KINETIC STUDY OF THE DISCOLORATION OF RHODAMINE B WITH PERSULFATE, IRON ACTIVATION

A. H. Mcheik, M. M. El Jamal

Chemistry Department, Faculty of Sciences (I), Lebanese University, Hariri Campus,

El Hadath, Lebanon
Email: mjamal@ul.edu.lb

Received 09 January 2013 Accepted 05 June 2013

ABSTRACT

The discoloration of the fluorescent dye, Rhodamine B (Rhd B) by persulphate (KSP), was investigated by spectro-photometric methods. The discoloration reaction of Rhd B by KSP is slow. It's first order with respect to Rhd B and KSP. It's effective at any pH and was found to increase linearly for pH< 4, temperature, and after addition of Fe(II) and Ag(I). The rate of discoloration remained the same after addition of Co(II), Ni(II), Cu(II) and carbonate salts, but decreased after addition of ethanol and chloride salts. The TLC experiment showed the formation of colored intermediates. The activation parameters (E_a , ΔG^{\sharp} , ΔH^{\sharp} and ΔS^{\sharp}) of the discoloration reaction of Rhd B by KSP in absence and in presence of Fe(II) were calculated.

Keywords: Rhodamine B, persulfate, metal activation, matrix effect, kinetic.

INTRODUCTION

Textile industry produces a large amount of dyed wastewater. The non-biodegradable nature of the spent dye baths constitutes serious environmental problems [1]. Rhodamine B (Rhd B), is one of the important xanthene cationic dyes, is largely applied in industry as a dye for paper, silk, wool, jute leather, cotton (Fig. 1). The xanthene dyes (Rose Bengal, Eosin Y and Rhodamine 6G,...) are used also as fluorescent markers in structural microscopic studies, photosensitizers, and laser dyes. They have long been known to be toxic to insects (insecticide). They are harmful if swallowed by human and animals, and cause irritation to the skin, eyes and lung [2-3].

The main technologies available for the treatment of dyes involve the transfer of the pollutant from a liquid phase to another phase, concentrating the dye on an adsorbent (for example) so that it can later be discarded in a landfill or incinerated [4-5]. This phase exchange is

not an ideal remedy. Destructive oxidation treatments provide more permanent solutions such as electrodegradation [6, 7], photodegradation using TiO₂ [8-13] are applied for removal of dyes (especially cationic dyes) from textile effluents.

Chemical oxidation of contaminants by oxidants has been studied to develop novel remediation technologies: KBrO₃ [14, 15], KClO₃ [16], O₃ [17], UV/O₃ [18], activated O₂ [19], Fenton's reagent [20 -21], photo Fenton [22], H₂O₂ [23] have been widely tested in laboratory for the degradation of basic dyes.

The use of persulfate (KSP) has recently been the focus of attention for an alternative oxidant in the chemical oxidation of contaminants [24-37]. It's one of the strongest oxidants known in aqueous solution (Table 1) [27, 38]. It has great capability for degrading numerous organic contaminants through free radicals (SO₄^{*} and OH*) generated in the persulfate system [30, 31]. It offers some advantages over other oxidants: Ease of storage and transport, high stability, high aqueous solubility

$$H_3C$$
 H_3C
 H_3C

Fig.1. Chemical structures of Rhodamine B, Fluorescein, and Rhodamine 110.

and relatively low cost.

Base activation [36, 37], UV-irradiation [25, 26, 35], thermal activation [34, 39-41], metal activation [42-46], and soil mineral activation [30] have been used to generate a sulfate radical (E°: 2.5- 3.1V), a stronger oxidant than persulfate (2.01 V) [41, 45]:

$$\begin{split} &S_2O_8^{2^-} + heat/UV \to 2SO_4^{-*} \\ &S_2O_8^{2^-} + 2M^{n+} \to 2SO_4^{-*} + 2M^{(n+1)+} \\ &SO_4^{-*} + H_2O \xrightarrow{\leftarrow} HO^* + H^+ + SO_4^{2^-} \end{split}$$

The present work is focused on the kinetic study of the discoloration of Rhd B (model molecule of xanthenes dye) with persulfate by UV- Visible spectrophotometry. The effect of various parameters such as initial pH, initial concentration of persulfate and Rhd B, transition metals concentration, and salts concentration was studied. A comparison between the degradation of Rhd B and Fluorescein (which has similar structure (Fig.1) with persulfate is undertaken to interpret better the results.

EXPERIMENTAL

Rhodamine B is used as purchased from Sigma (C₂₈H₃₁N₂O₃Cl, CAS: 81-88-9, MW: 479.02 g). The other chemical reagents used are from BDH (Analar or GPR). 20 mg L⁻¹(4.18×10⁻⁵ M) of Rhd B solution is prepared in distilled water in order to study its discoloration by KSP. The concentration of the dye in the reactional mixture is selected in such a way the absorbance of the dye at 560 nm followed Beer's law. The concentration of KSP in the reactional mixture (0.01 M) is selected in such a way the rate of the reaction is not very slow. The kinetic study of the discoloration of Rhd B by persulfate was carried on a double beam spectrophoto-meter, Specord 200 (Analytical Jena). 1 M of salt solutions (carbonate, chloride, and sulfate), and 0.1 M of transition

metal solutions ($Co(NO_3)_2$, $Ni(NO_3)_2$, and $CuSO_4$), were also prepared to study the matrix effect on the discoloration rate. $5x10^{-2}$ M of $AgNO_3$, $Fe^{II}(NH_4)_2(SO_4)_2$. 6 H_2O , and 10^{-2} M of the surfactant (SDS) were also prepared for the same reason. H_2SO_4 and NaOH solutions were used to study the effect of pH on the discoloration rate of Rhd B with KSP.

Kinetic Study

The order with respect to Rhd B was carried out in the presence of a large excess of persulfate. The reactional mixture was prepared as follows: x ml (2 < x < 4) of 20 mg L⁻¹ (0.93×10⁻⁵ M< [RhdB]_f < 1.85x10⁻⁵M) of Rhd B, (7- x) mL of H₂O and 2 mL of 5x10⁻² M KSP ([KSP]_f: 0.11 M).

The order with respect to persulfate was determined in the presence of different excess amounts of it, while the other parameters were kept constant: x mL of $5x10^{-2}$ M KSP (2 < x < 4) (7-x) mL of H_2O and 2 mL of 20 mg L^{-1} Rhd B ([RhdB] $_f$: 0.93×10^{-5} M). Quickly after the addition of the dye, the absorbance of the solution was recorded every 1 min. during 60 minutes at 560 nm (λ_{max} of Rhd B). The rate expression for the discoloration reaction is:

Rate =
$$k[S_2O_8^{2-}]^n[Rhd B]^m = k_{app}[Rhd B]^m$$

with $k_{app} = k[S_2O_8^{2-}]^n$

where k is the rate constant of the reaction, m and n are the pseudo order of the reaction with respect to Rhd B and $K_2S_2O_0$ respectively.

Effect of initial pH

The effect of pH on the degradation rate was studied in presence of several concentrations of NaOH or $\rm H_2SO_4$. The following mixtures were prepared for this purpose: 2 mL of 20 mg L⁻¹ Rhd B ([RhdB]_f: 0.93x10⁻⁵ M), 2mL of 5x10⁻² M of KSP (KSP]_f: 0.11 M), 5 mL of the acid

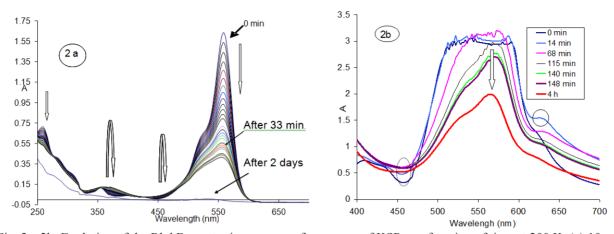


Fig. 2a, 2b: Evolution of the Rhd B spectra in presence of an excess of KSP as a function of time at 290 K. (a):10 mg L⁻¹ Rhd B (2.1x10⁻⁵ M), 2.5x10⁻² M KSP, spectrum taken every 1min during 1h). (b): 100 mg L⁻¹ Rhd B (21x10⁻⁵ M), 2.5x10⁻² M KSP.

or the base $(5x10^{-1} M - 5x10^{-3} M)$.

Metal activation

Concerning the effect of metal salts, several experiments were done at fixed concentration of KSP and Rhd B in the mixture as follow: 3 mL of 20 mg L⁻¹ Rhd B ([Rhd B]_f: 1.25x10⁻⁵ M), 2 mL of a predetermined concentration of KSP ranging from 5x10⁻² M to 10⁻³ M, x mL of Fe(II) (Fe^{II}(NH₄)₂(SO₄)₂). 6 H₂O, or Ag(I) solution of the same initial concentration of KSP and (5-x) mL of water.

Matrix effect

The effect of salt such as KCl, Na_2SO_4 , and Na_2CO_3 (1 M), on the rate constant was carried out at several concentrations of the mentioned salts by replacing some volume of water by the salt solution (x mL, with $0 \le x \le 5$), while keeping the concentration of the others reactants constant. The effect of the surfactant (SDS) and ethanol, on the discoloration rate were also studied in the same manner.

RESULTS AND DISCUSSION

UV-Visble spectroscopy and TLC Analysis

The visible spectrum of Rhd B decreases slowly with time especially the absorbance at 560 nm (λ_{max} of Rhd B). The general shape of Rhd B spectrum does not vary during its reaction with KSP. The reaction between Rhd B and KSP leads to total discoloration of Rhd B (Fig. 2a). The discoloration is very slow compared to that with crystal violet and malachite green [27]. The

slow decrease in the intense peaks of Rhd B at 560 and 260 nm is accompanied for the first time by an increase in the absorbance at 630, 470 and 370 nm (Fig. 2a, 2b). Then the decrease continues in the whole spectrum.

The TLC study (Silica gel – pure ethanol) shows the appearance of two spots other than that of Rhd B: the lower one, orange less intense than the upper one (violet) (Fig. 3). The two color intermediate products disappear with time, in favors of two others intermediates with higher R_s. The Rhd B spot persists even after long time (one day) but its color faded, and then it disappears completely. The same behavior is observed with the other spots. Finally the Rhd B/ KSP mixture become colorless. The discoloration of low concentration of Rhd B by KSP leads to the formation of colored intermediates undetectable clearly by UV-visible spectroscopy. The decrease in the whole UV - visible spectrum is probably a consequence of total mineralization of Rhd B. No blue shift (progressive decrease from 560 to 498 nm) is observed during the discoloration as observed during the degradation of Rhd B under visible light ir-

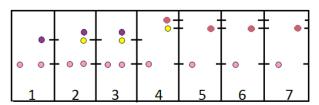


Fig. 3. TLC of the mixture (Rhd B + KSP) done at various moments during one day (from left to right) $(21x10^{-5} \text{ M} \text{ Rhd B}, 2.5x10^{-2} \text{ M KSP}, 290 \text{ K})$. The spot to the left corresponds to Rhd B.

radiation by Zn_xCd_{1-x}S/TiO₂ [47], and ZnCd/TiO₂ [48]. The degradation products of Rhd B are function of the oxidant used [49, 50]. There are several pathways for the Rhd B degradation:

N-deethylation (formation of Rhd B110), followed by deamination [49, 52].

Break in the chemical bond of xanthone molecule between the C_1 and C_{15} atoms (elimination of the phenyl ring containing COOH) leading to the formation of Pyronin G [50].

A break in the -O- bridge giving an intermediate having a similar structure to malachite green (increase in A_{630}).

Further degradation of these intermediate is possible [27, 51, 52].

Order with respect to Rhd B

Rhodamine B shows a strong absorption band at 560 nm and follows Beer's law for low concentrations (2–8 mg L⁻¹). The relation found between the absorbance A_{560} and Rhd B concentration in mg L⁻¹ is $A = 0.196 \times [RhdB]$ (mg L⁻¹) R²: 0.993). The order with respect to Rhd B is not zero according to the plot of A_{560} vs. time (Fig.4a). The functions ln A and 1/A vary linearly with time. The constant k_{obs} is the slope of the plot

$$\ln A_{560}$$
 or $\frac{1}{A_{560}}$ vs time $(k_{obs} \propto k_{app})$.

For all the experiments concerning the determination with respect to Rhd B, the correlation coefficient (R²) for order one is closer to one than that for 2nd order. Order one is confirmed by applying the differential method ($\ln V_o = 1.27 \times \ln[\text{Rhd B}]_o - 10.27$, R²: 1). The constant k_{obs} remains_the same with the increase in the dye concentration ($k_{obs} = 2.9 \times 10^{-4}$). Order one is also observed by discoloration of Rhd B by UV/KSP [29], by photo- Fenton reagent [22], and by photo-degradation in presence of TiO₂ [9].

Order with respect to KSP

The increase in KSP concentration increases the discoloration rate of Rhd B. For 10^{-3} M of KSP (in the mixture), the absorbance of Rhd B remained the same during 1 h, but with 10^{-2} M of KSP, the absorbance decreases significantly. The rate constant (k_{obs}) increases linearly with the increase in KSP concentration ($k_{obs} \times 10^4 = 200 \times [KSP]$, R²: 0.983). So the order with respect to KSP is one. Order one is also observed for the degradation of crystal violet by KSP [27].

Effect of initial pH

Initial pH of solution had great effect on Rhd B degradation rate during $UV/S_2O_8^{2-}$ system. Rhd B is stable in acidic and in basic medium, it does not loose its color with time in these mediums. Its maximum

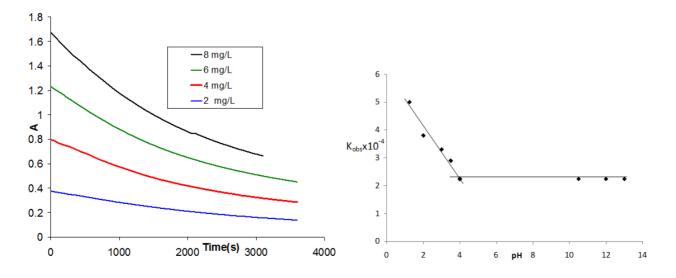


Fig. 4a: Variation of the absorbance of Rhd B (A_{560}) as a function of time for several concentrations of Rhd B in presence of 0.01 M of KSP at 290 K.

Fig. 4b: Effect of initial pH on the discoloration rate constant (k_{obs}). (6 mg L⁻¹ Rhd B (1.25x10⁻⁵ M), 0.01M KSP, 290 K).

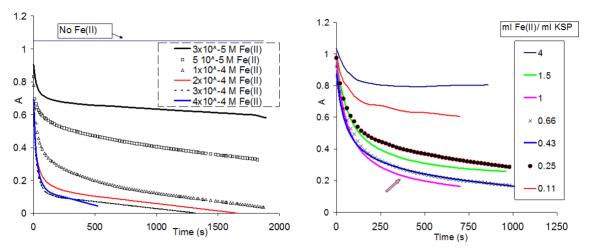


Fig. 5a. Effect of Fe(II) on the discoloration rate of Rhd B with KSP (6 mg L⁻¹ Rhd B (1.25x10⁻⁵ M), 1x10⁻³ M KSP, 290 K). Fig. 5b. Effect of the ratio [Fe(II)]/[KSP] on the discoloration rate of Rhd B with KSP (6 mg L⁻¹ Rhd B (1.25x10⁻⁵ M), $[KSP] = [Fe(II)] = 10^{-3} M$, 290 K).

wavelength remained the same whatever the pH. The results showed that, the discoloration rate constant does not vary in neutral and in basic medium, but increases linearly with the decrease in pH for pH \leq 4 (Fig. 4b). The radical SO_4^{*-} is rather stable and more numerous at low pH, while increasing system pH results in the transformation of SO_4^{*-} to hydroxyl radical OH* ($SO_4^{*-} + H_2O \xrightarrow{\longleftarrow} OH^* + SO_4^- + H^+$) [29]. The acidic medium has positive effect on the Rhd B discoloration due to shift to right in the equilibrium of the reaction mentioned. The obtained result is in accordance with that obtained by action of UV/ KSP on Rhd B [29] and by action of KSP on Orange G [45].

Matrix effect

Metal Activation

The chemical oxidation of organic pollutants by persulfate can be accelerated by metal ion activation, through enhancing sulfate radical generation. In the present case Ni(II), Co(II) and Cu(II) did not increase the discoloration rate constant, whereas the ions Ag(I) and Fe(II) accelerate it considerably (Fig. 5 and 6). For the same concentration of metal ion, Fe(II) decolorizes the mixture (Rhd B + KSP) faster than Ag(I). The discoloration behavior of Rhd B in presence of Fe(II)

is somehow special:

The control mixture (presence of Fe(II) without KSP) shows no discoloration of Rhd B. For $5x10^{-2}$ M (or $5x10^{-3}$ M) as initial concentration of KSP and Fe(II): The decrease in Fe(II) concentration in the medium increases the discoloration rate (result not shown). For the experiment done with $5x10^{-2}$ M of KSP and Fe(II), the mixtures with) Fe(II)/[KSP] ≥ 2 , the discoloration is completely stopped, but for other mixtures with low ratio (≤ 0.5), the color disappears completely in less than 1 minute.

The repetition of the experiment mentioned above with 5×10^{-3} M of KSP and 10^{-3} M of Fe(II) (instead of 5×10^{-3} M) as initial concentration showed different attitude: the discoloration rate increases with the increase in Fe(II) volume (0.3 ml < x < 4 ml or 3×10^{-5} M $- 4 \times 10^{-5}$ M) (Fig.5a). Linear relation is observed between k_{obs} and Fe(II) (R^2 : 0.97). This behavior is the opposite of that observed with higher concentrations of KSP and Fe(II).

Basing on this, we think about the existence of an optimum ratio of Fe(II)/[KSP]. This ratio corresponds to 1 when using 10^{-3} M as initial concentration of KSP and Fe(II) (2 ml of 10^{-3} M KSP + x ml of 10^{-3} M Fe(II)).

Table 1. The standards potentials of some oxidants used in advanced chemical oxidation [27, 38].

Oxidant	BrO ₃ / Br ₂	ClO ₃ ⁻ / Br ₂	H_2O_2/H_2O	O ₃ /O ₂	OH*	S ₂ O ₈ ²⁻ /SO ₄ ²⁻
E ^o (V)	1.51	1.46	1.76	2.07	2.38	2.01

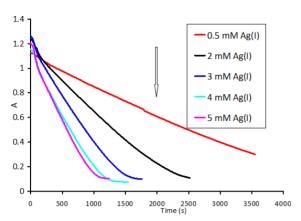


Fig. 6. Effect of Ag(I) concentration on the discoloration rate constant of Rhd B with KSP. (6 mg L^{-1} Rhd B (1.25x10⁻⁵ M), 0.01 M KSP, 290 K).

The optimum ratio of Fe(II)/[KSP] is also confirmed by varying the volume of 10⁻³ M KSP with respect to that of 10⁻³ M Fe(II) (Fig.5b). There is scavenger effect [27, 45].

Concerning the effect of Ag(I), the control mixture (presence of Ag(I) without KSP) shows no discoloration of Rhd B. For an initial concentration of KSP (5×10^{-2} M), the addition of x ml of 5×10^{-2} M Ag(I) ($0 \le x \le 5$), the discoloration of Rhd B became very fast, and increased linearly with the increase of Ag(I) concentration. The repetition of the experiment with an initial concentration of KSP and Ag(I) equal to 0.01 M gave the same results but the discoloration is slower ($k_{obs} = 0.226 \times [{\rm Ag}^+] + 2 \times 10^{-5}$, with R²: 0.982) (Fig. 6). No Scavenger effect is observed with Ag(I) as was the case with Fe(II). Similar result is observed during the

discoloration of crystal violet by KSP [27]. For another initial concentration of KSP and Ag(I) (2.5x10⁻³ M), the increase in Ag(I) concentration in the reactional mixture increases slightly the rate constant, whereas in the same conditions, the rate constant is higher with Fe(II). The order with respect to Rhd B in presence of Ag(I) becomes zero order, whereas it is rather 2 in presence of Fe(II), so the action of Ag(I) on the discoloration of Rhd B by KSP is different from that of Fe(II).

Effect of wastewater constituents

The high redox potential of sulfate free radical makes it very reactive in destroying organic contaminants. However, parallel reactions with other species, could result in scavenging of sulfate radical and could possibly limit its oxidation efficiency. Competition for sulfate radical could be from wastewater constituents such as chloride ions, carbonate and phosphate [45]. The equations below show the chemical reaction mechanism of persulfate with chlorides in aqueous solutions [31]:

$$SO_4^{*-} + Cl^- \leftrightarrow SO_4^{2-} + Cl^*$$

$$Cl^* + Cl^- \leftrightarrow Cl_2^{*-}$$

$$Cl_2^{*-} + Cl_2^{*-} \leftrightarrow Cl^- + Cl_3^{--}$$

When chloride concentrations were less than 0.05 M, presence of chloride ions had insignificant impacts on Rhd B degradation. However, beyond these concentration levels, Rhd B degradation rates reduced significantly with an increase in chloride concentrations (Fig. 7a) The presence of sulfate or carbonate even in high concentration (0.5 M) did not reduce the discoloration

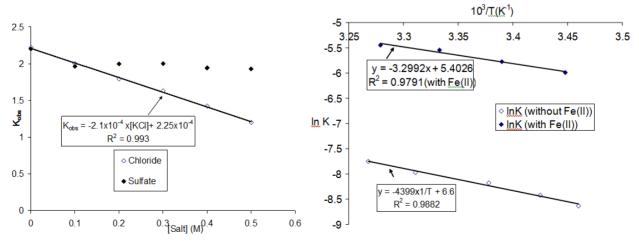


Fig. 7a. Effect of salts concentration on the degradation of Rhd B by KSP (6 mg L⁻¹ Rhd B (2.5x10⁻⁵ M), 0.01 M KSP, 290 K) Fig. 7b. Arrhenius plot of lnk_{obs} vs. 1/T. (12 mg L⁻¹ Rhd B (2.5x10⁻⁵ M) + 0.01 M K₂S₂O₈).

Table 2. Activation thermodynamic parameters of the degradation of Rhd B by persulfate. (12 mg L^{-1} Rhd B (2.5x10⁻⁵ M) + 0.01 M $K_2S_2O_8$) The values between () correspond to those in presence of 5x10⁻³ M Fe(II).

$E_a(\text{kJ mol}^{-1})$	ΔH^{\neq} (kJ mol ⁻¹)	ΔS^{\neq} (kJ.mol ⁻¹ K ⁻¹)	$\Delta G^{\#}_{298} (\text{kJ mol}^{-1})$
36.5 (27.3)	34 (24.8)	- 0.198 (-0.201)	93.0 (86.8)

rate significantly. Chloride was the ion that had the most negative effect. The reaction between sulfate and sulfate radical will generates sulfate radical. Similar results are observed with Lee et al. [31] and Soares et al. [12].

Effect ethanol and surfactant

Ethanol decreases the discoloration rate of Rhd B. Linear relation is observed between $k_{\rm obs}$ and the volume of pure ethanol added ($k_{\rm obs}=-0.373\times V({\rm ethanol})+1.87$ with R²:0.985). The presence of SDS in low concentration (1x10⁻⁴ M - 4x10⁻⁴ M) did not affect the discoloration rate of Rhd B by KSP. The increase in SDS concentration did not allow the measurement of the absorbance accurately due to the micelle formation.

Effect of Temperature

The increase in temperature $(15^{\circ}\text{C} - 33^{\circ}\text{C})$ increases the discoloration rate of Rhd B (Fig.7b). The activation parameters associated with the discoloration are calculated according to Arrhenius and Eyring equations [27] (Table 2). The activation energy in presence of Fe(II) is lower than in its absence. The decrease in the activation energy in presence of Fe(II) confirms the catalyst effect of Fe(II) (metal activation).

CONCLUSIONS

The discoloration of Rhd B is pseudo first order with respect to Rhd B and to persulfate. The discoloration increases with temperature, in acidic medium, after addition of Fe(II) and Ag(I), but decreases with the addition of NaCl and ethanol. There was no effect on the rate constant upon addition of Cu (II), Ni(II), or Co(II) and carbonate. Colored intermediate products are observed by TLC during the discoloration. The reaction of Rhd B with KSP leads probably to total mineralization of the dye.

REFERENCES

1. S. Wang, Z.H. Zhu, Effect of acid treatment of activated carbons on dye adsorption, Dyes and Pigments,

- 75, 2007, 306-314.
- E. R. Nestmann, G. R. Douglas, T. I. Matula, C. E. Grant, and D. J. Kowbel, Mutagenic Activity of Rhodamine Dyes and Their Impurities as Detected by Mutation Induction in Salmonella and DMA Damage in Chinese Hamster Ovary Cells, CANCER RESEARCH, 39, 1979, 4412-4417.
- 3. V.K. Gupta, Suhas, Application of low-cost adsorbents for dye removal- a review, J. Environ. Manage., 90, 8, 2009, 2313–2342.
- M. Rafatullah, O. Sulaiman, R. Hashima, and A. Ahmad, Adsorption of Methylene Blue on Low-Cost Adsorbents: A review, J. Hazard. Mat., 177, 2010, 70-80.
- 5. S. J. Allen, B. Koumanova, Decolourisation of water/wastewater using adsorption (review), J. Univ. Technol. Metall. (Sofia), 4, 3, 2005, 175-192.
- 6. Ricardo E. Palma-Goyes, Fernando L. Guzmán-Duque, Gustavo Peñuela, Ignacio González, Jose L. Nava, Ricardo A. Torres-Palma, Electrochemical degradation of crystal violet with BDD electrodes: Effect of electrochemical parameters and identification of organic by-products, Chemosphere, 81, 2010, 26-32.
- M. Panizza, A. Barbucci, R. Ricotti, G. Cerisola, Electrochemical Degradation of Methylene Blue, Sep. Purif. Technol., Sep. Purif. Technol., 54, 2007, 382-387.
- 8. E.G.L. Oliveira, J.J. Rodrigues Jr., H.P. de Oliveira, Influence of surfactant on the fast photodegradation of rhodamine B induced by TiO₂ dispersions in aqueous solution, Chem. Eng. J., 172, 2011, 96-101.
- P. Wilhelm, D. Stephan, Photodegradation of Rhodamine B in aqueous solution via SiO2@TiO₂ nano-spheres, J. of Photochem. and Photobiol. A: Chemistry, 185, 2007, 19-25.
- M. Aliabadi, T. Sagharigar, Photocatalytic Removal of Rhodamine B from Aqueous Solutions Using TiO₂ Nanocatalyst, J. Appl. Environ. Biol. Sci., 1, 12, 2011, 620-626.

- J. Li, L. Li, L. Zheng, Y. Xian, L. Jin, Photoelectrocatalytic degradation of rhodamine B using Ti/TiO₂ electrode prepared by Laser calcination method, Electrochim. Acta, 51, 2006, 4942-4949.
- B. E. T. Soares, M. A. Lansarin, and C. C. Moro, A Study of process variables for the Photocatalytic degradation of Rhodamine, Braz. J. Chem. Eng., 24, 01, 2007, 29-36
- 13. J.-M. Wu, T.-W. Zhang, Photodegradation of Rhodamine B in water assisted by titania films prepared through a novel procedure, J. of Photochem. and Photobiol. A: Chemistry, 162, 2004, 171-177
- A. H. Gemeay, G. R. El-Ghrabawy, A. B. Zaki, Kinetics of the oxidative decolorization of Reactive Blue-19 by acidic bromate in homogeneous and heterogeneous media, Dyes and Pigments, 73, 2007, 90 -97.
- X. Zhu, Y. Zhang, Inhibition kinetic determination of trace amount of iodide by the spectrophotometric method with Rhodamine B, Spectrochimica Acta Part A, 70, 2008, 510-513.
- 16. Y. Mohammed, A. C. Etonihu, V. A. Tsaku, Hexamethylpararosalinine (Crystal Violet) Oxidation by Chlorate Ions in Aqueous Acidic Medium: Kinetic Approach to the mechanism of reaction, Trakia J. of Sci., 9, 2, 2011, 1-7.
- E. Kusvuran, O. Gulnaz, A. Samil, Ö. Yildirim, Decolorization of malachite green, decolorization kinetics and stoichiometry of ozone-malachite green and removal of antibacterial activity with ozonation, Processes, J. of Haz. Mat., 186, 2011, 133-1043
- B. Cuiping, X. Xianfeng, G. Wenqi, F. Dexin, X. Mo, G. Zhongxue, X. Nian, Removal of Rhodamine
 B by ozone-based advanced oxidation process, Desalination, 278, 2011, 84-90
- K. J. Deng, F. Huang, D. Y. Wang, Z. H. Peng, Y. H. Zhou, Green Oxygenation Degradation of Rhodamine B by Using Activated Molecule Oxygen, Chinese Chem. Letters, 15, 10, 2004, 1223-1226.
- H.-J. Fan, S.-T. Huang, W.-H. Chung, J.-L. Jan, W.-Y. Lin, C.-C. Chen, Degradation pathways of crystal violet by Fenton and Fenton-like systems: Condition optimization and intermediate separation and identification, J. of Haz. Mat., 171, 2009, 1032-1044.
- 21. B. H. Hameed, T.W. Lee, Degradation of malachite green in aqueous solution by Fenton process, J. of

- Haz. Mat., 164, 2009, 468-472.
- 22. A. Kumar, M. Paliwal, R. Ameta, S. Ameta, Photochemical treatment of rhodamine-B wastewater by photo-Fenton reagent, Indian J. of Chem. Technol., 15, 2008, 7-11
- 23. J. C. Edwards, Investigation of Color Removal by Chemical Oxidation for Three Reactive Textile Dyes and Spent Textile Dye Wastewater, MS of Sci. in Environ. Sci. and Eng., Blacksburg, Virginia, July 18, 2000
- 24. S. Gokulakrishnan, P. Parakh, H. Prakash, Degradation of Malachite green by Potassium persulphate, its enhancement by 1,8-dimethyl-1,3,6,8,10,13-hexaazacyclotetradecane nickel(II) perchlorate complex, and removal of antibacterial activity, J. of Haz. Mat. 213-214, 2012, 19-27,
- 25. K. Djebbar, S. Aliouche, H. Chenini, T. Sehili, Decolourization process of an azoique dye (Congo red) by photochemical methods in homogeneous medium, Desalination 250, 2009, 76-86 (KSP, H₂O₂)
- 26. Y.-T. Lin, C. Liang, J.-H. Chen, Feasibility study of ultraviolet activated persulfate oxidation of phenol, Chemosphere, 82, 2011, 1168-1172.
- 27. L.M.A. Fayoumi, M.A. Ezzedine, H.H. Akel, M.M. El Jamal, Kinetic Study of the Degradation of Crystal Violet by K₂S₂O₈, Comparison with Malachite Green, Portugaliae Electrochimica Acta, 30, 2, 2012, 121-133.
- 28. P. A. Block, R. A. Brown, D. Robinson, Novel Activation Technologies for Sodium Persulfate In Situ Chemical Oxidation Proceedings of the Fourth International Conference on the Remediation of Chlorinated and Recalcitrant Compounds, 2004.
- 29. X. Chen, Z. Xue, Y. Yao, W. Wang, F. Zhu, and C. Hong, Oxidation Degradation of Rhodamine B in Aqueous by UV/S₂O₈ Treatment System, Hindawi Publishing Corporation International Journal of Photoenergy, Volume 2012, Article ID 754691, 5 P.
- 30. C. Liang, C.-F. Huang, Y.-J. Chen, Potential for activated persulfate degradation of BTEX contamination, water research, 42, 2008, 4091-4100
- 31. Y. Lee, S. Lo, J. Kuo, C. Hsieh, Decomposition of perfluorooctanoic acid by microwave activated persulfate: Effects of temperature, pH, and chloride ions, Front. Environ. Sci. Engin., 6, 1, 2012, 17-25.
- 32. C.S. Liu, K. Shih, C.X. Sun, F. Wang, Oxidative degradation of propachlor by ferrous and copper ion

- activated persulfate, Sci. of the Total Environment, 416, 2012, 507–512
- L. R. Bennedsen, J. Muff, E. G. Søgaard, Influence of chloride and carbonates on the reactivity of activated persulfate, Chemosphere, 86, 11, 2012, 1092-1097
- 34. K. Huang, Z. Zhao, G. E. Hoag, A. Dahmani, and P.A. Block, Degradation of volatile organic compounds with thermally activated persulfate oxidation, Chemosphere, 61, 2005, 551-560.
- 35. T. K. Lau, W. Chu, N. J. D. Graham, The Aqueous Degradation of Butylated Hydroxyanisole by UV/S₂O₈²⁻ Study of Reaction Mechanisms via Dimerization and Mineralization, Environ. Sci. Technol., 41, 2, 2007, 613–619.
- 36. M. A. Miraglio, Master of Sci. in Env. Eng., Washington State Univ., Dept. of Civil and Env. Eng., 2009, "Base - Activated persulfate treatment of contaminated soils with pH drift from alkaline to circumneutral".
- 37. O. S. Furman, A. L. Teel, R. J. Watts, Mechanism of base activation of persulfate, Environ. Sci. Technol., 44, 16, 2010, 6423-6428.
- A. J. Bard., R. Parsons, J. Jordan, Standard Potentials in Aqueous Solution; Marcel Dekker: New York and Basel, IUPAC 1985
- 39. K.C. Huang, R.A. Couttenye, G.E. Hoag, Kinetics of heat-assisted persulfate oxidation of methyl tert-butyl ether (MTBE), Chemosphere 49, 2002, 413-420.
- R.H. Waldemer, P.G. Tratnyek, R.L. Johnson, J.T. Nurmi, Oxidation of chlorinated ethenes by heatactivated persulfate: kinetics and products Environ. Sci. Technol., 41, 2007, 1010-1015.
- 41. C. Liang, C.J. Bruell, Thermally activated Persulfate oxidation of trichloroethylene: Experimental investigation of reaction order, Ind. Eng. Chem. Res., 47, 9, 2008, 2912-2918.
- 42. G.P. Anipsitakis, D.D. Dionysiou, Radical generation by the interaction of transition metals with common oxidants, Environ. Sci. &Technol., 38, 2004, 3705-3712.
- 43. M. H. Rasoulifard, S. M. M. D Mohammadi, A. Heidari, E. Farhangni, Degradation of acid red 14

- by silver ion-catalyzed peroxydisulfate oxidation in an aqueous solution TUBITAK, Turkish J. Eng. Env. Sci., 36, 2012, 73-80.
- 44. W.E Fristad, J. A. Klang, Silver(I)/persulfate oxidative decarboxylation of carboxylic acids Arylacetic acid dimerization., Tetrahedron Letters, 24, 22, 1983, 2219-2222.
- 45. X.-R. Xu, X.-Z. Li, Degradation of azo dye Orange G in aqueous solutions by persulfate with ferrous ion, Sep. Purific. Technol., 72, 2010, 105-111.
- 46. Oh, S.Y., S.G. Kang and P.C. Chiu. "Degradation of 2, 4-dinitrololuene by persulfate activated with zero valent iron, Sci. of the Total Environ., 408, 2010, 3464-3468.
- 47. L. Wenjuan, L. Danzhen, M. Sugang, Wei Chen, Xianzhi Fu, and Yu Shao Novel Approach To Enhance Photosensitized Degradation of Rhodamine B under Visible Light Irradiation by the Zn_xCd_{1-x}S/TiO₂ Nanocomposites, Environ. Sci. Technol., 45, 7, 2011, 2987–2993
- 48. T. Watanabe, T. Takizawa, K. Honda, Photocatalysis through excitation of adsorbates. 1. Highly efficient N-deethylation of rhodamine B adsorbed to CdS, J. Phys. Chem., 81, 1977, 1845-1851.
- 49. X. Hu, T. Mohamood, W. Ma, C. Chen, and J. Zhao, Oxidative Decomposition of Rhodamine B Dye in the Presence of VO₂⁺ and / or Pt(IV) under Visible Light Irradiation: N-Deethylation, Chromophore Cleavage, and Mineralization, J. Phys. Chem. B., 110, 2006, 26012-26018
- Y. Ma, X. Jin, M. Zhou, Z. Zhang, X. Teng, H. Chen, Chemiluminescence behavior based on oxidation reaction of rhodamine B with cerium (IV) in sulfuric acid medium, Anal. Chim. Acta, 489, 2003, 173–181
- 51. N. Khalfaoui, H. Boutoumi, H. Khalaf, N. Oturan and Mehmet A. Oturan, Electrochemical Oxidation of the Xanthene Dye Rhodamine 6G by Electrochemical Advanced Oxidation Using Pt and BDD Anodes, Current Organic Chem., 16, 2012, 2083-2090.
- 52. Y.-M. Zheng, R. F. Yunus, K.G. N. Nanayakkara, and J. P. Chen, Electrochemical Decoloration of Synthetic Wastewater Containing Rhodamine 6G: Behaviors and Mechanism, Ind. Eng. Chem. Res., 51, 2012, 5953–5960.