Community context determines competition vs. facilitation trade-offs in pollination systems

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16 Abstract

Plant-pollinator interactions are commonly viewed as mutualistic. However, 17 plants that share pollinators also compete for effective pollination. The 18 extent of this competition for pollination and its effects on pollination 19 service are poorly understood at the community level; many species interact simultaneously and multiple intertwined factors might play a role (abundance, visitation, traits). By examining comprehensive empirical data, we confirm that competition for pollination is pervasive and is only partially influenced by the number of shared pollinators. Furthermore, the factors that most influence the pollination service (abundance and visit effectiveness) also introduce a trade-off between the absolute amount of conspecific pollen received and the amount relative to heterospecific pollen. Most importantly, the ways plants balance these trade-offs depend on the community context, 28 as most species showed flexibility in the strategy they used to cope with 29 competition for pollination. Plant-pollinator mutualisms could be better understood as comprising mutualistic and competitive interactions.

12 Introduction

Animal pollination plays a disproportionally important role in food production and maintenance of global biodiversity [1-3]. Alongside the direct 34 benefits of mutualisms between plants and their pollinators, biodiversity can 35 be supported as a result of inter-species "facilitation"—the indirect positive feedback loops that exist between plant species that share pollinators or pollinators that share plants [4–7]. This inter-species facilitation is able to promote species coexistence by offsetting the effects of direct resource competition [8]. Indeed, studies at the core of our current theoretical knowledge of mutualisms predict that the maximum number of coexisting species is achieved when the number of shared mutualistic partners is maximised in a community [9]. 43 However, in natural communities, plants rarely share all available pollinators. One explanation is that plants are effectively competing for mutualistic partners because there is a trade-off between the benefits gained from maximising the number of partners and the costs of sharing them with other plant species 47 [10]. It has been clear for a long time that when competition for pollination is strong enough, these costs can make pollination detrimental to plants' fitness [11–13]. In fact, it can be sufficient to drive ecological differentiation [14,15]. However, it is less clear how exactly this ecological differentiation may, in turn, favour (or hinder) the pollination service. There are two main mechanisms through which competition for pollinators can affect plant fertilization [16]. The first is by changes in intraspecific pollen transfer. This happens, for example, when plants with more attractive flowers might reduce the number of visits to those less of attractive neighbouring plants, and hence reduce the amount of conspecific pollen deposited by animals [17]. The second is via interspecific pollen transfer and occurs when plants share pollination partners. In that case, even receiving a visit

might not necessarily translate into fertilization [18] because a focal plant might receive heterospecific pollen or because pollen from the focal plant 61 might be lost to different species. Naturally, the precise effects on plant 62 fitness of conspecific and heterospecific pollen deposition depend on the species involved (and are unknown for many plant species). However, there is substantial evidence supporting the link between pollen deposition and seed production; in particular for the detrimental effect of heterospecific 66 pollen deposition on plant fitness [19,20]. All together, provided pollen is 67 viable and compatible [21–23], the higher the quantity of conspecific pollen and its purity (relative to heterospecific pollen), the better the pollination service received by the focal plant. As such, measuring conspecific and 70 heterospecific pollen deposition provides a good indication of the potential 71 levels of facilitation and competition a plant population might experience. By definition, these two mechanisms of competition, intra and interspecific 73 pollen transfer, occur at the community scale. However, with few exceptions 74 [24,25], most of what we know about them and their relationship with key ecological factors is based on studies with two plant species. That is partly so because, at the community scale, the factors that determine the patterns of pollen deposition are tightly intertwined, operate simultaneously, and may 78 lead to emergent phenomena not observed at smaller scales [26]. For instance, 79 recent empirical evidence suggests that plants with flowering traits that are "original" relative to others in the community generally have fewer interaction partners [27]. This evidence is aligned with the notion that a species that 82 interacts with few species does so strongly with each of them whereas a species that interacts with a large number of species does so comparatively 84 weakly [28–30]. If evolutionary specialisation occurs by changing traits to focus on fewer but better partners [15], we should expect a reduction of competition for pollinators in species with "original" traits and an increase of competition in species with a large number of interaction partners [31,32].

Alternatively, it might also be the case that abundance (for example in terms of flower or pollen counts) is the dominant force driving pollen transfer [33]. 90 Abundant plant species might experience a dilution of available pollinators 91 [34,35] but might also receive more effective visits by capitalising on a larger share of both visits and the pollen carried by pollinators. In this case, a potential reduction in the absolute amount of conspecific pollen received could be compensated by an increase in the amount of conspecific pollen relative to heterospecific pollen. Altogether, it is clear that these ecological factors can indeed shape pollen deposition at the community level. However, their relative importance and how exactly they can minimise competition for pollination—or equivalently, maximise facilitation—in plant populations 99 (across species and communities) has not been determined. 100

Here, we investigate the balance between competition for pollination and 101 facilitation at the community scale using empirical data from eleven plant-102 pollinator communities in the Argentinian Pampas. First, we estimate the 103 extent to which the net outcome of plants sharing pollinators is facilitation 104 or competition. If facilitation is the primary outcome, we could expect 105 an overall positive relationship between the amount of conspecific pollen 106 deposited in stigmas and the amount of heterospecific pollen [24]. The higher 107 the deposition of conspecific pollen relative to heterospecific, the larger the 108 slope of this relationship. Second, we investigate the relative contribution 109 that four ecological factors have to the pollination service. Specifically, we hypothesize that the quantity and purity of conspecific pollen deposition 111 should decrease for plants that share many pollination partners. However, we 112 also hypothesize that other factors like the plant's functional originality, its 113 relative floral abundance, and the visit potential should have the potential to 114 compensate for this increase in competition. Third and finally, we examine how much the community context influences the way plants use these factors 116 to minimise competition/maximise facilitation. We use these four ecological 117

factors as a proxy of the realised pollination niche of a plant species in 118 their community. Although some plant species might be widespread across 119 multiple locations, competition for pollinators occurs at the community level 120 and hence we predict that pollination niche will be influenced by the local 121 context. If plant strategy is not flexible across the communities in our study, 122 we could expect niches to be more similar within species than across species. 123 On the contrary, if a species' strategy is flexible, we should expect plants to 124 be able to occupy different niches in each community such that intraspecific 125 competition could ultimately be minimised regardless of the context.

127 Methods

We collected data from eleven co-flowering plant communities and their pollinators in three locations in the Argentinian Pampas. In each location, we sampled two restored and two agricultural fragments, except in one located in the Flooding Pampas, where we were only able to sample one restored fragment due to the lack of available sites.

133 Competition vs. facilitation

Our first objective was to identify whether the net outcome of plants sharing 134 pollinators at the community level is facilitation or competition. To do 135 that we examine the relationship between heterospecific and conspecific 136 pollen deposited in plant stigmas [24]. This is because heterospecific pollen can be seen as an unavoidable cost that plants must pay for receiving 138 conspecific pollen from shared pollinators. A negative relationship (when 139 shared pollinators increase heterospecific pollen loads and decrease conspecific 140 loads) may indicate that competition for pollination is strong. In contrast, 141 a positive relationship is what would be observed when facilitation is the dominant outcome.

144 Data collection

In each of the studied communities, we quantified pollen deposition in a subset 145 of plant species between December 2010 and February 2011. This subset comprised between three and nine common insect-pollinated (entomophilous) plant species that were flowering during the sampling period. Based on 148 data from previous years [36], we chose plant species such that they cover 149 a wide range on a specialization-generalization gradient as well as a wide 150 range of abundances. In each of the selected plants, we removed all flowers 151 leaving only buds that were expected to go into florescence on the next day. 152 Two days after flowering, we collected all remaining flowers and counted the 153 number of conspecific and heterospecific pollen grains in their pistils. More 154 details can be found in [37]. 155

Data analysis

To evaluate the relationship between heterospecific and conspecific pollen, 157 we used a generalised linear mixed model (GLMM) with a Poisson error 158 distribution [24]. In this model, the number of conspecific pollen grains per 159 flower was the response variable and the number of heterospecific grains the 160 predictor variable (the log number of stigmas sampled was added as an offset term). The slope and intercept of this relationship were allowed to differ 162 between populations (plant species × community), which were modelled as 163 a random effect. Models were fitted using the function glmer from the R 164 package lme4 1.1-19 [38]. A positive slope for a plant population indicates 165 facilitation while a negative slope competition [24]. 166

Factors affecting quantity and purity of pollination service

Our second objective was to investigate the relative contribution that different ecological factors have on the mechanisms of pollinator-mediated competition.

Generally speaking, we expect that any factor that increases the amount 170 of conspecific pollen deposited in stigmas, both in quantity and purity 171 relative to heterospecific pollen, also has a positive effect on the pollination 172 service. Specifically, we investigated the effect of (i) a plant's number of shared pollinator species, (ii) a plant's abundance relative to the rest of 174 the community, (iii) the mean visit potential—a metric that combines the 175 amount and type of pollen carried by floral visitors and the number of visits 176 it receives from them, and (iv) the plant's functional originality [39]. See 177 Data Analysis section below for more details on these four factors.

179 Data collection

To obtain the number of shared pollinators for each species, we collected data 180 to construct qualitative and quantitative pollination networks. Qualitative 181 networks were constructed based on ten-hour observations of floral visits 182 in each fragment. Quantitative networks were constructed using two 50 m 183 randomly located transects in each fragment. We counted and collected all 184 floral visitors found in a 2 m wide strip while walking at a pace of 10 m 185 per minute [36,40]. We visited the transects each month between November 186 2010 and March 2011. To obtain floral abundance, we counted all units of 187 floral attraction found during an independent sampling of the same transects 188 used to construct the quantitative visitation networks. To estimate visit 189 potential, we need to construct pollen transfer networks in addition to the 190 visitation networks. To do this, we examined the pollen loads present on 191 the floral visitors collected [41]. When the pollen count on an individual 192 animal was estimated to be less than 2,000 grains, we identified every grain to the species level when possible and to pollen complexes when it was not. When the pollen count was above 2,000 grains, we classified approximately 195 50% of pollen and total pollen counts were extrapolated [42]. Finally, we 196 also recorded morphological traits that relate to plant type (herb, shrub, climber), life cycle (annual, perennial), flower colouration, phenology, and whether the species is native in the study region. More details can be found in Marrero *et al.* (2014 and 2017).

201 Data analysis

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222 as

To investigate the impact of ecological factors on pollination services, we 202 used two sets of linear mixed models (LMM) with bootstrap resampling. 203 The response variables for these model sets were the number of conspecific 204 and heterospecific pollen grains deposited per stigma in flowers open to 205 animal-mediated pollination. We used LMMs in which pollen loads were 206 log-transformed because these models offered a better fit than equivalent 207 GLMMs with Poisson (or quasi-Poisson) error structure. Models were fitted 208 using the R package nlme 3.1-131 [43]. 209 Because the amount of deposited pollen can vary widely across species, and 210 potentially also across communities, we evaluated two possible structures for 211 the random effects: one that includes a random intercept for plant species, and one that treats measures from species across different communities 213 independently. We selected the best random structure by comparing the 214 median Akaike Information Criterion for small samples (AICc). 215 As fixed predictors in the models, we included the four ecological factors 216 described above. Specifically, we calculated the number of shared pollinators 217 for each plant species by pooling data from the qualitative and quantitative 218 pollination networks. To calculate the plants' relative floral abundance in 219 their community, we aggregated floral counts for each species. We then 220

$$o_{ij} = \frac{v_{ij}}{v_i} \frac{p_{ji}}{p_j}$$

calculated the mean visit potential of pollinator species i to plant species j

, where v_{ij} is the observed number of visits by i to j, p_{ji} is the number of

pollen grains from j attached to i, v_i is the total number of visits performed by i, and p_j is the total number of grains carried by j. We log-transformed the number of shared pollinators, floral abundance, and visit potential before including them in the model.

Finally, functional originality is defined as the distance of a species from the 228 community trait average—the centroid of functional space of the community [27,39]. To include phenological variation, we treated floral abundance in 230 each of the survey months (November to March) as a "trait" in our analysis. 231 To account for the non-independence of floral counts and weight all traits 232 equally, we assigned a weight of 1/5 to these abundances (one for each month). 233 We scaled all traits prior to calculating the centroid of the functional space 234 and calculated the species-specific functional coordinates using the R package 235 FD 1.0-12 [44]. Finally, we scaled all four factors to have a zero mean and 236 unitary variance. 237

To estimate the coefficients, perform model selection, and quantify the 238 associated uncertainty, we used a combination of multi-model inference and 239 bootstrap resampling with 99 replicates. First, we performed model selection 240 using the AICc and determined the likelihood of each candidate model (a particular combination of predictors) by calculating the median $\Delta AICc$ (relative to the most likely model) for each bootstrap sample. As we wanted 243 model coefficients from more likely candidate models to carry more weight 244 in our results, we sampled the coefficients for our factors proportionally to 245 the likelihood of their candidate model. Finally, we used these distributions of the model coefficients to estimate their mean impact on the pollination service (in terms of quantity and purity of conspecific pollen deposition).

249 Flexibility of plant strategies

Our last objective was to tease apart the plant strategies that might reduce 250 competition and, importantly, how the community context influences these 251 strategies. If community context plays a relatively small role or the species 252 strategy is inflexible, we would expect plants of the same species to fill 253 similar pollination niches across different communities. Alternatively, if the 254 community plays a large role and/or plant strategies are flexible, we should 255 be able to observe differences in the realised niche a plant species occupies 256 across communities. To test this, we first used a principal component analysis (PCA) of the four ecological factors (number of shared pollinators, floral 258 abundance, visit potential, and trait originality). We scaled factors across 259 the whole study to ensure that the PCA space does not change according 260 to the species present in each community. We define a species' niche in 261 a community as its coordinates in PCA space. For each species that was 262 present in two or more communities, we then calculated (i) the median 263 distance between the strategies that the species uses in different communities 264 and (ii) the area of the convex hull defined by these points in the first two 265 principal components (only for species present in three or more communities). 266 We then compared the magnitude of these two metrics to those obtained with 99 Monte Carlo randomizations in which we replaced the strategy of the focal plant species by that of another randomly selected species in the 269 dataset. 270

$_{71}$ Results

272 Competition vs. facilitation

We first examined the degree of facilitation/competition in natural communities. Specifically, we examined the slope of the relationship between the

quantities of heterospecific and conspecific pollen deposited on stigmas of 275 each species in a community. Overall, we found that neither net facilita-276 tion nor net competition dominates (Table S1). Indeed, we found that the 277 proportions of species that experienced a statistically significant positive or negative relationship between heterospecific and conspecific pollen were 279 very similar (35% and 37%, respectively; Fig. 1; Table S2). Most of the 280 variation between plants in a facilitation-competition gradient was driven 281 by differences between species; however, we also observed some important 282 differences within species. For instance, six of the species sampled across multiple communities (Carduus acanthoides, Cirsium vulgare, Cypella her-284 bertii, Diplotaxis tenuifolia, Hirschfeldia incana, and Verbena intermedia) 285 showed evidence that the same species, can experience facilitation (positive 286 relationships) in some communities and competition (negative relationships) 287 in others. Interestingly, although there was wide variation in the slope of the relationship between heterospecific and conspecific pollen, almost all plants had more conspecific than heterospecific pollen deposited on their stigmas 290 (Fig. S1). 291

292 Factors affecting quantity and purity of pollination service

We then examined the potential roles played in pollen deposition by four 293 ecological factors (number of shared pollinators, abundance, mean visit potential, and functional originality). We found that our models of pollen 295 deposition had high explanatory power (the coefficient of determination R² 296 ranged between 0.76 and 0.93) although a large portion of the explanatory 297 power came from the random effects (Table S3). As determined by AICc, the 298 random structure best supported by the data was the one that fit an intercept for each species in each community (as opposed to a common intercept for 300 each species irrespective of the community to which they belong). This 301 structure was best for both the models of conspecific and heterospecific

relationship hetero-conspecific pollen \rightarrow facilitation predominates \rightarrow Mentha pulegium (9) Aloysia gratissima (3) Condalia microphylla (10) 0 Descurania argentina (12) 0 Nothoscordum euosimum (9) 0 Carduus acanthoides (50) 0 Physalis viscosa (5) Hirschfeldia incana (149) Gaillardia megapotamica (12) þ Diplotaxis tenuifolia (26) Solanum sisymbriifolium (3) Cypella herbertii (10) Prosopidastrum globosum (6) Baccharis pingraea (14) Turnera sidioides (30) Thelesperma megapotamicum (3) • Stemodia lanceolata (5) Nierembergia aristata (10) Cirsium vulgare (25) Verbena intermedia (30) Senecio pulcher (5) Dipsacus sp. (5) Sisyrinchium platense (14) Sphaeralcea crispa (5) Glandularia hookeriana (6) Lycium chilense (6) competition predominates

competition vs. facilitation

-0.25

-0.10

Figure 1: Competition vs. facilitation at the community level. Points indicate the estimated slopes for the relationship between the amount of heterospecific and conspecific pollen deposited per stigma across species in each of their communities. Statistically significant negative slopes (darkest shade) indicate that plants experience net competition for animal-mediated pollination, while significantly positive slopes (lightest shade) indicate that net facilitation is the predominant outcome. The number of plants analysed for each species is shown within brackets next to the species name. Error bars correspond to \pm the standard error (SE).

0.00

slope of species-community random effects β_{i}

0.10

0.25

0.50

303 pollen (Table S4).

Of the four factors we considered, we found that a plant's mean visit potential and relative floral abundance were the most important at predicting pollen deposition in plant stigmas (Fig. 2a). Surprisingly, the number of shared pollinators was comparatively unimportant, particularly for models of heterospecific pollen deposition, as it was only ever included in models with relatively large AICc values (Table S5).

We found that the relationship between each of the ecological factors and 310 pollen deposition was similar for both conspecific and heterospecific pollen. 311 That is, strategies that were associated with an increase in conspecific pollen 312 were also associated with an increase in heterospecific pollen deposition. 313 Specifically, the plants' mean visit potential had a positive effect on pollen 314 deposition (Fig. 2b). However, the effect size was slightly larger for het-315 erospecific than for conspecific pollen. This indicates that, although there 316 is a positive association between visit potential and the quantity of pollen 317 deposition, there is a negative relationship with its purity (Fig. 2c). In 318 contrast, a plants' relative floral abundance negatively affected its deposition 319 quantity, but the mean difference between the coefficients in the models 320 indicates a positive association with purity (Fig. 2c). The third most impor-321 tant factor, functional originality, had a positive, although comparatively 322 smaller, association with both the quantity and purity. Finally, the number 323 of shared pollinators had negative and neutral associations with conspecific 324 and heterospecific pollen, respectively, but these impacts were small when 325 compared to the other factors. Although the ecological factors were positively 326 correlated (Fig. S2), the collinearity between predictors did not qualitatively 327 affect our findings (Fig. S3). 328

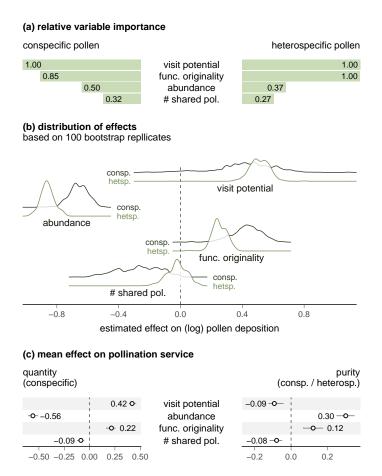


Figure 2: Effect of ecological factors on the pollination service. (a) The plant's visit potential and relative floral abundance are the most important factors determining the deposition of conspecific and heterospecific pollen. Meanwhile, the number of shared pollinators was generally less important. The graph shows the relative importance calculated as the sum of the Akaike weights of the candidate models that included the selected factor. (b) The association between ecological factors and heterospecific pollen tended to align with their association with conspecific pollen. Visit potential and functional originality had a positive association with pollen deposition, while floral abundance and the number of shared pollinators had a negative association. The plot shows the distribution of the effects (across 99 bootstrap replicates) of the four ecological factors for conspecific and heterospecific pollen. (c) The end result of these associations is that only the plants functional originality has a positive impact on both the quantity and purity of conspecific pollen deposition (relative to heterospecific pollen). The plot shows the model averaged mean effect (\pm SE of 99 bootstrap replicates).

Flexibility of plant strategies

We used a PCA of the analysed species to investigate whether plants' re-330 alised pollination niche is similar across communities or whether they are flexible and therefore a reflection of the community context. The first two 332 PCA components explained 75% of the total variance (Fig. 3a). The first 333 component was dominated by visit potential and relative abundance while 334 the second component was dominated by the number of shared pollinators 335 and the plant's functional originality. When we locate the species that were 336 sampled in more than one community in the first two PCA components (Fig. 337 3b), we observe that the niches of any given species do not tend to be close 338 to each other. Indeed, when we measured the median distance between the 339 plants' coordinates, we found that it was only significantly smaller than that 340 of randomisations for only two of the twelve species analysed (Fig. 3c).

Discussion

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Our results suggest that community context plays a central role in determining the net cost or benefit of sharing pollinators. First, we found that 344 pollinator sharing can lead to both net facilitative and net competitive effects 345 to plant species in a community. Second, we found that multiple ecological 346 factors can modulate the intensity of competition; however, conspecific and heterospecific pollen deposition are tightly coupled and this creates a clear trade-off between the quantity and purity of pollination [45]. Third, we 349 found that the way these factors influence competition is strongly shaped by 350 the community context. That is, if these factors collectively define a species' 351 pollination niche, the same plant species can occupy dramatically different 352 niches depending on the community to which it belongs.

The criterion we used to identify the effect of animal-mediated pollination

suggested that neither facilitation nor competition are dominant in polli-

(a) components' variance and variable contributions principal component analysis of ecological variables components' cumulative percentage of variance visit potential abundance func. originality component 1st 2nd 3rd 4th 25% 50% 75% 100% 0% contribution (b) plant realised niches in PCA space convex hulls of species niches across communities S. crispa 2nd component (18%) V. intermedia N. euosimum 0 1st component (57%) (c) flexibility of plant's strategies median distance between plant niches vs. randomisations → more flexible → S. crispa T. megapotamicum T. sidioides B. pingraea C. microphylla C. herbertii -0 C. acanthoide S. platense N. euosimun H. incana M. pulegium V. intermedia ← less flexible

Figure 3: The flexibility of plant strategies. (a) The two first components explain a large proportion of the total variance. (b) When plants that were sampled in more than one community are plotted in terms of these two components, we observe that their points, which represent the strategy (the particular combination of ecological factors) of that species in its community, do not seem to be grouped by plant species. (c) This was confirmed using Monte Carlo randomizations of the median distance between strategies of a plant species. Only two of the examined species had strategies that were less flexible than would be expected at random.

p-value

0.50

0.05

0.01

nation communities. A previous study showed that, in diverse pollination 356 communities, sharing pollinators does not necessarily translate to net com-357 petition [24]. The overall extent of facilitation/competition of pollination 358 communities can depend on factors like the adversity of the environment 359 [24,46] or how disturbances affect pollinator populations [47]. However, why 360 some species experience more competition than others was not understood. 361 Although we were unable to measure the fitness outcomes, here we show that 362 the coexistence of facilitative and competitive effects of animal-mediated 363 pollination can be at least partially explained by trade-offs in the pollination service—both the short and the long term. These trade-offs arise when plants 365 simultaneously maximise the deposition of conspecific pollen and minimise 366 that of heterospecific pollen. 367

In the short term, being a specialist and sharing no pollinators might reduce 368 competition [48] and hence be preferable, not only due to costs to male fitness 369 [16,49], but also because, as we show here, sharing pollinators reduces both 370 the quantity and purity of the conspecific pollen deposited. However, over 371 long periods of time, there could be a risk associated with a specialist plant 372 having few pollinators [50]. To ensure long-term survival, it is thus likely 373 that plants also need to balance this risk with the costs of sharing pollinators 374 [51]. One possible solution is to share pollinators and have original traits—as 375 we show that trait originality is generally beneficial to pollen deposition and 376 it is commonly thought that species that are further from others in trait space benefit from reduced competition. Yet, there are two possible caveats 378 to this strategy that highlight the interrelatedness of the ecological factors. 379 First, in a mutualism context, it is also possible that trait originality could 380 come at the cost of being less 'apparent' to pollinators [52]. Second, the 381 negative relationship between originality and generalism [32] has been shown to depend on plant abundance [27], with generalist species being able to 383 also have original traits only when they are abundant enough to provide a 384

valuable reward to make visiting worthwhile to pollinators.

Visit potential (high pollen and visits) and floral abundance, which were 386 the most important predictors of pollen deposition here, introduced an 387 even more explicit trade-off between gaining conspecific pollen and avoiding 388 heterospecific pollen. Receiving high visitation increases conspecific pollen 389 deposition but increases heterospecific pollen deposition to a greater extent even when the visitors are likely to carry a high proportion of conspecific 391 pollen [53]. Contrastingly, being abundant reduces the amount of conspecific 392 pollen deposited and simultaneously reduces heterospecific pollen at a faster 393 rate. Our results corroborate the importance that two-species studies have 394 ascribed to visitation and abundance [16,35,54], but they also suggest that 395 [55] balancing the pros and cons of sharing pollinators at the community 396 level is not trivial and might even be partially responsible for the diversity 397 of plant-pollinator communities [56]. 398

We observed, as expected, that the effects of pollen deposition can vary 399 widely among species. For instance, while the fitness of some plant species 400 can be hurt even by low amounts of heterospecific pollen, for others, fitness 401 can instead be limited by the amount of conspecific pollen [18,57]. Alter-402 natively, plant species can also differ substantially in the extent to which 403 self- vs. outcross-pollen differ in their value for fertilization. The difference 404 can be particularly relevant for species that are not self-fertile or those in 405 which self-fertilization is rarely effective due to a temporary separation in 406 the maturation of the sexes (dichogamy). 407

Importantly, we show here that the balances between costs and benefits are determined not only by species identity but also by the community to which plants belong. Specifically, most plant species appear to be flexible enough to adopt markedly different niches in different communities. From an evolutionary perspective, our results suggest that selection for a particular strategy might say something about the community in which a species has

typically inhabited during its evolutionary history. Furthermore, from a more 414 applied perspective, flowering plants are sometimes introduced to attract 415 pollinators on other nearby plants. On the one hand, our results suggest that 416 introduced plants that increase the relative originality of natives (e.g. through distinct phenology) might have positive effects [31]. On the other, because 418 different niches can lead to different outcomes across communities, our results 419 also highlight the difficulties involved in predicting whether the introduced 420 plant species will facilitate or compete with neighbours [58]. Other factors 421 that we were unable to measure (e.g. pollinator behaviour and densities or the spatial context) have also been shown to play a role in the outcome of 423 animal-mediated pollination [26,45,59,60]. Nevertheless, our results indicate 424 that the strategies a plant might use to successfully minimise competition for 425 pollination (or maximise facilitation) must be determined relative to other 426 species in the community, rather than an absolute property of the species itself. 428

Overall, pollination communities might not necessarily be the paradigmatic 429 mutualistic communities we often believe. Instead, we confirm that plant 430 competition is pervasive in pollination communities. The potential strategies 431 to minimise competition are likely to involve strong trade-offs in the pollina-432 tion service both in the short and long-term. Most remarkably, despite large 433 differences in the quantity and purity of conspecific pollen deposition among 434 species, our results provide solid evidence that competition for pollination is 435 inherently a community process and the same species can occupy different 436 niches in different communities. Many of the widely used theoretical mod-437 els of plant-pollinator communities do not account for the negative effects 438 of sharing pollinators [61]. However, achieving a better understanding of 439 species coexistence and how plant biodiversity is supported by pollination communities will require seeing them as both mutualistic and competitive communities [62]. 442

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