INTRODUCTION

1.1 RATIONALE: OVERVIEW OF THE NUTRIENT POLLUTION CHALLENGE

Human population is experiencing a continuous growth since the end of the Black Death in the XIV century (Biraben, 1980), which is at 7.8 billion as of 2020, and it is estimated to be at 9.7 billion and 10.9 billion by 2050 and 2100 respectively (United Nations, Department of Economic and Social Affairs, 2019). Population growth demands increasing amounts of food, which in turn requires an efficient food production system to ensure global food security. In this context, the development of different technical advancements has been a key factor to increase the productivity of the food production system. Notably, crucial developments were achieved in the late modern period¹, including the commercial production of phosphate in 1847 (Samreen & Kausar, 2019), the development of the Haber-Bosch process for the production of synthetic nitrogen-based fertilizers in 1913 (Smil, 1999), and the mechanization of agriculture and the development of the modern intensive farming in the XX century (Constable & Somerville, 2003; Nierenberg & Mastny, 2005).

Despite these advancements have increased the productivity of agriculture and farming industries, multiple environmental impacts associated with them emerges, including water scarcity, greenhouse gases emissions, nutrient pollution of waterbodies, and soil degradation, among others. These threats must be carefully addressed in order to avoid the depletion of natural resources and reach a sustainable food production system.

Focusing on the impacts derived from agriculture and farming on the nutrient cycles, it can be observed that the natural cycles of phosphorus and nitrogen have been altered by these activities (Bouwman et al., 2009). Large amounts of nutrients are released into the environment in the form of synthetic fertilizers and livestock manure. Nitrogen and phosphorus are accumulated in soils, creating a nutrient legacy that is further transported to waterbodies by runoff. This process results in the eutrophication of waterbodies, which can lead algal bloom episodes. Algal blooms are events

¹ The terminology used in this dissertation for the periodization of human history follows the English-language historiographical approach. It should be noted that the late modern period is referred to as the contemporary period in the European historiographical approaches.

resulting from the rapid increase of algae in a water system which can be promoted by an excess of nutrients in water. These episodes alter the normal functioning of aquatic ecosystems, since they cause hypoxia as a consequence of the aerobic degradation of algal biomass by bacteria. Moreover, some species of algae that cause algal blooms can releases toxins into the water systems. The main flows of nutrient releases into the environment by anthropogenic activities is shown in Figure 1.1.

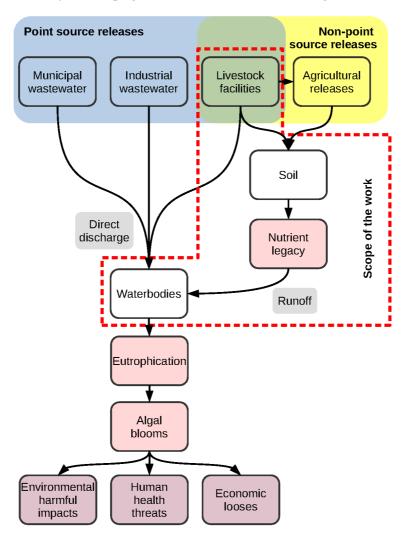


Figure 1.1: Main flows of nutrients released by anthropogenic activities.

In addition to the environmental problems, the use of nutrients for food production also raises geopolitical concerns since phosphorus is one to the most sensitive elements to depletion. Phosphorus is a non-renewable material whose reserves are expeted to be depleted in the next 50 to 100 years. Moreover, no substitute material is currently known

(Cordell et al., 2009). Conversely, synthetic nitrogen can be produced using the atmospheric N₂ as raw material through the Haber-Bosh process. However, nowadays this process relies on non-renewable energy sources, and therefore the production of synthetic nitrogen-based fertilizers is dependent on non-renewable resources as well.

Considering the two challenges described, i.e., nutrient pollution of waterbodies as a consequence of agricultural and farming activities, and the current dependency on non-renewable resources for the production of synthetic fertilizers, nutrient recovery and recycling is not only a desirable but also a necessary approach to develop a sustainable agricultural paradigm and ensure the global food security.

Attending to the nutrient releases from intensive livestock farming facilities, known as concentrated animal feeding operations (CAFOs)2, several manure management techniques are currently used. The land application of manure is a common technique that allows the recycling of nutrients as fertilizers for crops (Kellogg et al., 2000). However, the increase of intensive livestock farming generates vast amounts of waste generated by CAFOs, e.g., each adult cow generates between 28 and 39 kg of manure per day, and each adult pig generates around 11.5 kg of manure per day (USDA, 2009). Manure processing is commonly based on the separation of liquid and solid phases. The liquid phase can be treated in anaerobic and/or aerobic lagoons for organic matter and pathogens removal, as well as odor control (Tilley et al., 2014). The obtained liquid effluent can be used for irrigation and nutrient supplementation of crops. The solid phase can be composted for the degradation of organic matter and pathogens removal, resulting in a solid material called compost with a larger amount of nitrogen and phosphorus available for plants, which is result of the mineralization of nutrients previously contained in organic compounds. Since compost is also a good source of organic matter for crops, it is a valuable material suitable for sale (Tilley et al., 2014). However, both materials, the liquid effluent obtained from the lagoons and compost, are too bulky to be economically transported to nutrient deficient locations (Burns & Moody, 2002). As a result, livestock waste is usually spread in the surroundings of livestock facilities, at a detrimental cost of environment. This result in the gradual build-up of nutrients in soils, which might lead the harmful environmental impacts previously described.

A promising alternative for abating nutrient releases and reducing the environmental footprint of livestock industry is the implementation of processes for the recovery of phosphorus and nitrogen at CAFOs. At the

² CAFO is a regulatory term defined by the U.S. Environmental Protection Agency for large facilities where animals are kept and raised in confined situations (USDA, 2011). This term is used in this dissertation to denote the intensive livestock farming facilities studied.

time, that valuable nutrient-rich materials are obtained for the redistribution of phosphorus and nitrogen to nutrient-deficient areas. There exist a number of processes for nutrient recovery from livestock waste, which can be differentiated into those technologies oriented to phosphorus recovery, including struvite precipitation, calcium-based precipitates production, coagulation-flocculation, electrochemical processes, and systems based on solid-liquid separation; and processes focused on nitrogen recovery, such as stripping, membrane separation, waste drying coupled with ammonia scrubbing, and solid-liquid separation processes. We note that anaerobic digestion is an additional process that can be integrated for manure treatment if the generation of biomethane is pursued, and for increasing the amount of recoverable nutrients through the partial mineralization of nutrients contained in organic compounds. It must be noted that only phosphorus and nitrogen in inorganic compounds can be taken by plants, and therefore the recovery of inorganic nutrients will be the target of the processes studied in this thesis.

The multiple processes for the recovery of phosphorus and nitrogen from livestock waste differ in aspects such as recovery efficiency, processing capacity, capital and operating costs, and products obtained. Therefore, a detailed analysis of each CAFO must be performed in order to select the optimal nutrient recovery system attending to type factors such as the type and amount of waste to be processed, the environmental vulnerability to eutrophication of each region, the current or potential installation of anaerobic digestion systems, etc. Additionally, in the decision-making process these factors have to be prioritized, i.e., sorted by relevance, to select the most suitable nutrient recovery system for each particular facility. As example, more economical processes for nutrient recovery, whose recovery efficiencies are typically lower, could be installed in regions with a low risk of eutrophication. Conversely, regions at severe eutrophication risk require highly efficient nutrient recovery systems that may incur in larger investment and operating expenses. In order to perform a systematic evaluation of CAFOs and their context, we introduce a multi-criteria decision analysis (MCDA) framework integrating geospatial environmental data on eutrophication risk at the subbasin level and techno-economic information of the studied processes.

Attending to the regulatory aspect, nowadays most of the efforts for abating of nutrient releases into the environment and mitigating the eutrophication of waterbodies are focused on the limitation of fertilizer application in croplands. The application of fertilizer and manure for nitrogen supplementation in the European Union (EU) is currently regulated by the Nitrates Directive (91/676/EEC) (Grizzetti et al., 2021). Regarding the limitations for phosphorus application, these are defined at national level. Several

European countries have implemented phosphorus application standards based on the different crops and materials used as fertilizers, being generally more restrictive in Northwestern Europe (Amery & Schoumans,

In sum, it can be observed that nutrient application is limited either in the form of synthetic fertilizers or manure application. However, at present there is a lack of regulation regarding livestock waste treatment (Piot-Lepetit, 2011). In this regard, new efforts to promote the production and adoption of bio-fertilizers obtained from organic waste are being performed through the development of the "Integrated Nutrient Management Plan" (INMAP), which is part of the EU Farm-to-Fork strategy and part of the Circular Economy Action Plan. INMAP should propose actions to promote the recovery and recycling of nutrients, as well as the development of markets for recovered nutrients (Comission, 2020; ESSP, 2021). In this regard, a new regulation for fertilizer products has been released in 2019 (EU 2019/1009), moving struvite and other biofertilizers from the category of waste to fertilizers, establishing a regulatory framework for their use and trade.

In the United States, CAFOs are regulated under the Clean Water Act as point source waste discharges. This regulation sets the need of permits for discharging pollutants to water, which are called National Pollutant Discharge Elimination System (NPDES) permits, including nitrogen and phosphorus releases. These permits must include the necessary provisions for avoiding the harmful effects of the discharges on water and human health (US EPA, 2020b). The development and implementation of a Nutrient Management Plan (NMP) is a required element to obtain an NPDES permit. This document must identify the management practices to be implemented at each CAFO to protect natural resources from nutrient pollution. Land spreading of manure can also be regulated by the NPDES permits, establishing soil nutrient concentration limits and the yearly schedule for manure application. However, no specific methods or processes for waste treatment are defined under federal regulation (US EPA, 2020a). Regarding the use of the recovered nutrients, products obtained from nutrient recovery processes could be classified as waste ny the Clean Water Act, preventing the application of these materials on croplands (NACWA, 2014). However, the U.S. Environmental Protection Agency (US EPA) determined that, although these products could not be directly applied to land under the current regulation, they can be sold as a commodity to be outside of the Clean Water Act restrictions coverage (CNP, 2021). Moreover, US EPA acknowledges that highly refined and primarily inorganic products (such as struvite) could be outside of the scope of these restrictions (CNP, 2021). Nevertheless, further regulation is needed for defining the products obtained from nutrient recovery processes and to clearly state the conditions for their use as fertilizers on croplands.

Considering the previously described aspects, we note that the regulation of the products obtained from nutrient recovery systems is not totally developed yet either in the European Union and the United States, although important efforts are being performed in order to set a comprehensive regulatory framework for the recycling of phosphorus and nitrogen. Furthermore, no regulation regarding the implementation of nutrient recovery processes has been developed However, both regions have developed previous programs to study and promote the implementation of other technologies for the treatment of livestock and other organic waste. Particularly, the deployment of anaerobic digestion systems have received a considerable support from governmental agencies, resulting in programs such as AgSTAR in the US (US EPA, 2021), and BiogasAction (European Comission, 2021) and BIOGAS³ (BIOGAS₃ PROJECT, 2021) in Europe, among many others. These programs could be a guideline for the development of nutrient recovery plans at CAFOs. In this regard, we have studied the impact of the implementation of nutrient recovery systems in the economy of CAFOs, either considering the deployment of standalone nutrient recovery processes, or integrated systems combining nutrient recovery with anaerobic digestion for the production of electricity and biomethane. Moreover, incentive policies have been analyzed to minimize the negative impact of nutrient recovery on CAFOs economy using the Great Lakes area as case study. In addition, the fair distribution of monetary resources when limited budget is available has been studied using the Nash allocation scheme.

An overview of the main topics studied in this thesis can be observed in Figure 1.2. This work pretends to analyze strategies for promoting effective nutrient recycling addressing studies on the technical, environmental and economic dimensions involved, pursuing the development of sustainable food production paradigm.

1.2 APPROACHES FOR PROCESSES MODELING

Process modeling, defined as the mathematical modeling and simulation of systems, falls under the scope of the Process System Engineering (PSE) discipline. These systems include physical, chemical, and/or biological operations. Process modeling forms the foundation for other activities involved in the scope of PSE, including process design, optimal scheduling

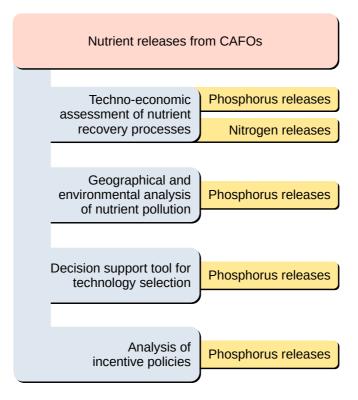


Figure 1.2: Main topics covered in this work.

and planning of the systems operations, and process control (Stephanopoulos & Reklaitis, 2011).

Different modeling techniques have been developed to mathematically describe and represent systems from different domains, including but not limited to the chemical, biochemical, agrochemical, food, and pharmaceutical domains of engineering (Pistikopoulos et al., 2021). An overview of the main modeling techniques is shown in the next sections based on the classification proposed by Martín and Grossmann (2012).

Short-cut methods 1.2.1

These type models are the most basic approach to process modeling. They are based on mass, energy, and momentum balances, and can be embedded in other models, such as supply chain models.

1.2.2 Rules of thumb

This approach is based on industrial operational data. It provides typical ranges for operating and design values, reflecting the actual parameters of the systems modeled. However, the use of these models is constrained by the availability of data. Compendiums of rules of thumb for different systems can be found in Couper et al. (2005), Hall (2012), Sadhukhan et al. (2014).

1.2.3 Dimensionless analysis

This methodology is based on dimensionless groups that describe the performance of a particular system. These models are able to capture the physical meaning of the modeled processes, and they are specially useful to capture scale-up and scale-down issues (Szirtes, 2007).

1.2.4 Mechanistic models

This approach relies on first principles for systems modeling, as short-cut models. However, mechanistic models rely in more detailed first principles such as the underlying chemistry, physics or biology that governs the behavior of a particular system. Chemical (Loeppert et al., 1995) and phase (Brignole & Pereda, 2013) equilibrium models, kinetic models (Buzzi-Ferraris & Manenti, 2009), population balances (Ramkrishna, 2000), and computer fluid dynamics (CFD) (Anderson & Wendt, 1995) fall under this category.

1.2.5 Surrogate models

These models aim at developed simplified models from data obtained from rigorous mechanistic models. This approach is widely used for embedding system models into other applications such as process control or supply chain design. Surrogate models building has been systematized into four steps, i.e., design of experiments (DOE), running the rigorous models at the sampling points designated by the DOE, construction of the surrogate model, and validation of the model obtained (Queipo et al., 2005).

Polynomial regression models, in which the relationship between the variables is expressed using a polynomial function, are one of the most basic types of surrogate models. In the case of polynomial regression models involving multiple variables, the optimal variables to be addressed

within the pool of variables considered can be determined by using machine learning-based tools such as ALAMO (Wilson & Sahinidis, 2017), ensuring an optimal trade-off between model accuracy and complexity. Other types of surrogate models are Kriging models, which estimate the relationship between variables as a sum of a linear model and a a stochastic Gaussian function representing the fluctuations of data (Quirante et al., 2015), and artificial neural networks (ANN), which are based on generating an input signal as the summation of all the weighted inputs, which is through nodes containing a transfer function. Nodes are connected by edges with assigned weights that adjust the signals transmitted between nodes. Nodes are structured in layers, in a way that nodes receive signals from nodes of the preceding layer, and if the output of the node is above a threshold value defined by the transfer function, sends the output signal to the next layer (Himmelblau, 2000).

1.2.6 Experimental correlations

As the surrogate models, experimental correlations are models built using data of the systems represented, but converserly to those one, experimental correlations are built using data from experimental results. Similarly to the rules of thumb, the accuracy of these models is limited by the availability of data, and they are only applicable to the range of operating conditions of the data used for constructing the model.

APPROACHES FOR DECISION-SUPPORT SYSTEMS

Decision-making activities require to analyze multiple relevant criteria for each course of action. Since criteria often conflict each other, each decision-making process requires the balancing of criteria, prioritizing some criteria over other through the use of some criteria weighting scheme. This procedure requires managing a vast amount of information of conflicting nature, leading to a complex decision-making process. Therefore, different approaches generally called multiple-criteria decision analysis (MCDA) have been developed to explicitly structure and solve decision problems. MDCA aim is to integrate criteria assessment with value judgment to analyze and compare the different available alternatives, identifying the best solution for the specific decision-making context studied. However, it must be highlighted that a certain grade of subjectivity might exist is several steps of MCDA, such as the choice of the set of criteria considered relevant for a particular problem. Therefore, the solution proposed by any MCDA approach must be analyzed considering the assumptions made

for building the problem. In sum, MCDA seeks to structure problems with multiple conflicting criteria, and providing justifiable and explainable solutions to guide decision-makers facing such problems. The solution of a multiple-criteria decision-making problem can be defined as a unique solution representing the most suitable alternative from the set of potential alternatives, or as a subset of satisfactory alternatives (Belton & Stewart, 2002).

An MCDA problem can be articulated in different stages, starting with the problem definition and structuring. At this stage, the goals, constraints, and stakeholders comprising the problem are defined, as well as the different solution alternatives. Based on this information, a model can be built for the assessment and comparison of alternatives. This stage includes the definition of the relevant criteria used for alternatives comparison, their relative priority, and the system for criteria evaluation. Finally, the information retrieved by the model can be used for making informed decisions.

Multi-criteria decision-making problems can be classified into Multi-Attribute Decision Analysis (MADA), which are discrete choice problems where the number of alternatives is finite, and Multi-Objective Decision Analysis (MODA), that are mathematical programming problems that consider infinite number of alternatives, as shown in Figure 1.3. However, we note that mathematical programming techniques are not limited to formulating and solving problems with infinite alternatives, but they can also be used for dealing with discrete decision-making problems (Giove et al., 2009).

1.3.1 Multi-Attribute Decision Analysis (MADA)

In the case of problems consisting of a finite number of alternatives, the suitability of each alternative to the problem given can be measured through its performance according to the multiple criteria considered. A large number of MCDA approaches have been (and are currently being) developed for discrete choice problems, including methods based on value functions (Multi-Attribute Value Theory methods, MAVT) and outranking methods.

1.3.1.1 Multi-Attribute Value Theory (MAVT)

INDICATOR-BASED METHODS Multi-Attribute Value Theory (MAVT) approaches are based on an indicator-based methodology for alternatives comparison. The relevant criteria considered in the decision-making process are normalized to a common scale to allow criteria comparison using

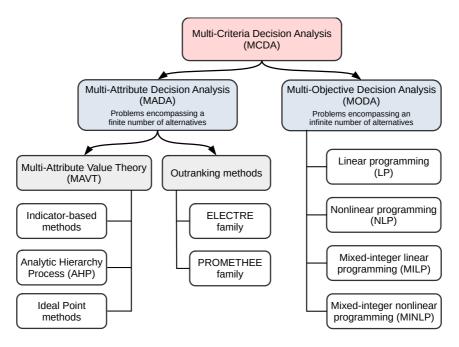


Figure 1.3: Classification of MCDA methods.

an utility or value function. A number of utility functions have been proposed in the literature, including standardization, min-max, and target utility functions (OECD and European Commission, 2008). The normalized criteria are weighted and aggregated to build a composite index, prioritizing some criteria over others. Different aggregation schemes have been proposed, providing different degrees of compensability between indicators, i.e. a deficit in one criteria can be fully, partially, or not compensated by a surplus in other criteria (Gasser et al., 2020). Additive weighting aggregation is a full compensatory method, while geometric and harmonic aggregation methods are partial compensation schemes. Other aggregation schemes include geometric averaging, which is a non-compensatory method, and the Choquet integral (Marichal & Roubens, 2000). The composite index obtained is a single numerical value that can be used to score and rank the proposed alternatives based on their suitability to the criteria considered.

A major source of uncertainty in indicator-based methods is the value of criteria weights. This issue can be addressed using the stochastic multicriteria acceptability analysis (SMAA) method. SMAA is a sensitivity analysis method that address the uncertainty of criteria weights value exploring the feasible space of weights through the Monte Carlo method. Further, details about the SMAA approach can be found in Tervonen and Lahdelma (2007).

In this thesis, and indicator-based methodology has been used to assess and select phosphorus recovery technologies based on technical, environmental, and economic criteria combined in a composite index, as it is shown in Chapter .

ANALYTIC HIERARCHY PROCESS (AHP) Analytic Hierarchy Process (AHP) decomposes the decision problem into multiple simpler sub-problems. These sub-problems are hierarchized and independently analyzed. The sub-problems are solved through the pairwise comparison of alternatives, obtaining numerical indexes that can be used to compare their performance. Finally a numerical weight (priority) is assigned to each element of the hierarchy, and they are used for aggregating the indexes obtained by each alternative at each element of the hierarchy in a final numerical value that can be used to score the overall performance of each alternative accordingly to the set of criteria considered (Saaty, 2000).

IDEAL POINT METHODS Ideal Point methods set an optimal solution, that represent a utopia point where all criteria values are optimal. The performance of each alternative is evaluated through a composite index, that can be constructed using the MAVT approach. The alternatives are ranked based on their relative distance relative to the optimal solution. One of the most common Ideal Point methods is TOPSIS (Hwang & Yoon, 1995).

1.3.1.2 Outranking methods

Outranking methods are based on the pairwise comparison of the alternatives for each criterion considered, determining the preferred alternative for each of the criteria. Preference information about all criteria is aggregated to establish evidence for selecting one alternative over another. These methods indicate the dominance of one alternative over another, but they do not quantify the performance gap between the alternatives compared (Giove et al., 2009). Some of the most popular outranking methods are ELECTRE I (Roy, 1968), II (Roy & Bertier, 1973), and III (Roy et al., 1978), and PROMETHEE (Vincke & Brans, 1985).

1.3.1.3 Multi-Objective Decision Analysis (MODA)

Problems consisting of an infinite number of solutions require multiobjective mathematical programming (optimization) techniques to be solve. These problems are subjected to a number of equality and/or inequality constraints restricting the solutions that are feasible. The multiple conflicting criteria are combined in an objective function. This objective function represents the improving level of the criteria, and it will be minimized or maximized for selecting the best solution that represents the optimal trade-off between the different conflicting criteria (Giove et al., 2009). In this thesis, this technique has been employed for determining the operating conditions of processes for the recovery of nutrient, energy and biomethane from livestock waste, as it is shown in Chapters 3 and 7. Other approach for solving multi-objective mathematical programming problems is to set a priori targets for different criteria, or combinations of criteria, that are considered satisfactory, obtaining the problem solution by minimizing the deviations from these goals. Mathematical programming problems can be also classified according to the use of linear or nonlinear equations, and continuous and/or discrete variables (Giove et al., 2009).

LINEAR PROGRAMMING (LP) Linear programming (LP) refers to those mathematical programming problems based on linear equations and continuous variables. A linear programming problem can be expressed as shown in Eq. 1.1, where x is a vector of dimension n, A is a $m \times n$ matrix, cis the *n* dimension vector of cost coefficients, and the right-hand side *b* is a vector of dimension m (Grossmann, 2021).

min
$$Z = c^T x$$

s.t. $Ax \le b$ (1.1)
 $x \ge 0$

The two most widely used methods to solve LP problems are the Simplex algorithm (Murty, 1983) and interior-point methods (Potra & Wright, 2000). The Simplex method is more efficient for solving problems with thousands of variables and constraints, while interior-point is more efficient on very large scale and sparse problems (Grossmann, 2021). These methods are implemented in solvers such as CPLEX (IBM, 2009), Gurobi (Gurobi Optimization, LLC, 2021), or XPRESS (FICO, 2021).

NONLINEAR PROGRAMMING (NLP) Nonlinear programming (NLP) refers to those mathematical programming problems containing nonlinear equations, either in the constrains or in the objective function, and continuous variables. A nonlinear programming problem can be expressed as shown in Eq. 1.2, where x is an n dimension vector, f(x) is the objective function of the problem, h(x) is the set of equality constraints and g(x) is the set of inequality constraints (Floudas, 1995).

min
$$f(x)$$

s.t. $h(x) = 0$
 $g(x) \le 0$
 $x \in X \subset \Re^n$

$$(1.2)$$

Some of the most common algorithms to solve NLP problems are successive quadratic programming (SQP), reduced gradient algorithms, and interior point methods.

SQP algorithms are based on the solution of quadratic programming subproblems. Each subproblem optimizes a quadratic model of the objective function subject to linearized constraints. In each of the iterations a search direction is determined reducing some merit function to ensure problem convergence (Gill et al., 2005). SNOPT is a solver based on this method (Gill et al., 2005).

Reduced gradient methods consider a linear approximation of the constraints and eliminate variables to reduce the dimension of the problem. The resulting problem is solved by applying the Newton's method. In each of the iterations, the reduced gradient is calculated, the search direction is determined, and finally a line search is performed minimizing the objective function. MINOS (Murtagh & Saunders, 1983) or CONOPT (Drud, 1985) are solvers based on this algorithm.

Interior point methods reformulate the original NLP problem by means of slack variables to replace the inequalities by equalities and the log-barrier function to handle the non-negativity of the *x* variables. The new problem is solve applying the Newton's method. IPOPT Wächter and Biegler, 2006 and KNITRO (Waltz & Nocedal, 2004) are solvers based on this approach

MIXED-INTEGER LINEAR PROGRAMMING (MILP) Mixed-integer linear programming (MILP) refers to those mathematical programming problems based on linear equations and containing discrete variables. A mixed-integer linear programming problem can be expressed as shown in Eq. 1.3, where *x* are continuous variables and *y* are discrete variables. Typically, discrete variables are binary variables (Grossmann, 2021).

min
$$Z = a^T x + b^T y$$

s.t. $Ax + By \le d$ (1.3)
 $x \ge 0$ (1.4)
 $y \in \{0,1\}^m$

A number of methods have been proposed to solve MILP problems, including cutting planes, Benders decomposition, branch and bound search, and branch and cut methods.

Cutting planes consist of a sequence of LP problems in which different cutting planes are generated to cut-off the solution of the relaxed LP. They reduce the feasible region of the linear relaxation of the original problem excluding those solutions that are feasible in the linear relaxation but not in the original MILP problem.

Benders decomposition is based on the generation of a lower and an upper bound of the solution of the MILP problem in each iteration. The upper bound is calculated from the primal problem, which correspond with the original problem where the binary variables have been fixed. Conversely, the lower bound is determined through a master problem, which is a LP problem derived from the original problem by means of the duality theory. Branch and bound method structure the problem in form of a binary tree that includes all possible combinations of binary variables. The tree is explored by solving the relaxed versions of the original problem. If the relaxation does not result in an integer solution (o or 1), it is necessary to go deeper into the solution tree to explore further combinations of the binary variables. If the result obtained is an integer, the next step is to return to the previous subproblem to explore the alternative branch. However, diverse procedures have been developed to discard certain branches, avoiding the need of exploring the whole tree and reducing the problem (Floudas, 1995).

Branch and cut methods combine branch and bound methods with cutting planes targeting a tighter lower bound. In the different nodes, the relaxed problem is solved. If the solution is not integer, the relaxing problem is solved by adding cutting planes in order to strengthen the lower bound (Grossmann, 2021). Gurobi (Gurobi Optimization, LLC, 2021) and CPLEX (IBM, 2009) are solvers based on this approach.

MIXED-INTEGER NONLINEAR PROGRAMMING (MINLP) Mixed-integer nonlinear programming (MINLP) refers to those mathematical programming problems containing nonlinear equations and discrete variables, typically, binary variables. A mixed-integer nonlinear programming problem can be expressed as shown in Eq 1.5, where x represents a vector of continuous variables, y is the vector of binary variables, h(x, y) and g(x, y)denote the equality and inequality constraints respectively. f(x) represents the objective function (Grossmann, 2021).

min
$$f(x,y)$$

s.t. $h(x,y) = 0$
 $g(x,y) \le 0$
 $x \in X \subseteq \Re^n$
 $y \in \{0,1\}^m$ (1.5)

Some algorithms for solving MINLP problems are the generalized Benders decomposition, outer approximation, and generalized cross decomposition.

Generalized Benders decomposition (GBD) is based on the generation of a lower and an upper bound of the solution of the MINLP problem in each iteration. Similarly to the Benders decomposition, the upper bound is calculated from the primal problem, which correspond with the original problem where the binary variables have been fixed. The lower bound is determined through a master problem, which is a LP problem derivated from the original problem by means of the duality theory. In addition, the master problem provides information about the binary variables to be fixed in the next iteration (Floudas, 1995).

Outer approximation (OA) provides a lower and an upper bound in each iteration. As the previous case, the upper bound is calculated from the primal problem. The lower bound is calculate from a master problem obtained based on an outer approximation, i.e., the nonlinear objective function and the constraints are linearized around the primal solution. Additionally, the master problem provides information about the binary variables to be fixed in the next iteration.

Generalized cross decomposition (GCD) is based on the generation of a primal problem that provides an upper bound of the solution and also the Lagrange multipliers for the dual subproblem. The dual problem is used to determine the lower bound of the problem, and provides a vector of binary variables to be fixed in the primal problem. The solution of the primal and dual problems go through convergence tests. If any of these test fails, a master problem is solved. This approach seeks to minimize the number of master problems to be solved since the computational requirements of the this problem are higher. This procedure is repeated at each iteration of the algorithm (Floudas, 1995).

1.4 APPROACHES FOR GEOSPATIAL ENVIRONMENTAL ASSESSMENT

The development of mitigation measures to reduce the environmental footprint of anthropogenic activities requires the previous understanding and quantification of the environmental impacts associated to each sector. This process, called environmental impact assessment (EIA), involves the analysis of multi-disciplinary information, including environmental, physical, geological, ecological, economic, and social data (Gharehbaghi & Scott-Young, 2018). Since EIA aims to evaluate the environmental impact of an activity on a particular geographical location, all these data have a common geographic component, becoming geospatial data.

Geospatial data can be managed and analyzed through specific systems denoted as geographic information system (GIS). GIS is a key tool for EIA that uses the geographic component of geospatial data as an integrative framework that provides the ability to analyze and map the descriptive information of the locations studied. The geographic component of data is the key element of GIS systems, since the spatial (or spatio-temporal) location is used as a key to relate other descriptive information. From the perspective of EIA, this information can be analyzed, interpreted, and mapped in order to determine the vulnerability level to a particular environmental threat at each location, find relationships between human activities and environmental damages, measure the performance of mitigation and remediation processes, etc.

As a result, the combination of GIS, EIA, and methods for the analysis of multi-dimensional information, such as MCDA, provides tools for the development of strategies to promote the transition to a sustainable paradigm for human growth. In this regard, the development of a sustainable, reliable, and resilient water, energy and food nexus is a major issue for food security and environmental protection.

THESIS OUTLINE

This dissertation is structured in three parts. Part I is devoted to the study of phosphorus management and recovery, Part II addresses a techno-economic assessment of the technologies for nitrogen recovery, and Part III conducts a techno-economic analysis for determining the optimal biomethane production process in order to integrate biogas production and nutrient recovery processes.

Part I - Phosphorus management and recovery

CHAPTER 3 - TECHNOLOGIES FOR PHOSPHORUS RECOVERY. chapter performs a review of the main processes for phosphorus recovery from livestock waste, identifying the most promising processes to be deployed at CAFOs using a mixed-integer nonlinear programming model.

CHAPTER 4 - ASSESSMENT OF PHOSPHORUS RECOVERY THROUGH STRUVITE PRECIPITATION. This chapter studies the mitigation of phosphorus releases through the deployment of struvite precipitation systems in the watersheds of the contiguous Unites States. Specific surrogate models to predict the production of struvite and calcium precipitates from cattle leachate were developed based on a detailed thermodynamic model. In addition, the variability in the organic waste composition is captured through a probability framework based on the Monte Carlo method.

CHAPTER 5 - GEOSPATIAL ENVIRONMENTAL AND TECHNO-ECONOMIC FRAMEWORK FOR SUSTAINABLE PHOSPHORUS MANAGEMENT AT LIVE-STOCK FACILITIES. This chapter presents a decision support framework, COW2NUTRIENT (Cattle Organic Waste to NUTRIent and ENergy Technologies), for the assessment and selection of phosphorus recovery technologies at CAFOs based on environmental information on nutrient pollution and techno-economic criteria. This framework combines eutrophication risk data at subbasin level and the techno-economic assessment of six state-of-the-art phosphorus recovery processes in a multi-criteria decision analysis (MCDA) model. We aimed to provide a useful framework for the selection of the most suitable P recovery system for each particular CAFO, and for designing and evaluating effective GIS-based incentives and regulatory policies to control and mitigate nutrient pollution of waterbodies.

CHAPTER 6 - ANALYSIS OF INCENTIVE POLICIES FOR PHOSPHORUS This chapter conduct a research on the design and analysis of incentive policies using the COW2NUTRIENT framework for the implementation of phosphorus recovery technologies at CAFOs minimizing the negative impact in the economic performance of CAFOs. Moreover, the fair allocation of monetary resources when the available budget is limited is studied using the Nash allocation scheme.

1.5.2 Part II - Nitrogen management and recovery

CHAPTER ?? - MULTI-SCALE TECHNO-ECONOMIC ASSESSMENT OF NI-TROGEN RECOVERY SYSTEMS FOR SWINE OPERATIONS. evaluates the main processes for nitrogen recovery at intensive swine operations. A multi-scale techno-economic analysis is performed to estimate the capital and operating costs for different treatment capacities, identifying the most promising processes.

1.5.3 Part III - Nitrogen management and recovery

CHAPTER 7 - OPTIMAL TECHNOLOGY SELECTION FOR THE BIOGAS This chapter performs a systematic UPGRADING TO BIOMETHANE. study of different biogas upgrading to biomethane processes in order to identify the optimal process attending to the particular characteristics of the biogas produced from livestock manure. Food waste and wastewater sludge are also included for comparison. We aimed to determine the optimal biomethane production processes for the potential combination of biomethane production and nutrient recovery processes into an integrated resources recovery facility.

BIBLIOGRAPHY

- Amery, F., & Schoumans, O. F. (2014). Agricultural phosphorus legislation in Europe. Institute for Agricultural; Fisheries Research (ILVO); Alterra Wageningen UR.
- Anderson, J. D., & Wendt, J. (1995). Computational fluid dynamics (Vol. 206). Springer.
- Belton, V., & Stewart, T. (2002). Multiple criteria decision analysis: an integrated approach. Springer Science & Business Media.
- BIOGAS₃ PROJECT. (2021). BIOGAS₃ PROJECT Sustainable small-scale biogas from agri-food waste for energy self-sufficiency [[Online; accessed 28-September-2021]].
- Biraben, J. N. (1980). An essay concerning mankind's demographic evolution. Journal of Human Evolution, 9(8), 655–663.
- Bouwman, A. F., Beusen, A. H. W., & Billen, G. (2009). Human alteration of the global nitrogen and phosphorus soil balances for the period 1970–2050. Global Biogeochemical Cycles, 23(4). https://doi.org/doi. org/10.1029/2009GB003576
- Brignole, E. A., & Pereda, S. (2013). *Phase equilibrium engineering*. Newnes. Burns, R. T., & Moody, L. B. (2002). Phosphorus recovery from animal manures using optimized struvite precipitation. Proceedings of Coagulants and Flocculants: Global market and technical opportunities for water treatment chemicals, 1-4.
- Buzzi-Ferraris, G., & Manenti, F. (2009). Kinetic models analysis. Chemical Engineering Science, 64(5), 1061–1074.
- CNP. (2021). Review of Applicability of EPA's Part 503 Biosolids Rule on Phosphorus Minerals recovered at Water Resource Recovery Facilities (tech. rep.). CNP.

- Comission, E. (2020). A new Circular Economy Action for a cleaner and more competitive Europe (tech. rep.). European Comission.
- Constable, G., & Somerville, B. (2003). A century of innovation: Twenty engineering achievements that transformed our lives. Chapter 7, Agricultural mechanization. Joseph Henry Press.
- Cordell, D., Drangert, J.-O., & White, S. (2009). The story of phosphorus: global food security and food for thought. Global environmental change, 19(2), 292-305.
- Couper, J. R., Penney, W. R., Fair, J. R., & Walas, S. M. (2005). Chemical process equipment: selection and design. Gulf professional publishing.
- Drud, A. (1985). CONOPT: A GRG code for large sparse dynamic nonlinear optimization problems. *Mathematical programming*, 31(2), 153–191.
- ESSP. (2021). ESPP input for the EU's "Integrated Nutrient Management Action Plan" (INMAP) (tech. rep.). European Sustainable Phosphorus Platform.
- European Comission. (2021). Biogas Action [[Online; accessed 28-September-
- FICO. (2021). Xpress-Optimizer Reference manual, 20.0 edition. http:// www.fico.com/xpress
- Floudas, C. A. (1995). Nonlinear and mixed-integer optimization: fundamentals and applications. Oxford University Press.
- Gasser, P., Suter, J., Cinelli, M., Spada, M., Burgherr, P., Hirschberg, S., Kadziński, M., & Stojadinović., B. (2020). Comprehensive resilience assessment of electricity supply security for 140 countries. Ecological Indicators, 109. https://doi.org/10.1016/j.ecolind.2019.105731
- Gharehbaghi, K., & Scott-Young, C. (2018). GIS as a vital tool for Environmental Impact Assessment and Mitigation, In IOP Conference Series: Earth and Environmental Science. IOP Publishing.
- Gill, P. E., Murray, W., & Saunders, M. A. (2005). SNOPT: An SQP algorithm for large-scale constrained optimization. SIAM review, 47(1), 99–131.
- Giove, S., Brancia, A., Satterstrom, F. K., & Linkov, I. (2009). Decision support systems and environment: Role of MCDA. In Decision support systems for risk-based management of contaminated sites (pp. 1-21). Springer.
- Grizzetti, B., Vigiak, O., Udias, A., Aloe, A., Zanni, M., Bouraoui, F., Pistocchi, A., Dorati, C., Friedland, R., De Roo, A., Benitez Sanz, C., Leip, A., & Bielza, M. (2021). How EU policies could reduce nutrient pollution in European inland and coastal waters. Global Environmental Change, 69, 102281.
- Grossmann, I. E. (2021). Advanced Optimization for Process Systems Engineering. Cambridge University Press.

- Gurobi Optimization, LLC. (2021). Gurobi Optimizer Reference Manual. https://www.gurobi.com
- Hall, S. M. (2012). *Rules of thumb for chemical engineers*. Butterworth-Heinemann.
- Himmelblau, D. M. (2000). Applications of artificial neural networks in chemical engineering. Korean journal of chemical engineering, 17(4), 373-392.
- Hwang, C., & Yoon, K. (1995). Multi attribute decision making-an introduction. Sage University Papers.
- IBM. (2009). V12.1: User's Manual for CPLEX (tech. rep.).
- Kellogg, R. L., Lander, C. H., Moffitt, D. C., & Gollehon, N. (2000). Manure Nutrients Relative to the Capacity of Cropland and Pastureland to Assimilate Nutrients: Spatial and Temporal Trends for the United States (tech. rep.). U.S. Department of Agriculture.
- Loeppert, R. H., Schwab, A. P., & Goldberg, S. (1995). Chemical equilibrium and reaction models. Soil Science Society of America.
- Marichal, J.-L., & Roubens, M. (2000). Determination of weights of interacting criteria from a reference set. European journal of operational Research, 124(3), 641-650.
- Martín, M., & Grossmann, I. E. (2012). BIOpt: A library of models for optimization of biofuel production processes. In I. D. L. Bogle & M. Fairweather (Eds.), 22nd European Symposium on Computer Aided Process Engineering (pp. 16–20). Elsevier.
- Murtagh, B. A., & Saunders, M. A. (1983). MINOS 5.0 User's Guide. (tech. rep.). Stanford Univ CA Systems Optimization Lab.
- Murty, K. G. (1983). Linear programming. Springer.
- NACWA. (2014). Issue Outline on Resource Recovery from Wastewater and Coverage of 40 C.F.R. Part 503 (tech. rep.). National Association of Clean Water Agencies.
- Nierenberg, D., & Mastny, L. (2005). Happier meals: Rethinking the global meat industry (Vol. 171). Worldwatch Institute.
- OECD and European Commission. (2008). Handbook on Constructing Composite Indicators. Methology and User Guide. OECD.
- Piot-Lepetit, I. (2011). *Agricultural externalities and environmental regulation*: The case of manure management and spreading land allocation. Bentham Science Publishers.
- Pistikopoulos, E. N., Barbosa-Povoa, A., Lee, J. H., Misener, R., Mitsos, A., Reklaitis, G. V., Venkatasubramanian, V., You, F., & Gani, R. (2021). Process systems engineering – The generation next? *Computers &* Chemical Engineering, 147, 107252.
- Potra, F. A., & Wright, S. J. (2000). Interior-point methods. *Journal of Compu*tational and Applied Mathematics, 124(1), 281-302.

- Queipo, N. V., Haftka, R. T., Shyy, W., Goel, T., Vaidyanathan, R., & Tucker, P. K. (2005). Surrogate-based analysis and optimization. Progress in aerospace sciences, 41(1), 1-28.
- Quirante, N., Javaloyes, J., & Caballero, J. A. (2015). Rigorous design of distillation columns using surrogate models based on K riging interpolation. AIChE Journal, 61(7), 2169–2187.
- Ramkrishna, D. (2000). Population balances: Theory and applications to particulate systems in engineering. Elsevier.
- Roy, B. (1968). Classement et choix en présence de points de vue multiples. Revue française d'informatique et de recherche opérationnelle, 2(8), 57-75.
- Roy, B. Et al. (1978). ELECTRE III: Un algorithme de classements fondé sur une représentation floue des préférences en présence de criteres multiples. Cahiers Centre d'Etudes de Recherche Opérationnelle.
- Roy, B., & Bertier, P. (1973). La Méthode ELECTRE II: Une application au média-planning.
- Saaty, T. L. (2000). Fundamentals of decision making and priority theory with the analytic hierarchy process (Vol. 6). RWS publications.
- Sadhukhan, J., Ng, K. S., & Hernandez, E. M. (2014). Biorefineries and chemical processes: design, integration and sustainability analysis. John Wiley & Sons.
- Samreen, S., & Kausar, S. (2019). Phosphorus Fertilizer: The Original and Commercial Sources. IntechOpen. https://doi.org/10.5772/ intechopen.82240
- Smil, V. (1999). Detonator of the population explosion. *Nature*, 400(6743), 415-415.
- Stephanopoulos, G., & Reklaitis, G. V. (2011). Process systems engineering: From Solvay to modern bio- and nanotechnology.: A history of development, successes and prospects for the future. Chemical Engineering Science, 66(19), 4272-4306.
- Szirtes, T. (2007). Applied dimensional analysis and modeling. Butterworth-Heinemann.
- Tervonen, T., & Lahdelma, R. (2007). Implementing stochastic multicriteria acceptability analysis. European Journal of Operational Research, 178(2), 500-513. https://doi.org/10.1016/j.ejor.2005.12.037
- Tilley, E., Ulrich, L., Lüthi, C., Reymond, P., & Zurbrügg, C. (2014). Compendium of sanitation systems and technologies. Eawag.
- United Nations, Department of Economic and Social Affairs. (2019). World Population Prospects 2019 [[Online; accessed 22-August-2021]].
- US EPA. (2020a). Elements of an NPDES Permit for a CAFO. NPDES Permit Writers' Manual for CAFOs (tech. rep.). U.S. Environmental Protection Agency.
- US EPA. (2020b). NPDES Permit Basics [[Online; accessed 24-August-2021]].

- US EPA. (2021). AgSTAR: Biogas Recovery in the Agriculture Sector [[Online; accessed 28-September-2021]].
- USDA. (2009). Agricultural Wate Management Field Handbook (tech. rep.). U.S. Department of Agriculture.
- USDA. (2011). Animal Feeding Operations (AFO) and Concentrated Animal Feeding Operations (CAFO) [[Online; accessed 10-August-
- Vincke, J. P., & Brans, P. (1985). A preference ranking organization method. The PROMETHEE method for MCDM. Management Science, 31(6), 647-656.
- Wächter, A., & Biegler, L. T. (2006). On the implementation of an interiorpoint filter line-search algorithm for large-scale nonlinear programming. Mathematical programming, 106(1), 25-57.
- Waltz, R. A., & Nocedal, J. (2004). KNITRO 2.0 User's Manual. Ziena Optimization, Inc.[en ligne] disponible sur http://www. ziena. com (September, 2010), 7, 33-34.
- Wilson, Z. T., & Sahinidis, N. V. (2017). The ALAMO approach to machine learning. Computers & Chemical Engineering, 106, 785–795.