

The Price Air Pollution
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Abstract

1 Introduction

Air pollution in China is a persistent issue, but how serious is it?

China's recent rise to an economic powerhouse, characterized by the emergence of mega cities and the increasing demand for energy, has resulted in alarmingly high levels of air pollution. According to Chan and Yao (2008) the primary drivers of rising levels of air pollutants are China's metropolitan areas. Typically, a large proportion of main direct and indirect emission origins such as power plants, biogenic, industrial and vehicular sources and domestic heating can be allocated to these areas. Air pollution in China has been a persistent issue, but how serious is it? Polluting the environment is not only problematic in terms of climate change, but also regarding public health. Maruyama Rentschler, Leonova, et al. (2022) state that approximately 1.41 billion individuals are exposed to PM2.5 levels higher than 5 $\mu\text{g}/\text{m}^3$. Out of these individuals, 53 percent of the total population, or about 0.765 billion people, are experiencing dangerous levels exceeding 35 $\mu\text{g}/\text{m}^3$. Song et al., 2017 report that the Global Burden of Diseases (GBD) project could assign around 1.6 million deaths in China to air pollution related causes alone in 2015. In addition, persistent and excessive air pollution can be a huge burden to the economy of China: A sustainable development is endangered because natural and human resources are being exposed to hazardous emissions.

With increasing health impairments of the population, costs of the public health system are rising. Due to the outdated structure of the social security system in China, which will be discussed in more detail in section 2, it is mainly the rural migrants in the urban areas who are suffering the most, as they are only insured in their domestic province and medical services are only reimbursed in part there. Thus, they have to cover the entire cost and pay "out-of-pocket" for any medical care or treatment they receive. China features a variety of regions, from rural, mountainous locations to coastal, heavily industrialized clusters. The impact of air pollution may differ between regions and their adjacent provinces.

Inspired by the paper "Does industrial air pollution drive health care expenditures? Spatial evidence from China" by Zeng and He (2019), our goal is to conduct an independent analysis, which builds on their research, identifies possible shortcomings and adds value to this research field. In their study Zeng and He (2019) employ a spatial lag model in order to disentangle the direct and indirect spatial effects of industrial air pollution on health care expenditures over the period of 2002 to 2014. Their findings reveal that industrial air pollution is a significant contributor to increased health care expenditures not only in the provinces where air pollution is emitted, but also in neighboring provinces. Furthermore, the authors assert that provincial health care expenditures are shaped by a combination of industrial air pollution, health reforms and other socioeconomic factors. These results highlight the importance of considering spatial effects and a range of contextual factors when analyzing the relationship between industrial air pollution and health care expenditures.

The following regional analysis should answer the research question:

"To what extent does air pollution and its spatial spillover effects influence health care expenditures in China?"

In section 2, first health impacts of air pollution on human health are discussed. This is followed by a brief overview of air pollution policies implemented in the observation period and the organization of the Chinese healthcare system. The chapter is intended to shed light on the chain of effects of air pollution on government expenditures. More specifically, exposure to air pollution of individuals has negative health effects on them. As a result of deteriorating public health, provincial health expenditures increase. Section 3 provides a brief description of the data used and a descriptive analysis. Subsequently, in section ?? the process of model finding and the testing procedure of this is described, while section ?? presents the findings of the analysis. Finally, in section ?? the most important points are summarized and shortcomings of our paper are discussed.

2 Literature review

2.1 Effects of air pollution on health

Analysing the effect of air pollution on health expenditures implicitly implies the aforementioned causal transition dependencies. Various studies have identified direct effects of exposure to air pollution on health.

Franklin, Brook, Pope III (2015) and Fiordelisi et al. (2017) demonstrate that the risk of cardiovascular diseases and the triggering of acute cardiac events is increased by PM air pollution. The pathways through which this occurs include the generation of proinflammatory or oxidative stress mediators in the lung that enter the systemic circulation, the direct infiltration of certain particles or components into the cardiovascular tissue, or an imbalance of the autonomous nervous system. In that context, Hoek et. al (2013) quantify the effect of PM_{2.5} long-term exposure by conducting a meta-analytic review of previous studies. Their pooled estimates indicate an increase in all-cause mortality and cardiovascular mortality of six percent and eleven percent, respectively, if increments of PM_{2.5} are increased by 10- $\mu\text{g}/\text{m}^3$. Equivalently, an increase in NO concentration of the same magnitude leads to an increased all-cause mortality of five percent.

Not only cardiovascular diseases can be enhanced or even caused by air pollution. Kim, et al. (2018) found direct impacts of polluted air conditions on lung diseases and pregnancy disorders. According to the authors, even prenatal or perinatal air pollutant exposure can be associated with reduced lung functions of children and can have long-term effects on respiratory health of adults. Additionally, a Swedish study conducted by Nyberg et al. (2000) focussed on the relationship between urban air pollution and lung cancer risk. The authors used data from national health registers and modeled air pollution exposure. The results show a significant correlation between higher NO pollution and increased incidence of lung cancer. To be specific, they attributed a higher relative lung cancer risk between 0.8 and 1.6 percent to attributable NO exposure.

A 2010 study from Leliveld et al. estimated that between 1.6 and 4.8 million worldwide premature deaths occur due to outdoor air pollution in 2010 alone, predominately as a result of PM 2.5 pollution. The largest share of these deaths could be attributed to Asia.

Additionally, Zhao et al. suggest that based on recent findings air pollution may be involved in the development of autoimmune diseases such as diabetes mellitus, multiple sclerosis, or rheumatoid arthritis. The authors argue that air pollution can cause imbalances in T cells, the production of proinflammatory cytokines, oxidative stress, local pulmonary inflammation and methylation changes. These effects are involved with initiating or aggravating autoimmune diseases.

To illustrate the magnitude of air pollution-related illnesses, the causes of death by illness patterns provide information. In 2019 the major premature death causes in China were cardiovascular diseases totaling 43 percent, followed by malignant neoplasms (26 percent), respiratory diseases (10 percent), unintentional injuries (6 percent) and neurological conditions (4 percent). Other conditions that accounted for between 3 and 1 percent of premature deaths, in descending order, were digestive diseases, genitourinary diseases, respiratory infections, diabetes mellitus and infectious and parasitic diseases.

Figure xyz illustrates the channel between air pollution and public health expenditures. As mentioned above, air pollution has a direct impact on health as harmful pollutants penetrate the blood circulation through the lungs and damage other organs.

2.2 Air pollutions policies from 2011 to 2018

Air pollution in China has been a major public health problem for many years. However in recent years the government has taken various measures to address the issue. One of the main focuses of these measures has been to reduce emissions of particulate matter, sulfur dioxide and nitrogen oxides. In 2013, prompted by a period of heavy smog in eastern China, the government introduced an "Air Pollution Control Action Plan" to combat air pollution, which included specific targets for reducing particulate matter, sulfur dioxide, and nitrogen oxide emissions. More specifically, targets included, among others, reducing PM₁₀ concentrations in cities by more than 10 percent and reducing PM_{2.5} concentrations in the Beijing-Tianjin-Hebei, Yangtze River Delta and Pearl River Delta regions of around 25, 20 and 15 percent, respectively.

Furthermore, the 13th Five-Year Plan (2016-2020) also set a target to reduce PM_{2.5} concentrations in areas heavily affected by air pollution by 18 percent by 2020. Targets have also been set for reducing sulfur dioxide and nitrogen oxide emissions by 15 percent and 15 percent, respectively, compared to 2015 levels. Huang, Pan, Guo, Li (2018) indicate in their analysis, in which they map the national air quality in 74 cities that the issued Air Pollution Prevention and Control Action Plan (APPCAP) in 2013 has shown effect. Between 2013

and 2017 PM2.5 concentrations have reduced in average by about 33 percent, and PM10 concentrations by about 28 percent. Sulfur dioxide concentrations have reduced as well with an average reduction of about 54 percent. Nitrogen Oxides emissions, however, have not significantly decreased.

In addition, new air quality standards were set in February 2012. These were initially enforced in large cities and their surrounding areas in the following years until they were implemented nationally in 2016. The new standards set stricter limits for air pollutants such as particulate matter, sulfur dioxide, nitrogen oxides and ozone to better protect public health. Chinese authorities have also introduced new monitoring and reporting technology to ensure that air quality is measured and monitored in every part of the country. Companies that violate the new standards can be fined or even shut down.

In summary, a number of measures have been taken by the Chinese government from 2011 to 2018 to reduce air pollution, including stricter emission standards for vehicles, power plants and industrial facilities, and the closure or upgrading of older, heavily polluting factories, which were partially effective.

2.3 Organization of the Chinese Healthcare System

In this chapter, we will discuss the structure and functioning of the Chinese health care system in more detail. To be more specific, we will first discuss the system by which social services (of any kind) are determined, namely the hukou or huji system. After that, the structure of the system will be discussed in more detail, as well as the services provided

available

to the individual. Before we delve deeper into the structure and allocation of the Chinese health care system, it is worth mentioning how social benefits are allocated in China. Liu, 2005 describes The Hukou (eng. "individual origin" or also Huji eng. "household origin") is the government's official residence registration. This was introduced in the 1950s with the aim of.

"...to maintain social order, protect the rights and interests of citizens, and serve the construction of socialism."

Because rural areas can further absorb and make use of surplus labor, the central government thought that the majority of the population should live there. Domestic migration of people was also viewed as dangerous because it could result in city overpopulation and endanger farm production. Domestic migration was likewise discouraged by the federal government. The central government closely regulated migration flows, and only recently have these limitations been loosened, in the era of Mao Zhedong. This doesn't imply that the hukou system is no longer in place; rather, it only means that, for instance, rural dwellers can move to a city but won't be eligible for any social benefits there and will have to make due with their own resources.

In the late 1990s, the years of the opening and globalization as Kanbur and Zhang (2005) describes it, China developed from a centralized to an open market economy, there have been several reforms in the country, including in the health care system.

According to Hougaard et al. (2011) no longer the central Chinese government but rather the different Provinces are responsible for the financing and administrating the health care system, instead of a . This, since funding is provided through taxation, in turn led to strong inequalities between the rich coastal regions and the impoverished rural regions in the east of the country. The Chinese health care system is characterized by the division of the population between two groups: The proportion of the population living in the urban and rural areas.

3 Data

For our analysis, we use data retrieved from the National Bureau of Statistics of China. The data was compiled from an extensive database. It provides data on government spending, air pollution, and various socioeconomic factors on a provincial level. In total we use yearly data from 2011-2019 from 31 provinces. In advance, we excluded autonomous regions such as Hong Kong, Macau or Taiwan, for which data were also available. The dependent variable of our analysis "Local Governments Expenditure on Medical and Health

Care" includes all conceivable kinds of government spending, including for example health care management and services expenditures, disease prevention and motherhood expenditures as well as rural health spending. The variable is measured in 100 million Yuan. Additionally, we extracted three different pollution variables, namely particulate matter, sulfur dioxide and nitrogen oxides emissions volumes, which are measured in 10000 tons. However, no information was provided on whether the data was smoothed regionally. Furthermore, include various socioeconomic control variables in our analysis. variables such as the population shares of different age groups (0-14, 15-65, 65-), which were collected through random sampling, the average disposable household income for each urban and rural population in the province, the coverage rate by forest in the region, and the proportions of urban and rural population in the region serve as important control variables in our analysis. In addition, we included other variables in our initial models, but at a certain point we no longer considered them important because the influences of this variable were either already covered by another variable or they offered very little explanatory value for our analysis. These include the consumer price index, the gross regional product, total local government expenditures und various key figures relating to the health system.

4 Methodology

As already discussed in section 2, the possibility of a spatial autoregressive process (SAR) is rather unlikely considering the Hukou system. We therefore excluded the possibility of a SAR component before using statistical testing to determine the final model.

We started the testing procedure by using the (Spatial) Hausmann Test to decide whether fixed or random effects suit the data better. The test strongly indicated the usage of a fixed effects model over time, which is in line with the intuition as China's provinces are heterogeneous and eliminating time constant features such as geographic differences seems like a reasonable approach.

In the following a Lagrange Multiplier Test to decide whether a spatial error model (SEM) should be used. We could not reject the H_0 of a OLS model. Therefore, we continued to use a spatial lag in X (SLX) model, whose lag components can directly be tested by t-tests. The inclusion of a SLX component comes from the fact that air pollution of one province is likely to influence air pollution of neighbouring provinces. But, as already discussed in section 3, it is not known whether the emission variables are only total emission in a provinces that were also created in this province or if some smoothing of the data was already applied such that some distance travelled by means of kernel distribution or similar is used in order to account the effect on neighbouring provinces. The introduction of spatial lags concerning the air pollution will not cause any harm other than potential multicollinearity issues if they are in fact redundant.

The next step was to check to possible presence of heteroscedastic and serial correlated errors. Firstly, a Breusch-Pagan test was used to test for heteroscedasticity. Here we could reject the Null-Hypothesis of homoscedastic errors. Secondly, a Durbin-Watson test was conducted to test for serial correlation. The test strongly suggested serial correlation. As a result, we used a robust variance covariance matrix in order to account for this error structure and improve the reliability of our results

For the identification of the model, a general-to-specific approach was utilized, where we started with all reasonable variables and potential lags and eliminated them iteratively with the use of t-tests.

As a robustness check of our results, we used a feasible generalized least square model. Very similar to the SLX model we use a robust variance covavariance matrix that is corrected for heteroscedasticity and serial correlation. Also, the most general model is assumed again and variables are eliminated one-by-one.

Some variables of interest are in monetary terms. To account for the mere growth in these variables due to increased prices we adjust every monetary variables (X) by the general consumer price index (CPI)

$$\tilde{X} = \frac{X}{CPI/100}.$$

The CPI is divided by 100 as it is reported with baseline value 100.

Additionally, on every variable a logarithmic transformation was used. The reason for this is twofold. To begin with it might further help with the heteroscedastic and serially correlated errors and therefore provide more reliant estimates. And it also helps to make the interpretation of the estimates more intuitively. One exception to this procedure is the share of working population, as this variables already presents a share of

the whole population, a further transformation would make its interpretation meaningless.

5 Results

As previously outlined, the models presented in the table below were obtained using the general to specific model selection procedure. Starting from the model with all controls, statistically insignificant explanatory control variables were subsequently removed one at a time. During this process, the sign as well as the significance of our pollution variables were very sensitive to small changes in the model, i.e., we observed large changes in obtained estimates over a single iteration of the general to specific selection procedure. Accordingly, the presented results are to be interpreted cautiously. The unstable coefficient estimates are most likely caused by a combination of highly collinear explanatory variables and the absence of important variables that drive health care expenditures in our model.

Nevertheless, some key takeaways can be made. We observe that in our model, which does not contain spatially lagged variables, that Particulate Matter Emissions have a significant effect while in the SLX model this is not the case. This might be due to the mentioned multicollinearity problems, as all pollution variables exhibit a positive spatial correlation. The negative and significant effect of Sulphur Gas emissions, is most likely also a result of the multicollinearity in our model. Especially, given that the spatial lag features a positive and significant coefficient, there is no sensible interpretation to be made. The key takeaway should thus be, even though we find a positive and significant effect of Particulate Matter Emission (fGLS) on Health Care Expenditure, the result is both unstable over iterations in our model selection procedure and also across model types. Further, given our modelling and identification strategy, we may not claim that the results represent a causal relationship between the examined variables.

6 Conclusion

Table 1: Results fGLS & SLX

	<i>Dependent variable:</i>	
	fGLS	SLX
log(Disposable Income per Capita Rural)	1.072*** (0.067)	
log(Disposable Income per Capita Urban)		1.180*** (0.081)
log(Forest Coverage Rate)	0.490*** (0.102)	0.287** (0.144)
log(Urban Population)	0.308** (0.145)	0.504*** (0.135)
log(Waste Gas Emissions Nitrogen)	−0.045 (0.035)	0.006 (0.041)
log(Waste Gas Emissions Particulate Matter)	0.063*** (0.024)	0.032 (0.023)
log(Waste Gas Emissions Sulphur)	−0.007 (0.016)	−0.063** (0.028)
log(Waste Gas Emissions Nitrogen Lag)		−0.111 (0.088)
log(Waste Gas Emissions Particulate Matter Lag)		0.117*** (0.040)
log(Waste Gas Emissions Sulphur Lag)		0.093** (0.041)
<i>Note:</i>		
*p<0.1; **p<0.05; ***p<0.01		