Community context determines competition / vs. facilitation trade-offs in pollination systems

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3 Abstract

Plant-pollinator interactions are commonly viewed as mutualistic interactions. When plant species share pollinators, however, interspecific pollen transfer occurs and plants can start competing compete for pollination. The extent of competition for pollination and its effects in the on pollination 27 service are not well understood at the community levelwhere; many species interact simultaneously and intertwined factors like multiple factors that might play a role (abundance, visitation, and traitsmight play a role traits) 30 are often intertwined. By examining comprehensive empirical data, we confirm that competition for pollination is pervasive and is only partially influenced by the number of shared pollinators. Furthermore, the factors that most 33 influence the pollination service the most (abundance and visit effectiveness) also introduce a trade-off between the absolute amount of conspecific pollen 35 received and the amount relative to heterospecific pollen. Most importantly, 36 the way ways plants balance these trade-offs are dependent depend on the community context, as most species showed flexibility on their strategy in the strategy they used to cope with competition for pollination. Achieving 39 a better understanding of plant-pollination plant-pollinator communities will require seeing them as both comprising mutualistic and competitive communities interactions.

Introduction

Animal pollination systems play plays a disproportionally important role in food production and maintenance of global biodiversity (Bascompte & Jordano 2007; Klein et al. 2007; Ollerton et al. 2011). Alongside the direct benefits of mutualisms between plants and their pollinators, biodiversity 47 can be supported as a result of the inter-species "facilitation"—the indirect positive feedback loops that exist between plant species that share pollinators or pollinators that share plants (Moeller 2004; Ghazoul 2006; 50 Molina-Montenegro et al. 2008; Liao et al. 2011). This inter-species facilitation is able to promote species coexistence by offsetting the effects of direct resource competition (Stachowicz 2001). Indeed, studies at the core of our current theoretical knowledge of mutualisms predict that the maximum number of coexisting species is achieved when the number of shared mutualistic partners is maximised in a community (Bastolla et al. 2009). However, in natural communities, plants rarely share all available pollinators. One explanation is that plants are effectively competing for mutualistic partners because there is a trade-off trade-off between the benefits gained from maximising the number of partners and the costs of sharing them with other plant species (Waser 1978). It has been clear for a long time that when competition for pollination is strong enough, these costs can make it pollination detrimental to plants' fitness (shown both experimentally and theoretically; Robertson 1895; Lewis 1961; Levin & Anderson 1970). In fact, it can be sufficient to drive ecological differentiation (Caruso 2000; for example by driving the evolution of both traits and reproductive strategies; Mitchell et al. 2009). He-However, it is less clear, however, how exactly ecological differentiation how exactly this ecological differentiation may, in turn, affect the trade-off between facilitation and competition that is involved in the favour (or hinder) the pollination service.

There are two main mechanisms through which competition for pollinators can affect plant fertilization (Morales & Traveset 2008). The first is by changes in intraspecific pollen transfer. This happens, for example, when plants with distinctively more attractive flowers might reduce the number of visits to those less of attractive neighbouring plants, and hence reduce the amount of conspecific pollen deposited by animals (Yang et al. 2011). The second is intraspecific via interspecific pollen transfer and occurs when plants share pollination partners. In that case, even receiving a visit might not necessarily translate into fertilization (Campbell & Motten 1985) because a focal plant might receive heterospecific pollen or because pollen from the 81 focal plant might be lost to different species. Generally speaking, provided pollen is viable and compatible (de Jong et al. 1992; Dafni & Firmage 2000; Ramsey & Vaughton 2000), the higher the amount quantity of conspecific pollen (both in absolute terms and and its purity (relative to heterospecific pollen), the better the pollination service received by the focal plant. These By definition, these two mechanisms of competition, by definition, occur 87 at the community scale. However, with few exceptions (Ratheke 1988; Lopezaraiza-Mikel et al. 2007; Hegland et al. 2009; Aizen & Rovere 2010; Tur et al. 2016), most of what we know about the deposition of conspecific and heterospecific pollen—and its relationship to competition vs. facilitation 91 in pollination systems is them and their relationship with key ecological factors is based on studies with two plant species. At-That is partly so because, at the community scale, the factors that determine the patterns of conspecific and heterospecific pollen deposition are tightly intertwined, operate simultaneously, and may lead to emergent phenomena not observed at smaller scales (Flanagan et al. 2011). For instance, recent empirical evidence suggests that plants with flowering traits that are "original" relative to others in the community generally have fewer interaction partners (Coux et al. 2016). This evidence is aligned with the

notion that a species which that interacts with few species does so strongly with each of them whereas a species that interacts with a large number 102 of species does so comparatively weakly (Bascompte et al. 2006; Vázquez 103 2007; Thébault & Fontaine 2008). If evolutionary specialisation 104 occurs by changing traits to focus on fewer but better partners (Caruso 2000), we should expect a reduction of competition for pollinators in species 106 with "original" traits and an increase of competition in species with a 107 large number of interaction partners (Gibson et al. 2012; Carvalheiro et al. 2014). Alternatively, it might also be the case that density abundance 109 (for example in terms of flower or pollen counts) is the dominant force 110 driving pollen transfer (Seifan et al. 2014). Abundant plant species might experience a dilution of available pollinators (Feinsinger 1987; Feldman et 112 al. 2004) but might also receive more effective visits by capitalising on a 113 larger share of both visits and the pollen carried by pollinators. In this 114 case, a potential reduction in the absolute amount of conspecific pollen 115 received could be compensated by an increase in the amount of conspecific 116 pollen relative to heterospecific pollen. Altogether, it is clear that some these 117 ecological factors can influence pollen deposition and that these factors, in the specific context of pollination, could be used to roughly define a 119 species realised niche in their community. However, the intertwined nature 120 of these factors makes it hard to determine whether they favour or hinder the pollination service indeed shape pollen deposition at the community level. 122 Moreover However, their relative importance has not yet been empirically 123 evaluated in natural plant communities and how exactly they can minimise competition for pollination—or equivalently, maximise facilitation—in plant 125 populations (across species and communities) has not been determined. 126 Here, we investigate the balance between competition for pollination and 127 facilitation at the community scale using empirical data from eleven plant-128 pollinator communities in the Argentinian Pampas. First, we estimate the

extent to which the net outcome of plants sharing pollinators is facilitation 130 or competition. If facilitation is the primary outcome, we could expect 131 two things. On the one hand, we could expect an overall positive relation-132 ship between the amount of conspecific pollen deposited in stigmas and the 133 amount of heterospecific pollen (Tur et al. 2016). The higher the deposition of conspecific pollen relative to heterospecific, the larger the slope 135 of that relationship. On the other, we could also expect a larger amount 136 of conspecific pollen deposited in stigmas when flowers are open to animal pollinators than what is deposited due to self-pollination. this relationship. 138 Second, we investigate the relative contribution that four ecological fac-139 tors have on to the pollination service. Specifically, we hypothesize that competition, measured as the quantity the quantity and purity of conspecific 141 pollen deposition and its purity (relative to heterospecific pollen), should 142 increase should decrease for plants that share many pollination partners. 143 However, we also hypothesize that other factors like the plant's functional 144 originality, its relative floral abundance, and the visit effectiveness should 145 have the potential to compensate for this increase in competition. Third and 146 finally, we examine how much does the community context influences the way plants use these factors to minimise competition.—/maximise facilitation. 148 We use these four ecological factors as a proxy of the realised pollination 149 niche of a plant species in their community. Although some plant species might be widespread across multiple locations, competition for pollinators 151 occurs at the community level and hence we predict the competition for 152 pollination niche to that pollination niche will be influenced by the local context. If plant strategy is not flexible across communities the communities 154 in our study, we could expect niches to be more similar within species than 155 across species in the study. On the contrary, if a species' strategy is flexible, 156 we should expect plants to be able to occupy different niches in each commu-157 nity such that competition could intraspecific competition could ultimately 158 be minimised regardless of the context.

$_{160}$ Methods

We collected data from eleven co-flowering plant communities and their pollinators in three locations in the Argentinian Pampas. These locations are 162 distributed across a precipitation gradient with a humid temperate climate 163 of the eastern Flooding Pampas becoming drier as one moves inland and 164 westward towards the Pampean Grasslands. This precipitation gradient generated a natural gradient of network architecture which allowed us to 166 examine the generality of our findings. In each location, we sampled two 167 restored and two agricultural fragments, except in one located in the Flooding 168 Pampas, where we were only able to sample one restored fragment due to 169 the lack of available sites. 170

Competition vs. facilitation

Our first objective was to identify whether the net outcome of plants sharing pollinators at the community level is facilitation or competition. We use two 173 complementary lines of evidence. The first line can be found via To do that 174 we examine the relationship between heterospecific and conspecific pollen 175 deposited in plant stigmas (Tur et al. 2016). This is because heterospecific 176 pollen can be seen as an unavoidable cost that plants must pay for receiving 177 conspecific pollen from shared pollinators. A negative relationship (when 178 shared pollinators increase heterospecific pollen loads and decrease conspecific 179 loads) may indicate that competition for pollination is strong. In contrast, 180 a positive relationship is what would be observed when facilitation is the 181 dominant outcome. 182 The second line of evidence can be obtained by examining the difference 183 between the amount of conspecific pollen deposited in flower stigmas with 184 and without animal-mediated pollination. Self-pollination can be favourable 185 to plant reproduction when competition for pollination is strong because it provides insurance against poor pollination service (Kalisz & Vogler 2003).

If competition is sufficiently strong, then the amount of conspecific pollen
deposited in flower stigmas when animal pollination is prevented should be
higher than when animal pollination is allowed. In contrast, the opposite
pattern may indicate that plants benefit from having pollen delivered
directly to the stigma by animals.

Data collection

In each of the studied communities, we quantified pollen deposition in a subset 194 of plant species between December 2010 and February 2011. This subset 195 comprised between three and nine common insect-pollinated (entomophilous) 196 plant species that were flowering during the sampling period (see Figure 1a). 197 We Based on data from previous years (Marrero et al. 2014), we chose plant 198 species such that they cover a wide range on a specialization-generalization 199 gradient as well as a wide range of abundances. In each of the selected 200 plants, we removed all flowers leaving only buds that were expected to go 201 into florescence on the next day. Approximately a quarter of these buds were bagged to prevent animal pollination. Two days after inflorescence, we 203 collected all remaining flowers '-and counted the number of conspecific and 204 heterospecific pollen grains in their pistils. More details can be found in 205 Marrero et al. (2016). 206

o7 Data analysis

To evaluate the relationship between heterospecific and conspecific pollen,
we used a generalised linear mixed model (GLMM) with a Poisson error
distribution (following Tur et al. 2016). In this model, the number of
conspecific pollen grains per flower was the response variable and the number of heterospecific grains the predictor (the variable (the log number of

stigmas sampled was added as an offset term). The slope and intercept of this relationship were allowed to differ between plant species within each 214 community populations (plant species × community), which were modelled 215 as a random effect. Models were fitted using the function glmer from 216 the R package lme4 1.1-19 (Bates et al. 2015). A positive slope for a plant-community combination might indicate plant population indicates 218 facilitation while a negative slope might indicate competition (Tur et al. 219 2016). 220 To compare the conspecific pollen deposited with and without animal-mediated 221 pollination, we followed the same approach as above. In this case, however, 222 the predictor was the treatment (whether the flower was bagged or unbagged) instead of the number of heterospecific grains. 224

Factors affecting quantity and purity of pollination service

Our second objective was to investigate the relative contribution of different 226 factors that describe the plant competition of pollination niche to the 227 pollination service that different ecological factors have on the mechanisms 228 of pollinator-mediated competition. Generally speaking, in the context of 229 competition for pollination, we expect that a any factor that increases the 230 amount of conspecific pollen deposited in stigmas, both in quantity and 231 purity relative to heterospecific pollen, also has a positive effect on the 232 pollination service. Specifically, we investigated the effect of (i) a plant's 233 number of shared pollinator species, (ii) a plants plant's abundance relative 234 to the rest of the community, (iii) the mean visit effectiveness—a metric that combines the share of pollen that a plant species is able to place on 236 each of their amount and type of pollen carried by floral visitors and the 237 number of visits it receives from them, and (iv) the plant's functional 238 originality (Laliberté & Legendre 2010). See Data Analysis section below for more details on these four variables factors.

241 Data collection

To obtain the number of shared pollinators for each species, we collected data to construct qualitative and quantitative pollination networks. Qualitative 243 networks were constructed based on two-hour observations of floral visits in each fragment. Quantitative networks were constructed using two 50 m 245 randomly located transects in each fragment. We counted and collected all floral visitors found in a 2 m wide strip while walking at a pace of 10 m per minute (Memmott 1999; Marrero et al. 2014). We visited the transects 248 each month between November 2010 and March 2011. To obtain plant 249 floral abundance, we counted all units of floral attraction found during an 250 independent sampling of the same transects used to construct the quantitative 251 visitation networks. To estimate visit effectiveness, we need to construct 252 pollen transfer networks in addition to the visitation networks. To do this, 253 we examined the pollen loads present on the floral visitors collected (Marrero et al. 2017). When the pollen count on an individual animal was estimated 255 to be less than 2,000 grains, we identified every grain to the species level when possible and to pollen complexes when it was not. When the pollen count was above 2,000 grains, we classified approx. approximately 50% of 258 pollen and total pollen counts were extrapolated (Bosch et al. 2009). Finally, 259 we also recorded morphological traits that relate to plant type (herb, shrub, 260 climber), life cycle (annual, perennial), flower colouration, phenology, and 261 whether the species is native in the study region. More details can be found 262 in Marrero et al. (2014 and 2017). 263

4 Data analysis

To investigate the impact of ecological factors on pollination service services, we used two sets of linear mixed models (LMM) with bootstrap resampling. The response variables for these model sets were the number of conspecific

and heterospecific pollen grains deposited per stigma in flowers open to 268 animal-mediated pollination. We used LMMs in which pollen loads were 269 log-transformed (Figure??) because these models offered a better fit than 270 a GLMM equivalent GLMMs with Poisson (or quasipoisson quasi-Poisson) 271 error structure. Models were fitted using the R package nlme 3.1-131 (Pinheiro et al. 2018). 273 Because the amount of deposited pollen can vary widely across species, and 274 potentially also across communities, we evaluated two possible structures for 275 the random effects: one that includes a random intercept for plant species, 276 and one that treats measures from species across different communities 277 independently. We selected the best random structure by comparing the 278 median Akaike Information Criterion for small samples (AICc). 279 As fixed predictors in the models, we included the four ecological variables 280 factors described above. Specifically, we calculated the number of shared 281 pollinators for each plant species by pooling data from the qualitative and 282 quantitative pollination networks. To calculate the plants' relative floral 283 abundance in their community, we simply aggregated floral counts for each 284 species. We then calculated the mean visit effectiveness . We define the 285

$$e_{ij} = \frac{v_{ij}}{v_i} \frac{p_{ji}}{p_j}$$

effectiveness of a visit by of pollinator species i to plant species j as

where v_{ij} is the observed number of visits by i to j, p_{ji} is the number of pollen grains from j attached to i, v_i is the total number of visits performed by i, and p_j is the total number of grains carried by j. Because We log-transformed the number of shared pollinators, floral abundance, visitation, and pollen production roughly followed a power-law distribution, we log-transformed these three predictors and visit effectiveness before including them in the model.

Finally, functional originality is defined as the distance of a species from 294 the community trait average—the centroid of functional space of the 295 community (Laliberté & Legendre 2010; Coux et al. 2016). To include 296 phenological variation, we treated plant floral abundance in each 297 of the survey months (November to March) as a "trait" in our analysis. To account for the non-independence of floral counts and weight all traits 299 equally, we assigned the abundances a weight of 1/5 to these abundances 300 (one for each month) compared to other functional traits. We scaled all traits prior to calculating the centroid of the functional space and calculated 302 the species-specific functional coordinates using the R package FD 1.0-12 303 (Laliberté et al. 2014). Finally, to facilitate comparison across the four continuous explanatory variables in our models, we rescaled them all we 305 scaled all four factors to have a zero mean and unitary variance. 306 To estimate the coefficients, perform model selection, and quantify the 307 associated uncertainty, we used a combination of multi-model inference and 308 bootstrap resampling with 99 replicates. First, we performed model selection using the AICc and determined the likelihood of each candidate model (a 310 particular combination of predicting variables predictors) by calculating the 311 median \triangle AICc (relative to the most likely model) for each bootstrap sample. As we wanted model coefficients from more likely candidate models to carry 313 more weight in our results, we sampled the coefficients for our variables 314 proportional factors proportionally to the likelihood of its their candidate 315 model. Finally, we used these distributions of the model coefficients to estimate their mean impact on the pollination service (in terms of quantity 317 and purity of conspecific pollen deposition). 318

Flexibility of plant strategies

Our last objective was to tease apart the plant strategies that might reduce competition and, importantly, how the community context influences these

strategies. If community context plays a relatively small role or the species strategy is inflexible, we would expect plants of the same species to fill similar 323 competition for pollination niches across different communities. Alternatively, 324 if the community plays a large role and plants /or plant strategies are flexible, 325 we should be able to observe differences in the realised niche a plant species occupies across communities. First, we To test this, we first used a principal 327 component analysis (PCA) of the four ecological variables factors (number of 328 shared pollinators, floral abundance, visit effectiveness, and trait originality). We scaled variables factors across the whole study to ensure that the PCA 330 space does not change according to the species present in each community. 331 We define a species' niche in a community as its coordinates in PCA space. 332 For each species that was present in two or more communities, we then 333 calculated (i) the median distance between the strategies that the species 334 uses in different communities and (ii) the area of the convex hull defined by 335 these points in the first two principal components (only for species present in 336 three or more communities). We then compared the magnitude of these two 337 metrics to those obtained with 99 Monte Carlo randomizations in which we 338 replaced the strategy of the focal plant species by that of another randomly 339 selected species in the dataset.

$_{\scriptscriptstyle{41}}$ Results

Competition vs. facilitation

We first examined the degree of facilitation/competition in natural communities. Specifically, we examined the slope of the relationship between the quantities of heterospecific and conspecific pollen deposited on stigmas of each species in a community. We found that overall neither facilitation nor Overall, we found that neither net facilitation nor net competition dominates (Table S1). Indeed, we found that the proportion proportions of

species that experienced a statistically significant positive or negative relationship between heterospecific and conspecific pollen than negative was 350 were very similar (35% and 37%, respectively; Figure 1a; Table S2). Most 351 of the variation between plants in a facilitation-competition gradient was 352 driven by differences between species; however, we also observed some important differences within species. For instance, six of the species sampled 354 across multiple communities (Carduus acanthoides, Cirsium vulgare, Cypella 355 herbertii, Diplotaxis tenuifolia, Hirschfeldia incana, and Verbena intermedia) showed evidence that the same species, can experience facilitation (positive 357 relationships) in some communities and competition (negative relationships) 358 in others. Interestingly, although there was a wide variation in the slope 359 of the relationship between heterospecific and conspecific pollen, almost all 360 plants had more conspecific than heterospecific pollen deposited on their 361 stigmas (Figure 16S1). 362 Second, we compared the amount of conspecific pollen deposited on flowers 363 that were open to animal-mediated pollinators with that of those that were closed. Here, we again found large differences between plants but with a 365 clear tendency to find more pollen in open flowers (Table ??). Specifically, 366 66% of plants had more pollen when open and 15% had less pollen when 367 open (the difference was statistically non-significant for the remaining 19%; 368 Figure 1c; Table ??). Variation within species was slightly smaller and 369 only four species (C. acanthoides, C. vulgare, H. incana, and Sisyrinchium 370 platense) had both positive and negative differences.

Factors affecting quantity and purity of pollination service

We then examined the potential roles played in pollen deposition by four ecological variables factors (number of shared pollinators, plant abundance, mean visit effectiveness, and functional originality) play in pollen deposition. We found that our models of pollen deposition had high explanatory power

competition vs. facilitation relationship hetero-conspecific pollen

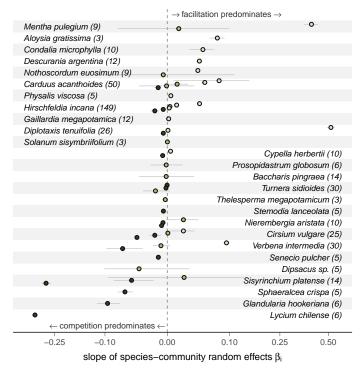


Figure 1: Competition vs. facilitation at the community level. (a) According to the first criterion we used, some species experience more competition for pollination than others. Points indicate the estimated slopes for the relationship between the amount of heterospecific and conspecific pollen deposited per stigma across species in each of their communities. Statistically significant negative slopes (darkest shade) indicate that plants experience net competition for animal-mediated pollination, while significantly positive slopes (lightest shade) indicate that net facilitation is the predominant outcome. The number of plants analysed for each species is shown within brackets next to the species name. (b) Despite the variety of slopes, overall plants had more conspecific than heterospecific pollen deposited in their stigmas. (c) The second criterion we used to distinguish between facilitation and competition shows similar patterns as the first one. Here we compare the amount of conspecific pollen deposited in plants' stigmas when they were open to animal-mediated pollination and when they were closed. Statistically significant negative differences (darkest shade) indicates competition while positive differences (lightest shade) indicates facilitation. In all plots, error Error bars correspond to \pm the standard error (SE).

(the coefficient of determination R² ranged between 0.76 and 0.93) although a large portion of the explanatory power came from the random effects (Table 378 S3). As determined by AICc, the random structure best supported by the 379 data was the one that fit an intercept for each species in each community (as 380 opposed to a common intercept for each species irrespective of the community 381 they belong to to which they belong). This structure was best for both the 382 models of conspecific and heterospecific pollen (Table S4). 383 Of the four variables, factors we considered, we found that a plant's mean 384 visit effectiveness and relative floral abundance were the most important 385 ecological variables at predicting pollen deposition in plant stigmas (Figure 386 2a). Surprisingly, the number of shared pollinators was comparatively 387 unimportant, particularly for models of heterospecific pollen deposition, as 388 it was only ever included in models with relatively large AICc values (Table S5). 390 We found that the relationship between each of the ecological factors and 391 pollen deposition was similar for both conspecific and heterospecific pollen. 392 That is, strategies that were associated with an increase in conspecific pollen 393 were also associated with an increase in heterospecific pollen deposition. 394 Specifically, the plants' mean visit effectiveness had a positive effect on 395 pollen deposition (Figure 2b). However, the effect size was slightly larger 396 for heterospecific than for conspecific pollen. This indicates that, although 397

That is, strategies that were associated with an increase in conspecific pollen were also associated with an increase in heterospecific pollen deposition. Specifically, the plants' mean visit effectiveness had a positive effect on pollen deposition (Figure 2b). However, the effect size was slightly larger for heterospecific than for conspecific pollen. This indicates that, although there is a positive association between visit effectiveness and the quantity of pollen deposition, there is a negative relationship with its purity (Figure 2c). In contrast, a plants' relative abundance had a negative effect on the pollen floral abundance negatively affected its deposition quantity, but the mean difference between the coefficients in the models indicates a positive association with the purity (Figure 2c). The third most important variable factor, functional originality, had a positive, although comparatively smaller, association with both the quantity and purity. Finally, the number

with conspecific and heterospecific pollen, respectively, but these impacts 407 were small when compared to the other variables. 408 Overall, collinearity did not affect our findings qualitatively. Although 409 relative abundance, the number of shared pollinators, and the visit 410 effectiveness were all-factors. Although the ecological factors were positively correlated (Figure S2), the effect each had on conspecific pollen was similar 412 among models that included all or just some of these three explanatory 413 variables collinearity between predictors did not qualitatively affect our 414 findings (Figure S3). One exception was visit-effectiveness, which exhibits 415 a positive association with the relative amount of conspecific pollen under 416 some variable combinations. Nevertheless, these differences were observed 417 only in model specifications with relatively low AICc support.

of shared pollinators had a negative and neutral association associations

Flexibility of plant strategies

We used a PCA of the analysed species to investigate whether plants' re-420 alised niche of competition for pollination pollination niche is similar across 421 communities or whether they are flexible, and therefore a reflection of the 422 community context. The first two PCA components explained 75% of the 423 total variance (Figure 3a). The first component was dominated by changes 424 in-visit effectiveness and relative abundance while the second component was 425 dominated by the number of shared pollinators and the plant's functional 426 originality. When we locate the species that were sampled in more than one 427 community in the first two PCA components (Figure 3b), we observe that the niches of any given species do not tend to be close to each other. Indeed, 429 when we measured the median distance between the plants' coordinates, we 430 found that it was only significantly smaller than that of randomisations for 431 only two of the twelve analysed species species analysed (Figure 3c).

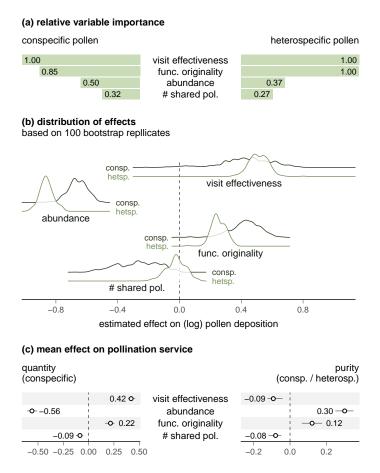
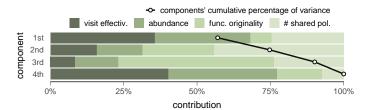
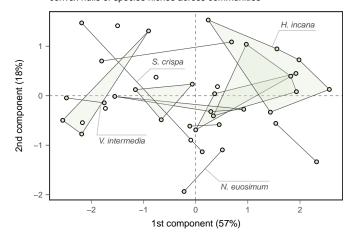


Figure 2: Effect of strategies to minimise competition in the pollination service. (a) The plant's visit effectiveness and relative floral abundance are the most important variables factors determining the deposition of conspecific and heterospecific pollen. Meanwhile, the number of shared pollinators was relatively unimportant generally less important. The graph shows the relative variable importance calculated as the sum of the Akaike weights of the candidate models that included the variable selected factor. (b) The association between ecological variables factors and heterospecific pollen tended to align with their association with conspecific pollen. Visit effectiveness and functional originality had a positive association with pollen deposition, while floral abundance and the number of shared pollinators had a negative association. The plot shows the distribution of the effects (across 99 bootstrap replicates) of the four ecological variables factors for conspecific and heterospecific pollen. (c) The end result of these associations is that only the plants' functional originality has a positive impact on both the quantity and purity of conspecific pollen deposition (relative to heterospecific pollen). The plot shows the model averaged mean effect (\pm SE of 99 bootstrap replicates).

(a) components' variance and variable contributions principal component analysis of ecological variables



(b) plant realised niches in PCA space convex hulls of species niches across communities



(c) flexibility of plant's strategies median distance between plant niches vs. randomisations

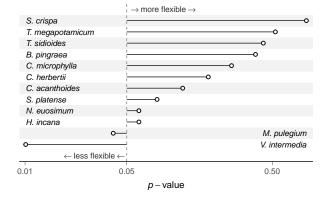


Figure 3: The flexibility of plant strategies. (a) The two first components explain a large proportion of the total variance. (b) When plants that were sampled in more than one community are plotted with in terms of these two components, we observe that their points,—which represent the strategy (the particular combination of ecological variables factors) of a that species in a its community—do not seem to be grouped by plant species. (c) This was confirmed using Monte Carlo randomizations of the median distance between strategies of a plant species. Only one two of the examined species had strategies that were more tightly connected less flexible than what would be expected at random.

3 Discussion

Our results suggest that community context plays a central role in determining the net cost or benefit of sharing pollinators. First, we found that 435 pollinator sharing can lead to both overall facilitative and net facilitative and 436 net competitive effects to plant species in a community. Second, we found 437 that multiple ecological factors can modulate the intensity of competition; however, conspecific and heterospecific pollen deposition are tightly coupled 439 and therefore there is this creates a clear trade-off between the quantity 440 and purity of pollination (Thomson et al. 2019). Third, we found that the way these factors influence competition is strongly shaped by the commu-442 nity context, this. That is, if collectively these factors are used to define a 443 niche of competition for pollination these factors collectively define a species' pollination niche, the same plant species can occupy dramatically different 445 niches depending on the community it belongs to which it belongs. 446 Both of the criteria The criterion we used to identify the effect of animal-447 mediated pollination suggested the same outcome: that neither facilitation 448 nor competition are dominant in pollination communities. A previous study showed that, in diverse pollination communities, sharing pollinators does 450 not necessarily translate to net competition (Tur et al. 2016). The overall 451 extent of facilitation/competition of pollination communities can depend 452 on factors like the adversity of the environment (Callaway et al. 2002; Tur 453 et al. 2016) or how disturbances affect pollinator populations (Stavert et 454 al. 2017). However, why some species experience more competition than 455 others was not understood. Here we Although we were unable to measure 456 the fitness outcomes, here we show that the coexistence of facilitative and 457 competitive effects of animal-mediated pollination can be at least partially 458 explained by the trade-offs plants have to undergo when maximising in the pollination service—both the short and the long term. These trade-offs arise 460 when plants simultaneously maximise the deposition of conspecific pollen

and simultaneously minimising minimise that of heterospecific pollenboth in the short and the long term. 463 In the short term, being a specialist and sharing no pollinators might reduce competition (Muchhala et al. 2010) and hence be preferable, not only due to 465 costs to male fitness (Morales & Traveset 2008; Muchhala & Thomson 2012), 466 but also because, as we show that here, sharing pollinators reduces both the quantity and purity of the conspecific pollen deposited. However, over 468 long periods of time, there could be a risk associated with a specialist plant 469 having no pollinators. Hence, is likely that to ensure long term survival, few 470 pollinators (Ricketts 2004). To ensure long-term survival, it is thus likely 471 that plants also need to balance this risk with the costs of sharing pollinators 472 (Aizen et al. 2012). One possible solution is to share pollinators and have 473 original traits—as we show that trait originality is generally beneficial to pollen deposition and it 's-is commonly thought that species that are further 475 from others in trait space benefit from reduced competition. Yet, there are two possible caveats to this strategy that highlight the interrelatedness of the ecological factors. First, in a mutualism context, it is also possible that 478 trait originality could come at the cost of being less 'apparent' to pollinators 479 (Reverté et al. 2016). Second, the negative relationship between originality 480 and generalism (Carvalheiro et al. 2014) has been shown depend on the 481 plant's to depend on plant abundance (Coux et al. 2016), with generalist 482 species being able to also have original traits only when they are abundant 483 enough to provide a valuable reward to make it worth-visiting worthwhile to pollinators. 485 Visit effectiveness (high pollen and visits) and floral abundance, which 486 were the most important predictors of pollen deposition, involved here, 487 introduced an even more explicit trade-off between gaining conspecific pollen 488 and avoiding heterospecific pollen. On the one hand, receiving Receiving 489 high visitation increases conspecific pollen deposition but increases het-

erospecific pollen deposition to a greater extent—even when the visitors 491 are likely to carry a high proportion of conspecific pollen (Fang & Huang 492 2016). On the other Contrastingly, being abundant reduces the amount of 493 heterospecific pollen deposited but this comes at the expense of a noticeable 494 reduction in the amount of conspecific pollenconspecific pollen deposited and simultaneously reduces heterospecific pollen at a faster rate. Our re-496 sults corroborate the importance that two-species studies have ascribed to 497 visitation and abundance (Feldman et al. 2004; Morales & Traveset 2008; Muñoz & Cavieres 2008). But importantly, but they also suggest that 490 (because visitation, pollen production and abundance are usually correlated; 500 Sargent & Otto 2006) balancing the pros and cons of sharing pollinators at the community level is not trivial and might even be partially responsible 502 for the diversity of plant-pollinator communities (Benadi & Pauw 2018). 503 We observed, as expected, that the effects of pollen deposition can vary 504 widely among species. For instance, while the fitness of some plant species 505 can be hurt even by low amounts of heterospecific pollen, for othersspecies, fitness can instead be limited by the amount of conspecific pollen (Campbell 507 & Motten 1985; Arceo-Gómez et al. 2019). Alternatively, plant species can 508 also differ substantially on in the extent to which self- vs. outcross-pollen differ in their value for fertilization. The difference can be particularly 510 relevant for species that are not self-fertile or those in which self-fertilization 511 is rarely effective due to a temporary separation in the maturation of the 512 sexes (dichogamy). Importantly, here we show we show here that the balances between cost costs and benefits are determined not only by species identity but also on 515 the community plants belong to by the community to which plants belong. 516 Specifically, most plant species appear to be flexible enough to adopt 517 markedly different niches in different communities. Although there are 518 many exceptions, some plant families (Asteraceae for example) are often

generalists in their communities while others (Orchidaceae) are known to 520 be more commonly specialised (Johnson & Steiner 2000). From an evolu-521 tionary perspective, our results suggest that that selection for a particular 522 strategy might say something about the community in which a species has 523 typically inhabited during its evolutionary history. Furthermore, from a more applied perspective, flowering plants are sometimes introduced to 525 attract pollinators on other nearby plants. On the one hand, our results 526 suggest that introduced plants that increase the relative originality of natives (e.g. through distinct phenology) might have positive effects (Gibson et al. 528 2012). On the other, because different niches can lead to different outcomes 520 across communities, our results also highlight the difficulties involved in 530 predicting whether the introduced plant species will facilitate or compete 531 with neighbours (Bartomeus et al. 2008). Other factors that we were unable 532 to measure (e.g. pollinator behaviour and densities or the spatial context) 533 have also been shown to play a role in the outcome of animal-mediated 534 pollination (Cariveau & Norton 2009; Flanagan et al. 2011; Ye et al. 2014; 535 Thomson et al. 2019). Nevertheless, our results indicate that the strategies 536 a plant might use to successfully minimise competition for pollination must be (or maximise facilitation) must be determined relative to other species 538 in the community, rather than an absolute property of the species itself. 539 Overall, pollination communities might not necessarily be the paradigmatic 540 mutualistic communities we often believe. Instead, we confirm that plant competition is pervasive in pollination communities. The potential strategies to minimise competition are likely to involve strong trade-offs in the polli-543 nation service both in the short and long-term. Most remarkably, despite large differences on in the quantity and purity of conspecific pollen deposition among species, our results provide solid evidence that competition for pollination is inherently a community process and the same species can occupy different niches in different communities. Many of the widely used

theoretical models of plant-pollinator communities do not account for the negative effects of sharing pollinators (but see Rohr *et al.* 2014 and similar). However, achieving a better understanding of species coexistence and how plant biodiversity is supported by pollination communities will require seeing them as both mutualistic and competitive communities (Johnson & Bronstein 2019).

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