# Title: Maintenance of community function through compensation breaks down over time in a desert rodent community

# Authors: Renata M. Diaz1 and S. K. Morgan Ernest2

1. School of Natural Resources and Environment, University of Florida, Gainesville, FL. Corresponding author. renata.diaz@weecology.org

2. Department of Wildlife Ecology and Conservation, University of Florida, Gainesville, FL.

**Original submission:** This submission analyzes long-term data on rodent community abundance and energy use from the Portal Project. Sections of this timeseries have been analyzed in numerous other publications, but this is the first to analyze data from 2007-2020 on compensation on experimental and control plots.

**No prior publication:** This submission is posted as a preprint on bioRxiv at <https://www.biorxiv.org/content/10.1101/2021.10.01.462799v1>.

**Animal welfare:** Rodent censuses were conducted with IACUC approval, most recently under protocol 201808839\_01 at the University of Florida.

**Open research:** All data and code to reproduce these analyses are archived on Zenodo at <https://doi.org/10.5281/zenodo.5544362> and <https://doi.org/10.5281/zenodo.5539881>.

**Analytic methods:** All analyses were conducted in R version 4.0.3.

# Abstract

Understanding the ecological processes that maintain community function in systems experiencing species loss, and how these processes change over time, is key to understanding the relationship between community structure and function and predicting how communities may respond to perturbations in the Anthropocene. Using a 30-year experiment on desert rodents, we show that the impact of species loss on community-level energy use has changed repeatedly and dramatically over time, due to the addition of new species to the community, and a reduction in functional redundancy among the same set of species. Although strong compensation, initially driven by the dispersal of functionally redundant species to the local community, occurred in this system from 1996-2010, since 2010, compensation has broken down due to decreasing functional overlap within the same set of species. Simultaneously, long-term changes in sitewide community composition due to niche complementarity have decoupled the dynamics of compensation from the overall impact of species loss on community-level energy use. Shifting, context-dependent compensatory dynamics, such as those demonstrated here, highlight the importance of explicitly long-term, metacommunity, and eco-evolutionary perspectives on the link between species-level fluctuations and community function in a changing world.

# Key words: community function, compensation, zero-sum dynamic, environmental fluctuations, functional redundancy

# Introduction

Determining the extent to which community-level properties are affected by species loss, and how and why this changes over time, is key for understanding how communities are structured and how community function may respond to future perturbations (Gonzalez and Loreau 2009). When species are lost from a community, their contributions to community function (e.g. productivity or resource use) are also directly lost (Bannar-Martin et al. 2018). Community function may be maintained, however, if the remaining species in a community perform similar functions in the new community context as the lost species did originally, and compensate for the decline in function directly caused by species loss - i.e., functional redundancy (Walker 1992, 1995; Ernest and Brown 2001; Rosenfeld 2002; Gonzalez and Loreau 2009). When compensation via functional redundancy occurs among consumers with a common resource base, it is consistent with a zero-sum competitive dynamic, in which resources not used by one species are readily absorbed by competitors, and any increases in the abundance of one species must come at the direct expense of others (Van Valen 1973; Ernest et al. 2008).

Because the response of system-level function to species loss is partially determined by degree of functional redundancy in a community, processes that cause functional redundancy to change over time can have important consequences for the long-term maintenance of ecosystem function following species loss. Shifts in species composition (e.g., colonization events) may buffer community function against species loss, if a community gains species that are functionally similar to the species that are lost (Ernest and Brown 2001; Leibold et al. 2017). The ability of colonization to supply functionally redundant species depends on the species (and traits) present in the broader metacommunity, and on the rate of dispersal supplying appropriate species to local communities (Leibold et al. 2017).

Even without the addition of new species and traits, however, functional redundancy within a consistent set of coexisting species may fluctuate over time. While, in theory, functional redundancy may occur via the special case of complete niche neutrality (where functionally redundant species do not differ in any aspects of the niche; Hubbell 2001), it may also occur in niche-structured systems that contain species that share some traits but differ along other niche axes (Thibault et al. 2010). In these systems, if functionally-similar, but non-identical, species respond to environmental change in similar ways, functional overlap can be maintained or even strengthened. However, if niche differences cause species to respond differently to changing conditions, the degree of functional overlap between those species may decline, resulting in a breakdown in compensation (Loreau 2004; Fetzer et al. 2015). As the dynamics of both metacommunity processes and changing environmental conditions produce changes in functional redundancy in a community over time, the extent to which a community can compensate for species loss - and the strength of zero-sum competition for limiting resources - may also be dynamic and context-dependent.

Despite logical conceptual support, and evidence from experimental microcosms (Fetzer et al. 2015), there is little empirical documentation of how, and through which mechanisms, temporal changes in functional redundancy modulate the effect of species loss on ecosystem function in natural assemblages. Although relatively plentiful, observational data cannot unambiguously detect compensation through functional redundancy, and even short-term experiments may not be sufficient to capture temporal variation in compensation (Ernest and Brown 2001; Houlahan et al. 2007). In contrast, long-term manipulative experiments are uniquely suited to address this question. In long-term experiments in which key species are removed from a community over an extended period of time, the impact of species loss on community function can be directly quantified by comparing community function between complete and manipulated assemblages. As metacommunity dynamics and environmental conditions shift over time, long-term monitoring can reveal how these processes contribute to changes in functional redundancy and ecosystem function across different time periods. Due to the financial and logistical resources required to maintain and monitor whole-community manipulations over long timescales, these experiments are rare in natural systems representative of realistic evolutionary, geographic, and environmental constraints (Hughes et al. 2017).

Here, we use a 30-year experiment on desert rodents to investigate how shifts in functional redundancy alter the effect of species loss on community function over time. In this study, kangaroo rats (*Dipodomys* spp*.*), which are the largest and competitively dominant species in the rodent community, have been removed from a subset of experimental plots to explore how the loss of key species affects community function, measured as community-level metabolic flux (“total energy use”, or *Etot*; Ernest et al. 2019) or total biomass. For systems of consumers with a shared resource base, such as the granivorous rodents that make up this community, *Etot* reflects the total amount of resources being processed by an assemblage, and total biomass directly reflect standing biomass. Both are important metrics of community function (Lawton 1994; Ernest and Brown 2001). Long-term monitoring of this experiment has documented repeated shifts in the habitat and species composition of this system, resulting in distinct time periods characterized by different habitat conditions and configurations of the rodent community (Christensen et al. 2018). Abrupt reorganization events in community composition occurred in 1996 and in 2010, associated with the establishment and subsequent decline of the pocket mouse *Chaetodipus baileyi*. *C. baileyi* is similar in size, and presumably other traits, to kangaroo rats, and its establishment in 1996 drove a pronounced increase in compensation due to functional redundancy between *C. baileyi* and kangaroo rats (Ernest and Brown 2001; Thibault et al. 2010). Over the course of this experiment, shifting environmental conditions have caused the habitat at the study site to transition from desert grassland to scrub, driving a shift in baseline rodent community composition away from kangaroo rats and favoring other, smaller, granivores (Brown et al. 1997; Ernest et al. 2008). By making comparisons across these time periods, we explored how changes in community composition, and functional overlap among the same set of species, have contributed to changes in how species loss has impacted *Etot* over time*.*

# Methods

## The Portal Project

The Portal Project consists of a set of 24 fenced experimental plots located approximately 7 miles east of Portal, AZ, USA, on unceded land of the Chiricahua Apache. Beginning in 1977, kangaroo rats (*Dipodomys spectabilis, D. merriami,* and *D. ordii*) have been experimentally excluded from a subset of these plots (exclosures), while all other rodents are allowed access through small holes cut in the plot fencing. Control plots, with larger holes, are accessible to all rodents, including kangaroo rats. Rodents on all plots are censused via monthly live-trapping, and each individual captured is identified to species and weighed. For additional details on the site and methodology of the Portal Project, see Ernest et al. (2019).

## Data

We used data for control and exclosure plots from February 1988 until January 2020. The experimental treatments for some plots have changed over time, and we used the subset of plots that have had the same treatments for the longest period of time (Ernest et al. 2019). Four control plots, and five exclosure plots, met these critera. In order to achieve a balanced sample, we randomly selected four exclosure plots for analysis. We broke the timeseries into three time periods defined by the major transitions in the rodent community surrounding the establishment of *C. baileyi* in 1996-1997 and its subsequent decline in 2009-2010 (Ernest and Brown 2001; Thibault et al. 2010; Christensen et al. 2018). We defined the first time period as from February 1988 to June 1997, which was the first census period in which *C. baileyi* was captured on all exclosure plots at the site (following Bledsoe and Ernest, 2019). The second time period spans from July 1997 until January 2010, which is the midpoint of the 95% credible interval for the timing of the last reorganization event as estimated in Christensen et al. (2018). The last time period spans from February 2010 until January 2020, when data collection was interrupted by the COVID-19 pandemic. For each individual rodent captured, we estimated the individual-level metabolic rate using the allometric scaling relationship between individual body mass and metabolic rate 5.69 \* (*m*0.75), where *m* is body mass in grams (Nagy et al. 1999; see White et al. 2004 for details). We calculated treatment and species-level energy use as the sum of the appropriate individuals’ metabolic rates, and total biomass as the sum of their raw body mass measurements. All data were accessed using the R package *portalr* (Christensen et al. 2019).

## Rodent community energy use

For all variables, we combined data for all plots within a treatment in each monthly census period and calculated treatment-level means. This is necessary to calculate compensation, and we treated other variables in the same way to maintain consistency. To measure the overall impact of kangaroo rat removal on overall community function, we calculated a “total energy ratio” as the ratio of treatment-level *Etot* for kangaroo-rat exclosure plots relative to unmanipulated control plots, i.e. *EtotE/EtotC* where *EtotE* and *EtotC* are total energy use on exclosures and controls, respectively (Thibault et al 2010; Bledsoe and Ernest 2019). The total function ratio is distinct from compensation, which we defined as the proportion of the energy made available by removing kangaroo rats from the community that is taken up via compensatory increases in energy use by small granivores (all granivorous species other than kangaroo rats; *Baiomys taylori, C. baileyi, Chaetodipus hispidus, Chaetodipus intermedius, Chaetodipus penicillatus, Perognathus flavus, Peromyscus eremicus, Peromyscus leucopus, Peromyscus maniculatus, Reithrodontomys fulvescens, Reithrodontomys megalotis,* and *Reithrodontomys montanus*). We calculated this as (*SGE - SGC)/KRC*, where *SGE* and *SGC* are the amount of energy used by small granivores (SG) on exclosure and control plots, respectively, and *KRC* is the amount of energy used by kangaroo rats (KR) on control plots (Ernest and Brown 2001). To compare these variables across time periods, we used generalized least squares models (the R package *nlme*; Pinheiro et al. 2020) of the form (*SGE – SGC)/KRC* ~ *time period*, for compensation, and *EtotE/EtotC* ~ *time period*, for the total function ratio. We included a continuous-time autoregressive temporal autocorrelation term to account for temporal autocorrelation between values from monthly census periods within each multi-year time period, which improved model fit considerably (for details of model selection, see Appendix S1; for code detailing these analyses, see the online repositories, below). To evaluate change in baseline community composition over time, we calculated the proportion of treatment-level energy use accounted for by kangaroo rats on control plots in each census period (e.g. *KRC/EtotC*). Because proportional abundance is bounded 0-1 and is therefore not appropriate for generalized least squares, we compared values across time periods using a generalized linear model with a binomial link function of the form *KRC/EtotC ~ time period*. Finally, we calculated the proportional energy use accounted for by *C. baileyi* (CB)*,* specifically, on exclosure and control plots in each census period (e.g. *CBE/EtotE* and *CBC/EtotC*, respectively). *C. baileyi* was not present at the site prior to 1996, and we restricted the analysis of *C. baileyi* proportional energy use to July 1997-2020. We explored whether *C. baileyi* proportional energy use differed between treatment types, and whether there was an interaction between treatment type and time period on *C. baileyi* proportional energy use (for details, see S1). We found no evidence of an interaction between treatment type and time period, and therefore compared *C. baileyi* proportional energy use over time and across treatments using a binomial generalized linear model of the form *CBE/EtotE ~ time period + treatment*). For all models, wecalculated estimated means and 95% confidence or credible intervals for time-period (and, for *C. baileyi* proportional energy use, treatment) level values, and contrasts between time periods (and, for *C. baileyi* proportional energy use, treatments), using the R package *emmeans* (Lenth 2021). For biomass, we repeated these analyses substituting biomass values for energy use values throughout.

# Results

    The impact of kangaroo rat removal on community function has changed repeatedly over time, through a combination of abrupt shifts in compensation associated with *C. baileyi*, and long-term changes in baseline community composition sitewide (Figure 1). These dynamics are qualitatively identical whether function is measured as total energy use (Figure 1) or total biomass (Appendix S3 Figure 1). The first shift occurred around 1996, when *C. baileyi* - which had not previously been observed at the site - established in the community (Figure 1D; see also Ernest and Brown 2001). *C. baileyi* rapidly became dominant on exclosure plots and dramatically increased compensation (Figure 1B). From 1996-2010, small granivores increased to compensate for 55% of kangaroo rat energy use on control plots (95% interval 46-63%), from an average of 19% from 1988-1996 (95% interval 9-29%; contrast *p* < 0.001; for complete results of all models, see Appendix 1) from 1996-2010. With *C. baileyi*’s addition to the community, the total function ratio (on exclosures relative to controls; Figure 1A) increased from 30% (20-39%) to 68% (60-77%, contrast *p* < 0.001). In the second shift, beginning around 2010, *C. baileyi*’s abundance sitewide dropped precipitously (Figure 1D). *C. baileyi*’s proportional energy use dropped from an average of 72% (70-75%) to 25% (22-28%, contrast *p* < 0.001) on exclosure plots, and from 12% (10-14%) to essentially 0 on control plots (contrast *p* < 0.001). Other species of small granivore did not make compensatory gains to offset the decline in *C. baileyi* (Figure 1B)*.* As a result, compensation declined from an average of 55% (46-63%) to 22% (12-32%, contrast *p* < 0.001), a level not significantly different from the 19% (9-29%, contrast *p* = .9) observed prior to *C. baileyi*’s establishment at the site. Somewhat paradoxically, while the total function ratio also dropped following *C. baileyi*’s decline, from an average of 68% (60-77%) from 1996-2010 to 46% (37-56%, contrast *p* = 0.002) from 2010-2020, it remained higher than its average of 30% (20-39%, contrast *p* = 0.04) from 1988-1996 (Figure 1A). Over the course of the experiment, rodent community composition shifted sitewide, such that in later years, kangaroo rats have accounted for a lower proportion of baseline *Etot* than they did at the beginning of the study (Figure 1C). From 1988-1996, kangaroo rats accounted for 92% (90-94%) of *Etot* on controls; in later time periods, this dropped to an average of approximately 70% (1988-1996 compared to later time periods, *p* < 0.001; 1996-2010 and 2020-2020 not significantly different, *p =* .86). Because the proportion of *Etot* directly lost to kangaroo rat removal was smaller from 2010-2020 than from 1988-1996, the total function ratio was higher from 2010-2020 than it was from 1988-1996 - even though there was not a detectable difference between the two time periods in the proportion of lost energy being offset through compensation.

# Discussion

    The dynamics of rodent community energy use at Portal illustrate that the role of functional redundancy in buffering community function against species loss fluctuates over time, due to changes in both species composition and in the degree of functional overlap among the same species. The 1996 increase in compensation, driven by *C. baileyi*’s establishment at the site, was a clear and compelling instance of colonization from the regional species pool overcoming a dispersal constraint on functional redundancy (Ernest and Brown 2001; Leibold et al 2017). Although the small granivore species originally present in the community did not possess the traits necessary to compensate for kangaroo rats, *C. baileyi* supplied those traits and substantially, but incompletely, restored community function on exclosure plots. In contrast, following the community reorganization event in 2010, *C. baileyi* remained present in the community, butceased to operate as a partial functional replacement for kangaroo rats. This is consistent with fluctuating conditions modulating the degree of functional redundancy between similar, but non-identical, competitors. Kangaroo rats and *C. baileyi* are relatively similar in size and are demonstrably capable of using similar resources. However, *C. baileyi* prefers different, shrubbier microhabitats than kangaroo rats, and the two groups have been observed to replace each other in adjacent habitats (Rosenzweig and Winakur 1969; M’Closkey 1982; Price 1978). We suggest that this study site, which has historically been dominated by kangaroo rats, constitutes marginal habitat for *C. baileyi*, and that, while conditions from 1996-2010 aligned sufficiently with *C. baileyi*’s requirements to create appreciable functional redundancy between kangaroo rats and *C. baileyi*, conditions since 2010 have caused this redundancy to break down. *C. baileyi*’s decline occurred immediately following a period of low plant productivity and low rodent abundance community-wide (Appendix 2; Christensen et al. 2018). In the decade after, the site experienced two long and severe droughts interspersed with an exceptionally wet period (Appendix 2; Christensen et al. 2018). These extreme conditions may themselves have limited *C. baileyi’s* fitness at the site, or the community-wide low abundance event may have temporarily overcome incumbency effects and triggered a community reorganization tracking longer-term habitat shifts (Thibault and Brown 2008; Christensen et al. 2018). Regardless of the proximate cause of *C. baileyi*’s decline, the fact that *C. balieyi* remains in the community, but no longer compensates for kangaroo rats, illustrates that changing conditions can have profound effects on community function by modulating the degree of functional redundancy within a consistent set of species.

While changes in compensation over time have contributed to changes in community function in this system, changes in compensation alone do not fully account for the long-term changes in the overall impact of kangaroo rat removal on *Etot*. Specifically, although the ratio of *Etot* on exclosure plots relative to control plots declined coinciding with the breakdown in compensation associated with *C. baileyi*, it remained higher than the levels observed prior to 1996 (Figure 1A). This difference in exclosure *Etot* between the first and last time periods cannot be explained by an increase in compensation, as compensation from 2010-2020 approximated pre-1996 levels (Figure 1B). Rather, the increase in *Etot* on exclosure plots relative to control plots was the result of a long-term decrease in the contribution of kangaroo rats to *Etot* sitewide. Because kangaroo rats accounted for a smaller proportion of *Etot* on control plots from 2010-2020 than they did prior to 1996, their removal had a smaller impact on community function – even though there was not an increase in the degree to which small granivores compensated for their absence. In fact, the relatively consistent levels of compensation achieved in the decades preceding and following *C. baileyi*’s dominance at the site suggest a relatively stable, and limited, degree of functional overlap between kangaroo rats and the original small granivores (i.e., excluding *C. bailyei*). Niche complementarity, combined with changing habitat conditions, may partially explain how small granivores have come to account for a larger proportion of community *Etot* over time, but have not increased their capacity to exploit resources usually controlled by kangaroo rats. It is well-documented that, while kangaroo rats readily forage in open microhabitats where predation risk can be relatively high, smaller granivores rely on vegetation cover as an antipredator defense and preferentially forage in sheltered microhabitats (Kelt 2011). Over the course of this experiment, the habitat at this study site has transitioned from an arid grassland to a shrubland (Brown et al. 1997). As the landscape availability of open and sheltered microhabitats shifted, small granivores may have gained access to a larger proportion of resources and increased their share of *Etot* sitewide. However, kangaroo rats may have continued to use resources predominately located in the remaining open areas – resources that would have remained inaccessible to smaller granivores, even on plots where kangaroo rats were removed. The long-term reduction in the impact of kangaroo rat removal on community function, driven by niche complementarity and consistent niche partitioning, contrasts with the temporary compensatory dynamic driven by functional redundancy with *C. baileyi*. Although changes in the overall effect of species loss are sometimes treated interchangeably with compensation (e.g. Ernest and Brown 2001 compared to Thibault et al. 2010, Bledsoe and Ernest 2019), especially at short timescales, it is important to recognize that multiple distinct pathways modulate the long-term impacts of species loss on community function. Particularly in strongly niche-structured systems, such as the desert granivores studied here, complementarity effects and fluctuations in functional redundancy may occur simultaneously, with complex and potentially counterintuitive outcomes for community-level function.

Overall, the decadal-scale changes in energy use among the Portal rodents underscore the importance of long-term metacommunity dynamics to the maintenance of community function following species loss (see also Leibold et al. 2017). Although a single colonization event may be enough to allow for temporary compensation via functional redundancy, as conditions fluctuate, species that are capable of compensating under some conditions may no longer perform that function (see also Isbell et al. 2011). Particularly if limiting similarity prevents similar competitors from specializing on precisely the same habitats (Rosenfeld 2002), this type of temporary, context-dependent compensation may be common. To maintain compensation for species loss over time, multiple colonization events, supplying species that are functionally redundant under different conditions, may be required. Depending on dispersal rates and the composition of regional species pools, repeated colonization events tracking changing conditions may be unlikely or even impossible. At Portal, dispersal limitation introduced a 20-year lag between the initiation of the experiment and the beginning of energetic compensation driven by *C. baileyi*. Theoretically, another species capable of compensating for kangaroo rats, and better-suited to conditions at the site since 2010, could colonize the site and restore compensation – but it is unclear whether such a species exists or how long it might take for it to disperse to the site. More generally, as ecosystems globally undergo reductions in habitat connectivity and regional beta diversity, and move into novel climatic spaces, maintenance of community function through functional redundancy may become increasingly rare and fragile (Dornelas et al. 2014; Williams and Jackson 2007).

Finally, the long-term variability in functional redundancy documented here adds important nuance to our understanding of how zero-sum dynamics operate in natural assemblages. Theories invoking zero-sum dynamics, and tests for compensatory dynamics in empirical data, often treat a zero-sum dynamic as a strong and temporally consistent constraint on population dynamics (Hubbell 2001; Houlahan et al. 2007). In this framing, any resources made available through species loss should immediately be taken up by other species. This is not consistent with the dynamics that occur at Portal, as there have been extended periods of time when there are substantial resources available on exclosure plots that are not being used by any rodent species. Rather, these results are more consistent with a zero-sum constraint operating at metacommunity, or, more broadly, evolutionary scales (Van Valen 1973; Terry and Rowe 2015; Leibold et al. 2017). Over short timescales, or within a closed local assemblage, niche differences may weaken zero-sum effects, especially under fluctuating conditions. However, over larger temporal and spatial scales, dispersal or evolution may supply new species equipped to take advantage of available resources - either through functional redundancy with species that have been lost, or niche complementarity allowing them to exploit novel niches. Moving forward, a long-term, metacommunity, and even macroevolutionary approach may be necessary to fully understand how zero-sum constraints manifest in community dynamics; how functional redundancy and niche complementarity jointly contribute to dynamics of community function; and how, and when, community-level properties are maintained in the face of species extinctions and changing conditions over time.

# Acknowledgements

The Portal Project has been supported by numerous grants from the U.S. National Science Foundation, most recently NSF DEB-1929730, and has been sustained over the decades via the dedicated efforts of dozens of researchers and volunteers. RMD was supported in part by the National Science Foundation Graduate Research Fellowship under Grant No. DGE-1315138 and DGE-1842473.

# Literature cited

Bannar‐Martin, K. H., C. T. Kremer, S. K. M. Ernest, M. A. Leibold, H. Auge, J. Chase, S. A. J. Declerck, N. Eisenhauer, S. Harpole, H. Hillebrand, F. Isbell, T. Koffel, S. Larsen, A. Narwani, J. S. Petermann, C. Roscher, J. S. Cabral, and S. R. Supp. 2018. Integrating community assembly and biodiversity to better understand ecosystem function: the Community Assembly and the Functioning of Ecosystems (CAFE) approach. Ecology Letters 21:167–180.

Bledsoe, E. K., and S. K. M. Ernest. 2019. Temporal changes in species composition affect a ubiquitous species’ use of habitat patches. Ecology 100:e02869.

Brown, J. H., T. J. Valone, and C. G. Curtin. 1997. Reorganization of an arid ecosystem in response to recent climate change. Proceedings of the National Academy of Sciences 94:9729–9733.

Christensen, E. M., D. J. Harris, and S. K. M. Ernest. 2018. Long-term community change through multiple rapid transitions in a desert rodent community. Ecology 99:1523–1529.

Christensen, E. M., G. M. Yenni, H. Ye, J. L. Simonis, E. K. Bledsoe, R. M. Diaz, S. D. Taylor, E. P. White, and S. K. M. Ernest. 2019. portalr: an R package for summarizing and using the Portal Project Data. Journal of Open Source Software 4:1098.

Dornelas, M., N. J. Gotelli, B. McGill, H. Shimadzu, F. Moyes, C. Sievers, and A. E. Magurran. 2014. Assemblage Time Series Reveal Biodiversity Change but Not Systematic Loss. Science 344:296–299.

Ernest, S. K. M., and J. H. Brown. 2001. Delayed Compensation for Missing Keystone Species by Colonization. Science 292:101–104.

Ernest, S. K. M., J. H. Brown, K. M. Thibault, E. P. White, and J. R. Goheen. 2008. Zero Sum, the Niche, and Metacommunities: Long‐Term Dynamics of Community Assembly. The American Naturalist 172:E257–E269.

Ernest, S. K. M., G. M. Yenni, G. Allington, E. K. Bledsoe, E. M. Christensen, R. M. Diaz, K. Geluso, J. R. Goheen, Q. Guo, E. Heske, D. Kelt, J. M. Meiners, J. Munger, C. Restrepo, D. A. Samson, M. R. Schutzenhofer, M. Skupski, S. R. Supp, K. Thibault, S. Taylor, E. White, H. Ye, D. W. Davidson, J. H. Brown, and T. J. Valone. 2020. The Portal Project: a long-term study of a Chihuahuan desert ecosystem. bioRxiv:332783.

Fetzer, I., K. Johst, R. Schäwe, T. Banitz, H. Harms, and A. Chatzinotas. 2015. The extent of functional redundancy changes as species’ roles shift in different environments. Proceedings of the National Academy of Sciences 112:14888–14893.

Gonzalez, A., and M. Loreau. 2009. The Causes and Consequences of Compensatory Dynamics in Ecological Communities. Annual Review of Ecology, Evolution, and Systematics 40:393–414.

Houlahan, J. E., D. J. Currie, K. Cottenie, G. S. Cumming, S. K. M. Ernest, C. S. Findlay, S. D. Fuhlendorf, R. D. Stevens, T. J. Willis, I. P. Woiwod, and S. M. Wondzell. 2007. Compensatory dynamics are rare in natural ecological communities. Proceedings of the National Academy of Sciences. 104(9): 3273-3277.

Hubbell, S. P. 2001. The Unified Neutral Theory of Biodiversity and Biogeography (MPB-32). Princeton University Press.

Hughes, B. B., R. Beas-Luna, A. K. Barner, K. Brewitt, D. R. Brumbaugh, E. B. Cerny-Chipman, S. L. Close, K. E. Coblentz, K. L. de Nesnera, S. T. Drobnitch, J. D. Figurski, B. Focht, M. Friedman, J. Freiwald, K. K. Heady, W. N. Heady, A. Hettinger, A. Johnson, K. A. Karr, B. Mahoney, M. M. Moritsch, A.-M. K. Osterback, J. Reimer, J. Robinson, T. Rohrer, J. M. Rose, M. Sabal, L. M. Segui, C. Shen, J. Sullivan, R. Zuercher, P. T. Raimondi, B. A. Menge, K. Grorud-Colvert, M. Novak, and M. H. Carr. 2017. Long-Term Studies Contribute Disproportionately to Ecology and Policy. BioScience 67:271–281.

Isbell, F., V. Calcagno, A. Hector, J. Connolly, W. S. Harpole, P. B. Reich, M. Scherer-Lorenzen, B. Schmid, D. Tilman, J. van Ruijven, A. Weigelt, B. J. Wilsey, E. S. Zavaleta, and M. Loreau. 2011. High plant diversity is needed to maintain ecosystem services. Nature 477:199–202.

Kelt, D. A. 2011. Comparative ecology of desert small mammals: a selective review of the past 30 years. Journal of Mammalogy 92:1158–1178.

Lawton, J. H. 1994. What Do Species Do in Ecosystems? Oikos 71:367–374.

Leibold, M. A., J. M. Chase, and S. K. M. Ernest. 2017. Community assembly and the functioning of ecosystems: how metacommunity processes alter ecosystems attributes. Ecology 98:909–919.

Lenth, R. V. 2021. emmeans: Estimated Marginal Means, aka Least-Squares Means.

Loreau, M. 2004. Does functional redundancy exist? Oikos 104:606–611.

M’Closkey, R. T. 1982. The principle of equal opportunity: a test with desert rodents. Canadian Journal of Zoology 60:1968–1972.

Nagy, K. A., I. A. Girard, and T. K. Brown. 1999. Energetics of Free-Ranging Mammals, Reptiles, and Birds. Annual Review of Nutrition 19:247–277.

Pinheiro, J., D. Bates, S. DebRoy, D. Sarkar, and R Core Team. 2020. nlme: Linear and Nonlinear Mixed Effects Models.

Price, M. V. 1978. The Role of Microhabitat in Structuring Desert Rodent Communities. Ecology:13.

R Core Team. 2020. R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna, Austria.

Rosenfeld, J. S. 2002. Functional redundancy in ecology and conservation. Oikos 98:156–162.

Rosenzweig, M. L., and J. Winakur. 1969. Population Ecology of Desert Rodent Communities: Habitats and Environmental Complexity. Ecology 50:558–572.

Terry, R. C., and R. J. Rowe. 2015. Energy flow and functional compensation in Great Basin small mammals under natural and anthropogenic environmental change. Proceedings of the National Academy of Sciences 112:9656–9661.

Thibault, K. M., and J. H. Brown. 2008. Impact of an extreme climatic event on community assembly. Proceedings of the National Academy of Sciences of the United States of America 105:3410–3415.

Thibault, K. M., S. K. M. Ernest, and J. H. Brown. 2010. Redundant or complementary? Impact of a colonizing species on community structure and function. Oikos 119:1719–1726.

Van Valen, L. 1973. A new evolutionary law. Evolutionary Theory 1:1–30.

Walker, B. 1995. Conserving Biological Diversity through Ecosystem Resilience. Conservation Biology 9:747–752.

Walker, B. H. 1992. Biodiversity and Ecological Redundancy. Conservation Biology 6:18–23.

White, E. P., S. K. M. Ernest, and K. M. Thibault. 2004. Trade‐offs in Community Properties through Time in a Desert Rodent Community. The American Naturalist 164:670–676.

Williams, J. W., and S. T. Jackson. 2007. Novel climates, no-analog communities, and ecological surprises. Frontiers in Ecology and the Environment 5:475–482.

# Figure legends

**Figure 1.** Dynamics of energy use and rodent community composition over time. Lines represent the ratio of energy use on exclosure plots to control plots (a), 6-month moving averages of energetic compensation (b), and the share of community energy use accounted for by kangaroo rats on control plots (c), and by *C. baileyi* (d), on control (gold) and exclosure (blue) plots. Dotted vertical lines mark the boundaries between time periods used for statistical analysis. Horizontal lines are time-period estimates from generalized least squares (a, b) and generalized linear (c, d) models, and the semitransparent envelopes mark the 95% confidence or credible intervals.

# Figure 1

