

# Draft information

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## **Requests to co-authors**

- Title suggestions/feedback?
- Feedback primarily on abstract, introduction and discussion prior to submission.

## **Changes since the last draft**

- Adjusted language away from competition for pollination towards pollination service and pollen deposition across the paper.
- Removed the section where we calculated competition/facilitation in ecological communities
- Other minor adjustments to address comments by reviews received when submitted to Proceedings B
- Rewrote abstract
- Updated title

## **To-do**

- Update supplementary information

# The trade-offs of pollinator sharing: pollination service is determined by the community context

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# Abstract

A fundamental feature of pollination systems is the indirect facilitation and competition that arises when plants species share pollinators. ~~When plants share pollinators, the pollination service can be influenced;~~ not only by how many partners plant species share, but also by multiple intertwined factors like abundance, visitation, or traits. These factors inherently operate at the community level. However, most of our understanding of how these factors may affect the pollination service is based on systems of up to a handful of species. By examining comprehensive empirical data in eleven natural communities, here we show that the pollination service is ~~surprisingly~~ only partially influenced by the number of shared pollinators. The factors that most influence the pollination service (abundance and visit effectiveness) also introduce a trade-off between the absolute amount of conspecific pollen received and the amount relative to heterospecific pollen. Importantly, the ways plants appear to balance these trade-offs depend strongly on the community context, as most species showed flexibility in the strategy they used to cope with competition for pollination.

**Keywords:** interspecific pollen transfer, pollen deposition, pollination costs and benefits, pollination network, pollination niche, and pollinator sharing

## Introduction

Animal pollination plays a disproportionately important role in food production and maintenance of global biodiversity (Klein et al. 2007, Bascompte and Jordano 2007, Ollerton et al. 2011). At a pairwise level, the mutually beneficial relationship between plants and pollinators underpins the pollination service. At a community level, sometimes involving hundreds of species, both plant and pollinator species are connected in a myriad of indirect connections when pollination partners are shared. These indirect connections can dramatically alter the quality of the pollination service that plants receive because they determine how conspecific and heterospecific pollen is transferred across the community (Morales and Traveset 2008). Generally speaking, there is a trade-off between the benefits gained from maximising the number of partners and the costs of sharing them with other plant species (Waser 1978). However, due to the large number of factors that operate at the community level, we do not understand how sharing pollinators affect the pollination service beyond systems with more than a handful of species. Here we investigate how pollinator sharing affects pollen transfer *in natural communities* and how it compares to other factors known to play a role in community dynamics like abundance, traits, and visitation patterns.

There are two main mechanisms through which sharing pollinators can affect plant fertilisation (Morales and Traveset 2008). The first is by changes in intraspecific pollen transfer. Intraspecific pollen transfer happens, for example, when plants with more attractive flowers might reduce the number of visits to those less of attractive neighbouring plants, and hence reduce the amount of *conspecific pollen* deposited by animals (Yang et al. 2011). The second is via interspecific pollen transfer. In that case, even receiving a visit might not necessarily translate into fertilisation (Campbell and Motten 1985) because a focal plant might receive *heterospecific pollen* or because pollen from the focal plant might be lost to different species. Naturally, the precise effects on female or male plant fitness of conspecific and heterospecific pollen deposition depend on the species involved (and are unknown for many plant species).

57 However, even for species well adapted to pollinator sharing, receiving foreign pollen on  
58 stigmas or losing pollen to foreign stigmas, is neutral at best. Indeed, there is substantial  
59 evidence supporting the idea that heterospecific pollen deposition can be detrimental to seed  
60 production and plant fitness (Ashman and Arceo-Gómez 2013, Arceo-Gómez and Ashman  
61 2016).



62 All together, provided pollen is viable and compatible (de Jong et al. 1992, Dafni and Firmage  
63 2000, Ramsey and Vaughton 2000), the higher the quantity of conspecific pollen and its  
64 purity (relative to heterospecific pollen), the better the pollination service received by the  
65 focal plant. As such, measuring conspecific and heterospecific pollen deposition provides  
66 a good indication of the potential levels of facilitation and competition a plant population  
67 might experience.

68 By definition, intra- and interspecific pollen transfer, occur at the community scale. However,  
69 with few exceptions (Aizen and Rovere 2010, Tur et al. 2016), most of what we know about  
70 pollen transfer and its relationship with key ecological factors are based on studies with two  
71 plant species. That is partly so because, at the community scale, the factors that determine  
72 the patterns of pollen deposition are tightly intertwined, operate simultaneously, and may  
73 lead to emergent phenomena not observed at smaller scales (Flanagan et al. 2011). For  
74 instance, recent empirical evidence suggests that plants with flowering traits that are “original”  
75 relative to others in the community generally have fewer interaction partners (Coux et al.  
76 2016).



77 This evidence aligns with the notion that a species that interacts with few species does so  
78 strongly with each of them, whereas a species that interacts with a large number of species  
79 does so comparatively weakly (Bascompte et al. 2006, Vázquez et al. 2007, Thébault and  
80 Fontaine 2008). If evolutionary specialisation occurs by changing traits to focus on fewer but  
81 better partners (Caruso 2000), we should expect a reduction of competition for pollinators in  
82 species with “original” traits and an increase of competition in species with a large number

83 of interaction partners (Gibson et al. 2012, Carvalheiro et al. 2014). Alternatively, it might  
84 also be the case that abundance (for example, in terms of flower or pollen counts) is the  
85 dominant force driving pollen transfer (Seifan et al. 2014). Abundant plant species might  
86 experience a dilution of available pollinators (Feinsinger 1987, Feldman et al. 2004) but might  
87 also receive more effective visits by capitalising on a larger share of both visits and the pollen  
88 carried by pollinators (?). In this case, a potential reduction in the absolute amount of  
89 conspecific pollen received could be compensated by an increase in the amount of conspecific  
90 pollen relative to heterospecific pollen. Altogether, it is clear that these ecological factors can  
91 indeed shape pollen deposition at the community level. However, we still do not understand  
92 their relative importance and **the trade-offs** that might exist between them.

93 Here, we investigate pollen deposition dynamics at the community scale using empirical data  
94 from eleven plant-pollinator communities in the Argentinian Pampas. First, we investigate the  
95 relative contribution that four ecological factors have to the pollination service. Specifically,  
96 we hypothesise that there are trade-offs on how these factors affect the quantity and purity of  
97 conspecific pollen deposition. While quantity and purity should decrease for plants that share  
98 many pollination partners, other factors like the plant's functional originality, its relative  
99 floral abundance, and its visitation patterns should have the potential to compensate for  
100 this decrease. Second, we examine how these four factors that might affect pollen deposition  
101 can change across communities where species are present. Because these factors may affect  
102 the pollination service in contrasting ways, and a species' role is relative to other species in  
103 the community, we predict that species present in multiple communities should be flexible  
104 enough to compete for pollinators under different community contexts.

## Methods

We collected data from eleven co-flowering plant communities and their pollinators in three locations in the Argentinian Pampas. In each location, we sampled two restored and two agricultural fragments, except in one located in the Flooding Pampas, where we were only able to sample one restored fragment due to the lack of available sites.

### Factors affecting quantity and purity of pollination service

Our first objective was to investigate the relative contribution that different ecological factors have on pollen deposition. Generally speaking, we expect that any factor that increases the amount of conspecific pollen deposited in stigmas, both in quantity and purity relative to heterospecific pollen, also has a positive effect on the pollination service. Specifically, we investigated the effect of (i) a plant's number of shared pollinator species, (ii) a plant's abundance relative to the rest of the community, (iii) the mean visit potential—a metric that combines the amount and type of pollen carried by floral visitors and the number of visits it receives from them—and (iv) the plant's functional originality (Laliberté and Legendre 2010). See *Data Analysis* section below for more details on these four factors.

### Data collection

In each of the studied communities, we quantified pollen deposition in a subset of plant species between December 2010 and February 2011. This subset comprised between three and nine common insect-pollinated (entomophilous) plant species that were flowering during the sampling period. Based on data from previous years (Marrero et al. 2014), we chose plant species such that they cover a wide range on a specialization-generalization gradient as well as a wide range of abundances. In each of the selected plants, we removed all flowers leaving only buds that were expected to go into florescence on the next day. Two days after flowering,

we collected all remaining flowers and counted the number of conspecific and heterospecific pollen grains in their pistils. More details can be found in Marrero et al. (2016).

To obtain the number of shared pollinators for each species, we collected data to construct qualitative and quantitative pollination networks. Qualitative networks were constructed based on ten-hour observations of floral visits in each fragment. Quantitative networks were constructed using two 50 m randomly located transects in each fragment. We counted and collected all floral visitors found in a 2 m wide strip while walking at a pace of 10 m per minute (Memmott 1999, Marrero et al. 2014). We visited the transects each month between November 2010 and March 2011. To obtain floral abundance, we counted all units of floral attraction found during an independent sampling of the same transects used to construct the quantitative visitation networks. To estimate visit potential, we need to construct pollen transfer networks in addition to the visitation networks. To do this, we examined the pollen loads present on the floral visitors collected (Marrero et al. 2017). When the pollen count on an individual animal was estimated to be less than 2,000 grains, we identified every grain to the species level when possible and to pollen complexes when it was not. When the pollen count was above 2,000 grains, we classified approximately 50% of pollen and total pollen counts were extrapolated (Bosch et al. 2009). Finally, we also recorded morphological traits that relate to plant type (herb, shrub, climber), life cycle (annual, perennial), flower colouration, phenology, and whether the species is native in the study region. More details can be found in Marrero *et al.* (2014 and 2017).

## Data analysis

To investigate the impact of ecological factors on pollination services, we used two sets of linear mixed models (LMM) with bootstrap resampling. The response variables for these model sets were the number of conspecific and heterospecific pollen grains deposited per stigma in flowers open to animal-mediated pollination. We used LMMs in which pollen loads



were log-transformed because these models offered a better fit than equivalent GLMMs with Poisson (or quasi-Poisson) error structure. Models were fitted using the R package `nlme` 3.1-131 (Pinheiro et al. 2018).

Because the amount of deposited pollen can vary widely across species, and potentially also across communities, we evaluated two possible structures for the random effects: one that includes a random intercept for plant species, and one that treats measures from species across different communities independently. We selected the best random structure by comparing the median Akaike Information Criterion for small samples (AICc).

As fixed predictors in the models, we included the four ecological factors described above. Specifically, we calculated the number of shared pollinators for each plant species by pooling data from the qualitative and quantitative pollination networks. To calculate the plants' relative floral abundance in their community, we aggregated floral counts for each species. We then calculated the mean visit potential of pollinator species  $i$  to plant species  $j$  as

$$o_{ij} = \frac{v_{ij} p_{ji}}{v_i p_j}$$

, where  $v_{ij}$  is the observed number of visits by  $i$  to  $j$ ,  $p_{ji}$  is the number of pollen grains from  $j$  attached to  $i$ ,  $v_i$  is the total number of visits performed by  $i$ , and  $p_j$  is the total number of grains carried by  $j$ . We log-transformed the number of shared pollinators, floral abundance, and visit potential before including them in the model.

Finally, functional originality is defined as the distance of a species from the community trait average—the centroid of functional space of the community (Laliberté and Legendre 2010, Coux et al. 2016). To include phenological variation, we treated floral abundance in each of the survey months (November to March) as a “trait” in our analysis. To account for the non-independence of floral counts and weight all traits equally, we assigned a weight of 1/5 to these abundances (one for each month). We scaled all traits before calculating the centroid of the functional space and calculated the species-specific functional coordinates using the R

package FD 1.0-12 (Laliberté et al. 2014). Finally, we scaled all four factors to have a zero mean and unitary variance.

To estimate the coefficients, perform model selection, and quantify the associated uncertainty, we used a combination of multi-model inference and bootstrap resampling with 99 replicates. First, we performed model selection using the AICc and determined the likelihood of each candidate model (a particular combination of predictors) by calculating the median  $\Delta\text{AICc}$  (relative to the most likely model) for each bootstrap sample. As we wanted model coefficients from more likely candidate models to carry more weight in our results, we sampled the coefficients for our factors proportionally to the likelihood of their candidate model. Finally, we used these distributions of the model coefficients to estimate their mean impact on the pollination service (in terms of quantity and purity of conspecific pollen deposition).

## Flexibility of plant strategies

Our second objective was to tease apart whether and how these factors that might affect pollen deposition might change across communities species are present. If community context plays a relatively small role, or species are inflexible in regards to these factors, we would expect plants of the same species to fill similar pollination niches across different communities. Alternatively, if the community plays a significant role and plant species are flexible, we should be able to observe differences in the realised niche a plant species occupies across communities. To test this, we first used a principal component analysis (PCA) of the four ecological factors (number of shared pollinators, floral abundance, visit potential, and trait originality). We scaled factors across the whole study to ensure that the PCA space does not change according to the species present in each community. We define a species' niche in a community as its coordinates in PCA space. For each species that was present in two or more communities, we then calculated (i) the median distance between the strategies that the species uses in different communities and (ii) the area of the convex hull defined by

these points in the first two principal components (only for species present in three or more communities). We then compared the magnitude of these two metrics to those obtained with 99 Monte Carlo randomizations in which we replaced the strategy of the focal plant species by that of another randomly selected species in the dataset.

## Results

### Factors affecting quantity and purity of pollination service

We ~~then~~ examined the potential roles played in pollen deposition by four ecological factors (number of shared pollinators, abundance, mean visit potential, and functional originality). We found that our models of pollen deposition had high explanatory power (the coefficient of determination  $R^2$  ranged between 0.76 and 0.93) although a large portion of the explanatory power came from the random effects (Table S3). As determined by AICc, the random structure best supported by the data was the one that fit an intercept for each species in each community (as opposed to a common intercept for each species irrespective of the community to which they belong). This structure was best for both the models of conspecific and heterospecific pollen (Table S4).

Of the four factors we considered, we found that a plant's mean visit potential and relative floral abundance were the most important at predicting pollen deposition in plant stigmas (Fig. 1a). Surprisingly, the number of shared pollinators was comparatively unimportant, particularly for models of heterospecific pollen deposition, as it was only ever included in models with relatively large AICc values (Table S5).

[Figure 1 about here.]

We found that the relationship between each of the ecological factors and pollen deposition was similar for both conspecific and heterospecific pollen. That is, strategies that were associated

with an increase in conspecific pollen were also associated with an increase in heterospecific pollen deposition. Specifically, the plants' mean visit potential had a positive effect on pollen deposition (Fig. 1b). However, the effect size was slightly larger for heterospecific than for conspecific pollen. This larger effect indicates that, although there is a positive association between visit potential and the quantity of pollen deposition, there is a negative relationship with its purity (Fig. 1c). In contrast, a plants' relative floral abundance negatively affected its deposition quantity, but the mean difference between the coefficients in the models indicates a positive association with purity (Fig. 1c). The third most important factor, functional originality, had a positive, although comparatively smaller, association with both the quantity and purity. Finally, the number of shared pollinators had negative and neutral associations with conspecific and heterospecific pollen, respectively, but these impacts were small when compared to the other factors. Although the ecological factors were positively correlated (Fig. S2), the collinearity between predictors did not qualitatively affect our findings (Fig. S3).

## Flexibility of plant strategies

We used a PCA of the ecological factors–species matrix to investigate whether plants' “strategy” towards pollen deposition is similar across communities or whether they are flexible and therefore a reflection of the community context. The first two PCA components explained 75% of the total variance (Fig. 2a). The first component was dominated by visit potential and relative abundance while the second component was dominated by the number of shared pollinators and the plant's functional originality. When we locate the species that were sampled in more than one community in the first two PCA components (Fig. 2b), we observe that the niches of any given species do not tend to be close to each other. Indeed, when we measured the median distance between the plants' coordinates, we found that it was ~~only~~ significantly smaller than that of randomisations for only two of the twelve species analysed (Fig. 2c).

[Figure 2 about here.]

## Discussion

Our results suggest that community context plays a central role in determining the pollen deposition dynamics and ultimately the net cost or benefit of sharing pollinators. First, we found that multiple ecological factors can modulate the quality of the pollination service; however, conspecific and heterospecific pollen deposition are tightly coupled and this creates a clear trade-off between the quantity and purity of pollination (Thomson et al. 2019). Second, we found that the way these factors shape pollen deposition for a species could be dramatically different across communities. For instance, while in a particular community a plant species could show high levels of pollinator sharing and relatively low trait differentiation, the same species in another community can have relatively high trait differentiation and low levels of pollinator sharing. Our findings highlight that trade-offs can at least partially explain the coexistence of facilitative and competitive effects of animal-mediated pollination in the pollination service—both the short and the long term.

The trade-offs involved in attaining high-quality pollination service (and more broadly between facilitation and competition) are likely to arise when plants simultaneously maximise the deposition of conspecific pollen and minimise that of heterospecific pollen. In the short term, being a specialist and sharing no pollinators might reduce competition (Muchhala et al. 2010) and hence be preferable, not only due to costs to male fitness (Morales and Traveset 2008, Muchhala and Thomson 2012), but also because, as we show here, sharing pollinators reduces both the quantity and purity of the conspecific pollen deposited. However, over long periods of time, there could be a risk associated with a specialist plant having few pollinators (Ricketts 2004). To ensure long-term survival, it is thus likely that plants also need to balance this risk with the costs of sharing pollinators (Aizen et al. 2012). One possible solution is

to share pollinators *and* have original traits—as we show that trait originality is generally beneficial to pollen deposition and it is commonly thought that species that are further from others in trait space benefit from reduced competition. Yet, there are two possible caveats to this strategy that highlight the interrelatedness of the ecological factors. First, in a mutualism context, it is also possible that trait originality could come at the cost of being less ‘apparent’ to pollinators (Reverté et al. 2016). Second, the negative relationship between originality and generalism (Carvalho et al. 2014) has been shown to depend on plant abundance (Coux et al. 2016), with generalist species being able also to have original traits only when they are abundant enough to provide a valuable reward to make visiting worthwhile to pollinators.

Visit potential (high pollen and visits) and floral abundance, which were the most important predictors of pollen deposition here, introduced an even more explicit trade-off between gaining conspecific pollen and avoiding heterospecific pollen. Receiving high visitation increases conspecific pollen deposition but increases heterospecific pollen deposition to a greater extent—even when the visitors are likely to carry a high proportion of conspecific pollen (Fang and Huang 2016). Contrastingly, being abundant reduces the amount of conspecific pollen deposited and simultaneously reduces heterospecific pollen at a faster rate. Our results corroborate the importance that two-species studies have ascribed to visitation and abundance (Feldman et al. 2004, Muñoz and Cavieres 2008, Morales and Traveset 2008), but they also suggest that (because visitation, pollen production and abundance are usually correlated; Sargent and Otto 2006) balancing the pros and cons of sharing pollinators at the community level is not trivial and might even be partially responsible for the diversity of plant-pollinator communities (Benadi and Pauw 2018).

We observed, as expected, that the effects of pollen deposition can vary widely among species. For instance, while the fitness of some plant species can be hurt even by low amounts of heterospecific pollen, for others, fitness can instead be limited by the amount of conspecific pollen (Campbell and Motten 1985, Arceo-Gómez et al. 2019). Alternatively, plant species

can also differ substantially in the extent to which self- vs. outcross-pollen differ in their value for fertilization. The difference can be particularly relevant for species that are not self-fertile or those in which self-fertilization is rarely effective due to a temporary separation in the maturation of the sexes (dichogamy).

Importantly, we show here that the balances between costs and benefits are determined not only by species identity but also by the community to which plants belong. Specifically, most plant species appear to be flexible enough to adopt markedly different niches in different communities. From an evolutionary perspective, our results suggest that selection for a particular strategy might say something about the community in which a species has typically inhabited during its evolutionary history. Furthermore, from a more applied perspective, flowering plants are sometimes introduced to attract pollinators on other nearby plants. On the one hand, our results suggest that introduced plants that increase the relative originality of natives (e.g. through distinct phenology) might have positive effects (Gibson et al. 2012). On the other, because different niches can lead to different outcomes across communities, our results also highlight the difficulties involved in predicting whether the introduced plant species will facilitate or compete with neighbours (Bartomeus et al. 2008). Other factors that we were unable to measure (e.g. pollinator behaviour and densities or the spatial context) have also been shown to play a role in the outcome of animal-mediated pollination (Cariveau and Norton 2009, Flanagan et al. 2011, Ye et al. 2014, Thomson et al. 2019). Nevertheless, our results indicate that the strategies a plant might use to successfully minimise competition for pollination (or maximise facilitation) must be determined relative to other species in the community, rather than an absolute property of the species itself.

Overall, using empirical data on pollen deposition, we show that at the community level sharing pollinators has a smaller effect on pollen deposition than ~~what~~ we expected based on experimental studies with a handful species. Other factors that underpin community dynamics (abundance, traits, visitation) also influence patterns of pollination quantity and purity. The

interrelatedness of these factors and the flexibility of species to position themselves within these factors means that their contributions to the quality of the pollination service should not and cannot be understood in isolation. All factors we analysed involve substantial trade-offs in pollen deposition in the short and likely also in the longer term. These trade-offs emphasise the competitive nature of pollination. However, many of the widely used theoretical models of plant-pollinator communities do not account for the adverse effects of sharing pollinators (but see Rohr et al. 2014 and similar). We propose that achieving a better understanding of species coexistence and how pollination support plant biodiversity will require seeing them as both mutualistic and competitive communities (Johnson and Bronstein 2019).

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# List of Figures

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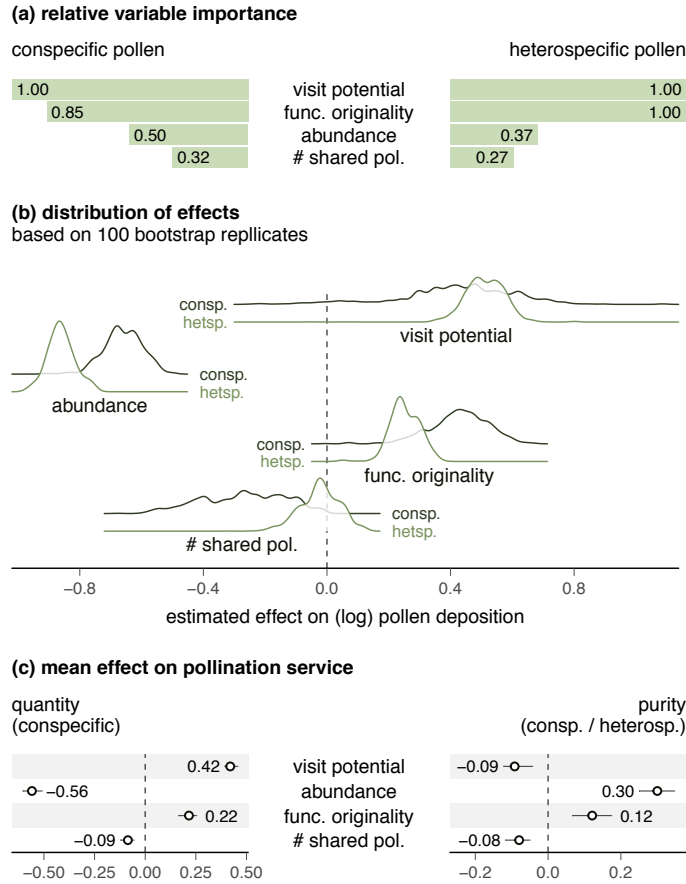


Figure 1: Effect of ecological factors on the pollination service. (a) The plant's visit potential and relative floral abundance are the most important factors determining the deposition of conspecific and heterospecific pollen. Meanwhile, the number of shared pollinators was generally less important. The graph shows the relative importance calculated as the sum of the Akaike weights of the candidate models that included the selected factor. (b) The association between ecological factors and heterospecific pollen tended to align with their association with conspecific pollen. Visit potential and functional originality had a positive association with pollen deposition, while floral abundance and the number of shared pollinators had a negative association. The plot shows the distribution of the effects (across 99 bootstrap replicates) of the four ecological factors for conspecific and heterospecific pollen. (c) The end result of these associations is that only the plants' functional originality has a positive impact on both the quantity and purity of conspecific pollen deposition (relative to heterospecific pollen). The plot shows the model averaged mean effect ( $\pm$  SE of 99 bootstrap replicates).



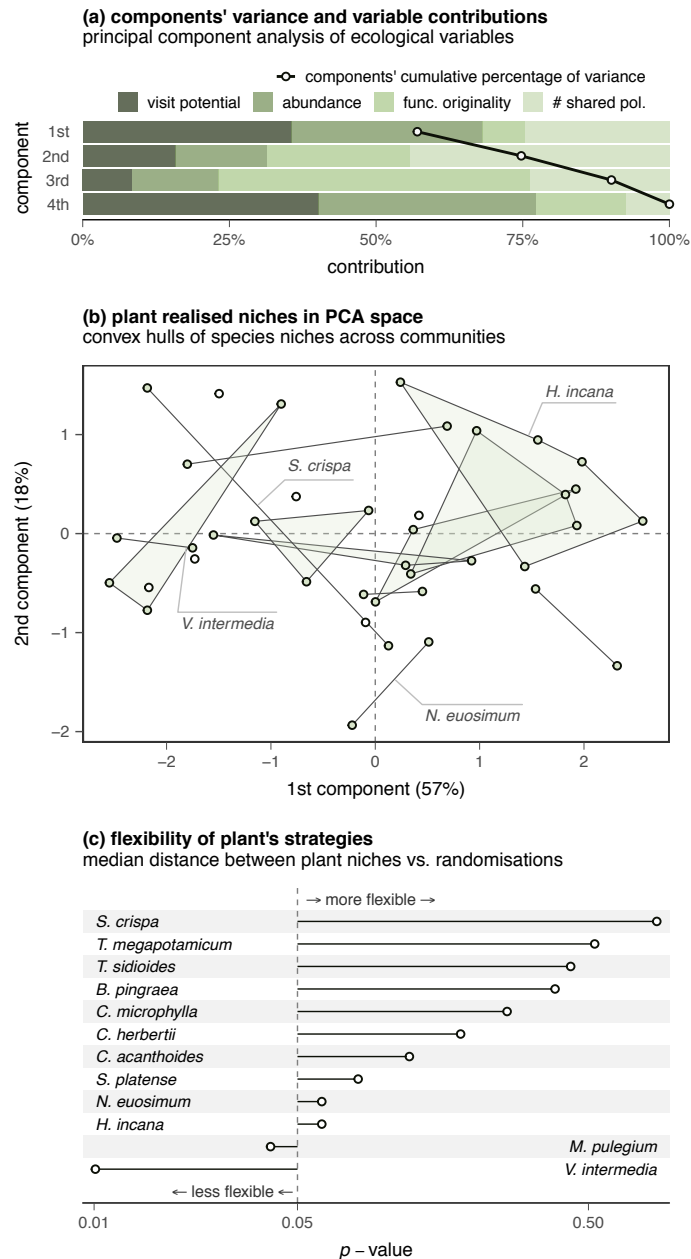


Figure 2: The flexibility of plant strategies. (a) The two first components explain a large proportion of the total variance. (b) When plants that were sampled in more than one community are plotted in terms of these two components, we observe that their points, which represent the strategy (the particular combination of ecological factors) of that species in its community, do not seem to be grouped by plant species. (c) This was confirmed using Monte Carlo randomizations of the median distance between strategies of a plant species. Only two of the examined species had strategies that were less flexible than would be expected at random.

# Community context determines competition vs. facilitation trade-offs in pollination systems

Supplementary information

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Table S1: Summary of the model used to analyse the relationship between heterospecific and conspecific pollen

predictor	estimate	S.E.	z-value
<b>fixed component</b>			
(Intercept)	4.976	0.279	17.862
heterospecific	0.008	0.017	0.474
<b>random component (species:community)</b>			
S.D. random intercept	1.964	-	-
S.D. random slope	0.120	-	-

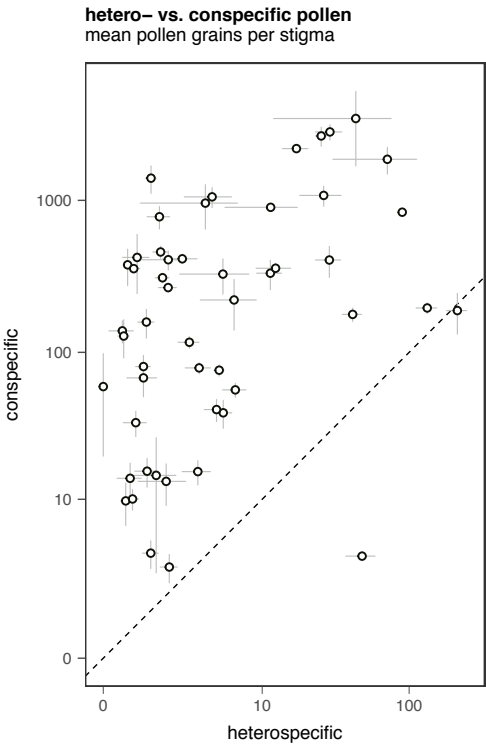


Figure S1: Despite the variation in these slopes, plants overall had more conspecific than heterospecific pollen deposited in their stigmas.

Table S2: The slope of the relationship between heterospecific and conspecific pollen for each species in their community (fixed effect + conditional effect). Community names are constructed by location - agricultural/restored - fragment number.

species name	community	slope	S.E.
<i>Aloysia gratissima</i>	Anquilóo - reserve - 2	0.0746	0.0144
<i>Baccharis pingraea</i>	San Claudio - reserve - 1	-0.0012	0.0359
<i>Carduus acanthoides</i>	Anquilóo - agricultural - 2	0.0116	0.0147
<i>Carduus acanthoides</i>	San Claudio - agricultural - 1	-0.0106	0.0040
<i>Carduus acanthoides</i>	San Claudio - agricultural - 2	0.0518	0.0044
<i>Carduus acanthoides</i>	San Claudio - reserve - 1	0.0781	0.0710
<i>Carduus acanthoides</i>	San Claudio - reserve - 2	-0.0008	0.0359
<i>Cirsium vulgare</i>	Anquilóo - agricultural - 2	-0.0401	0.0025
<i>Cirsium vulgare</i>	Las Chilcas - reserve - 1	0.0007	0.0012
<i>Cirsium vulgare</i>	San Claudio - agricultural - 2	0.0197	0.0158
<i>Cirsium vulgare</i>	San Claudio - reserve - 1	-0.0149	0.0076
<i>Condalia microphylla</i>	Anquilóo - reserve - 1	0.0487	0.0200
<i>Cypella herbertii</i>	Las Chilcas - agricultural - 2	0.0037	0.0002
<i>Cypella herbertii</i>	Las Chilcas - reserve - 1	-0.0052	0.0001
<i>Descurania argentina</i>	Anquilóo - agricultural - 2	0.0429	0.0048
<i>Diploaxis tenuifolia</i>	Anquilóo - reserve - 1	0.0008	0.0004
<i>Diploaxis tenuifolia</i>	Anquilóo - reserve - 2	0.5173	0.0270
<i>Diploaxis tenuifolia</i>	San Claudio - reserve - 2	-0.0045	0.0001
<i>Dipsacus</i> sp.	San Claudio - reserve - 2	-0.0368	0.0648
<i>Gaillardia megapotamica</i>	Anquilóo - reserve - 2	0.0016	0.0004
<i>Glandularia hookeriana</i>	Anquilóo - reserve - 2	-0.0942	0.0244
<i>Hirschfeldia incana</i>	Anquilóo - agricultural - 1	-0.0045	0.0013
<i>Hirschfeldia incana</i>	Anquilóo - agricultural - 2	-0.0148	0.0057
<i>Hirschfeldia incana</i>	San Claudio - agricultural - 1	0.0110	0.0020
<i>Hirschfeldia incana</i>	San Claudio - agricultural - 2	0.0031	0.0023
<i>Hirschfeldia incana</i>	San Claudio - reserve - 1	0.0022	0.0002
<i>Hirschfeldia incana</i>	San Claudio - reserve - 2	0.0432	0.0020
<i>Lycium chilense</i>	Anquilóo - reserve - 2	-0.3355	0.0087
<i>Mentha pulegium</i>	Las Chilcas - agricultural - 2	0.0136	0.0866
<i>Mentha pulegium</i>	Las Chilcas - reserve - 1	0.3973	0.0388
<i>Nierembergia aristata</i>	Anquilóo - agricultural - 1	0.0197	0.0217
<i>Nierembergia aristata</i>	Anquilóo - reserve - 1	-0.0065	0.0016
<i>Nierembergia aristata</i>	Anquilóo - reserve - 2	-0.0048	0.0011
<i>Nothoscordum euosimum</i>	Las Chilcas - agricultural - 1	0.0405	0.0034
<i>Nothoscordum euosimum</i>	Las Chilcas - agricultural - 2	-0.0045	0.1162
<i>Physalis viscosa</i>	Anquilóo - agricultural - 1	0.0041	0.0005
<i>Prosopidastrum globosum</i>	Anquilóo - reserve - 2	-0.0012	0.0194
<i>Senecio pulcher</i>	Las Chilcas - agricultural - 1	-0.0104	0.0007
<i>Sisyrinchium platense</i>	Las Chilcas - agricultural - 1	-0.2850	0.0203
<i>Sisyrinchium platense</i>	Las Chilcas - agricultural - 2	-0.0487	0.0324
<i>Sisyrinchium platense</i>	Las Chilcas - reserve - 1	0.0206	0.1143
<i>Solanum sisymbriifolium</i>	San Claudio - agricultural - 1	0.0002	0.0004
<i>Sphaeralcea crispa</i>	Anquilóo - reserve - 1	-0.0601	0.0133
<i>Stemodia lanceolata</i>	Las Chilcas - agricultural - 1	-0.0044	0.0001
<i>Thelesperma megapotamicum</i>	Anquilóo - agricultural - 1	-0.0022	0.0025
<i>Turnera sidioides</i>	Anquilóo - agricultural - 1	-0.0002	0.0001
<i>Turnera sidioides</i>	Anquilóo - agricultural - 2	-0.0140	0.0170
<i>Turnera sidioides</i>	Anquilóo - reserve - 2	-0.0014	0.0002
<i>Verbena intermedia</i>	Anquilóo - reserve - 2	-0.0643	0.0327
<i>Verbena intermedia</i>	San Claudio - agricultural - 2	0.0932	0.0071
<i>Verbena intermedia</i>	San Claudio - reserve - 2	-0.0073	0.0101

Table S3: The coefficient of determination  $R^2$  of the most parsimonious pollen deposition models (those with the lowest AICc). The marginal coefficient of determination describes the proportion of variance explained by just the fixed effects.

conditional $R^2_{(c)}$			marginal $R^2_{(m)}$		
mean	min	max	mean	min	max
<b>conspecific pollen</b>					
0.91	0.87	0.93	0.09	0.06	0.14
<b>heterospecific pollen</b>					
0.80	0.76	0.87	0.27	0.21	0.35

Table S4: Comparison of the two random structures we considered for the models of conspecific and heterospecific pollen deposition. The table shows median  $\Delta$ AIC values of 99 bootstrap resamples of the data. The 5th and 95th percentile are shown inside square brackets. Communities are defined by individual fragments but ignore the hierarchical arrangement of sampling sites.

random structure	$\Delta$ AIC	
	median	C.I.
<b>conspecific pollen</b>		
1   plant sp. * community	0.0	[0, 0]
1   plant sp.	30.7	[8.2, 58.1]
<b>heterospecific pollen</b>		
1   plant sp. * community	0.0	[0, 0]
1   plant sp.	44.6	[19.3, 88.4]

Table S5: Comparison of the different fixed structures we considered for the models of conspecific and heterospecific pollen deposition. The table shows median  $\Delta\text{AIC}$  values of 99 bootstrap resamples of the data. The 5th and 95th percentile are shown inside square brackets.

fixed structure	$\Delta\text{AIC}$	
	median	C.I.
<b>conspecific pollen</b>		
~ abundance + visit potential	0.0	[0, 0]
~ abundance + visit potential + func. originality	0.9	[0.4, 1.3]
~ abundance + visit potential + # shared pol.	1.9	[1.6, 2.1]
~ abundance + visit potential + # shared pol. + func. originality	2.2	[1.6, 2.8]
~ visit potential + func. originality	2.8	[2.1, 3.8]
~ visit potential + # shared pol. + func. originality	3.6	[2.3, 4.6]
~ visit potential	118.3	[75.3, 178.7]
~ visit potential + # shared pol.	119.0	[76, 179.9]
~ abundance	189.7	[150.1, 239.7]
~ abundance + func. originality	191.6	[151.7, 241.6]
~ abundance + # shared pol.	191.7	[151.9, 241.7]
~ func. originality	192.5	[152.9, 242.2]
~ abundance + # shared pol. + func. originality	193.7	[153.6, 243.6]
~ # shared pol. + func. originality	193.7	[154.6, 243.7]
~ # shared pol.	351.8	[293.5, 419.9]
<b>heterospecific pollen</b>		
~ abundance + visit potential	0.0	[0, 0]
~ abundance + visit potential + func. originality	1.1	[0.5, 1.5]
~ abundance + visit potential + # shared pol.	2.1	[1.9, 2.1]
~ abundance + visit potential + # shared pol. + func. originality	3.1	[2.6, 3.5]
~ visit potential + func. originality	11.9	[10, 13.9]
~ visit potential + # shared pol. + func. originality	13.2	[11.2, 15.2]
~ visit potential	67.5	[53.4, 87.5]
~ visit potential + # shared pol.	68.4	[54.2, 88.7]
~ abundance + # shared pol.	206.9	[160.6, 251.5]
~ abundance	207.6	[162.8, 251.7]
~ abundance + func. originality	208.6	[163.2, 252.6]
~ abundance + # shared pol. + func. originality	208.6	[162.2, 253.2]
~ func. originality	214.3	[168.3, 258.7]
~ # shared pol. + func. originality	216.3	[170.3, 260.6]
~ # shared pol.	336.0	[282.6, 391.5]

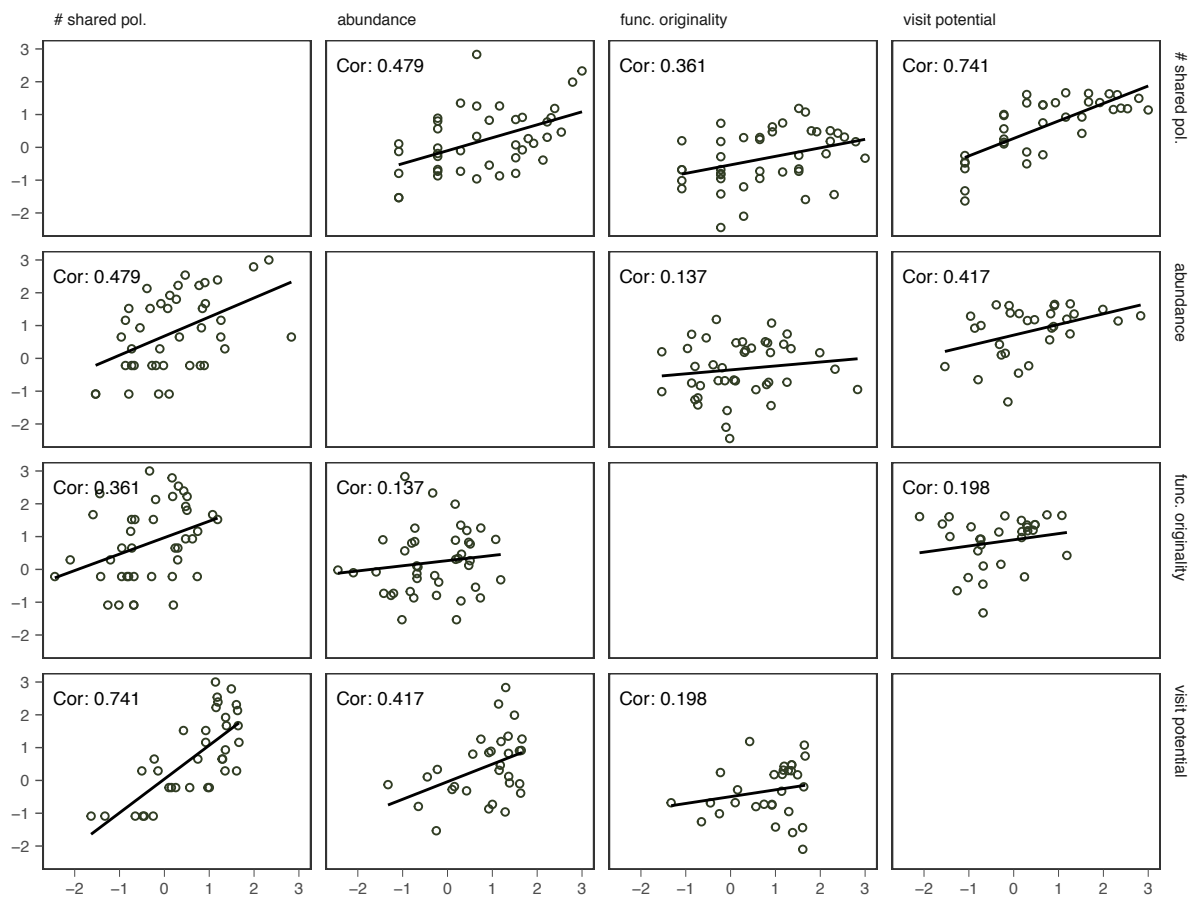


Figure S2: Correlation between the explanatory variables included in the statistical models.

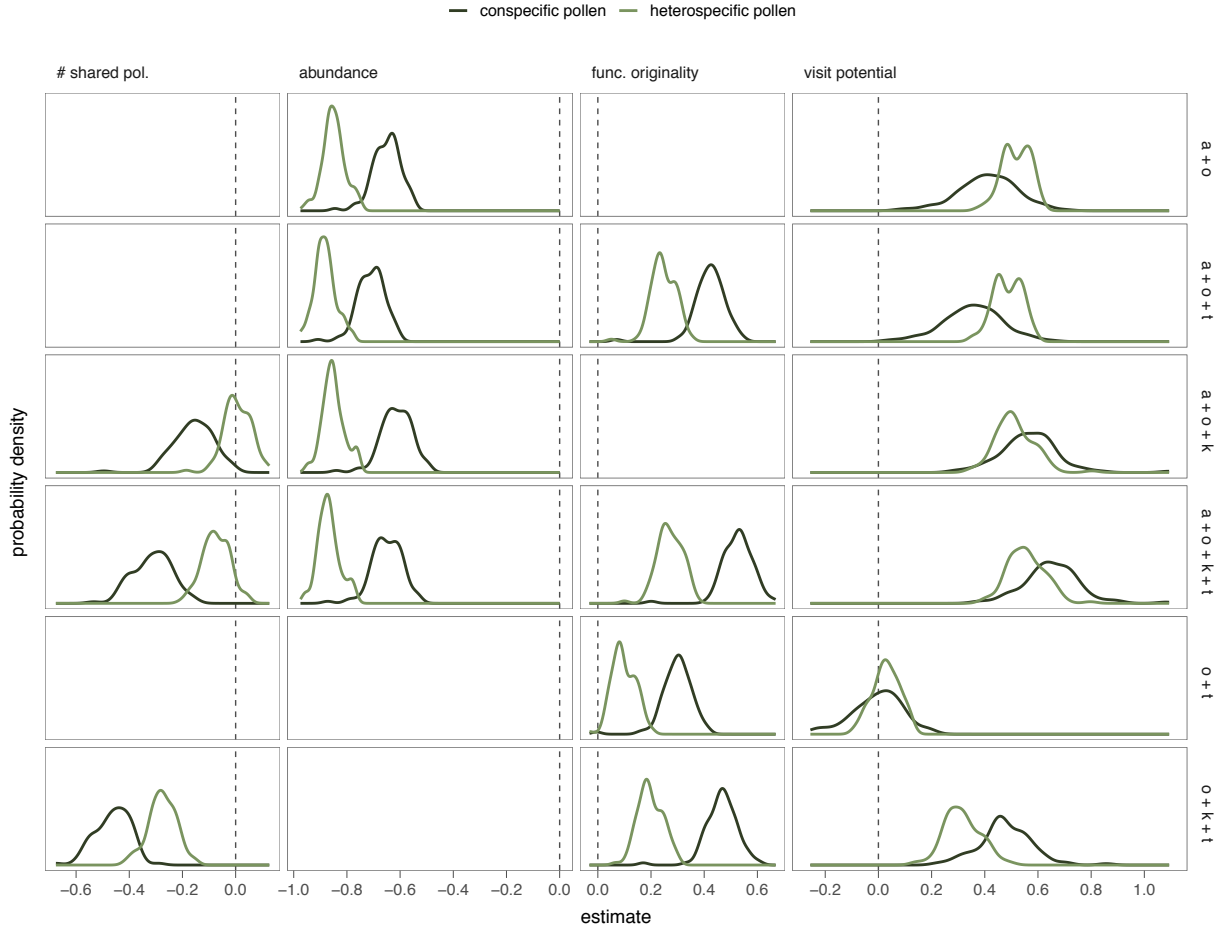


Figure S3: Distribution of effect estimates for models of conspecific and heterospecific pollen density gain. Model formulas have been abbreviated:  $a$  for abundance,  $k$  for the number of shared pollinators,  $o$  for the visit potential, and  $t$  for functional originality. Only candidate formulas with a  $\Delta AICc < 4$  for either conspecific or heterospecific pollen are shown. Models candidates are arranged in decreasing order of support. Although relative abundance, the number of shared pollinators, and the visit potential were all positively correlated, the effect each had on conspecific pollen was similar among models that included all or just some of these three explanatory variables. One exception was visit potential, which exhibits a positive association with the relative amount of conspecific pollen under some variable combinations. Nevertheless, these differences were observed only in model specifications with relatively low AICc support.