365 | | | | | | | | | | | |
| 1999 | 0.301 | 0.535 | 0.869 | 1.078 | 1.157 | 1.097 | 1.456 | | | | | | | | | | |
| 2000 | 0.275 | 0.574 | 0.911 | 1.109 | 1.003 | 1.211 | | | | | | | | | | | |
| 2001 | | 0.581 | 0.886 | 1.098 | 1.105 | 1.290 | | | | | | | | | | | |
| 2002 | 0.156 | 0.468 | 1.035 | 1.406 | 1.444 | 1.371 | | 1.937 | | | | | | | | | |
| 2003 | 0.345 | 0.544 | 1.223 | 1.327 | 1.507 | 1.422 | | | | | | | | | | | |
| 2004 | 0.142 | 0.523 | 0.963 | 1.429 | 1.528 | 1.721 | | | | | | | | | | | |
| 2005 | 0.213 | 0.509 | 1.012 | 1.050 | 1.034 | 1.316 | 1.940 | 2.516 | 1.734 | | | | | | | | |
| 2006 | 0.086 | 0.304 | 0.565 | 0.869 | 1.216 | 1.346 | 1.263 | 1.773 | 1.656 | 2.851 | | | | | | | |
| 2007 | 0.048 | 0.167 | 0.642 | 1.062 | 1.289 | 1.603 | 1.548 | 2.768 | 3.977 | | | | | | | | |
| 2008 | 0.105 | 0.320 | 0.817 | 1.119 | 1.296 | 1.285 | 1.744 | 5.263 | | | | | | | | | |
| 2009 | 0.057 | 0.315 | 0.803 | 1.194 | 1.338 | 1.381 | 1.544 | 2.142 | 1.739 | | | | | | | | |
| 2010 | | 0.282 | 0.952 | 1.059 | 1.448 | 1.528 | 1.449 | 3.196 | | | | | | | | | |
| 2011 | 0.084 | 0.322 | 0.873 | 1.341 | 1.328 | 1.497 | 1.631 | 1.834 | 2.221 | | | | | | | | |

Table A.48. Total catch-at-age (numbers, 000s of fish) of Gulf of Maine Atlantic cod from 1982 to 2011 with an age 9⁺ group. *Only ages 1 through the 9⁺ group are used as assessment model inputs.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------|-------|-------|--------|--------|--------|--------|-------|-------|------|-------|
| 1982 | 1.3 | 448.8 | 2926.5 | 2287.2 | 1430.7 | 748.8 | 65.9 | 94.1 | 72.6 | 90.1 |
| 1983 | 13.6 | 597.5 | 2462.0 | 2913.2 | 1201.6 | 704.0 | 452.7 | 50.0 | 62.5 | 56.2 |
| 1984 | 18.3 | 370.3 | 2129.6 | 1675.9 | 1643.6 | 437.5 | 219.6 | 105.6 | 9.5 | 53.4 |
| 1985 | 67.1 | 505.7 | 1944.3 | 2405.1 | 1151.8 | 738.1 | 161.4 | 107.2 | 48.4 | 33.2 |
| 1986 | 17.8 | 760.7 | 1747.0 | 2747.8 | 992.0 | 279.3 | 202.7 | 48.0 | 38.2 | 47.5 |
| 1987 | 100.7 | 281.8 | 2018.3 | 1568.3 | 1574.5 | 345.4 | 89.4 | 81.0 | 14.5 | 37.5 |
| 1988 | 3.4 | 415.1 | 1542.8 | 2086.6 | 1156.9 | 447.7 | 67.4 | 25.6 | 26.2 | 9.3 |
| 1989 | 0.0 | 166.4 | 1247.2 | 2385.1 | 1651.9 | 521.1 | 87.1 | 70.3 | 9.4 | 19.6 |
| 1990 | 0.0 | 65.5 | 812.5 | 5547.8 | 2717.6 | 541.4 | 189.1 | 29.7 | 36.4 | 43.3 |
| 1991 | 3.3 | 121.6 | 499.6 | 942.7 | 5561.3 | 1037.9 | 150.7 | 55.5 | 26.0 | 15.8 |
| 1992 | 23.8 | 370.3 | 830.1 | 867.6 | 502.1 | 2190.0 | 226.2 | 80.2 | 6.0 | 5.5 |
| 1993 | 26.6 | 105.9 | 512.3 | 2149.0 | 944.7 | 103.3 | 497.1 | 41.6 | 11.3 | 0.0 |
| 1994 | 11.7 | 124.0 | 201.9 | 1525.6 | 1294.2 | 266.3 | 66.2 | 74.2 | 28.7 | 7.9 |
| 1995 | 11.6 | 78.9 | 319.5 | 1321.8 | 1260.4 | 221.7 | 29.9 | 6.5 | 18.2 | 2.8 |
| 1996 | 22.1 | 37.5 | 111.6 | 627.7 | 2003.9 | 405.9 | 36.7 | 4.0 | 0.5 | 1.6 |
| 1997 | 1.5 | 69.1 | 137.5 | 519.6 | 467.8 | 869.2 | 72.5 | 5.5 | 2.3 | 1.0 |
| 1998 | 0.9 | 5.9 | 171.1 | 492.3 | 628.9 | 152.8 | 205.9 | 28.7 | 5.2 | 2.3 |
| 1999 | 0.1 | 73.9 | 90.9 | 347.8 | 336.6 | 172.3 | 53.7 | 59.5 | 12.4 | 1.1 |
| 2000 | 0.0 | 24.8 | 485.0 | 556.5 | 813.7 | 176.6 | 85.2 | 12.5 | 10.5 | 0.0 |
| 2001 | 0.0 | 0.6 | 394.0 | 1163.8 | 684.4 | 385.5 | 106.6 | 57.2 | 8.3 | 11.6 |
| 2002 | 0.0 | 16.8 | 41.6 | 374.9 | 912.6 | 323.8 | 163.5 | 66.4 | 28.1 | 20.3 |
| 2003 | 22.9 | 44.9 | 125.6 | 167.8 | 582.1 | 706.1 | 186.0 | 75.7 | 29.2 | 26.8 |
| 2004 | 0.2 | 149.4 | 105.9 | 609.3 | 259.7 | 407.4 | 251.6 | 68.4 | 33.0 | 27.4 |
| 2005 | 1.5 | 23.5 | 180.1 | 159.6 | 945.8 | 89.2 | 246.6 | 109.1 | 28.5 | 31.7 |
| 2006 | 0.2 | 19.2 | 59.1 | 426.6 | 290.1 | 461.7 | 30.3 | 79.7 | 39.0 | 27.3 |
| 2007 | 0.4 | 12.2 | 108.5 | 299.4 | 976.4 | 137.4 | 230.2 | 7.9 | 19.2 | 22.0 |
| 2008 | 0.4 | 12.2 | 130.5 | 598.4 | 707.4 | 780.5 | 86.4 | 110.6 | 4.0 | 16.6 |
| 2009 | 0.1 | 10.7 | 101.5 | 622.5 | 1093.3 | 477.9 | 304.8 | 20.9 | 30.5 | 9.6 |
| 2010 | 0.2 | 8.2 | 83.6 | 394.5 | 888.5 | 668.3 | 164.3 | 71.7 | 11.2 | 7.6 |
| 2011 | 0.7 | 8.7 | 60.5 | 322.2 | 589.6 | 573.9 | 339.9 | 34.9 | 38.4 | 9.4 |

Table A.49. Mean weights-at-age (kg) of the total catch Gulf of Maine Atlantic cod from 1982 to 2011 an age 9⁺ group. Mean catch weights-at-age in the 9⁺ group were estimated using a numbers weighted approach. Cells shaded grey were imputed using a 5-year centered moving average, cells shaded red were imputed using a time series average. *Only ages 1 through the 9⁺ group are used as assessment model inputs.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------|-------|-------|-------|-------|-------|-------|-------|--------|--------|--------|
| 1982 | 0.012 | 0.356 | 0.858 | 1.514 | 2.606 | 5.067 | 7.065 | 9.620 | 9.771 | 15.664 |
| 1983 | 0.024 | 0.224 | 0.768 | 1.542 | 2.418 | 3.808 | 6.055 | 6.071 | 10.317 | 13.325 |
| 1984 | 0.001 | 0.234 | 0.653 | 1.478 | 2.678 | 3.609 | 5.540 | 8.368 | 10.138 | 14.828 |
| 1985 | 0.039 | 0.206 | 0.733 | 1.404 | 2.819 | 4.658 | 5.884 | 8.502 | 11.244 | 13.676 |
| 1986 | 0.005 | 0.277 | 0.501 | 1.698 | 2.774 | 4.778 | 6.504 | 8.109 | 10.206 | 14.646 |
| 1987 | 0.004 | 0.154 | 0.642 | 1.323 | 3.090 | 4.668 | 7.259 | 10.036 | 11.099 | 14.582 |
| 1988 | 0.003 | 0.122 | 0.577 | 1.666 | 2.360 | 5.205 | 5.200 | 6.193 | 10.103 | 12.993 |
| 1989 | 0.046 | 0.236 | 0.752 | 1.518 | 2.959 | 4.282 | 5.980 | 9.276 | 12.519 | 20.913 |
| 1990 | 0.021 | 0.193 | 0.811 | 1.349 | 2.141 | 4.474 | 7.721 | 10.820 | 11.750 | 18.718 |
| 1991 | 0.014 | 0.236 | 1.113 | 1.601 | 2.281 | 3.894 | 7.144 | 10.429 | 12.261 | 14.031 |
| 1992 | 0.023 | 0.055 | 1.033 | 1.530 | 2.747 | 2.976 | 5.587 | 10.921 | 10.483 | 14.483 |
| 1993 | 0.021 | 0.081 | 0.690 | 1.748 | 2.150 | 4.420 | 5.670 | 9.817 | 13.673 | 15.701 |
| 1994 | 0.022 | 0.058 | 0.730 | 1.712 | 3.085 | 3.251 | 6.335 | 7.684 | 12.542 | 11.846 |
| 1995 | 0.027 | 0.103 | 1.288 | 1.591 | 2.649 | 5.090 | 6.865 | 11.466 | 13.128 | 22.443 |
| 1996 | 0.033 | 0.100 | 1.293 | 2.096 | 2.260 | 3.462 | 7.558 | 11.728 | 14.455 | 16.269 |
| 1997 | 0.017 | 0.064 | 1.351 | 2.128 | 3.022 | 3.074 | 4.699 | 9.000 | 12.156 | 16.938 |
| 1998 | 0.008 | 0.202 | 1.071 | 1.931 | 2.633 | 3.972 | 4.255 | 7.122 | 12.118 | 16.676 |
| 1999 | 0.052 | 0.222 | 0.635 | 1.723 | 2.777 | 3.892 | 5.670 | 6.704 | 9.811 | 12.279 |
| 2000 | 0.030 | 0.282 | 1.081 | 2.150 | 3.316 | 4.325 | 5.898 | 5.352 | 9.331 | 12.680 |
| 2001 | 0.045 | 0.316 | 0.890 | 2.176 | 3.144 | 4.666 | 6.140 | 7.273 | 9.072 | 9.559 |
| 2002 | 0.032 | 0.185 | 0.795 | 1.797 | 2.906 | 3.792 | 6.132 | 6.969 | 8.808 | 12.205 |
| 2003 | 0.038 | 0.202 | 0.809 | 1.843 | 2.378 | 3.654 | 5.112 | 7.649 | 9.191 | 12.058 |
| 2004 | 0.025 | 0.111 | 0.483 | 1.606 | 2.965 | 3.547 | 5.350 | 7.220 | 9.764 | 13.303 |
| 2005 | 0.027 | 0.126 | 0.558 | 1.625 | 2.401 | 4.233 | 4.502 | 6.349 | 8.002 | 12.549 |
| 2006 | 0.071 | 0.289 | 0.648 | 1.493 | 2.932 | 3.357 | 4.463 | 5.562 | 7.430 | 12.146 |
| 2007 | 0.025 | 0.220 | 0.744 | 1.731 | 2.922 | 3.735 | 4.771 | 6.167 | 7.302 | 12.394 |
| 2008 | 0.085 | 0.247 | 0.862 | 2.179 | 2.818 | 3.530 | 3.988 | 5.819 | 7.528 | 12.044 |
| 2009 | 0.032 | 0.337 | 0.911 | 2.153 | 3.126 | 3.575 | 4.368 | 5.959 | 8.000 | 12.887 |
| 2010 | 0.023 | 0.264 | 1.200 | 1.995 | 3.203 | 3.914 | 4.447 | 5.708 | 8.730 | 11.612 |
| 2011 | 0.086 | 0.329 | 0.933 | 2.056 | 2.874 | 3.870 | 4.839 | 5.717 | 5.953 | 12.984 |

Table A.50. Mean January 1/spawning stock weights-at-age (kg) of Gulf of Maine Atlantic cod from 1982 to 2011 an age 9⁺ group. Weights were estimated from catch weights using Rivard (1980, 1982) approach. Cells shaded grey were imputed using a 5-year centered moving average, cells shaded red were imputed using a time series average. *Only ages 1 through the 9⁺ group are used as assessment model inputs.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------|-------|-------|-------|-------|-------|-------|-------|-------|--------|--------|
| 1982 | 0.002 | 0.241 | 0.594 | 1.165 | 2.127 | 4.635 | 7.622 | 9.289 | 9.695 | 15.664 |
| 1983 | 0.008 | 0.050 | 0.501 | 1.114 | 1.894 | 3.136 | 5.539 | 6.549 | 9.962 | 13.325 |
| 1984 | 0.000 | 0.075 | 0.372 | 1.019 | 2.021 | 2.952 | 4.593 | 7.118 | 7.845 | 14.828 |
| 1985 | 0.015 | 0.014 | 0.403 | 0.910 | 2.013 | 3.532 | 4.608 | 6.863 | 9.700 | 13.676 |
| 1986 | 0.001 | 0.104 | 0.316 | 1.077 | 1.917 | 3.670 | 5.504 | 6.908 | 9.315 | 14.646 |
| 1987 | 0.001 | 0.028 | 0.406 | 0.777 | 2.273 | 3.574 | 5.889 | 8.079 | 9.487 | 14.582 |
| 1988 | 0.000 | 0.022 | 0.293 | 0.980 | 1.709 | 4.010 | 4.927 | 6.705 | 10.069 | 12.993 |
| 1989 | 0.022 | 0.027 | 0.292 | 0.887 | 2.179 | 3.172 | 5.578 | 6.945 | 8.799 | 20.913 |
| 1990 | 0.006 | 0.095 | 0.431 | 0.937 | 1.742 | 3.627 | 5.750 | 8.043 | 10.440 | 18.718 |
| 1991 | 0.007 | 0.071 | 0.450 | 1.083 | 1.689 | 2.846 | 5.654 | 8.972 | 11.518 | 14.060 |
| 1992 | 0.012 | 0.028 | 0.476 | 1.215 | 2.026 | 2.564 | 4.629 | 8.832 | 10.453 | 14.483 |
| 1993 | 0.012 | 0.046 | 0.191 | 1.254 | 1.702 | 3.449 | 4.083 | 7.388 | 12.219 | 15.708 |
| 1994 | 0.010 | 0.038 | 0.236 | 1.003 | 2.244 | 2.571 | 5.294 | 6.601 | 11.095 | 11.846 |
| 1995 | 0.012 | 0.051 | 0.275 | 0.946 | 2.021 | 3.934 | 4.722 | 8.526 | 10.045 | 22.443 |
| 1996 | 0.022 | 0.060 | 0.356 | 1.462 | 1.784 | 2.971 | 6.185 | 8.967 | 12.844 | 16.357 |
| 1997 | 0.005 | 0.049 | 0.391 | 1.466 | 2.407 | 2.571 | 3.973 | 8.245 | 11.940 | 16.938 |
| 1998 | 0.002 | 0.059 | 0.256 | 1.445 | 2.245 | 3.423 | 3.558 | 5.739 | 10.442 | 16.676 |
| 1999 | 0.022 | 0.044 | 0.343 | 1.196 | 2.237 | 3.139 | 4.752 | 5.301 | 8.351 | 12.279 |
| 2000 | 0.009 | 0.120 | 0.461 | 1.063 | 2.257 | 3.422 | 4.773 | 5.508 | 7.882 | 12.661 |
| 2001 | 0.023 | 0.097 | 0.456 | 1.305 | 2.420 | 3.851 | 5.091 | 6.513 | 6.912 | 9.538 |
| 2002 | 0.012 | 0.089 | 0.465 | 1.050 | 2.249 | 3.247 | 5.296 | 6.514 | 7.924 | 12.152 |
| 2003 | 0.022 | 0.089 | 0.346 | 1.053 | 1.742 | 2.977 | 4.118 | 6.837 | 8.011 | 12.023 |
| 2004 | 0.011 | 0.066 | 0.351 | 0.971 | 2.110 | 2.620 | 4.199 | 5.908 | 8.627 | 13.288 |
| 2005 | 0.008 | 0.060 | 0.248 | 0.821 | 1.654 | 3.338 | 3.841 | 5.758 | 7.593 | 12.546 |
| 2006 | 0.043 | 0.089 | 0.295 | 0.808 | 1.890 | 2.467 | 4.076 | 4.912 | 6.744 | 12.137 |
| 2007 | 0.009 | 0.124 | 0.450 | 0.925 | 1.771 | 3.005 | 3.723 | 5.020 | 6.329 | 12.394 |
| 2008 | 0.046 | 0.085 | 0.420 | 1.117 | 1.888 | 2.892 | 3.630 | 5.147 | 6.803 | 12.040 |
| 2009 | 0.014 | 0.171 | 0.480 | 1.248 | 2.283 | 2.908 | 3.658 | 4.735 | 6.735 | 12.878 |
| 2010 | 0.006 | 0.100 | 0.589 | 1.168 | 2.328 | 3.198 | 3.685 | 4.778 | 7.153 | 11.612 |
| 2011 | 0.084 | 0.087 | 0.492 | 1.353 | 1.972 | 3.262 | 4.114 | 4.788 | 5.751 | 12.995 |

Table A.51. Summary of vessels and trawl doors used in the Northeast Fisheries Science Center (NEFSC) spring and fall surveys from 1963 to 2012. All survey indices are standardized to Albatross IV, Polyvalent door equivalents. *Note, the spring survey did not begin until 1968, 2012 fall survey data are not available at time of this report.

| Year | Spring | Autumn | Door |
|------|--------------------------|--------------------------|--------------|
| 1963 | | Albatross IV | BMV |
| 1964 | | Albatross IV | BMV |
| 1965 | | Albatross IV | BMV |
| 1966 | | Albatross IV | BMV |
| 1967 | | Albatross IV | BMV |
| 1968 | Albatross IV | Albatross IV | BMV |
| 1969 | Albatross IV | Albatross IV | BMV |
| 1970 | Albatross IV | Albatross IV | BMV |
| 1971 | Albatross IV | Albatross IV | BMV |
| 1972 | Albatross IV | Albatross IV | BMV |
| 1973 | Albatross IV | Albatross IV | BMV |
| 1974 | Albatross IV | Albatross IV | BMV |
| 1975 | Albatross IV | Albatross IV | BMV |
| 1976 | Albatross IV | Albatross IV | BMV |
| 1977 | Albatross IV | Delaware II | BMV |
| 1978 | Albatross IV | Delaware II | BMV |
| 1979 | Albatross IV/Delaware II | Albatross IV/Delaware II | BMV |
| 1980 | Albatross IV/Delaware II | Delaware II | BMV |
| 1981 | Delaware II | Albatross IV/Delaware II | BMV |
| 1982 | Delaware II | Albatross IV | BMV |
| 1983 | Albatross IV | Albatross IV | BMV |
| 1984 | Albatross IV | Albatross IV | BMV |
| 1985 | Albatross IV | Albatross IV | Polyvalent |
| 1986 | Albatross IV | Albatross IV | Polyvalent |
| 1987 | Albatross IV/Delaware II | Albatross IV | Polyvalent |
| 1988 | Albatross IV | Albatross IV/Delaware II | Polyvalent |
| 1989 | Delaware II | Delaware II | Polyvalent |
| 1990 | Delaware II | Delaware II | Polyvalent |
| 1991 | Delaware II | Delaware II | Polyvalent |
| 1992 | Albatross IV | Albatross IV | Polyvalent |
| 1993 | Albatross IV | Delaware II | Polyvalent |
| 1994 | Delaware II | Albatross IV | Polyvalent |
| 1995 | Albatross IV | Albatross IV | Polyvalent |
| 1996 | Albatross IV | Albatross IV | Polyvalent |
| 1997 | Albatross IV | Albatross IV | Polyvalent |
| 1998 | Albatross IV | Albatross IV | Polyvalent |
| 1999 | Albatross IV | Albatross IV | Polyvalent |
| 2000 | Albatross IV | Albatross IV | Polyvalent |
| 2001 | Albatross IV | Albatross IV | Polyvalent |
| 2002 | Albatross IV | Albatross IV | Polyvalent |
| 2003 | Delaware II | Albatross IV | Polyvalent |
| 2004 | Albatross IV | Albatross IV | Polyvalent |
| 2005 | Albatross IV | Albatross IV | Polyvalent |
| 2006 | Albatross IV | Albatross IV | Polyvalent |
| 2007 | Albatross IV | Albatross IV | Polyvalent |
| 2008 | Albatross IV | Albatross IV | Polyvalent |
| 2009 | Henry B. Bigelow | Henry B. Bigelow | PolyIce oval |
| 2010 | Henry B. Bigelow | Henry B. Bigelow | PolyIce oval |
| 2011 | Henry B. Bigelow | Henry B. Bigelow | PolyIce oval |
| 2012 | Henry B. Bigelow | | PolyIce oval |

Table A.52. Summary of survey calibration coefficients for converting survey index values to Albatross IV, Polyvalent door equivalent units.

| Calibration type | Index | Length (cm) | Calibration coefficient | Lower 95% CI | Upper 95% CI | Source |
|-----------------------------|---------------------|-------------|-------------------------|--------------|--------------|------------------------|
| Deleware II to Albatross IV | Biomass (weight) | NA | 0.670 | 0.530 | 0.870 | Forrester et al., 1997 |
| | Abundance (numbers) | NA | 0.790 | 0.690 | 0.940 | |
| BMV door to Polyvalent door | Biomass (weight) | NA | 1.620 | 1.370 | 1.940 | Miller et al. 2010 |
| | Abundance (numbers) | NA | 1.560 | 1.330 | 1.880 | |
| Bigelow to Albatross IV | Biomass (weight) | NA | 1.580 | 0.906 | 1.643 | Brooks et al. 2010 |
| | | ≤ 20 | 5.724 | 4.166 | 7.864 | |
| | | 21 | 5.600 | 4.094 | 7.661 | |
| | | 22 | 5.477 | 4.022 | 7.458 | |
| | | 23 | 5.353 | 3.950 | 7.256 | |
| | | 24 | 5.230 | 3.877 | 7.054 | |
| | | 25 | 5.106 | 3.805 | 6.852 | |
| | | 26 | 4.983 | 3.733 | 6.651 | |
| | | 27 | 4.859 | 3.660 | 6.451 | |
| | | 28 | 4.736 | 3.588 | 6.251 | |
| | | 29 | 4.612 | 3.515 | 6.052 | |
| | | 30 | 4.489 | 3.442 | 5.854 | |
| | | 31 | 4.365 | 3.369 | 5.657 | |
| | | 32 | 4.242 | 3.295 | 5.460 | |
| | | 33 | 4.118 | 3.221 | 5.265 | |
| | | 34 | 3.995 | 3.147 | 5.071 | |
| | | 35 | 3.871 | 3.072 | 4.879 | |
| | | 36 | 3.748 | 2.996 | 4.688 | |
| | Abundance (numbers) | 37 | 3.624 | 2.919 | 4.499 | Brooks et al. 2010 |
| | | 38 | 3.501 | 2.841 | 4.313 | |
| | | 39 | 3.377 | 2.762 | 4.130 | |
| | | 40 | 3.254 | 2.680 | 3.950 | |
| | | 41 | 3.130 | 2.596 | 3.774 | |
| | | 42 | 3.007 | 2.509 | 3.604 | |
| | | 43 | 2.883 | 2.417 | 3.440 | |
| | | 44 | 2.760 | 2.320 | 3.284 | |
| | | 45 | 2.636 | 2.216 | 3.136 | |
| | | 46 | 2.513 | 2.105 | 2.999 | |
| | | 47 | 2.389 | 1.986 | 2.874 | |
| | | 48 | 2.266 | 1.860 | 2.760 | |
| | | 49 | 2.142 | 1.726 | 2.659 | |
| | | 50 | 2.019 | 1.586 | 2.569 | |
| | | 51 | 1.895 | 1.442 | 2.491 | |
| | | 52 | 1.772 | 1.295 | 2.423 | |
| | | 53 | 1.648 | 1.147 | 2.368 | |
| | | ≥ 54 | 1.602 | 1.092 | 2.350 | |

Table A.53. Summary of the differences in survey protocol from the FSV Albatross IV survey (2008 and earlier) and FSV Henry B. Bigelow (2009 - present). Adapted from Brooks et al. (2010).

| Measure | FSV Henry B Bigelow | FSV Albatross IV |
|--|---|---|
| Tow speed | 3.0 knots SOG | 3.8 knots SOG |
| Tow duration | 20min | 30 mins |
| Headrope height | 3.5-4m | 1-2m |
| Ground gear (cookies, rock hoppers, etc.) | Rockhopper Sweep Total Length-25.5m Center- 8.9m length, 16" rockhoppers. Wings- 8.2m each 14" rockhoppers | Roller Sweep Total Length-24.5m Center-5m length, 16" rollers. Wings- 9.75m each, 4" cookies. |
| Mesh | Poly webbing Forward Portion of trawl (jibs, upper and lower wing ends, 1 st &2 nd side panels, 1 st bottom belly)12cm,4mm Square aft to codend:6cm, 2.5mm Codend: 12cm, 4mm dbl. Codend Liner: 2.54cm, knotless | Nylon webbing Body of trawl= 12.7cm Codend- 11.5cm Liner (codend and aft portion of top belly)-1.27cm knotless |
| Net design | 4 Seam, 3 Bridle | Yankee 36 (recent years) |
| Door type | 550 kg PolyIce oval | 450 kg polyvalent |
| Other comments | Wing End to Door distance= 36.5m | Wing End to Door Distance= 9m |

Table A.54. Summary of the sampling of Northeast Fisheries Science Center (NEFSC) Gulf of Maine offshore survey strata broken down by survey (spring/fall) and time of day (day/night) between 1963 and spring 2011. The day/night classification is based on sunrise/sunset (zenith angle of $90^{\circ}50'$). *Note that the spring survey did not begin until 1968.

| Year | Strata sampled | | | | Tows sampled | | | |
|------|----------------|-------|------|-------|--------------|-------|------|-------|
| | Spring | | Fall | | Spring | | Fall | |
| | Day | Night | Day | Night | Day | Night | Day | Night |
| 1963 | | | 8 | 9 | | | 22 | 35 |
| 1964 | | | 10 | 9 | | | 15 | 32 |
| 1965 | | | 10 | 9 | | | 25 | 23 |
| 1966 | | | 9 | 9 | | | 22 | 21 |
| 1967 | | | 8 | 10 | | | 19 | 30 |
| 1968 | 8 | 10 | 9 | 10 | 27 | 23 | 19 | 31 |
| 1969 | 9 | 9 | 9 | 10 | 25 | 26 | 18 | 33 |
| 1970 | 6 | 9 | 10 | 10 | 17 | 35 | 21 | 32 |
| 1971 | 10 | 9 | 10 | 10 | 28 | 29 | 20 | 35 |
| 1972 | 10 | 9 | 8 | 9 | 28 | 27 | 24 | 31 |
| 1973 | 10 | 9 | 8 | 10 | 23 | 25 | 20 | 34 |
| 1974 | 10 | 8 | 9 | 9 | 29 | 18 | 28 | 29 |
| 1975 | 8 | 7 | 8 | 9 | 25 | 27 | 27 | 38 |
| 1976 | 8 | 9 | 7 | 10 | 30 | 34 | 17 | 38 |
| 1977 | 10 | 10 | 8 | 10 | 37 | 30 | 26 | 45 |
| 1978 | 10 | 10 | 10 | 9 | 37 | 29 | 54 | 66 |
| 1979 | 9 | 9 | 10 | 10 | 44 | 28 | 56 | 73 |
| 1980 | 10 | 8 | 10 | 10 | 26 | 24 | 23 | 28 |
| 1981 | 10 | 9 | 10 | 10 | 34 | 18 | 27 | 26 |
| 1982 | 9 | 9 | 10 | 10 | 32 | 21 | 21 | 33 |
| 1983 | 10 | 7 | 8 | 9 | 34 | 19 | 19 | 29 |
| 1984 | 9 | 10 | 7 | 9 | 31 | 19 | 20 | 31 |
| 1985 | 9 | 9 | 9 | 10 | 27 | 20 | 17 | 33 |
| 1986 | 9 | 10 | 7 | 9 | 25 | 27 | 19 | 34 |
| 1987 | 8 | 7 | 9 | 9 | 28 | 19 | 23 | 28 |
| 1988 | 10 | 9 | 8 | 9 | 35 | 19 | 23 | 29 |
| 1989 | 8 | 10 | 8 | 8 | 27 | 24 | 20 | 31 |
| 1990 | 9 | 10 | 8 | 10 | 23 | 29 | 23 | 29 |
| 1991 | 10 | 9 | 9 | 10 | 29 | 21 | 20 | 33 |
| 1992 | 10 | 9 | 9 | 10 | 29 | 23 | 21 | 30 |
| 1993 | 9 | 9 | 9 | 9 | 27 | 23 | 24 | 27 |
| 1994 | 10 | 9 | 8 | 10 | 35 | 18 | 18 | 32 |
| 1995 | 10 | 9 | 9 | 10 | 27 | 26 | 20 | 37 |
| 1996 | 10 | 9 | 10 | 9 | 27 | 25 | 25 | 27 |
| 1997 | 10 | 10 | 8 | 10 | 30 | 23 | 24 | 28 |
| 1998 | 10 | 10 | 9 | 10 | 39 | 36 | 33 | 34 |
| 1999 | 9 | 10 | 9 | 10 | 29 | 23 | 33 | 37 |
| 2000 | 9 | 9 | 9 | 10 | 30 | 22 | 21 | 31 |
| 2001 | 10 | 9 | 9 | 9 | 33 | 19 | 27 | 27 |
| 2002 | 10 | 10 | 10 | 10 | 29 | 26 | 27 | 22 |
| 2003 | 7 | 9 | 10 | 9 | 23 | 29 | 19 | 32 |
| 2004 | 10 | 8 | 8 | 9 | 32 | 18 | 21 | 27 |
| 2005 | 10 | 6 | 9 | 9 | 32 | 19 | 21 | 30 |
| 2006 | 10 | 10 | 8 | 9 | 33 | 26 | 25 | 33 |
| 2007 | 10 | 10 | 9 | 9 | 27 | 23 | 23 | 30 |
| 2008 | 10 | 9 | 10 | 10 | 30 | 21 | 21 | 32 |
| 2009 | 10 | 9 | 9 | 8 | 39 | 31 | 22 | 31 |
| 2010 | 8 | 10 | 9 | 9 | 34 | 30 | 22 | 29 |
| 2011 | 8 | 9 | | | 28 | 25 | | |

Table A.55. Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl survey indices for Gulf of Maine Atlantic cod from 1963 to 2012. *Note: the spring survey did not begin until 1968, 2012 fall survey data not available at time of this report.

| Year | Abundance (numbers/tow) | | Biomass (kg/tow) | |
|------|-------------------------|-------|------------------|--------|
| | Spring | Fall | Spring | Fall |
| 1963 | | 5.914 | | 17.950 |
| 1964 | | 4.015 | | 22.799 |
| 1965 | | 4.500 | | 12.089 |
| 1966 | | 3.720 | | 12.838 |
| 1967 | | 2.602 | | 9.313 |
| 1968 | 5.329 | 4.374 | 17.480 | 19.437 |
| 1969 | 3.215 | 2.758 | 13.100 | 15.154 |
| 1970 | 2.191 | 4.905 | 11.089 | 16.442 |
| 1971 | 1.429 | 4.361 | 7.004 | 16.529 |
| 1972 | 2.057 | 9.301 | 8.031 | 12.988 |
| 1973 | 7.525 | 4.452 | 18.807 | 8.764 |
| 1974 | 2.902 | 4.328 | 7.419 | 8.959 |
| 1975 | 2.512 | 6.143 | 6.039 | 8.619 |
| 1976 | 2.782 | 2.148 | 7.556 | 6.740 |
| 1977 | 3.872 | 3.073 | 8.541 | 10.199 |
| 1978 | 2.050 | 5.773 | 7.697 | 12.899 |
| 1979 | 3.644 | 3.142 | 7.555 | 13.927 |
| 1980 | 2.155 | 7.035 | 6.232 | 14.202 |
| 1981 | 4.832 | 2.349 | 10.650 | 7.533 |
| 1982 | 3.763 | 7.769 | 8.616 | 15.919 |
| 1983 | 3.912 | 2.786 | 10.962 | 8.416 |
| 1984 | 3.667 | 2.449 | 6.143 | 8.735 |
| 1985 | 2.517 | 2.821 | 7.645 | 8.264 |
| 1986 | 1.957 | 1.950 | 3.476 | 4.715 |
| 1987 | 1.083 | 2.996 | 1.976 | 3.394 |
| 1988 | 3.127 | 5.903 | 3.603 | 6.616 |
| 1989 | 2.112 | 4.553 | 2.424 | 4.535 |
| 1990 | 2.362 | 2.986 | 3.077 | 4.912 |
| 1991 | 2.393 | 1.252 | 2.891 | 2.782 |
| 1992 | 2.435 | 1.434 | 8.627 | 2.448 |
| 1993 | 2.507 | 1.232 | 5.875 | 1.003 |
| 1994 | 1.271 | 2.130 | 2.428 | 2.737 |
| 1995 | 1.930 | 2.008 | 2.432 | 3.665 |
| 1996 | 2.465 | 1.327 | 5.427 | 2.352 |
| 1997 | 2.192 | 0.872 | 5.616 | 1.872 |
| 1998 | 1.710 | 0.843 | 4.180 | 1.501 |
| 1999 | 2.301 | 1.807 | 5.090 | 3.505 |
| 2000 | 3.083 | 2.604 | 3.211 | 4.652 |
| 2001 | 2.147 | 1.980 | 6.215 | 7.324 |
| 2002 | 3.724 | 5.328 | 10.934 | 24.659 |
| 2003 | 3.677 | 2.529 | 9.495 | 5.988 |
| 2004 | 0.981 | 3.533 | 2.412 | 4.906 |
| 2005 | 1.765 | 1.338 | 2.701 | 2.897 |
| 2006 | 1.363 | 3.594 | 2.702 | 4.229 |
| 2007 | 12.393 | 1.992 | 15.811 | 2.714 |
| 2008 | 7.990 | 3.460 | 10.823 | 5.307 |
| 2009 | 3.599 | 3.447 | 7.161 | 5.845 |
| 2010 | 1.296 | 0.948 | 3.336 | 2.572 |
| 2011 | 0.894 | 0.990 | 2.133 | 2.647 |
| 2012 | 0.893 | | 1.645 | |
| Avg | 2.978 | 3.342 | 6.806 | 8.337 |
| Min | 0.893 | 0.843 | 1.645 | 1.003 |
| Max | 12.393 | 9.301 | 18.807 | 24.659 |

Table A.56. Coefficients of variation (CV) for the Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl survey indices for Gulf of Maine cod from 1963 to 2012. *Note: the spring survey did not begin until 1968, 2012 fall survey data not available at time of this report.

| Year | Abundance (numbers/tow) | | Biomass (kg/tow) | |
|------|-------------------------|-------|------------------|-------|
| | Spring | Fall | Spring | Fall |
| 1963 | | 0.250 | | 0.391 |
| 1964 | | 0.412 | | 0.496 |
| 1965 | | 0.274 | | 0.273 |
| 1966 | | 0.217 | | 0.227 |
| 1967 | | 0.223 | | 0.219 |
| 1968 | 0.127 | 0.181 | 0.153 | 0.198 |
| 1969 | 0.328 | 0.152 | 0.329 | 0.217 |
| 1970 | 0.214 | 0.318 | 0.237 | 0.248 |
| 1971 | 0.190 | 0.205 | 0.211 | 0.307 |
| 1972 | 0.208 | 0.535 | 0.233 | 0.199 |
| 1973 | 0.328 | 0.151 | 0.415 | 0.267 |
| 1974 | 0.188 | 0.260 | 0.199 | 0.201 |
| 1975 | 0.222 | 0.226 | 0.249 | 0.153 |
| 1976 | 0.181 | 0.197 | 0.166 | 0.214 |
| 1977 | 0.269 | 0.124 | 0.208 | 0.126 |
| 1978 | 0.191 | 0.188 | 0.207 | 0.151 |
| 1979 | 0.234 | 0.112 | 0.176 | 0.128 |
| 1980 | 0.171 | 0.261 | 0.182 | 0.153 |
| 1981 | 0.194 | 0.224 | 0.205 | 0.233 |
| 1982 | 0.219 | 0.636 | 0.223 | 0.670 |
| 1983 | 0.263 | 0.170 | 0.225 | 0.188 |
| 1984 | 0.443 | 0.220 | 0.324 | 0.334 |
| 1985 | 0.202 | 0.176 | 0.223 | 0.354 |
| 1986 | 0.314 | 0.230 | 0.197 | 0.228 |
| 1987 | 0.257 | 0.308 | 0.314 | 0.234 |
| 1988 | 0.211 | 0.349 | 0.281 | 0.232 |
| 1989 | 0.184 | 0.223 | 0.207 | 0.181 |
| 1990 | 0.249 | 0.190 | 0.280 | 0.204 |
| 1991 | 0.251 | 0.267 | 0.240 | 0.246 |
| 1992 | 0.317 | 0.213 | 0.374 | 0.243 |
| 1993 | 0.223 | 0.259 | 0.347 | 0.263 |
| 1994 | 0.223 | 0.309 | 0.216 | 0.292 |
| 1995 | 0.273 | 0.301 | 0.257 | 0.325 |
| 1996 | 0.240 | 0.254 | 0.275 | 0.249 |
| 1997 | 0.168 | 0.299 | 0.192 | 0.307 |
| 1998 | 0.344 | 0.346 | 0.324 | 0.287 |
| 1999 | 0.242 | 0.181 | 0.320 | 0.193 |
| 2000 | 0.221 | 0.306 | 0.155 | 0.332 |
| 2001 | 0.311 | 0.271 | 0.327 | 0.279 |
| 2002 | 0.203 | 0.578 | 0.215 | 0.686 |
| 2003 | 0.223 | 0.307 | 0.368 | 0.251 |
| 2004 | 0.256 | 0.327 | 0.293 | 0.214 |
| 2005 | 0.241 | 0.065 | 0.248 | 0.228 |
| 2006 | 0.203 | 0.301 | 0.249 | 0.188 |
| 2007 | 0.665 | 0.368 | 0.540 | 0.277 |
| 2008 | 0.716 | 0.389 | 0.609 | 0.285 |
| 2009 | 0.531 | 0.535 | 0.491 | 0.429 |
| 2010 | 0.243 | 0.233 | 0.264 | 0.304 |
| 2011 | 0.279 | 0.304 | 0.201 | 0.336 |
| 2012 | 0.187 | | 0.209 | |
| Avg | 0.265 | 0.274 | 0.270 | 0.270 |
| Min | 0.127 | 0.065 | 0.153 | 0.126 |
| Max | 0.716 | 0.636 | 0.609 | 0.686 |

Table A.57. Northeast Fisheries Science Center (NEFSC) spring survey abundance indices-at-age (numbers/tow) from 1970 to 2012 for Gulf of Maine Atlantic cod. Age data are not available prior to 1970.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 | Age17 |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1970 | 0.000 | 0.159 | 0.124 | 0.053 | 0.098 | 0.290 | 0.475 | 0.589 | 0.073 | 0.045 | 0.076 | 0.133 | 0.059 | 0.000 | 0.018 | 0.000 | 0.000 | |
| 1971 | 0.000 | 0.069 | 0.109 | 0.099 | 0.280 | 0.086 | 0.096 | 0.280 | 0.207 | 0.142 | 0.050 | 0.013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1972 | 0.053 | 0.300 | 0.153 | 0.499 | 0.208 | 0.205 | 0.052 | 0.083 | 0.119 | 0.300 | 0.027 | 0.017 | 0.026 | 0.000 | 0.017 | 0.000 | 0.000 | |
| 1973 | 0.000 | 0.053 | 4.273 | 0.917 | 0.614 | 0.384 | 0.144 | 0.106 | 0.186 | 0.276 | 0.186 | 0.072 | 0.113 | 0.112 | 0.088 | 0.000 | 0.000 | |
| 1974 | 0.164 | 0.311 | 0.081 | 1.534 | 0.177 | 0.231 | 0.082 | 0.000 | 0.064 | 0.038 | 0.089 | 0.043 | 0.037 | 0.000 | 0.016 | 0.000 | 0.035 | |
| 1975 | 0.012 | 0.094 | 0.707 | 0.095 | 1.139 | 0.246 | 0.073 | 0.000 | 0.006 | 0.025 | 0.028 | 0.026 | 0.062 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1976 | 0.000 | 0.052 | 0.253 | 1.114 | 0.150 | 0.870 | 0.131 | 0.056 | 0.038 | 0.000 | 0.036 | 0.000 | 0.054 | 0.027 | 0.000 | 0.000 | 0.000 | |
| 1977 | 0.000 | 0.068 | 0.264 | 0.460 | 2.015 | 0.139 | 0.775 | 0.000 | 0.114 | 0.000 | 0.000 | 0.000 | 0.000 | 0.031 | 0.000 | 0.000 | 0.006 | |
| 1978 | 0.000 | 0.070 | 0.083 | 0.297 | 0.383 | 0.764 | 0.084 | 0.226 | 0.013 | 0.108 | 0.000 | 0.022 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1979 | 0.044 | 0.426 | 1.407 | 0.186 | 0.470 | 0.301 | 0.549 | 0.094 | 0.104 | 0.013 | 0.031 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1980 | 0.070 | 0.037 | 0.500 | 0.436 | 0.123 | 0.294 | 0.226 | 0.337 | 0.000 | 0.105 | 0.026 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1981 | 0.000 | 1.091 | 0.619 | 0.850 | 1.335 | 0.318 | 0.304 | 0.080 | 0.144 | 0.091 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1982 | 0.014 | 0.357 | 1.040 | 0.498 | 0.737 | 0.848 | 0.083 | 0.135 | 0.000 | 0.040 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1983 | 0.013 | 0.610 | 0.968 | 1.042 | 0.453 | 0.336 | 0.250 | 0.060 | 0.000 | 0.071 | 0.033 | 0.017 | 0.045 | 0.000 | 0.016 | 0.000 | 0.000 | |
| 1984 | 0.000 | 0.151 | 1.309 | 0.987 | 0.853 | 0.229 | 0.047 | 0.090 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1985 | 0.000 | 0.029 | 0.238 | 0.676 | 0.612 | 0.707 | 0.094 | 0.109 | 0.026 | 0.026 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1986 | 0.000 | 0.537 | 0.259 | 0.767 | 0.218 | 0.075 | 0.046 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1987 | 0.000 | 0.030 | 0.471 | 0.191 | 0.222 | 0.075 | 0.000 | 0.068 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.015 | 0.000 | 0.000 | 0.000 | |
| 1988 | 0.029 | 0.719 | 0.926 | 0.791 | 0.283 | 0.205 | 0.099 | 0.036 | 0.020 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1989 | 0.000 | 0.025 | 0.609 | 0.712 | 0.630 | 0.069 | 0.068 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1990 | 0.000 | 0.009 | 0.233 | 1.325 | 0.669 | 0.076 | 0.032 | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1991 | 0.000 | 0.028 | 0.077 | 0.233 | 1.750 | 0.247 | 0.041 | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1992 | 0.000 | 0.050 | 0.247 | 0.223 | 0.248 | 1.368 | 0.213 | 0.073 | 0.000 | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1993 | 0.000 | 0.201 | 0.507 | 0.804 | 0.364 | 0.084 | 0.446 | 0.055 | 0.023 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1994 | 0.000 | 0.015 | 0.316 | 0.407 | 0.201 | 0.083 | 0.053 | 0.142 | 0.009 | 0.027 | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1995 | 0.000 | 0.037 | 0.187 | 1.165 | 0.321 | 0.147 | 0.034 | 0.000 | 0.011 | 0.000 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1996 | 0.000 | 0.057 | 0.022 | 0.586 | 1.355 | 0.385 | 0.060 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1997 | 0.000 | 0.159 | 0.139 | 0.390 | 0.271 | 0.874 | 0.244 | 0.115 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1998 | 0.000 | 0.018 | 0.228 | 0.359 | 0.513 | 0.143 | 0.408 | 0.021 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1999 | 0.000 | 0.166 | 0.342 | 0.726 | 0.351 | 0.305 | 0.134 | 0.266 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | |
| 2000 | 0.026 | 1.173 | 0.737 | 0.438 | 0.485 | 0.099 | 0.092 | 0.011 | 0.022 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2001 | 0.000 | 0.029 | 0.355 | 0.683 | 0.510 | 0.342 | 0.065 | 0.097 | 0.055 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2002 | 0.000 | 0.340 | 0.045 | 0.548 | 1.584 | 0.606 | 0.342 | 0.185 | 0.057 | 0.017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2003 | 0.000 | 0.075 | 0.825 | 0.059 | 0.718 | 1.072 | 0.387 | 0.340 | 0.081 | 0.082 | 0.030 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2004 | 0.000 | 0.136 | 0.045 | 0.230 | 0.116 | 0.208 | 0.213 | 0.011 | 0.011 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2005 | 0.000 | 0.029 | 0.739 | 0.081 | 0.623 | 0.011 | 0.138 | 0.128 | 0.015 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2006 | 0.028 | 0.184 | 0.237 | 0.434 | 0.049 | 0.197 | 0.023 | 0.126 | 0.069 | 0.000 | 0.015 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2007 | 0.000 | 0.100 | 3.422 | 3.077 | 4.446 | 0.437 | 0.796 | 0.075 | 0.041 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2008 | 0.000 | 0.079 | 1.165 | 3.930 | 1.582 | 1.099 | 0.053 | 0.082 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2009 | 0.000 | 0.063 | 0.279 | 1.050 | 1.135 | 0.600 | 0.438 | 0.008 | 0.022 | 0.000 | 0.004 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2010 | 0.000 | 0.059 | 0.279 | 0.335 | 0.197 | 0.229 | 0.113 | 0.043 | 0.016 | 0.010 | 0.005 | 0.000 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2011 | 0.000 | 0.005 | 0.024 | 0.140 | 0.383 | 0.189 | 0.086 | 0.033 | 0.035 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2012 | 0.000 | 0.069 | 0.105 | 0.224 | 0.243 | 0.159 | 0.051 | 0.036 | 0.004 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |

Table A.58. Northeast Fisheries Science Center (NEFSC) spring survey biomass indices-at-age (weight/tow) from 1970 to 2012 for Gulf of Maine Atlantic cod. Age data are not available prior to 1970. *Note, biomass indices are not used in the current assessment.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 | Age17 | | |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1970 | 0.000 | 0.007 | 0.037 | 0.034 | 0.154 | 0.715 | 2.274 | 3.140 | 0.626 | 0.390 | 0.605 | 1.840 | 0.950 | 0.000 | 0.318 | 0.000 | 0.000 | 0.000 | | |
| 1971 | 0.000 | 0.014 | 0.055 | 0.133 | 0.623 | 0.384 | 0.343 | 1.786 | 1.767 | 1.073 | 0.656 | 0.170 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | | |
| 1972 | 0.000 | 0.014 | 0.054 | 0.827 | 0.522 | 0.738 | 0.284 | 0.516 | 0.914 | 3.161 | 0.256 | 0.208 | 0.268 | 0.000 | 0.270 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1973 | 0.000 | 0.002 | 0.769 | 0.892 | 1.780 | 1.434 | 0.652 | 0.765 | 1.156 | 2.874 | 2.127 | 0.914 | 1.627 | 1.837 | 1.979 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1974 | 0.002 | 0.011 | 0.015 | 1.056 | 0.478 | 1.310 | 0.655 | 0.000 | 0.470 | 0.176 | 1.213 | 0.402 | 0.527 | 0.000 | 0.289 | 0.000 | 0.815 | 0.000 | 0.000 | |
| 1975 | 0.000 | 0.003 | 0.180 | 0.098 | 2.161 | 0.954 | 0.512 | 0.000 | 0.052 | 0.250 | 0.566 | 0.166 | 1.097 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1976 | 0.000 | 0.005 | 0.061 | 0.794 | 0.253 | 2.727 | 0.728 | 0.608 | 0.438 | 0.000 | 0.451 | 0.000 | 0.958 | 0.532 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1977 | 0.000 | 0.008 | 0.086 | 0.359 | 2.132 | 0.321 | 3.710 | 0.000 | 1.134 | 0.000 | 0.000 | 0.000 | 0.000 | 0.666 | 0.000 | 0.000 | 0.126 | 0.000 | 0.000 | |
| 1978 | 0.000 | 0.009 | 0.039 | 0.338 | 0.695 | 2.398 | 0.480 | 1.738 | 0.134 | 1.613 | 0.000 | 0.253 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1979 | 0.000 | 0.033 | 0.568 | 0.254 | 0.926 | 0.918 | 2.248 | 0.721 | 0.741 | 0.184 | 0.464 | 0.498 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1980 | 0.000 | 0.002 | 0.175 | 0.563 | 0.263 | 1.019 | 0.875 | 1.880 | 0.000 | 1.072 | 0.383 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1981 | 0.000 | 0.137 | 0.285 | 0.937 | 3.306 | 1.289 | 1.869 | 0.605 | 1.220 | 1.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1982 | 0.000 | 0.038 | 0.456 | 0.672 | 1.901 | 3.511 | 0.339 | 1.085 | 0.000 | 0.439 | 0.176 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1983 | 0.000 | 0.057 | 0.448 | 1.536 | 1.138 | 1.718 | 1.672 | 0.682 | 0.000 | 1.134 | 0.526 | 0.306 | 1.283 | 0.000 | 0.462 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1984 | 0.000 | 0.011 | 0.752 | 1.412 | 2.176 | 1.133 | 0.204 | 0.455 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1985 | 0.000 | 0.001 | 0.101 | 0.898 | 1.658 | 3.035 | 0.518 | 0.663 | 0.342 | 0.429 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.000 | 0.046 | 0.125 | 1.199 | 0.644 | 0.268 | 0.358 | 0.474 | 0.000 | 0.000 | 0.000 | 0.000 | 0.362 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1987 | 0.000 | 0.002 | 0.164 | 0.139 | 0.574 | 0.230 | 0.000 | 0.432 | 0.061 | 0.000 | 0.000 | 0.000 | 0.000 | 0.373 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1988 | 0.000 | 0.036 | 0.162 | 0.821 | 0.489 | 1.035 | 0.548 | 0.177 | 0.191 | 0.145 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.000 | 0.001 | 0.111 | 0.518 | 1.151 | 0.182 | 0.461 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.000 | 0.001 | 0.057 | 1.042 | 1.357 | 0.263 | 0.210 | 0.147 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.000 | 0.002 | 0.015 | 0.204 | 2.083 | 0.376 | 0.104 | 0.108 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.000 | 0.003 | 0.112 | 0.225 | 0.713 | 5.715 | 1.204 | 0.494 | 0.000 | 0.161 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.000 | 0.012 | 0.164 | 1.100 | 0.714 | 0.321 | 2.341 | 0.589 | 0.258 | 0.000 | 0.377 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.000 | 0.001 | 0.061 | 0.348 | 0.467 | 0.210 | 0.150 | 0.804 | 0.060 | 0.098 | 0.229 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.000 | 0.004 | 0.045 | 0.794 | 0.411 | 0.415 | 0.135 | 0.000 | 0.032 | 0.000 | 0.597 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.000 | 0.004 | 0.007 | 1.054 | 2.802 | 1.269 | 0.291 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.000 | 0.010 | 0.062 | 0.553 | 0.719 | 2.581 | 0.914 | 0.777 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.000 | 0.001 | 0.102 | 0.427 | 1.043 | 0.461 | 1.849 | 0.136 | 0.161 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.000 | 0.015 | 0.115 | 0.722 | 0.683 | 0.953 | 0.768 | 1.482 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.353 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.000 | 0.093 | 0.322 | 0.454 | 1.204 | 0.409 | 0.489 | 0.052 | 0.190 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.000 | 0.003 | 0.168 | 0.756 | 1.395 | 1.452 | 0.581 | 0.876 | 0.793 | 0.000 | 0.190 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.000 | 0.024 | 0.014 | 0.642 | 4.305 | 1.963 | 2.061 | 1.113 | 0.753 | 0.060 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.000 | 0.009 | 0.163 | 0.048 | 1.141 | 2.852 | 1.544 | 1.964 | 0.535 | 0.920 | 0.282 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.000 | 0.006 | 0.016 | 0.196 | 0.294 | 0.763 | 0.936 | 0.043 | 0.043 | 0.117 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.000 | 0.001 | 0.156 | 0.084 | 1.084 | 0.030 | 0.549 | 0.716 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.000 | 0.013 | 0.062 | 0.343 | 0.091 | 0.611 | 0.142 | 0.686 | 0.602 | 0.000 | 0.153 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.000 | 0.009 | 1.329 | 2.694 | 7.333 | 1.337 | 2.581 | 0.308 | 0.221 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.000 | 0.004 | 0.466 | 4.137 | 2.619 | 2.734 | 0.299 | 0.565 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.000 | 0.002 | 0.146 | 1.513 | 2.346 | 1.560 | 1.260 | 0.065 | 0.222 | 0.000 | 0.046 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.000 | 0.005 | 0.099 | 0.403 | 0.551 | 0.881 | 0.522 | 0.318 | 0.171 | 0.103 | 0.113 | 0.000 | 0.171 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.000 | 0.000 | 0.011 | 0.164 | 0.657 | 0.510 | 0.302 | 0.193 | 0.295 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.000 | 0.006 | 0.054 | 0.291 | 0.500 | 0.391 | 0.166 | 0.180 | 0.041 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table A.59. Northeast Fisheries Science Center (NEFSC) fall survey abundance indices-at-age (numbers/tow) from 1970 to 2011 for Gulf of Maine Atlantic cod. Age data are not available prior to 1970.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 | Age17 |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1970 | 0.743 | 0.938 | 0.254 | 0.520 | 0.336 | 0.487 | 0.424 | 0.836 | 0.130 | 0.090 | 0.037 | 0.037 | 0.073 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1971 | 1.334 | 0.207 | 0.224 | 0.190 | 0.607 | 0.444 | 0.509 | 0.222 | 0.280 | 0.193 | 0.031 | 0.040 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1972 | 0.031 | 5.663 | 1.118 | 1.595 | 0.181 | 0.072 | 0.122 | 0.031 | 0.121 | 0.351 | 0.000 | 0.000 | 0.000 | 0.016 | 0.000 | 0.000 | 0.000 | |
| 1973 | 0.638 | 0.327 | 2.146 | 0.179 | 0.540 | 0.191 | 0.055 | 0.018 | 0.039 | 0.182 | 0.122 | 0.000 | 0.000 | 0.016 | 0.000 | 0.000 | 0.000 | |
| 1974 | 0.265 | 1.131 | 0.267 | 1.922 | 0.125 | 0.276 | 0.000 | 0.052 | 0.036 | 0.066 | 0.000 | 0.120 | 0.000 | 0.000 | 0.069 | 0.000 | 0.000 | |
| 1975 | 0.006 | 0.223 | 3.028 | 0.139 | 2.354 | 0.250 | 0.105 | 0.020 | 0.000 | 0.000 | 0.000 | 0.006 | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1976 | 0.000 | 0.209 | 0.216 | 0.578 | 0.104 | 0.835 | 0.044 | 0.099 | 0.000 | 0.000 | 0.063 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1977 | 0.000 | 0.046 | 0.446 | 0.456 | 1.151 | 0.133 | 0.604 | 0.024 | 0.083 | 0.021 | 0.061 | 0.000 | 0.022 | 0.026 | 0.000 | 0.000 | 0.000 | |
| 1978 | 0.241 | 1.411 | 0.359 | 1.141 | 0.661 | 1.450 | 0.101 | 0.269 | 0.012 | 0.082 | 0.000 | 0.019 | 0.000 | 0.028 | 0.000 | 0.000 | 0.000 | |
| 1979 | 0.000 | 0.364 | 0.617 | 0.131 | 0.696 | 0.319 | 0.754 | 0.056 | 0.135 | 0.000 | 0.053 | 0.000 | 0.000 | 0.005 | 0.013 | 0.000 | 0.000 | |
| 1980 | 0.027 | 1.319 | 2.558 | 1.664 | 0.518 | 0.236 | 0.402 | 0.192 | 0.022 | 0.012 | 0.000 | 0.049 | 0.000 | 0.014 | 0.000 | 0.000 | 0.022 | |
| 1981 | 0.010 | 0.581 | 0.399 | 0.469 | 0.509 | 0.092 | 0.081 | 0.081 | 0.099 | 0.000 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1982 | 0.000 | 0.835 | 3.264 | 2.476 | 0.971 | 0.222 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1983 | 0.000 | 0.305 | 0.905 | 0.757 | 0.267 | 0.250 | 0.219 | 0.000 | 0.000 | 0.000 | 0.018 | 0.028 | 0.037 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1984 | 0.000 | 0.513 | 0.418 | 0.586 | 0.384 | 0.196 | 0.194 | 0.062 | 0.000 | 0.016 | 0.000 | 0.000 | 0.045 | 0.035 | 0.000 | 0.000 | 0.000 | |
| 1985 | 0.218 | 0.445 | 0.917 | 0.627 | 0.201 | 0.246 | 0.064 | 0.000 | 0.034 | 0.070 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1986 | 0.000 | 0.394 | 0.404 | 0.626 | 0.368 | 0.073 | 0.041 | 0.000 | 0.000 | 0.045 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1987 | 0.128 | 0.570 | 1.388 | 0.586 | 0.198 | 0.125 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1988 | 0.000 | 1.889 | 2.366 | 1.069 | 0.367 | 0.146 | 0.000 | 0.044 | 0.000 | 0.011 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1989 | 0.000 | 0.145 | 2.468 | 1.458 | 0.283 | 0.138 | 0.053 | 0.000 | 0.009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1990 | 0.000 | 0.057 | 0.218 | 1.788 | 0.611 | 0.255 | 0.048 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1991 | 0.009 | 0.144 | 0.151 | 0.230 | 0.621 | 0.075 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1992 | 0.059 | 0.289 | 0.448 | 0.144 | 0.041 | 0.327 | 0.126 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1993 | 0.031 | 0.210 | 0.575 | 0.361 | 0.017 | 0.000 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1994 | 0.032 | 0.184 | 0.909 | 0.816 | 0.093 | 0.051 | 0.000 | 0.045 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1995 | 0.008 | 0.068 | 0.308 | 1.226 | 0.304 | 0.082 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1996 | 0.029 | 0.122 | 0.379 | 0.231 | 0.516 | 0.050 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1997 | 0.000 | 0.297 | 0.091 | 0.165 | 0.168 | 0.151 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1998 | 0.050 | 0.085 | 0.342 | 0.110 | 0.185 | 0.041 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1999 | 0.025 | 0.432 | 0.375 | 0.590 | 0.244 | 0.122 | 0.019 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2000 | 0.008 | 0.540 | 0.981 | 0.399 | 0.492 | 0.140 | 0.010 | 0.000 | 0.034 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2001 | 0.018 | 0.000 | 0.171 | 0.720 | 0.478 | 0.356 | 0.124 | 0.092 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2002 | 0.000 | 0.269 | 0.104 | 0.333 | 2.683 | 1.070 | 0.750 | 0.077 | 0.043 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2003 | 0.542 | 0.461 | 0.186 | 0.216 | 0.518 | 0.451 | 0.071 | 0.062 | 0.000 | 0.011 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2004 | 1.369 | 0.661 | 0.172 | 0.577 | 0.254 | 0.250 | 0.149 | 0.057 | 0.023 | 0.010 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2005 | 0.034 | 0.153 | 0.378 | 0.078 | 0.456 | 0.023 | 0.090 | 0.082 | 0.023 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2006 | 0.064 | 1.241 | 0.599 | 1.007 | 0.252 | 0.293 | 0.037 | 0.053 | 0.036 | 0.000 | 0.000 | 0.014 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2007 | 0.011 | 0.136 | 0.863 | 0.395 | 0.496 | 0.023 | 0.067 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2008 | 0.165 | 0.650 | 1.227 | 1.060 | 0.189 | 0.139 | 0.000 | 0.000 | 0.000 | 0.010 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2009 | 0.020 | 0.660 | 2.096 | 0.314 | 0.277 | 0.045 | 0.035 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2010 | 0.008 | 0.094 | 0.132 | 0.290 | 0.288 | 0.092 | 0.023 | 0.013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.006 | 0.000 | 0.000 | 0.000 | |
| 2011 | 0.036 | 0.060 | 0.091 | 0.210 | 0.304 | 0.175 | 0.078 | 0.005 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |

Table A.60. Northeast Fisheries Science Center (NEFSC) fall survey biomass indices-at-age (weight/tow) from 1970 to 2011 for Gulf of Maine Atlantic cod. Age data are not available prior to 1970. *Note, biomass indices are not used in the current assessment.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16 | Age17 |
|------|-------|-------|-------|-------|--------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1970 | 0.005 | 0.187 | 0.152 | 0.732 | 1.291 | 1.467 | 2.626 | 5.792 | 1.125 | 0.780 | 0.493 | 0.443 | 1.349 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1971 | 0.333 | 0.050 | 0.269 | 0.321 | 1.769 | 2.138 | 2.743 | 1.519 | 2.520 | 2.357 | 0.644 | 0.337 | 1.531 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1972 | 0.000 | 0.769 | 0.832 | 3.572 | 0.647 | 0.264 | 0.813 | 0.208 | 1.480 | 4.078 | 0.000 | 0.000 | 0.000 | 0.000 | 0.323 | 0.000 | 0.000 | |
| 1973 | 0.006 | 0.036 | 0.984 | 0.374 | 2.282 | 0.919 | 0.322 | 0.178 | 0.235 | 2.076 | 1.128 | 0.000 | 0.000 | 0.224 | 0.000 | 0.000 | 0.000 | |
| 1974 | 0.000 | 0.086 | 0.133 | 2.515 | 0.344 | 1.778 | 0.000 | 0.419 | 0.456 | 0.814 | 0.000 | 1.602 | 0.000 | 0.000 | 0.812 | 0.000 | 0.000 | |
| 1975 | 0.000 | 0.056 | 1.328 | 0.144 | 5.392 | 0.695 | 0.587 | 0.169 | 0.000 | 0.000 | 0.000 | 0.095 | 0.154 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1976 | 0.000 | 0.073 | 0.182 | 0.678 | 0.154 | 3.230 | 0.328 | 0.963 | 0.000 | 0.000 | 1.133 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1977 | 0.000 | 0.009 | 0.237 | 0.565 | 2.121 | 0.506 | 3.589 | 0.188 | 0.929 | 0.298 | 1.031 | 0.000 | 0.300 | 0.427 | 0.000 | 0.000 | 0.000 | |
| 1978 | 0.004 | 0.285 | 0.264 | 1.559 | 1.500 | 4.493 | 0.408 | 2.047 | 0.143 | 1.260 | 0.000 | 0.361 | 0.000 | 0.577 | 0.000 | 0.000 | 0.000 | |
| 1979 | 0.000 | 0.140 | 0.542 | 0.347 | 2.328 | 1.744 | 5.123 | 0.573 | 1.607 | 0.000 | 1.042 | 0.000 | 0.000 | 0.000 | 0.122 | 0.360 | 0.000 | |
| 1980 | 0.001 | 0.427 | 1.836 | 3.159 | 1.589 | 1.580 | 2.409 | 1.228 | 0.338 | 0.216 | 0.000 | 0.702 | 0.000 | 0.291 | 0.000 | 0.000 | 0.427 | |
| 1981 | 0.000 | 0.135 | 0.440 | 0.993 | 2.249 | 0.516 | 0.656 | 0.676 | 1.225 | 0.000 | 0.644 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1982 | 0.000 | 0.412 | 4.594 | 6.161 | 3.224 | 1.528 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1983 | 0.000 | 0.072 | 0.979 | 1.310 | 0.957 | 1.222 | 2.154 | 0.000 | 0.000 | 0.000 | 0.266 | 0.648 | 0.809 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1984 | 0.000 | 0.147 | 0.422 | 1.345 | 1.419 | 1.287 | 1.465 | 0.705 | 0.000 | 0.302 | 0.000 | 0.000 | 0.907 | 0.737 | 0.000 | 0.000 | 0.000 | |
| 1985 | 0.003 | 0.093 | 0.967 | 1.568 | 0.780 | 1.842 | 0.663 | 0.000 | 0.691 | 1.657 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1986 | 0.000 | 0.137 | 0.284 | 1.563 | 1.229 | 0.576 | 0.332 | 0.000 | 0.000 | 0.595 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1987 | 0.001 | 0.086 | 0.900 | 0.881 | 0.714 | 0.813 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1988 | 0.000 | 0.331 | 1.586 | 1.982 | 1.172 | 0.877 | 0.000 | 0.388 | 0.000 | 0.117 | 0.165 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1989 | 0.000 | 0.040 | 1.011 | 1.715 | 0.771 | 0.676 | 0.204 | 0.000 | 0.119 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1990 | 0.000 | 0.013 | 0.094 | 1.718 | 1.565 | 1.234 | 0.238 | 0.052 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1991 | 0.000 | 0.025 | 0.108 | 0.392 | 1.592 | 0.404 | 0.000 | 0.260 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1992 | 0.001 | 0.062 | 0.400 | 0.178 | 0.109 | 1.100 | 0.599 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1993 | 0.001 | 0.026 | 0.295 | 0.553 | 0.061 | 0.000 | 0.068 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1994 | 0.001 | 0.053 | 0.482 | 1.226 | 0.324 | 0.331 | 0.000 | 0.320 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1995 | 0.001 | 0.009 | 0.270 | 1.958 | 0.793 | 0.586 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1996 | 0.001 | 0.035 | 0.274 | 0.508 | 1.246 | 0.289 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1997 | 0.000 | 0.045 | 0.082 | 0.291 | 0.772 | 0.681 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1998 | 0.001 | 0.016 | 0.258 | 0.206 | 0.607 | 0.185 | 0.227 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 1999 | 0.001 | 0.131 | 0.380 | 1.239 | 0.941 | 0.671 | 0.143 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2000 | 0.000 | 0.119 | 0.849 | 0.774 | 1.821 | 0.497 | 0.098 | 0.000 | 0.495 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2001 | 0.001 | 0.000 | 0.129 | 1.310 | 1.301 | 2.228 | 1.124 | 0.981 | 0.000 | 0.250 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2002 | 0.000 | 0.038 | 0.101 | 0.730 | 10.977 | 5.659 | 5.789 | 0.642 | 0.723 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2003 | 0.014 | 0.172 | 0.122 | 0.497 | 1.402 | 2.361 | 0.443 | 0.538 | 0.000 | 0.174 | 0.000 | 0.266 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2004 | 0.002 | 0.083 | 0.108 | 0.978 | 0.878 | 1.125 | 0.666 | 0.485 | 0.192 | 0.191 | 0.198 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2005 | 0.001 | 0.017 | 0.171 | 0.124 | 0.985 | 0.134 | 0.312 | 0.542 | 0.231 | 0.380 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2006 | 0.001 | 0.257 | 0.288 | 1.031 | 0.432 | 1.021 | 0.221 | 0.269 | 0.510 | 0.000 | 0.000 | 0.199 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2007 | 0.001 | 0.023 | 0.455 | 0.402 | 1.310 | 0.098 | 0.426 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2008 | 0.014 | 0.206 | 1.246 | 2.105 | 0.469 | 0.754 | 0.000 | 0.000 | 0.000 | 0.134 | 0.379 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2009 | 0.001 | 0.366 | 2.461 | 1.057 | 1.246 | 0.480 | 0.234 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |
| 2010 | 0.000 | 0.032 | 0.155 | 0.515 | 1.125 | 0.440 | 0.106 | 0.046 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.154 | 0.000 | 0.000 | |
| 2011 | 0.001 | 0.017 | 0.086 | 0.372 | 0.706 | 0.802 | 0.385 | 0.054 | 0.224 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | |

Table A.61. Massachusetts Department of Marine Fisheries (MADMF) spring and fall survey indices from 1978 to 2012 for Gulf of Maine Atlantic cod. *Note: 2012 fall survey data not available at time of this report.

| Year | Abundance (numbers/tow) | | Biomass (kg/tow) | |
|------|-------------------------|---------|------------------|-------|
| | Spring | Fall | Spring | Fall |
| 1978 | 47.887 | 156.060 | 11.058 | 1.515 |
| 1979 | 96.559 | 8.924 | 14.276 | 1.052 |
| 1980 | 65.979 | 12.531 | 14.509 | 1.286 |
| 1981 | 69.406 | 9.291 | 18.689 | 3.638 |
| 1982 | 25.842 | 6.125 | 12.161 | 0.659 |
| 1983 | 54.850 | 1.676 | 18.746 | 0.092 |
| 1984 | 10.330 | 10.548 | 7.240 | 0.133 |
| 1985 | 8.455 | 2.871 | 4.765 | 0.070 |
| 1986 | 24.089 | 2.750 | 7.841 | 0.249 |
| 1987 | 17.206 | 313.148 | 7.865 | 0.348 |
| 1988 | 22.242 | 8.872 | 7.703 | 0.366 |
| 1989 | 52.244 | 4.150 | 17.346 | 0.218 |
| 1990 | 32.409 | 12.708 | 15.879 | 0.758 |
| 1991 | 13.699 | 7.483 | 8.730 | 0.480 |
| 1992 | 16.924 | 27.496 | 8.766 | 0.272 |
| 1993 | 92.659 | 51.500 | 5.861 | 1.353 |
| 1994 | 16.358 | 49.019 | 4.334 | 1.998 |
| 1995 | 23.364 | 4.678 | 3.993 | 0.807 |
| 1996 | 12.961 | 7.007 | 3.152 | 0.083 |
| 1997 | 17.887 | 1.456 | 2.500 | 0.014 |
| 1998 | 27.570 | 4.335 | 3.250 | 0.360 |
| 1999 | 161.058 | 8.005 | 8.997 | 0.308 |
| 2000 | 50.771 | 0.679 | 20.604 | 0.272 |
| 2001 | 41.844 | 49.555 | 26.445 | 0.757 |
| 2002 | 24.338 | 3.299 | 11.158 | 3.995 |
| 2003 | 1120.371 | 122.284 | 10.984 | 1.850 |
| 2004 | 131.589 | 57.620 | 8.147 | 5.580 |
| 2005 | 193.262 | 40.350 | 10.402 | 0.207 |
| 2006 | 1077.030 | 7.505 | 9.177 | 1.939 |
| 2007 | 61.576 | 7.918 | 8.430 | 0.077 |
| 2008 | 482.100 | 7.549 | 12.229 | 2.379 |
| 2009 | 480.516 | 5.042 | 4.489 | 0.807 |
| 2010 | 8.075 | 2.022 | 5.645 | 1.400 |
| 2011 | 59.064 | 2.610 | 4.519 | 1.355 |
| 2012 | 11.465 | | 2.276 | |
| Avg | 132.914 | 29.914 | 9.776 | 1.079 |
| Min | 8.075 | 0.679 | 2.276 | 0.014 |
| Max | 1120.371 | 313.148 | 26.445 | 5.580 |

Table A.62. Coefficients of variation (CV) for the Massachusetts Department of Marine Fisheries (MADMF) spring and fall bottom trawl survey indices of Gulf of Maine Atlantic cod between 1978 and 2012. *Note: 2012 fall survey data not available at time of this report.

| Year | Abundance (numbers/tow) | | Biomass (kg/tow) | |
|------|-------------------------|-------|------------------|-------|
| | Spring | Fall | Spring | Fall |
| 1978 | 0.147 | 0.322 | 0.138 | 0.555 |
| 1979 | 0.278 | 0.260 | 0.219 | 0.377 |
| 1980 | 0.124 | 0.266 | 0.128 | 0.345 |
| 1981 | 0.207 | 0.422 | 0.265 | 0.453 |
| 1982 | 0.221 | 0.321 | 0.175 | 0.690 |
| 1983 | 0.166 | 0.338 | 0.153 | 0.569 |
| 1984 | 0.289 | 0.189 | 0.259 | 0.444 |
| 1985 | 0.206 | 0.308 | 0.194 | 0.396 |
| 1986 | 0.552 | 0.304 | 0.354 | 0.864 |
| 1987 | 0.221 | 0.173 | 0.271 | 0.186 |
| 1988 | 0.206 | 0.240 | 0.237 | 0.436 |
| 1989 | 0.268 | 0.064 | 0.342 | 0.456 |
| 1990 | 0.288 | 0.262 | 0.341 | 0.413 |
| 1991 | 0.219 | 0.263 | 0.122 | 0.543 |
| 1992 | 0.287 | 0.076 | 0.321 | 0.340 |
| 1993 | 0.340 | 0.245 | 0.270 | 0.237 |
| 1994 | 0.227 | 0.513 | 0.241 | 0.787 |
| 1995 | 0.262 | 0.316 | 0.225 | 0.690 |
| 1996 | 0.218 | 0.365 | 0.305 | 0.426 |
| 1997 | 0.240 | 0.243 | 0.250 | 0.456 |
| 1998 | 0.261 | 0.260 | 0.468 | 0.486 |
| 1999 | 0.369 | 0.552 | 0.261 | 0.452 |
| 2000 | 0.391 | 0.379 | 0.459 | 0.387 |
| 2001 | 0.435 | 0.474 | 0.536 | 0.545 |
| 2002 | 0.096 | 0.596 | 0.390 | 0.812 |
| 2003 | 0.507 | 0.478 | 0.219 | 0.466 |
| 2004 | 0.459 | 0.299 | 0.278 | 0.399 |
| 2005 | 0.223 | 0.415 | 0.197 | 0.412 |
| 2006 | 0.337 | 0.398 | 0.181 | 0.460 |
| 2007 | 0.274 | 0.275 | 0.251 | 0.665 |
| 2008 | 0.204 | 0.417 | 0.215 | 0.443 |
| 2009 | 0.352 | 0.416 | 0.187 | 0.431 |
| 2010 | 0.234 | 0.449 | 0.456 | 0.471 |
| 2011 | 0.534 | 0.328 | 0.424 | 0.246 |
| 2012 | 0.274 | | 0.401 | |
| Avg | 0.283 | 0.330 | 0.278 | 0.481 |
| Min | 0.096 | 0.064 | 0.122 | 0.186 |
| Max | 0.552 | 0.596 | 0.536 | 0.864 |

Table A.63. Massachusetts Department of Marine Fisheries (MADMF) spring survey abundance indices-at-age (numbers/tow) from 1982 to 2012 for Gulf of Maine Atlantic cod. Age data are not available prior to 1982.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 |
|------|----------|--------|--------|--------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1982 | 1.668 | 13.218 | 6.649 | 2.921 | 1.024 | 0.216 | 0.049 | 0.046 | 0.050 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.718 | 30.253 | 17.570 | 4.710 | 0.347 | 1.121 | 0.075 | 0.023 | 0.033 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.257 | 1.898 | 5.090 | 2.101 | 0.751 | 0.147 | 0.086 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 1.569 | 1.670 | 2.695 | 2.024 | 0.498 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 1.075 | 18.031 | 3.376 | 0.903 | 0.582 | 0.100 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.725 | 8.622 | 5.376 | 2.045 | 0.168 | 0.147 | 0.053 | 0.000 | 0.000 | 0.070 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 1.895 | 10.409 | 6.750 | 1.927 | 1.211 | 0.016 | 0.033 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.298 | 21.463 | 22.947 | 6.868 | 0.513 | 0.108 | 0.048 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 4.930 | 4.972 | 5.938 | 14.182 | 2.149 | 0.155 | 0.083 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.355 | 5.331 | 2.295 | 1.801 | 3.669 | 0.249 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 1.506 | 4.379 | 5.699 | 3.444 | 0.484 | 1.301 | 0.066 | 0.044 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 80.090 | 2.842 | 6.100 | 2.509 | 0.879 | 0.166 | 0.074 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 4.627 | 5.406 | 3.883 | 1.703 | 0.608 | 0.131 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 11.998 | 5.985 | 2.420 | 2.408 | 0.525 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 8.843 | 0.777 | 0.497 | 0.955 | 1.590 | 0.299 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 12.431 | 2.910 | 1.035 | 0.920 | 0.190 | 0.383 | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 23.481 | 1.487 | 0.924 | 0.779 | 0.637 | 0.034 | 0.211 | 0.017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 143.000 | 11.832 | 2.407 | 2.275 | 0.735 | 0.630 | 0.036 | 0.127 | 0.017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 2.151 | 35.360 | 6.995 | 2.371 | 2.316 | 0.784 | 0.663 | 0.059 | 0.073 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 25.987 | 0.084 | 4.998 | 4.710 | 3.448 | 1.961 | 0.323 | 0.227 | 0.106 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.924 | 19.340 | 0.220 | 1.379 | 1.145 | 0.561 | 0.318 | 0.111 | 0.253 | 0.025 | 0.049 | 0.000 | 0.012 | 0.000 | 0.000 |
| 2003 | 1094.105 | 17.109 | 5.496 | 0.439 | 1.938 | 0.937 | 0.221 | 0.074 | 0.014 | 0.025 | 0.000 | 0.014 | 0.000 | 0.000 | 0.000 |
| 2004 | 116.135 | 8.927 | 1.882 | 2.627 | 0.361 | 1.083 | 0.455 | 0.076 | 0.029 | 0.000 | 0.014 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 179.479 | 5.524 | 4.141 | 0.795 | 1.955 | 0.263 | 0.663 | 0.243 | 0.094 | 0.105 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 1053.701 | 9.992 | 7.139 | 3.930 | 0.525 | 1.532 | 0.109 | 0.057 | 0.000 | 0.017 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 49.323 | 3.776 | 3.078 | 2.303 | 2.163 | 0.343 | 0.519 | 0.025 | 0.046 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 456.954 | 7.275 | 10.336 | 3.242 | 2.287 | 1.695 | 0.155 | 0.155 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 466.098 | 8.907 | 2.350 | 1.654 | 1.045 | 0.348 | 0.112 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 1.165 | 2.415 | 1.393 | 1.423 | 0.819 | 0.678 | 0.129 | 0.000 | 0.000 | 0.000 | 0.052 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 55.378 | 0.326 | 1.001 | 0.621 | 0.933 | 0.558 | 0.139 | 0.086 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 6.239 | 3.368 | 0.671 | 0.446 | 0.304 | 0.415 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table A.64. Massachusetts Department of Marine Fisheries (MADMF) spring survey biomass indices-at-age (weight/tow) from 1981 to 2012 for Gulf of Maine Atlantic cod. Age data are not available prior to 1982. *Note: biomass indices are not used in the current assessment.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1982 | 0.001 | 1.539 | 3.012 | 3.230 | 2.081 | 1.212 | 0.248 | 0.315 | 0.523 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.001 | 2.497 | 6.811 | 4.805 | 0.567 | 2.669 | 0.787 | 0.105 | 0.506 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.000 | 0.197 | 2.112 | 2.722 | 1.414 | 0.546 | 0.249 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.016 | 0.213 | 1.393 | 2.022 | 1.120 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.001 | 3.062 | 1.528 | 1.437 | 1.320 | 0.364 | 0.129 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.001 | 0.491 | 3.030 | 1.618 | 0.539 | 0.583 | 0.535 | 0.000 | 0.000 | 1.069 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.002 | 0.311 | 2.263 | 2.343 | 2.472 | 0.101 | 0.210 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.001 | 1.543 | 7.795 | 6.497 | 0.851 | 0.402 | 0.257 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.005 | 0.262 | 2.430 | 9.278 | 2.831 | 0.513 | 0.560 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.000 | 0.607 | 0.759 | 2.013 | 4.702 | 0.648 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.001 | 0.215 | 2.545 | 2.594 | 0.683 | 2.232 | 0.363 | 0.133 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.090 | 0.104 | 2.162 | 1.918 | 0.908 | 0.470 | 0.208 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.001 | 0.425 | 1.083 | 1.434 | 1.024 | 0.366 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.010 | 0.288 | 0.955 | 1.948 | 0.722 | 0.071 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.003 | 0.063 | 0.212 | 0.770 | 1.607 | 0.498 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.008 | 0.212 | 0.575 | 0.851 | 0.324 | 0.509 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.017 | 0.093 | 0.360 | 0.846 | 1.119 | 0.086 | 0.688 | 0.042 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.073 | 1.114 | 1.166 | 2.580 | 1.521 | 1.831 | 0.123 | 0.524 | 0.066 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.002 | 3.323 | 3.263 | 3.238 | 4.704 | 2.196 | 2.893 | 0.326 | 0.659 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.018 | 0.004 | 2.350 | 7.397 | 8.089 | 5.370 | 1.655 | 0.832 | 0.731 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.001 | 0.750 | 0.051 | 1.303 | 2.230 | 1.690 | 1.650 | 0.660 | 1.879 | 0.184 | 0.511 | 0.000 | 0.250 | 0.000 | 0.000 |
| 2003 | 0.342 | 1.137 | 1.190 | 0.213 | 3.650 | 2.904 | 0.718 | 0.402 | 0.094 | 0.187 | 0.000 | 0.148 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.050 | 0.345 | 0.720 | 2.127 | 0.635 | 2.321 | 1.241 | 0.288 | 0.238 | 0.000 | 0.182 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.081 | 0.192 | 0.734 | 0.804 | 3.244 | 0.821 | 2.195 | 1.270 | 0.554 | 0.507 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.997 | 0.484 | 0.824 | 2.232 | 0.596 | 3.138 | 0.210 | 0.271 | 0.000 | 0.149 | 0.275 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.026 | 0.212 | 0.530 | 1.553 | 3.057 | 0.795 | 2.002 | 0.096 | 0.159 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.283 | 0.468 | 2.859 | 2.421 | 3.146 | 1.716 | 0.531 | 0.805 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.221 | 0.429 | 0.468 | 1.443 | 1.091 | 0.472 | 0.365 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.000 | 0.144 | 0.320 | 0.921 | 1.339 | 1.684 | 0.689 | 0.000 | 0.000 | 0.000 | 0.548 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.021 | 0.015 | 0.291 | 0.540 | 1.361 | 1.393 | 0.442 | 0.309 | 0.147 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.003 | 0.299 | 0.341 | 0.461 | 0.396 | 0.745 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table A.65. Massachusetts Department of Marine Fisheries (MADMF) fall survey abundance indices-at-age (numbers/tow) from 1981 to 2011 for Gulf of Maine Atlantic cod. Age data are not available prior to 1982. *Note: this survey index is not used in the current assessment.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 |
|------|---------|--------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1982 | 4.571 | 1.023 | 0.476 | 0.004 | 0.026 | 0.026 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 1.339 | 0.257 | 0.021 | 0.030 | 0.030 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 10.286 | 0.148 | 0.081 | 0.017 | 0.000 | 0.016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 2.536 | 0.301 | 0.010 | 0.010 | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 1.883 | 0.464 | 0.375 | 0.000 | 0.029 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 312.047 | 1.075 | 0.000 | 0.019 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 5.490 | 3.136 | 0.225 | 0.022 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 3.940 | 0.038 | 0.114 | 0.030 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 7.735 | 4.233 | 0.525 | 0.150 | 0.038 | 0.028 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 5.043 | 1.950 | 0.398 | 0.013 | 0.066 | 0.013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 26.408 | 0.980 | 0.071 | 0.000 | 0.000 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 49.188 | 1.735 | 0.397 | 0.148 | 0.033 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 40.006 | 4.943 | 3.622 | 0.415 | 0.034 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 2.933 | 1.080 | 0.333 | 0.312 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 6.921 | 0.049 | 0.012 | 0.000 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 1.429 | 0.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 3.273 | 0.619 | 0.293 | 0.071 | 0.079 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 5.793 | 2.066 | 0.123 | 0.000 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.046 | 0.423 | 0.176 | 0.021 | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 49.115 | 0.090 | 0.123 | 0.149 | 0.051 | 0.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.913 | 1.103 | 0.069 | 0.223 | 0.317 | 0.349 | 0.197 | 0.094 | 0.034 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 119.856 | 0.557 | 1.404 | 0.120 | 0.176 | 0.094 | 0.076 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 40.235 | 14.123 | 0.589 | 1.534 | 0.258 | 0.659 | 0.198 | 0.024 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 39.090 | 0.779 | 0.439 | 0.042 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.870 | 3.825 | 2.066 | 0.542 | 0.063 | 0.096 | 0.043 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 7.593 | 0.167 | 0.107 | 0.052 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.810 | 2.974 | 2.539 | 0.865 | 0.099 | 0.262 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 2.808 | 0.938 | 0.586 | 0.590 | 0.069 | 0.017 | 0.034 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.209 | 0.401 | 0.354 | 0.801 | 0.181 | 0.075 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.953 | 0.546 | 0.396 | 0.306 | 0.327 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table A.66. Massachusetts Department of Marine Fisheries (MADMF) fall survey biomass indices-at-age (weight/tow) from 1981 to 2011 for Gulf of Maine Atlantic cod. Age data are not available prior to 1982. *Note: this survey index is not used in the current assessment.

| Year | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 |
|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| 1982 | 0.026 | 0.212 | 0.293 | 0.009 | 0.044 | 0.075 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.002 | 0.027 | 0.005 | 0.029 | 0.029 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.010 | 0.024 | 0.038 | 0.014 | 0.000 | 0.047 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.004 | 0.025 | 0.007 | 0.007 | 0.027 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.003 | 0.063 | 0.101 | 0.000 | 0.082 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.237 | 0.085 | 0.000 | 0.019 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.013 | 0.245 | 0.092 | 0.016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.004 | 0.008 | 0.100 | 0.032 | 0.074 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.008 | 0.332 | 0.153 | 0.093 | 0.067 | 0.106 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.008 | 0.220 | 0.118 | 0.007 | 0.080 | 0.048 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.050 | 0.104 | 0.027 | 0.000 | 0.000 | 0.092 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.744 | 0.149 | 0.177 | 0.234 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.032 | 0.651 | 0.990 | 0.285 | 0.041 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.009 | 0.301 | 0.205 | 0.270 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.050 | 0.007 | 0.002 | 0.000 | 0.024 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.009 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.029 | 0.075 | 0.113 | 0.065 | 0.078 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.091 | 0.115 | 0.058 | 0.000 | 0.044 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.001 | 0.096 | 0.127 | 0.037 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.032 | 0.003 | 0.060 | 0.241 | 0.228 | 0.193 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.002 | 0.079 | 0.022 | 0.273 | 0.766 | 1.000 | 0.922 | 0.557 | 0.373 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.217 | 0.067 | 0.407 | 0.077 | 0.347 | 0.393 | 0.342 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2004 | 0.093 | 0.710 | 0.162 | 1.179 | 0.369 | 2.082 | 0.879 | 0.107 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.020 | 0.037 | 0.117 | 0.032 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2006 | 0.005 | 0.300 | 0.517 | 0.469 | 0.067 | 0.309 | 0.272 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.009 | 0.023 | 0.024 | 0.021 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.011 | 0.248 | 0.834 | 0.651 | 0.182 | 0.454 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.030 | 0.128 | 0.137 | 0.341 | 0.085 | 0.026 | 0.060 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.003 | 0.074 | 0.145 | 0.586 | 0.456 | 0.137 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.005 | 0.051 | 0.166 | 0.654 | 0.386 | 0.094 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table A.67. Summary of maturity samples (individual fish) taken by the Maine – New Hampshire inshore groundfish survey by region and year.

| Year | Total maturity samples by region | | | | | Total | |
|------|----------------------------------|----|----|----|----|-------|--|
| | Region | | | | | | |
| | 1 | 2 | 3 | 4 | 5 | | |
| 2001 | 12 | | 2 | | 19 | 33 | |
| 2002 | 50 | 8 | | | 2 | 60 | |
| 2003 | 6 | 1 | 1 | 1 | 1 | 10 | |
| 2004 | 35 | 17 | 2 | 3 | 49 | 106 | |
| 2005 | 114 | 34 | 32 | 15 | 69 | 264 | |
| 2006 | 148 | 36 | 7 | 12 | 24 | 227 | |
| 2007 | 189 | 14 | 5 | 6 | 80 | 294 | |
| 2008 | 117 | 30 | 8 | 3 | 47 | 205 | |
| 2009 | 127 | 9 | 7 | 8 | 58 | 209 | |
| 2010 | 167 | 20 | 6 | | 40 | 233 | |
| 2011 | 44 | 30 | 23 | 14 | 60 | 171 | |

Table A.68. Proportion of mature fish observed by the Maine – New Hampshire inshore groundfish survey by region and year.

| Year | Proportion mature by region | | | | | |
|---------|-----------------------------|------|------|------|------|-------|
| | Region | | | | | Total |
| | 1 | 2 | 3 | 4 | 5 | |
| 2001 | 0.83 | | 1.00 | | 1.00 | 0.94 |
| 2002 | 0.90 | 1.00 | | | 1.00 | 0.92 |
| 2003 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.86 | 0.88 | 1.00 | 0.67 | 0.90 | 0.88 |
| 2005 | 0.91 | 0.85 | 0.94 | 1.00 | 1.00 | 0.94 |
| 2006 | 0.95 | 0.94 | 0.86 | 0.92 | 0.83 | 0.93 |
| 2007 | 0.87 | 0.86 | 1.00 | 1.00 | 1.00 | 0.91 |
| 2008 | 0.89 | 0.93 | 1.00 | 1.00 | 1.00 | 0.93 |
| 2009 | 0.70 | 0.89 | 1.00 | 0.88 | 1.00 | 0.81 |
| 2010 | 0.82 | 0.95 | 1.00 | | 1.00 | 0.87 |
| 2011 | 0.73 | 0.97 | 1.00 | 1.00 | 1.00 | 0.92 |
| Average | 0.86 | 0.93 | 0.98 | 0.93 | 0.98 | 0.91 |

Table A.69. Gulf of Maine Atlantic cod commercial otter trawl landings per unit effort index (LPUE) from 1982 to 2011 (from Palmer 2012b).

| Year | Estimate | CV | Lower 95% CI | Upper 95% CI |
|------|----------|------|--------------|--------------|
| 1982 | 1.000 | | | |
| 1983 | 0.990 | 0.04 | 0.911 | 1.075 |
| 1984 | 0.684 | 0.04 | 0.628 | 0.745 |
| 1985 | 0.594 | 0.04 | 0.546 | 0.645 |
| 1986 | 0.468 | 0.04 | 0.432 | 0.508 |
| 1987 | 0.316 | 0.04 | 0.291 | 0.344 |
| 1988 | 0.321 | 0.04 | 0.295 | 0.348 |
| 1989 | 0.494 | 0.05 | 0.451 | 0.541 |
| 1990 | 0.712 | 0.05 | 0.652 | 0.778 |
| 1991 | 0.721 | 0.04 | 0.661 | 0.787 |
| 1992 | 0.375 | 0.04 | 0.344 | 0.408 |
| 1993 | 0.324 | 0.04 | 0.297 | 0.353 |
| 1994 | 0.142 | 0.04 | 0.131 | 0.154 |
| 1995 | 0.163 | 0.04 | 0.151 | 0.175 |
| 1996 | 0.222 | 0.04 | 0.207 | 0.240 |
| 1997 | 0.257 | 0.04 | 0.238 | 0.278 |
| 1998 | 0.268 | 0.04 | 0.248 | 0.288 |
| 1999 | 0.127 | 0.04 | 0.118 | 0.138 |
| 2000 | 0.372 | 0.04 | 0.345 | 0.400 |
| 2001 | 0.601 | 0.04 | 0.558 | 0.647 |
| 2002 | 0.580 | 0.04 | 0.538 | 0.625 |
| 2003 | 0.558 | 0.04 | 0.518 | 0.602 |
| 2004 | 0.493 | 0.04 | 0.456 | 0.533 |
| 2005 | 0.504 | 0.04 | 0.466 | 0.545 |
| 2006 | 0.709 | 0.04 | 0.655 | 0.768 |
| 2007 | 0.879 | 0.04 | 0.811 | 0.953 |
| 2008 | 1.477 | 0.04 | 1.365 | 1.600 |
| 2009 | 1.621 | 0.04 | 1.495 | 1.758 |
| 2010 | 1.986 | 0.05 | 1.818 | 2.170 |
| 2011 | 1.475 | 0.04 | 1.362 | 1.598 |

Table A.70. Gulf of Maine Atlantic cod recreational VTR landings per unit effort index (LPUE) from 1994 to 2011 (from Wood 2012).

| Year | LPUE index | CV | Lower 95% CI | Upper 95% CI |
|-------------|-------------------|-----------|---------------------|---------------------|
| 1994 | 1.00 | | | |
| 1995 | 0.930 | 0.05 | 0.84 | 1.03 |
| 1996 | 0.821 | 0.05 | 0.74 | 0.91 |
| 1997 | 0.580 | 0.05 | 0.53 | 0.64 |
| 1998 | 0.541 | 0.05 | 0.49 | 0.60 |
| 1999 | 0.625 | 0.05 | 0.57 | 0.69 |
| 2000 | 0.747 | 0.05 | 0.68 | 0.82 |
| 2001 | 0.843 | 0.05 | 0.77 | 0.93 |
| 2002 | 0.568 | 0.05 | 0.52 | 0.62 |
| 2003 | 0.567 | 0.05 | 0.52 | 0.62 |
| 2004 | 0.497 | 0.05 | 0.45 | 0.55 |
| 2005 | 0.430 | 0.05 | 0.39 | 0.47 |
| 2006 | 0.256 | 0.05 | 0.23 | 0.28 |
| 2007 | 0.264 | 0.05 | 0.24 | 0.29 |
| 2008 | 0.309 | 0.05 | 0.28 | 0.34 |
| 2009 | 0.394 | 0.05 | 0.36 | 0.43 |
| 2010 | 0.520 | 0.05 | 0.47 | 0.57 |
| 2011 | 0.436 | 0.05 | 0.40 | 0.48 |

Table A.71. Example of Lorenzen-based age varying estimates of natural mortality (M) based on the average weight-at-age of Gulf of Maine Atlantic cod. Age-specific M estimates were re-scaled based on an assumption of age-invariant constant $M = 0.2$.

| Age | Average weight (kg) | Standard deviation (kg) | Natural mortality (M) | Cumulative survival | Rescaled natural mortality (M_{adj}) | Rescaled cumulative survival |
|-----|---------------------|-------------------------|-----------------------|---------------------|--|------------------------------|
| 1 | 0.116 | 0.062 | 0.86 | | 0.48 | |
| 2 | 0.387 | 0.099 | 0.60 | 0.42 | 0.33 | 0.62 |
| 3 | 1.068 | 0.172 | 0.44 | 0.23 | 0.25 | 0.44 |
| 4 | 2.000 | 0.222 | 0.36 | 0.15 | 0.20 | 0.35 |
| 5 | 3.186 | 0.506 | 0.32 | 0.10 | 0.18 | 0.28 |
| 6 | 4.723 | 0.972 | 0.28 | 0.08 | 0.16 | 0.24 |
| 7 | 6.705 | 1.420 | 0.25 | 0.06 | 0.14 | 0.20 |
| 8 | 9.029 | 1.826 | 0.23 | 0.04 | 0.13 | 0.18 |
| 9 | 11.441 | 2.091 | 0.21 | 0.04 | 0.12 | 0.15 |
| 10 | 13.770 | 2.601 | 0.20 | 0.03 | 0.11 | 0.14 |
| 11 | 18.404 | 3.490 | 0.18 | 0.02 | 0.10 | 0.12 |

Table A.72. Ratio of NEFSC spring survey proportions-at-age to fishery proportion-at-age. Cells shaded blue indicate where the survey proportion-at-age was greater than observed in the fishery. Cells shaded grey indicate where no information was available from either the survey or the fishery and no comparison could be made. Cells shaded white indicate where the fishery proportion-at-age was greater relative to the survey.

| Year | Age5 | Age6 | Age7 | Age8 | Age9+ |
|-------------------|------|-----------|-----------|-----------|------------|
| 1982 | 1.1 | 1.2 | 1.4 | No survey | 0.5 |
| 1983 | 0.8 | 0.9 | 1.9 | No survey | 5.2 |
| 1984 | 1.2 | 0.5 | 1.9 | No survey | No survey |
| 1985 | 1.1 | 0.7 | 1.2 | 0.6 | 0.9 |
| 1986 | 0.9 | 0.8 | 2.7 | No survey | 1.3 |
| 1987 | 0.7 | No survey | 2.8 | 2.6 | 1.3 |
| 1988 | 0.7 | 2.2 | 2.1 | 1.2 | 3.3 |
| 1989 | 0.7 | 4.0 | No survey | No survey | No survey |
| 1990 | 0.9 | 1.1 | 4.0 | No survey | No survey |
| 1991 | 1.0 | 1.1 | 1.4 | No survey | No survey |
| 1992 | 0.9 | 1.4 | 1.4 | No survey | 3.1 |
| 1993 | 0.8 | 0.9 | 1.4 | 2.1 | No fishery |
| 1994 | 0.4 | 1.1 | 2.6 | 0.4 | 7.6 |
| 1995 | 0.8 | 1.4 | No survey | 0.8 | 12.8 |
| 1996 | 1.0 | 1.7 | No survey | No survey | No survey |
| 1997 | 0.8 | 2.6 | 16.1 | No survey | No survey |
| 1998 | 0.6 | 1.3 | 0.5 | 2.6 | No survey |
| 1999 | 0.7 | 1.0 | 1.9 | No survey | 4.4 |
| 2000 | 0.7 | 1.4 | 1.1 | 2.7 | No fishery |
| 2001 | 0.9 | 0.6 | 1.7 | 6.6 | 1.0 |
| 2002 | 0.9 | 1.0 | 1.4 | 1.0 | 0.4 |
| 2003 | 0.8 | 1.1 | 2.3 | 1.4 | 2.3 |
| 2004 | 0.9 | 1.5 | 0.3 | 0.6 | 0.6 |
| 2005 | 0.2 | 1.0 | 2.0 | 0.9 | No survey |
| 2006 | 0.6 | 1.1 | 2.3 | 2.6 | 0.8 |
| 2007 | 1.0 | 1.1 | 2.9 | 0.7 | No survey |
| 2008 | 1.1 | 0.5 | 0.6 | No survey | No survey |
| 2009 | 1.0 | 1.1 | 0.3 | 0.6 | 0.3 |
| 2010 | 0.7 | 1.5 | 1.3 | 3.1 | 7.2 |
| 2011 | 1.0 | 0.7 | 2.7 | 2.6 | No survey |
| Cells ≥ 1 | 5.0 | 20.0 | 23.0 | 11.0 | 10.0 |
| Total | 31.0 | 30.0 | 28.0 | 19.0 | 18.0 |
| Fraction ≥ 1 | 0.16 | 0.67 | 0.82 | 0.58 | 0.56 |

Table A.73. Ratio of NEFSC fall survey proportions-at-age to fishery proportion-at-age. Cells shaded blue indicate where the survey proportion-at-age was greater than observed in the fishery. Cells shaded grey indicate where no information was available from either the survey or the fishery and no comparison could be made. Cells shaded white indicate where the fishery proportion-at-age was greater relative to the survey.

| Year | Age5 | Age6 | Age7 | Age8 | Age9+ |
|-------------------|-----------|-----------|-----------|-----------|------------|
| 1982 | 1.4 | No survey | No survey | No survey | No survey |
| 1983 | 0.9 | 1.2 | No survey | No survey | 3.5 |
| 1984 | 0.7 | 1.3 | 0.9 | No survey | 2.7 |
| 1985 | 0.9 | 1.0 | No survey | 1.9 | 5.5 |
| 1986 | 1.0 | 0.8 | No survey | No survey | 3.7 |
| 1987 | 1.6 | No survey | No survey | No survey | No survey |
| 1988 | 0.9 | No survey | 4.7 | No survey | 6.6 |
| 1989 | 0.9 | 2.1 | No survey | 3.4 | No survey |
| 1990 | 1.3 | 0.7 | 0.9 | No survey | No survey |
| 1991 | 1.0 | No survey | 5.3 | No survey | No survey |
| 1992 | 0.8 | 3.1 | No survey | No survey | No survey |
| 1993 | No survey | 1.3 | No survey | No survey | No fishery |
| 1994 | 0.9 | No survey | 2.8 | No survey | No survey |
| 1995 | 1.1 | 1.1 | No survey | No survey | No survey |
| 1996 | 1.1 | No survey | No survey | No survey | No survey |
| 1997 | 1.1 | No survey | No survey | No survey | No survey |
| 1998 | 1.5 | 0.8 | No survey | No survey | No survey |
| 1999 | 1.5 | 0.7 | No survey | No survey | No survey |
| 2000 | 1.2 | 0.2 | No survey | 5.0 | No fishery |
| 2001 | 0.9 | 1.1 | 1.5 | No survey | 1.9 |
| 2002 | 1.0 | 1.4 | 0.4 | 0.5 | No survey |
| 2003 | 1.1 | 0.6 | 1.4 | No survey | 1.4 |
| 2004 | 1.0 | 0.9 | 1.3 | 1.1 | 1.2 |
| 2005 | 0.5 | 0.8 | 1.6 | 1.7 | 1.4 |
| 2006 | 0.9 | 1.8 | 1.0 | 1.4 | 0.7 |
| 2007 | 0.8 | 1.3 | No survey | No survey | No survey |
| 2008 | 1.0 | No survey | No survey | No survey | 11.0 |
| 2009 | 1.0 | 1.2 | No survey | No survey | No survey |
| 2010 | 0.9 | 1.0 | 1.3 | No survey | 5.7 |
| 2011 | 1.1 | 0.8 | 0.5 | 2.8 | No survey |
| Cells ≥ 1 | 14.0 | 12.0 | 8.0 | 7.0 | 11.0 |
| Total | 30.0 | 23.0 | 14.0 | 9.0 | 13.0 |
| Fraction ≥ 1 | 0.47 | 0.52 | 0.57 | 0.78 | 0.85 |

Table A.74. Summary of the Gulf of Maine Atlantic cod ASAP model formulation used to build a ‘bridge’ from the SAW 53 ASAP base model (SAW53_BASE) to the 2011 update of the same model (SAW55_BASE).

| Step | Model | Description | Type | Software version | Years | Catch | Selectivity blocks | Time of spawning | Stock recruit | Survey selectivity | NEFSC | | MADMF Spring | | | | | | | | | | | | | |
|------|--------------------|--|------|------------------|-----------|---|----------------------|------------------|---------------|--|----------|------|--------------|--|--|--|--|--|--|--|--|--|--|--|--|--|
| | | | | | | | | | | | Spring | Fall | | | | | | | | | | | | | | |
| 1 | SAW53_BASE | Base model from SAW53 | ASAP | v2.0.21 | 1982-2010 | Single fleet | 1982-1990, 1991-2010 | April 1 (0.25) | Mean | NEFSC, flat topped (6+), MADMF double logistic | Ages 1-9 | | | | | | | | | | | | | | | |
| 2 | SAW55_B1 | Update recreational catch and catch WAA (MRIP adjustments) | | | | Single fleet; update recreational catch and catch WAA to account for MRIP adjustments | | | | | | | | | | | | | | | | | | | | |
| 3 | SAW55_B2 | Update commercial and recreational discards to account for discard mortality, update catch WAA | | | | Single fleet; update commercial and recreational discards and catch WAA to account for differential discard mortality | | | | | | | | | | | | | | | | | | | | |
| 4 | SAW55_B3 | Update stock WAA | | | 1982-2011 | Single fleet | 1982-1990, 1991-2011 | | | | | | | | | | | | | | | | | | | |
| 5 | SAW55_B4 | Update maturity ogive | | | | | | | | | | | | | | | | | | | | | | | | |
| 6 | SAW55_B5 | Update MADMF spring survey index and timing (April->May) | | v3.0.8 | | | | | | | | | | | | | | | | | | | | | | |
| 7 | SAW55_B6 | Add 2011 data | | | | | | | | | | | | | | | | | | | | | | | | |
| 8 | SAW55_BASE | Update software | | | | | | | | | | | | | | | | | | | | | | | | |
| 9 | SAW55_BASE_100MORT | 100% discard mortality | | | | | | | | | | | | | | | | | | | | | | | | |

Table A.75. Summary Gulf of Maine Atlantic cod model results from the ‘bridge building’ exercise performed to update the SAW 53 ASAP base model (SAW53_BASE) to the 2011 update of the same model (SAW55_BASE). Differences in model formulations are summarized in Table A.74.

| Model | SAW53_BASE | SAW55_B1 | SAW55_B2 | SAW55_B3 | SAW55_B4 | SAW55_B5 | SAW55_B6 | SAW55_BASE | SAW55_BASE_100MORT |
|---|-----------------------|--|--|------------------|-----------------------|----------------------------|---------------|-----------------|------------------------|
| Model description | Base model from SAW53 | Update recreational catch and catch WAA (MRIP adjustments) | Update commercial and recreational discards to account for discard mortality, update catch WAA | Update stock WAA | Update maturity ogive | Update MADMF spring survey | Add 2011 data | Update software | 100% discard mortality |
| Number of parameters | 99 | 99 | 99 | 99 | 99 | 99 | 101 | 101 | 101 |
| Objective function | 2467 | 2486 | 2471 | 2471 | 2471 | 2466 | 2554 | 2554 | 2570 |
| Components of objective function | Recruit devs | 286 | 285 | 282 | 282 | 282 | 293 | 293 | 296 |
| | Suvey age comps | 831 | 830 | 831 | 831 | 825 | 860 | 860 | 859 |
| | Catch age comps | 378 | 399 | 383 | 383 | 384 | 395 | 395 | 412 |
| | Index fit | 764 | 766 | 771 | 771 | 771 | 794 | 794 | 790 |
| | Catch fit | 208 | 206 | 204 | 204 | 204 | 211 | 211 | 213 |
| RMSE | Fleet 1 | 0.24 | 0.25 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 | 0.25 |
| | Index 1 | 1.05 | 1.07 | 1.12 | 1.12 | 1.12 | 1.14 | 1.14 | 1.09 |
| | Index 2 | 0.91 | 0.94 | 0.97 | 0.97 | 0.97 | 0.97 | 0.97 | 0.94 |
| | Index 3 | 1.07 | 1.09 | 1.15 | 1.15 | 1.15 | 1.13 | 1.13 | 1.07 |
| | Recruit devs | 1.28 | 1.04 | 1.08 | 1.08 | 1.08 | 1.42 | 1.42 | 1.04 |
| SSB₁₉₈₂ (mt) | 23,675 | 22,697 | 23,153 | 23,153 | 23,240 | 23,243 | 23,320 | 23,320 | 22,847 |
| SSB₂₀₁₀ (mt) | 11,868 | 11,172 | 11,515 | 11,515 | 11,877 | 11,814 | 12,746 | 12,746 | 12,984 |
| SSB₂₀₁₁ (mt) | | | | | | | 11,874 | 11,874 | 11,403 |
| F_{mult, 2010} | | 1.14 | 0.85 | 0.67 | 0.67 | 0.67 | 0.62 | 0.62 | 0.74 |
| F_{mult, 2011} | | 1.14 | | | | | 0.59 | 0.59 | 0.75 |
| Mohn's rho (5 year peel) | SSB | 0.22 | 0.25 | 0.47 | 0.47 | 0.47 | 0.33 | 0.33 | 0.14 |
| | F _{mult} | -0.22 | -0.24 | -0.35 | -0.35 | -0.35 | -0.25 | -0.25 | -0.14 |
| | Age 1 N | 0.15 | 0.18 | 0.30 | 0.30 | 0.28 | 0.37 | 0.37 | 0.23 |

Table A.76. Summary Gulf of Maine Atlantic cod model estimated fishery and survey selectivity parameters results from the ‘bridge building’ exercise performed to update the SAW 53 ASAP base model (SAW53_BASE) to the 2011 update of the same model (SAW55_BASE). Differences in model formulations are summarized in Table A.74.

| Model | SAW53_BASE | | SAW55_B1 | | SAW55_B2 | | SAW55_B3 | | SAW55_B4 | | SAW55_B5 | | SAW55_B6 | | SAW55_BASE | | SAW55_BASE_100MORT | |
|---------------|-----------------------|-------|--|-------|--|-------|------------------|-------|-----------------------|-------|----------------------------|-------|---------------|-------|-----------------|-------|------------------------|----|
| Description | Base model from SAW53 | | Update recreational catch and catch WAA (MRIP adjustments) | | Update commercial and recreational discards to account for discard mortality, update catch WAA | | Update stock WAA | | Update maturity ogive | | Update MADMF spring survey | | Add 2011 data | | Update software | | 100% discard mortality | |
| | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV | Selectivity | CV |
| Fleet block 1 | 1 | 0.05 | 0.17 | 0.05 | 0.16 | 0.04 | 0.18 | 0.04 | 0.18 | 0.04 | 0.18 | 0.04 | 0.18 | 0.04 | 0.18 | 0.05 | 0.16 | |
| | 2 | 0.28 | 0.10 | 0.29 | 0.10 | 0.25 | 0.10 | 0.25 | 0.10 | 0.25 | 0.10 | 0.25 | 0.10 | 0.25 | 0.10 | 0.29 | 0.10 | |
| | 3 | 0.58 | 0.10 | 0.58 | 0.10 | 0.57 | 0.09 | 0.57 | 0.09 | 0.57 | 0.09 | 0.57 | 0.09 | 0.57 | 0.09 | 0.58 | 0.10 | |
| | 4 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | |
| | 5 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 6 | 0.77 | 0.26 | 0.76 | 0.26 | 0.77 | 0.25 | 0.77 | 0.25 | 0.77 | 0.25 | 0.78 | 0.25 | 0.78 | 0.25 | 0.77 | 0.26 | |
| | 7 | 0.99 | 0.39 | 0.99 | 0.39 | 1.00 | 0.38 | 1.00 | 0.38 | 1.00 | 0.38 | 1.00 | 0.14 | 1.00 | 0.15 | 1.00 | 0.39 | |
| | 8 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | |
| | 9 | 0.31 | 0.47 | 0.31 | 0.47 | 0.33 | 0.46 | 0.33 | 0.46 | 0.33 | 0.46 | 0.33 | 0.45 | 0.33 | 0.45 | 0.31 | 0.47 | |
| Fleet block 2 | 1 | 0.02 | 0.17 | 0.02 | 0.16 | 0.02 | 0.18 | 0.02 | 0.18 | 0.02 | 0.18 | 0.02 | 0.18 | 0.02 | 0.18 | 0.02 | 0.16 | |
| | 2 | 0.11 | 0.10 | 0.10 | 0.10 | 0.07 | 0.11 | 0.07 | 0.11 | 0.07 | 0.11 | 0.07 | 0.11 | 0.07 | 0.11 | 0.10 | 0.10 | |
| | 3 | 0.40 | 0.08 | 0.38 | 0.08 | 0.33 | 0.08 | 0.33 | 0.08 | 0.33 | 0.08 | 0.33 | 0.08 | 0.32 | 0.08 | 0.38 | 0.08 | |
| | 4 | 0.84 | 0.08 | 0.84 | 0.08 | 0.82 | 0.07 | 0.82 | 0.07 | 0.82 | 0.07 | 0.82 | 0.07 | 0.79 | 0.07 | 0.82 | 0.07 | |
| | 5 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | |
| | 6 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 7 | 0.90 | 0.20 | 0.91 | 0.19 | 0.95 | 0.17 | 0.95 | 0.17 | 0.95 | 0.17 | 0.95 | 0.17 | 0.92 | 0.17 | 0.87 | 0.19 | |
| | 8 | 0.88 | 0.33 | 0.89 | 0.32 | 0.90 | 0.29 | 0.90 | 0.29 | 0.90 | 0.29 | 0.89 | 0.29 | 0.88 | 0.27 | 0.86 | 0.31 | |
| | 9 | 0.67 | 0.54 | 0.68 | 0.54 | 0.70 | 0.51 | 0.70 | 0.51 | 0.70 | 0.51 | 0.70 | 0.51 | 0.77 | 0.50 | 0.77 | 0.53 | |
| Index 1 | 1 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | 0.04 | 0.19 | |
| | 2 | 0.12 | 0.16 | 0.12 | 0.16 | 0.14 | 0.16 | 0.14 | 0.16 | 0.14 | 0.16 | 0.14 | 0.16 | 0.14 | 0.16 | 0.12 | 0.16 | |
| | 3 | 0.26 | 0.16 | 0.26 | 0.16 | 0.30 | 0.15 | 0.30 | 0.15 | 0.30 | 0.15 | 0.30 | 0.15 | 0.30 | 0.15 | 0.26 | 0.15 | |
| | 4 | 0.46 | 0.15 | 0.47 | 0.15 | 0.51 | 0.15 | 0.51 | 0.15 | 0.51 | 0.15 | 0.51 | 0.15 | 0.51 | 0.15 | 0.48 | 0.15 | |
| | 5 | 0.71 | 0.15 | 0.72 | 0.15 | 0.75 | 0.15 | 0.75 | 0.15 | 0.75 | 0.15 | 0.75 | 0.15 | 0.75 | 0.15 | 0.72 | 0.15 | |
| | 6 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 7 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 8 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 9 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| Index 2 | 1 | 0.14 | 0.22 | 0.14 | 0.21 | 0.17 | 0.15 | 0.17 | 0.15 | 0.17 | 0.15 | 0.17 | 0.15 | 0.17 | 0.15 | 0.14 | 0.21 | |
| | 2 | 0.33 | 0.21 | 0.34 | 0.21 | 0.40 | 0.14 | 0.40 | 0.14 | 0.40 | 0.14 | 0.40 | 0.14 | 0.40 | 0.14 | 0.34 | 0.20 | |
| | 3 | 0.51 | 0.21 | 0.52 | 0.21 | 0.59 | 0.14 | 0.59 | 0.14 | 0.59 | 0.14 | 0.59 | 0.14 | 0.59 | 0.14 | 0.52 | 0.20 | |
| | 4 | 0.82 | 0.21 | 0.83 | 0.21 | 0.89 | 0.14 | 0.89 | 0.14 | 0.89 | 0.14 | 0.89 | 0.14 | 0.89 | 0.14 | 0.83 | 0.20 | |
| | 5 | 0.97 | 0.21 | 0.97 | 0.21 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 0.98 | 0.20 | |
| | 6 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 7 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 8 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| | 9 | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | 1.00 | | |
| Index 3 | A50 ascend | 0.00 | 3000.09 | 0.00 | 3000.09 | 0.00 | 3000.00 | 0.00 | 3000.00 | 0.00 | 3000.18 | 0.00 | 3000.10 | 0.00 | 3000.30 | 0.00 | 3000.10 | |
| | Slope ascend | 10.00 | | 10.00 | | 10.00 | | 10.00 | | 10.00 | | 10.00 | | 10.00 | | 10.00 | | |
| | A50 descend | 0.00 | 3000.42 | 0.00 | 2999.92 | 0.00 | 2999.98 | 0.00 | 2999.98 | 0.00 | 2999.54 | 0.00 | 2999.96 | 0.00 | 2994.57 | 0.00 | 3000.00 | |
| | Slope descend | 4.22 | 0.22 | 4.10 | 0.21 | 3.54 | 0.17 | 3.54 | 0.17 | 3.54 | 0.17 | 3.51 | 0.18 | 3.50 | 0.18 | 4.29 | 0.22 | |

Table A.77. Summary of Gulf of Maine Atlantic cod ASAP model configurations that explored alternate formulations of a two-selectivity block structure by altering the starting point of the second selectivity block.

| Model | SAW55_BASE_1987 | SAW55_BASE_1989 | SAW55_BASE | SAW55_BASE_1993 | | | | | |
|-------------------------------|-----------------|-----------------|------------|-----------------|------|------|------|------|------|
| Block 2 start year | 1987 | 1989 | 1991 | 1993 | | | | | |
| Parameters | 101 | 101 | 101 | 101 | | | | | |
| Objective function | 2544 | 2544 | 2554 | 2551 | | | | | |
| Catch RMSE | 0.28 | 0.29 | 0.29 | 0.28 | | | | | |
| SSB1982 (mt) | 23551 | 23390 | 23320 | 23269 | | | | | |
| SSB2011 (mt) | 11921 | 11967 | 11874 | 11888 | | | | | |
| F_{age5, 2011} | 0.58 | 0.58 | 0.59 | 0.60 | | | | | |
| Block 1 | Age1 | 0.06 | 0.21 | 0.04 | 0.19 | 0.04 | 0.18 | 0.04 | 0.18 |
| | Age2 | 0.37 | 0.13 | 0.31 | 0.11 | 0.25 | 0.10 | 0.22 | 0.12 |
| | Age3 | 0.74 | 0.13 | 0.67 | 0.11 | 0.57 | 0.09 | 0.52 | 0.12 |
| | Age4 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 0.90 | 0.13 |
| | Age5 | 1.00 | | 1.00 | | 1.00 | | 1.00 | |
| | Age6 | 0.86 | 0.30 | 0.83 | 0.27 | 0.78 | 0.25 | 0.80 | 0.23 |
| | Age7 | 1.00 | 0.00 | 1.00 | 0.03 | 1.00 | 0.07 | 0.99 | 0.33 |
| | Age8 | 1.00 | 0.01 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 |
| | Age9+ | 0.32 | 0.56 | 0.33 | 0.49 | 0.33 | 0.45 | 0.30 | 0.45 |
| | Age1 | 0.02 | 0.16 | 0.02 | 0.18 | 0.02 | 0.18 | 0.01 | 0.20 |
| Block 2 | Age2 | 0.08 | 0.09 | 0.07 | 0.10 | 0.07 | 0.11 | 0.06 | 0.12 |
| | Age3 | 0.33 | 0.07 | 0.33 | 0.07 | 0.32 | 0.08 | 0.31 | 0.08 |
| | Age4 | 0.81 | 0.07 | 0.81 | 0.07 | 0.79 | 0.07 | 0.78 | 0.07 |
| | Age5 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.00 |
| | Age6 | 1.00 | | 1.00 | | 1.00 | | 1.00 | |
| | Age7 | 0.89 | 0.17 | 0.91 | 0.17 | 0.92 | 0.17 | 0.90 | 0.18 |
| | Age8 | 0.84 | 0.27 | 0.86 | 0.27 | 0.88 | 0.27 | 0.87 | 0.28 |
| | Age9+ | 0.79 | 0.48 | 0.79 | 0.51 | 0.77 | 0.50 | 0.79 | 0.48 |

Table A.78. Summary of Gulf of Maine Atlantic cod ASAP model configurations that fit the Massachusetts Department of Marine Fisheries (MADMF) spring survey using both parametric (double logistic; e.g., SAW55_BASE) and non-parametric (at-age; e.g. all other models) approaches.

| Model | SAW55_BASE | SAW55_BASE_FIXED_MADMF_AGE1_9 | SAW55_BASE_FIXED_MADMF_AGE1_6 |
|----------------------------------|-----------------|-------------------------------|-------------------------------|
| Parameters | 101 | 103 | 100 |
| Objective function | 2554 | 2552 | 2543 |
| Maximum gradient | 1.6E-03 | 3.9E-05 | 1.8E-03 |
| Components of objective function | Suvey age comps | 860 | 858 |
| | Catch age comps | 395 | 395 |
| | Index fit | 794 | 795 |
| | Catch fit | 211 | 211 |
| | Recruit devs | 293 | 294 |
| RMSE | Catch | 0.29 | 0.29 |
| | Index1 | 1.14 | 1.14 |
| | Index2 | 0.97 | 0.97 |
| | Index3 | 1.13 | 1.14 |
| | Index total | 1.08 | 1.09 |
| | Recruit devs | 1.42 | 1.43 |
| Mean age RMSE | Fleet1 | 1.34 | 1.34 |
| | Index1 | 1.50 | 1.50 |
| | Index2 | 1.74 | 1.73 |
| | Index3 | 1.37 | 1.36 |
| SSB1982 (mt) | | 23320 | 23152 |
| SSB2011 (mt) | | 11874 | 11669 |
| Fage5, 2011 | | 0.59 | 0.60 |
| | | | 0.60 |

Table A.79. Summary of Gulf of Maine Atlantic cod ASAP model configurations which explored various configurations of a three-selectivity block model. The SAW55_BASE model is the two-block reference model.

| Model | SAW55_BASE | SAW55_3BLOCK | SAW55_3BLOCK_DL | SAW55_3BLOCK_SL | SAW55_3BLOCK_SL_1989 | SAW55_3BLOCK_SL_MADMF_1_6 | SAW55_3BLOCK_BASE |
|----------------------------------|-----------------|--------------|-----------------|-----------------|----------------------|---------------------------|-------------------|
| Selectivity blocks | 2 | 3 | 3 | 3 | 3 | 3 | 3 |
| Year splits | 1991 | 1991, 2005 | 1991, 2005 | 1991, 2004 | 1989, 2005 | 1989, 2005 | 1989, 2005 |
| Parameters | 101 | 109 | 97 | 91 | 91 | 93 | 93 |
| Objective function | 2554 | 2538 | 2544 | 2548 | 2536 | 2524 | 2055 |
| Components of objective function | Suvey age comps | 860 | 856 | 858 | 861 | 858 | 846 |
| | Catch age comps | 395 | 383 | 386 | 388 | 378 | 390 |
| | Index fit | 794 | 796 | 796 | 796 | 797 | 794 |
| | Catch fit | 211 | 211 | 211 | 211 | 211 | 210 |
| | Recruit devs | 293 | 293 | 293 | 293 | 293 | 59 |
| RMSE | Catch | 0.29 | 0.27 | 0.27 | 0.28 | 0.28 | 0.21 |
| | Index1 | 1.14 | 1.15 | 0.15 | 1.15 | 1.15 | 1.13 |
| | Index2 | 0.97 | 1.00 | 1.00 | 1.01 | 1.01 | 0.97 |
| | Index3 | 1.13 | 1.15 | 1.14 | 1.15 | 1.14 | 1.14 |
| | Recruit devs | 1.42 | 1.40 | 1.40 | 1.10 | 1.11 | 1.51 |
| Mean age RMSE | Fleet1 | 1.34 | 1.10 | 1.11 | 1.16 | 0.98 | 0.98 |
| | Index1 | 1.50 | 1.44 | 1.45 | 1.44 | 1.44 | 1.02 |
| | Index2 | 1.74 | 1.61 | 1.62 | 1.61 | 1.62 | 1.18 |
| | Index3 | 1.37 | 1.38 | 1.38 | 1.38 | 1.40 | 1.06 |
| SSB ₁₉₈₂ (mt) | 23,320 | 22,992 | 23,103 | 22,410 | 22,546 | 22,446 | 22,036 |
| SSB ₂₀₁₁ (mt) | 11,874 | 12,069 | 12,184 | 11,971 | 12,020 | 11,841 | 9,903 |
| F _{age5, 2011} | 0.59 | 0.63 | 0.63 | 0.63 | 0.62 | 0.64 | 0.78 |

Table A.80. Summary of Gulf of Maine Atlantic cod ASAP model selectivity parameter estimates and the corresponding coefficients of variation (CV) from model configurations which explored various configurations of a three-selectivity block model. The SAW55_BASE model is the two-block reference model.

| Model | SAW55_BASE | SAW55_3BLOCK | SAW55_3BLOCK_DL | SAW55_3BLOCK_SL | SAW55_3BLOCK_SL_1989 | SAW55_3BLOCK_SL_MADMF_1_6 | SAW55_3BLOCK_BASE | | | | | |
|--------------------|------------|--------------|-----------------|-----------------|----------------------|---------------------------|-------------------|------|------|------|------|------|
| Selectivity blocks | 2 | 3 | 3 | 3 | 3 | 3 | 3 | | | | | |
| Year splits | 1991 | 1991, 2005 | 1991, 2005 | 1991, 2004 | 1989, 2005 | 1989, 2005 | 1989, 2005 | | | | | |
| Block 1 | A50% up | | 2.69 | 0.05 | 2.60 | 0.04 | 2.34 | 0.04 | 2.35 | 0.04 | 2.33 | 0.05 |
| | Slope up | | 0.56 | 0.08 | 0.53 | 0.08 | 0.46 | 0.09 | 0.46 | 0.09 | 0.46 | 0.10 |
| | A50% down | | 8.96 | 2.36 | | | | | | | | |
| | Slope down | | 0.06 | 517.24 | | | | | | | | |
| Block 2 | A50% up | | 3.50 | 0.04 | 3.30 | 0.03 | 3.30 | 0.02 | 3.30 | 0.02 | 3.32 | 0.02 |
| | Slope up | | 0.56 | 0.05 | 0.56 | 0.05 | 0.56 | 0.05 | 0.56 | 0.05 | 0.56 | 0.05 |
| | A50% down | | 0.00 | 3001.61 | | | | | | | | |
| | Slope down | | 7.31 | 0.48 | | | | | | | | |
| Block 3 | A50% up | | 3.76 | 0.03 | 3.75 | 0.03 | 3.75 | 0.03 | 3.76 | 0.03 | 3.77 | 0.04 |
| | Slope up | | 0.52 | 0.07 | 0.52 | 0.07 | 0.52 | 0.07 | 0.52 | 0.07 | 0.53 | 0.07 |
| | A50% down | | 9.00 | 0.00 | | | | | | | | |
| | Slope down | | 0.01 | 3000.08 | | | | | | | | |
| Block 1 | Age1 | 0.04 | 0.18 | 0.04 | 0.18 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | |
| | Age2 | 0.25 | 0.10 | 0.24 | 0.10 | 0.22 | 0.24 | 0.32 | 0.32 | 0.32 | 0.33 | |
| | Age3 | 0.57 | 0.09 | 0.56 | 0.09 | 0.64 | 0.68 | 0.81 | 0.80 | 0.80 | 0.81 | |
| | Age4 | 1.00 | 0.00 | 1.00 | 0.00 | 0.91 | 0.93 | 0.97 | 0.97 | 0.97 | 0.97 | |
| | Age5 | 1.00 | | 1.00 | | 0.98 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age6 | 0.78 | 0.25 | 0.78 | 0.25 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age7 | 1.00 | 0.07 | 1.00 | 0.05 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age8 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age9+ | 0.33 | 0.45 | 0.33 | 0.45 | 0.32 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age1 | 0.02 | 0.18 | 0.02 | 0.19 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | |
| Block 2 | Age2 | 0.07 | 0.11 | 0.09 | 0.12 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | |
| | Age3 | 0.32 | 0.08 | 0.39 | 0.09 | 0.37 | 0.37 | 0.37 | 0.37 | 0.37 | 0.36 | |
| | Age4 | 0.79 | 0.07 | 0.87 | 0.09 | 0.83 | 0.78 | 0.78 | 0.78 | 0.78 | 0.77 | |
| | Age5 | 1.00 | 0.00 | 1.00 | 0.00 | 1.00 | 0.96 | 0.95 | 0.95 | 0.95 | 0.95 | |
| | Age6 | 1.00 | | 1.00 | | 0.96 | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 | |
| | Age7 | 0.92 | 0.17 | 0.97 | 0.22 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age8 | 0.88 | 0.27 | 0.94 | 0.38 | 0.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age9+ | 0.77 | 0.50 | 0.66 | 0.66 | 0.72 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| Block 3 | Age1 | | 0.00 | 0.54 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | |
| | Age2 | | 0.03 | 0.25 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | |
| | Age3 | | 0.18 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | |
| | Age4 | | 0.58 | 0.17 | 0.61 | 0.62 | 0.62 | 0.62 | 0.61 | 0.61 | 0.61 | |
| | Age5 | | 0.89 | 0.18 | 0.91 | 0.92 | 0.92 | 0.91 | 0.91 | 0.91 | 0.91 | |
| | Age6 | | 1.00 | | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 | 0.98 | |
| | Age7 | | 0.81 | 0.29 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age8 | | 0.83 | 0.44 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |
| | Age9+ | | 0.77 | 0.73 | 0.50 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | |

Table A.81. Summary of the sensitivity runs conducted on the final Gulf of Maine Atlantic cod ASAP model, SAW55_3BLOCK_BASE.

| Run description | Flat top fleet selectivity | Domed fleet selectivity |
|------------------------------------|-----------------------------------|--|
| Base run | SAW55_3BLOCK_BASE | SAW55_3BLOCK_BASE_DOME |
| 100% discard mortality | SAW55_3BLOCK_BASE_100MORT | SAW55_3BLOCK_BASE_DOME_100MORT |
| M split | SAW55_3BLOCK_BASE_M_SPLIT | SAW55_3BLOCK_BASE_DOME_M_SPLIT |
| M split and 100% discard mortality | SAW55_3BLOCK_BASE_M_SPLIT_100MORT | SAW55_3BLOCK_BASE_DOME_M_SPLIT_100MORT |

Table A.82. Summary of Gulf of Maine Atlantic cod ASAP model configurations which explored various configurations of the final base model SAW55_3BLOCK_BASE under fleet flat topped-selectivity assumptions.

| Model | SAW55_3BLOCK_BASE | SAW55_3BLOCK_BASE_100MORT | SAW55_3BLOCK_BASE_M_SPLIT | SAW55_3BLOCK_BASE_M_SPLIT_M6 | SAW55_3BLOCK_BASE_M_SPLIT_100MORT |
|----------------------------------|--|--------------------------------------|--|--|---|
| Selectivity blocks | 3 | 3 | 3 | 3 | 3 |
| Year splits | 1989, 2005 | 1989, 2005 | 1989, 2005 | 1989, 2005 | 1989, 2005 |
| Parameters | 93 | 93 | 93 | 93 | 93 |
| Objective function | 2055 | 2041 | 2047 | 2069 | 2039 |
| Maximum gradient | 9.2E-05 | 9.0E-04 | 3.6E-05 | 3.7E-04 | 9.5E-05 |
| Components of objective function | Suvey age comps | 602 | 602 | 601 | 604 |
| | Catch age comps | 390 | 378 | 390 | 394 |
| | Index fit | 794 | 789 | 786 | 798 |
| | Catch fit | 210 | 212 | 210 | 210 |
| | Recruit devs | 59 | 60 | 60 | 63 |
| | | | | | 61 |
| RMSE | Catch | 0.21 | 0.18 | 0.15 | 0.25 |
| | Index1 | 1.13 | 1.08 | 0.99 | 0.98 |
| | Index2 | 0.97 | 0.93 | 0.98 | 1.25 |
| | Index3 | 1.14 | 1.07 | 1.03 | 1.11 |
| | Index total | 1.08 | 1.03 | 1.00 | 1.12 |
| | Recruit devs | 1.51 | 1.46 | 1.26 | 1.31 |
| Mean age RMSE | Fleet1 | 0.96 | 0.94 | 0.96 | 0.99 |
| | Index1 | 1.02 | 1.00 | 1.02 | 1.04 |
| | Index2 | 1.18 | 1.21 | 1.17 | 1.19 |
| | Index3 | 1.06 | 1.05 | 1.07 | 1.10 |
| SSB1982 (mt) | | 22036 | 22052 | 21531 | 20884 |
| SSB2011 (mt) | | 9903 | 9521 | 10221 | 10256 |
| Fage5,2011 | | 0.78 | 0.97 | 0.82 | 1.00 |
| Mohn's rho (5 year peel) | SSB | 0.40 | 0.26 | -0.01 | -0.08 |
| | Fmult | -0.27 | -0.19 | 0.06 | 0.14 |
| | Age 1 N | 0.76 | 0.57 | 0.24 | 0.12 |
| Comments | Based on SAW55_3BLOCK_FINAL, but rec. dev. lambda=0.2 and took out MADMF spring age 7-9. | Assumption of 100% discard mortality | 2 block natural mortality: 1982-88 = 0.2, 1989-1990 = linear ramp, 2003-2011 = 0.4 | 2 block natural mortality: 1982-88 = 0.2, 1989-1990 = linear ramp, 2003-2011 = 0.6 | Assumption of 100% discard mortality and 2 block natural mortality: 1982-88 = 0.2, 1989-2002 = linear ramp, 2003-2011 = 0.4 |

Table A.83. Summary of Gulf of Maine Atlantic cod ASAP model configurations which explored various configurations of the final base model SAW55_3BLOCK_BASE under fleet domed-selectivity assumptions.

| Model | SAW55_3BLOCK_BASE | SAW55_3BLOCK_BASE_DOME | SAW55_3BLOCK_BASE_DOME_100MORT | SAW55_3BLOCK_BASE_DOME_M_SPLIT | SAW55_3BLOCK_BASE_DOME_M_SPLIT_100MORT |
|----------------------------------|--|---|---|---|---|
| Selectivity blocks | 3 | 3 | 3 | 3 | 3 |
| Year splits | 1989, 2005 | 1989, 2005 | 1989, 2005 | 1989, 2005 | 1989, 2005 |
| Parameters | 93 | 107 | 107 | 107 | 107 |
| Objective function | 2055 | 2048 | 2034 | 2040 | 2031 |
| Maximum gradient | 9.2E-05 | 4.5E-05 | 1.5E-04 | 4.4E-05 | 9.3E-04 |
| Components of objective function | Suvey age comps | 602 | 600 | 600 | 598 |
| | Catch age comps | 390 | 385 | 373 | 386 |
| | Index fit | 794 | 794 | 789 | 786 |
| | Catch fit | 210 | 210 | 212 | 210 |
| | Recruit devs | 59 | 59 | 60 | 60 |
| | | | | | 61 |
| RMSE | Catch | 0.21 | 0.21 | 0.18 | 0.15 |
| | Index1 | 1.13 | 1.12 | 1.07 | 0.98 |
| | Index2 | 0.97 | 0.97 | 0.93 | 0.98 |
| | Index3 | 1.14 | 1.14 | 1.08 | 1.04 |
| | Index total | 1.08 | 1.08 | 1.03 | 1.00 |
| | Recruit devs | 1.51 | 1.51 | 1.45 | 1.26 |
| Mean age RMSE | Fleet1 | 0.96 | 0.94 | 0.92 | 0.94 |
| | Index1 | 1.02 | 1.03 | 1.00 | 1.02 |
| | Index2 | 1.18 | 1.17 | 1.20 | 1.16 |
| | Index3 | 1.06 | 1.06 | 1.05 | 1.07 |
| SSB1982 (mt) | | 22036 | 23156 | 22941 | 21531 |
| SSB2011 (mt) | | 9903 | 10017 | 9597 | 10221 |
| Fage5,2011 | | 0.78 | 0.77 | 0.99 | 0.82 |
| Mohn's rho (5 year peel) | SSB | 0.40 | 0.47 | 0.28 | 0.00 |
| | F _{mult} | -0.27 | -0.29 | -0.19 | 0.05 |
| | Age 1 N | 0.76 | 0.78 | 0.65 | 0.15 |
| Comments | Based on SAW55_3BLOCK_FINAL, but rec. dev. lambda=0.2 and took out MADMF spring age 7-9. | Allow for domed shaped commercial selectivity | Allow for domed shaped commercial selectivity and assume 100% discard mortality | Allow for domed shaped commercial selectivity with 2 block natural mortality: 1982-88 = 0.2, 1989-2002 = linear ramp, 2003-2011 = 0.4 | Allow for domed shaped commercial selectivity with 2 block natural mortality: 1982-88 = 0.2, 1989-2002 = linear ramp, 2003-2011 = 0.4 and assume 100% discard mortality |

Table A.84. Comparison of the fleet and index selectivity parameters and the corresponding coefficients of variation (CV) from the Gulf of Maine Atlantic cod ASAP SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models.

| Block/Index | Parameter | SAW55_3BLOCK_BASE | | SAW55_3BLOCK_BASE_M_SPLIT | |
|--------------|-----------|--------------------|------|---------------------------|------|
| | | Parameter estimate | CV | Parameter estimate | CV |
| 1982-1988 | A50% up | 2.33 | 0.05 | 2.33 | 0.05 |
| | Slope up | 0.46 | 0.10 | 0.45 | 0.09 |
| 1989-2004 | A50% up | 3.32 | 0.02 | 3.35 | 0.02 |
| | Slope up | 0.56 | 0.05 | 0.53 | 0.05 |
| 2005-2011 | A50% up | 3.77 | 0.04 | 3.82 | 0.03 |
| | Slope up | 0.53 | 0.07 | 0.51 | 0.07 |
| NEFSC spring | Age1 | 0.04 | 0.24 | 0.03 | 0.24 |
| | Age2 | 0.13 | 0.20 | 0.11 | 0.20 |
| | Age3 | 0.27 | 0.19 | 0.24 | 0.19 |
| | Age4 | 0.48 | 0.19 | 0.46 | 0.19 |
| | Age5 | 0.68 | 0.20 | 0.67 | 0.20 |
| | Age6 | 1.00 | | 1.00 | |
| | Age7 | 1.00 | | 1.00 | |
| | Age8 | 1.00 | | 1.00 | |
| | Age9+ | 1.00 | | 1.00 | |
| NEFSC fall | Age1 | 0.15 | 0.26 | 0.12 | 0.26 |
| | Age2 | 0.33 | 0.25 | 0.29 | 0.25 |
| | Age3 | 0.50 | 0.25 | 0.47 | 0.25 |
| | Age4 | 0.73 | 0.25 | 0.72 | 0.25 |
| | Age5 | 0.86 | 0.27 | 0.87 | 0.27 |
| | Age6 | 1.00 | | 1.00 | |
| | Age7 | 1.00 | | 1.00 | |
| | Age8 | 1.00 | | 1.00 | |
| | Age9+ | 1.00 | | 1.00 | |
| MADMF spring | Age1 | 1.00 | | 1.00 | |
| | Age2 | 0.73 | 0.15 | 0.81 | 0.15 |
| | Age3 | 0.64 | 0.18 | 0.77 | 0.18 |
| | Age4 | 0.64 | 0.23 | 0.81 | 0.23 |
| | Age5 | 0.63 | 0.34 | 0.83 | 0.33 |
| | Age6 | 0.57 | 0.58 | 0.76 | 0.58 |

Table A.85. Gulf of Maine Atlantic cod January 1 biomass (mt) and spawning stock biomass (SSB, mt) from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models.

| Year | SAW55_3BLOCK_BASE | | SAW55_3BLOCK_BASE_M_SPLIT | |
|------|------------------------|-----------------------------|---------------------------|-----------------------------|
| | January 1 biomass (mt) | Spawning stock biomass (mt) | January 1 biomass (mt) | Spawning stock biomass (mt) |
| 1982 | 38,309 | 22,036 | 37,911 | 21,531 |
| 1983 | 28,575 | 16,343 | 28,332 | 15,989 |
| 1984 | 23,081 | 13,454 | 22,944 | 13,186 |
| 1985 | 21,526 | 12,380 | 21,459 | 12,177 |
| 1986 | 20,540 | 11,537 | 20,536 | 11,390 |
| 1987 | 19,892 | 11,211 | 19,955 | 11,130 |
| 1988 | 20,231 | 11,621 | 20,529 | 11,694 |
| 1989 | 27,593 | 15,516 | 28,621 | 15,894 |
| 1990 | 34,120 | 19,988 | 35,959 | 20,821 |
| 1991 | 28,227 | 17,253 | 29,861 | 18,062 |
| 1992 | 18,989 | 10,842 | 20,446 | 11,473 |
| 1993 | 14,097 | 7,575 | 15,633 | 8,229 |
| 1994 | 13,050 | 6,988 | 15,026 | 7,930 |
| 1995 | 13,276 | 7,975 | 15,939 | 9,442 |
| 1996 | 13,512 | 8,371 | 16,756 | 10,245 |
| 1997 | 11,364 | 7,091 | 14,959 | 9,176 |
| 1998 | 9,959 | 6,268 | 13,971 | 8,621 |
| 1999 | 10,577 | 6,812 | 15,878 | 9,778 |
| 2000 | 15,003 | 9,070 | 22,822 | 12,976 |
| 2001 | 18,755 | 11,885 | 28,082 | 17,222 |
| 2002 | 17,077 | 11,951 | 25,502 | 17,208 |
| 2003 | 14,334 | 10,005 | 20,786 | 13,966 |
| 2004 | 12,646 | 8,594 | 18,344 | 11,878 |
| 2005 | 11,038 | 7,213 | 15,800 | 9,831 |
| 2006 | 10,852 | 6,752 | 16,075 | 9,311 |
| 2007 | 14,311 | 8,725 | 20,846 | 11,693 |
| 2008 | 16,670 | 10,282 | 22,921 | 13,297 |
| 2009 | 18,506 | 11,457 | 24,493 | 14,332 |
| 2010 | 17,178 | 11,141 | 21,184 | 12,979 |
| 2011 | 14,728 | 9,903 | 16,312 | 10,221 |

Table A.86. Gulf of Maine Atlantic cod fully recruited fishing mortality (F_{full}) from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models.

| Year | SAW55_3BLOCK_BASE | SAW55_3BLOCK_BASE_M_SPLIT |
|------|----------------------------------|----------------------------------|
| | Fully recruited F (F_{full}) | Fully recruited F (F_{full}) |
| 1982 | 0.73 | 0.75 |
| 1983 | 0.87 | 0.89 |
| 1984 | 0.78 | 0.80 |
| 1985 | 0.91 | 0.93 |
| 1986 | 0.83 | 0.85 |
| 1987 | 0.82 | 0.83 |
| 1988 | 0.62 | 0.62 |
| 1989 | 0.92 | 0.93 |
| 1990 | 1.13 | 1.13 |
| 1991 | 1.26 | 1.23 |
| 1992 | 1.35 | 1.31 |
| 1993 | 1.53 | 1.46 |
| 1994 | 1.45 | 1.32 |
| 1995 | 0.99 | 0.86 |
| 1996 | 1.03 | 0.85 |
| 1997 | 0.92 | 0.72 |
| 1998 | 0.82 | 0.61 |
| 1999 | 0.48 | 0.35 |
| 2000 | 0.62 | 0.45 |
| 2001 | 0.72 | 0.51 |
| 2002 | 0.57 | 0.40 |
| 2003 | 0.67 | 0.48 |
| 2004 | 0.68 | 0.50 |
| 2005 | 0.92 | 0.70 |
| 2006 | 0.78 | 0.60 |
| 2007 | 0.75 | 0.60 |
| 2008 | 0.94 | 0.77 |
| 2009 | 0.98 | 0.83 |
| 2010 | 0.87 | 0.79 |
| 2011 | 0.86 | 0.90 |

Table A.87. Gulf of Maine Atlantic cod fishing mortality-at-age from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) model.

| Year | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------|------|------|------|------|------|------|------|------|-------|
| 1982 | 0.04 | 0.24 | 0.60 | 0.71 | 0.73 | 0.73 | 0.73 | 0.73 | 0.73 |
| 1983 | 0.05 | 0.29 | 0.71 | 0.85 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 |
| 1984 | 0.04 | 0.26 | 0.63 | 0.76 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 |
| 1985 | 0.05 | 0.30 | 0.74 | 0.89 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 |
| 1986 | 0.04 | 0.27 | 0.68 | 0.81 | 0.83 | 0.83 | 0.83 | 0.83 | 0.83 |
| 1987 | 0.04 | 0.27 | 0.67 | 0.80 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 |
| 1988 | 0.03 | 0.20 | 0.51 | 0.61 | 0.62 | 0.62 | 0.62 | 0.62 | 0.62 |
| 1989 | 0.01 | 0.08 | 0.33 | 0.71 | 0.87 | 0.91 | 0.92 | 0.92 | 0.92 |
| 1990 | 0.02 | 0.10 | 0.41 | 0.87 | 1.08 | 1.12 | 1.13 | 1.13 | 1.13 |
| 1991 | 0.02 | 0.11 | 0.46 | 0.97 | 1.20 | 1.25 | 1.26 | 1.26 | 1.26 |
| 1992 | 0.02 | 0.12 | 0.49 | 1.04 | 1.29 | 1.34 | 1.35 | 1.35 | 1.35 |
| 1993 | 0.02 | 0.13 | 0.55 | 1.18 | 1.46 | 1.52 | 1.53 | 1.53 | 1.53 |
| 1994 | 0.02 | 0.13 | 0.52 | 1.12 | 1.38 | 1.44 | 1.45 | 1.45 | 1.45 |
| 1995 | 0.02 | 0.09 | 0.36 | 0.77 | 0.95 | 0.98 | 0.99 | 0.99 | 0.99 |
| 1996 | 0.02 | 0.09 | 0.37 | 0.79 | 0.98 | 1.02 | 1.03 | 1.03 | 1.03 |
| 1997 | 0.01 | 0.08 | 0.33 | 0.71 | 0.88 | 0.91 | 0.92 | 0.92 | 0.92 |
| 1998 | 0.01 | 0.07 | 0.30 | 0.63 | 0.78 | 0.82 | 0.82 | 0.82 | 0.82 |
| 1999 | 0.01 | 0.04 | 0.17 | 0.37 | 0.46 | 0.48 | 0.48 | 0.48 | 0.48 |
| 2000 | 0.01 | 0.05 | 0.22 | 0.48 | 0.59 | 0.62 | 0.62 | 0.62 | 0.62 |
| 2001 | 0.01 | 0.06 | 0.26 | 0.55 | 0.68 | 0.71 | 0.72 | 0.72 | 0.72 |
| 2002 | 0.01 | 0.05 | 0.21 | 0.44 | 0.54 | 0.56 | 0.57 | 0.57 | 0.57 |
| 2003 | 0.01 | 0.06 | 0.24 | 0.52 | 0.64 | 0.67 | 0.67 | 0.67 | 0.67 |
| 2004 | 0.01 | 0.06 | 0.25 | 0.52 | 0.65 | 0.67 | 0.68 | 0.68 | 0.68 |
| 2005 | 0.01 | 0.03 | 0.17 | 0.56 | 0.83 | 0.90 | 0.92 | 0.92 | 0.92 |
| 2006 | 0.00 | 0.03 | 0.15 | 0.47 | 0.71 | 0.77 | 0.78 | 0.78 | 0.78 |
| 2007 | 0.00 | 0.03 | 0.14 | 0.45 | 0.68 | 0.74 | 0.75 | 0.75 | 0.75 |
| 2008 | 0.01 | 0.03 | 0.18 | 0.57 | 0.85 | 0.92 | 0.93 | 0.94 | 0.94 |
| 2009 | 0.01 | 0.03 | 0.19 | 0.59 | 0.89 | 0.96 | 0.97 | 0.98 | 0.98 |
| 2010 | 0.00 | 0.03 | 0.17 | 0.53 | 0.79 | 0.86 | 0.87 | 0.87 | 0.87 |
| 2011 | 0.00 | 0.03 | 0.16 | 0.52 | 0.78 | 0.84 | 0.86 | 0.86 | 0.86 |

Table A.88. Gulf of Maine Atlantic cod January 1 numbers-at-age (000s) from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) model. Summary statistics reported (i.e., median, mean and geometric mean) include only the years 1982-2009, which was the recruitment series used in the reference points determination and stock projections.

| Year | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------------------------------|--------|--------|--------|-------|-------|------|------|------|-------|
| 1982 | 10,579 | 12,112 | 5,124 | 2,988 | 1,708 | 157 | 222 | 148 | 232 |
| 1983 | 11,545 | 8,337 | 7,799 | 2,313 | 1,198 | 673 | 62 | 87 | 149 |
| 1984 | 11,450 | 9,033 | 5,128 | 3,143 | 809 | 411 | 231 | 21 | 81 |
| 1985 | 8,912 | 9,002 | 5,727 | 2,228 | 1,203 | 305 | 154 | 87 | 38 |
| 1986 | 14,069 | 6,958 | 5,466 | 2,236 | 750 | 397 | 100 | 51 | 41 |
| 1987 | 15,005 | 11,031 | 4,337 | 2,277 | 814 | 268 | 142 | 36 | 33 |
| 1988 | 27,950 | 11,771 | 6,901 | 1,823 | 838 | 294 | 97 | 51 | 25 |
| 1989 | 4,279 | 22,155 | 7,860 | 3,409 | 814 | 369 | 129 | 42 | 33 |
| 1990 | 4,224 | 3,453 | 16,754 | 4,623 | 1,377 | 278 | 122 | 42 | 25 |
| 1991 | 7,479 | 3,398 | 2,564 | 9,128 | 1,585 | 384 | 74 | 32 | 18 |
| 1992 | 7,445 | 6,004 | 2,494 | 1,332 | 2,826 | 390 | 90 | 17 | 12 |
| 1993 | 9,665 | 5,968 | 4,374 | 1,255 | 386 | 640 | 84 | 19 | 6 |
| 1994 | 3,254 | 7,726 | 4,280 | 2,063 | 316 | 74 | 115 | 15 | 4 |
| 1995 | 3,451 | 2,604 | 5,579 | 2,077 | 552 | 65 | 14 | 22 | 4 |
| 1996 | 2,741 | 2,782 | 1,957 | 3,193 | 791 | 176 | 20 | 4 | 8 |
| 1997 | 4,503 | 2,208 | 2,084 | 1,105 | 1,183 | 243 | 52 | 6 | 4 |
| 1998 | 3,939 | 3,634 | 1,669 | 1,223 | 445 | 402 | 80 | 17 | 3 |
| 1999 | 7,865 | 3,184 | 2,770 | 1,015 | 531 | 166 | 145 | 29 | 7 |
| 2000 | 4,693 | 6,391 | 2,500 | 1,905 | 572 | 274 | 84 | 73 | 18 |
| 2001 | 1,170 | 3,805 | 4,959 | 1,636 | 967 | 260 | 121 | 37 | 40 |
| 2002 | 5,171 | 947 | 2,927 | 3,134 | 770 | 399 | 104 | 48 | 31 |
| 2003 | 1,904 | 4,196 | 738 | 1,952 | 1,654 | 367 | 186 | 48 | 37 |
| 2004 | 6,304 | 1,542 | 3,241 | 474 | 951 | 713 | 154 | 78 | 36 |
| 2005 | 3,922 | 5,106 | 1,190 | 2,077 | 230 | 408 | 298 | 64 | 47 |
| 2006 | 6,590 | 3,195 | 4,050 | 819 | 976 | 82 | 135 | 98 | 36 |
| 2007 | 5,296 | 5,373 | 2,546 | 2,859 | 418 | 393 | 31 | 51 | 50 |
| 2008 | 4,513 | 4,319 | 4,286 | 1,808 | 1,487 | 173 | 154 | 12 | 39 |
| 2009 | 3,532 | 3,676 | 3,423 | 2,938 | 840 | 520 | 56 | 49 | 16 |
| 2010 | 2,177 | 2,876 | 2,910 | 2,329 | 1,333 | 283 | 163 | 17 | 20 |
| 2011 | 1,175 | 1,774 | 2,285 | 2,020 | 1,127 | 495 | 99 | 56 | 13 |
| 1982-2009 median recruitment | 5,234 | | | | | | | | |
| 1982-2009 mean recruitment | 7,195 | | | | | | | | |
| 1982-2009 geometric mean | 5,792 | | | | | | | | |

Table A.89. Gulf of Maine Atlantic cod fishing mortality-at-age from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) model.

| Year | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------|------|------|------|------|------|------|------|------|-------|
| 1982 | 0.04 | 0.24 | 0.61 | 0.73 | 0.75 | 0.75 | 0.75 | 0.75 | 0.75 |
| 1983 | 0.05 | 0.29 | 0.73 | 0.87 | 0.89 | 0.89 | 0.89 | 0.89 | 0.89 |
| 1984 | 0.04 | 0.26 | 0.65 | 0.78 | 0.80 | 0.80 | 0.80 | 0.80 | 0.80 |
| 1985 | 0.05 | 0.30 | 0.76 | 0.91 | 0.93 | 0.93 | 0.93 | 0.93 | 0.93 |
| 1986 | 0.04 | 0.28 | 0.69 | 0.83 | 0.85 | 0.85 | 0.85 | 0.85 | 0.85 |
| 1987 | 0.04 | 0.27 | 0.68 | 0.81 | 0.83 | 0.83 | 0.83 | 0.83 | 0.83 |
| 1988 | 0.03 | 0.20 | 0.50 | 0.60 | 0.62 | 0.62 | 0.62 | 0.62 | 0.62 |
| 1989 | 0.01 | 0.07 | 0.32 | 0.72 | 0.89 | 0.92 | 0.93 | 0.93 | 0.93 |
| 1990 | 0.01 | 0.08 | 0.39 | 0.87 | 1.08 | 1.12 | 1.13 | 1.13 | 1.13 |
| 1991 | 0.01 | 0.09 | 0.42 | 0.95 | 1.18 | 1.22 | 1.23 | 1.23 | 1.23 |
| 1992 | 0.02 | 0.10 | 0.45 | 1.01 | 1.25 | 1.30 | 1.31 | 1.31 | 1.31 |
| 1993 | 0.02 | 0.11 | 0.50 | 1.13 | 1.40 | 1.45 | 1.46 | 1.46 | 1.46 |
| 1994 | 0.02 | 0.10 | 0.45 | 1.02 | 1.26 | 1.31 | 1.32 | 1.32 | 1.32 |
| 1995 | 0.01 | 0.06 | 0.29 | 0.67 | 0.82 | 0.86 | 0.86 | 0.86 | 0.86 |
| 1996 | 0.01 | 0.06 | 0.29 | 0.66 | 0.82 | 0.85 | 0.85 | 0.85 | 0.85 |
| 1997 | 0.01 | 0.05 | 0.25 | 0.56 | 0.69 | 0.72 | 0.72 | 0.72 | 0.72 |
| 1998 | 0.01 | 0.04 | 0.21 | 0.47 | 0.58 | 0.60 | 0.60 | 0.61 | 0.61 |
| 1999 | 0.00 | 0.03 | 0.12 | 0.27 | 0.33 | 0.35 | 0.35 | 0.35 | 0.35 |
| 2000 | 0.01 | 0.03 | 0.15 | 0.35 | 0.43 | 0.44 | 0.45 | 0.45 | 0.45 |
| 2001 | 0.01 | 0.04 | 0.17 | 0.39 | 0.49 | 0.50 | 0.51 | 0.51 | 0.51 |
| 2002 | 0.00 | 0.03 | 0.14 | 0.31 | 0.38 | 0.40 | 0.40 | 0.40 | 0.40 |
| 2003 | 0.01 | 0.04 | 0.17 | 0.37 | 0.46 | 0.48 | 0.48 | 0.48 | 0.48 |
| 2004 | 0.01 | 0.04 | 0.17 | 0.39 | 0.48 | 0.50 | 0.50 | 0.50 | 0.50 |
| 2005 | 0.00 | 0.02 | 0.12 | 0.41 | 0.64 | 0.69 | 0.70 | 0.70 | 0.70 |
| 2006 | 0.00 | 0.02 | 0.10 | 0.35 | 0.55 | 0.60 | 0.60 | 0.60 | 0.60 |
| 2007 | 0.00 | 0.02 | 0.10 | 0.35 | 0.55 | 0.59 | 0.60 | 0.60 | 0.60 |
| 2008 | 0.00 | 0.02 | 0.13 | 0.45 | 0.70 | 0.76 | 0.77 | 0.77 | 0.77 |
| 2009 | 0.00 | 0.02 | 0.14 | 0.49 | 0.76 | 0.82 | 0.83 | 0.83 | 0.83 |
| 2010 | 0.00 | 0.02 | 0.13 | 0.47 | 0.72 | 0.78 | 0.79 | 0.79 | 0.79 |
| 2011 | 0.00 | 0.02 | 0.15 | 0.53 | 0.82 | 0.88 | 0.89 | 0.90 | 0.90 |

Table A.90. Gulf of Maine Atlantic cod January 1 numbers-at-age (000s) from 1982 to 2011 as estimated from the ASAP SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) model. Summary statistics reported (i.e., median, mean and geometric mean) include only the years 1982-2009, which was the recruitment series used in the reference points determination and stock projections.

| Year | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9+ |
|------------------------------|--------|--------|--------|-------|-------|------|------|------|-------|
| 1982 | 10,904 | 12,271 | 5,154 | 2,969 | 1,670 | 153 | 213 | 141 | 218 |
| 1983 | 11,913 | 8,594 | 7,873 | 2,293 | 1,170 | 647 | 59 | 82 | 139 |
| 1984 | 11,755 | 9,322 | 5,263 | 3,119 | 786 | 393 | 217 | 20 | 74 |
| 1985 | 9,226 | 9,242 | 5,886 | 2,251 | 1,172 | 290 | 145 | 80 | 35 |
| 1986 | 14,398 | 7,205 | 5,585 | 2,256 | 741 | 378 | 93 | 47 | 37 |
| 1987 | 15,699 | 11,292 | 4,478 | 2,296 | 808 | 261 | 133 | 33 | 29 |
| 1988 | 30,113 | 12,322 | 7,054 | 1,864 | 835 | 289 | 93 | 47 | 22 |
| 1989 | 4,771 | 23,893 | 8,250 | 3,492 | 835 | 369 | 127 | 41 | 31 |
| 1990 | 4,931 | 3,824 | 18,087 | 4,867 | 1,381 | 278 | 119 | 41 | 23 |
| 1991 | 9,056 | 3,865 | 2,797 | 9,773 | 1,618 | 373 | 72 | 31 | 16 |
| 1992 | 9,558 | 7,019 | 2,777 | 1,445 | 2,973 | 393 | 87 | 17 | 11 |
| 1993 | 13,358 | 7,328 | 4,964 | 1,382 | 409 | 662 | 83 | 18 | 6 |
| 1994 | 4,846 | 10,020 | 5,023 | 2,299 | 341 | 77 | 118 | 15 | 4 |
| 1995 | 5,602 | 3,605 | 6,872 | 2,418 | 627 | 73 | 16 | 24 | 4 |
| 1996 | 4,862 | 4,149 | 2,532 | 3,830 | 930 | 206 | 23 | 5 | 9 |
| 1997 | 8,555 | 3,530 | 2,858 | 1,387 | 1,454 | 302 | 65 | 7 | 4 |
| 1998 | 7,851 | 6,159 | 2,431 | 1,621 | 577 | 529 | 107 | 23 | 4 |
| 1999 | 16,584 | 5,603 | 4,235 | 1,421 | 730 | 232 | 209 | 42 | 11 |
| 2000 | 10,157 | 11,638 | 3,849 | 2,649 | 765 | 369 | 116 | 104 | 26 |
| 2001 | 2,553 | 7,048 | 7,856 | 2,304 | 1,308 | 348 | 165 | 52 | 58 |
| 2002 | 11,472 | 1,753 | 4,690 | 4,562 | 1,075 | 556 | 145 | 69 | 46 |
| 2003 | 4,316 | 7,730 | 1,152 | 2,770 | 2,269 | 497 | 253 | 66 | 52 |
| 2004 | 14,342 | 2,877 | 5,001 | 655 | 1,279 | 959 | 206 | 105 | 49 |
| 2005 | 8,744 | 9,556 | 1,859 | 2,825 | 298 | 531 | 391 | 84 | 62 |
| 2006 | 14,456 | 5,845 | 6,287 | 1,110 | 1,257 | 106 | 179 | 131 | 49 |
| 2007 | 11,031 | 9,668 | 3,856 | 3,814 | 522 | 486 | 39 | 66 | 66 |
| 2008 | 8,695 | 7,378 | 6,377 | 2,340 | 1,797 | 203 | 180 | 14 | 48 |
| 2009 | 6,254 | 5,812 | 4,844 | 3,762 | 996 | 595 | 63 | 56 | 19 |
| 2010 | 3,511 | 4,179 | 3,810 | 2,830 | 1,547 | 313 | 176 | 19 | 22 |
| 2011 | 1,749 | 2,347 | 2,742 | 2,240 | 1,190 | 503 | 96 | 53 | 12 |
| 1982-2009 median recruitment | 9,392 | | | | | | | | |
| 1982-2009 mean recruitment | 10,214 | | | | | | | | |
| 1982-2009 geometric mean | 9,007 | | | | | | | | |

Table A.91. Summary of the Gulf of Maine Atlantic cod 2011 point estimates and their corresponding 90% probability intervals for the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models.

| Model | SSB ₂₀₁₁ (mt) | B ₂₀₁₁ (mt) | F _{full} |
|--------------------------|--------------------------|--------------------------|--------------------|
| SAW55_3BLOCK_BASE | 9,903 (7,644 - 13,503) | 14,728 (11,890 - 19,149) | 0.86 (0.53 - 1.05) |
| SAW55_3BLOCK_BASE_MSPLIT | 10,221 (7,943 - 13,676) | 16,312 (13,173 - 20,771) | 0.90 (0.57 - 1.09) |

Table A.92. Summary of Gulf of Maine Atlantic cod ASAP model configurations which explored assessment starting years of 1932 with Beverton-Holt stock recruit functions fit internally within the model. The two model configurations that were explored were the SAW55_3BLOCK_BASE_1932_F2N1_NOPRIOR_BH ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NOPRIOR_BH (M -ramp).

| Model | SAW55_3BLOCK_BASE_1932_F2N1_NOPRIOR_BH | SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NOPRIOR_BH |
|----------------------------------|---|---|
| Starting year | 1932 | 1932 |
| Parameters | 186 | 186 |
| Objective function | 3626 | 3643 |
| Maximum gradient | 2.0E-04 | 3.3E-04 |
| Components of objective function | Suvey age comps | 814 |
| | Catch age comps | 390 |
| | Index fit | 1146 |
| | Catch fit | 651 |
| | Recruit devs | 625 |
| RMSE | Catch | 0.409 |
| | Index1 | 1.13 |
| | Index2 | 1.11 |
| | Index3 | 1.27 |
| | Index total | 1.16 |
| | Recruit devs | 1.06 |
| Mean age RMSE | Fleet1 | 0.99 |
| | Index1 | 0.97 |
| | Index2 | 1.15 |
| | Index3 | 1.08 |
| SSBStart (mt) | | 13,382 |
| SSB ₁₉₈₂ (mt) | | 23,715 |
| SSB ₂₀₁₁ (mt) | | 9,316 |
| FMULT, 2011 | | 0.92 |
| Mohn's rho (5 year peel) | SSB | 0.11 |
| | F _{mult} | -0.09 |
| | Age 1 N | 0.17 |
| | Alpha | 8218.1 |
| | Beta | 4949.0 |
| | SSB0 (mt) | 165,840 |
| | R0 (000s) | 7,980 |
| | Steepness | 0.90 (0.05) |
| | Fmsy | 0.26 (0.04) |
| | SSBm _{sy} (mt) | 42,769 (0.13) |
| MSY (mt) | | 11,081 (0.13) |
| Comments | 1932-1981 selectivity block fixed at A50=2 and slope up=0.6 | 1932-1981 selectivity block fixed at A50=2 and slope up=0.6 |

Table A.93. Inputs to the Gulf of Maine Atlantic cod yield per recruit (YPR) analysis for the ASAP $M = 0.2$ (ASAP SAW55_3BLOCK_BASE) and M -ramp (ASAP_3BLOCK_BASE_M_SPLIT) models.

| Age | Catch weights (kg) | Stock weights (kg) | Fishery selectivity ($M = 0.2$) | Fishery selectivity (M -ramp) | Fraction mature | Natural mortality |
|-----|--------------------|--------------------|-----------------------------------|----------------------------------|-----------------|-------------------|
| 1 | 0.31 | 0.12 | 0.01 | 0.00 | 0.08 | 0.20 |
| 2 | 1.01 | 0.52 | 0.03 | 0.03 | 0.26 | 0.20 |
| 3 | 2.07 | 1.26 | 0.19 | 0.17 | 0.59 | 0.20 |
| 4 | 3.07 | 2.19 | 0.61 | 0.59 | 0.84 | 0.20 |
| 5 | 3.79 | 3.12 | 0.91 | 0.91 | 0.95 | 0.20 |
| 6 | 4.55 | 3.82 | 0.98 | 0.99 | 0.99 | 0.20 |
| 7 | 5.79 | 4.77 | 1.00 | 1.00 | 1.00 | 0.20 |
| 8 | 7.56 | 6.55 | 1.00 | 1.00 | 1.00 | 0.20 |
| 9 | 12.49 | 12.50 | 1.00 | 1.00 | 1.00 | 0.20 |

Table A.94. Yield per recruit proxy reference points for Gulf of Maine Atlantic cod under both the $M = 0.2$ (ASAP SAW55_3BLOCK_BASE) and M -ramp (ASAP_3BLOCK_BASE_M_SPLIT) models.

| Model | F_{MSY} (proxy) | F_{MSY} | SSB_{MSY} proxy (mt) | MSY proxy (mt) | Median age1 recruitment | SSB hinge (mt) | Hinge year |
|---------------------------------------|----------------------|-----------|--------------------------|--------------------------|----------------------------|------------------|---------------|
| $M = 0.2$ (SAW55_3BLOCK_BASE) | $F_{40\%}$ | 0.18 | 54,743 (40,207 - 73,354) | 9,399 (6,806 - 13,153) | 5,254 | 6,300 | 1998 |
| M -ramp (SAW55_3BLOCK_BASE_M_SPLIT) | $F_{40\%}$ | 0.18 | 80,200 (64,081 - 99,972) | 13,786 (10,900 - 17,329) | 9,446 | 7,900 | 1994 |

Table A.95. Short-term projections (3 years) for Gulf of Maine Atlantic cod under an assumed harvest of 75% F_{MSY} based on the ASAP $M = 0.2$ and M -ramp models. The M-ramp projections were conducted under two assumptions of natural mortality: 0.2 and 0.4.

*Note, the projections have not been adjusted for retrospective bias.

| Year | Input | ASAP, 1982 BASE | | | ASAP, 1982 M-RAMP | | | | | |
|------|---------------|---|-----------------------------|------------|---|-----------------------------|------------|---|-----------------------------|------------|
| | | M=0.2 | | | M=0.2 | | | M=0.4 | | |
| | | F _{msy} = 0.18, B _{msy} = 54,743 mt | | | F _{msy} = 0.18, B _{msy} = 80,200 mt | | | F _{msy} = 0.18, B _{msy} = 80,200 mt | | |
| | | Rebuild year at 75% F_{MSY} = 2022 | | | Rebuild year at 75% F_{MSY} = 2022 | | | NO REBUILD at 75% F_{MSY} | | |
| | | Catch (mt) | Spawning stock biomass (mt) | F_{full} | Catch (mt) | Spawning stock biomass (mt) | F_{full} | Catch (mt) | Spawning stock biomass (mt) | F_{full} |
| 2011 | Model result | 6,830 | 9,903 | 0.86 | 6,830 | 10,221 | 0.90 | 6,830 | 10,221 | 0.90 |
| 2012 | Assumed catch | 3,767 | 8,995 | 0.46 | 3,767 | 8,196 | 0.52 | 3,767 | 7,711 | 0.58 |
| 2013 | Projection | 1,249 | 9,406 | 0.14 | 1,142 | 9,163 | 0.14 | 822 | 6,927 | 0.14 |
| 2014 | Projection | 1,503 | 12,143 | 0.14 | 1,563 | 13,916 | 0.14 | 935 | 8,875 | 0.14 |
| 2015 | Projection | 2,030 | 16,802 | 0.14 | 2,582 | 22,124 | 0.14 | 1,313 | 12,234 | 0.14 |

Table A.96. Results of consequence analysis of Gulf of Maine cod; column and row headers indicate ‘true’ state of nature and basis of management action (75% F_{MSY} for 2013 – 2015) under assumed states of nature; cells provide projections of SSB and fully recruited fishing mortality for ‘true’ states of nature for catch set according to assumed state of nature; diagonals (shaded) indicate that management actions were correctly specified for the state of nature.

| Management actions - catches in 2013-2015 | Year | Input | ASAP, 1982 start, M=0.2 | | | ASAP, 1982 start, M-ramp (project M=0.2) | | | ASAP, 1982 start, M-ramp (project M=0.4) | | |
|---|------|---------------|---|----------|-------------------|--|----------|-------------------|--|----------|-------------------|
| | | | SSB _{MSY} = 54,743 mt; MSY=9,399 mt; F _{MSY} = 0.18 | | | SSB _{MSY} = 80,200 mt; MSY=13,786 mt; F _{MSY} = 0.18 | | | SSB _{MSY} = 80,200 mt; MSY=13,786 mt; F _{MSY} = 0.18 | | |
| | | | Catch (mt) | SSB (mt) | F _{full} | Catch (mt) | SSB (mt) | F _{full} | Catch (mt) | SSB (mt) | F _{full} |
| ASAP, 1982 start, M=0.2 | 2011 | Result | 6,830 | 9,903 | 0.86 | 6,830 | 10,221 | 0.90 | 6,830 | 10,221 | 0.90 |
| | 2012 | Assumed catch | 3,767 | 8,995 | 0.46 | 3,767 | 8,195 | 0.52 | 3,767 | 7,711 | 0.58 |
| | 2013 | Projection | 1,249 | 9,406 | 0.14 | 1,249 | 9,137 | 0.15 | 1,249 | 6,834 | 0.21 |
| | 2014 | Projection | 1,503 | 12,143 | 0.14 | 1,503 | 13,825 | 0.13 | 1,503 | 8,432 | 0.24 |
| | 2015 | Projection | 2,030 | 16,802 | 0.14 | 2,030 | 22,210 | 0.11 | 2,030 | 11,428 | 0.23 |
| ASAP, 1982 start, M-ramp (project M=0.2) | 2011 | Result | 6,830 | 9,903 | 0.86 | 6,830 | 10,221 | 0.90 | 6,830 | 10,221 | 0.90 |
| | 2012 | Assumed catch | 3,767 | 8,995 | 0.46 | 3,767 | 8,196 | 0.52 | 3,767 | 7,711 | 0.58 |
| | 2013 | Projection | 1,142 | 9,425 | 0.12 | 1,142 | 9,163 | 0.14 | 1,142 | 6,858 | 0.19 |
| | 2014 | Projection | 1,563 | 12,221 | 0.14 | 1,563 | 13,916 | 0.14 | 1,563 | 8,498 | 0.24 |
| | 2015 | Projection | 2,582 | 16,800 | 0.17 | 2,582 | 22,124 | 0.14 | 2,582 | 11,344 | 0.30 |
| ASAP, 1982 start, M-ramp (project M=0.4) | 2011 | Result | 6,830 | 9,903 | 0.86 | 6,830 | 10,221 | 0.90 | 6,830 | 10,221 | 0.90 |
| | 2012 | Assumed catch | 3,767 | 8,995 | 0.46 | 3,767 | 8,195 | 0.52 | 3,767 | 7,711 | 0.58 |
| | 2013 | Projection | 822 | 9,493 | 0.09 | 822 | 9,226 | 0.10 | 822 | 6,927 | 0.14 |
| | 2014 | Projection | 935 | 12,645 | 0.08 | 935 | 14,319 | 0.08 | 935 | 8,875 | 0.14 |
| | 2015 | Projection | 1,313 | 17,969 | 0.08 | 1,313 | 23,276 | 0.06 | 1,313 | 12,234 | 0.14 |

Table A.97. Status of 2013 spawning stock biomass and fishing mortality of Gulf of Maine cod; column and row headings indicate ‘true’ state of nature and basis of management action respectively; cells indicate 2013 stock status resulting from application of management actions under assumed state of nature (rows) to ‘true’ state of nature.

| Management actions - catches in 2013-2015 | States of Nature | | |
|---|--|--|--|
| | ASAP, 1982 start, M=0.2 | ASAP, 1982 start, M-ramp (project M=0.2) | ASAP, 1982 start, M-ramp (project M=0.4) |
| ASAP, 1982 start, M=0.2 | Overfished, overfishing is not occurring | Overfished, overfishing is not occurring | Overfished, overfishing is occurring |
| ASAP, 1982 start, M-ramp (project M=0.2) | Overfished, overfishing is not occurring | Overfished, overfishing is not occurring | Overfished, overfishing is occurring |
| ASAP, 1982 start, M-ramp (project M=0.4) | Overfished, overfishing is not occurring | Overfished, overfishing is not occurring | Overfished, overfishing is not occurring |

Figures

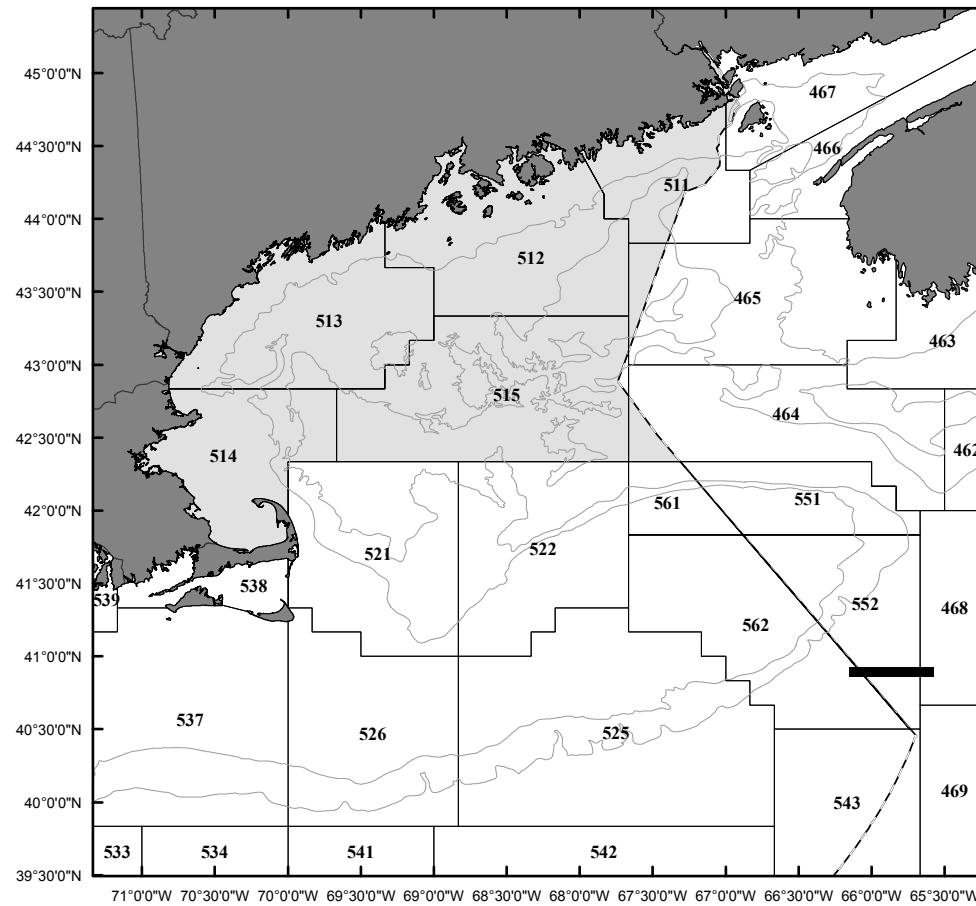


Figure A.1. Map of the Gulf of Maine Atlantic cod (*Gadus morhua*) management and assessment area (shaded grey). The United States exclusive economic zone (EEZ) is defined by the dashed line. Within the Gulf of Maine region, this line is informally referred to as the “Hague Line”.

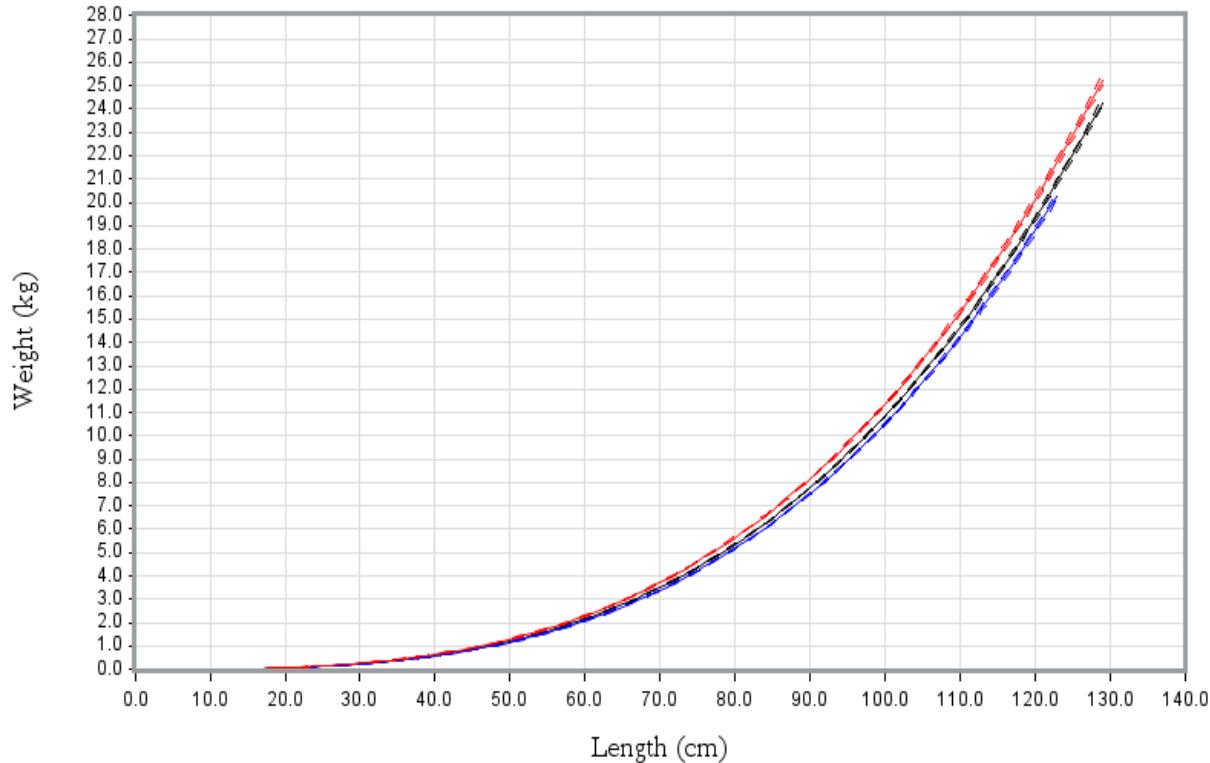
LW relationship: Cod, GOM

1992 - 2010

Annual: alpha=0.000005132, beta=3.1625

Spring: alpha=0.000004714, beta=3.1741

Fall: alpha=0.000006178, beta=3.1322



*Dashed lines indicate 95% confidence intervals

Figure A.2. Gulf of Maine Atlantic cod seasonal and annual length-weight relationships as estimated from NEFSC bottom trawl survey data.

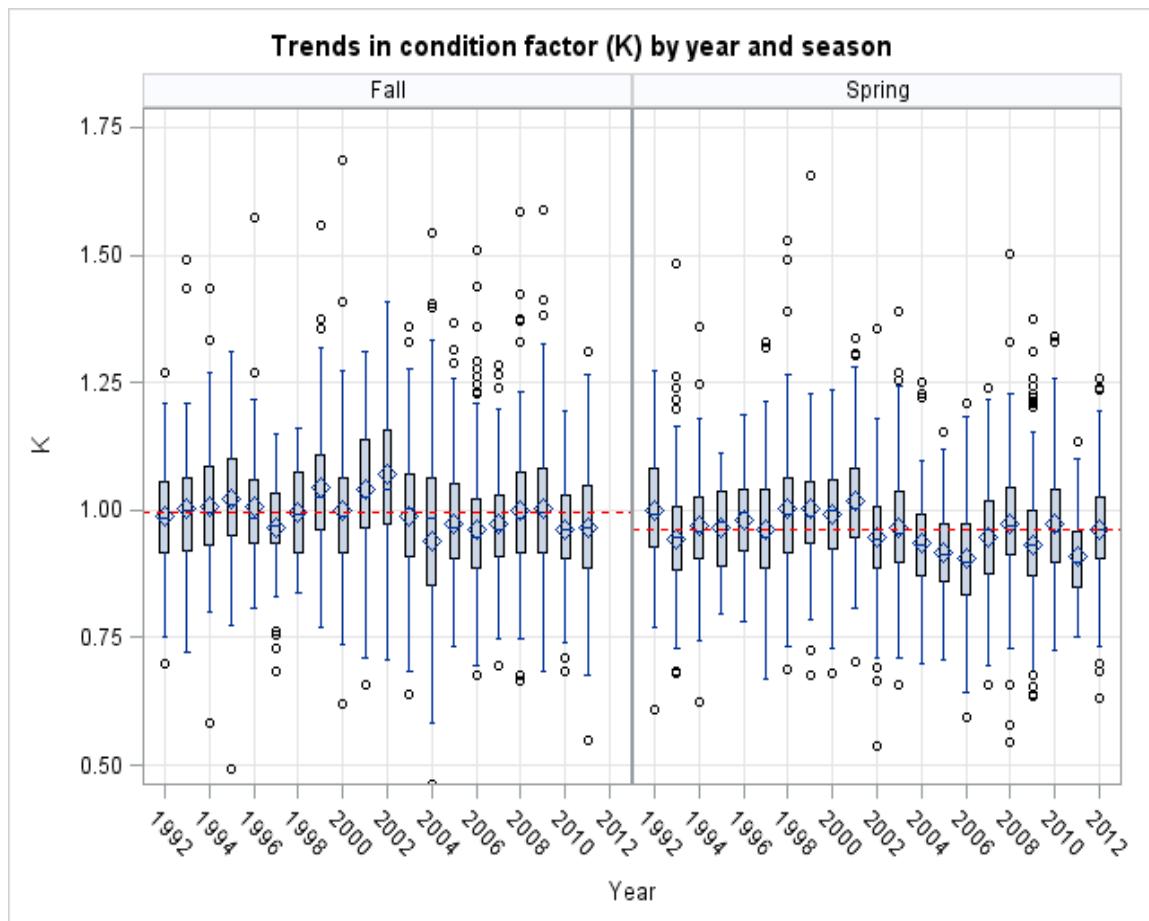


Figure A.3. Annual trends in the seasonal condition factor of Gulf of Maine Atlantic cod based on length and weight data collected from the NEFSC bottom trawl survey.

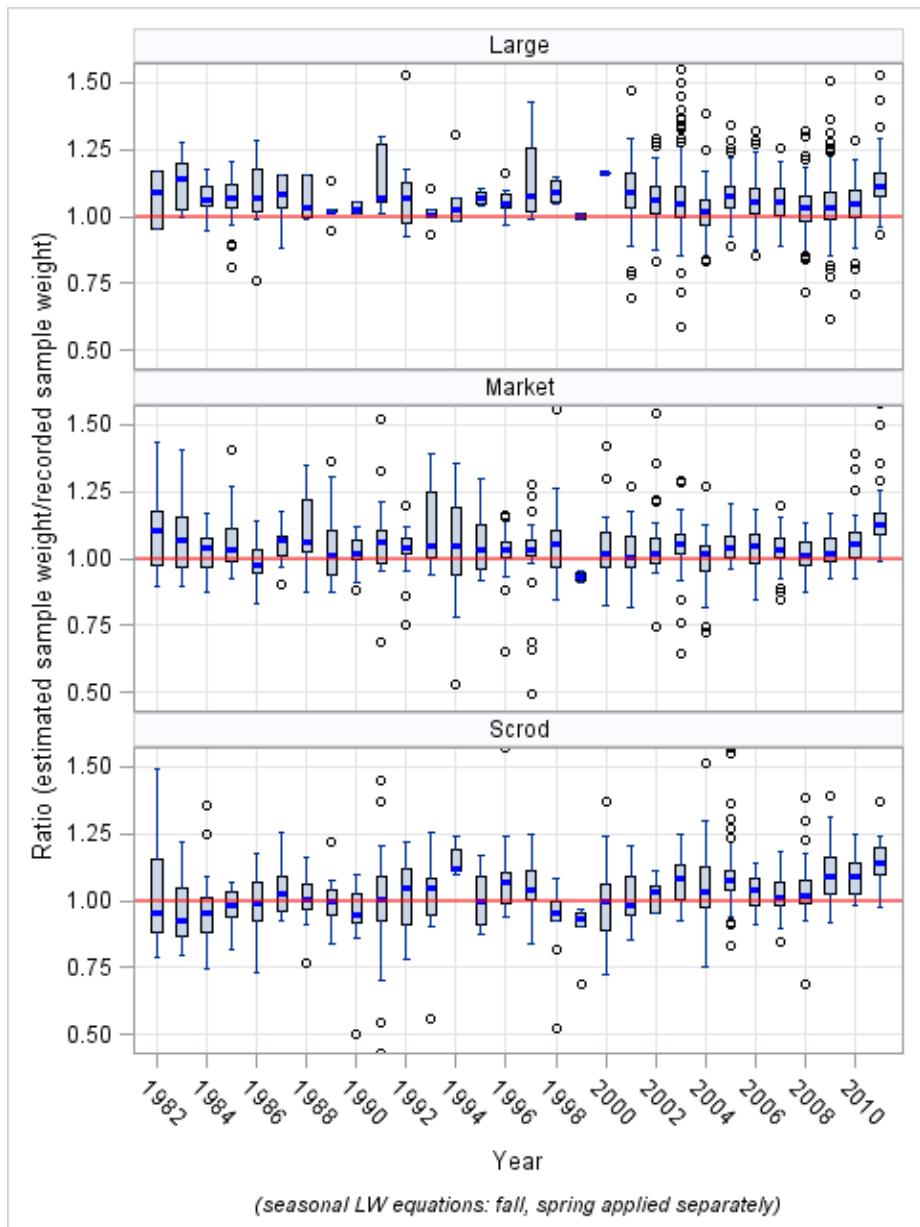


Figure A.4. Distribution of the ratios of estimate commercial biological sample weights to the recorded sample weight by market category and year using the established gutted-to-live conversion factor of 1.17. Estimated sample weights were obtained by applying the season (spring, fall) length weight equation to the recorded length distribution of the sample. The solid red line indicates the 1.0 equality line.

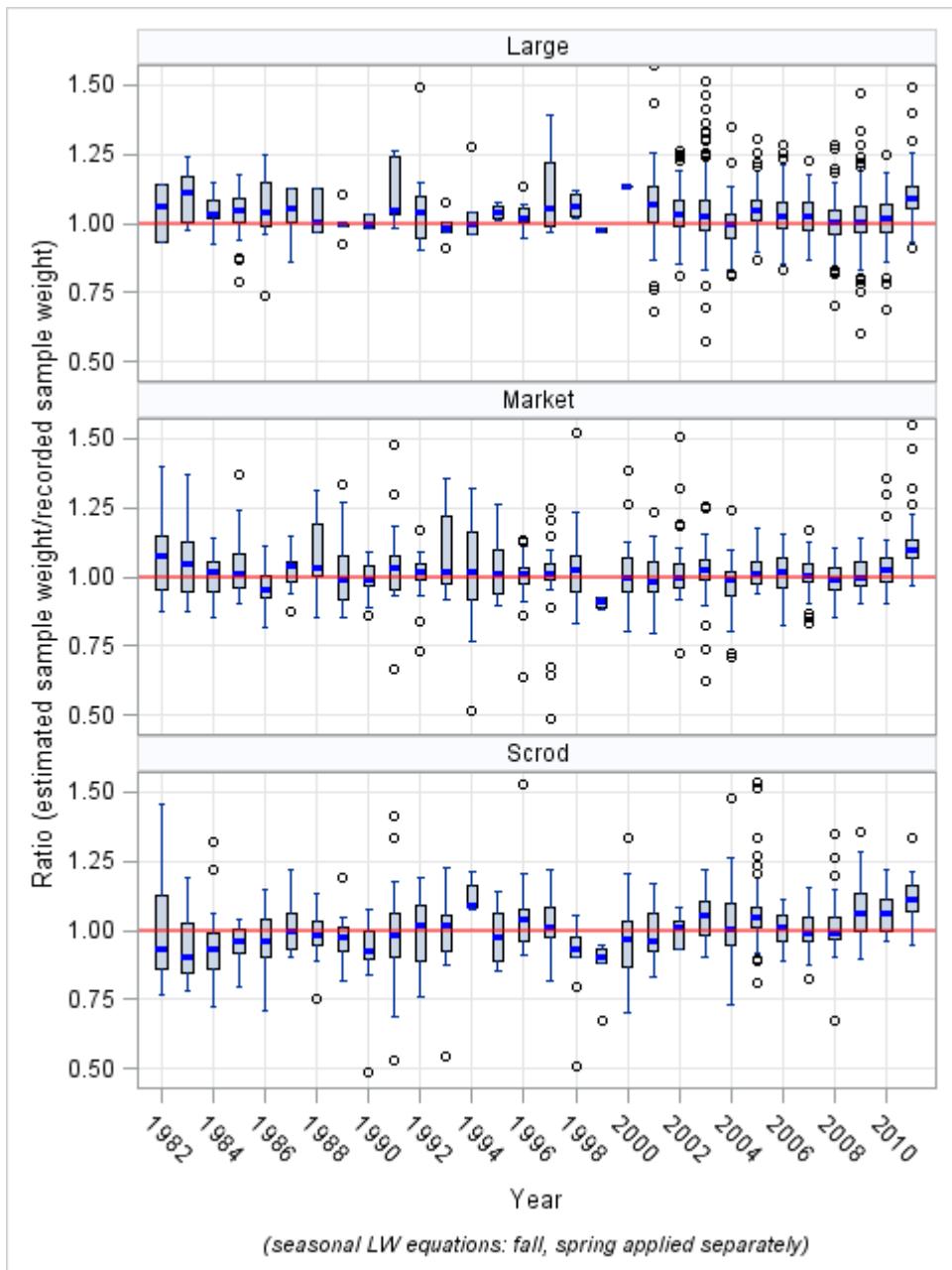


Figure A.5. Distribution of the ratios of estimate commercial biological sample weights to the recorded sample weight by market category and year using a preliminary gutted-to-live conversion factor of 1.20. Estimated sample weights were obtained by applying the season (spring, fall) length weight equation to the recorded length distribution of the sample. The solid red line indicates the 1.0 equality line.

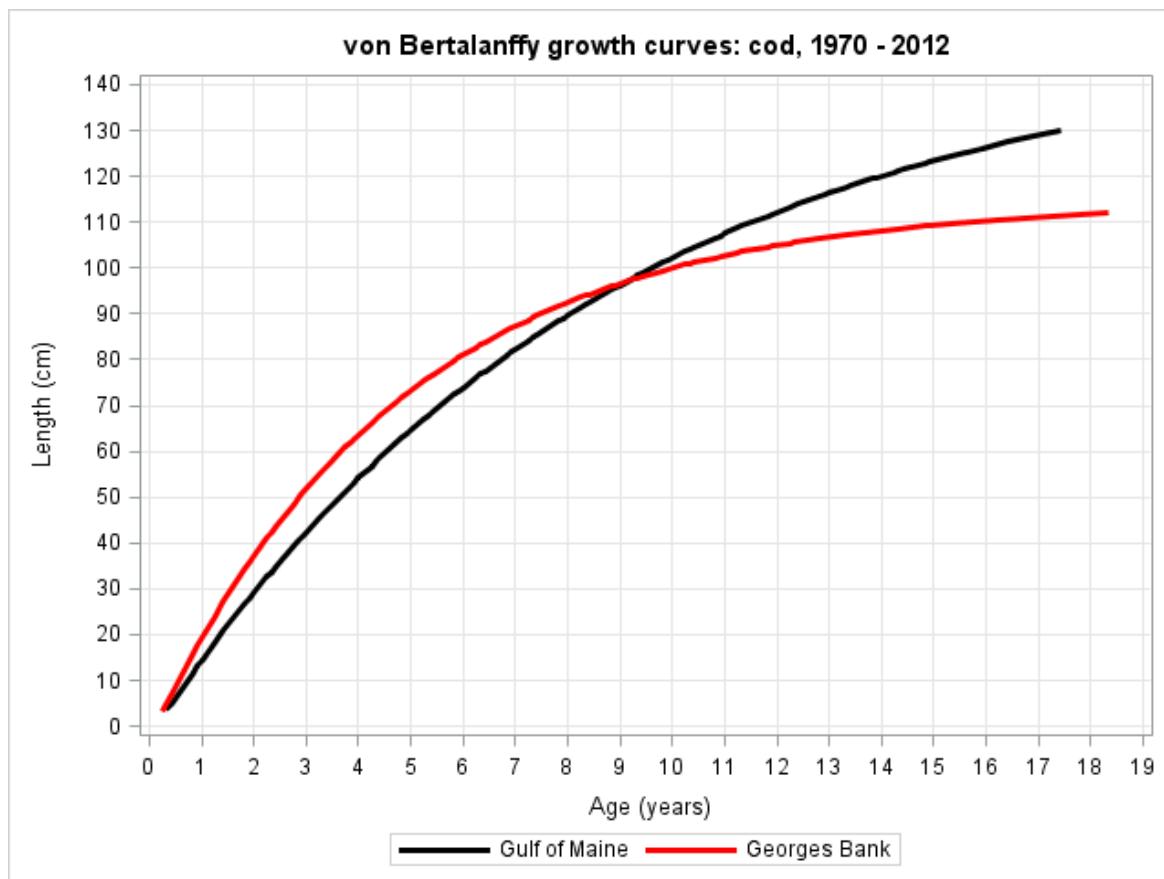


Figure A.6. Comparison of von Bertalanffy growth curves for the Gulf of Maine and Georges Banks Atlantic cod stocks as estimated from data collected from the Northeast Fisheries Science Center bottom trawl survey between 1970 and 2011.

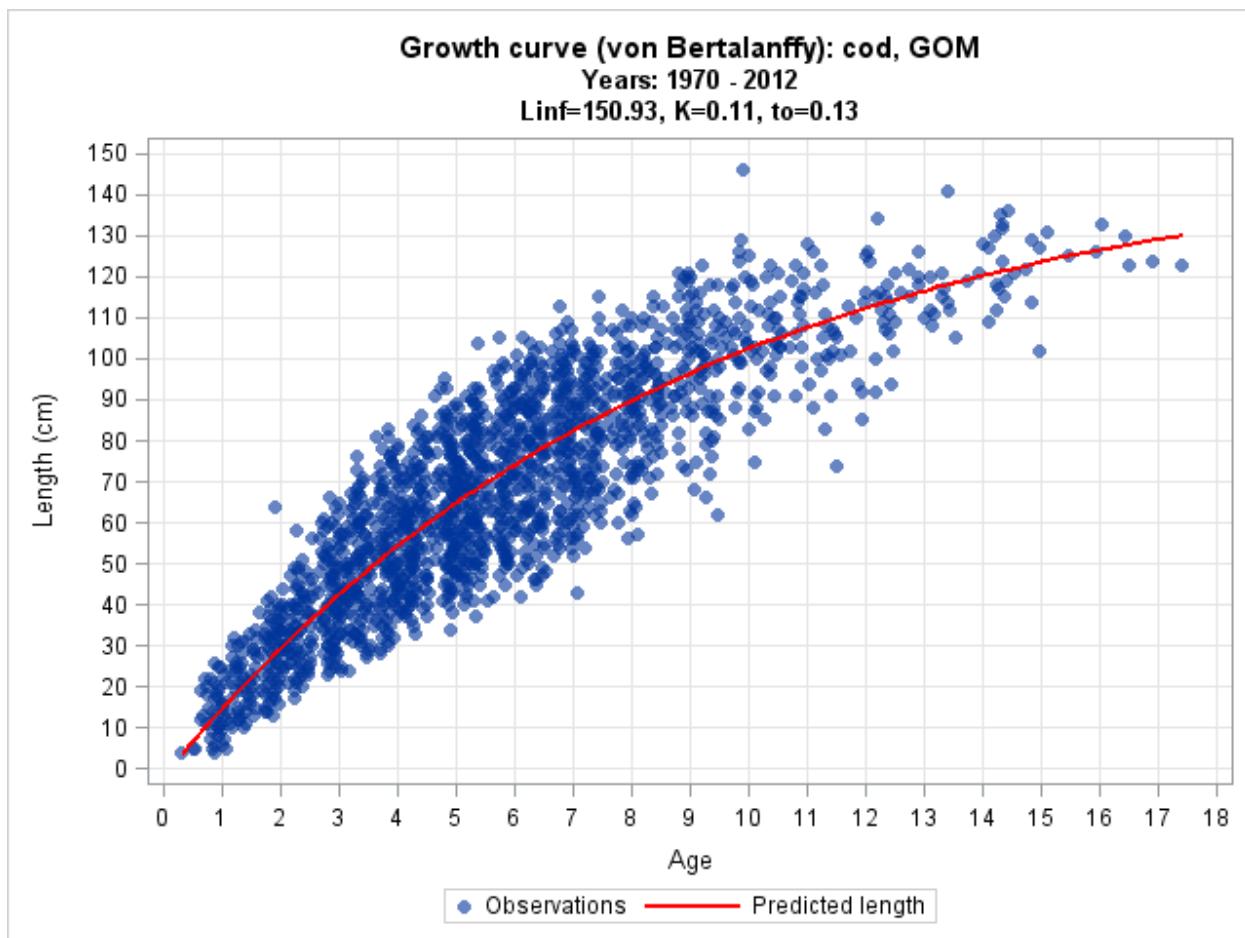


Figure A.7. Gulf of Maine Atlantic cod von Bertalanffy growth curve estimated from data collected from the Northeast Fisheries Science Center bottom trawl survey between 1970 and 2011. Growth parameters estimated for the Gulf of Maine stock were: $L_{inf}=150.93$ cm, $K=0.11$, $t_0=0.13$.

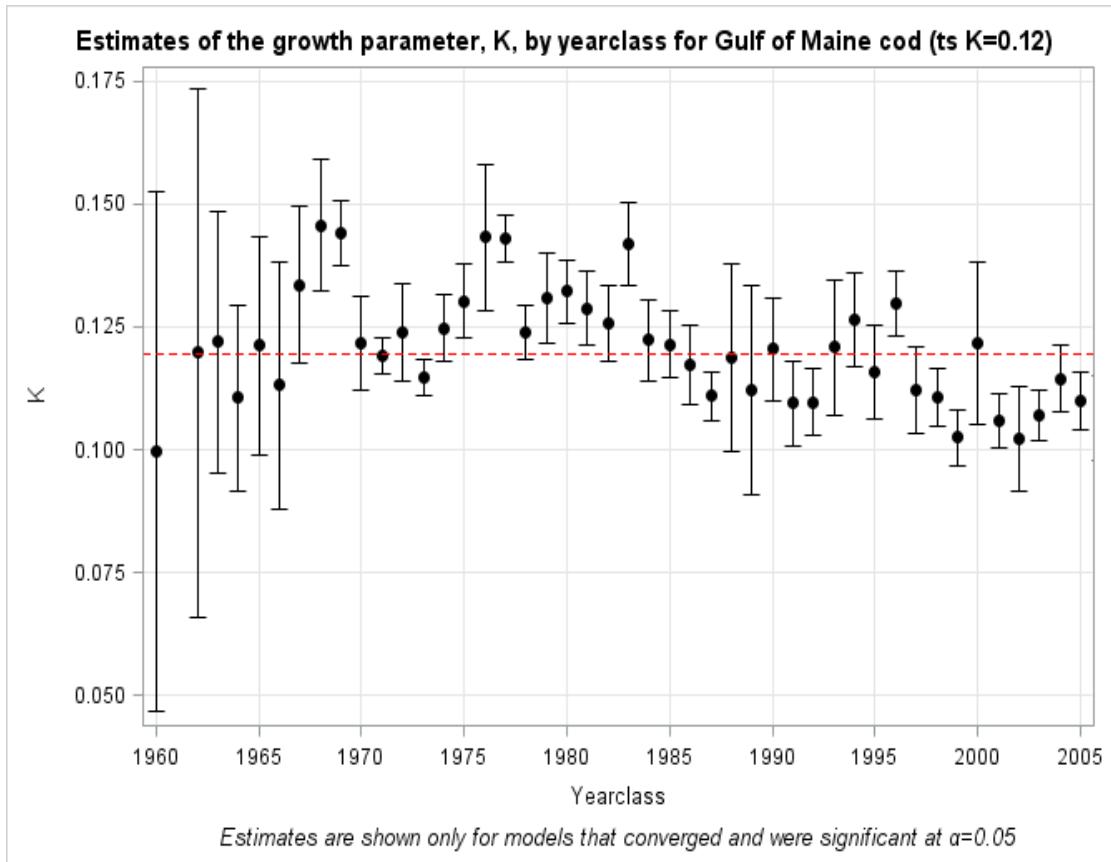


Figure A.8. Trends in the growth parameter , K , by year class for Gulf of Maine Atlantic cod. Errors bars indicate 95% confidence intervals about the parameter estimate. The dashed red line corresponds to the average cohort K estimate for the entire time series (0.12).

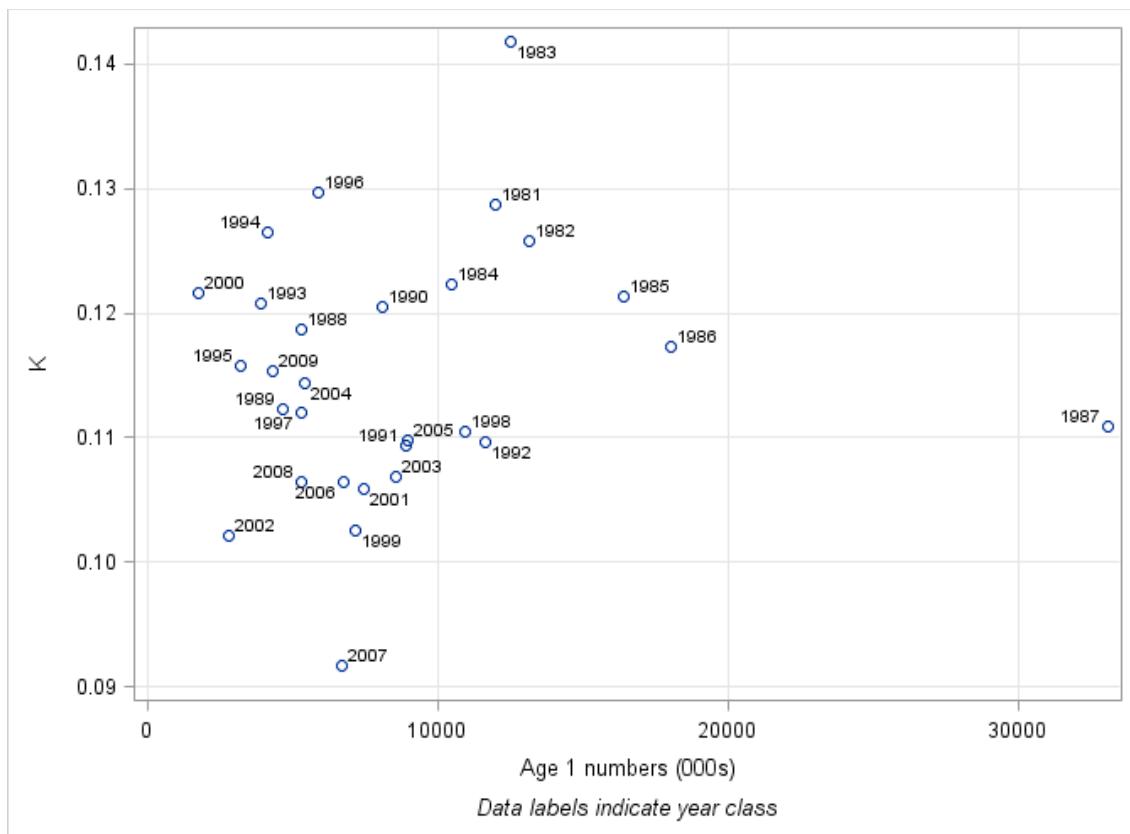


Figure A.9. Scatter plot of the growth parameter, K , relative to year class strength (age 1 numbers) for Gulf of Maine Atlantic cod. Age 1 recruitment estimates are based on the 2011 assessment of the Gulf of Maine cod stock (NEFSC 2012).

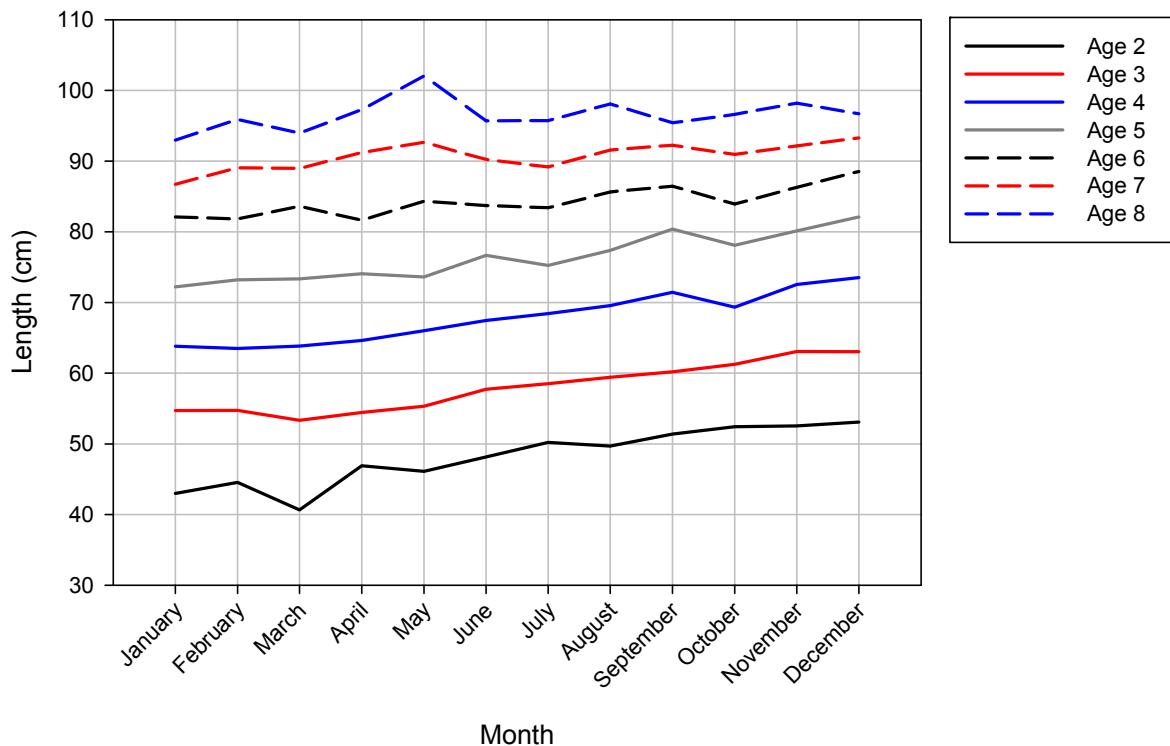


Figure A.10. Mean length-at-age of Altantic cod by month. Estimated from commercial port samples taken between 1981 and 2010.

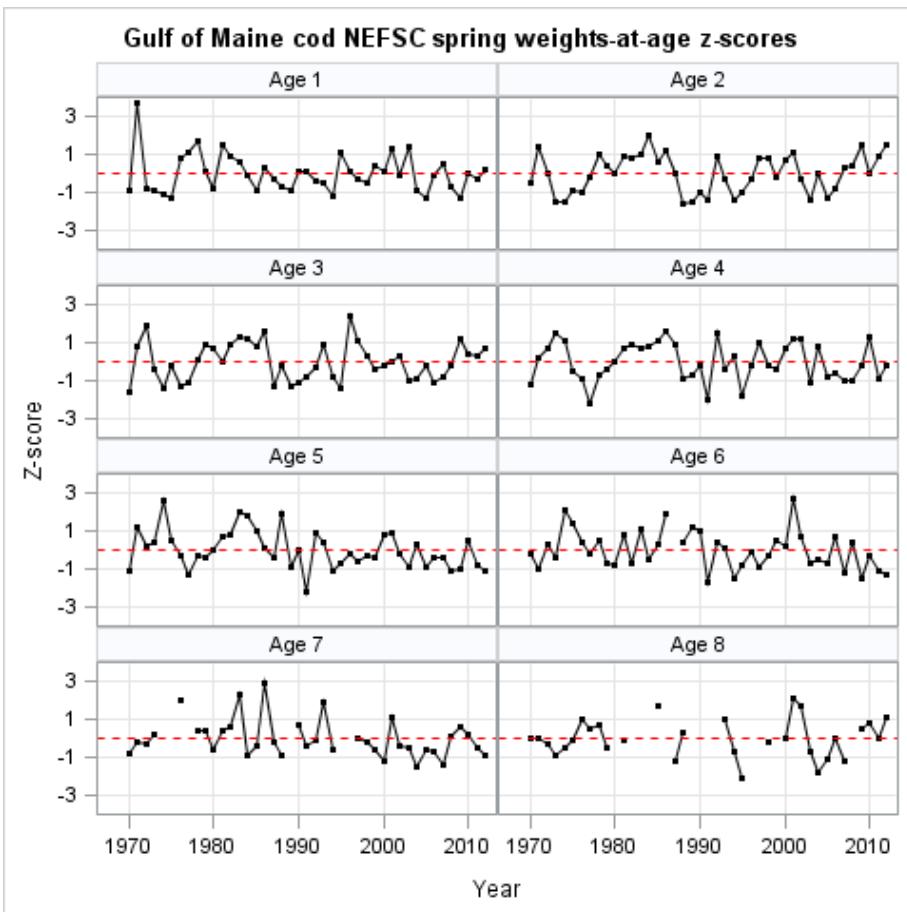


Figure A.11 Average spring survey weights-at-age of Gulf of Maine Atlantic cod ages 1-8 from 1982 to 2012. Survey weights are based on the average weight-at-age of cod sampled from the Northeast Fisheries Science Center spring bottom trawl survey. Average weights are presented as z-scores ($[x-\mu]/\sigma$).

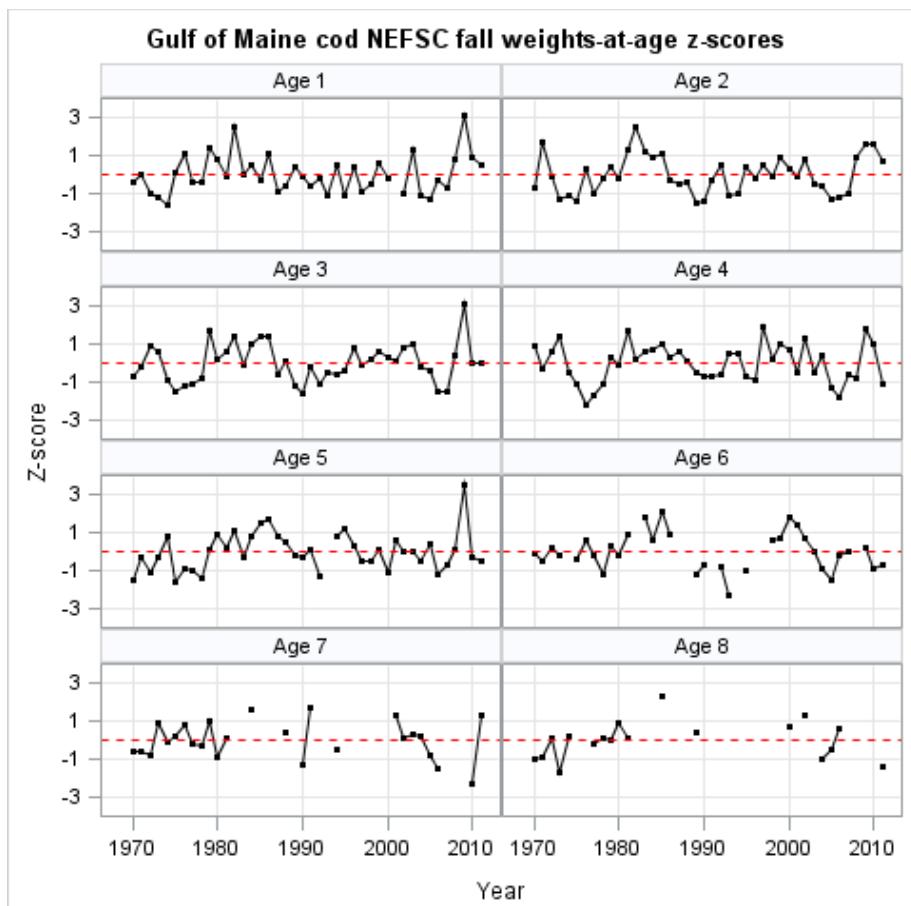


Figure A.12. Average fall survey weights-at-age of Gulf of Maine Atlantic cod ages 1-8 from 1982 to 2011. Survey weights are based on the average weight-at-age of cod sampled from the Northeast Fisheries Science Center fall bottom trawl survey. Average weights are presented as z-scores ($[x-\mu]/\sigma$).

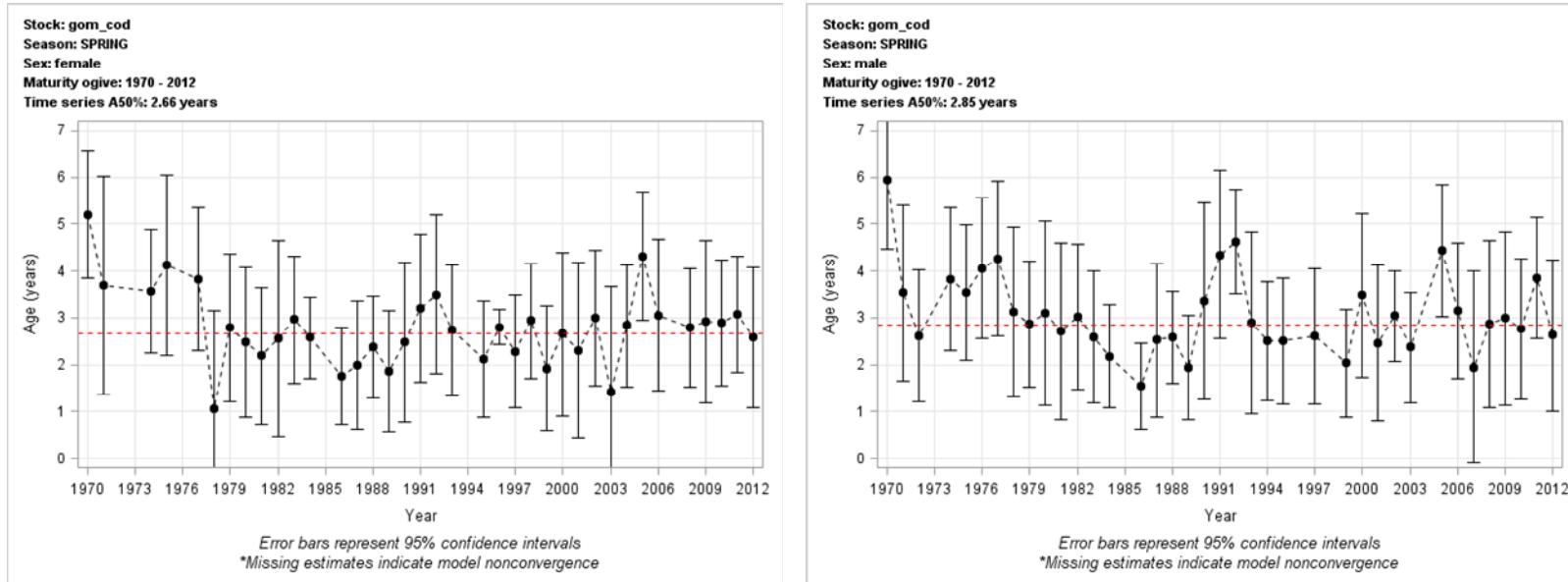


Figure A.13. Annual average age-at-50% maturity ($A_{50\%}$) and corresponding 95% confidence intervals for female (left panels) and male (right panels) Gulf of Maine Atlantic cod from 1970 to 2012. Average maturity has been estimated from data collected from the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey. Years in which the $A_{50\%}$ could not be estimated are omitted from the plots.

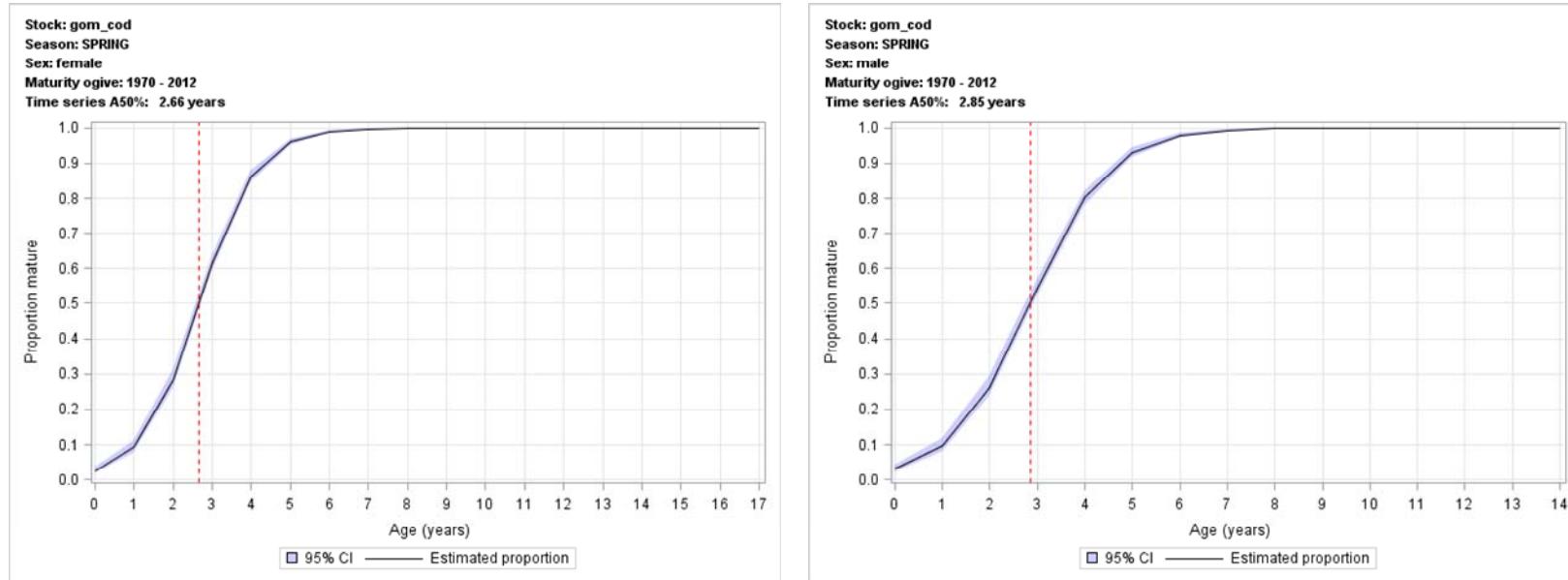


Figure A.14. Age-based maturity ogives for female (left) and male (right) Gulf of Maine Atlantic cod based on time series averages of maturity and age information collected from the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey from 1970 to 2012. The dashed red line indicates the age at 50% maturity ($A_{50\%}$).

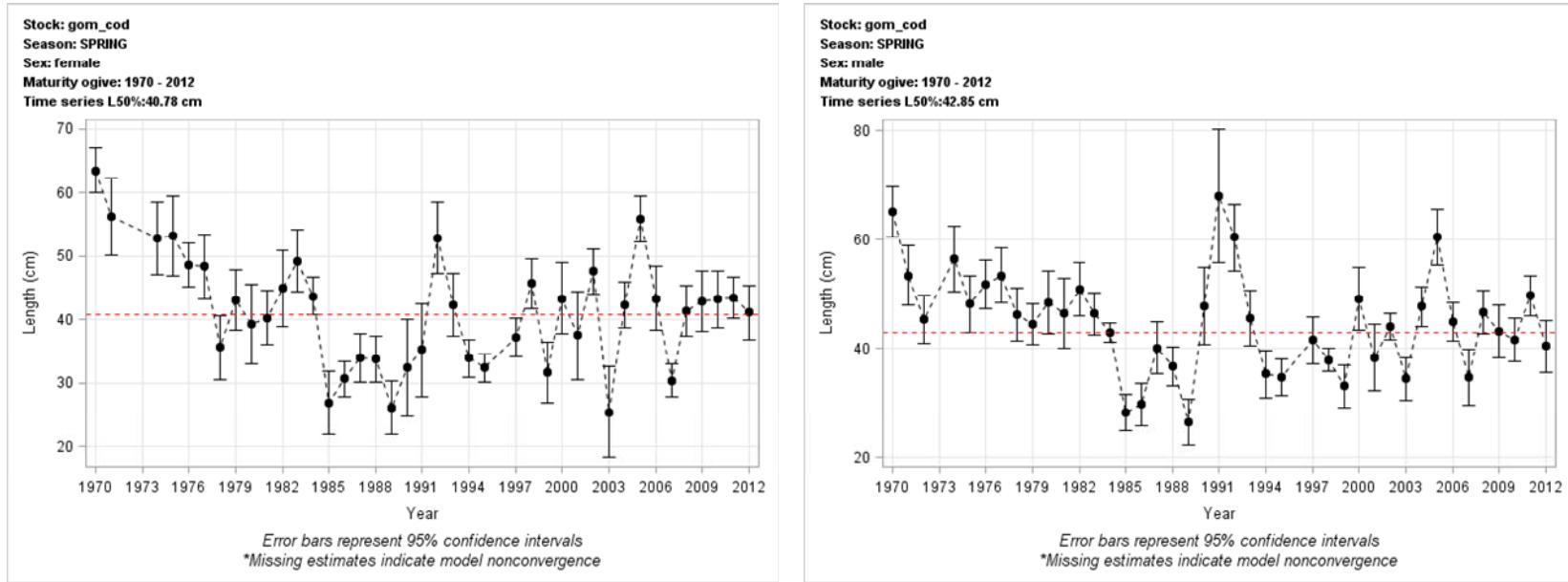


Figure A.15. Annual average length-at-50% maturity ($L_{50\%}$) and corresponding 95% confidence intervals for female (left panels) and male (right panels) Gulf of Maine Atlantic cod from 1970 to 2012. Average maturity has been estimated from data collected from the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey. Years in which the $L_{50\%}$ could not be estimated are omitted from the plots.

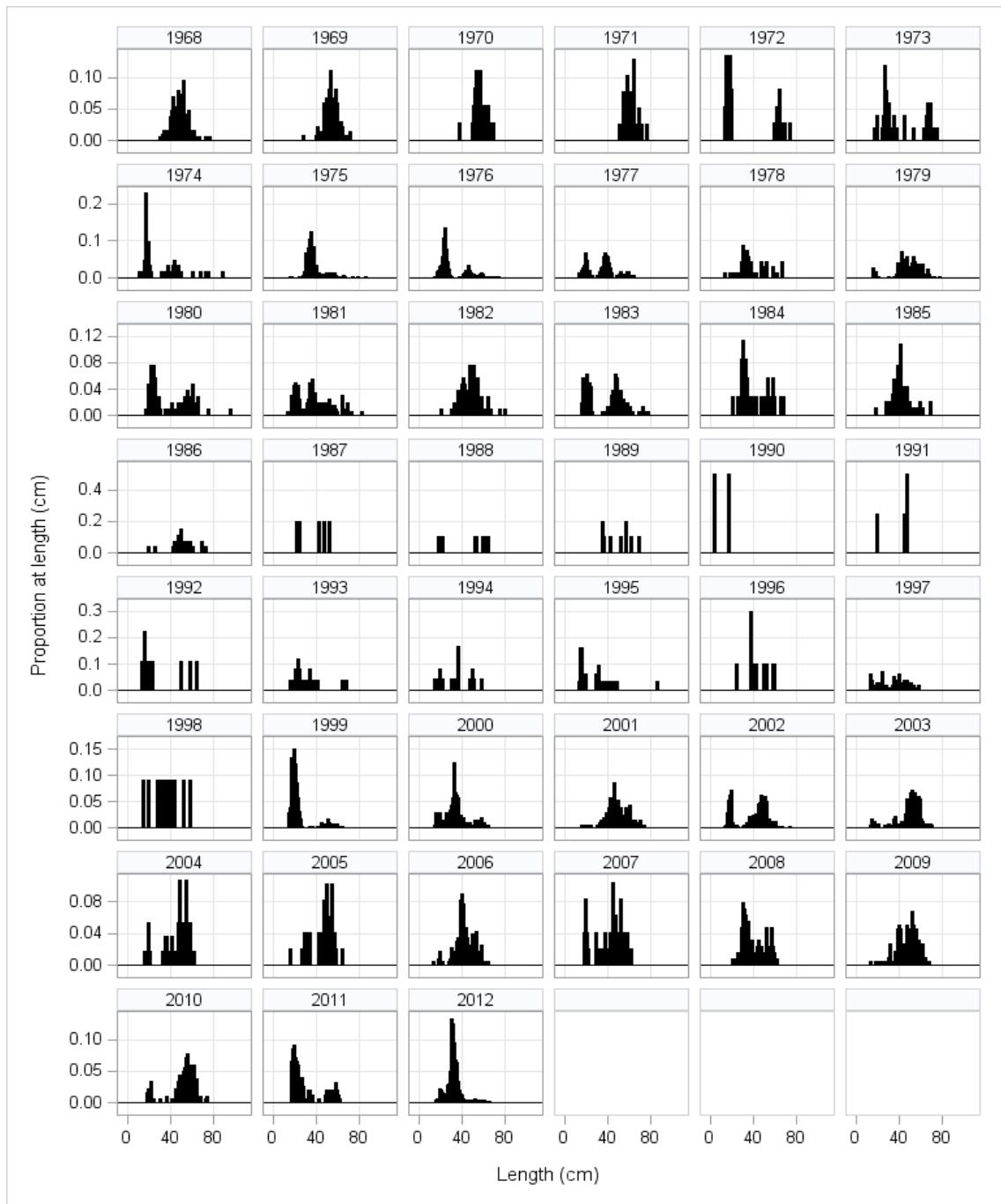


Figure A.16. Gulf of Maine Atlantic cod proportion at length observed in the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey between 1968 and 2012.

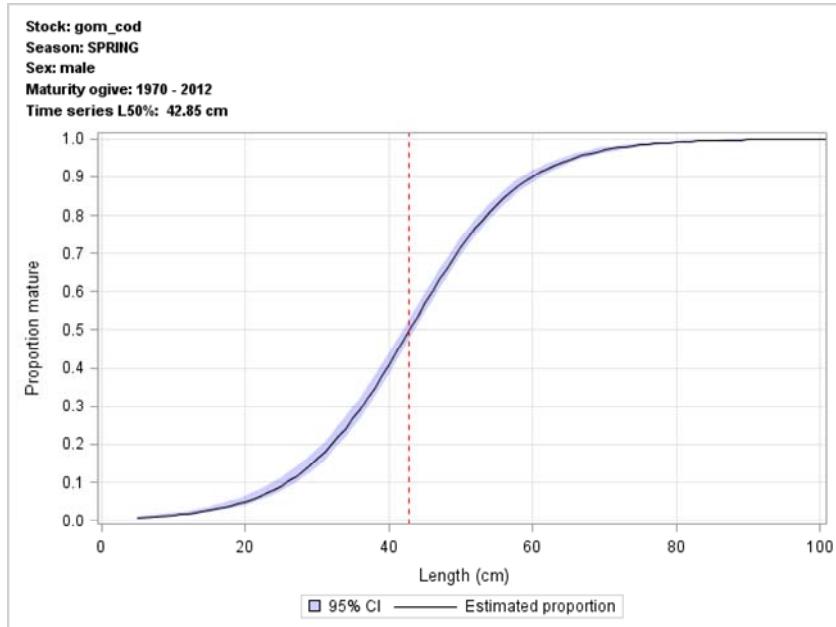
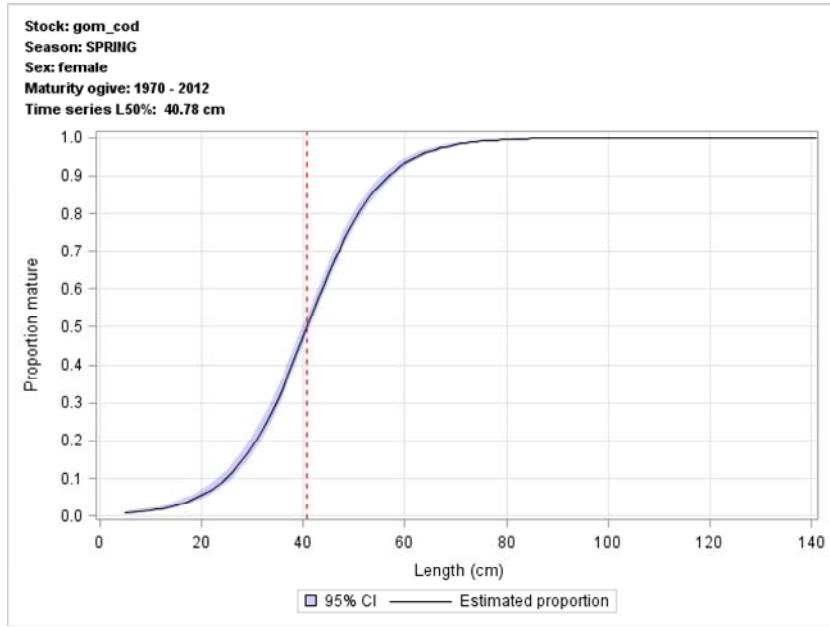


Figure A.17. Length-based maturity ogives for female (left) and male (right) Gulf of Maine Atlantic cod based on time series averages of maturity and length information collected from the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey from 1970 to 2012. The dashed red line indicates the length at 50% maturity ($L_{50\%}$).

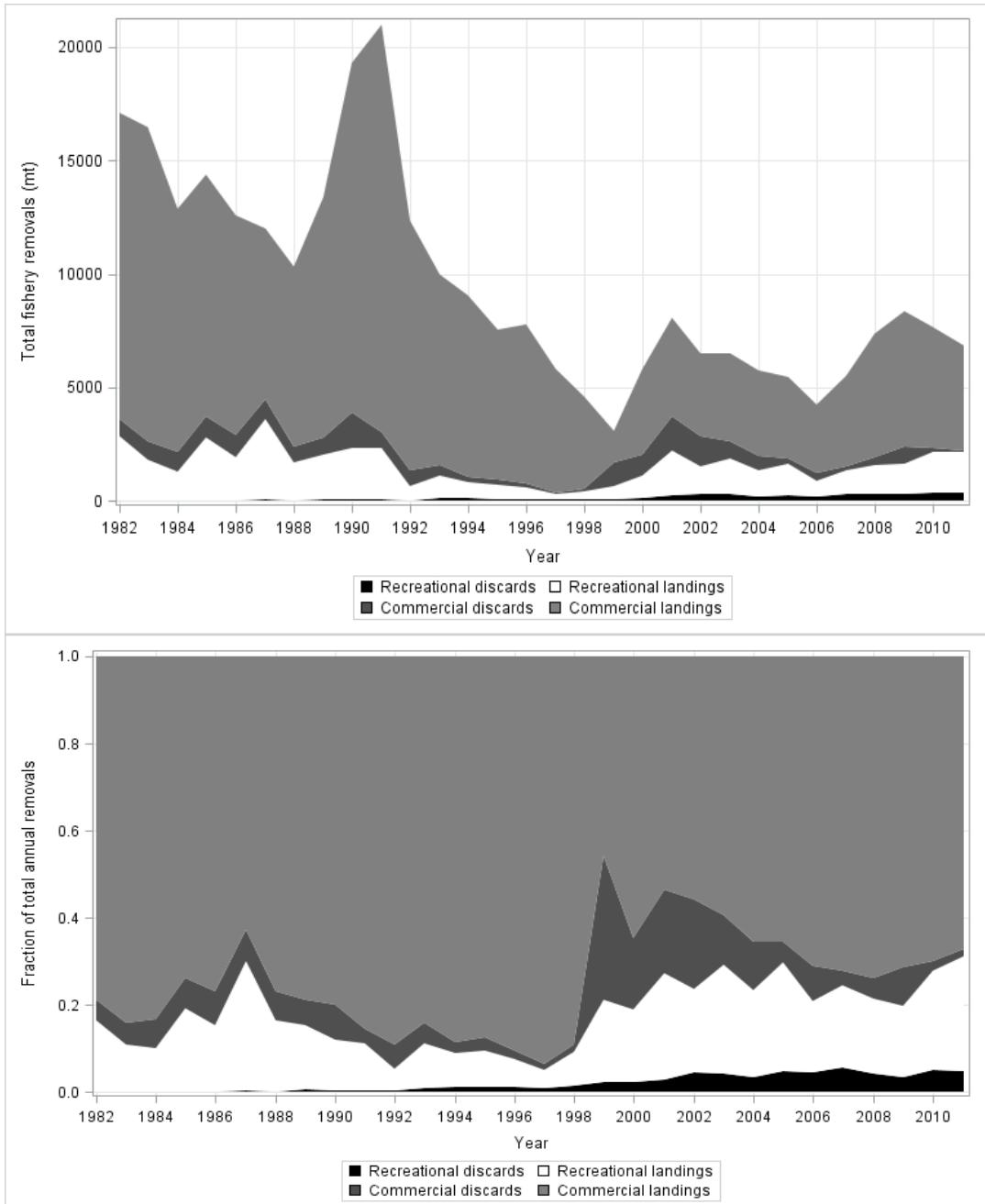


Figure A.18. Total (top) and fractional (as a fraction of the total, bottom) catch of Gulf of Maine Atlantic cod from 1982 to 2011 by fleet (commercial and recreational) and disposition (landed, discarded).

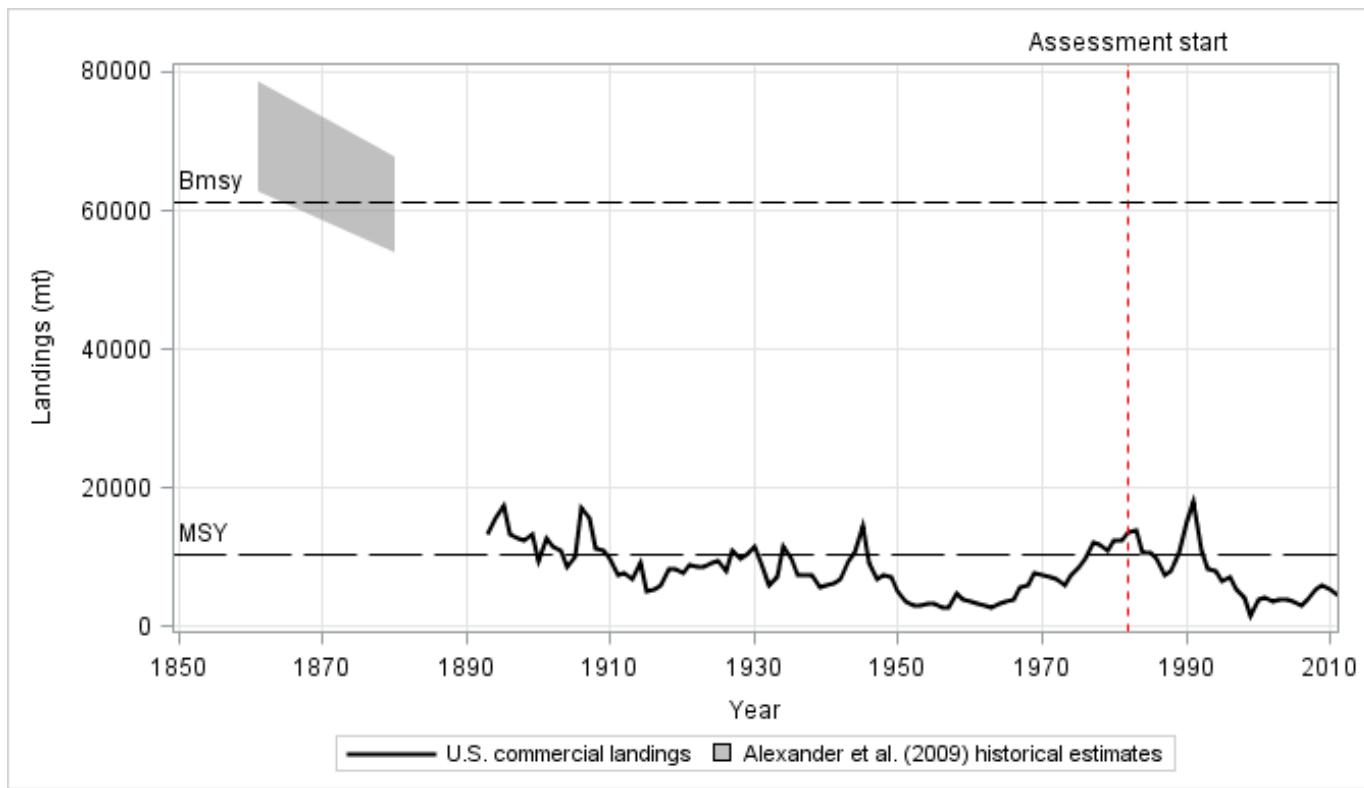


Figure A.19. United States commercial landings of Gulf of Maine Atlantic cod from 1861 to 2011. The grey-shaded polygon represents estimates of landings in 1861 and 1870 using two different conversion factors for converting cured salted cod to live fish (Alexander et al. 2009). Biological reference points ($B_{MSY} = 61,218$ mt, $MSY = 10,392$ mt) from the most recent assessment (NEFSC 2012) are shown by the dashed lines.

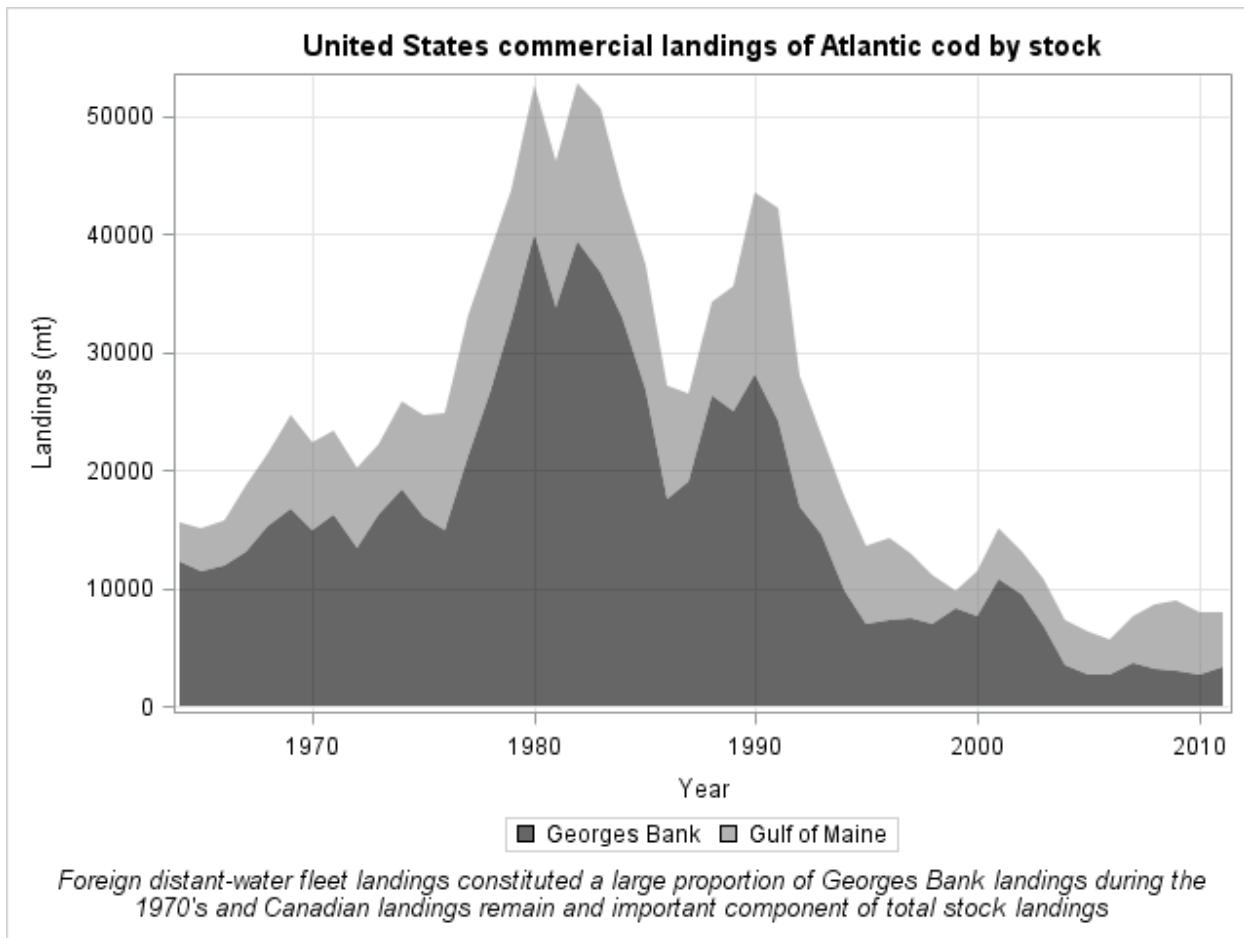


Figure A.20. Total United States commercial landings of Gulf of Maine and Georges Bank Atlantic cod from 1964 to 2011.

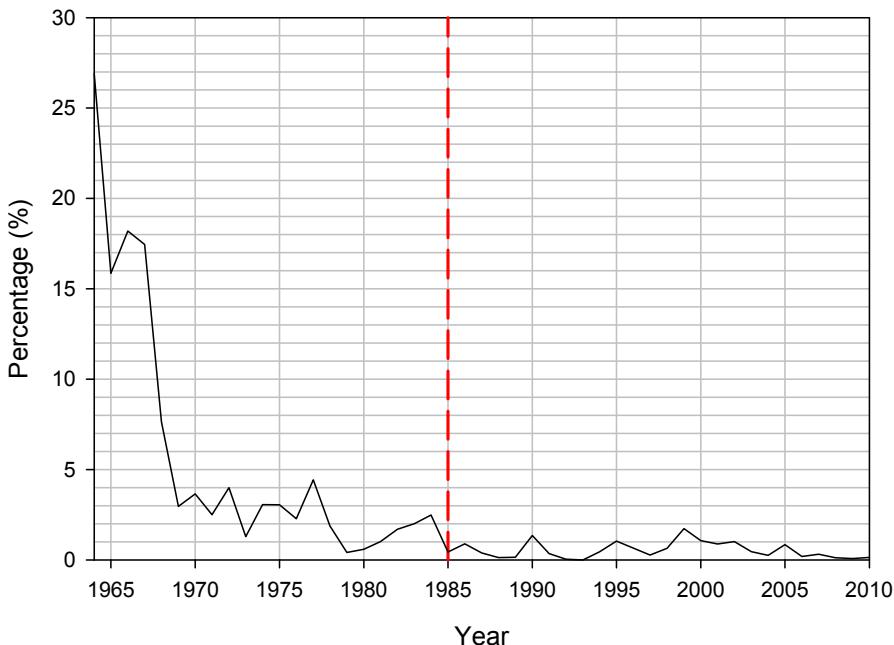


Figure A.21. Percentage of total commercial landings of Gulf of Maine Atlantic cod from statistical areas 464, 465 and 467 between 1964 and 2010. The Hague Line, which formally defined the Exclusive Economic Zones of the United States and Canada was adopted on October 12, 1984 (dashed red line).

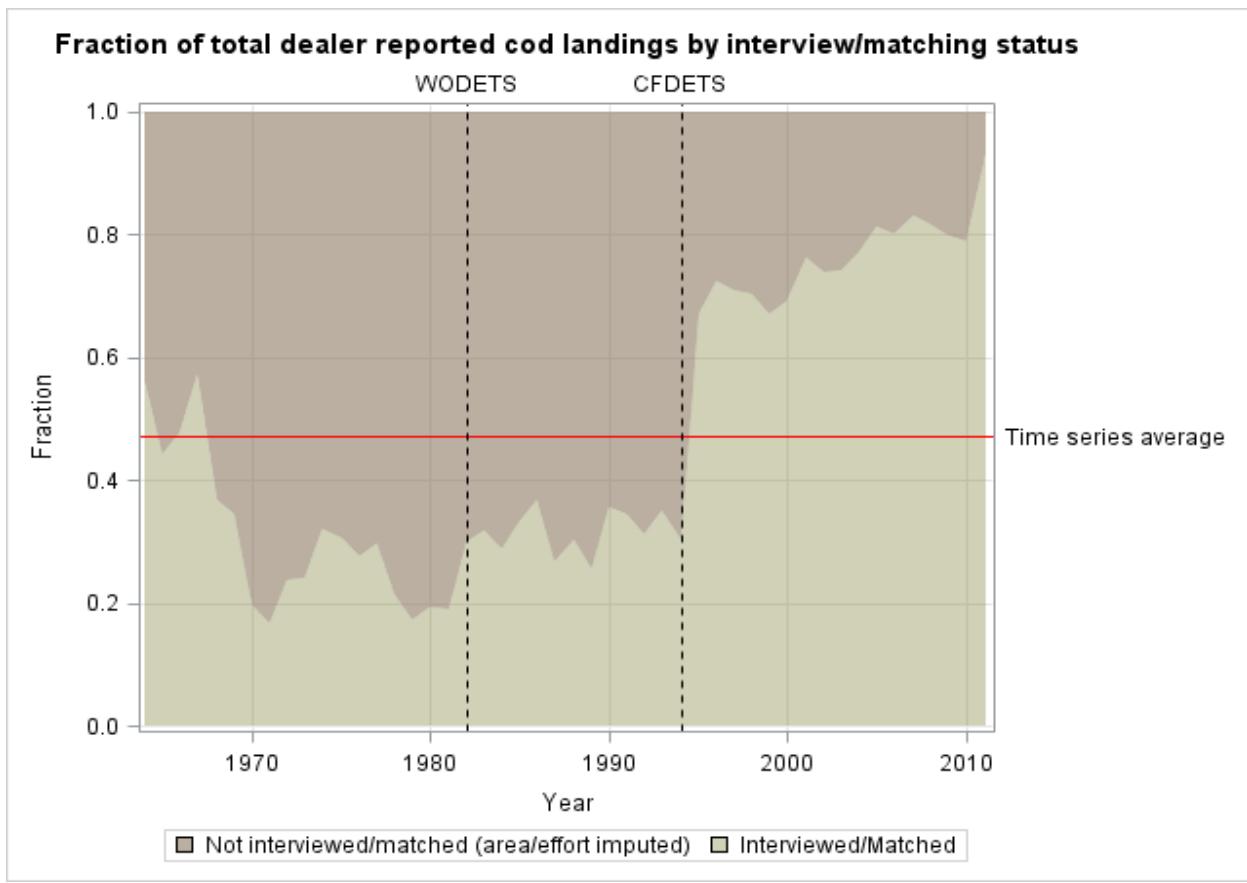


Figure A.22. Fraction of the Gulf of Maine Atlantic cod commercial landings from either interviewed trips (1964-1994) or those trips that could be directly matched to a vessel trip report (1994-2011). The red line indicates the time series average fraction of landings from interviewed/matched trips.

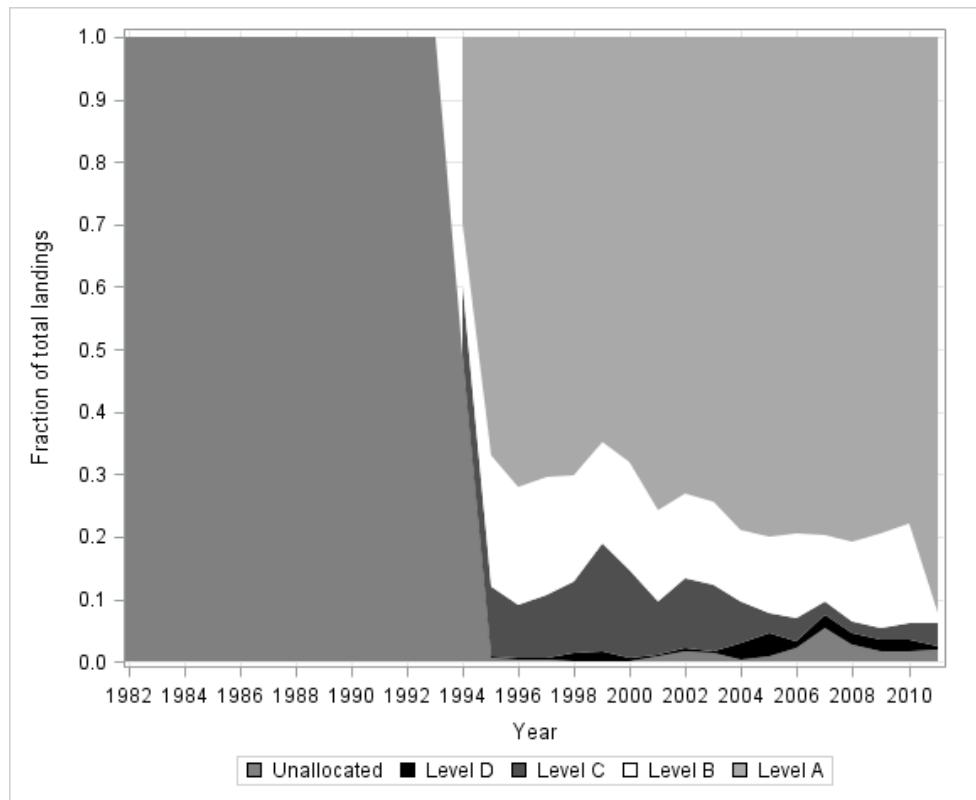


Figure A.23. Fraction of the Gulf of Maine Atlantic cod commercial landings by allocation level between 1982 and 2011. Prior to 1994 landings were allocated based on a port interview process. From 1994 onward landings were allocated to statistical area and gear type based on a standardized allocation scheme described in Wigley et al. (2008).

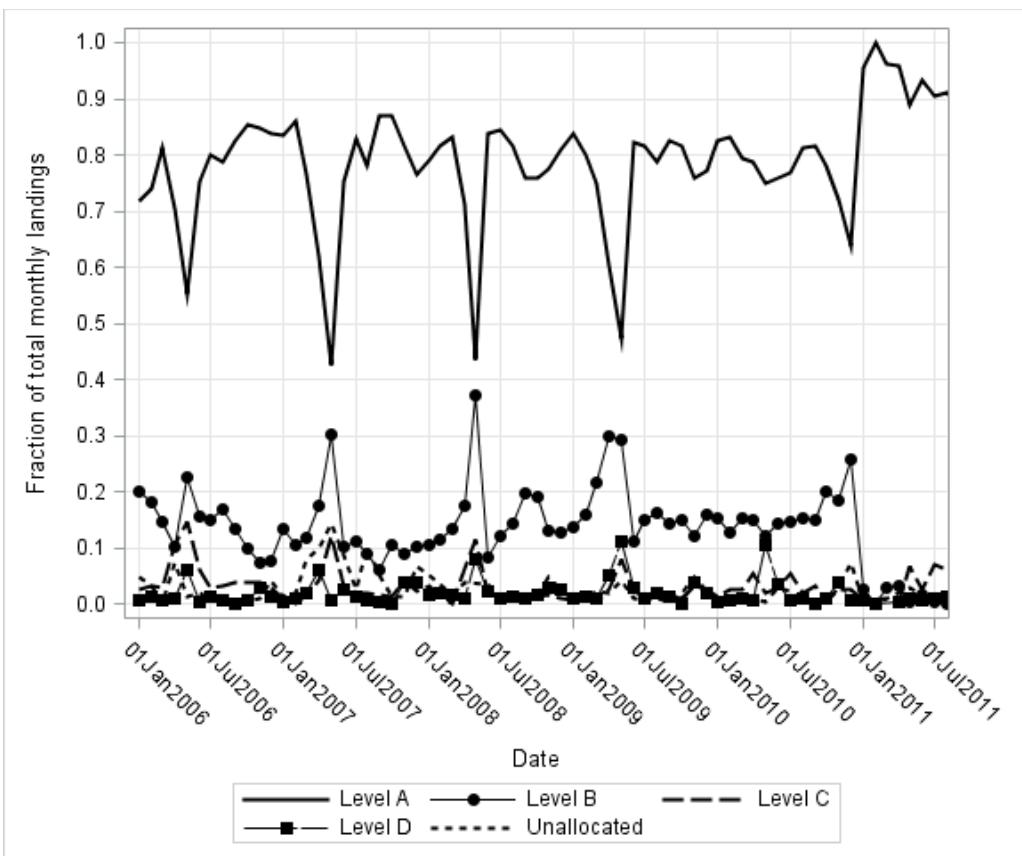


Figure A.24. Fraction of the Gulf of Maine Atlantic cod commercial landings by allocation level between 2006 and 2011 by month.

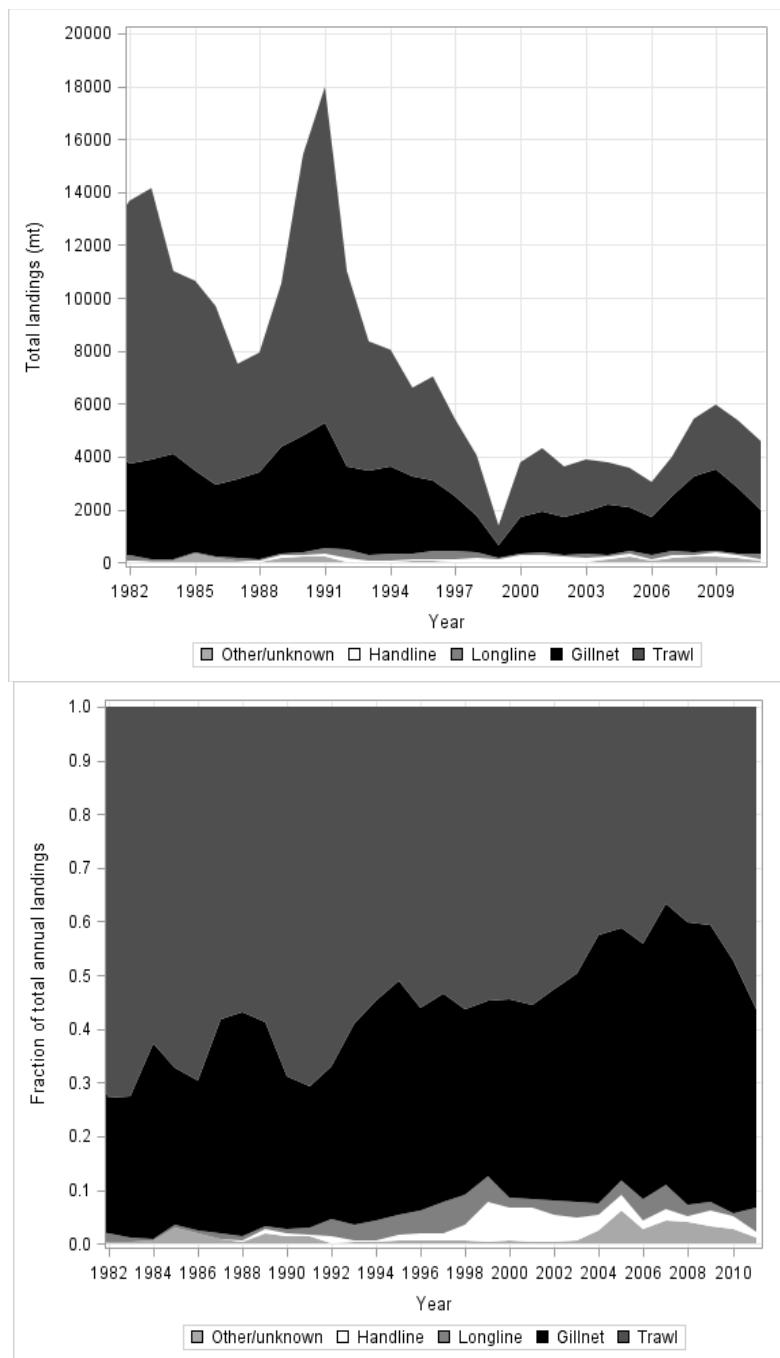


Figure A.25. Total (top) and fractional (as a fraction of the total, bottom) commercial landings of Gulf of Maine Atlantic cod by gear from 1982 to 2011.

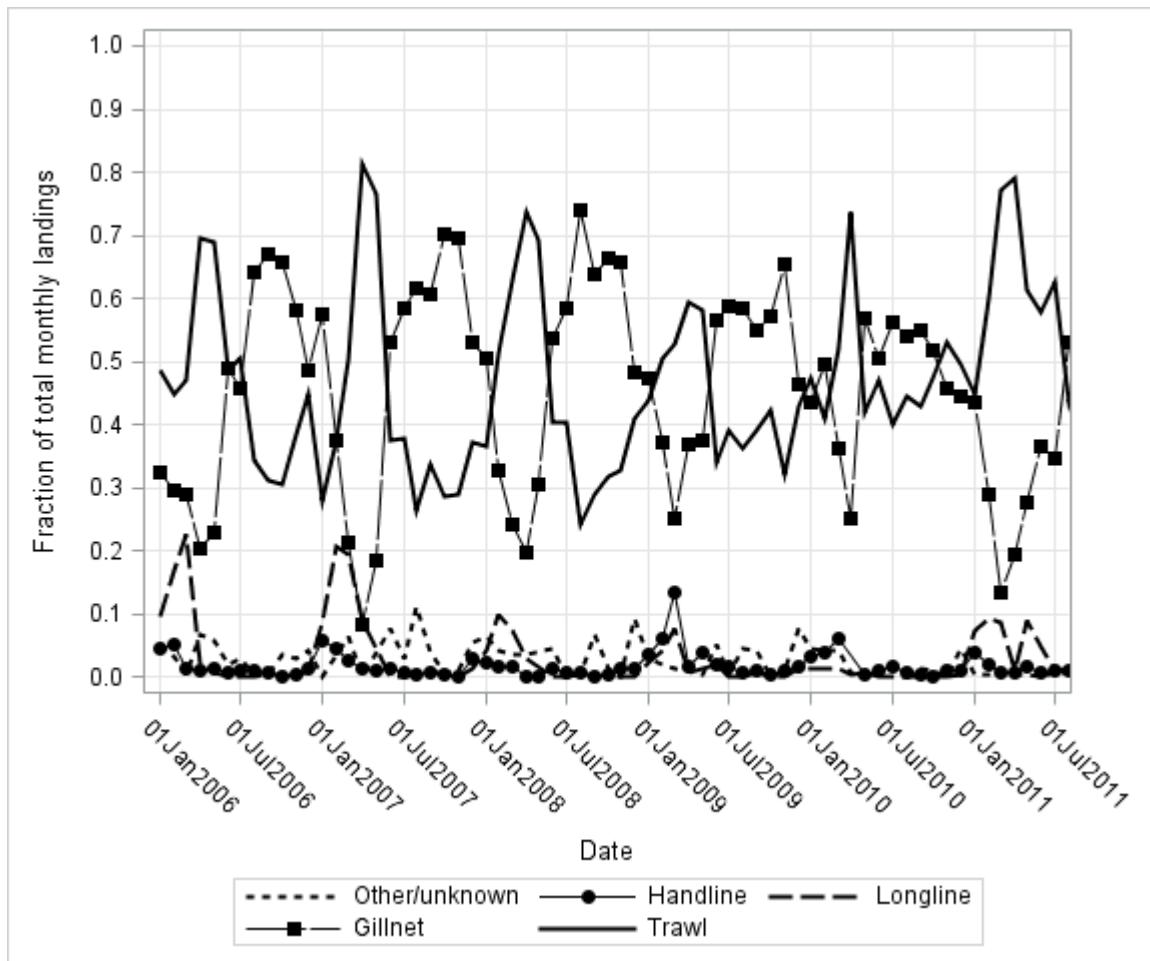


Figure A.26. Monthly commercial landings patterns (as a fraction of the total landings) of Gulf of Maine Atlantic cod by gear from 2006 to 2011.

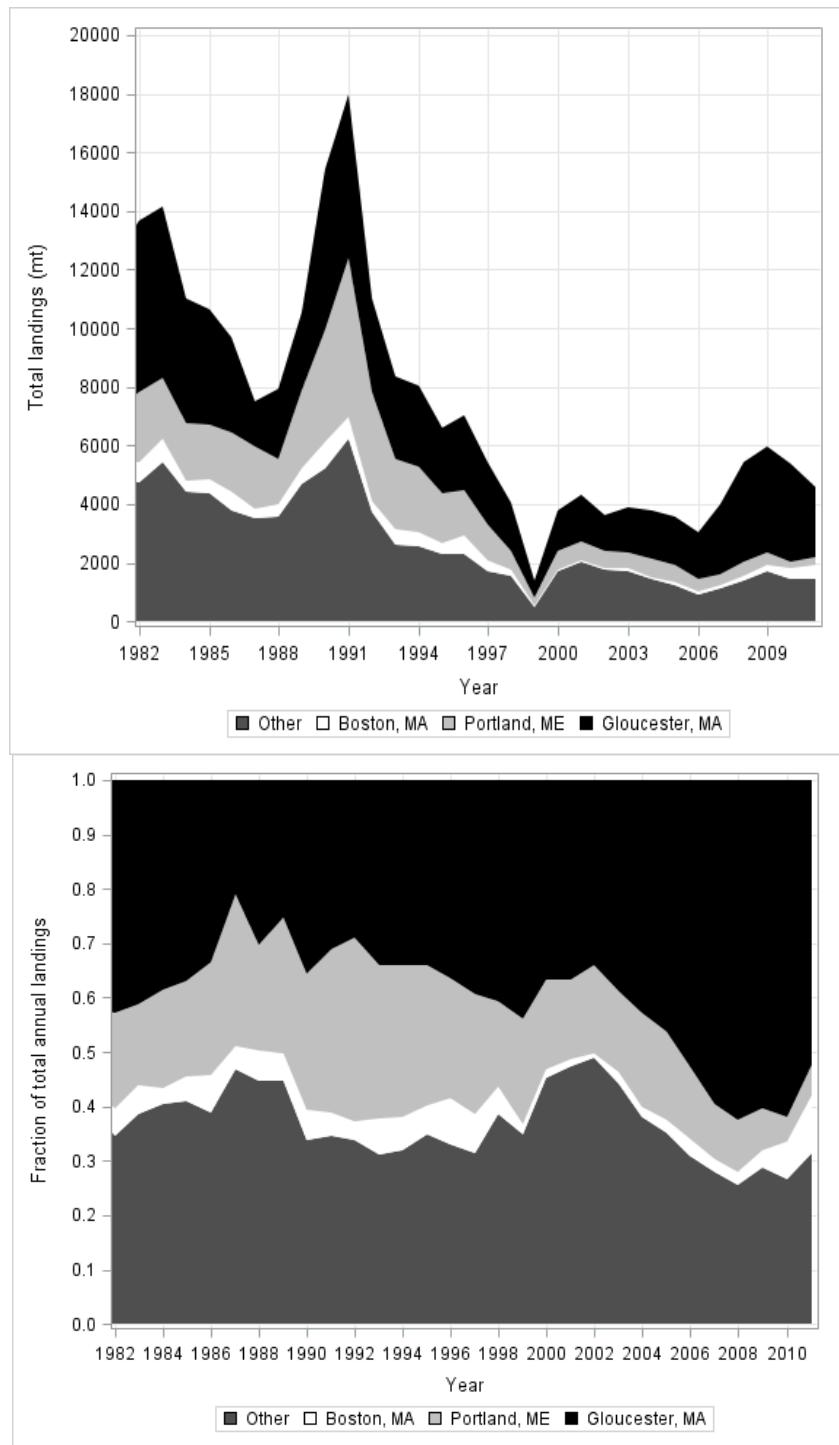


Figure A.27. Total (top) and fractional (as a fraction of the total, bottom) commercial landings of Gulf of Maine Atlantic cod by port from 1982 to 2011.

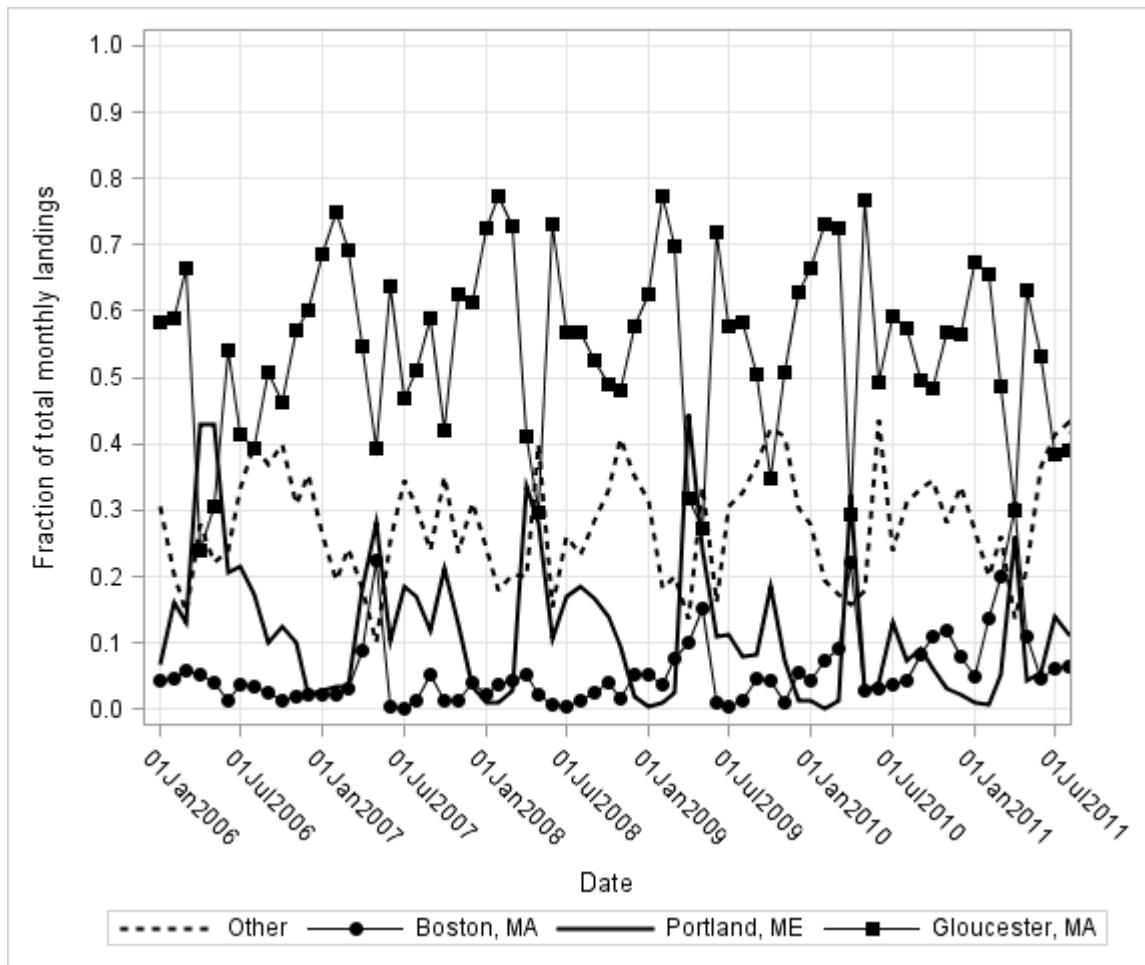


Figure A.28. Monthly commercial landings patterns (as a fraction of the total landings) of Gulf of Maine Atlantic cod by port from 2006 to 2011.

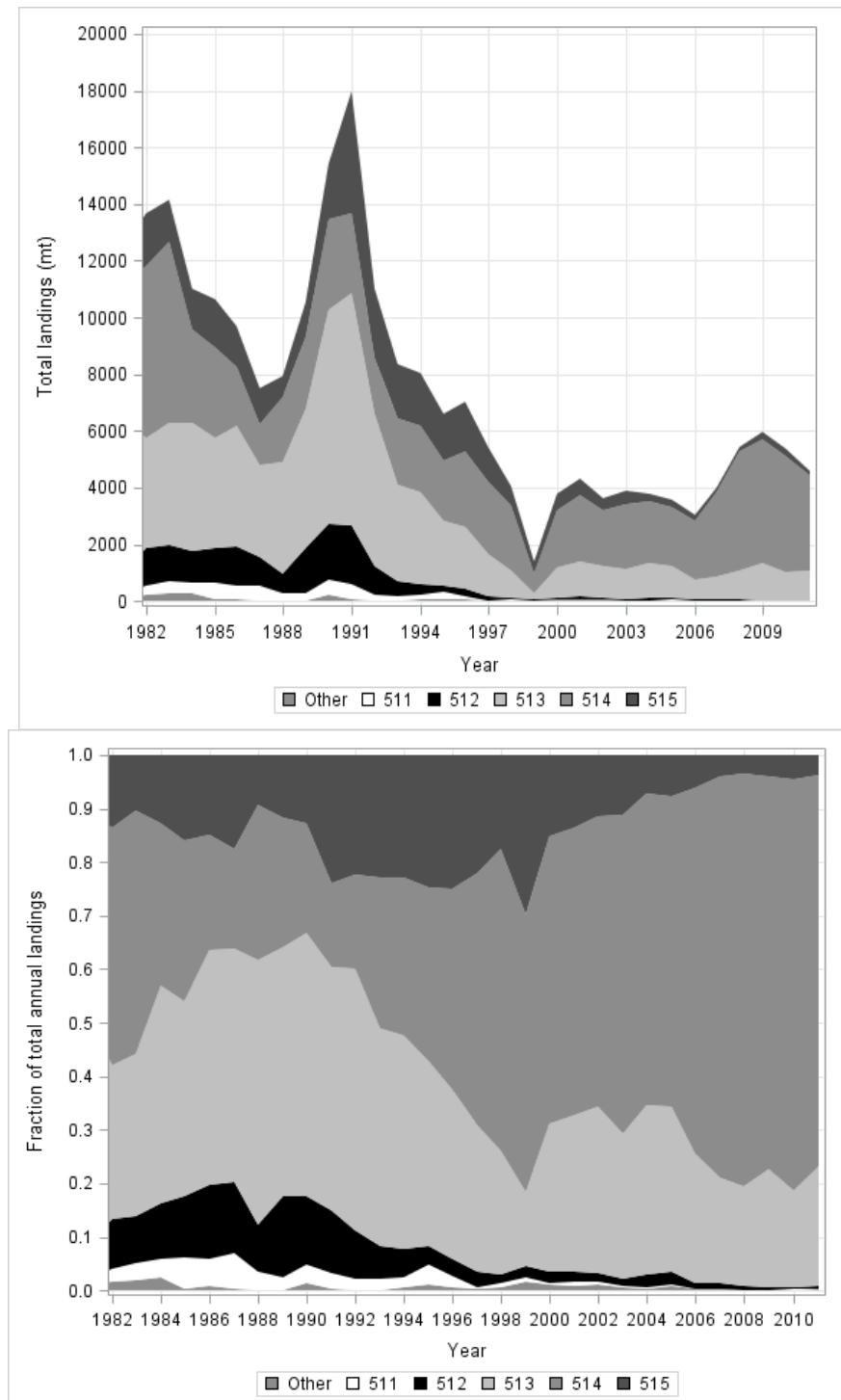


Figure A.29. Total (top) and fractional (as a fraction of the total, bottom) commercial landings of Gulf of Maine Atlantic cod by statistical area from 1982 to 2011.

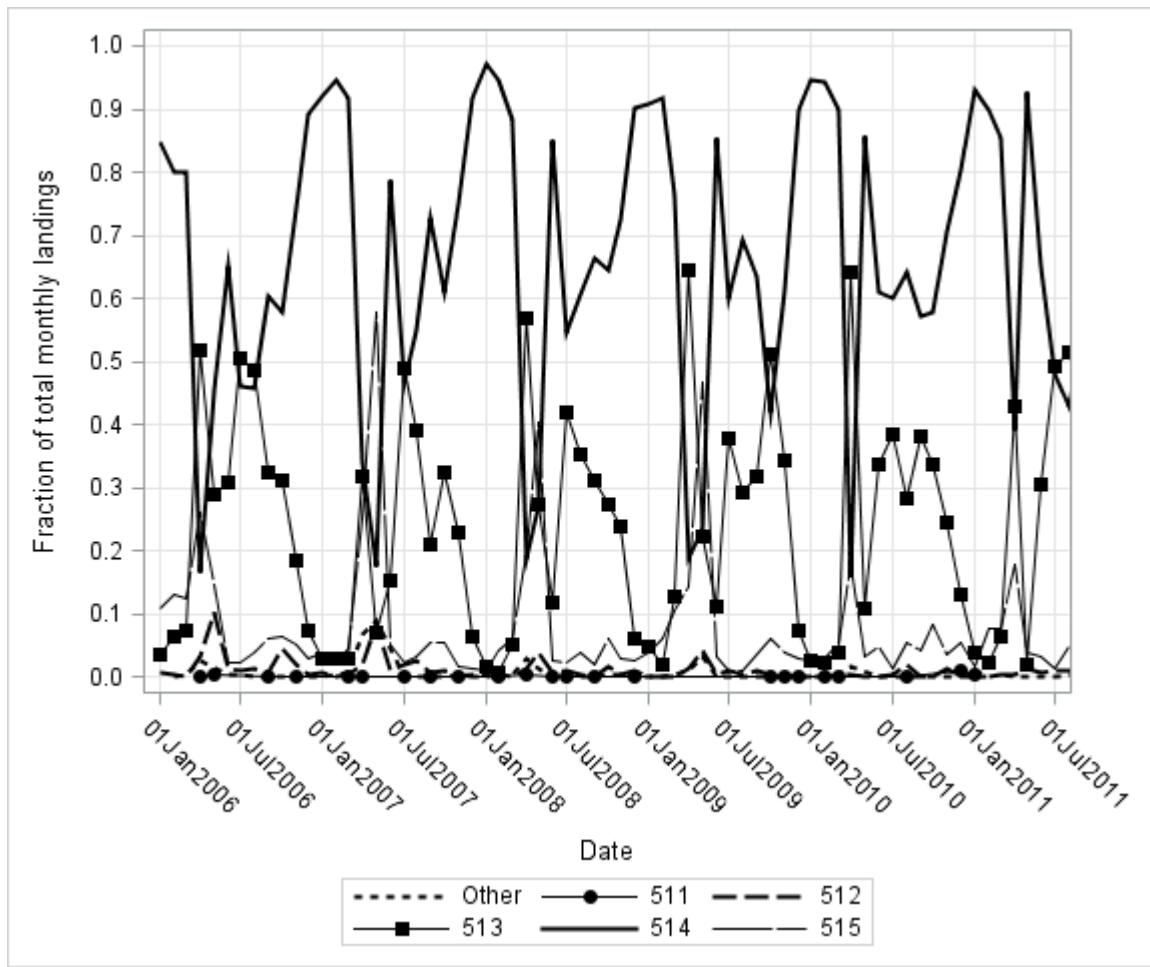


Figure A.30. Monthly commercial landings patterns (as a fraction of the total landings) of Gulf of Maine Atlantic cod by statistical area from 2006 to 2011.

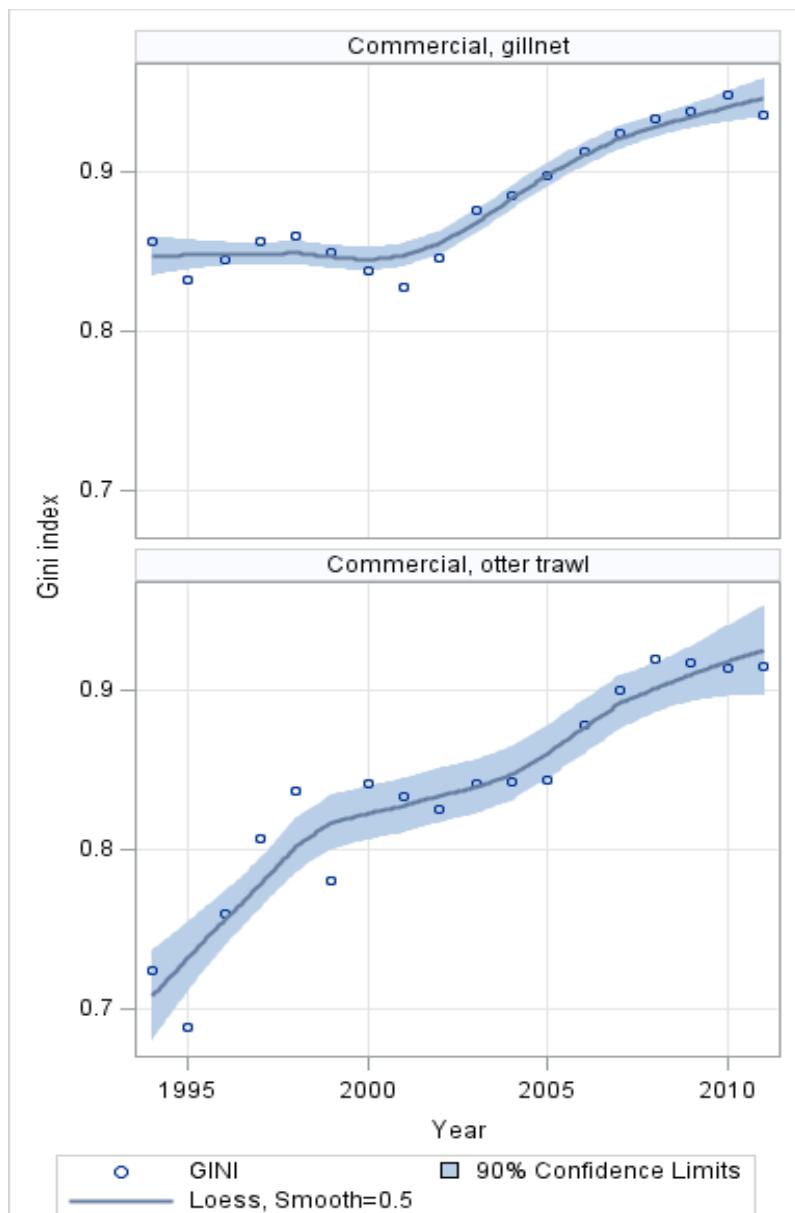


Figure A.31. Gini indices for the commercial otter trawl (050) and sink gillnet (100) fleets from 1994-2011. Indices were based on the spatial distribution of the retained catch reported on vessel trip reports.

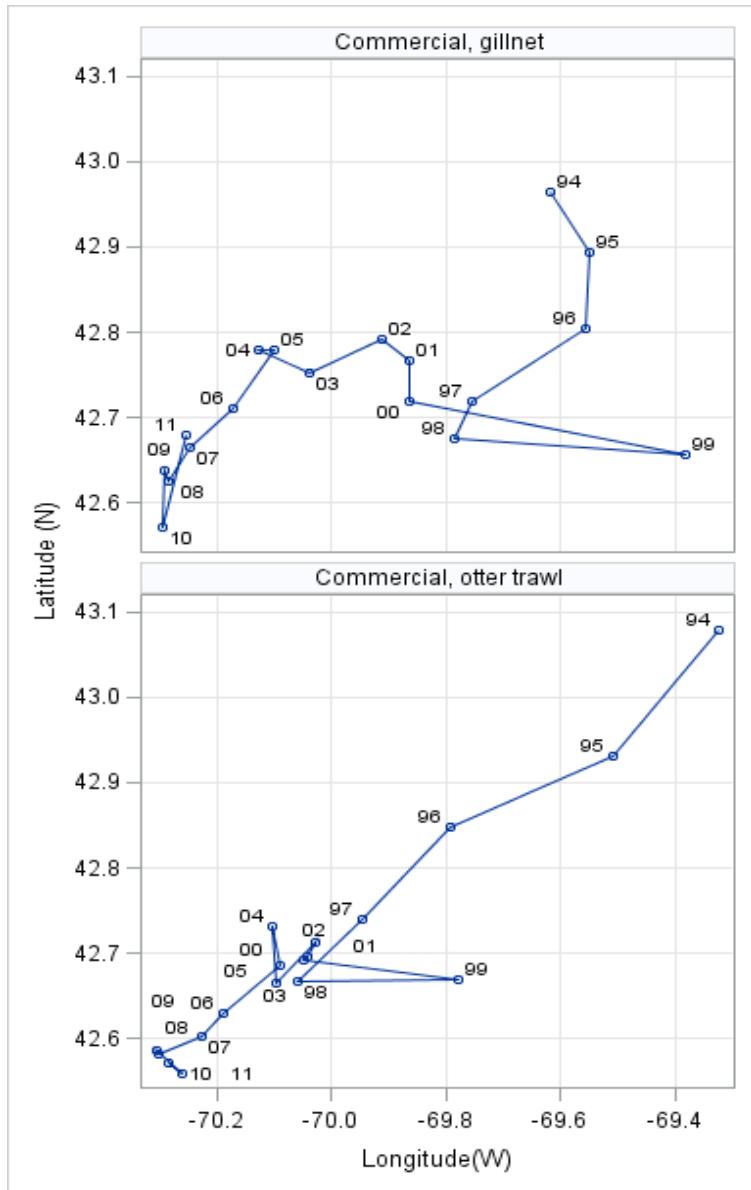


Figure A.32. Landings-weighted mean location (centroid) of Gulf of Maine cod catch by the commercial gillnet (GNS) and otter trawl (OTF) fleets from 1994 to 2011. Centroids were based on the spatial distribution of the retained catch reported on vessel trip reports.

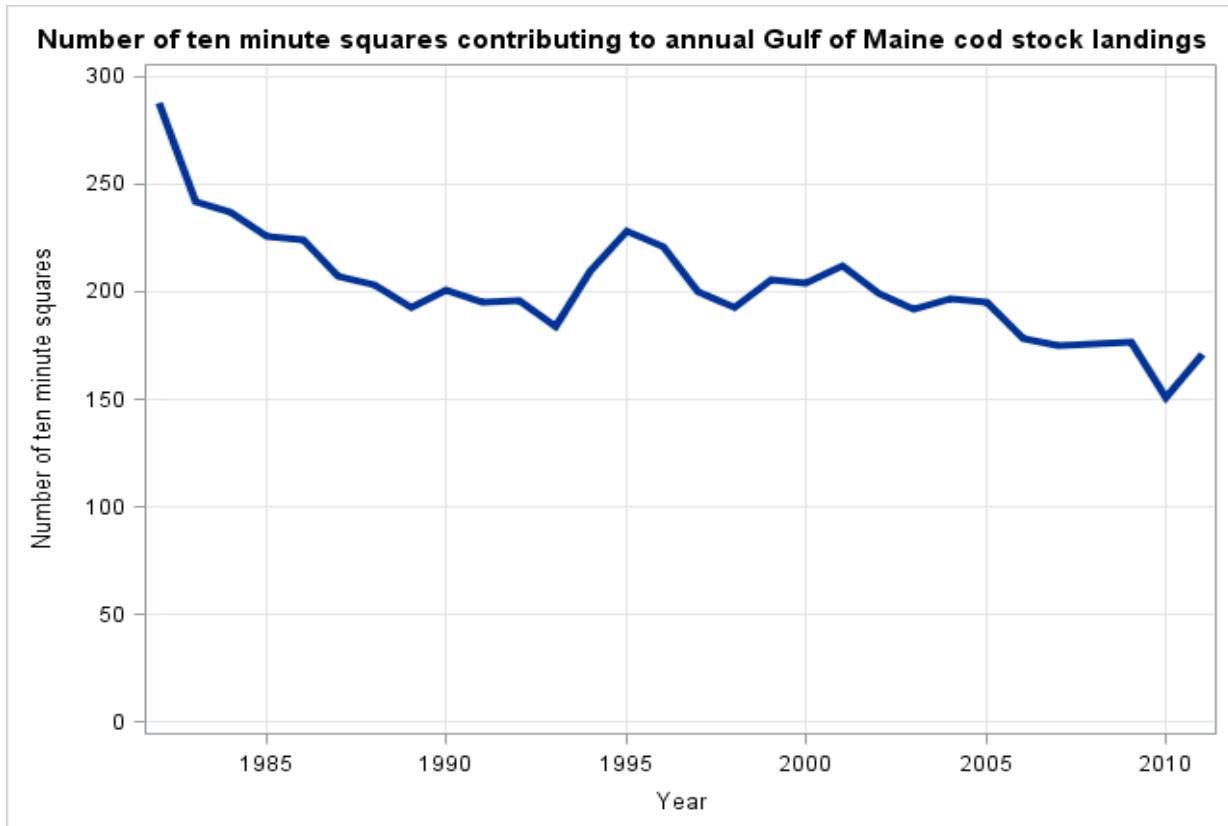


Figure A.33. Number of ten minute squares contributing to the annual landings of Gulf of Maine Atlantic cod between 1982 and 2011.

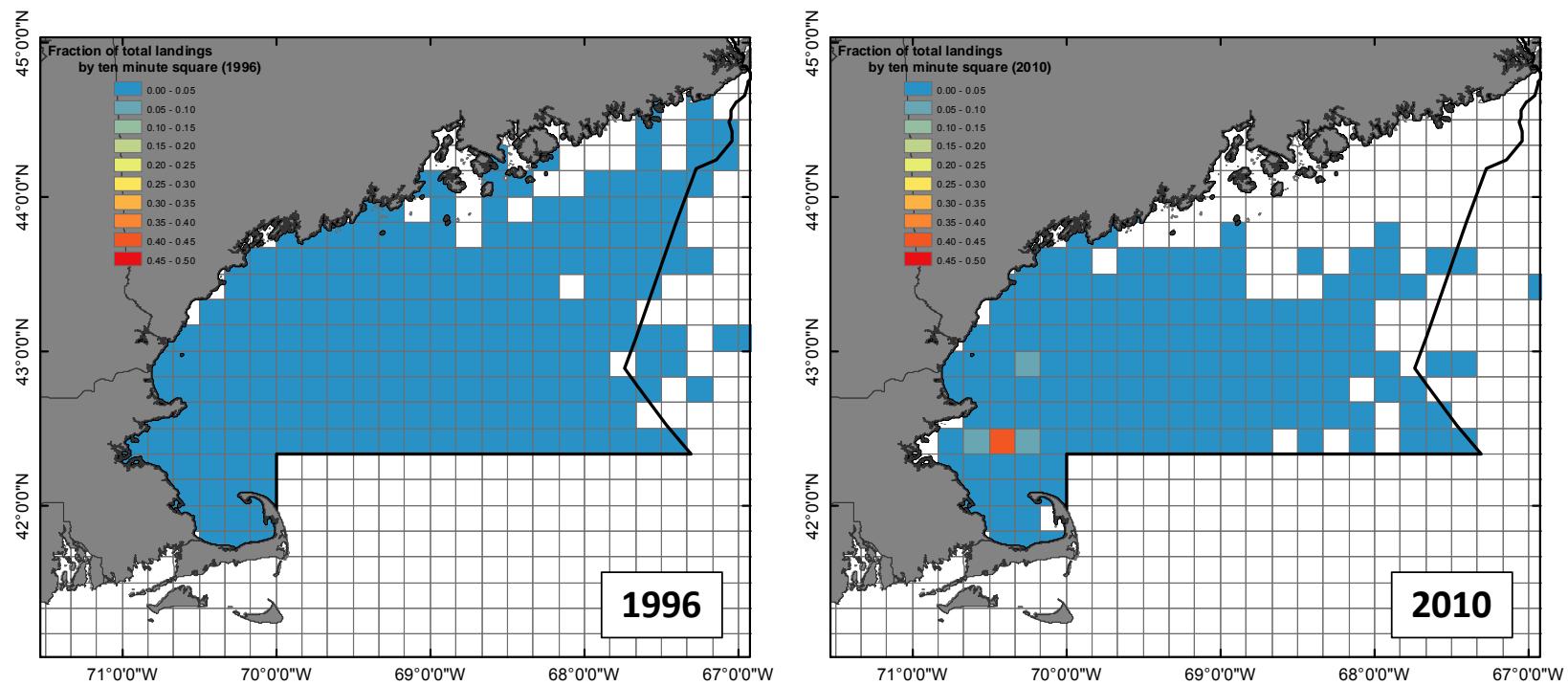


Figure A.34. Comparison of the fraction of annual landings per ten minute square in 1996 (left) to the distribution in 2010 (right).

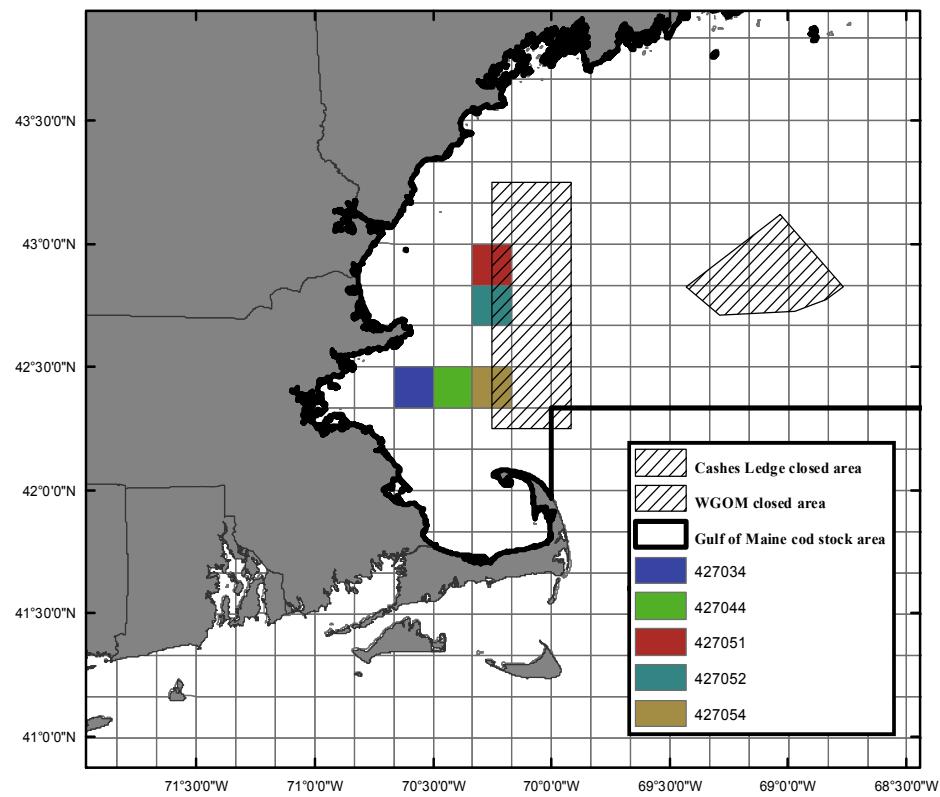


Figure A.35. Location of the top 5 ten minute squares with respect to the fraction of annual commercial landings of Gulf of Maine Atlantic cod between 1994 and 2011.

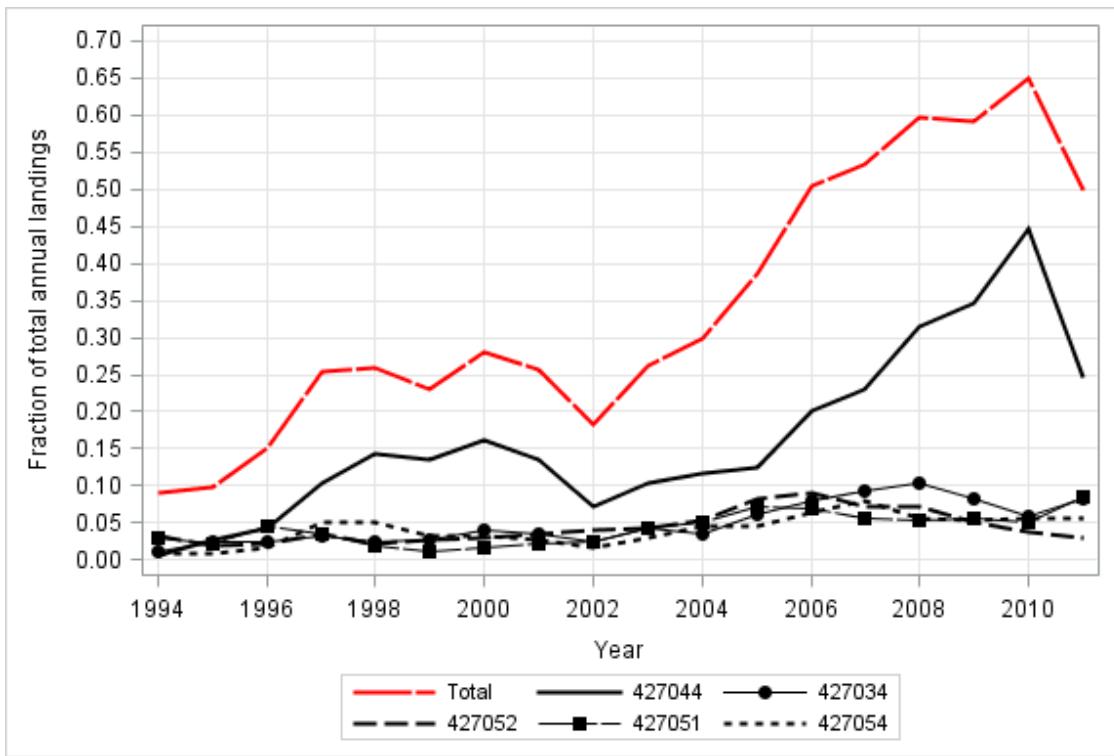


Figure A.36. Contribution of the top 5 ten minute squares to the annual commercial landings of Gulf of Maine Atlantic cod between 1994 and 2011.

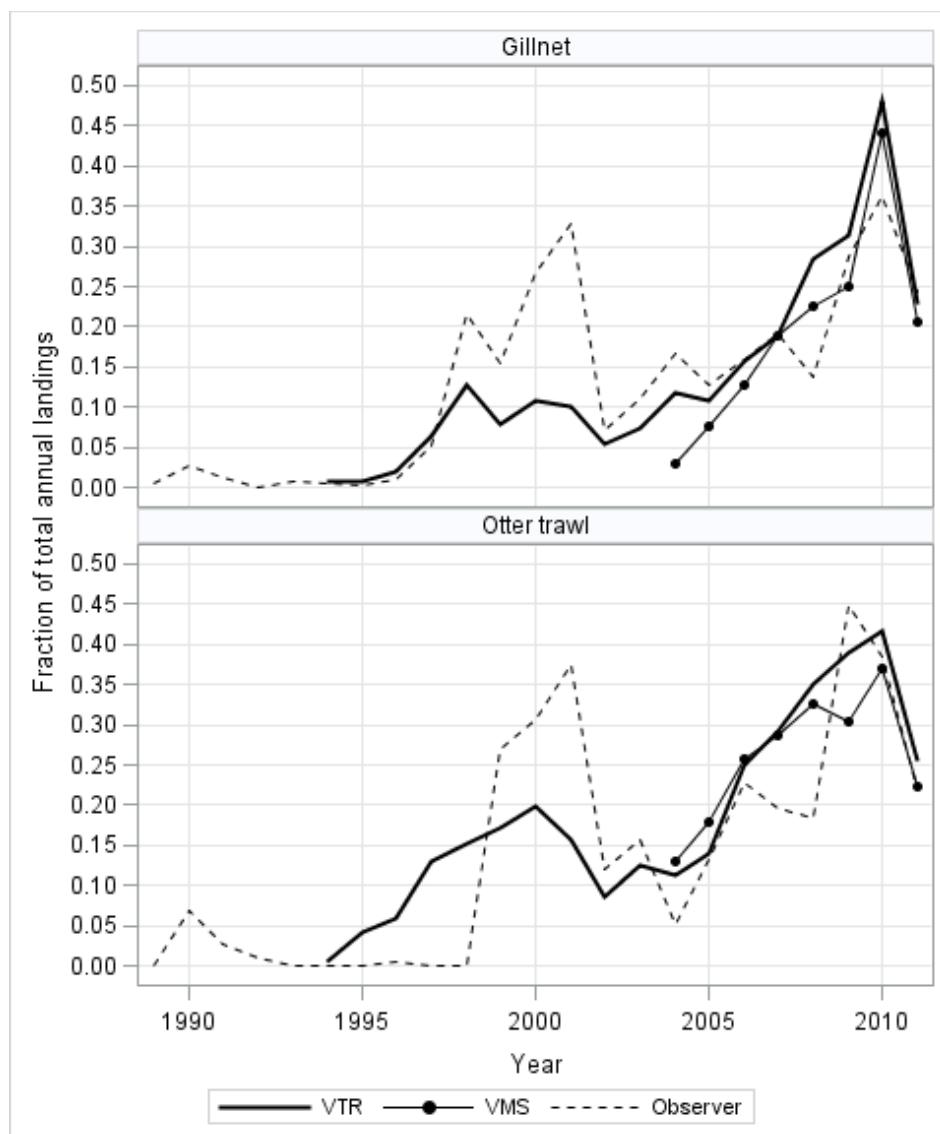


Figure A.37. Fraction of the total annual Gulf of Maine cod commercial gillnet and otter trawl landings from ten minute square 427044. Fractional landings have been calculated using three data sources: vessel trip reports (VTR), vessel monitoring data (VMS), and data collected by at-sea observers (Observers). Not all data sources are available for all years.

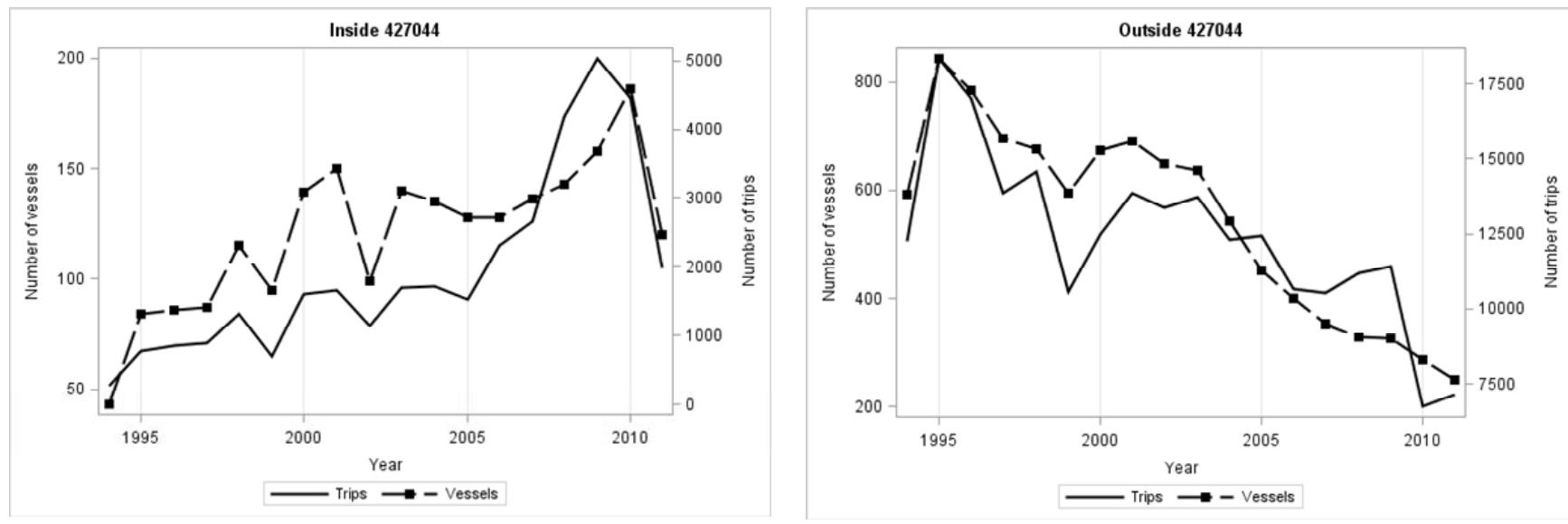


Figure A.38. Number of vessels and trips landing Gulf of Maine cod both inside (top) and outside (bottom) ten minute square 427044 between 1994 and 2011.

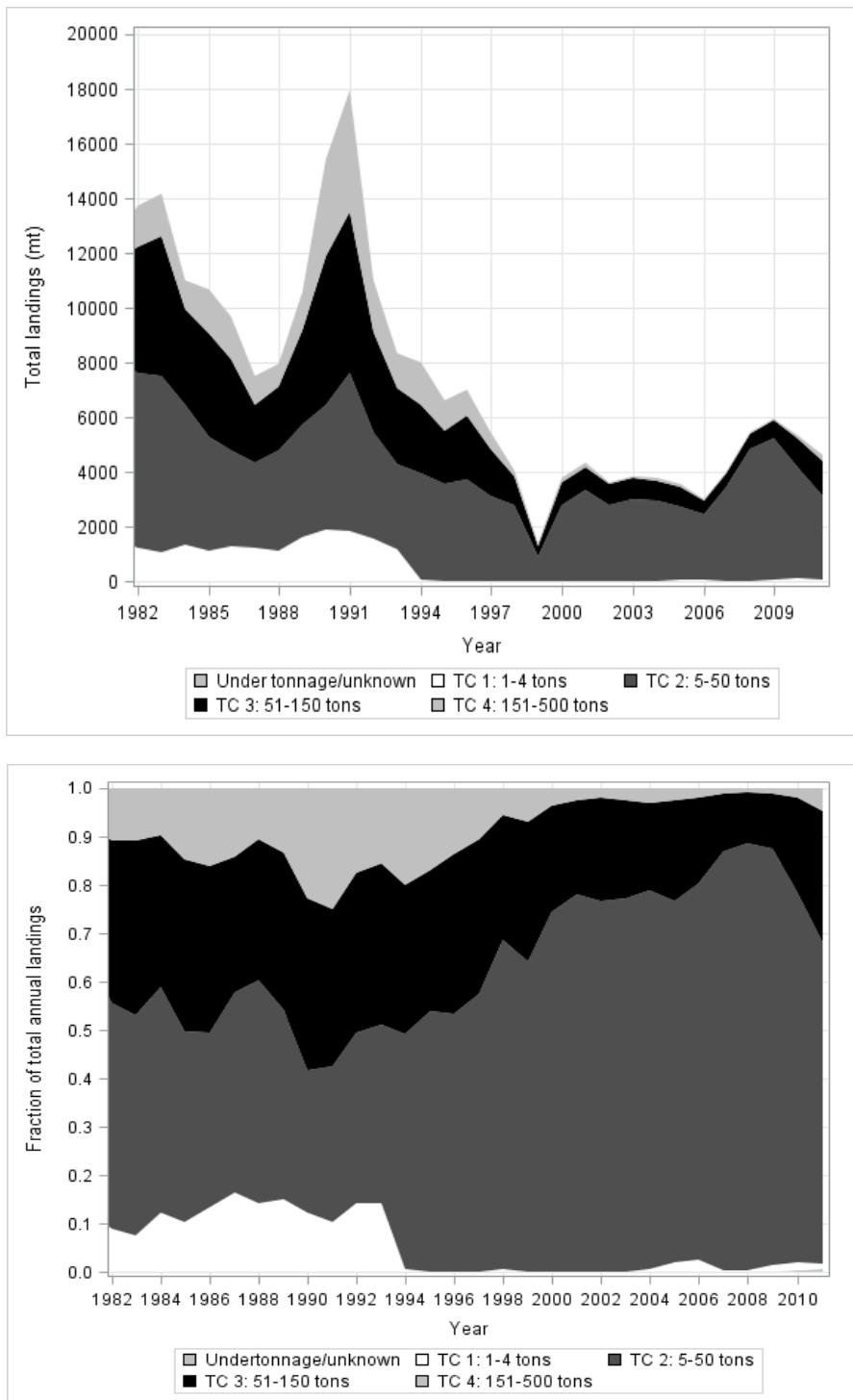


Figure A.39. Total (top) and fractional (as a fraction of the total, bottom) commercial landings of Gulf of Maine Atlantic cod by vessel ton class from 1982 to 2011.

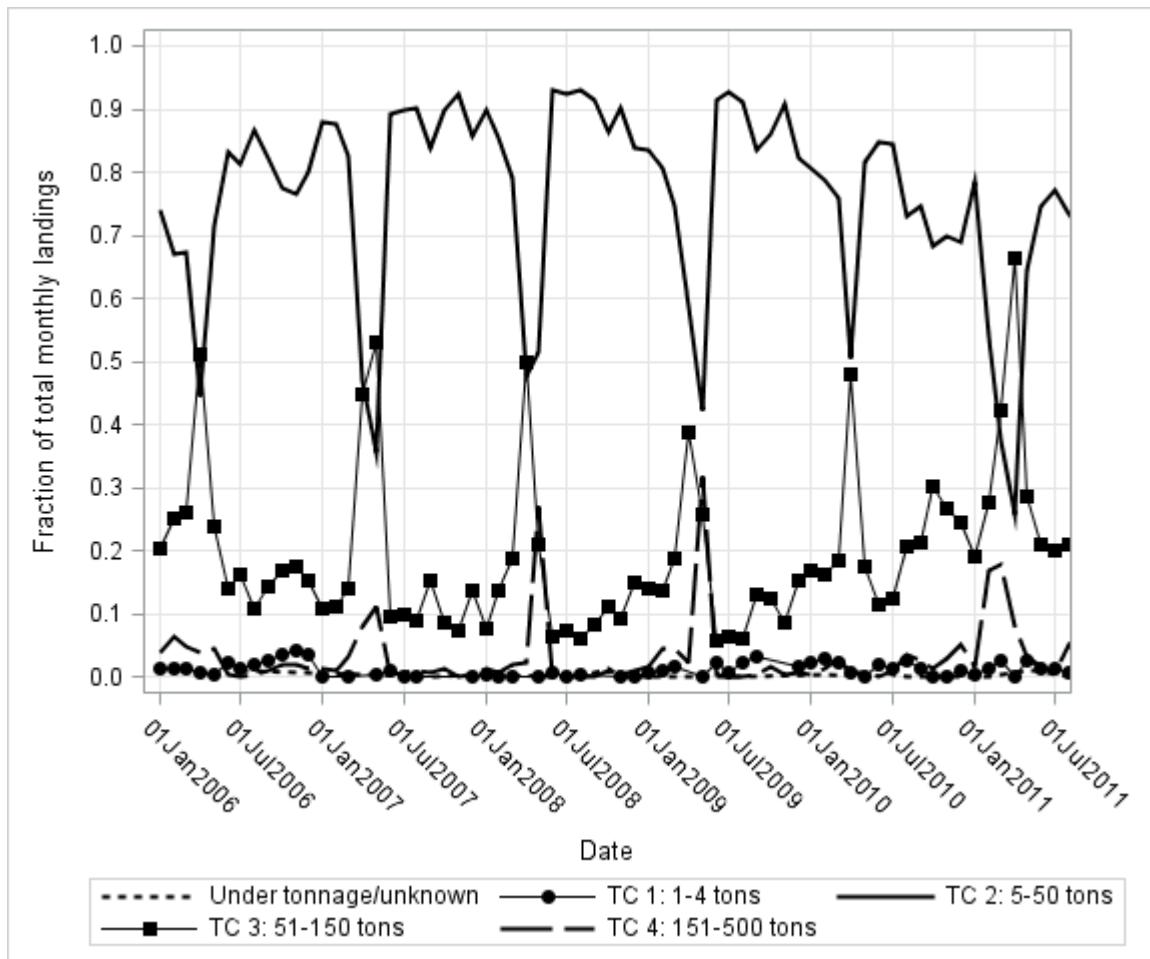


Figure A.40. Monthly commercial landing patterns (as a fraction of the total landings) of Gulf of Maine Atlantic cod by ton class from 2006 to 2011.

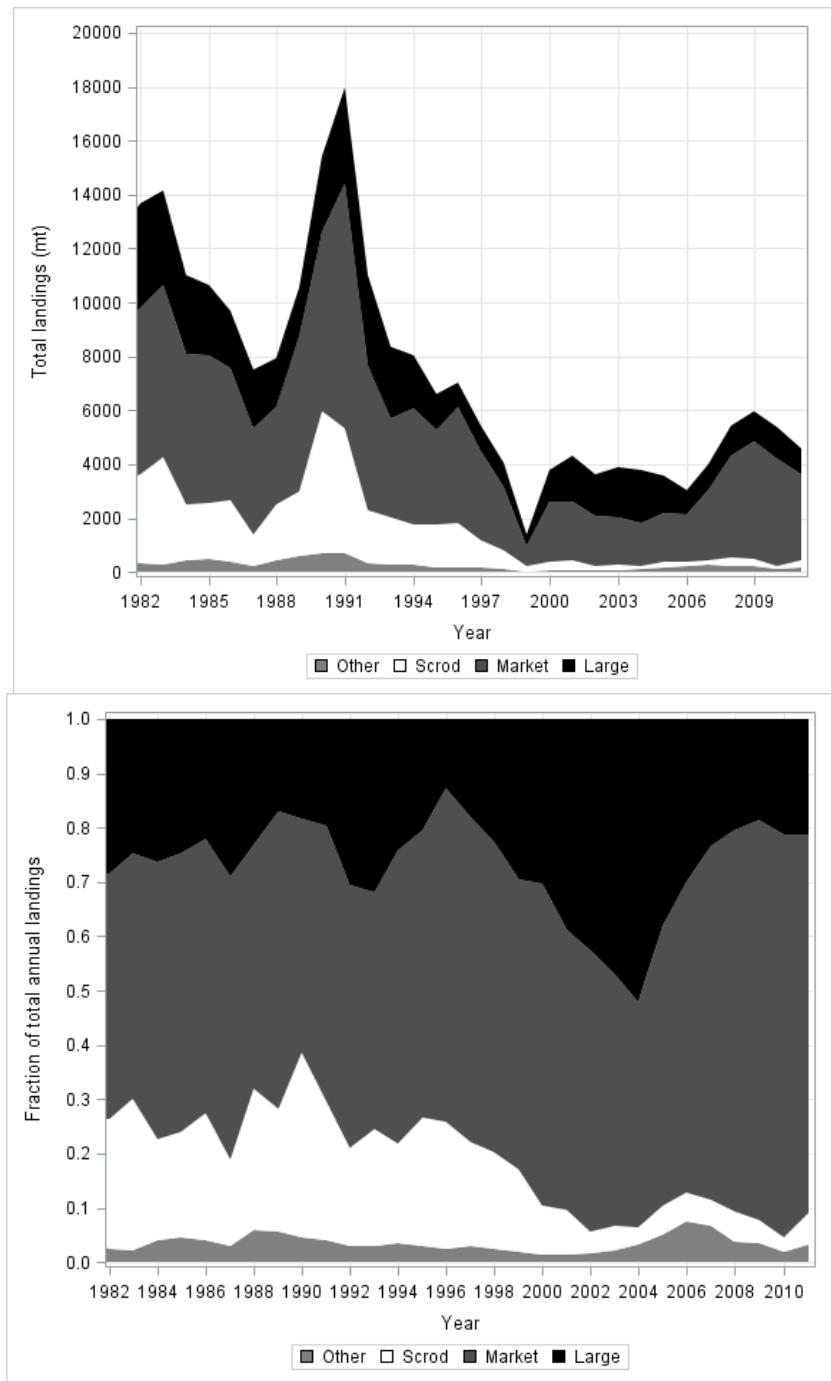


Figure A.41. Total (top) and fractional (as a fraction of the total, bottom) commercial landings of Gulf of Maine Atlantic cod by market category from 1982 to 2011.

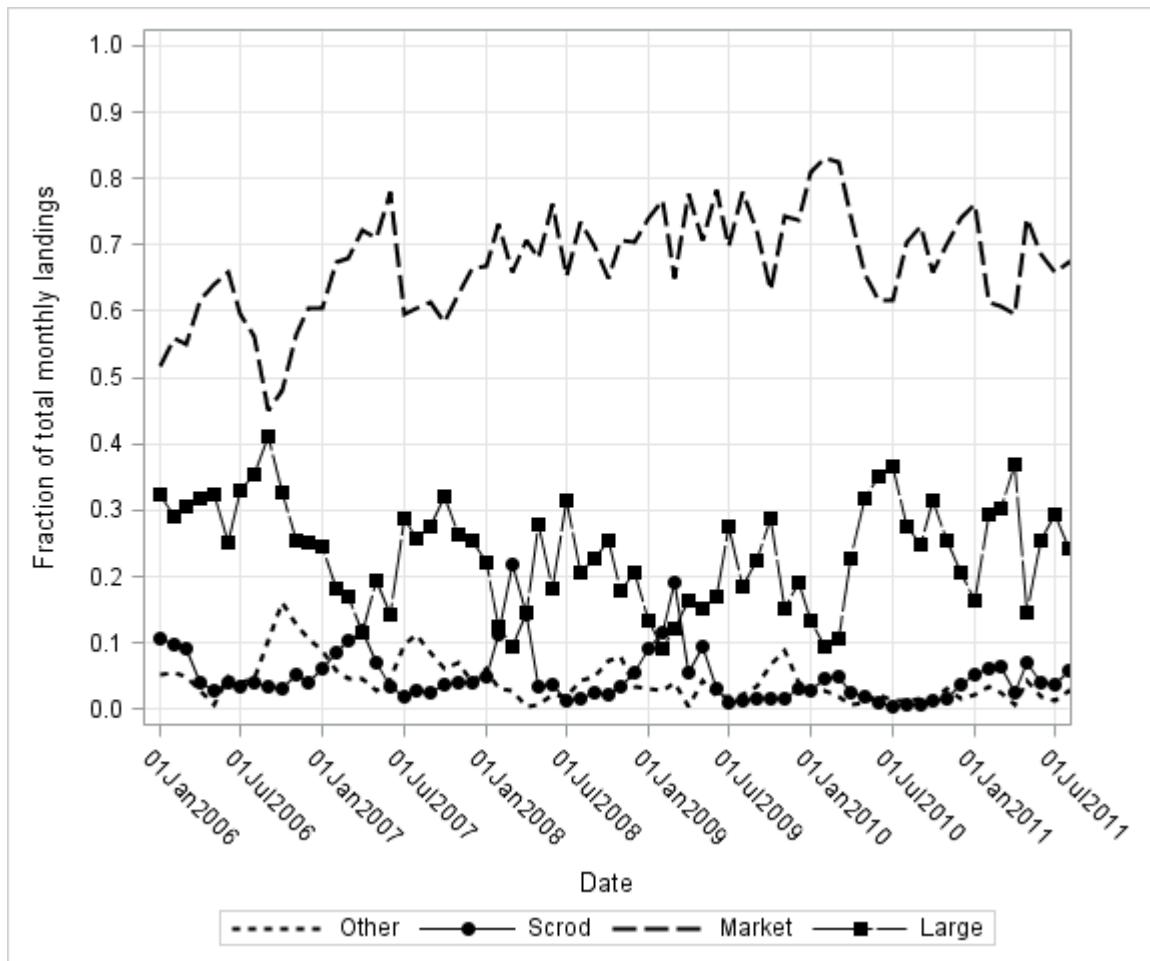


Figure A.42. Monthly commercial landing patterns (as a fraction of the total landings) of Gulf of Maine Atlantic cod by market category from 2006 to 2011.

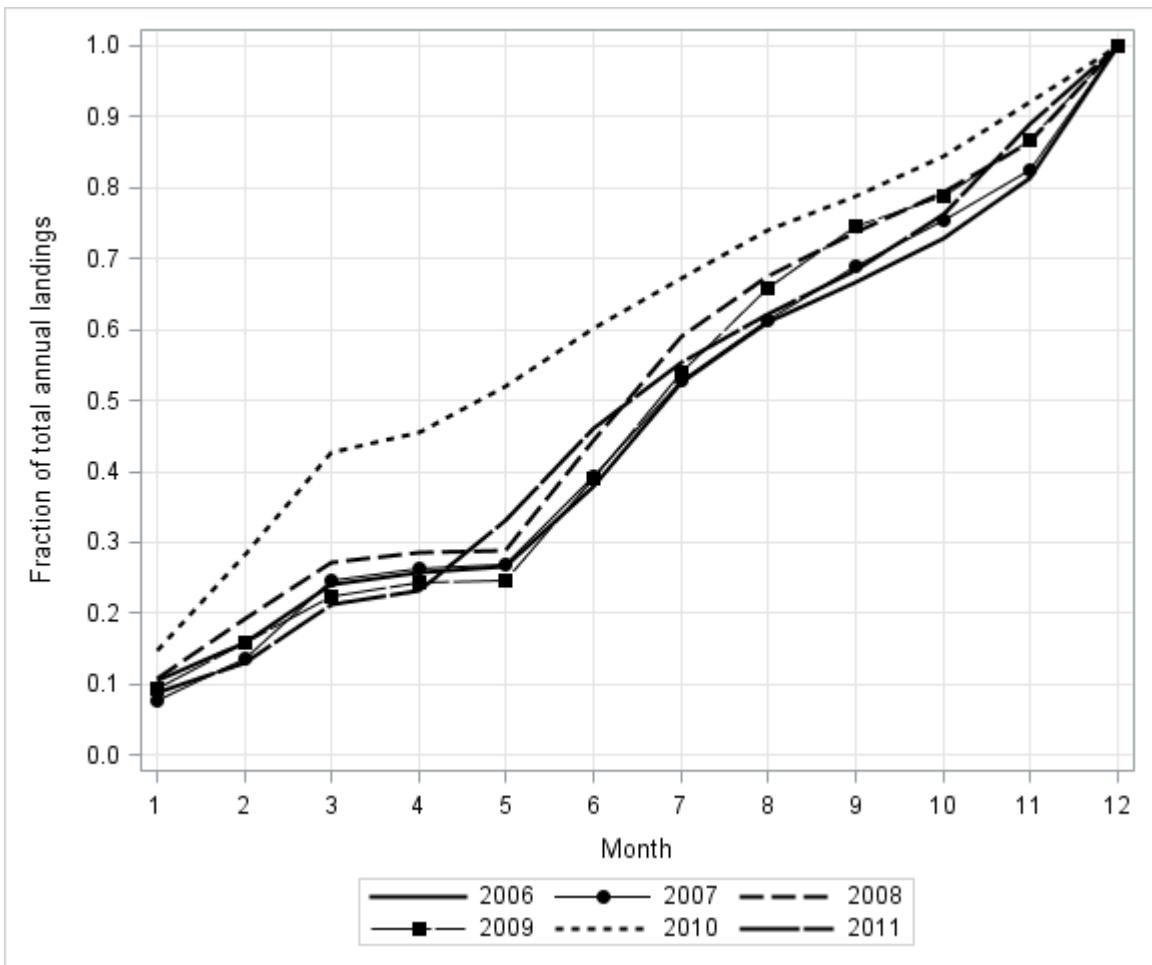


Figure A.43. Cumulative monthly commercial landings of Gulf of Maine Atlantic cod by year from 2006 to 2011.

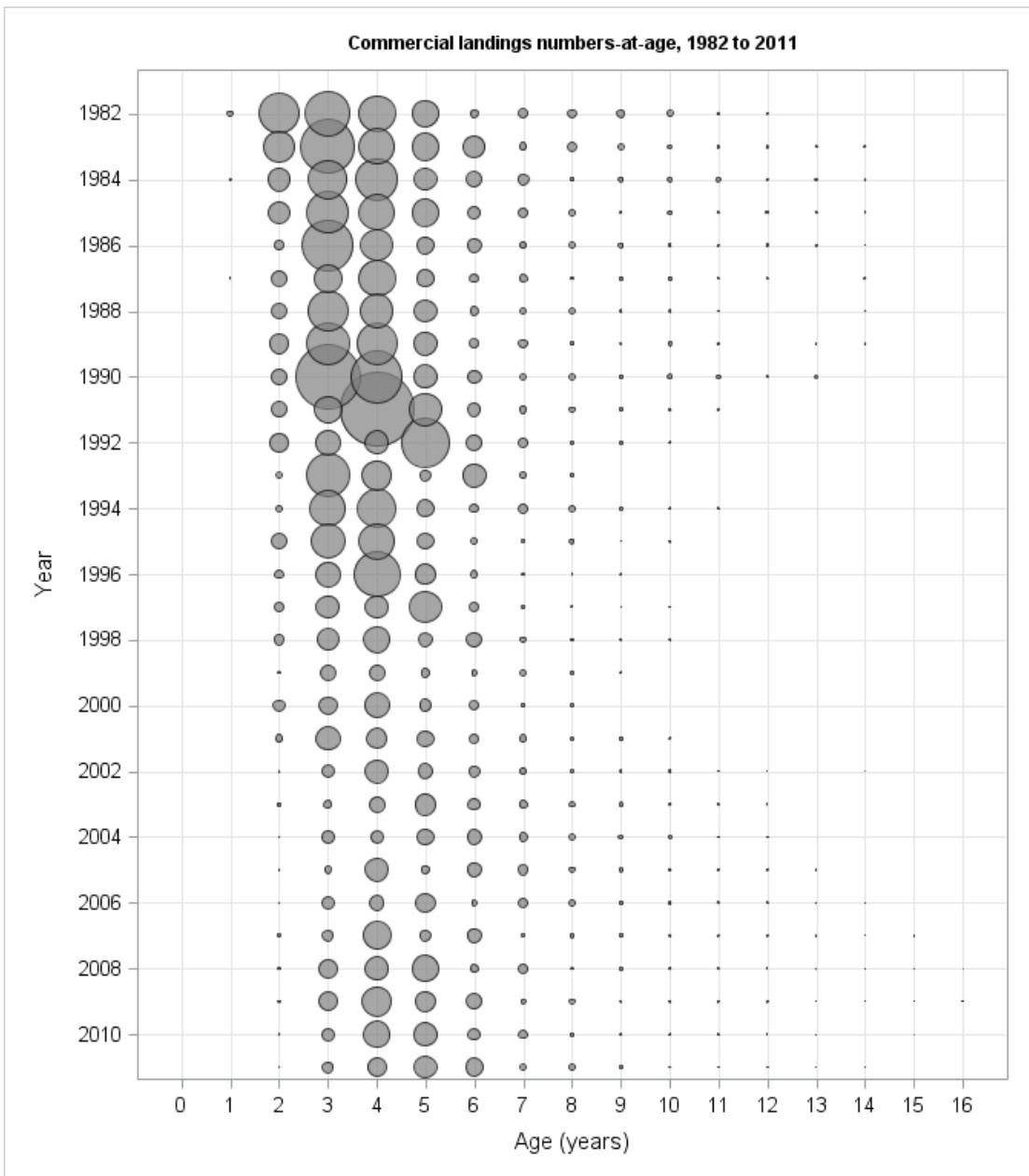


Figure A.44. Commercial landings-at-age of Gulf of Maine Atlantic cod from 1982 to 2011.

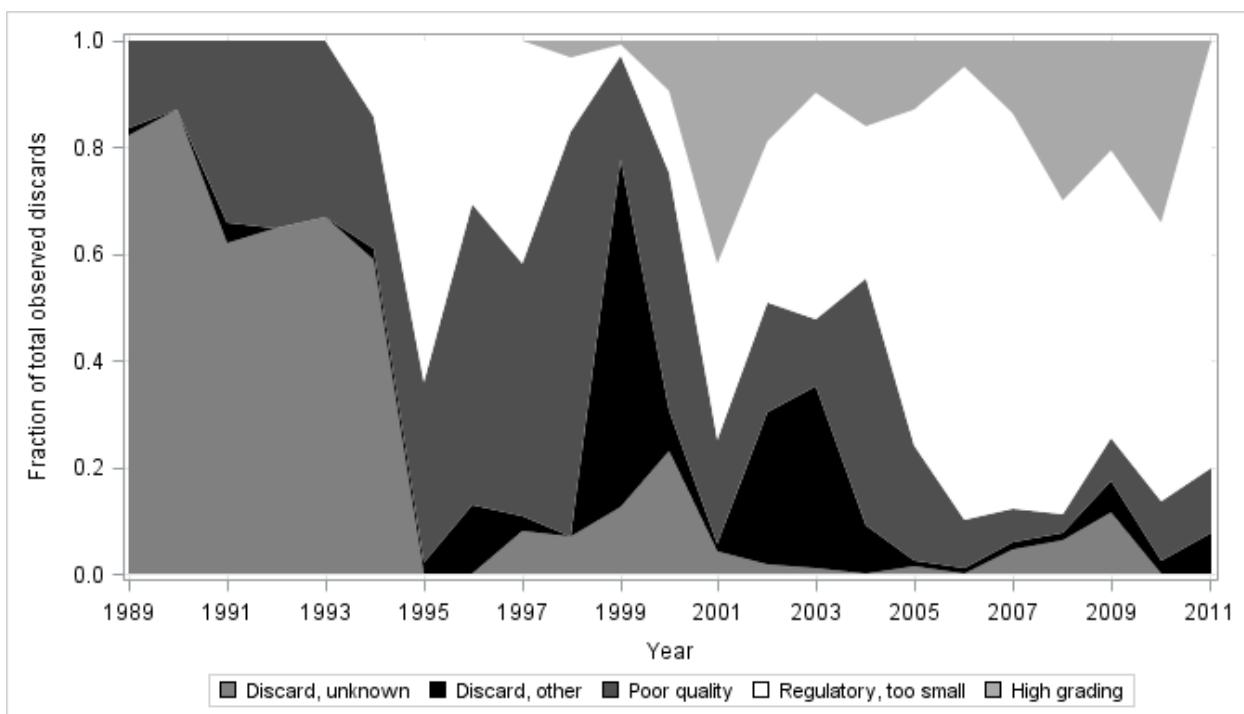


Figure A.45. Discard reasons for Gulf of Maine Atlantic cod as recorded by fisheries observers between 1989 and 2011.

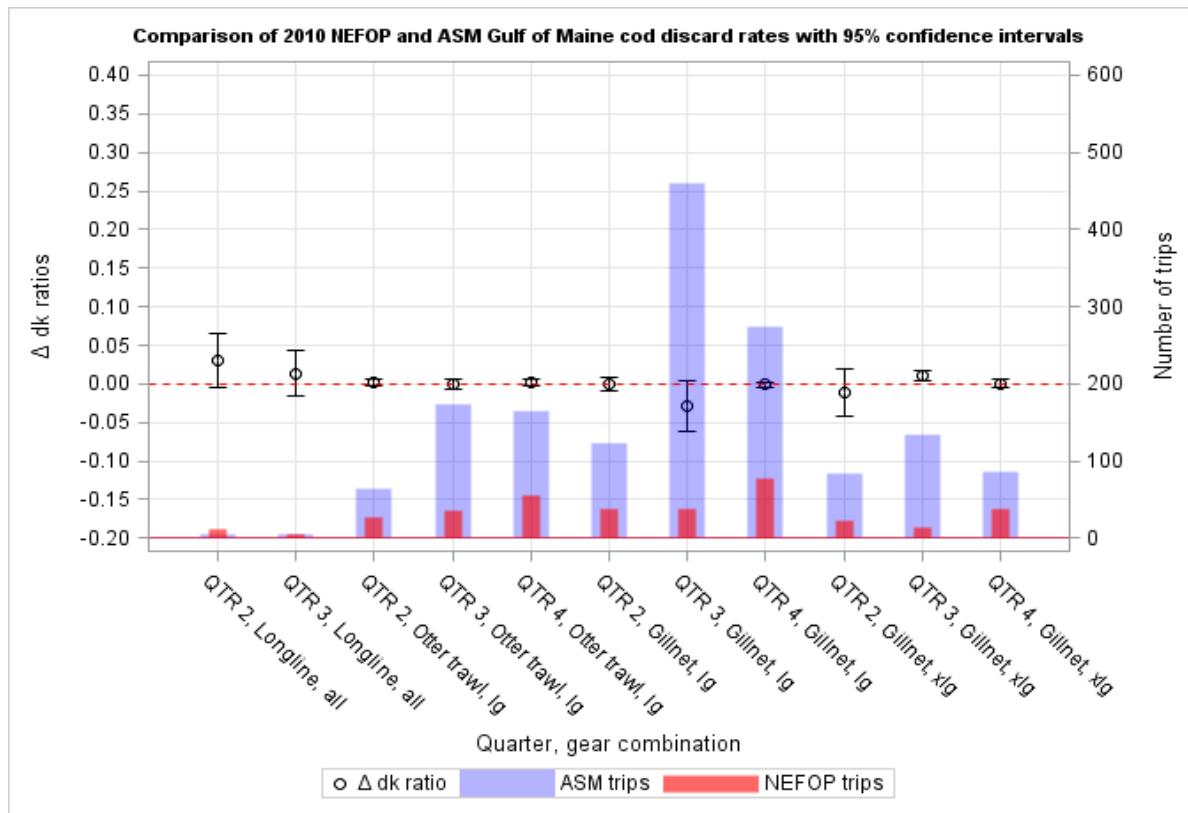


Figure A.46. Differences between the 2010 Gulf of Maine Atlantic cod discard rates estimated from data collected by groundfish at-sea monitors (ASMs) and certified observers showing 95% confidence intervals (top panel) and the number of trips included in each analysis (bottom panel) broken down by gear-mesh combination and quarter (adapted from Wigley et al. 2012).

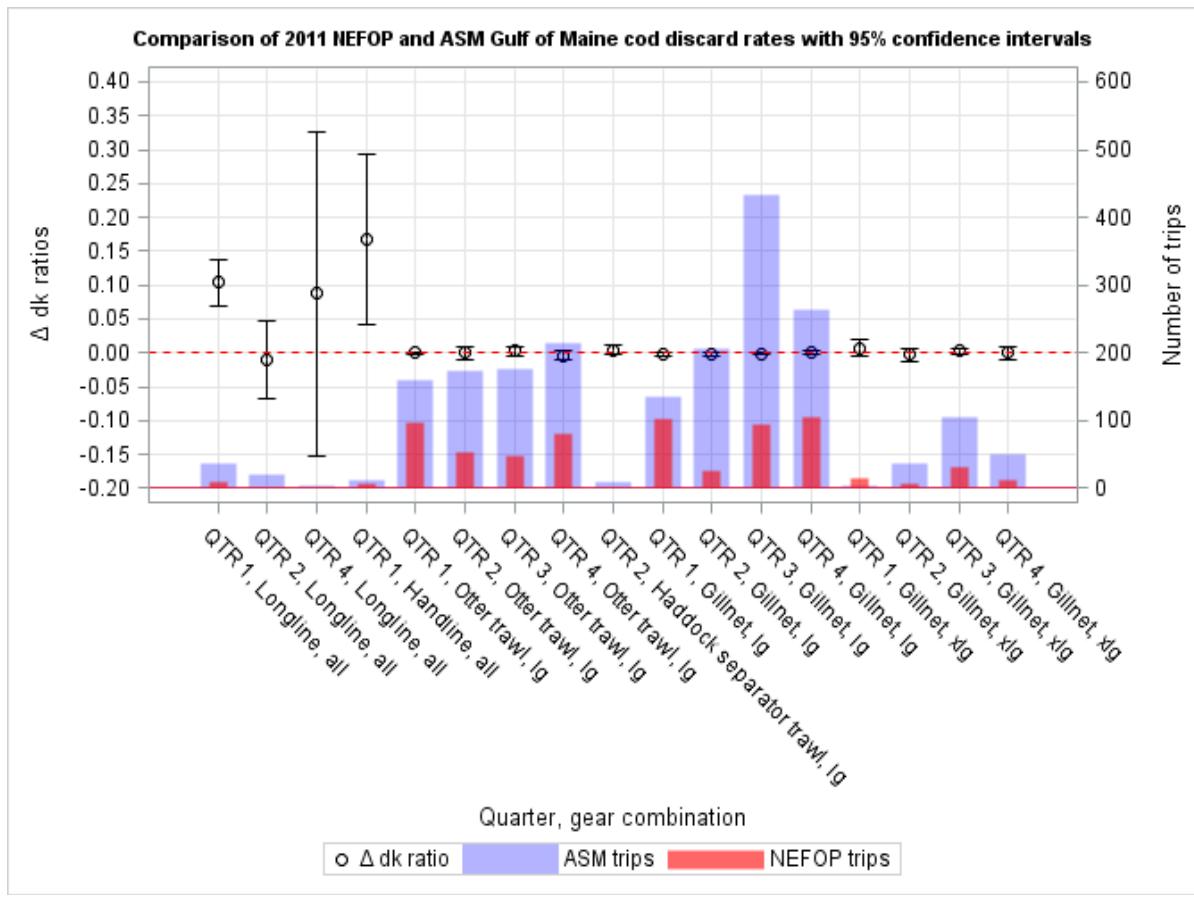


Figure A.47. Differences between the 2011 Gulf of Maine Atlantic cod discard rates estimated from data collected by groundfish at-sea monitors (ASMs) and certified observers showing 95% confidence intervals (top panel) and the number of trips included in each analysis (bottom panel) broken down by gear-mesh combination and quarter (adapted from Wigley et al. 2012).

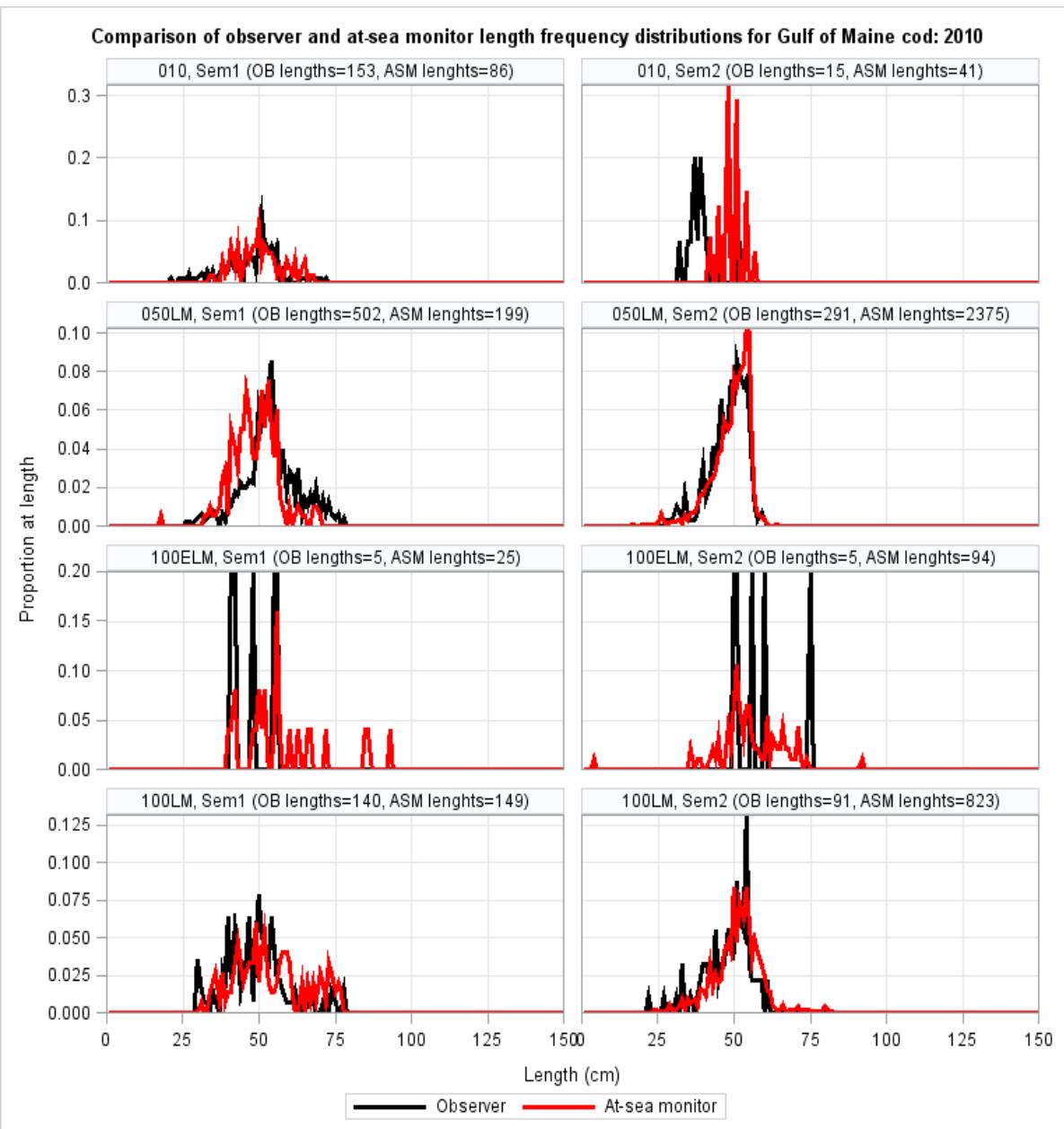


Figure A.48. Length frequency distributions of Gulf of Maine Atlantic cod commercials discards estimated from data collected by groundfish at-sea monitors (ASMs) and certified observers in 2010.

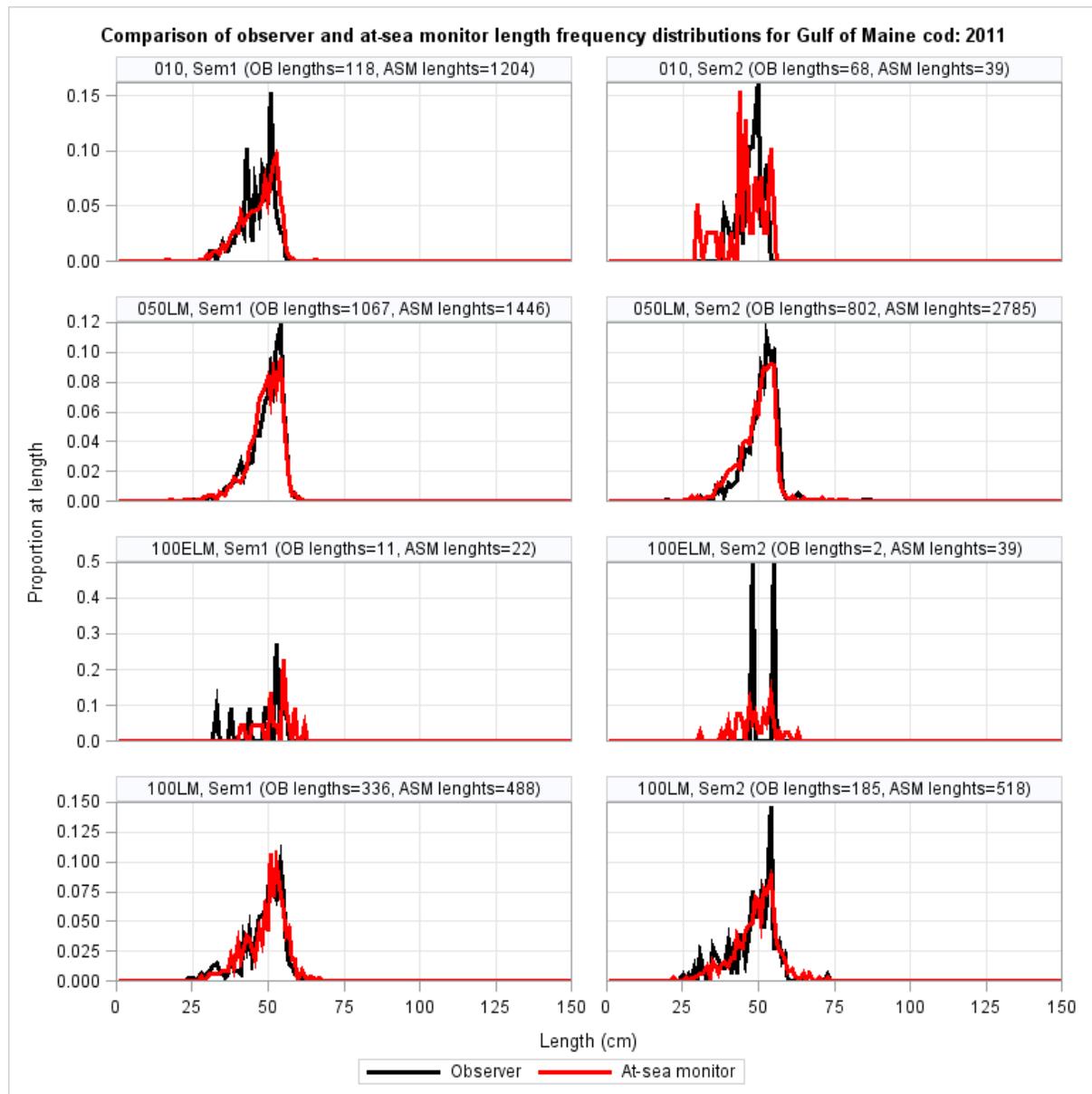


Figure A.49. Length frequency distributions of Gulf of Maine Atlantic cod commercials discards estimated from data collected by groundfish at-sea monitors (ASMs) and certified observers in 2011.

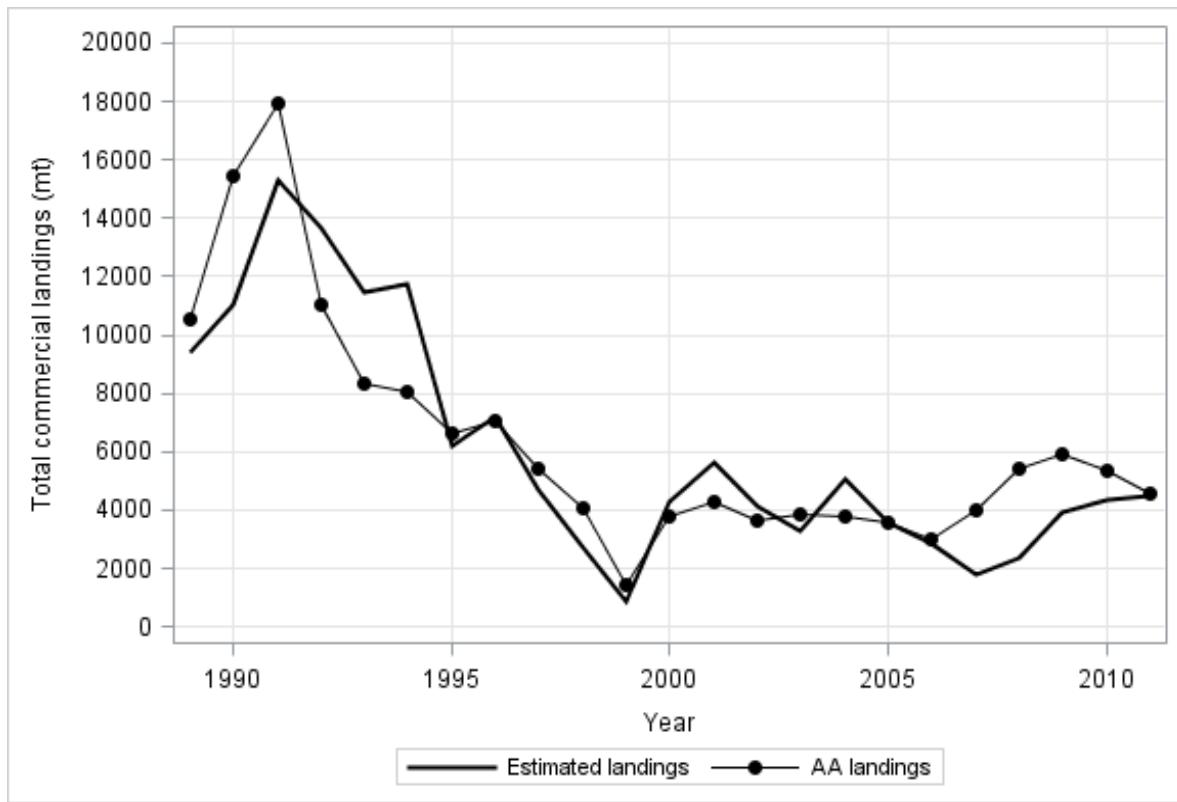


Figure A.50. Comparison of Gulf of Maine Atlantic cod landings estimates generated using the Standardized Bycatch Reporting Methodology (SBRM, Wigley et al. 2007) combined ratio approach to the stock landings from the Commercial Fisheries Database AA tables. Landings are shown only for longline, gillnet and otter trawl gears; all gear types not included in the discard estimation procedure were considered ‘other’ gear types and excluded. The comparison provides a cross validation of both the discard estimation and landings allocation procedure.

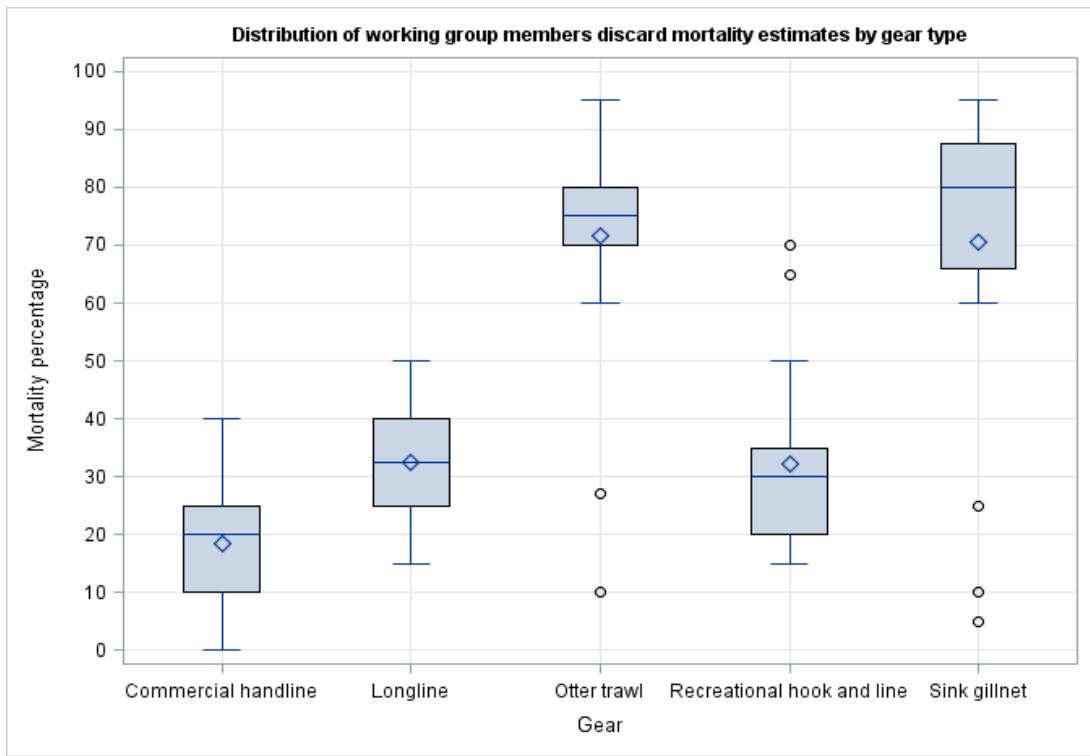


Figure A.51. Box plot distribution of the discard survival estimates by gear type developed by the Discard Mortality Working Group (expressed as percent mortality; NEFSC 2012b). Median estimates (horizontal blue line within the interquartile boxes) were used to adjust discard estimates in the current assessment.

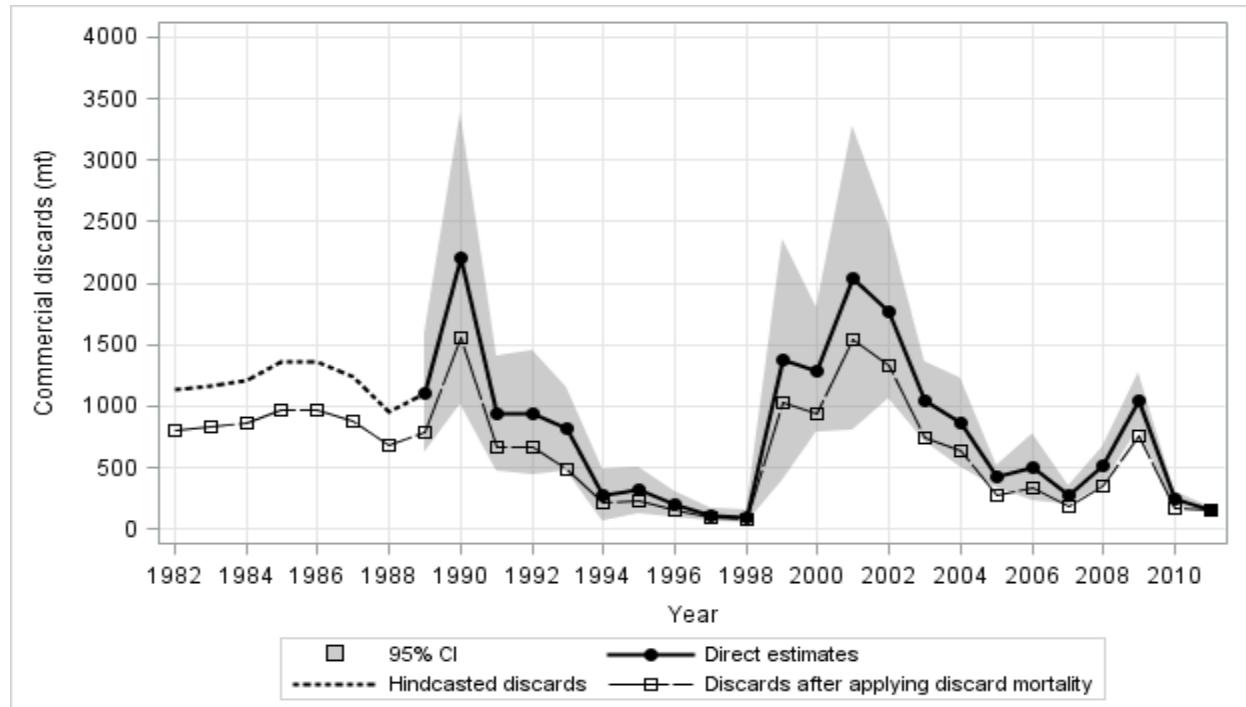


Figure A.52. Impacts of the revised discard mortality estimates on the estimates of commercial discards in terms of biomass (mt).

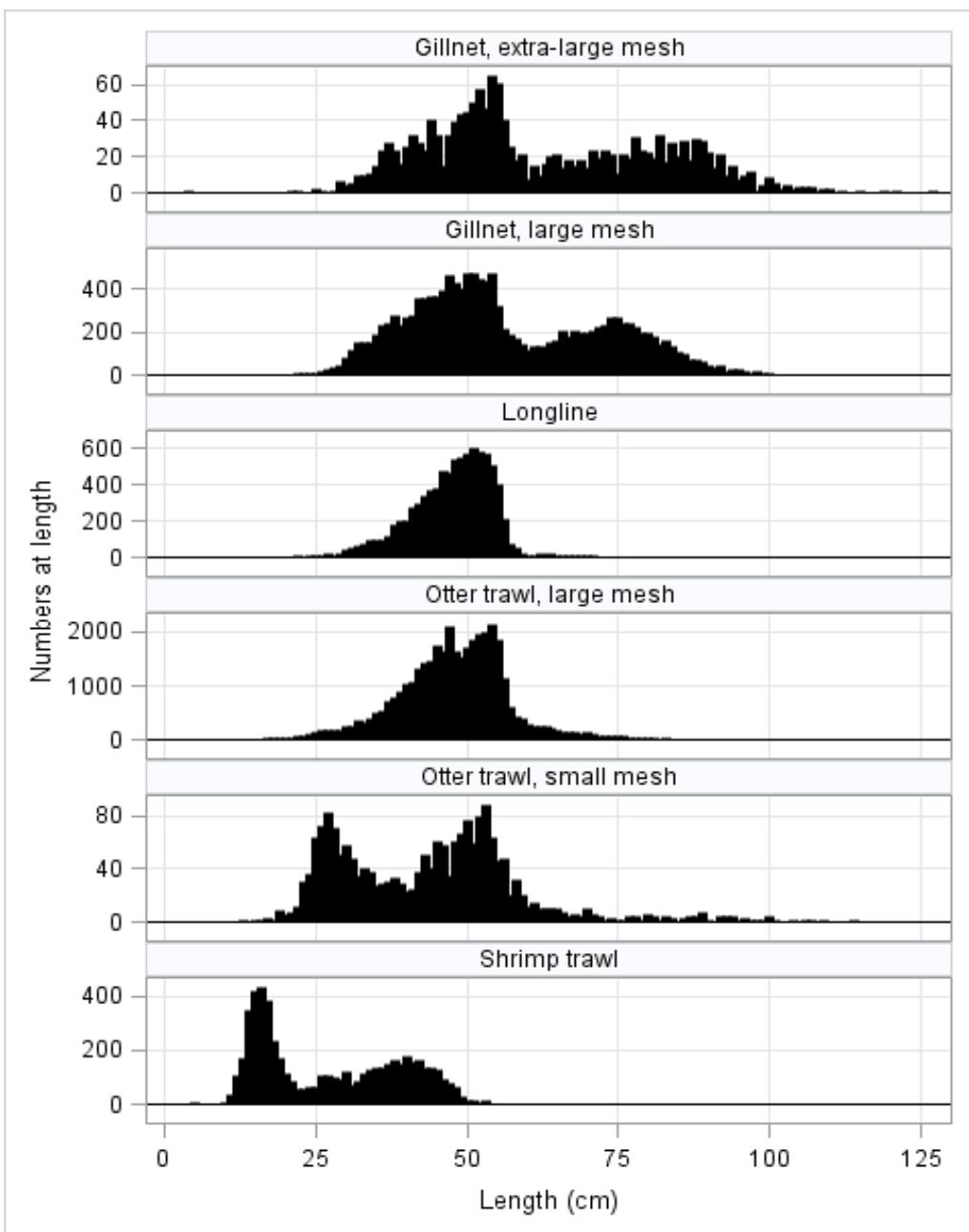


Figure A.53. Aggregate length frequency distributions, by gear type, of Gulf of Maine Atlantic cod discarded in the commercial fishery between 1989 and 2011.

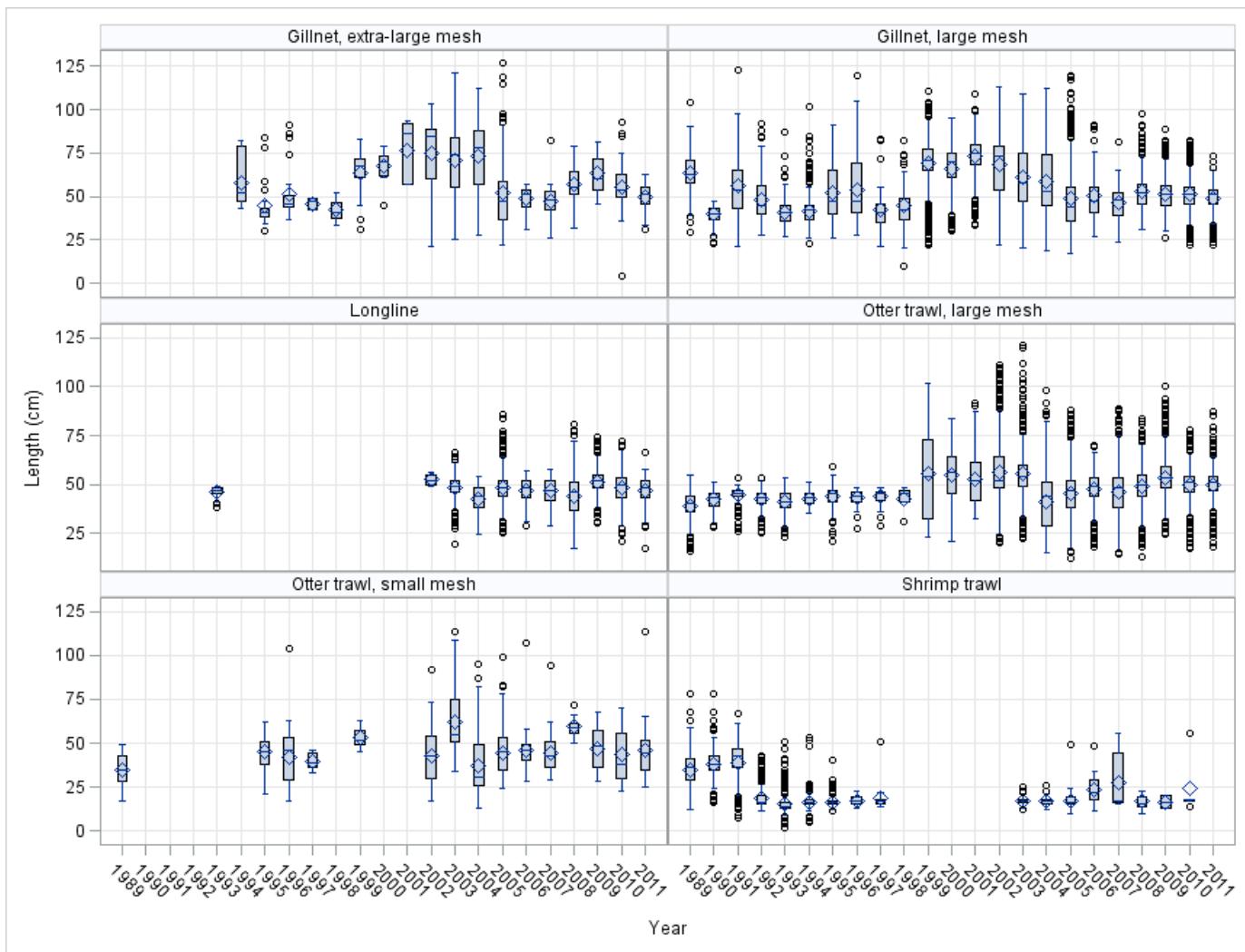


Figure A.54. Box plots showing the length distribution of Gulf of Maine Atlantic cod discarded by the commercial fishery by gear type between 1989 and 2011. Missing years indicate that there were either no observed trips for that gear in the Gulf of Maine or no cod were observed to have been discarded.

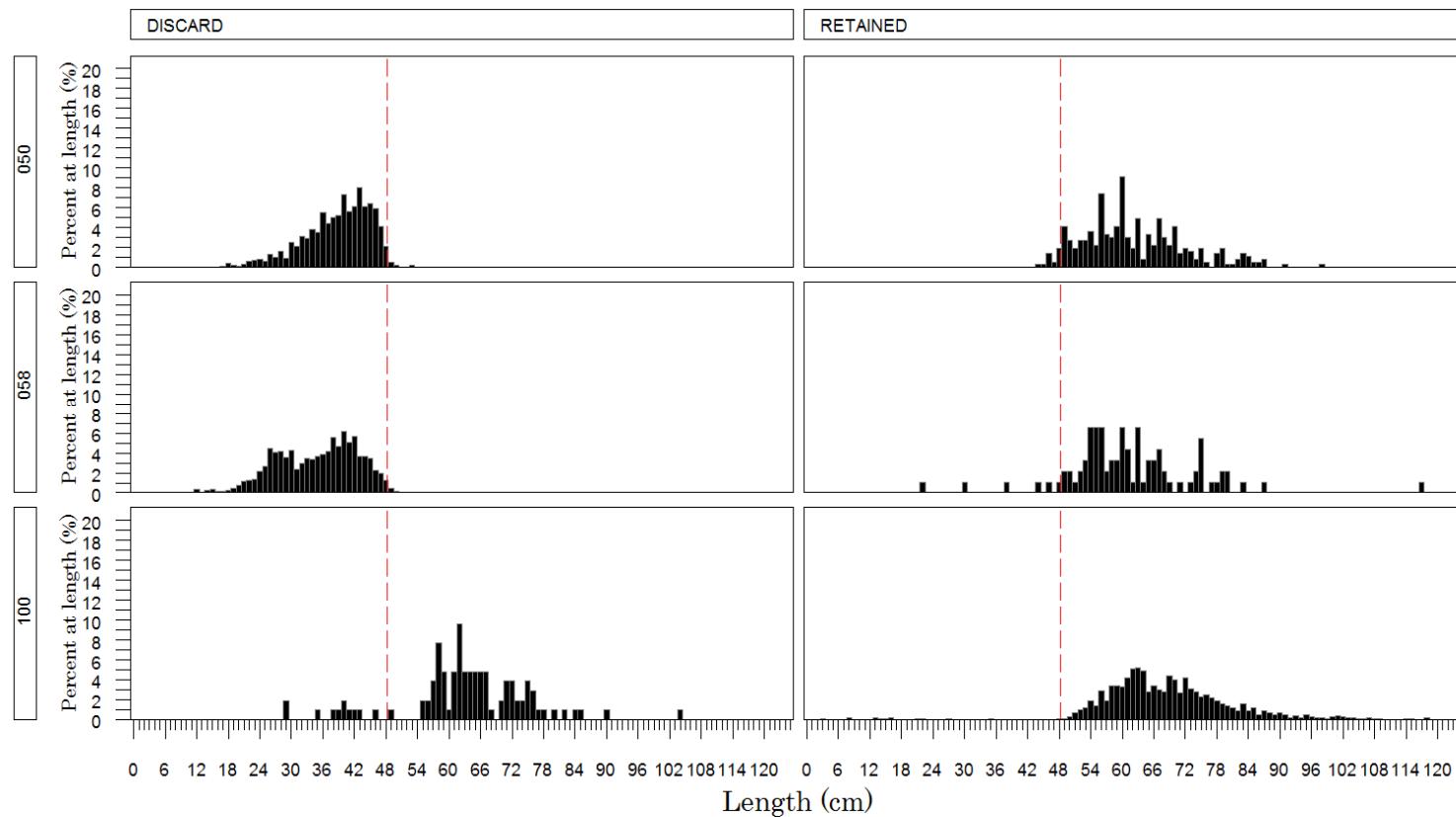


Figure A.55. Example of the length frequency distributions of Gulf of Maine Atlantic cod observed caught in the commercial fishery by large mesh otter trawl (050), shrimp trawl (058) and large mesh sink gillnet (100) gear in 1989. The 1989 – 1996 commercial minimum retention size of 19 inches (48.3 cm) is indicated by a dashed red line.

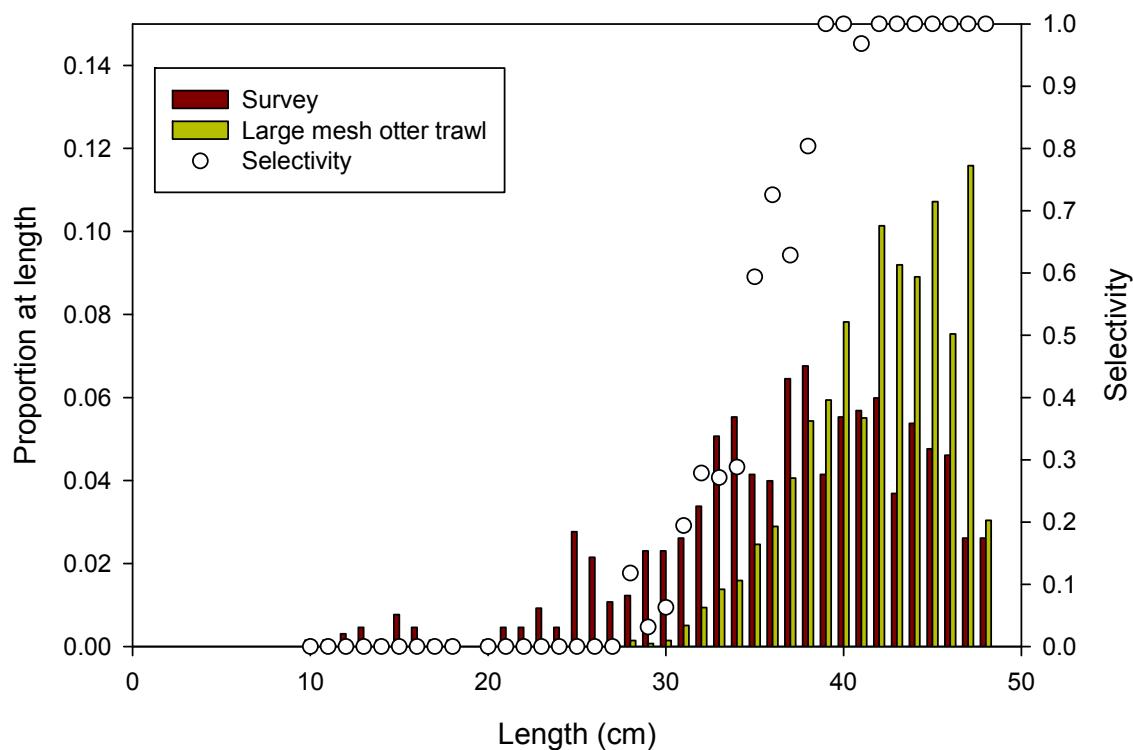


Figure A.56. Example of applying the survey-filter method to estimate the selectivity-at-length of fishing gears for Gulf of Maine Atlantic cod. In this example the proportion caught at length by large mesh otter trawl are compared to the proportion caught at-length from the Northeast Fishery Science Center spring and fall surveys (combined) to estimate the selectivity-at-length of large mesh otter trawl.

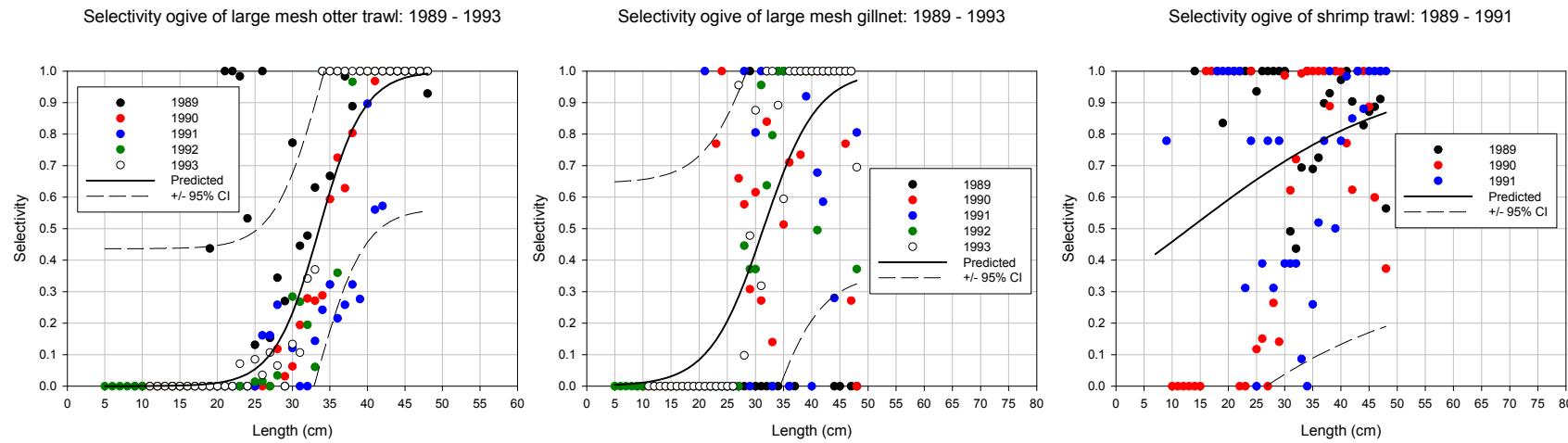
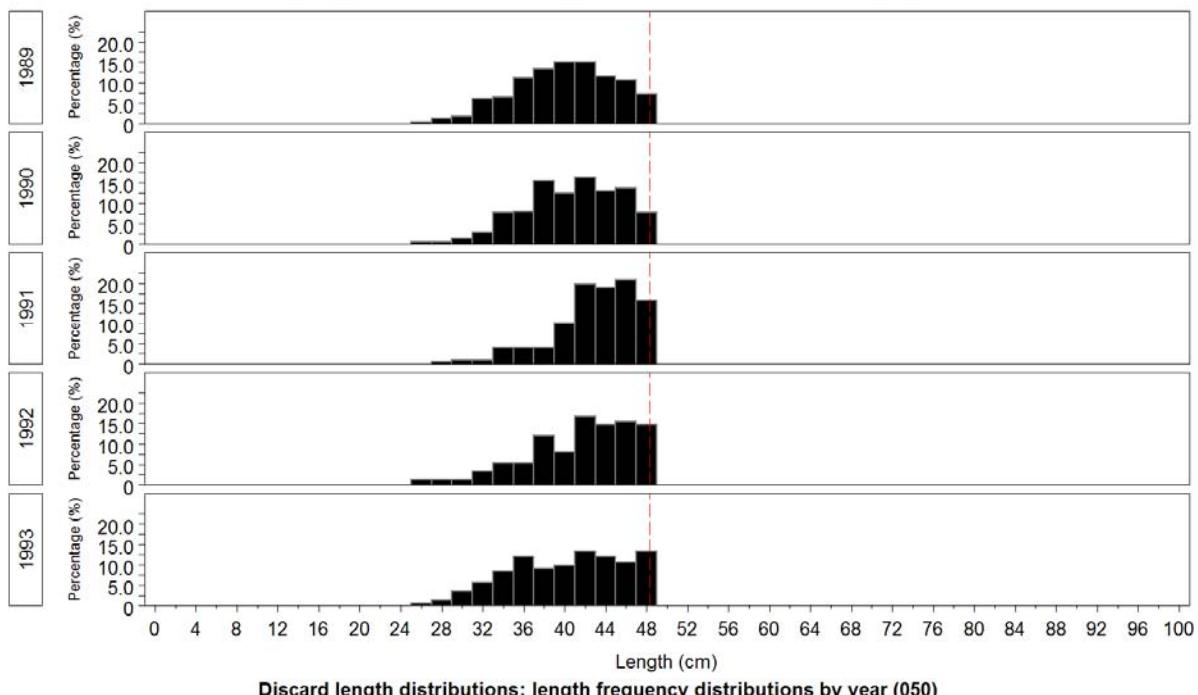


Figure A.57. Estimated selectivity ogives for large mesh otter trawl, large mesh sink gillnet and shrimp trawl and the corresponding 95% confidence intervals (CI) for Gulf of Maine Atlantic cod. Selectivity ogives were estimated from logistic fits to the aggregated annual estimates of selectivity-at-length.

Survey length distributions: after application of 050 discard selectivity ogives



Discard length distributions: length frequency distributions by year (050)

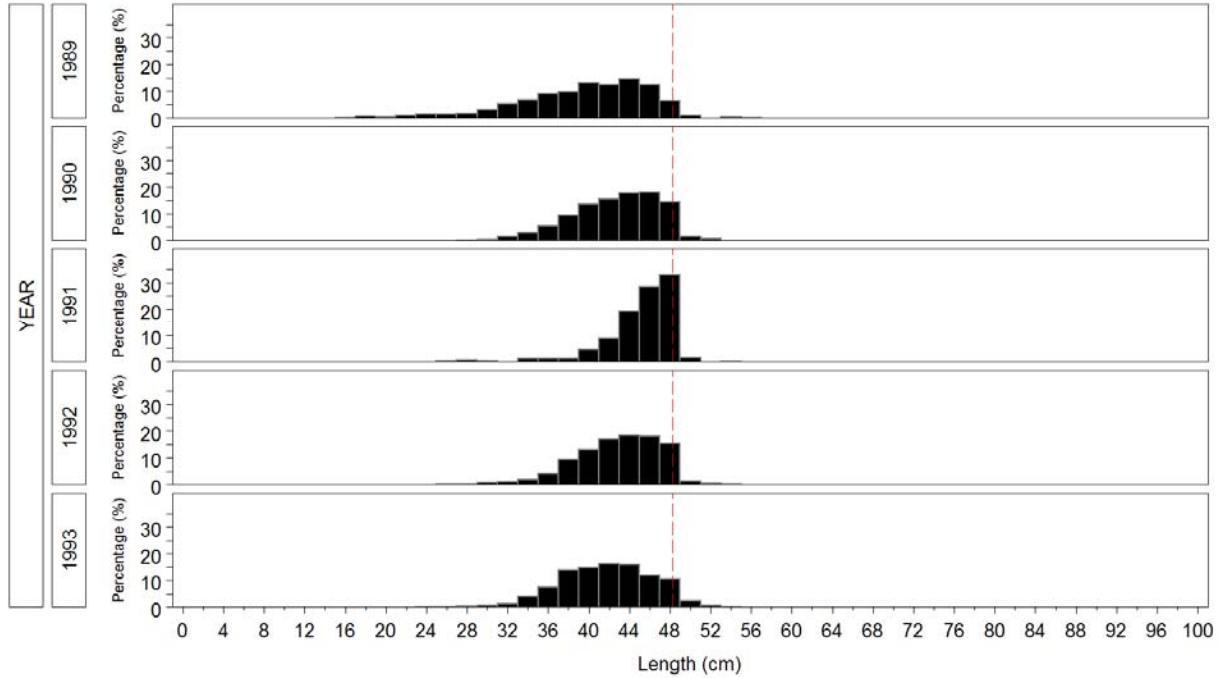
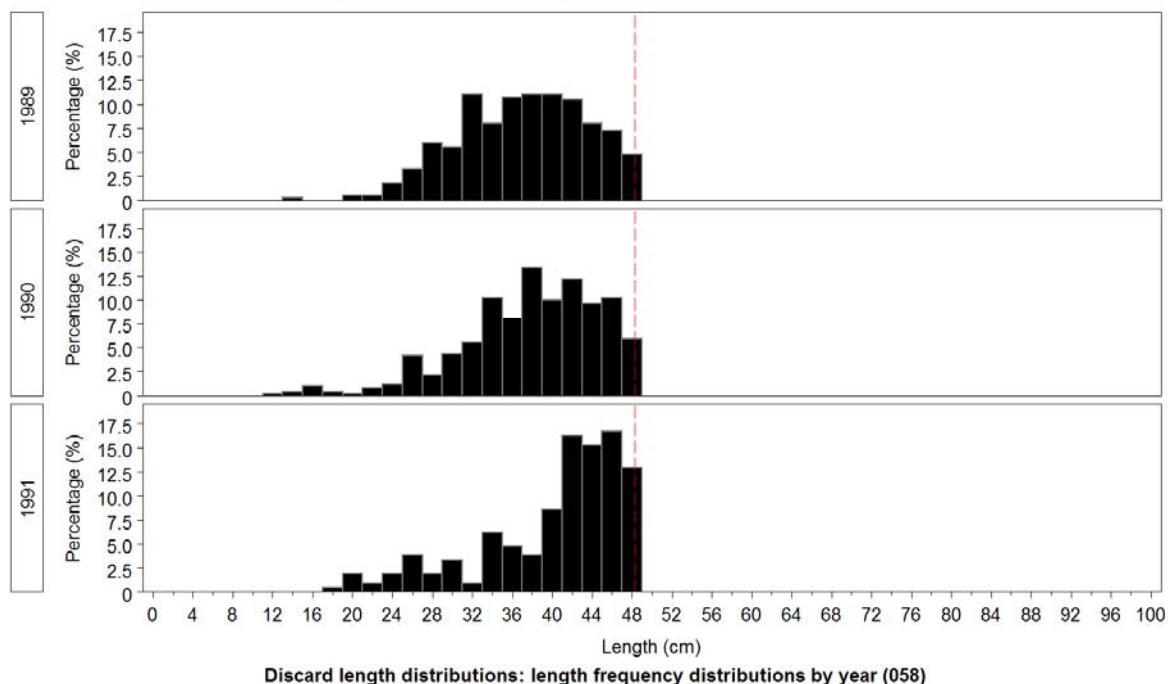


Figure A.58. Comparison of the survey filter-based estimates (top) of discards-at-length for large mesh otter trawl gear to the direct observer observations (bottom) from 1989 to 1993 for Gulf of Maine Atlantic cod. The dashed red line represents the commercial minimum retention size of 19 inches (48.3 cm) from 1989 to 1996.

Survey length distributions: after application of 058 discard selectivity ogives



Discard length distributions: length frequency distributions by year (058)

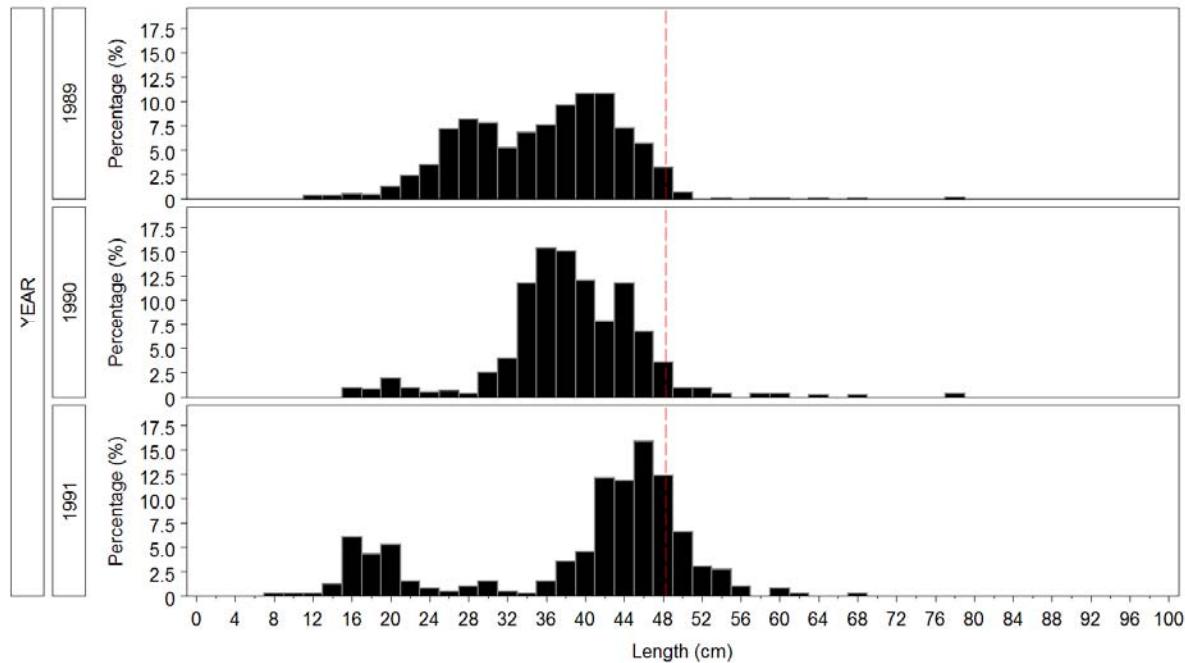


Figure A.59. Comparison of the survey filter-based estimates (top) of discards-at-length for shrimp trawl gear to the direct observer observations (bottom) from 1989 to 1991 for Gulf of Maine Atlantic cod. The dashed red line represents the commercial minimum retention size of 19 inches (48.3 cm) from 1989 to 1996.

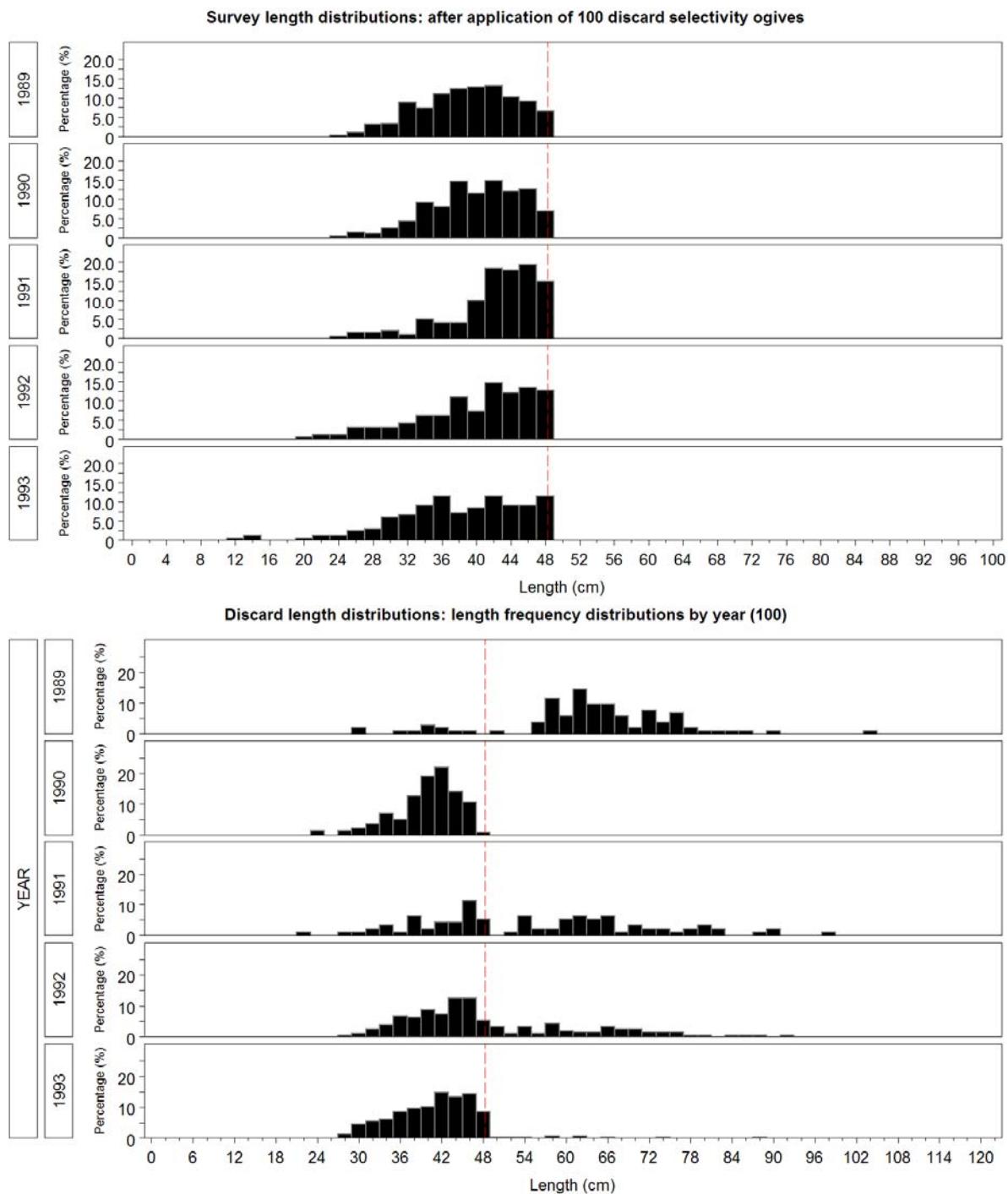


Figure A.60. Comparison of the survey filter-based estimates (top) of discards-at-length for large mesh sink gillnet gear to the direct observer observations (bottom) from 1989 to 1993 for Gulf of Maine Atlantic cod. The dashed red line represents the commercial minimum retention size of 19 inches (48.3 cm) from 1989 to 1996.

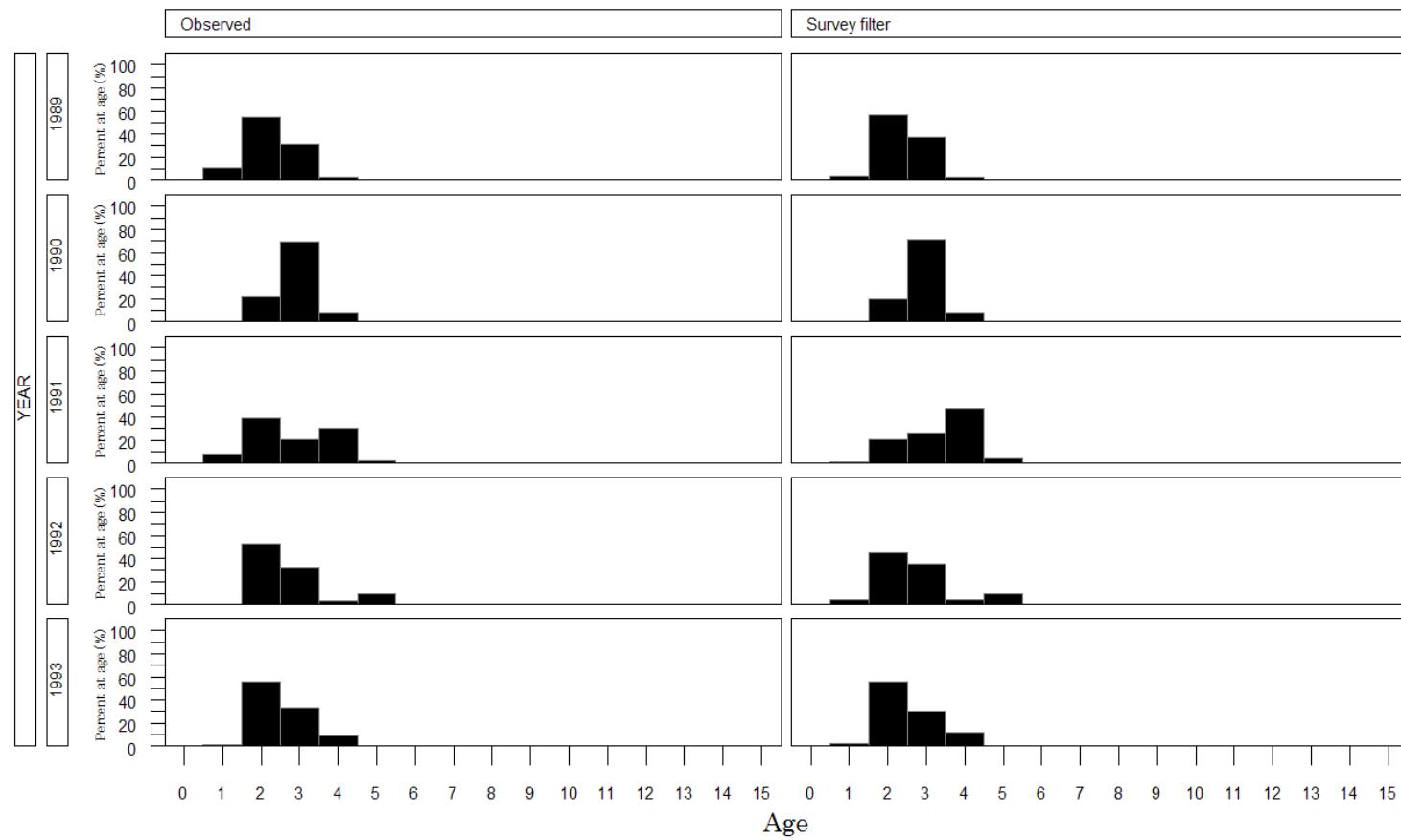


Figure A.61. Comparison of the survey filter-based estimates (right) of numbers-at-age for large mesh otter trawl gear to the direct observer observations (left) from 1989 to 1993 for Gulf of Maine Atlantic cod.

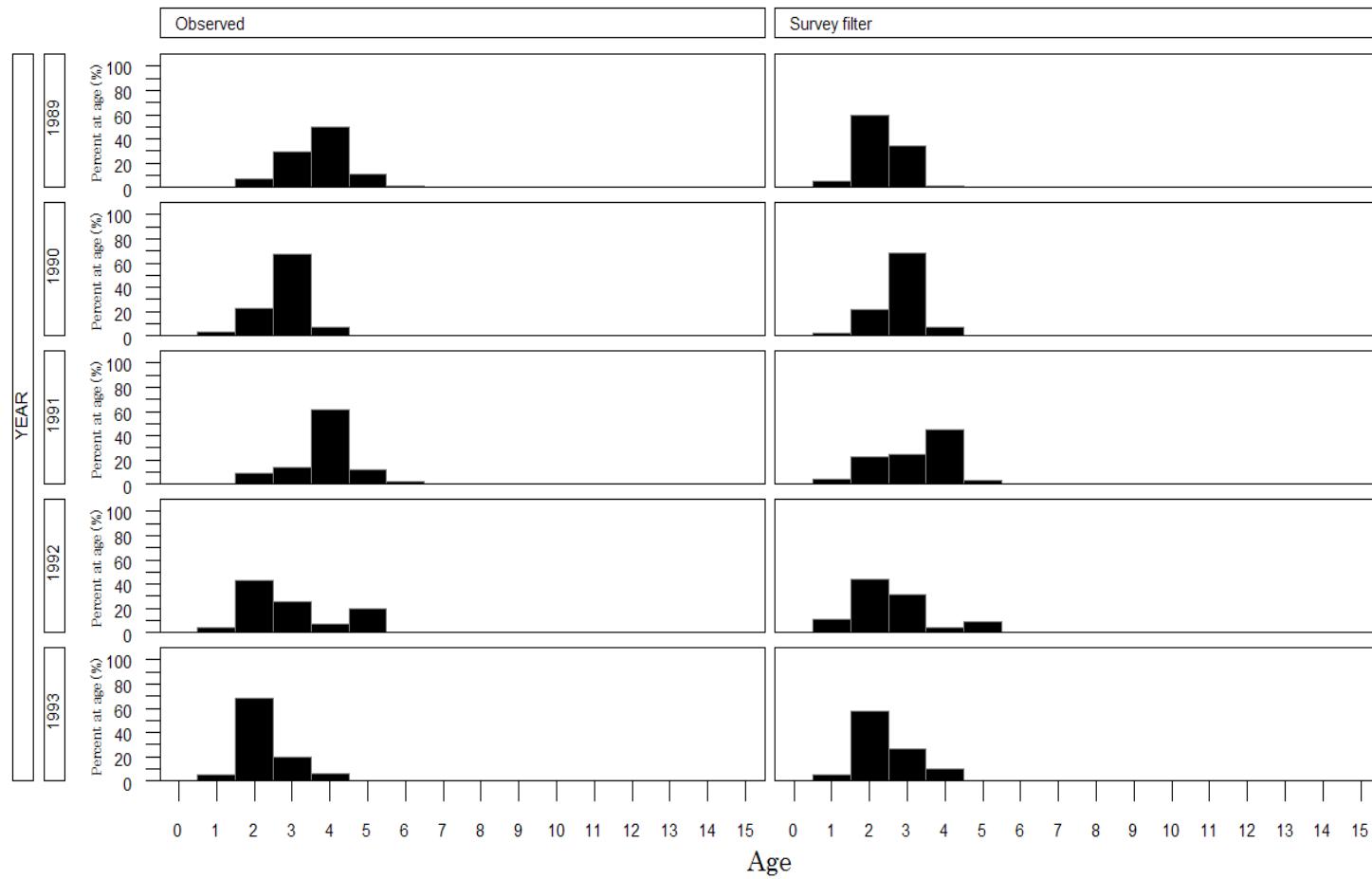


Figure A.62. Comparison of the survey filter-based estimates (right) of numbers-at-age for large mesh sink gillnet gear to the direct observer observations (left) from 1989 to 1993 for Gulf of Maine Atlantic cod.

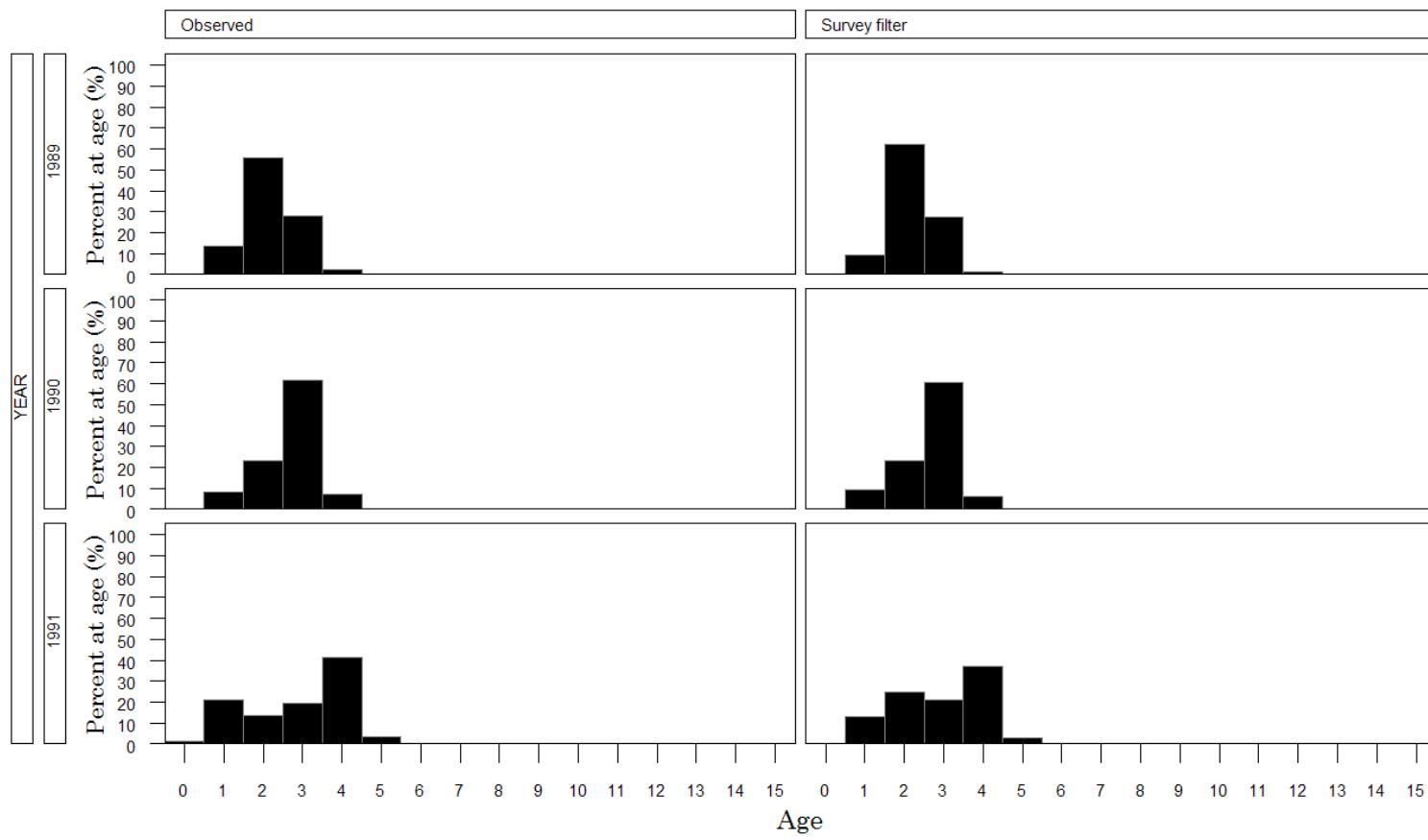


Figure A.63. Comparison of the survey filter-based estimates (right) of numbers-at-age for shrimp trawl gear to the direct observer observations (left) from 1989 to 1991 for Gulf of Maine Atlantic cod.

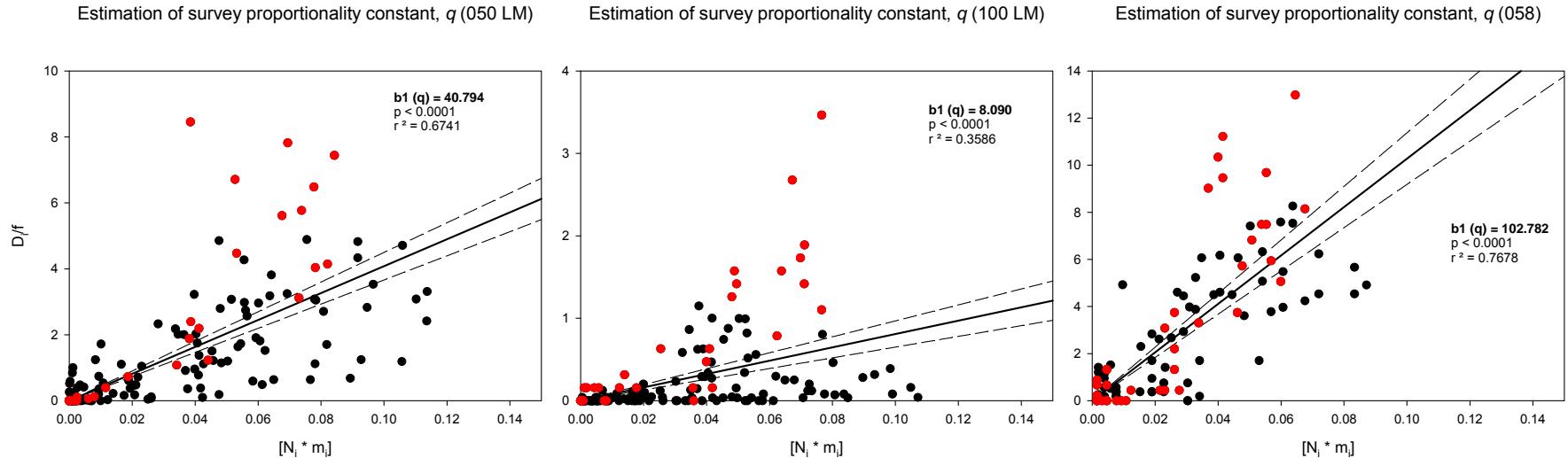


Figure A.64. Plots of the relationship by gear type between fraction of fish observed discarded-at-length (D_i/f) and the estimated number at length from the survey-filter method ($N_i \cdot m_i$) for Gulf of Maine Atlantic cod. Large mesh otter trawl (050 LM), large mesh sink gillnet (100 LM) and shrimp trawl gear (058) are shown. The slope of the relationship (q) is the proportionality constant required to expand the survey-filter estimates of numbers at length to estimates of total discards at length. The dots colored red represent observations from 1990.

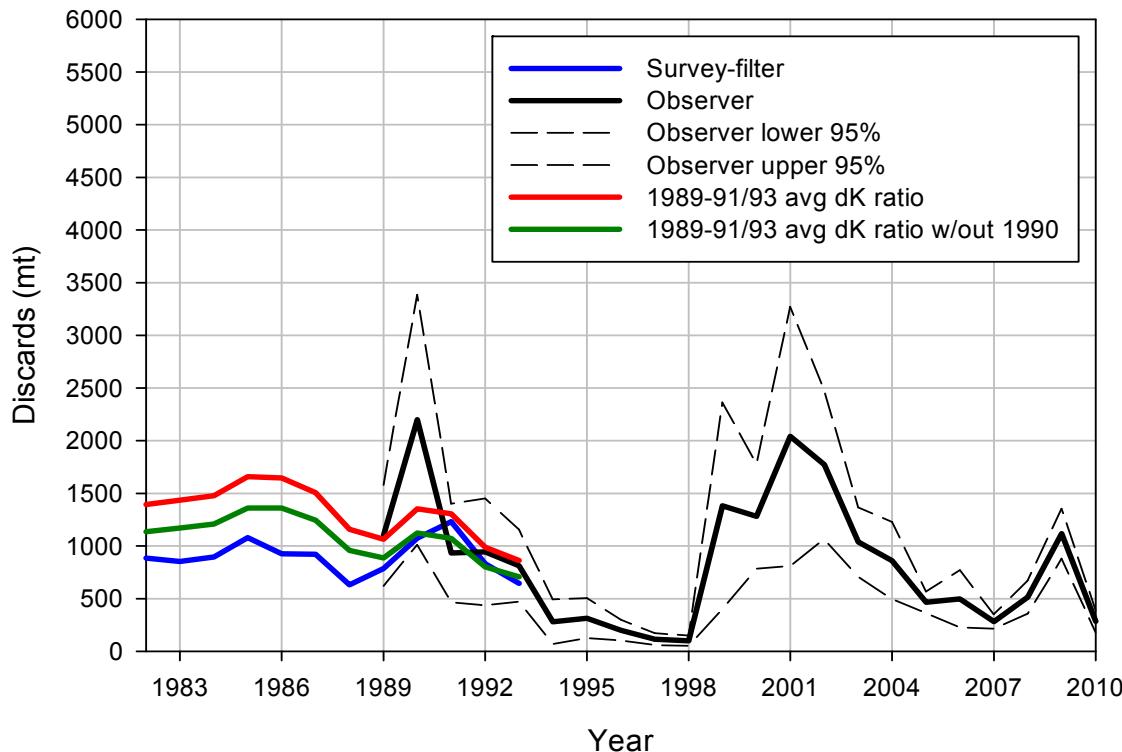


Figure A.65. Comparison of three different methods for achieving hindcasted estimates of Gulf of Maine Atlantic cod commercial discards from 1982 to 1988. (1) The survey-filter method uses the proportionality constant (q) multiplied by an index of fishing effort (total retained catch, K_{all}) to estimate total discards (blue line). (2) Use of the average ratio of discarded cod to total retained catch (d_{cod}/k_{all}) from 1989 to 1993 multiplied by total retained catch (K_{all} , red line). (3) Use of the average ratio of discarded cod to total retained catch (d_{cod}/k_{all}) from 1989 to 1993, excluding 1990, multiplied by total retained catch (K_{all} , green line). The ‘observer’ line shows the direct estimates of discards from 1989 to 2010 achieved using the Standardized Bycatch Reporting Methodology (Wigley et al. 2007) and the corresponding 95% confidence intervals.

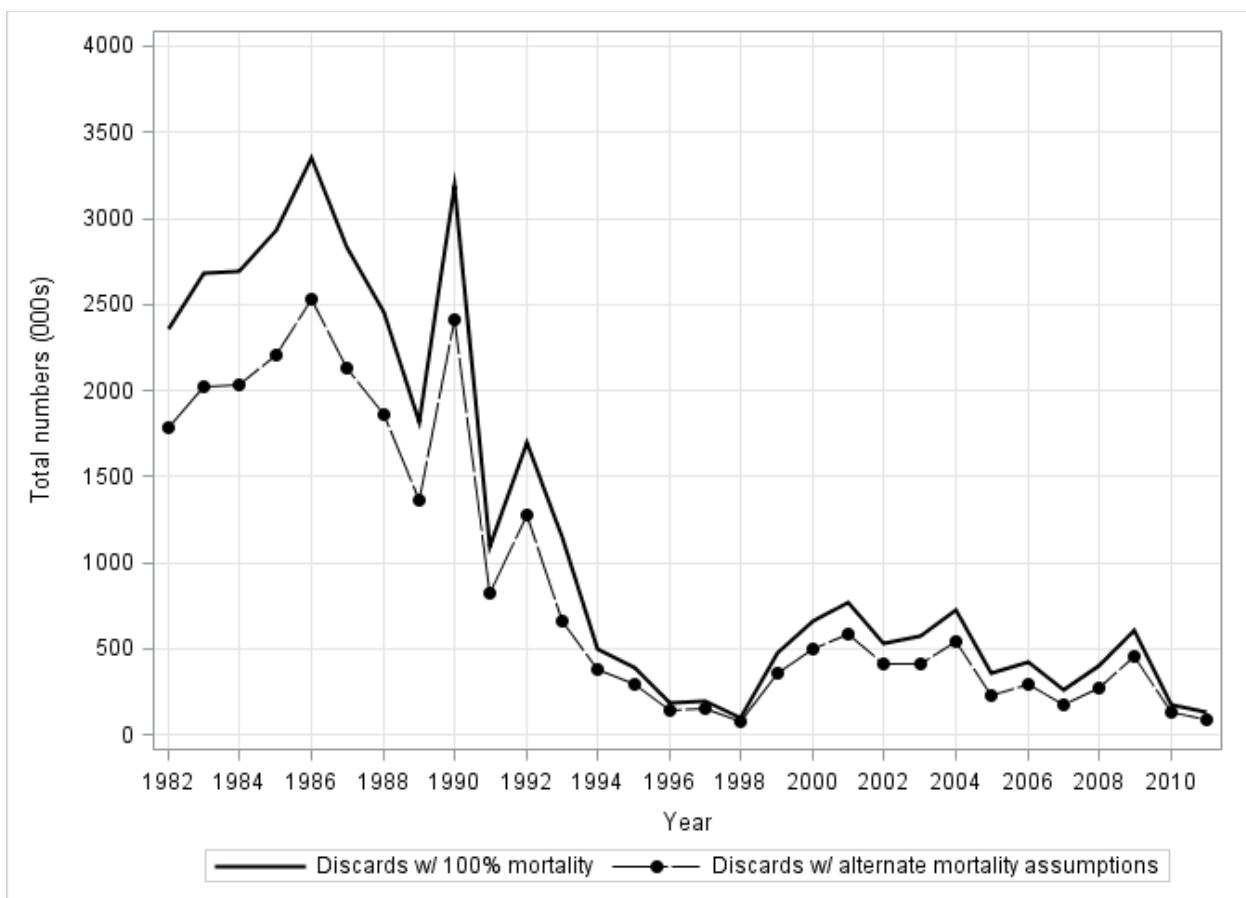


Figure A.66. Impacts of the revised discard mortality estimates on the estimates of Gulf of Maine Atlantic cod commercial discards in terms of numbers of fish (thousands).

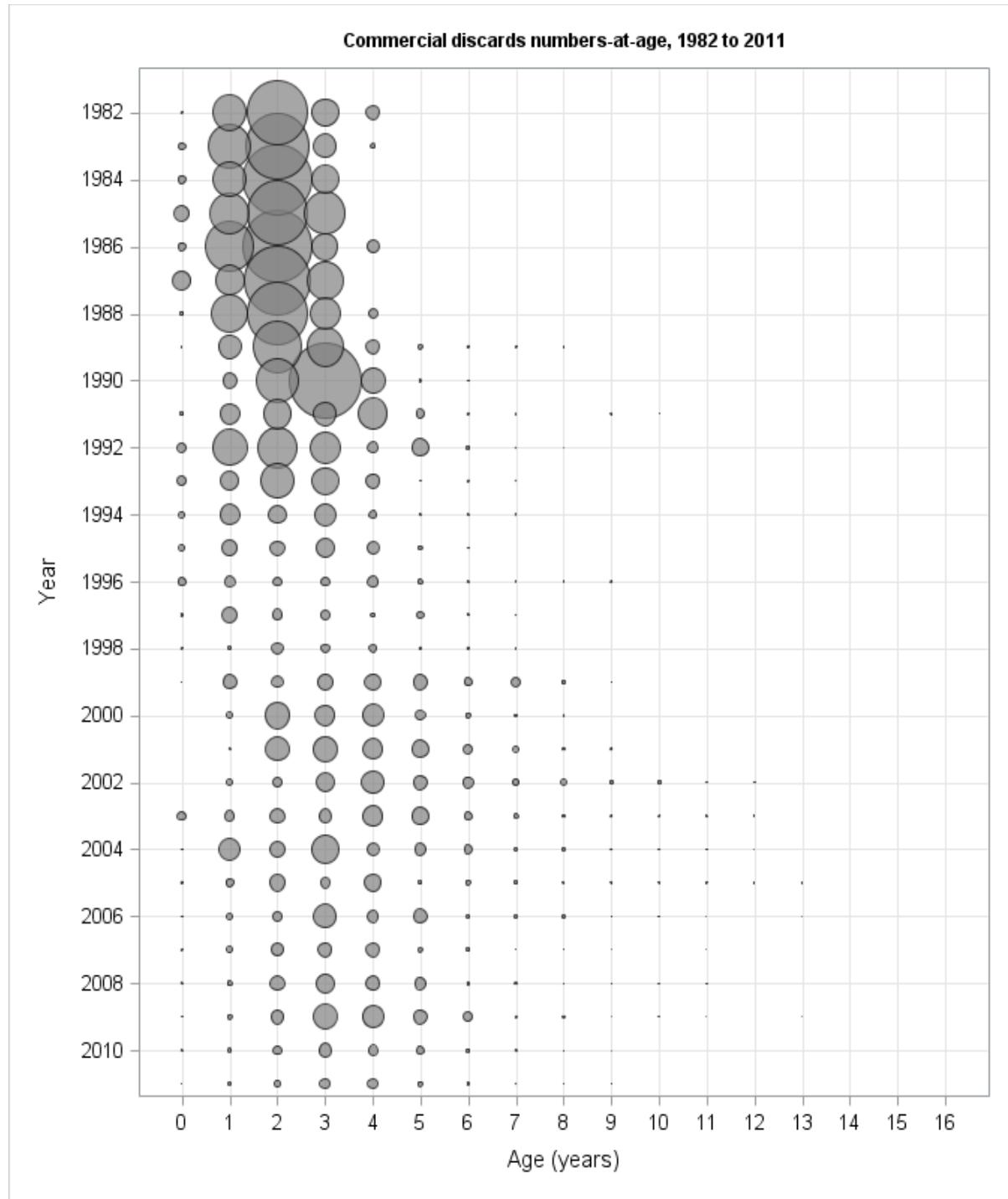


Figure A.67. Commercial discards-at-age of Gulf of Maine Atlantic cod from 1982 to 2011.

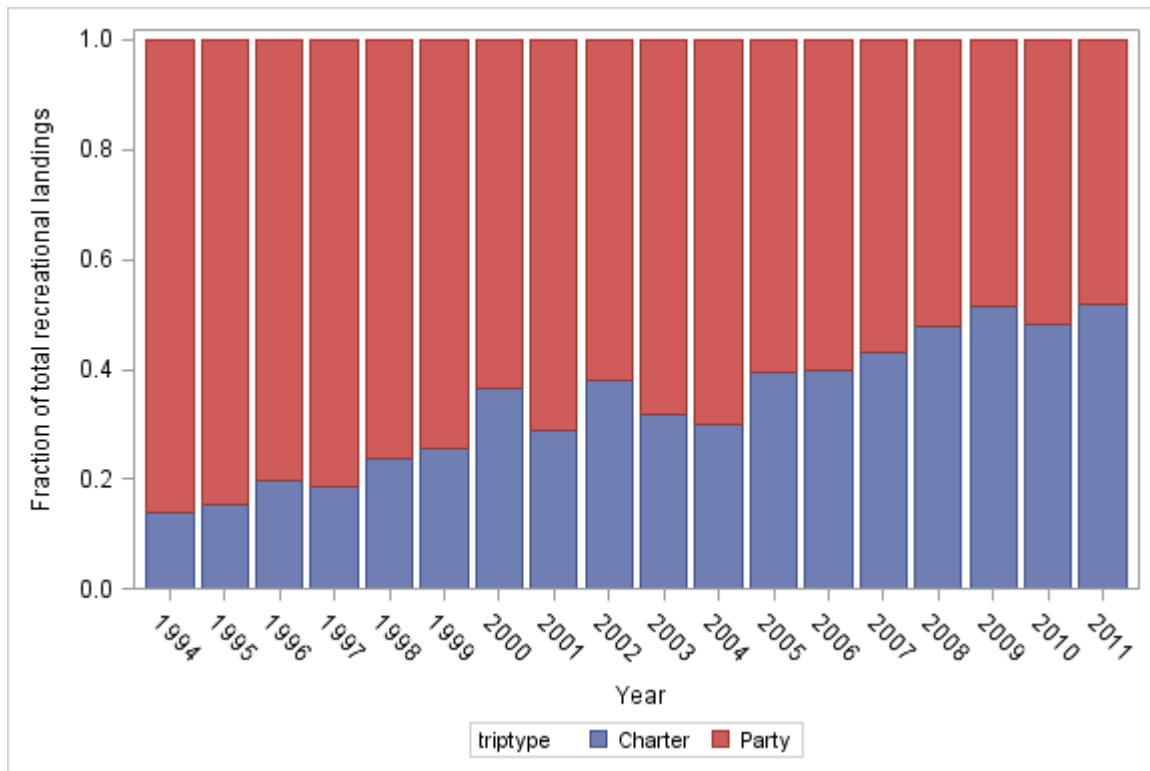


Figure A.68. Fraction of the total annual VTR-reported recreational Gulf of Maine Atlantic cod catch, by trip type, from 1994-2011.

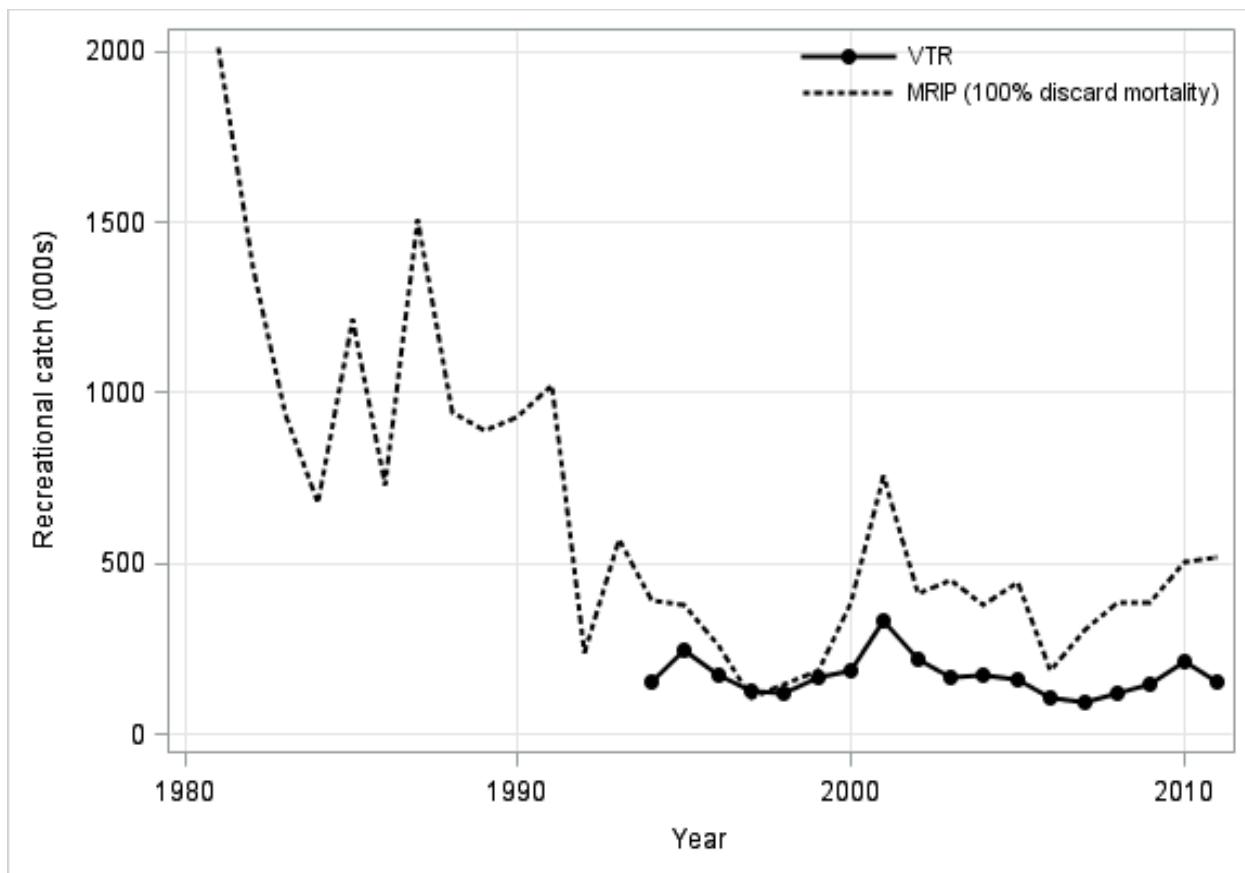


Figure A.69. Comparison of Gulf of Maine Atlantic cod recreational landings estimates derived through the Marine Recreational Information Program (MRIP) to recreational landings reported on Vessel Trip Reports (VTRs) between 1981 and 2011. *Note: VTR data collection began in 1994.



Figure A.70. Gini indices for the recreational charter and party boat fleets from 1994-2011. Indices were based on the spatial distribution of the retained catch reported on vessel trip reports.

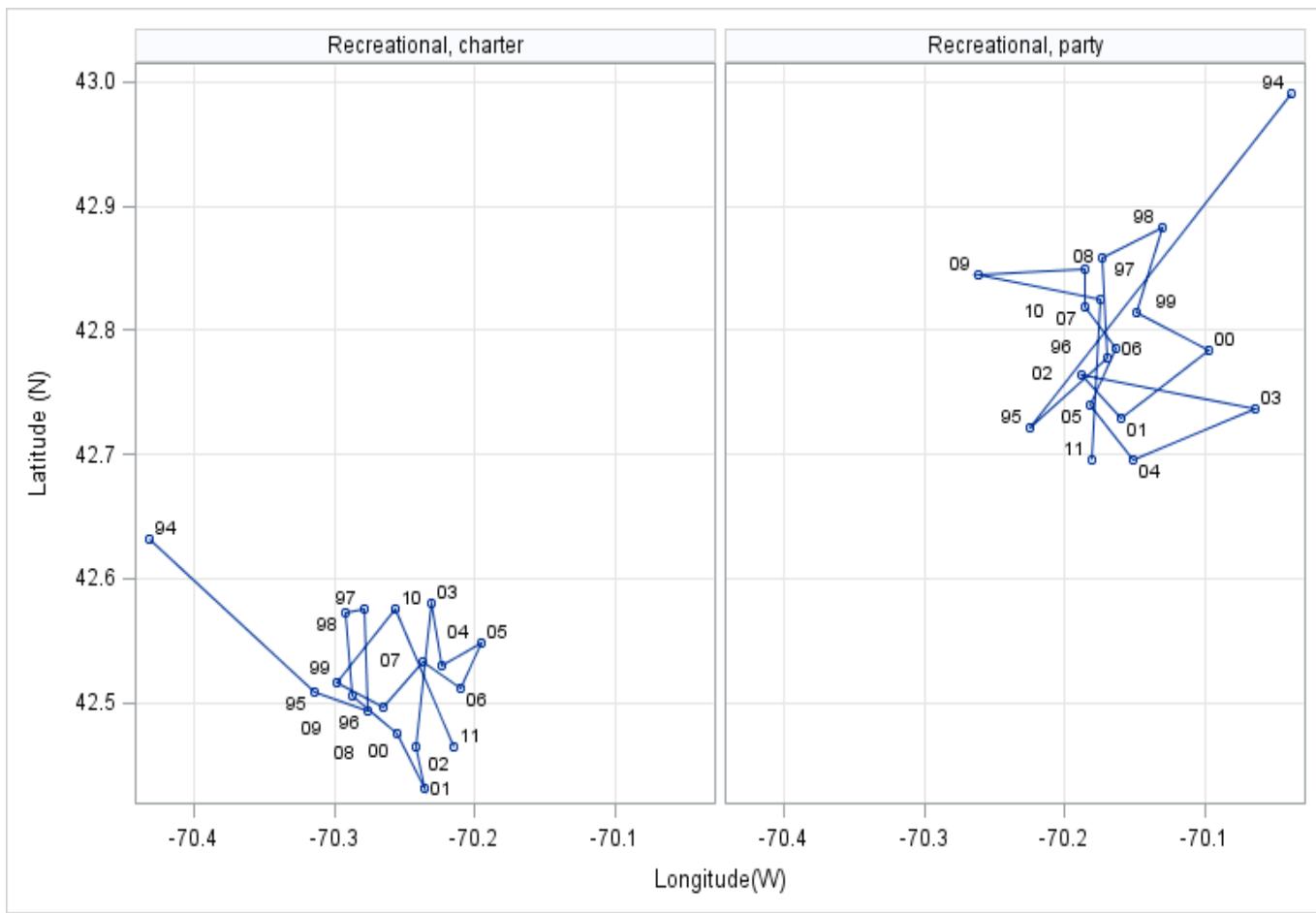


Figure A.71. Landings-weighted mean location (centroid) of Gulf of Maine cod catch by the recreational charter and party boat fleets from 1994 to 2011. Centroids were based on the spatial distribution of the retained catch reported on vessel trip reports.

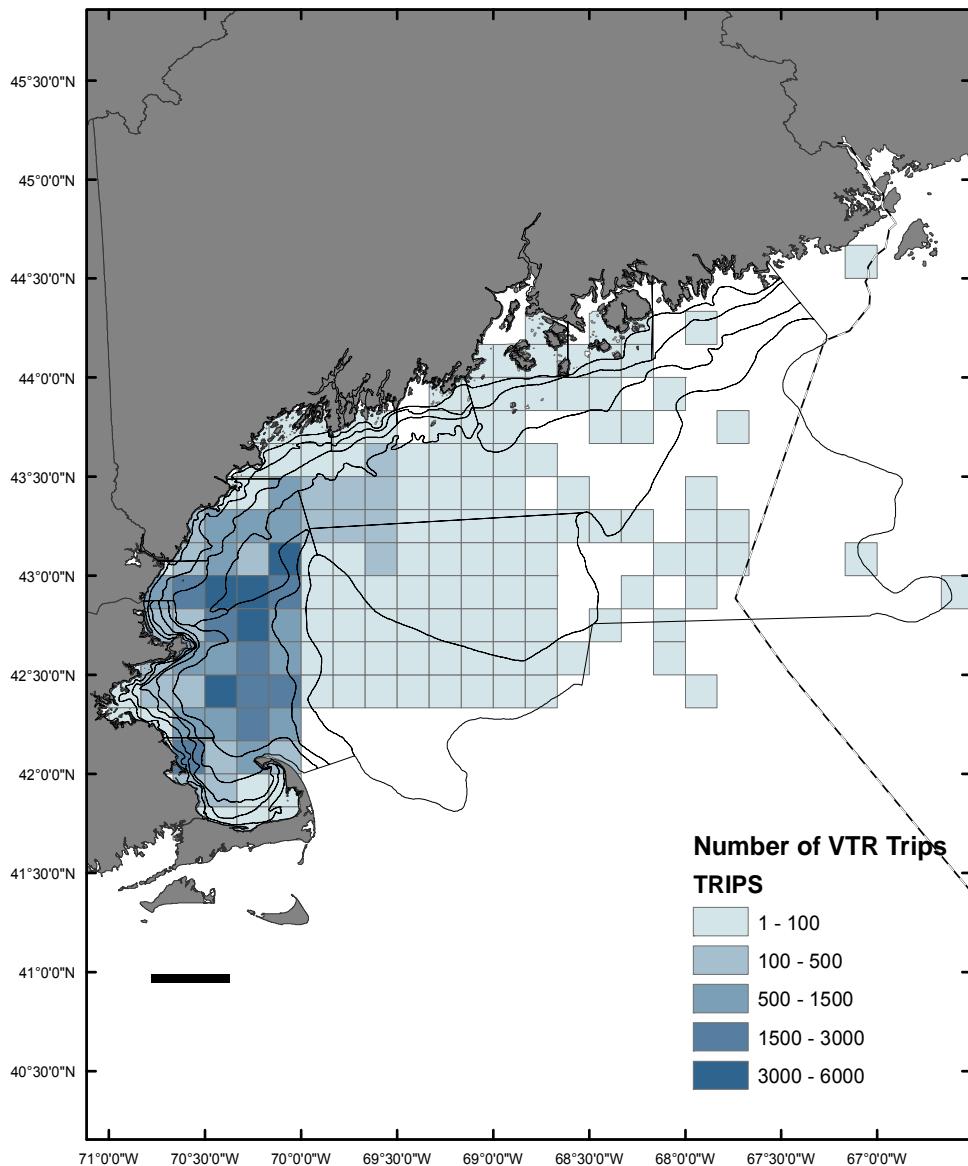


Figure A.72. Spatial distribution of recreational effort on trips reported catching Gulf of Maine Atlantic cod between 1994 and 2011 as determined from vessel trip reports (VTRs). VTR-based recreation effort has been binned to ten minute squares and overlaid on the Northeast Fisheries Science Center bottom trawl survey sampling strata

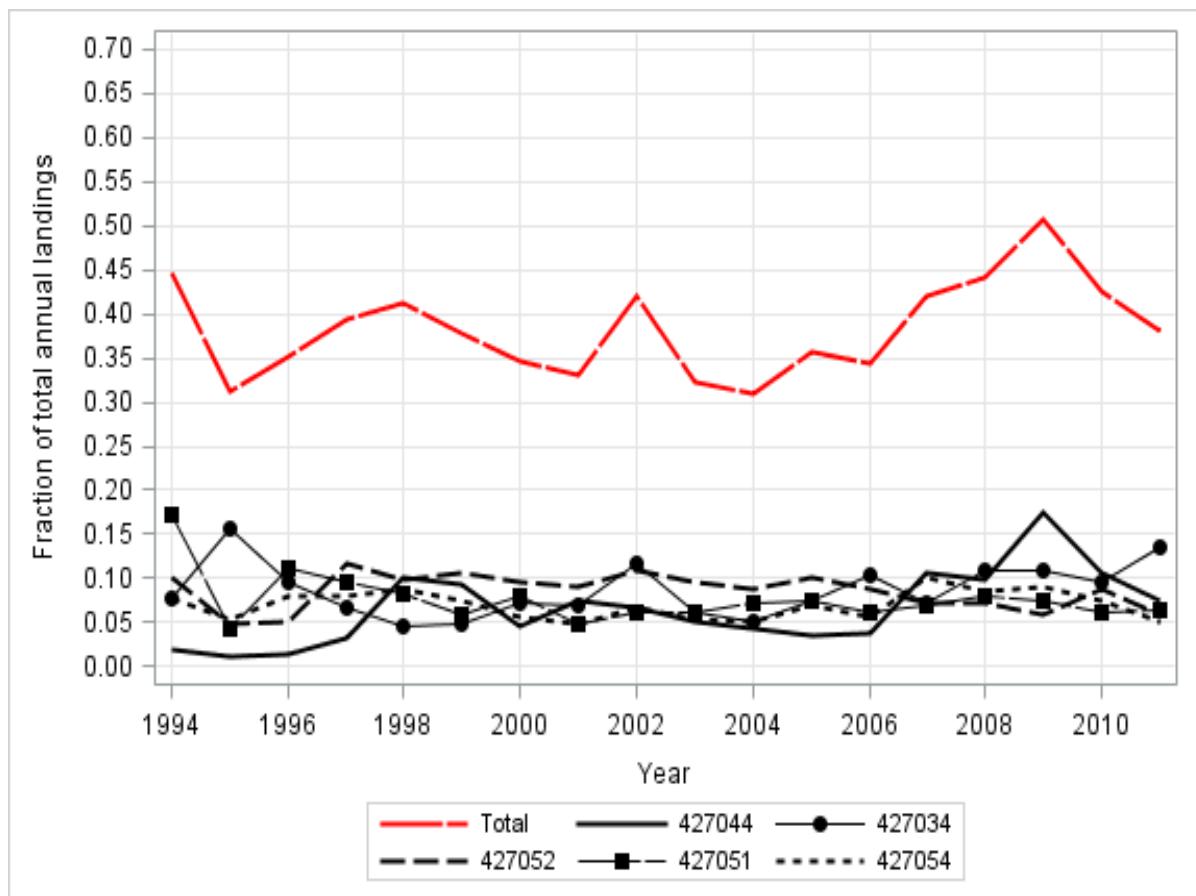


Figure A.73. Recreational utilization of the top ten minute squares between 1994 and 2011 expressed as an annual fraction of the total retained catch of Gulf of Maine Atlantic cod reported on vessel trip reports.

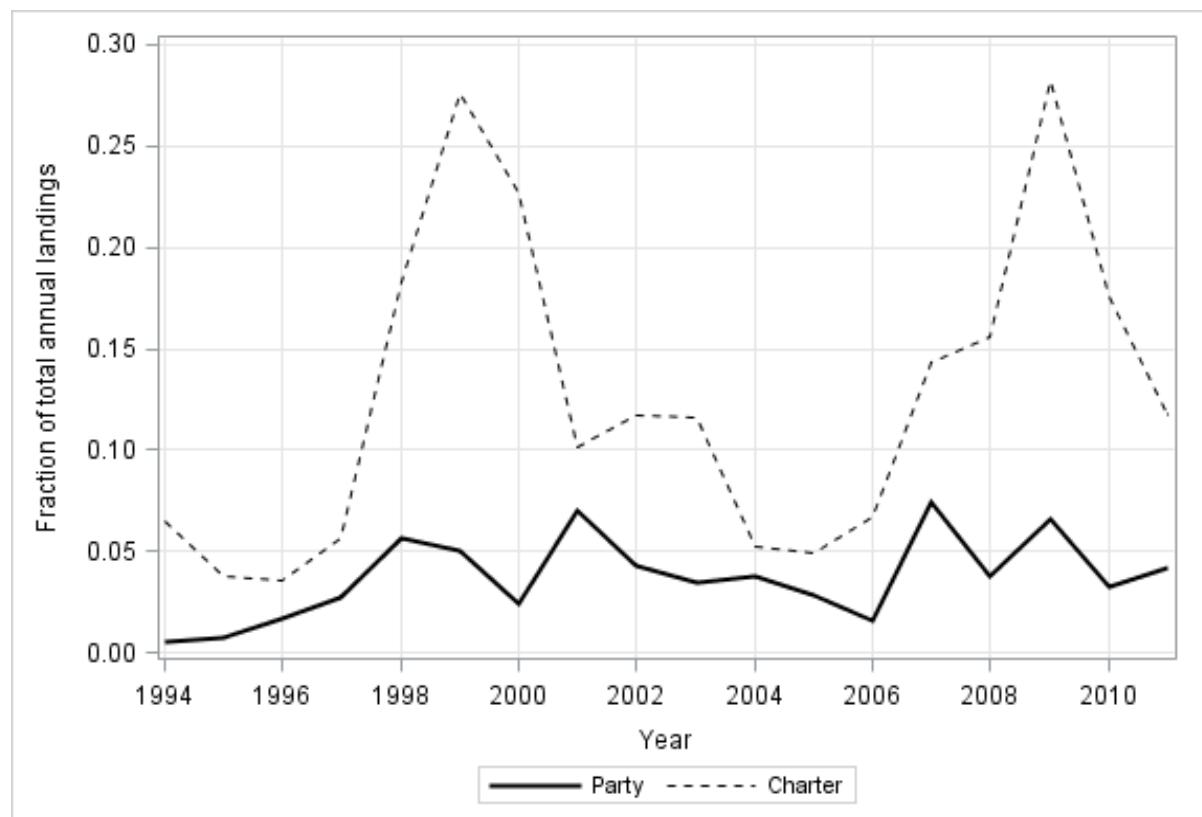


Figure A.74. Utilization of ten minute square 427044 by the recreational charter and party boat fishery between 1994 and 2011 expressed as annual fraction of the total retained catch of Gulf of Maine Atlantic cod reported on vessel trip reports.

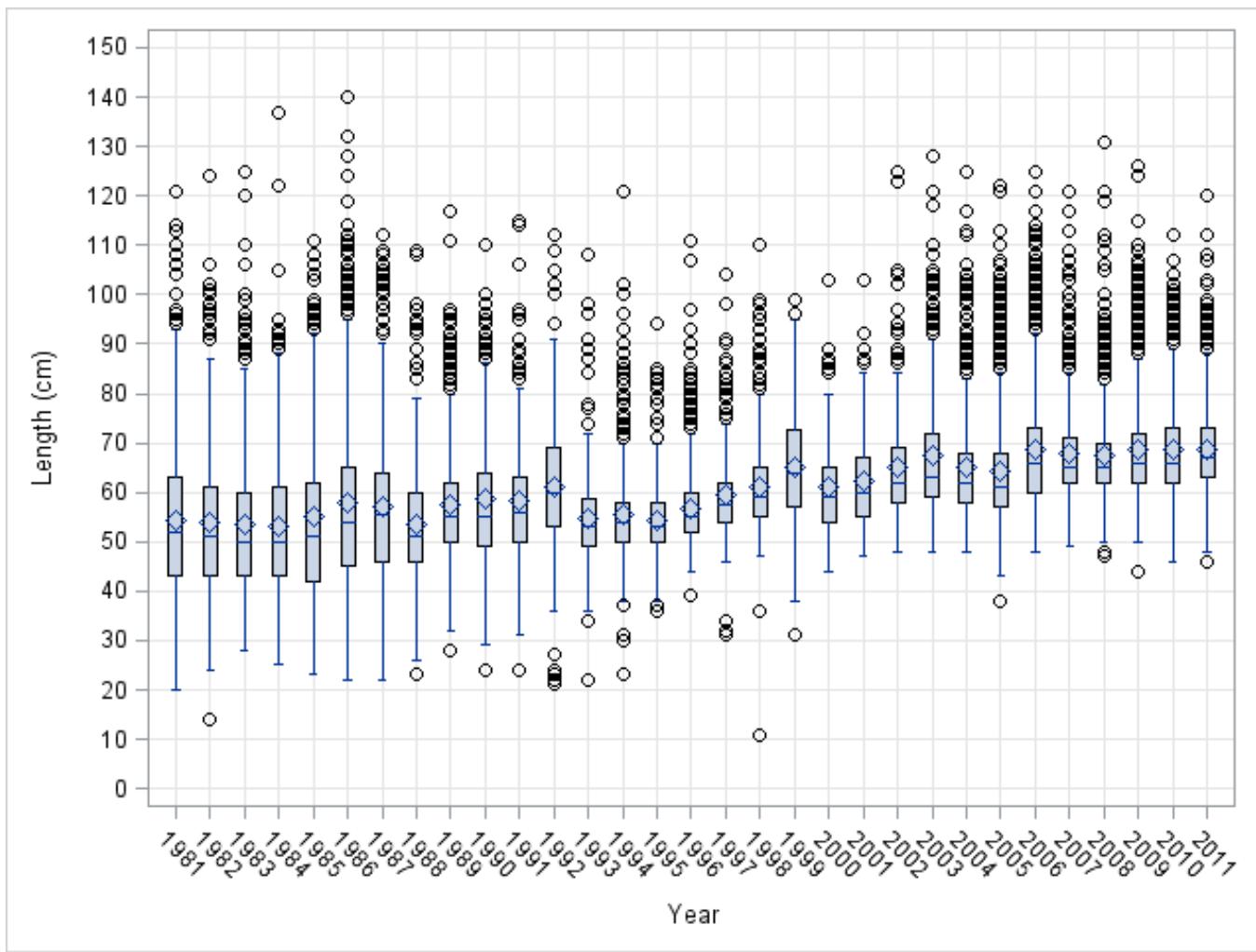


Figure A.75. Box plots showing the length distribution of Gulf of Maine Atlantic cod recreational harvest (AB1 catch) between 1981 and 2011.

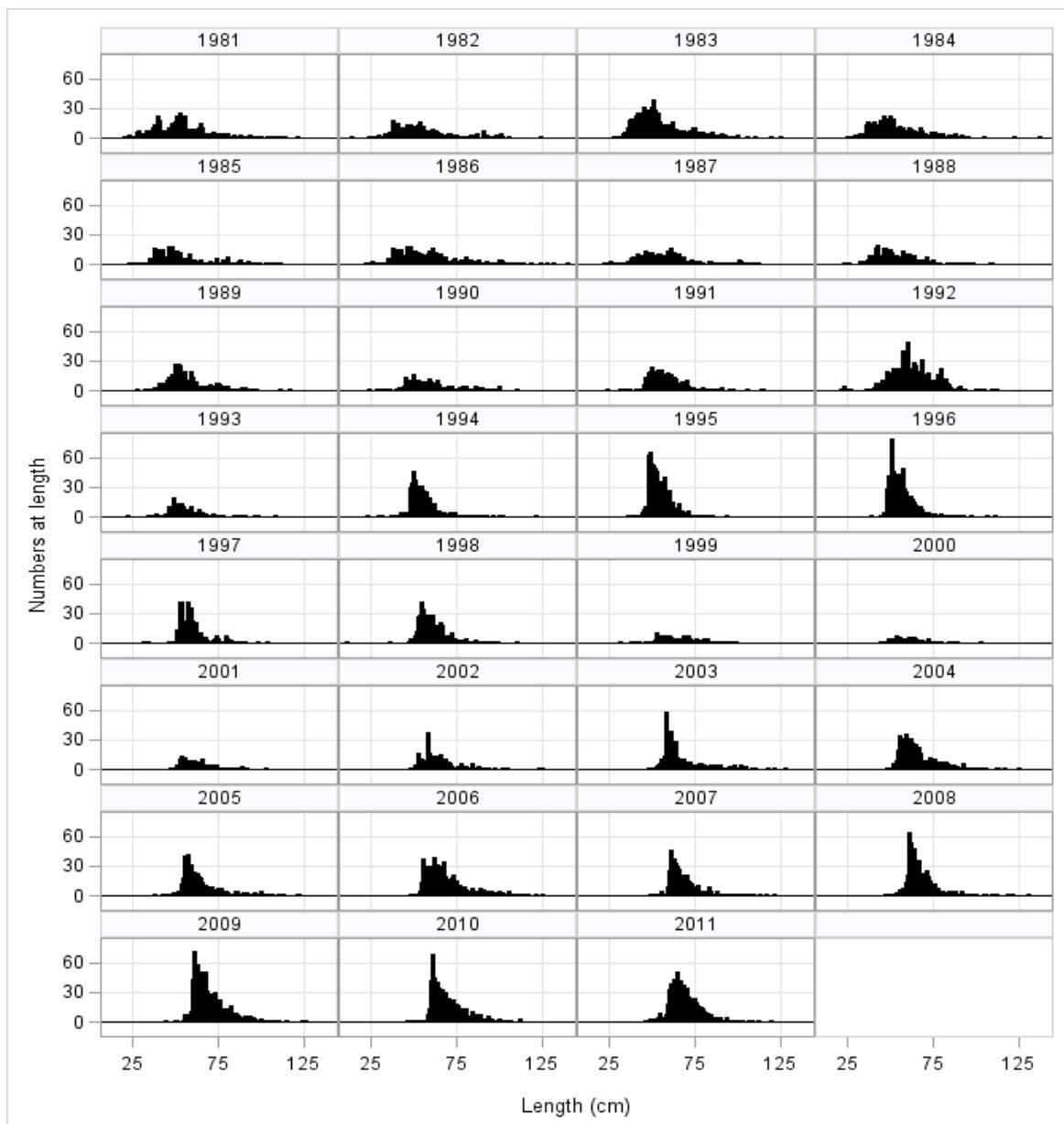


Figure A.76. Length frequency distribution of Gulf of Maine Atlantic cod recreational harvest (AB1 catch) between 1981 and 2011.

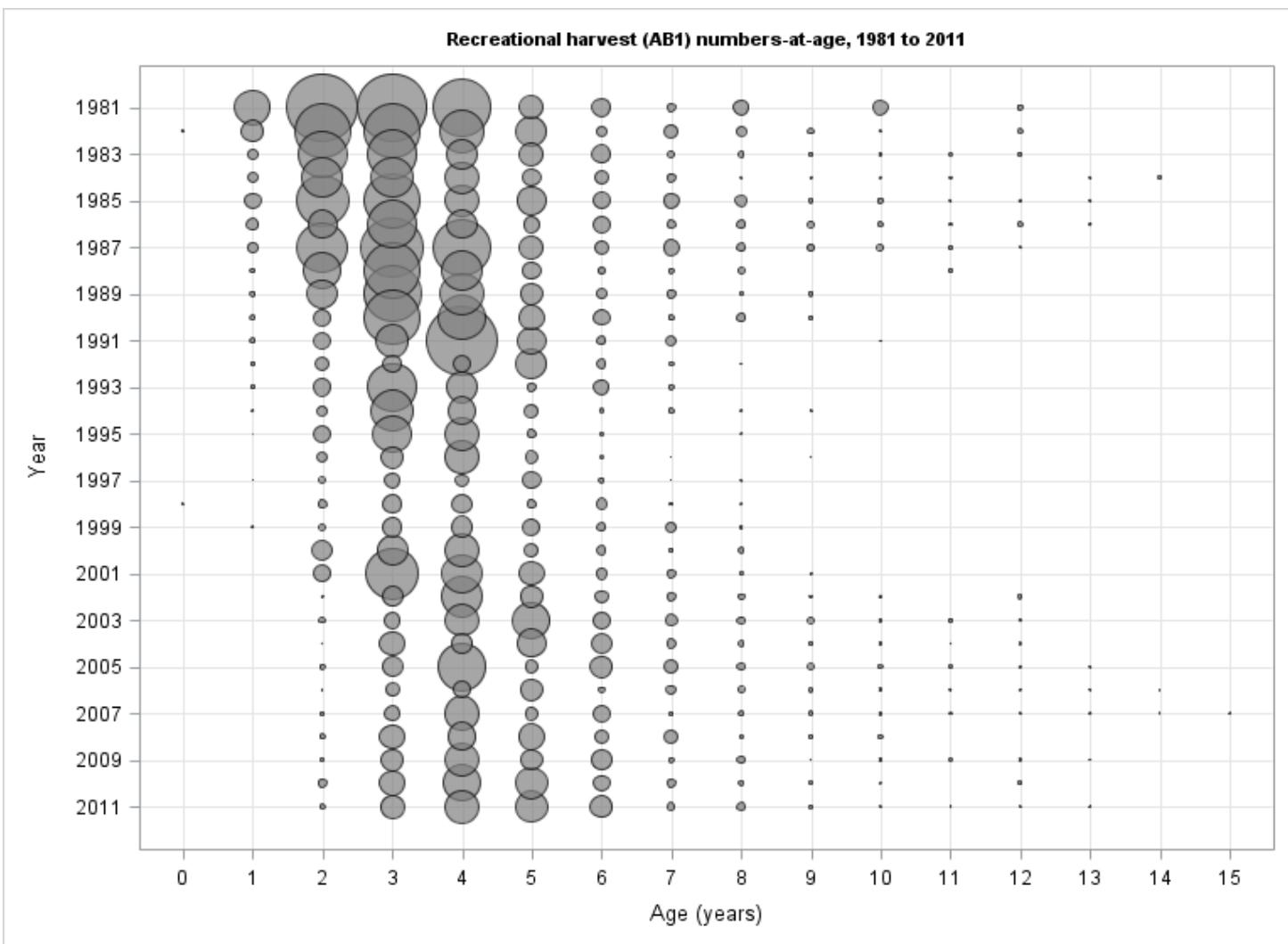


Figure A.77. Recreational landings-at-age of Gulf of Maine Atlantic cod from 1981 to 2011.

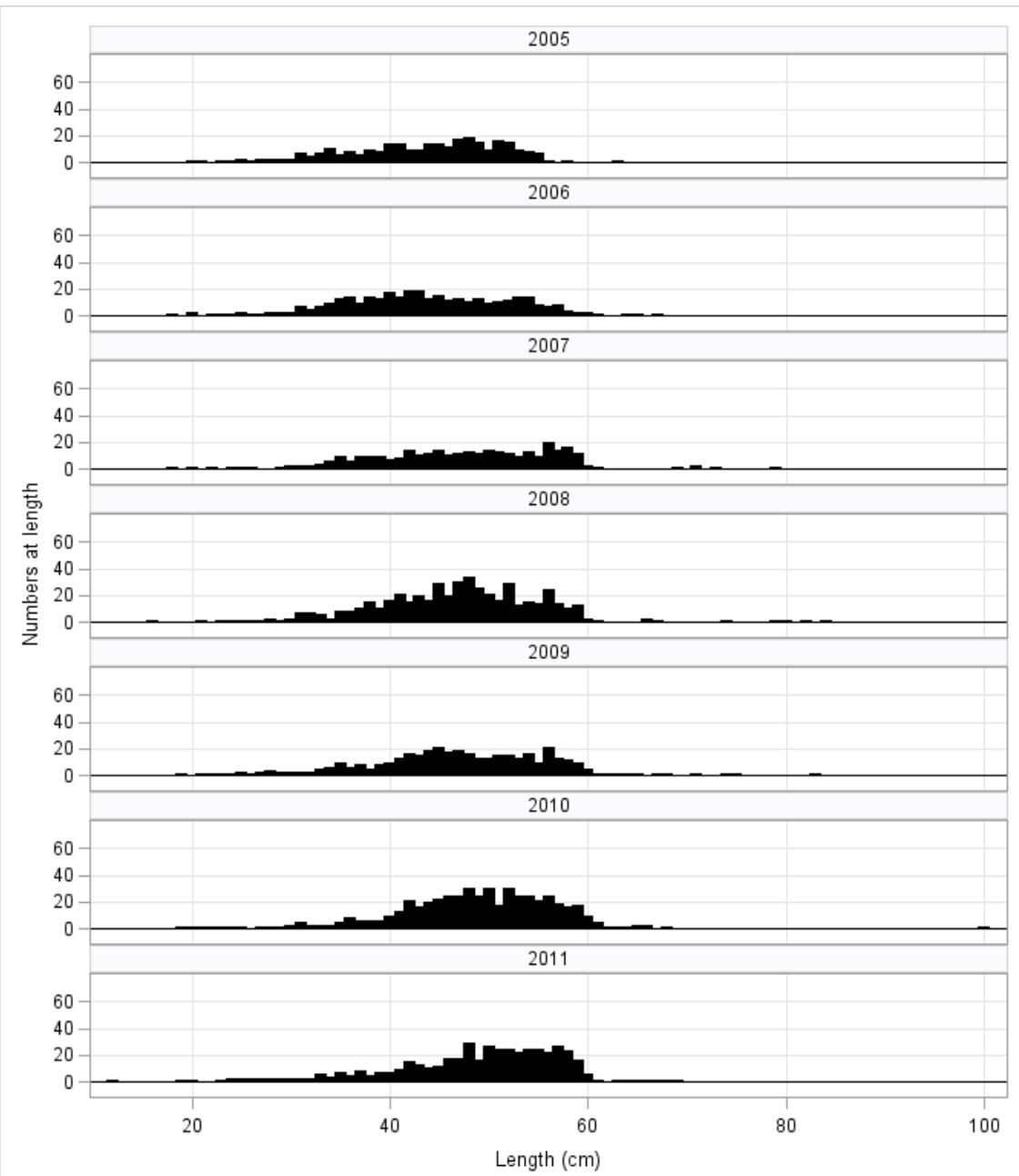


Figure A.78. Annual length frequency distributions of Gulf of Maine Atlantic cod discarded in the recreational fishery between 2005 and 2011. *No sampling of recreational discards occurred prior to 2005.*

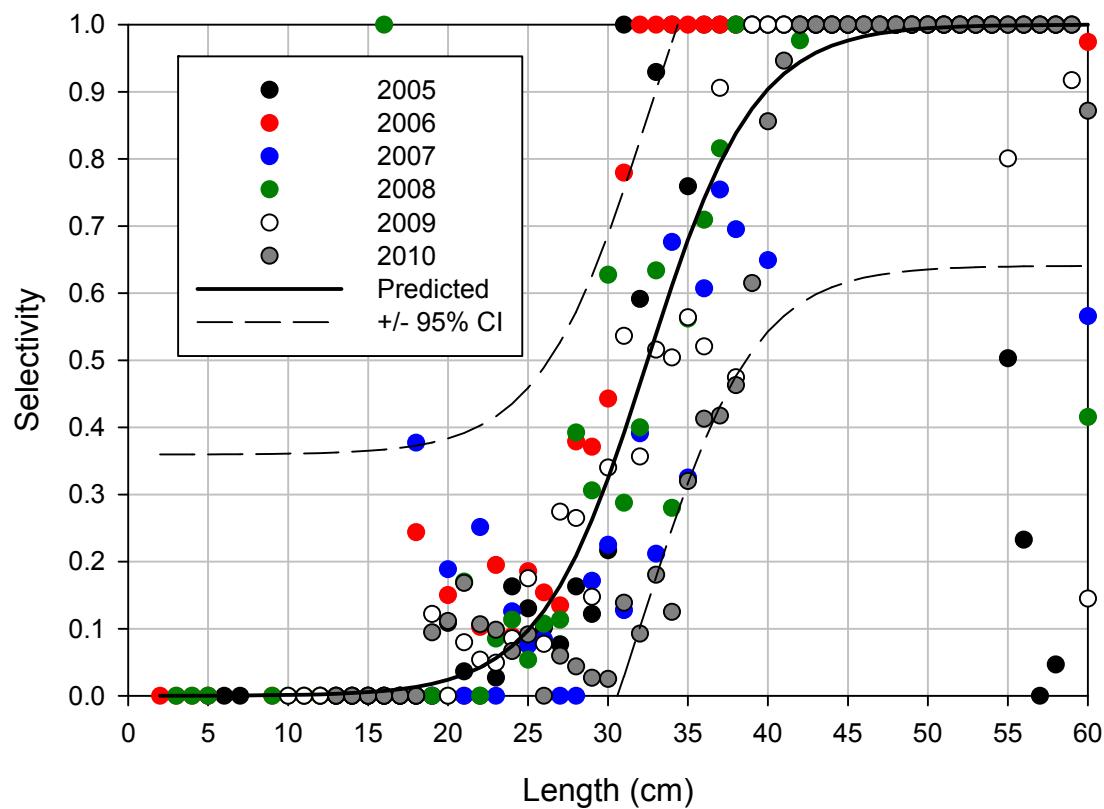
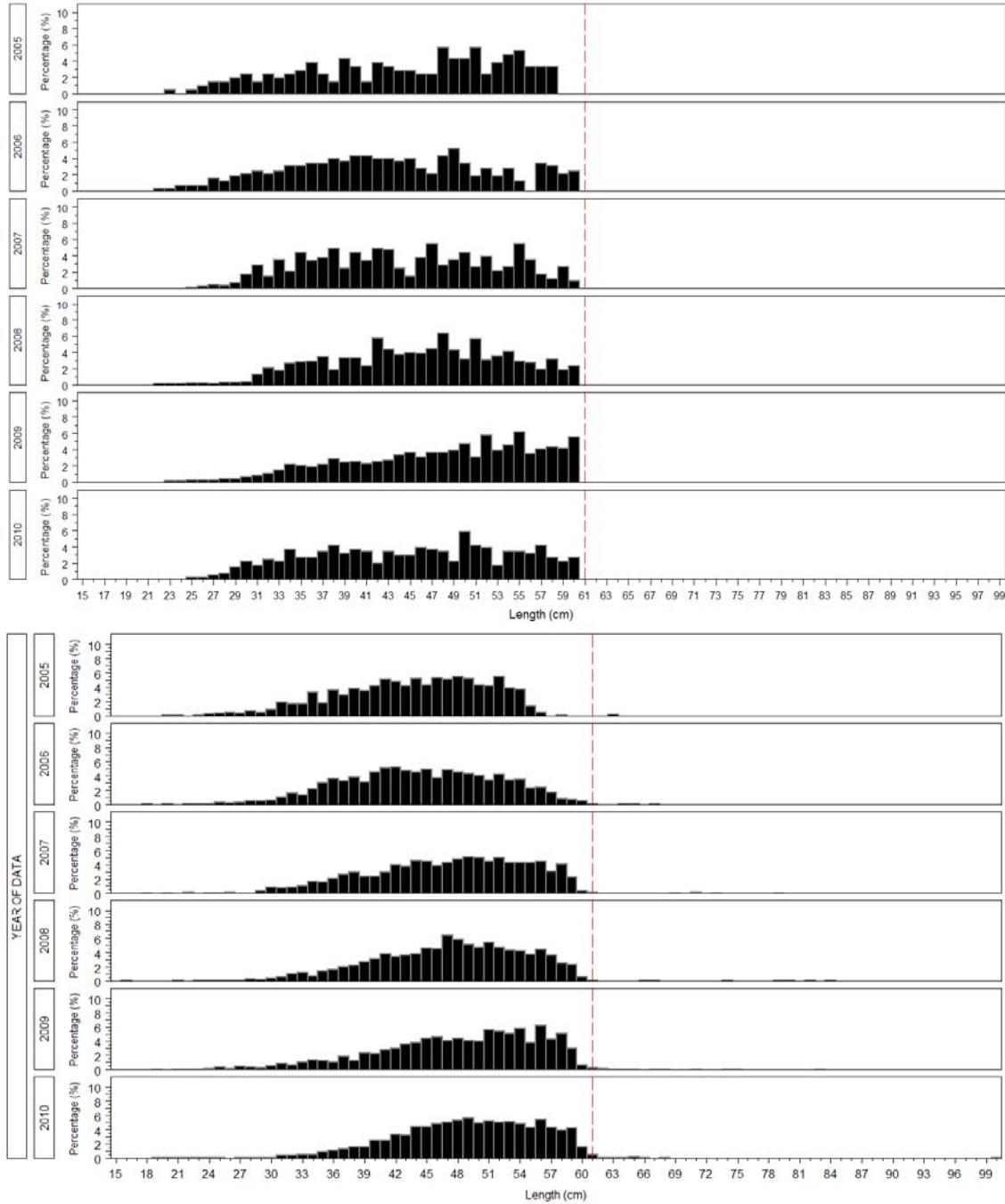


Figure A.79. Estimated selectivity ogive for the recreational fishery and the corresponding 95% confidence interval (CI) for Gulf of Maine Atlantic cod. The selectivity ogive was estimated from the logistic fits to the aggregated annual estimates of selectivity-at-length.



Figures A.80. Comparison of recreational discard length frequency distributions estimated using the survey filter approach (top) to those generated from the B2 sampling of the I9 catch (bottom) between 2005 and 2010 for Gulf of Maine Atlantic cod. The dashed red line represents the recreational minimum retention size of 24 inches (61.0 cm) from May 1, 2006-2010. The minimum retention size from January 1, 2005 to May 1, 2006 was 23 inches (58.4cm).

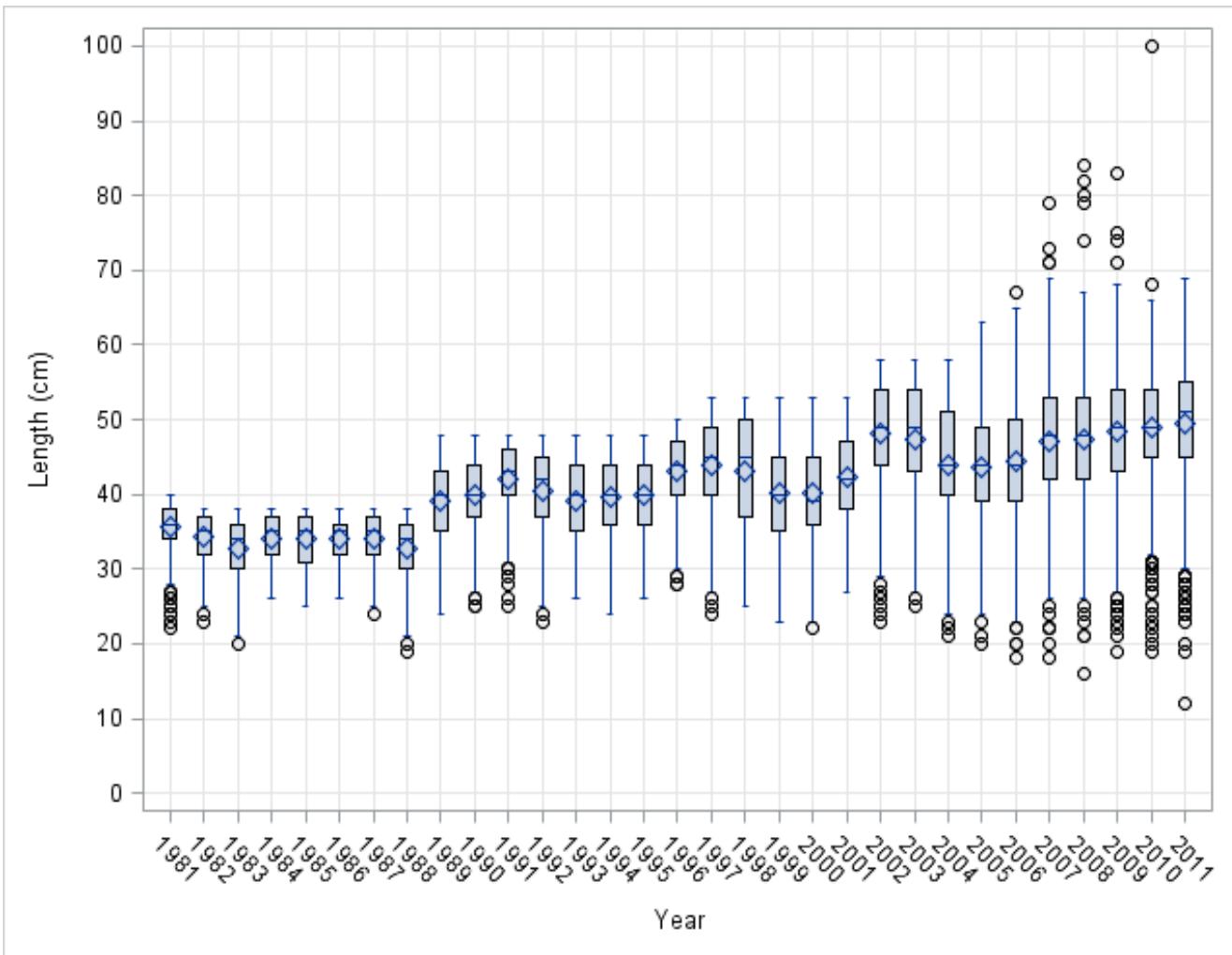


Figure A.81. Box plots showing the length distribution of Gulf of Maine Atlantic cod recreational releases (B2 catch) between 1981 and 2011.

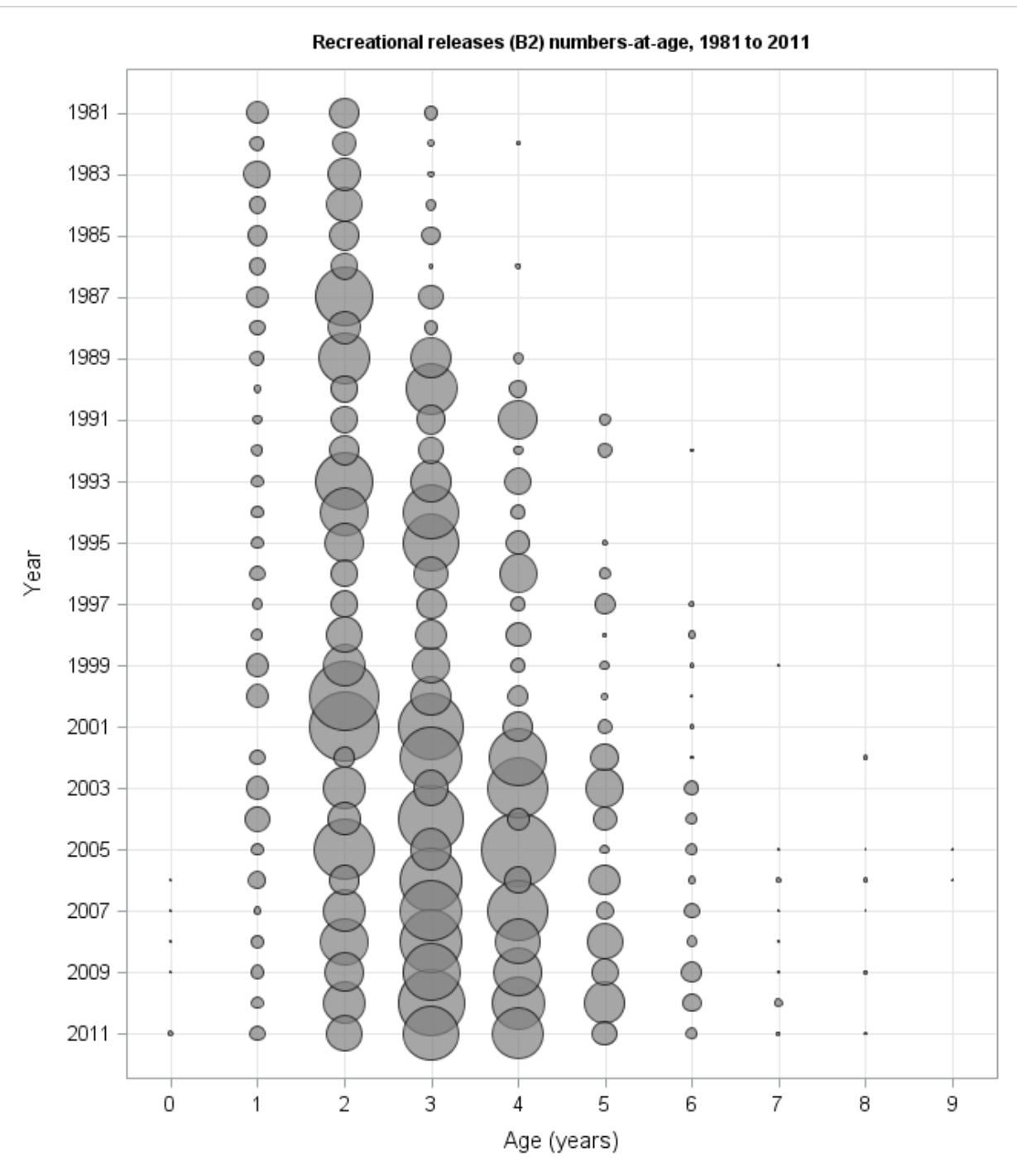


Figure A.82. Recreational discards-at-age of Gulf of Maine Atlantic cod from 1981 to 2011.

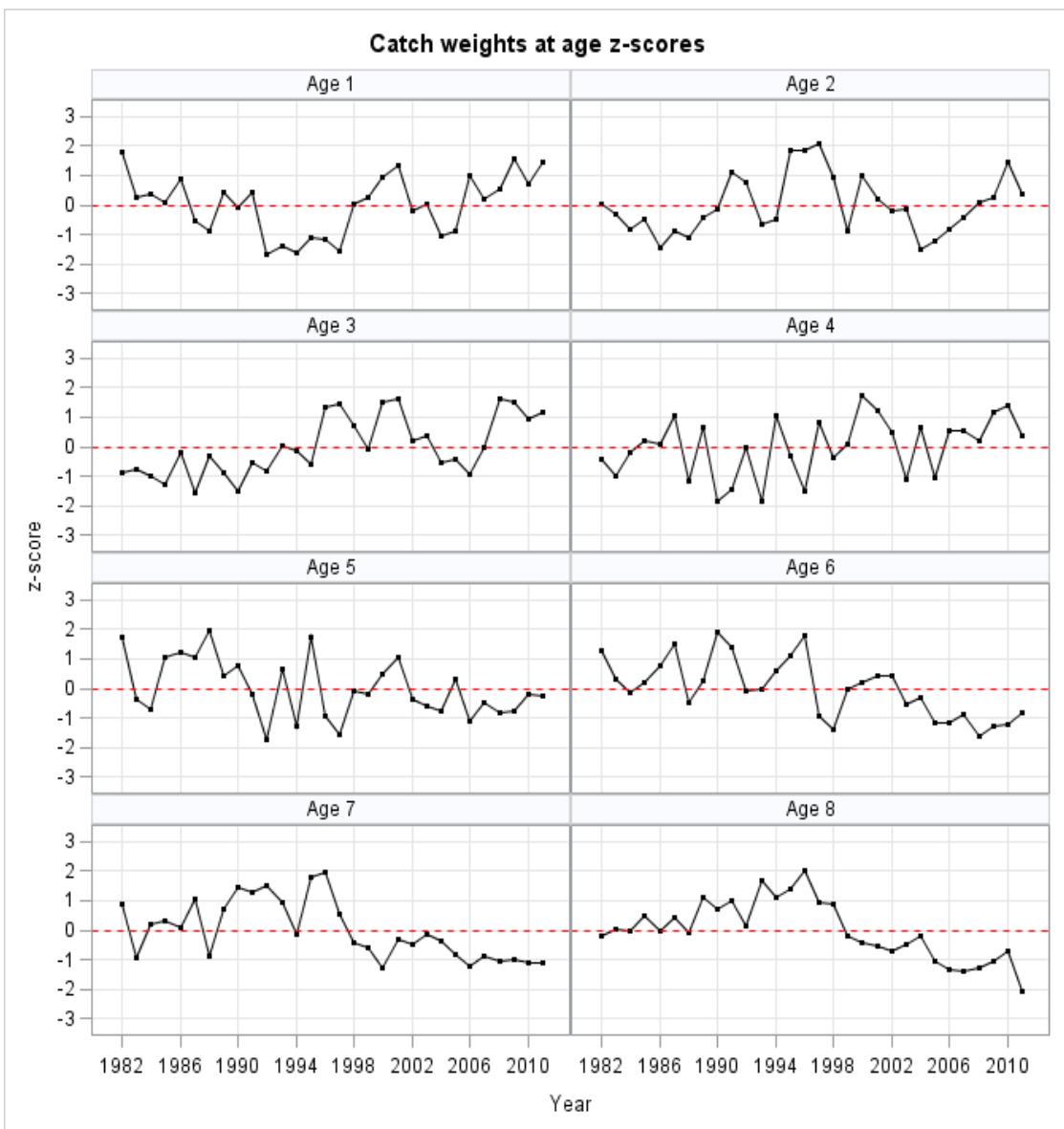


Figure A.83. Average catch weights-at-age of age1-8 Gulf of Maine Atlantic cod from 1982 to 2011. Weights-at-age were estimated using a number weighted average of commercial landing, commercial discard, recreational landings, and recreational discards weights-at-age. Average weights are presented as z-scores ($[x-\mu]/\sigma$).

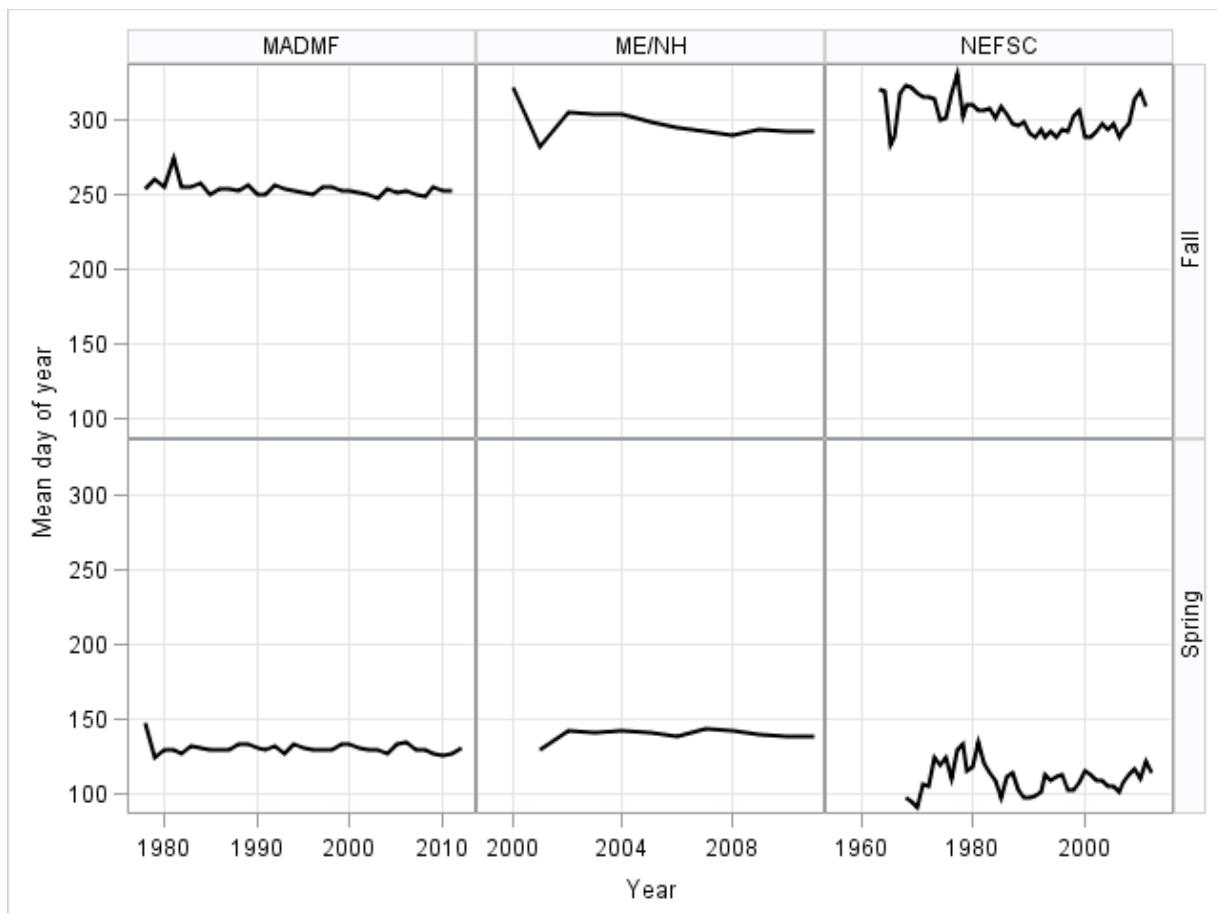


Figure A.84. Mean day of the year of sampling in the Gulf of Maine by each of the three ongoing regional bottom trawl surveys: Northeast Fisheries Science Center (NEFSC), Massachusetts Department of Marine Fisheries (MADMF) and the Maine – New Hampshire inshore bottom trawl survey (ME/NH). Days are expressed as Julian days (e.g., January 1 is day 1 and December 31 is day 365/66).

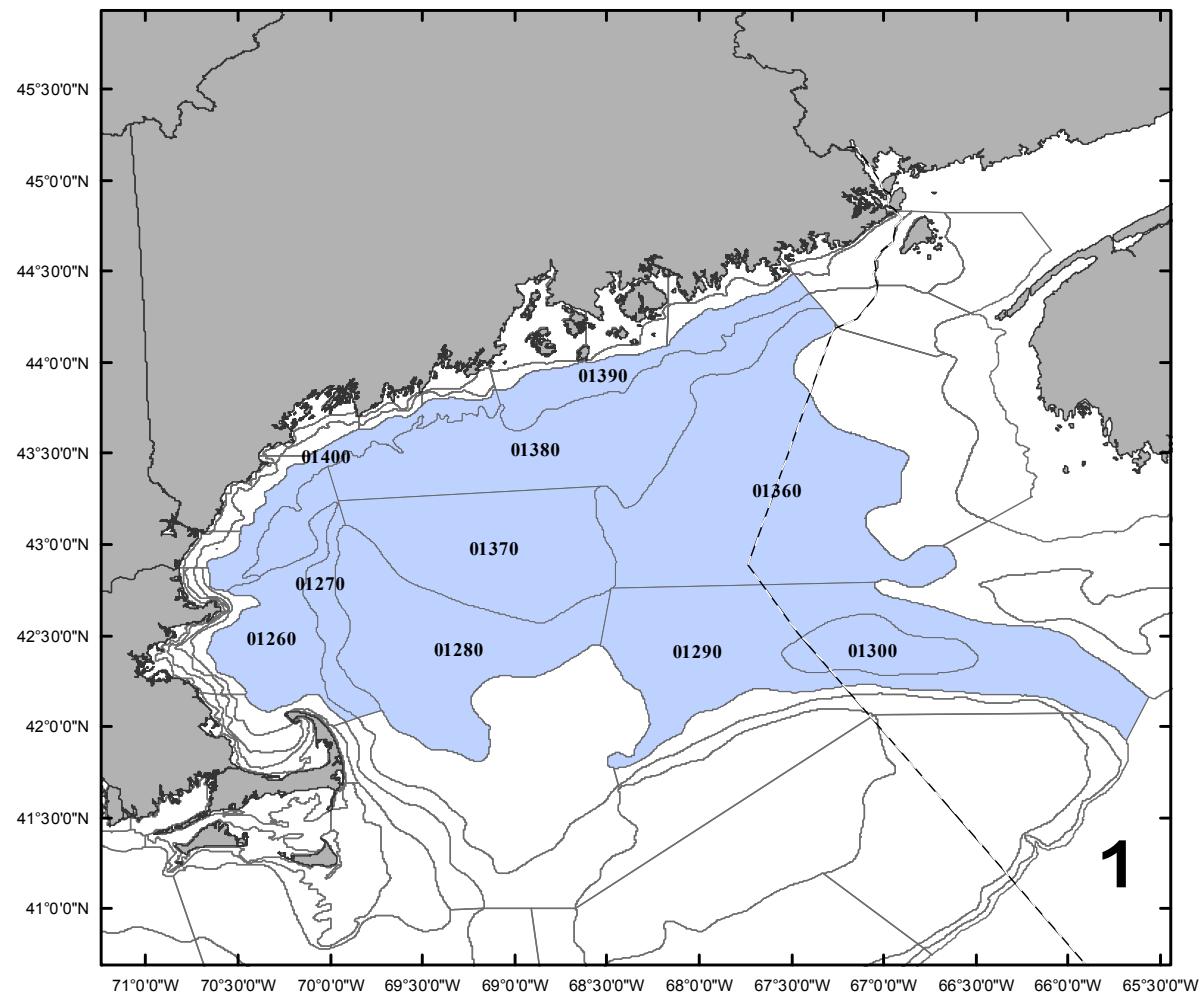


Figure A.85. Map of the Northeast Fisheries Science Center (NEFSC) bottom trawl offshore survey strata included in the Gulf of Maine Atlantic cod stock assessment (shaded blue).

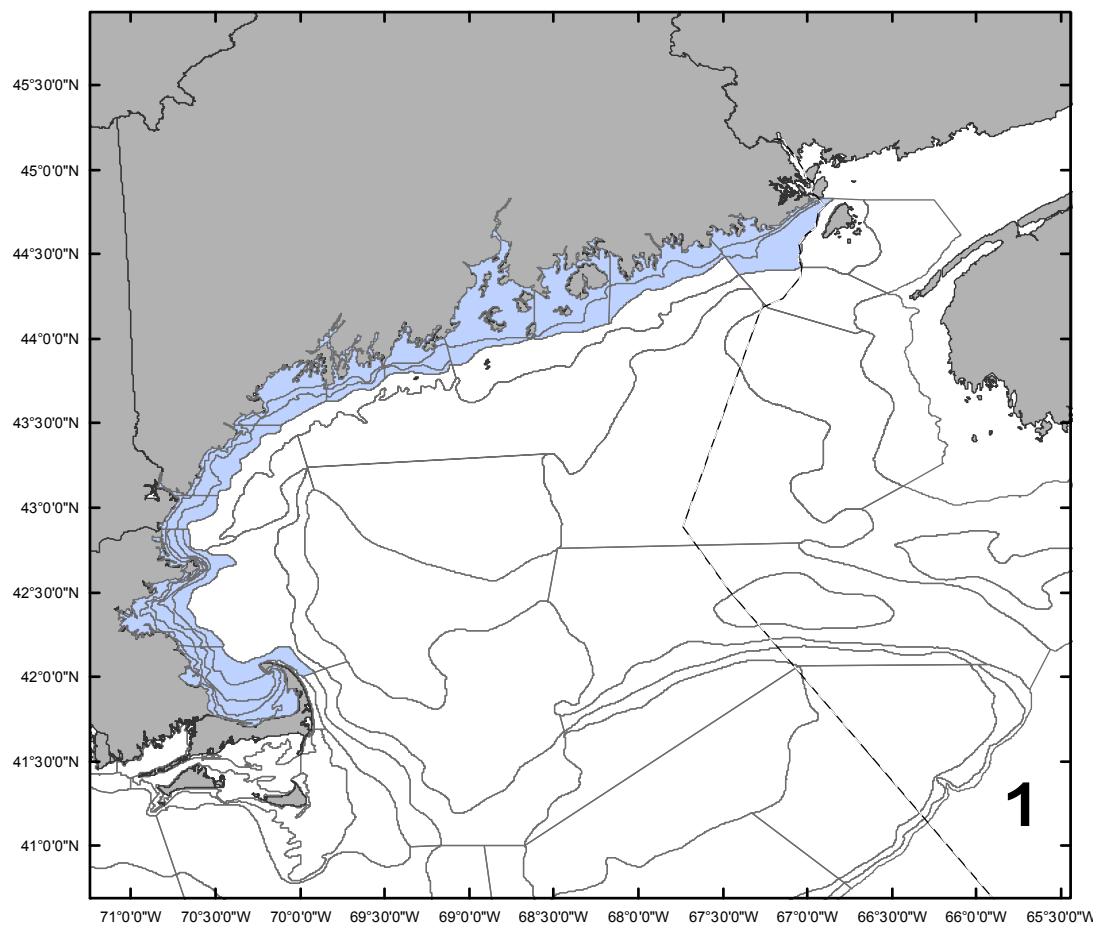


Figure A.86. Map of the Northeast Fisheries Science Center (NEFSC) bottom trawl inshore survey strata in the Gulf of Maine region (shaded blue).

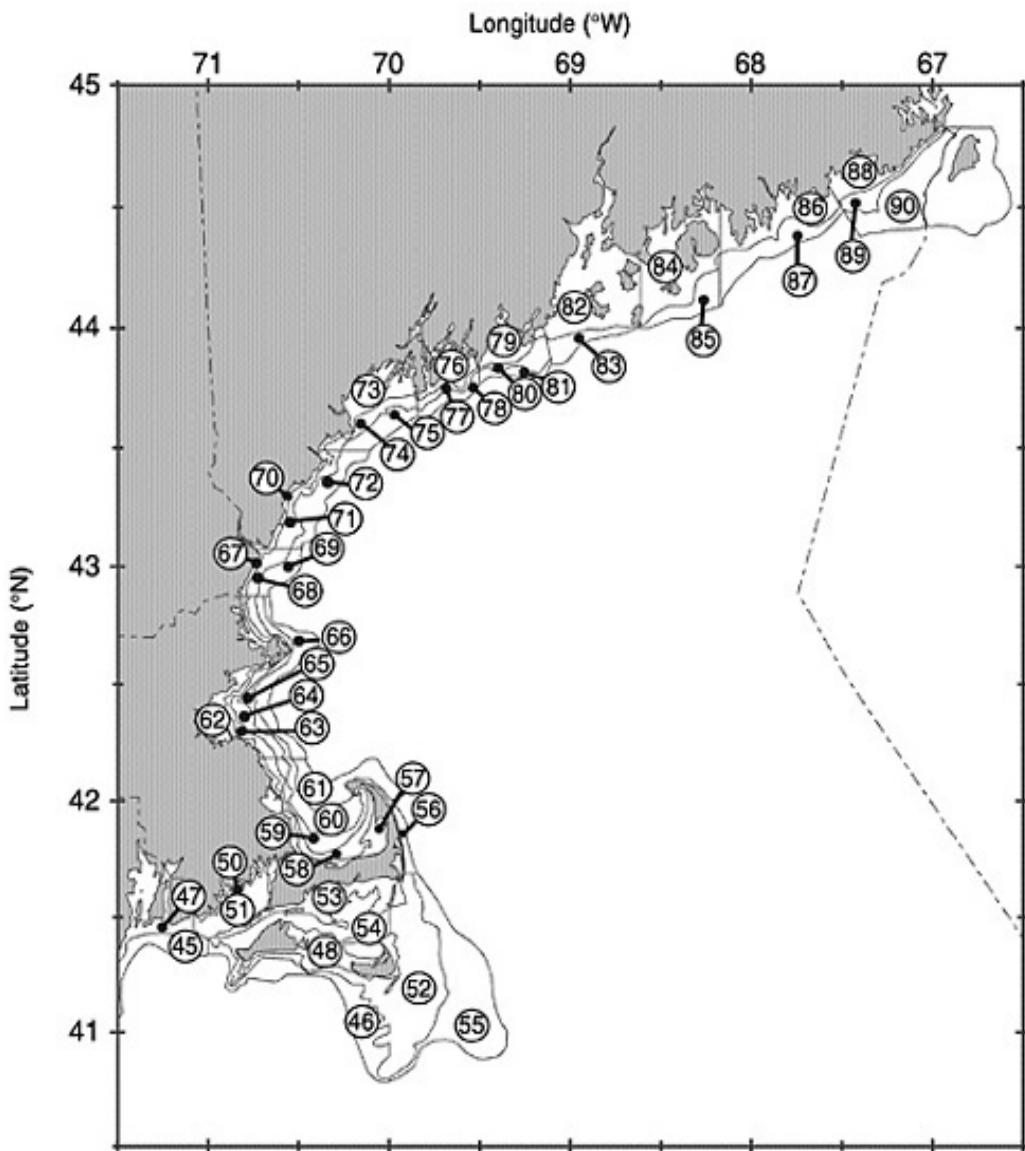


Figure A.87. Map identifying the inshore survey strata of the Northeast Fisheries Science Center (NEFSC) bottom trawl survey. *Note the survey strata are identified using their 2-digit labels. Strata identifiers are five-digit identifiers beginning with a two-digit prefix and one-digit suffix (e.g., the full identifier for inshore strata 66 is 03660).

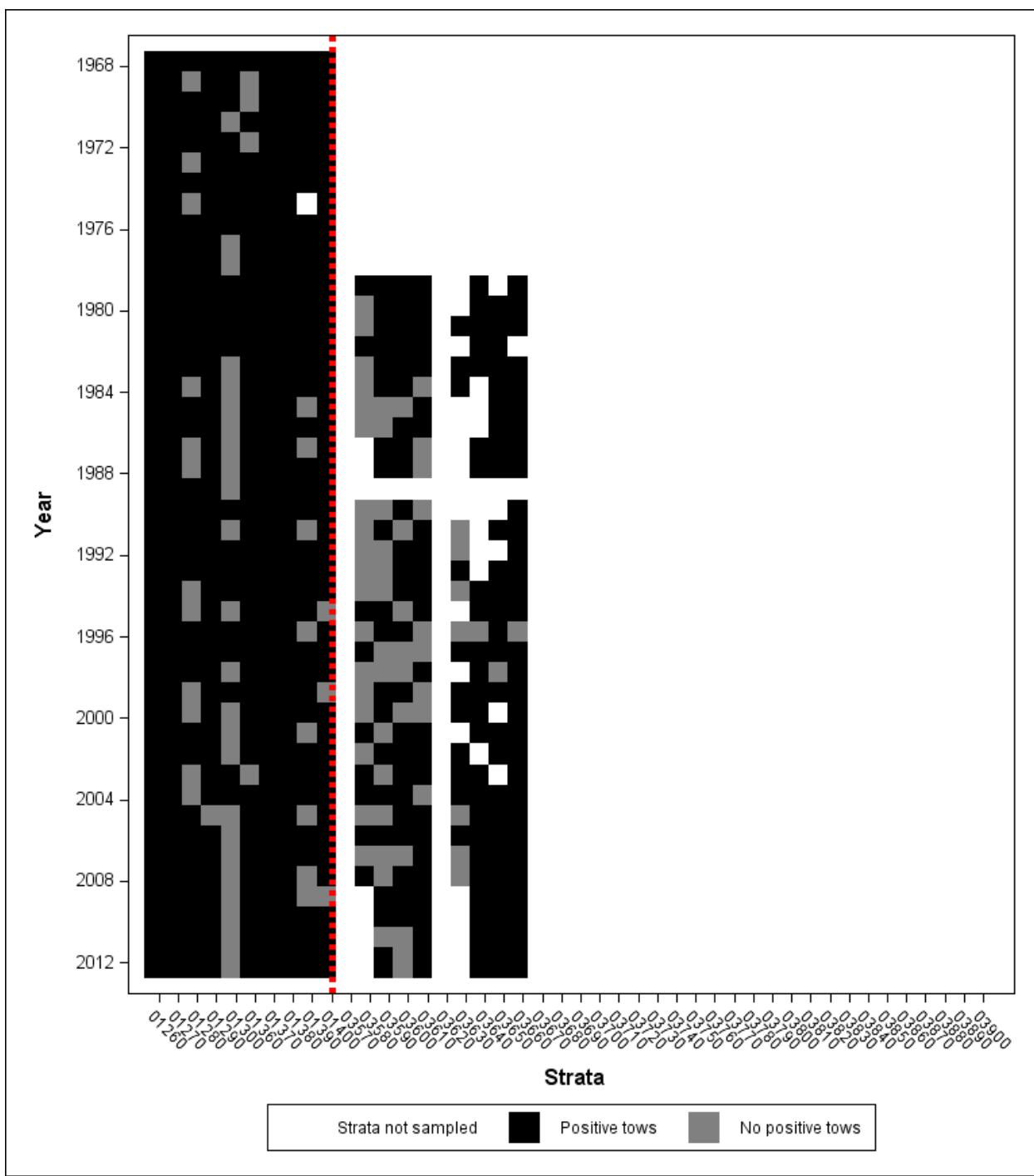


Figure A.88. Sampling summary of offshore (01 prefix) and inshore (03 prefix) strata in the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey from 1968-2012. Positive and negative tows are indicated with respect to catches of Gulf of Maine Atlantic cod.

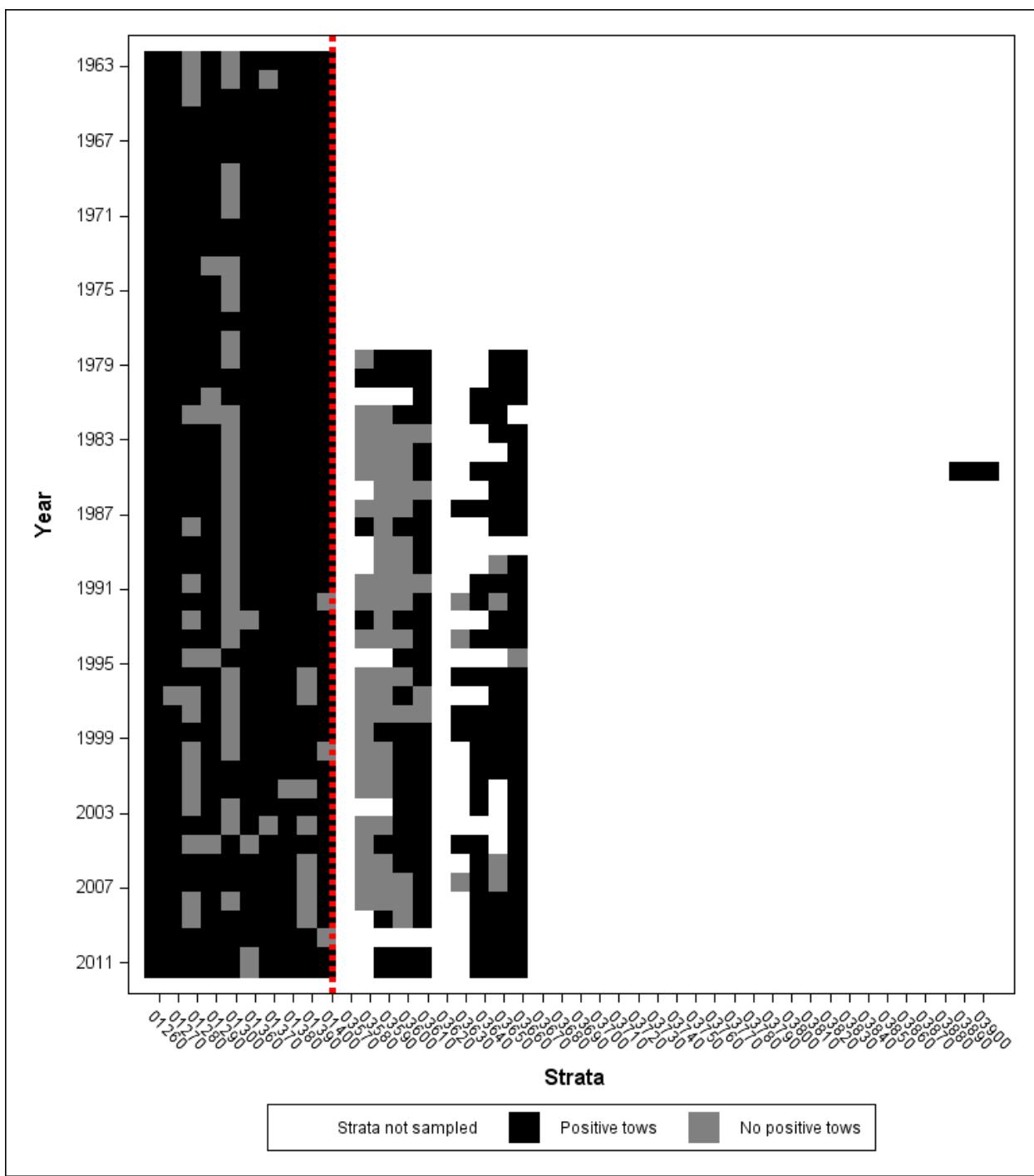


Figure A.89. Sampling summary of offshore (01 prefix) and inshore (03 prefix) strata in the Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey from 1963-2011. Positive and negative tows are indicated with respect to catches of Gulf of Maine cod.

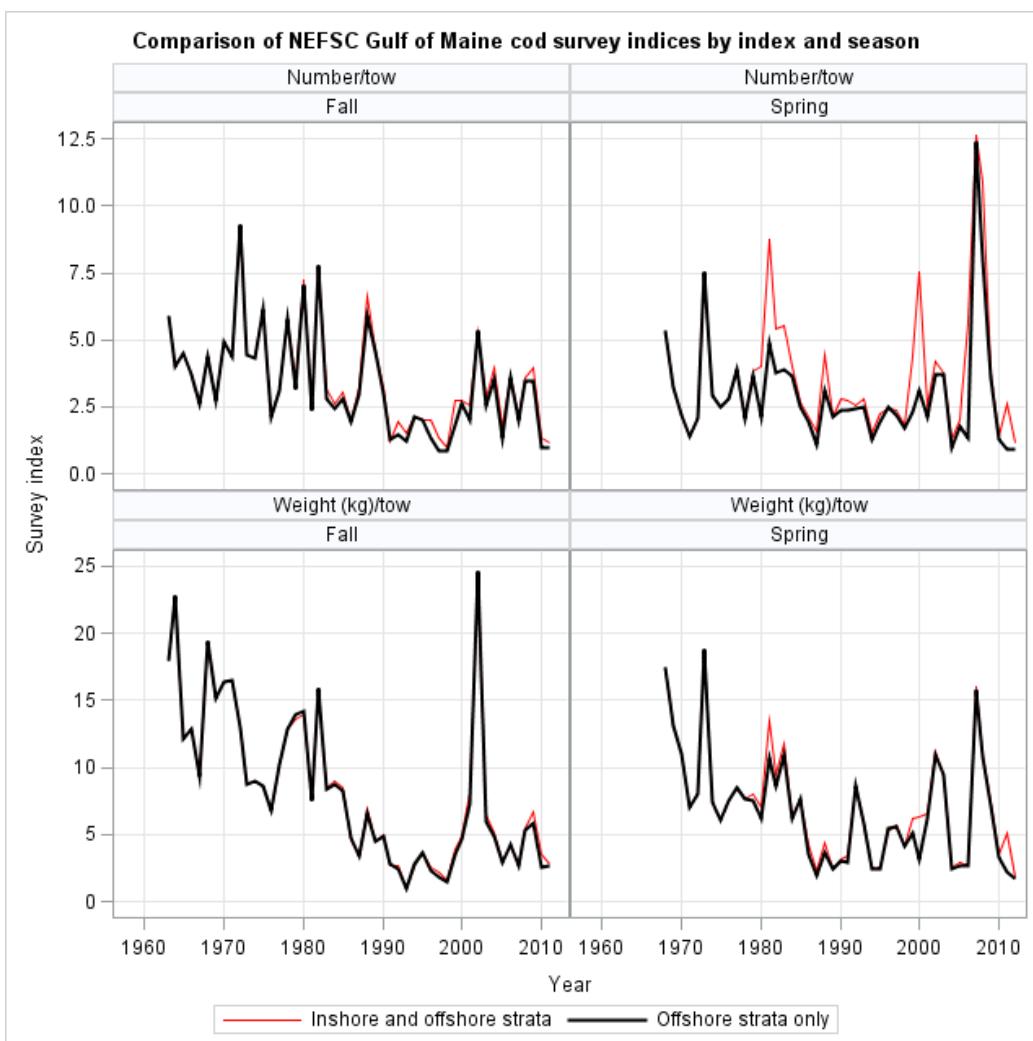


Figure A.90. Comparison of Northeast Fisheries Science Center (NEFSC) bottom trawl survey indices for Gulf of Maine Atlantic cod calculated using offshore strata (black) and both inshore and offshore survey strata (red).

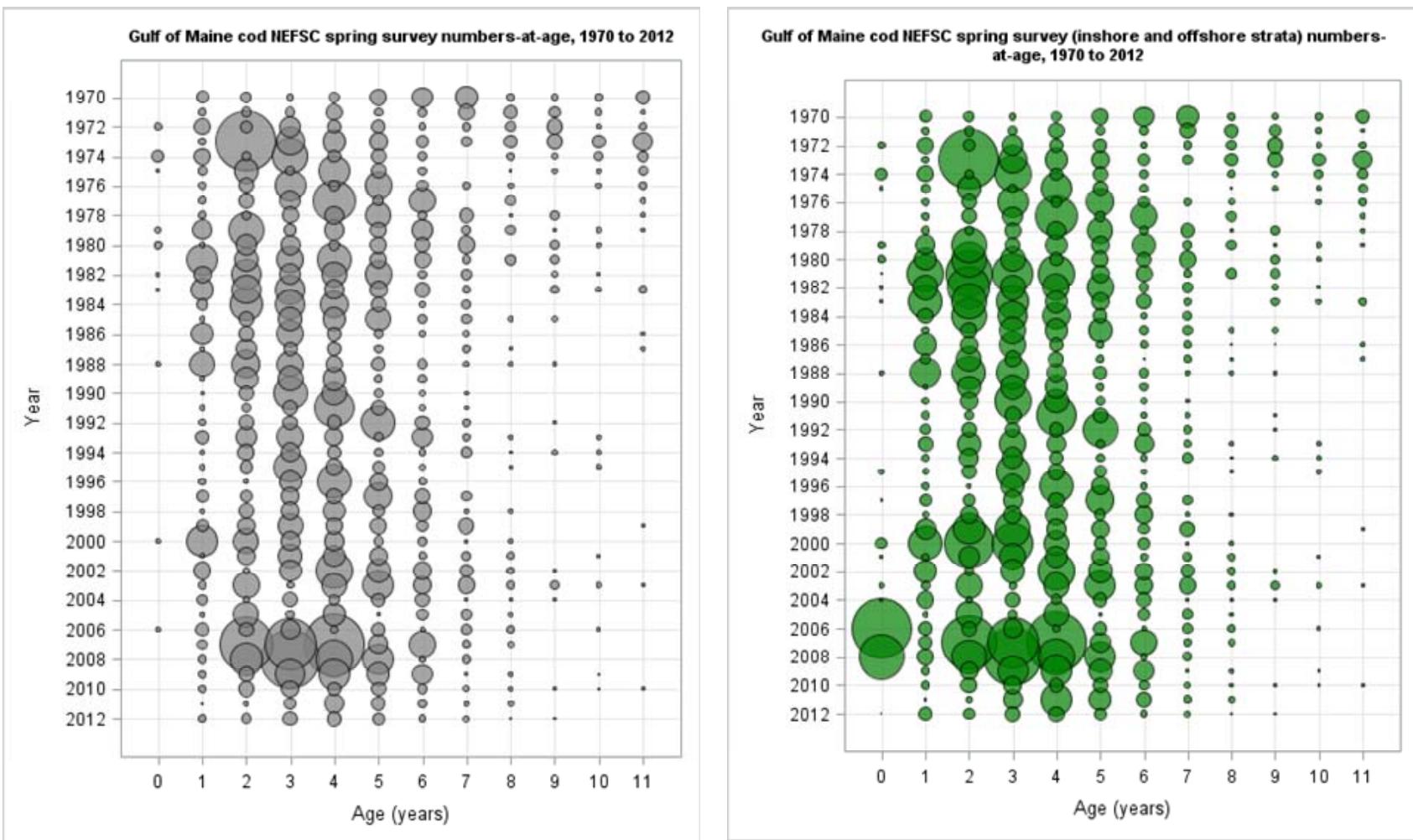


Figure A.91. Comparison of Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey numbers at age indices for Gulf of Maine Atlantic cod calculated using offshore strata (grey) and both inshore and offshore survey strata (green).

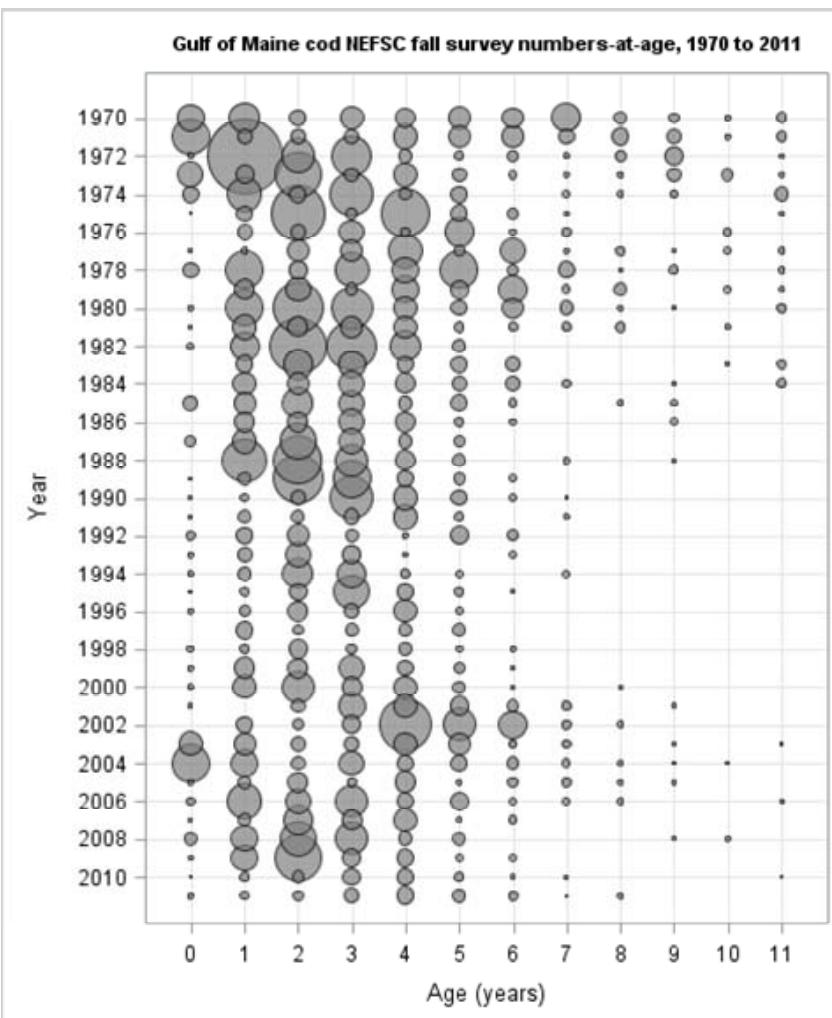


Figure A.92. Comparison of Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey numbers at age indices for Gulf of Maine Atlantic cod calculated using offshore strata (grey) and both inshore and offshore survey strata (green).

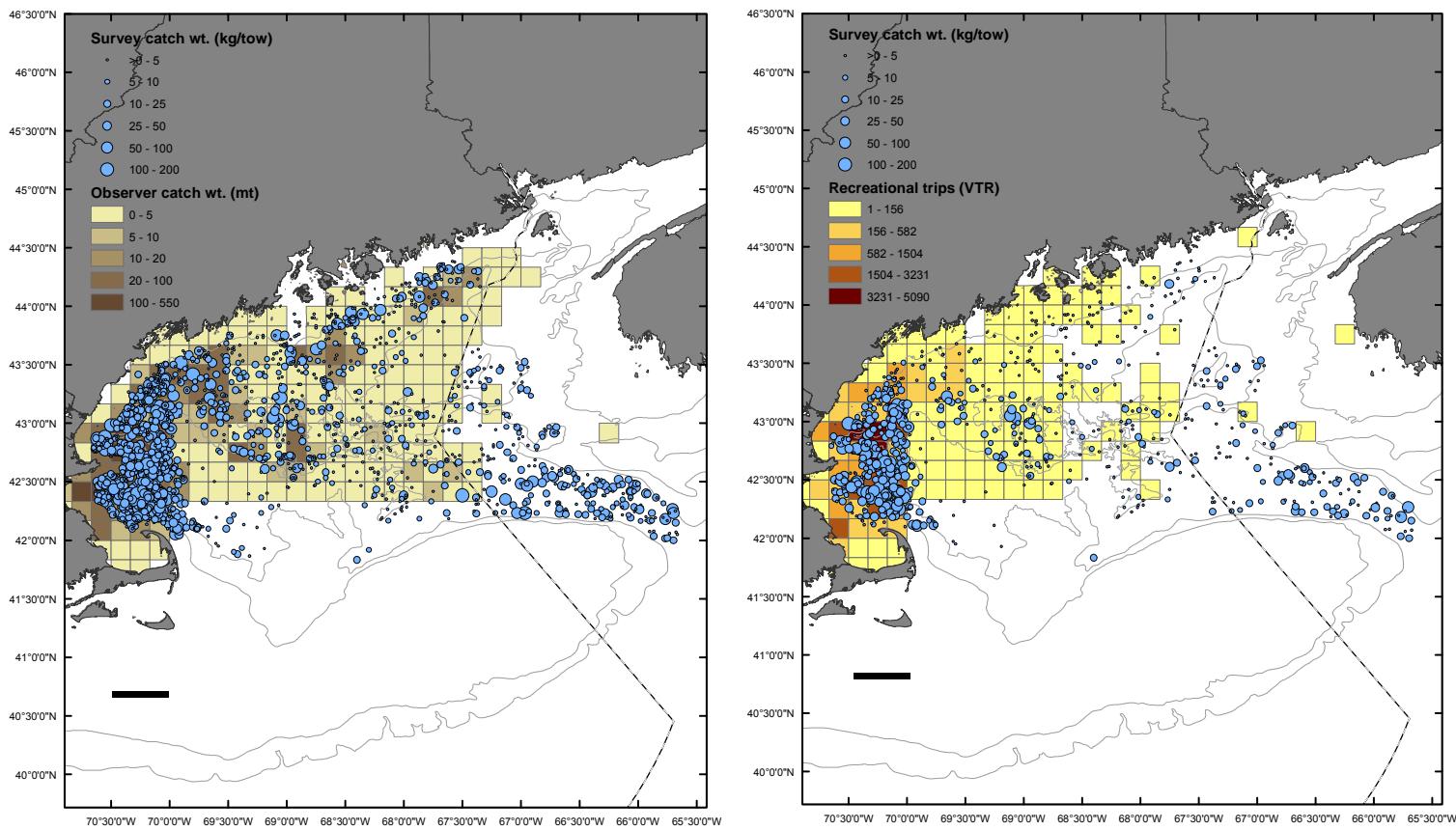


Figure A.93. Spatial overlap of survey catches (kg/tow) of Gulf of Maine Atlantic cod from the Northeast Fisheries Science Center (NEFSC) bottom trawl survey (spring and fall combined) and commercial and recreational fishing effort. On the left, NEFSC survey catches from 1989 – 2010 are overlaid on total observed cod catch (landings and discards) binned to ten minute squares from the same time period. On the right, NEFSC survey catches from 1994 – 2010 are overlaid on the number of VTR-reported recreational trips that caught cod binned to ten minute squares. *Note the different time periods used in each plot.

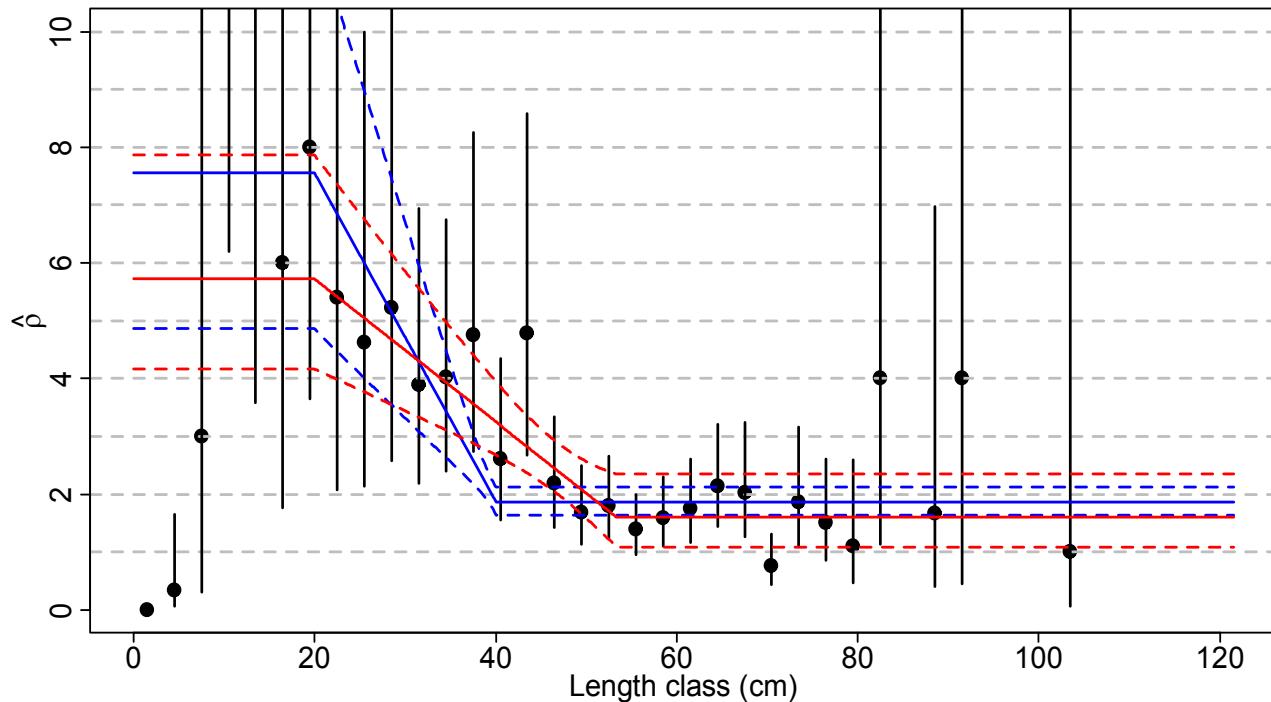


Figure A.94. Beta-binomial-based estimates of calibration factors and corresponding 95% confidence intervals by length class (3 cm bins) for Atlantic cod. The black points and vertical bars represent results where different calibration factors are estimated for each length class. The blue lines represent results from a segmented regression model where the two points connecting the segments are known (20 and 40 cm) and the red lines represent results from a segmented regression model where the first point (20 cm) is known but the second is estimated. Segmented regression fits are based on data from fish ≥ 20 cm (from Brooks et al. 2010).

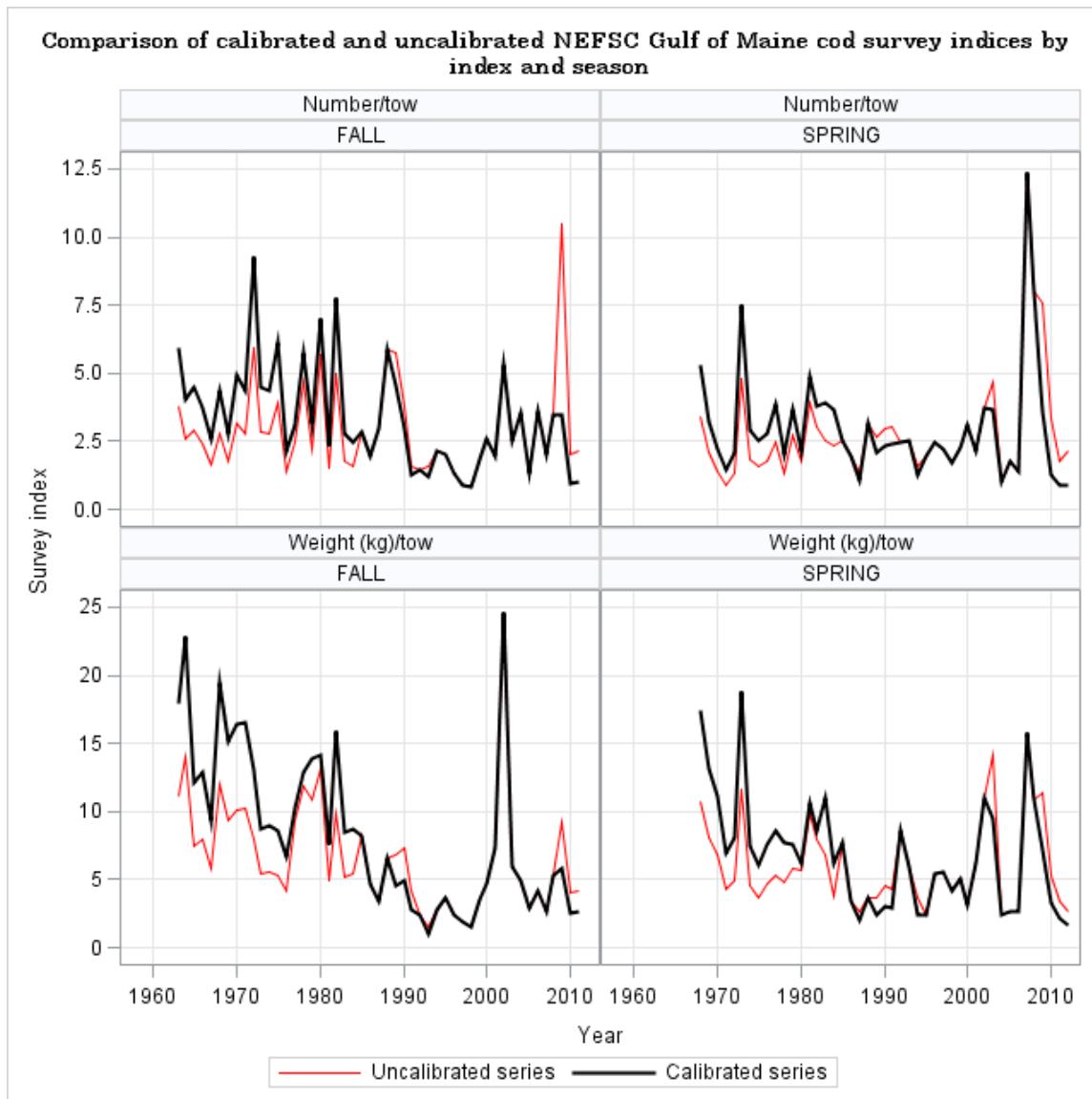
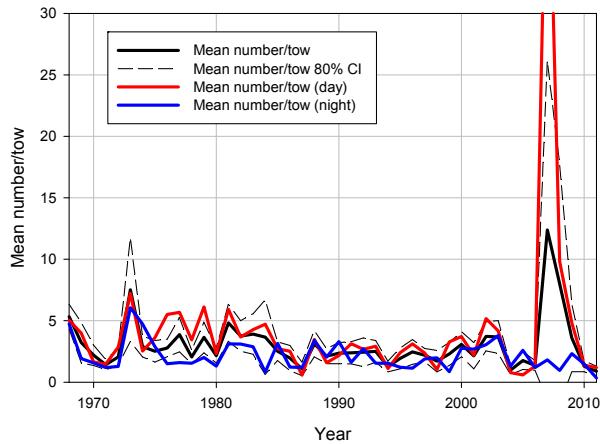
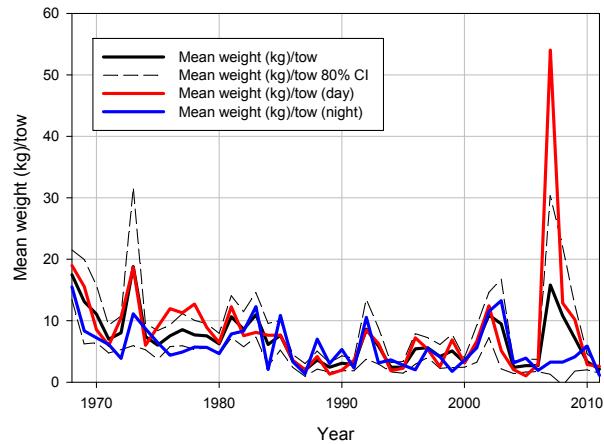


Figure A.95. Northeast Fisheries Science Center spring (right panels) and fall (left panels) survey indices of abundance (top panels) and biomass (bottom panels) showing both raw (unconverted) and vessel, door and survey converted indices over time for Gulf of Maine Atlantic cod.

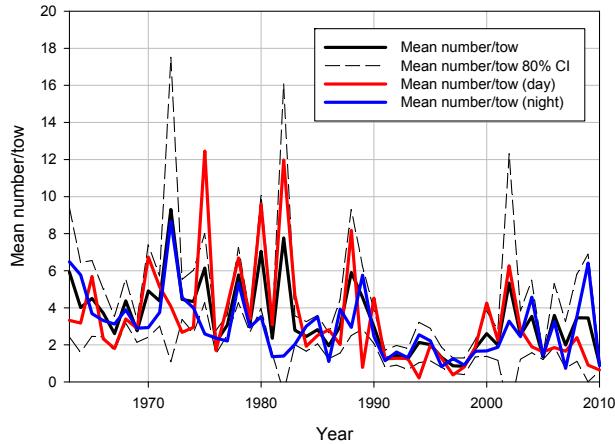
NEFSC spring survey: day/night comparisons of abundance



NEFSC spring survey: day/night comparisons of biomass



NEFSC fall survey: day/night comparisons of abundance



NEFSC fall survey: day/night comparisons of biomass

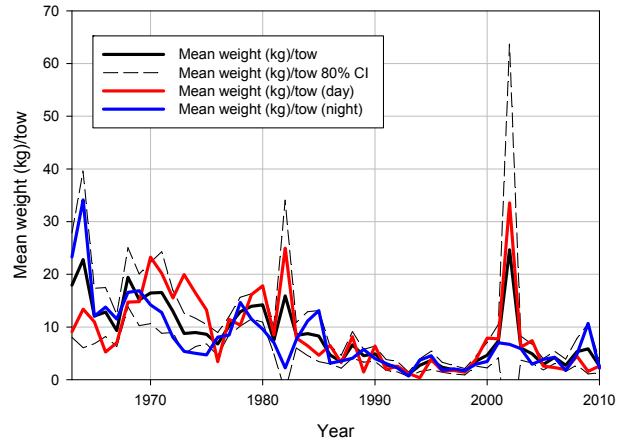


Figure A.96. Northeast Fisheries Science Center spring (top panels) and fall (bottom panels) survey indices of abundance (left panels) and biomass (right panels) broken down by day- and night-only tows compared to the aggregate index (day and night tows combined) and its associated 80% confidence interval (CI) for Gulf of Maine Atlantic cod.

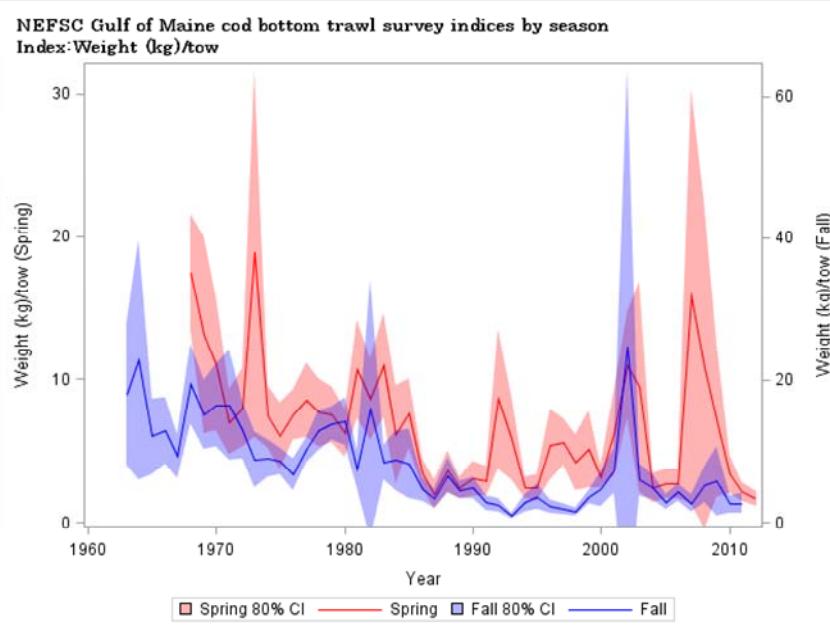
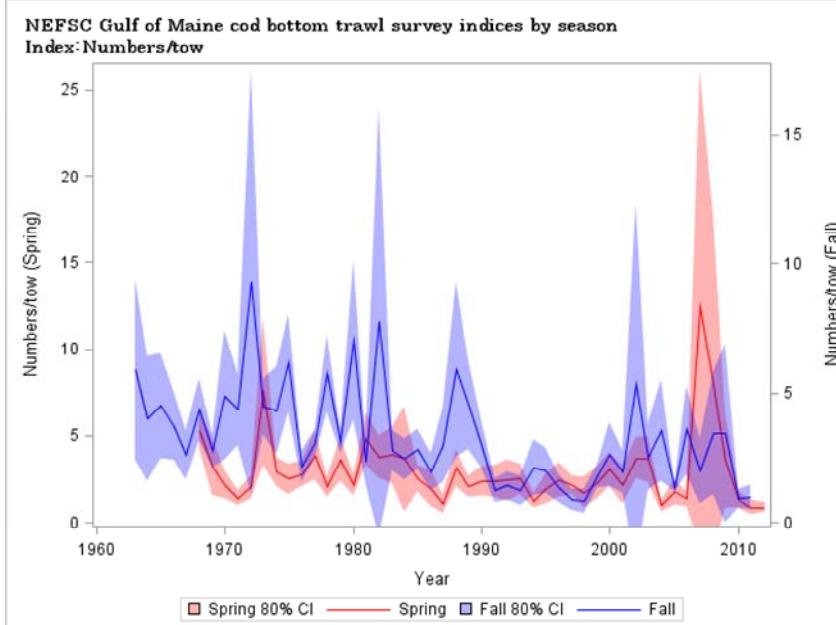


Figure A.97. Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl survey abundance (left) and biomass (right) indices for Gulf of Maine Atlantic cod from 1963 to 2012. *Note, the spring survey did not begin until 1968, 2012 fall survey data not available at time of this report.

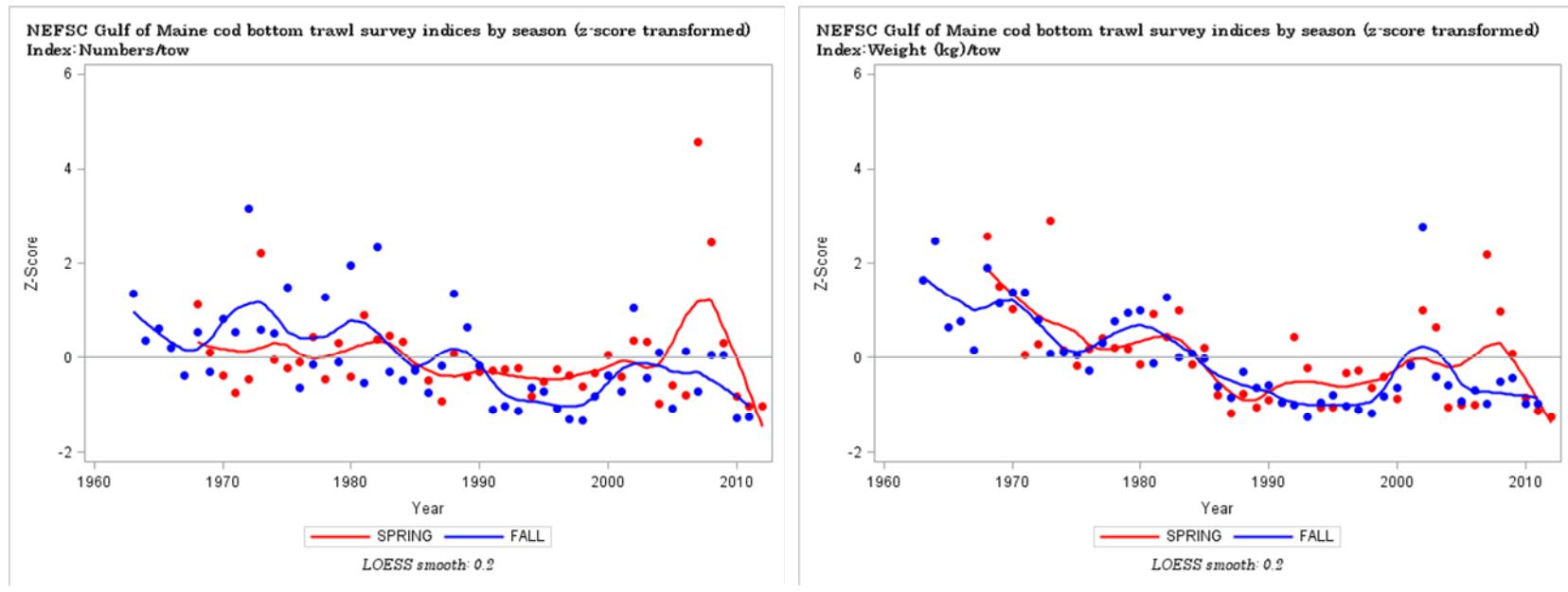


Figure A.98. Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl survey abundance (left) and biomass (right) indices for Gulf of Maine Atlantic cod from 1963 to 2012 expressed as z-scores ($[x-\mu]/\sigma$). *Note, the spring survey did not begin until 1968, 2012 fall survey data not available at time of this report.

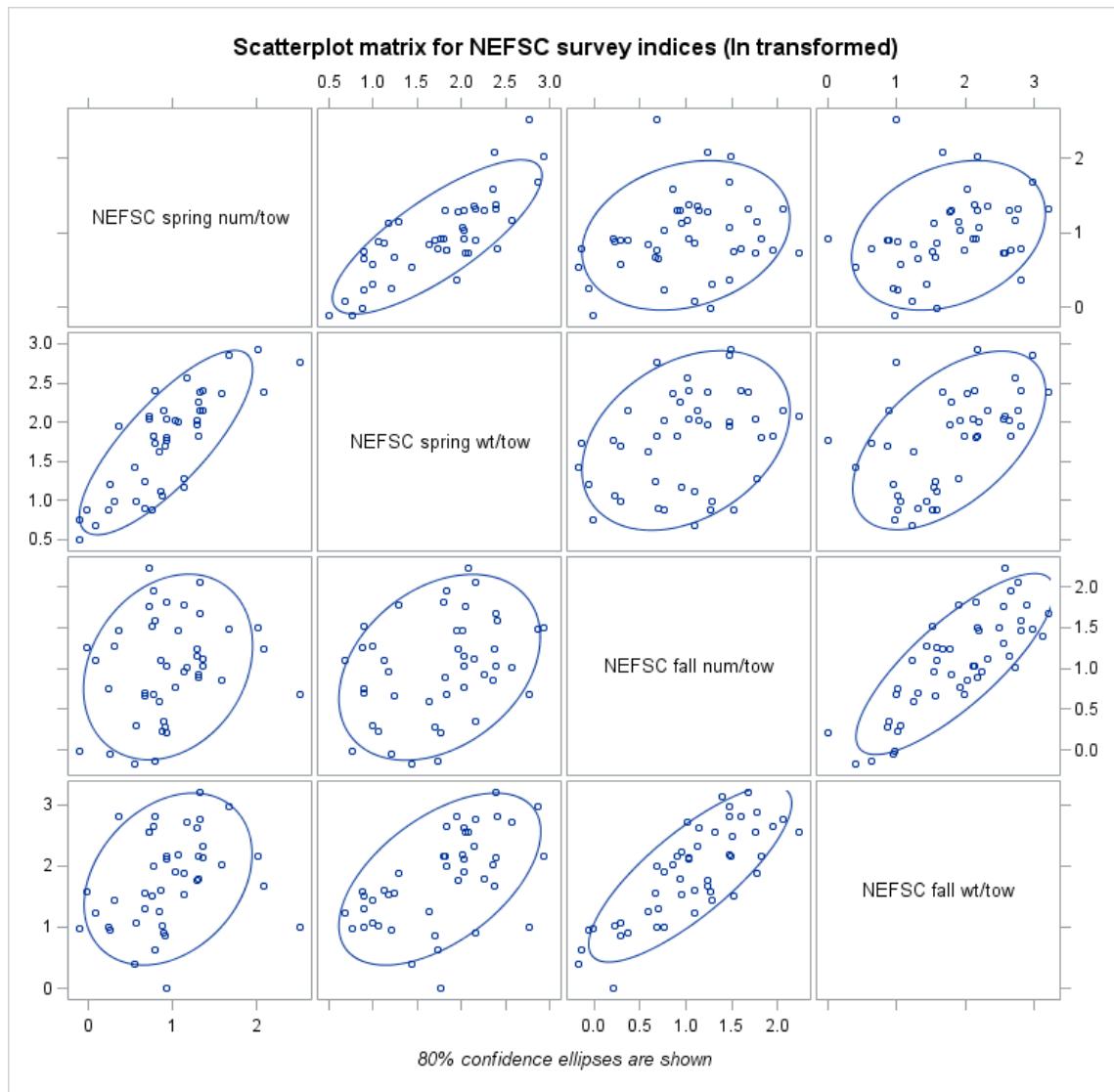


Figure A.99. Scatter plots showing the level of agreement between Northeast Fisheries Science Center (NEFSC) bottom trawl survey indices (log transformed) of Gulf of Maine Atlantic cod. 80% confidence ellipses are shown.

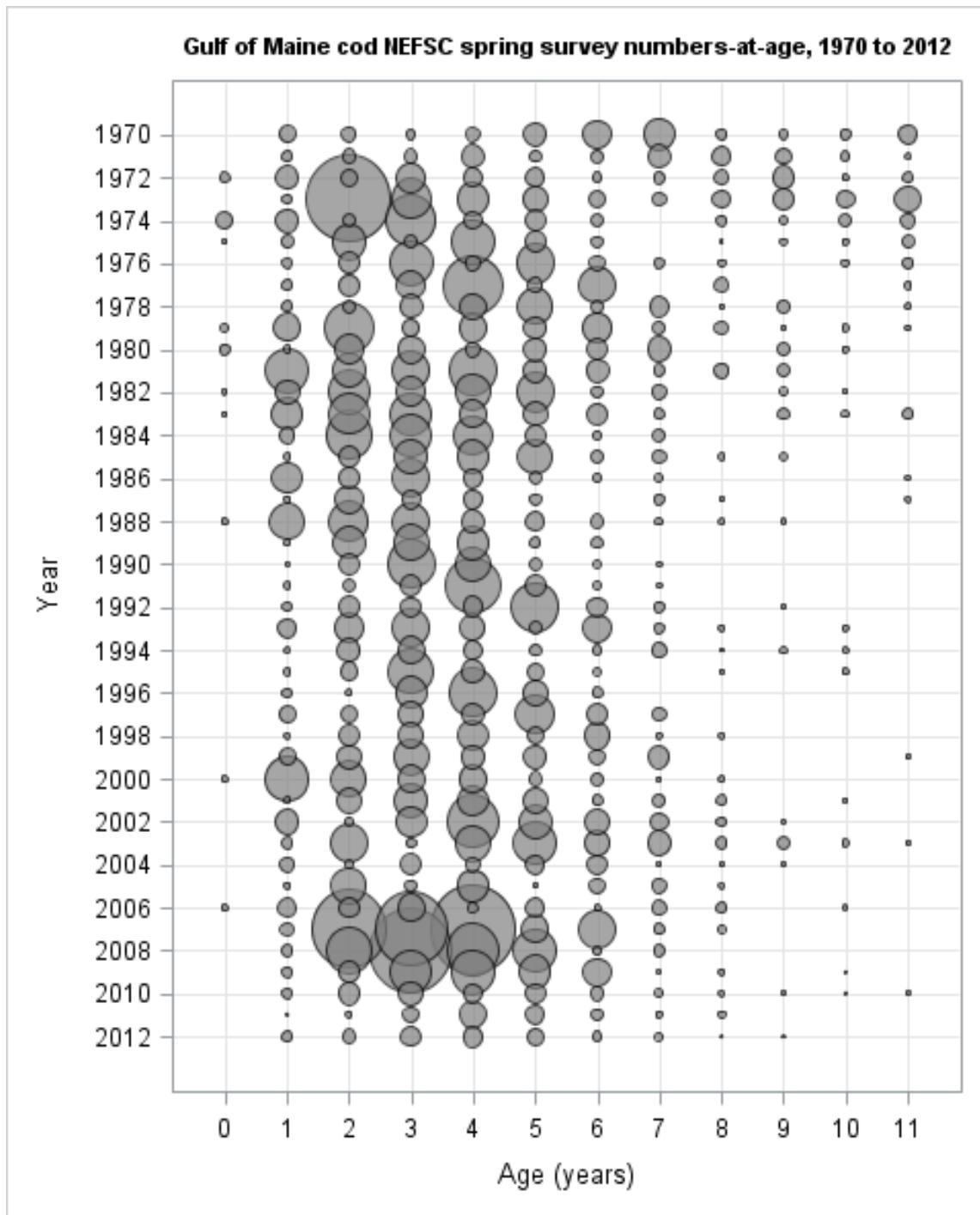


Figure A.100. Numbers-at-age from NEFSC spring bottom trawl survey from 1970 to 2012 for Gulf of Maine Atlantic cod. *Note that age 11 is a plus group.

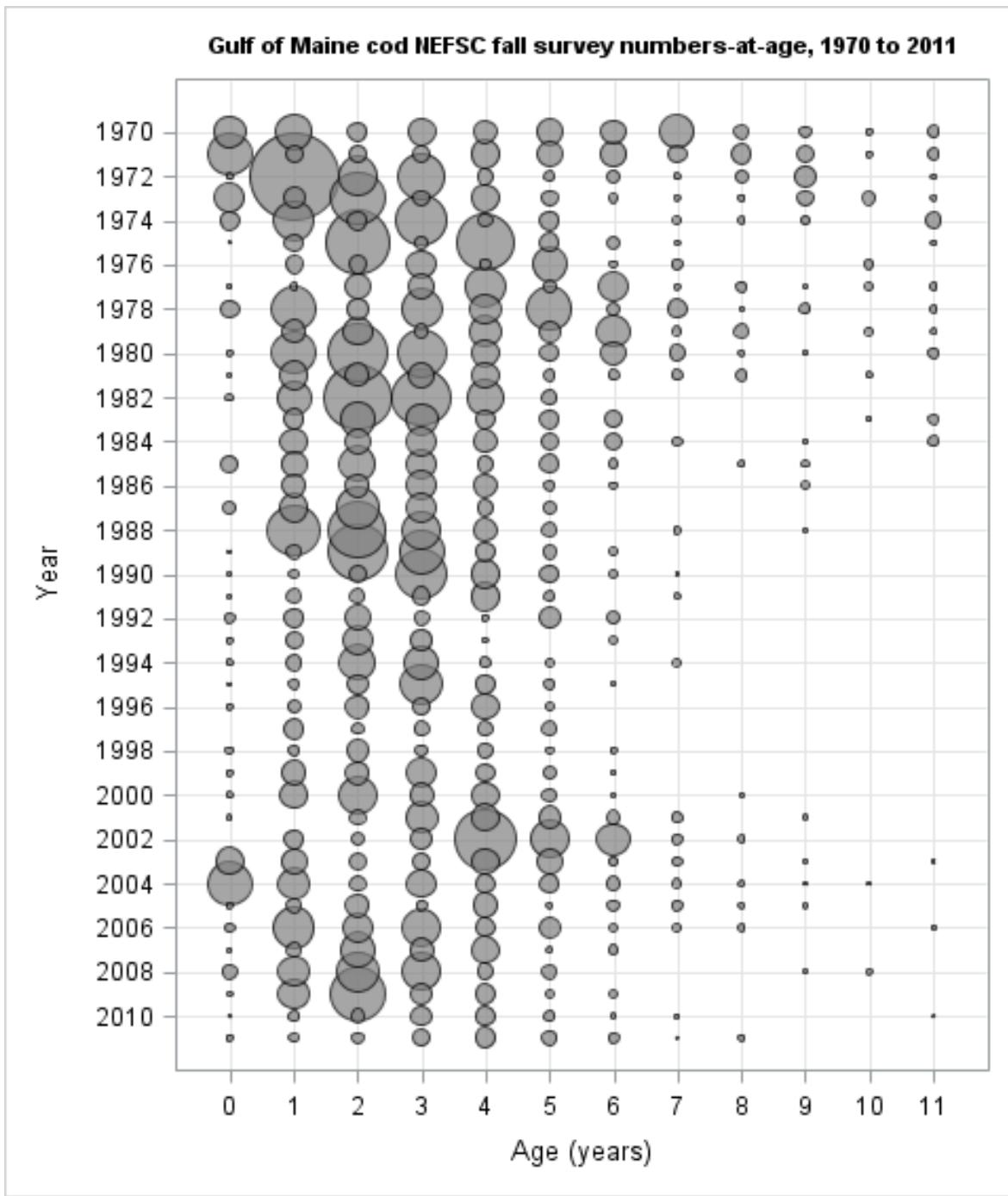


Figure A.101. Numbers-at-age from NEFSC fall bottom trawl survey from 1970 to 2011 for Gulf of Maine Atlantic cod. *Note that age 11 is a plus group.

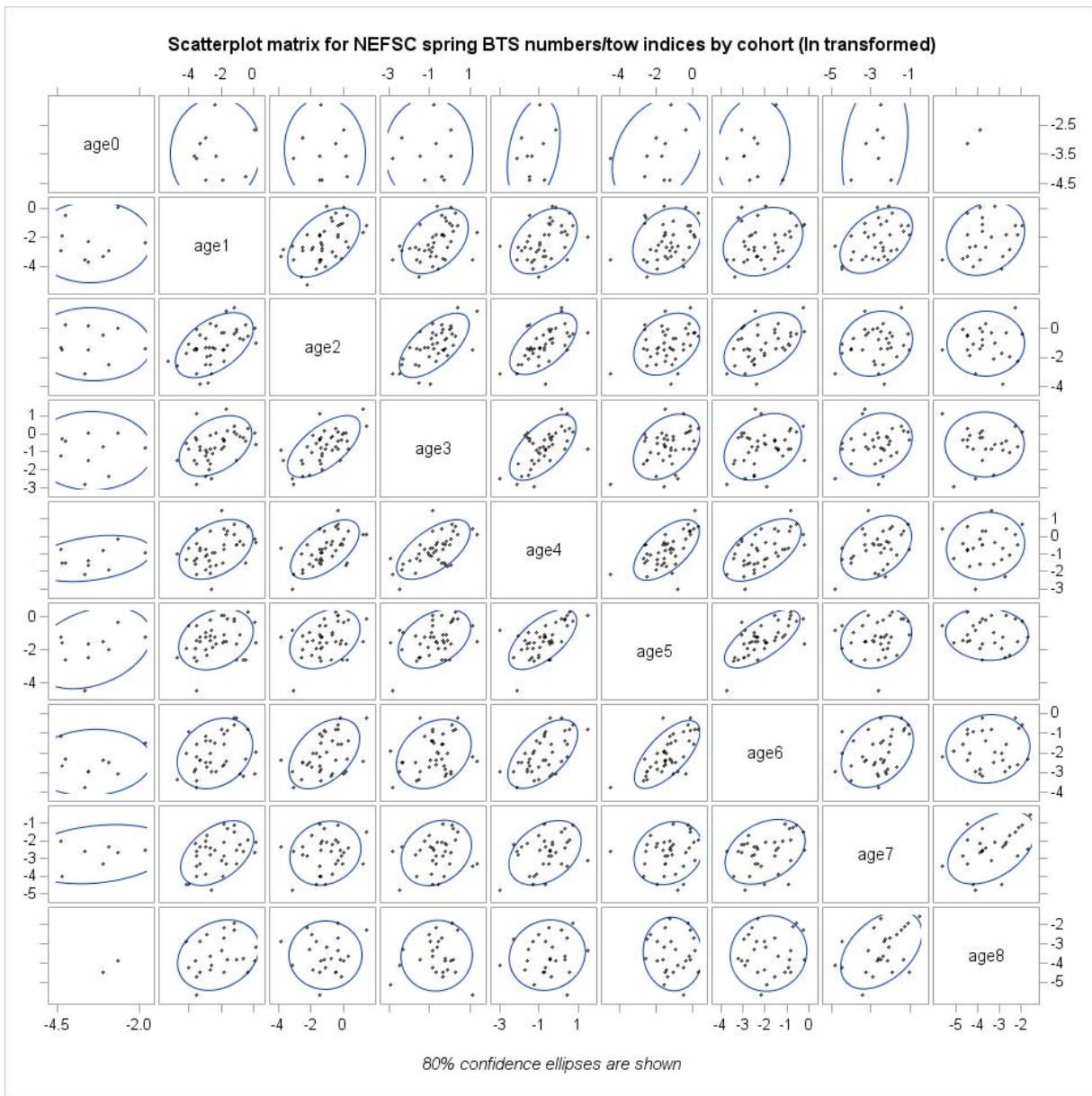


Figure A.102. Scatter plots showing the level of agreement between Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey Gulf of Maine Atlantic cod indices at age (log transformed) on a cohort basis. 80% confidence ellipses are shown.

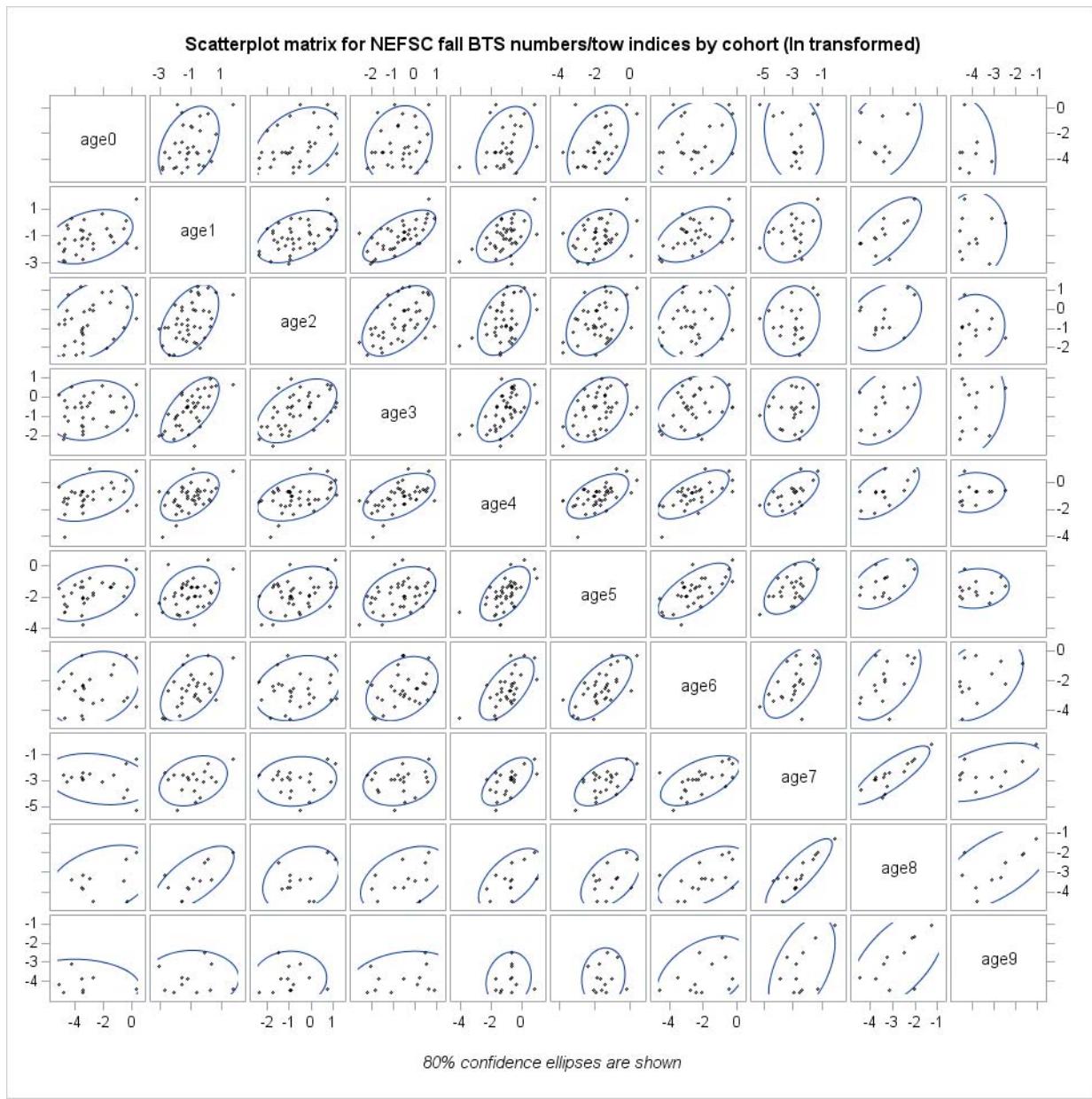


Figure A.103. Scatter plots showing the level of agreement between Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey Gulf of Maine Atlantic cod indices at age (log transformed) on a cohort basis. 80% confidence ellipses are shown.

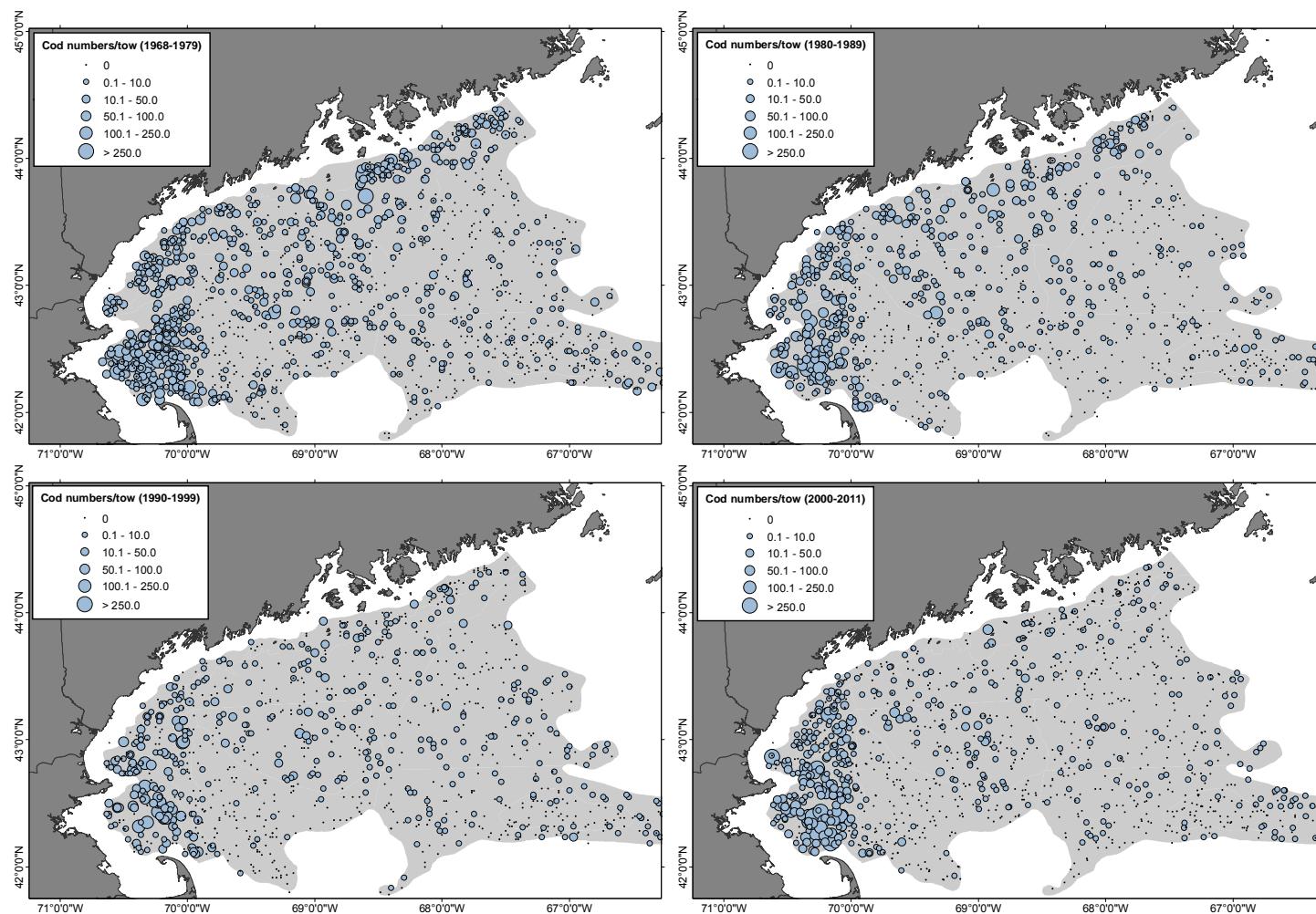


Figure A.104. Spatial distribution of Gulf of Maine Atlantic cod catches (numbers/tow) from the Northeast Fisheries Science Center spring bottom trawl survey from 1968 – 2011. Periods are as follows: 1968 – 1979 (top left), 1980 – 1989 (top right), 1990 – 1999 (bottom left), 2000-2011 (bottom right).

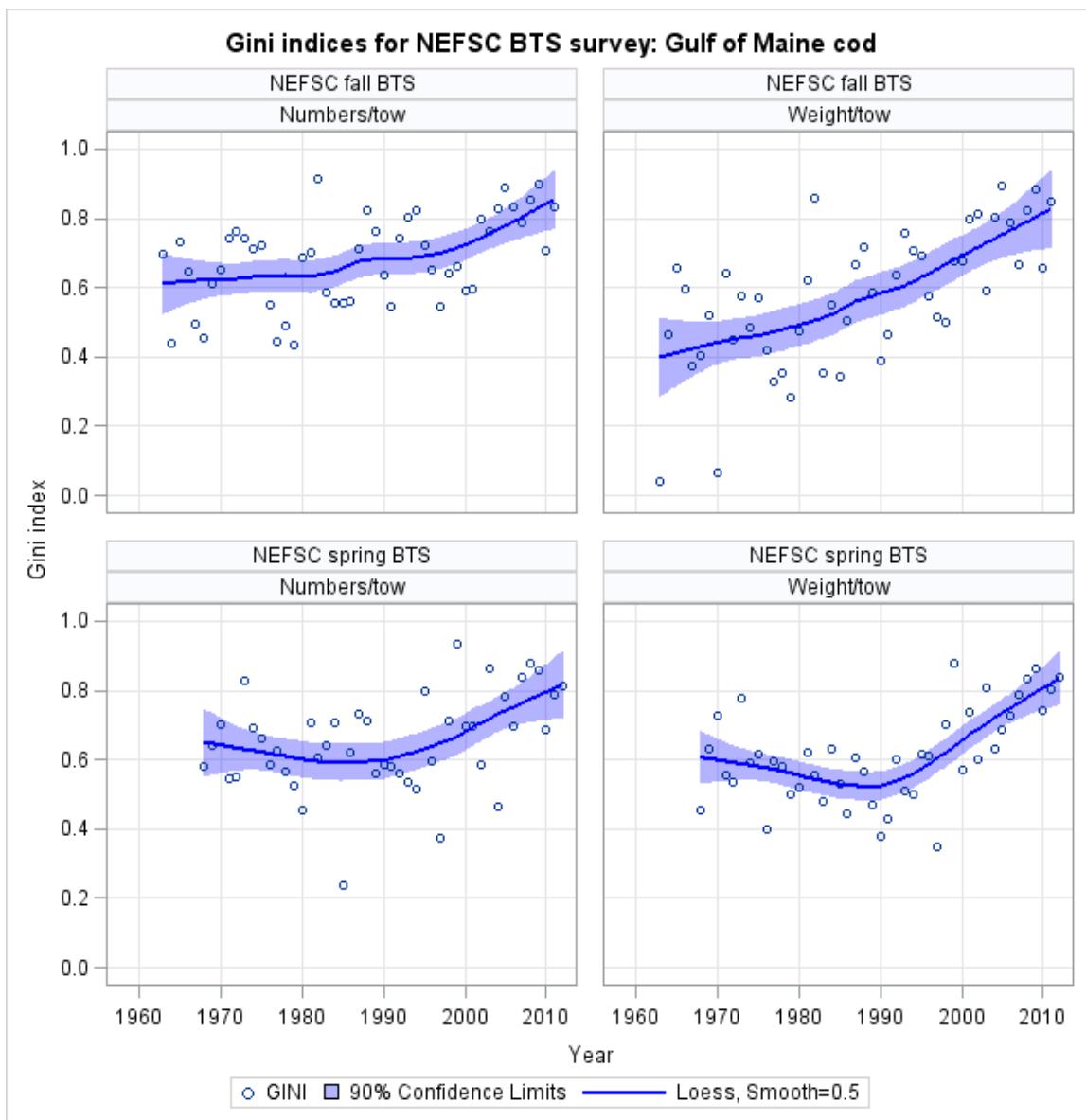


Figure A.105. Gini indices for Gulf of Maine Atlantic cod from the Northeast Fisheries Science Center (NEFSC) fall (top) and spring (bottom) bottom trawl surveys in terms of abundance (numbers/tow, left) and biomass (kg/tow, right). A loess smooth has been fit to the data with smoothing parameter of 0.5. The loess smooth is shown by the solid blue line along with the corresponding 90% confidence interval.

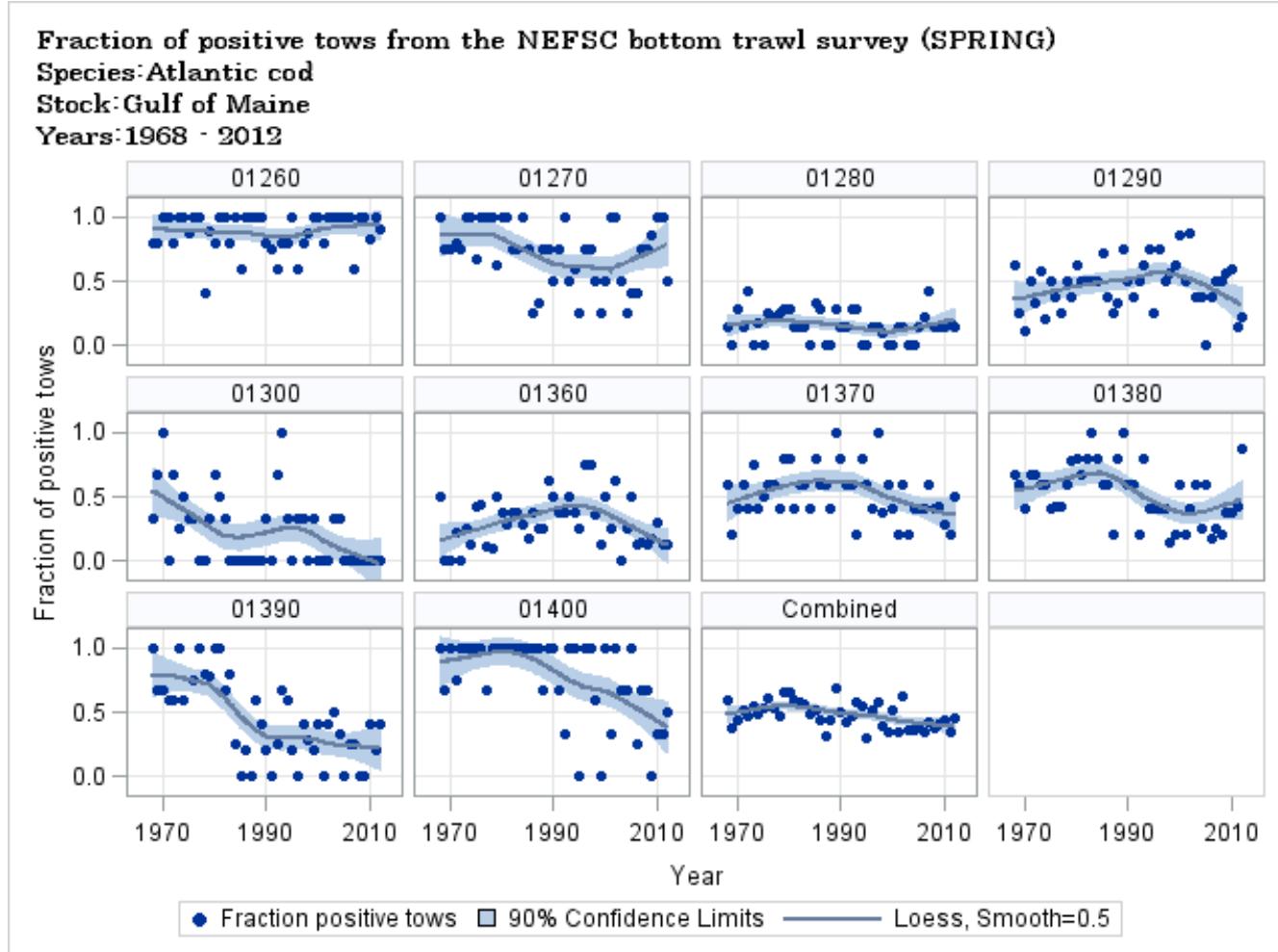


Figure A.106. Fraction of Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey tows with positive catches of Gulf of Maine Atlantic cod by strata from 1968-2012.

Fraction of positive tows from the NEFSC bottom trawl survey (FALL)

Species: Atlantic cod

Stock: Gulf of Maine

Years: 1963 - 2011

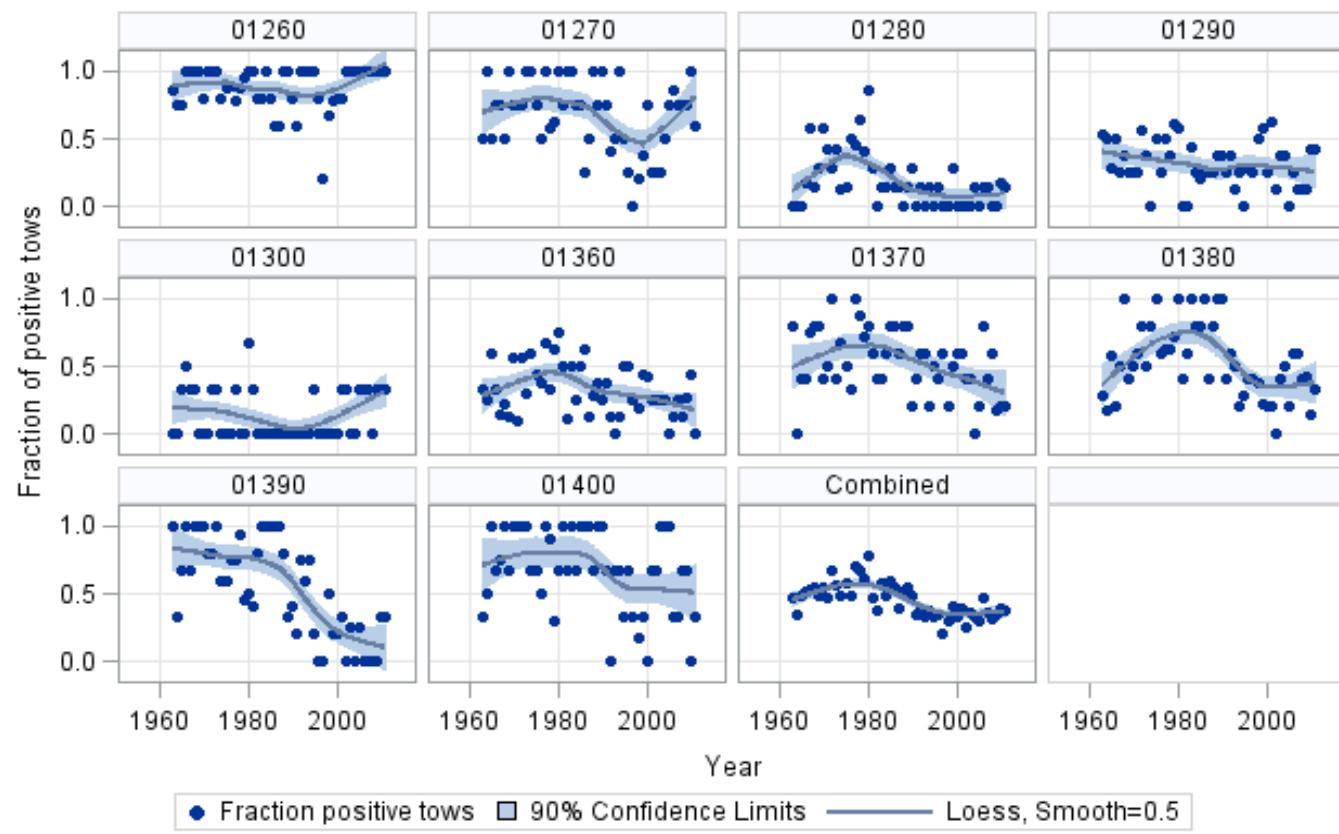


Figure A.107. Fraction of Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey tows with positive catches of Gulf of Maine Atlantic cod by strata from 1963-2012.

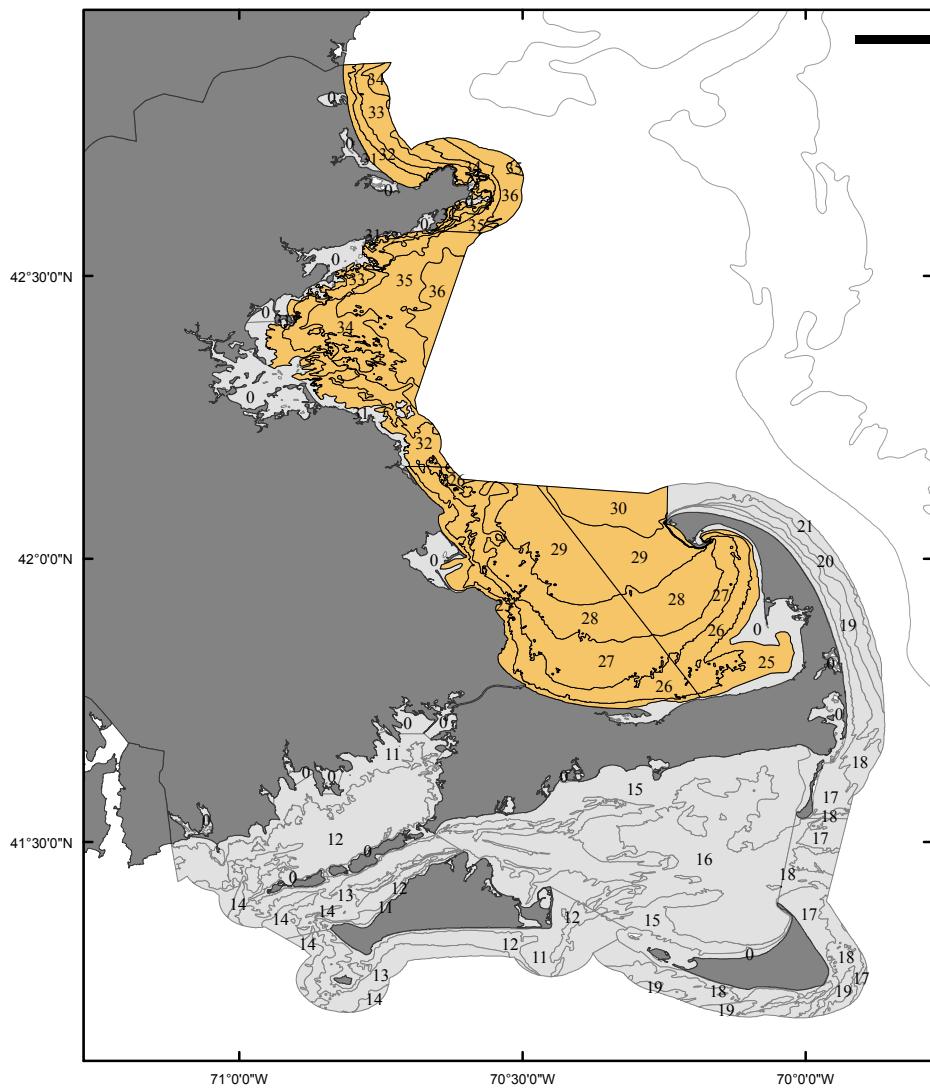


Figure A.108. Map of the Massachusetts Department of Marine Fisheries (MADMF) bottom trawl survey strata included in the Gulf of Maine Atlantic cod stock assessment (shaded orange).

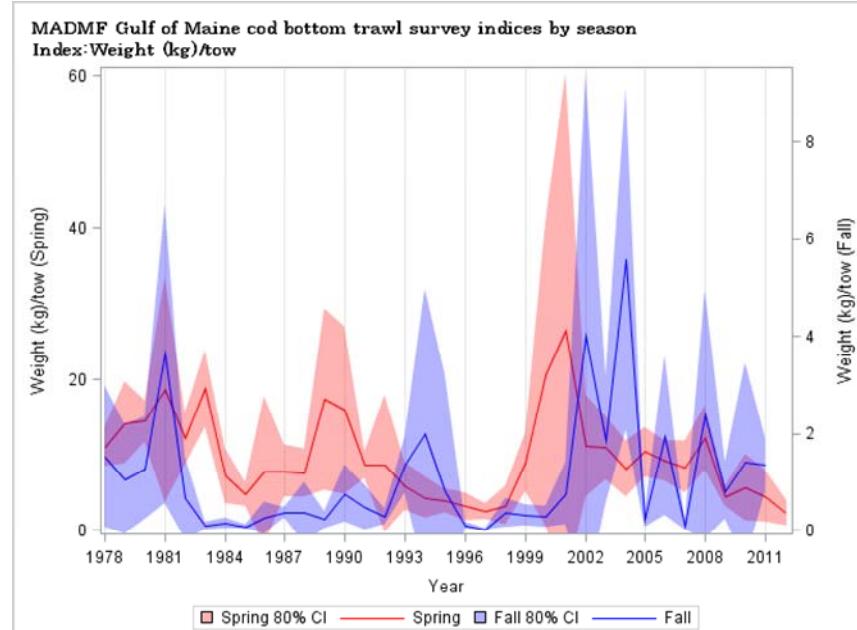
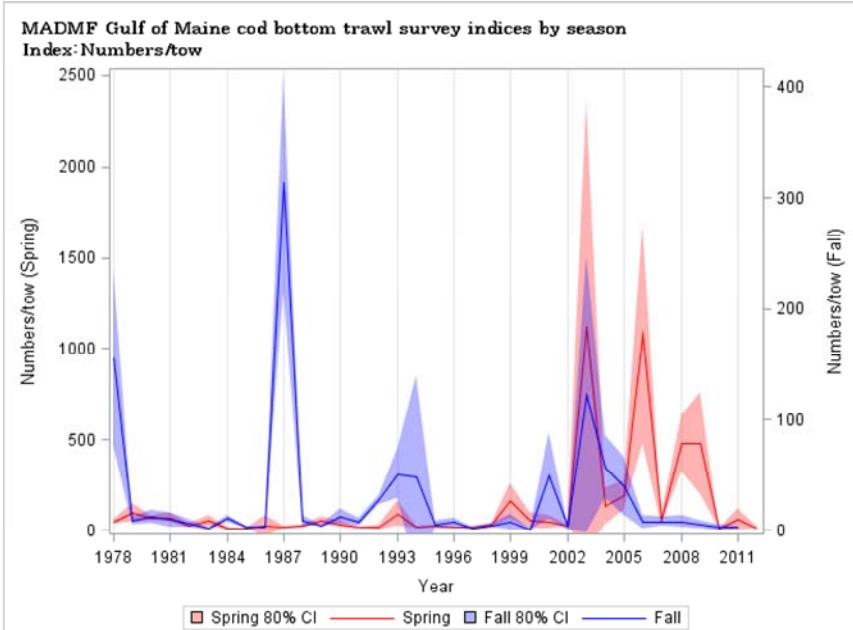


Figure A.109. Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl survey abundance (top) and biomass (bottom) indices of Gulf of Maine Atlantic cod from 1978 to 2012. *Note, 2012 fall survey data not available at time of this report.

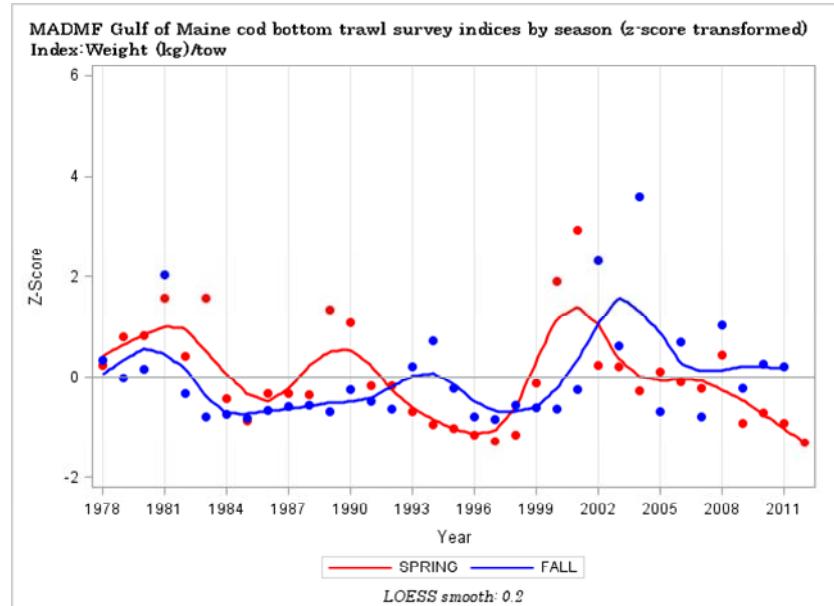
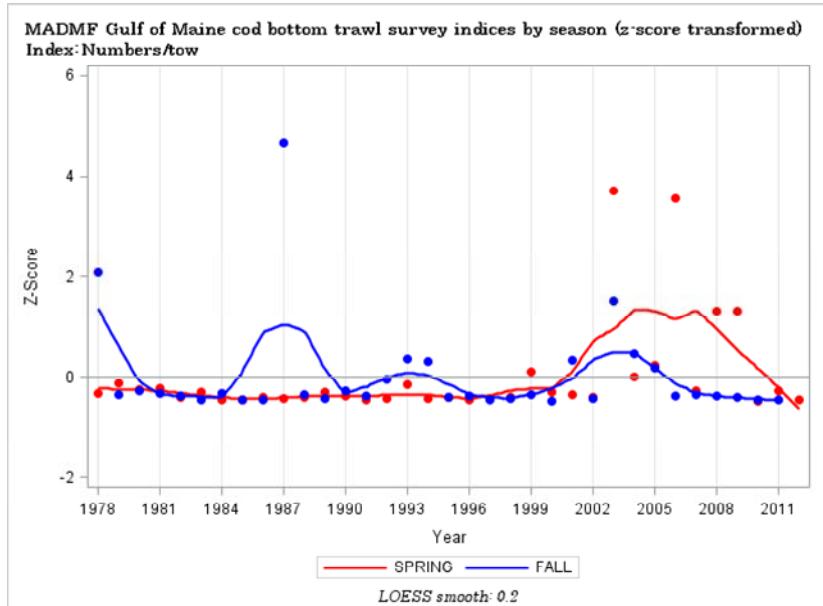


Figure A.110. Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl survey abundance (top) and biomass (bottom) indices of Gulf of Maine Atlantic cod from 1978 to 2012 expressed as z-scores ($[x-\mu]/\sigma$). *Note, 2012 fall survey data not available at time of this report.

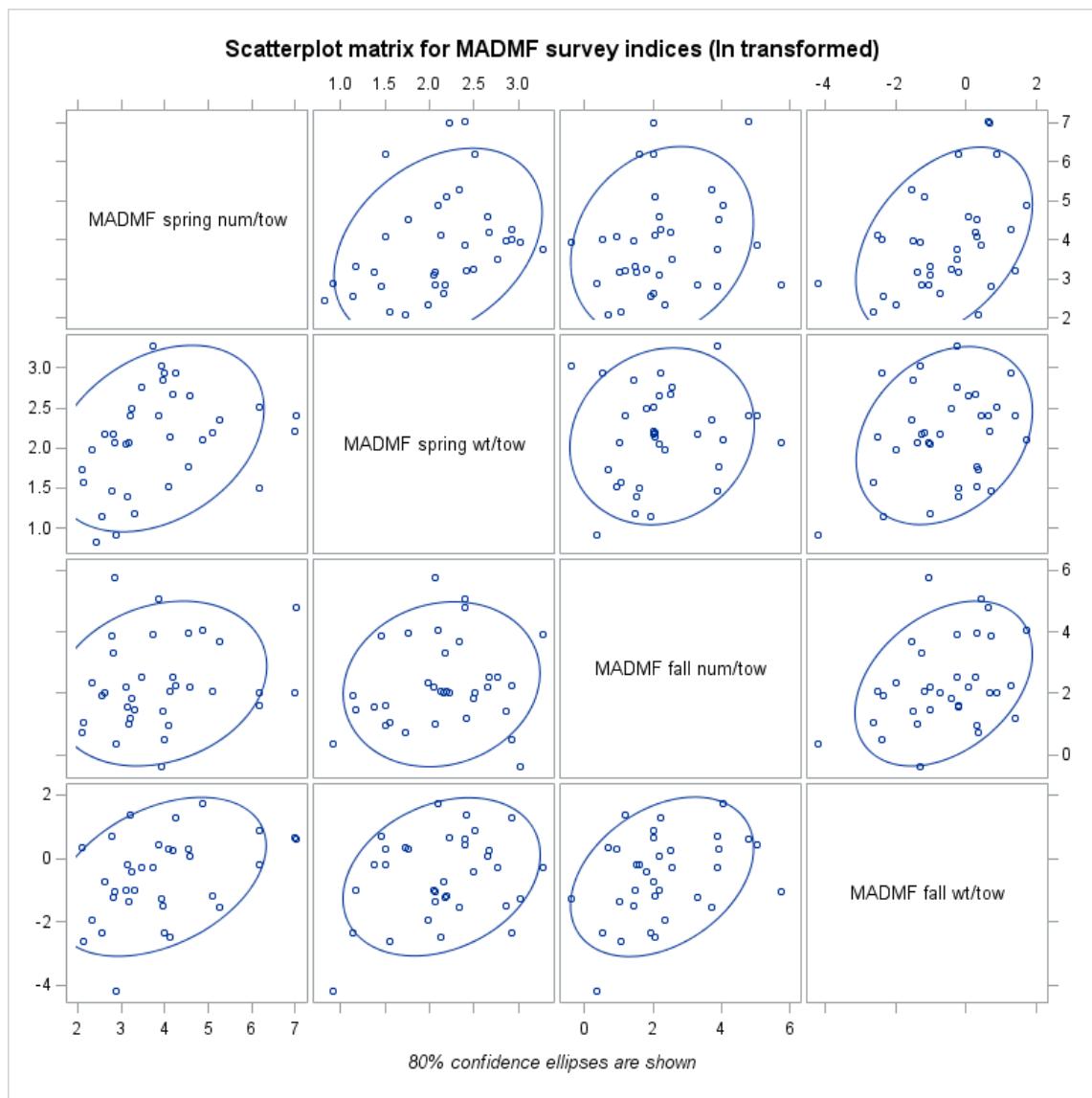


Figure A.111. Scatter plots showing the level of agreement between the Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl survey Gulf of Maine Atlantic cod indices (log transformed). 80% confidence ellipses are shown.

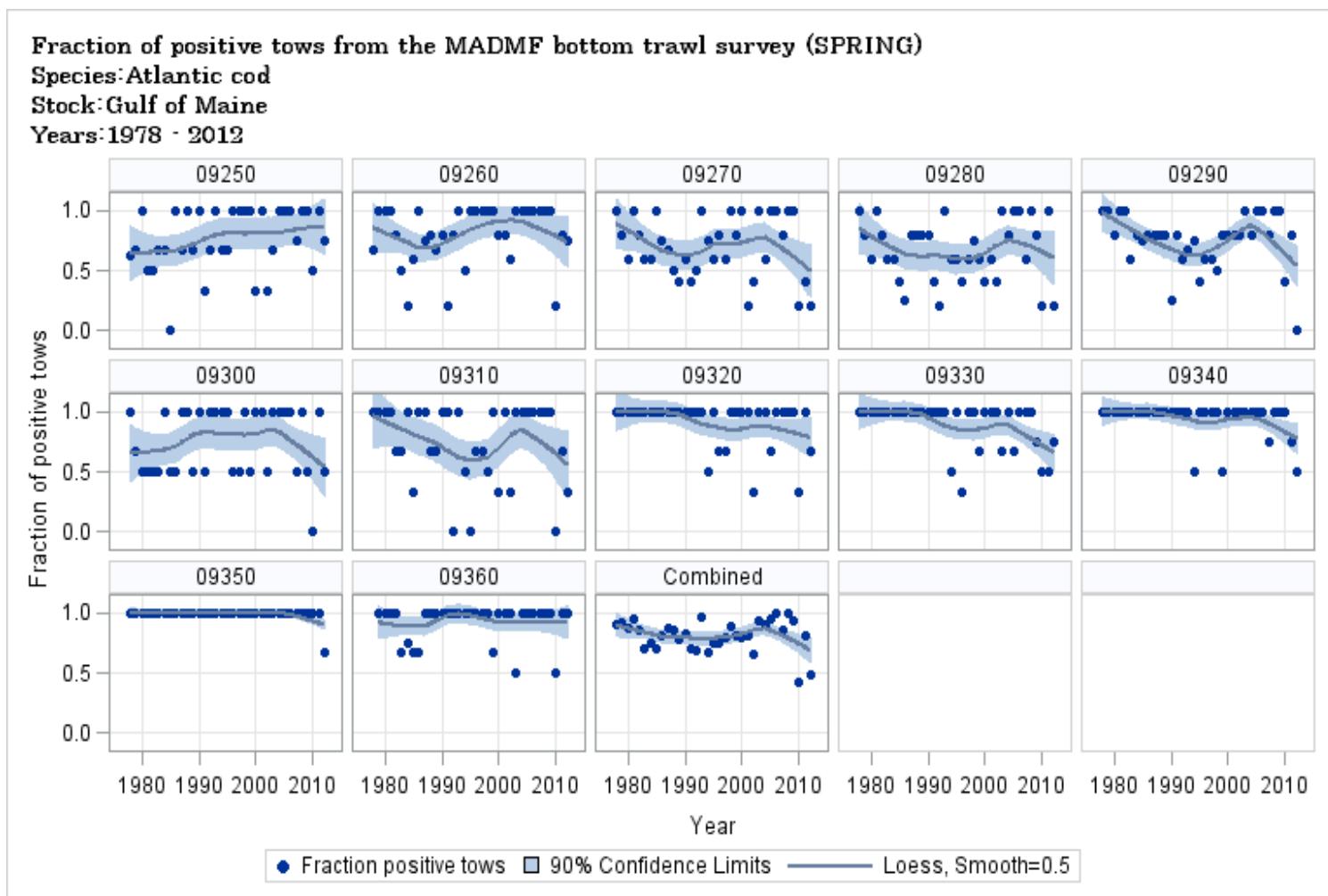


Figure A.112. Fraction of Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl survey survey tows with positive catches of Gulf of Maine Atlantic cod by strata from 1978-2012.

Fraction of positive tows from the MADMF bottom trawl survey (FALL)

Species: Atlantic cod

Stock: Gulf of Maine

Years: 1978 - 2011

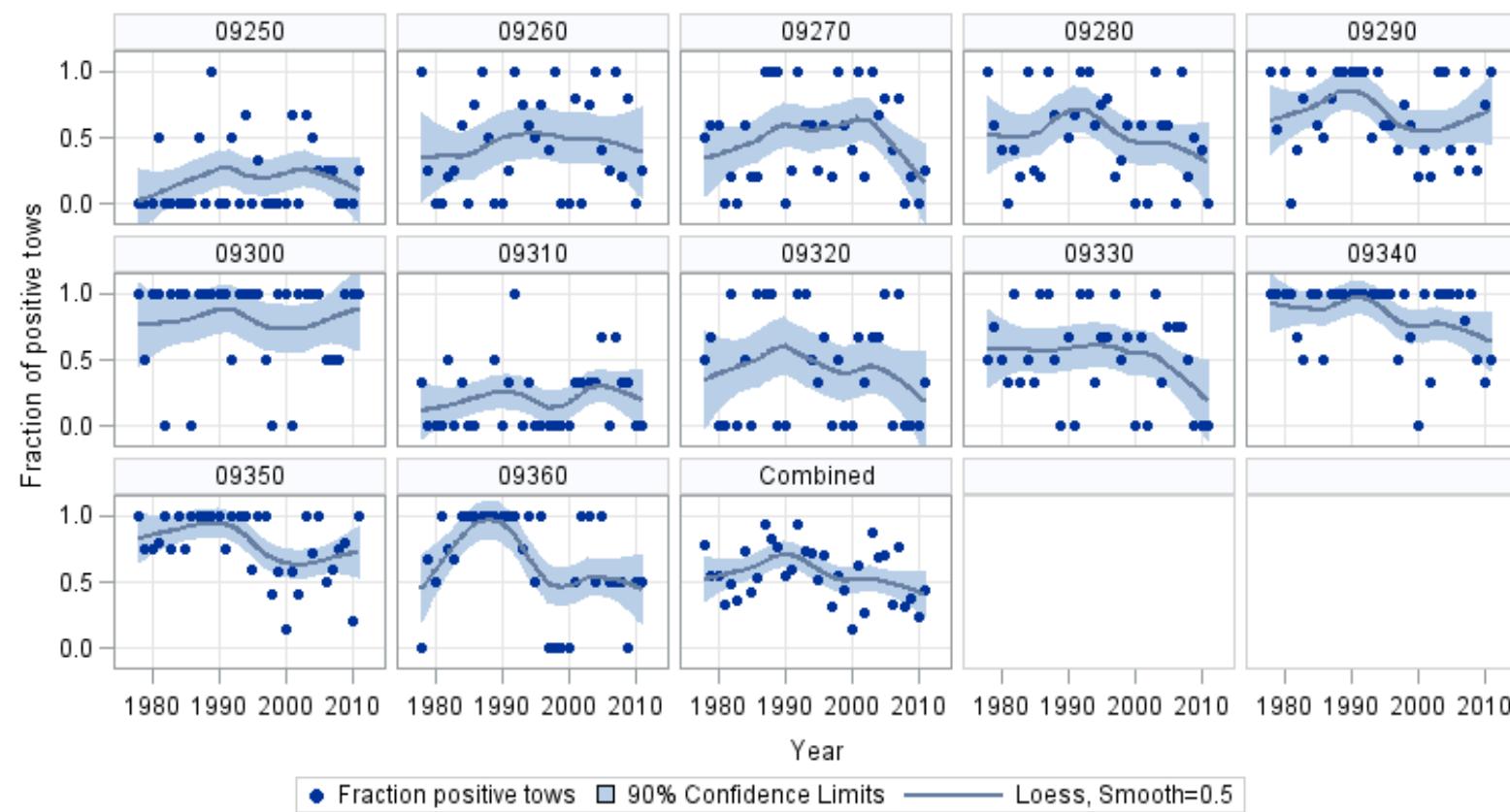


Figure A.113. Fraction of Massachusetts Department of Marine Fisheries (MADMF) fall bottom trawl survey survey tows with positive catches of Gulf of Maine Atlantic cod by strata from 1978-2011.

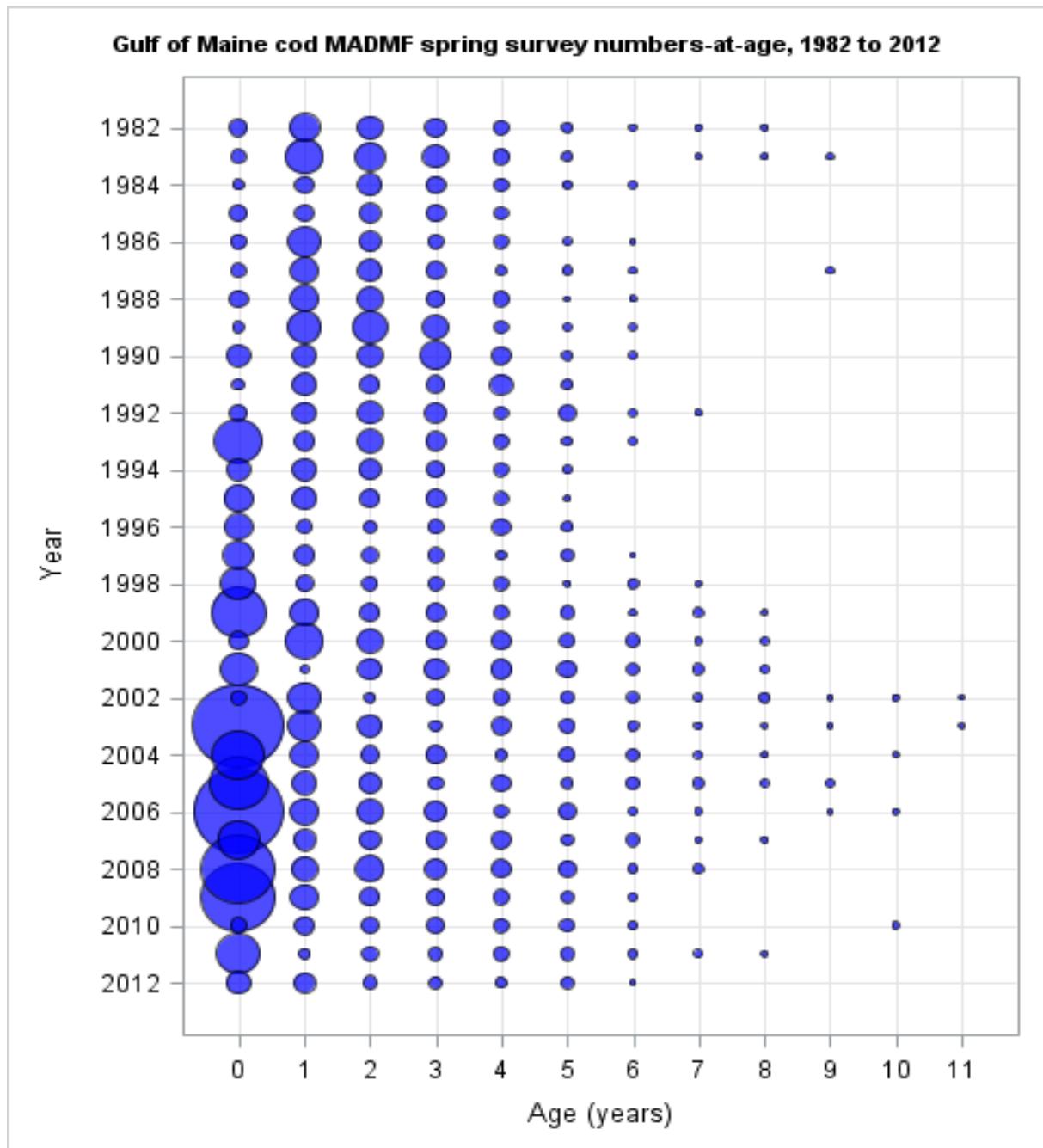


Figure A.114. Gulf of Maine cod numbers-at-age from the Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl survey, 1982-2012. There was insufficient age information available from the MADMF spring survey prior to 1982. *Note that age 11 is a plus group.

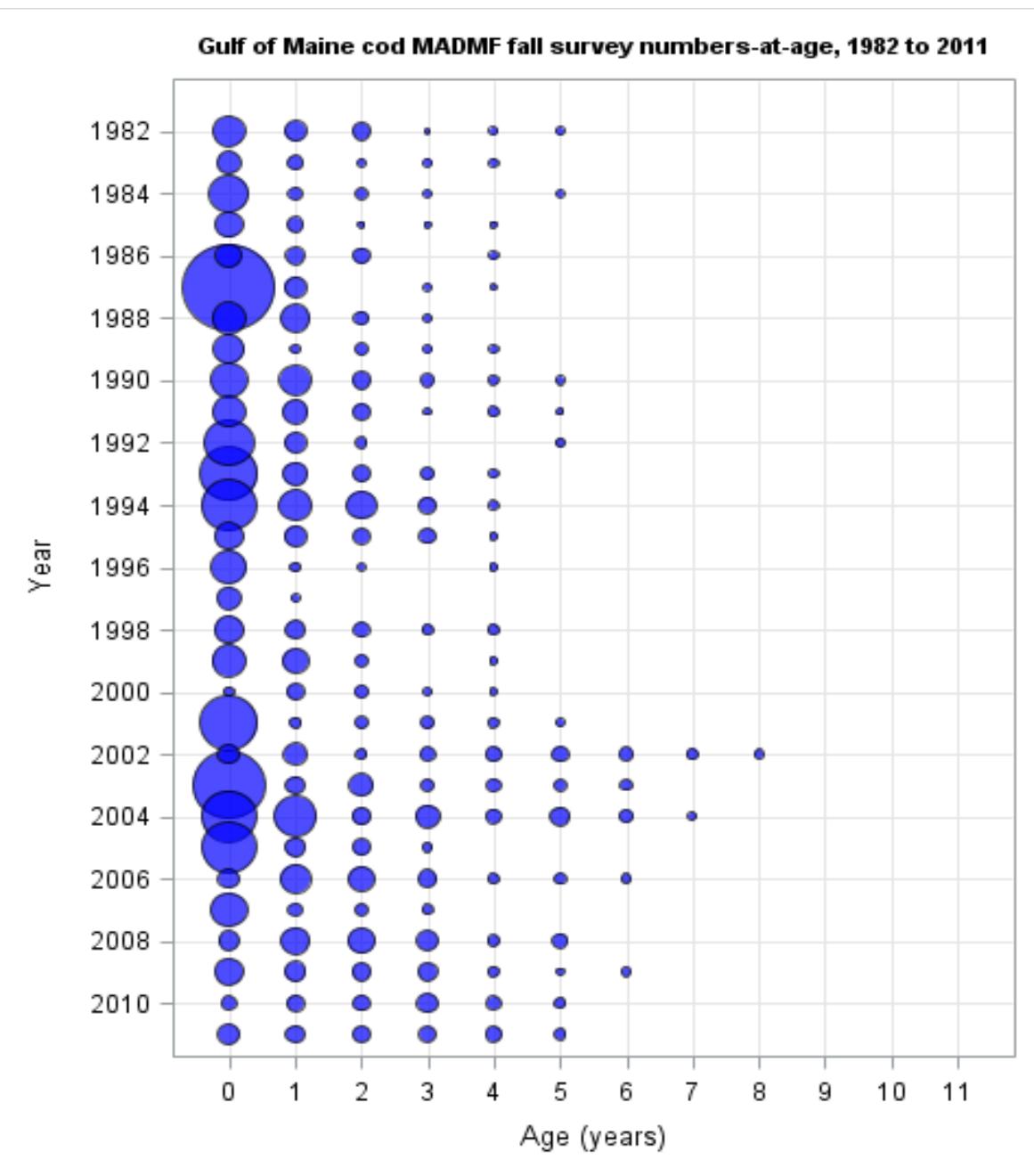


Figure A.115. Gulf of Maine cod numbers-at-age from the Massachusetts Department of Marine Fisheries (MADMF) fall bottom trawl survey, 1982-2011. There was insufficient age information available from the MADMF fall survey prior to 1982. *Note that age 11 is a plus group.

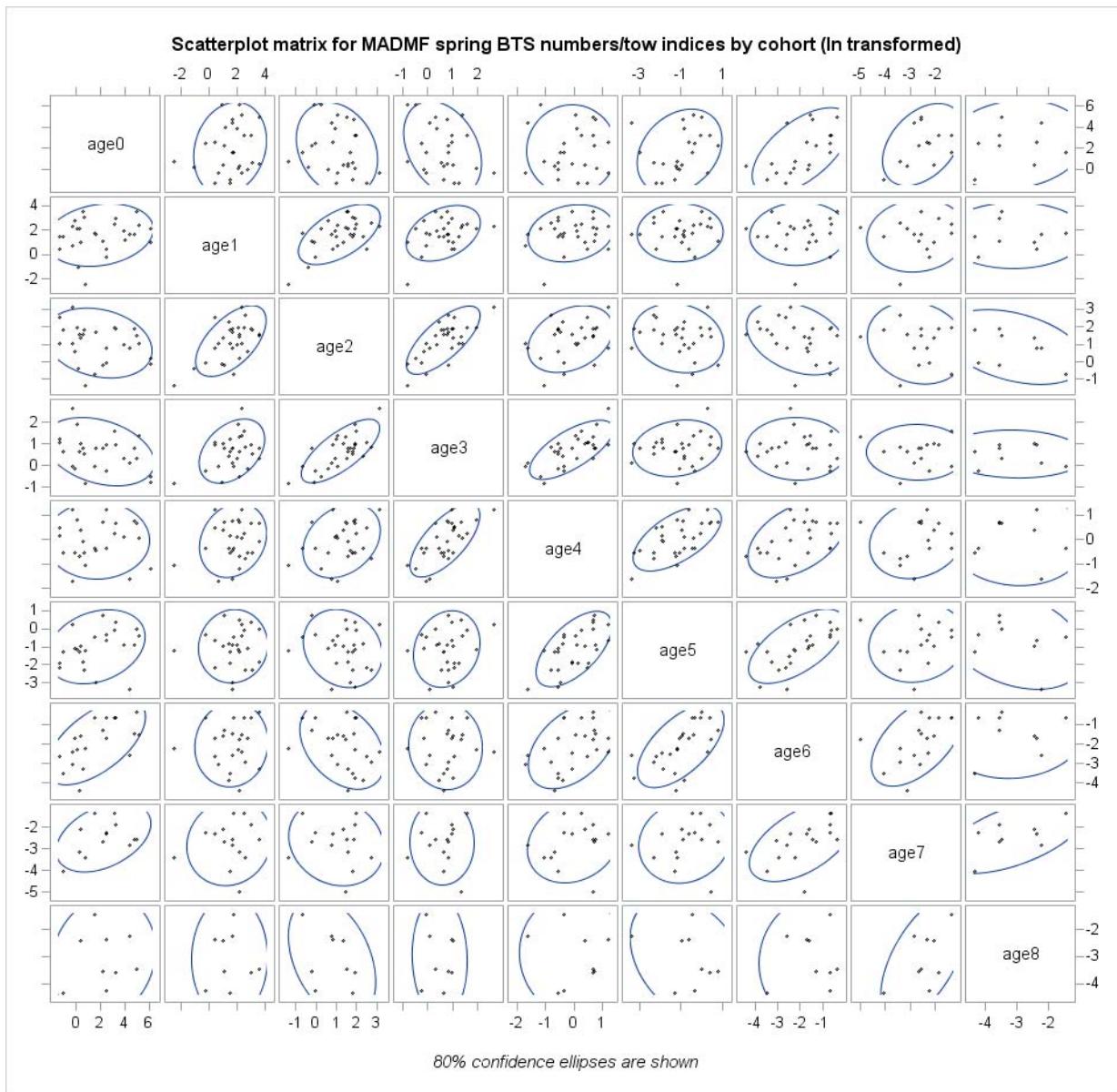


Figure A.116. Scatter plots showing the level of agreement between Massachusetts Department of Marine Fisheries (MADMF) spring bottom trawl Gulf of Maine Atlantic cod survey indices at age (log transformed) on a cohort basis. 80% confidence ellipses are shown.

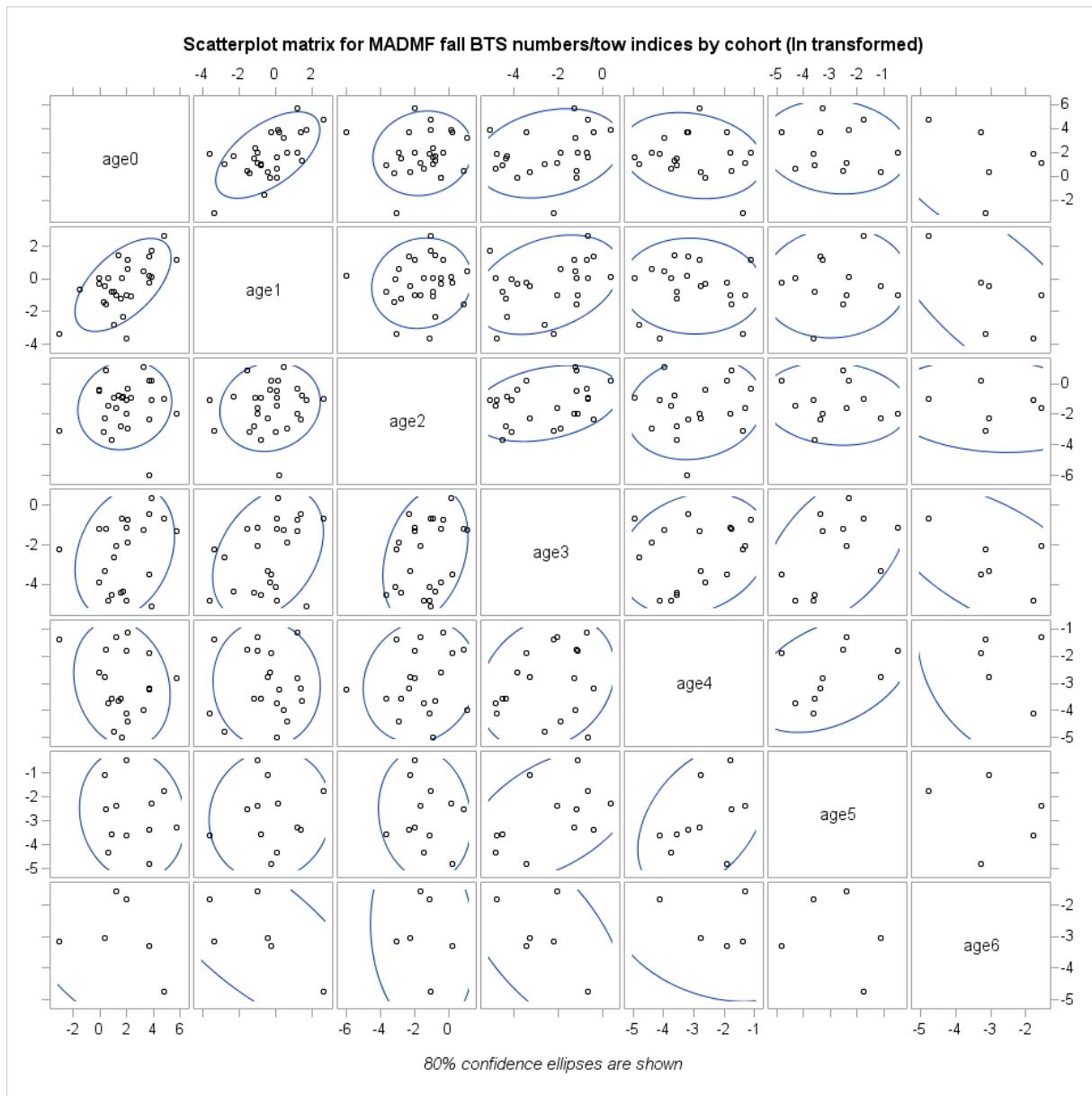


Figure A.117. Scatter plots showing the level of agreement between Massachusetts Department of Marine Fisheries (MADMF) fall bottom trawl Gulf of Maine Atlantic cod survey indices at age (log transformed) on a cohort basis. 80% confidence ellipses are shown.

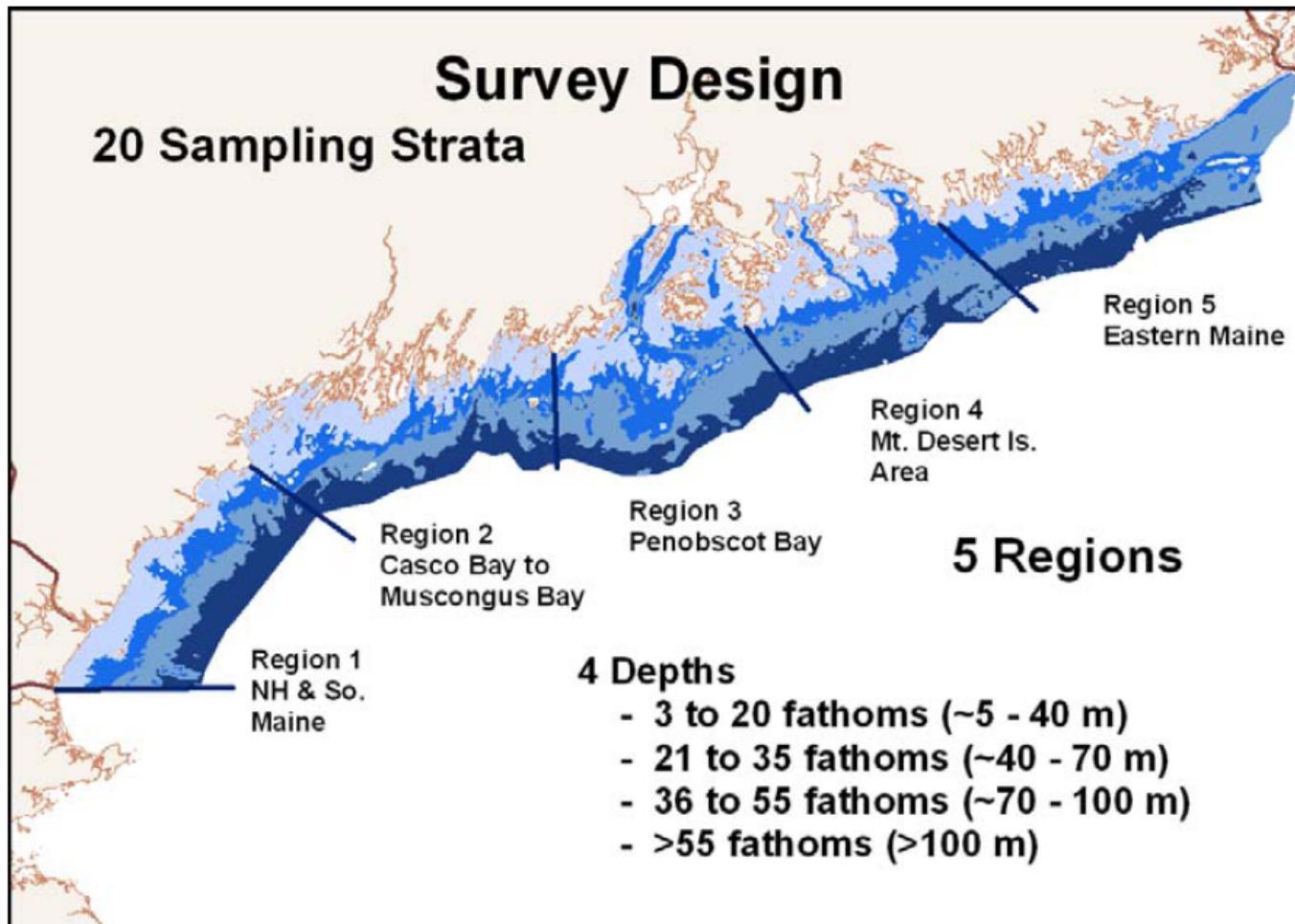


Figure A. 118. Map of the Maine – New Hampshire inshore groundfish trawl survey strata set (map from Sherman et al. 2005).

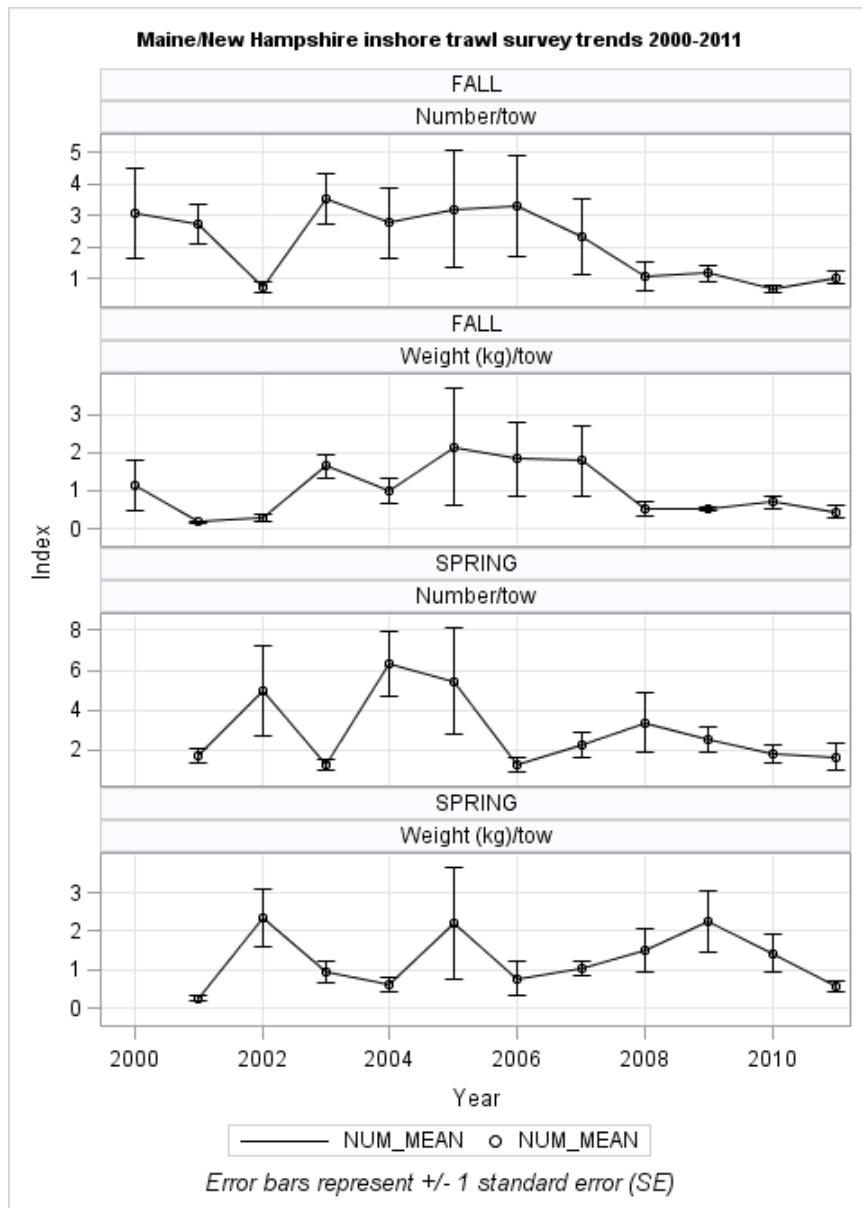


Figure A.119. Maine – New Hampshire inshore groundfish trawl survey spring and fall survey abundance from 1978 to 2011 for Gulf of Maine Atlantic cod. Bars indicate ± 1 standard error (SE). Data provided by S. Sherman (pers. comm.).

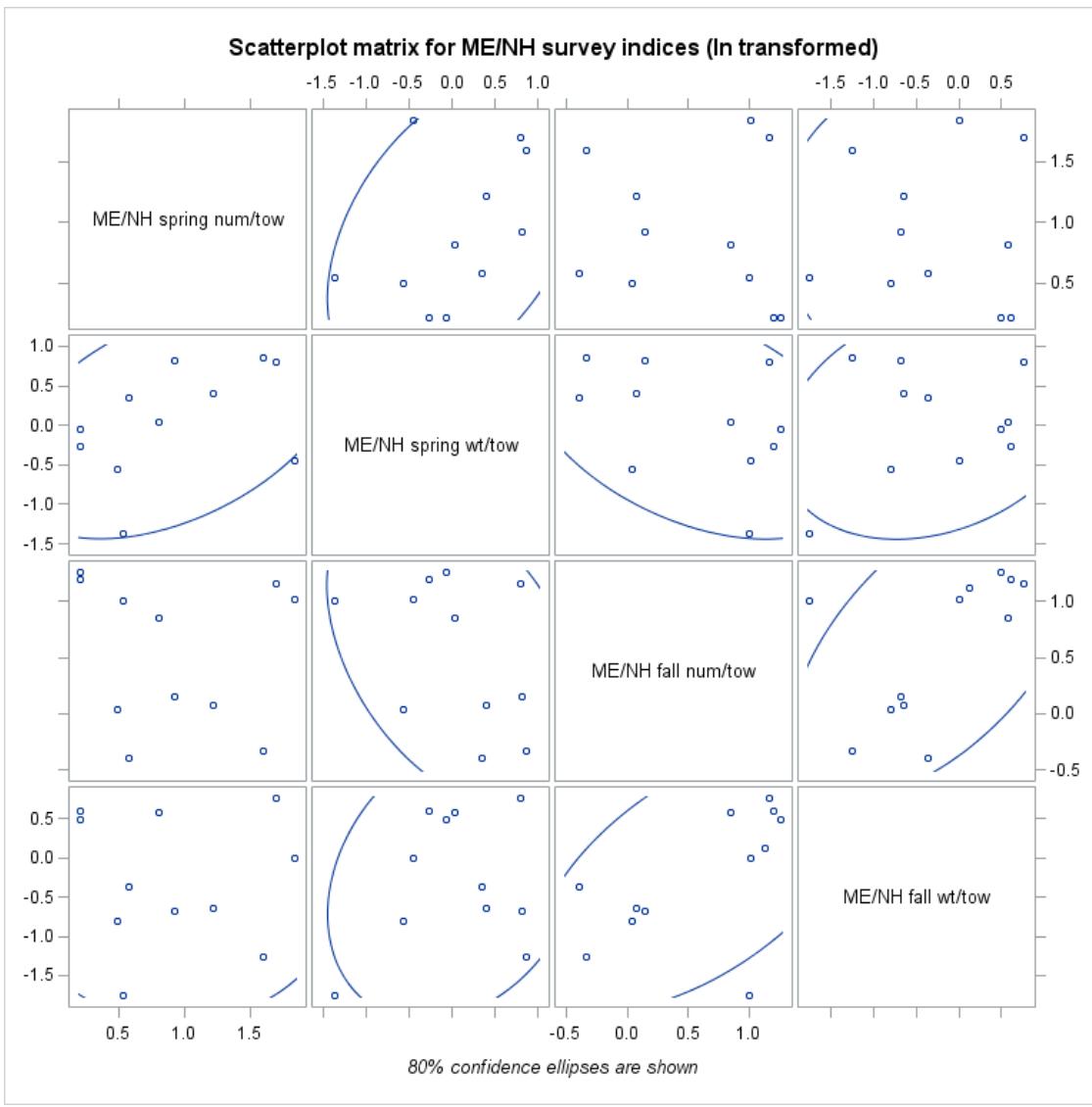


Figure A.120. Scatter plots showing the level of agreement between the Maine – New Hampshire (ME/NH) inshore groundfish trawl survey Gulf of Maine Atlantic cod indices (log transformed). 80% confidence ellipses are shown.

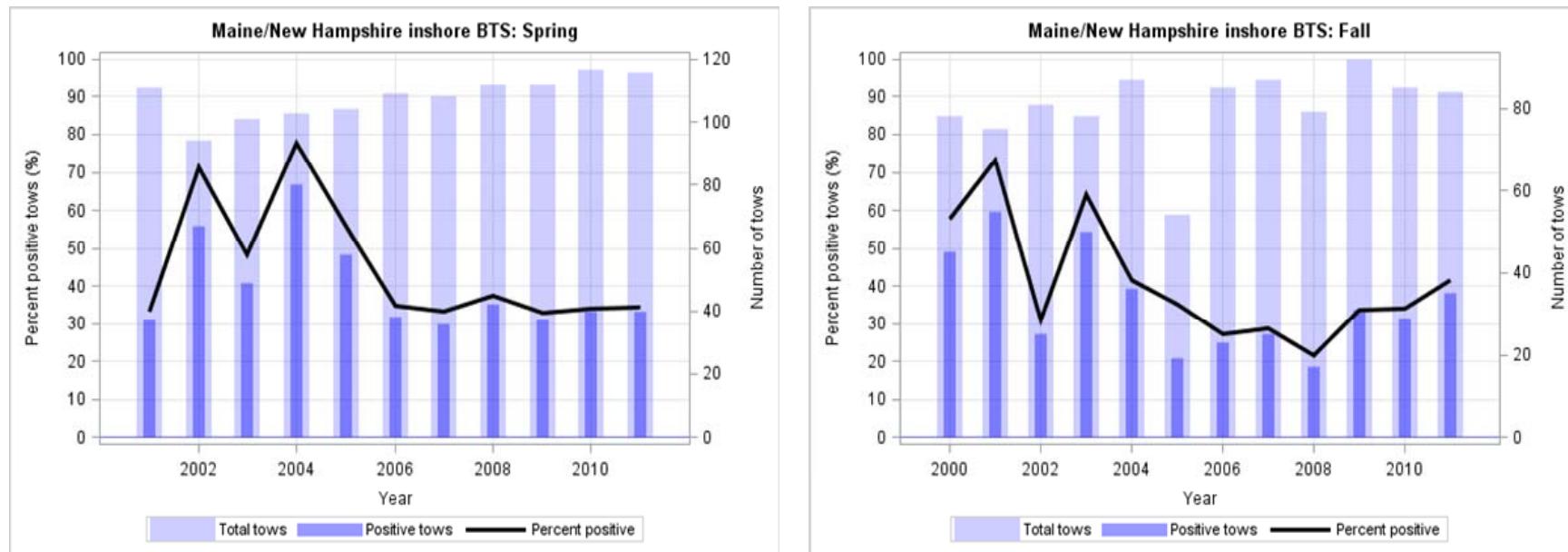


Figure A.121. Fraction of Maine – New Hampshire inshore groundfish trawl survey Tows with positive catches of Gulf of Maine Atlantic cod from 2000-2011.

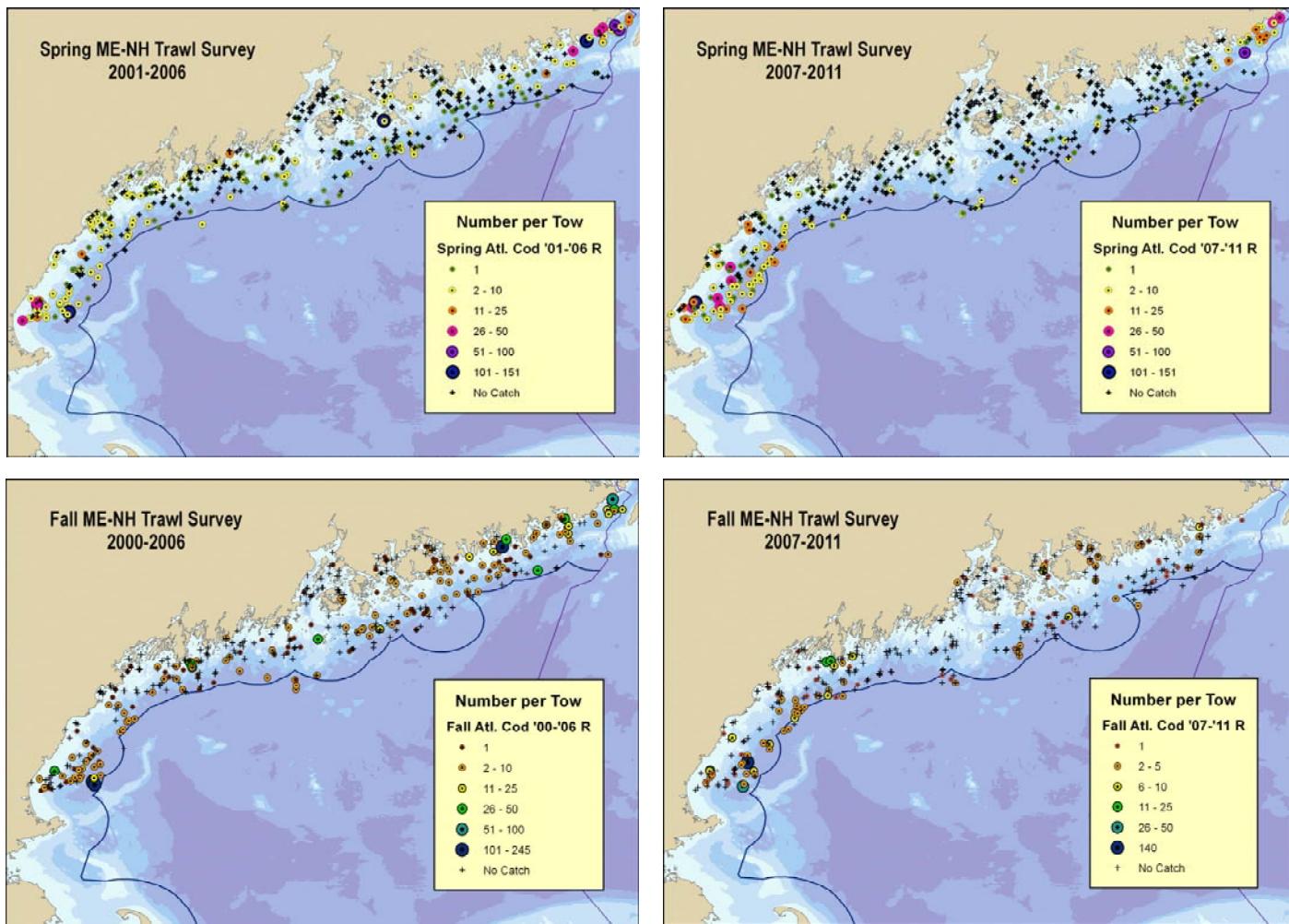


Figure A.122. Spatial distribution of Gulf of Maine Atlantic cod catches (numbers/tow) from the spring (top) and fall (bottom) Maine – New Hampshire inshore groundfish trawl survey from 2001-2006 (top) and 2007-2011 (left). Map provided by S. Sherman (pers. comm.).

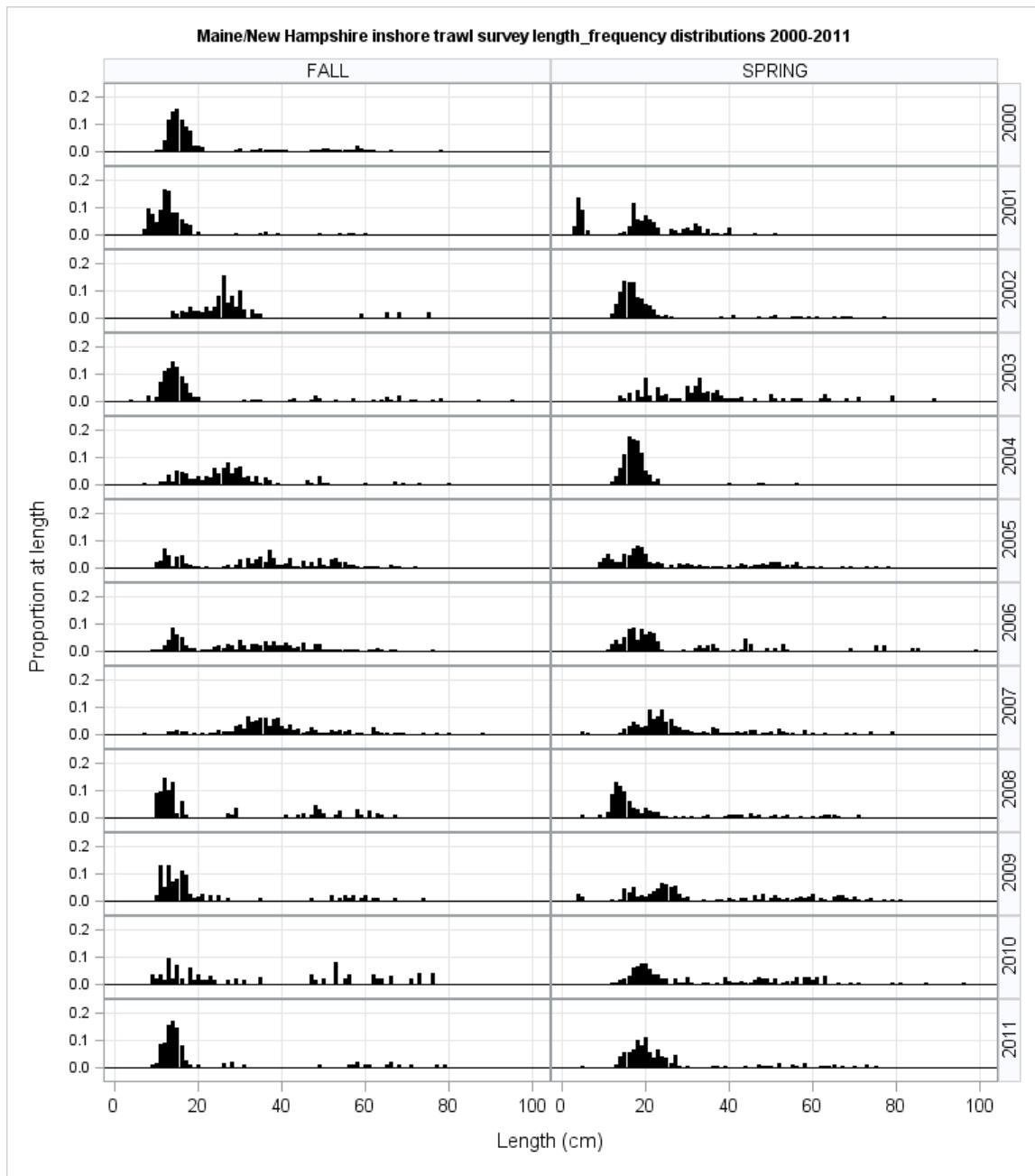


Figure A.123. Length distributions of Gulf of Maine Atlantic cod sampled in the Maine – New Hampshire inshore groundfish trawl spring (top) and fall (bottom) surveys from 2006 to 2009.

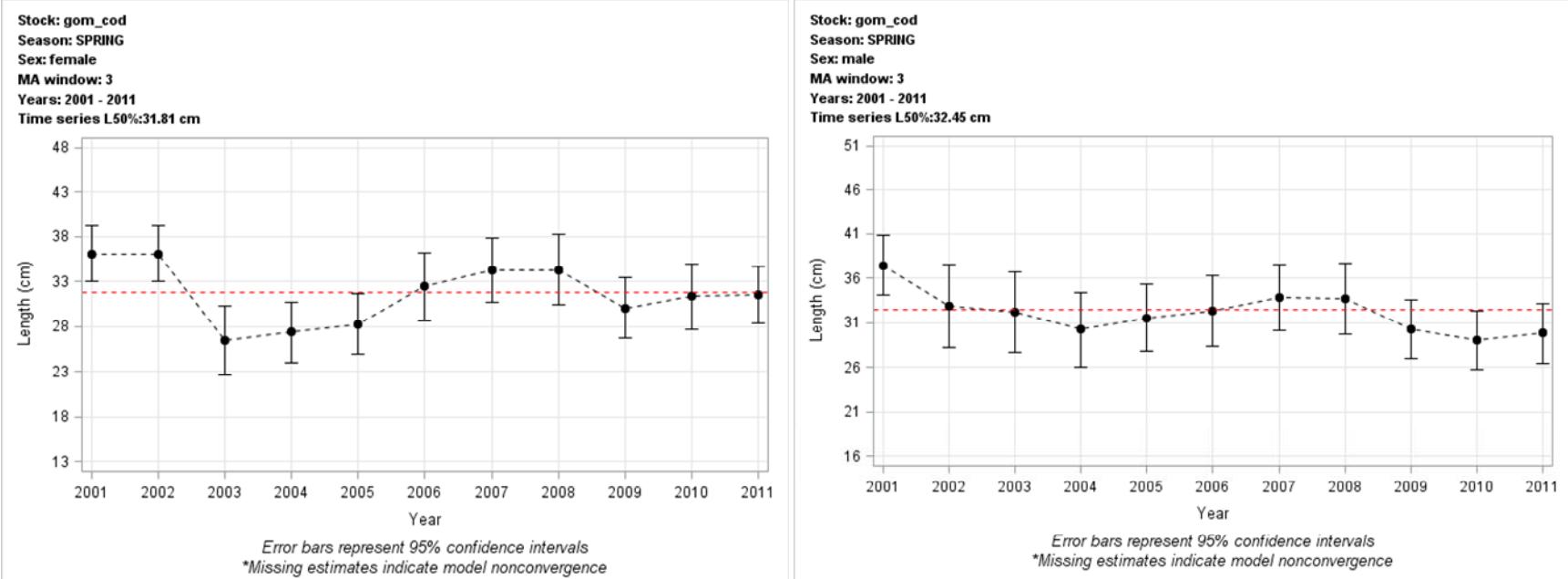
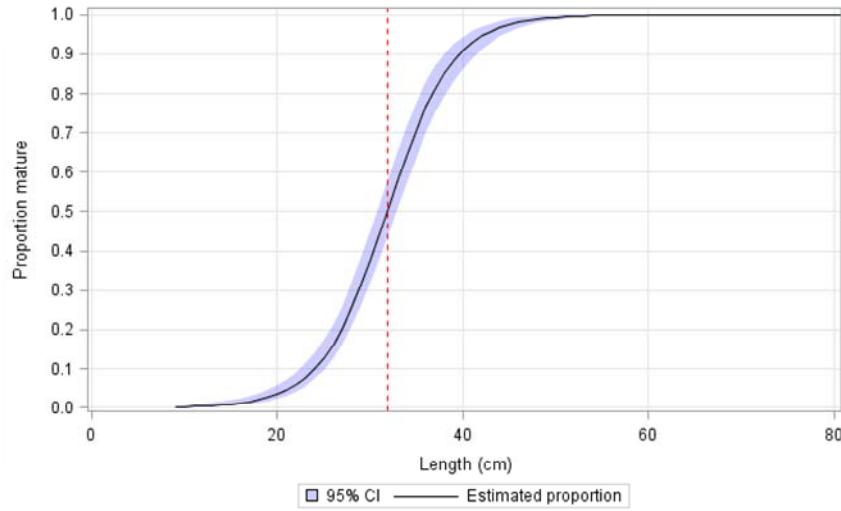


Figure A.124. Annual average length-at-50% maturity (L50) and corresponding 95% confidence intervals for female (left panels) and male (right panels) Gulf of Maine Atlantic cod from 2001 to 2011. Average maturity has been estimated from data collected from the Maine – New Hampshire spring inshore groundfish trawl survey. Years in which maturity ogives could not be estimated are omitted from the plots.

Stock: gom_cod
Season: SPRING
Sex: female
Maturity ogive: 2001 - 2011
Time series L50%: 31.8 cm



Stock: gom_cod
Season: SPRING
Sex: male
Maturity ogive: 2001 - 2011
Time series L50%: 32.5 cm

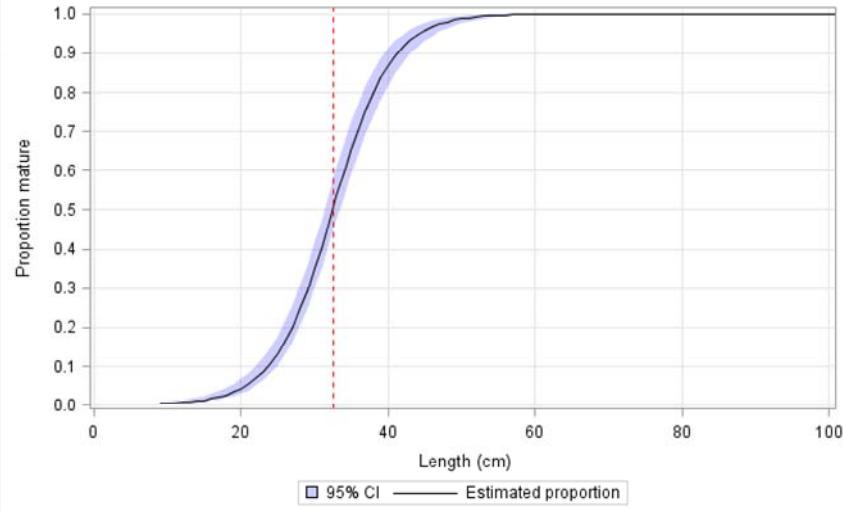


Figure A.125. Annual Length-based maturity ogives for female (left) and male (right) Gulf of Maine Atlantic cod based on time series averages of maturity and length information collected from the Maine – New Hampshire spring inshore groundfish trawl survey between 2001 and 2011. The dashed red line indicates the length at 50% maturity.

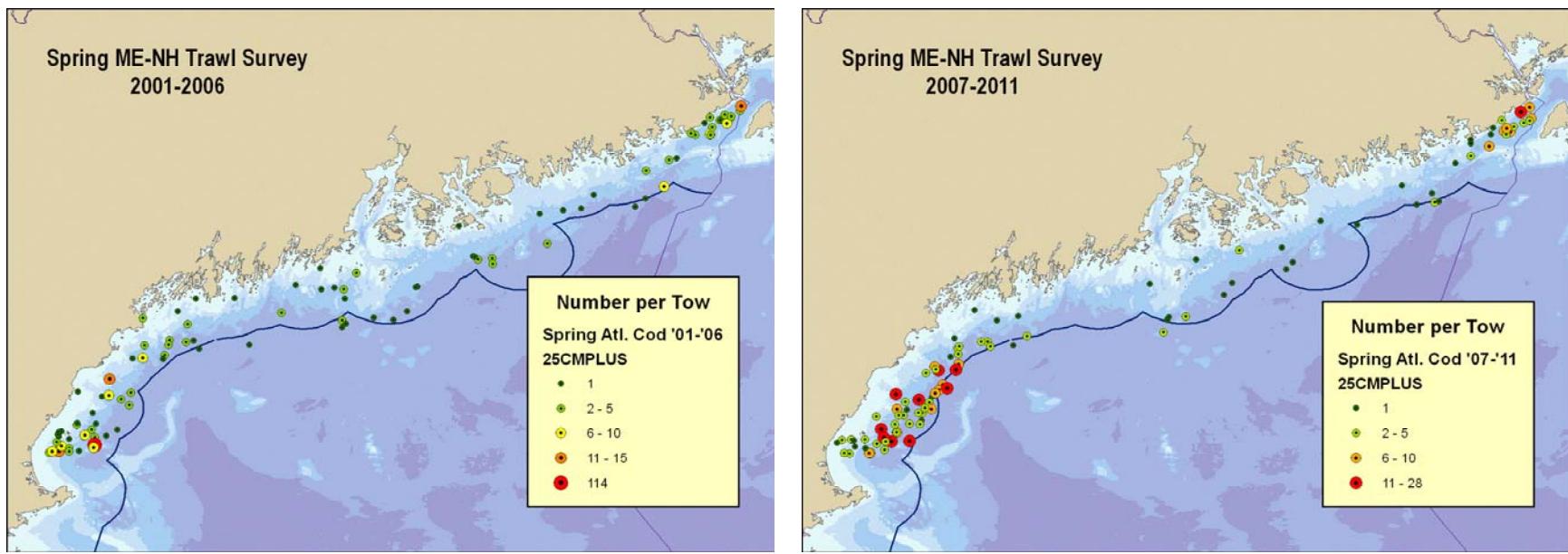


Figure A.126. Distribution of fish ≥ 25 cm from the Maine – New Hampshire spring inshore groundfish trawl survey from 2001-2006 (left) and 2007-2011 (right).

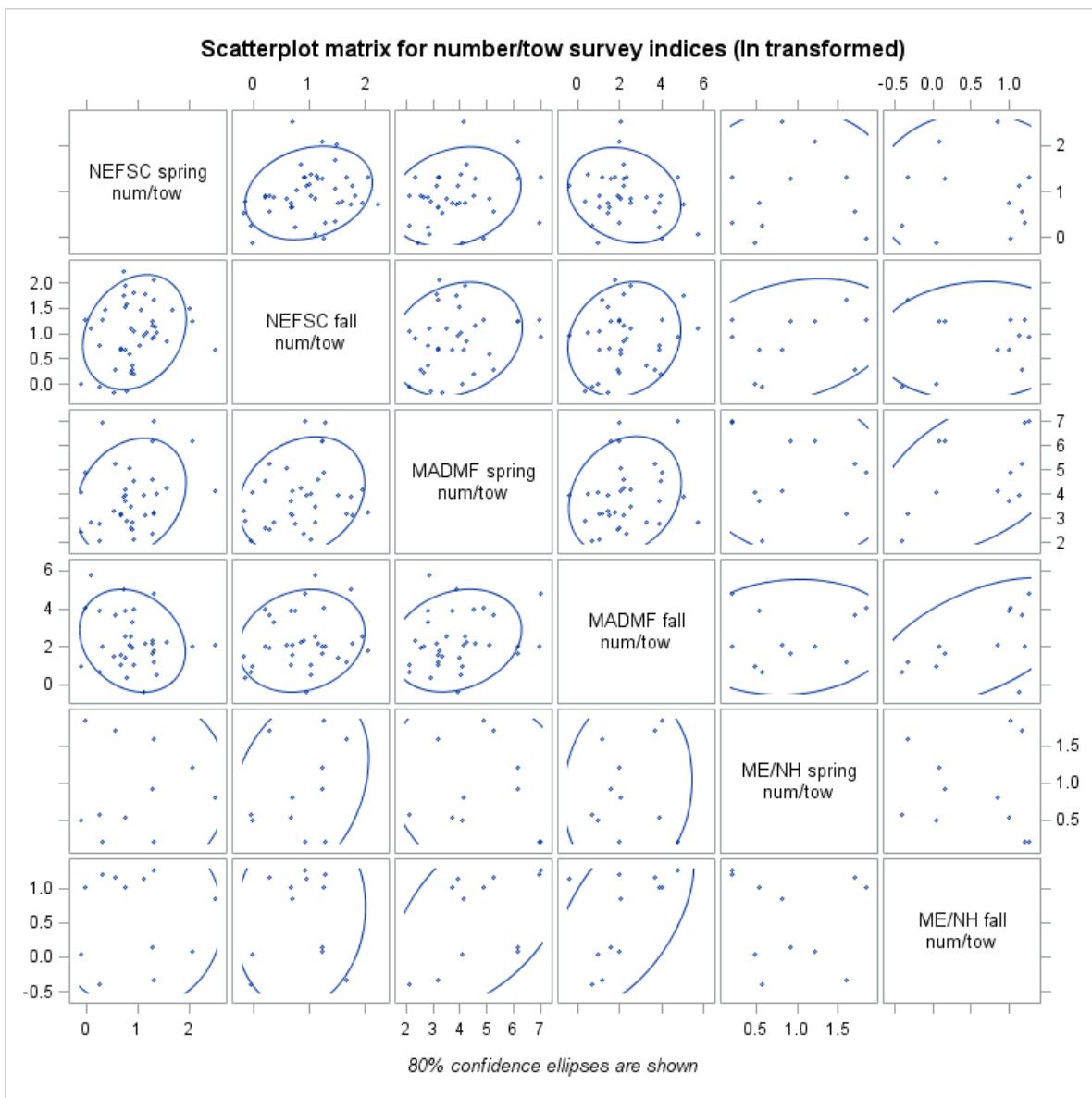


Figure A.127. Scatter plots showing the level of agreement between the Northeast Fisheries Science Center (NEFSC), Massachusetts Department of Marine Fisheries (MADMF) and the Maine – New Hampshire (ME/NH) inshore groundfish trawl survey Gulf of Maine Atlantic cod abundance (numbers/tow) indices (log transformed). 80% confidence ellipses are shown.

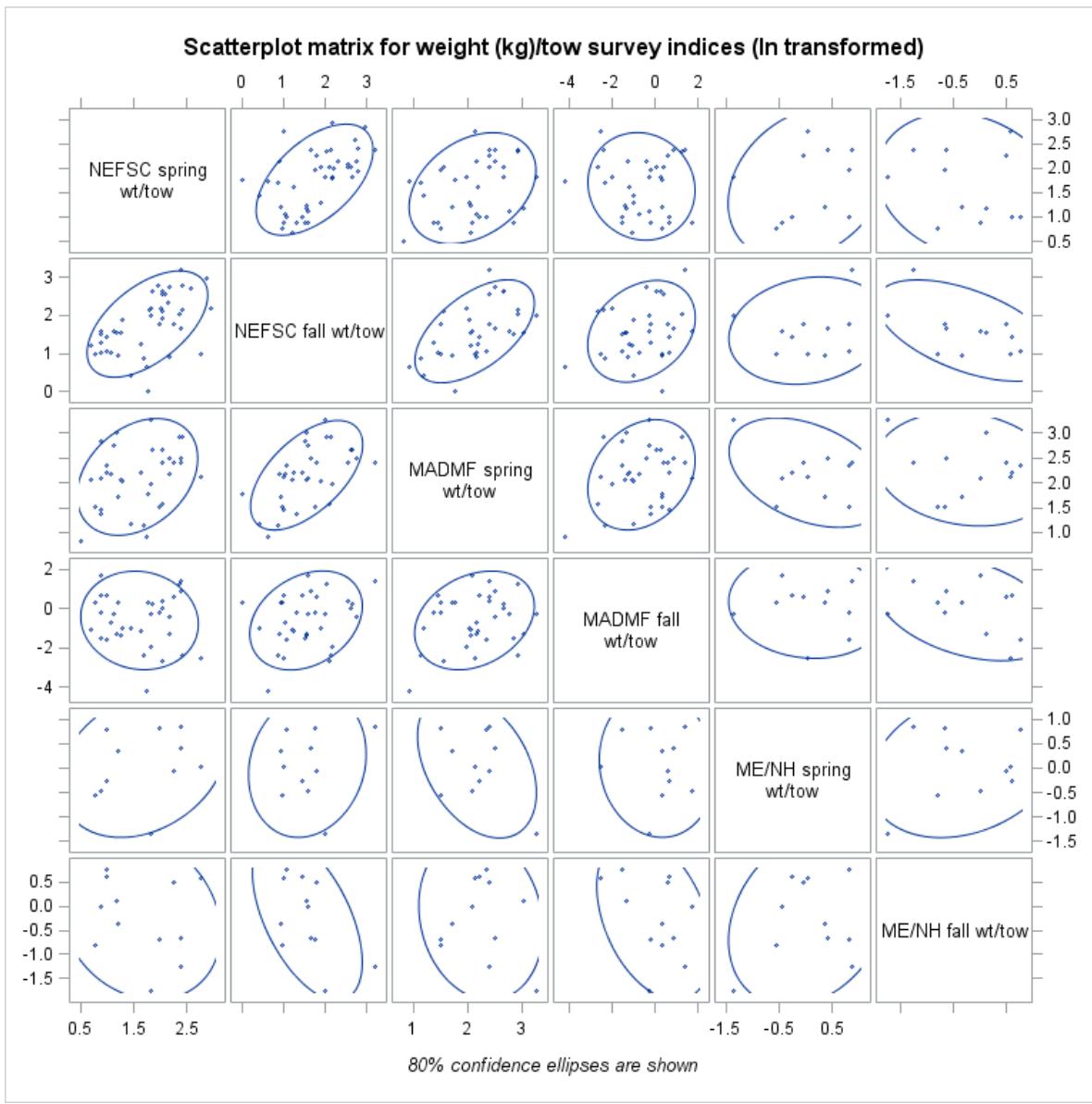


Figure A.128. Scatter plots showing the level of agreement between the Northeast Fisheries Science Center (NEFSC), Massachusetts Department of Marine Fisheries (MADMF) and the Maine – New Hampshire (ME/NH) inshore groundfish trawl survey Gulf of Maine Atlantic cod biomass (weight/tow) indices (log transformed). 80% confidence ellipses are shown.

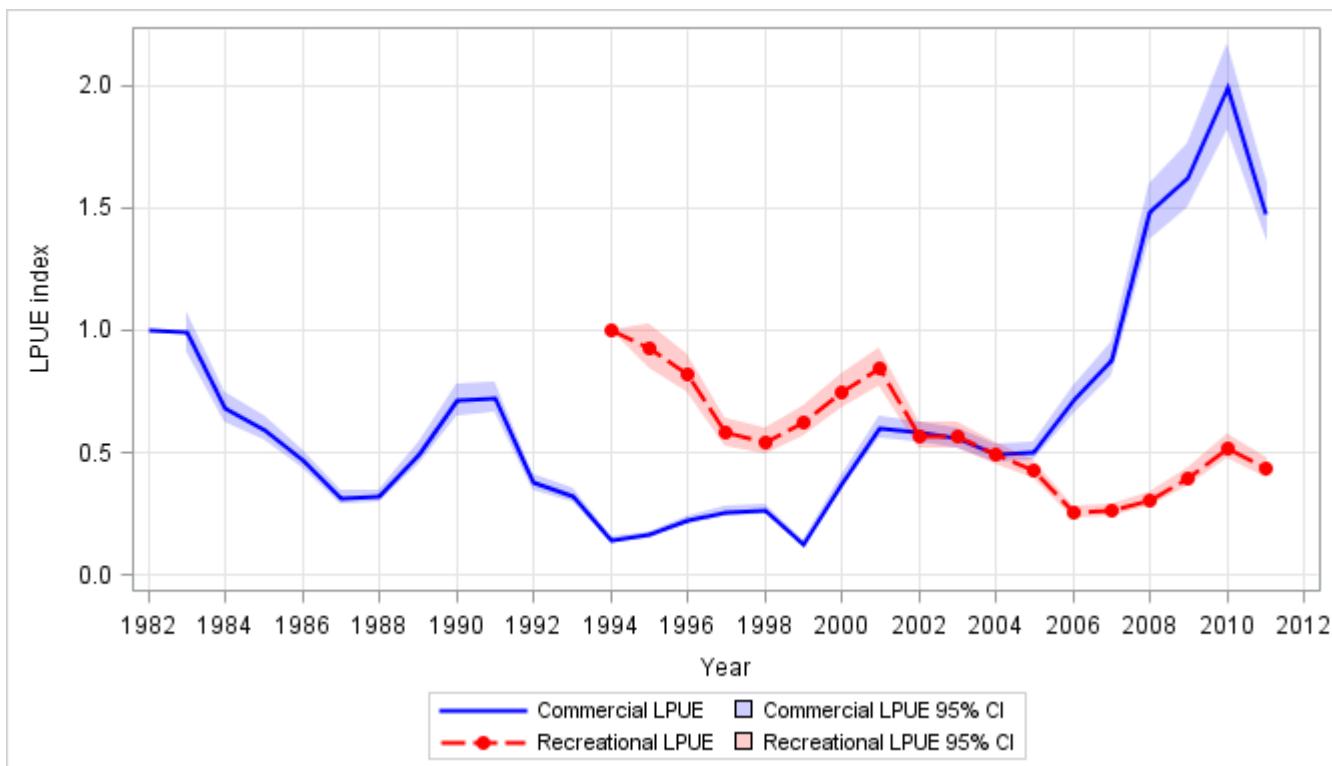


Figure A.129. Commercial otter trawl and recreational landings per unit effort (LPUE) indices for Gulf of Maine Atlantic cod. The development of the commercial otter trawl LPUE index is described in Palmer (2012b). The development of the recreational LPUE index is described in Wood (2012).

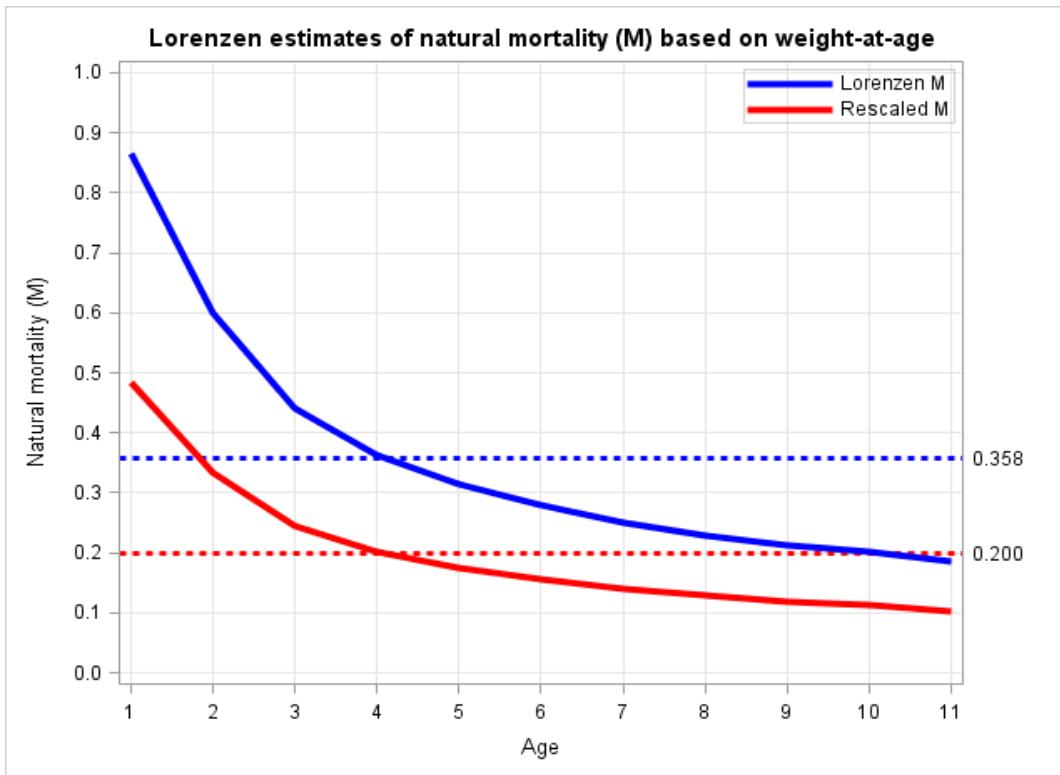


Figure A.130. Example of Lorenzen (1996) based estimates of natural mortality (M) at age based on time series average of stock weights at age. The blue line indicates the unadjusted Lorenzen estimate of natural mortality. The red line has been rescaled based on a constant M assumption of 0.2.

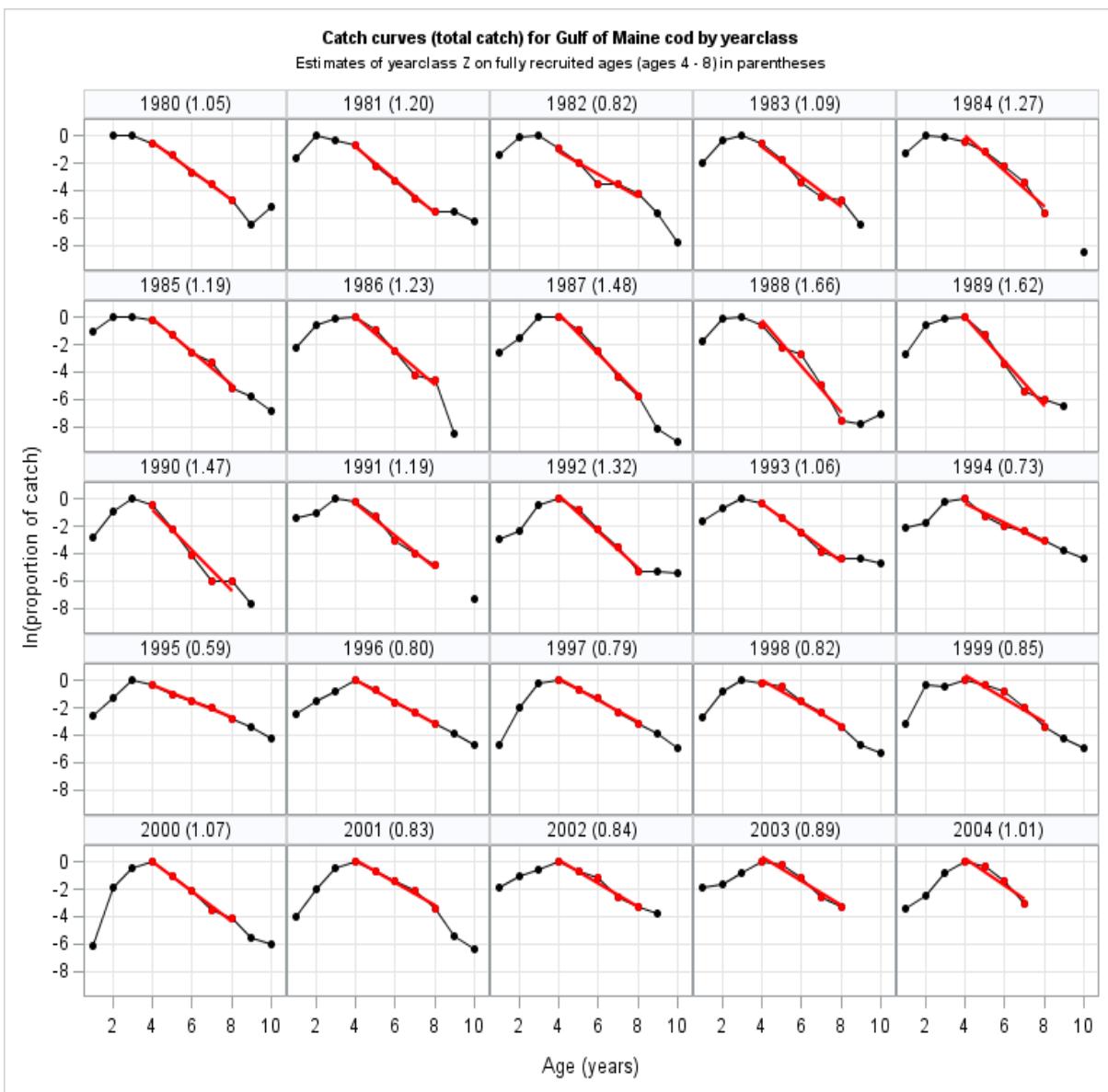


Figure A.131. Gulf of Maine Atlantic cod year class curves computed on ages 4-8 (red circles) log-transformed catch (commercial and recreational landings and discards). The corresponding slope of each regression line is shown next to the year class label above each plot.

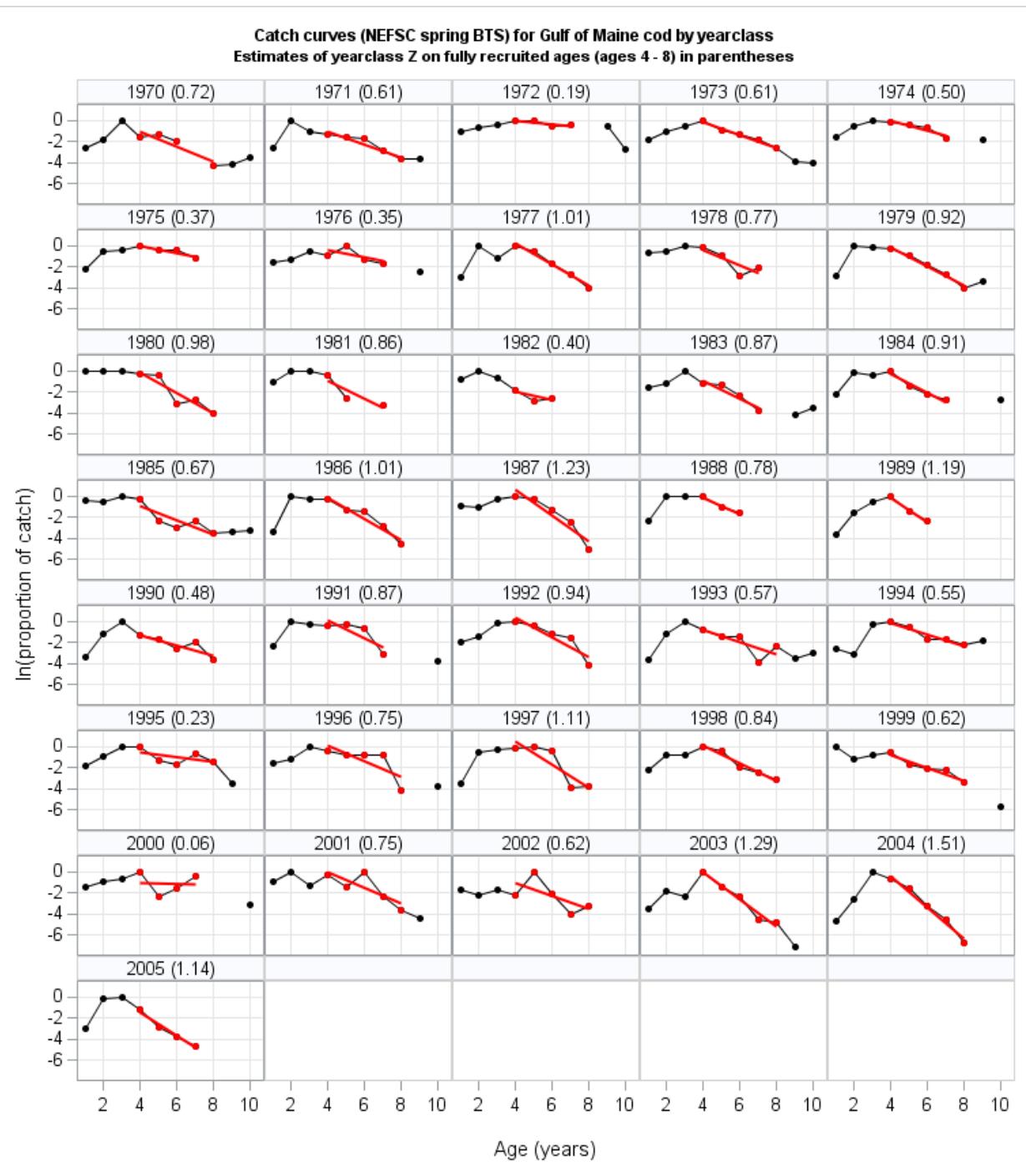


Figure A.132. Gulf of Maine Atlantic cod year class curves computed on ages 4-8 (red circles) log-transformed Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey abundance (numbers/tow) indices. The corresponding slope of each regression line is shown next to the year class label above each plot.

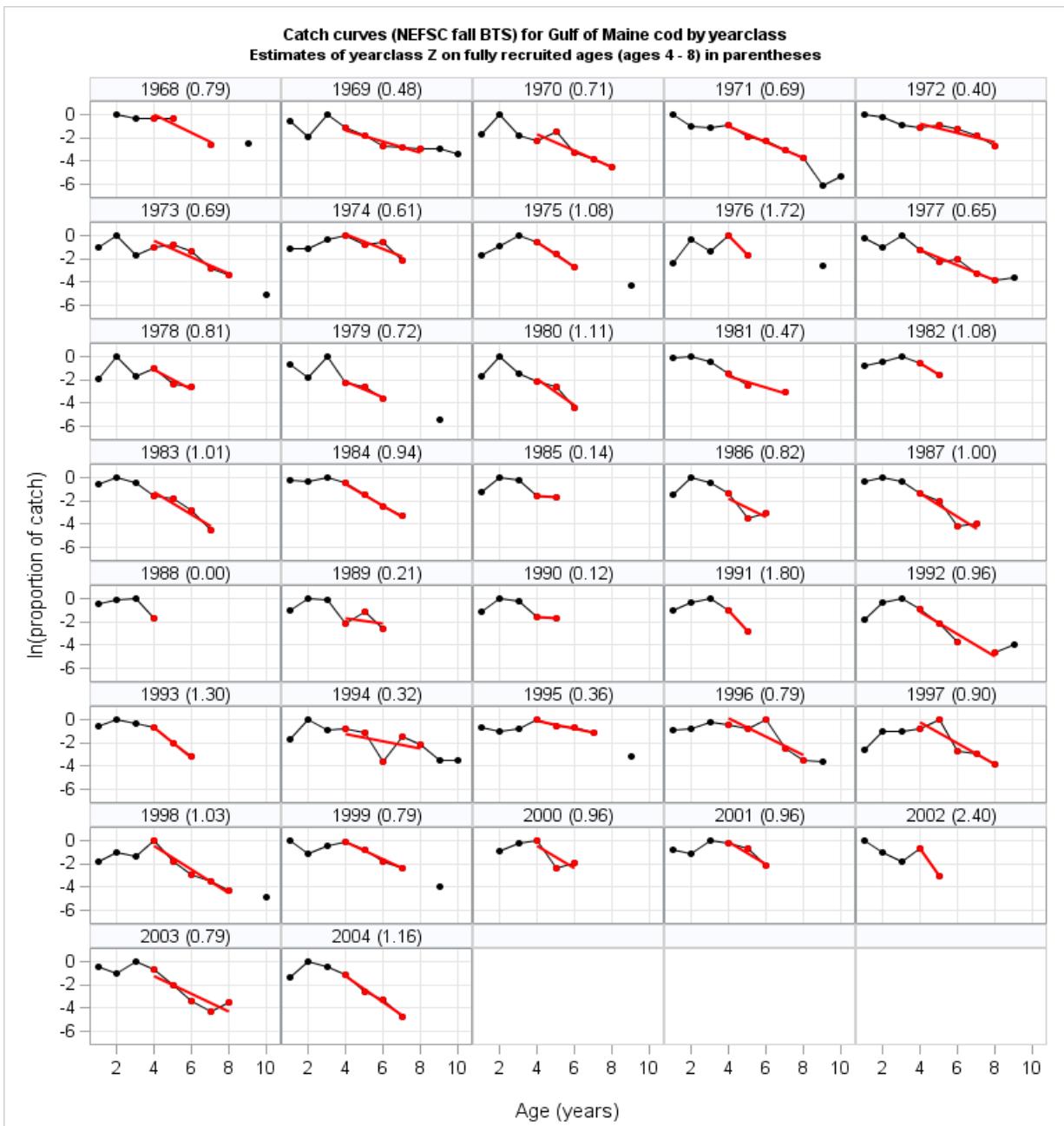


Figure A.133. Gulf of Maine Atlantic cod year class curves computed on ages 4-8 (red circles) log-transformed Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey abundance (numbers/tow) indices. The corresponding slope of each regression line is shown next to the year class label above each plot.

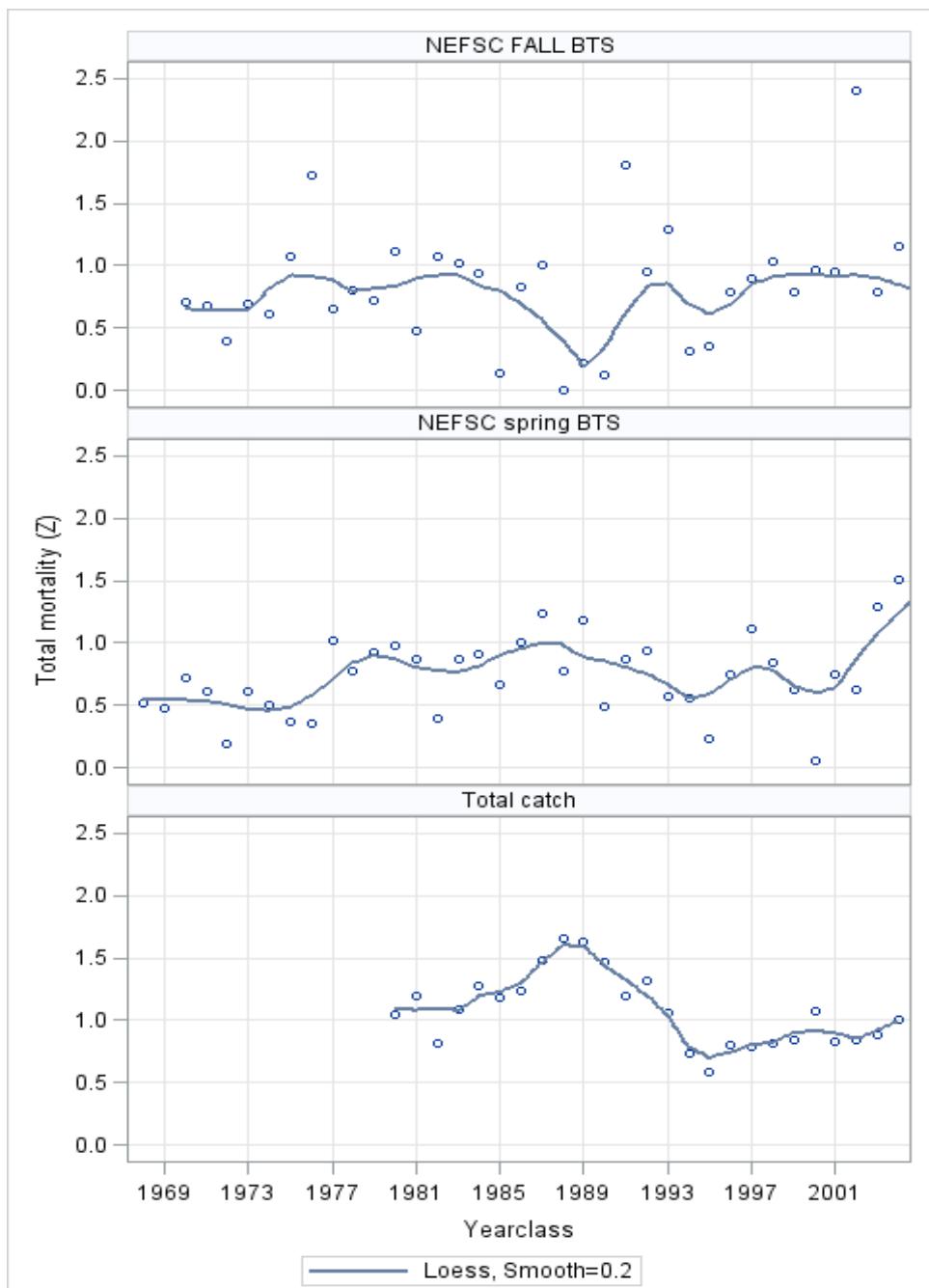


Figure A.134. Plots of the annual estimates of Gulf of Maine Atlantic cod total mortality (Z) as estimated from the year class curve analyses for total catch and Northeast Fisheries Science Center (NEFSC) spring and fall bottom trawl surveys.

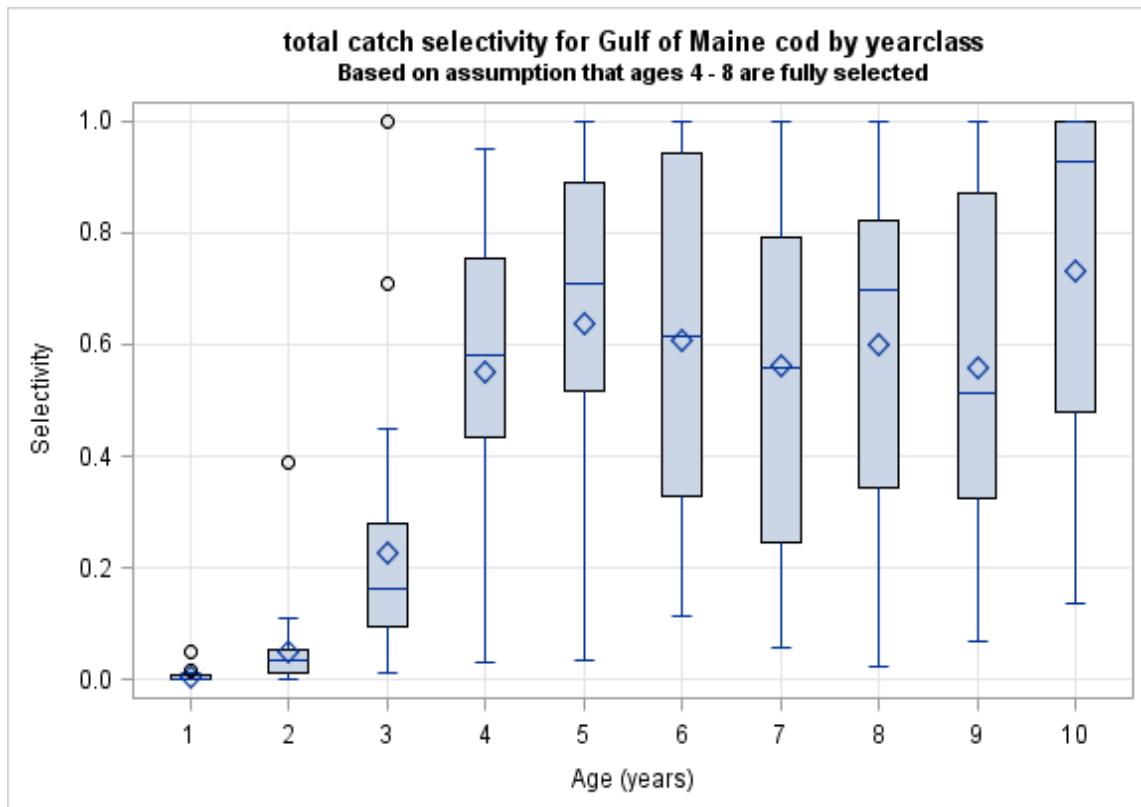


Figure A.135. Box plot distribution of the residuals fits to the Gulf of Maine Atlantic cod year class linear regression relationship by age from the total catch year class curve analysis.

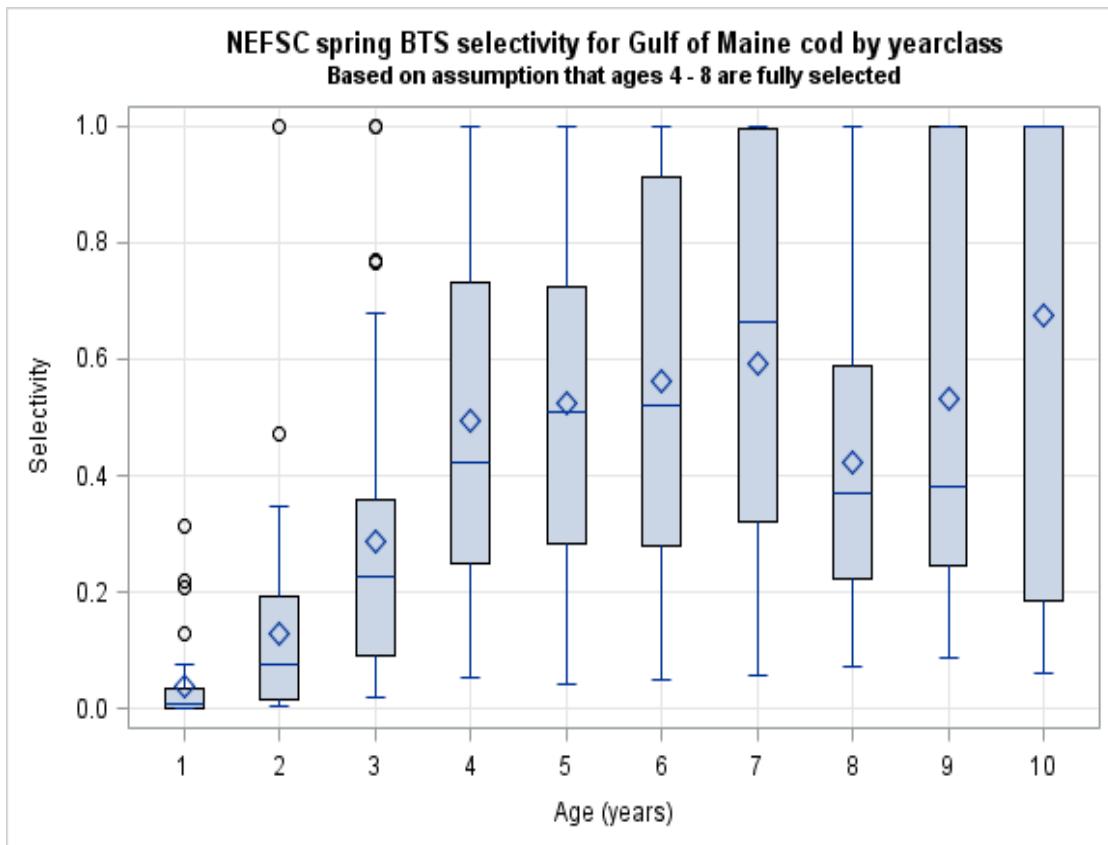


Figure A.136. Box plot distribution of the residuals fits to the Gulf of Maine Atlantic cod year class linear regression relationship by age from the Northeast Fisheries Science Center (NEFSC) spring bottom trawl survey year class curve analysis.

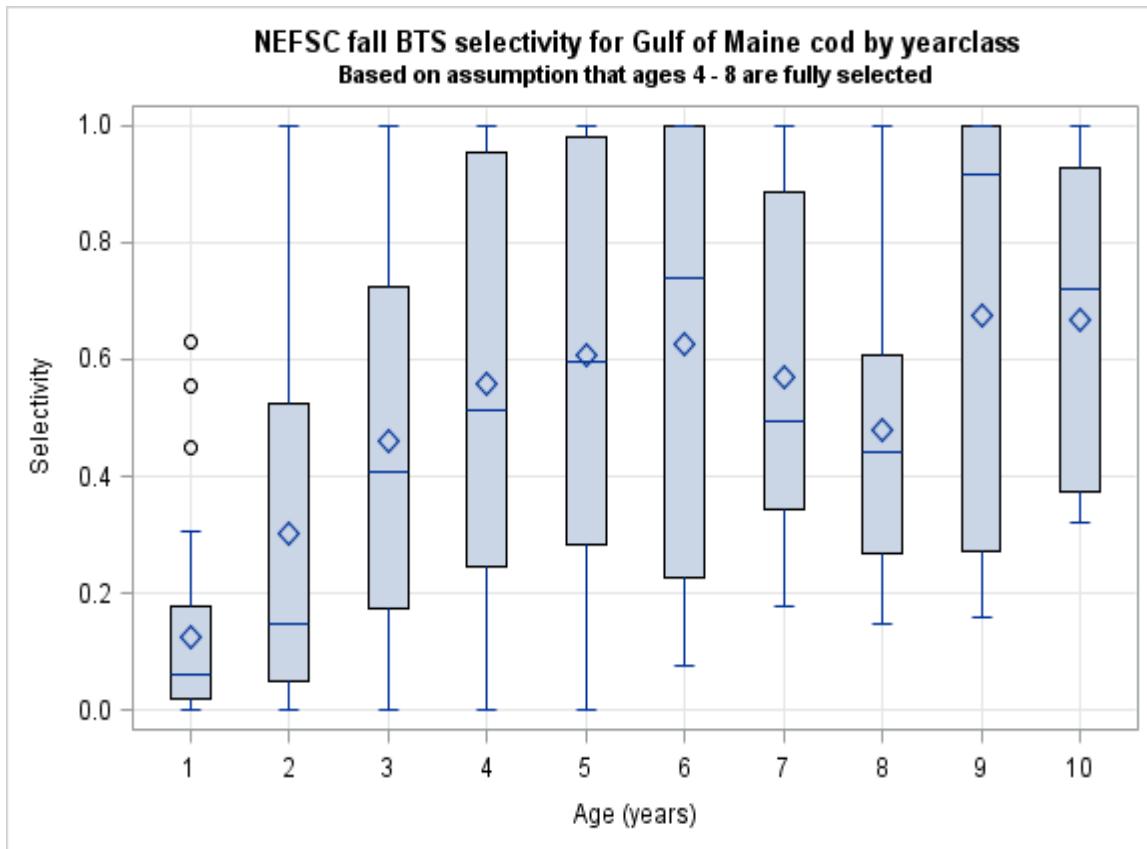


Figure A.137. Box plot distribution of the residuals fits to the Gulf of Maine Atlantic cod year class linear regression relationship by age from the Northeast Fisheries Science Center (NEFSC) fall bottom trawl survey year class curve analysis.

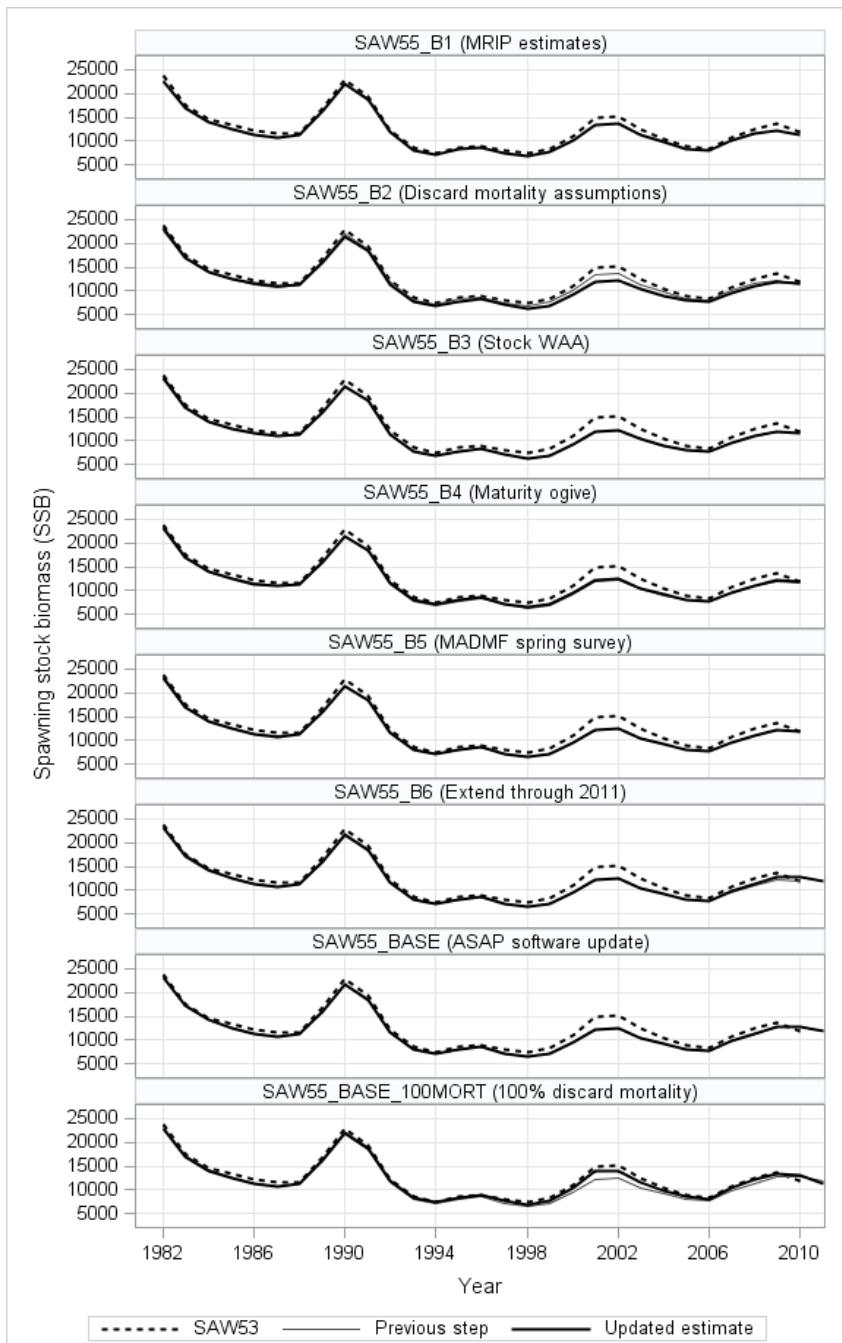


Figure A.138. Summary of the impacts on the time series of spawning stock biomass resulting from the update of the SAW 53 Gulf of Maine cod ASAP model with new data. In each plot, the SAWQ 53 model results are shown by a dashed line with the model results based on the updated data input shown by a solid black line. The solid grey line indicates the model results from the previous step such that the impacts can be understood not only compared to the SAW 53 model, but also to the previous step.

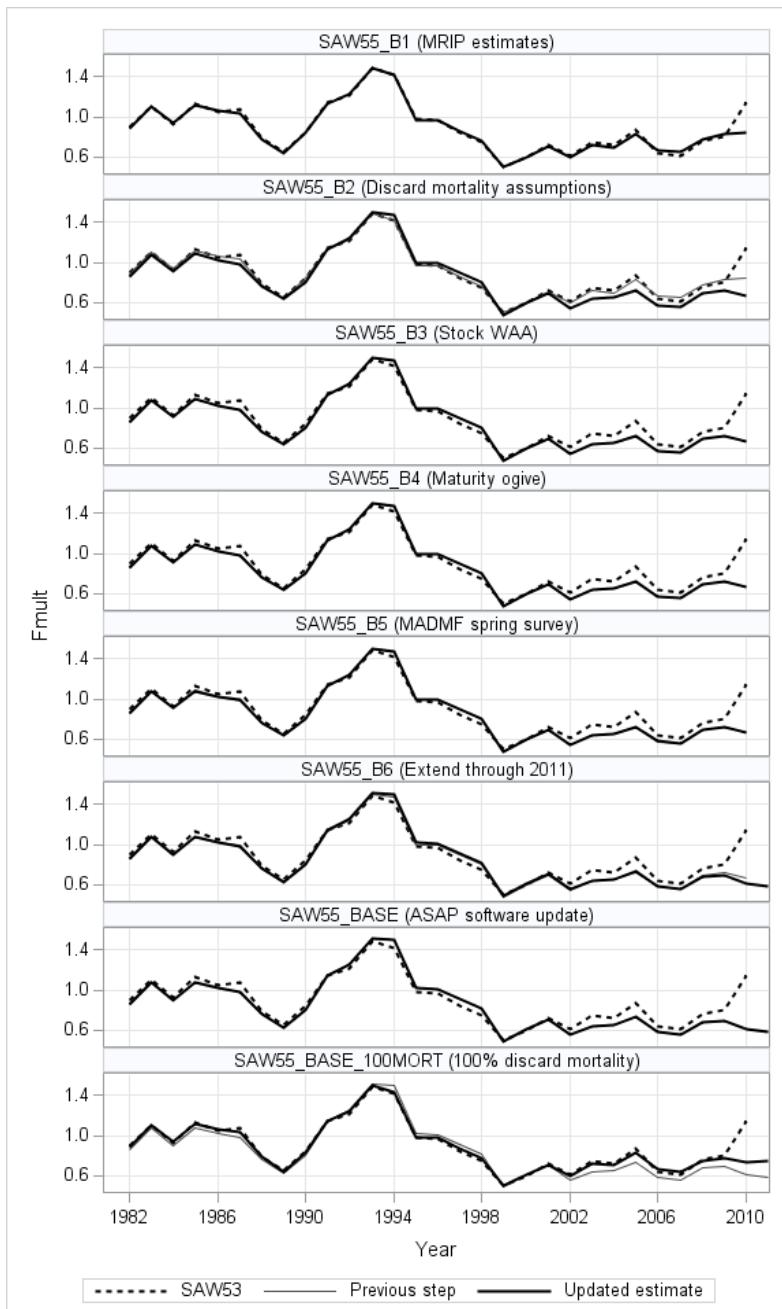


Figure A.139. Summary of the impacts on the time series of fishing mortality (age 5) resulting from the update of the SAW 53 Gulf of Maine Atlantic cod ASAP model with new data. In each plot, the SAWQ 53 model results are shown by a dashed line with the model results based on the updated data input shown by a solid black line. The solid grey line indicates the model results from the previous step such that the impacts can be understood not only compared to the SAW 53 model, but also to the previous step.

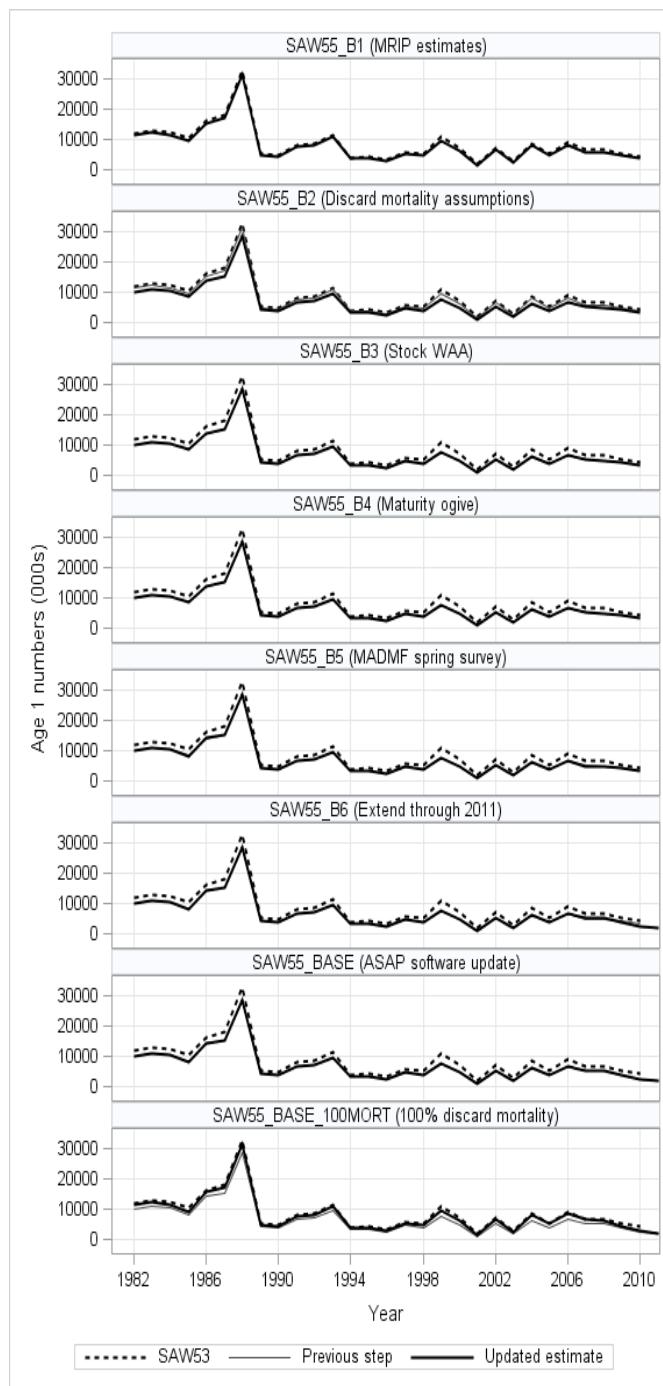


Figure A.140. Summary of the impacts on the time series of age 1 recruitment resulting from the update of the SAW 53 Gulf of Maine Atlantic cod ASAP model with new data. In each plot, the SAWQ 53 model results are shown by a dashed line with the model results based on the updated data input shown by a solid black line. The solid grey line indicates the model results from the previous step such that the impacts can be understood not only compared to the SAW 53 model, but also to the previous step.

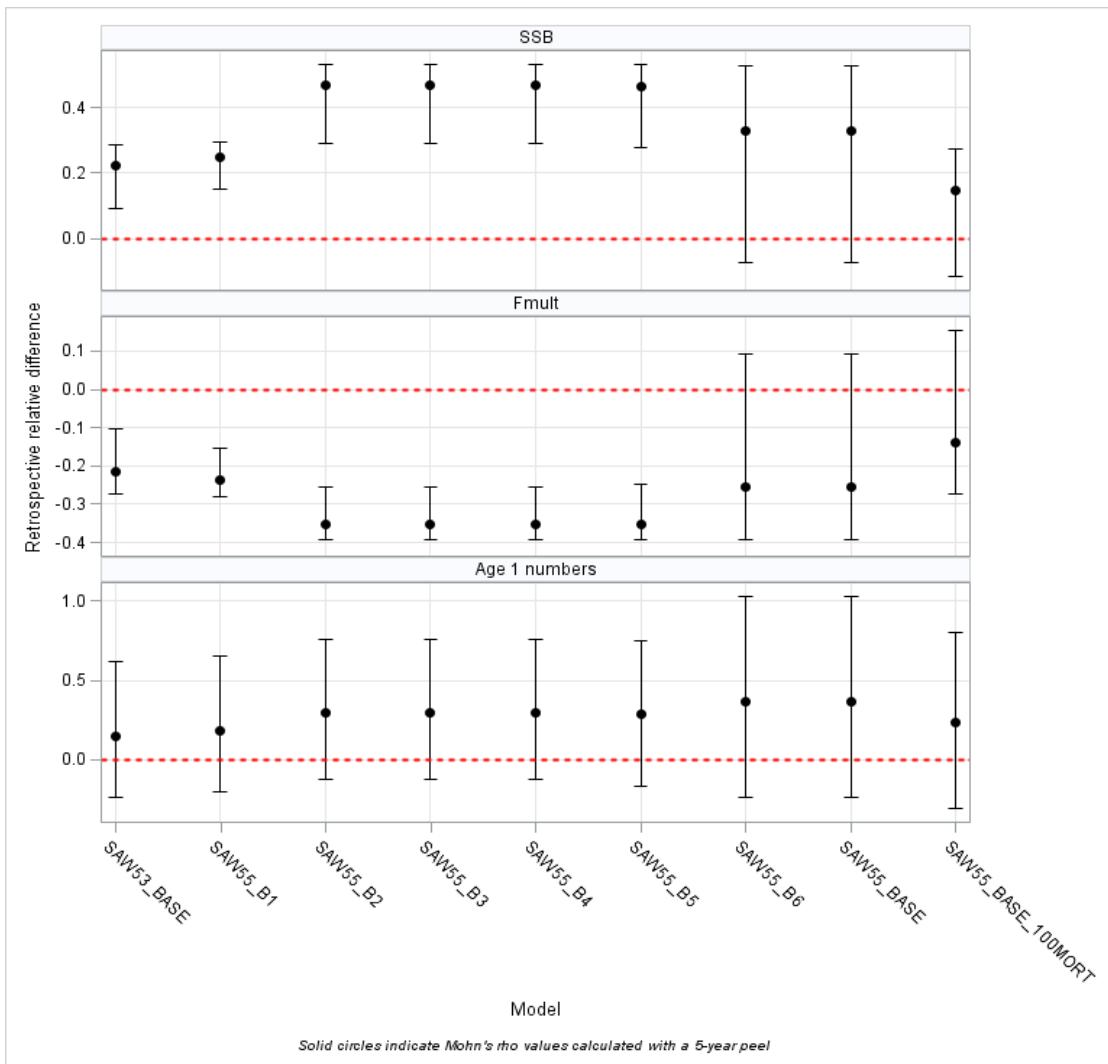


Figure A.141. Summary of the Mohn's rho values (dots) and minimum and maximum observed relative difference resulting from a five year retrospective peel for the eight model runs considered in the update of the SAW 53 Gulf of Maine cod assessment model.

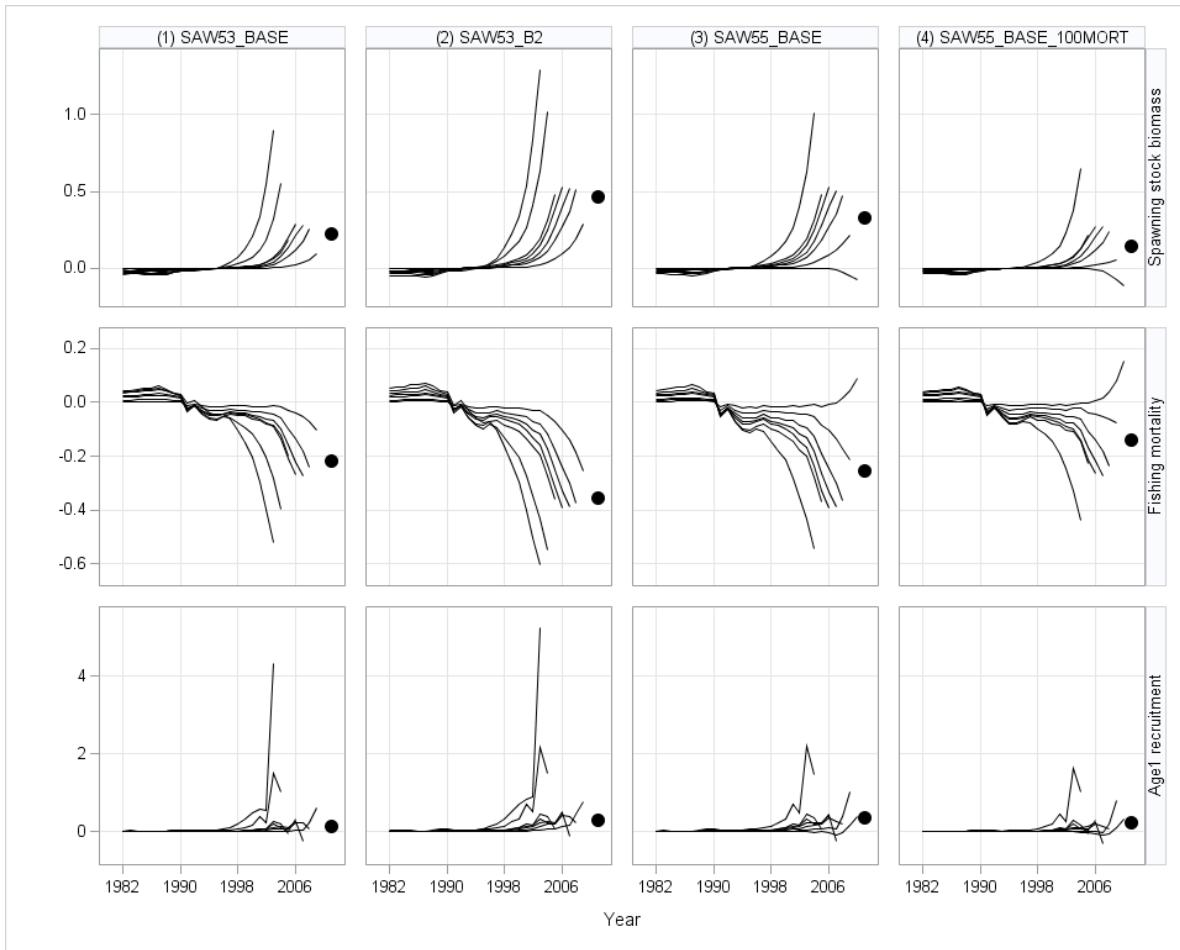


Figure A.142. ASAP BASE model retrospective patterns for the Gulf of Maine Atlantic cod SAW53 model (SAW53_BASE), SAW 53 model after application of revised discard mortality rates (SAW53_B2), updated SAW 55 base model (SAW55_BASE) and the SAW 55 base model under an assumption of 100% discard mortality (SAW55_BASE_100MORT). The black circles indicate the Mohn's rho value based on a five year retrospective peel.

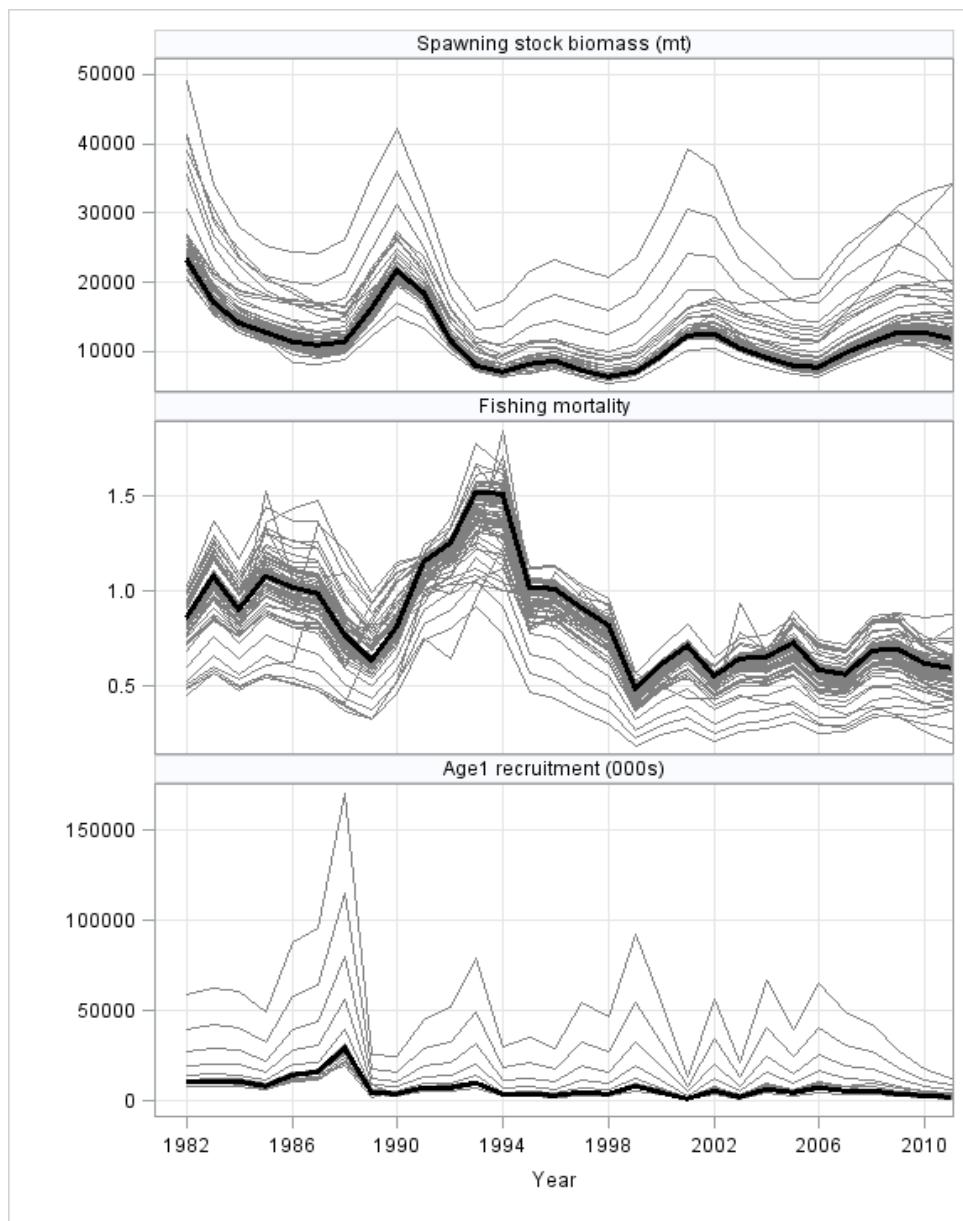


Figure A.143. Estimates of Gulf of Maine Atlantic cod spawning stock biomass (top), average age 5 fishing mortality (middle) and age 1 recruitment (bottom) from the 100-plus ASAP sensitivity runs. The results of the SAW55_BASE model are shown by a solid black line. A full description of the major sensitivity runs that were conducted can be found in Appendix A.6.

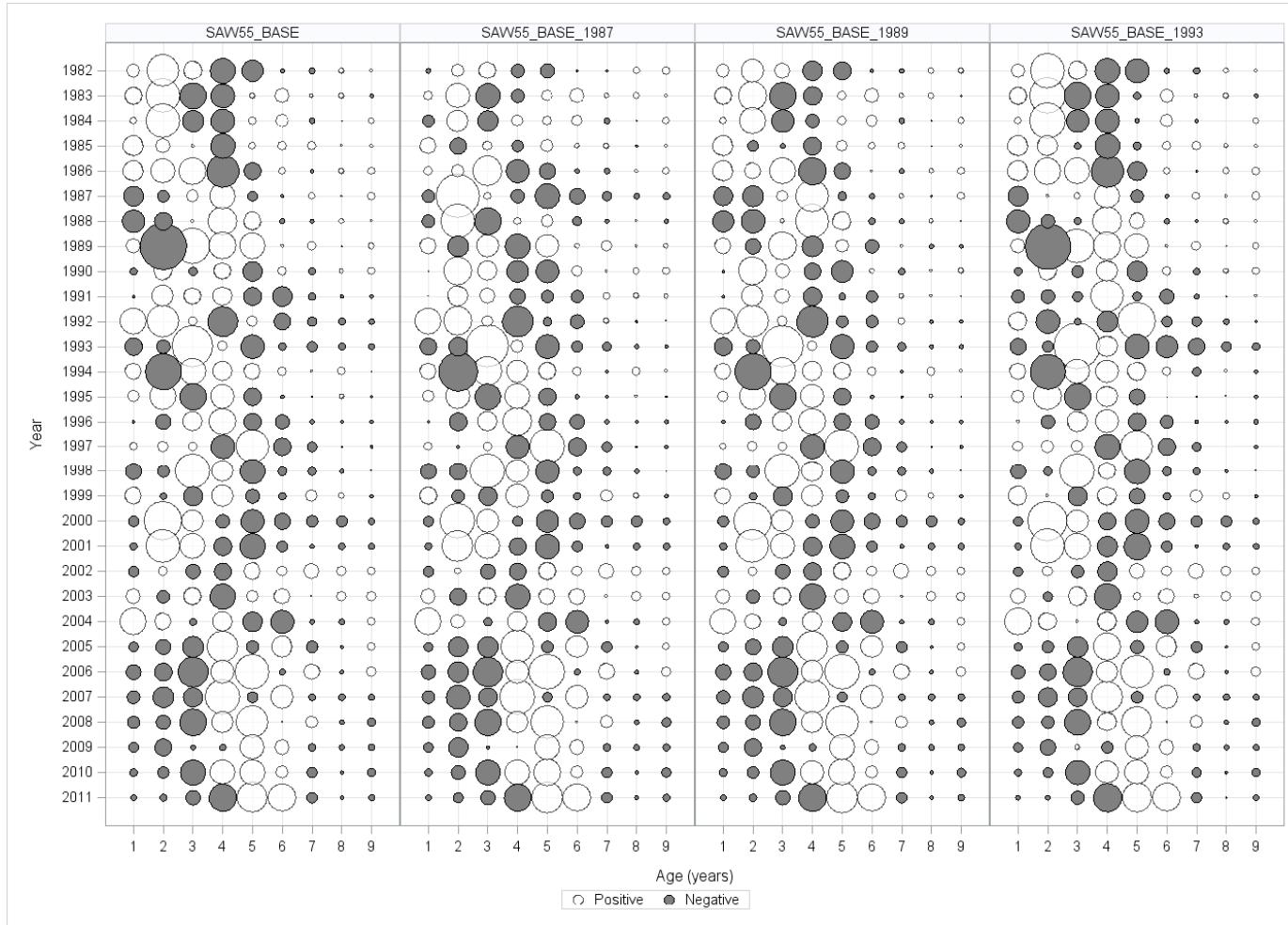


Figure A.144. ASAP model residuals for the fits to the fishery catch-at-age of Gulf of Maine Atlantic cod for four different 2-selectivity block models. The models vary by the transition year between blocks 1 and 2 (i.e., year in which block 2 begins). In all sensitivity runs the transition year is indicated by the model suffix; for the SAW55_BASE model block 2 begins in 1991.

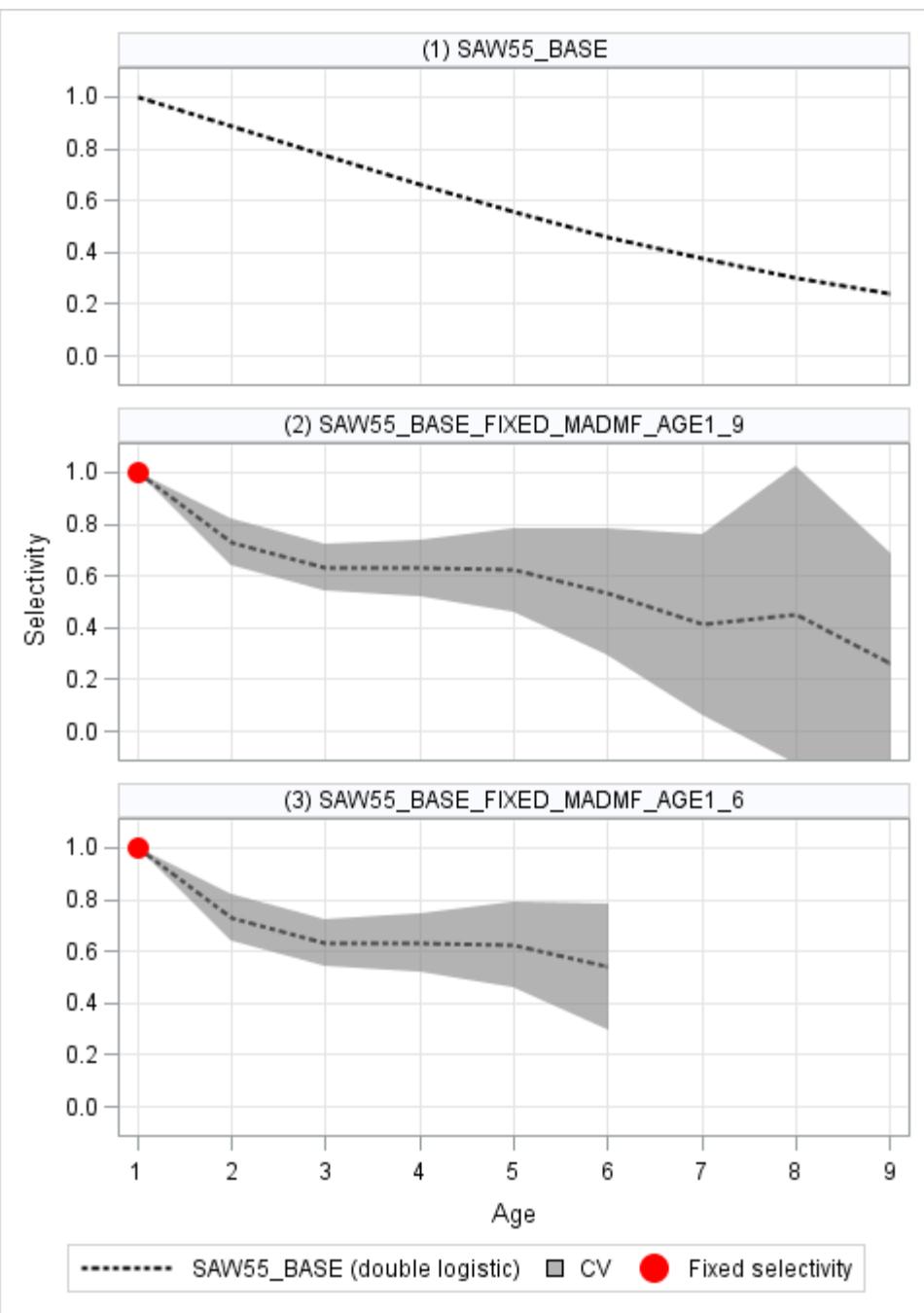


Figure A.145. Model estimated Gulf of Maine Atlantic cod selectivity at age for the Massachusetts Department of Marine Fisheries (MADMF) spring survey based on ASAP model explorations of fitting the survey using both parametric (double logistic, SAW55_BASE) and non-parametric (at-age, all other models) approaches.

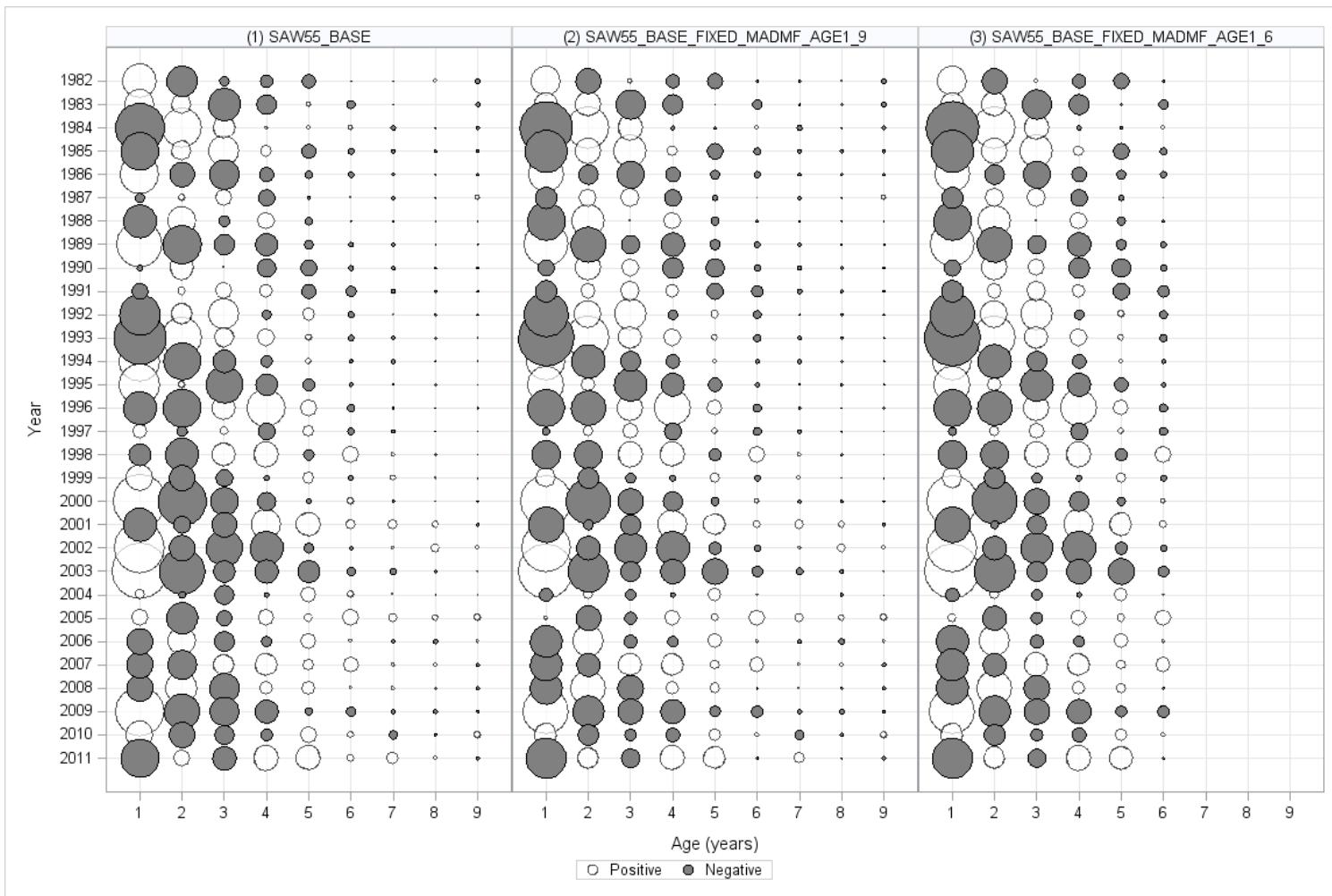


Figure A.146. Residual plots from the fitting of Massachusetts Department of Marine Fisheries (MADMF) spring survey Gulf of Maine Atlantic cod indices at age from ASAP model explorations using both parametric (double logistic, SAW55_BASE) and non-parametric (at-age, all other models) approaches.

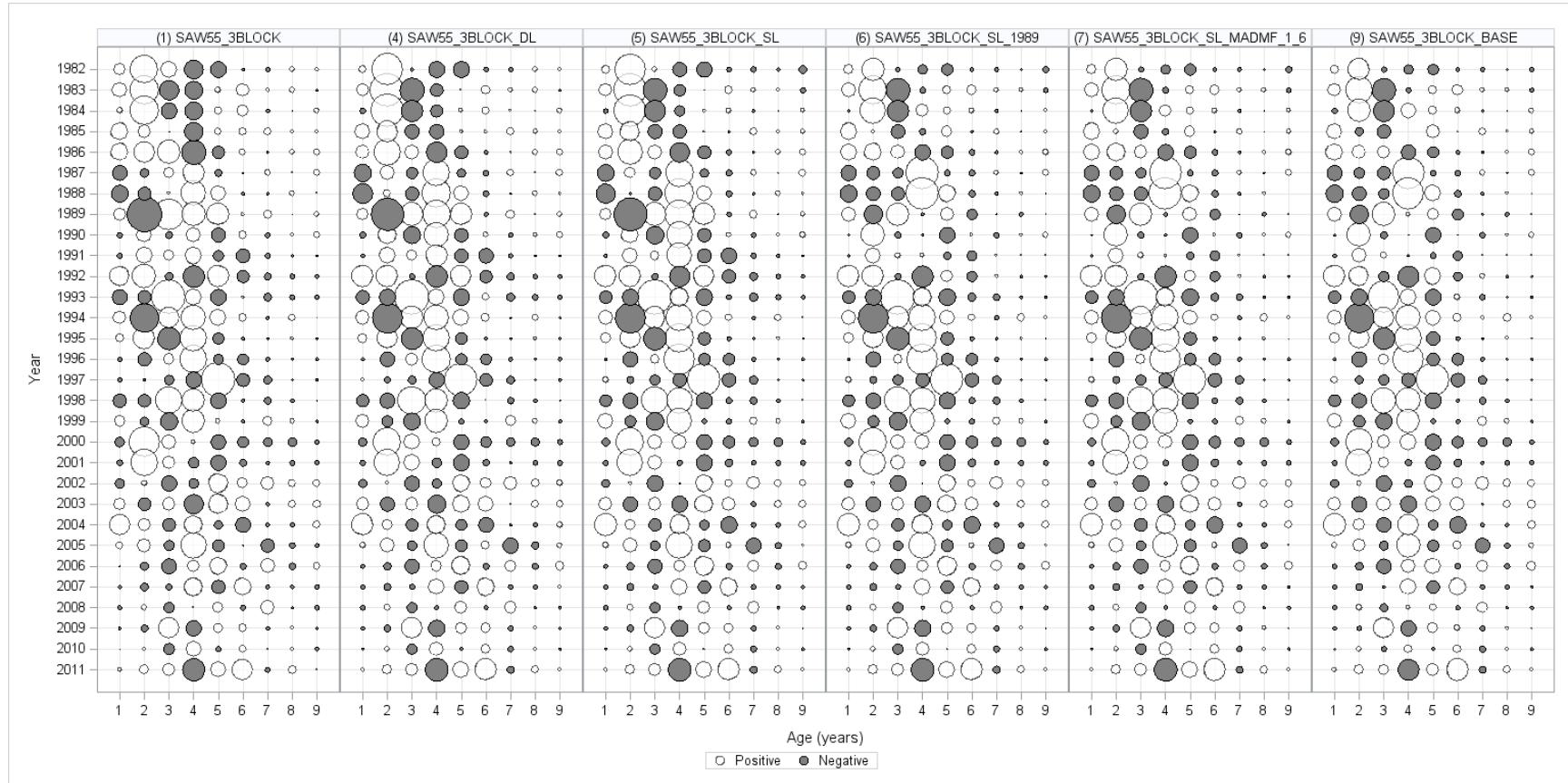


Figure A.147. ASAP model residuals for the fits to the fishery catch-at-age of Gulf of Maine Atlantic cod for six different 3-selectivity block models.

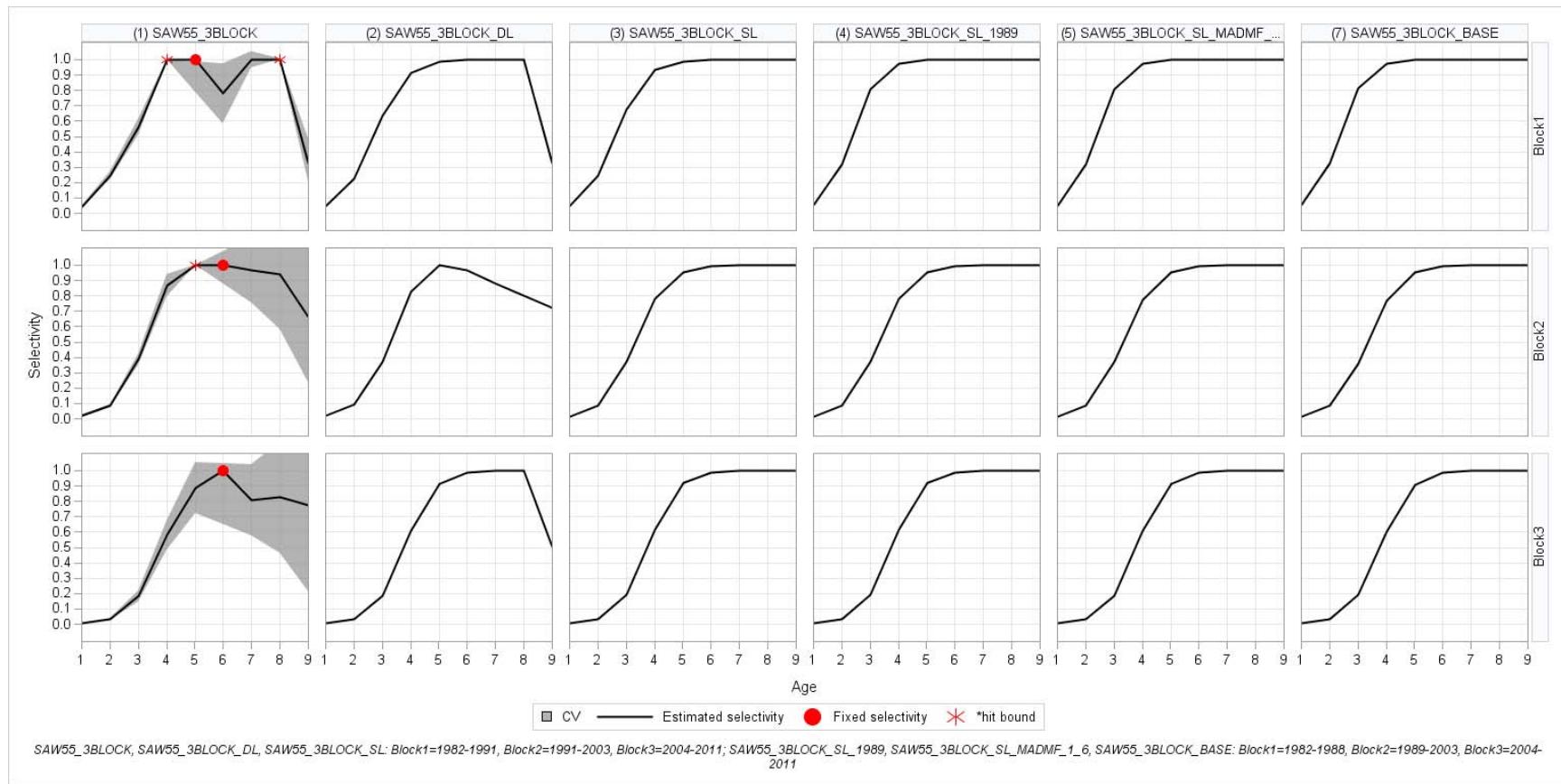


Figure A.148. ASAP estimated fishery selectivities for Gulf of Maine Atlantic cod from six different 3-selectivity block models.

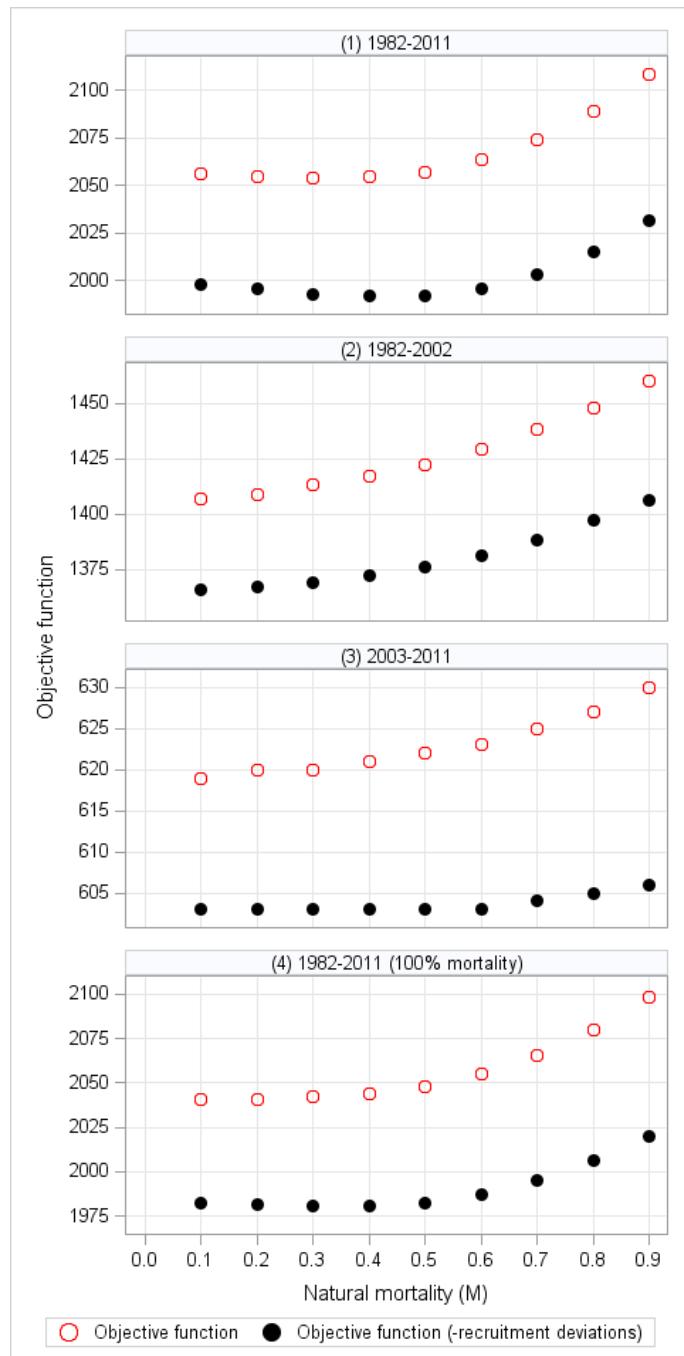


Figure A.149. Response of the model objective function to profiling over a range of Gulf of Maine cod natural mortality values. Four different models configurations are explored: (1) 1982-2011, (2) 1982-2002, (3) 2003-2011, and (4) 1982-2011 under assumption of 100% discard mortality.

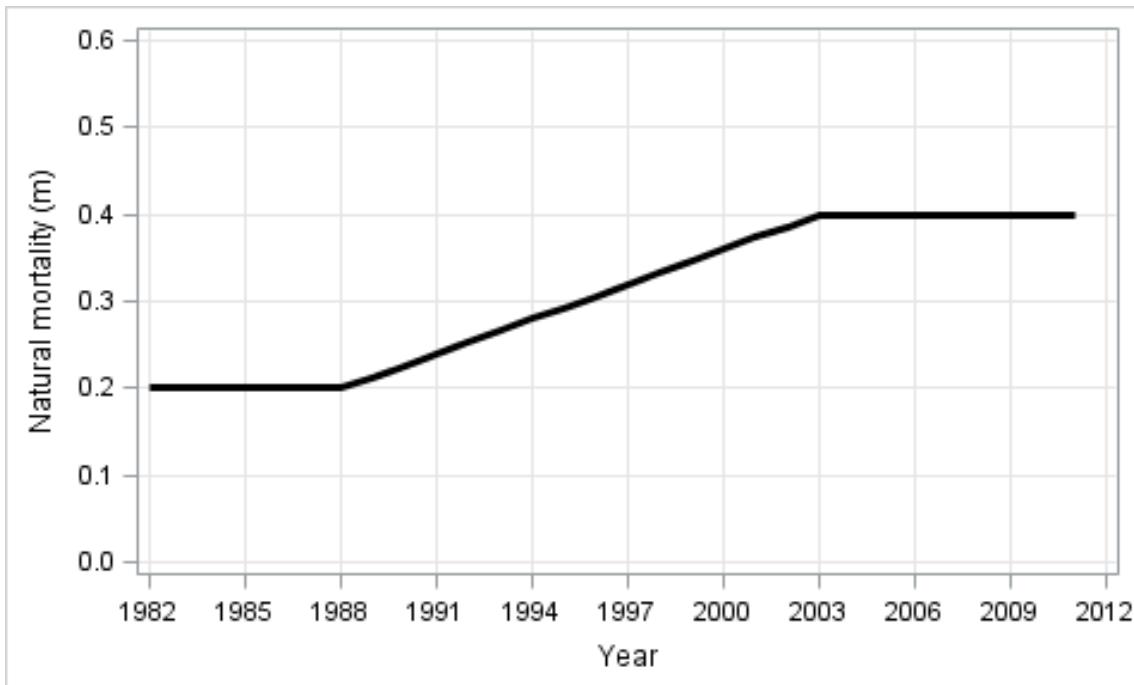


Figure A.150. Time series of natural mortality used in the Gulf of Maine cod natural mortality ramp assessment models.

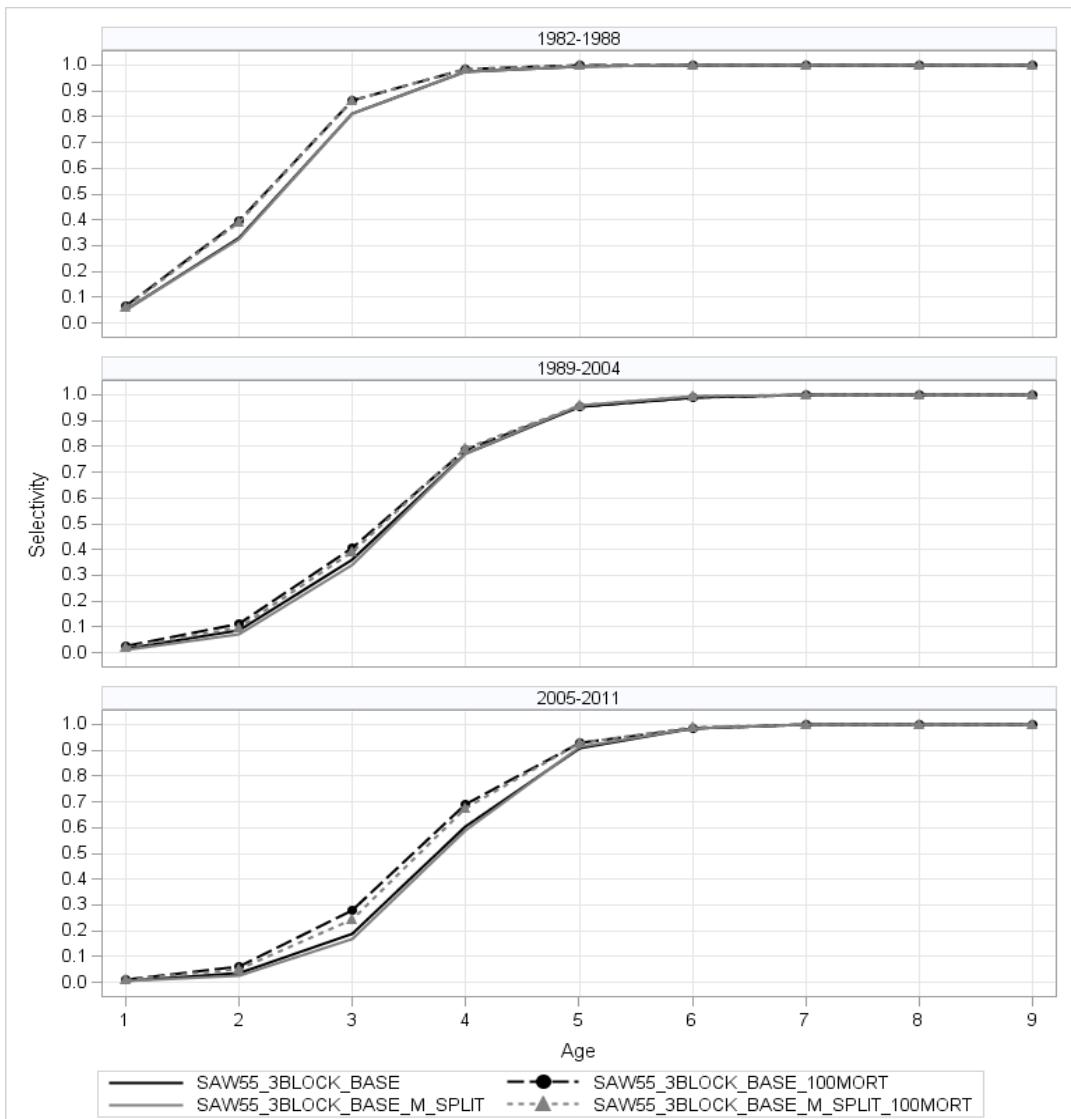


Figure A.151. Gulf of Maine Atlantic cod fishery selectivity blocks for block 1 (1982-1988), block 2 (1989-2004) and block 3 (2005-2011) estimated by sensitivity runs of the ASAP SAW55_3BLOCK_BASE model.

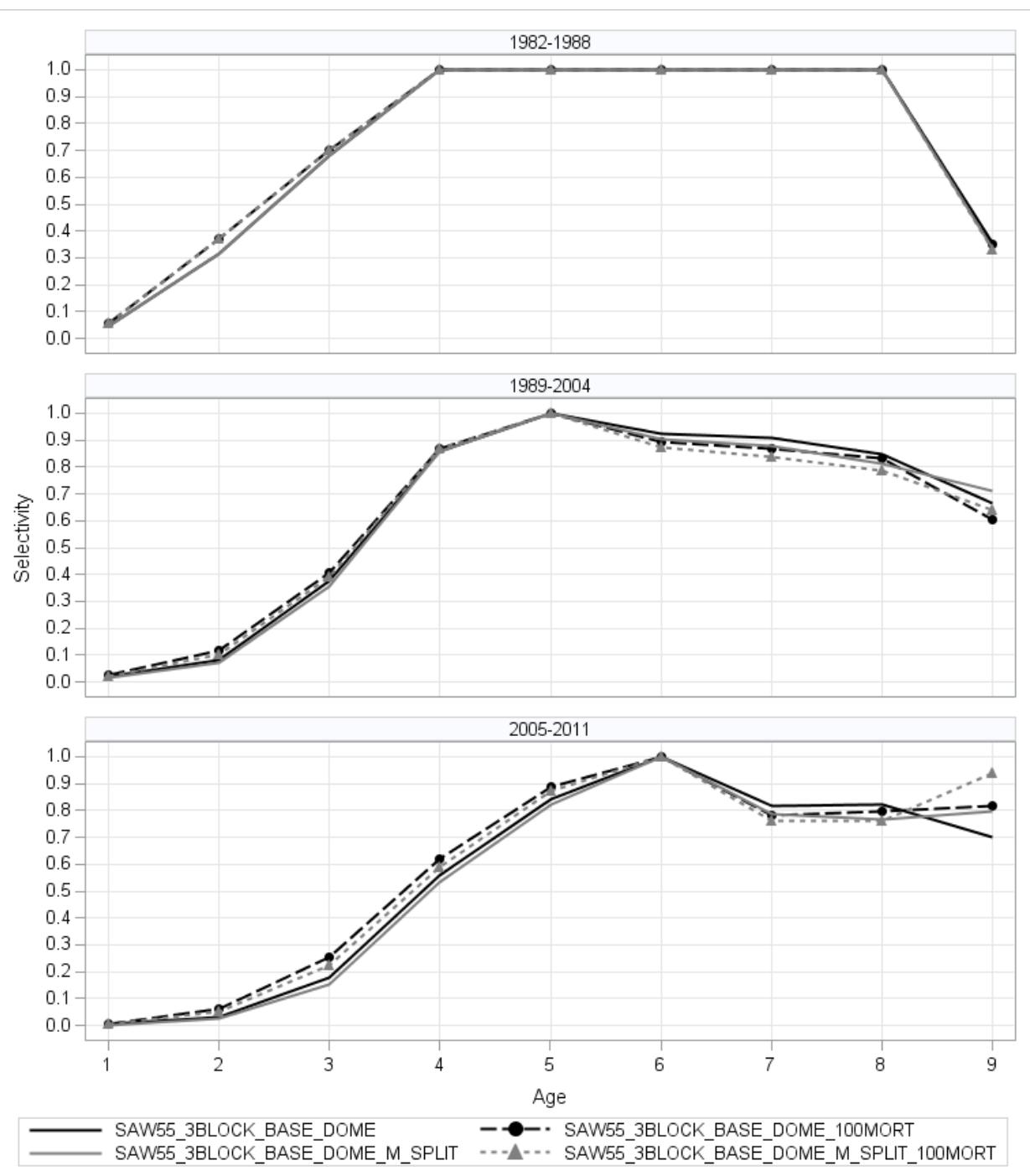


Figure A.152. Gulf of Maine Atlantic cod fishery selectivity blocks for block 1 (1982-1988), block 2 (1989-2004) and block 3 (2005-2011) estimated by sensitivity runs of the ASAP SAW55_3BLOCK_BASE_DOME model.

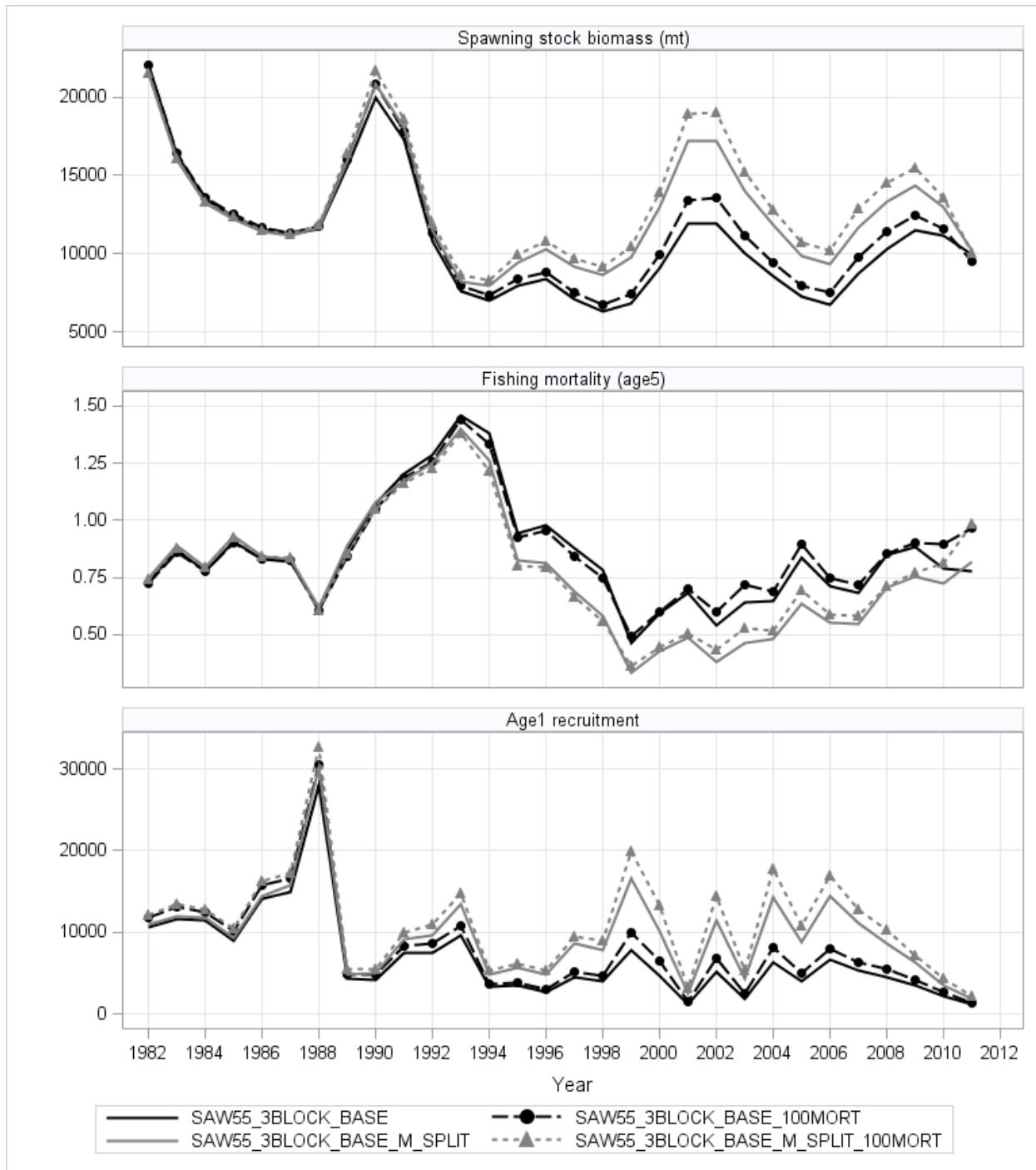


Figure A.153. Gulf of Maine Atlantic cod spawning stock biomass, age 5 fishing mortality and age 1 recruitment estimated by sensitivity runs of the ASAP SAW55_3BLOCK_BASE model.

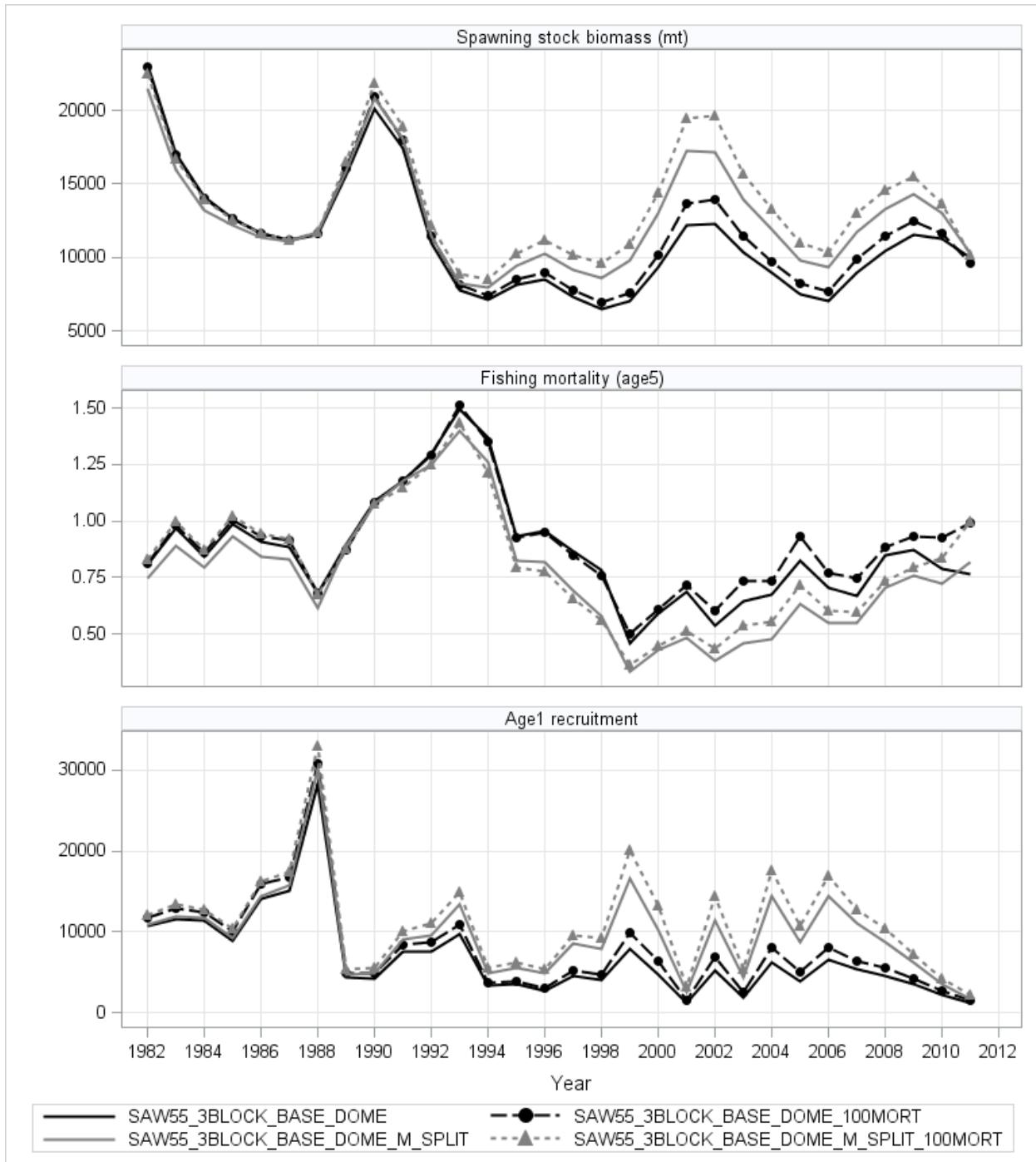


Figure A.154. Gulf of Maine Atlantic cod spawning stock biomass, age 5 fishing mortality and age 1 recruitment estimated by sensitivity runs of the ASAP SAW55_3BLOCK_BASE_DOME model.

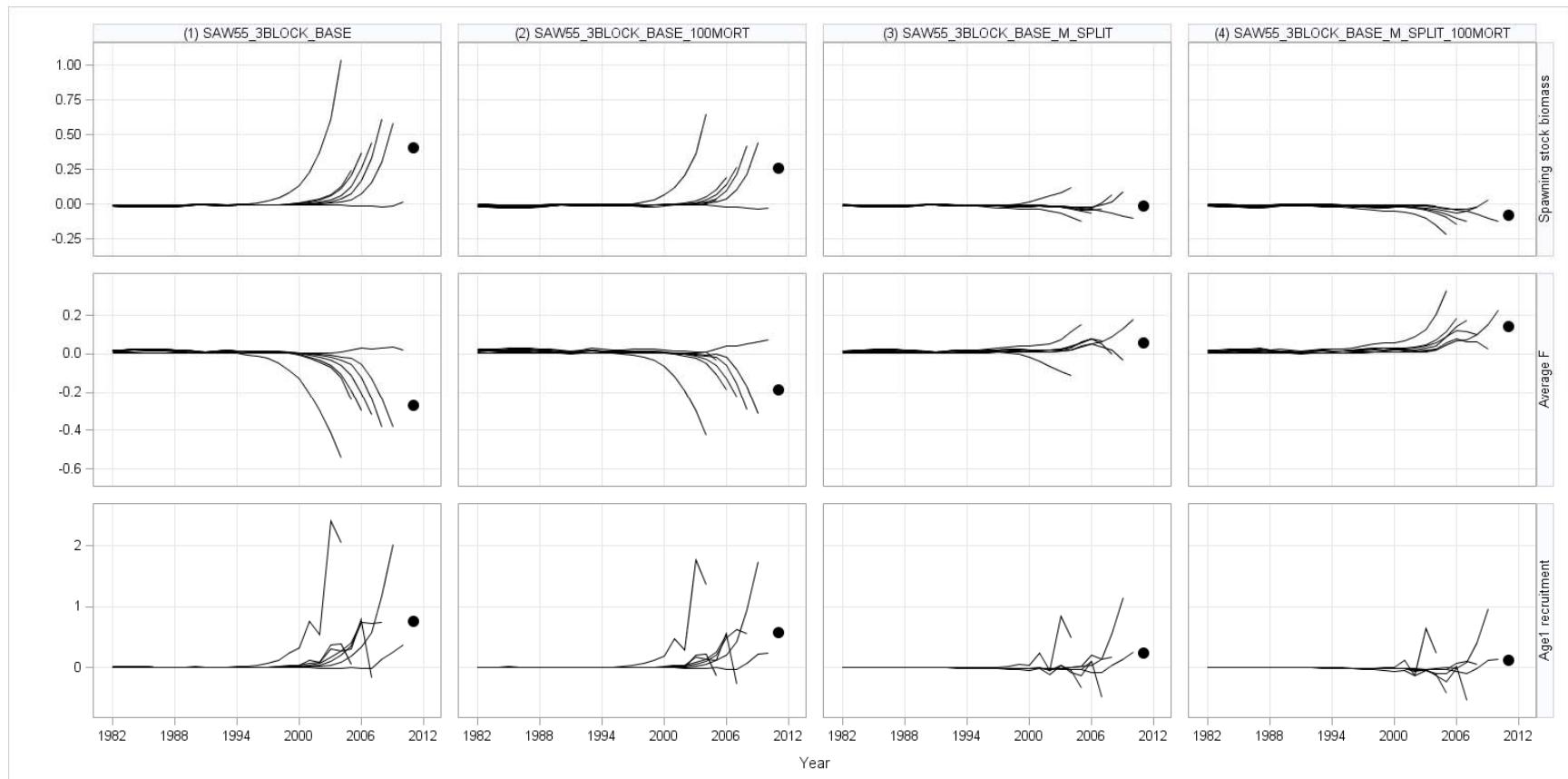


Figure A.155. Model retrospective patterns for sensitivity runs of the Gulf of Maine Atlantic ASAP SAW55_3BLOCK_BASE model. The black circles indicates the Mohn's rho value based on a five year retrospective peel.

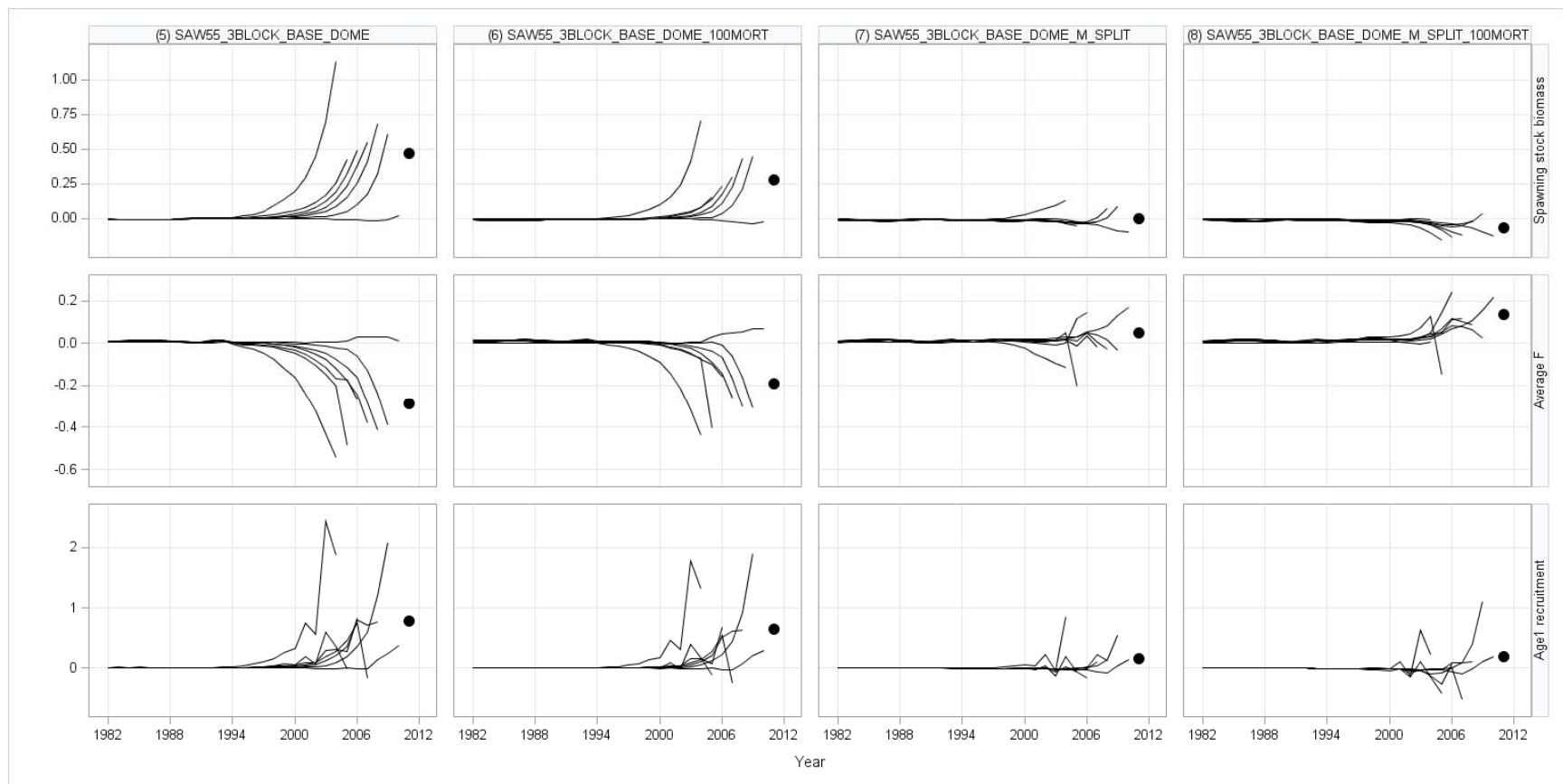


Figure A.156. Model retrospective patterns for sensitivity runs of the Gulf of Maine Atlantic ASAP SAW55_3BLOCK_BASE_DOME model. The black circles indicate the Mohn's rho value based on a five year retrospective peel.

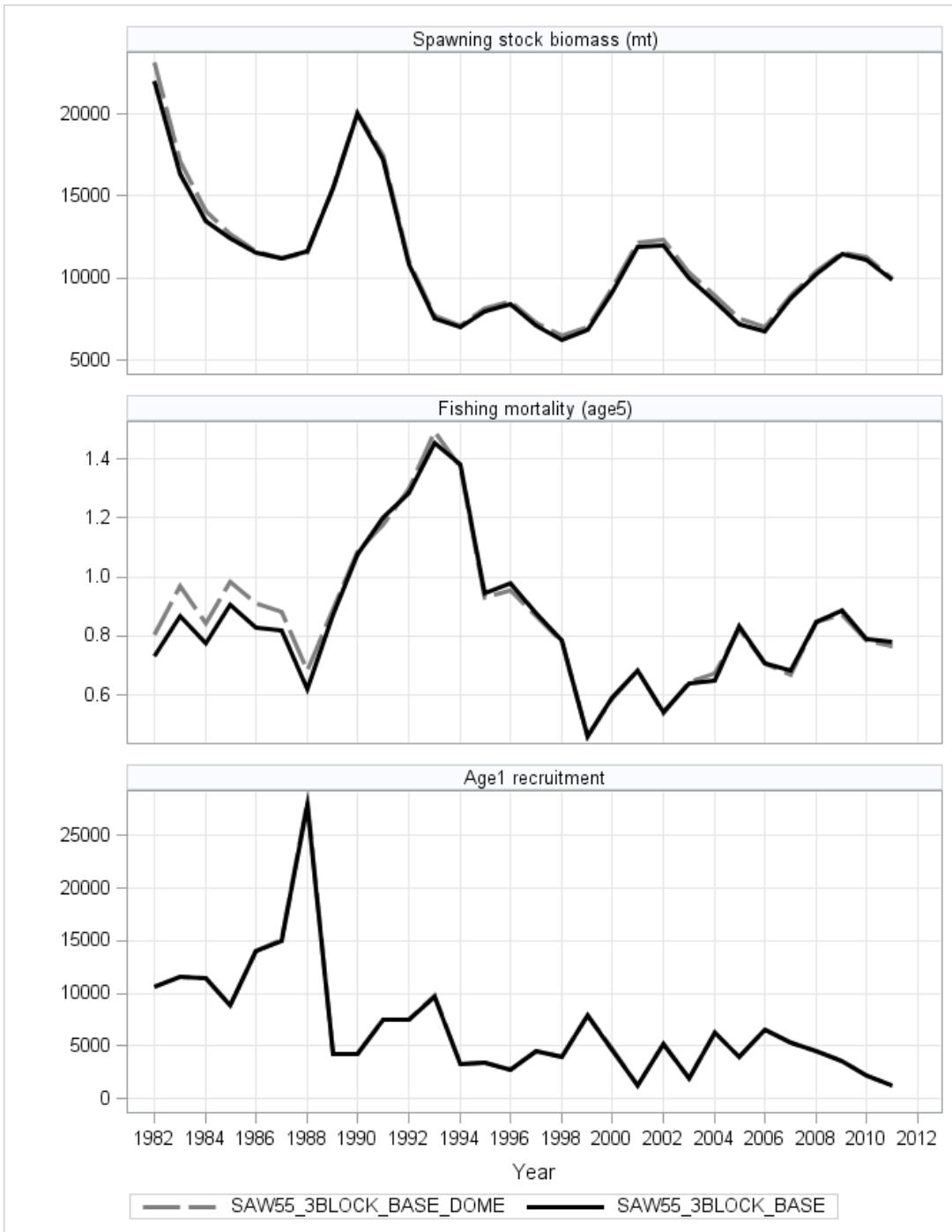


Figure A.157. Comparison of the Gulf of Maine Atlantic cod spawning stock biomass, age 5 fishing mortality and age 1 recruitment from the ASAP SAW55_3BLOCK_BASE and SAW55_3BLOCK_BASE_DOME models.

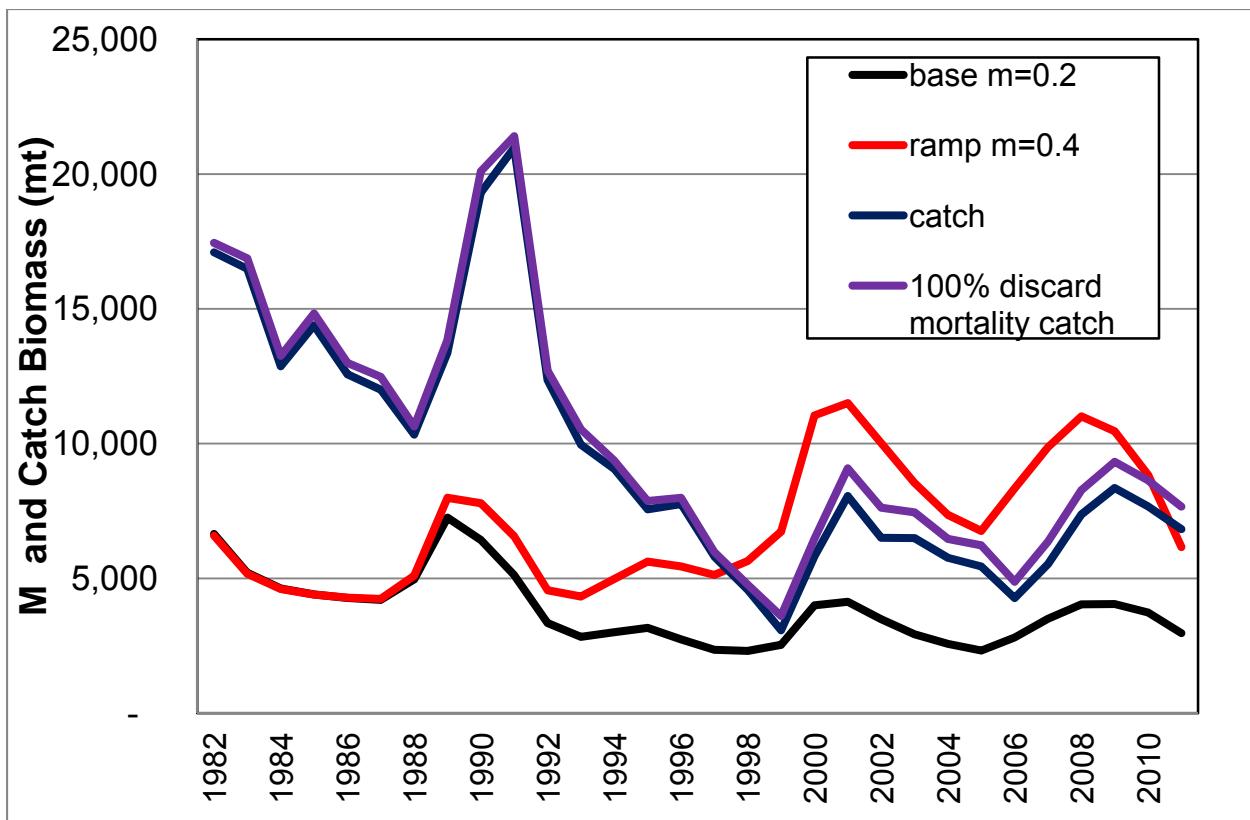


Figure A.158. Comparison of fishery (catch) and natural mortality removals under the $M_{0.2}$ (SAW55_3BLOCK_BASE) and M_{Ramp} (SAW55_3BLOCK_BASE_M_SPLIT) models. Fishery removals are shown using both the revised discard mortality assumptions and the 100% discard mortality assumption.

Fleet 1 Catch (Catch)

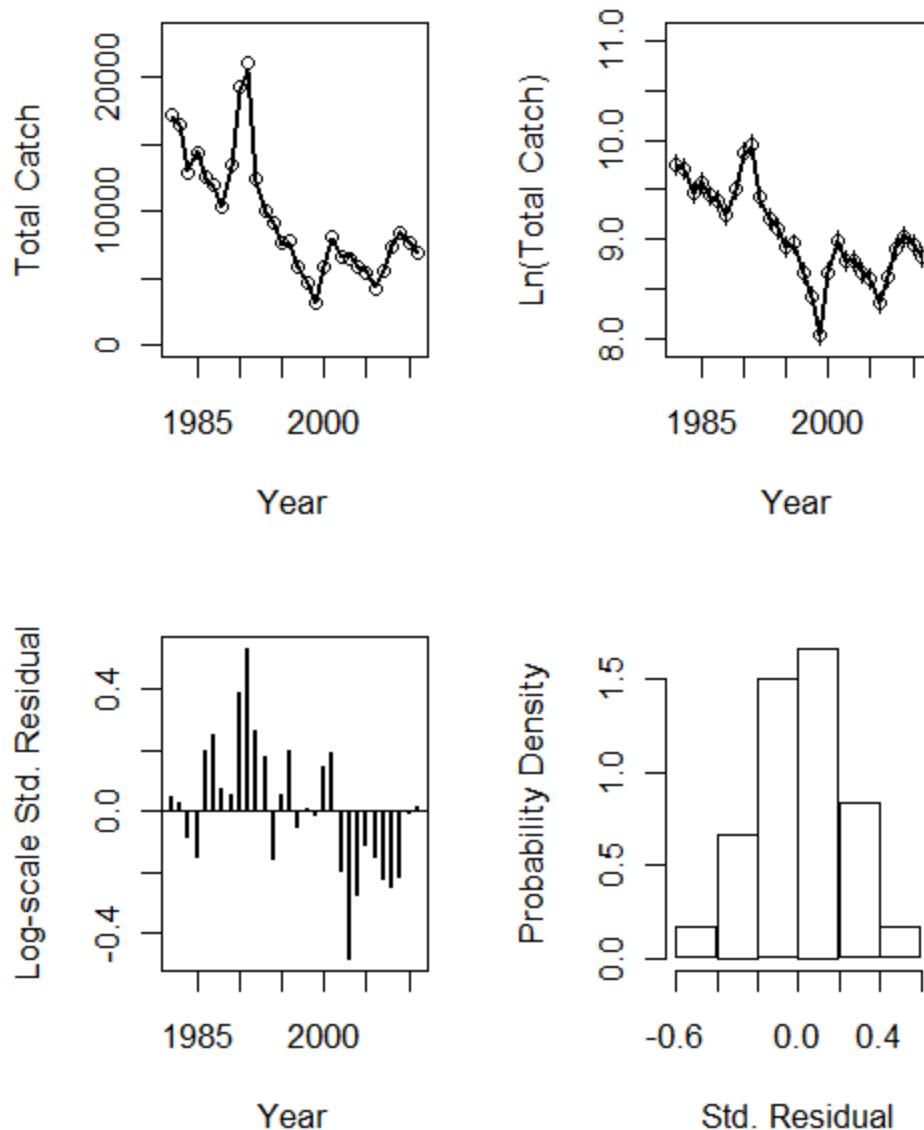


Figure A.159. ASAP SAW55_3BLOCK_BASE model fit to the total Gulf of Maine Atlantic cod fishery catch (Fleet 1).

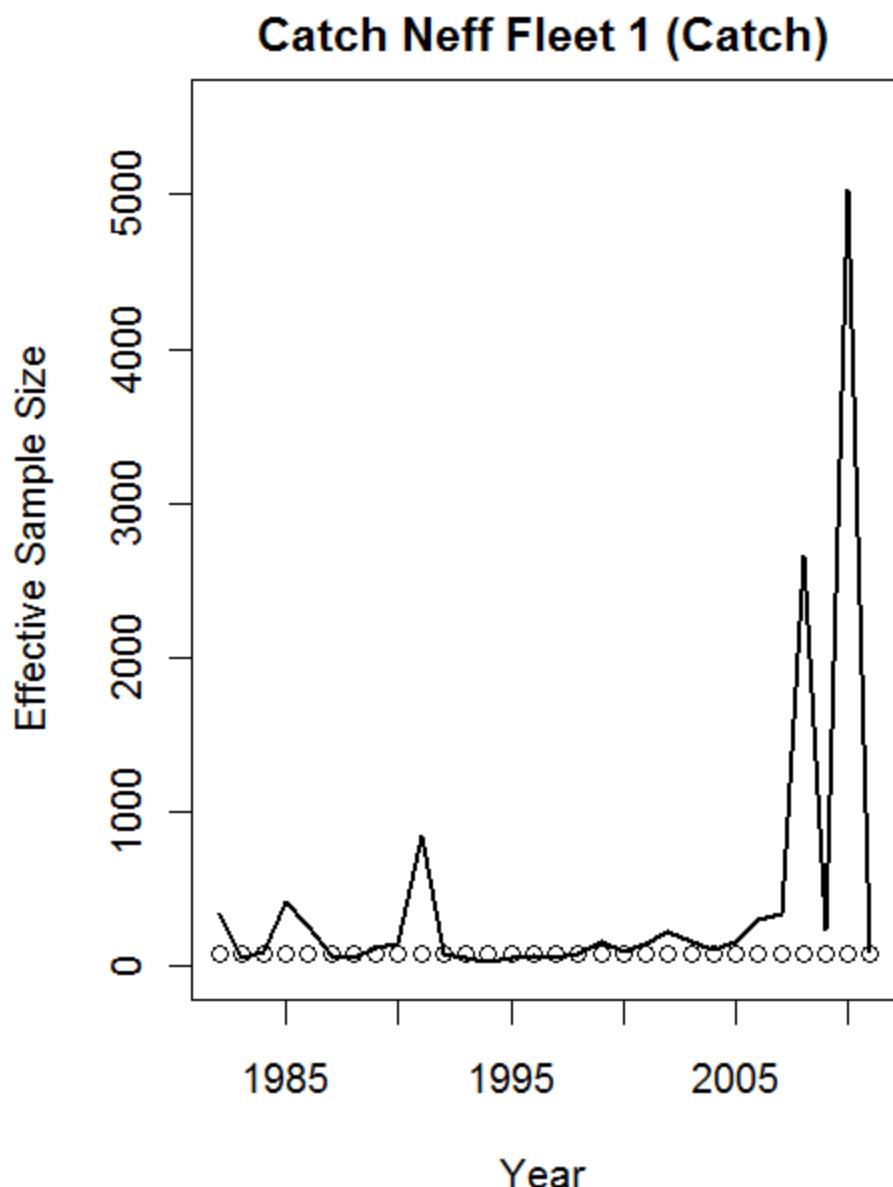


Figure A.160. ASAP SAW55_3BLOCK_BASE model comparison of input effective sample size versus the model estimated effective sample size for the Gulf of Maine Atlantic cod fishery catch.

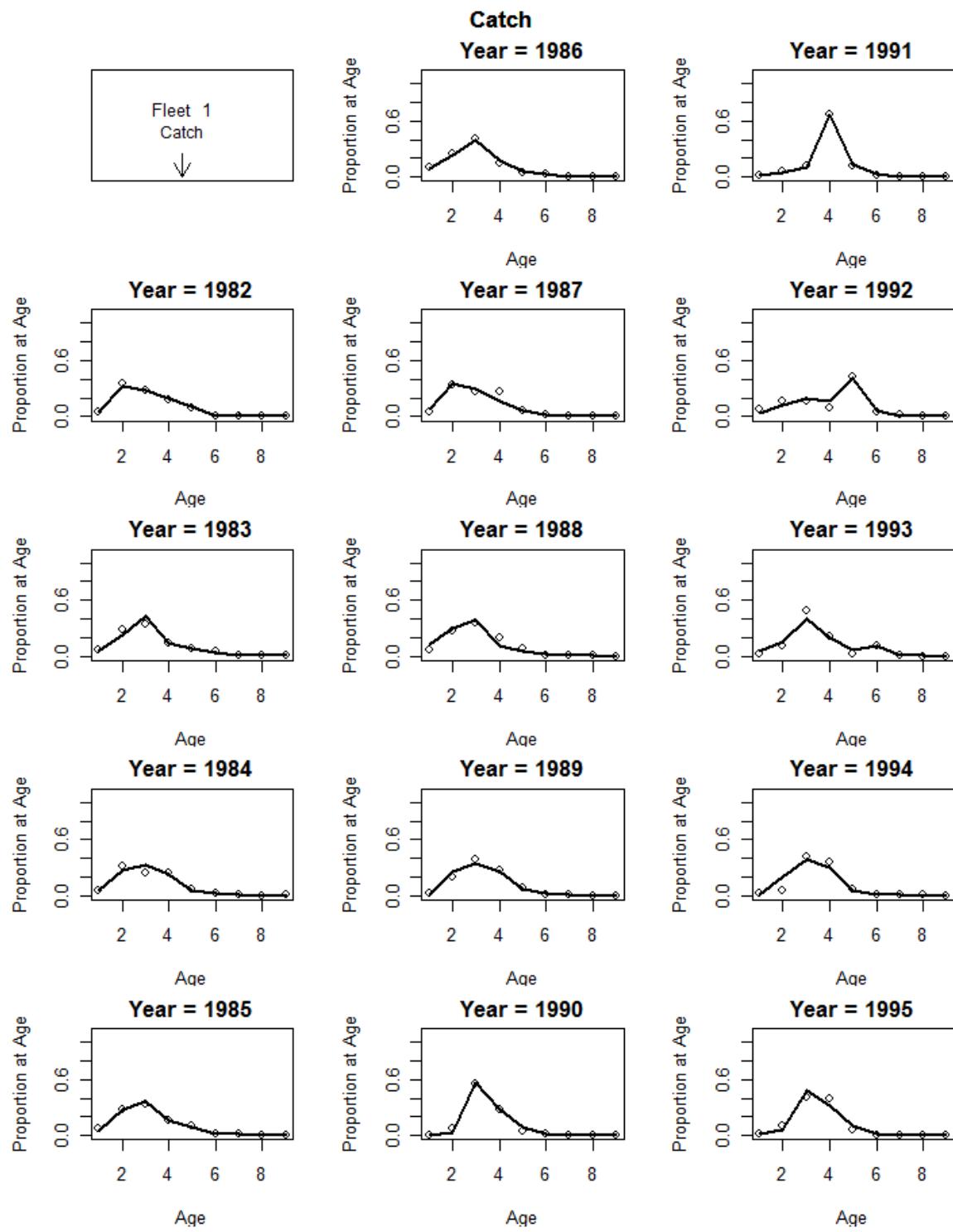


Figure A.161.a. Comparison of the ASAP SAW55_3BLOCK_BASE estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

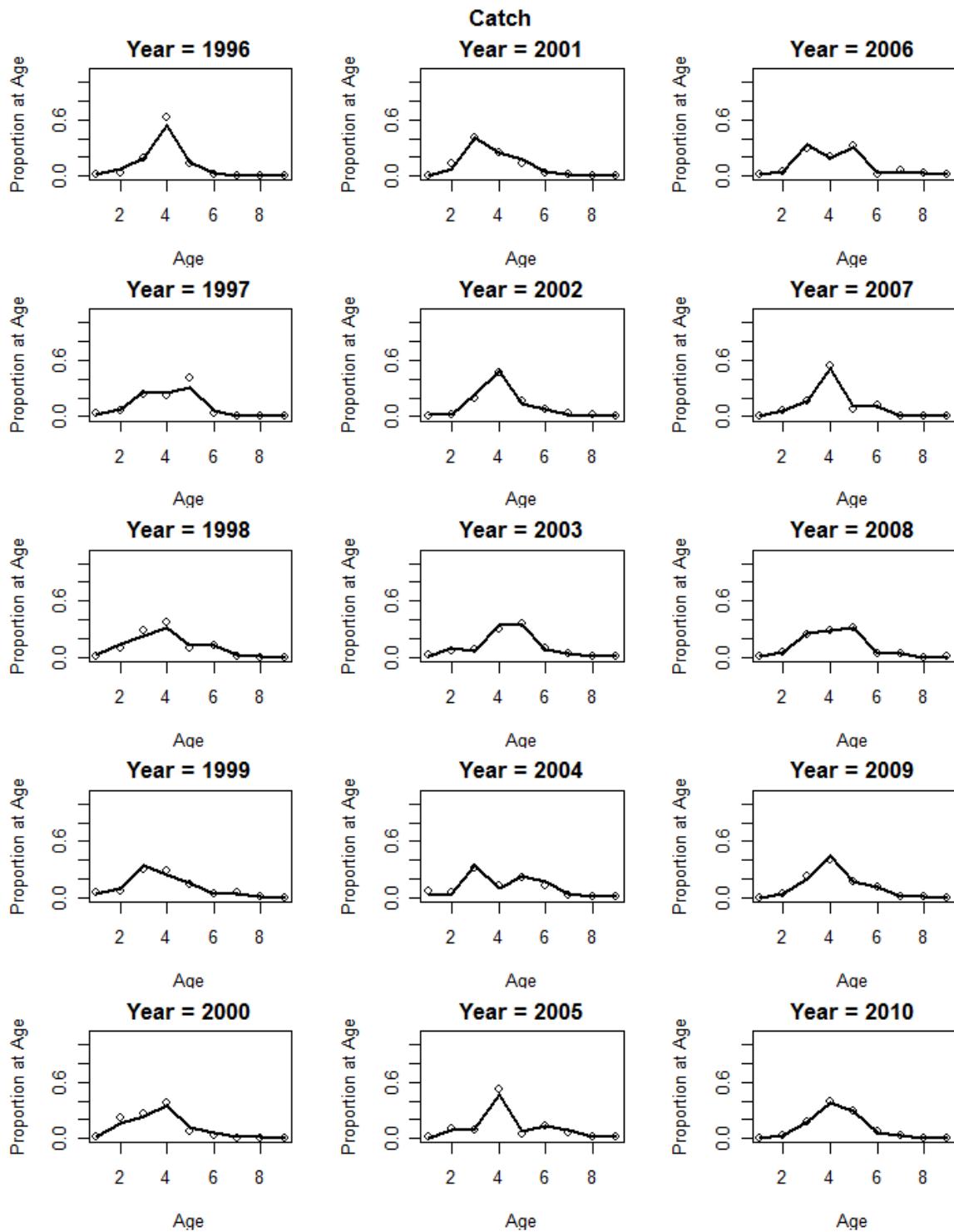


Figure A.161.b. Comparison of the ASAP SAW55_3BLOCK_BASE estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

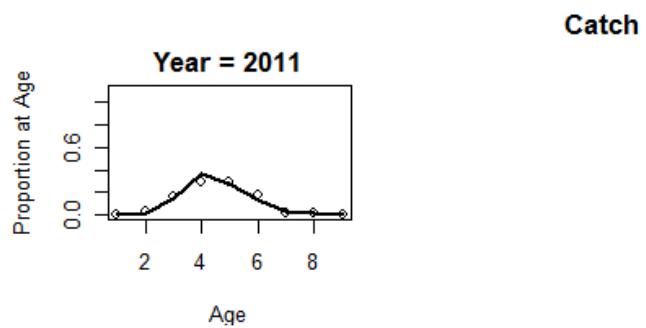


Figure A.161.c. Comparison of the ASAP SAW55_3BLOCK_BASE estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

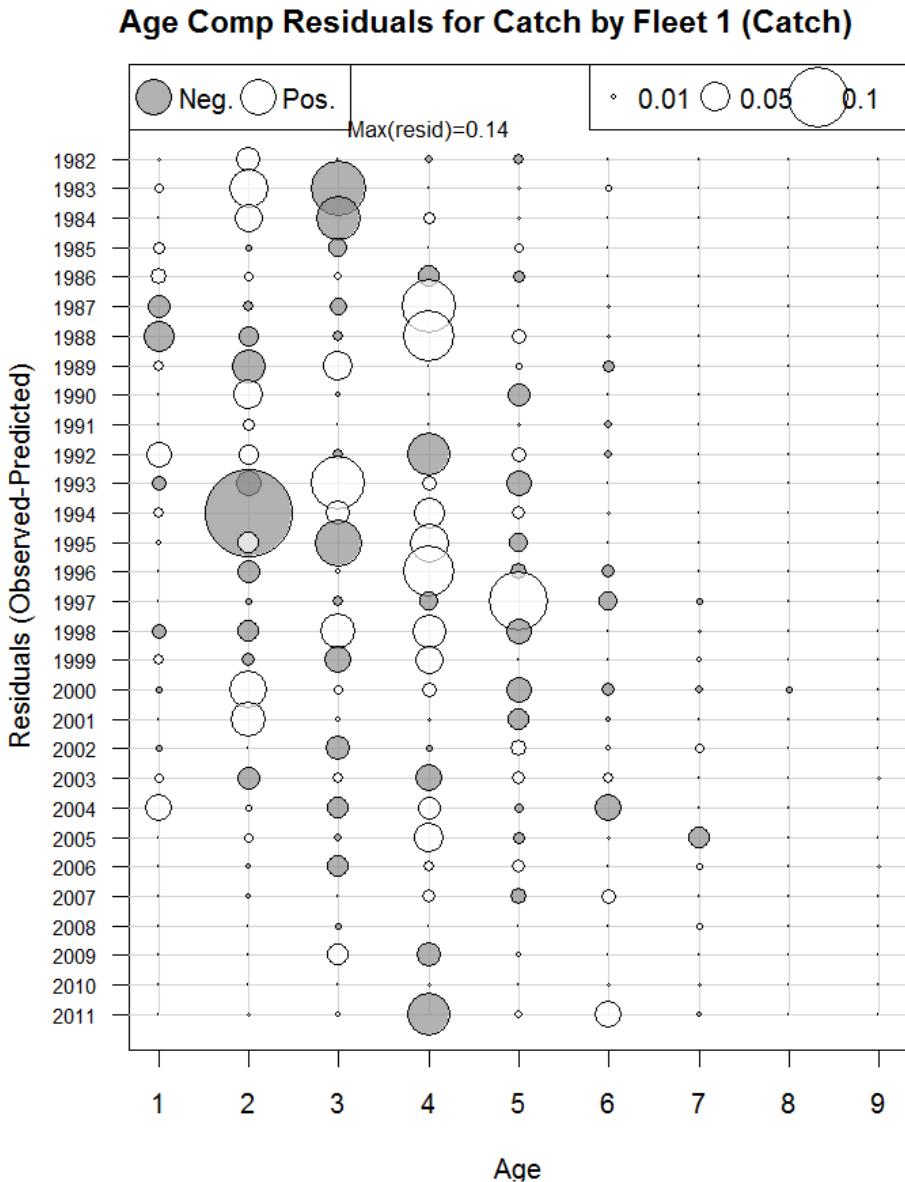


Figure A.162. ASAP SAW55_3BLOCK_BASE model fit residuals for the fishery (Fleet 1) catch-at-age of Gulf of Maine Atlantic cod.

Catch Fleet 1 (Catch) ESS = 80

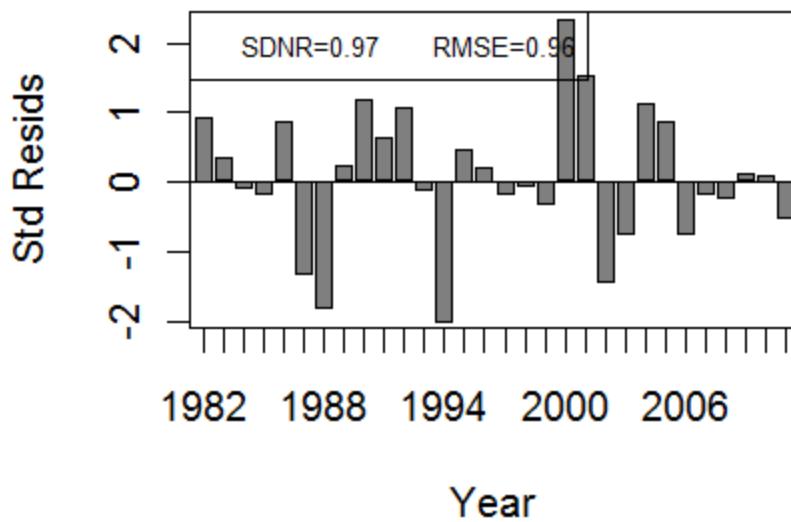
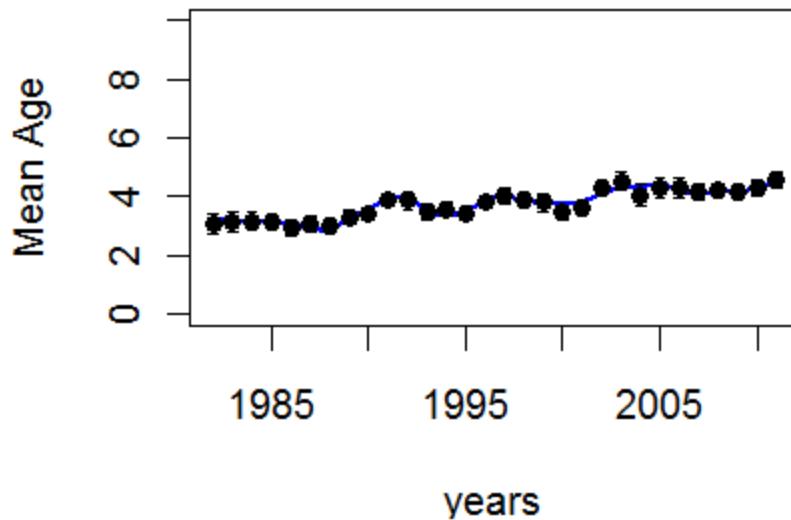


Figure A.163. ASAP SAW55_3BLOCK_BASE predicted mean age of Gulf of Maine Atlantic cod in the fishery catch (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

Fleet 1 (Catch)

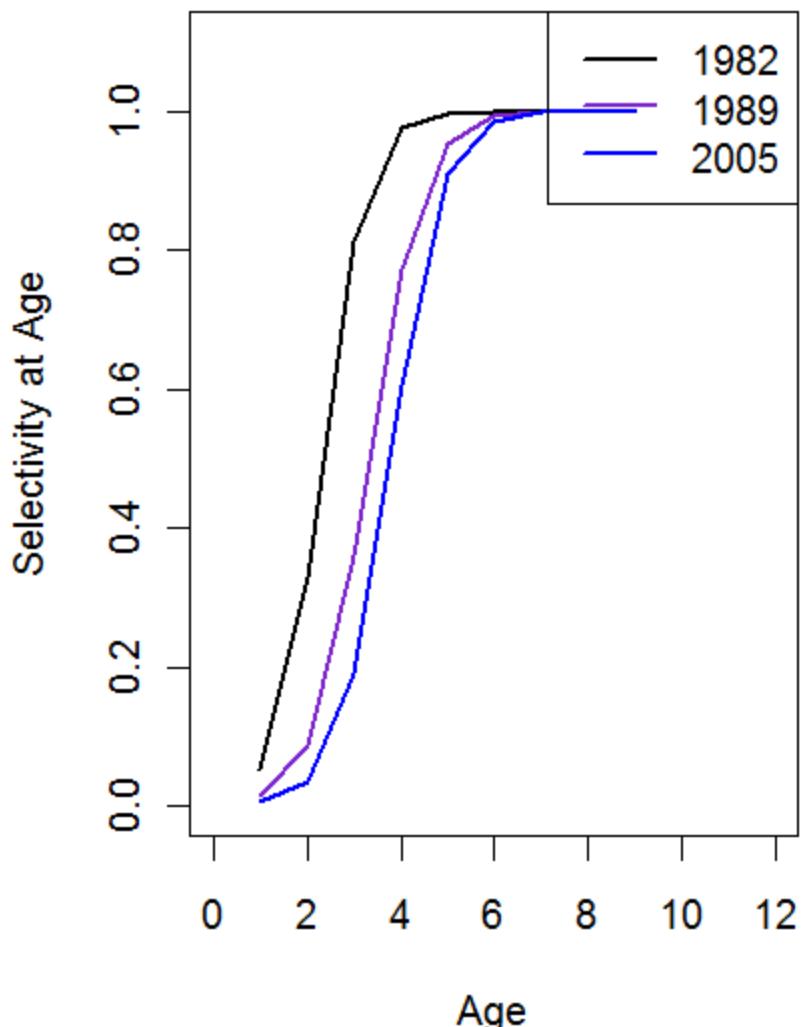


Figure A.164. Gulf of Maine Atlantic cod fishery selectivity blocks for block 1 (1982-1989), block 2 (1990-2004) and block 3 (2005-2011) estimated by the ASAP SAW55_3BLOCK_BASE model.

Index 1 (NEFSCspring)

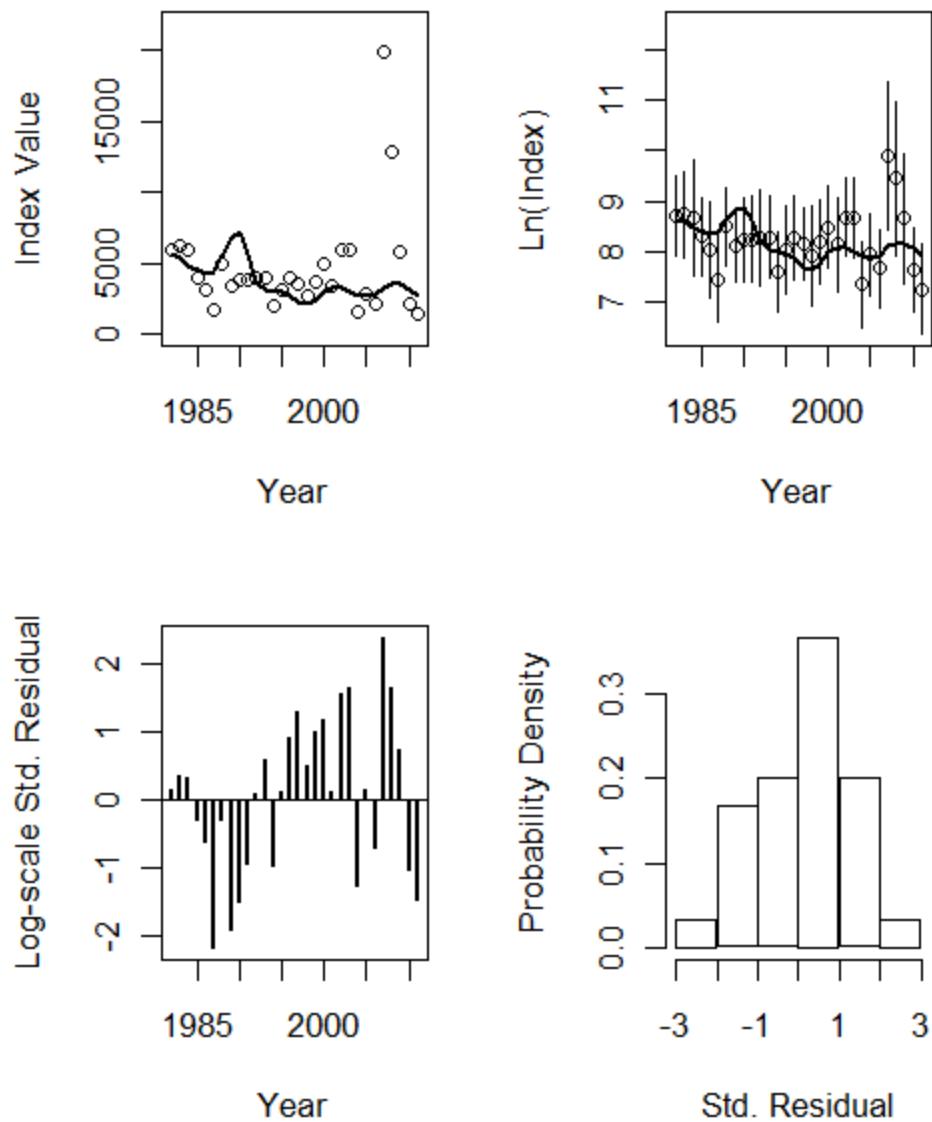


Figure A.165. ASAP SAW55_3BLOCK_BASE model fit to the NEFSC Gulf of Maine Atlantic cod spring (Index 1) survey.

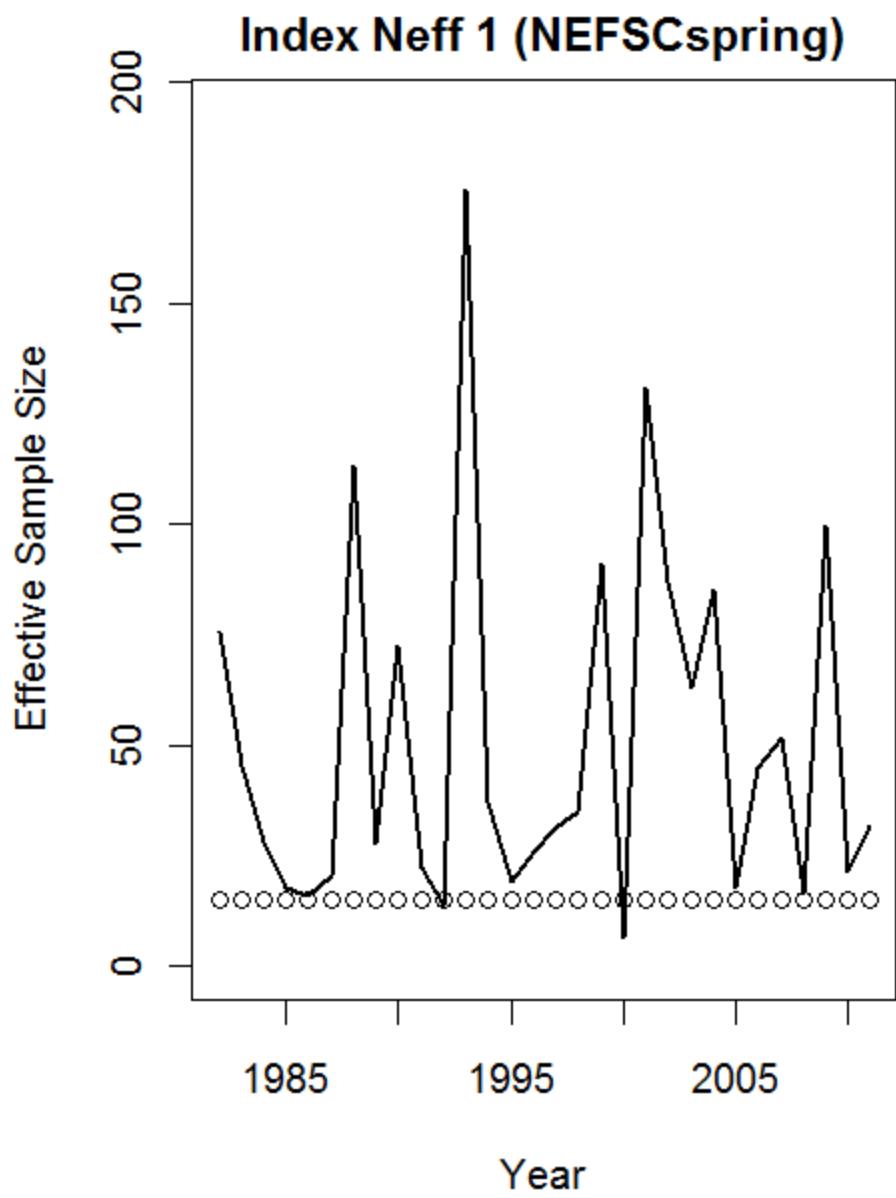


Figure A.166. ASAP SAW55_3BLOCK_BASE model comparison of input effective sample size versus the model estimated effective sample size for the NEFSC spring (Index 1) Gulf of Maine Atlantic cod index.

Age Comp Residuals for Index 1 (NEFSCspring)

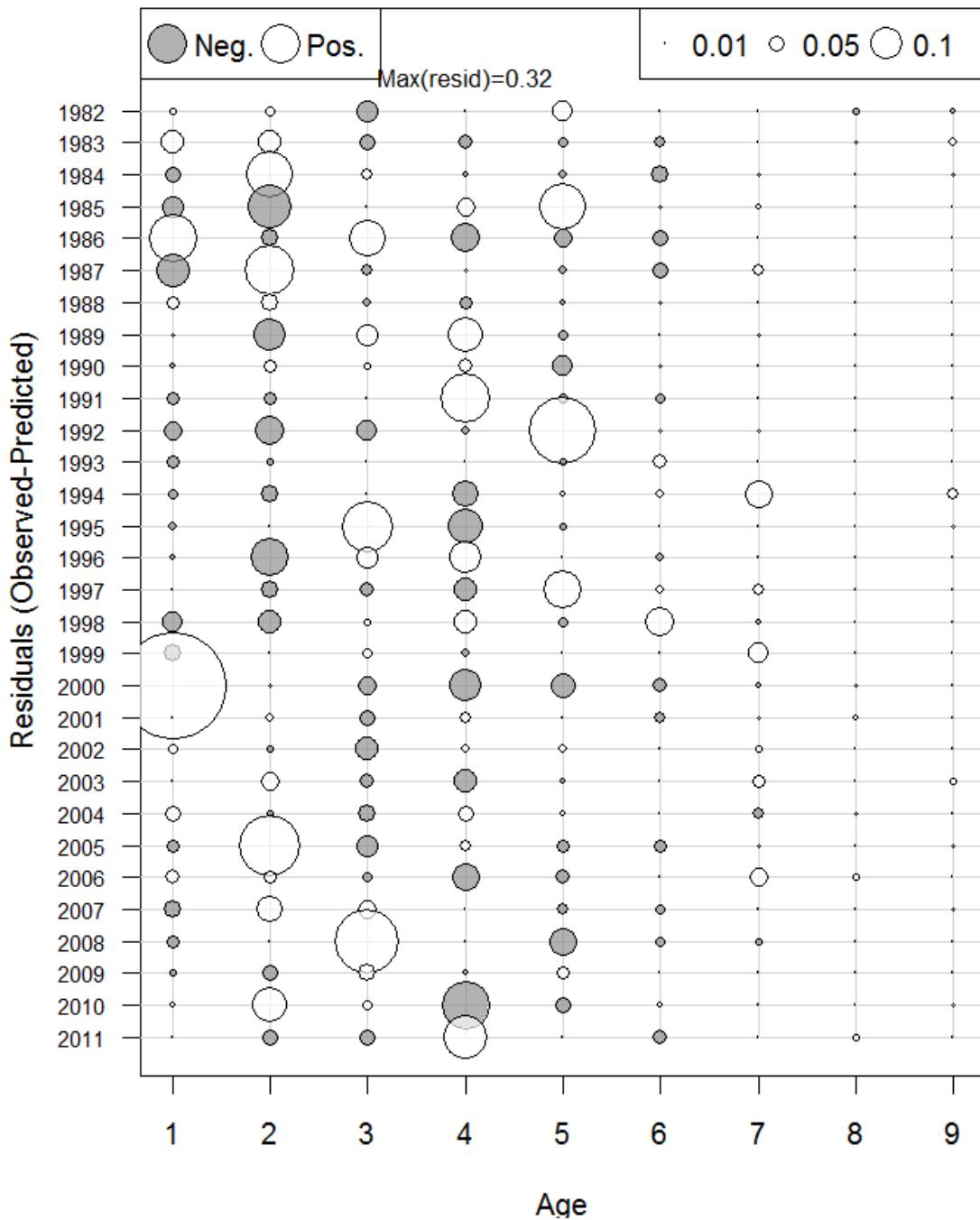


Figure A.167. ASAP SAW55_3BLOCK_BASE model fit residuals for the NEFSC spring survey (Index 1) Gulf of Maine Atlantic cod age composition.

Index 1 (NEFSCspring) ESS = 15

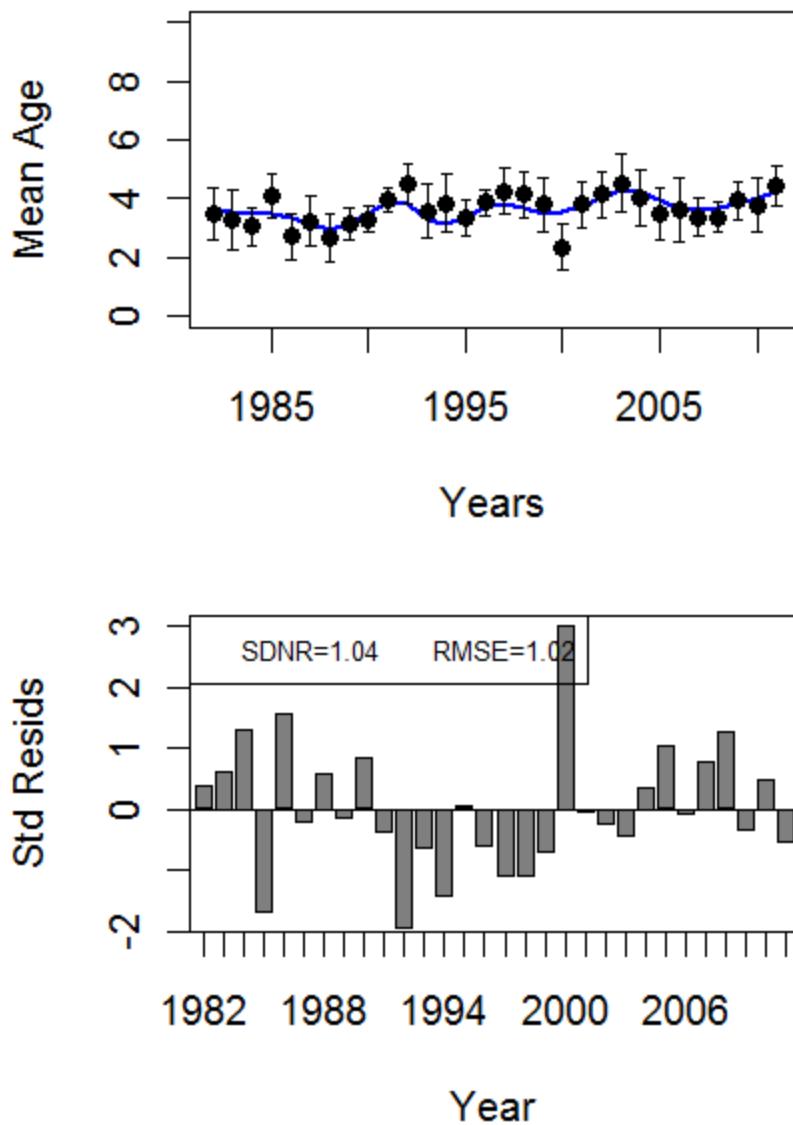


Figure A.168. ASAP SAW55_3BLOCK_BASE predicted mean age of Gulf of Maine Atlantic cod in the NEFSC spring (Index 1) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

Index 2 (NEFSCfall)

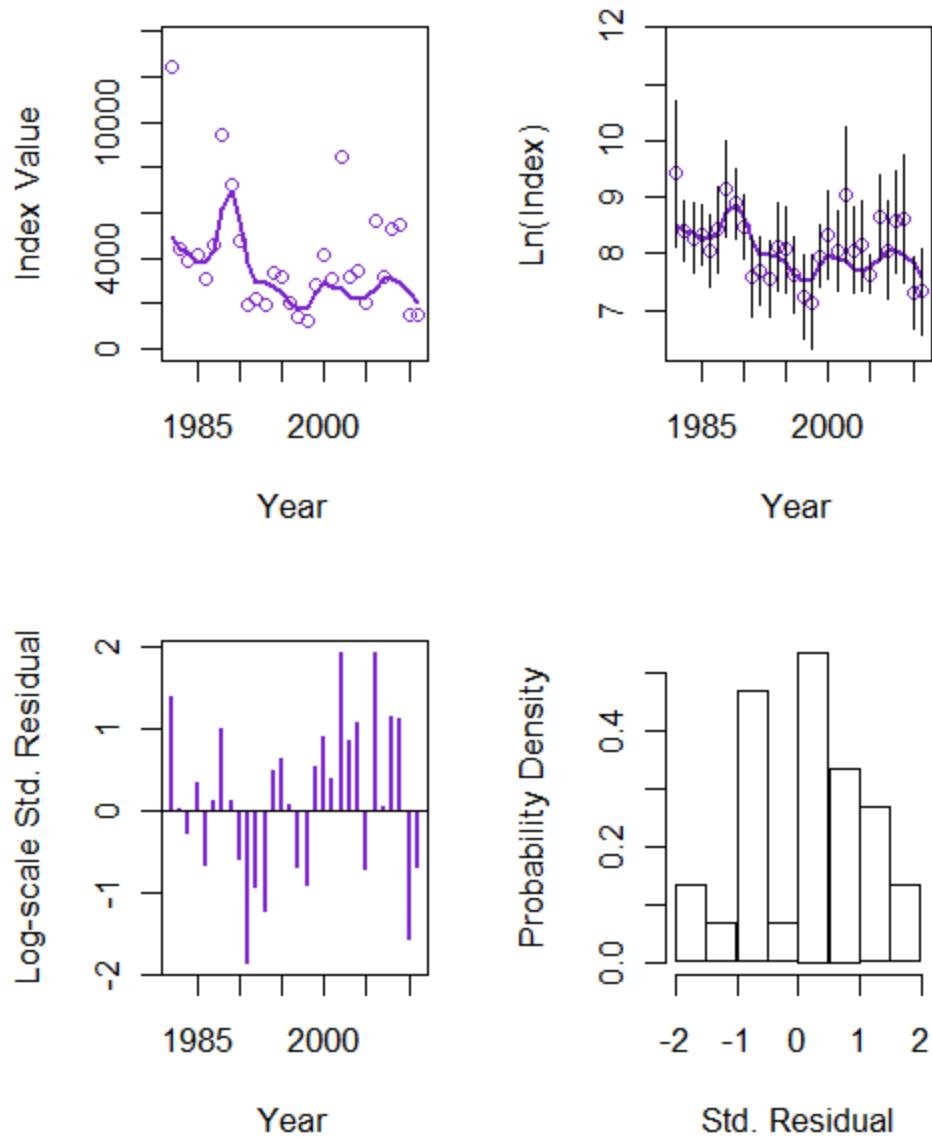


Figure A.169. ASAP SAW55_3BLOCK_BASE model fit to the NEFSC fall (Index 2) survey Gulf of Maine Atlantic cod index.

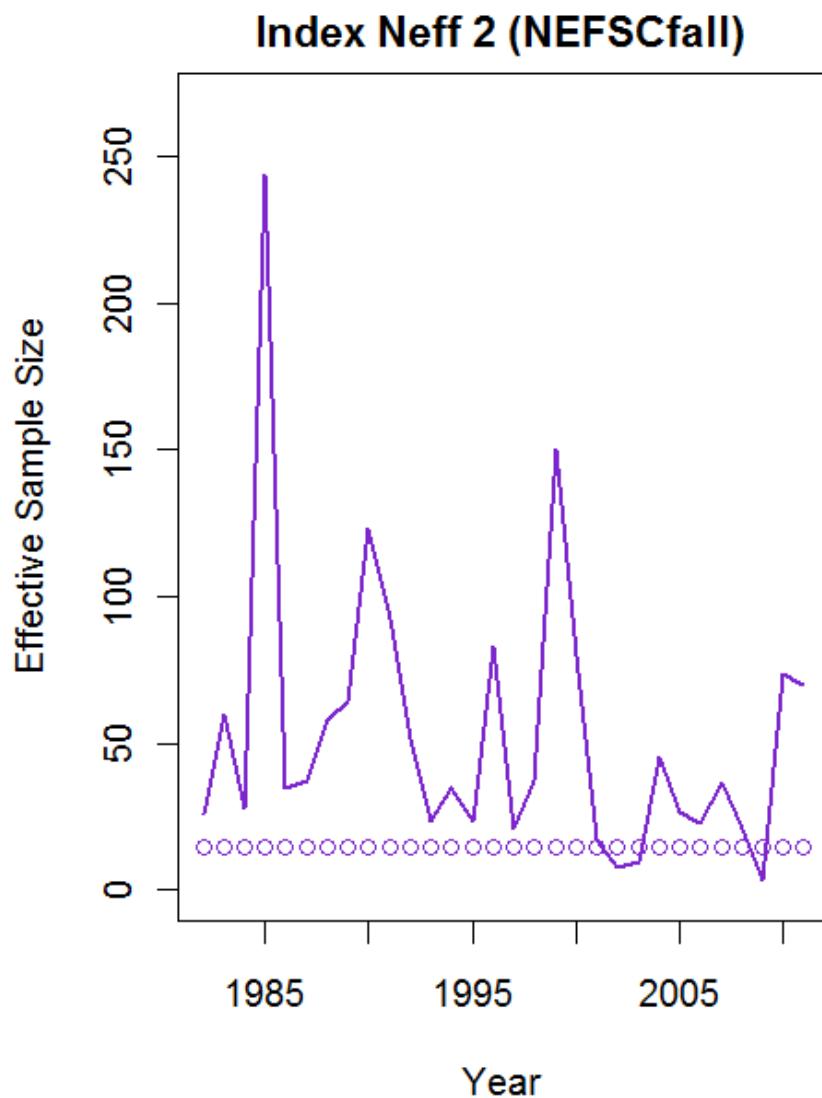


Figure A.170. ASAP SAW55_3BLOCK_BASE model comparison of input effective sample size versus the model estimated effective sample size for the NEFSC fall (Index 2) survey Gulf of Maine Atlantic cod index.

Age Comp Residuals for Index 2 (NEFSCfall)

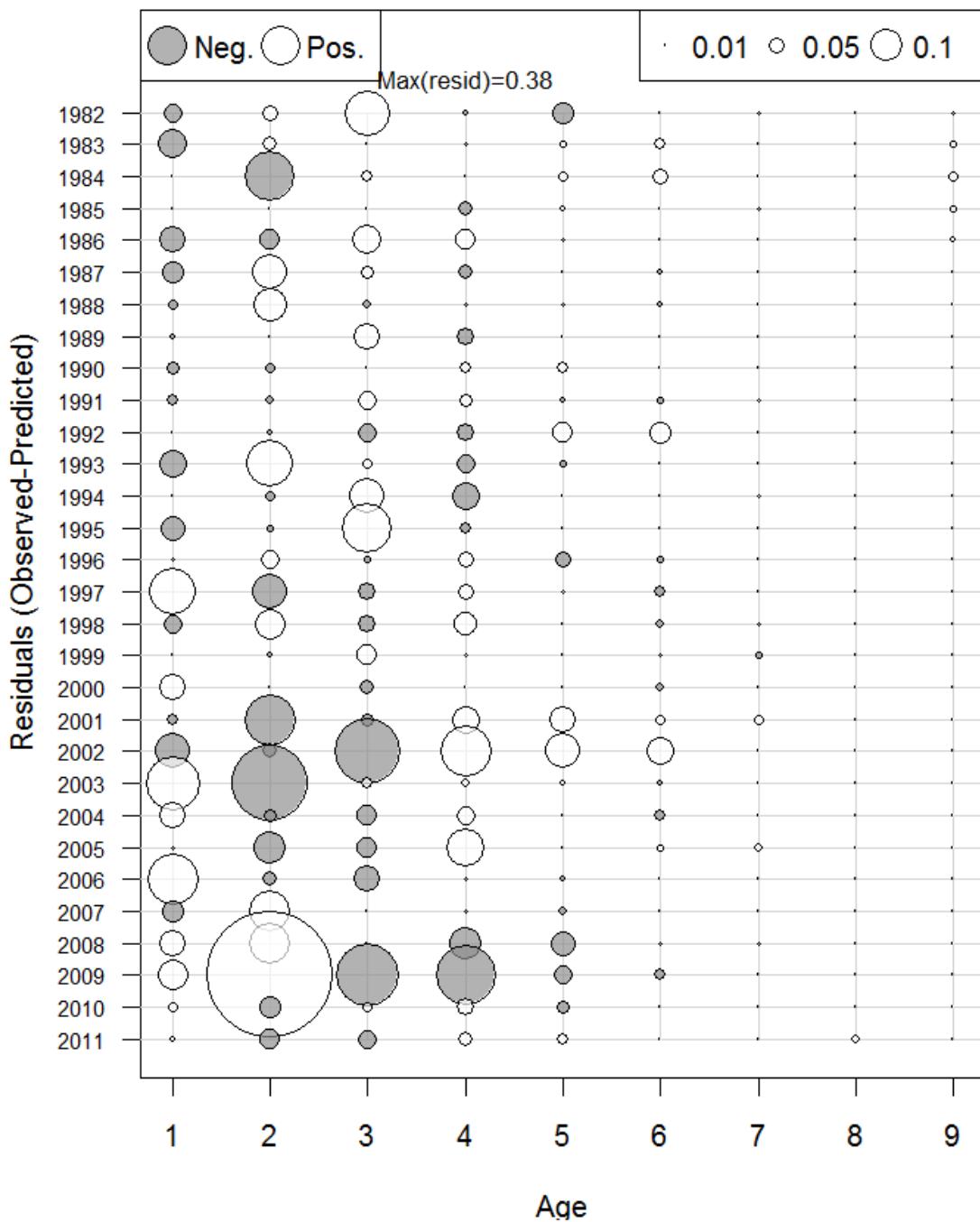


Figure A.171. ASAP SAW55_3BLOCK_BASE model fit residuals for the NEFSC fall survey (Index 2) Gulf of Maine Atlantic cod age composition.

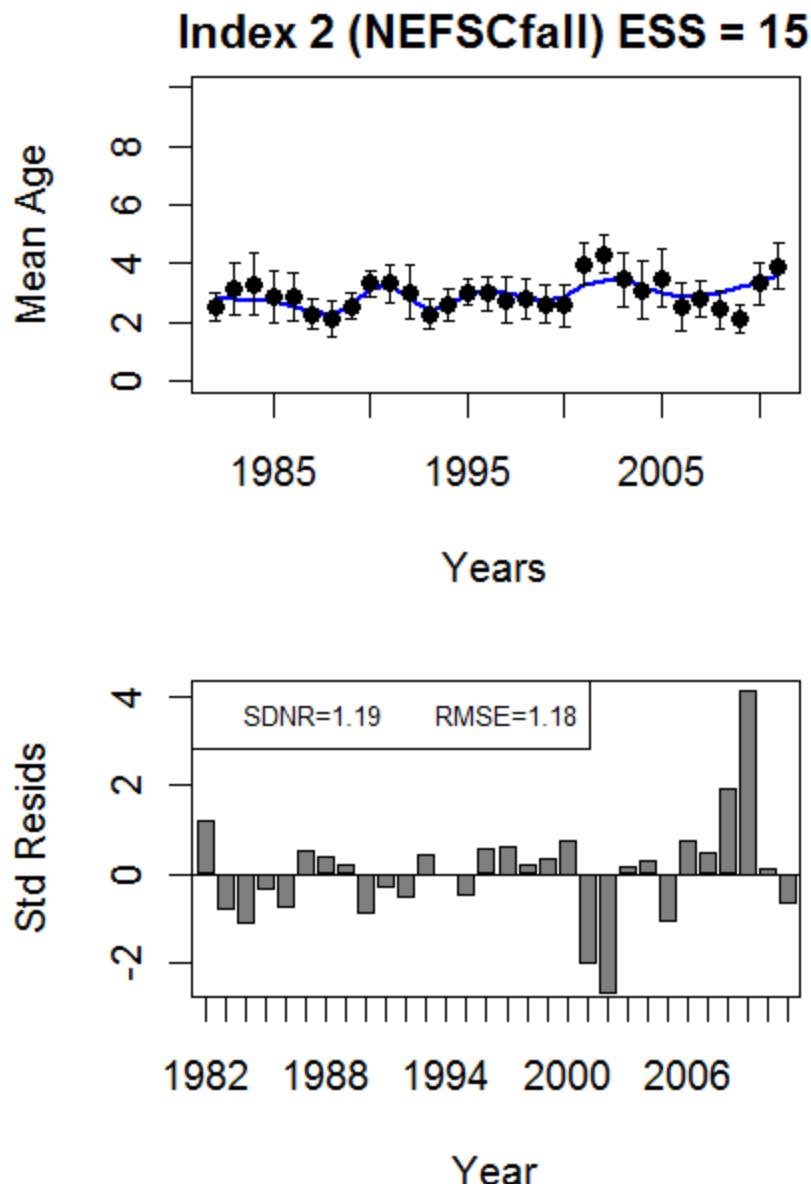


Figure A.172. ASAP SAW55_3BLOCK_BASE predicted mean age of Gulf of Maine Atlantic cod in the NEFSC fall (Index 2) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

Index 3 (MAspring)

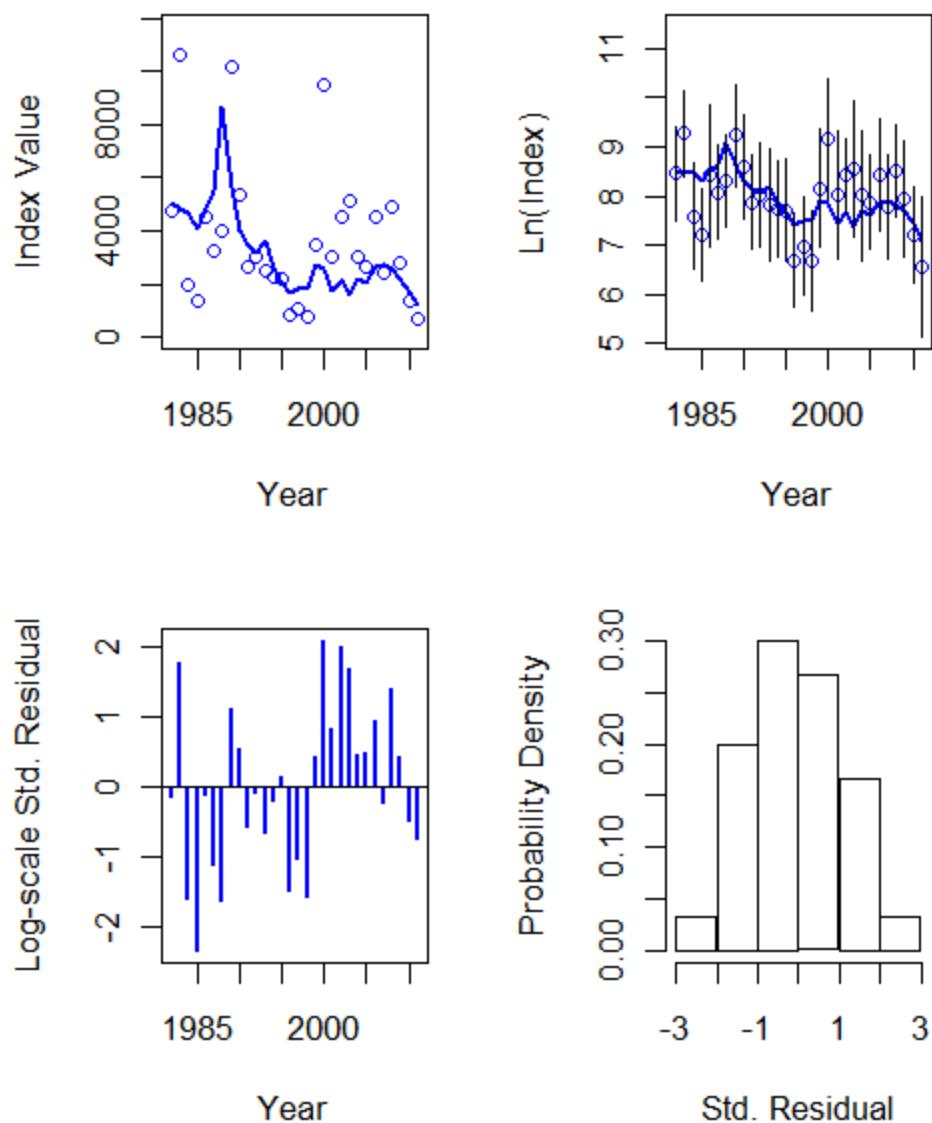


Figure A.173. ASAP SAW55_3BLOCK_BASE model fit to the MADMF spring (Index 3) survey Gulf of Maine Atlantic cod index.

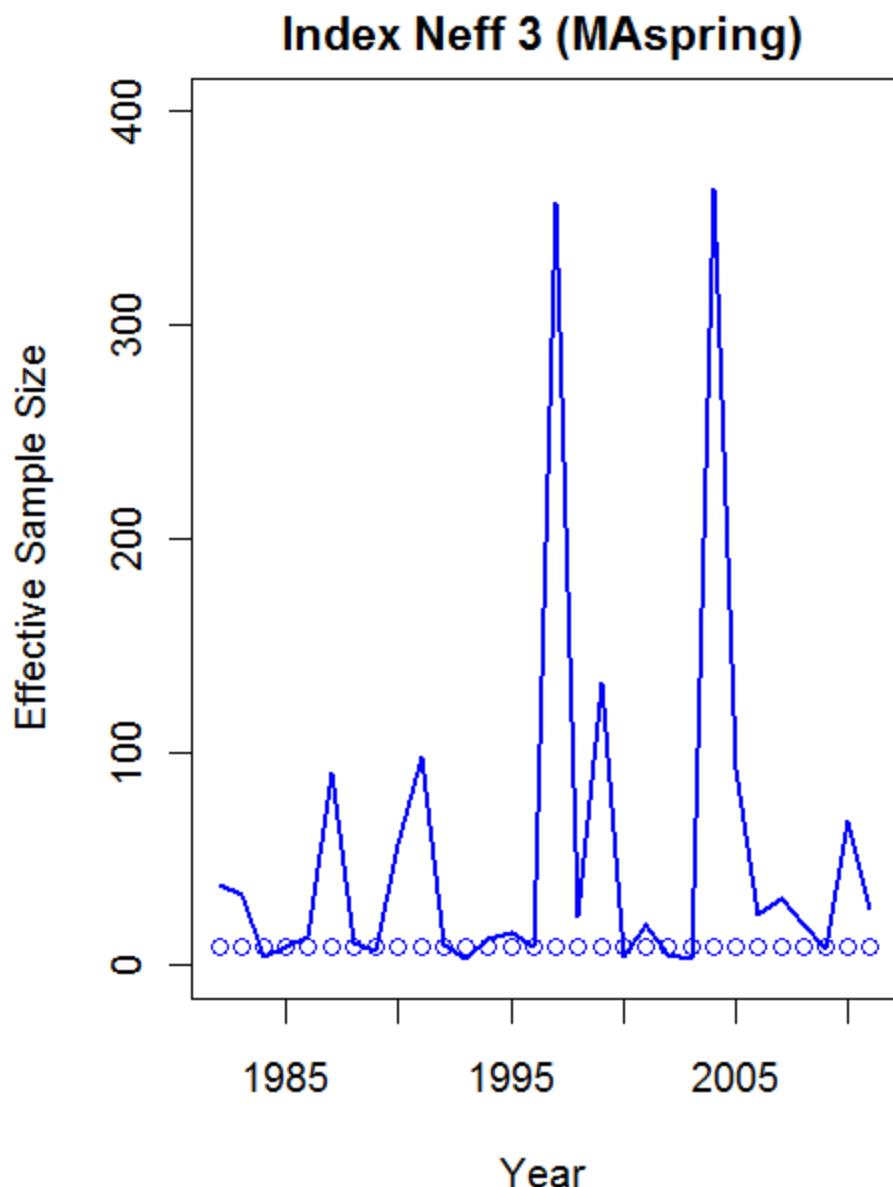


Figure A.174. ASAP SAW55_3BLOCK_BASE model comparison of input effective sample size versus the model estimated effective sample size for the MADMF spring (Index 3) survey Gulf of Maine Atlantic cod index.

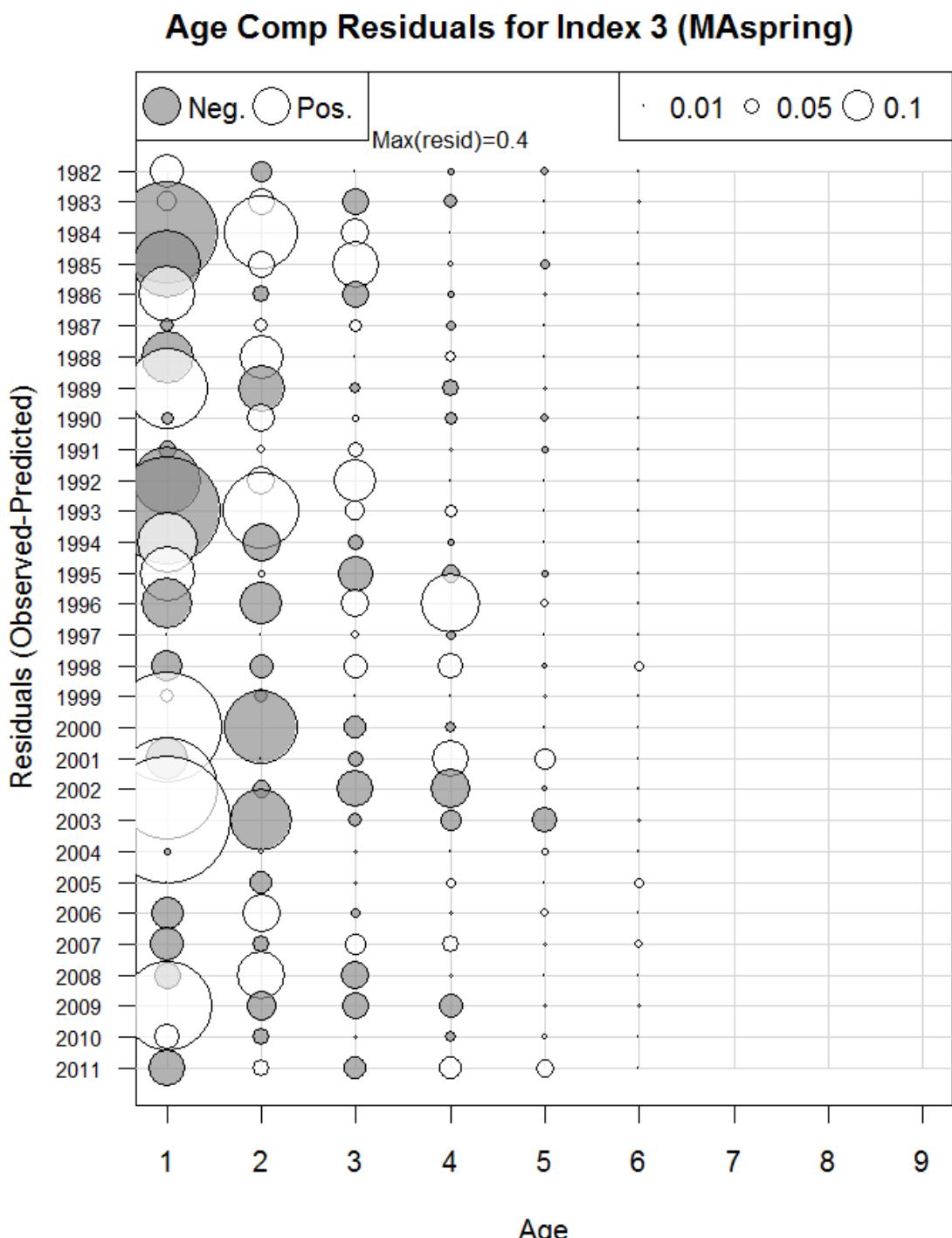


Figure A.175. ASAP SAW55_3BLOCK_BASE model fit residuals for the MADMF spring survey (Index 3) Gulf of Maine cod age composition.

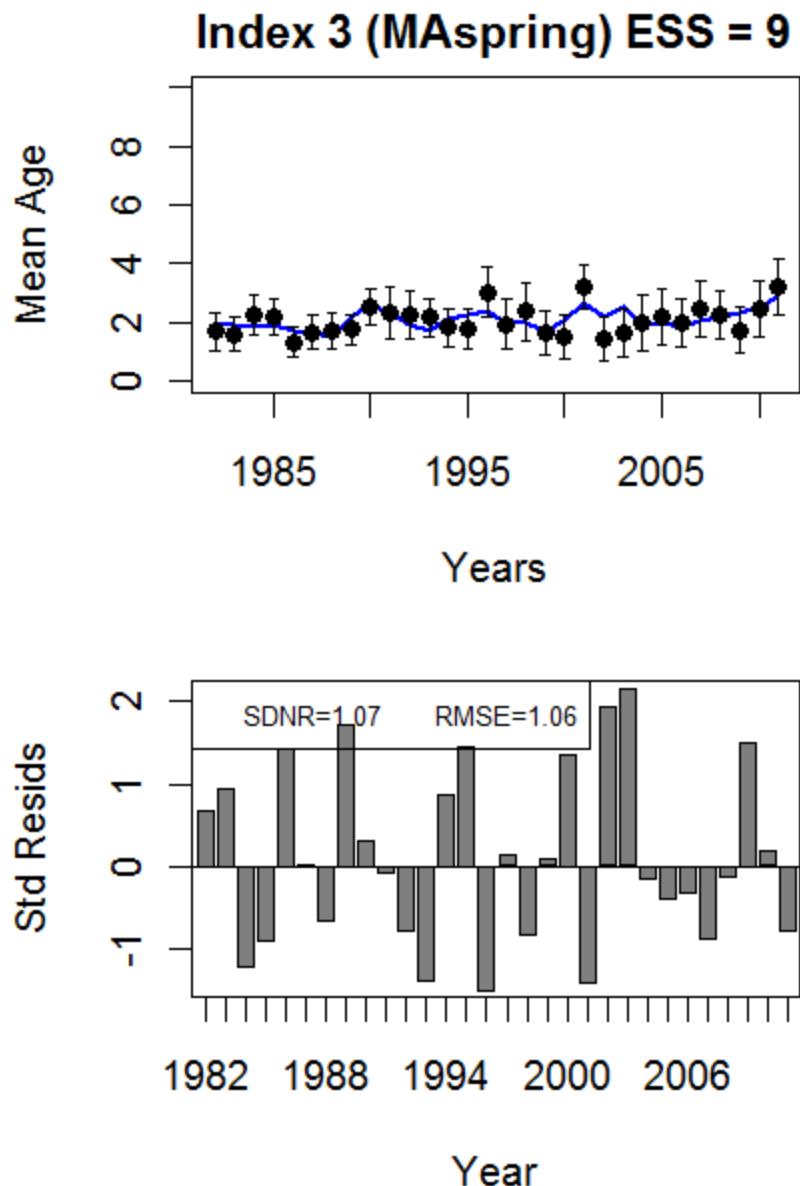


Figure A.176. ASAP SAW55_3BLOCK_BASE predicted mean age of Gulf of Maine Atlantic cod in the MADMF spring (Index 3) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

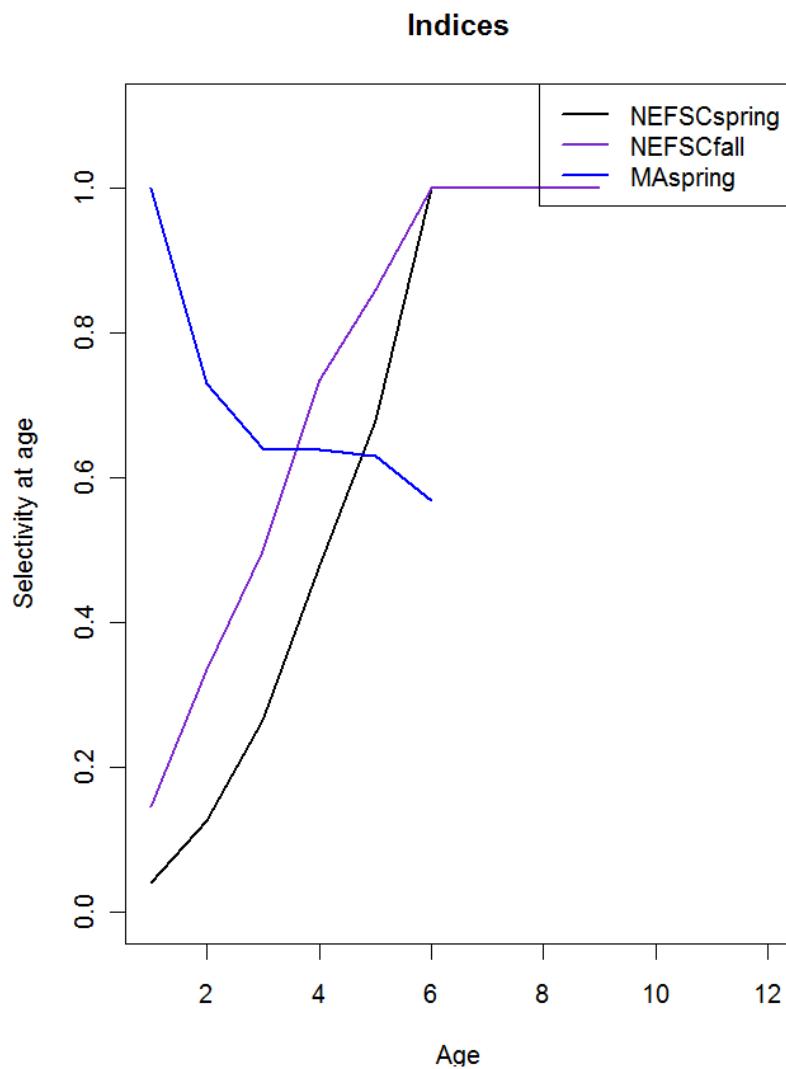


Figure A.177. Gulf of Maine Atlantic cod selectivity-at-age for the NEFSC spring (Index 1), fall (Index 2) and MADMF spring (Index 3) surveys from the ASAP SAW55_3BLOCK_BASE model.

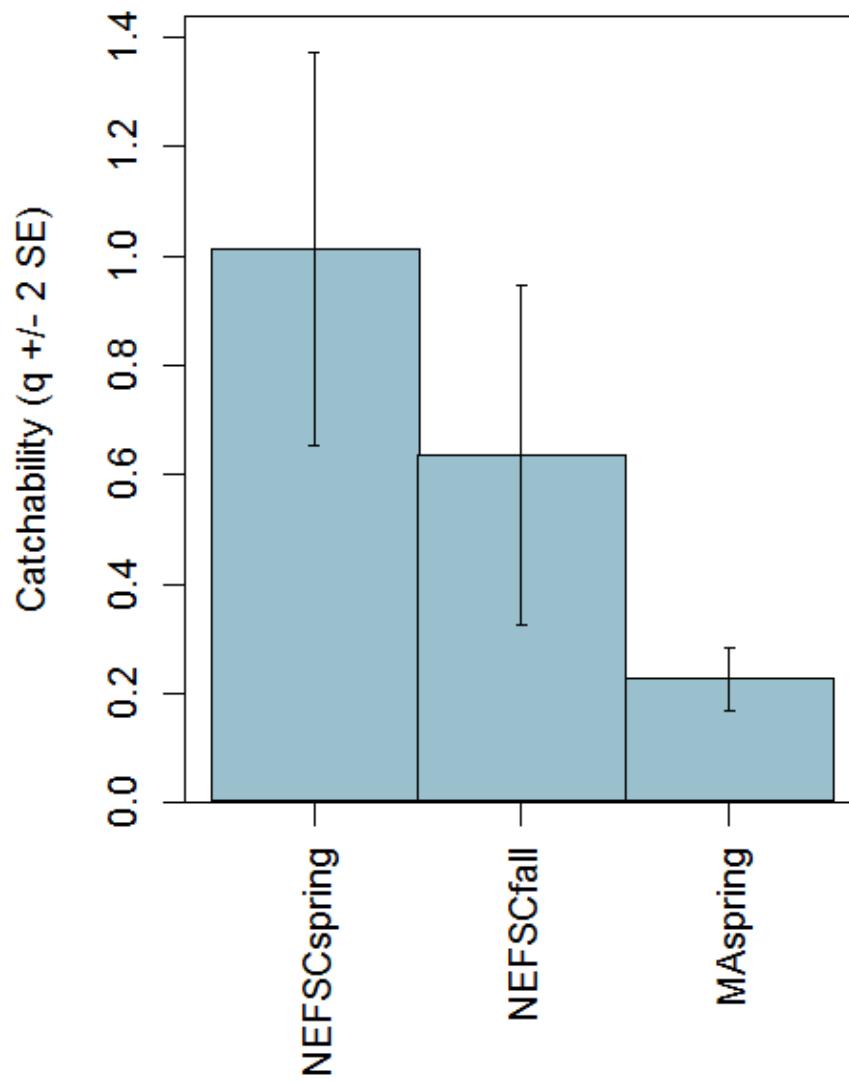


Figure A.178. Gulf of Maine Atlantic cod survey catchability, q , for the NEFSC spring (Index 1), fall (Index 2) and MADMF spring (Index 3) surveys from the ASAP SAW55_3BLOCK_BASE model.

q Retro for NEFSCspring Index

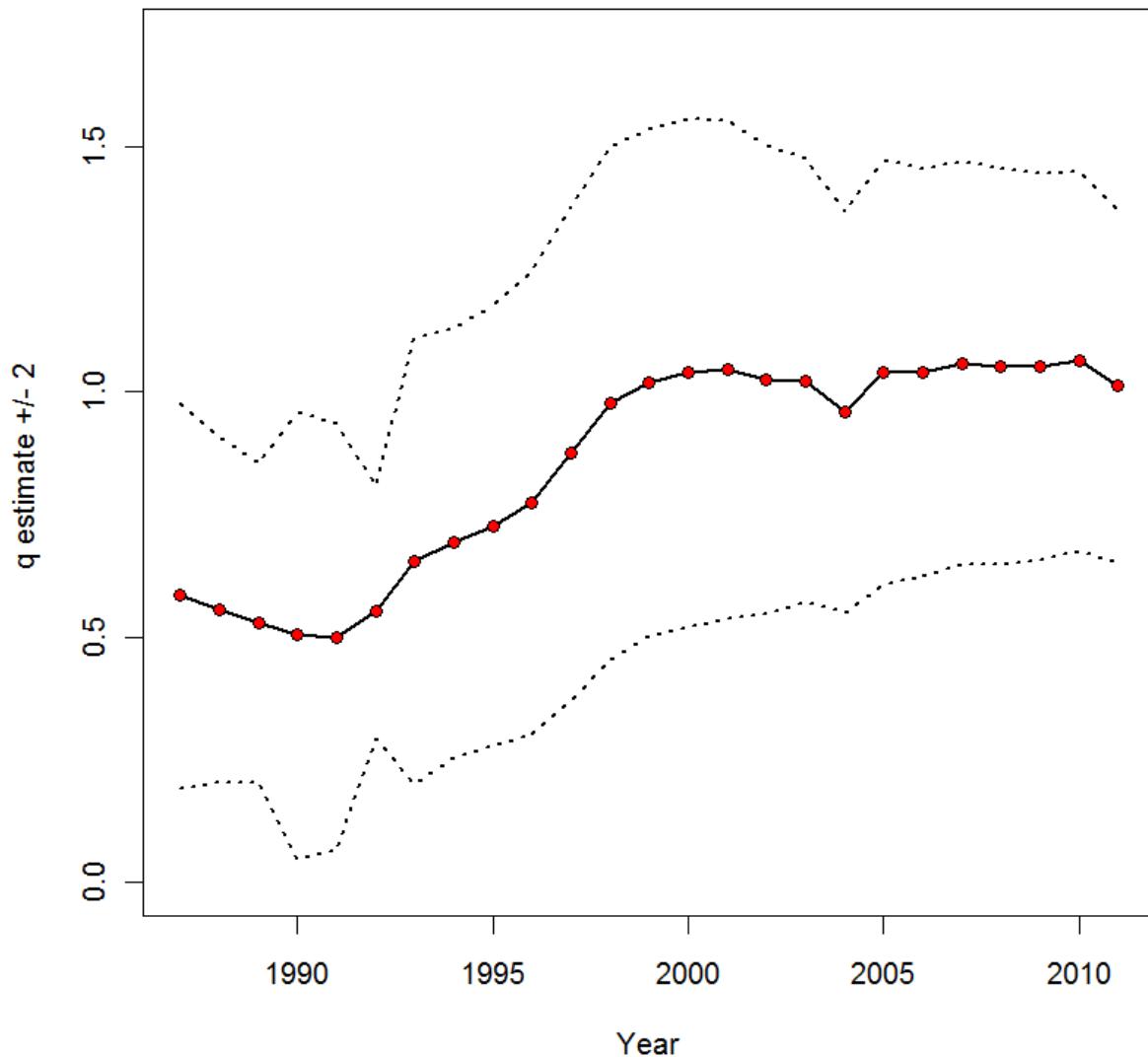


Figure A.179. Retrospective changes in the NEFSC spring survey catchability estimates, q , as years are removed from the ASAP SAW55_3BLOCK_BASE model.

q Retro for NEFSCfall Index

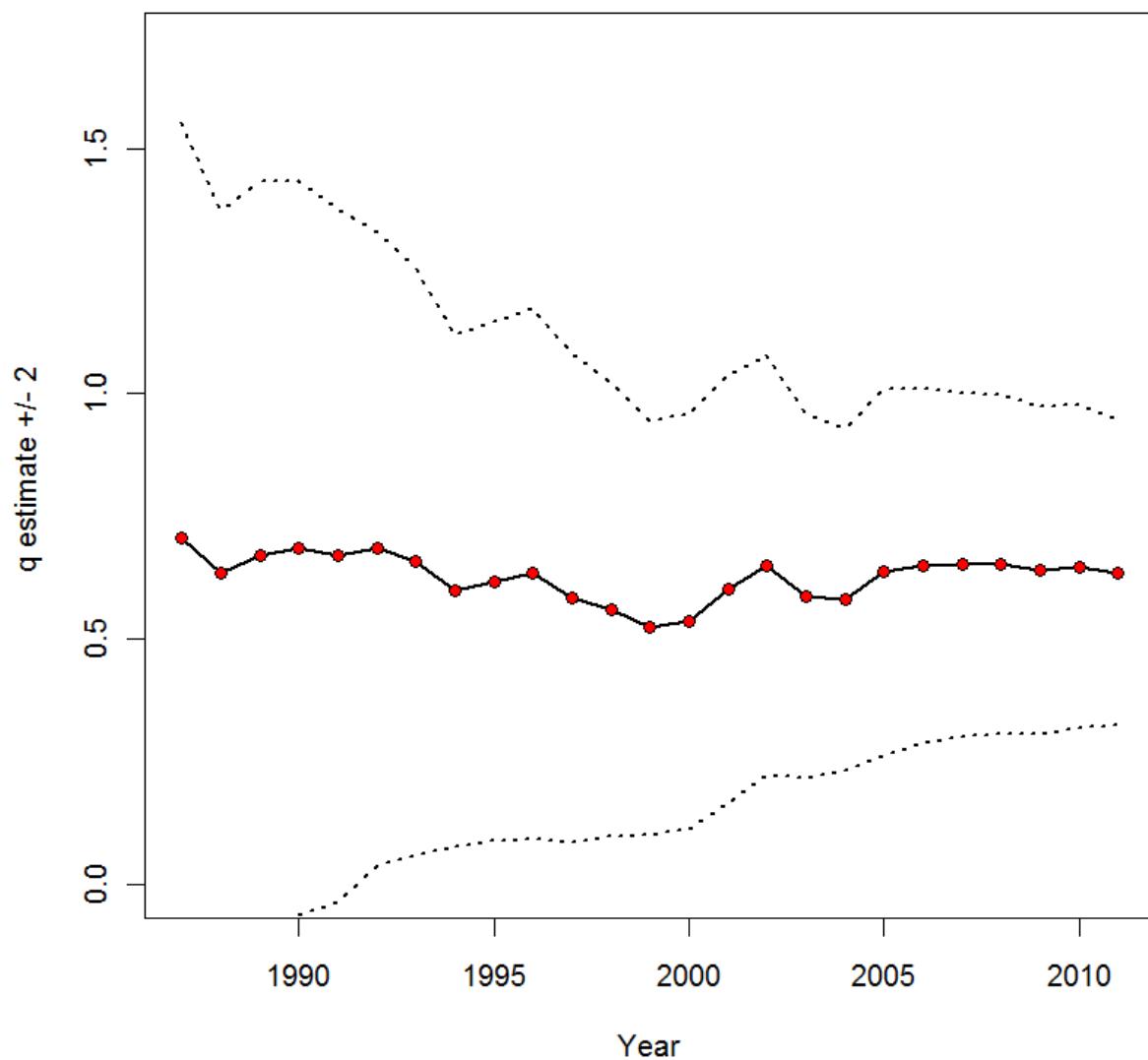


Figure A.180. Retrospective changes in the NEFSC fall survey catchability estimates, q , as years are removed from the ASAP SAW55_3BLOCK_BASE model.

q Retro for MAspring Index

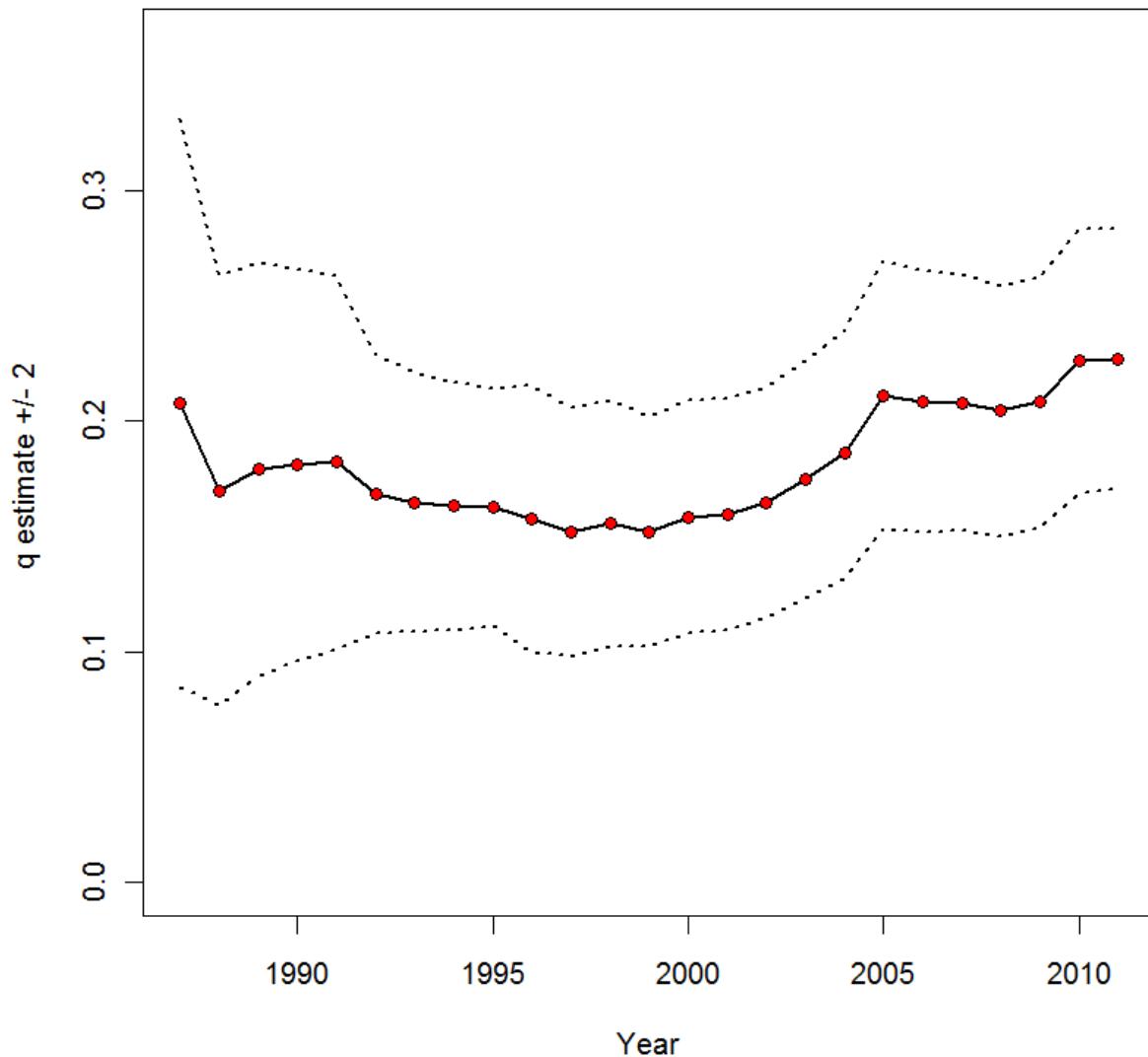


Figure A.181. Retrospective changes in the MADMF spring survey catchability estimates, q , as years are removed from the ASAP SAW55_3BLOCK_BASE model.

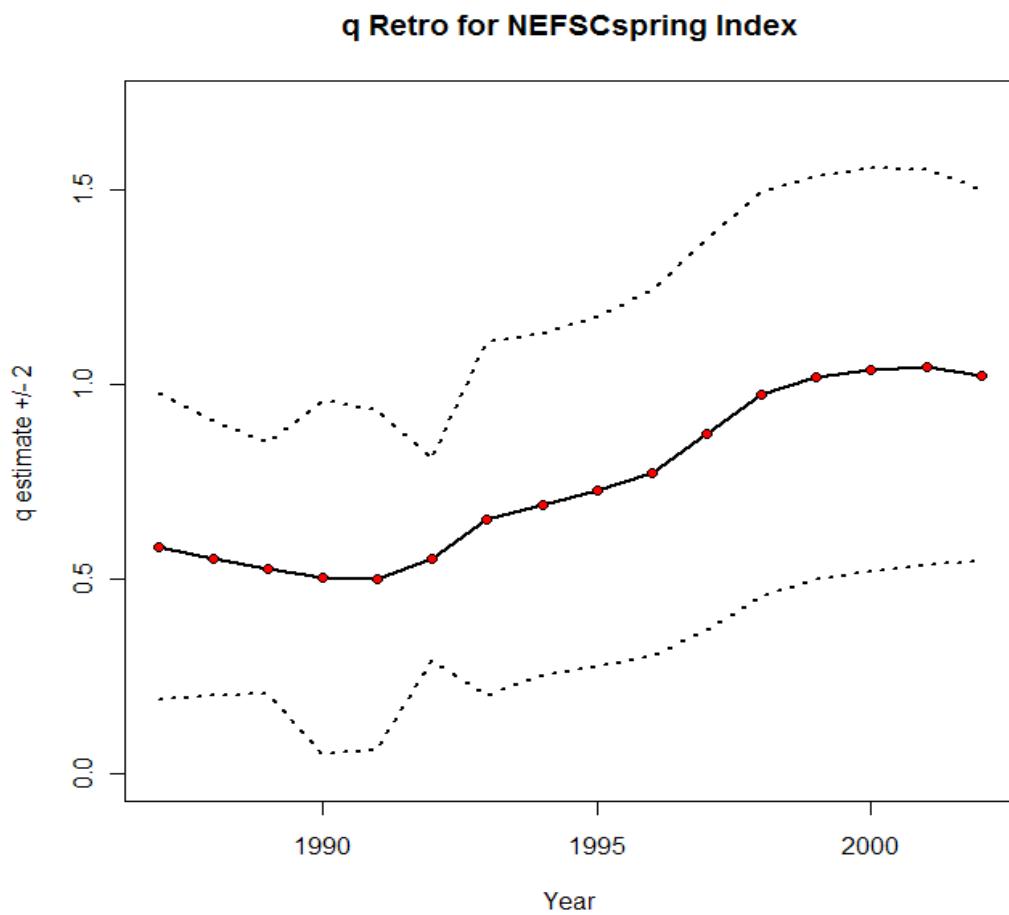


Figure A.182. Retrospective changes in the NEFSC spring survey catchability estimates, q , as years are removed from the ASAP SAW55_3BLOCK_BASE_1982_2002 model.

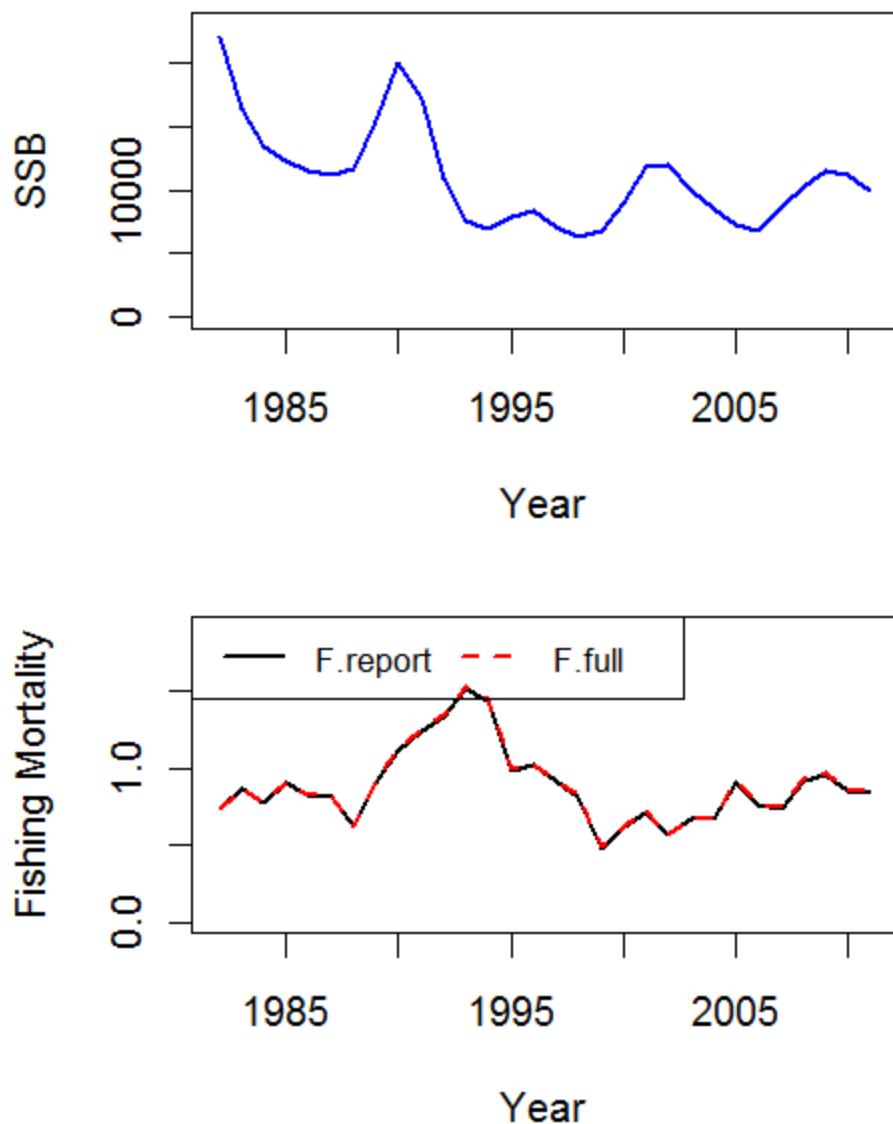


Figure A.183. ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine Atlantic cod spawning stock biomass (SSB) and fishing mortality (F_{full} = fully recruited fishing mortality, F_{report} = fishing mortality on age 5).

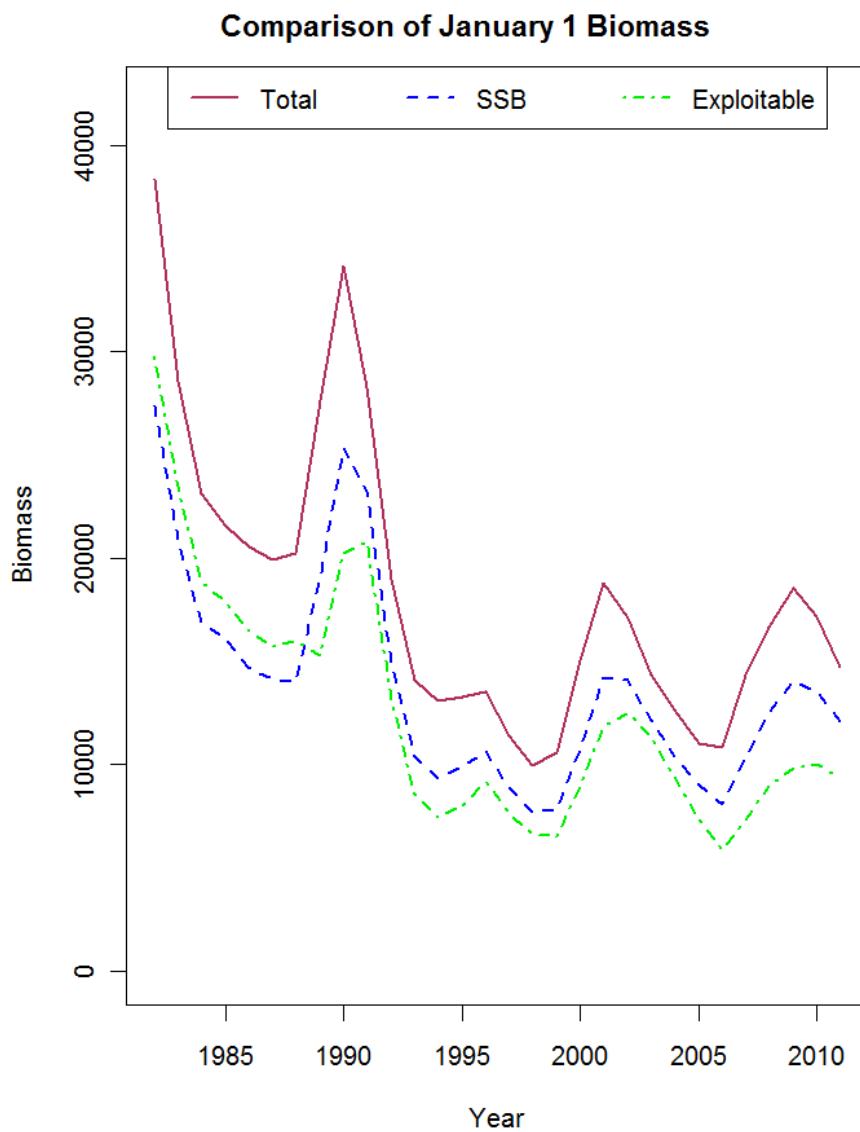


Figure A.184. Comparison of ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine Atlantic cod January 1 biomass after application of maturity ogive (SSB) and fleet selectivity ogives (exploitable).

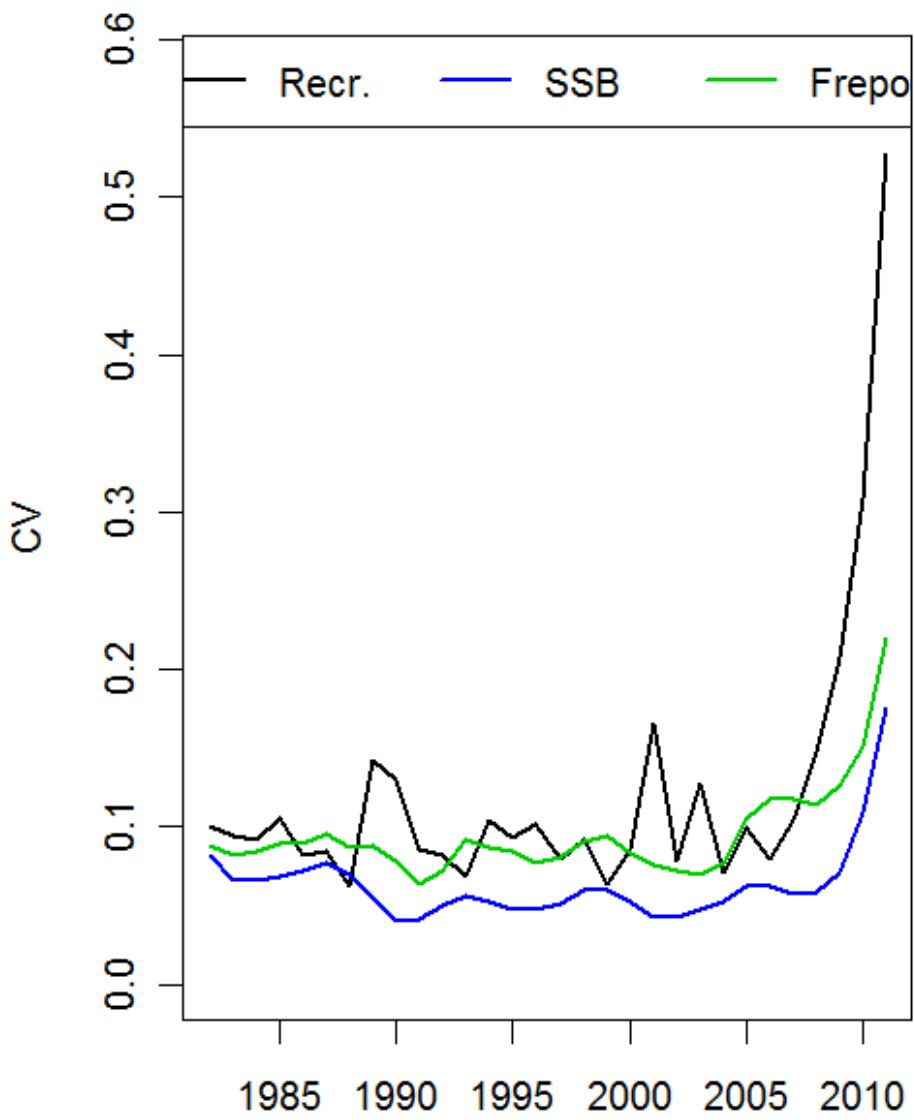


Figure A.185. Coeffecients of variation (CV) for the ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine Atlantic cod spawning stock biomass (SSB), average fishing mortality ($F_{report} = age\ 5\ fishing\ mortality$) and age 1 recruitment.

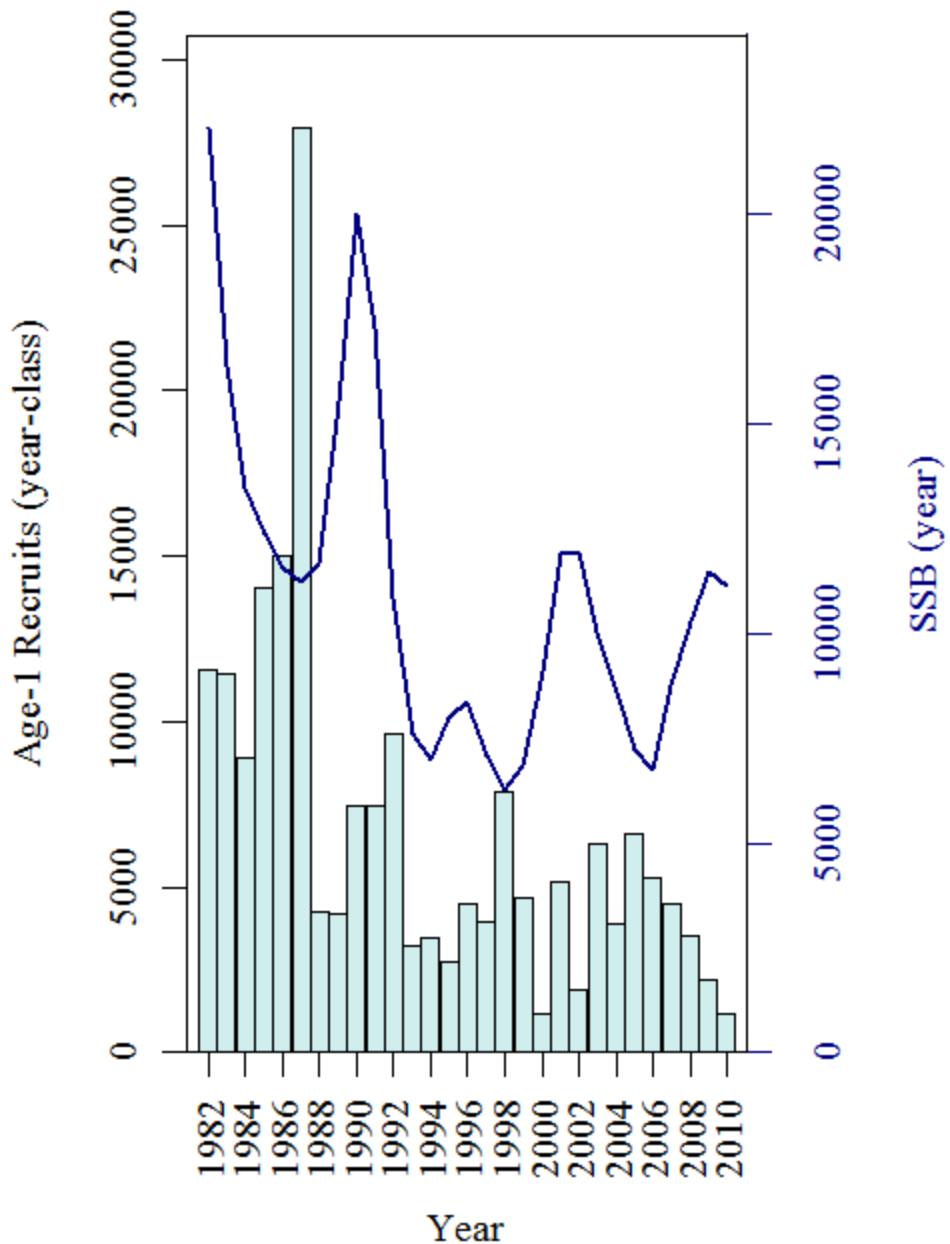


Figure A.186. ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine cod spawning stock biomass (SSB; solid blue line) and lagged age 1 recruitment (light blue bars).

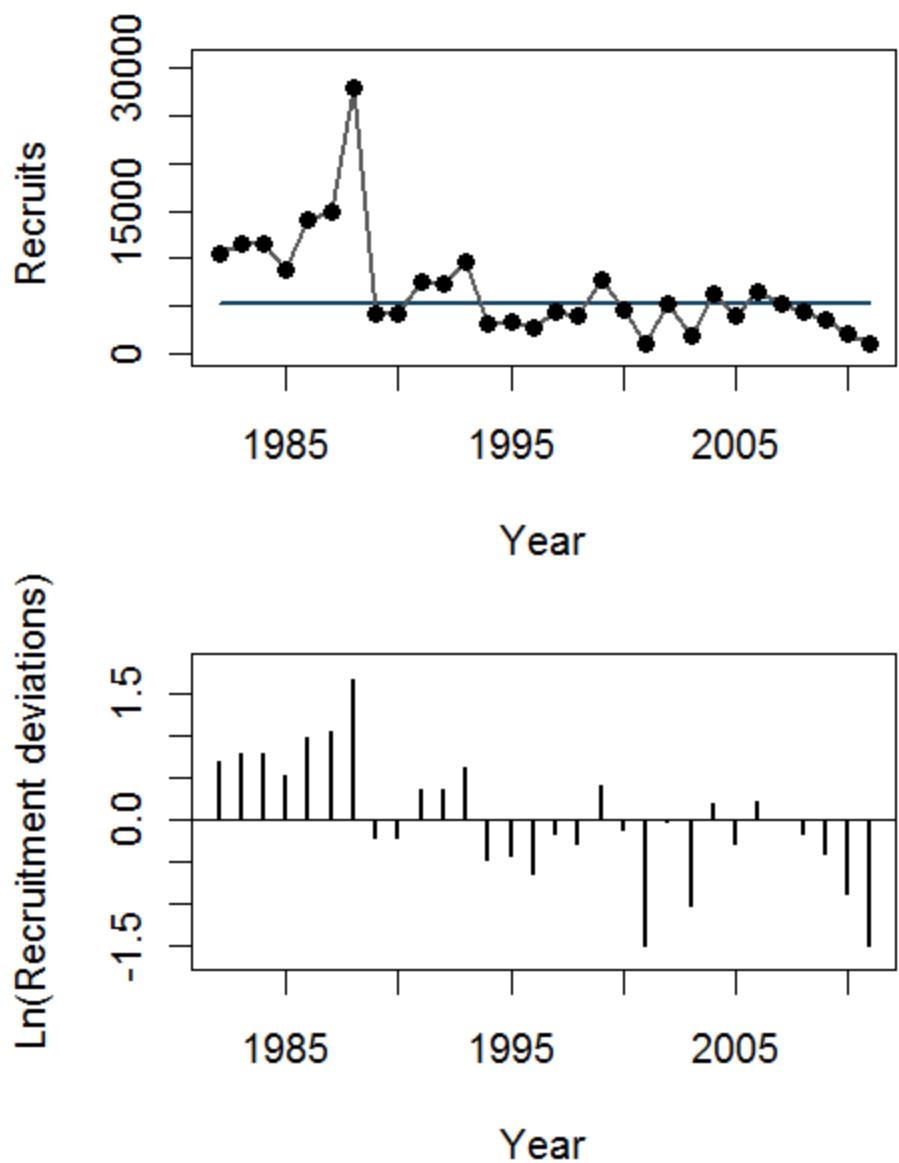


Figure A.187. ASAP SAW55_3BLOCK_BASE estimated Gulf of Maine Atlantic cod age 1 recruitment and recruitment residuals from the geometric mean.

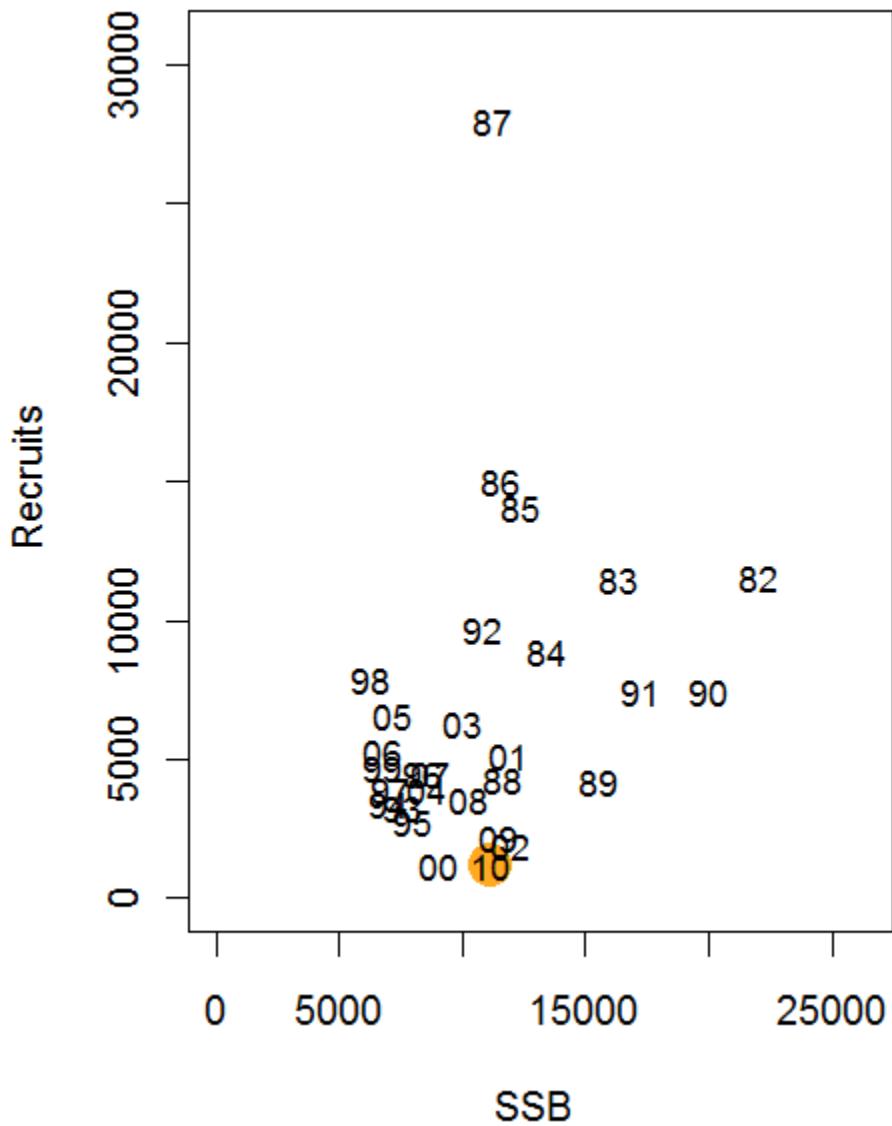


Figure A.188. Scatterplot of ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine Atlantic cod spawning stock biomass (SSB) versus recruitment at age 1 (000s). The symbol for each observation is the last two digits of the year (e.g., 88 indicated age 1 estimates of the 1987 year class). The most recent recruitment estimate is highlighted by an orange circle.

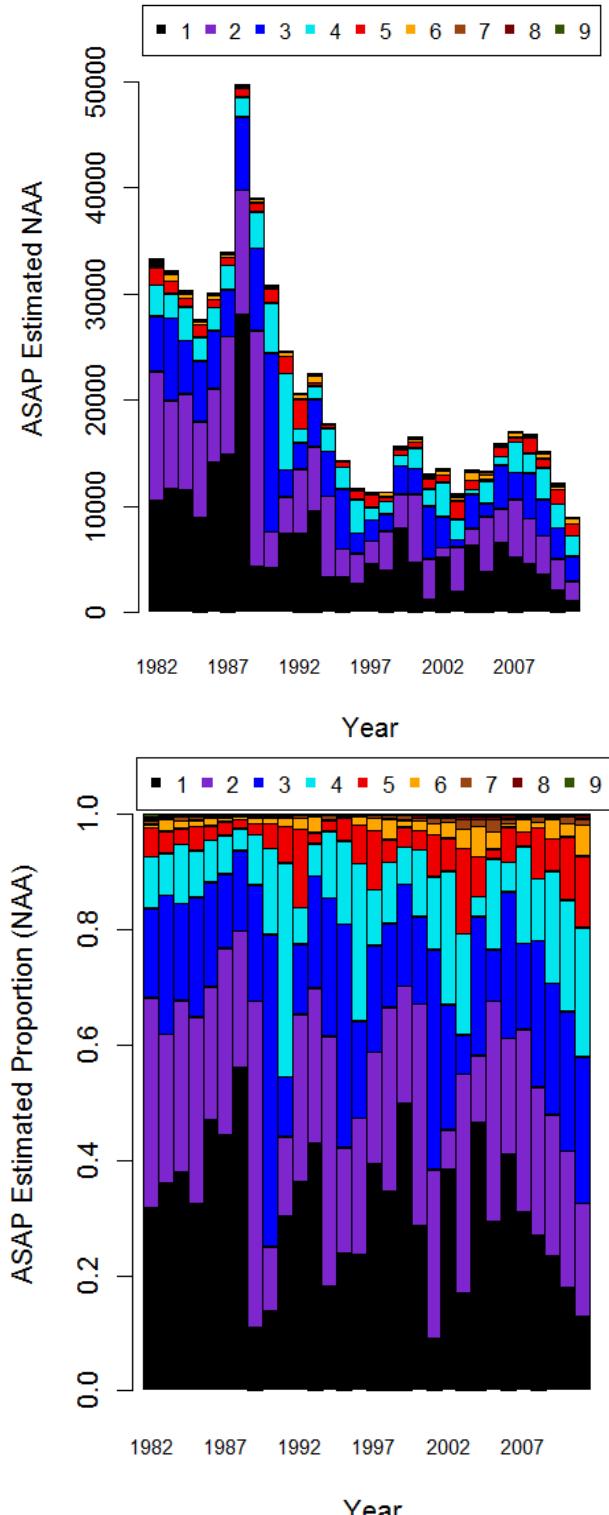


Figure A.189. ASAP SAW55_3BLOCK_BASE model estimates of Gulf of Maine Atlantic cod numbers-at-age in absolute (top) numbers (000s) and relative (bottom) terms.

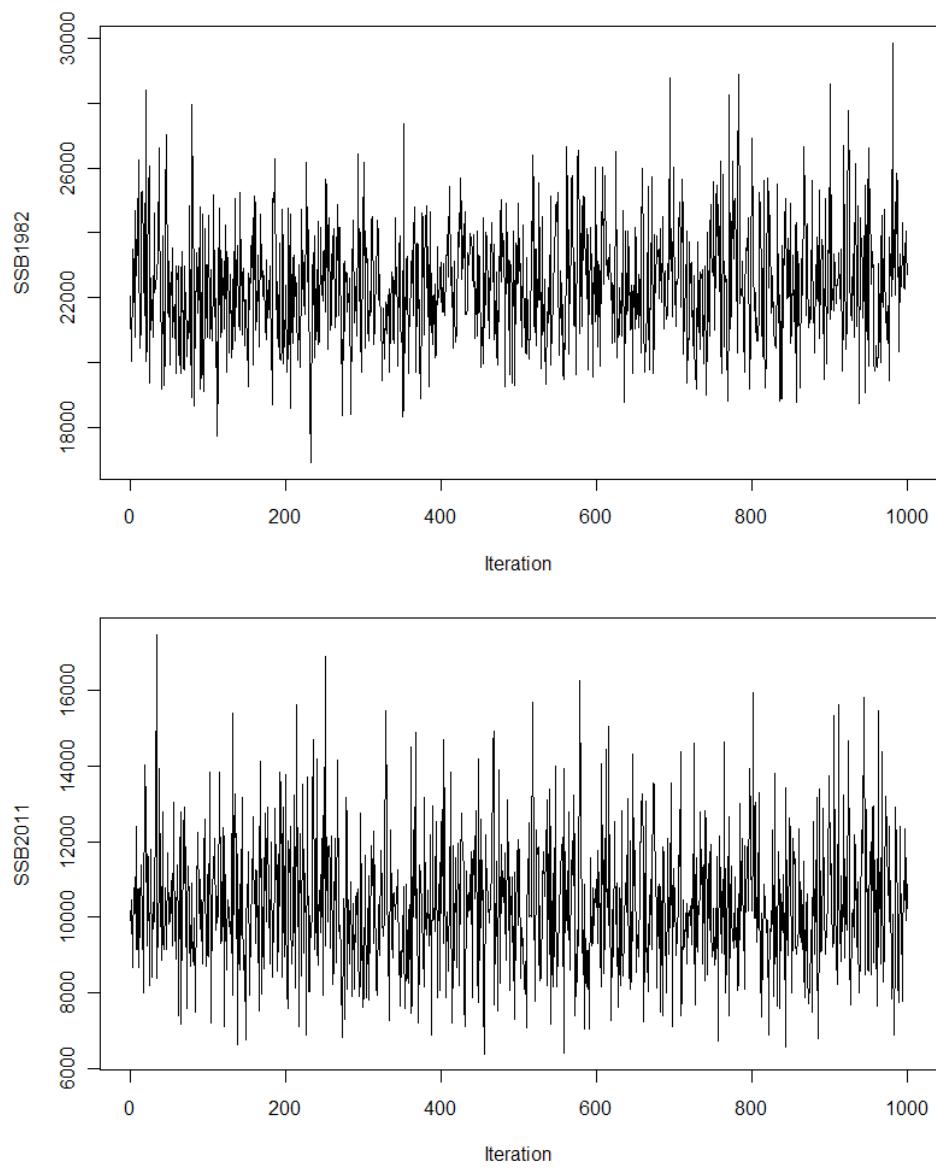


Figure A.190. Trace of MCMC chains for Gulf of Maine Atlantic cod 1982 and 2011 spawning stock biomass, showing good mixing (ASAP SAW55_3BLOCK_BASE model). Each chain had initial length of 1,000,000 and was thinned at a rate of one out of every 1000th resulting in a final chain length of 1000.

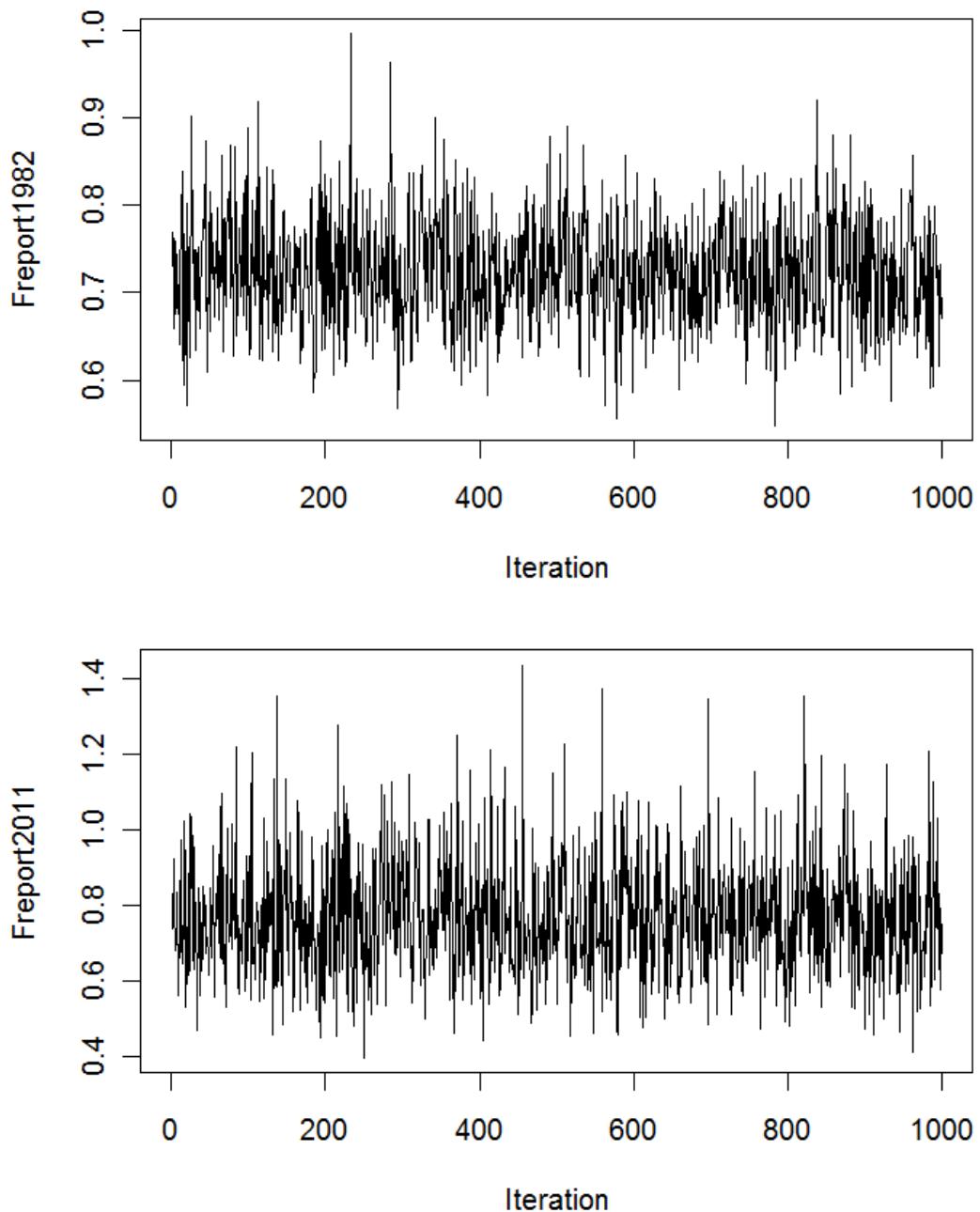


Figure A.191. Trace of MCMC chains for Gulf of Maine Atlantic cod 1982 and 2011 fishing mortality at age 5 (Freport), showing good mixing (ASAP SAW55_3BLOCK_BASE model). Each chain had initial length of 1,000,000 and was thinned at a rate of one out of every 1000th resulting in a final chain length of 1000.

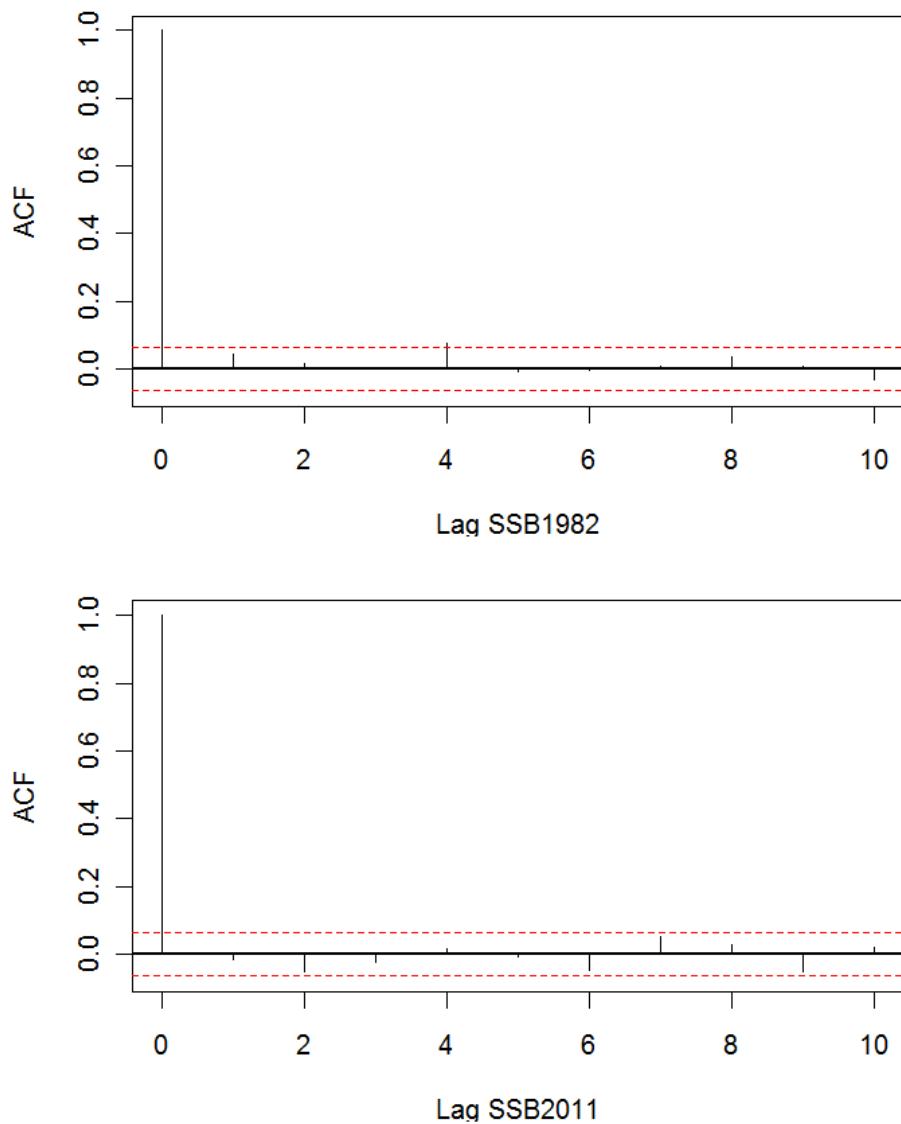


Figure A.192. Autocorrelation within the 1982 and 2011 Gulf of Maine Atlantic cod spawning stock biomass (SSB) MCMC chains from the ASAP SAW55_3BLOCK_BASE model.

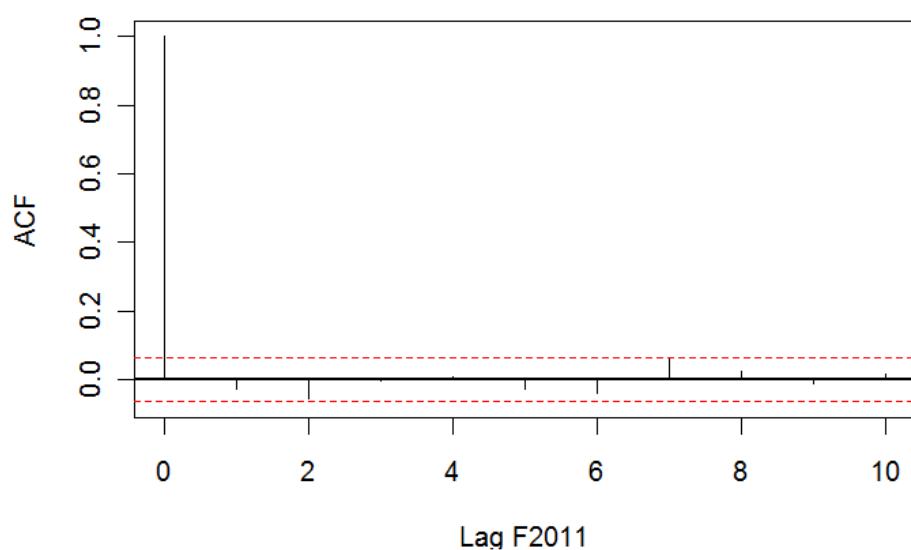
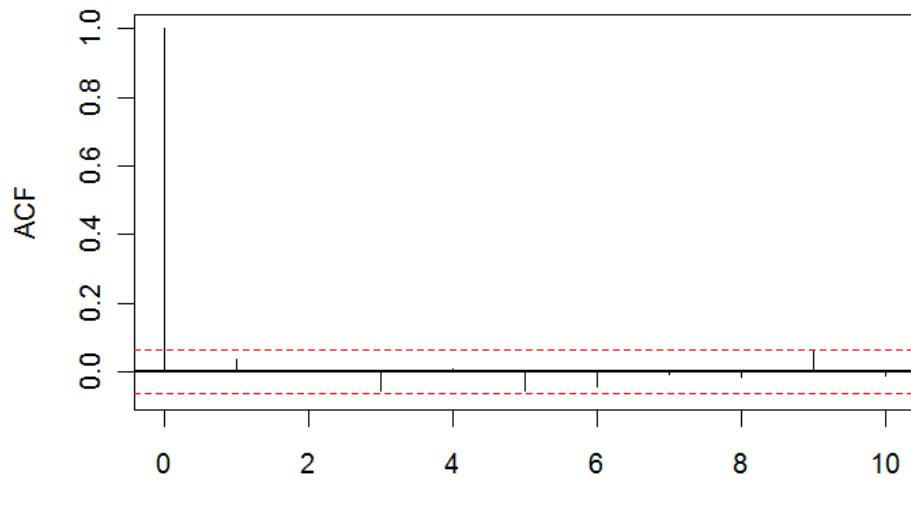


Figure A.193. Autocorrelation within the 1982 and 2011 Gulf of Maine Atlantic cod fishing mortality at age 5 (Freport) MCMC chains from the ASAP SAW55_3BLOCK_BASE model.

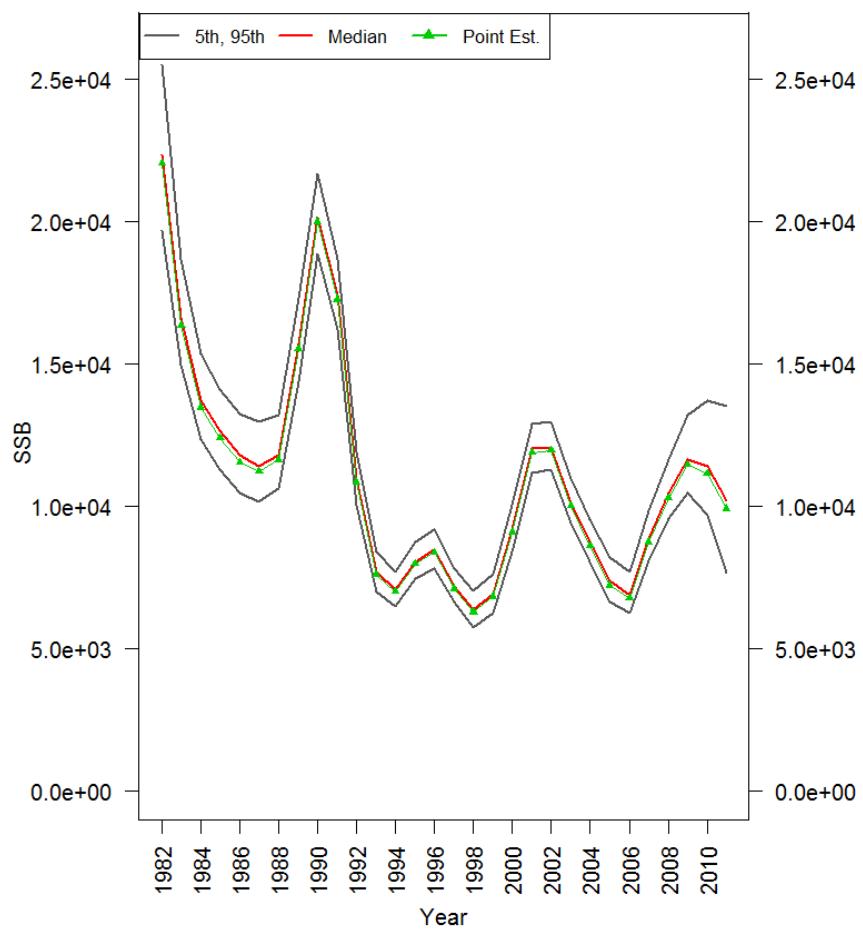


Figure A.194. 90% probability interval for Gulf of Maine Atlantic cod spawning stock biomass (SSB) from the ASAP SAW55_3BLOCK_BASE model. The median value is in red, while the 5th and 95th percentiles are in dark grey. The point estimate from the base model (joint posterior modes) is shown in the thin green line with filled triangles.

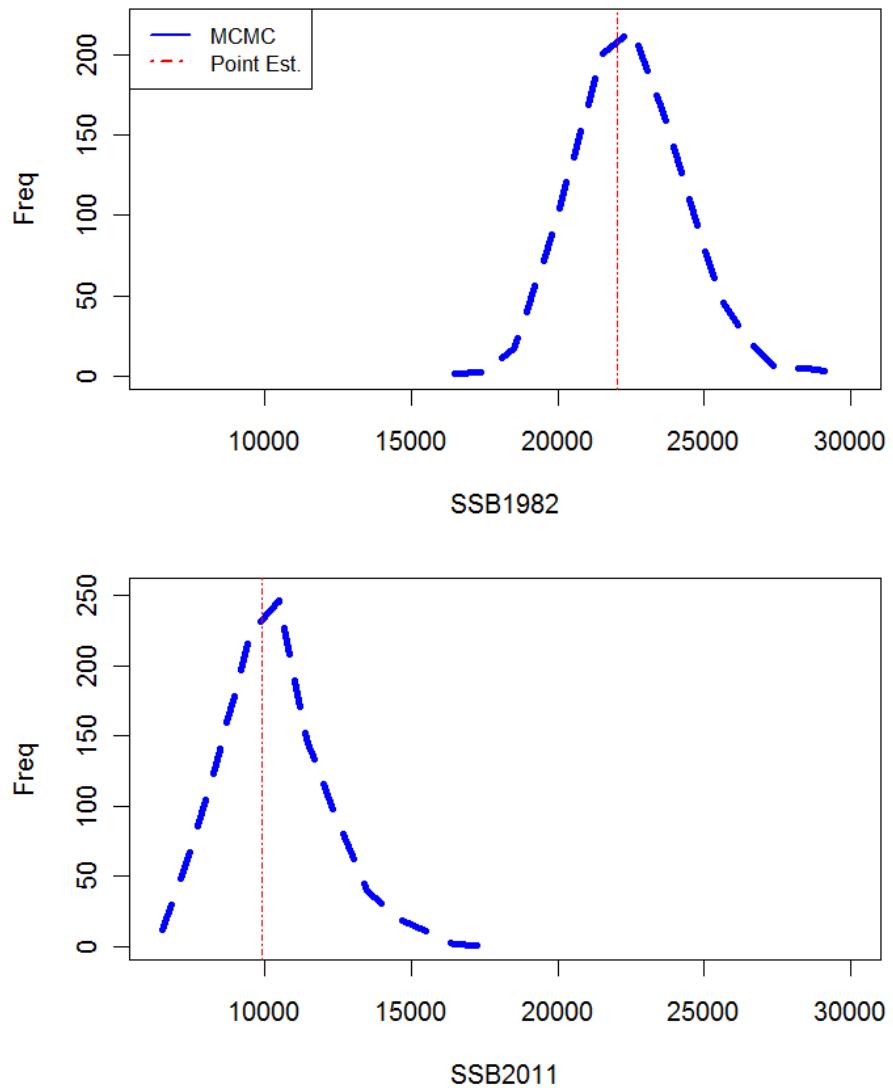


Figure A.195. MCMC distribution of Gulf of Maine Atlantic cod spawning stock biomass in 1982 and 2011 estimated from the ASAP SAW55_3BLOCK_BASE model. The model point estimate is indicated by the dashed red line.

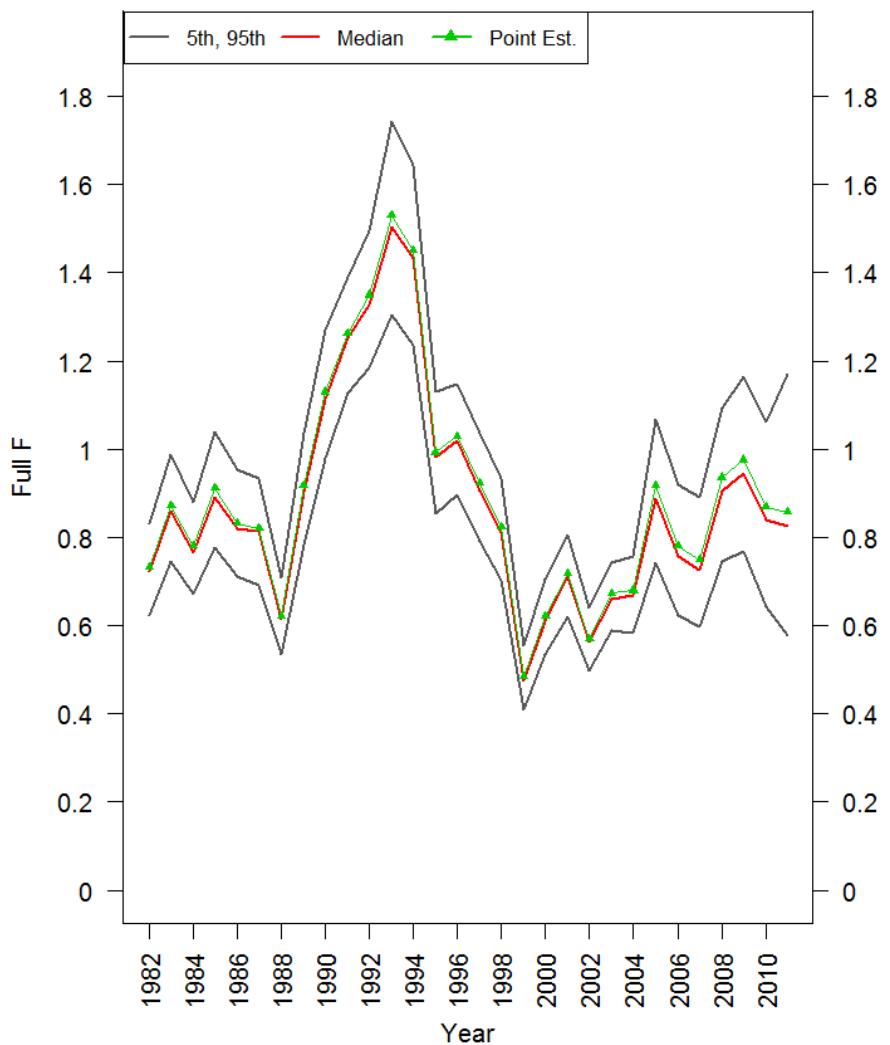


Figure A.196. 90% probability interval for Gulf of Maine Atlantic fully recruited fishing mortality (Full F) from the ASAP SAW55_3BLOCK_BASE model. The median value is in red, while the 5th and 95th percentiles are in dark grey. The point estimate from the base model (joint posterior modes) is shown in the thin green line with filled triangles.

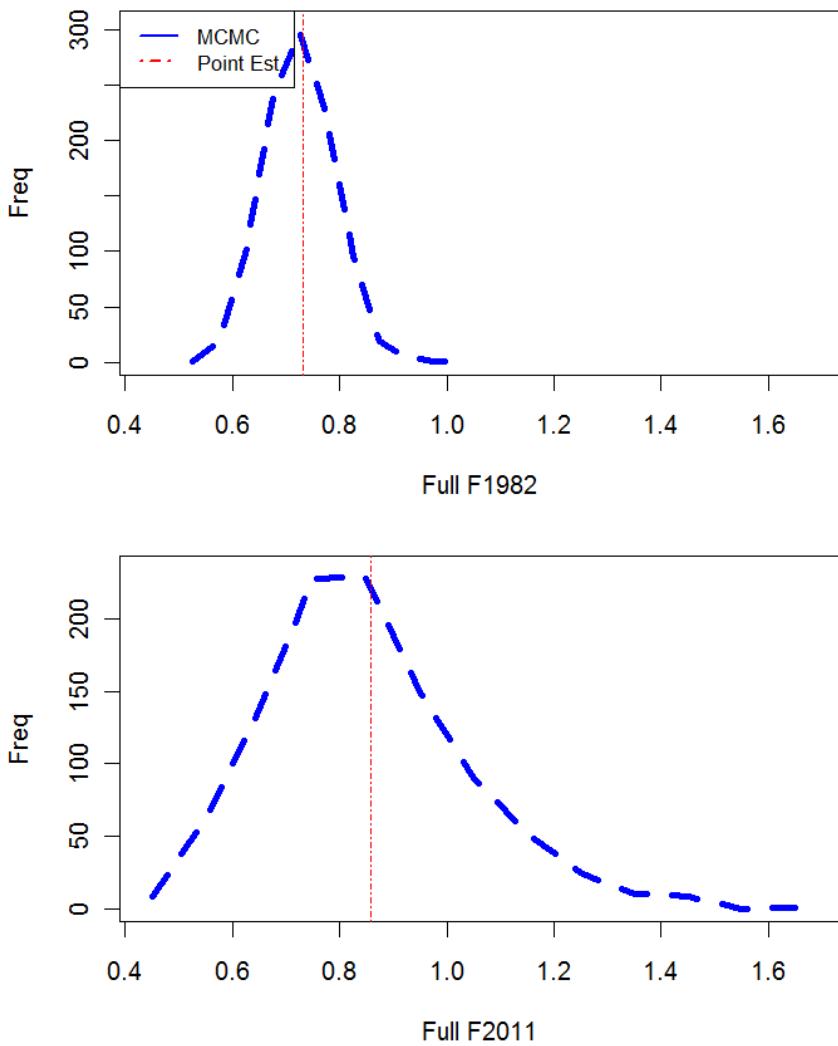


Figure A.197. MCMC distribution of Gulf of Maine Atlantic cod fully recruited fishing mortality (Full F) in 1982 and 2011 estimated from the ASAP SAW55_3BLOCK_BASE model. The model point estimate is indicated by the dashed red line.

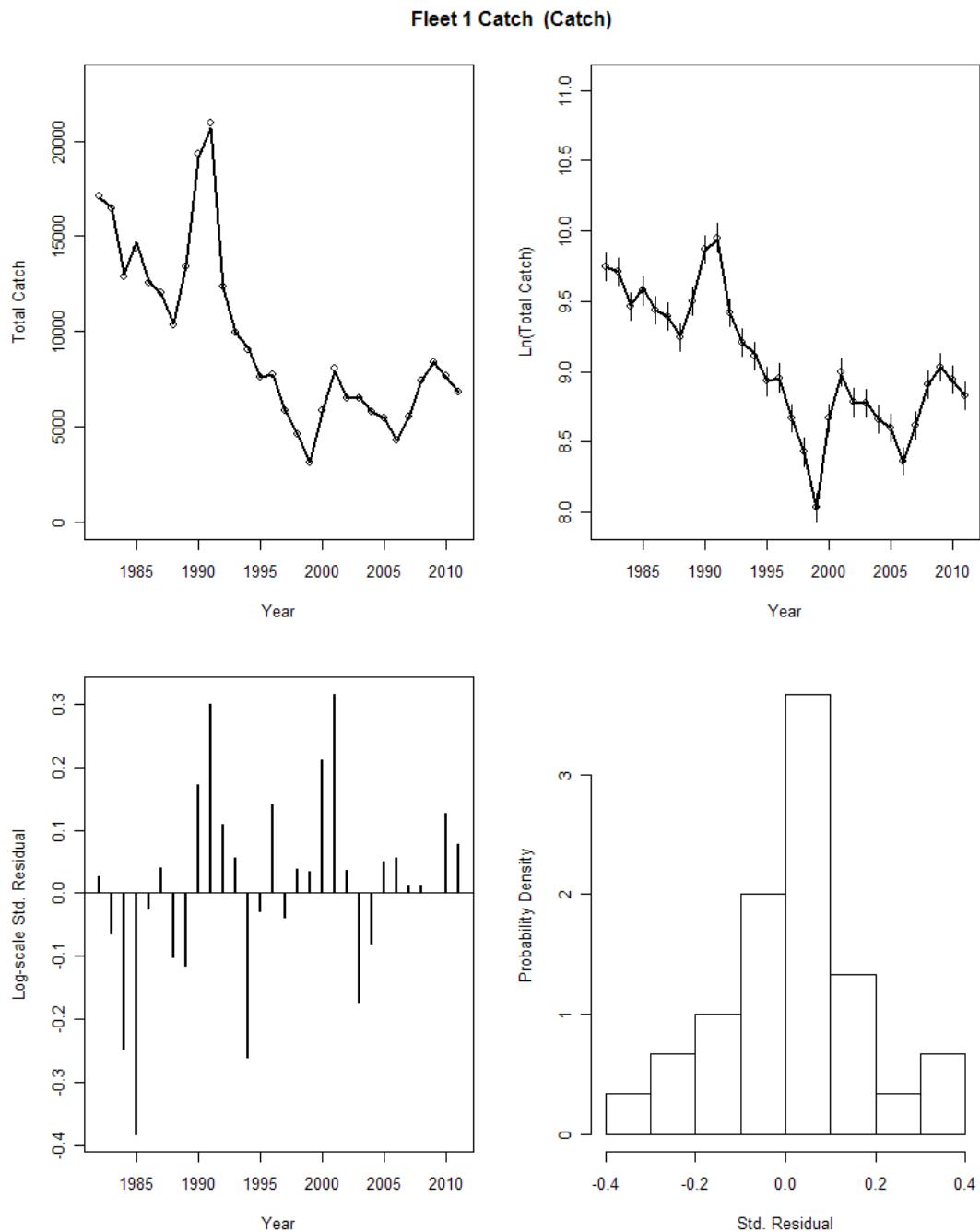


Figure A.198. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit to the total Gulf of Maine Atlantic cod fishery catch (Fleet 1).

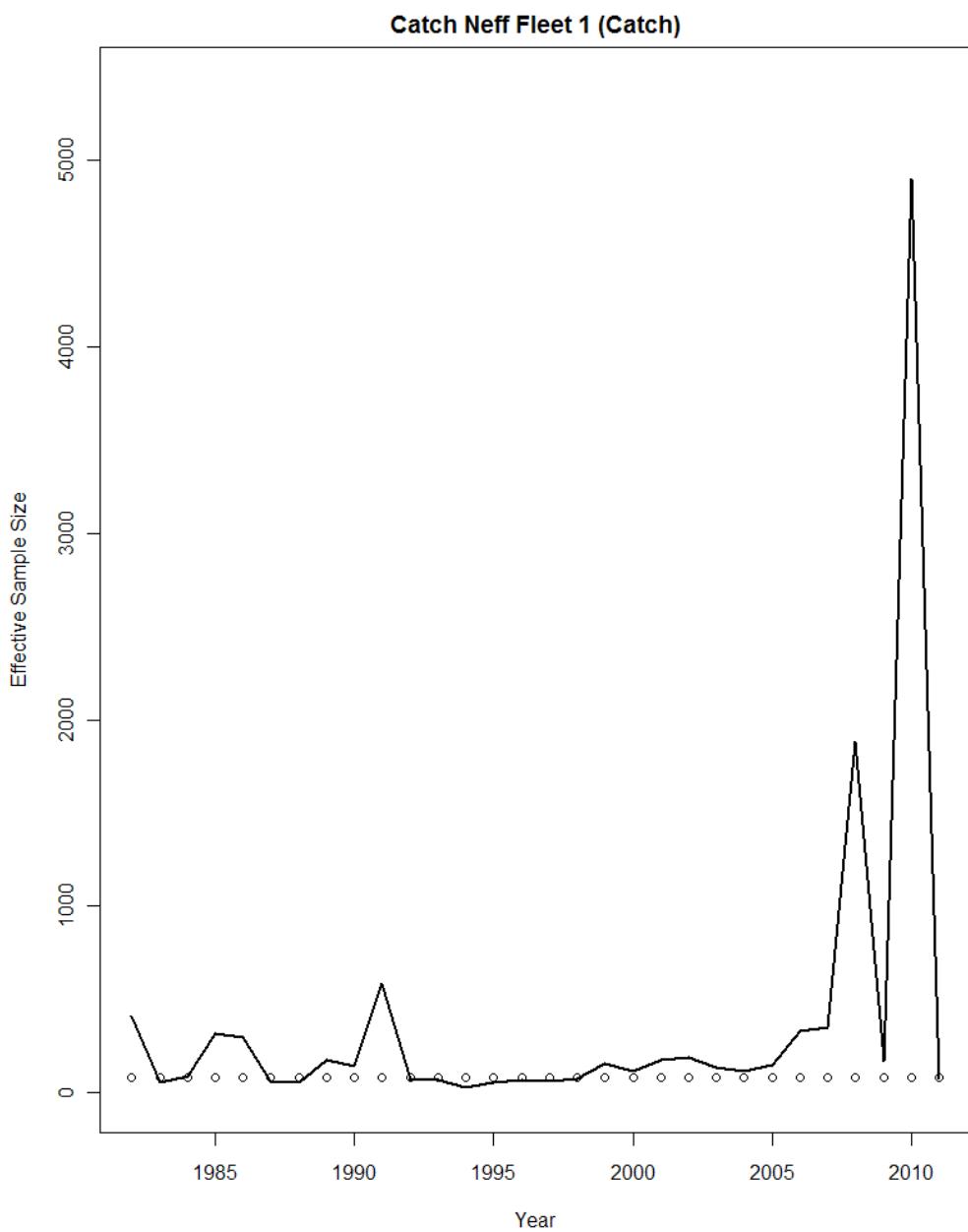


Figure A.199. ASAP SAW55_3BLOCK_BASE_M_SPLIT model comparison of input effective sample size versus the model estimated effective sample size for the Gulf of Maine Atlantic cod fishery catch.

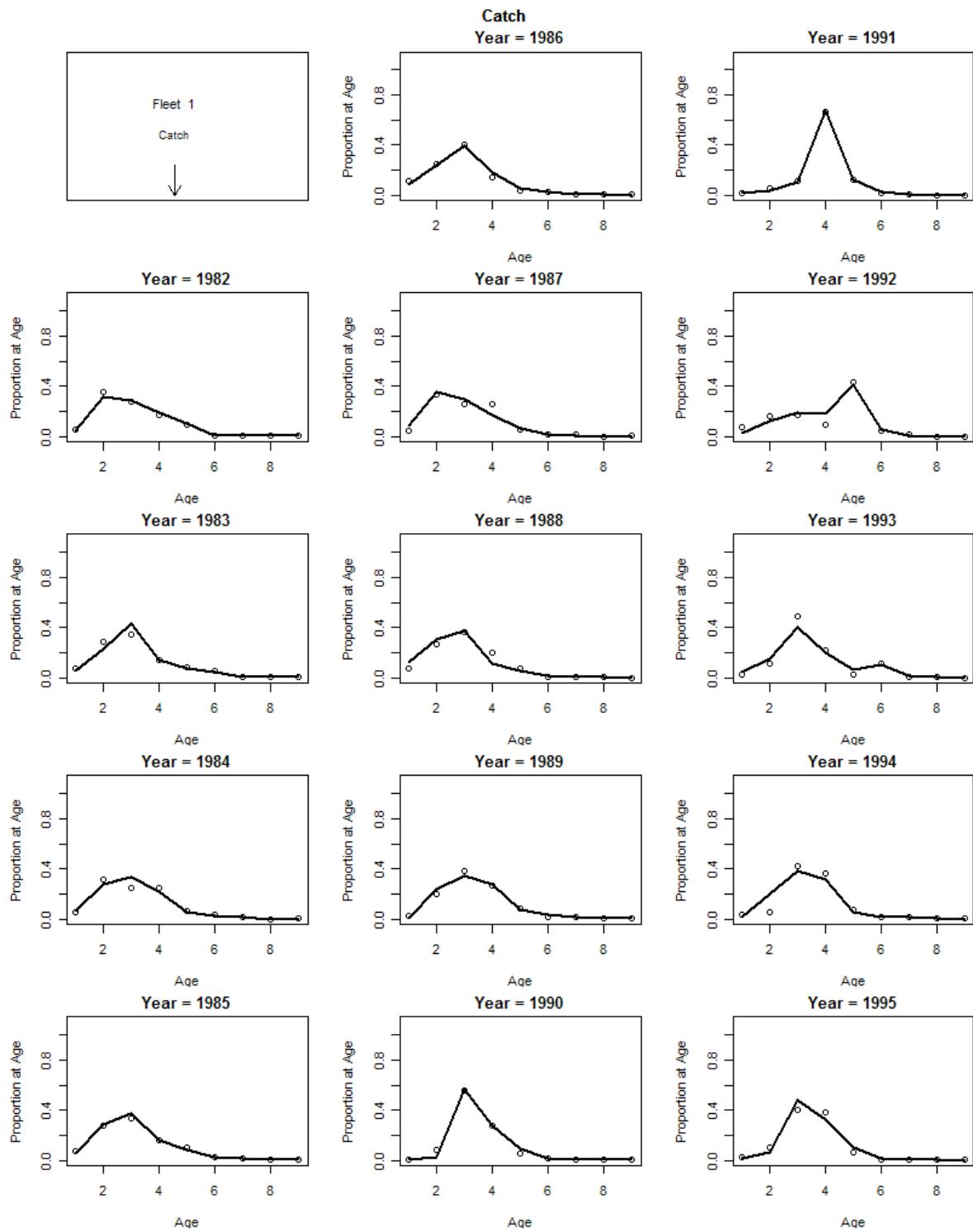


Figure A.200.a. Comparison of the ASAP SAW55_3BLOCK_BASE_M_SPLIT estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

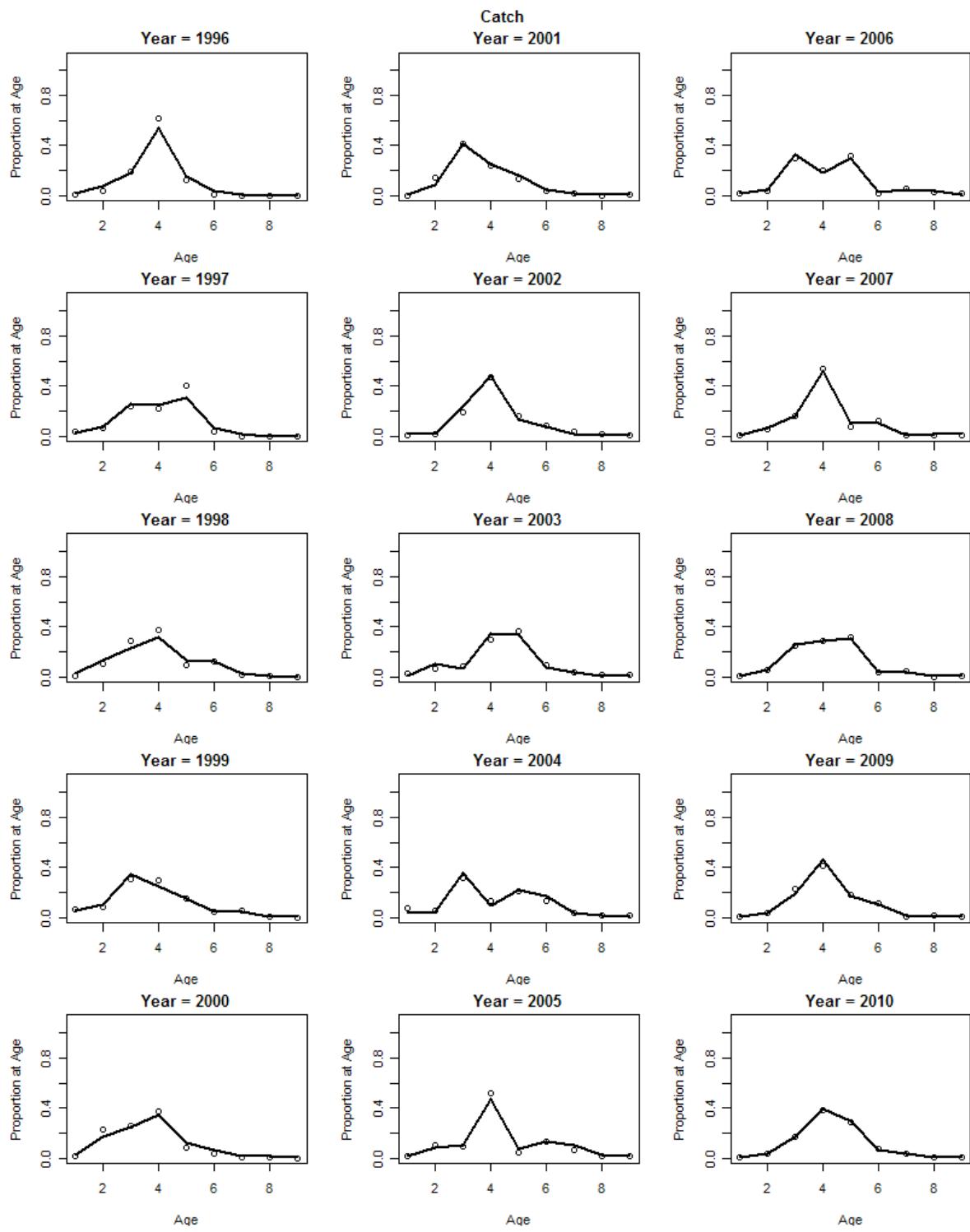


Figure A.200.b. Comparison of the ASAP SAW55_3BLOCK_BASE_M_SPLIT estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

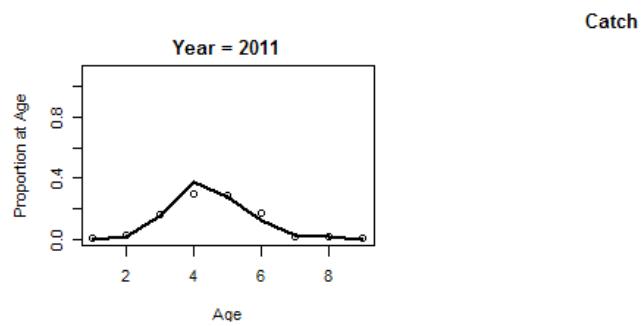


Figure A.200.c. Comparison of the ASAP SAW55_3BLOCK_BASE_M_SPLIT estimates of Gulf of Maine Atlantic cod proportion-at-age in the fishery to the data estimates.

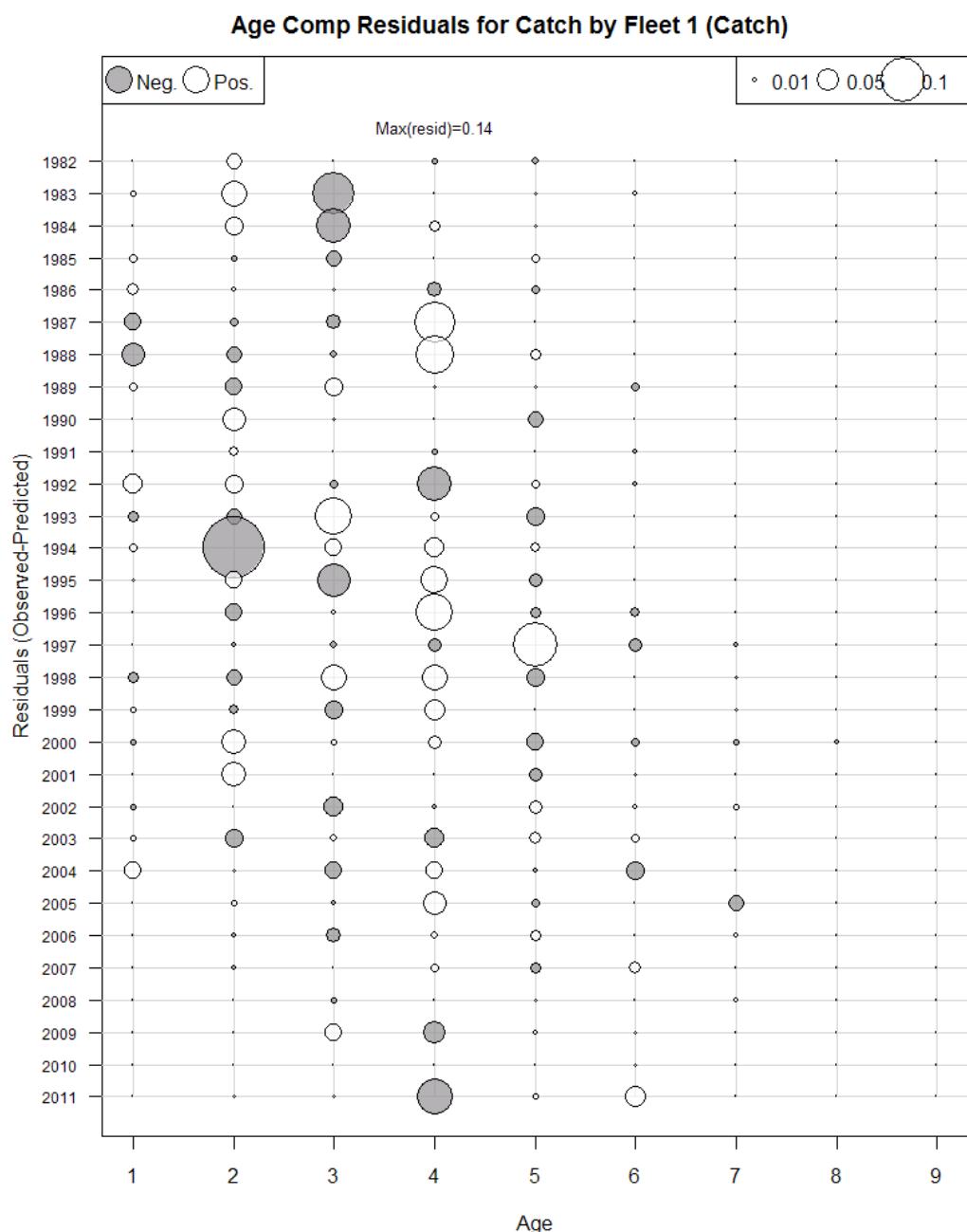


Figure A.201. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit residuals for the fishery (Fleet 1) catch-at-age of Gulf of Maine Atlantic cod.

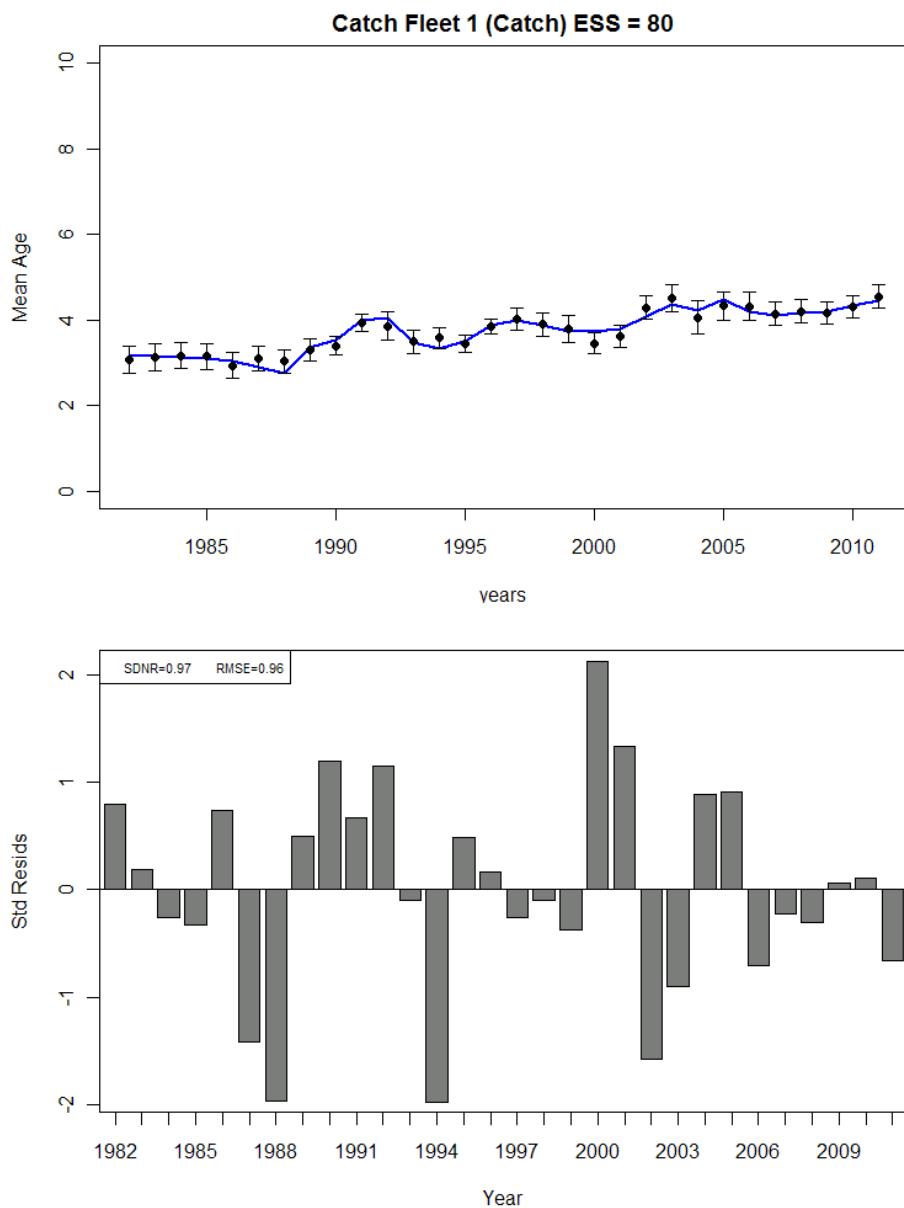


Figure A.202. ASAP SAW55_3BLOCK_BASE_M_SPLIT predicted mean age of Gulf of Maine Atlantic cod in the fishery catch (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

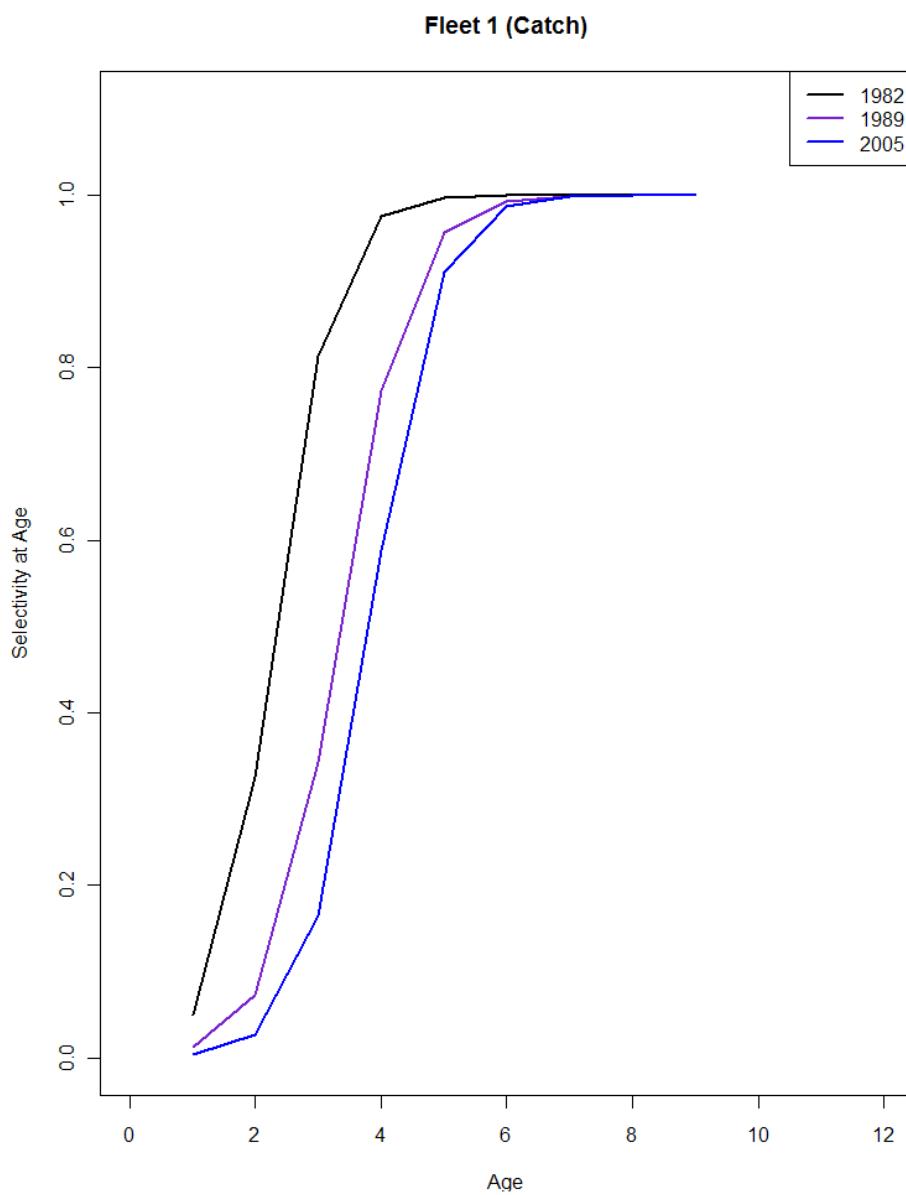


Figure A.203. Gulf of Maine Atlantic cod fishery selectivity blocks for block 1 (1982-1989), block 2 (1990-2004) and block 3 (2005-2011) estimated by the ASAP SAW55_3BLOCK_BASE_M_SPLIT model.

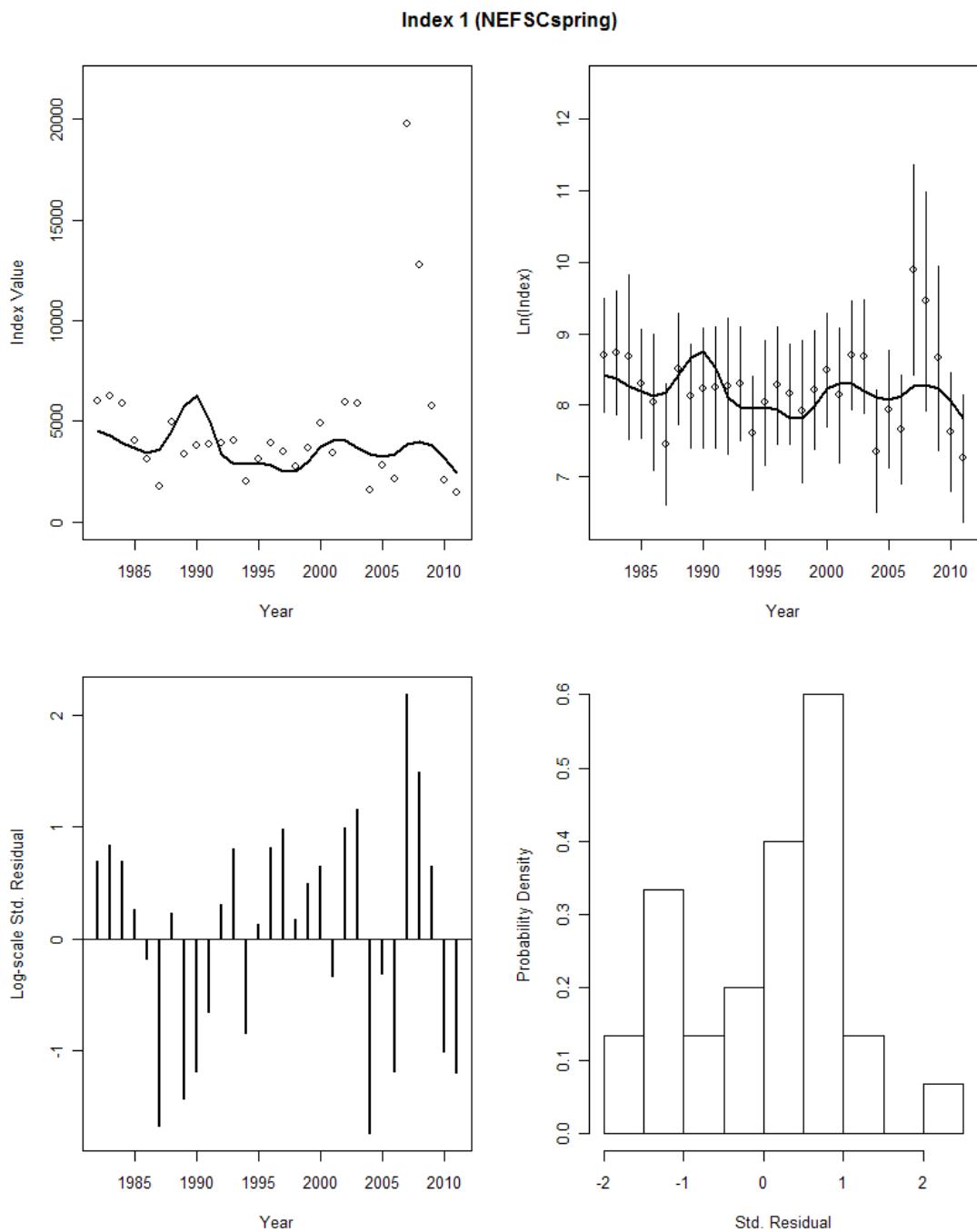


Figure A.204. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit to the NEFSC Gulf of Maine Atlantic cod spring (Index 1) survey.

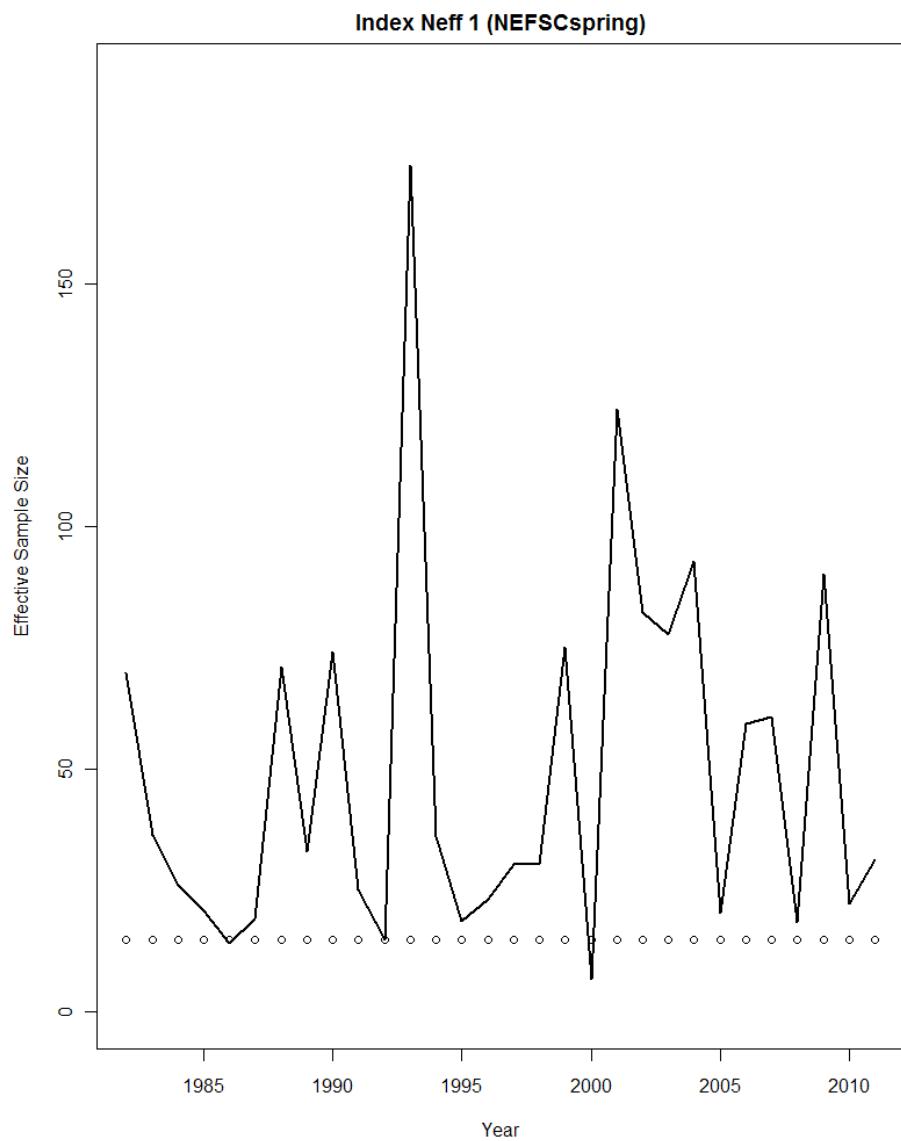


Figure A.205. ASAP SAW55_3BLOCK_BASE_M_SPLIT model comparison of input effective sample size versus the model estimated effective sample size for the NEFSC spring (Index 1) Gulf of Maine Atlantic cod index.

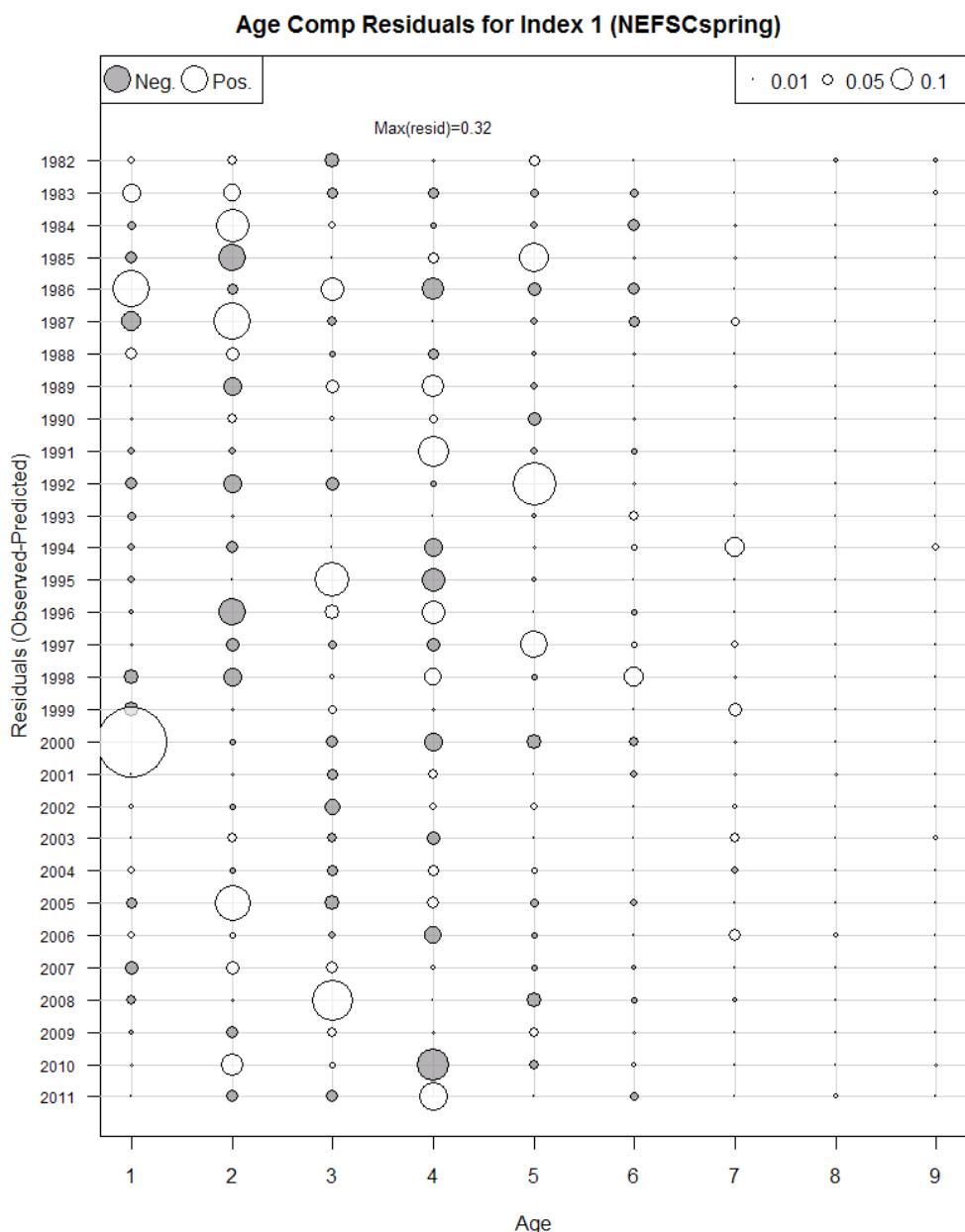


Figure A.206. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit residuals for the NEFSC spring survey (Index 1) Gulf of Maine Atlantic cod age composition.

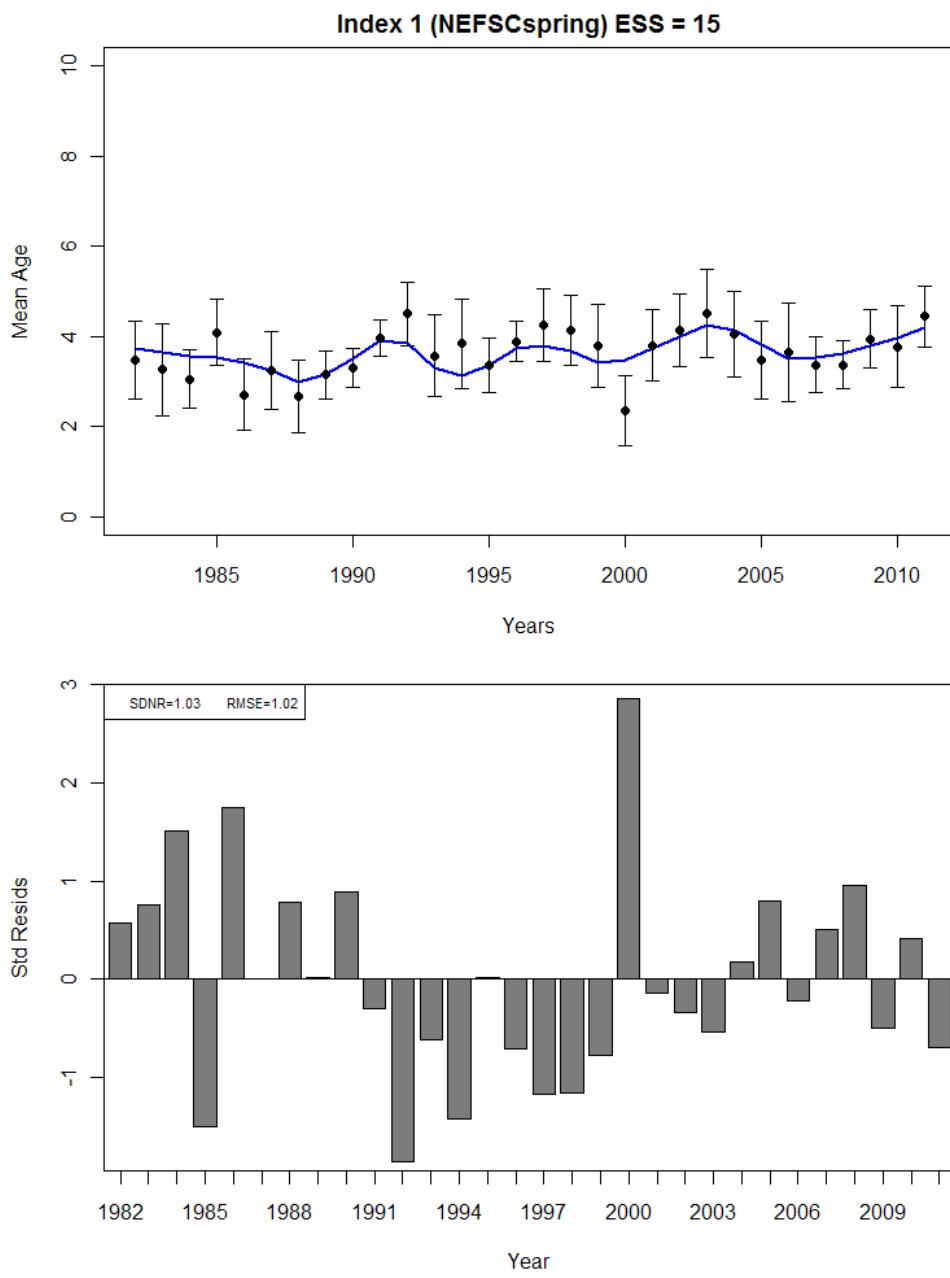


Figure A.207. ASAP SAW55_3BLOCK_BASE_M_SPLIT predicted mean age of Gulf of Maine Atlantic cod in the NEFSC spring (Index 1) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

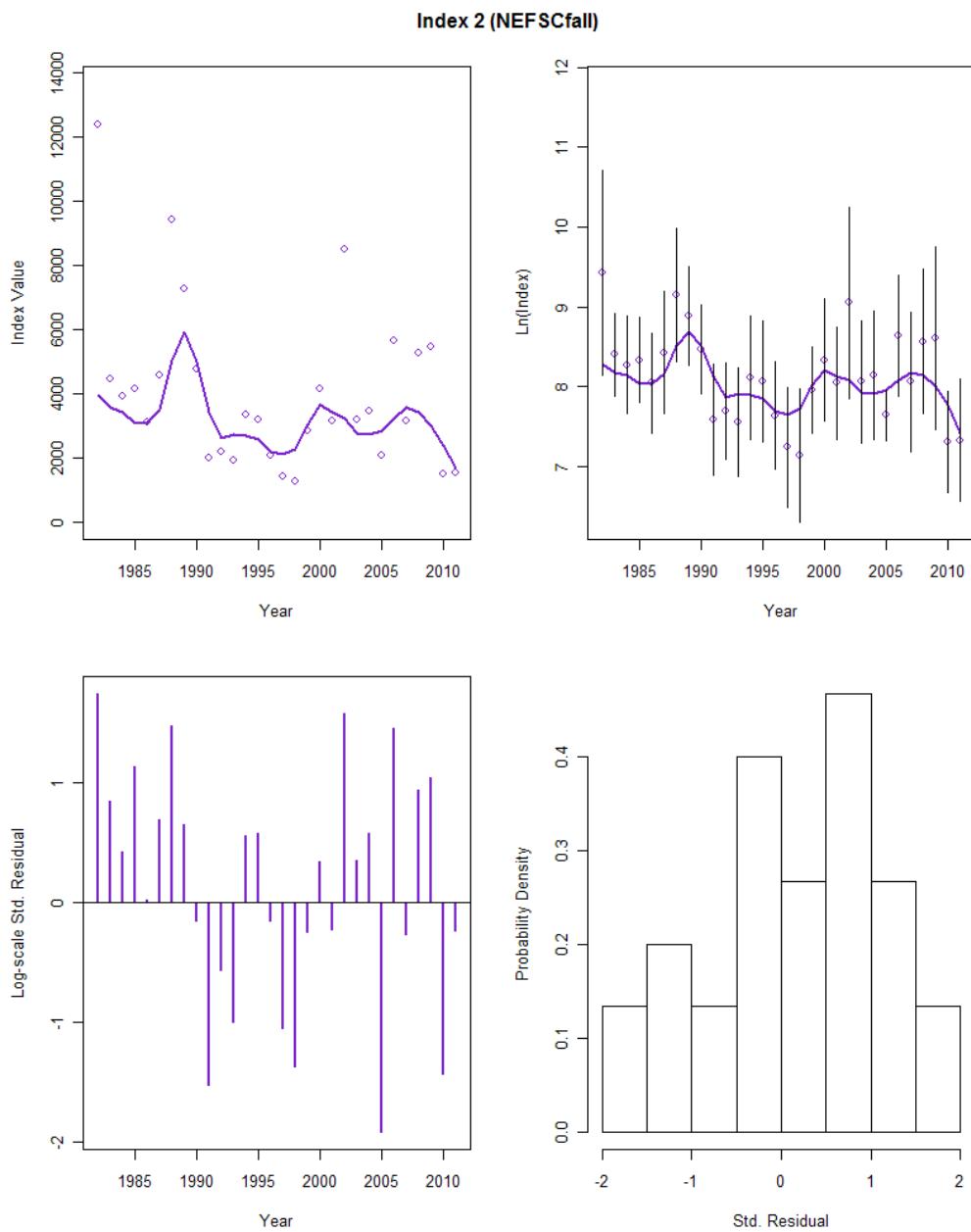


Figure A.208. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit to the NEFSC fall (Index 2) survey Gulf of Maine Atlantic cod index.

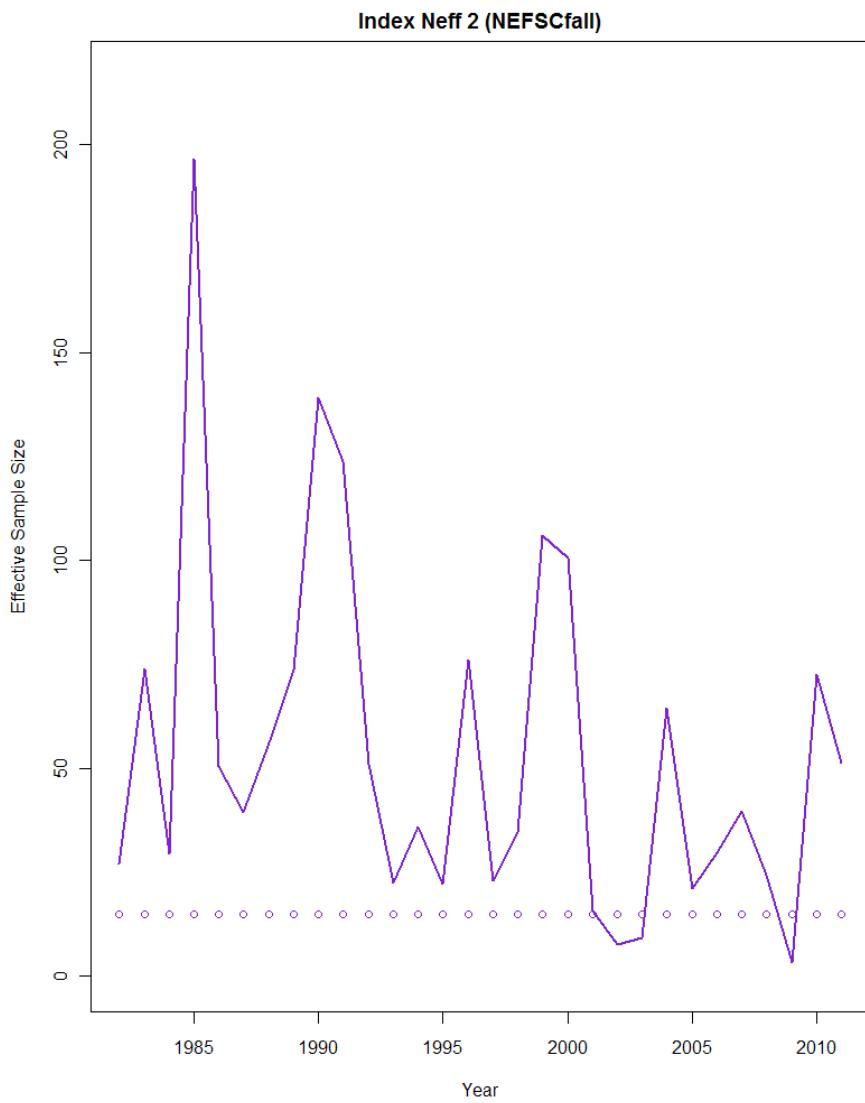


Figure A.209. ASAP SAW55_3BLOCK_BASE_M_SPLIT model comparison of input effective sample size versus the model estimated effective sample size for the NEFSC fall (Index 2) survey Gulf of Maine Atlantic cod index.

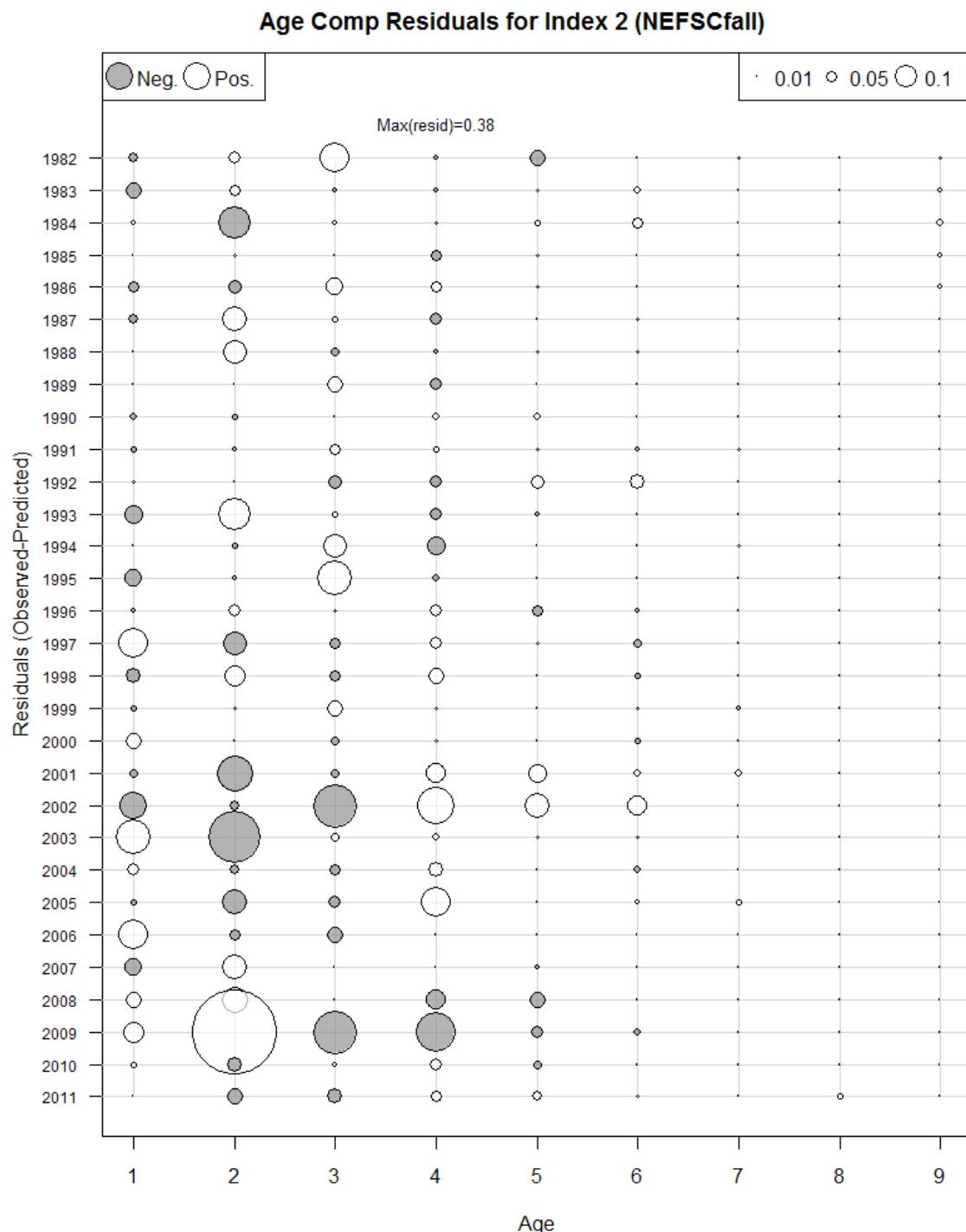


Figure A.210. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit residuals for the NEFSC fall survey (Index 2) Gulf of Maine Atlantic cod age composition.

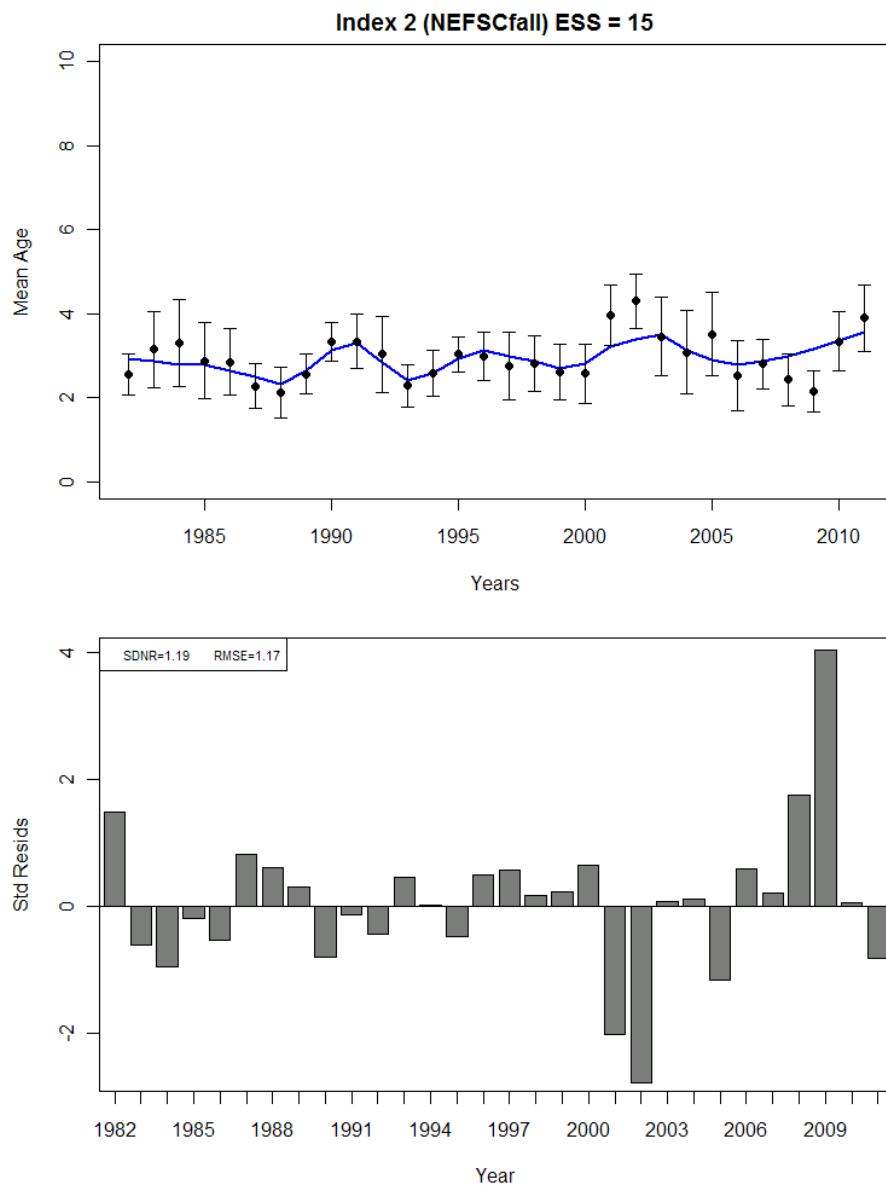


Figure A.211. ASAP SAW55_3BLOCK_BASE_M_SPLIT predicted mean age of Gulf of Maine Atlantic cod in the NEFSC fall (Index 2) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

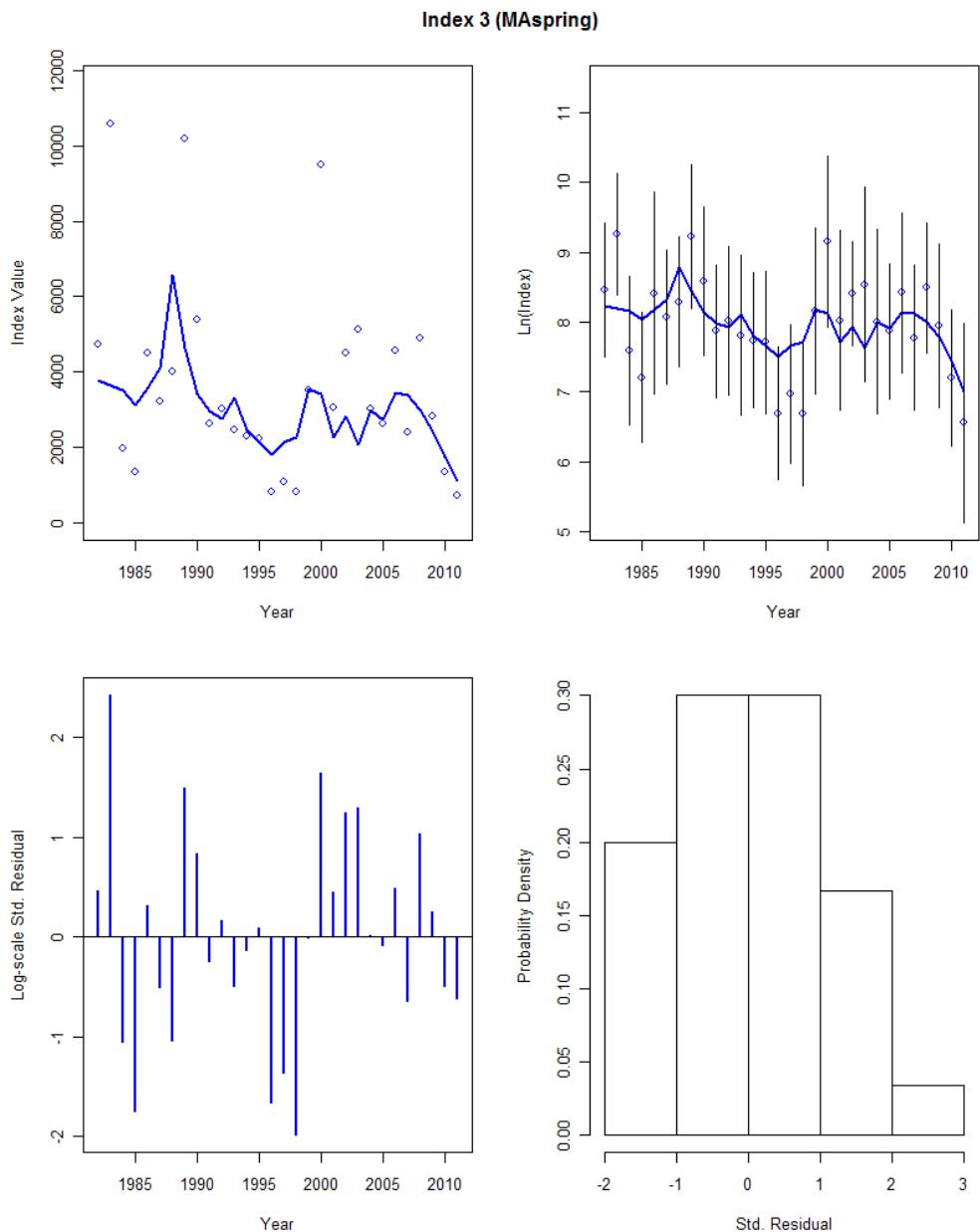


Figure A.212. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit to the MADMF spring (Index 3) survey Gulf of Maine Atlantic cod index.

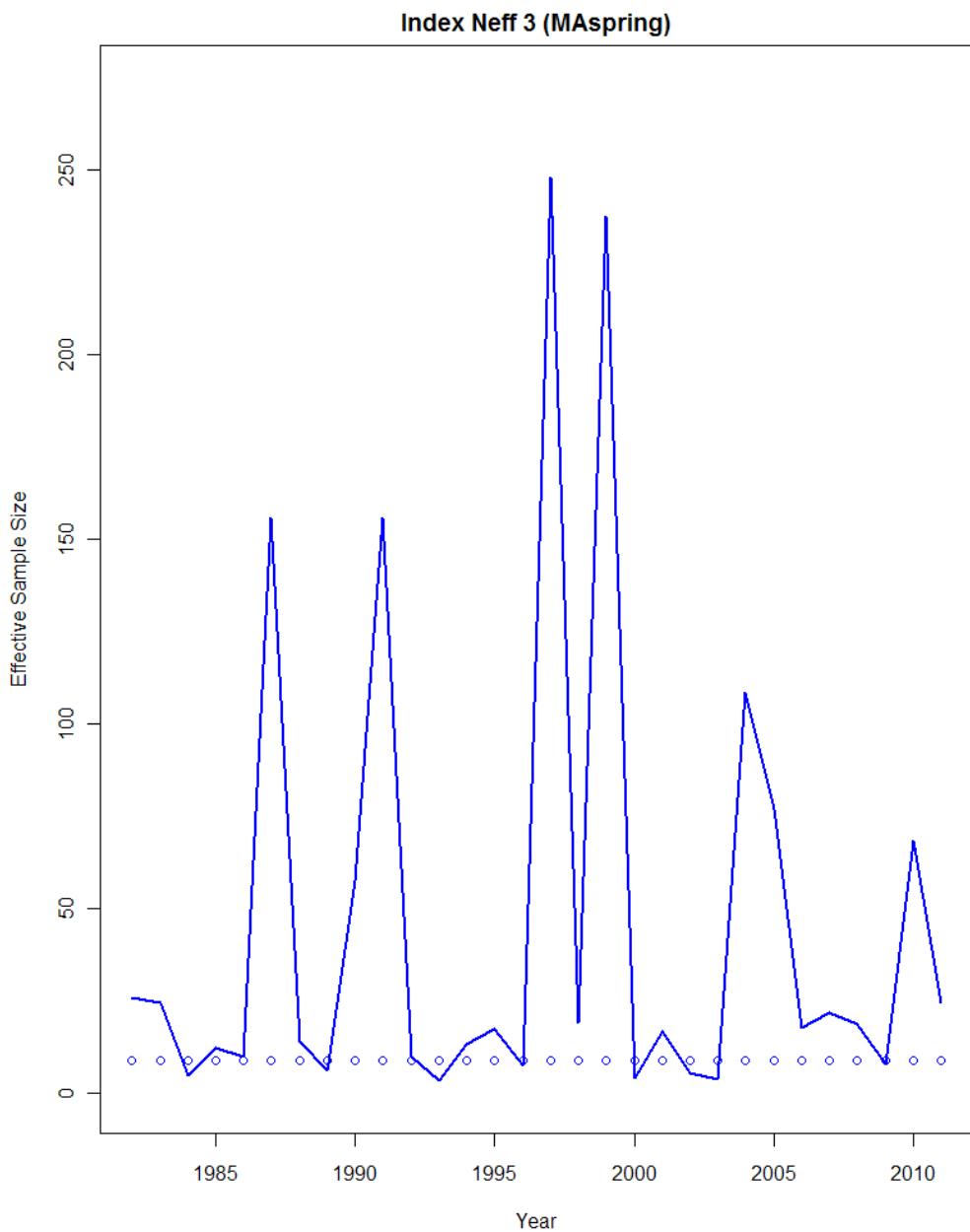


Figure A.213. ASAP SAW55_3BLOCK_BASE_M_SPLIT model comparison of input effective sample size versus the model estimated effective sample size for the MADMF spring (Index 3) survey Gulf of Maine Atlantic cod index.

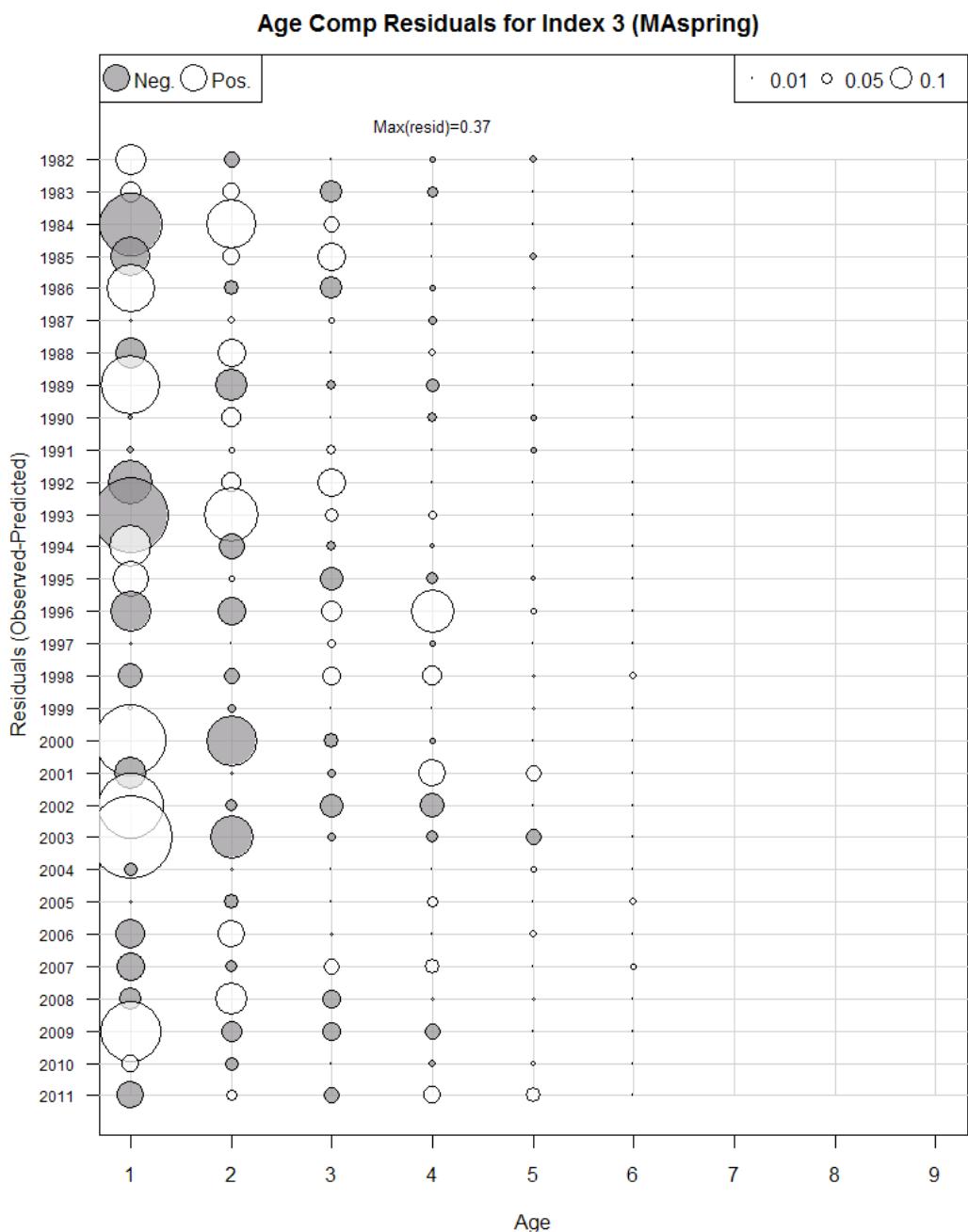


Figure A.214. ASAP SAW55_3BLOCK_BASE_M_SPLIT model fit residuals for the MADMF spring survey (Index 3) Gulf of Maine cod age composition.

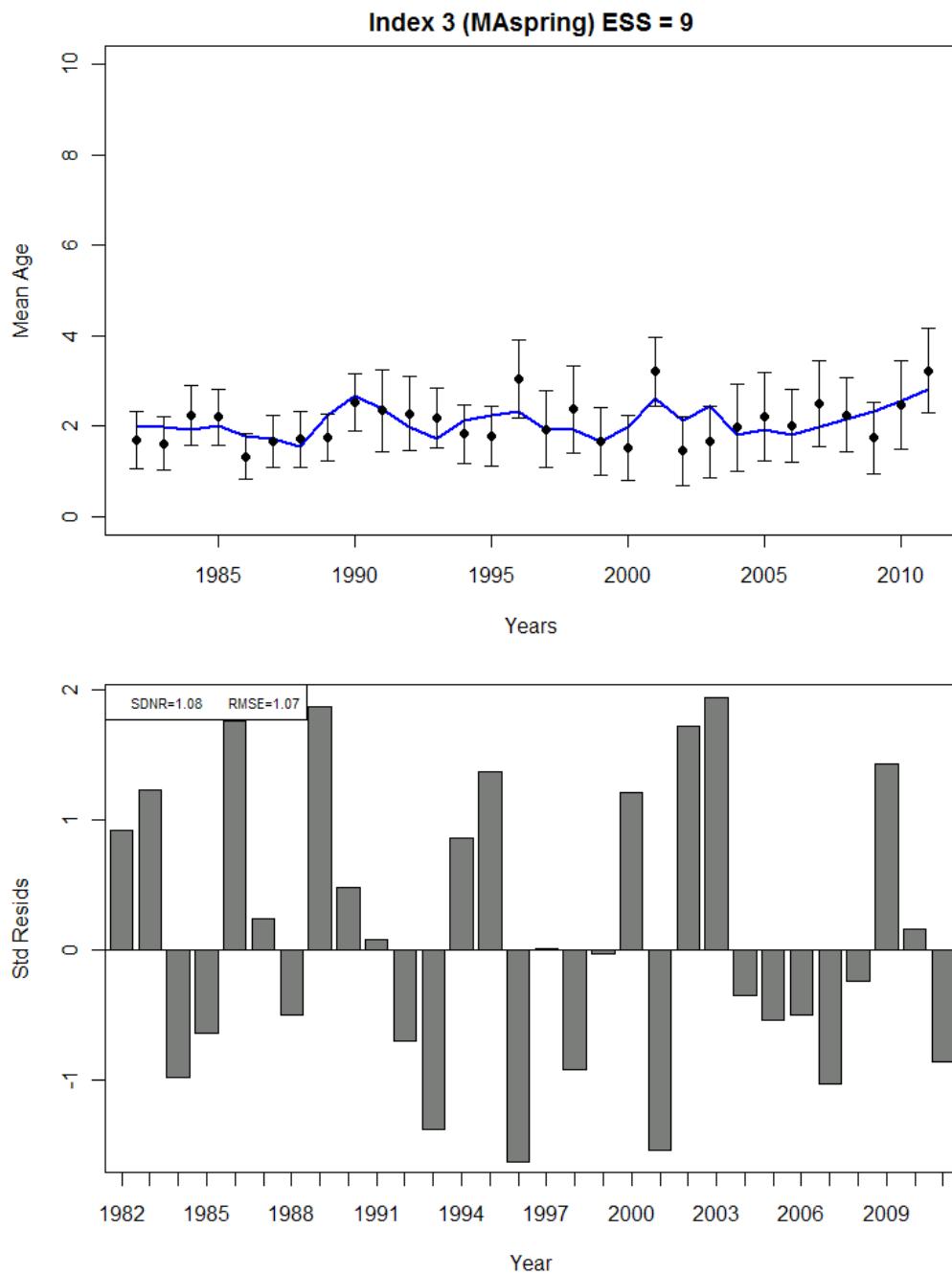


Figure A.215. ASAP SAW55_3BLOCK_BASE_M_SPLIT predicted mean age of Gulf of Maine Atlantic cod in the MADMF spring (Index 3) survey (blue line) compared to observed mean age (top plot) and the residuals about the mean (bottom plot).

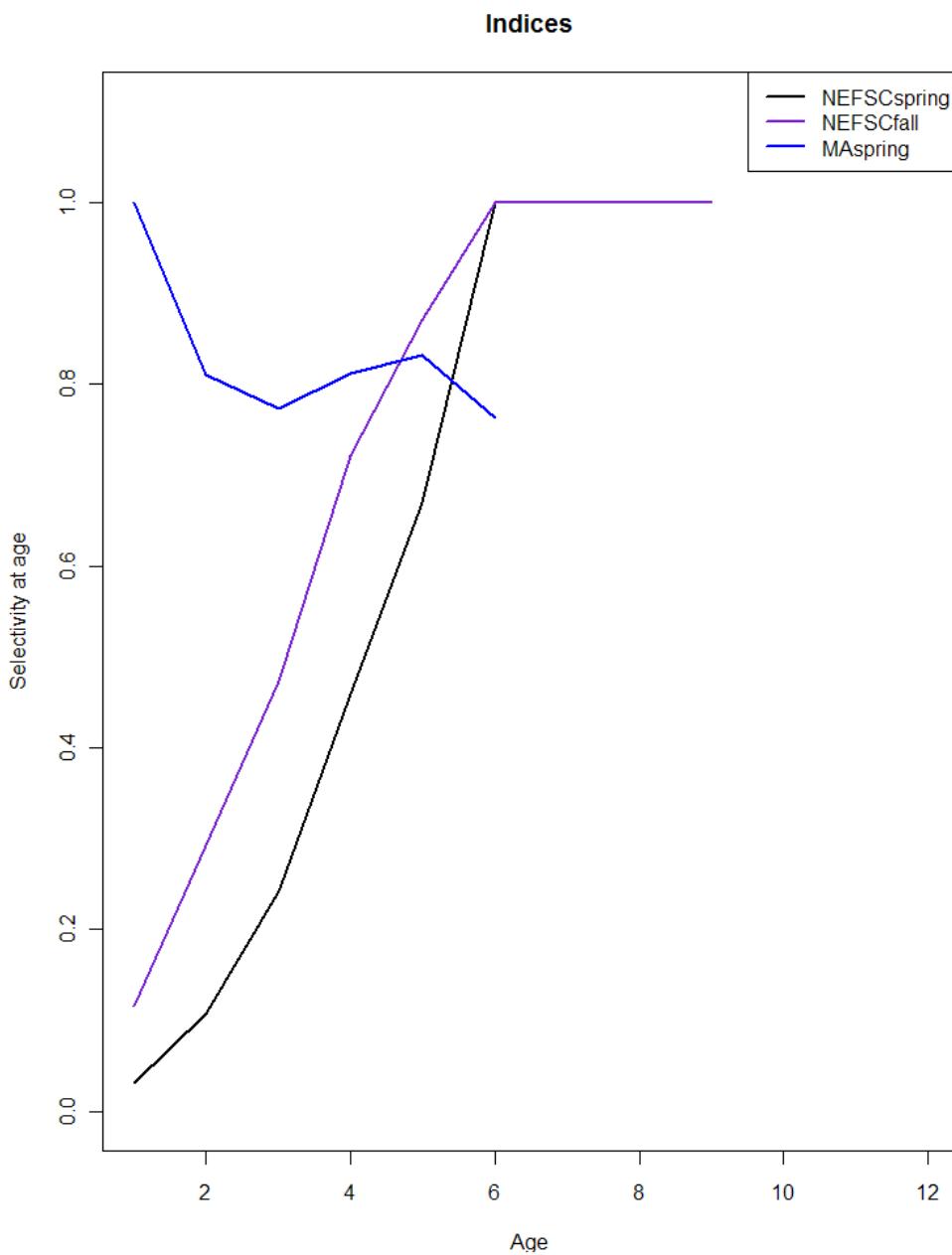


Figure A.216. Gulf of Maine Atlantic cod selectivity-at-age for the NEFSC spring (Index 1), fall (Index 2) and MADMF spring (Index 3) surveys from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model.

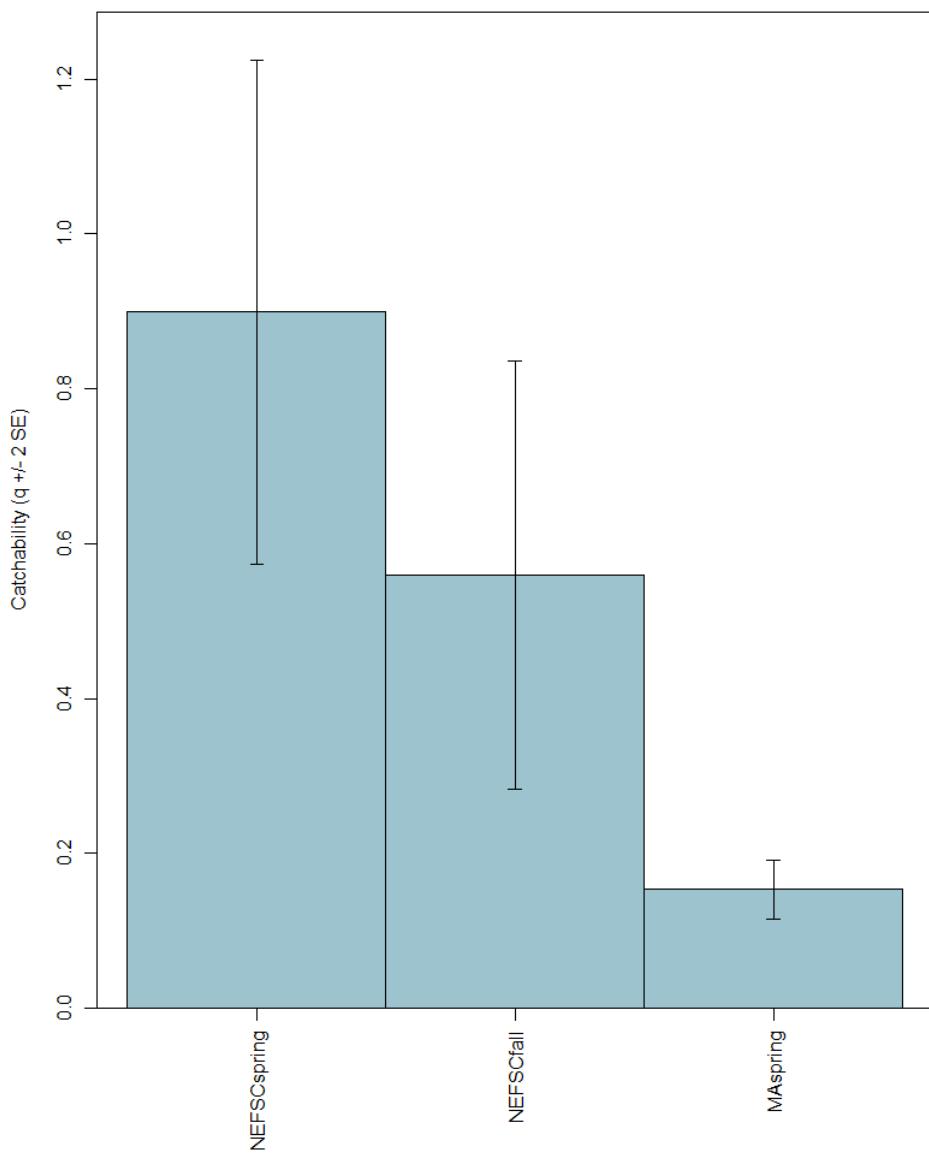


Figure A.217. Gulf of Maine Atlantic cod survey catchability, q , for the NEFSC spring (Index 1), fall (Index 2) and MADMF spring (Index 3) surveys from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model.

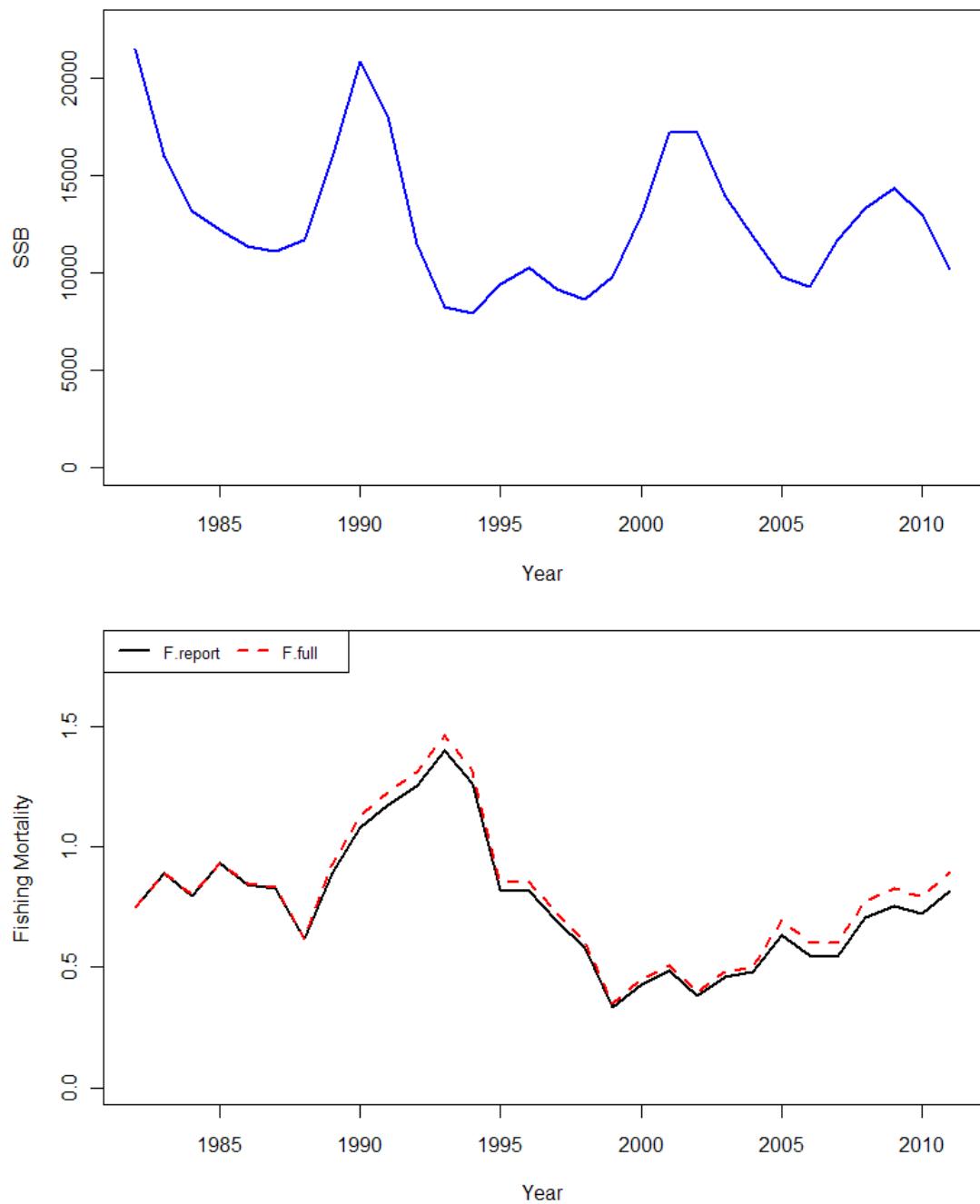


Figure A.218. ASAP SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine Atlantic cod spawning stock biomass (SSB) and fishing mortality (F_{full} = fully recruited fishing mortality, F_{report} = fishing mortality on age 5).

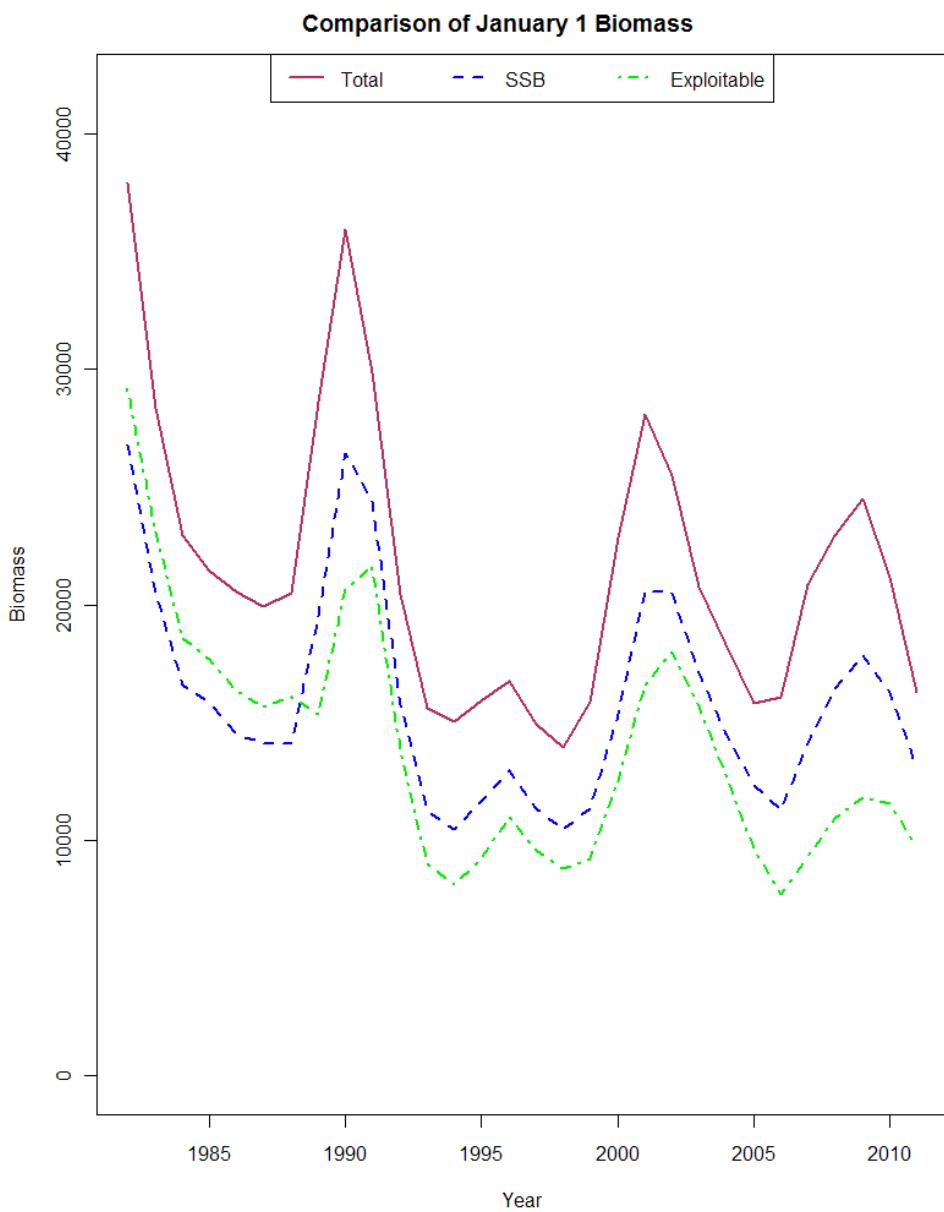


Figure A.219. Comparison of ASAP SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine Atlantic cod January 1 biomass after application of maturity ogive (SSB) and fleet selectivity ogives (exploitable).

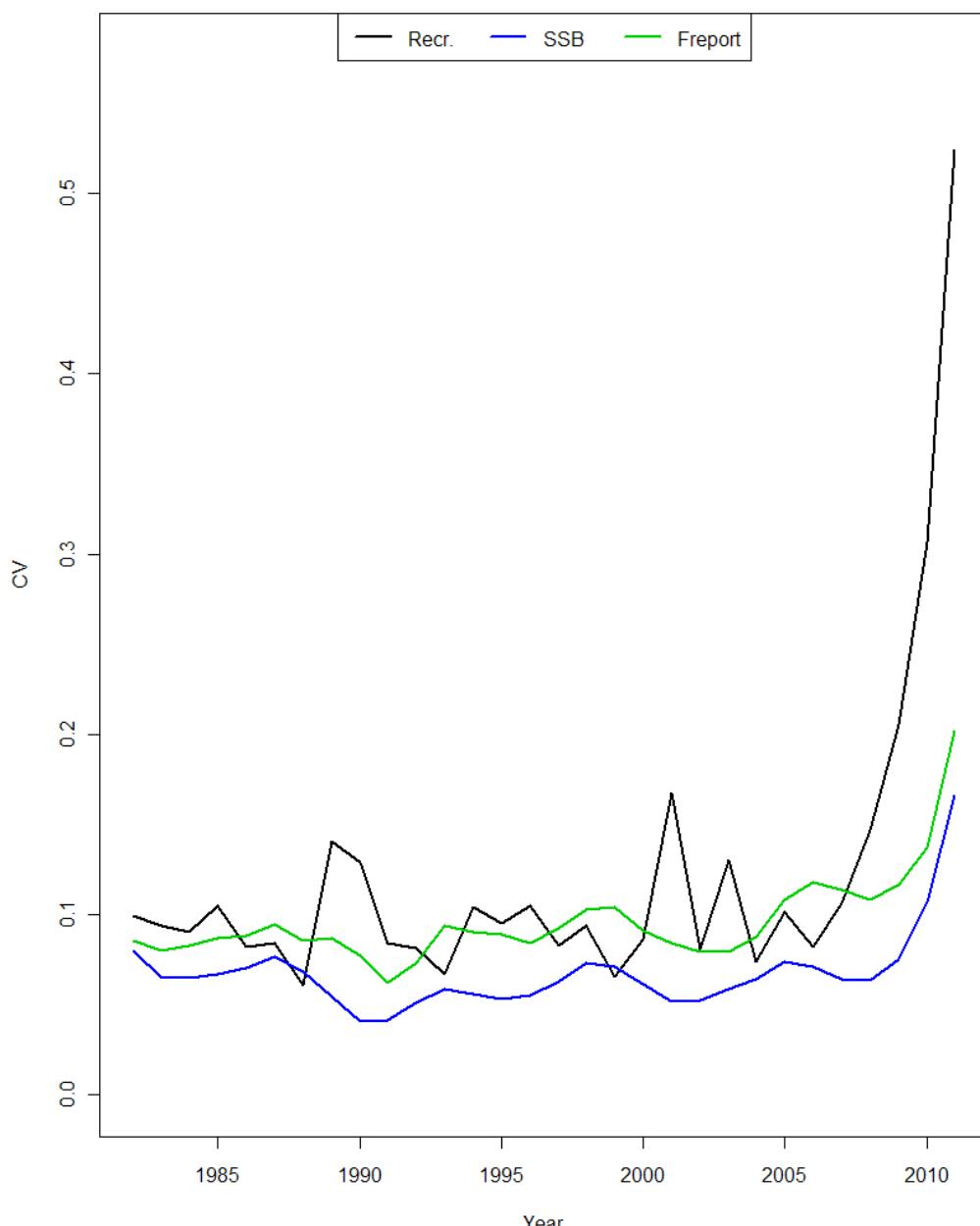


Figure A.220. Coeffecients of variation (CV) for the ASAP
SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine Atlantic cod spawning
stock biomass (SSB), average fishing mortality ($F_{\text{report}} = \text{age } 5 \text{ fishing mortality}$) and age 1
recruitment.

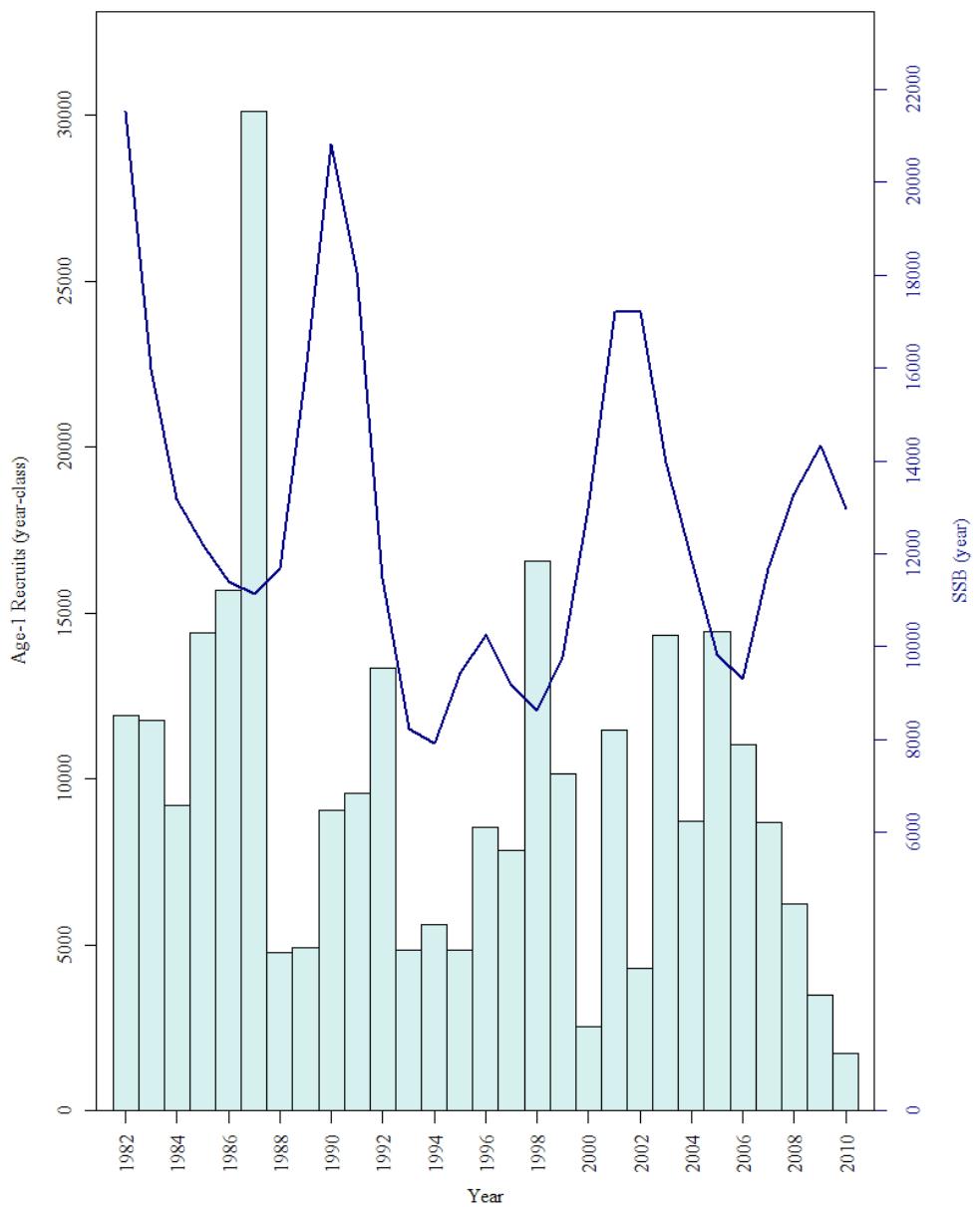


Figure A.221. ASAP SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine cod spawning stock biomass (SSB; solid blue line) and lagged age 1 recruitment (light blue bars).

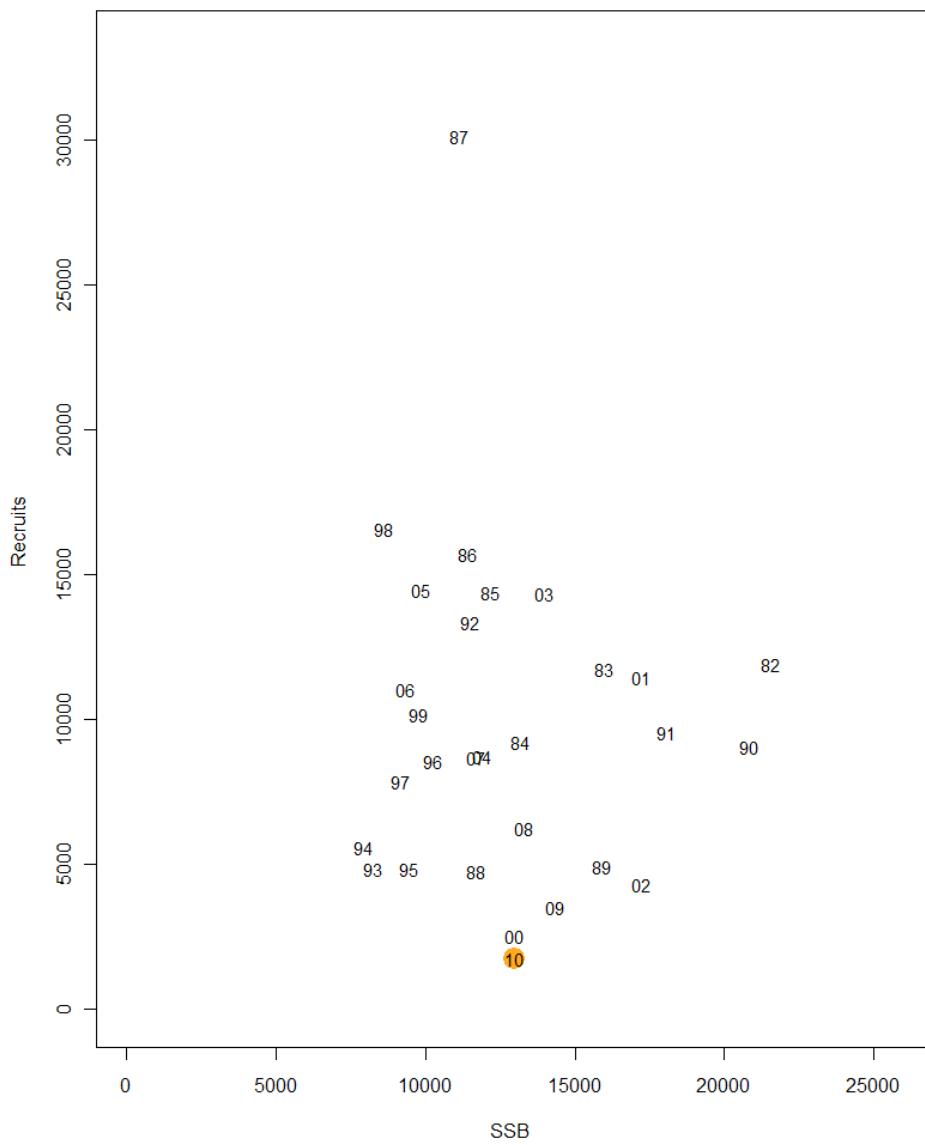


Figure A.222. Scatterplot of ASAP SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine Atlantic cod spawning stock biomass (SSB) versus recruitment at age 1 (000s). The symbol for each observation is the last two digits of the year (e.g., 88 indicated age 1 estimates of the 1987 year class). The most recent recruitment estimate is highlighted by an orange circle.

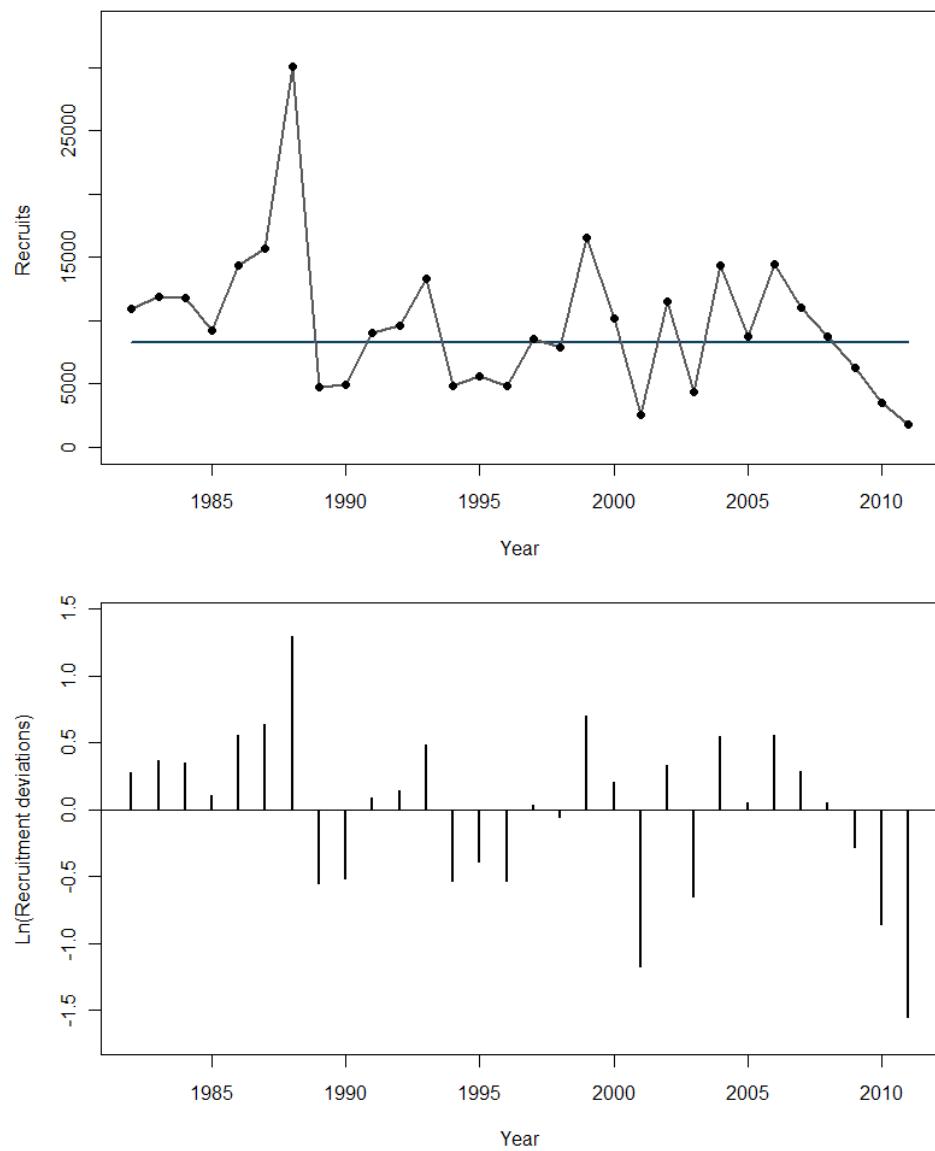


Figure A.223. ASAP SAW55_3BLOCK_BASE_M_SPLIT estimated Gulf of Maine Atlantic cod age 1 recruitment and recruitment residuals from the geometric mean.

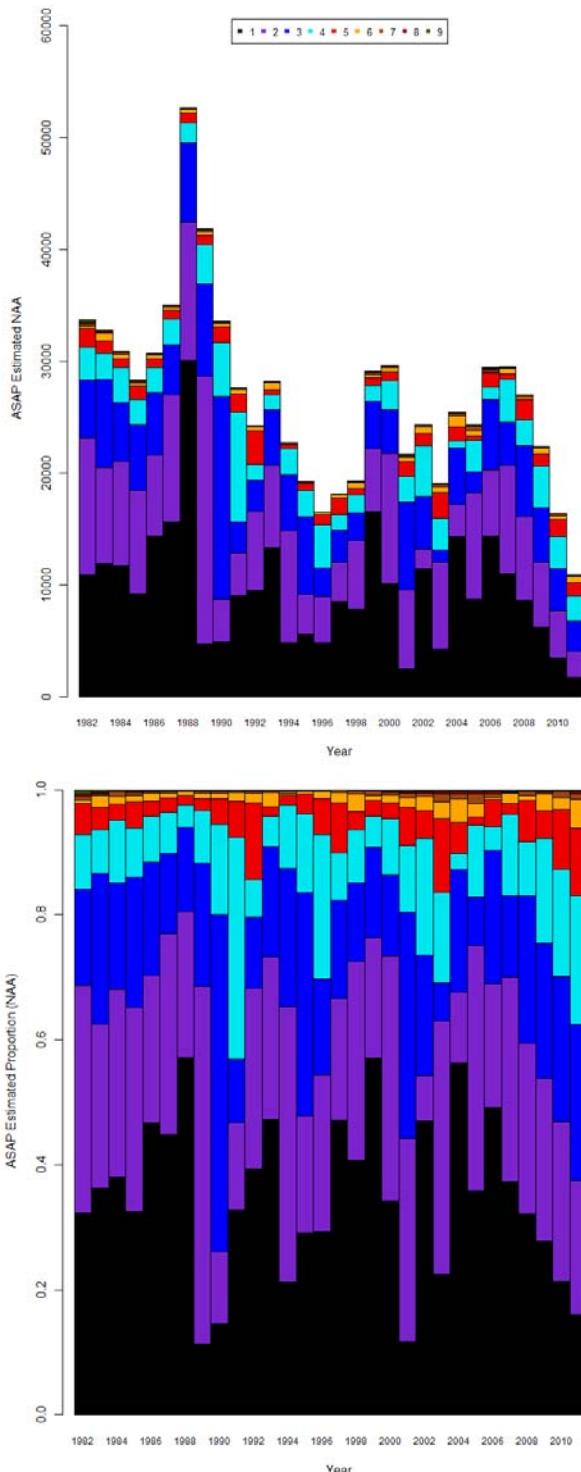


Figure A.224. ASAP SAW55_3BLOCK_BASE_M_SPLIT model estimates of Gulf of Maine Atlantic cod numbers-at-age in absolute (top) numbers (000s) and relative (bottom) terms.

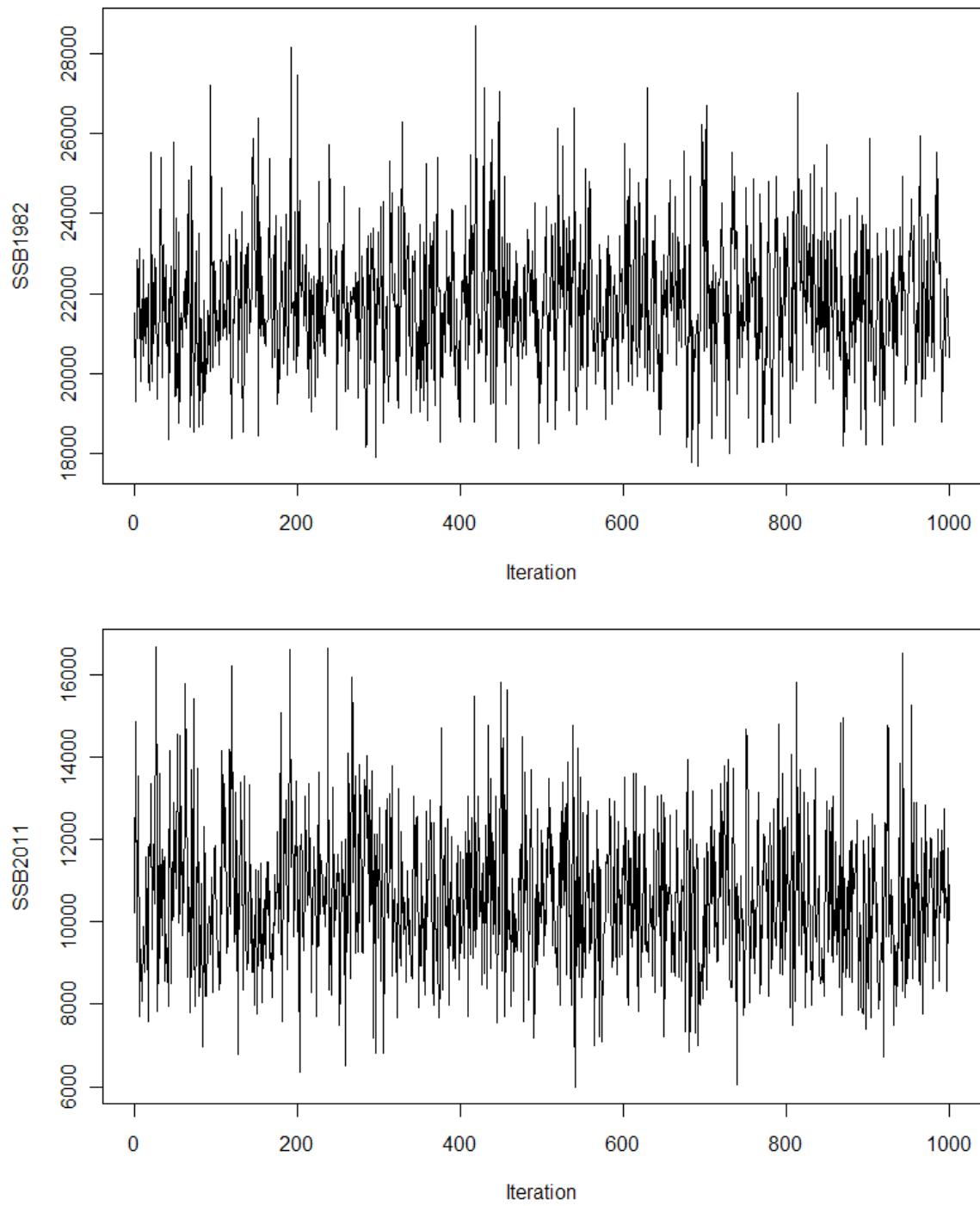


Figure A.225. Trace of MCMC chains for Gulf of Maine Atlantic cod 1982 and 2011 spawning stock biomass, showing good mixing (ASAP SAW55_3BLOCK_BASE_M_SPLIT model). Each chain had initial length of 100,000 and was thinned at a rate of one out of every 1000th resulting in a final chain length of 1000.

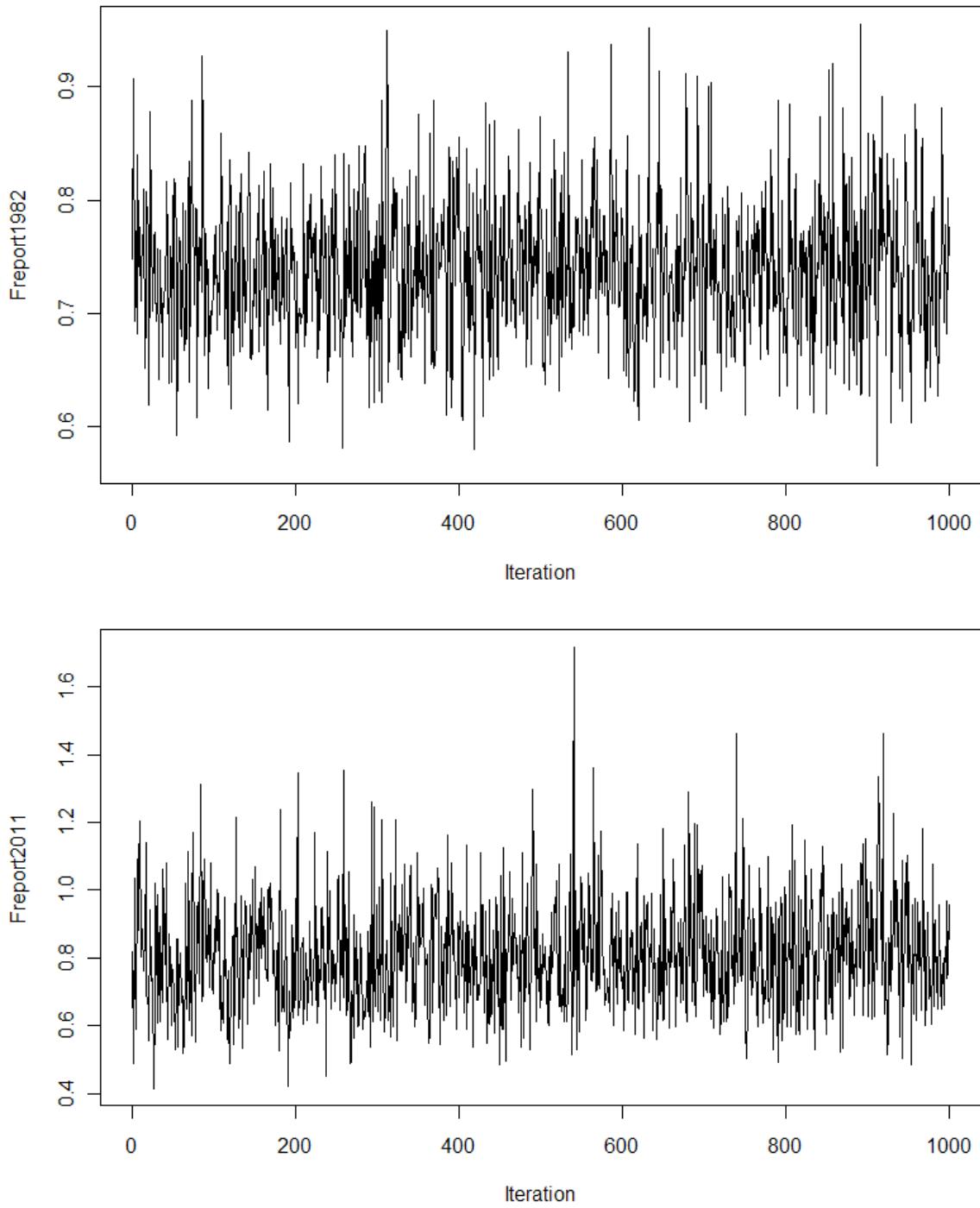


Figure A.226. Trace of MCMC chains for Gulf of Maine Atlantic cod 1982 and 2011 fishing mortality at age 5 (Freport), showing good mixing (ASAP SAW55_3BLOCK_BASE_M_SPLIT model). Each chain had initial length of 100,000 and was thinned at a rate of one out of every 1000th resulting in a final chain length of 1000.

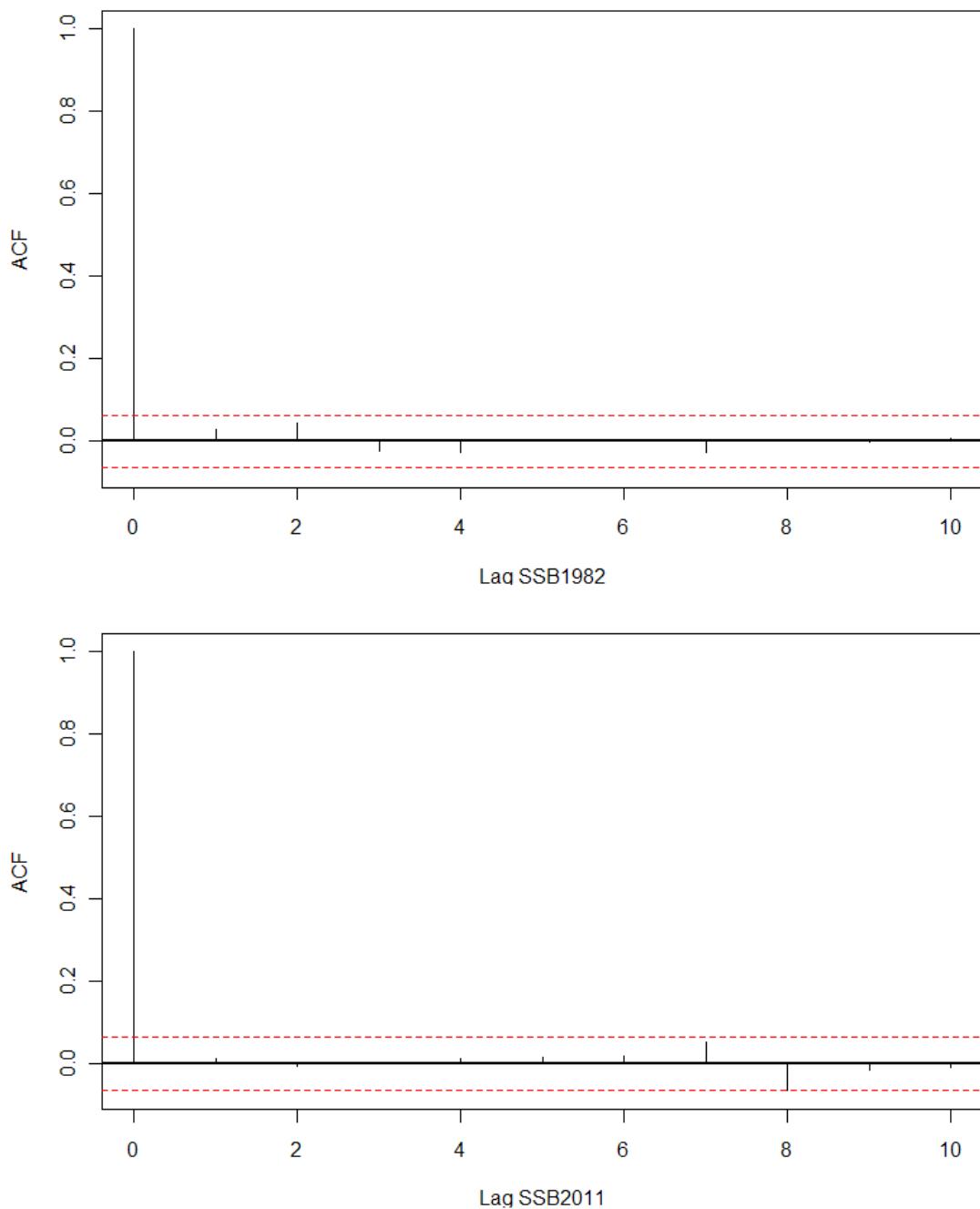


Figure A.227. Autocorrelation within the 1982 and 2011 Gulf of Maine Atlantic cod spawning stock biomass (SSB) MCMC chains from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model.

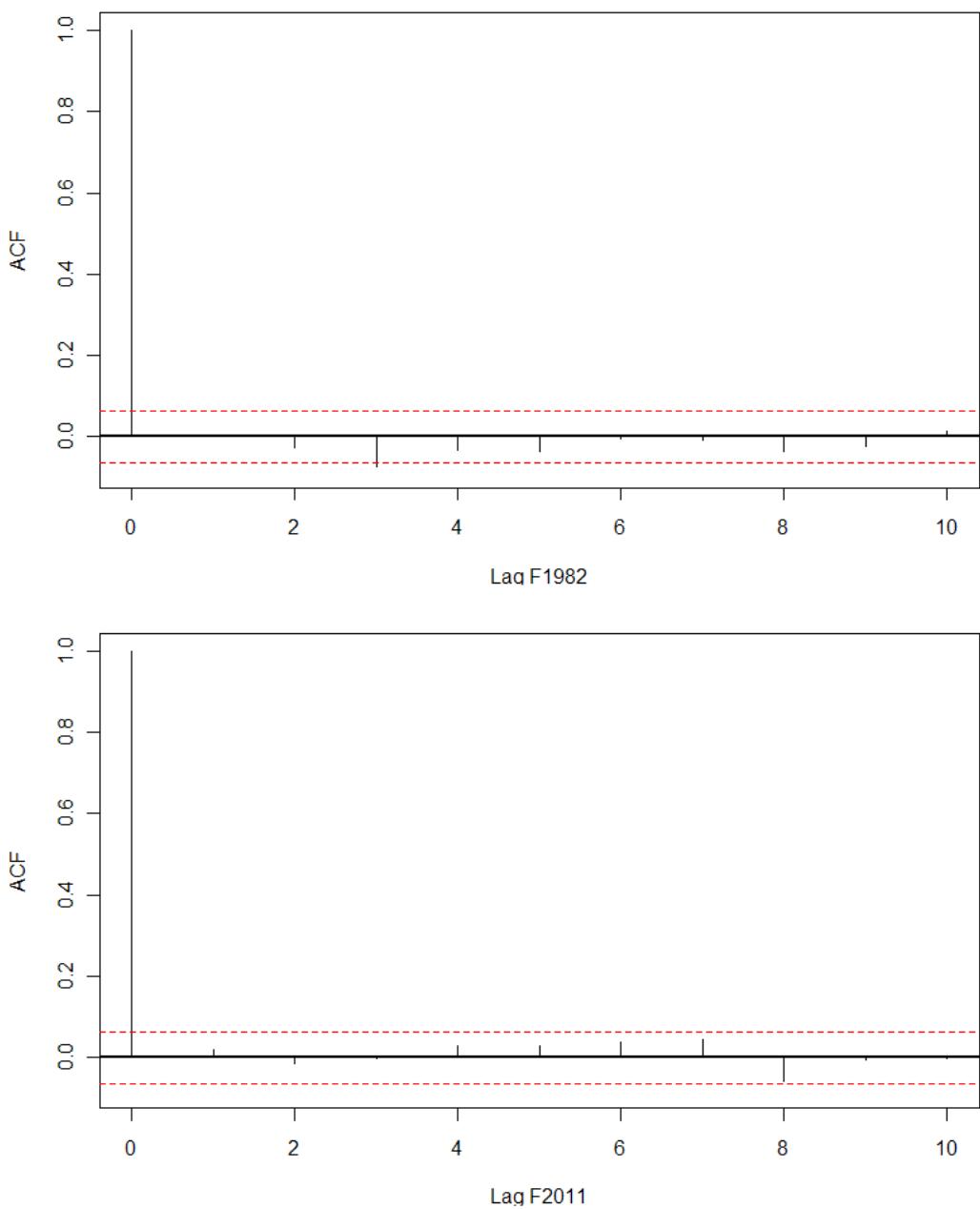


Figure A.228. Autocorrelation within the 1982 and 2011 Gulf of Maine Atlantic cod fishing mortality at age 5 (Freport) MCMC chains from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model.

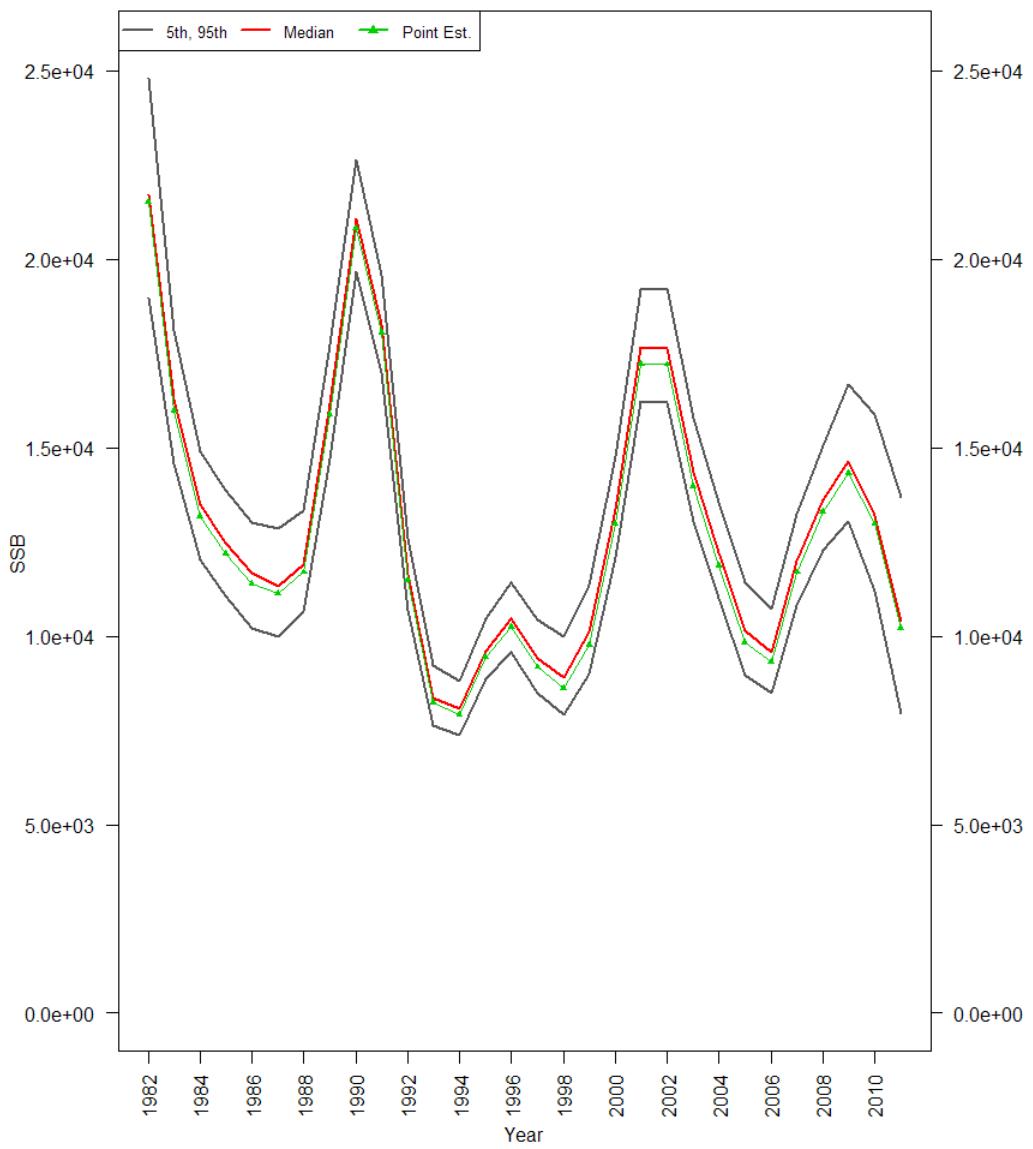


Figure A.229. 90% probability interval for Gulf of Maine Atlantic cod spawning stock biomass (SSB) from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model. The median value is in red, while the 5th and 95th percentiles are in dark grey. The point estimate from the base model (joint posterior modes) is shown in the thin green line with filled triangles.

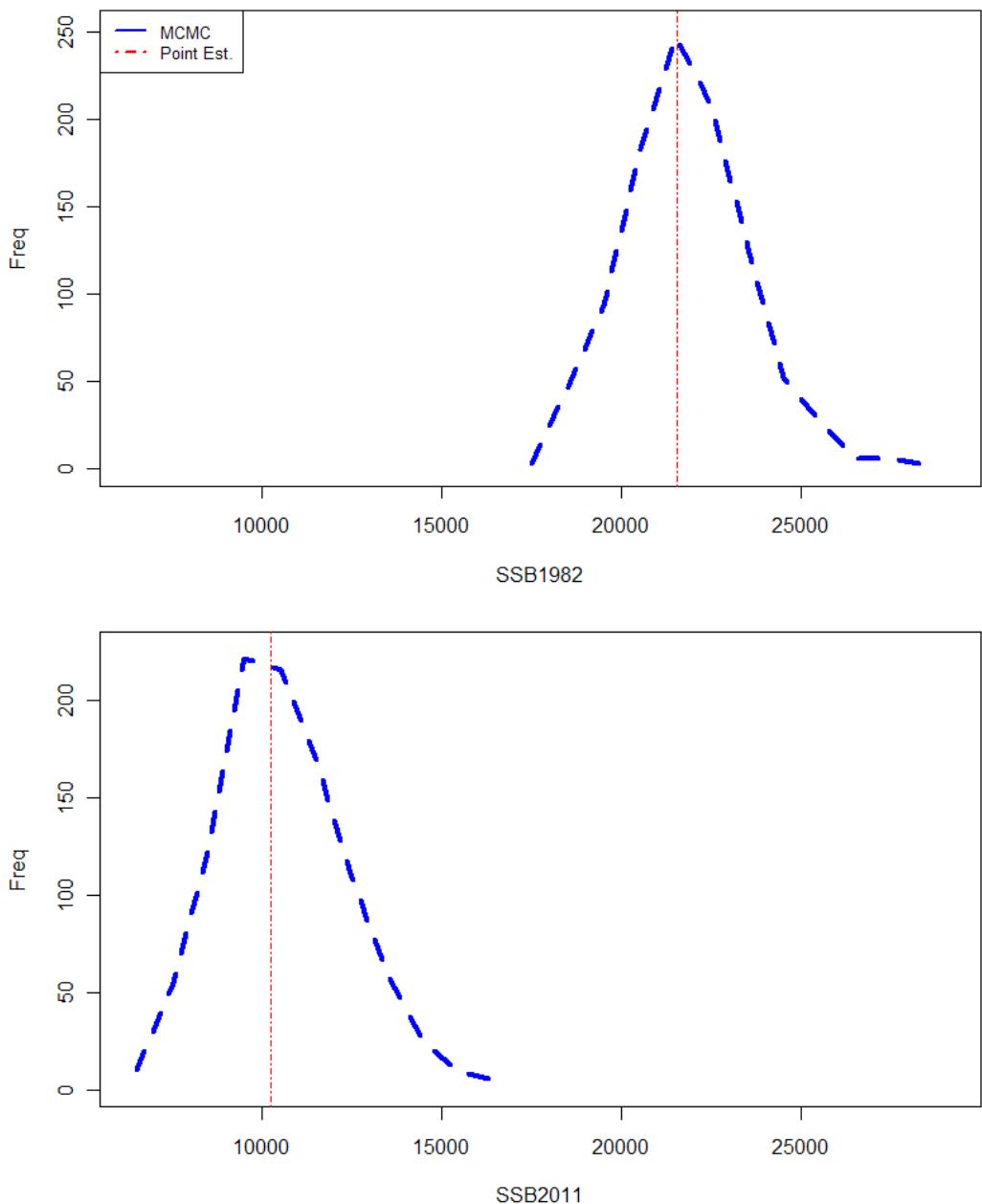


Figure A.230. MCMC distribution of Gulf of Maine Atlantic cod spawning stock biomass in 1982 and 2011 estimated from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model. The model point estimate is indicated by the dashed red line.

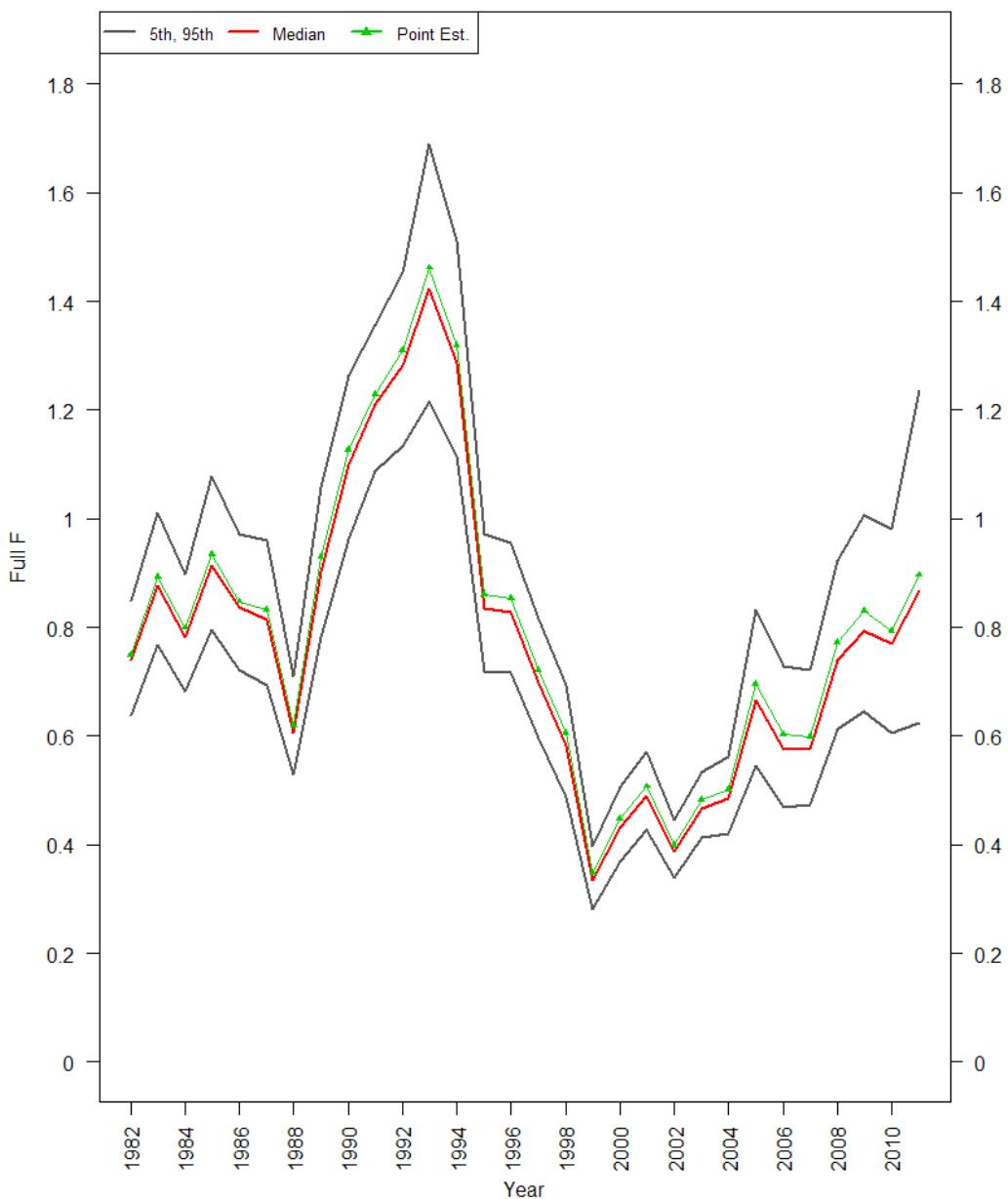


Figure A.231. 90% probability interval for Gulf of Maine Atlantic fully recruited fishing mortality (Full F) from the ASAP SAW55_3BLOCK_BASE_M_SPLIT model. The median value is in red, while the 5th and 95th percentiles are in dark grey. The point estimate from the base model (joint posterior modes) is shown in the thin green line with filled triangles.

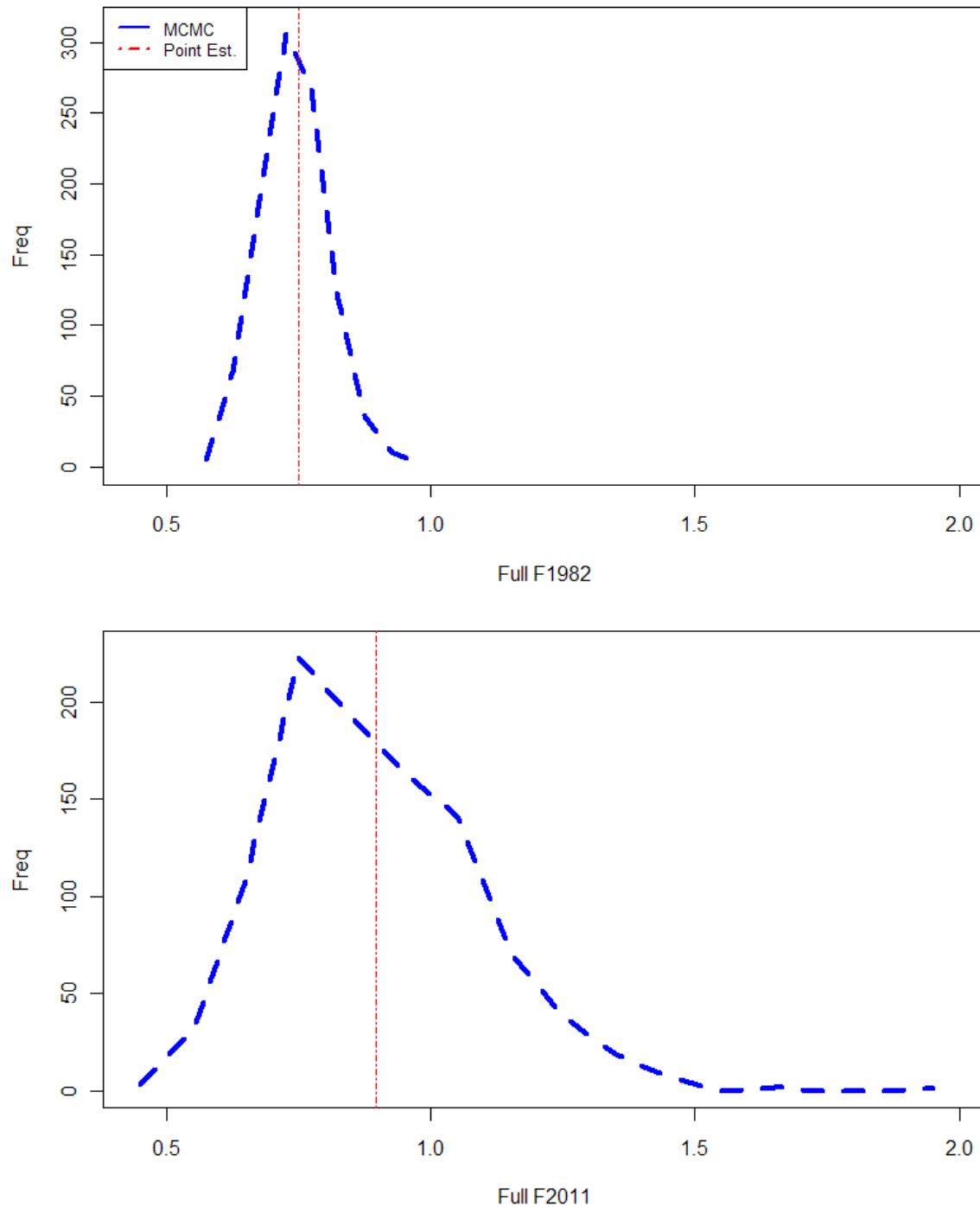


Figure A.232. MCMC distribution of Gulf of Maine Atlantic cod fully recruited fishing mortality (Full F) in 1982 and 2011 estimated from the ASAP SAW55_3BLOCK_BASE model. The model point estimate is indicated by the dashed red line.

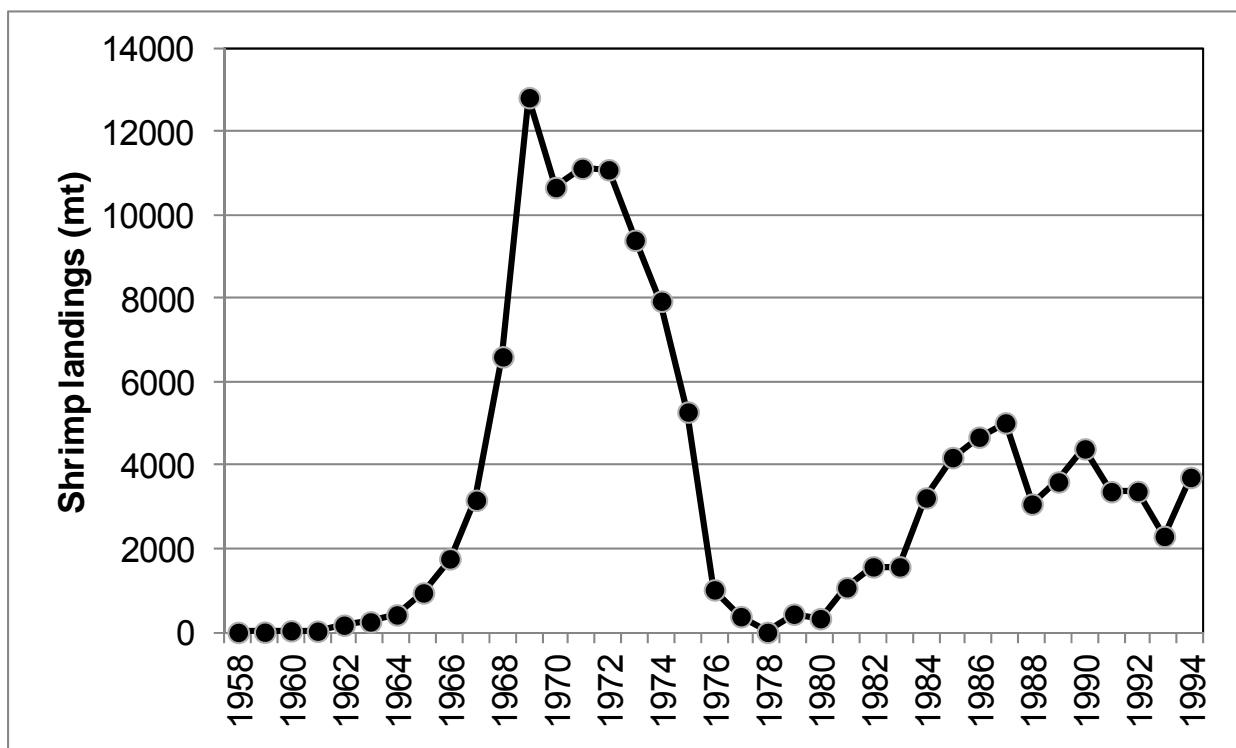


Figure A.233. Northern shrimp landings (mt) between 1958 and 1994.

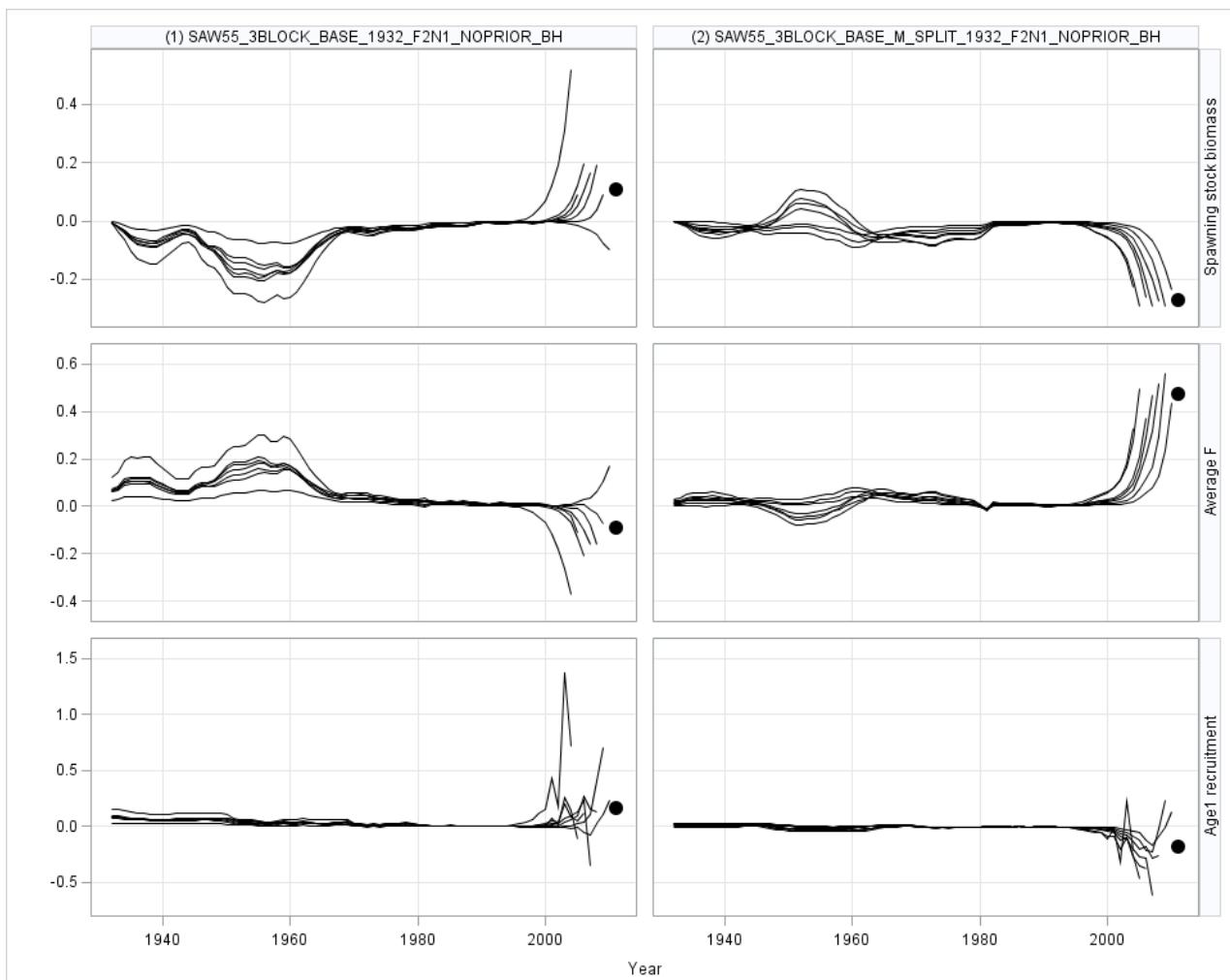


Figure A.234. Model retrospective patterns for sensitivity runs of the Gulf of Maine Atlantic ASAP SAW55_3BLOCK_BASE_1932_F2N1_NOPRIOR_BH, and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NOPRIOR_BH models. The black circles indicates the Mohn's rho value based on a five year retrospective peel.

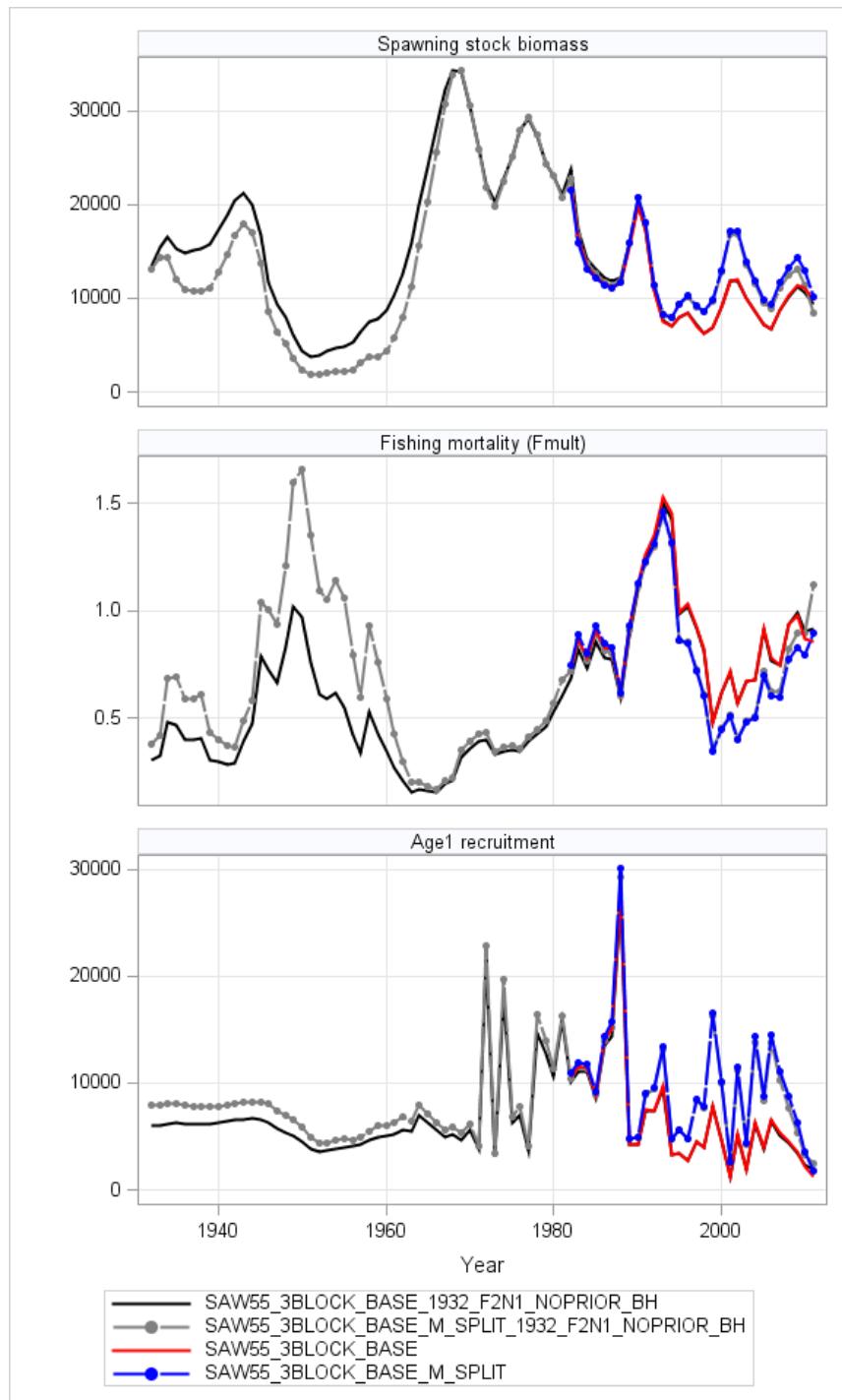


Figure A.235. Comparison of the Gulf of Maine Atlantic cod spawning stock biomass, fully recruited fishing mortality and age 1 recruitment from the ASAP SAW55_3BLOCK_BASE, SAW55_3BLOCK_BASE_M_SPLIT, SAW55_3BLOCK_BASE_1932_F2N1_NOPRIOR_BH, and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NOPRIOR_BH models.

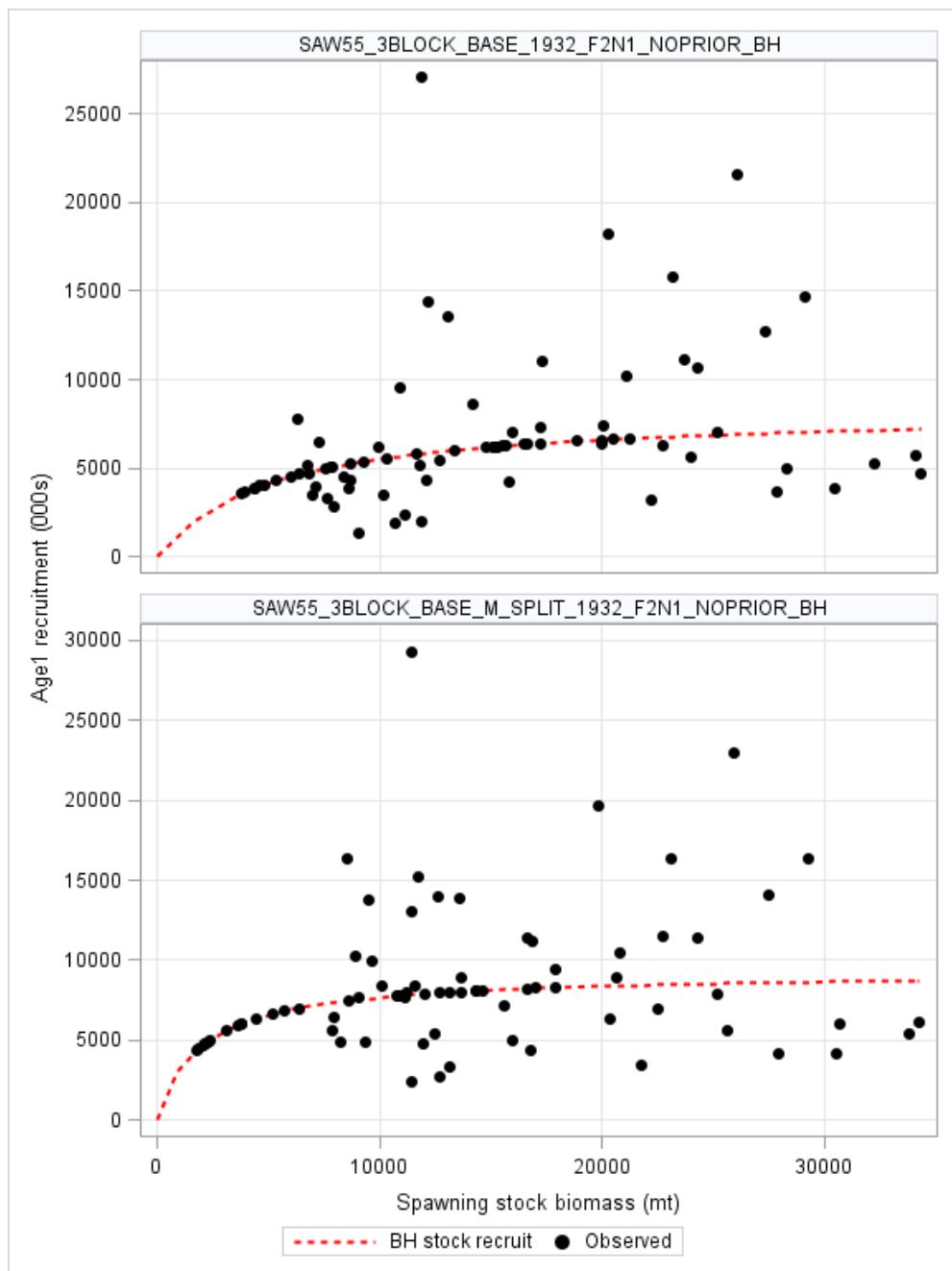


Figure A.236. Estimated Beverton Holt stock recruitment relationships for the ASAP SAW55_3BLOCK_BASE_1932_F2N1_NORPIOR_BH, and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NORPIOR_BH models.

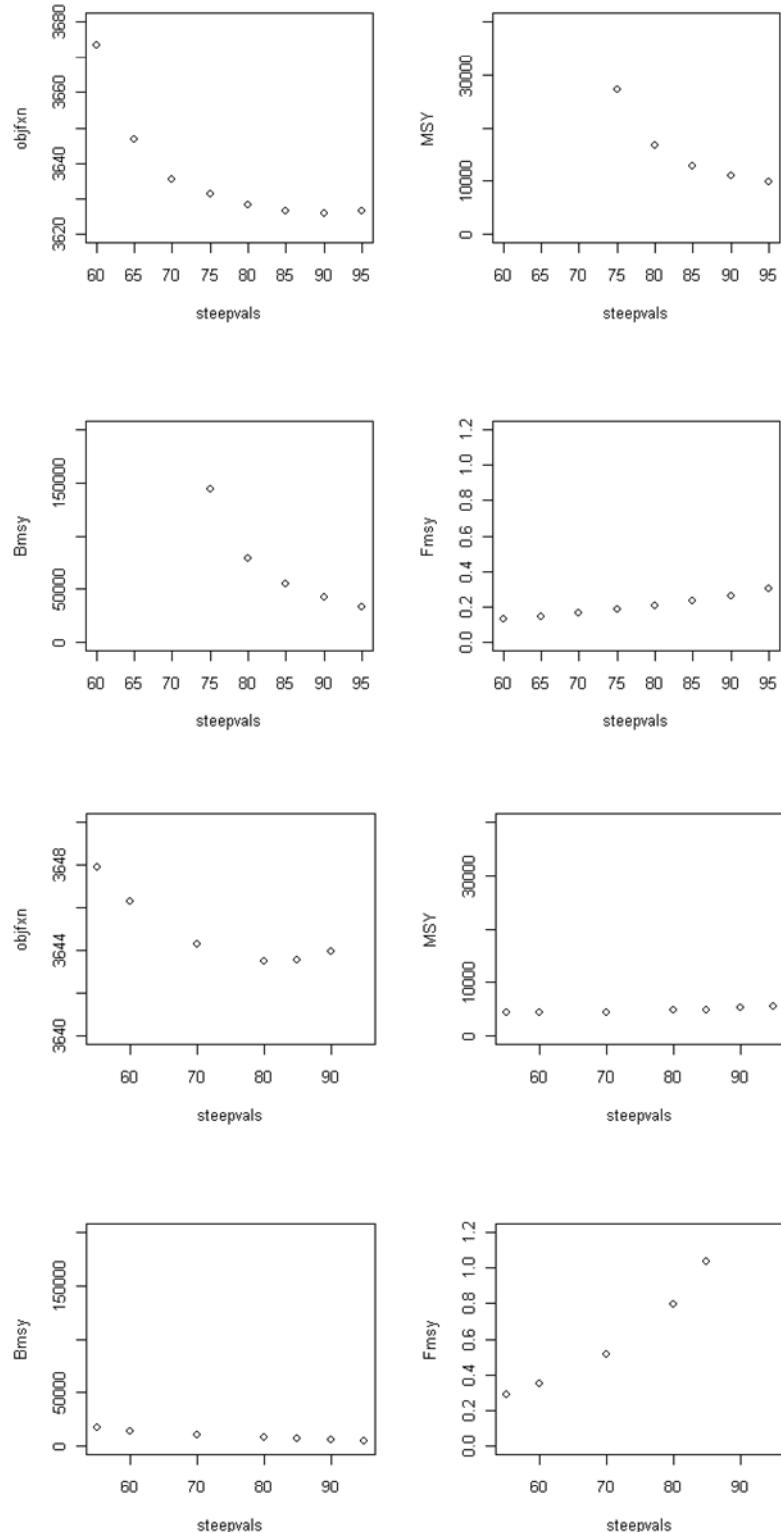


Figure A.237. Impacts of profiling over the Beverton and Holt steepness on the reference point estimates for the ASAP SAW55_3BLOCK_BASE_1932_F2N1_NORPIOR_BH (top four plots), and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NORPIOR_BH (bottom four plots) models.

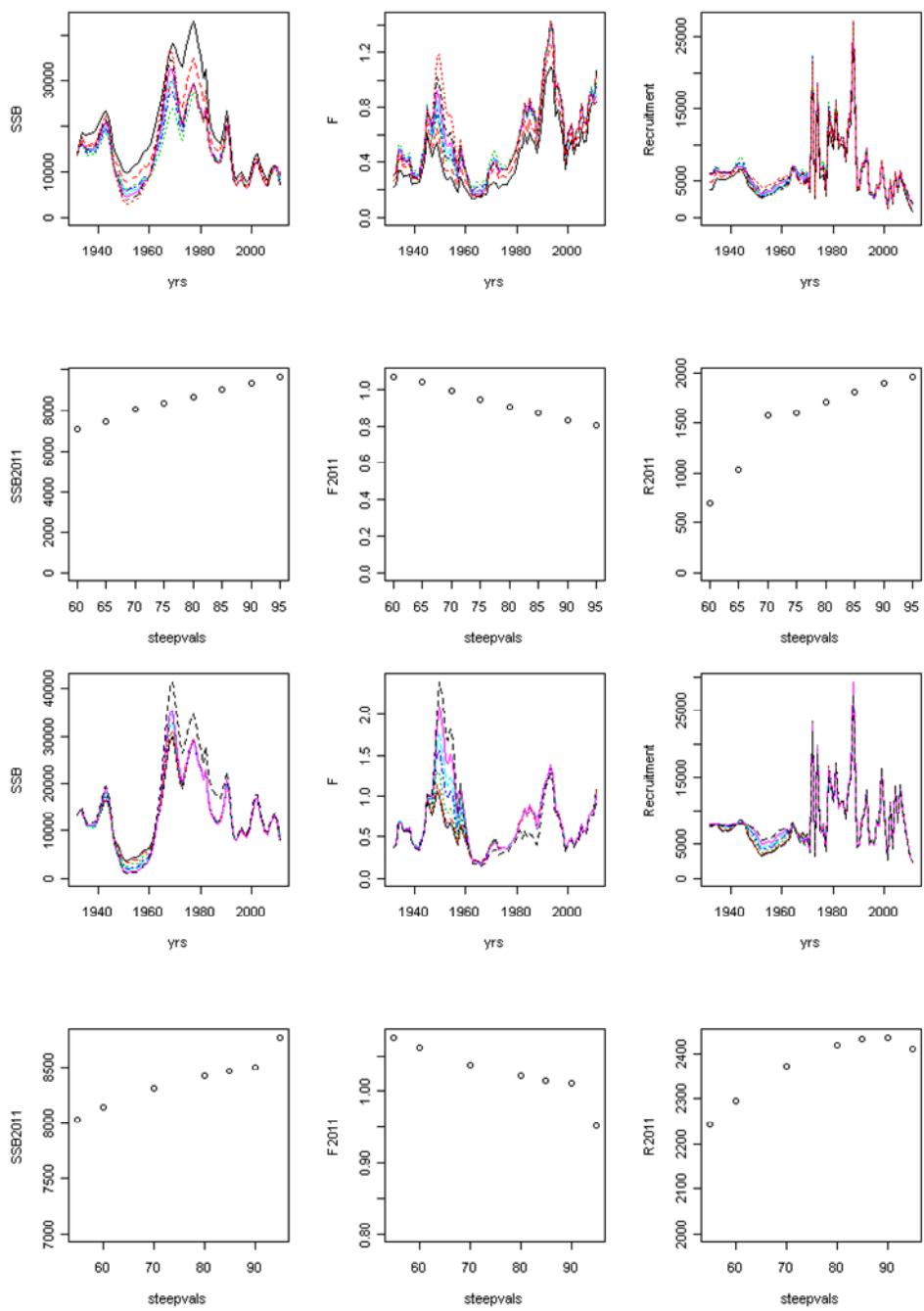


Figure A.238. Impacts of profiling over the Beverton and Holt steepness parameter on estimates of spawning stock biomass, fishing mortality and recruitment for the ASAP SAW55_3BLOCK_BASE_1932_F2N1_NORPIOR_BH (top six plots), and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NORPIOR_BH (bottom six plots) models.

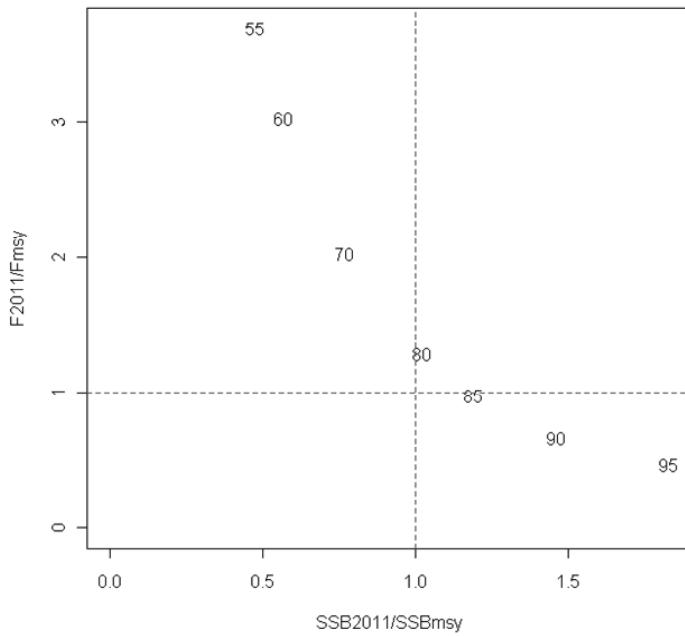
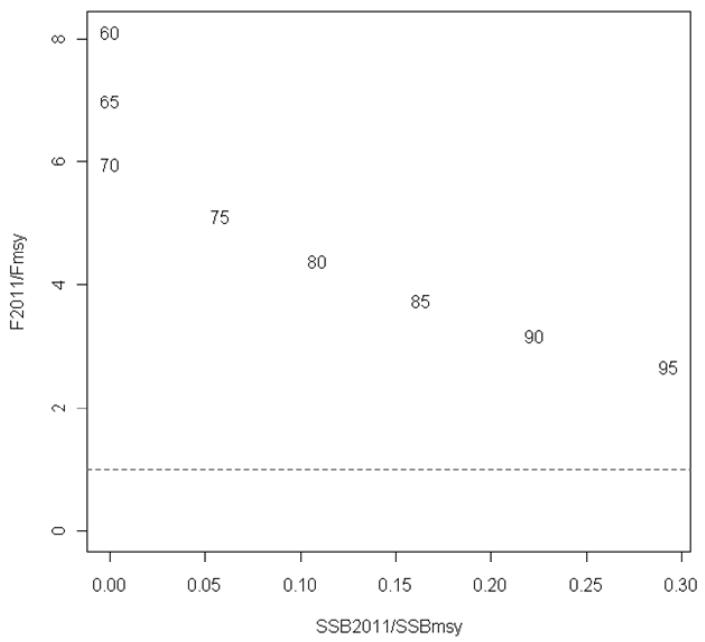


Figure A.239. Impacts of profiling over the Beverton and Holt steepness on stock status for the ASAP SAW55_3BLOCK_BASE_1932_F2N1_NORPIOR_BH (top plot), and SAW55_3BLOCK_BASE_M_SPLIT_1932_F2N1_NORPIOR_BH (bottom plot) models.

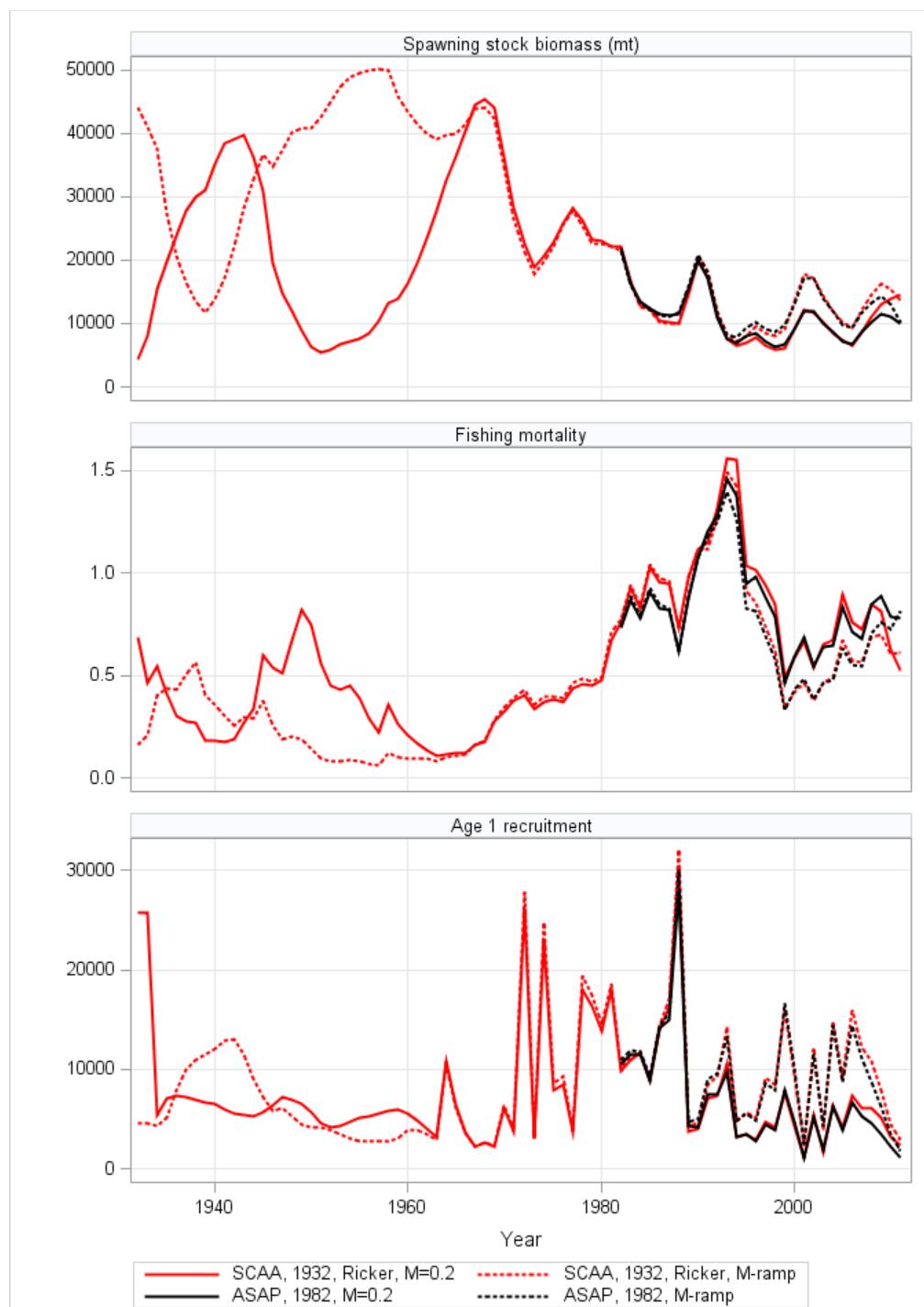


Figure A.240. Comparison of the Gulf of Maine Atlantic cod spawning stock biomass, fishing mortality and age 1 recruitment estimated by the ASAP (1982 start) and SCAA (1932 start) models with assumptions of constant natural mortality, M , of 0.2 and a ramped M changing from 0.2 to 0.4.

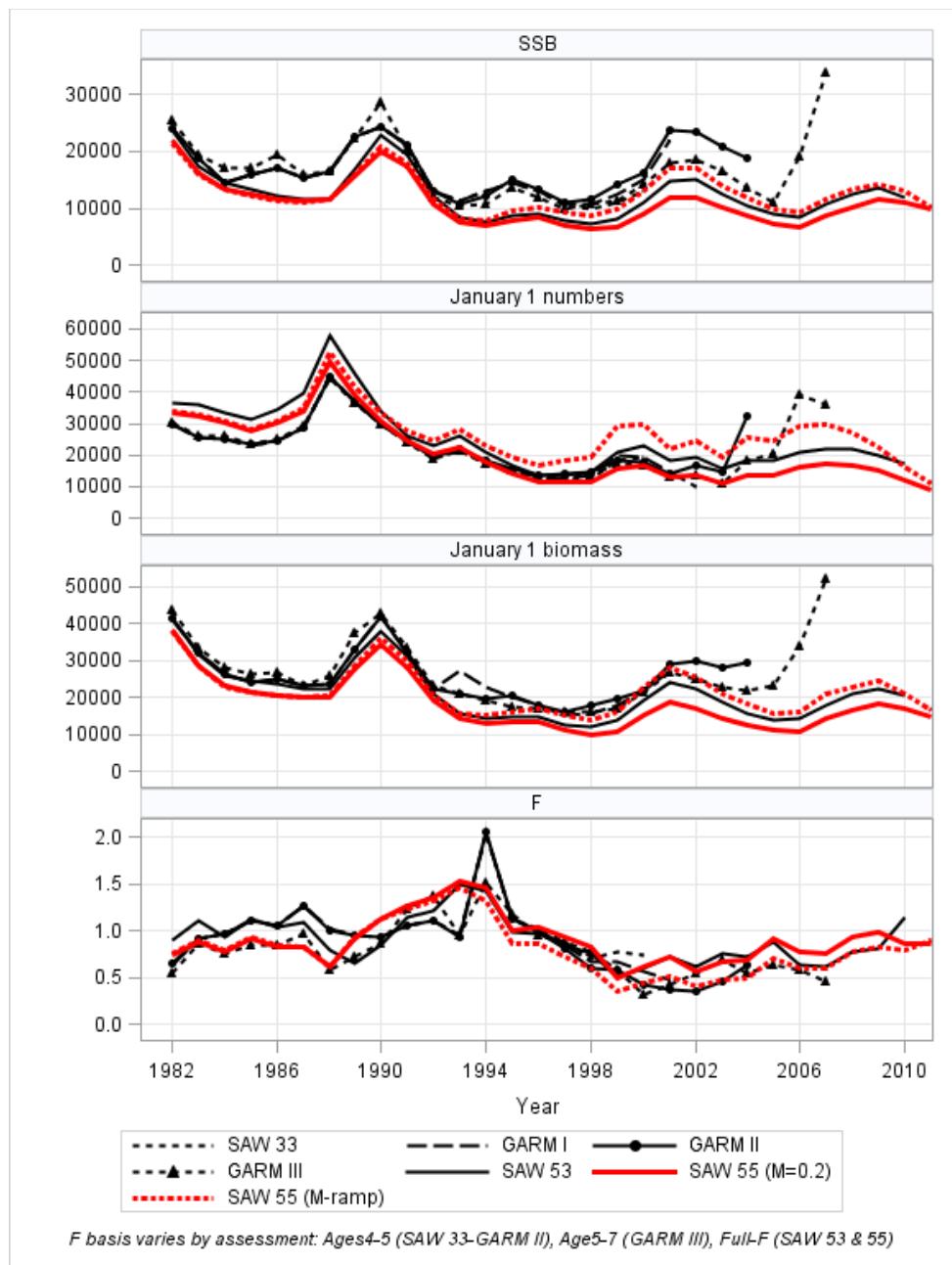


Figure A.241. Comparison of estimates of average spawning stock biomass (SSB), January 1 stock numbers, January 1 stock biomass, and fishing mortality from previous Gulf of Maine Atlantic cod stock assessments including estimates from the ASAP SAW55_3BLOCK_BASE ($M = 0.2$) and SAW55_3BLOCK_BASE_M_SPLIT (M -ramp) models. *Note that the ages included in the average F calculation are not constant across assessments.

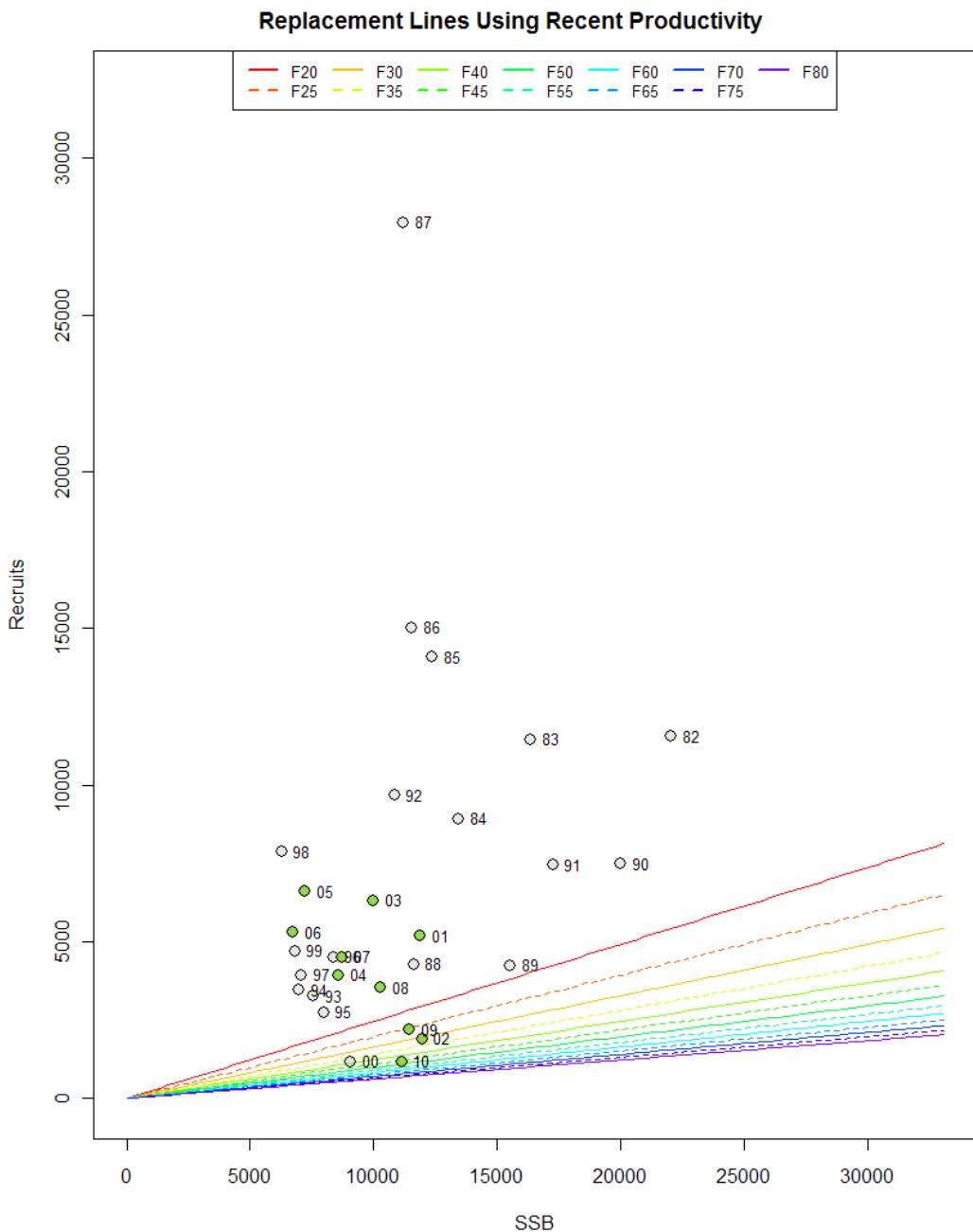


Figure A.242. Comparison Gulf of Maine Atlantic cod replacement lines under a range of percent spawner per recruit values based on an assumption of $M = 0.2$ (based on SAW55_3BLOCK_BASE model). The most recent ten years of recruitment observations (2001-2010) are highlighted green.

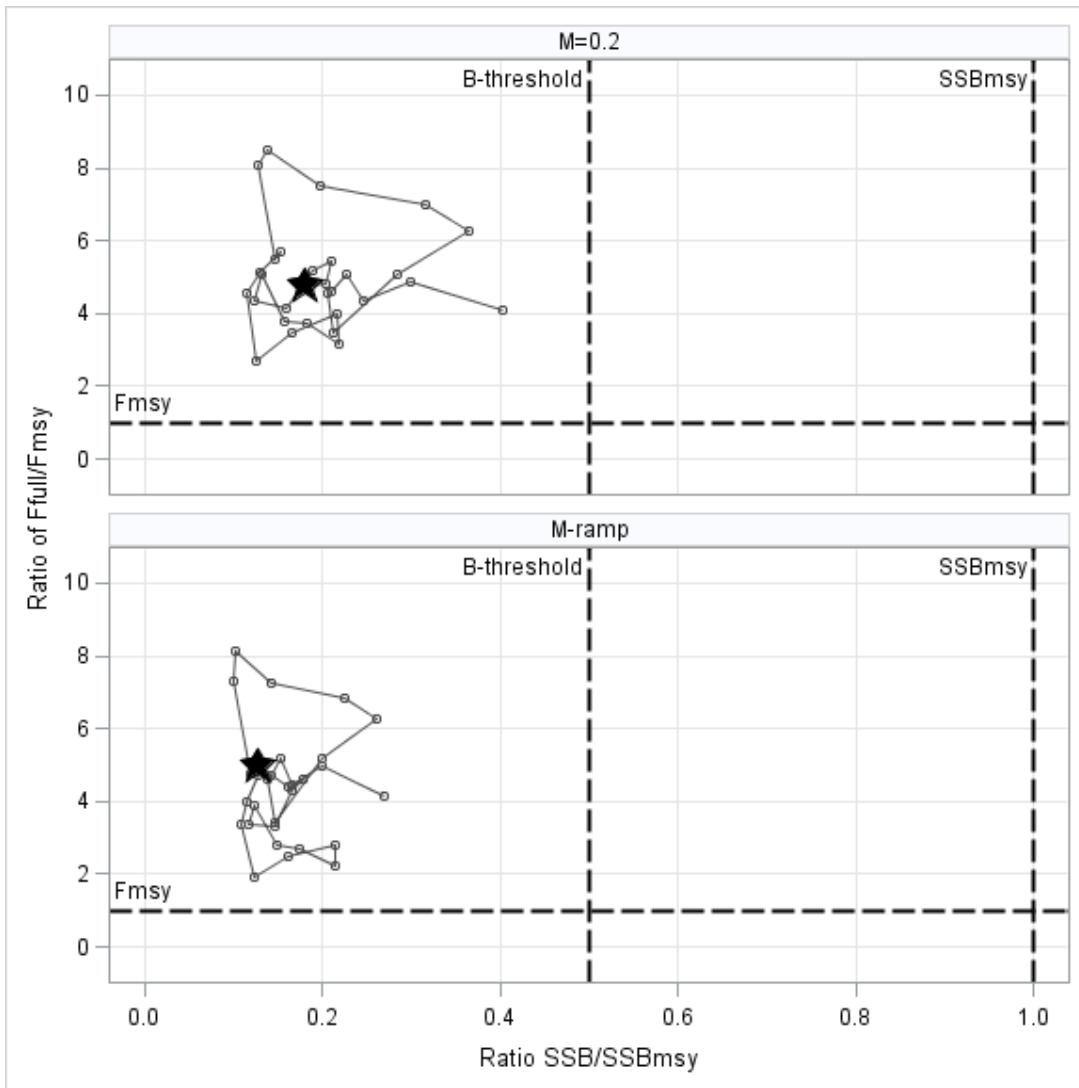


Figure A.243. Time series plot of the Gulf of Maine Atlantic cod fully selected fishing mortality/2011 F_{MSY} ratio relative to the spawning stock biomass/2011 SSB_{MSY} ratio from 1982 to 2011. The 2011 data point is indicated by a black star. Results are shown for both the $M = 0.2$ (top) and M -ramp (bottom) models.

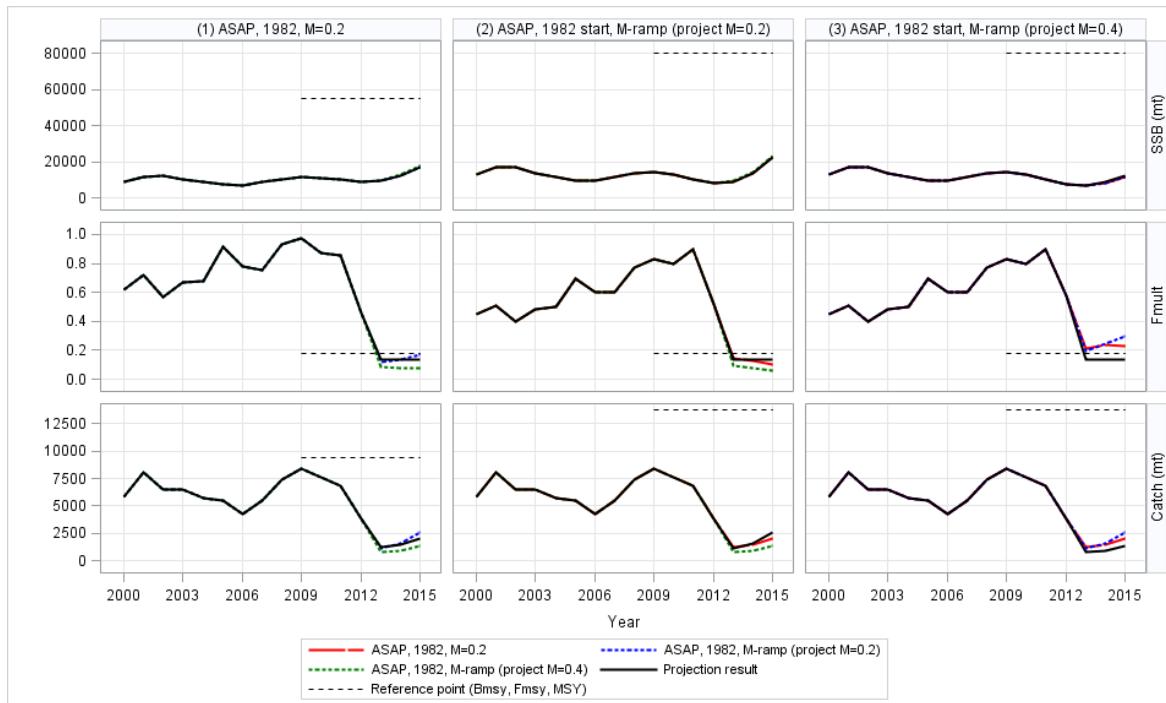


Figure A.244. Trends in Gulf of Maine cod SSB (top row), fully recruited fishing mortality (middle row) and catch (bottom row) during 2000 – 2015; column headers indicate ‘true’ state of nature; cells provide trend in indicator under ‘true’ state of nature when catch during projection period (based on 75% F_{MSY} is correctly specified (black) and mis-specified (red: ASAP, 1982, $M = 0.2$; blue: ASAP, 1982, M -ramp (project at $M = 0.2$); green: ASAP, 1982, M -ramp (project at $M = 0.4$); MSY proxy reference points indicated in dashed line on each plot.