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GENOMICS PAPER

Time-Series Sequences for Evolutionary Inferences

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Introduction

The application of genomics to evolutionary biology has provided unprecedented power and resolution to investigate processes like mutation (Nachman and Crowell 2000; Rozhok and Degregori 2019), genetic drift (Whitney and Garland 2010; Funk et al. 2016), gene flow (Gallego-García et al. 2019; LaCava et al. 2021), and natural selection (Brauer et al. 2016; Martins et al. 2018). Evolutionary genomics has historically used contemporary or single-timepoint samples to study microevolutionary processes that most often depend on idealized model assumptions (e.g., Wright-Fisher model) (Butlin 2008; Hoban et al. 2016). For example, genome scanning methods aim to detect natural selection by assuming that the impact of selection can be discerned from the effects of neutral evolutionary processes on genetic differentiation (Hoban et al. 2016). However, different demographic scenarios can also produce signatures of selection, leading to false positives (Lotterhos and Whitlock 2015; Haasl and Payseur 2016). Similarly, when using measures of genetic differentiation to estimate gene flow, most models assume that population size is large enough that genetic drift is negligible—that is, not driving neutral divergence between populations (Whitlock and McCauley 1999; Ma et al. 2015).

Overall, violating model assumptions (e.g., neglecting the role of drift) can lead to erroneous results that compound into incorrect inferences (Waples 1998; Strasburg and Rieseberg 2010). On one hand, simulation and methodological research may attempt to remove or account for some of these assumptions (Bradburd et al. 2013; Lotterhos and Whitlock 2015;

Ma et al. 2015; Haasl and Payseur 2016). On the other hand, to model processes and predict changes over time, some researchers have employed space-for-time substitutions, which assume that the drivers of biotic change such as changes in allele frequency, species abundance, and community structure between segregated populations (spatially or environmentally) also drive biotic changes across time (Pickett 1989; Fitzpatrick and Keller 2015). Space-for-time substitutions often either use a "different aged sites" or "spatiotemporal turnover" framework (Wogan and Wang 2018). In the "different aged sites" framework, researchers infer rates of evolutionary change using populations of different ages (i.e., different colonization times) or populations that have experienced an environmental condition for different lengths of time as a substitute for longitudinal samples. For example, Terekhanova et al. (2014) used freshwater lakes of different ages to quantify the strength of positive selection on alleles related to adaptation to freshwater environments in the threespine stickleback, Gasterosteus aculeatus. Terekhanova et al. (2014) used populations of different ages as time-series snapshots to investigate the temporal dynamics of local adaptation. Under a "spatiotemporal turnover" framework, populations along an environmental gradient (e.g., temperature, urbanization) are used to isolate the impact of abiotic variables on population differences (e.g., allele frequencies, population size) to predict population responses to environmental change through time (Wogan and Wang 2018). Bay et al. (2018)used genomic data from the yellow warbler, Setophaga petechia, to quantify the association between genomic variation and climate before applying these associations in a "spatiotempoI772 A. A. Snead and F. Alda

ral turnover" framework to predict potential population declines with future climate change. Both Terekhanova et al. (2014) and Bay et al. (2018) used single-timepoint samples to make inferences about temporal changes in allele frequencies or populations, albeit with different approaches. The key similarity is the use of contemporary samples from populations that are spatially segregated with different conditions or histories to make inferences about biotic changes across time.

Researchers often use space-for-time substitutions to investigate the drivers of change over time. While Wogan and Wang (2018) identified multiple scenarios in which the substitution was valid, they acknowledged scenarios where the method was inappropriate, highlighting the need for further work validating its application. Furthermore, direct conversions between space (e.g., distance, environments) and time, which are critical for quantifying the rate of evolutionary response, are system-specific (Bergland et al. 2014) due to differences between taxa like generation time or population size. Temporal genomics is an attractive alternative to space-for-time substitutions that applies timeseries sequences to evolutionary biology using data derived from either sampling at multiple timepoints (Vega et al. 2017; Mei et al. 2020), historical samples (Tracy and Jamieson 2011; Chattopadhyay et al. 2019), or fossils (Mitchell et al. 2016; van der Valk et al. 2021). As we will discuss below, time-series genomic data from either repeated sampling events or preserved specimens combined with contemporary sampling can also be used to disentangle the impact of multiple evolutionary forces (Buffalo and Coop 2019; Buffalo and Coop 2020).

The application of time-series genetic data is not restricted to fundamental evolutionary questions (Waters et al. 2015; Vega et al. 2017; von Seth et al. 2021). Longstanding conservation programs have used temporal analyses of genetic data to inform their conservation programs (Wasko et al. 2004; Fisch et al. 2013) and evaluate the efficacy of actionable interventions (La Haye et al. 2017; Jensen et al. 2018). For example, Kotzé et al. (2019) compared genetic samples of Cape mountain zebra, Equus zebra zebra, from African nature reserves collected between 1999 and 2001 with modern samples collected in 2016 to identify decreases in genetic diversity and population structure that could result in inbreeding depression and a reduction of adaptive evolutionary potential. Similarly, Glover et al. (2012) analyzed 3,049 samples of Atlantic salmon, Salmo salar, collected between 1970 and 2010 from 21 Norwegian populations and demonstrated that historic population structure was preserved despite large quantities of farmraised salmon escaping into the wild, each year. These results indicated that escapees experience low to moderate reproductive success, providing critical information regarding the impact of commercial salmon producers on the genetic variation of wild populations. Wellestablished conservation programs are therefore an excellent source of long-term temporal samples that enable improved conservation efforts.

Unfortunately, the economic and time constraints of scientific research hinder carrying out longitudinal studies in natural populations that are not already the focus of long-term research. However, natural history collections may offer alternative sources of historical samples that can be used strategically to develop temporal genetic and genomic studies (Lopez et al. 2020; Card et al. 2021). Thanks to recent advances in highthroughput sequencing and DNA extraction protocols, it is possible to assemble genomic data from degraded (Bi et al. 2013; Gamba et al. 2016; Tsai et al. 2020) and formalin-fixed tissues through nondestructive sampling (Tin et al. 2014; Hykin et al. 2015; Ruane and Austin 2017; Nakahama 2021), enabling researchers to harness natural history collections while preserving specimens for future research (Ruane 2021). In the current issue, Cohen et al. (2022) use samples from museum specimens and contemporary samples from wild populations to show that Colorado potato beetles, Leptinotarsa decemlineata, underwent a population expansion in the 19th century, and genome scans suggest that agricultural field populations have experienced strong selection on known insecticide resistance genes. von Thaden et al. (2021) compared modern samples collected through a long-term monitoring program and historical museum specimens of European wildcat, Felis silvestris, and demonstrated that the metapopulation structure previously attributed to postglacial differentiation was better explained by a human-induced bottleneck followed by re-expansion from isolated refugia. Overall, natural history collections enable researchers to detect recent evolutionary responses without having to establish long-term studies on existing natural populations.

Natural history collections, advancing methodologies, and lower sequencing costs have propelled the field of temporal genomics. The field is diverse, with both applied (Wasko et al. 2004; Vega et al. 2017; Kotzé et al. 2019) and fundamental (Bergland et al. 2014; Machugh et al. 2017) research focused on short (Kawecki et al. 2012; Lenski et al. 2015; Durland et al. 2021) and long (Brunel et al. 2020; Der Sarkissian et al. 2020) timescales. On longer timescales, ancient DNA (aDNA) research focuses on processes that act over thousands (Machugh et al. 2017; Yang et al. 2020) to millions (van der Valk et al. 2021) of years. Loog et al. (2020) used contemporary and ancient samples of the grey wolf, *Canis lupus*, spanning 50,000 years to trace modern wolf ancestry back to an expansion from Beringia

(the land bridge spanning Asia and North America). On a shorter timescale, Perry et al. (2022), this issue, quantify gene expression in brown bears, Ursus arctos, in different tissue types to investigate differences in gene isoform expression among seasons of the hibernation cycle. By sampling repeatedly within a year, the authors uncovered specific changes in gene expression that relate to hibernation phenotypes, providing insights into the molecular mechanisms that regulate complex animal behaviors. Furthermore, temporal methodologies can use samples across development to better understand changes in gene expression throughout an organism's life. In this issue, Oomen et al. (2022) used experimental tanks of Atlantic cod (Gadus morhua) larvae at different temperatures and sampled them throughout development to examine the temporal dynamics of gene expression under thermal stress. Through this special issue, we aim to highlight the diversity of temporal genomic research while promoting the extension of time-series methodologies to other "omics" approaches (e.g., metabolomics, proteomics, epigenomics). This introduction provides a brief overview of the application of temporal genomics to study microevolution, and its implementation within conservation. We also contextualize the studies within this special issue to inspire future research with time-series molecular data.

Quantifying evolutionary forces

Stochastic evolutionary forces (mutation, gene flow, and genetic drift) and natural selection can be distinguished based on whether deterministic shifts in population allele frequencies occur. While mutation is not random (Martincorena et al. 2012; Sung et al. 2012), it is often considered neutral because mutations do not necessarily have an immediate fitness benefit. Similarly, gene flow is not random either. A host of features, including individual performance, dispersal capacity, and behavior, can impact an individual's dispersal probability (Edelaar and Bolnick 2012). Compared to mutation, and gene flow, genetic drift, the change in allele frequencies across generations driven by chance events independent of phenotype, is random (Wright 1931; Wright 1984). Conversely, natural selection acts on heritable phenotypic variation across generations to drive population trait values toward an environmentally dependent optimum, often resulting in local adaptation. Hence, natural selection drives deterministic shifts that increase the frequency of alleles that confer high fitness in that particular environment. The relative strength of these processes is both species- and populationspecific, driven by a combination of genomic, demographic, and environmental factors. Directly quantifying the strength of these evolutionary forces with sequence data requires time-series genomic data because single-timepoint studies cannot estimate the rate of allele frequency change outside of idealized population genetic models. Temporal genomics enables researchers to directly estimate rates of evolutionary change by comparing allele frequencies across time (Kelson et al. 2020; Buffalo and Coop 2019; van der Valk et al. 2019; 2020). Furthermore, different evolutionary processes can generate similar genomic patterns (Barber et al. 2011; Simões et al. 2016; Wolf and Ellegren 2017) that are indistinguishable without repeated sampling events (Buffalo and Coop 2019; Chen et al. 2019; Buffalo and Coop 2020). For example, areas within the genome with high genetic divergence are often referred to as speciation genes and often assumed to be driven by disruptive selection; however, similar patterns of divergence could be attributed to background selection on sites in linkage disequilibrium (Wolf and Ellegren 2016). Similarly, high rates of gene flow from adjacent populations can increase the frequency of alleles at a particular locus and be misattributed to selection (Chen et al. 2019). Therefore, temporal genomics can provide unprecedented insights into the relative contribution of each evolutionary force while disentangling their simultaneous impacts.

Mutation

Only mutation is capable of generating new genetic variation, and the rate at which it occurs varies significantly across taxa, cell types, life stages, and regions of the genome (Martincorena et al. 2012; Sung et al. 2012). Somatic mutations occur within non-gametic cells, are not inherited, and thus do not contribute to the population genetic variability. Germline mutations occur within gametes and can be inherited, resulting in genetic variability that can persist across generations. Therefore, the direct quantification of mutation rates can be arduous as it requires comprehensive sampling across many generations. Researchers have tried to circumvent time-series sequence data extracted from individuals to infer mutation rates (Eyre-Walker and Keightley 1999; Kumar and Subramanian 2002) utilizing, for example, cell cultures (Milholland et al. 2017) that may not be representative of natural reproduction. However, direct quantification of mutation rates requires time-series genomic data from individuals to accurately measure the introduction and potential fixation of mutations across generations.

Experimental evolution has capitalized on model organisms (e.g., *Drosophila melanogaster*, *Escherichia coli*, *Saccharomyces cerevisiae*) with short generation times (20 min–7 days) to quantify mutation rates by repeatedly sampling populations under controlled laboratory conditions (Haag-Liautard et al. 2008; Wielgoss et al.

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2011; Zhu et al. 2014; Ramiro et al. 2020). The application of temporal genomics to such experimental evolution studies enables researchers to detect mutations and explore their evolutionary consequences (Cutter 2005; Segrè et al. 2006; Khan et al. 2011). Furthermore, experimental evolution with repeated genomic sampling is instrumental in the study of environmental drivers (e.g., temperature) of changes in mutation rate (Waldvogel and Pfenninger 2012) and heterogeneity of mutation rates across the genome (Lang and Murray 2011; Raynes and Sniegowski 2014). One limitation of time-series genomics in studying mutation rates within natural populations is the ability to identify mutations with confidence. Researchers can use experimental evolution to sequence ancestral populations with high resolution and compare them to subsequent generations to detect novel mutations. However, detecting novel mutations in the wild requires intensive genetic sampling of the focal and neighboring populations to ensure that newly detected alleles are truly new mutations rather than undetected standing genetic variation or private alleles from adjacent populations.

One unparalleled example of the application of temporal genomics in the wild is the case of the COVID-19 pandemic. The advancement and accessibility of highthroughput sequencing techniques have been instrumental in the surveillance and management of the pandemic: tracking the spread of the virus and identifying new mutations and variants of concern, which, in turn, has facilitated the development of clinical tests and predicting the efficacy of vaccines against viral variants (Knyazev et al. 2022). Furthermore, the democratization of viral sequencing has offered free access to millions of sequences worldwide (by May 2022, 10,851,095 h-CoV-19 genomes had been submitted to the GSAID database https://www.gisaid.org/). Analysis of these sequences have enabled scientists to witness the temporal dynamics of viral genome evolution in real-time, even within a single host, and how these changes affect the transcriptome, translatome, and immune response (Pathak et al. 2022).

Gene flow

Gene flow can increase genetic variation in a population by introducing novel or previously lost alleles from adjacent populations (Consuegra et al. 2005; Frankham 2015). High levels of gene flow, on the other hand, can decrease genetic variation across populations through the loss of private alleles and genetic homogenization (Baillie et al. 2016). Among other things, the magnitude of gene flow depends on the geographical distance and environmental differences between populations (Hand et al. 2015; Sacks et al. 2016). Observed patterns of gene

flow could be the product of historic or contemporary environmental conditions; however, single-timepoint samples are not able to discern between the two as the modern genetic structure could be driven by events at either timescale. Furthermore, inferences into the abiotic drivers of gene flow are further complicated by the fact that population genetic structure requires time to reach equilibrium after a change in gene flow patterns (Cushman et al. 2006; Anderson and Meikle 2010; Anderson et al. 2010; García et al. 2021); genetic equilibrium is the population genetic structure that is stable at the current pattern of gene flow (e.g., magnitude, directionality). Landguth et al. (2010) simulated different gene flow patterns (e.g., nearest-neighbor dispersal, long-distance dispersal, panmixia) to determine how many generations it takes to detect the genetic signal of a landscape barrier. They showed significant lag time between barrier establishment and any quantifiable genetic effects using F_{ST} , a metric that took >200 generations to reach 50% of its equilibrium value. This could indicate that F_{ST} performs poorly as a metric to measure how recent landscape change affects population connectivity (Landguth et al. 2010). On the contrary, they showed relatively short lag times (1-15 generations) between barrier establishment and the emergence of population genetic structure measured by other metrics like Dps or G'_{ST} . After a barrier disappears, its genetic effects may still be detected after 100 generations (Landguth et al. 2010). Therefore, longitudinal sampling is useful to assess whether populations are at equilibrium and, if gene flow changes between timepoints, whether the change is due to an environmental shift or adjusting to a new equilibrium.

Gene flow estimates based only on single-timepoint samples may fail to detect rapid changes in connectivity patterns, but temporal genetic and genomic studies can identify changes in historical connectivity across a range of timescales (Therkildsen et al. 2013; Holmes 2015; Pascual et al. 2016; Anderson et al. 2020; Mamoozadeh et al. 2020). In this issue, Byerly et al. (2022) use contemporary and historic samples to explore potential changes in historic and contemporary connectivity and diversity between two populations of roseate terns, Sterna dougallii. They show that both populations were historically and contemporarily isolated and have not suffered significant changes in their genetic diversity, suggesting temporal stability and that recent environmental shifts are not responsible for limiting gene flow between populations. Conversely, in the alpine chipmunk, Tamias alpinus, Bi et al. (2013) observed a reduction in gene flow between populations over time, indicating temporal variability in gene flow, potentially induced by climate change (Bi et al. 2013). Without temporal data, it would be challenging to

determine the relative roles of historical events vs. recent landscape or environmental changes on genetic connectivity. Furthermore, some studies incorporate historical and contemporary data from abiotic factors (e.g., paleoclimatic, Landsat, land-use change) into analyses of gene flow, which can better resolve how specific environmental factors affect population genetic structure. For example, Kelson et al. (2020) showed that the permeability of partial barriers to migration in river populations of rainbow trout, Oncorhynchus mykiss, varied temporally as a result of interannual differences in precipitation and high elevation streamflow. Hence, employing time-series genomic data enables researchers to disentangle historical and contemporary drivers of population genetic connectivity and the temporal stability of population structure.

Genetic drift

In contrast to evolutionary forces that generate genetic variation, genetic drift decreases genetic diversity within populations and increases variation between them through stochastic changes in allele frequencies. Given that the strength of genetic drift increases as population size declines (Wright 1931; Wright 1984) and strong drift rapidly erodes genetic diversity (Harper et al. 2006), researchers use patterns of genetic diversity to infer demographic histories and the processes driving population changes. Most current methods rely on the coalescent theory, which posits that the probability that two alleles coalesce or occur in the same individual is inversely proportional to population size and directly proportional to time (in generations). Coalescent models enable researchers to estimate the time to the most recent common ancestor and changes in population size through time (Marchi et al. 2021). However, in many species, genetic diversity is weakly correlated with current population size (Leffler et al. 2012; Prado-Martinez et al. 2013; Díez-del-Molino et al. 2018). Current methods, coalescent-based or otherwise, using contemporary genetic diversity may not always accurately infer changes in population size because multiple demographic histories could drive similar observed patterns thereby limiting the power to quantify environmental drivers of demographic change. Being unable to identify the potential causes of population size declines restricts our ability to predict changes in genetic diversity, and the amount of standing genetic variation is critical to the ability of species to respond to changing selective pressures (Barrett and Schluter 2008; Willi and Hoffmann 2009; Razgour et al. 2019).

Time-series sequence data provides researchers the opportunity to directly compare allele frequencies before and after an environmentally driven demographic shift and quantify the associated change in genetic diversity, or the severity of a bottleneck, thereby giving a proxy for the strength of genetic drift (van der Valk et al. 2019; Sánchez-Barreiro et al. 2021). Temporal genomics along with abiotic data can isolate the environmental factors, either natural or anthropogenic, that drove the demographic change and genetic drift. For example, Hoeck et al. (2010) used historical and recent samples of Galápagos mockingbird species, Mimus spp., across islands of varying sizes to directly quantify the impact of patch size on genetic diversity. They found that smaller islands with smaller populations experienced more substantial genetic drift and sharper decreases in genetic diversity even though diversity across the archipelago remained stable. Time-series genomic data enables researchers to discern between potential demographic histories by calibrating models with multiple data points throughout time. Additionally, the temporal genomic framework enables researchers to develop and test the efficacy of single-timepoint methodologies at inferring the demographic history of a population. In this issue, Reid and Pinsky (2022) combine forward genetic simulations with coalescent simulations to compare the accuracy with which temporal and single-timepoint methods can reconstruct demographic changes. They find that temporal and some single-timepoint methods can accurately estimate demographic history, thereby providing critical methodological considerations for researchers investigating recent changes in population size. Globally, population genetic diversity has decreased by ~6% since the industrial revolution (Leigh et al. 2019). Thus, identifying specific factors driving this loss of genetic variation is critical to mitigating localized extinction and predicting future adaptive potential (Ørsted et al. 2019). In this context, temporal genomics with appropriate environmental data offers more accurate insights into the abiotic drivers of population declines and the relative role of genetic drift in predicting the future of declining populations.

Natural selection

Depending on its type (e.g., positive, diversifying, fluctuating), natural selection can decrease or increase genetic variation in a population. The effects of natural selection can be disrupted by genetic drift as drift drives stochastic changes in allele frequencies regardless of fitness consequences. Because the strength of genetic drift is inversely proportional to population size, drift will be more effective at disrupting natural selection in small populations. In other words, selection can drive more rapid shifts in phenotypic traits in large populations where stochastic mortality (i.e., allele loss) does

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not have as great an impact on population allelic composition (Eyre-Walker 2006; Messer and Petrov 2013). Individuals are continuously exposed to changing environmental conditions that alter the selective landscape. However, detecting selection can be challenging as current methods depend on comparing genetic differentiation at candidate genomic variants against a background of putatively neutral loci, even though allele frequencies at these neutral loci can be impacted by gene flow and genetic drift. While detecting selection is essential, quantifying the strength of selection is critical to estimating the relative impact of environmental drivers on population genomic composition. Assessing the strength of selection has been historically restricted to phenotypic studies because quantifying selection requires repeated sampling to track changes over time (Kingsolver et al. 2001; Ousterhout et al. 2018). Therefore, temporal genomic data can move genomics from identifying bouts of selection to quantifying its strength, similar to temporal phenotypic studies.

Rather than comparing population allele frequencies against a putatively neutral background, temporal genomic data offer greater resolution by quantifying allele frequency changes across time to detect loci under selection (Buffalo and Coop 2019; Buffalo and Coop 2020). Time-series genomic data enable researchers to directly estimate the strength of selection on specific loci by measuring the magnitude of allele frequency change (Bertram 2021; Gompert et al. 2021). In this issue, Pfenninger and Foucault (2022) integrated timeseries environmental, phenotypic, and genomic data across 3 years to identify signals of positive selection in natural populations of harlequin flies, Chironomus riparius. They detected strong positive selection and identified multiple independent allele clusters that segregated temporally and that were strongly correlated with environmental fluctuations. These results suggest that harlequin fly populations can rapidly respond to environmental change through polygenic adaptation. While Pfenninger and Foucault (2022) demonstrate that recurrent abiotic changes can drive temporally heterogeneous selection, rare and dramatic environmental perturbations can also result in intense selection pressures (Lee et al. 2017; Donihue et al. 2018). Using temporal data from wild populations of green anole lizards, Anolis carolinensis, along a north-south transect, Campbell-Staton et al. (2017) demonstrated that an extreme winter weather event drove rapid allele frequency shifts accompanied by changes in gene expression and thermal tolerance. Time-series sequence data further enable researchers to identify the effect of fluctuating selection on population allele frequencies that otherwise would be undetected using singletimepoint data (Bergland et al. 2014; Durland et al. 2021). Therefore, temporal genomics offers a robust framework to detect specific loci under selection, quantify the strength selection, and evaluate the genomic mechanisms responsible for phenotypic change, while investigating the consequences of temporally heterogeneous selection landscapes.

Disentangling interactions

Multiple neutral and selective processes operate simultaneously. For example, the combination of high gene flow with spatially heterogeneous selective pressures can maintain genetic variation in a population (McDonald and Yeaman 2018). The magnitude of gene flow also mitigates the impact of genetic drift, with high migration rates counteracting the effects of drift (Blanquart et al. 2012). Interactions between evolutionary forces can further influence genomic architecture; for example, the balance between selection and gene flow can influence the effect size of loci on trait values and linkage between adaptive alleles (Yeaman and Whitlock 2011). In addition to simultaneous interactions, past processes dictate current evolutionary responses through changes in baseline allele frequencies that shift the distribution of phenotypes upon which contemporary evolutionary forces act. Decreases in genetic diversity due to genetic drift that may limit adaptive responses to future climatic shifts are typical examples of this interconnectedness across time (Bay et al. 2018; Pacific et al. 2015). Similarly, strong selection can fix specific alleles that may become detrimental in the future if the environment changes (Zimova et al. 2016; Tillotson et al. 2019). High levels of gene flow, on the contrary, can increase genetic diversity while attenuating rates of local adaptation and population genetic divergence in the short term (Frankham 2015; Baillie et al. 2016); however, increases in genetic diversity could preserve a population's ability to respond adaptively to the future stressors. Temporal genomics offers a generalizable framework that, rather than comparing loci across the genome or relying on the coalescent model, accounts for previous evolutionary forces by analyzing the difference between past and current allele frequen-

When available, time-series genomic data can disentangle the impact of evolutionary forces on allele frequency change with greater precision than single-timepoint studies. Chen et al. (2019) used an extensive pedigree analysis along with a genomic time series to tease apart the relative roles of gene flow and selection on allele frequencies in wild populations of the Florida scrub Jay, *Aphelocoma coerulescens*. Without temporal data, the large genetic contributions from immigrants would have been misattributed to selection. Similarly,

Gompert et al. (2021) used repeat genomic sampling to reveal that even limited gene flow can increase genetic diversity in largely isolated patches of Lycaeides butterflies. Evolutionary forces also can act upon each other. Mutation rate may be under variable selection as a factor of genome size and the strength of genetic drift (Lynch 2010). Sprouffske et al. (2018) tested how mutation rate may affect adaptive evolution using strains of *E*. *coli* with different mutation rates to evaluate their ability to adapt (i.e., populations grow) to novel chemical environments. While genetic diversity increased in populations consisting of strains with high mutation rates, this diversity only facilitated local adaptation when mutation rates were moderate, suggesting that high mutation rates stifle local adaptation to novel environments. Similarly, Wielgoss et al. (2011) demonstrated that selection can drive an increase in mutability until mutation rates reach a rate at which natural selection can no longer purge detrimental alleles at a rate fast enough to maintain the population. Therefore, temporal genomics can disentangle evolutionary forces and investigate complex interactions between them.

Conservation

The utility of genetics in conservation biology (i.e., conservation genetics) has been long recognized (Hedrick and Miller 1992; Frankham 1995; Hedrick 2001), with studies commonly focused on estimating wildlife management units (Balakrishnan et al. 2003; Rutkowski et al. 2017), inbreeding (Krakowski et al. 2003; Parra-Olea et al. 2012), and genetic connectivity (Meffe and Vrijenhoek 1988; Unfried et al. 2013). Established conservation programs offer long-term data and unique and detailed insights into temporal patterns of population genetic structure and the impact of conservation decisions on managed populations. In the past decade, conservation genomics has complemented conservation genetics and increased our ability to infer the impact of neutral processes (e.g., gene flow, genetic drift) on allele frequencies (Campbell et al. 2015; Steane et al. 2015) while measuring adaptive genetic variation (Vandersteen Tymchuk et al. 2010; Oh et al. 2019). Because of the historical cost of obtaining genomic data, genetic data are commonly used in temporal studies (Lancaster et al. 2006; Draheim et al. 2018). Temporal conservation genetic studies with controlled sampling regimes often use putatively neutral microsatellite markers. Per locus, microsatellites are more informative than binary single-nucleotide polymorphisms (SNPs), but the larger number of analyzed genomic markers allows for more resolution than microsatellites (Zimmerman et al. 2020; Hauser et al. 2021). Nevertheless, similar inferences can be drawn

from microsatellites and SNPs (Lemopoulos et al. 2019; Hauser et al. 2021), and under certain circumstances, microsatellite data may identify phylogenetic clades (Mesak et al. 2014) and reconstruct parentage relationships (Hauser et al. 2021) better than SNPs. Microsatellites can be highly polymorphic, allowing parentage relationships to be assigned with higher accuracy compared to biallelic SNPs. In this issue, Gray et al. (2022) use microsatellite data from the Floreana tortoise, Chelonoidis niger, captive breeding program established in 2011 to evaluate program efficacy and provide recommendations for other captive breeding programs. They demonstrated that breeding success is biased toward females with low levels of inbreeding, and that relatedness between individuals of a breeding pair did not impact breeding success. Further, this study highlights that employing microsatellite data can be more informative to some temporal biological questions than genomic data (Gray et al. 2022).

Conservation plans often depend on precise estimates of demography and connectivity, but as discussed above (see "gene flow and genetic drift" sections) estimates based solely on single timepoints have limitations. Furthermore, applying genomics to conservation programs is challenging (Shafer et al. 2015). For example, genetic diversity is commonly used as a proxy for (effective) population size even though there is no established correlation between measures of genetic diversity and Red list status according to the International Union for Conservation of Nature (IUCN) (Díezdel-Molino et al. 2018). Long-standing conservation programs can capitalize on repeated tissue sampling and use population monitoring to validate temporal methodologies. Indeed, temporal genetic and genomic studies commonly harness samples or data from conservation programs (Chen et al. 2016; Jensen et al. 2018) or charismatic species (Sánchez-Barreiro et al. 2021; Viluma et al. 2022). For instance, Aguillon et al. (2017) explored the genomic consequences of sexspecific limited dispersal in a population of Florida scrub jay, *A. coerulescens*, at Archbold Biological Station in Venus, Florida that had been monitored for decades (Woolfenden and Fitzpatrick 1984; Schoech et al. 1991). Similarly, the grey wolf, Canis lupus, has been the focus of conservation efforts in Scandinavia since the 1980s (Flagstad et al. 2003; Seddon et al. 2005; Bensch et al. 2006; Kardos et al. 2017) and the extensive sampling efforts conducted on this charismatic mammal enabled Viluma et al. (2022) to directly quantify the temporal pattern of genomic erosion due to inbreeding and drift. Sánchez-Barreiro et al. (2021) used samples spanning almost two centuries (1845-2012) of two subspecies of the white rhinoceros, Ceratotherium simum simum and C. simum cottoni, to quantify the genomic conse1778 A. A. Snead and F. Alda

quences of their rapidly declining populations. Importantly, their direct quantification of genetic erosion provides critical inferences for conservation managers attempting to preserve species' genetic diversity (Díezdel-Molino et al. 2018; Sánchez-Barreiro et al. 2021).

Furthermore, time-series genomic data can serve to better evaluate the efficacy of conservation actions by directly quantifying the impact of management decisions on the genetic diversity of endangered species or populations (Chen et al. 2016). As conservation genomics leverages long-term demographic, pedigree, or genetic data from species conservation programs, we would like to stress the importance of open data, collaborative, and citizen science for the advancement of temporal genomics. In this context, long-term investments and efforts to increase communication and collaborations among scientists, scientific collections, and conservation programs would help to alleviate the financial and time constraints of these studies (Card et al. 2021; Nakahama 2021; Ruane 2021; Poo et al. 2022).

Conclusions

Genomic studies based on single time point samples provide critical insights into microevolutionary processes but come with their own limitations (Funk et al. 2016; Martins et al. 2018; Rozhok and Degregori 2019; LaCava et al. 2021). Time-series genomic data, on the other hand, offer greater resolution and certainty about complex interactions between evolutionary forces. Temporal genomics further enables researchers to disentangle the impact of evolutionary forces on population allele frequencies by circumventing assumptions from idealized population genetic models such as genetic equilibrium and neutral genomic backgrounds (Buffalo and Coop 2019; Buffalo and Coop 2020). Hence, temporal genomic can be used to provide inferences with fewer caveats and greater certainty. Already, numerous temporal genomic studies have significantly advanced our understanding of mutation (Raynes and Sniegowski 2014; Ramiro et al. 2020), gene flow (Pascual et al. 2016; Anderson et al. 2020), genetic drift (van der Valk et al. 2019; Sánchez-Barreiro et al. 2021), and natural selection (Bertram 2021; Gompert et al. 2021). Increased affordability of high-throughput sequencing and methodological advances to isolate DNA from museum specimens (Bi et al. 2013; Hykin et al. 2015) are giving more researchers the opportunity to adopt temporal approaches in their studies. Temporal genomics has the potential to rapidly improve our understanding of microevolutionary processes and their temporal dynamics while providing conservation managers with robust methods to guide and evaluate their conservation strategies. Noteworthily, temporal analyses are not restricted to genomic (or transcriptomic) data. In this issue, Snead and Clarke (2022) highlight the importance of integrating multiple types of 'omics approaches (i.e., genomics, transcriptomics, epigenomics, proteomics, and metabolomics) within temporal frameworks to provide a more holistic investigation of evolutionary processes. Therefore, regardless of sequence type, temporal analyses permit researchers to directly quantify molecular changes to investigate fundamental topics within evolutionary biology while opening new avenues and expanding the application of genomics to real-world problems.

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Conflict of interest

The authors declare no conflicts of interest.

Data availability statement

No data or analyses were included within this work.

References

Aguillon SM, Fitzpatrick JW, Bowman R, Schoech SJ, Clark AG, Coop G, Chen N. 2017. Deconstructing isolation-by-distance: the genomic consequences of limited dispersal. PLoS Genet 13:e1006911.

Anderson CD, Epperson BK, Fortin MJ, Holderegger R, James PMA, Rosenberg MS, Scribner KT, Spear S. 2010. Considering spatial and temporal scale in landscape-genetic studies of gene flow. Mol Ecol 19:3565–75.

Anderson CS, Meikle DB. 2010. Genetic estimates of immigration and emigration rates in relation to population density and forest patch area in *Peromyscus leucopus*. Conserv Genet 11:1593–605.

Anderson G, Lal M, Stockwell B, Hampton J, Smith N, Nicol S, Rico C. 2020. No population genetic structure of Skipjack Tuna (*Katsuwonus pelamis*) in the tropical western and central Pacific assessed using Single nucleotide polymorphisms. Front Mar Sci 1102.

Baillie SM, Muir AM, Scribner K, Bentzen P, Krueger CC. 2016. Loss of genetic diversity and reduction of genetic distance among lake trout *Salvelinus namaycush* ecomorphs, Lake Superior 1959 to 2013. J Great Lakes Res 42:204–16.

- Balakrishnan CN, Monfort SL, Gaur A, Singh L, Sorenson MD. 2003. Phylogeography and conservation genetics of Eld's deer (*Cervus eldi*). Mol Ecol 12:1–10.
- Barber BR, Unmack PJ, Pérez-Losada M, Johnson JB, Crandall KA. 2011. Different processes lead to similar patterns: a test of codivergence and the role of sea level and climate changes in shaping a southern temperate freshwater assemblage. BMC Evol Biol 11:1–9.
- Barrett RDH, Schluter D. 2008. Adaptation from standing genetic variation. Trends Ecol Evol 23:38–44.
- Bay RA, Harrigan RJ, Underwood VL, Gibbs HL, Smith TB, Ruegg K. 2018. Genomic signals of selection predict climatedriven population declines in a migratory bird. Science 359:83–86.
- Bensch S, Andrén H, Hansson B, Pedersen HC, Sand H, Sejberg D, Wabakken P, Åkesson M, Liberg O. 2006. Selection for heterozygosity gives hope to a wild population of inbred wolves. PLoS One 1:e72.
- Bergland AO, Behrman EL, O'Brien KR, Schmidt PS, Petrov DA. 2014. Genomic evidence of rapid and stable adaptive oscillations over seasonal time scales in *Drosophila*. PLoS Genet 10:e1004775.
- Bertram J. 2021. Allele frequency divergence reveals ubiquitous influence of positive selection in *Drosophila*. PLoS Genet 17:e1009833.
- Bi K, Linderoth T, Vanderpool D, Good JM, Nielsen R, Moritz C. 2013. Unlocking the vault: next-generation museum population genomics. Mol Ecol 22:6018–32.
- Blanquart F, Gandon S, Nuismer SL. 2012. The effects of migration and drift on local adaptation to a heterogeneous environment. J Evol Biol 25:1351–63.
- Bradburd GS, Ralph PL, Coop GM. 2013. Disentangling the effects of geographic and ecological isolation on genetic differentiation. Evolution 67:3258–73.
- Brauer CJ, Hammer MP, Beheregaray LB. 2016. Riverscape genomics of a threatened fish across a hydroclimatically heterogeneous river basin. Mol Ecol 25:5093–113.
- Brunel S, Andrew Bennett E, Cardin L, Garraud D, Emam HB, Beylier A, Boulestin B, Chenal F, Ciesielski E, Convertini F et al. 2020. Ancient genomes from present-day France unveil 7,000 years of its demographic history. Proc Natl Acad Sci 117:12791–8.
- Buffalo V, Coop G. 2019. The linked selection signature of rapid adaptation in temporal genomic data. Genetics 213: 1007.
- Buffalo V, Coop G. 2020. Estimating the genome-wide contribution of selection to temporal allele frequency change. Proc Natl Acad Sci 117:20672–80.
- Butlin RK. 2010. Population genomics and speciation. Genetica 138:409–18.
- Byerly P, Chesser RT, Fleischer RC, McInerney N, Przelomska NAS, Leberg PL. 2022. Museum genomics provide evidence for persistent genetic differentiation in a threatened seabird species in the Western Atlantic. Interg Comp Biol (doi:10.109 3/icb/icac107).
- Campbell NR, Harmon SA, Narum SR. 2015. Genotyping-in-Thousands by sequencing (GT-seq): a cost effective SNP genotyping method based on custom amplicon sequencing. Mol Ecol Resour 15:855–67.
- Campbell-Staton SC, Cheviron ZA, Rochette N, Catchen J, Losos JB, Edwards S V. 2017. Winter storms drive rapid phenotypic,

- regulatory, and genomic shifts in the green anole lizard. Science 357:495–8.
- Card DC, Shapiro B, Giribet G, Moritz C, Edwards S V. 2021. Museum genomics. Annu Rev Genet 55:633–59.
- Chattopadhyay B, Garg KM, Mendenhall IH, Rheindt FE. 2019. Historic DNA reveals Anthropocene threat to a tropical urban fruit bat. Curr Biol 29:R1299–300.
- Chen N, Cosgrove EJ, Bowman R, Fitzpatrick JW, Clark AG. 2016. Genomic consequences of population decline in the endangered Florida scrub-Jay. Curr Biol 26:2974–9.
- Chen N, Juric I, Cosgrove EJ, Bowman R, Fitzpatrick JW, Schoech SJ, Clark AG, Coop G. 2019. Allele frequency dynamics in a pedigreed natural population. Proc Natl Acad Sci 116:2158–64
- Cohen Z, François O, Schoville S. 2022. Museum genomics of an agricultural super-pest, the Colorado Potato Beetle, *Leptinotarsa decemlineata *;(Chrysomelidae), provides evidence of adaptation from standing variation. Integr Comp Biol (doi:10.1093/icb/icac137).
- Consuegra S, Verspoor E, Knox D, García De Leániz C. 2005. Asymmetric gene flow and the evolutionary maintenance of genetic diversity in small, peripheral Atlantic salmon populations. Conserv Genet 6:823–42.
- Cushman SA, McKelvey KS, Hayden J, Schwartz MK. 2006. Gene flow in complex landscapes: testing multiple hypotheses with causal modeling. Am Nat 168:486–99.
- Cutter AD. 2005. Mutation and the experimental evolution of outcrossing in *Caenorhabditis elegans*. J Evol Biol 18: 27–34.
- Der Sarkissian C, Möller P, Hofman CA, Ilsøe P, Rick TC, Schiøtte T, Sørensen MV, Dalén L, Orlando L. 2020. Unveiling the ecological applications of ancient DNA from mollusk shells. Front Ecol Evol 8:37
- Díez-del-Molino D, Sánchez-Barreiro F, Barnes I, Gilbert MTP, Dalén L. 2018. Quantifying temporal genomic erosion in endangered species. Trends Ecol Evol 33:176–85.
- Donihue CM, Herrel A, Fabre AC, Kamath A, Geneva AJ, Schoener TW, Kolbe JJ, Losos JB. 2018. Hurricane-induced selection on the morphology of an island lizard. Nature 560:88–91.
- Draheim HM, Moore JA, Fortin MJ, Scribner KT. 2018. Beyond the snapshot: landscape genetic analysis of time series data reveal responses of American black bears to landscape change. Evol Appl 11:1219–30.
- Durland E, De Wit P, Langdon C. 2021. Temporally balanced selection during development of larval Pacific oysters (*Crassostrea gigas*) inherently preserves genetic diversity within offspring. Proc Royal Soc B Biol Sci 288: 20203223.
- Edelaar P, Bolnick DI. 2012. Non-random gene flow: an underappreciated force in evolution and ecology. Trends Ecol Evol 27:659–65.
- Eyre-Walker A, Keightley PD. 1999. High genomic deleterious mutation rates in hominids. Nature 397:344–7.
- Eyre-Walker A. 2006. The genomic rate of adaptive evolution. Trends Ecol Evol 21:569–75.
- Fisch KM, Ivy JA, Burton RS, May B. 2013. Evaluating the performance of captive breeding techniques for conservation hatcheries: a case study of the delta smelt captive breeding program. J Hered 104:92–104.
- Fitzpatrick MC, Keller SR. 2015. Ecological genomics meets community-level modelling of biodiversity: mapping the ge-

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nomic landscape of current and future environmental adaptation. Ecology Letters 18:1–16.

- Flagstad Walker CW, Vilà C, Sundqvist AK, Fernholm B, Hufthammer AK, Wiig Koyola I, Ellegren H. 2003. Two centuries of the Scandinavian wolf population: patterns of genetic variability and migration during an era of dramatic decline. Mol Ecol 12:869–80.
- Frankham R. 1995. Conservation genetics. Annu Rev Genet 29:305–27.
- Frankham R. 2015. Genetic rescue of small inbred populations: meta-analysis reveals large and consistent benefits of gene flow. Mol Ecol 24:2610–8.
- Funk WC, Lovich RE, Hohenlohe PA, Hofman CA, Morrison SA, Sillett TS, Ghalambor CK, Maldonado JE, Rick TC, Day MD et al. 2016. Adaptive divergence despite strong genetic drift: genomic analysis of the evolutionary mechanisms causing genetic differentiation in the island fox (*Urocyon littoralis*). Mol Ecol 25:2176–94.
- Gallego-García N, Forero-Medina G, Vargas-Ramírez M, Caballero S, Shaffer HB. 2019. Landscape genomic signatures indicate reduced gene flow and forest-associated adaptive divergence in an endangered neotropical turtle. Mol Ecol. 28: 2757– 71.
- Gamba C, Hanghøj K, Gaunitz C, Alfarhan AH, Alquraishi SA, Al-Rasheid KAS, Bradley DG, Orlando L. 2016. Comparing the performance of three ancient DNA extraction methods for high-throughput sequencing. Mol Ecol Resour 16:459–69.
- García J, Morán-Ordóñez A, García JT, Calero-Riestra M, Alda F, Sanz J, Suárez-Seoane S. 2021. Current landscape attributes and landscape stability in breeding grounds explain genetic differentiation in a long-distance migratory bird. Animal Conservation 24:120–34.
- Glover KA, Quintela M, Wennevik V, Besnier F, Sørvik AGE, Skaala Ø. 2012. Three decades of farmed escapees in the wild: a spatio-temporal analysis of Atlantic salmon population genetic structure throughout Norway. PLoS One 7: e43129.
- Gompert Z, Springer A, Brady M, Chaturvedi S, Lucas LK. 2021. Genomic time-series data show that gene flow maintains high genetic diversity despite substantial genetic drift in a butterfly species. Mol Ecol 30:4991–5008.
- Gray R, Fusco N, Miller J, Tapia W, Mariani C, Caccone A, Jensen E. 2022. Temporal monitoring of the Floreana Island Galapagos giant tortoise captive breeding program. Integr Comp Biol (doi:10.1093/icb/icac129).
- Haag-Liautard C, Coffey N, Houle D, Lynch M, Charlesworth B, Keightley PD. 2008. Direct estimation of the mitochondrial DNA mutation rate in *Drosophila melanogaster*. PLoS Biol 6:e204.
- Haasl RJ, Payseur BA. 2016. Fifteen years of genomewide scans for selection: trends, lessons and unaddressed genetic sources of complication. Mol Ecol 25:5–23.
- Hand BK, Lowe WH, Kovach RP, Muhlfeld CC, Luikart G. 2015. Landscape community genomics: understanding ecoevolutionary processes in complex environments. Trends Ecol Evol 30:161–8.
- Harper GL, Maclean N, Goulson D. 2006. Analysis of museum specimens suggests extreme genetic drift in the adonis blue butterfly (*Polyommatus bellargus*). Biol J Linn Soc 88: 447–52.

Hauser SS, Athrey G, Leberg PL. 2021. Waste not, want not: microsatellites remain an economical and informative technology for conservation genetics. Ecol Evol 11:15800–14.

- Hedrick PW, Miller PS. 1992. Conservation genetics: techniques and fundamentals. Ecol Appl. 2:30–46.
- Hedrick PW. 2001. Conservation genetics: where are we now? Trends Ecol Evol 16:629–36.
- Hoban S, Kelley JL, Lotterhos KE, Antolin MF, Bradburd G, Lowry DB, Poss ML, Reed LK, Storfer A, Whitlock MC. 2016. Finding the genomic basis of local adaptation: pitfalls, practical solutions, and future directions. Am Nat 188: 379–97.
- Hoeck PEA, Bollmer JL, Parker PG, Keller LF. 2010. Differentiation with drift: a spatio-temporal genetic analysis of Galápagos mockingbird populations (*Mimus* spp.). Philos Trans R Soc B Biol Sci 365:1127–38.
- Holmes I. 2015. Temporal population genetic instability in range-edge western toads, *Anaxyrus boreas*. J Hered 106: 45–56
- Hykin SM, Bi K, McGuire JA. 2015. Fixing formalin: a method to recover genomic-scale DNA sequence data from formalin-fixed museum specimens using high-throughput sequencing. PLoS One 10:e0141579.
- Jensen EL, Edwards DL, Garrick RC, Miller JM, Gibbs JP, Cayot LJ, Tapia W, Caccone A, Russello MA. 2018. Population genomics through time provides insights into the consequences of decline and rapid demographic recovery through head-starting in a Galapagos giant tortoise. Evol Appl 11: 1811–21.
- Kardos M, Åkesson M, Fountain T, Flagstad Ø, Liberg O, Olason P, Sand H, Wabakken P, Wikenros C, Ellegren H. 2018. Genomic consequences of intensive inbreeding in an isolated wolf population. Nat Ecol Evol 2:124–31.
- Kawecki TJ, Lenski RE, Ebert D, Hollis B, Olivieri I, Whitlock MC. 2012. Experimental evolution. Trends Ecol Evol 27:547–60
- Kelson SJ, Miller MR, Thompson TQ, SM O'Rourke, Carlson SM. 2020. Temporal dynamics of migration-linked genetic variation are driven by streamflows and riverscape permeability. Mol Ecol 29:870–85.
- Khan AI, Dinh DM, Schneider D, Lenski RE, Cooper TF. 2011. Negative epistasis between beneficial mutations in an evolving bacterial population. Science 332:1193–6.
- Kingsolver JG, Hoekstra HE, Hoekstra JM, Berrigan D, Vignieri SN, Hill CE, Hoang A, Gibert P, Beerli P. 2001. The strength of phenotypic selection in natural populations. Am Nat. 157:245–61
- Knyazev S, Chhugani K, Sarwal V, Ayyala R, Singh H, Karthikeyan S, Deshpande D, Baykal PI, Comarova Z, Lu A, Porozov Y. 2022. Unlocking capacities of genomics for the COVID-19 response and future pandemics. Nat Methods 19:374–80.
- Kotzé A, Smith RM, Moodley Y, Luikart G, Birss C, Van Wyk AM, Grobler JP, Dalton DL. 2019. Lessons for conservation management: monitoring temporal changes in genetic diversity of Cape mountain zebra (*Equus zebra zebra*). PLoS One 14:e0220331.
- Krakowski J, Aitken SN, El-Kassaby YA. 2003. Inbreeding and conservation genetics in whitebark pine. Conserv Genet 4:581–93.

- Kumar S, Subramanian S. 2002. Mutation rates in mammalian genomes. Proc Natl Acad Sci 99:803–8.
- La Haye MJJ, Reiners TE, Raedts R, Verbist V, Koelewijn HP. 2017. Genetic monitoring to evaluate reintroduction attempts of a highly endangered rodent. Conserv Genet 18:877–92.
- LaCava MEF, Gagne RB, Gustafson KD, Oyler-McCance S, Monteith KL, Sawyer H, Kauffman MJ, Thiele DJ, Ernest HB. 2021. Functional connectivity in a continuously distributed, migratory species as revealed by landscape genomics. Ecography 44:987–99.
- Lancaster ML, Gemmell NJ, Negro S, Goldsworthy S, Sunnucks P. 2006. Ménage à trois on Macquarie Island: hybridization among three species of fur seal (*Arctocephalus* spp.) following historical population extinction. Mol Ecol 15:3681–92.
- Landguth EL, Cushman SA, Schwartz MK, McKelvey KS, Murphy M, Luikart G. 2010. Quantifying the lag time to detect barriers in landscape genetics. Mol Ecol 19:4179–91.
- Lang GI, Murray AW. 2011. Mutation rates across budding yeast chromosome VI are correlated with replication timing. Genome Biol Evol 3:799–811.
- Lee CE, Remfert JL, Opgenorth T, Lee KM, Stanford E, Connolly JW, Kim J, Tomke S. 2017. Evolutionary responses to crude oil from the Deepwater Horizon oil spill by the copepod *Eurytemora affinis*. Evol Appl 10:813–28.
- Leffler EM, Bullaughey K, Matute DR, Meyer WK, Ségurel L, Venkat A, Andolfatto P, Przeworski M. 2012. Revisiting an old riddle: what determines genetic diversity levels within species? PLoS Biol 10:e1001388.
- Leigh DM, Hendry AP, Vázquez-Domínguez E, Friesen VL. 2019. Estimated six per cent loss of genetic variation in wild populations since the industrial revolution. Evol Appl 12:1505–12.
- Lemopoulos A, Prokkola JM, Uusi-Heikkilä S, Vasemägi A, Huusko A, Hyvärinen P, Koljonen M-L, Koskiniemi J, Vainikka A. 2019. Comparing RADseq and microsatellites for estimating genetic diversity and relatedness—Implications for brown trout conservation. Ecol Evol 9:2106–20.
- Lenski RE, Wiser MJ, Ribeck N, Blount ZD, Nahum JR, Jeffrey Morris J, Zaman L, Turner CB, Wade BD, Maddamsetti R et al. 2015. Sustained fitness gains and variability in fitness trajectories in the long-term evolution experiment with *Escherichia coli*. Proc R Soc B Biol Sci 282: 20152292.
- Loog L, Thalmann O, Sinding MHS, Schuenemann VJ, Perri A, Germonpré M, Bocherens H, Witt KE, Samaniego Castruita JA, Velasco MS et al. 2020. Ancient DNA suggests modern wolves trace their origin to a Late Pleistocene expansion from Beringia. Mol Ecol 29:1596–610.
- Lopez L, Turner KG, Bellis ES, Lasky JR. 2020. Genomics of natural history collections for understanding evolution in the wild. Mol Ecol Resour 20:1153–60
- Lotterhos KE, Whitlock MC. 2015. The relative power of genome scans to detect local adaptation depends on sampling design and statistical method. Mol Ecol 24:1031–46.
- Lynch M. 2010. Evolution of the mutation rate. Trends Genet 26:345–52.
- Ma L, Ji Y-J, Zhang D-X. 2015. Statistical measures of genetic differentiation of populations: rationales, history and current states. Curr Zool 61:886–97.
- Machugh DE, Larson G, Orlando L. 2017. Taming the past: Ancient DNA and the study of animal domestication. Annu Rev Anim Biosci 5:329–51.

- Mamoozadeh NR, Graves JE, McDowell JR. 2020. Genome-wide SNPs resolve spatiotemporal patterns of connectivity within striped marlin (*Kajikia audax*), a broadly distributed and highly migratory pelagic species. Evol Appl 13:677–98.
- Marchi N, Schlichta F, Excoffier L. 2021. Demographic inference. Curr Biol 31:R276–9.
- Martincorena I, Seshasayee ASN, Luscombe NM. 2012. Evidence of non-random mutations rates suggests an evolutionary rish management strategy. Nature 485:95–8.
- Martins K, Gugger PF, Llanderal-Mendoza J, González-Rodríguez A, Fitz-Gibbon ST, Zhao JL, Rodríguez-Correa H, Oyama K, Sork VL. 2018. Landscape genomics provides evidence of climate-associated genetic variation in Mexican populations of *Quercus rugosa*. Evol Appl 11:1842–58.
- McDonald TK, Yeaman S. 2018. Effect of migration and environmental heterogeneity on the maintenance of quantitative genetic variation: a simulation study. J Evol Biol 31:1386–99.
- Meffe GK, Vrijenhoek RC. 1988. Conservation genetics in the management of desert fishes. Conserv Biol 2:157–69.
- Mei L, Chen M, Shang Y, Tang G, Tao Y, Zeng L, Huang B, Li Z, Zhan S, Wang C. 2020. Population genomics and evolution of a fungal pathogen after releasing exotic strains to control insect pests for 20 years. ISME J 14:1422–34.
- Mesak F, Tatarenkov A, Earley RL, Avise JC. 2014. Hundreds of SNPs vs. dozens of SSRs: which dataset better characterizes natural clonal lineages in a self-fertilizing fish? Front Ecol Evol 2:74
- Messer PW, Petrov DA. 2013. Frequent adaptation and the McDonald-Kreitman test. Proc Natl Acad Sci 110:8615–20.
- Milholland B, Dong X, Zhang L, Hao X, Suh Y, Vijg J. 2017. Differences between germline and somatic mutation rates in humans and mice. Nat Commun 8:1–8.
- Mitchell KJ, Scanferla A, Soibelzon E, Bonini R, Ochoa J, Cooper A. 2016. Ancient DNA from the extinct South American giant glyptodont *Doedicurus sp.* (Xenarthra: *Glyptodontidae*) reveals that glyptodonts evolved from Eocene armadillos Mol Ecol 25:3499–508.
- Nachman MW, Crowell SL. 2000. Estimate of the mutation rate per nucleotide in humans. Genetics 156: 297–304.
- Nakahama N. 2021. Museum specimens: an overlooked and valuable material for conservation genetics. Ecol Res 36: 13–23.
- Oh KP, Aldridge CL, Forbey JS, Dadabay CY, Oyler-Mccance SJ, Baer C. 2019. Conservation Genomics in the Sagebrush Sea: population divergence, demographic history, and local adaptation in sage-grouse (*Centrocercus* spp.). Genome Biol Evol. 11:2023–34.
- Oomen RA, Knutsen H, Olsen EM, Jentoft S, Stenseth NC, Hutchings JA. 2022. Warming accelerates the onset of the molecular stress response and increases mortality of larval Atlantic cod. Interg Comp Biol (doi:10.1093/icb/icac145).
- Ørsted M, Hoffmann AA, Sverrisdóttir E, Nielsen KL, Kristensen TN. 2019. Genomic variation predicts adaptive evolutionary responses better than population bottleneck history. PLos Genet. 15:e1008205.
- Ousterhout BH, Graham SR, Hasik AZ, Serrano M, Siepielski AM. 2018. Past selection impacts the strength of an aquatic trophic cascade. Func Ecol 32:1554–62.
- Pacific M, Foden WB, Visconti P, Watson JEM, Butchart SHM, Kovacs KM, Scheffers BR, Hole DG, Martin TG, Akçakaya HR

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et al. 2015. Assessing species vulnerability to climate change. Nat Clim Change 5:2015–224.

- Parra-Olea G, Zamudio KR, Recuero E, Aguilar-Miguel X, Huacuz D, Zambrano L. 2012. Conservation genetics of threatened Mexican axolotls (*Ambystoma*). Anim Conser 15: 61–72.
- Pascual M, Palero F, García-Merchán VH, Macpherson E, Robainas-Barcia A, Mestres F, Roda T, Abelló P. 2016. Temporal and spatial genetic differentiation in the crab *Liocarci*nus depurator across the Atlantic-Mediterranean transition. Sci Reports 6:1–10.
- Pathak AK, Mishra GP, Uppili B, Walia S, Fatihi S, Abbas T, Banu S, Ghosh A, Kanampalliwar A, Jha A et al. 2022. Spatiotemporal dynamics of intra-host variability in SARS-CoV-2 genomes. Nucleic Acids Res 50:1551–61.
- Perry BW, Armstrong EE, Robbins CT, Jansen HT, Kelley JL. 2022. Temporal analysis of gene expression and isoform switching in brown bears (*Ursus arctos*). Integr Comp Biol (doi: 10.1093/icb/icac093).
- Pfenninger M, Foucault Q. 2022. Population genomic time series data of a natural population suggests adaptive tracking of environmental changes. Integr Comp Biol (doi:10.1093/icb/icac 098).
- Pickett STA. 1989. Long-term studies in ecology: space-for-time substitution as an alternative to long-term studies. New York (NY): Springer. p. 110–35.
- Poo S, Whitfield SM, Shepack A, Watkins-Colwell GJ, Nelson G, Goodwin J, Bogisich A, Brennan PL, D'Agostino J, Koo MS et al. 2022. Bridging the research gap between live collections in zoos and preserved collections in natural history museums. Bioscience 72:449–60.
- Prado-Martinez J, Sudmant PH, Kidd JM, Li H, Kelley JL, Lorente-Galdos B, Veeramah KR, Woerner AE, O'Connor TD, Santpere G et al. 2013. Great ape genetic diversity and population history. Nature 499:471–5.
- Ramiro RS, Durão P, Bank C, Gordo I. 2020. Low mutational load and high mutation rate variation in gut commensal bacteria. PLoS Biol 18:e3000617.
- Raynes Y, Sniegowski PD. 2014. Experimental evolution and the dynamics of genomic mutation rate modifiers. Heredity 113:375–80.
- Razgour O, Forester B, Taggart JB, Bekaert M, Juste J, Ibáñez C, Puechmaille SJ, Novella-Fernandez R, Alberdi A, Manel S. 2019. Considering adaptive genetic variation in climate change vulnerability assessment reduces species range loss projections. Proc Natl Acad Sci 116: 10418–23.
- Reid BN, Pinsky ML. 2022. Simulation-based evaluation of methods, data types, and temporal sampling schemes for detecting recent population declines. Interg Comp Biol (doi:10.1093/icb/icac144).
- Rozhok A, Degregori J. 2019. Somatic maintenance impacts the evolution of mutation rate. BMC Evol Biol 19:1–17.
- Ruane S, Austin CC. 2017. Phylogenomics using formalin-fixed and 100+ year-old intractable natural history specimens. Mol Ecol Resour 17: 1003–8.
- Ruane S. 2021. New data from old specimens. Ichthyol Herpetol 109: 392–6.
- Rutkowski R, Zawadzka D, Suchecka E, Merta D. 2017. Conservation genetics of the capercaillie in Poland—Delineation of conservation units. PLoS One. 12:e0174901.

Sacks BN, Brazeal JL, Lewis JC. 2016. Landscape genetics of the nonnative red fox of California. Ecol Evol 6: 4775–91.

- Sánchez-Barreiro F, Gopalakrishnan S, Ramos-Madrigal J, Westbury M V., de Manuel M, Margaryan A, Ciucani MM, Vieira FG, Patramanis Y, Kalthoff DC et al. 2021. Historical population declines prompted significant genomic erosion in the northern and southern white rhinoceros (*Ceratotherium simum*). Mol Ecol 30:6355–69.
- Schoech SJ, Mumme RL, Moore MC. 1991. Reproductive endocrinology and mechanisms of breeding inhibition in cooperatively breeding Florida scrub jays (*Aphelocoma c. coerulescens*). Condor 93:354–64.
- Seddon JM, Parker HG, Ostrander EA, Ellegren H. 2005. SNPs in ecological and conservation studies: a test in the Scandinavian wolf population. Mol Ecol 14:503–11.
- Segrè A V., Murray AW, Leu JY. 2006. High-resolution mutation mapping reveals parallel experimental evolution in yeast. PLoS Biol 4:e256
- Shafer ABA, Wolf JBW, Alves PC, Bergström L, Bruford MW, Brännström I, Colling G, Dalén L, De Meester L, Ekblom R et al. 2015. Genomics and the challenging translation into conservation practice. Trends Ecol Evol 30:78–87.
- Simões M, Breitkreuz L, Alvarado M, Baca S, Cooper JC, Heins L, Herzog K, Lieberman BS. 2016. The evolving theory of evolutionary radiations. Trends Ecol Evol 31:27–34.
- Snead AA, Clarke RD. 2022. The biological hierarchy, time, and temporal 'omics in evolutionary biology: a perspective. Integr Comp Biol (doi:10.1093/icb/icac138).
- Sprouffske K, Aguilar-Rodríguez J, Sniegowski P, Wagner A. 2018. High mutation rates limit evolutionary adaptation in *Escherichia coli*. PLoS Genet 14:e1007324.
- Steane DA, Potts BM, McLean E, Collins L, Prober SM, Stock WD, Vaillancourt RE, Byrne M. 2015. Genome-wide scans reveal cryptic population structure in a dry-adapted eucalypt. Tree Genet Genomes 11:1–14.
- Strasburg JL, Rieseberg LH. 2010. How robust are "isolation with migration" analyses to violations of the IM model? A simulation study. Mol Biol Evol 27:297–310.
- Sung W, Ackerman MS, Miller SF, Doak TG, Lynch M. 2012. Drift-barrier hypothesis and mutation-rate evolution. Proc Natl Acad Sci 109:18488.
- Terekhanova NV, Logacheva MD, Penin AA, Neretina TV, Barmintseva AE, Bazykin GA, Kondrashov AS, Mugue NS. 2014. Fast evolution from precast bricks: genomics of young freshwater populations of threespine stickleback *Gasterosteus aculeatus*. PLoS Genet 10: e1004696.
- Therkildsen NO, Hemmer-Hansen J, Hedeholm RB, Wisz MS, Pampoulie C, Meldrup D, Bonanomi S, Retzel A, Olsen SM, Nielsen EE. 2013. Spatiotemporal SNP analysis reveals pronounced biocomplexity at the northern range margin of Atlantic cod *Gadus morhua*. Evol Appl 6:690–705.
- Tillotson MD, Barnett HK, Bhuthimethee M, Koehler ME, Quinn TP. 2019. Artifical selection on reproductive timming in hatchery salmon drives a phenological shift and potential maladaptation to climate change. Evol Appl 12: 1344–59.
- Tin MMY, Economo EP, Mikheyev AS. 2014. Sequencing degraded DNA from non-destructively sampled museum specimens for RAD-tagging and low-coverage shotgun phylogenetics. PLoS One 9:e96793.

- Tracy LN, Jamieson IG. 2011. Historic DNA reveals contemporary population structure results from anthropogenic effects, not pre-fragmentation patterns. Conser Genet 12:517–26.
- Tsai WLE, Schedl ME, Maley JM, McCormack JE. 2020. More than skin and bones: comparing extraction methods and alternative sources of DNA from avian museum specimens. Mol Ecol Resour 20:1220–7.
- Unfried TM, Hauser L, Marzluff JM. 2013. Effects of urbanization on song sparrow (*Melospiza melodia*) population connectivity. Conser Genet 14:41–53.
- van der Valk T, Díez-del-Molino D, Marques-Bonet T, Guschanski K, Dalén L. 2019. Historical genomes reveal the genomic consequences of recent population decline in eastern gorillas. Curr Biol 29:165–170.e6.
- van der Valk T, Pečnerová P, Díez-del-Molino D, Bergström A, Oppenheimer J, Hartmann S, Xenikoudakis G, Thomas JA, Dehasque M, Sağlıcan E et al. 2021. Million-year-old DNA sheds light on the genomic history of mammoths. Nature 591:265–9.
- Vandersteen Tymchuk W, O'Reilly P, Bittman J, MacDonald D, Schulte P. 2010. Conservation genomics of Atlantic salmon: variation in gene expression between and within regions of the Bay of Fundy. Mol Ecol 19:1842–59.
- Vega R, Vázquez-Domínguez E, White TA, Valenzuela-Galván D, Searle JB. 2017. Population genomics applications for conservation: the case of the tropical dry forest dweller *Peromyscus melanophrys*. Conserv Genet 18:313–26.
- Viluma A, Flagstad Ø, Åkesson M, Wikenros C, Sand H, Wabakken P, Ellegren H. 2022. Whole-genome resequencing of temporally stratified samples reveals substantial loss of haplotype diversity in the highly inbred Scandinavian wolf population. Genome Res 32:449–58.
- von Seth J, Dussex N, Díez-del-Molino D, van der Valk T, Kutschera VE, Kierczak M, Steiner CC, Liu S, Gilbert MTP, Sinding MHS et al. 2021. Genomic insights into the conservation status of the world's last remaining Sumatran rhinoceros populations. Nat Commun 12:1–11.
- von Thaden A, Cocchiararo B, Mueller SA, Reiners TE, Reinert K, Tuchscherer I, Janke A, Nowak C. 2021. Informing conservation strategies with museum genomics: long-term effects of past anthropogenic persecution on the elusive European wildcat. Ecol Evol 11:17932–51.
- Waldvogel A-M, Pfenninger M. 2021. Temperature dependence of spontaneous mutation rates. Genome Res 31: 1582–9.
- Waples RS. 1998. Separating the wheat from the chaff: patterns of genetic differentiation in high gene flow species. Am Genet Assoc 438:438–50.
- Wasko AP, Martins C, Oliveira C, Senhorini JA, Foresti F. 2004. Genetic monitoring of the Amazonian fish matrinchã

- (*Brycon cephalus*) using RAPD markers: insights into supportive breeding and conservation programmes. J Appl Ichthyol 20:48–52.
- Waters CD, Hard JJ, Brieuc MSO, Fast DE, Warheit KI, Waples RS, Knudsen CM, Bosch WJ, Naish KA, Waters D. 2015. Effectiveness of managed gene flow in reducing genetic divergence associated with captive breeding. Evol Appl 8: 956–71.
- Whitlock MC, McCauley DE. 1999. Indirect measures of gene flow and migration: $F_{ST} \neq 1/(4Nm+1)$. Heredity. 82:117–25.
- Whitney KD, Garland T. 2010. Did genetic drift drive increases in genome complexity? PLoS Genet 6:e1001080.
- Wielgoss S, Schneider D, Barrick JE, Tenaillon O, Cruveiller S, Chane-Woon-Ming B, Médigue C, Lenski RE. 2011. Mutation rate inferred from synonymous substitutions in a long-term evolution experiment with *Escherichia coli*. G3 Genes Genom Genet 1:183–6.
- Willi Y, Hoffmann AA. 2009. Demographic factors and genetic variation influence population persistence under environmental change. J Evol Biol 22:124–33.
- Wogan GOU, Wang IJ. 2018. The value of space-for-time substitution for studying fine-scale microevolutionary processes. Ecography 41:1456–68.
- Wolf JB, Ellegren H. 2017. Making sense of genomic islands of differentiation in light of speciation. Nat Rev Genet 18:87–100.
- Woolfenden GE, Fitzpatrick JW. 1984. The Florida scrub jay: demography of a cooperative-breeding bird. Princeton (NJ): Princeton University Press.p. 11–51.
- Wright S. 1931. Evolution in Mendelian populations. Genetics 16:97.
- Wright S. 1984. Evolution and the genetics of populations, Volume 2: theory of gene frequencies. Vol. 2. Chicago (IL): University of Chicago Press.p. 169–221.
- Yang MA, Fan X, Sun B, Chen C, Lang J, Ko YC, Tsang CH, Chiu H, Wang T, Bao Q et al. 2020. Ancient DNA indicates human population shifts and admixture in northern and southern China. Science 369:282–8.
- Yeaman S, Whitlock MC. 2011. The genetic architecture of adaptation under migration-selection balance. Evolution 65:1897–911.
- Zhu YO, Siegal ML, Hall DW, Petrov DA. 2014. Precise estimates of mutation rate and spectrum in yeast. Proc Natl Acad Sci 111:E2310–8.
- Zimmerman SJ, Aldridge CL, Oyler-McCance SJ. 2020. An empirical comparison of population genetic analyses using microsatellite and SNP data for a species of conservation concern. BMC Genomics 21:1–16.
- Zimova M, Mills LS, Nowak JJ. 2016. High fitness costs of climate change-induced camouflage mismatch. Ecol Lett 19:299–307.