

Community context determines competition vs. facilitation trade-offs in pollination systems

E. Fernando Cagua¹ (efc29@uclive.ac.nz)

Hugo J. Marrero² (hugomarrero@gmail.com)

Jason M. Tylianakis¹ (jason.tylianakis@canterbury.ac.nz)

Daniel B. Stouffer¹ (daniel.stouffer@canterbury.ac.nz)

¹ Centre for Integrative Ecology, School of Biological Sciences, University of
Canterbury, Private Bag 4800, Christchurch 8041, New Zealand

² Centro de Recursos Naturales Renovables de las Zonas Semiáridas, CON-
ICET, Camino de la Carrindanga Km. 7, 8000 Bahía Blanca, Argentina

Running title: Competition for pollination vs. facilitation

Keywords: abundance, floral visitation, functional originality, interspecific
pollen transfer, plant-pollinator interactions, pollen deposition, pollination
costs and benefits, pollination network, pollination niche, and pollinator
sharing

Type of article: Letter.

Number of words: 153 in abstract; 4,511 in main text.

Number of displays: 3 figures; 0 tables; 0 text boxes.

Number of references: 42

Author for correspondence: E. Fernando Cagua (+64 20 4026 8153).

Data accessibility: Data supporting the results will be accessible in an
appropriate data repository after publication.

Author contributions: EFC wrote the manuscript and performed all
analyses with input from all authors; HJM collected the data; all authors
contributed to the study design.

20 Abstract

21 Plant-pollinator interactions are commonly viewed as mutualistic. However,
22 plants that share pollinators also compete for effective pollination. The
23 extent of this competition for pollination and its effects on pollination
24 service are poorly understood at the community level; many species interact
25 simultaneously and multiple intertwined factors might play a role (abundance,
26 visitation, traits). By examining comprehensive empirical data, we confirm
27 that competition for pollination is pervasive and is only partially influenced
28 by the number of shared pollinators. Furthermore, the factors that most
29 influence the pollination service (abundance and visit effectiveness) also
30 introduce a trade-off between the absolute amount of conspecific pollen
31 received and the amount relative to heterospecific pollen. Most importantly,
32 the ways plants balance these trade-offs depend on the community context,
33 as most species showed flexibility in the strategy they used to cope with
34 competition for pollination. Plant-pollinator mutualisms could be better
35 understood as comprising mutualistic and competitive interactions.

36 Introduction

37 Animal pollination plays a disproportionately important role in food produc-
38 tion and maintenance of global biodiversity (Bascompte & Jordano 2007;
39 Klein *et al.* 2007; Ollerton *et al.* 2011). Alongside the direct benefits of mu-
40 tualisms between plants and their pollinators, biodiversity can be supported
41 as a result of inter-species “facilitation”—the indirect positive feedback loops
42 that exist between plant species that share pollinators or pollinators that
43 share plants (Moeller 2004; Ghazoul 2006; Molina-Montenegro *et al.* 2008;
44 Liao *et al.* 2011). This inter-species facilitation is able to promote species
45 coexistence by offsetting the effects of direct resource competition (Stachow-
46 icz 2001). Indeed, studies at the core of our current theoretical knowledge
47 of mutualisms predict that the maximum number of coexisting species is
48 achieved when the number of shared mutualistic partners is maximised in a
49 community (Bastolla *et al.* 2009).

50 However, in natural communities, plants rarely share all available pollinators.
51 One explanation is that plants are effectively competing for mutualistic part-
52 ners because there is a trade-off between the benefits gained from maximising
53 the number of partners and the costs of sharing them with other plant species
54 (Waser 1978). It has been clear for a long time that when competition for
55 pollination is strong enough, these costs can make pollination detrimental to
56 plants’ fitness (shown both experimentally and theoretically; Robertson 1895;
57 Lewis 1961; Levin & Anderson 1970). In fact, it can be sufficient to drive
58 ecological differentiation (Caruso 2000; for example by driving the evolution
59 of both traits and reproductive strategies; Mitchell *et al.* 2009). However, it
60 is less clear how exactly this ecological differentiation may, in turn, favour
61 (or hinder) the pollination service.

62 There are two main mechanisms through which competition for pollinators
63 can affect plant fertilization (Morales & Traveset 2008). The first is by

64 changes in intraspecific pollen transfer. This happens, for example, when
65 plants with more attractive flowers might reduce the number of visits to
66 those less of attractive neighbouring plants, and hence reduce the amount
67 of *conspecific pollen* deposited by animals (Yang *et al.* 2011). The second
68 is via interspecific pollen transfer and occurs when plants share pollination
69 partners. In that case, even receiving a visit might not necessarily translate
70 into fertilization (Campbell & Motten 1985) because a focal plant might
71 receive *heterospecific pollen* or because pollen from the focal plant might be
72 lost to different species. Generally speaking, provided pollen is viable and
73 compatible (de Jong *et al.* 1992; Dafni & Firmage 2000; Ramsey & Vaughton
74 2000), the higher the quantity of conspecific pollen and its purity (relative
75 to heterospecific pollen), the better the pollination service received by the
76 focal plant. By definition, these two mechanisms of competition occur at the
77 community scale. However, with few exceptions (Aizen & Rovere 2010; Tur
78 *et al.* 2016), most of what we know about them and their relationship with
79 key ecological factors is based on studies with two plant species.

80 That is partly so because, at the community scale, the factors that determine
81 the patterns of pollen deposition are tightly intertwined, operate simultane-
82 ously, and may lead to emergent phenomena not observed at smaller scales
83 (Flanagan *et al.* 2011). For instance, recent empirical evidence suggests
84 that plants with flowering traits that are “original” relative to others in the
85 community generally have fewer interaction partners (Coux *et al.* 2016).
86 This evidence is aligned with the notion that a species that interacts with few
87 species does so strongly with each of them whereas a species that interacts
88 with a large number of species does so comparatively weakly (Bascompte *et*
89 *al.* 2006; Vázquez *et al.* 2007; Thébault & Fontaine 2008). If evolutionary
90 specialisation occurs by changing traits to focus on fewer but better partners
91 (Caruso 2000), we should expect a reduction of competition for pollinators in
92 species with “original” traits and an increase of competition in species with

93 a large number of interaction partners (Gibson *et al.* 2012; Carvalheiro *et al.*
94 2014). Alternatively, it might also be the case that abundance (for example in
95 terms of flower or pollen counts) is the dominant force driving pollen transfer
96 (Seifan *et al.* 2014). Abundant plant species might experience a dilution of
97 available pollinators (Feinsinger 1987; Feldman *et al.* 2004) but might also
98 receive more effective visits by capitalising on a larger share of both visits and
99 the pollen carried by pollinators. In this case, a potential reduction in the
100 absolute amount of conspecific pollen received could be compensated by an
101 increase in the amount of conspecific pollen relative to heterospecific pollen.
102 Altogether, it is clear that these ecological factors can indeed shape pollen
103 deposition at the community level. However, their relative importance and
104 how exactly they can minimise competition for pollination—or equivalently,
105 maximise facilitation—in plant populations (across species and communities)
106 has not been determined.

107 Here, we investigate the balance between competition for pollination and
108 facilitation at the community scale using empirical data from eleven plant-
109 pollinator communities in the Argentinian Pampas. First, we estimate the
110 extent to which the net outcome of plants sharing pollinators is facilitation or
111 competition. If facilitation is the primary outcome, we could expect an overall
112 positive relationship between the amount of conspecific pollen deposited in
113 stigmas and the amount of heterospecific pollen (Tur *et al.* 2016). The higher
114 the deposition of conspecific pollen relative to heterospecific, the larger the
115 slope of this relationship. Second, we investigate the relative contribution
116 that four ecological factors have to the pollination service. Specifically, we
117 hypothesize that the quantity and purity of conspecific pollen deposition
118 should decrease for plants that share many pollination partners. However, we
119 also hypothesize that other factors like the plant’s functional originality, its
120 relative floral abundance, and the visit effectiveness should have the potential
121 to compensate for this increase in competition. Third and finally, we examine

122 how much the community context influences the way plants use these factors
123 to minimise competition/maximise facilitation. We use these four ecological
124 factors as a proxy of the realised pollination niche of a plant species in
125 their community. Although some plant species might be widespread across
126 multiple locations, competition for pollinators occurs at the community level
127 and hence we predict that pollination niche will be influenced by the local
128 context. If plant strategy is not flexible across the communities in our study,
129 we could expect niches to be more similar within species than across species.
130 On the contrary, if a species' strategy is flexible, we should expect plants to
131 be able to occupy different niches in each community such that intraspecific
132 competition could ultimately be minimised regardless of the context.

133 **Methods**

134 We collected data from eleven co-flowering plant communities and their
135 pollinators in three locations in the Argentinian Pampas. In each location,
136 we sampled two restored and two agricultural fragments, except in one
137 located in the Flooding Pampas, where we were only able to sample one
138 restored fragment due to the lack of available sites.

139 **Competition vs. facilitation**

140 Our first objective was to identify whether the net outcome of plants sharing
141 pollinators at the community level is facilitation or competition. To do that
142 we examine the relationship between heterospecific and conspecific pollen
143 deposited in plant stigmas (Tur *et al.* 2016). This is because heterospecific
144 pollen can be seen as an unavoidable cost that plants must pay for receiving
145 conspecific pollen from shared pollinators. A negative relationship (when
146 shared pollinators increase heterospecific pollen loads and decrease conspecific
147 loads) may indicate that competition for pollination is strong. In contrast,

148 a positive relationship is what would be observed when facilitation is the
149 dominant outcome.

150 **Data collection**

151 In each of the studied communities, we quantified pollen deposition in a subset
152 of plant species between December 2010 and February 2011. This subset
153 comprised between three and nine common insect-pollinated (entomophilous)
154 plant species that were flowering during the sampling period (see Fig. 1a).
155 Based on data from previous years (Marrero *et al.* 2014), we chose plant
156 species such that they cover a wide range on a specialization-generalization
157 gradient as well as a wide range of abundances. In each of the selected
158 plants, we removed all flowers leaving only buds that were expected to go
159 into florescence on the next day. Two days after flowering, we collected all
160 remaining flowers and counted the number of conspecific and heterospecific
161 pollen grains in their pistils. More details can be found in Marrero *et al.*
162 (2016).

163 **Data analysis**

164 To evaluate the relationship between heterospecific and conspecific pollen,
165 we used a generalised linear mixed model (GLMM) with a Poisson error
166 distribution (following Tur *et al.* 2016). In this model, the number of
167 conspecific pollen grains per flower was the response variable and the number
168 of heterospecific grains the predictor variable (the log number of stigmas
169 sampled was added as an offset term). The slope and intercept of this
170 relationship were allowed to differ between populations (plant species \times
171 community), which were modelled as a random effect. Models were fitted
172 using the function `glmer` from the R package `lme4` 1.1-19 (Bates *et al.*
173 2015). A positive slope for a plant population indicates facilitation while a

174 negative slope competition (Tur *et al.* 2016).

175 **Factors affecting quantity and purity of pollination service**

176 Our second objective was to investigate the relative contribution that different
177 ecological factors have on the mechanisms of pollinator-mediated competition.
178 Generally speaking, we expect that any factor that increases the amount
179 of conspecific pollen deposited in stigmas, both in quantity and purity
180 relative to heterospecific pollen, also has a positive effect on the pollination
181 service. Specifically, we investigated the effect of (*i*) a plant's number of
182 shared pollinator species, (*ii*) a plant's abundance relative to the rest of the
183 community, (*iii*) the mean visit effectiveness—a metric that combines the
184 amount and type of pollen carried by floral visitors and the number of visits
185 it receives from them, and (*iv*) the plant's functional originality (Laliberté &
186 Legendre 2010). See *Data Analysis* section below for more details on these
187 four factors.

188 **Data collection**

189 To obtain the number of shared pollinators for each species, we collected data
190 to construct qualitative and quantitative pollination networks. Qualitative
191 networks were constructed based on ten-hour observations of floral visits
192 in each fragment. Quantitative networks were constructed using two 50 m
193 randomly located transects in each fragment. We counted and collected all
194 floral visitors found in a 2 m wide strip while walking at a pace of 10 m per
195 minute (Memmott 1999; Marrero *et al.* 2014). We visited the transects each
196 month between November 2010 and March 2011. To obtain floral abundance,
197 we counted all units of floral attraction found during an independent sampling
198 of the same transects used to construct the quantitative visitation networks.
199 To estimate visit effectiveness, we need to construct pollen transfer networks

200 in addition to the visitation networks. To do this, we examined the pollen
201 loads present on the floral visitors collected (Marrero *et al.* 2017). When the
202 pollen count on an individual animal was estimated to be less than 2,000
203 grains, we identified every grain to the species level when possible and to
204 pollen complexes when it was not. When the pollen count was above 2,000
205 grains, we classified approximately 50% of pollen and total pollen counts were
206 extrapolated (Bosch *et al.* 2009). Finally, we also recorded morphological
207 traits that relate to plant type (herb, shrub, climber), life cycle (annual,
208 perennial), flower colouration, phenology, and whether the species is native
209 in the study region. More details can be found in Marrero *et al.* (2014 and
210 2017).

211 **Data analysis**

212 To investigate the impact of ecological factors on pollination services, we
213 used two sets of linear mixed models (LMM) with bootstrap resampling.
214 The response variables for these model sets were the number of conspecific
215 and heterospecific pollen grains deposited per stigma in flowers open to
216 animal-mediated pollination. We used LMMs in which pollen loads were
217 log-transformed because these models offered a better fit than equivalent
218 GLMMs with Poisson (or quasi-Poisson) error structure. Models were fitted
219 using the R package `nlme` 3.1-131 (Pinheiro *et al.* 2018).

220 Because the amount of deposited pollen can vary widely across species, and
221 potentially also across communities, we evaluated two possible structures for
222 the random effects: one that includes a random intercept for plant species,
223 and one that treats measures from species across different communities
224 independently. We selected the best random structure by comparing the
225 median Akaike Information Criterion for small samples (AICc).

226 As fixed predictors in the models, we included the four ecological factors

described above. Specifically, we calculated the number of shared pollinators for each plant species by pooling data from the qualitative and quantitative pollination networks. To calculate the plants' relative floral abundance in their community, we aggregated floral counts for each species. We then calculated the mean visit effectiveness of pollinator species i to plant species j as

$$e_{ij} = \frac{v_{ij} p_{ji}}{v_i p_j}$$

, where v_{ij} is the observed number of visits by i to j , p_{ji} is the number of pollen grains from j attached to i , v_i is the total number of visits performed by i , and p_j is the total number of grains carried by j . We log-transformed the number of shared pollinators, floral abundance, and visit effectiveness before including them in the model.

Finally, functional originality is defined as the distance of a species from the community trait average—the centroid of functional space of the community (Laliberté & Legendre 2010; Coux *et al.* 2016). To include phenological variation, we treated floral abundance in each of the survey months (November to March) as a “trait” in our analysis. To account for the non-independence of floral counts and weight all traits equally, we assigned a weight of 1/5 to these abundances (one for each month). We scaled all traits prior to calculating the centroid of the functional space and calculated the species-specific functional coordinates using the R package FD 1.0-12 (Laliberté *et al.* 2014). Finally, we scaled all four factors to have a zero mean and unitary variance.

To estimate the coefficients, perform model selection, and quantify the associated uncertainty, we used a combination of multi-model inference and bootstrap resampling with 99 replicates. First, we performed model selection using the AICc and determined the likelihood of each candidate model (a particular combination of predictors) by calculating the median ΔAICc (relative to the most likely model) for each bootstrap sample. As we wanted model coefficients from more likely candidate models to carry more weight

255 in our results, we sampled the coefficients for our factors proportionally to
256 the likelihood of their candidate model. Finally, we used these distributions
257 of the model coefficients to estimate their mean impact on the pollination
258 service (in terms of quantity and purity of conspecific pollen deposition).

259 **Flexibility of plant strategies**

260 Our last objective was to tease apart the plant strategies that might reduce
261 competition and, importantly, how the community context influences these
262 strategies. If community context plays a relatively small role or the species
263 strategy is inflexible, we would expect plants of the same species to fill
264 similar pollination niches across different communities. Alternatively, if the
265 community plays a large role and/or plant strategies are flexible, we should
266 be able to observe differences in the realised niche a plant species occupies
267 across communities. To test this, we first used a principal component analysis
268 (PCA) of the four ecological factors (number of shared pollinators, floral
269 abundance, visit effectiveness, and trait originality). We scaled factors across
270 the whole study to ensure that the PCA space does not change according
271 to the species present in each community. We define a species' niche in
272 a community as its coordinates in PCA space. For each species that was
273 present in two or more communities, we then calculated (*i*) the median
274 distance between the strategies that the species uses in different communities
275 and (*ii*) the area of the convex hull defined by these points in the first two
276 principal components (only for species present in three or more communities).
277 We then compared the magnitude of these two metrics to those obtained
278 with 99 Monte Carlo randomizations in which we replaced the strategy of
279 the focal plant species by that of another randomly selected species in the
280 dataset.

281 Results

282 Competition vs. facilitation

283 We first examined the degree of facilitation/competition in natural commu-
284 nities. Specifically, we examined the slope of the relationship between the
285 quantities of heterospecific and conspecific pollen deposited on stigmas of
286 each species in a community. Overall, we found that neither net facilitation
287 nor net competition dominates (Table S1). Indeed, we found that the
288 proportions of species that experienced a statistically significant positive
289 or negative relationship between heterospecific and conspecific pollen were
290 very similar (35% and 37%, respectively; Fig. 1; Table S2). Most of the
291 variation between plants in a facilitation-competition gradient was driven
292 by differences *between* species; however, we also observed some important
293 differences *within* species. For instance, six of the species sampled across
294 multiple communities (*Carduus acanthoides*, *Cirsium vulgare*, *Cypella her-*
295 *bertii*, *Diplotaxis tenuifolia*, *Hirschfeldia incana*, and *Verbena intermedia*)
296 showed evidence that the same species, can experience facilitation (positive
297 relationships) in some communities and competition (negative relationships)
298 in others. Interestingly, although there was wide variation in the slope of the
299 relationship between heterospecific and conspecific pollen, almost all plants
300 had more conspecific than heterospecific pollen deposited on their stigmas
301 (Fig. S1).

302 Factors affecting quantity and purity of pollination service

303 We then examined the potential roles played in pollen deposition by four
304 ecological factors (number of shared pollinators, abundance, mean visit
305 effectiveness, and functional originality). We found that our models of pollen
306 deposition had high explanatory power (the coefficient of determination R^2
307 ranged between 0.76 and 0.93) although a large portion of the explanatory

power came from the random effects (Table S3). As determined by AICc, the random structure best supported by the data was the one that fit an intercept for each species in each community (as opposed to a common intercept for each species irrespective of the community to which they belong). This structure was best for both the models of conspecific and heterospecific pollen (Table S4).

Of the four factors we considered, we found that a plant's mean visit effectiveness and relative floral abundance were the most important at predicting pollen deposition in plant stigmas (Fig. 2a). Surprisingly, the number of shared pollinators was comparatively unimportant, particularly for models of heterospecific pollen deposition, as it was only ever included in models with relatively large AICc values (Table S5).

We found that the relationship between each of the ecological factors and pollen deposition was similar for both conspecific and heterospecific pollen. That is, strategies that were associated with an increase in conspecific pollen were also associated with an increase in heterospecific pollen deposition. Specifically, the plants' mean visit effectiveness had a positive effect on pollen deposition (Fig. 2b). However, the effect size was slightly larger for heterospecific than for conspecific pollen. This indicates that, although there is a positive association between visit effectiveness and the quantity of pollen deposition, there is a negative relationship with its purity (Fig. 2c). In contrast, a plants' relative floral abundance negatively affected its deposition quantity, but the mean difference between the coefficients in the models indicates a positive association with purity (Fig. 2c). The third most important factor, functional originality, had a positive, although comparatively smaller, association with both the quantity and purity. Finally, the number of shared pollinators had negative and neutral associations with conspecific and heterospecific pollen, respectively, but these impacts were small when compared to the other factors. Although the ecological factors

337 were positively correlated (Fig. S2), the collinearity between predictors did
338 not qualitatively affect our findings (Fig. S3).

339 **Flexibility of plant strategies**

340 We used a PCA of the analysed species to investigate whether plants' realised
341 pollination niche is similar across communities or whether they are flexible
342 and therefore a reflection of the community context. The first two PCA
343 components explained 75% of the total variance (Fig. 3a). The first com-
344 ponent was dominated by visit effectiveness and relative abundance while
345 the second component was dominated by the number of shared pollinators
346 and the plant's functional originality. When we locate the species that were
347 sampled in more than one community in the first two PCA components (Fig.
348 3b), we observe that the niches of any given species do not tend to be close
349 to each other. Indeed, when we measured the median distance between the
350 plants' coordinates, we found that it was only significantly smaller than that
351 of randomisations for only two of the twelve species analysed (Fig. 3c).

352 **Discussion**

353 Our results suggest that community context plays a central role in deter-
354 mining the net cost or benefit of sharing pollinators. First, we found that
355 pollinator sharing can lead to both net facilitative and net competitive effects
356 to plant species in a community. Second, we found that multiple ecological
357 factors can modulate the intensity of competition; however, conspecific and
358 heterospecific pollen deposition are tightly coupled and this creates a clear
359 trade-off between the quantity and purity of pollination (Thomson *et al.*
360 2019). Third, we found that the way these factors influence competition
361 is strongly shaped by the community context. That is, if these factors
362 collectively define a species' pollination niche, the same plant species can

363 occupy dramatically different niches depending on the community to which
364 it belongs.

365 The criterion we used to identify the effect of animal-mediated pollination
366 suggested that neither facilitation nor competition are dominant in polli-
367 nation communities. A previous study showed that, in diverse pollination
368 communities, sharing pollinators does not necessarily translate to net com-
369 petition (Tur *et al.* 2016). The overall extent of facilitation/competition
370 of pollination communities can depend on factors like the adversity of the
371 environment (Callaway *et al.* 2002; Tur *et al.* 2016) or how disturbances
372 affect pollinator populations (Stavert *et al.* 2017). However, why some
373 species experience more competition than others was not understood. Al-
374 though we were unable to measure the fitness outcomes, here we show that
375 the coexistence of facilitative and competitive effects of animal-mediated
376 pollination can be at least partially explained by trade-offs in the pollination
377 service—both the short and the long term. These trade-offs arise when plants
378 simultaneously maximise the deposition of conspecific pollen and minimise
379 that of heterospecific pollen.

380 In the short term, being a specialist and sharing no pollinators might reduce
381 competition (Muchhala *et al.* 2010) and hence be preferable, not only due to
382 costs to male fitness (Morales & Traveset 2008; Muchhala & Thomson 2012),
383 but also because, as we show here, sharing pollinators reduces both the
384 quantity and purity of the conspecific pollen deposited. However, over long
385 periods of time, there could be a risk associated with a specialist plant having
386 few pollinators (Ricketts 2004). To ensure long-term survival, it is thus likely
387 that plants also need to balance this risk with the costs of sharing pollinators
388 (Aizen *et al.* 2012). One possible solution is to share pollinators *and* have
389 original traits—as we show that trait originality is generally beneficial to
390 pollen deposition and it is commonly thought that species that are further
391 from others in trait space benefit from reduced competition. Yet, there are

two possible caveats to this strategy that highlight the interrelatedness of the ecological factors. First, in a mutualism context, it is also possible that trait originality could come at the cost of being less ‘apparent’ to pollinators (Reverté *et al.* 2016). Second, the negative relationship between originality and generalism (Carvalho *et al.* 2014) has been shown to depend on plant abundance (Coux *et al.* 2016), with generalist species being able to also have original traits only when they are abundant enough to provide a valuable reward to make visiting worthwhile to pollinators.

Visit effectiveness (high pollen and visits) and floral abundance, which were the most important predictors of pollen deposition here, introduced an even more explicit trade-off between gaining conspecific pollen and avoiding heterospecific pollen. Receiving high visitation increases conspecific pollen deposition but increases heterospecific pollen deposition to a greater extent—even when the visitors are likely to carry a high proportion of conspecific pollen (Fang & Huang 2016). Contrastingly, being abundant reduces the amount of conspecific pollen deposited and simultaneously reduces heterospecific pollen at a faster rate. Our results corroborate the importance that two-species studies have ascribed to visitation and abundance (Feldman *et al.* 2004; Morales & Traveset 2008; Muñoz & Cavieres 2008), but they also suggest that (because visitation, pollen production and abundance are usually correlated; Sargent & Otto 2006) balancing the pros and cons of sharing pollinators at the community level is not trivial and might even be partially responsible for the diversity of plant-pollinator communities (Benadi & Pauw 2018).

We observed, as expected, that the effects of pollen deposition can vary widely among species. For instance, while the fitness of some plant species can be hurt even by low amounts of heterospecific pollen, for others, fitness can instead be limited by the amount of conspecific pollen (Campbell & Motten 1985; Arceo-Gómez *et al.* 2019). Alternatively, plant species can also differ

substantially in the extent to which self- vs. outcross-pollen differ in their value for fertilization. The difference can be particularly relevant for species that are not self-fertile or those in which self-fertilization is rarely effective due to a temporary separation in the maturation of the sexes (dichogamy). Importantly, we show here that the balances between costs and benefits are determined not only by species identity but also by the community to which plants belong. Specifically, most plant species appear to be flexible enough to adopt markedly different niches in different communities. From an evolutionary perspective, our results suggest that selection for a particular strategy might say something about the community in which a species has typically inhabited during its evolutionary history. Furthermore, from a more applied perspective, flowering plants are sometimes introduced to attract pollinators on other nearby plants. On the one hand, our results suggest that introduced plants that increase the relative originality of natives (e.g. through distinct phenology) might have positive effects (Gibson *et al.* 2012). On the other, because different niches can lead to different outcomes across communities, our results also highlight the difficulties involved in predicting whether the introduced plant species will facilitate or compete with neighbours (Bartomeus *et al.* 2008). Other factors that we were unable to measure (e.g. pollinator behaviour and densities or the spatial context) have also been shown to play a role in the outcome of animal-mediated pollination (Cariveau & Norton 2009; Flanagan *et al.* 2011; Ye *et al.* 2014; Thomson *et al.* 2019). Nevertheless, our results indicate that the strategies a plant might use to successfully minimise competition for pollination (or maximise facilitation) must be determined relative to other species in the community, rather than an absolute property of the species itself.

Overall, pollination communities might not necessarily be the paradigmatic mutualistic communities we often believe. Instead, we confirm that plant competition is pervasive in pollination communities. The potential strategies

450 to minimise competition are likely to involve strong trade-offs in the pollina-
451 tion service both in the short and long-term. Most remarkably, despite large
452 differences in the quantity and purity of conspecific pollen deposition among
453 species, our results provide solid evidence that competition for pollination is
454 inherently a community process and the same species can occupy different
455 niches in different communities. Many of the widely used theoretical models
456 of plant-pollinator communities do not account for the negative effects of
457 sharing pollinators (but see Rohr *et al.* 2014 and similar). However, achiev-
458 ing a better understanding of species coexistence and how plant biodiversity
459 is supported by pollination communities will require seeing them as both
460 mutualistic and competitive communities (Johnson & Bronstein 2019).

461 **Acknowledgements**

462 We thank Jamie Stavert, Bernat Bramon Mora, Laís Maia, and Michelle
463 Marraffini for feedback and valuable discussions. We also thank Cátedra de
464 Botánica General, Facultad de Agronomía, Universidad de Buenos Aires,
465 the Agrasar and Bordeu families, and the University of Buenos Aires, for
466 logistical support and permission to conduct this study at estancias Anquilóo,
467 Las Chilcas and San Claudio, respectively. Fieldwork was supported by
468 grants PICT 08-12504 and 0851. EFC acknowledges the support from
469 the University of Canterbury Doctoral Scholarship and a New Zealand
470 International Doctoral Research Scholarship administered by New Zealand
471 Education.DBS and JMT acknowledge the support of Rutherford Discovery
472 Fellowships (RDF-13-UOC-003 and UOC-1002) and the Marsden Fund
473 Council (UOC-1705), administered by the Royal Society of New Zealand Te
474 Apārangi.

References

- Aizen, M.A. & Roever, A.E. (2010). Reproductive interactions mediated by flowering overlap in a temperate hummingbird-plant assemblage. *Oikos*, 119, 696–706.
- Aizen, M.A., Sabatino, M. & Tylianakis, J.M. (2012). Specialization and Rarity Predict Nonrandom Loss of Interactions from Mutualist Networks. *Science*, 335, 1486–1489.
- Arceo-Gómez, G., Kaczorowski, R.L., Patel, C. & Ashman, T.-L. (2019). Interactive effects between donor and recipient species mediate fitness costs of heterospecific pollen receipt in a co-flowering community. *Oecologia*.
- Bartomeus, I., Vilà, M. & Santamaría, L. (2008). Contrasting effects of invasive plants in plant-pollinator networks. *Oecologia*, 155, 761–770.
- Bascompte, J. & Jordano, P. (2007). Plant-Animal Mutualistic Networks: The Architecture of Biodiversity. *Annual Review of Ecology, Evolution, and Systematics*, 38, 567–593.
- Bascompte, J., Jordano, P. & Olesen, J.M. (2006). Asymmetric Coevolutionary Networks Facilitate Biodiversity Maintenance. *Science*, 312, 431–433.
- Bastolla, U., Fortuna, M.A., Pascual-García, A., Ferrera, A., Luque, B. & Bascompte, J. (2009). The architecture of mutualistic networks minimizes competition and increases biodiversity. *Nature*, 458, 1018–1020.
- Bates, D., Mächler, M., Bolker, B. & Walker, S. (2015). Fitting Linear Mixed-Effects Models Using lme4. *Journal of Statistical Software*, 67, 1–48.
- Benadi, G. & Pauw, A. (2018). Frequency dependence of pollinator visitation rates suggests that pollination niches can allow plant species coexistence. *Journal of Ecology*, 106, 1892–1901.
- Bosch, J., Martín González, A.M., Rodrigo, A. & Navarro, D. (2009). Plant-

pollinator networks: Adding the pollinator's perspective. *Ecology Letters*,
12, 409–419.

Callaway, R.M., Brooker, R.W., Choler, P., Kikvidze, Z., Lortie, C.J. &
Michalet, R. *et al.* (2002). Positive interactions among alpine plants increase
with stress. *Nature*, 417, 844–848.

Campbell, D.R. & Motten, A.F. (1985). The Mechanism of Competition for
Pollination between Two Forest Herbs. *Ecology*, 66, 554–563.

Cariveau, D.P. & Norton, A.P. (2009). Spatially contingent interactions
between an exotic and native plant mediated through flower visitors. *Oikos*,
118, 107–114.

Caruso, C.M. (2000). Competition for Pollination Influences Selection on
Floral Traits of *Ipomopsis aggregata*. *Evolution*, 54, 1546–1557.

Carvalho, L.G., Biesmeijer, J.C., Benadi, G., Fründ, J., Stang, M. &
Bartomeus, I. *et al.* (2014). The potential for indirect effects between
co-flowering plants via shared pollinators depends on resource abundance,
accessibility and relatedness. *Ecology Letters*, 17, 1389–1399.

Coux, C., Rader, R., Bartomeus, I. & Tylianakis, J.M. (2016). Linking
species functional roles to their network roles. *Ecology Letters*, 19, 762–770.

Dafni, A. & Firmage, D. (2000). Pollen viability and longevity: Practical,
ecological and evolutionary implications. *Plant systematics and evolution*,
222, 113–132.

Fang, Q. & Huang, S.-Q. (2016). A paradoxical mismatch between in-
terspecific pollinator moves and heterospecific pollen receipt in a natural
community. *Ecology*, 97, 1970–1978.

Feinsinger, P. (1987). Effects of plant species on each others pollination: Is
community structure influenced? *Trends in Ecology & Evolution*, 2, 123–126.

Feldman, T.S., Morris, W.F. & Wilson, W.G. (2004). When can two plant

species facilitate each other's pollination? *Oikos*, 105, 197–207.

Flanagan, R.J., Mitchell, R.J. & Karron, J.D. (2011). Effects of multiple competitors for pollination on bumblebee foraging patterns and *Mimulus Ringens* reproductive success. *Oikos*, 120, 200–207.

Ghazoul, J. (2006). Floral diversity and the facilitation of pollination. *Journal of Ecology*, 94, 295–304.

Gibson, M.R., Richardson, D.M. & Pauw, A. (2012). Can floral traits predict an invasive plant's impact on native plant-pollinator communities? *Journal of Ecology*, 100, 1216–1223.

Johnson, C.A. & Bronstein, J.L. (2019). Coexistence and competitive exclusion in mutualism. *Ecology*, e02708.

de Jong, T.J., Waser, N.M., Price, M.V. & Ring, R.M. (1992). Plant size, geitonogamy and seed set in *Ipomopsis aggregata*. *Oecologia*, 89, 310–315.

Klein, A.-M., Vaissiere, B.E., Cane, J.H., Steffan-Dewenter, I., Cunningham, S.A. & Kremen, C. *et al.* (2007). Importance of pollinators in changing landscapes for world crops. *Proceedings of the Royal Society B: Biological Sciences*, 274, 303–313.

Laliberté, E. & Legendre, P. (2010). A distance-based framework for measuring functional diversity from multiple traits. *Ecology*, 91, 299–305.

Laliberté, E., Legendre, P. & Shipley, B. (2014). *FD: Measuring functional diversity from multiple traits, and other tools for functional ecology*. R package version 1.0-12.

Levin, D.A. & Anderson, W.W. (1970). Competition for pollinators between simultaneously flowering species. *The American Naturalist*, 104, 455–467.

Lewis, H. (1961). Experimental Sympatric Populations of *Clarkia*. *The*

553 *American Naturalist*, 95, 155–168.

554 Liao, K., Gituru, R.W., Guo, Y.-H. & Wang, Q.-F. (2011). The presence of co-
555 flowering species facilitates reproductive success of *Pedicularis Monbeigiana*
556 (Orobanchaceae) through variation in bumble-bee foraging behaviour. *Annals*
557 *of Botany*, 108, 877–884.

558 Marrero, H.J., Medan, D., Zarlavsky, G. & Torretta, J. (2016). Agricultural
559 land management negatively affects pollination service in Pampean agro-
560 ecosystems. *Agriculture, Ecosystems & Environment*, 218, 28–32.

561 Marrero, H.J., Torretta, J. & Medan, D. (2014). Effect of land use intensi-
562 fication on specialization in plant-floral visitor interaction networks in the
563 Pampas of Argentina. *Agriculture, Ecosystems & Environment*, 188, 63–71.

564 Marrero, H.J., Torretta, J.P., Vázquez, D.P., Hodara, K. & Medan, D.
565 (2017). Exotic plants promote pollination niche overlap in an agroecosystem.
566 *Agriculture, Ecosystems & Environment*, 239, 304–309.

567 Memmott, J. (1999). The structure of a plant-pollinator food web. *Ecology*
568 *Letters*, 2, 276–280.

569 Mitchell, R.J., Flanagan, R.J., Brown, B.J., Waser, N.M. & Karron, J.D.
570 (2009). New frontiers in competition for pollination. *Annals of Botany*, 103,
571 1403–1413.

572 Moeller, D.A. (2004). Facilitative interactions among plants via shared
573 pollinators. *Ecology*, 85, 3289–3301.

574 Molina-Montenegro, M.A., Badano, E.I. & Cavieres, L.A. (2008). Positive
575 interactions among plant species for pollinator service: Assessing the “magnet
576 species” concept with invasive species. *Oikos*, 117, 1833–1839.

577 Morales, C.L. & Traveset, A. (2008). Interspecific pollen transfer: Magnitude,
578 prevalence and consequences for plant fitness. *Critical Reviews in Plant*

579 *Sciences*, 27, 221–238.

580 Muchhala, N. & Thomson, J.D. (2012). Interspecific competition in pol-
 581 lination systems: Costs to male fitness via pollen misplacement: Pollen
 582 misplacement. *Functional Ecology*, 26, 476–482.

583 Muchhala, N., Brown, Z., Armbruster, W.S. & Potts, M.D. (2010). Compe-
 584 tition Drives Specialization in Pollination Systems through Costs to Male
 585 Fitness. *The American Naturalist*, 176, 732–743.

586 Muñoz, A.A. & Cavieres, L.A. (2008). The presence of a showy invasive
 587 plant disrupts pollinator service and reproductive output in native alpine
 588 species only at high densities: Invasive impacts on native species pollination.
 589 *Journal of Ecology*, 96, 459–467.

590 Ollerton, J., Winfree, R. & Tarrant, S. (2011). How many flowering plants
 591 are pollinated by animals? *Oikos*, 120, 321–326.

592 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D. & R Core Team. (2018).
 593 *Nlme: Linear and Nonlinear Mixed Effects Models*. R package version
 594 3.1-137.

595 Ramsey, M. & Vaughton, G. (2000). Pollen quality limits seed set in
 596 *Burchardia Umbellata* (Colchicaceae). *American Journal of Botany*, 87,
 597 845–852.

598 Reverté, S., Retana, J., Gómez, J.M. & Bosch, J. (2016). Pollinators show
 599 flower colour preferences but flowers with similar colours do not attract
 600 similar pollinators. *Annals of Botany*, 118, 249–257.

601 Ricketts, T.H. (2004). Tropical Forest Fragments Enhance Pollinator Activity
 602 in Nearby Coffee Crops. *Conservation Biology*, 18, 1262–1271.

603 Robertson, C. (1895). The philosophy of flower seasons, and the phaenological
 604 relations of the entomophilous flora and the anthophilous insect fauna. *The*

605 *American Naturalist*, 29, 97–117.

606 Rohr, R.P., Saavedra, S. & Bascompte, J. (2014). On the structural stability
607 of mutualistic systems. *Science*, 345, 1253497.

608 Sargent, R.D. & Otto, S.P. (2006). The role of local species abundance in
609 the evolution of pollinator attraction in flowering plants. *The American*
610 *Naturalist*, 167, 67–80.

611 Seifan, M., Hoch, E.-M., Hanoteaux, S. & Tielbörger, K. (2014). The
612 outcome of shared pollination services is affected by the density and spatial
613 pattern of an attractive neighbour. *Journal of Ecology*, 102, 953–962.

614 Stachowicz, J.J. (2001). Mutualism, facilitation, and the structure of ecolog-
615 ical communities. *BioScience*, 51, 235–246.

616 Stavert, J.R., Pattemore, D.E., Gaskett, A.C., Beggs, J.R. & Bartomeus,
617 I. (2017). Exotic species enhance response diversity to land-use change but
618 modify functional composition. *Proceedings of the Royal Society B: Biological*
619 *Sciences*, 284, 20170788.

620 Thébault, E. & Fontaine, C. (2008). Does asymmetric specialization differ
621 between mutualistic and trophic networks? *Oikos*, 117, 555–563.

622 Thomson, J.D., Fung, H.F. & Ogilvie, J.E. (2019). Effects of spatial pattern-
623 ing of co-flowering plant species on pollination quantity and purity. *Annals*
624 *of Botany*, 123, 303–310.

625 Tur, C., Sáez, A., Traveset, A. & Aizen, M.A. (2016). Evaluating the effects
626 of pollinator-mediated interactions using pollen transfer networks: Evidence
627 of widespread facilitation in south Andean plant communities. *Ecology*
628 *Letters*, 19, 576–586.

629 Vázquez, D.P., Melián, C.J., Williams, N.M., Blüthgen, N., Krasnov, B.R. &
630 Poulin, R. (2007). Species Abundance and Asymmetric Interaction Strength

631 in Ecological Networks. *Oikos*, 116, 1120–1127.

632 Waser, N.M. (1978). Interspecific pollen transfer and competition between
633 co-occurring plant species. *Oecologia*, 36, 223–236.

634 Yang, S., Ferrari, M.J. & Shea, K. (2011). Pollinator behavior mediates
635 negative interactions between two congeneric invasive plant species. *The*
636 *American Naturalist*, 177, 110–118.

637 Ye, Z.-M., Dai, W.-K., Jin, X.-F., Gituru, R.W., Wang, Q.-F. & Yang, C.-F.
638 (2014). Competition and facilitation among plants for pollination: Can
639 pollinator abundance shift the plant-plant interactions? *Plant Ecology*, 215,
640 3–13.

641 List of Figures

642	1	Competition vs. facilitation at the community level. Points	
643		indicate the estimated slopes for the relationship between	
644		the amount of heterospecific and conspecific pollen deposited	
645		per stigma across species in each of their communities. Sta-	
646		tistically significant negative slopes (darkest shade) indicate	
647		that plants experience net competition for animal-mediated	
648		pollination, while significantly positive slopes (lightest shade)	
649		indicate that net facilitation is the predominant outcome. The	
650		number of plants analysed for each species is shown within	
651		brackets next to the species name. Error bars correspond to	
652		\pm the standard error (SE).	29

653 2 Effect of ecological factors on the pollination service. (a)

654 The plant's visit effectiveness and relative floral abundance

655 are the most important factors determining the deposition of

656 conspecific and heterospecific pollen. Meanwhile, the num-

657 ber of shared pollinators was generally less important. The

658 graph shows the relative importance calculated as the sum

659 of the Akaike weights of the candidate models that included

660 the selected factor. (b) The association between ecological

661 factors and heterospecific pollen tended to align with their

662 association with conspecific pollen. Visit effectiveness and

663 functional originality had a positive association with pollen

664 deposition, while floral abundance and the number of shared

665 pollinators had a negative association. The plot shows the

666 distribution of the effects (across 99 bootstrap replicates) of

667 the four ecological factors for conspecific and heterospecific

668 pollen. (c) The end result of these associations is that only

669 the plants' functional originality has a positive impact on

670 both the quantity and purity of conspecific pollen deposition

671 (relative to heterospecific pollen). The plot shows the model

672 averaged mean effect (\pm SE of 99 bootstrap replicates). . . . 30

673	3	The flexibility of plant strategies. (a) The two first compo-	
674		nents explain a large proportion of the total variance. (b)	
675		When plants that were sampled in more than one community	
676		are plotted in terms of these two components, we observe that	
677		their points, which represent the strategy (the particular com-	
678		bination of ecological factors) of that species in its community,	
679		do not seem to be grouped by plant species. (c) This was	
680		confirmed using Monte Carlo randomizations of the median	
681		distance between strategies of a plant species. Only two of the	
682		examined species had strategies that were less flexible than	
683		would be expected at random.	31

competition vs. facilitation
relationship hetero–conspecific pollen

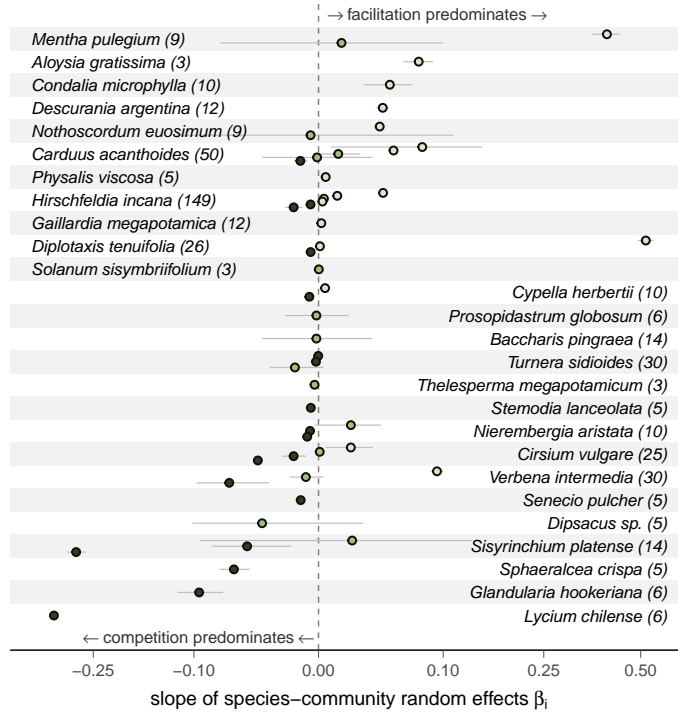


Figure 1: Competition vs. facilitation at the community level. Points indicate the estimated slopes for the relationship between the amount of heterospecific and conspecific pollen deposited per stigma across species in each of their communities. Statistically significant negative slopes (darkest shade) indicate that plants experience net competition for animal-mediated pollination, while significantly positive slopes (lightest shade) indicate that net facilitation is the predominant outcome. The number of plants analysed for each species is shown within brackets next to the species name. Error bars correspond to \pm the standard error (SE).

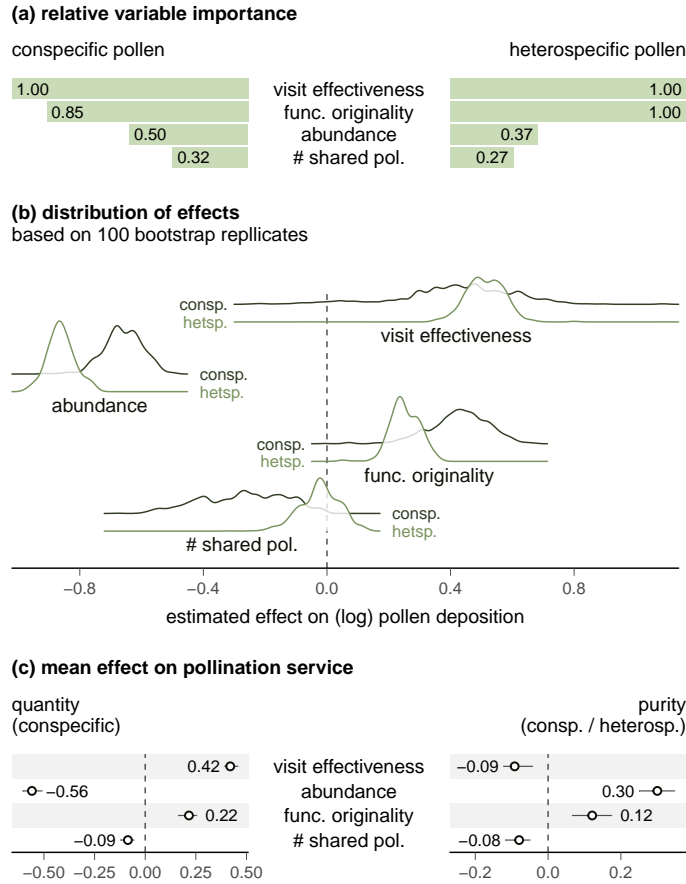


Figure 2: Effect of ecological factors on the pollination service. (a) The plant's visit effectiveness and relative floral abundance are the most important factors determining the deposition of conspecific and heterospecific pollen. Meanwhile, the number of shared pollinators was generally less important. The graph shows the relative importance calculated as the sum of the Akaike weights of the candidate models that included the selected factor. (b) The association between ecological factors and heterospecific pollen tended to align with their association with conspecific pollen. Visit effectiveness and functional originality had a positive association with pollen deposition, while floral abundance and the number of shared pollinators had a negative association. The plot shows the distribution of the effects (across 99 bootstrap replicates) of the four ecological factors for conspecific and heterospecific pollen. (c) The end result of these associations is that only the plants' functional originality has a positive impact on both the quantity and purity of conspecific pollen deposition (relative to heterospecific pollen). The plot shows the model averaged mean effect (\pm SE of 99 bootstrap replicates).

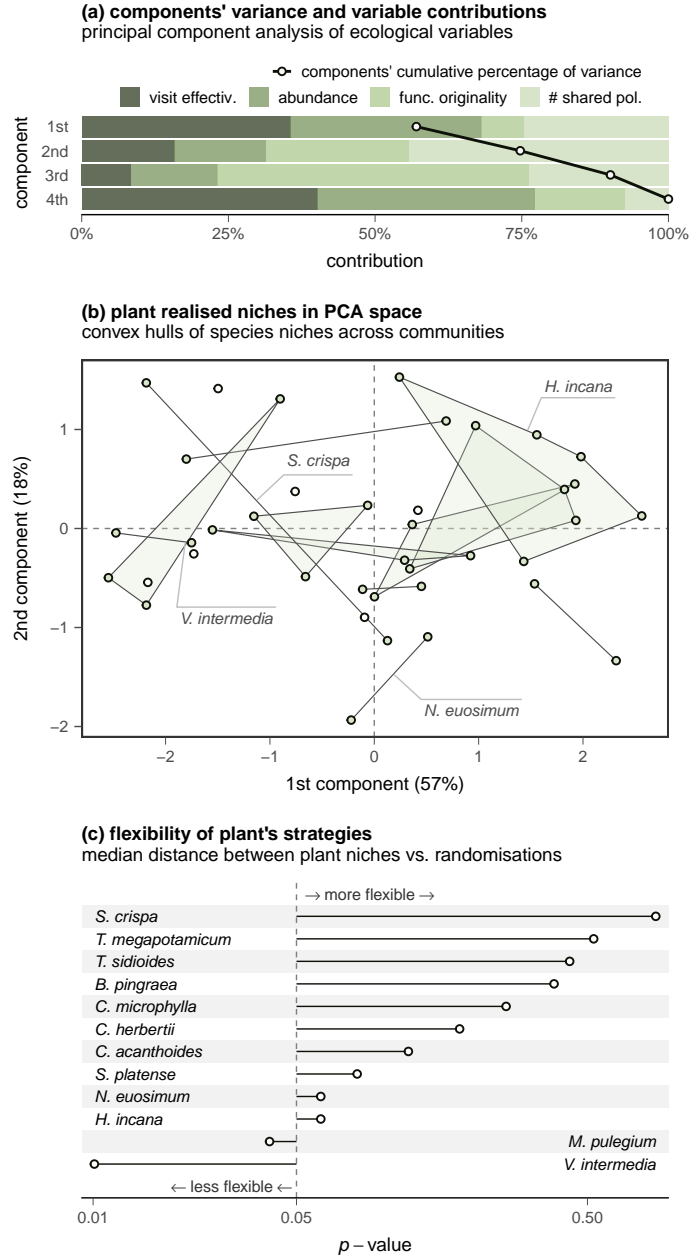


Figure 3: The flexibility of plant strategies. (a) The two first components explain a large proportion of the total variance. (b) When plants that were sampled in more than one community are plotted in terms of these two components, we observe that their points, which represent the strategy (the particular combination of ecological factors) of that species in its community, do not seem to be grouped by plant species. (c) This was confirmed using Monte Carlo randomizations of the median distance between strategies of a plant species. Only two of the examined species had strategies that were less flexible than would be expected at random.