

Linking evolutionary and ecological theory illuminates non-equilibrium biodiversity

Abstract

Whether or not biodiversity dynamics tend toward stable equilibria remains an unsolved question in ecology and evolution with important implications for our understanding of diversity and its conservation. Phylo/population genetic models and macroecological theory represent two primary lenses through which we view biodiversity. While phylo/population genetics provide an averaged view of changes in demography and diversity over timescales of generations to geological epochs, macroecology provides an ahistorical description of commonness and rarity across contemporary co-occurring species. Our goal is leverage advances in community-wide high throughput sequencing technology, specifically metabarcoding, to combine evolutionary and macroecological approaches to gain novel insights into the non-equilibrium nature of biodiversity. We help guide near future research with a call for bioinformatic advances and an outline of quantitative predictions made possible by our approach.

Non-equilibrium, inference, and theory in ecology and evolution

The idea of an ecological and evolutionary equilibrium has pervaded studies of biodiversity from geological to ecological, and from global to local (Sepkoski 1984; Chesson 2000; Hubbell 2001; Tilman 2004; Rabosky 2009; Harte 2011). The consequences of non-equilibrium dynamics for biodiversity are not well understood and the need to understand them is critical with anthropogenic pressures forcing biodiversity into states of rapid transition (Blonder *et al.* 2015). Non-equilibrial processes could profoundly inform conservation in ways only just beginning to be explored (Wallington *et al.* 2005).

Biodiversity theories based on assumptions of equilibrium, both mechanistic (Chesson 2000; Hubbell 2001; Tilman 2004) and statistical (**see the Glossary**) (Pueyo *et al.* 2007; Harte 2011) have found success in predicting ahistorical patterns of diversity such as the species abundance distribution (Hubbell 2001; Harte 2011; White *et al.* 2012) and the species area relationship (Hubbell 2001; Harte 2011). These theories assume a macroscopic equilibrium in terms of these coarse-grained metrics, as opposed to focusing on details of species identity (such as in Blonder *et al.* (2015)), although macroscopic and microscopic approaches are not mutually exclusive. Nonetheless, the equilibrium assumed by these theories is not realistic (Ricklefs 2006), and many processes, equilibrial or otherwise, can generate the same macroscopic, ahistorical predictions (McGill *et al.* 2007).

Tests of equilibrial ecological theory alone will not allow us to identify systems out of equilibrium, nor permit us to pinpoint the mechanistic causes of any non-equilibrial processes. The dynamic natures of evolutionary innovation and landscape change suggest that ecological theory could be greatly enriched by synthesizing its insights with inference from population genetic theory that explicitly accounts for history. This would remedy two shortfalls of equilibrial theory: 1) if theory fits observed ahistorical patterns but the implicit dynamical assumptions were wrong, we would make the wrong conclusion about the equilibrium of the system; 2) if theories do not fit the data we cannot know why unless we have a perspective on the temporal dynamics underlying those data.

No efforts to date have tackled these challenges. We propose that combining insights from ecological theory and inference of evolutionary and demographic change from genetic data will allow us to understand and

predict the consequences of non-equilibrial processes in governing the current and future states of ecological assemblages. The advent of high throughput sequencing-enabled metabarcoding has made unprecedented data available about the biodiversity of lineages from microbes to arthropods (Taberlet *et al.* 2012; Ji *et al.* 2013; Zhou *et al.* 2013; Bohmann *et al.* 2014; Gibson *et al.* 2014; Dodsworth 2015; Leray & Knowlton 2015; Linard *et al.* 2015; Shokralla *et al.* 2015; Venkataraman *et al.* 2015; Liu *et al.* 2016). These metabarcoding data have great potential to yield synthetic insight across ecology and evolution. However, to draw such insights we need a new tool set of bioinformatic methods (**Box box:dry**) and meaningful predictions (**section ref sec:pred**) grounded in theory to make use of those data. We present the foundation of this methodological tool set here.

Ecological theories and non-equilibrium

Neutral and statistical theories in ecology focus on macroscopic patterns, and equilibrium is presumed to be relevant to those patterns. Our goal throughout is not to validate neutral or statistical theories—quite the opposite, we propose new data dimensions, namely genetics, to help better test alternative hypotheses against these null theories, thereby gaining insight into what non-neutral and non-statistical mechanisms are at play in systems of interest.

Non-neutral and non-statistical models (e.g., Tilman (2004); Chesson (2000)) also invoke ideas of equilibrium in their derivation. However, these equilibria focus on the micro-scale details of species interactions, which could in fact lead to instability and non-equilibrium at larger scales (cite: romGEB, critical transition in food web stuff from Beth, random matrix theory). Thus, detailed species interactions could in fact be drivers of non-equilibrium and thus interesting hypotheses to test as alternatives to neutral/statistical models.

To use neutral/statistical theories as null models, we need a robust measure of goodness of fit. The emerging consensus is that likelihood-based test statistics should be preferred (Baldrige *et al.* 2016). The “exact test” of Etienne (2007) has been extended by Rominger & Merow (2017) into a simple z-score which can parsimoniously describe the goodness of fit between theory and pattern. We advocate its use in our proposed framework.

The unified neutral theory of biodiversity (UNTB) (Hubbell 2001) is a useful null because it assumes that one mechanism—demographic drift—drives community assembly. Equilibrium occurs when homogeneous stochastic processes of birth, death, speciation and immigration have reached stationarity. Thus neutrality in ecology is analogous to neutral drift in population genetics (Hubbell 2001) (also cite molecular neutral theory).

Rather than assuming any one mechanism dominates the assembly of populations into a community, statistical theories assume all mechanisms could be valid, but their unique influences have been lost to the enormity of the system and thus the outcome of assembly is a community in statistical equilibrium [Harte (2011); pueyo2007]. The mechanistic agnosticism is what makes statistical theories useful nulls. These statistical theories are also consistent with niche-based equilibria (Pueyo *et al.* 2007; Neill *et al.* 2009) if the complicated, individual or population level models with many mechanistic drivers were to be upscaled to entire communities.

The maximum entropy theory of ecology (METE) (Harte 2011) derives its predictions by condensing the many bits of mechanistic information down into ecological state variables and then mathematically maximizing information entropy conditional on those state variables. METE can predict multiple ahistorical patterns, including distributions of species abundance, body size, spatial aggregation, and trophic links (Harte 2011; Rominger *et al.* 2015; Rominger & Merow 2017), making for a stronger null theory (McGill 2003). However, multiple dynamics can still map to this handful of metrics (McGill *et al.* 2007) and while extensive testing often supports METE’s predictions [Harte (2011); white2012; xiao2015] at single snapshots in time, METE fails to match observed patterns in disturbed and rapidly evolving communities [Rominger *et al.* (2015); harte2011]. Just like deviations from UNTB, we cannot know the cause of these departures from theoretical predictions without adding tests of theories and metrics that capture temporal dynamics.

⁸⁷ **Data and Code Availability**

⁸⁸ All data and code needed to reproduce the results of this manuscript are available at <https://github.com/ajrominger/EcoEvoNonStat> and a detailed description of the analytical approach is available in the supplement.
⁸⁹

References

- Baldrige E, Harris DJ, Xiao X, White EP (2016) An extensive comparison of species-abundance distribution models. *PeerJ*, **4**, e2823.
- Blonder B, Nogués-Bravo D, Borregaard MK, *et al.* (2015) Linking environmental filtering and disequilibrium to biogeography with a community climate framework. *Ecology*, **96**, 972–985.
- Bohmann K, Evans A, Gilbert MTP, *et al.* (2014) Environmental DNA for wildlife biology and biodiversity monitoring. *Trends in Ecology & Evolution*, **29**, 358–367.
- Chesson P (2000) Mechanisms of maintenance of species diversity. *Annu. Rev. Ecol. Syst.*, **31**, 343–366.
- Dodsworth S (2015) Genome skimming for next-generation biodiversity analysis. *Trends in plant science*, **20**, 525–527.
- Etienne RS (2007) A neutral sampling formula for multiple samples and an ‘exact’ test of neutrality. *Ecology letters*, **10**, 608–618.
- Gibson J, Shokralla S, Porter TM, *et al.* (2014) Simultaneous assessment of the macrobiome and microbiome in a bulk sample of tropical arthropods through DNA metasytematics. *Proceedings of the National Academy of Sciences*, **111**, 8007–8012.
- Harte J (2011) *The maximum entropy theory of ecology*. Oxford University Press.
- Hubbell SP (2001) *The unified neutral theory of biodiversity and biogeography*. Princeton University Press.
- Ji Y, Ashton L, Pedley SM, *et al.* (2013) Reliable, verifiable and efficient monitoring of biodiversity via metabarcoding. *Ecology letters*, **16**, 1245–1257.
- Leray M, Knowlton N (2015) DNA barcoding and metabarcoding of standardized samples reveal patterns of marine benthic diversity. *Proceedings of the National Academy of Sciences*, **112**, 2076–2081.
- Linard B, Crampton-Platt A, Gillett CP, Timmermans MJ, Vogler AP (2015) Metagenome skimming of insect specimen pools: Potential for comparative genomics. *Genome biology and evolution*, **7**, 1474–1489.
- Liu S, Wang X, Xie L, *et al.* (2016) Mitochondrial capture enriches mito-DNA 100 fold, enabling PCR-free mitogenomics biodiversity analysis. *Molecular ecology resources*, **16**, 470–479.
- McGill B (2003) Strong and weak tests of macroecological theory. *Oikos*, **102**, 679–685.
- McGill BJ, Etienne RS, Gray JS, *et al.* (2007) Species abundance distributions: Moving beyond single prediction theories to integration within an ecological framework. *Ecol. Lett.*, **10**, 995–1015.
- Neill C, Daufresne T, Jones CG (2009) A competitive coexistence principle? *Oikos*, **118**, 1570–1578.
- Pueyo S, He F, Zillio T (2007) The maximum entropy formalism and the idiosyncratic theory of biodiversity. *Ecology Letters*, **10**, 1017–1028.
- Rabosky DL (2009) Ecological limits and diversification rate: Alternative paradigms to explain the variation in species richness among clades and regions. *Ecol. Lett.*, **12**, 735–743.
- Ricklefs RE (2006) The unified neutral theory of biodiversity: Do the numbers add up? *Ecology*, **87**, 1424–1431.
- Rominger AJ, Goodman KR, Lim JY, *et al.* (2015) Community assembly on isolated islands: Macroecology meets evolution. *Glob. Ecol. Biogeogr.*
- Rominger AJ, Merow C (2017) meteR: An r package for testing the maximum entropy theory of ecology. *Methods in Ecology and Evolution*, **8**, 241–247.
- Sepkoski JJ (1984) A kinetic model of phanerozoic taxonomic diversity. III. Post-Paleozoic families and mass extinctions. *Paleobiology*, **10**, 246–267.

131 Shokralla S, Porter TM, Gibson JF, *et al.* (2015) Massively parallel multiplex DNA sequencing for specimen
132 identification using an illumina MiSeq platform. *Scientific reports*, **5**, 9687.

133 Taberlet P, Coissac E, Pompanon F, Brochmann C, Willerslev E (2012) Towards next-generation biodiversity
134 assessment using DNA metabarcoding. *Molecular Ecology*, **21**, 2045–2050.

135 Tilman D (2004) Niche tradeoffs, neutrality, and community structure: A stochastic theory of resource
136 competition, invasion, and community assembly. *Proc. Natl. Acad. Sci. U. S. A.*, **101**, 10854–10861.

137 Venkataraman A, Bassis CM, Beck JM, *et al.* (2015) Application of a neutral community model to assess
138 structuring of the human lung microbiome. *MBio*, **6**.

139 Wallington TJ, Hobbs RJ, Moore SA (2005) Implications of current ecological thinking for biodiversity
140 conservation: A review of the salient issues. *Ecol. Soc.*, **10**.

141 White EP, Thibault KM, Xiao X (2012) Characterizing species abundance distributions across taxa and
142 ecosystems using a simple maximum entropy model. *Ecology*, **93**, 1772–1778.

143 Zhou X, Li Y, Liu S, *et al.* (2013) Ultra-deep sequencing enables high-fidelity recovery of biodiversity for
144 bulk arthropod samples without PCR amplification. *Gigascience*, **2**, 4.