1	<u>Title</u>
2	Urbanization Differentially Affects the Distributions of Two Sympatric Congeners with Similar
3	Ecological Niches
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5	Running Title
6	Comparison of Cardinalis Urban Distributions
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#### **Conflict of Interest**

We have no conflicts of interest.

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#### Abstract

Urbanization has altered organisms and ecosystems around the world and will continue to do so into the foreseeable future. Although avian responses to urbanization at the community level have been well characterized, we lack species-level studies examining differences in spatial distributions in response to urbanization. We tested for differences in spatial distributions of two congeneric passerine bird species across an urban ecosystem: northern cardinals (Cardinalis cardinalis) and pyrrhuloxia (Cardinalis sinuatus), two closely related songbird species that overlap in distribution. We developed and deployed a new method for testing differences in spatial distributions between species using MaxENT, eBird, and structured bird-survey data across two regions: the state of Arizona and across the Tucson metropolitan area. We then investigated potential urban-environmental factors that may similarly or differentially influence spatial distribution of (a) northern cardinal populations in two urban centers in the state (Tucson and Phoenix) and (b) both species in and around Tucson. We found that northern cardinals occur both further north in the state of Arizona and further into the city of Tucson than pyrrhuloxia. We also found that pyrrhuloxia are excluded from urban areas by habitat availability, but that they are positively associated with human development in the areas that they do reside. High-intensity

urbanization has limited urban northern cardinal distributions, and their affinity for areas near open water may have facilitated their northward expansion across the state. Species distributions in response to habitat urbanization differ between even closely related species with very similar niches. Further research into the morphological, physiological, behavioral, and evolutionary differences between pyrrhuloxia and northern cardinals in Arizona may reveal the mechanisms that facilitate urban adaptation, expansion, or avoidance of some species but not of others.

## **Key Words (6-10, arranged alphabetically)**

Avian, comparative, ecology, MaxENT, niche, spatial, urbanization

#### **Main Text**

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### **Introduction**

Human activities have rapidly changed natural landscapes throughout the globe over the last two centuries, and these changes continue to exert extreme pressures on free-ranging organisms (Allan et al. 2017, Elhacham et al. 2020, Gerten et al. 2019, Pörtner et al. 2023, Rosenberg et al. 2019, Seto et al. 2012, Watson et al. 2018). Cities contain novel resource distributions, temperature gradients, and ecological communities alongside unique anthropogenic disturbances that are driven by socioeconomic factors like wealth inequities (Chamberlain et al. 2019, Chen et al. 2021, Jenerette 2011, Kinzig et al. 2005, Schell et al. 2020, Sepp et al. 2017, Seress and Liker 2015). Not all species respond similarly to these changes, with some expanding (Clark 2017), contracting (Muñoz et al. 2021), or shifting their ranges (Arnold et al. 2021, Żmihorski et al. 2020) in response to human development. Urbanization is broadly associated with declines in species richness (Afrifa et al. 2022, Chen et al. 2023, Haight et al. 2023, Hensley et al. 2019, Knapp et al. 2021, Lerman et al. 2021, Sol et al. 2020, Vasquez et al. 2022, Warren et al. 2019), but we lack an understanding of the more fine-scale, species-specific mechanisms underlying these patterns. The majority of comparative work into the effects of urbanization on species distributions is focused largely on comparisons of generalist versus specialist species (Abilhoa et al. 2017, Callaghan et al. 2019 and 2020, Devictor et al. 2007, Luck et al. 2010) or native versus introduced species (Humphrey et al. 2023, Lerman et al. 2020, Mills et al. 1989, Tsang et al. 2019). Comparative genetic, morphological, and behavioral studies have revealed that urbanization can have quite different effects on even closely related species, and those differences can illuminate some of the mechanisms underlying species responses to urbanization (Fusco et al. 2021), e.g. differences in

migration and dispersal rate (Markowski et al. 2021), body size (McNew et al. 2017), and feeding preferences (De León et al. 2018). To our knowledge, no study to date has compared spatial and habitat distributions between species with shared niches and evolutionary histories to test for differences in their responses to urbanization.

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Species-distribution models can be constructed by using occurrence data of the species and spatial data of relevant environmental variables across the region of interest (Sillero et al. 2021). The probable spatial distribution of the species based on the association between occurrence and the environmental variables is called either a spatial distribution model (SDM) or an environmental niche model (ENM; see Peterson & Soberón 2012 for a terminology discussion). Contemporary methods for constructing these models include generalized additive models, maximum entropy models, random forest models, regularized regression models, and others, but MaxENT (a maximum entropy modeling approach, Phillips et al. 2004, 2006) is both widely used and among the top-performing spatial-distribution modeling approaches (Valavi et al. 2022). Although MaxENT has often been used to model species ranges under various human activities such as climate change (Nameer 2020), it has only rarely been applied to urban areas (but see Davis et al. 2012, Ito et al. 2020, Préau et al. 2018, Sallam et al. 2017, and Wiese et al. 2019), and never with a specific focus on differences in species distributions across urban areas. Similarly, MaxENT has not commonly been employed to compare species distributions (Espinosa et al. 2018) and has never been used within a hypothesis-testing framework to identify regions where species differ in their predicted occupancies. The software ENMTools implements tests to determine if two species have identical distributions or if they have more similar distributions than would be expected by chance (Warren et al. 2010 and 2021), but no test exists to determine areas of significant difference between species.

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#### Study system

Northern cardinals (Cardinalis cardinalis) and pyrrhuloxia (Cardinalis sinuatus) are two songbird species (Order Passeriformes: Family Cardinalidae) that are similarly distributed throughout much of the Sonoran Desert in the USA and Mexico and were estimated to have diverged around 6 million years ago (Provost et al. 2018, Barker et al. 2015, Hooper and Price 2017, Jetz et al. 2012, Kaiya Provost pers. comm). The southwestern northern cardinal subspecies (C. c. igneus) is a distinct population that is thought to have diverged from the nominate subspecies ~2.4 million years ago (Smith et al. 2011). However, there is not perfect geographic overlap between the two Cardinalis species. The range of C. c. igneus extends further north than that of C. sinuatus, and while both species are observed around the Tucson metropolitan area (pers. obs.), only northern cardinals are commonly seen around the Phoenix metropolitan area, although both are much more sparsely distributed than the eastern USA population of northern cardinals (Halkin et al. 2021). The northern expansion of the eastern U.S. population of northern cardinals, C. c. cardinalis, is believed to have only occurred after European colonization as a result of human-driven land-use changes (Halkin et al. 2021). The historic range of northern cardinals in the Sonoran Desert region was possibly much more similar to that of pyrrhuloxia, with water and land use changes driving the northern expansion of this subspecies as well. In this study, we combined community-science data (eBird; Fuller 2020, Sullivan et al. 2009) and structured survey data (Tucson Bird Count: Turner 2003; Central Arizona Phoenix Long Term Ecological Research Station Bird Survey, Warren et al. 2023) to compare the

distributions of northern cardinals and pyrrhuloxia across two regions: the state of Arizona and

across the Tucson metropolitan area. We also investigated the extent to which various biotic, bioclimatic, and socioeconomic environmental factors predicted distributions of both species in Tucson, and in northern cardinals between two cities in Arizona (Tucson and Phoenix). We outline the specific hypotheses and predictions tested in this study in Table 1. To our knowledge, this is the first comparative study of spatial distributions of congeneric species across an urban landscape, and our methods can provide a foundation for future investigations into the mechanisms underlying species responses to human disturbances.

## **Table 1: Hypotheses and predictions.**

Test #	Description
H10.	The two species do not differ in their distributions across an urban environment.
H1A.	Given the pattern that generalist species with broader ecological niches tend to
	persist in cities better than their specialist counterparts (Warren et al. 2019), we
	predict that the species that has a broader statewide range (northern cardinals) will
	also have a broader range in the urban environment compared to their congener
	(pyrrhuloxia).

- H20. The same environmental variables will predict the distributions of a given species in the Tucson area and across the state of Arizona.
- H2A. Given general, observed/reported differences in the ranges and densities of these birds in urban versus rural environments in Tucson, the environmental variables that predict the distributions of the two species in an urban environment will be different from those across the broader range.

- H30. The environmental variables that predict the distributions of northern cardinals will not differ between cities (i.e. Phoenix v. Tucson).
- H3A. Given the differences in size, history, growth rates, and human population densities of Phoenix and Tucson, the environmental variables that predict the distributions of northern cardinals will differ between two cities.

#### Methods

#### Species Occurrence Data

To determine presence of both bird species across Arizona, we used observational data from the Tucson Bird Count (TBC; 2001-2020, Turner 2003), the Central Arizona Phoenix Long Term Ecological Research Station Bird Survey (CAP LTER; 2000-2020, Warren et al. 2023) and eBird (2017-2021, Sullivan et al. 2009). We used all years available from TBC, and we filtered eBird data to keep only 5 years of data because eBird has increased in popularity over time, and some of the datasets from earlier years may have been more biased by cultural differences in accessibility of eBird (Grade et al. 2022, Perkins 2020). We filtered all datasets to retain observations only during the breeding season, which we conservatively approximated in both species to be April and May (Halkin et al. 2021, Tweit and Thompson 2020, pers. obs.; statewide northern cardinal N = 9488, pyrrhuloxia N = 4139; Tucson northern cardinal N = 1991, pyrrhuloxia N = 1218; Phoenix northern cardinal N = 1119). For maxent analyses, we also filtered to keep only one observation per raster cell (1 km²) of each species (Johnston et al. 2021;

statewide northern cardinal N = 1338, pyrrhuloxia N = 751; Tucson northern cardinal N = 381, pyrrhuloxia N = 335; Phoenix northern cardinal N = 199).

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#### Environmental Data:

As predictors of species distributions, we used the following environmental variables: elevation (United States Geological Survey National Land Cover Database (USGS NLCD) Digital Elevation Model; USGS 2020), 19 bioclimatic variables (WorldClim database, Table S1.1, Fick and Hijmans 2017), tree cover (2016 USGS NLCD Tree Canopy Cover file, Homer et al. 2020; the 2019 USGS NLCD Tree Canopy Cover file is not yet available at the time of these analyses, as of May 10, 2023), percent developed imperviousness (2019 NLCD Percent Developed Imperviousness (CONUS) file, Dewitz and U.S. Geological Survey, 2021), landcover variables (2019 USGS NLCD, Table S1.2, Dewitz and U.S. Geological Survey, 2021), and Median Household Income by census tract (U.S. Census Bureau 2020). These variables are commonly used in MaxENT models of avian distributions (i.e. Jenkins and Ha 2022). We prepared and analyzed our data in R Statistical Software (v4.1.0; R Core Team 2021) using the packages 'dismo' (Hijmans et al. 2011), 'raster' (Higmans and Van Etten 2012), 'rgdal' (Bivand et al. 2015), 'rgeos' (Bivand et al. 2017), 'ENMTools' (Warren et al. 2010 and 2021), 'fnn' (Beygelzimer et al. 2015), and 'leaflet' (Graul and Graul 2016). We split the NLCD file into separate tiff files, each representing 1 of the 20 variables in the NLCD file, excluding the 4 that exclusively pertain to Alaska, 1 that is irrelevant to the low desert (perennial ice/snow), and all 4 variables relating to urban development (Developed, Open Space; Developed, Low Intensity; Developed, Medium Intensity; and Developed, High Intensity). We excluded the urban-development variables because they are categorical

representations of the percent of impervious surfaces in an area, which would be redundant with and less informative than the NLCD Percent Developed Impervious file. The cells in each of the files generated from the NLCD land-cover file represented either the presence of that variable with a 0 value, or the distance from that cell to the nearest cell containing that variable in meters. These represented distance to open water, barren land, deciduous forest, evergreen forest, mixed forest, shrub/scrub, grassland/herbaceous, pasture/hay, cultivated crops, woody wetlands, and emergent herbaceous wetlands.

We reduced the bioclimatic files to raster files representing their second and third principal components across the Arizona region because the original files were highly correlated with each other (Tables S1.3, S1.4, and S1.5). We excluded the first principal component because it was nearly perfectly correlated with the elevation file, and likely just represented variation in climate due to elevation. None of our final files had a final Pearson's correlation coefficient above 0.8 across the state of Arizona or the Tucson and Phoenix regions (Tables S1.6, S1.7, and S1.8).

Prior to analysis, each tiff file was reprojected to the World Geodetic System 84 (WGS 84) coordinate reference system, cropped to the boundaries of the state of Arizona, and resampled across the lowest resolution file, which were the bioclim variables with a 30-arc-second resolution (~1 km²). We converted each file to an ASCII file, which is required for input into MaxENT. We then cropped these ASCII files for an analysis across the city of Tucson using the extent of a minimum longitude of -111.183682, a max of -110.720903, a minimum latitude of 32.034553, and a maximum latitude of 32.554540, and across the city of Phoenix using the extent of a minimum longitude of -112.584727, a max of -111.425540, a minimum latitude of 33.089419, and a maximum latitude of 33.885028. These were selected by determining the

boundaries of the urban area from the US Census urban-area spatial file (U.S. Census Bureau 2020).

# Analyses:

For the MaxENT analysis of each species across the entire state of Arizona, we randomly selected 10,000 points across the region for use as the background environmental conditions and randomly selected 50% of the observations for training data and used the remaining for model testing (Feng et al. 2017). Our analyses across Tucson and Phoenix used the same methods but only used 2,500 background points. All MaxENT models used the default modeling settings. We replicated these methods using subsets of observational data across the state using only eBird data, across Tucson using only eBird or only TBC data, and across Phoenix using only eBird or only CAP LTER data and we found similar results (data not shown). We only present the model that used the entire available data for observations. This produced our empirical models representing the distributions of northern cardinals and pyrrhuloxia across the state of Arizona and across the city of Tucson, and the distribution of northern cardinals across the city of Phoenix.

We applied three tests of niche similarity, which test the null hypothesis that the two species distributions are randomly sampled from the same distribution of environmental variables and are effectively the same (Graham et al. 2004). We report these as D, I, and a rank correlation test, which test for significant differences in range of the species, and as D env, I env, and rank correlation env, which test for significant differences in the environmental niche of the species (Warren et al. 2008). These three test statistics have been shown to produce similar

results, but we present the results of all three for consistency and comparability across the literature.

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To test where the two species differed in their distributions across the city of Tucson, we removed the species designations associated with each observation in the dataset of raw observations. Then, we randomly assigned each of the observations to one of the two species, keeping the number of total observations for each species equal to the true number of observations of that species. We then ran the MaxENT model using the same parameters as we did for our empirical analysis, and then subtracted the pseudoreplicate model for the spatial distribution of the probability of occurrence of pyrrhuloxia from the pseudoreplicate model for the spatial distribution of the probability of occurrence of northern cardinals. We repeated this process 1,000 times with different random permutations of the data to generate a null distribution of the differences between the distributions of the two species given the number of observations of each species. Finally, we subtracted the empirical model of the distribution of pyrrhuloxia from the empirical model of the distribution of northern cardinals and compared this file to the 1,000 null files. We used a significance level of 0.05, so we considered any cell in the empirical difference file to be significant if it demonstrated an absolute difference between the probabilities of the two species that is greater than 950 of the null models. Every nonsignificant cell was converted to 0, which appears as regions of no difference in the figure. We repeated this process using models across the state of Arizona to identify regions across the state where the two species differ. We visualized the spatial distribution of the significant differences in probability of occurrences of the two species in QGIS (QGIS 2023) and added major interstates to the map for visual reference (U.S. Census Bureau 2021).

We also compared permutation-importance values of the empirical MaxENT models. We used an arbitrary cutoff of a minimum of 5% to determine which variables contributed to the model, and then compared between species and regions to identify factors that differ in determining the distributions of the species.

Finally, to test for differences in the environmental variables associated with the distribution of each species, we extracted values from the raster files of each environmental variable for each observation of either species. We then ran t-tests on each environmental variable to compare for differences between species across the entire state of Arizona and across the Tucson region.

## Results

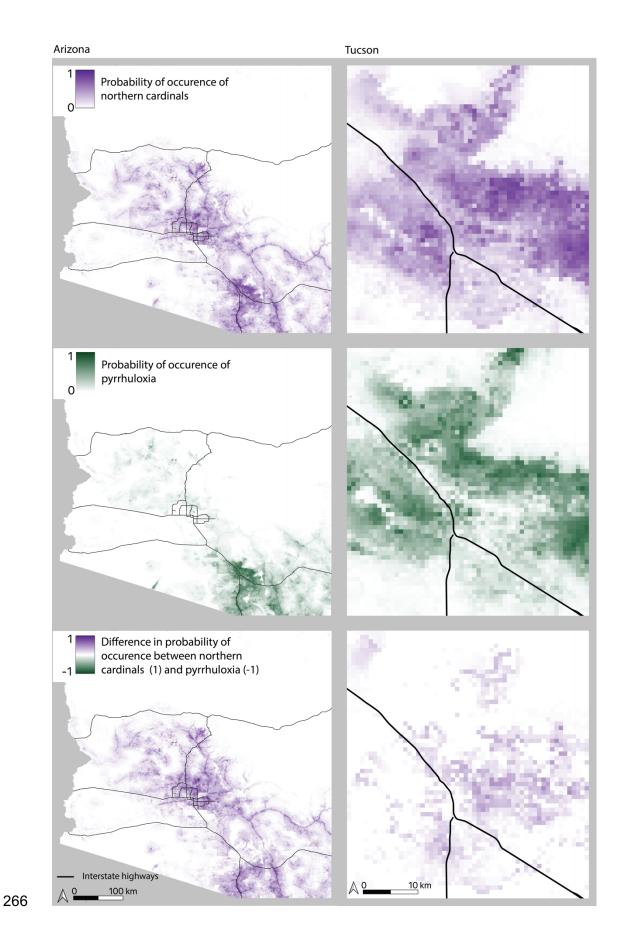
Distributions of pyrrhuloxia and northern cardinals were significantly different across both Tucson and across the state of Arizona in all test statistics except for rank correlation of the Tucson model, which was nearly significant (p-value = 0.05; Table S1.9). Northern cardinals had a higher probability of occurrence in the city of Tucson than pyrrhuloxia and had a higher probability of occurrence than pyrrhuloxia in the regions north of Tucson in the statewide analysis (Fig. 1). However, the environmental niches were not significantly different between the two species in any of the test statistics, although several were nearly significant (Table S1.9).

# Figure 1: Predicted probabilities of occurrence of Cardinalis species across Tucson and across

#### <u>Arizona</u>

MaxENT models of the predicted probability of occurrence of northern cardinals in purple (top) and pyrrhuloxia in green (center), and the regions of significant difference between the two

(bottom). Models across the Arizona region are on the left column and models across the Tucson
 region are on the right column. Interstates and major highways are displayed as a spatial
 reference. Each pixel represents a square kilometer.



## Differences between species across Tucson

Five environmental variables contributed to the models both species' distributions across the city of Tucson (Table 2). These were the second and third principal components of the climate variables (northern cardinals: 19.8%, 18%; pyrrhuloxia: 20.1%, 7.3% respectively), elevation (northern cardinals: 5.1%, pyrrhuloxia: 8.7%), distance to evergreen forests (NLCD 42; northern cardinals: 20.9%, pyrrhuloxia: 16.1%), and distance to cultivated crops (NLCD 82; northern cardinals: 7.7%, pyrrhuloxia: 5.9%). None contributed only to the model of northern cardinals, and four contributed only to the model of pyrrhuloxia. They were surface imperviousness (8.3%), distance to barren land (NLCD 31, 7.2%), distance to deciduous forest (NLCD 41, 5.6%), and distance to grassland/herbaceous (NLCD 71, 5.2%).

# Table 2: Permutation Importance of Environmental Variables in MaxENT Models

Variables representing the permutation-importance scores for northern cardinals and pyrrhuloxia from each empirical MaxENT model are shown. All units are percentages. Any variable with a permutation-importance score above 5% is considered to have significantly contributed to the model, and significant values are bolded.

	Tucson		<u>Arizona</u>		<u>Phoenix</u>
<u>Variable</u>	Permutation	Permutation	Northern Cardinal Permutation importance	Permutation Permutation	Northern Cardinal Permutation importance
ClimPC2	19.8	20.1	1.3	0.3	1.6
ClimPC3	18	7.3	5.9	35.7	2.6
<u>Elev</u>	5.1	8.7	3.8	0.7	0.5
imperviousness	4.3	8.3	27.6	21.8	23.4
MedianHouseholdInco me	2.7	1.2	4	1.6	4.2
canopy	0.6	0	9.4	5.6	1.1
NLCD 11: Open Water	3.2	4.6	10.3	7.3	15.8

NLCD 31: Barren Land	2.8	7.2	2.2	1.5	2.8
NLCD 41: Deciduous Forest	2.4	5.6	2.7	2.3	5.1
NLCD 42: Evergreen Forest	20.9	16.1	2.6	0.8	11.6
NLCD 52: Shrub/Scrub	2.4	2.5	0	0	4.3
NLCD 71: Grassland/Herbaceous	4.8	5.2	0.8	1.1	11.2
NLCD 81: Pasture/Hay	1.7	2.4	1.1	0.6	2.5
NLCD 82: Cultivated Crops	7.7	5.9	22.4	18.9	10.2
NLCD 90: Woody Wetlands	3.4	5	5.9	1.6	3.1

Across Tucson, northern cardinals were more likely to be observed at areas of greater percent impervious surface and greater percent canopy cover compared to pyrrhuloxia (Table 3). Northern cardinals were also closer to open water, barren land, deciduous forest, evergreen forest, and woody wetlands (NLCD 11, 31, 41, 42, 90), and further from shrub/scrub (NLCD 52) compared to pyrrhuloxia (Table 3).

However, in the response curves of the MaxENT model for Tucson (Fig. S1.1), which depict relationship between each variable in the model and probability of presence of the species, the distribution of pyrrhuloxia was positively related to percent surface imperviousness, whereas the relationship between northern cardinal distribution and impervious surface area increased until 80% imperviousness, then steeply declined. Northern cardinals showed an affinity for proximity to open water, while pyrrhuloxia showed no relationship. Northern cardinals showed a negative relationship with barren land, and pyrrhuloxia showed a positive relationship with it. Northern cardinals showed an affinity for deciduous and evergreen forests, while pyrrhuloxia showed a negative relationship with the former but a positive relationship with the latter. Both species showed complex relationships with woody wetlands.

# Table 3: Comparison of Environmental Variable Means for Northern Cardinal vs.

# **Pyrrhuloxia Across Tucson**

Means, standard deviations, and results of the t-test for the difference in means between the species are shown for each variable. Elevation is in meters. All units for NLCD files are in meters. Climate temperatures are in Celsius and precipitation values are in millimeters. A p-value of < 0.05 is considered significant, and rows with significant p-values are bolded.

Environmental variable	Northern cardinal mean	Northern cardinal SD	Pyrrhuloxia mean	Pyrrhuloxia SD	<u>T</u>	<u>DF</u>	<u>P</u>
Canopy	0.69	3.47	0.16	1.01	3.55	727	< 0.01
Elevation	826.54	84.23	831.11	81.04	-0.89	1004	0.37
Imperviousness	22.77	25.81	18.72	23.40	-205.85	529	< 0.01
Median Household Income	\$79665.67	\$21707.97	\$80802.45	\$21663.16	-0.85	987	0.40
NLCD 11: Open Water	2585.76	1664.41	2875.35	1767.62	-2.71	954	0.01
NLCD 31: Barren Land	1197.14	997.49	1337.19	1232.12	-1.99	866	0.05
NLCD 41: Deciduous Forest	8782.49	7450.85	10951.92	8083.57	-4.48	942	< 0.01
NLCD 42: Evergreen Forest	6572.56	5870.97	7660.52	6903.74	-2.71	895	0.01
NLCD 52: Shrub/Scrub	87.33	204.22	44.96	118.38	4.24	991	< 0.01
NLCD 71: Grassland/Herbaceous	2731.93	1590.36	2720.45	1634.67	0.12	972	0.91
NLCD 81: Pasture/Hay	12400.45	5715.35	13066.78	5646.68	-1.89	992	0.06
NLCD 82: Cultivated Crops	7445.95	3992.82	7650.06	3902.69	-0.84	997	0.40
NLCD 90: Woody Wetlands	4358.03	2616.70	4673.02	2538.61	-1.97	1000	0.05
Clim 1: Annual Mean Temperature	20.14	0.55	20.10	0.52	1.32	1017	0.19
Clim 2: Mean Diurnal Range	15.35	0.59	15.30	0.63	1.41	953	0.16
Clim 3: Isothermality	45.61	0.83	45.51	0.91	1.81	937	0.07
Clim 4: Temperature Seasonality	733.62	12.47	732.98	11.78	0.85	1011	0.40
Clim 5: Max Temp of Warmest Month	37.30	0.80	37.24	0.76	1.34	1008	0.18
Clim 6: Min Temp of Coldest Month	3.65	0.42	3.63	0.43	0.64	966	0.52
Clim 7: Temperature Annual	33.65	0.82	33.61	0.84	0.93	974	0.35

Range							
Clim 8: Mean Temp of Wettest Quarter	28.68	0.66	28.62	0.61	1.53	1022	0.13
Clim 9: Mean Temp of Driest Quarter	23.43	0.65	23.39	0.60	1.22	1020	0.22
Clim 10: Mean Temp of Warmest Quarter	29.22	0.68	29.15	0.61	1.72	1027	0.09
Clim 11: Mean Temp of Coldest Quarter	11.50	0.46	11.44	0.43	2.01	1011	0.05
Clim 12: Annual Precipitation	338.88	31.53	340.41	29.86	-0.80	1010	0.42
Clim 13: Precipitation of Wettest Month	60.74	4.71	60.95	4.40	-0.73	1016	0.47
Clim 14: Precipitation of Driest Month	5.59	0.64	5.59	0.61	-0.17	1008	0.87
Clim 15: Precipitation Seasonality	60.66	3.15	60.49	3.23	0.90	973	0.37
Clim 16: Precipitation of Wettest Quarter	154.39	12.75	154.68	12.05	-0.38	1011	0.71
Clim 17: Precipitation of Driest Quarter	21.60	2.21	21.61	2.09	-0.04	1011	0.97
Clim 18: Precipitation of Warmest Quarter	125.57	10.69	125.57	10.17	0.00	1009	1.00
Clim 19: Precipitation of Coldest Quarter	85.65	10.42	86.59	10.09	-1.48	1001	0.14

#### Differences between species across Arizona

In species-distribution models across the state of Arizona, five variables contributed to the models of both species (Table 2): the second principal component of the climate variables (northern cardinals: 5.9%, pyrrhuloxia: 35.7%), elevation (northern cardinals: 27.6%, pyrrhuloxia: 21.8%), median household income (northern cardinals: 9.4%, pyrrhuloxia: 5.6%), NLCD 11 (distance to open water; northern cardinals: 10.3%, pyrrhuloxia: 7.3%), and NLCD 82 (distance to cultivated crops; northern cardinals: 22.4%, pyrrhuloxia: 18.9%). For the model of northern cardinals, NLCD 90 (distance to woody wetlands) also contributed (5.9%). No variables were significant only for the model of pyrrhuloxia.

Across the state of Arizona, northern cardinals were more likely to be observed at sites of significantly greater surface imperviousness and at lower elevation compared to pyrrhuloxia (Table 4). Northern cardinals were also closer to pasture/hay and woody wetlands, but further from deciduous forest, evergreen forest, shrub/scrub, and cultivated crops, compared to pyrrhuloxia (Table 4).

In response curves output by the MaxENT model for Arizona, the two species followed very similar trends for all variables except for the third principal component of the climate variables, for which northern cardinals had a positive relationship and pyrrhuloxia had a complicated but negatively trending relationship (Fig. S1.2).

For climate variables across Arizona (Table 4), compared to pyrrhuloxia, northern cardinals were found at higher Mean Diurnal Range (Clim 2), Temperature Seasonality (Clim 4), Max Temperature of Warmest Month (Clim 5), Temperature Annual Range (Clim 7), Mean Temperature of Warmest Quarter (Clim 10), Precipitation of Driest Quarter (Clim 17), and Precipitation of Coldest Quarter (Clim 19) (Clim 2, 4, 5, 7, 10, 17, 19). Northern cardinals were found at lower Isothermality (Clim 3), Mean Temperature of Wettest Quarter (Clim 8), Precipitation of Wettest Month (Clim 13), Precipitation Seasonality (Clim 15), and Precipitation of Warmest Quarter (Clim 18).

# Table 4: Comparison of Environmental Variable Means for Northern Cardinal vs.

#### Pyrrhuloxia Across Arizona

Means, standard deviations, and results of the t-test for the difference in means between the species are shown for each variable. Elevation is in meters. All units for NLDC files are in

meters. Climate temperatures are in Celsius and precipitation values are in millimeters. A p-value of < 0.05 is considered significant, and rows with significant p-values are bolded.

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Environmental variable	Northern cardinal mean	Northern cardinal SD	Pyrrhuloxia mean	Pyrrhuloxia SD	<u>T</u>	<u>DF</u>	<u>P</u>
Canopy	1.10	4.15	0.79	3.95	1.95	2129	0.05
Elevation	934.16	294.09	1008.02	278.77	-6.69	2135	< 0.01
Imperviousness	14.55	22.91	11.66	20.56	3.48	2239	< 0.01
Median Household Income	\$68947.52	\$27263.93	\$69803.92	\$23140.01	-0.90	2346	0.37
NLCD 11: Open Water	2943.13	2896.10	3071.06	2527.69	-1.24	2293	0.22
NLCD 31: Barren Land	3205.36	3753.89	2961.45	3584.97	1.72	2121	0.09
NLCD 41: Deciduous Forest	18475.40	22291.06	10610.19	13780.84	11.94	2858	< 0.01
NLCD 42: Evergreen Forest	7527.39	9341.36	6468.55	7794.03	3.26	2379	0.01
NLCD 52: Shrub/Scrub	67.72	153.58	39.46	94.87	6.15	2859	< 0.01
NLCD 71: Grassland/Herbaceous	1792.84	1642.24	1836.66	1859.83	-0.63	1830	0.53
NLCD 81: Pasture/Hay	10850.86	8794.86	12422.67	8553.75	-4.69	2087	< 0.01
NLCD 82: Cultivated Crops	8392.41	7861.74	7942.76	6408.66	1.67	2424	0.01
NLCD 90: Woody Wetlands	3231.30	3261.02	4470.37	3716.73	-8.94	1820	< 0.01
Clim 1: Annual Mean Temperature	18.91	1.99	18.82	1.77	1.37	2259	0.17
Clim 2: Mean Diurnal Range	16.40	1.30	16.24	1.37	3.02	1950	< 0.01
Clim 3: Isothermality	46.83	2.65	47.36	2.46	-5.40	2177	< 0.01
Clim 4: Temperature Seasonality	733.63	46.77	714.27	34.30	12.78	2619	< 0.01
Clim 5: Max Temp of Warmest Month	37.04	2.21	36.34	1.86	9.05	2360	< 0.01
Clim 6: Min Temp of Coldest Month	2.03	2.05	2.09	1.93	-0.79	2144	0.43
Clim 7: Temperature Annual Range	35.01	1.87	34.25	1.51	11.92	2438	< 0.01

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21.10

Clim 8: Mean Temp of Wettest Quarter

Clim 9: Mean Temp of

6.01

2.17

26.87

22.03

2.64

1.85

-14.05

-0.40

2881

2334

< 0.01

0.69

Driest Quarter							
Clim 10: Mean Temp of Warmest Quarter	28.03	2.27	27.66	2.07	4.46	2206	< 0.01
Clim 11: Mean Temp of Coldest Quarter	10.28	1.85	10.36	1.56	-1.33	2353	0.19
Clim 12: Annual Precipitation	376.10	76.07	373.24	64.39	1.07	2352	0.29
Clim 13: Precipitation of Wettest Month	69.06	23.27	76.20	21.63	-8.28	2172	< 0.01
Clim 14: Precipitation of Driest Month	5.44	1.47	5.45	0.90	-0.15	2861	0.88
Clim 15: Precipitation Seasonality	61.57	14.34	68.91	11.42	-15.13	2468	< 0.01
Clim 16: Precipitation of Wettest Quarter	169.27	49.80	185.04	44.74	-8.73	2236	< 0.01
Clim 17: Precipitation of Driest Quarter	25.57	8.43	23.43	4.21	9.17	2940	< 0.01
Clim 18: Precipitation of Warmest Quarter	137.06	46.55	153.33	40.70	-9.79	2289	< 0.01
Clim 19: Precipitation of Coldest Quarter	96.64	22.85	88.31	15.50	11.70	2744	< 0.01

# Differences between northern cardinals in Tucson versus Phoenix

In comparisons of the models of northern cardinal distributions between the cities of Tucson and Phoenix (Table 2, Figs. 1, S1.3), two environmental variables – distance to evergreen forests (20.9% and 11.6% respectively) and to cultivated crops (7.7% and 10.2% respectively) – contributed to the models across both cities. Four additional environmental variables contributed to the Phoenix model of northern cardinals but not to the Tucson model: surface imperviousness (23.4%), distance from open water (NLCD 11, 15.8%), from deciduous forest (NLCD 41, 5.1%), and from grassland/herbaceous (NLCD 71, 11.2%). Three variables only contributed to the models in Tucson: climate PC2 (19.8%), climate PC3 (18%), and elevation (5.1%).

In response curves output by the MaxENT model for each city, the distribution of northern cardinals differed in many variables between Tucson and Phoenix, but had similar

response curves for percent surface imperviousness, and distances to barren land, deciduous forest, and evergreen forest (Fig. S1.3).

Thirty of the thirty-two environmental variables associated with the distribution of northern cardinals differed significantly between the cities of Tucson and Phoenix (Table S1.10); the only ones that did not differ were surface imperviousness and distance from shrub/scrub (NLCD 52).

#### **Discussion**

We demonstrated that two closely related cardinal species do not have significantly different environmental niches, but that they differ in their probability of occurrence in both the highly urbanized center of Tucson and across the recently developed northern extent of Arizona. Northern cardinals have a broader range of occurrence than pyrrhuloxia in both regions.

However, we also present evidence that pyrrhuloxia may be excluded from the urban center due to other environmental variables besides urbanization alone. Our findings further demonstrate that urbanization has similar effects on the same species in different cities (northern cardinals in both Tucson and Phoenix) but has different effects on even closely related species in the same city (pyrrhuloxia and northern cardinals in Tucson), and that fine-scale species differences underlie the ability of a species to persist in urban areas. Our work emphasizes the need for species-specific studies to inform urban planning, specifically suggesting that urban residents and planners can reduce the negative impacts of human development on cardinal species by integrating a mixture of native canopy cover, shrub/scrub, and open ecosystems into the urban landscape. We also demonstrate that the impacts of human land-use extend beyond urbanization.

## Are responses to urbanization similar between closely related species?

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We found that effects of urbanization were similar within a species in different cities, but not between closely related species within the same city. Though neither species were present in the highest regions of surface imperviousness, northern cardinals were more likely to be observed in Tucson in areas with more impervious surface than were pyrrhuloxia. Interestingly pyrrhuloxia had a positive relationship with percent surface imperviousness in the MaxENT model response curves from Tucson. Pyrrhuloxia also seem to be choosier in their habitat selection, since more variables are important in explaining the likelihood of presence of pyrrhuloxia in the Tucson area compared to northern cardinals. This suggests that pyrrhuloxia may be excluded from the city of Tucson by other environmental factors besides surface imperviousness, such as natural habitat availability (e.g. shrub/scrub and open habitat, Gould 1960), but that they may have an affinity for human development that occurs within their fundamental niche. Other studies of avian community structure have found differences in urban occupancies of closely related speces (e.g. Davis et al. 2012, Leveau et al. 2017), which suggests that species level differences may underlie urban community structures in many contexts. Our results emphasize that the effects of human development differ between even closely related species, with mid-to-high intensity urbanization having a greater impact on northern cardinals and low-intensity human development and resource management along the suburban outskirts shaping the distribution of pyrrhuloxia, but with both species being largely impacted by habitat availability. Much of our understanding of avian responses to urbanization comes from only a few species (Fidino and Magel 2017), and our findings here demonstrate the need for species specific investigations to uncover the factors underlying species responses to urbanization. Further studies into the morphological, physiological, and genetic mechanisms that allow

northern cardinals to persist in regions of higher urbanization than pyrrhuloxia will provide important clarity into the stressors affecting these species in Tucson.

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Are responses to urbanization similar between the same species (northern cardinals) in different cities?

Despite differences in every other climate and land-use variable aside from canopy cover and distance from shrub/scrub, northern cardinals demonstrated a consistent response to urbanization in both Phoenix and Tucson. The consistency of the relationship between this species with canopy and shrub/scrub variables highlights the importance of these aspects of their environment to their persistence. These findings align with previous fieldwork that found northern cardinals require both open habitat and dense foliage within their nesting territories (Gould 1960). However, comparative studies of urban assemblages between cities in the southwestern U.S. found that overall avian communities differ between urban sites in cities (Hensley et al. 2019). Differences in percent impervious surface between cities may explain this seeming contradiction, since despite maintaining a similar association with urbanization between sites, the model of northern cardinal distribution found more regions that excluded northern cardinals in the urban center of Phoenix than the Tucson model found (Figures 1, S4). Our findings show that this species retains specific habitat needs across climate conditions that could be addressed with land use management plans along mid-to-high intensity urban areas that create interconnected patches of open space with dense foliage throughout the urban landscape. Several studies have found parallel responses in genetic variation to urbanization across multiple cities within the same species (Mueller et al. 2013, Mueller et al. 2020, Salmón et al. 2021, Winchell et al. 2023), and we show that spatial distributions are also consistently affected by urbanization.

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### What drives differences in species responses to urbanization?

We found differences between these species in their associations with land-cover variables in the city, and these differences affected the responses of these species to urbanization. Northern cardinals were found closer to open water and at higher levels of both impervious surface and canopy cover. The distribution of open water in the state of Arizona is heavily engineered, especially in urban areas, with irrigation systems, dams, and man-made lakes creating a novel pattern of water availability for the native wildlife (Colby and Jacobs 2007), and city planning in the desert creates novel matrixes of canopy coverage, with some sparse regions and some very dense regions (Nelson et al. 2021). This suggests that underlying differences in preferences for canopy density may allow northern cardinals to persist in the city, but not pyrrhuloxia, who demonstrate a greater affinity for shrub/scrub than northern cardinals (Gould 1960). Our work here expands upon previous findings in Phoenix that show effects of landscaping on avian communities (Warren et al. 2019), and we demonstrate the role of species differences in shaping urban wildlife communities. Urban planning and landscaping that better integrate native shrub/scrub ecologies into the city may help reduce the exclusionary effects of urbanization on pyrrhuloxia. Our findings echo those from other systems that have found effects of various urban planning strategies on avian community structure (Benitez et al. 2021), but we emphasize the importance of understanding species specific associations with ecological variables within each urban area.

Our findings also illuminate climatic differences in ecological niches that might allow for differences in statewide distributions of these two species. Twelve climate variables differed between the species across the state of Arizona. Northern cardinals were found in areas with

higher fluctuations in temperature at the daily, seasonal, and annual scales and with greater precipitation in the driest and coldest parts of the year. In contrast, pyrrhuloxia were found in areas with more precipitation in the wettest month and warmest quarter, which is the summer monsoon season. Pyrrhuloxia are also found in areas with higher mean temperatures during the wettest quarter, but northern cardinals were found in areas with higher mean temperature of the warmest quarter and maximum temperature of the warmest month. Although climate played an important role in shaping northern cardinal and pyrrhuloxia distributions at both citywide and statewide models, no climate variables differed significantly between observations of the two species across the city of Tucson. The overall combination of climatic factors may therefore be more deterministic for the occupancy of an area by either one of these species and could be highly influenced by the climates of regions they historically occupied. For instance, evidence from several desert species show the effects of Pleistocene environmental processes on contemporary patterns of genetic variation (Provost et al. 2022), and our methods in this study did not integrate historical climate data associated with these species. It is also possible that the two species depend on plants that are sensitive to different climatic phenologies, although no differential ecological specializations have yet been identified. Plant communities in Arizona are rapidly changing in response to climate change (Brusca et al. 2013) and if *Cardinalis* species show differences in associations with particular plants, this could explain differences in species responses both in urban areas and across the state.

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#### Human impacts beyond urbanization

Only one land-use variable influenced the models of both species in all analyses: proximity to cultivated crops. This suggests that human land use beyond just city development

plays a large role in shaping contemporary species distributions. Many studies of the effects of urbanization on avian wildlife compare populations along an urban to rural gradient, but our findings contribute to a growing body of evidence that suggest that rural land use changes may affect species alongside urbanization (e.g. Kumar and Kaur Kler 2021, Lazarina et al. 2020). Much of the research on cropland has focused on its effects on grassland species, which are often declining due to habitat loss (Pool et al. 2014, Scholtz et al. 2017). However, much like urbanization, croplands also facilitates range expansion of some species, with evidence of species differences within families (Veech et al. 2010), and further research into the complex effects of anthropogenic land use changes on avian community dynamics is needed at multiple scales beyond urbanization.

In the model of the entire Arizona region, three additional variables associated with human-influenced land use changes affected the distribution of both species: surface imperviousness, canopy cover, and proximity to open water. Proximity to open water also significantly contributed to the model of northern cardinal distribution across Phoenix, but not across Tucson, which further supports the notion that open water, and therefore the human engineering of it, is increasingly important for northern cardinals along the northern extent of their range in the state of Arizona. This aligns with evidence that riverways have historically been important for the ranges of northern cardinals (Smith et al. 2011). And the distribution of cultivated crops shapes much of the open space in areas along the outskirts of human development and within the habitats of these species. However, these variables also affected pyrrhuloxia distributions across Arizona, and observation locations of the two species only differed significantly from each other in their means of distance to cultivated crops, with pyrrhuloxia found closer to croplands than northern cardinals. This suggests that land-use change

resulting from human activities underlies contemporary distributions of both species, leaving open questions about how and why northern cardinals have experienced a northward range expansion but not pyrrhuloxia. We did not differentiate between types of cropland, and crop diversity affects avian diversity in agricultural regions (Katuwal et al. 2022, Marcacci et al. 2021). Cropland management strategies could be differentially affecting these species. The difference in the response curves of the species to the climate variables, and only to climate variables, across the state of Arizona suggests that differences in the fundamental climate niches of the two species may have permitted northern cardinals but not pyrrhuloxia to expand their range in response to human development.

### **Summary**

In conclusion, we demonstrate a novel method using permutations of species observations and MaxENT models to test for differences in the spatial distributions of species with similar niches. We found significant differences in urban occupancy between two congeneric species that share very similar ecological niches, and we also identify several anthropogenic and ecological variables associated with these distributional differences. The finding that environmental predictors of urban occupancy differ between northern cardinals and pyrrhuloxia in Arizona suggests that species differences in nesting and territory habitats have permitted northern cardinals to better adapt to areas of high human disturbance compared to pyrrhuloxia and highlights fertile ground for future comparative urban research. Our work emphasizes the importance of comparative studies in urban ecology as a method of understanding mechanisms underlying differences in species responses to human disturbances.

518 **Conflict of Interest** 519 The authors declare no conflicts of interest. 520 521 **Data Availability** 522 All data are made available on Dryad at (link to come). The scripts used to run these analyses are 523 available on github at: https://github.com/dannyjackson/Spatial Github 524 525 **Biosketches** 526 Danny Jackson researches the effects of urbanization on the spatial ecology, morphology, 527 physiology, and genetic variation of native species. This work represents a component of their 528 dissertation research at Arizona State University. 529 530 Kevin McGraw studies urban impacts on phenotypic variation – including coloration, disease, 531 and behavior – in birds. 532 533 **Author Contributions** 534 Danny Jackson led the development and implementation of this project and wrote the paper. 535 Kevin McGraw edited the paper and provided valuable feedback and advice throughout the 536 design and execution of the research. 537 538

## 539 References

- Abilhoa, V., & Amorin, R. (2017). Effects of urbanization on the avian community in a southern

  Brazilian city. *Revista Brasileira de Ornitologia*, 25, 31-39.
- 542 Afrifa, J. K., Monney, K. A., & Deikumah, J. P. (2022). Effects of urban land-use types on
- avifauna assemblage in a rapidly developing urban settlement in Ghana. *Urban*
- 544 *Ecosystems*, 1-13.
- 545 Allan, J. R., Venter, O., & Watson, J. E. (2017). Temporally inter-comparable maps of terrestrial
- wilderness and the Last of the Wild. *Scientific data*, 4(1), 1-8.
- Arnold, Z. J., Wenger, S. J., & Hall, R. J. (2021). Not just trash birds: Quantifying avian
- diversity at landfills using community science data. *PLoS One, 16*(9), e0255391.
- Barker, F. K., Burns, K. J., Klicka, J., Lanyon, S. M., & Lovette, I. J. (2015). New insights into
- New World biogeography: An integrated view from the phylogeny of blackbirds,
- cardinals, sparrows, tanagers, warblers, and allies. *The Auk: Ornithological Advances*,
- 552 *132*(2), 333-348.
- Beygelzimer, A., Kakadet, S., Langford, J., Arya, S., Mount, D., Li, S., & Li, M. S. (2015).
- Package 'fnn'.
- Bivand, R., Keitt, T., Rowlingson, B., Pebesma, E., Sumner, M., Hijmans, R., ... & Bivand, M.
- R. (2015). Package 'rgdal'. Bindings for the Geospatial Data Abstraction Library.
- 557 Available online: https://cran. r-project. org/web/packages/rgdal/index. Html.
- Bivand, R., Rundel, C., Pebesma, E., Stuetz, R., Hufthammer, K. O., & Bivand, M. R. (2017).
- Package 'rgeos'. *The comprehensive R archive network (CRAN)*.
- Bufkin, D. (1981). From mud village to modern metropolis: The urbanization of Tucson. The
- *Journal of Arizona History*, *22*(1), 63-98.

- Callaghan, C. T., Benedetti, Y., Wilshire, J. H., & Morelli, F. (2020). Avian trait specialization is
- negatively associated with urban tolerance. *Oikos*, *129*(10), 1541-1551.
- Callaghan, C. T., Major, R. E., Wilshire, J. H., Martin, J. M., Kingsford, R. T., & Cornwell, W.
- K. (2019). Generalists are the most urban-tolerant of birds: a phylogenetically controlled
- analysis of ecological and life history traits using a novel continuous measure of bird
- responses to urbanization. Oikos, 128(6), 845-858.
- Chamberlain, D. E., Henry, D. A., Reynolds, C., Caprio, E., & Amar, A. (2019). The relationship
- between wealth and biodiversity: A test of the Luxury Effect on bird species richness in
- the developing world. *Global change biology*, 25(9), 3045-3055.
- 571 Chen, X., Li, F., Li, X., Hu, Y., & Hu, P. (2021). Quantifying the demographic distribution
- characteristics of ecological space quality to achieve urban agglomeration sustainability.
- *Environmental Research Letters, 16*(9), 094025.
- 574 Chen, Y., Zhao, P., Xu, Q., Qu, B., Li, D., Clement, S., & Li, L. (2023). Relating biodiversity
- with health disparities of human population: An ecological study across the United
- 576 States. *One Health*, 16, 100548.
- 577 Clark, C. J. (2017). eBird records show substantial growth of the Allen's Hummingbird
- (Selasphorus sasin sedentarius) population in urban Southern California. *The Condor:*
- 579 *Ornithological Applications, 119*(1), 122-130.
- Colby, B. G., & Jacobs, K. L. (Eds.). (2007). Arizona water policy: management innovations in
- 581 *an urbanizing, arid region.* Routledge.
- Davis, A. Y., Belaire, J. A., Farfan, M. A., Milz, D., Sweeney, E. R., Loss, S. R., & Minor, E. S.
- 583 (2012). *Green infrastructure and bird diversity across an urban socioeconomic gradient.*
- 584 *Ecosphere*, 3(11), 1-18.

585 De León, L. F., Sharpe, D. M., Gotanda, K. M., Raeymaekers, J. A., Chaves, J. A., Hendry, A. 586 P., & Podos, J. (2019). Urbanization erodes niche segregation in Darwin's 587 finches. Evolutionary Applications, 12(7), 1329-1343. 588 Devictor, V., Julliard, R., Couvet, D., Lee, A., & Jiguet, F. (2007). Functional homogenization 589 effect of urbanization on bird communities. Conservation Biology, 21(3), 741-751. 590 Dewitz, J., and U.S. Geological Survey, 2021, National Land Cover Database (NLCD) 2019 591 Products (ver. 2.0, June 2021): U.S. Geological Survey data release, 592 doi:10.5066/P9KZCM54. 593 Elhacham, E., Ben-Uri, L., Grozovski, J., Bar-On, Y. M., & Milo, R. (2020). Global human-594 made mass exceeds all living biomass. *Nature*, 588(7838), 442-444. 595 Elith, J., Phillips, S. J., Hastie, T., Dudík, M., Chee, Y. E., & Yates, C. J. (2011). A statistical 596 explanation of MaxEnt for ecologists. *Diversity and distributions*, 17(1), 43-57. 597 Espinosa, C. C., Trigo, T. C., Tirelli, F. P., da Silva, L. G., Eizirik, E., Queirolo, D., ... & de 598 Freitas, T. R. (2018). Geographic distribution modeling of the margay (Leopardus wiedii) 599 and jaguarundi (Puma yagouaroundi): a comparative assessment. Journal of Mammalogy, 600 99(1), 252-262. 601 Feng, X., Gebresenbet, F., & Walker, C. (2017). Shifting from closed-source graphical-interface 602 to open-source programming environment: a brief tutorial on running Maxent in R (No. 603 e3346v1). PeerJ Preprints. 604 Fick, S.E. and R.J. Hijmans, 2017. WorldClim 2: new 1km spatial resolution climate surfaces for 605 global land areas. *International Journal of Climatology* 37 (12): 4302-4315.

606 Fuller, L. (2020). Community Science: Why we do it, and why we call it that. eBird Pacific 607 Northwest. https://ebird.org/pnw/news/community-science-why-we-do-it-and-why-we-608 call-it-that. 609 Fusco, N. A., Carlen, E. J., & Munshi-South, J. (2021). Urban landscape genetics: are biologists 610 keeping up with the pace of urbanization?. Current Landscape Ecology Reports, 6, 35-45. 611 Gould, P. J. (1961). Territorial relationships between Cardinals and Pyrrhuloxias. The 612 Condor, 63(3), 246-256. 613 Graham, C. H., Ron, S. R., Santos, J. C., Schneider, C. J., & Moritz, C. (2004). Integrating 614 phylogenetics and environmental niche models to explore speciation mechanisms in 615 dendrobatid frogs. Evolution, 58(8), 1781-1793. 616 Graul, C., & Graul, M. C. (2016). Package 'leafletR'. Interactive Web-Maps Based on the Leaflet 617 JavaScript Library. Available online: https://github.com/chgrl/leafletR. 618 Halkin, S. L., Shustack, D. P., DeVries, M. S., Jawor, J. M., and Linville, S. U. (2021). Northern 619 Cardinal (Cardinalis cardinalis), version 2.0. Birds of the World (P. G. Rodewald and B. 620 K. Keeney, Editors). Cornell Lab of Ornithology, Ithaca, NY, USA. 621 https://doi.org/10.2173/bow.norcar.02 622 Hijmans, R.J, Phillips, S., Leathwick, J. and Elith, J. (2011), Package 'dismo'. Available online 623 at: http://cran.r-project.org/web/packages/dismo/index.html. 624 Hijmans, R. J., & Van Etten, J. (2012). Geographic analysis and modeling with raster data. R 625 package version 2.0-12. 626 Homer, C., Dewitz, J., Jin, S., Xian, G., Costello, C., Danielson, P., ... & Riitters, K. (2020). 627 Conterminous United States land cover change patterns 2001–2016 from the 2016

628	national land cover database. ISPRS Journal of Photogrammetry and Remote Sensing,
629	<i>162</i> , 184-199.
630	Hope, D., Gries, C., Casagrande, D., Redman, C. L., Grimm, N. B., & Martin, C. (2006). Drivers
631	of spatial variation in plant diversity across the Central Arizona-Phoenix
632	ecosystem. Society and Natural Resources, 19(2), 101-116.
633	Humphrey, J. E., Haslem, A., & Bennett, A. F. (2023). Housing or habitat: what drives patterns
634	of avian species richness in urbanized landscapes?. Landscape Ecology, 1-19.
635	Jenerette, G. D., Harlan, S. L., Stefanov, W. L., & Martin, C. A. (2011). Ecosystem services and
636	urban heat riskscape moderation: water, green spaces, and social inequality in Phoenix,
637	USA. Ecological applications, 21(7), 2637-2651.
638	Jenkins, C. L., & Ha, H. (2022). A maximum entropy and GIS approach to predict potential
639	habitat for northern bobwhites in the Black Belt prairie physiographic region of
640	Alabama. Ecological Informatics, 69, 101662.
641	Kumar, S., & Kler, T. K. (2021). Avian diversity at Beas River conservation reserve under
642	urbanization and intensive agriculture in Punjab, India. Biological Diversity: Current
643	Status and Conservation Policies, 1.
644	Gerten, C., Fina, S., & Rusche, K. (2019). The sprawling planet: simplifying the measurement of
645	global urbanization trends. Frontiers in Environmental Science, 7, 140.
646	Grade, A. M., Chan, N. W., Gajbhiye, P., Perkins, D. J., & Warren, P. S. (2022). Evaluating the
647	use of semi-structured crowdsourced data to quantify inequitable access to urban
648	biodiversity: A case study with eBird. PLoS One, 17(11), e0277223.

649 Hensley, C. B., Trisos, C. H., Warren, P. S., MacFarland, J., Blumenshine, S., Reece, J., & Katti, 650 M. (2019). Effects of urbanization on native bird species in three southwestern US Cities. Frontiers in Ecology and Evolution, 7, 71. 651 652 Hooper, D. M., & Price, T. D. (2017). Chromosomal inversion differences correlate with range 653 overlap in passerine birds. *Nature Ecology & Evolution*, 1(10), 1526-1534. 654 Ito, H., Hayakawa, K., Ooba, M., & Fujii, T. (2020). Analysis of habitat area for endangered 655 species using maxent by urbanization in Chiba, Japan. GEOMATE Journal, 18(68), 94-100. 656 657 Jetz, W., Thomas, G. H., Joy, J. B., Hartmann, K., & Mooers, A. O. (2012). The global diversity 658 of birds in space and time. *Nature*, 491(7424), 444-448. 659 Johnston, Alison, et al. "Analytical guidelines to increase the value of community science data: 660 An example using eBird data to estimate species distributions." *Diversity and* Distributions (2021). 661 Knapp, S., Aronson, M. F., Carpenter, E., Herrera-Montes, A., Jung, K., Kotze, D. J., ... & Hahs, 662 663 A. K. (2021). A research agenda for urban biodiversity in the global extinction crisis. 664 BioScience, 71(3), 268-279. 665 Kinzig, A. P., Warren, P., Martin, C., Hope, D., & Katti, M. (2005). The effects of human socioeconomic status and cultural characteristics on urban patterns of biodiversity. 666 667 Ecology and Society, 10(1). 668 Lazarina, M., Tsianou, M. A., Boutsis, G., Andrikou-Charitidou, A., Karadimou, E., & 669 Kallimanis, A. S. (2020). Urbanization and human population favor species richness of

670

alien birds. Diversity, 12(2), 72.

671 Lerman, S. B., Narango, D. L., Andrade, R., Warren, P. S., Grade, A. M., & Straley, K. (2020). 672 Wildlife in the city: human drivers and human consequences. Urban Ecology: Its Nature 673 and Challenges (pp. 37-66). Wallingford UK: CABI. 674 Lerman, S. B., Narango, D. L., Avolio, M. L., Bratt, A. R., Engebretson, J. M., Groffman, P. M., 675 ... & Trammell, T. L. (2021). Residential yard management and landscape cover affect 676 urban bird community diversity across the continental USA. Ecological Applications, 677 31(8), e02455. Luck, G. W., & Smallbone, L. T. (2011). The impact of urbanization on taxonomic and 678 679 functional similarity among bird communities. Journal of Biogeography, 38(5), 894-906. 680 Mabry, J. B., & Thiel, J. H. (1995). A Thousand Years of Irrigation in Tucson. Archaeology in 681 Tucson, 9(4), 1-6. 682 Muñoz, P., García-Rodríguez, A., & Sandoval, L. (2021). Urbanization, habitat extension and 683 spatial pattern, threaten a Costa Rican endemic bird. Revista de Biología Tropical, 69(1), 170-180. 684 685 Markowski, M., Minias, P., Bańbura, M., Glądalski, M., Kaliński, A., Skwarska, J., ... & 686 Bańbura, J. (2021). Genetic structure of urban and non-urban populations differs between 687 two common parid species. Scientific Reports, 11(1), 1-10. 688 Mills, G. S., Dunning Jr, J. B., & Bates, J. M. (1989). Effects of urbanization on breeding bird 689 community structure in southwestern desert habitats. *The Condor*, 91(2), 416-428. 690 Nameer, P. O. (2020). The expanding distribution of the Indian Peafowl (Pavo cristatus) as an 691 indicator of changing climate in Kerala, southern India: A modelling study using 692 MaxEnt. Ecological Indicators, 110, 105930.

693 Nelson, J. R., Grubesic, T. H., Miller, J. A., & Chamberlain, A. W. (2021). The equity of tree 694 distribution in the most ruthlessly hot city in the United States: Phoenix, Arizona. Urban Forestry & Urban Greening, 59, 127016. 695 696 Perkins, D. J. (2020). Blind Spots in Citizen Science Data: Implications of Volunteer Biases in 697 eBird Data. 698 Peterson, A. T., & Soberón, J. (2012). Species distribution modeling and ecological niche 699 modeling: getting the concepts right. Natureza & Conservação, 10(2), 102-107. 700 Phillips, S. J., Dudík, M., & Schapire, R. E. (2004, July). A maximum entropy approach to 701 species distribution modeling. In *Proceedings of the twenty-first international conference* 702 on Machine learning (p. 83). 703 Phillips, S. J., R. P. Anderson, and R. E. Schapire. 2006. Maximum entropy modeling of species 704 geographic distributions. Ecological Modelling 190:231-259. 705 Pörtner, H. O., Scholes, R. J., Arneth, A., Barnes, D. K. A., Burrows, M. T., Diamond, S. E., ... 706 & Val, A. L. (2023). Overcoming the coupled climate and biodiversity crises and their 707 societal impacts. Science, 380(6642), eabl4881. 708 Préau, C., Trochet, A., Bertrand, R., & Isselin-Nondereu, F. (2018). Modeling potential 709 distributions of three European amphibian species comparing ENFA and 710 Maxent. Herpetological Conservation and Biology, 13(1), 91-104. 711 Provost, K. L., Mauck III, W. M., & Smith, B. T. (2018). Genomic divergence in allopatric 712 Northern Cardinals of the North American warm deserts is linked to behavioral 713 differentiation. Ecology and Evolution, 8(24), 12456-12478. 714 R Core Team (2022). R: A language and environment for statistical computing. R Foundation for 715 Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.

- Rosenberg, K. V., Dokter, A. M., Blancher, P. J., Sauer, J. R., Smith, A. C., Smith, P. A., ... &
- 717 Marra, P. P. (2019). Decline of the North American avifauna. Science, 366(6461), 120-
- 718 124.
- 719 QGIS.org, 2023. QGIS Geographic Information System. QGIS Association. http://www.qgis.org
- 720 Sallam, M. F., Michaels, S. R., Riegel, C., Pereira, R. M., Zipperer, W., Lockaby, B. G., &
- Koehler, P. G. (2017). Spatio-temporal distribution of vector-host contact (VHC) ratios
- and ecological niche modeling of the West Nile virus mosquito vector, Culex
- quinquefasciatus, in the City of New Orleans, LA, USA. *International Journal of*
- *Environmental Research and Public Health, 14*(8), 892.
- Schell, C. J., Dyson, K., Fuentes, T. L., Des Roches, S., Harris, N. C., Miller, D. S., ... &
- Lambert, M. R. (2020). The ecological and evolutionary consequences of systemic
- racism in urban environments. *Science*, 369(6510).
- 728 Sepp, T., McGraw, K. J., Kaasik, A., & Giraudeau, M. (2018). A review of urban impacts on
- avian life-history evolution: Does city living lead to slower pace of life?. *Global Change*
- 730 *Biology*, 24(4), 1452-1469.
- 731 Seress, G., & Liker, A. (2015). Habitat urbanization and its effects on birds. Acta Zoologica
- 732 Academiae Scientiarum Hungaricae, 61(4), 373-408.
- 733 Seto, K. C., Güneralp, B., & Hutyra, L. R. (2012). Global forecasts of urban expansion to 2030
- and direct impacts on biodiversity and carbon pools. *Proceedings of the National*
- 735 Academy of Sciences, 109(40), 16083-16088.
- 736 Sillero, N., Arenas-Castro, S., Enriquez-Urzelai, U., Vale, C. G., Sousa-Guedes, D., Martínez-
- Freiría, F., ... & Barbosa, A. M. (2021). Want to model a species niche? A step-by-step
- guideline on correlative ecological niche modelling. *Ecological Modelling*, 456, 109671.

739 Smith, B. T., Escalante, P., Hernández Baños, B. E., Navarro-Sigüenza, A. G., Rohwer, S., & 740 Klicka, J. (2011). The role of historical and contemporary processes on phylogeographic structure and genetic diversity in the Northern Cardinal, Cardinalis cardinalis. BMC 741 742 evolutionary biology, 11(1), 1-12. 743 Sol, D., Trisos, C., Múrria, C., Jeliazkov, A., González-Lagos, C., Pigot, A. L., ... & Pavoine, S. 744 (2020). The worldwide impact of urbanisation on avian functional diversity. *Ecology* 745 Letters, 23(6), 962-972. 746 Sullivan, B. L., Wood, C. L., Iliff, M. J., Bonney, R. E., Fink, D., & Kelling, S. (2009). eBird: A 747 citizen-based bird observation network in the biological sciences. Biological 748 conservation, 142(10), 2282-2292. 749 Tsang, T. P., Dyer, E. E., & Bonebrake, T. C. (2019). Alien species richness is currently 750 unbounded in all but the most urbanized bird communities. *Ecography*, 42(8), 1426-751 1435. 752 Turner, W. R. (2003). Citywide biological monitoring as a tool for ecology and conservation in 753 urban landscapes: the case of the Tucson Bird Count. Landscape and Urban Planning, *65*(3), 149-166. 754 755 Tweit, R. C. and Thompson, C. W. (2020). Pyrrhuloxia (Cardinalis sinuatus), version 1.0. Birds 756 of the World (A. F. Poole and F. B. Gill, Editors). Cornell Lab of Ornithology, Ithaca, NY, USA. https://doi.org/10.2173/bow.pyrrhu.01 757 758 U.S. Census Bureau (2020). Decennial Census Official Publications. Retrieved from 759 [https://www.census.gov/programs-surveys/decennial-census/decade/decennial-760 publications.1900.html].

- 761 U.S. Census Bureau (2020). Decennial Census Official Publications. Retrieved from
- 762 [https://catalog.data.gov/dataset/tiger-line-shapefile-2020-nation-u-s-2010-urban-areas].
- 763 U.S. Census Bureau (2021). Tiger/line shapefile, 2019, Nation, U.S., Primary Roads National
- Shapefile. Catalog. Retrieved from [https://catalog.data.gov/dataset/tiger-line-shapefile-
- 765 2019-nation-u-s-primary-roads-national-shapefile]
- 766 U.S. Geological Survey, 2020, 3D Elevation Program 1-Meter Resolution Digital Elevation
- Model (published 20200606), URL <a href="https://www.usgs.gov/the-national-map-data-">https://www.usgs.gov/the-national-map-data-</a>
- 768 <u>delivery</u>
- Valavi, R., Guillera-Arroita, G., Lahoz-Monfort, J. J., & Elith, J. (2022). Predictive performance
- of presence-only species distribution models: a benchmark study with reproducible code.
- 771 Ecological Monographs, 92(1), e01486.
- Vasquez, A. V., & Wood, E. M. (2022). Urban parks are a refuge for birds in park-poor
- areas. Frontiers in Ecology and Evolution, 10, 1048.
- Warren, D. L., Glor, R. E., & Turelli, M. (2008). Environmental niche equivalency versus
- conservatism: quantitative approaches to niche evolution. *Evolution*, 62(11), 2868-2883.
- Warren, D. L., Glor, R. E., & Turelli, M. (2010). ENMTools: a toolbox for comparative studies
- of environmental niche models. *Ecography*, 33(3), 607-611.
- Warren, D. L., Matzke, N. J., Cardillo, M., Baumgartner, J. B., Beaumont, L. J., Turelli, M., ... &
- Dinnage, R. (2021). ENMTools 1.0: An R package for comparative ecological
- 780 biogeography. *Ecography*, 44(4), 504-511.
- Warren, P.S., Lerman, S.B., Bateman, H., Katti, M., and Shochat, E. (2023) Point-count bird
- 782 censusing: long-term monitoring of bird abundance and diversity in central Arizona-
- Phoenix, ongoing since 2000 ver 22. *Environmental Data Initiative*.

784	https://doi.org/10.6073/pasta/cfb2ea74d8f4f5bddb2061b8e54b8bf3 (Accessed 2023-05-
785	04).
786	Warren, P. S., Lerman, S. B., Andrade, R., Larson, K. L., & Bateman, H. L. (2019). The more
787	things change: species losses detected in Phoenix despite stability in bird-socioeconomic
788	relationships. Ecosphere, 10(3), e02624.
789	Watson, J. E., Venter, O., Lee, J., Jones, K. R., Robinson, J. G., Possingham, H. P., & Allan, J.
790	R. (2018). Protect the last of the wild. Nature, 563, 27-30.
791	Western Regional Climate Center. WRCC. (2006, July 28). Retrieved May 4, 2023, from
792	https://wrcc.dri.edu/Climate/summaries.php
793	Wiese, D., Escalante, A. A., Murphy, H., Henry, K. A., & Gutierrez-Velez, V. H. (2019).
794	Integrating environmental and neighborhood factors in MaxEnt modeling to predict
795	species distributions: A case study of Aedes albopictus in southeastern Pennsylvania.
796	PloS one, 14(10), e0223821.
797	Yang, Z., Dominguez, F., Gupta, H., Zeng, X., & Norman, L. (2017). Potential impacts of the
798	continuing urbanization on regional climate: The developing Phoenix-Tucson "Sun
799	Corridor". In Water Bankruptcy in the Land of Plenty (pp. 179-193). CRC Press.
800	Żmihorski, M., Kowalski, M., Cichocki, J., Rubacha, S., Kotowska, D., Krupiński, D., & Pärt
801	T. (2020). The use of socio-economy in species distribution modelling: Features of rural
802	societies improve predictions of barn owl occurrence. Science of The Total Environment,
803	<i>741</i> , 140407.
804	