

Local Climate Zone classification using random forests: A case study of Hong Kong

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1 Introduction

1.1 Background

Urban Heat Islands, where urban areas are warmer than the neighboring rural areas, are a major concern as the population worldwide increases and urbanizes (United Nations et al., 2019). This is for four primary reasons: increased energy consumption, elevated emissions of air pollutants and greenhouse gasses, compromised human health and comfort, and impaired water quality (US EPA, 2014). A multitude of processes combine to cause Urban Heat Islands. Broadly urban built structures hold more heat than the structures and vegetation in surrounding areas (Hibbard et al., 2017). In particular, cities have an overabundance of impervious surfaces, along with a lack of vegetation (Hibbard et al., 2017). Knowledge about microclimates within cities, especially in the context of prospective sites for climate risk adaptation efforts, has the potential to ease some of these disturbances (Lempert et al., 2018) .

Unfortunately, information on these sites can be difficult to obtain. Though there is a plethora of satellite imagery, the current methods to classify this imagery are not always accurate. (Yokoya et al., 2018). Historically the focus was on the broad categories of urban

vs. rural, which does not give enough information about the nuanced climate within a large urban area (Stewart & Oke, 2012). Local Climate Zone (LCZ) classification was created by Stewart and Oke (2012) (Figure 1) to alleviate this problem, but it often requires a significant investment from individuals with specialist knowledge to successfully classify a city (Bechtel et al., 2015). Therefore, machine learning methods that can classify LCZ efficiently and accurately are of great interest right now.

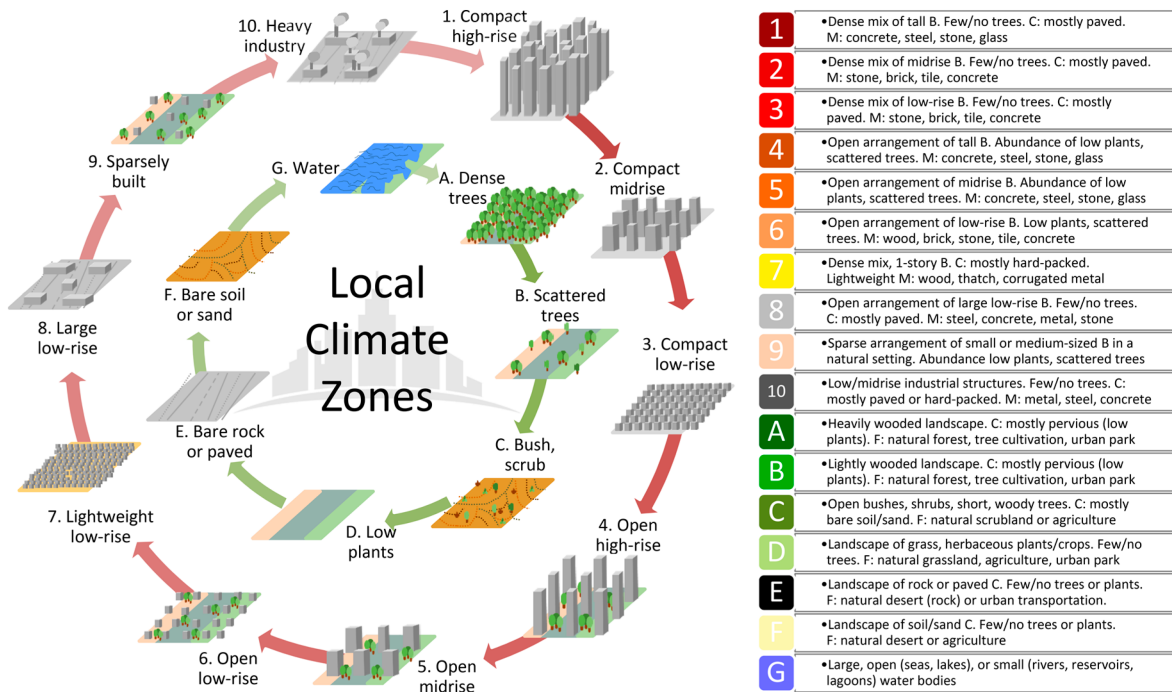


Figure 1: Local Climate Zone classes. Originally from Stewart and Oke (2012) and remade by Bechtel et al. (2017). Copyright CC-BY 4.0

1.2 Objective

The goal of this project is to recreate aspects of the article “Comparison between convolutional neural networks and random forest for local climate zone classification in mega urban areas using Landsat images” (Yoo et al., 2019), where methods for predicting LCZ classes

for four large cities throughout the world were compared. To do so, a small training dataset from the 2017 Institute of Electrical and Electronics Engineers (IEEE) Geoscience and Remote Sensing Society (GRSS) Data Fusion Contest (Tuia et al., 2017) was used as ground truth for LCZ classes. This was combined with Landsat 8 satellite input data to create a series of models, which were then compared to a larger, full LCZ layer for each city and assessed for accuracy. The primary types of models considered in the Yoo et al. (2019) work were random forests and convolutional neural networks. However, for this project the focus will be only on random forests. In addition, rather than four cities, this investigation will focus on just Hong Kong. This city was chosen because each LCZ that is classified has at least 4 polygons. Finally, here, a classification scheme like the one used by the World Urban Database and Access Portal Tools (WUDAPT) project, denoted as Scheme 1 (S1) in Yoo et al. (2019) will be the focus, with comparisons between accuracy using a variety of tuning parameters.

All code and higher resolution images for this project can be found on GitHub at <https://github.com/erickabsmith/masters-project-lcz-classification>.

2 Methods

2.1 Data

The LCZ reference data from the contest set was taken from the WUDAPT database, checked for correctness, and then provided as a 100m resolution raster layer. The LCZ classes are numbered 1-17 in the layer, rather than 1-10 and A-G. The first step I took was randomly

dividing the polygons of each class in the reference data into training and testing groups. This required some preprocessing since I only had access to the data in raster form and without any sort of polygon identification. Pixels within the same polygon cannot be put into both training and validation groups because it will artificially increase accuracy metrics (Zhen et al., 2013).

The Landsat 8 data I used includes four different dates, or “scenes,” each of which were downloaded from the USGS EarthExplorer portal (Table 1). Then the atmospheric (band 9) and panchromatic (band 8) bands were left out, and the data were resampled using the area weighted average to 100 m x 100 m grids (Yokoya et al., 2018). To get these data into a usable format they were loaded into R by band (each band is a separate raster layer) and then stacked with the LCZ reference data and converted to a dataframe, ready to be used as input data. This slightly differs to Yoo et al. (2019), who collected their own Landsat data.

Table 1: Acquisition Dates of Each Landsat 8 Scene

| Scene | Date |
|-------|-------------|
| 1 | 29-Nov-2013 |
| 2 | 15-Oct-2014 |
| 3 | 16-Nov-2014 |
| 4 | 18-Oct-2015 |

All 9 available bands of all 4 Landsat scenes amounted to 36 input variables. Each pixel is an observation, with 279,312 pixels total over the almost 2,800 square kilometers that are contained within the area of interest. The LCZ reference data has 179 polygons, which cover 8,846 pixels. This is approximately 88.5 square kilometers and 3% of the total area of interest. The LCZ reference data was randomly split into training and test sets by LCZ class and polygon. The resulting groups are not exactly equal in pixel numbers, but I determined

that some aspect of randomization of polygons takes precedence over equality in pixels due to the potential for artificially high accuracy values, as mentioned before. In order to get as close as possible a barplot was made showing the number of pixels in training and test, by class, so that equality in pixel number could be explored without having knowledge of polygon groupings. Sampling was repeated until these bars looked marginally even. (Table 2).

Table 2: Delineation of training and test data by polygon and pixel.

| Local Climate Zone | Train | Test |
|-------------------------------|----------|----------|
| Class 1: Compact high-rise | 13 (295) | 13 (336) |
| Class 2: Compact mid-rise | 6 (117) | 5 (62) |
| Class 3: Compact low-rise | 7 (185) | 7 (141) |
| Class 4: Open high-rise | 10 (275) | 9 (398) |
| Class 5: Open mid-rise | 4 (79) | 4 (47) |
| Class 6: Open low-rise | 6 (60) | 7 (60) |
| Class 7: Lightweight low-rise | 0 (0) | 0 (0) |
| Class 8: Large low-rise | 4 (90) | 5 (47) |
| Class 9: Sparsely built | 0 (0) | 0 (0) |
| Class 10: Heavy Industry | 4 (107) | 5 (112) |
| Class 11: Dense trees | 7 (762) | 7 (854) |
| Class 12: Scattered trees | 6 (194) | 7 (213) |
| Class 13: Bush, scrub | 4 (459) | 5 (232) |
| Class 14: Low plants | 6 (346) | 6 (222) |
| Class 15: Bare rock or paved | 0 (0) | 0 (0) |
| Class 16: Bare soil or sand | 0 (0) | 0 (0) |
| Class 17: Water | 5 (1266) | 5 (1113) |

^a Number of polygons is listed first, with number of pixels in parentheses.

2.2 Random Forests

Random forests consist of many decision trees. A decision tree can be used for classification or regression. Here I will focus on classification since my goal is to predict LCZ class, which is a categorical variable. Decision trees put each observation through a series of conditional statements, or splits, in order to group it with other observations that are similar. These similar groups are expected to have similar values for the predictor variable. Since the true value of the predictor variable is known for the observations in the training dataset, it is

possible to measure the accuracy of each prospective split, and therefore to build an optimal tree.

Each point in which the data could potentially be split into two groups is called a node. The first split is called the root node. Nodes are selected such that each one uses the conditional statement which subsets the data into the best possible split. When any more subsetting does not increase accuracy, the path ends, and this is called a leaf node. Any nodes between the root node and leaf nodes are called internal nodes.

Splits are typically evaluated by Gini impurity or entropy.

$$\text{Gini Impurity} = I_G(t) = 1 - \sum_{i=1}^C p(i|t)^2$$

$$\text{Entropy} = I_H(t) = - \sum_{i=1}^C p(i|t) \log_2 p(i|t)$$

Where i is a class in the predictor variable, ranging from 1 to C . C is the total number of classes represented for a particular node, t . $p(i|t)$ is the proportion of samples that belong to each i , for a particular node t .

A Gini Impurity, or entropy, of 0 indicates a completely homogeneous group, which cannot be improved upon. These metrics are used for comparison: all possible variables and thresholds within those variables are tried and their Gini impurity or entropy calculated, the variable and threshold with the best (lowest) value selected for the node. This is done recursively at each node until no split offers a decrease in the metric of choice. The resulting collection of nodes is a decision tree.

Decision trees perform poorly with new samples. Including a threshold value for the accu-

racy metric can help, as it keeps trees more simple, but they are still prone to overfitting. Collecting them into a random forest addresses this issue. True to its name, a random forest is a collection of decision trees that have a component of randomness. The predictive aspect of a random forest does not only rely on one tree, it is determined by totaling up the decisions, or “votes,” that all of the trees make.

Each tree is created individually and starts with bootstrapping the training dataset. Bootstrapping resampling is done with replacement, and results in about $1/3$ of the observations being excluded from the sample. For a given tree, the observations included in the sample are “inbag,” while the $1/3$ of observations that are not selected are called “out-of-bag” (OOB). These groupings allow for validation of the model without use of an external dataset.

Then the process is similar to the one described above for decision trees. However, for each node only a subset of the variables are randomly selected and used as candidates to split up the bootstrapped data. Not all the variables are used at each node because doing so reduces correlation and introduces randomness into the model. Whichever variable best splits the data will be the variable that is kept at that node in that specific tree. This is either done recursively until no more splits are beneficial (just as in a regular decision tree), or until minimum size of the node is reached. To create the next tree the process is the same, but starts completely over with a new bootstrap sample of the training data. This is repeated for a chosen number of trees.

This choice of the number of trees to create is a tuning parameter. Too many trees can be computationally expensive, but too few can create a model that is poor at prediction. Another important tuning parameter is the choice of how many variables to randomly select

at each node. The default value for classification is typically the square root of the number of variables, and for regression it is the number of variables divided by three. Values for the minimum size of terminal nodes, which is the minimum number of observations required to make a leaf node, can also be varied. A large minimum size means smaller trees and faster completion of the random forest. This is an alternate strategy for controlling depth, which, in the `randomForest` package in R, cannot be directly controlled. Depth indicates the number of splits a tree has. Greater depth gives more information about the data. Another tuning parameter with a similar approach is the maximum number of terminal nodes a tree can have. Without setting this parameter, trees are grown as large as possible.

Tuning parameters are usually adjusted based off of OOB error rates. This uses the random forest to classify the OOB samples for each individual decision tree. A given observation will be put through each individual tree in the forest for which it is OOB. The predictions from each tree are then combined either by direct vote (for categorical variables) or an average (for quantitative variables) and this becomes the final predicted result for that observation. This result is compared with the ground truth and counted as either correct or incorrect. This is repeated for each observation in the dataset and the overall proportion of incorrect predictions is called the OOB error, which is an important tool for evaluating performance of the model. In this analysis I use both OOB error and the metrics listed in the Accuracy Assessment section for evaluation.

Similarly, to use a random forest to make predictions for a new set of input values, the observations are fed into each decision tree individually and tallied up. Since there isn't a known ground truth value for each observation and each observation was not included

in the creation of the model, all trees are used. The final voted upon predictions are the predictions of the random forest. This overall process is called bagging, because it is the action of bootstrapping the data to create each tree and using the aggregate to make a decision. Bagging is useful because it reduces variance without introducing bias. For this analysis, the number of trees parameter was varied individually. This is limited in that it does not explore any potential interactions, but beneficial in that it is more clear the individual contribution of this parameter. I chose it as parameter of interest because reducing it as much as possible can have a large impact on time and computational resources required.

For this work, random forests were created in R using the `randomForest` function, within the `randomForest` package (Liaw & Wiener, 2002).

2.3 Accuracy Assessment

In line with the methods used in our reference paper and the remote sensing field, accuracy metrics will be based on predictions for the test dataset and will include the following:

$$\text{Overall Accuracy} = OA = \frac{\text{number of correctly classified reference sites}}{\text{total number of reference sites}}$$

OA_{urb} and OA_{nat} will be used, which are the same as overall OA but only includes the urban and natural classes, respectively.

$$UA(z) = \frac{\text{number of correctly identified pixels in class } z}{\text{total number of pixels identified as class } z}$$

$$PA(z) = \frac{\text{number of correctly identified pixels in class } z}{\text{number of pixels truly in class } z}$$

UA is a measure of user's accuracy, which is also called precision or positive predictive value.

PA is the measure of producer's accuracy, also known as recall or sensitivity. The harmonic mean of UA and PA gives the F_1 score, which is a measure of the model's accuracy. An F_1 Score closer to 1 indicates a model that has both low false positives and low false negatives.

$$F_1 \text{ Score} = 2 * \frac{UA * PA}{UA + PA}$$

3 Results

3.1 Varying the Parameter for Number of Trees

The parameter for the number of trees was initially varied between 5 and 500 at intervals of 5. The resulting overall accuracy metrics indicate a leveling off around 125 trees (Figure 2). There's also a clear distinction between accuracy in urban vs. natural classes, with natural classes having a much higher overall accuracy.

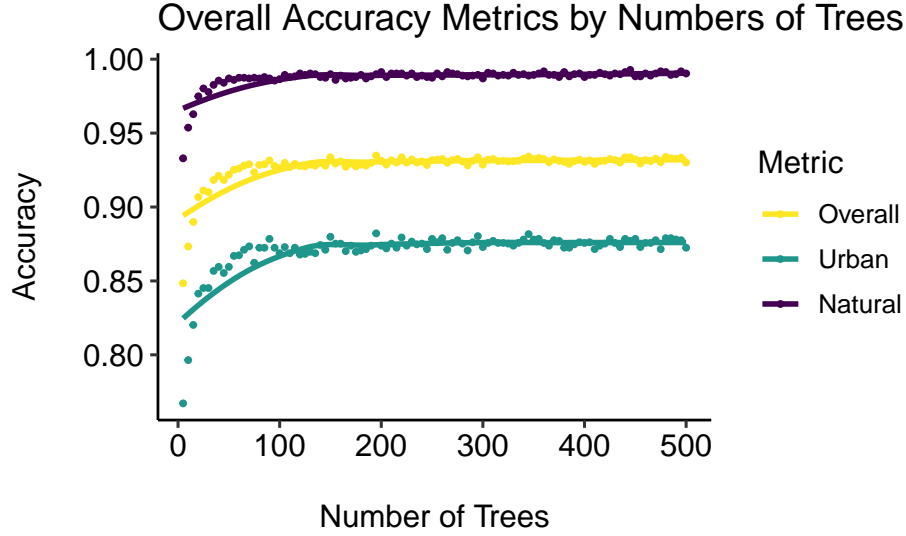


Figure 2: The increase in OA metrics levels off around 125 trees. Urban classes (1-10) have much lower accuracy than natural classes (11-17). These metrics were calculated based on the out-of-bag dataset.

The plot of F-1 scores (Figure 3) explores each class individually. It's clear that there are three approximate groupings. The dense trees; scattered trees; bush, scrub; low plants; and water classes (11, 12, 13, 14, and 17, respectively) have the highest F-1 scores throughout, staying in the high 0.9 range almost irrespective of number of random forest trees, though scattered trees (12) stands out as a little lower than the others. The open low-rise, large low-rise, and heavy industry classes (6, 8, and 10, respectively) have much lower F-1 scores which are around 0.5 to 0.7. Finally, in the middle are the compact high-rise, compact mid-rise, compact low-rise, open high-rise, and open mid-rise classes (1-5, respectively), with F-1 scores hovering around 0.7 to 0.9.

F-1 Score by Class for 5 to 500 Trees

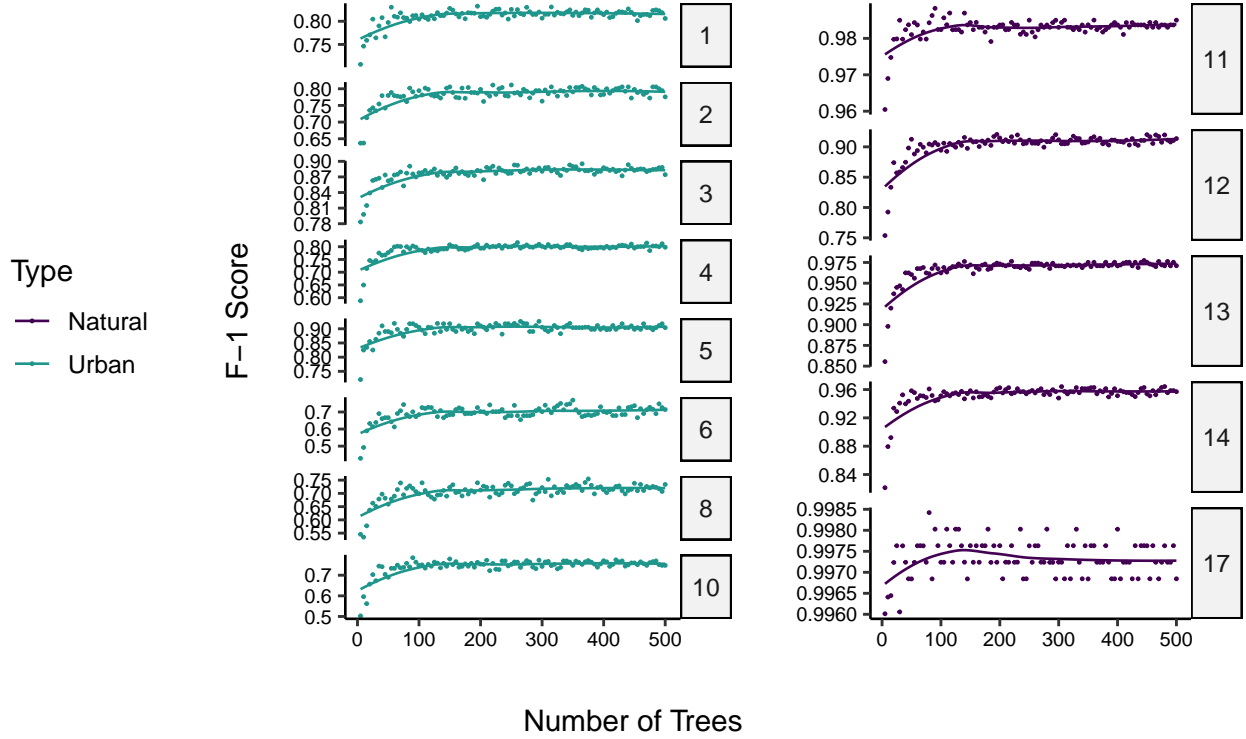


Figure 3: The variation between LCZ classes in F-1 score can be seen. As the number of trees in the random forest increases, F-1 score also increases, until around 100 trees. These metrics were calculated based on the out-of-bag dataset.

Almost all of the classes' F-1 scores appear to respond to increased numbers of trees, to an extent. This response seems to stop entirely once the number of trees reaches 100, but to explore it more thoroughly I ran another simulation varying the number of trees between 25 and 2500 at intervals of 25, but there was little to no effect on the F-1 score by class.

3.2 Predicting on the Test Dataset

3.2.1 Validation Metrics

OA and F-1 metrics dropped dramatically upon applying the random forest to the test data (Figure 4). The F-1 Score for class 17, Water, remained high, but since water has a very characteristic signature this is not surprising. Classification for classes 2 (compact midrise), 5 (open midrise), 8 (large low-rise), and 14 (low plants) performed especially poorly.

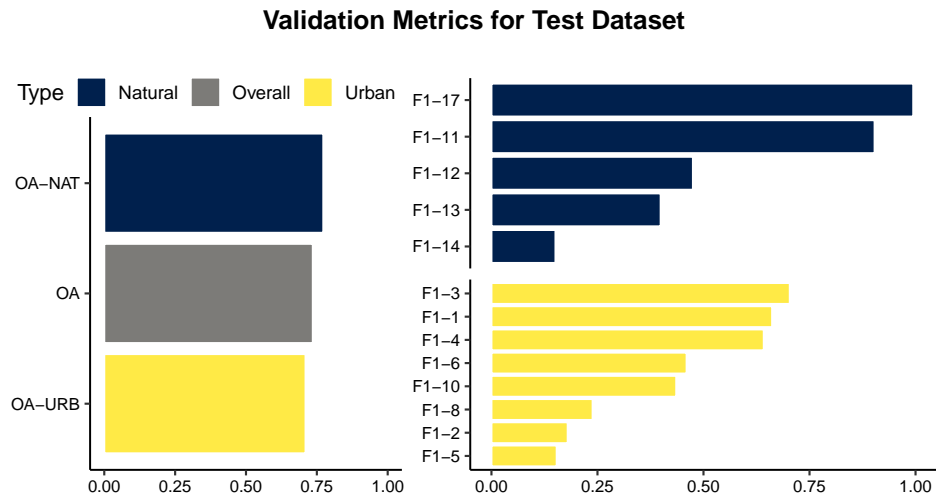


Figure 4: Accuracy among random forest predictions for the test dataset varied widely, but was lower than expected for F-1 scores, which do not seem to agree with the OA metrics. Classes 2, 5, 8, and 14 have particularly low F-1 Scores

3.2.2 Importance Measures

In general there is not a clear pattern in which bands or scenes proved to be the best predictors based on mean decrease in Gini Impurity (Figure 5). However, bands 7, 10, and 11 in Scene 4 were particularly useful. Scene 4 overall seems to contribute the most effectively to the model, surpassing the other scenes in each of the bands except for 4, 5, and 6.

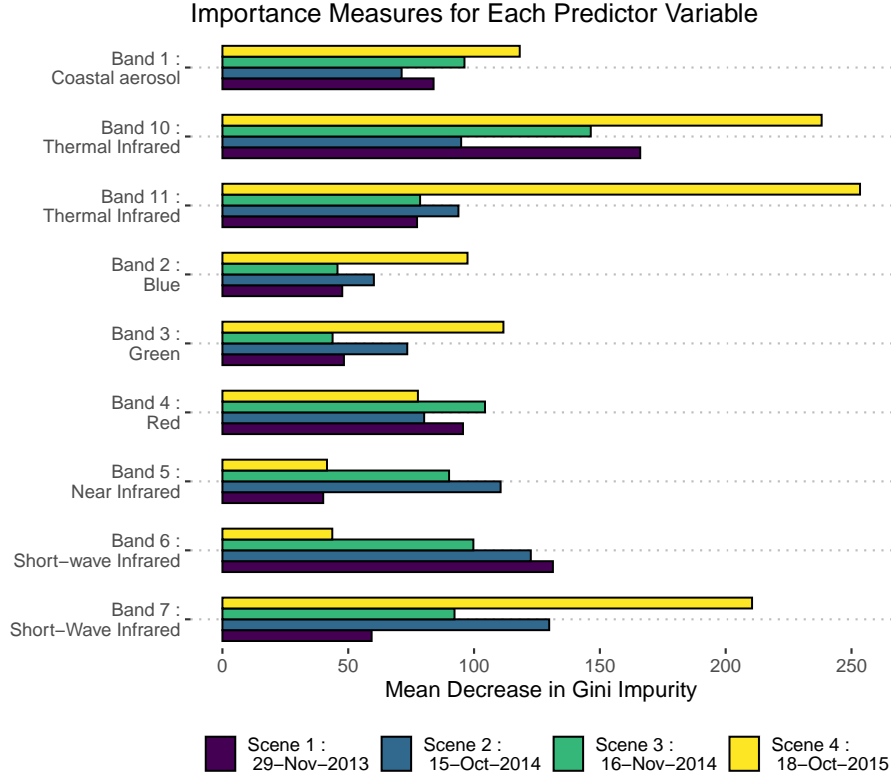


Figure 5: There is not a clear pattern in Mean Decrease for Gini Impurity between the different bands and scenes, though there is some indication that bands in scene 4 were particularly effective as predictors.

3.3 A Full Prediction

Despite the decreased overall accuracy for the random forest based on the test data, it is the best fitting model based on the training data, and therefore the one I used for the full prediction of LCZ classes throughout the Hong Kong area of interest (Figure 6).

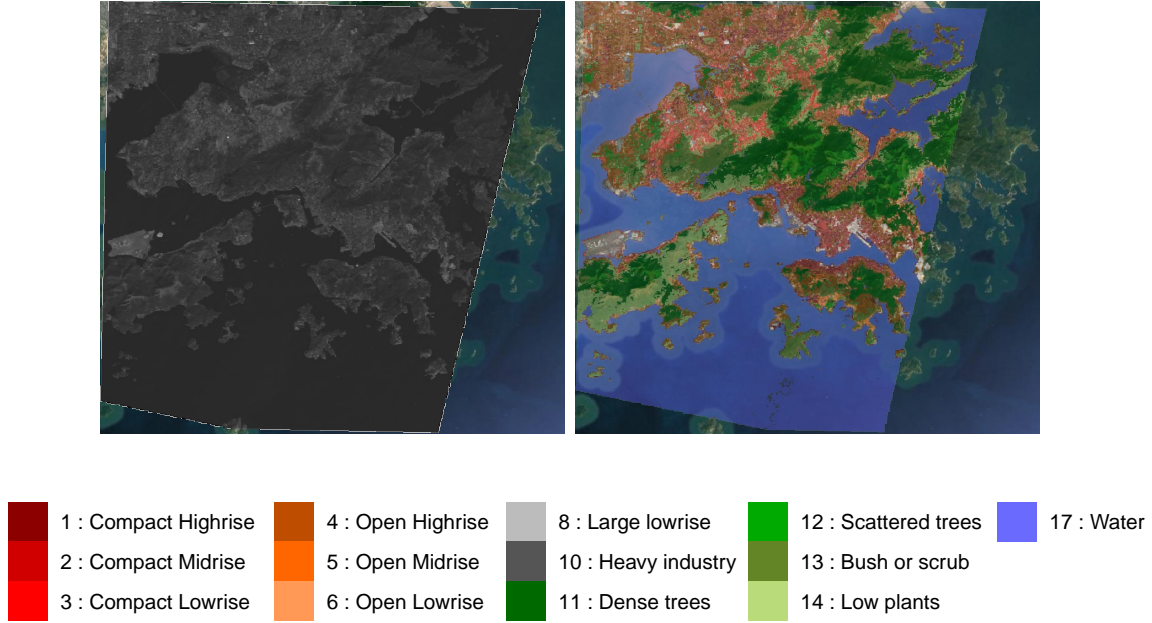


Figure 6: Imagery of the area of interest. Each has a basemap of satellite reference imagery. Top Left: Only satellite reference. Top Right: One Landsat 8 Scene. Bottom: A fully predicted LCZ map.

4 Discussion

The results of this analysis point to two primary issues in the current method for using random forests to classify LCZ classes. The first is the large decrease in accuracy between prediction for the OOB sample as compared to the that for the test dataset. This is not necessarily surprising considering the spatial autocorrelation inherently present in data like these, but it is concerning. OOB error is often used to tune the model parameters and these results suggest that it causes overconfidence in model success. Even when using the OA and F_1 metrics on the OOB sample there were much higher accuracy estimates than I saw with the test dataset. This points to a clear solution, which is that there should be an independent set of data used for building the model that is separate from both the testing and training

data. This solution is not particularly practical due to the already low availability of ground truth data.

The second issue I observed is that use of an aggregate measure like OA can mask very low accuracy in specific LCZ class categories, which were seen in the F_1 by class scores. For example, water has an accuracy of almost 100% by every metric tested. It also takes up a large proportion of the training and test data, as well as the overall area of interest. As has been mentioned, the signature for water is quite distinctive (CITE). Therefore it is reasonable to suggest that the high rate of correct predictions is not due to the suitability of random forest classification for LCZ classes, but rather it may be due to the general ease of classifying water with Landsat 8 imagery, despite method used. I don't think this is unique to the Hong Kong area as it's possible that any dominant class could monopolize the predictions.

That being said, there are a number of other potential reasons this discrepancy in accuracy between classes and between OOB samples as compared to test data, could be occurring. The uneven distribution of polygons in different classes in the training dataset is likely an important contributor. The small proportion of training and test pixels relative to the entire area of interest also may be cause for concern. However, the latter is not actually tested for our accuracy methods due to lack of a fully classified ground truth layer. Without this layer it is not possible to test the transferability of these models to other urban areas, but based on the classification success within Hong Kong, it is likely that transferability is poor.

In terms of my results varying the number of trees the random forest, there is an upper limit to how accurate the model can be. That upper limit may be in part due to the

appropriateness of the method, but I postulate that the quality and amount of data are the true culprits. This is a limitation of my current analysis and I suggest it as an area for future study. It would be valuable to understand what level of initial classification commitment is necessary for a robust LCZ analysis. Another limitation is the use of only one tuning parameter. I expect further experimentation of other important tuning parameters, as well as investigation of any interactions between them, may point toward solutions that do not require more investment in initial classification.

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