- ¹ Phenological responses to climate mediate seedling competition with
- an invasive woodland herb
- D.M. Buonaiuto ^{1,2,3a}, E.M. Wolkovich^{2,3,4}
- 4 Author affiliations:
- ⁵ Department of Environmental Conservation, University of Massachusetts, Amherst, Massachusetts, USA.
- 6 ORCID: 0000-0003-4022-2591
- ⁷ Arnold Arboretum of Harvard University, Boston, Massachusetts, USA.
- ⁸ Department of Organismic and Evolutionary Biology, Harvard University, Cambridge, Massachusetts, USA
- ⁹ Forest & Conservation Sciences, Faculty of Forestry, University of British Columbia, Vancouver, British
- 10 Columbia, Canada

12

^aCorresponding author: 617.823.0687; dbuonaiuto@umass.edu

Abstract

- 1. Invasive plants are often characterized by rapid germination and precocious phenology. Theory suggests that early germination may provide invaders with competitive advantage over slower germinating natives, but the relative contribution of rapid germination vs. other intrinsic competitive traits to the success of invaders is poorly understood. Depending on the relationship between germination and competition, shifts in germination phenology due to climate change may increase the dominance of invaders or buffer communities against their impacts.
- 20 2. We investigated the link between temperature variation, germination phenology and competitive inter21 actions with a sequence of controlled environment experiments. First, we evaluated the relationships
 22 between temperature variation and germination phenology for two North American herbaceous species,
 23 the invasive *Hesperis matronalis* and native *Cryptotaenia canadensis*. We then leveraged temperature24 response differences to manipulate the relative germination phenology of these taxa and quantified the
 25 effects of their phenological differences on competition.
- 3. Seeds of the invasive H. matronalis germinated rapidly, reaching 50% germination in under ten days 26 in all treatment combinations. C. candensis did not reach 50% germination with less than seven weeks 27 of cold stratification. However, with more than 10 weeks of cold stratification and low (20/10°C) 28 incubation temperatures, the germination phenology of C. canadensis was well matched to that of 29 H. matronalis. When grown together, we found that precocious germination phenology doubled the 30 competitive impact of H. matronalis relative to its other intrinsic competitive traits. Phenological 31 advantage of just two-three days relative to C. canadensis was enough to secure competitive dominance 32 at the seedling stage. 33
- 4. Synthesis. This study revealed that the mechanistic link between the germination phenology and competitive success of an invasive plant can be strongly mediated by climate sensitivity differences between introduced and native species. Climate change will likely exacerbate these differences, especially in regions where warming reduces the cold stratification. Our findings suggest that phenological diversity

- in native plant communities is an important property of invasion resistance. The relationship between
 environmental variation, germination dynamics and competition provide a path forward for forecasting climate change impacts on seasonal community assembly, and highlights the need to incorporate
 phenological diversity in restorations.
- Keywords: competition, climate change, germination, invasion, phenology, priority effects, stratification

43 Introduction

- A central tenet of community assembly theory is that the order of arrival of species mediates inter-specific interactions and can dictate the trajectory of community structure (Fukami, 2015). These historical contingencies, known as priority effects, alter the structure and function of communities, driving communities to long-term alternate stable states (Fukami & Nakajima, 2011). Yet in many ecosystems, plant communities must re-assemble each year after a period of dormancy. In these communities, priority effects are the products of phenology, the timing of seasonal life cycle events, rather than the timing of the arrival of propagules, which occurs prior to the dormant season in many cases (Rudolf, 2019; Howe & Smallwood, 1982; Baskin & Baskin, 1988a). Invasive plants are often characterized by rapid germination and precocious phenology under a wide variety of environmental conditions (Gioria et al., 2018; Gioria & Pyšek, 2017; Wolkovich & Cleland, 2011; Smith, 2013). By contrast, native plants tend to exhibit more constrained germination cues (Marushia et al., 2010; Wainwright & Cleland, 2013; Van Clef & Stiles, 2001). In many temperate systems, seeds of native plants are dispersed with deep physiological dormancy, requiring prolonged exposure to specific environmental conditions, such as cold stratification (cool temperatures of $0-10^{\circ}$ C), to break dormancy and stimulate germination 57 (ten Brink et al., 2013; Cavieres & Sierra-Almeida, 2017; Bradford & Nonogaki, 2007). These differences in germination physiology can yield strong differences in the relative germination phenology of invasive and native plants, with invaders germinating well before their native competitors (Gioria & Pyšek, 2017). This difference in relative germination timing among species, which we refer to as **phenological** advantage, can contribute significantly to the competitive abilities, and ultimately invasion success, of invasive plants. By allowing them to begin drawing down seasonal resources and modifying their environment before their native competitors emerge (Kardol et al., 2013), invaders gain a competitive advantage through a seasonal priority effect (Wainwright et al., 2011).
- Despite the growing interest in seasonal priority effects, it has been difficult to quantify their overall contri-

bution to the competitive success of invaders. Germination is notoriously difficult to monitor in the field, and 67 rapid phenology often co-varies with other competitive traits (Dickson et al., 2012; Milbau et al., 2003; Hao et al., 2009). Because of these difficulties, many experiments vary phenological advantage by sowing com-69 peting seeds at different time intervals (Young et al., 2017). While these experiments have provided strong evidence that phenological advantage—on the order of just days to weeks—can yield substantial priority effects (Weidlich et al., 2020), their experimental set-up is difficult to translate into natural communities in which priority effects are mediated by climate, and difficult to use for forecasting. Understanding the role that phenological advantage plays in mediating the dynamics of interspecific competition is critical for predicting and managing the structure and function of plant communities in the face of anthropogenic climate change. Due to interspecific differences in responses to environmental variation, sustained alterations to environmental 76 conditions are already shifting community-wide patterns of germination (Walck et al., 2011). If patterns of germination are indeed tightly linked to the competitive dynamics of communities, then phenological re-78 organization is likely to shift the strength of species' interactions, change patterns of invasion, and strongly influence biological filtering of plant communities. In this study, we generated contrasting levels of phenological advantage among two woodland herbaceous 81 species (the North American invasive Hesperis matronalis and native Cryptotaenia canadensis) by leveraging 82 their differences in germination timing in response to environmental variation. First, we performed a series of 83 germination assays in controlled environments under varying temperature regimes to estimate a realistic range of climate-driven variation in phenological advantage. We then used competition trials under contrasting environmental conditions to indirectly manipulate the phenological advantage between these two taxa and quantify the contribution of seasonal priority effects to their competitive dynamics. By linking climate variation, phenological advantage, and seasonal priority effects, our study has important implications for how anthropogenic climate change will alter phenological assembly and, in turn, plant community interactions in

the decades to come.

Materials and Methods

Focal species

For this study, we focused on a pair of woodland herbaceous species. Dames Rocket (Hesperis matronalis) is a herbaceous biennial/perennial species in the Brasicaeceae family, originally from Eurasia, and introduced to North America in the 19th century (Francis et al., 2009). It can rapidly invade meadows, forest edges and woodlands, forming thick stands and excluding native vegetation (Francis et al., 2009). It is currently listed as a noxious or invasive weed is several states and provinces in the United States and Canada (Susko & Hussein, 2008). Honewort (Cryptotania canadensis) is a herbaceous perennial in the Apiaceae family, native to forests and woodlands of eastern North America (Hawkins et al., 2007). The habitat overlap of these two species suggests that they may compete in nature. While their habitat requirements may be similar, the two species display a substantially different germination niche, making them a suitable model for our study. C. canadensis seeds are classified with deep physiological dormancy and require a substantial period of cold moist 102 stratification to release dormancy and initiate germination (Baskin & Baskin, 1988b). While some reports 103 suggests that cold stratification enhances germination in H. matronalis at low incubation temperatures, 104 several studies have demonstrated that fresh and after-ripened (dry-stored) seeds of H. matronalis are capable 105 of rapid and complete germination at a wide range of temperatures (Susko & Hussein, 2008).

107 Experiment I: Germination Assays

To investigate the relationship between environmental variation and relative germination timing among species, we obtained seeds of *C. canadensis* from Prairie Moon Nursery (Winona, MN) and seeds of *H. matronalis* from American Meadows (Shelburne, VT). We performed germination assays in the growth facilities of the Arnold Arboretum in Boston, Massachusetts, USA (42.3074° N, 71.1208° W). We assigned seeds to a fully-crossed set of twenty experimental treatments; 10 levels of cold stratification duration (0.2,4,5,6,7,8,9,11,13 weeks at 4°C) and two levels of incubation temperature (warm—25°C:15°C (day/night),

 $114 \quad cool - 20^{\circ}C:10^{\circ}C \quad (day/night)).$

Prior to applying experimental treatments we performed a "float test" in which all seeds were placed in distilled water, and unfilled seeds (floating) were removed from the experiment (Baskin & Baskin, 2014). We imbibed the remaining seeds in distilled water for 24 hours and then placed 20 seeds for each species/ 117 treatment combination in petri dishes on moist pool-filter sand, with three replicates per treatment. For the 118 cold stratification treatments, we wrapped petri dishes in aluminum foil to prevent light exposure and placed 119 them in a growth chamber at 4°C. After each stratification interval, we transferred the petri dishes to their 120 assigned incubation chamber for 25 days, moistening the germination substrate as necessary to maintain 121 maximum saturation of the medium without flooding the seeds. We checked for new germinates every two 122 days, defining a seed as germinated when its radical or cotyledon tissue was visible (Baskin & Baskin, 2014). 123 We assessed the viability of any seeds that did not germinate in the 25 day incubation period by performing 124 a "crush test" in which we applied pressure to the intact seed to evaluate its condition (Baskin & Baskin, 125 2014). We excluded any seeds deemed unviable from all subsequent analyses. Due to the staggering of our 126 stratification treatments the experiment took place between 27 August - 12 December 2018. 127

128 Experiment II: Competition Trials

To quantify the contribution of seasonal priority effects to interspecific competition dynamics we performed competition trials under controlled conditions in a research greenhouse at the Arnold Arboretum from October 2020 - February 2021. We planted seeds of *C. canadensis* and *H. matronalis* into 8.9 cm square pots, employing a response surface design where we varied both the overall density of seeds and proportion of each species in each pot (Inouye, 2001). High and low density treatments consisted of 14 and 8 seeds respectively. Proportion treatments were 100:0%, 25:75%, 50:50%, 75:25%, 0:100% (*C. canadensis:H. matronalis*). Each density by proportion treatment was replicated six times.

We randomly assigned half of the pots to low (45 days) and high (72 days) cold stratification treatments in dark growth chambers at 4°C. We staggered the start of the treatments, so that at the conclusion of the

cold stratification, all pots were transferred to a heated greenhouse maintained at 15-25°C with 14 hours of supplemental light. Germination was observed daily from 24 December 2020 - 13 January 2021 and every two days from 15 January 2021 - 01 February 2021. The locations of each pot in the greenhouse were randomly reassigned every three days to minimize any blocking effects on germination or growth.

After 35 days we added 1 tsp per gallon of water of Peters 20-10-20 liquid feed fertilizer to all pots. After 62 days, we harvested the above-ground biomass from all pots, dried it for 48 hours at 60°C, and recorded the dry weight of each species/pot.

145 Statistical analysis

146 Germination Assays

To assess interspecific differences in the relationship between germination rate and temperature variability,
we fit a Bayesian mixed-effect accelerated failure time model (AFT, Onofri et al., 2010) using a Weibull
distribution for the likelihood function. We included weeks of stratification and incubation temperature and
their interaction with species as fixed effects. The model written below is modified from Onofri et al. (2010).

 $t_{50} = t_0 \phi$

where t_{50} is time to 50% germination and corresponds to the germination times of a references seed lot (t_0) multiplied by and "acceleration factor" (ϕ) . The acceleration factor is a product of the experimental treatments of our study through the equation:

$$\phi = exp(\beta_{sp}X_{sp} + \beta_{strat}X_{strat} + \beta_{inc}X_{inc} + \beta_{strat}X_{species}X_{strat} * X_{sp} + \beta_{inc}X_{species}X_{inc} * X_{sp})$$

where X_{sp} , X_{strat} and X_{inc} are the species, stratification and incubation treatment levels in our experiment, and β_{sp} , β_{strat} , β_{inc} , $\beta_{stratXspecies}$ and $\beta_{incXspecies}$ are the estimated effects on ϕ for adding an additional week or stratification or degree of incubation for each species respectively.

The AFT modeling framework let us robustly compare germination timing (t50, time to 50% germination)

even among treatments with different final germination percentages by accounting for viable seeds that did not germinate during our incubation window (Soltani *et al.*, 2015; Onofri *et al.*, 2010). One drawback of this approach is that this class of models assumes that all viable seeds will eventually germinate, which we would not expect in nature. For this reason, we considered any estimated t50 values greater than 40 days to indicate that seeds would not reach 50% germination under those conditions.

55 Competition trials

We quantified phenological advantage between species by subtracting the mean germination time of H.

matronalis from that of C. canadensis in each pot. This allowed us to evaluate the effect phenological

advantage with a regression design (Cottingham et al., 2005), with advantage values ranging from -1.3 to 9.5

(C. candensis mean germination time 1.3 days earlier to 9.5 days later than H. matronalis).

For each pot, we calculated the relative growth rate difference (RGRD) among species using the equation below, modified from Connolly & Wayne (2005).

RGRD =
$$ln(\frac{Y_{Cc}}{y_{Cc}}) - ln(\frac{Y_{Hm}}{y_{Hm}})$$

172

174

180

182

where Y_{Hm} and Y_{Cc} are the final biomass of the species at the end of the experiment and y_{Hm} and y_{Cc} are the initial biomass of the seeds planted at the outset of the experiment. For this calculation we obtained estimates of seed mass for our focal species from the Kew Gardens Seed Information Database (Kew, 2022).

We then modeled the effect of seedling density of C. canadensis, H. matronalis and phenological advantage using on RGRD. The model is written below:

181
$$RGRD_i = N(\alpha + \beta_{Hm}n_{Hm} + \beta_{Cc}n_{Cc} + \beta_{pri}MGT, \sigma^2_{RGRD})$$

where β_{Hm} and β_{Cc} are known as the species influence parameters—representing the intrinsic competitive 183 ability of each species—or the estimated effect of increasing the seedling density of each species by one 184 individual on the RGRD (Connolly & Wayne, 2005). n_{Hm} and n_{Cc} are the number of germinated individuals 185 of H. matronalis and C. canadensis respectively. The relationship between the species influence parameters 186 and n_{Hm} and n_{Cc} indicate that these parameter estimates are dependent on both the number of seeds planted 187 (i.e. the seed bank in each pot) and the fraction of those seeds that germinated. This formulation allows 188 our model to partition the effects of interspecific differences in germination phenology from the effects of 189 interspecific differences in germination fraction under each environmental treatment. β_{pri} is the priority 190 effect, or the effect of increasing the difference in mean germination time between H.matronalis and C. 191 canadensis by one day. In this formulation, α is an un-interpretable intercept (Connolly & Wayne, 2005). 192

193 Model Implementation

We fit all models using the R package "brms" (Bürkner, 2018). We ran the model on four chains with 4000 iterations and a 3000 iteration warm-up for a total of 4000 posterior draws for each parameter, using weakly informative priors. We validated model performance by obtaining \hat{R} values between 1 and 1.01, high effective sample sizes and no divergent transitions. For all models we report the mean posterior estimate along with 90% uncertainty intervals (I₉₀).

$_{94}$ Results

55 Germination advantage

H. matronalis reached 50% germination in under ten days for all environmental treatments, always exceeding 75% germination regardless of environmental conditions (Fig. 1, 2, Tab. S1). Increasing cold stratification duration and incubation temperature only marginally enhanced the germination rate of this species (Fig. 2). By contrast, increasing incubation temperature had a delaying effect on the germination rate of C.

canadensis, suggesting that the mean 20°C temperature of our warm incubation treatment is supra-optimal for the species (Fig. 2). Without sufficient cold stratification (<7 weeks for low incubation and <10 weeks for high incubation temperatures), seeds of *C. canadensis* did not reach 50% germination during the duration of our experiment (Fig. 1, 2, Tab. S1). However, under high levels of cold stratification germination rates of *C. canadensis* began to converge on those of *H. matronialis*, and at levels of stratification >10 weeks and low incubation, the germination rate and fraction of *C. canadensis* was well matched to that of *H. matronalis* (Fig. 1, 2, Tab. S1).

207 Germination priority effects

In the absence of phenological advantage, the influence of adding one seedling of H. matronalis to a pot community was almost 4X less than adding one C. canadensis seedling on the pot-level RGRD (represented by the species' influence parameters—H. matronalis (β_{Hm}): 0.13, I_{90} : 0.08,0.17, C. canadensis (β_{Cc}): -0.40, I_{90} : -0.46, -0.35). Each day increase in the phenological advantage of H. matronalis had approximately the same influence on shifting the community biomass composition towards H. matronalis as adding an individual of that species to the community (seasonal priority effect (β_{pri}): 0.15, I_{90} : 0.09, 0.20, Fig. 3, Tab. S2). Together, these results suggest that H. matronalis will come to dominate the community biomass composition unless C. canadensis is at high relative abundance or the phenological advantage of H. matronalis is small.

Discussion

Environmental drivers of seasonal priority effects

Our results identify a natural mechanism—species' differential sensitivity to temperature—that can generate seasonal priority effects, and thus reshape the competitive landscape between native and invasive species.

In the absence of phenological advantage, the intrinsic competitive abilities of each species in our study suggest that the native species, *C. canadensis*, is the stronger competitor (Fig. 3). However, the influence of

one day of phenological advantage for H. matronalis virtually doubled its influence on the final community 222 composition, suggesting that seasonal priority effects play a major role in the competitive success of H. 223 matronalis (Fig. 3). Our results indicate that C. canadensis can compete with the invasive H. matronalis 224 at high relative abundance levels and/or when phenological advantage is low, joining a growing body of 225 experiments demonstrating that relative germination phenology can function as a seasonal priority effect, 226 enhancing the performance of the earliest germinating species at the expense of later germinants (Koerner 227 et al., 2008; Dickson et al., 2012; Ross & Harper, 1972). 228 Our findings advance decades of work on seasonal priority effects by connecting them directly to environmental 229 variation and extending them to forest systems. While research on the importance of seasonal priority effects 230 has increased in recent years, the vast majority of studies have focused on dry grassland communities, and 231 typically introduce priority effects artificially — by staggered planting of competitors (Young et al., 2017; Wei-232 dlich et al., 2020). In contrast, our experiment employed a more ecologically realistic mechanism—generating 233 variation in phenological advantage indirectly by leveraging interspecific differences in germination responses to cold stratification. Our results suggest that seasonal priority effects can strongly influence seedling com-235 petition even in environment where germination phenology is primarily under temperature control. The results of our pair-wise competition trial suggests that seasonal priority effects, manifested through rapid germination phenology and propagule pressure, are mechanistically related to the competitive dominance,

Seasonal priority effects and anthropogenic climate change

and—ultimately—invasion success of *H. matronalis*.

The implications of our study for the role of climate variability in mediating seasonal priority effects is twofold. First, our results suggest that interannual climate variability should generate both among- and withinseason variation in competition strength among species, potentially driving species coexistence via the storage
effect (Chesson, 2003). Second, the key role we observed of climate in generating germination advantage and
therefore seasonal priority effects suggests that sustained alteration to historic patterns of climate variability,

246 like those driven by anthropogenic climate change, are likely to alter the dynamics of competing seedlings.

247 Changing patterns of phenological assembly have downstream effects on the structure and function of plant

248 communities.

In our study, the phenological advantage of *H. matronalis* was maximized under lower stratification treat-

250 ments and warmer incubation temperatures. This suggests that the warming temperatures associated with

251 anthropogenic climate change may increase the magnitude of seasonal priority effects, largely due to the

delay of germination in more climate-sensitive native species like C. canadensis. Interestingly, the difference

in phenological advantage among our focal species was much higher in our germination assays than in our

4 competition trials even at comparable levels of stratification. There are likely several explanations for these

255 differences.

260

First, we used different metrics of germination speed, time to 50% germination (t50) and mean germination

time in each experiment. While the metrics are related and often confused, there are important differences

between them that make one or the other more appropriate for the two types of experiments we ran (see

Supporting Information: "Measures of germination speed"). Second, the incubation temperatures in our

greenhouse competition trials were more variable than in our growth chamber germination assays. The lower

germination fractions we observed in C. canadensis under greenhouse conditions suggest that the temperature

range was likely supra-optimal for this species, and the lower germination fraction increased the difference

between t50 estimates and mean germination time measurements. Finally, we conducted germination assays

and competition trials in different growth media (filter sand vs. potting soil), which have different moisture

retention and light transmissible capacities. Germination media can affect germination rates (Baskin &

Baskin, 2014), which may further explain differences among our two experiments.

Despite these differences, in both experiments increasing cold stratification advanced the germination phe-

nology of C. canadensis and weakly that of H. matronalis, resulting in weaker phenological advantage for H.

matronalis at higher stratification levels. (Fig. 2, Fig. S1) suggesting that the relationships between cold

stratification and germination phenology we observed were robust.

Climate change may also increase the risk of precocious phenology (Inouye, 2000). While in our experiment, 271 there was no cost to germinating too early, in natural systems, selection usually limits early activity. Across 272 environments, optimum germination phenology is likely driven by a trade-off between maximizing the length 273 of the growing season and the risk of exposure to damaging environmental episodes when germinating too 274 early (Augspurger & Salk, 2017). For example, in dry grassland ecosystems, the precocious germination of 275 invasives has a substantial cost if water availability is too low (Wainwright et al., 2011). In temperate forest 276 ecosystems, the primary risk of early phenology is damage from late season frost (Kollas et al., 2014) and 277 climate change is also altering the timing and frequency of frost events (Ma et al., 2019). Understanding how 278 climate change will reshape this tradeoff between seasonal priority effects and frost risk is a critical next step 279 for understanding plant community interactions in an era of global change.

While we found that seasonal priority effects impacted competition among seedlings, our experiment was 281 not able to quantify the role of seasonal priority effects in influencing the long-term, among-year dynamics 282 of these perennial species. Many studies suggest that these short term priority effects many be transient, though several studies that used staggered planting methods at similar scale to the phenological advantage we 284 observed in our trials saw the influence of these initial priority effects on community composition maintained several seasons later (Vaughn & Young, 2015; Young et al., 2017; Torrez et al., 2017). In perennial communities, these long terms dynamics are even more difficult to assess. Many perennial herbs, C. canadensis included, rely heavily on vegetative reproduction (Hawkins et al., 2005), and competition between ramets, and between ramets and seedlings, may also impact species interactions in the long-term. Understanding how phenological differences across life stages of long-lived perennial plants affects competition is an important need for predicting how communities may be impacted by interannual environmental variation and climate 291 change. 292

Because we found that seasonal priority effects mediated competition for early ontological stages (germination, seedling), our findings may be most relevant to global change biology in the context of native plant
establishment; whether colonizing new areas due to range shifts, recovering from novel disturbances or eco-

logical restoration. In fact, there has been a growing call to increase phenological diversity in restoration 296 planning (Hess et al., 2019). Studies have found that including early active species in plantings can suppress 297 the abundance of invaders in both grassland (Cleland et al., 2013) and forest ecosystems (Schuster et al., 298 2020). At the same time, restoration mixes tend to lack species which fill the early season phenological 299 niche (Havens & Vitt, 2016). The results of our study suggest that minimizing the priority effect conferred 300 to invasive species due to rapid germination and early phenology by including species with similar, early 301 phenological traits could be a powerful tool for managing plant invasions and restoring native ecosystems in 302 an era of global change. 303

Conclusion

By leveraging the differential germination sensitivities to environmental cues of two competing species to manipulate phenological advantage between them, we were able to quantify the contribution of seasonal priority 306 effects gained through rapid phenology on the competitive ability of the invasive species H. matronalis. We 307 found that priority effects were approximately as strong as the intrinsic competitive traits of H. matronalis 308 in influencing its competitive dominance over the native forest herb Cryptotaenia canadensis, suggesting sea-309 sonal priority effects mechanistically increase the invasion success of H. matronalis. Variation in germination 310 phenology was strongly mediated by differences in how species respond to temperature cues, suggesting that 311 sustained climate change will alter patterns of phenological advantage, potentially strengthening the seasonal 312 priority effects of invaders as climate warms. Our findings highlight the important role of phenological diver-313 sity in the invasion resistance of native plant communities, implying that measures of phenological diversity 314 should be incorporated into plant community assessments and ecological restoration. 315

316 Author Contributions

- DMB and EMW conceived of this manuscript; DMB designed and executed the experiments, collected the
- data and performed the analyses; DMB led the writing of the manuscript. All authors contributed critically
- to drafts and gave final approval for submission.

320 Acknowledgements

- special thanks to K.J Woodruff, L. Toomey and the rest of the Arnold Arboretum research greenhouses staff
- for helping to manage the growth facilities as well as maintain and repeat these experiments throughout lab
- closures, quarantines and all other Covid19-related challenges.

324 Conflict of Interest

The authors declare no conflict of interest.

Data Availability

- Data from the germination assays and competition trials, and associated modeling code will be made available
- at the time of publication at KNB (https://knb.ecoinformatics.org/).

$\mathbf{References}$

- Augspurger, C.K. & Salk, C.F. (2017) Constraints of cold and shade on the phenology of spring ephemeral herb species. *Journal of Ecology* **105**, 246–254.
- Baskin, C. & Baskin, J. (2014) Seeds: Ecology, Biogeography, and Evolution of Dormancy and Germination.
 Elsevier Inc.
- Baskin, C.C. & Baskin, J.M. (1988a) Germination ecophysiology of herbaceous plant species in a temperate region. *American Journal of Botany* **75**, 286–305.

- Baskin, J.M. & Baskin, C.C. (1988b) The ecological life cycle of cryptotaenia canadensis (l.) dc. (umbelliferae), a woodland herb with monocarpic ramets. *The American Midland Naturalist* **119**, 165–173.
- Bradford, K.J.K.J.. & Nonogaki, H. (2007) Seed development, dormancy, and germination. Blackwell Pub., Oxford, OX, UK; Ames, Iowa, USA.
- Bürkner, P.C. (2018) Advanced bayesian multilevel modeling with the r package brms. R Journal 10, 395-411.
- Cavieres, L. & Sierra-Almeida, A. (2017) Assessing the importance of cold-stratification for seed germination in alpine plant species of the high-andes of central chile. *Perspectives in Plant Ecology, Evolution and Systematics* **30**.
- Chesson, P. (2003) Quantifying and testing coexistence mechanisms arising from recruitment fluctuations.
 Theoretical Population Biology 64, 345–357.
- Cleland, E.E., Larios, L. & Suding, K.N. (2013) Strengthening invasion filters to reassemble native plant communities: Soil resources and phenological overlap. *Restoration Ecology* 21, 390–398.
- Connolly, J. & Wayne, P. (2005) Assessing determinants of community biomass composition in two-species plant competition studies. *Oecologia* **142**, 450–457.
- Cottingham, K.L., Lennon, J.T. & Brown, B.L. (2005) Knowing when to draw the line: designing more informative ecological experiments. Frontiers in Ecology and the Environment 3, 145–152.
- Dickson, T.L., Hopwood, J.L. & Wilsey, B.J. (2012) Do priority effects benefit invasive plants more than native plants? an experiment with six grassland species. *Biological Invasions* 14, 2617–2624.
- Francis, A., Cavers, P.B. & Warwick, S.I. (2009) The biology of canadian weeds. 140. hesperis matronalis l.

 Canadian Journal of Plant Science 89, 191–206.
- Fukami, T. (2015) Historical contingency in community assembly: Integrating niches, species pools, and priority effects. *Annual Review of Ecology, Evolution, and Systematics* **46**, 1–23.
- Fukami, T. & Nakajima, M. (2011) Community assembly: alternative stable states or alternative transient states? *Ecology Letters* **14**, 973–984.
- Gioria, M. & Pyšek, P. (2017) Early bird catches the worm: germination as a critical step in plant invasion.

 Biological Invasions 19, 1055–1080.
- Gioria, M., Pyšek, P. & Osborne, B.A. (2018) Timing is everything: does early and late germination favor invasions by herbaceous alien plants? *Journal of Plant Ecology* 11, 4–16.
- Hao, J.H., Qiang, S., Liu, Q.Q. & Cao, F. (2009) Reproductive traits associated with invasiveness in conyza sumatrensis. *Journal of Systematics and Evolution* 47, 245–254.
- Havens, K. & Vitt, P. (2016) The importance of phenological diversity in seed mixes for pollinator restoration.

 Natural Areas Journal 36, 531–537.
- Hawkins, T.S., Baskin, J.M. & Baskin, C.C. (2005) Life cycles and biomass allocation in seed- and rametderived plants of cryptotaenia canadensis (apiaceae), a monocarpic species of eastern north america. *Cana*dian Journal of Botany 83, 518–528.
- Hawkins, T.S., Baskin, J.M. & Baskin, C.C. (2007) Seed morphology, germination phenology, and capacity to form a seed bank in six herbaceous layer apiaceae species of the eastern deciduous forest. *Castanea* **72**, 8–14.
- Hess, M.C.M., Mesléard, F. & Buisson, E. (2019) Priority effects: Emerging principles for invasive plant species management. *Ecological Engineering* 127, 48–57.

- 376 Howe, H.F. & Smallwood, J. (1982) Ecology of seed dispersal 13, 201–228.
- Inouye, B. (2001) Response surface experimental designs for investigating interspecific competition. *Ecology* 82, 2696–2706.
- Inouye, D.W. (2000) The ecological and evolutionary significance of frost in the context of climate change.

 Ecology Letters 3, 457–463.
- Kardol, P., Souza, L. & Classen, A.T. (2013) Resource availability mediates the importance of priority effects in plant community assembly and ecosystem function. *Oikos* **122**, 84–94.
- Kew, R.B.G. (2022) Seed information database (sid). version 7.1.
- Koerner, C., Stoecklin, J., Reuther-Thiebaud, L. & Pelaez-Riedl, S. (2008) Small differences in arrival time influence composition and productivity of plant communities. *New Phytologist* 177, 698–705.
- Kollas, C., Körner, C. & Randin, C.F. (2014) Spring frost and growing season length co-control the cold range limits of broad-leaved trees. *Journal of Biogeography* 41, 773–783.
- Ma, Q., Huang, J.G., Hänninen, H. & Berninger, F. (2019) Divergent trends in the risk of spring frost damage to trees in europe with recent warming. *Global Change Biology* **25**, 351–360.
- Marushia, R.G., Cadotte, M.W. & Holt, J.S. (2010) Phenology as a basis for management of exotic annual plants in desert invasions. *Journal of Applied Ecology* 47, 1290–1299.
- Milbau, A., Nijs, I., Van Peer, L., Reheul, D. & De Cauwer, B. (2003) Disentangling invasiveness and invasibility during invasion in synthesized grassland communities. *New Phytologist* **159**, 657–667.
- Onofri, A., Gresta, F. & Tei, F. (2010) A new method for the analysis of germination and emergence data of weed species. Weed Research 50, 187–198.
- Ross, M.A. & Harper, J.L. (1972) Occupation of biological space during seedling establishment 60, 77–88.
- Rudolf, V.H.W. (2019) The role of seasonal timing and phenological shifts for species coexistence. *Ecology*Letters 22, 1324–1338.
- Schuster, M.J., Wragg, P.D., Williams, L.J., Butler, E.E., Stefanski, A. & Reich, P.B. (2020) Phenology matters: Extended spring and autumn canopy cover increases biotic resistance of forests to invasion by common buckthorn (rhamnus cathartica). Forest Ecology and Management 464, 118067.
- Smith, L.M. (2013) Extended leaf phenology in deciduous forest invaders: mechanisms of impact on native communities. *Journal of Vegetation Science* **24**, 979–987.
- Soltani, E., Ghaderi-Far, F., Baskin, C.C. & Baskin, J.M. (2015) Problems with using mean germination time to calculate rate of seed germination. *Australian Journal of Botany* **63**, 631–635.
- Susko, D.J. & Hussein, Y. (2008) Factors affecting germination and emergence of dame's rocket (hesperis matronalis). Weed Science **56**, 389–393.
- ten Brink, D.J., Hendriksma, H.P. & Bruun, H.H. (2013) Habitat specialization through germination cueing: a comparative study of herbs from forests and open habitats. *Annals of Botany* **111**, 283–292.
- Torrez, V., Mergeay, J., Meester, L.D., Honnay, O., Helsen, K. & Standish, R. (2017) Differential effects of dominant and subordinate plant species on the establishment success of target species in a grassland restoration experiment. *Applied vegetation science* **20**, 363–375.
- Van Clef, M. & Stiles, E.W. (2001) Seed longevity in three pairs of native and non-native congeners: Assessing invasive potential. *Northeastern Naturalist* 8, 301–310.

- Vaughn, K.J. & Young, T.P. (2015) Short-term priority over exotic annuals increases the initial density and longer-term cover of native perennial grasses. *Ecological Applications* **25**, 791–799.
- Wainwright, C.E. & Cleland, E.E. (2013) Exotic species display greater germination plasticity and higher germination rates than native species across multiple cues. *Biological Invasions* **15**, 2253–2264.
- Wainwright, C.E., Wolkovich, E.M. & Cleland, E.E. (2011) Seasonal priority effects: implications for invasion and restoration in a semi-arid system. *Journal of Applied Ecology* **49**, 234–241.
- Walck, J.L., Hidayati, S.N., Dixon, K.W., Thompson, K. & Poschlod, P. (2011) Climate change and plant regeneration from seed. *Global Change Biology* **17**, 2145–2161.
- Weidlich, E.W.A., Nelson, C.R., Maron, J.L., Callaway, R.M., Delory, B.M. & Temperton, V.M. (2020)
 Priority effects and ecological restoration. *Restoration Ecology* n/a, e13317.
- Wolkovich, E.M. & Cleland, E.E. (2011) The phenology of plant invasions: a community ecology perspective.

 Frontiers in Ecology and the Environment 9, 287–294.
- Young, T.P., Stuble, K.L., Balachowski, J.A. & Werner, C.M. (2017) Using priority effects to manipulate competitive relationships in restoration. *Restoration Ecology* 25, S114–S123.

Figures Figures

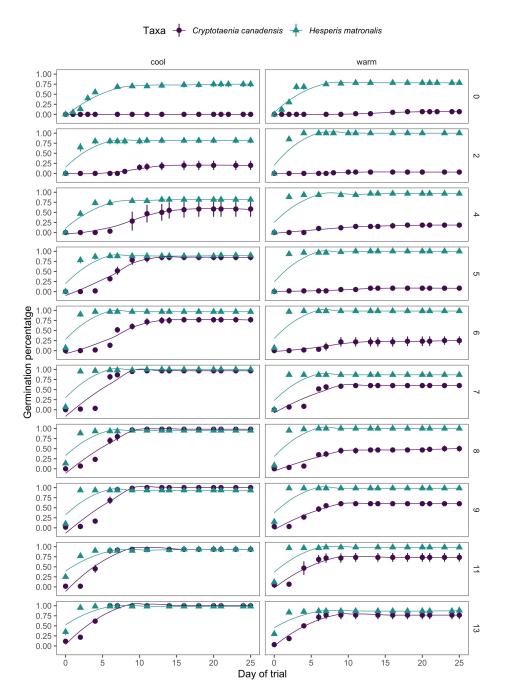


Figure 1: Cumulative germination rates of H. matronalis and C.canadensis under cool (20/10° C day/night) and warm (25/15° C day/night) incubation temperatures for each stratification treatment (number of weeks at 4° C, right axis) in our germination assays. Points depict the mean cumulative germination at each observation time, and vertical bars depict the standard error across replicates. Lines represents a generalized trend for cumulative germination, fit with a loess function. See Fig. 2 for model estimates from these data.

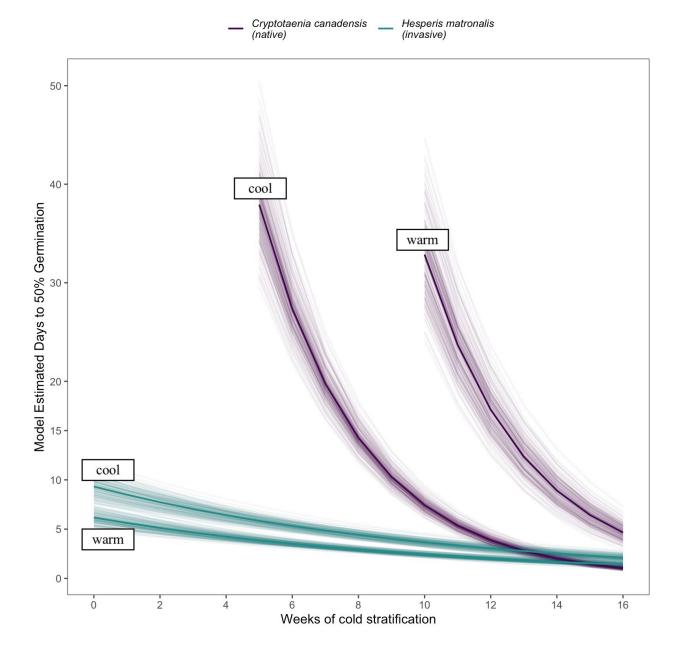


Figure 2: The effects of weeks of cold stratification at 4° C on the time to 50% germination of *Cryptotaenia* canadensis and *Hesperis matronalis* under warm (20/10° C day/night) and cool (25/15° C day/night) incubation conditions, estimated with accelerated failure time model. We show here only stratification treatment levels which allowed both species to reach 50% germination in less that 40 days. The solid lines depict the mean estimate, while lighter lines depict uncertainty.

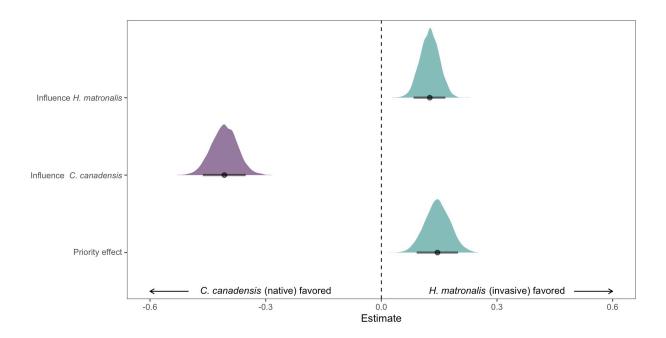


Figure 3: Estimated effects of species' abundance (intrinsic competitive ability parameters) and phenological advantage (seasonal priority effects) on the relative growth rate difference between *H. matronalis* and *C. canadensis*. Negative parameter estimates indicate the community biomass composition shifts to favor *C. candensis* while positive estimate towards dominance by *H. matronalis*. The points indicate the mean estimated effect of each parameter and bars the 90% uncertainty intervals. The full posterior distribution for each parameter is also depicted as an additional measure of uncertainty.