

Life Cycle Assessment of Distributed Plastic Recycling via Additive Manufacturing

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Abstract

Distributed recycling via additive manufacturing (DRAM) emphasizes an emerging approach to locally recycle waste plastic directly by 3D-printing it into valuable products. During the last decade, major progress has been reported to validate the technical feasibility, and economic viability of the system. Although, significant research has been made to validate the technical feasibility and economic viability of DRAM system, the environmental evaluation of such system is still at its early stage. This study expands on the environmental assessment of plastic recycling network system (DRAM) to accelerate the implementation of such systems for waste plastic management extending the lifetime of plastic material. This work is based on a university operational physical demonstrator in Nancy France. It aims to evaluate the environmental impact of a DRAM unit as a basis scenario compared to two scenarios of the traditional centralized recycling value chain. To achieve this goal this work makes a comparative environmental performance using a life cycle assessment (LCA) to evaluate the multidimensional potential impacts of virgin and recycled polylactic acid. Four impact categories were considered : climate change, potential eutrophication, resource depletion, and ion radiation. For the first three of these impact categories, the results show a minimum 97% reduction in environmental impact compared to a virgin supply chain. On the other hand, the amount of ion radiation equals approximately 2.8 times the emissions of the virgin scenario due mainly to the specific situation in France, where most of the energy is produced by nuclear means. These results provide an environmental overview of the benefits and disadvantages of developing a DRAM system in a specific context.

Keywords : *LCA ; Distributed Recycling ; Plastic ; Additive Manufacturing, Circular Economy*

1. Introduction

- ¹ Since the early 20th century, the invention of plastic, or synthetic organic polymers, has changed the landscape of various industrial sectors. Production increased at a compound annual growth rate of 8.4%, rising from 2Mt in 1950 to 368Mt in 2019 ([Geyer et al., 2017](#)).
⁴ This versatile material stands out thanks to its easy processing and handling in shape, color,

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5 texture, thermal and barrier properties, and its mechanical and chemical resistance (Andradey
6 and Neal, 2009 ; Thompson et al., 2009). Thanks to these properties, today 39.6% of the
7 demand comes from the packaging industry, followed by the construction and automotive
8 industries with 20.4% and 9.6% of the production share respectively (Plastic Europe -,
9 2020). Unfortunately, the main problem is associated with multiple environmental damages
10 throughout its life cycle. Terrestrial, aquatic, and atmospheric ecosystems are not exempt
11 from the externalities of this innovation, which represent a major issue (Kumar et al., 2021).
12 Micro-, meso-, and nano-plastic pollution contribute to the detriment of ecosystem services
13 such as the ability to sequester carbon (Wang et al., 2022), soil productivity (Zhang et al.,
14 2022), and eutrophication (Vuori and Ollikainen, 2022).

15 Therefore, the adequate use and disposal of plastics became an urgent problem, which
16 is characterized by high complexity and multifaceted feedback loops. As a consequence, a
17 systemic view of the current entire plastics value chain is needed, including petrochemical
18 companies (de Vargas Mores et al., 2018; Iles and Martin, 2013), converters (Paletta et al.,
19 2019), brand owners or manufacturers (Gong et al., 2020 ; Ma et al., 2020), retailers and
20 consumers (Confente et al., 2020 ; Filho et al., 2021 ; Friedrich, 2020), and recycling operators
21 (Huysveld et al., 2019 ; Pazienza and De Lucia, 2020), as well as the influences of policy-makers
22 on wider economic and societal changes (Paletta et al., 2019). The European Union (EU)
23 intends to develop a circular economy (CE) based on a production and consumption model
24 with key activities such as “sharing,” “reducing,” “reusing,” “repairing,” “renewing,” and
25 “recycling” the existing materials and products as many times as possible, in order to create
26 added value by extending the life cycle of products (European Commission, 2018 ; Matthews
27 et al., 2020). As part of the European Green Pact presented on March 20, 2020, there are plans
28 to establish an action plan involving the circular economy, mainly promoting the development
29 of sustainable products, reducing waste, and empowering citizens as key players (European
30 Commission, 2018).

31 Considering the French context, a target was established that by 2025 all plastic waste
32 should be recycled, but currently recycling statistics in France are only reaching levels close
33 to 25%.

34 Despite these ambitious objectives, plastics recycling has historically been an expensive process
35 due to the inherent separate collection, transportation, processing, and remanufacturing
36 (Hopewell et al., 2009 ; Singh et al., 2017). The economies of scale have been leveraged to
37 reduce these costs with centralized and global recycling chains (Kreiger et al., 2013 ; Kreiger
38 and Pearce, 2013). Nevertheless, in order to carry out this recycling system, multiple steps
39 need to be accomplished that integrate the sorting phase, long-distance transport, waste
40 treatment, and remanufacturing. The high costs of these processes and the low selling price
41 (mainly due to the dependence of the recycled plastic price on the petroleum and virgin prices)
42 seldom generate benefits and often require costly public subsidies (Hamilton and Steven,
43 2019). In addition, these centralized plastic manufacturing and recycling lines lead to soil,
44 water, and air pollution (Arena et al., 2003 ; Carlsson Reich, 2005). In addition to the current
45 problems in the plastic recycling network, we can highlight that supply chains in general are
46 under increasing pressure from various stakeholders to make decisions from a sustainable
47 perspective ; in other words, based on economic, environmental, and social objectives (Hassini
48 et al., 2012).

49 Additive manufacturing technology (also known as 3D printing) enables the potential
50 of distributed manufacturing (DM) for products of high added value (Bonnín Roca et al.,
51 2019 ; Petersen and Pearce, 2017 ; Woern and Pearce, 2017). Nowadays, the accessibility of

freely available designs has increased significantly, together with the development of open-source technologies and the supply of raw materials (virgin and recycled filaments) for 3D printing (Hunt et al., 2015). It has opened the path to design distributed and more flexible manufacturing systems that will integrated in a more symbiotic manner within the urban areas close to the raw materials sink, that is our domestic waste (Herman2020 ?) For example, recent studies such as Zhong and Pearce (2018) demonstrated that the coupling of an OS extruder (recyclebot) and RepRap 3D printer “*brings a traditional industrial system into a single small home, business or community center*”. Furthermore, various studies in the literature show the technical feasibility of this distributed plastic recycling approach. More recently, Santander et al. (2020) demonstrates how of poly lactic acid (PLA) waste from a local network of secondary schools could be designed and implemented to be treated in a single small size and open-source recycling facility. The authors, explored the economic and environmental scenarios of this configuration through an optimization approach. However, the study did not include the multidimensional environmental impacts as only emissions of CO_2 equivalent were considered.

From an environmental point of view, the DRAM and DM approaches have been also evaluated mainly using the Life Cycle Assessment (LCA) approach. For example, Kreiger and Pearce (2013) and Kreiger et al. (2014) explored the environmental benefit of distributed recycling using open-source extruders (RecycleBots), which have been used to obtain 3D printing filament from post-consumer goods. Their results show that significant savings could be obtained. However, their research is limited to the consideration of energy and CO_2 emissions as environmental indicators. More recently, Kerdlap et al. (2021) through a simulation approach, quantified the plastic life-cycle environmental impact of small-scale sorting and recycling systems in comparison to traditional large-scale centralized systems. However, it does not consider the use of Open Source Technologies. Therefore, major efforts need to be made in order to evaluate, in a holistic way, the environmental impacts of the global DRAM value chain. Thus, this paper aims to contribute to this field through the evaluation of multidimensional environmental impacts from the implementation of distributed recycling via an additive manufacturing approach in a territory. Specifically, an environmental evaluation using life cycle assessment (LCA) is conducted, comparing a distributed plastic recycling system to produce 3D filament with a traditional production system of virgin plastic filament for 3D printing. From this evaluation, the environmental impacts (positives or negatives) of implementing DRAM have been analyzed and discussed.

This article is structured as follows. Section 2 outlines an overview of the environmental impacts studies in the DRAM context. Section 3 presents the system (case study) evaluated. Section 4 materials and methods, where life cycle analysis methodology is explained. Section 5 presents the life cycle analysis performed. Section 6 presents the discussion of results. Finally, Section 7 presents the main conclusions and recommendations for future works.

2. Overview DRAM system and environmental impacts studies

Distributed manufacturing is defined as the decentralization of production through the installation of multiple production factories with similar technology distributed geographically (Bonnín Roca et al., 2019). It is characterized by local production that thrives on the synergy of the emerging capabilities of digital manufacturing, information, and communication technologies, and the peer-to-peer production approach (Kostakis et al., 2018 ; Kostakis, 2013 ; Pavlo et al., 2018). Indeed, DM offers the possibility to decentralize production structures,

the flexibility to reflect local customer needs, lower logistics costs, shorter lead times, and lower environmental impacts (Petersen and Pearce, 2017; Woern and Pearce, 2017). Based on the DM paradigm, a new possibility of plastic recycling supported by additive manufacturing, called distributed recycling by additive manufacturing (DRAM), has emerged in the literature (Cruz Sanchez et al., 2020; Hart et al., 2018). Promoted by the development of 3D printing in an open-source context, DRAM is proposed to provide recycled plastic feedstock to the various 3D printers in a DM context. This recycled plastic can take the form of a filament (Fused Filament Fabrication), and recent works have dealt with the validation of a granular form Fused Deposition Modeling (Alexandre et al., 2020; Justino Netto et al., 2021) FDM (Fused Deposition Modeling) is currently the most widely used AM technology for plastic recycling (Kuclourya2022 ?). The 3D printing feedstock is then obtained via plastic recycling on a local scale using open-source machines such as shredders and extruders (Zhong and Pearce, 2018).

The main feature of DRAM is the reduction of the impact in the collection phase, favoring shorter and simpler supply chains (Despeisse et al., 2017; Garmulewicz et al., 2018). Several works have focused on the validation of the DRAM approach from a technical (López et al., 2022 ; Mohammed et al., 2022), economic, and environmental perspective (Cañado et al., 2022 ; Wu et al., 2022). Cruz Sanchez et al. (2020) conducted a systematic literature review to examine the level of development of the different DRAM stages from a technical perspective. Their results show that significant progress has been made in the stages of compounding, feedstock, printing and quality assessment. However, they also show that little work has been done for the preparation and recovery stages in spite of this.

From an environmental point of view, the DRAM and DM approaches have been evaluated mainly using the Life Cycle Assessment (LCA) approach. LCA is one of the most widely used environmental impact assessment methodologies. LCA corresponds to a methodology standardized by ISO 14040 and ISO 14044., which has even been used for environmental regulations in different parts of the world. In the context of DM/DRAM-related research, LCA has been applied in various ways. For example, Kreiger and Pearce (2013) conducted a life cycle assessment, in terms of energy consumed and emissions involved, to compare centralized manufacturing and distributed manufacturing using RepRaps (3D printers) for the distributed production of goods. The results showed that the use of poly lactic acid (PLA) in a distributed manufacturing context reduces energy demand and system emissions, which can be greatly diminished if a solar photovoltaic (PV) array is used. Later, several works related to plastic recycling using 3D printing has focused on demonstrating the technical capability of this technology to perform mechanical recycling. In environmental terms, the advantages that can be obtained with the use of 3D printing to produce specific products have been evaluated in comparison with their conventional mode of production (Top2022 ?). On the other hand, several studies have been carried out to demonstrate the environmental benefits of distributed recycling compared to centralized recycling.

Kreiger et al. (2014) explored the environmental benefit of distributed recycling using open-source extruders (RecycleBots), which have been used to obtain 3D printing filament from post-consumer goods. Focusing on the use of High Density Poly-Ethylene (HDPE) for the material, they performed an LCA of energy consumption and CO_2 emissions to compare distributed recycling to standard centralized recycling. Their results showed that distributed recycling of HDPE uses less energy than the best-case scenario investigated for centralized recycling, and it can achieve savings of over 80%. In terms of the different existing options for end-of-life (EOL) treatment of plastic waste, mechanical recycling via AM has shown positive

¹⁴⁴ results compared to incineration and landfill if it is closed loop.

¹⁴⁵ Table X represents in summary the different articles that have worked on environmental
¹⁴⁶ assessment using LCA in the DRAM concept either from a distributed recycling point of view,
¹⁴⁷ additive manufacturing or the whole of distributed recycling via additive manufacturing. The
¹⁴⁸ table organizes the previous works according to the level of approach that allows identifying
¹⁴⁹ which was their subject of study ; the comparison made identifying if it is DM vs CM or
¹⁵⁰ DR vs CR ; the use of recycled material in their studies, the identification of the DRAM
¹⁵¹ concept in their methodology ; the economic aspect and finally the identification of some of
¹⁵² the categories of impacts considered in the LCA carried out.

TABLE 1: Overview of the studies on environmental impact about DRAM

Source	Scope level	Comparaison	Recycled plastic	DR	AM	Economic aspect	GDP	APO	GWP	ADP	FDP	HTP	MAE	PAE	TFEFEP/MTEP	WDP	IRP_HE
Kreiger and Pearce (2013)	Home based manufacturing	DM vs CM	x		x	x	x		x		x		x				
Kreiger and Pearce (2014)	Home based manufacturing	DRAM vs CR	x		x	x		x		x		x					
Cerdas et al. (2017)	Product manufacturing	DM vs CM			x	x		x	x	x	x	x	x	x	x	x	x
Gaikwad et al., (2018)	Product manufacturing	DRAM vs CM	x		x	x			x								
Zhao et al., (2018)	Product manufacturing	DRAM vs CR	x		x	x			x		x		x		x	x	x
Santander et al.(2020)	City manufacturing	DRAM vs CM	x		x	x	x			x							
Kerdlap et al. (2021)	Regional recycling system	DR vs CR	x		x				x		x		x		x	x	x
Garcia et al. (2021)	Product manufacturing	DM vs CM	x		x		x		x								
Top et al. (2022)	Product manufacturing	DM vs CM			x				x								

153 In conclusion, even though different studies have been conducted aiming to validate the
154 distributed recycling approach from a technical, economic, and environmental perspective,
155 only the work of Kreiger et al. (2014) is focused on the environmental assessment of distributed
156 recycling for 3D printing purposes. However, their research is limited to the consideration of
157 energy and CO_2 emissions as environmental indicators. Therefore, major efforts need to be
158 made in order to evaluate, in a holistic way, the environmental impacts of the global DRAM
159 value chain. Thus the contribution of this research lies in the evaluation of environmental
160 impacts from the implementation of distributed recycling via an additive manufacturing
161 approach in a territory. Specifically, an environmental evaluation using life cycle assessment
162 (LCA) is conducted, comparing a distributed plastic recycling system to produce 3D filament
163 with a traditional production system of virgin plastic filament for 3D printing. From this
164 evaluation, the environmental impacts (positives or negatives) of implementing DRAM have
165 been analyzed and discussed.

166 The technological efficiency in environmental terms of AM and DR has already been
167 proven. This work seeks to analyze the environmental impact of a localized recycling system
168 oriented to the production of filament to meet part of the demand in a specific city. even
169 though different studies have been conducted aiming to validate the distributed recycling
170 approach from a technical, economic, and environmental perspective, only the work of Kreiger
171 et al. (2014) is focused on the environmental assessment of distributed recycling for 3D
172 printing purposes. However, their research is limited to the consideration of energy and CO_2
173 emissions as environmental indicators. This is the first study that seeks to analyze using LCA
174 methodology the DRAM concept as a new production system that allows to respond in part
175 to the demand for plastics and to complement the current recycling systems. In conclusion,
176 even though different studies have been conducted aiming to validate the distributed recycling
177 approach from a technical, economic, and environmental perspective, only the work of Kreiger
178 et al. (2014) is focused on the environmental assessment of distributed recycling for 3D
179 printing purposes. However, their research is limited to the consideration of energy and CO_2
180 emissions as environmental indicators.

181 3. Case study : The Lorraine Fab Living Lab

182 In order to achieve the objectives of this research, a case study with the following
183 characteristics was selected :

- 184 1. Existence of a favorable context to implement a DRAM recycling system. A favorable
185 context is defined as a considerable amount of plastic waste to be treated, as well as
186 initiatives for the widespread use of 3D printing.
- 187 2. Existence of a space dedicated to the recycling of plastic for 3D printing.

188 Under these considerations, the selected context for this study was the Lorraine Fab living
189 Lab (LF2L)², an innovation space located in Nancy, France. This university laboratory has
190 been selected mainly for the following reasons :

- 191 (1) Innovation spaces such as Fablabs, Maker spaces, design factories among others have proven
192 to be favorable environments for eco-innovations facilitating the implementation of circular
193 economy strategies (Coskun et al., 2022). Previous studies show that these collaborative
194 environments foster sustainable experimental learning, provide methodologies and tools for

2. <https://lf2l.fr/projects/green-fablab/>

195 the co-creation of circular solutions, drive the transition toward sustainable smart cities,
196 foster the creation of new sustainable business models, and facilitate knowledge exchange
197 on circular solutions ([Kasmi et al., 2021](#)). (2) Since 2014, the LF2L has been studying the
198 possibility of recycling, in their installations, PLA for reuse in 3D printing. The pilot recycling
199 process present in this center has been developed in the research work of Cruz Sanchez et
200 al. ([2017](#)), and the possibility of implementing this recycling process in the region is being
201 evaluated. (3) An investment program has been launched by the Grand Est region to promote
202 the implementation of Fablabs, and consequently the use of 3D printers, in the schools and
203 high schools of the region ([Canopé, 2022](#)). The goal of this investment plan is to ensure that
204 all the schools and high schools in the region will be equipped with this technology in the
205 near future. This corresponds to the future scenario evaluated in the work of Santander et
206 al. ([2020](#)). However, in contrast to Santander et al. ([2020](#)), in our research work a complete
207 environmental evaluation of the scenario is carried out. Consequently, this case study has been
208 selected because of their experience in experimenting with the DRAM strategy, as well as the
209 availability of technical and economic data and the aforementioned scientific publications that
210 provide details on its local implementation. For this study, the context described above has
211 been simplified in terms of geographical scale, and the following assumptions are considered.

- 212 — • Consideration is given to only one type of plastic waste to be recycled. Specifically,
213 PLA has been considered. This is mainly because PLA is one of the most used plastics
214 in 3D printing ([Bikas et al., 2016](#)). For the plastic waste sources, PLA waste from
215 schools and high schools (who have 3D printers) has been considered. PLA has been
216 selected as feedstock material due to the fact that (1) in Europe, PLA is considered a
217 “miscellaneous product” ([Maga et al., 2019](#)) because of its classification as a Bioplastic
218 and because, at least in Europe, there is no defined recycling strategy for these types
219 of plastics so they are usually sent to a landfill or are incinerated ; and (2) PLA is one
220 of the most widely used plastics in 3D printing.
- 221 — The recycling system modeled does not consider the sorting, separation, and cleaning
222 process because the collected material corresponds to non-contaminated waste : for
223 example, discarded 3D printing parts used for prototyping.
- 224 — From a geographical point of view, only schools and high schools in the Lorraine region
225 of France have been considered, and the route of recovery and delivery considered is
226 obtained in the work of Santander et al. ([2020](#)).
- 227 — Each school and high school requires 1 kg of filament per month.
- 228 — The 3D printing activities carried out in these establishments have the specific purpose
229 of making product prototypes and mock-ups, which allows them to generate testing
230 activities, design evaluations, functional evaluations, and corrections. Therefore, after
231 a short lifetime, 3D printing can be a source of significant amounts of plastic waste
232 due to printed parts that do not possess the desired quality, unused raw materials, or
233 products that have already fulfilled their life cycle ([Jaafarnia et al., 2021](#)).

234 4. Material and methods

235 The chosen methodology is Life Cycle Assessment (LCA) because, unlike other analytical
236 methodologies with an environmental focus, such as Material Flow Analysis (MFA), Substances
237 Flow Analyses, and Environmental Risk, it allows us to evaluate the environmental impacts
238 of the system value chain to be modeled ([Mahmud et al., 2021](#) ; [Pryshlakivsky and Searcy,](#)
239 [2021](#)).

240 The LCA represents a different tool and different techniques created to determine in an
241 effective and fast way the results to help with managerial decision-making in environmental
242 terms and in the deepening of the sustainable development policy ([Guinée, 2002](#)). LCA is
243 a structured, comprehensive and internationally standardized method in accordance with
244 ISO 14040 and 14044. It defines all the steps for its use : goal and scope definition, inventory
245 analysis, impact assessment, and interpretation ([Pennington et al., 2004](#)).

246 LCA is a tool for quantitative evaluation of materials, energy flows, and the potential
247 impact of products, services, or technologies ([Dehghanian and Mansour, 2009](#)). The analysis
248 takes into account the entire life cycle of a product : from resource extraction, through
249 production, use, and recycling, to the disposal of the remaining waste ([IES, 2010](#)). LCA is
250 considered a legitimate environmental methodology that enables systems analysis for waste
251 policy and strategy ([Gontard et al., 2022](#)). The four main stage for a life cycle assessment
252 are :

- 253 1. **Goal and Scope Definition** : This is the first stage and serves to orient the study
254 bases. It defines the main objectives of the life cycle assessment, the target audience of
255 this report, the functional unit that is the reference point from which the potential
256 environmental impacts will be obtained, the limits of the system under study, the
257 categories of environmental impacts to be evaluated, and the hypotheses to be used in
258 different stages of the LCA.
- 259 2. **Inventory Analysis** : Inventory is the stage in which the flows are quantified. It sets
260 out the database used, the energy and material input, the calculations performed, and
261 how the system was modeled.
- 262 3. **Impact Assessment** : This stage presents the software used and the calculation
263 methodology used to transform flows and characterize them in the impact categories
264 evaluated. From this characterization, the impact profile of the system under study is
265 obtained.
- 266 4. **Interpretation** : This is the conclusion of steps 2 and 3, presenting the results obtained
267 from the hypotheses used, the considerations, and the functional study that has been
268 defined. The phases of the life cycle that have the most impact are identified, and
269 sensitivity analyses can be carried out to evaluate the behavior of the systems according
270 to the variation of certain parameters. The inter-phase analysis stage is fundamental
271 for decision-making, as it identifies critical points and provides a basis for future
272 improvements.

273 The LCA methodology is mostly used in an iterative way, allowing a better definition of
274 the objectives to be achieved and the system to be analyzed. In the following sections, an
275 LCA is presented for DRAM using the Lorraine Fab Living Lab as a case study.

276 5. LCA application and results

277 5.1. Goal and Scope Definition

278 The main objective of this life cycle assessment is to compare the potential impacts
279 produced by a DRAM chain with the impacts produced by a traditional chain of virgin plastic
280 filament for 3D printing.

281 5.1.1. *The functional unit*

282 As mentioned above, this study will compare the environmental impact of two PLA 3D
 283 filament supply systems. By way of context for this study, the 3D filament will be supplied to
 284 all schools and high schools located in the city of Nancy, France. To compare two or more
 285 systems, it is necessary to define a functional unit that represents the service delivered by
 286 both systems. In this case the main service consists of the following :

287 “*A monthly delivery of 1 kg of ‘standard’ plastic filament (PLA) to each school and high
 288 school in Nancy during ten months of the year.*”

289 The 10 months correspond to the normal operation of a school from September to June.

290 5.1.2. *System boundary*

291 Figure 1 shows the two systems compared in this study and the processes integrated into
 292 the life cycle assessment. As can be observed, there are three different boundaries represented
 293 by the colors green, blue, and red.

294 The first boundary (green box) corresponds to the Biosphere, which represents everything
 295 related to nature, such as raw materials from natural resources, ecosystems, and solar energy.
 296 The second boundary (dark blue box) corresponds to the Technosphere limit, representing
 297 human activity (e.g. use of electricity, fuel, etc.). Here, we can observe all the material flows
 298 that are considered for the life cycle assessment, the flows between processes, or the flows
 299 that are part of the functional unit mentioned above. Finally, the boundary in a light blue
 300 color represents the limit of the services and processes taken into consideration in this study.
 301 On the one hand, in the upper process flow, we can observe the processes considered in the
 302 system to produce filament from virgin plastic. The process starts with PLA production (the
 303 whole production process detailed by ([Vink2003 ?](#))) which is followed by PLA transportation,
 304 filament production, and finally product delivery. On the other hand, Fig. 1 shows the recycling
 305 process enclosed in the orange box. The process starts with the collection of waste produced
 306 by schools and high schools, then the plastic recycling process, the production of filament,
 307 and the delivery of filament are carried out.

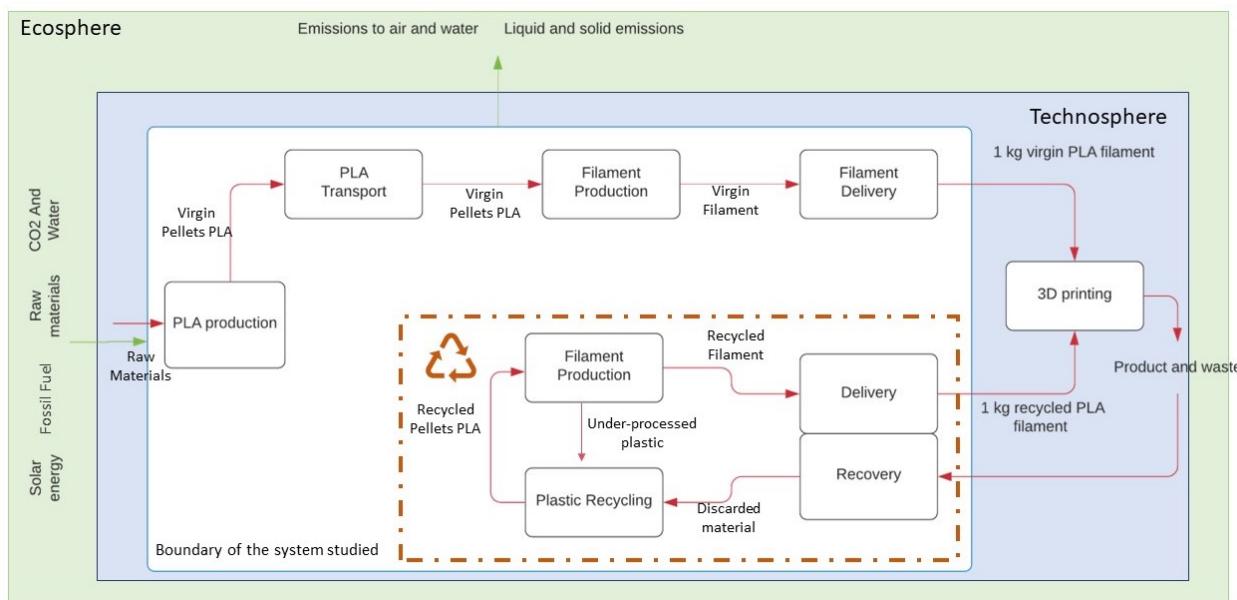


FIGURE 1: System limits for life cycle assessment

308 5.1.3. *Selected impact categories*

309 The objective of performing a life cycle assessment is to determine and evaluate the
310 potential environmental impacts produced at different stages in the life cycle of a product,
311 service, activity, or process.

312 However, it is quite difficult to identify, a priori, the categories of impacts in which the
313 system under study is most detrimental (climate change, eutrophication, etc.). To resolve,
314 it was decided to analyse relevant literature of LCA applied on the bioplastics and additive
315 manufacturing separately.

316 These two domains were considered because they can be considered closely related to a
317 DRAM system. First, the production of virgin bioplastics can be confronted with the recycling
318 of bioplastics. Secondly, the integration of these materials in additive manufacturing as a
319 viable solution for plastic waste management.

320 Concerning the plastic perspective, Bishop et al. (2021) compared the impact of bioplastics
321 (such as PLA) regarding the petrochemical plastics considering the impact categories across
322 44 relevant articles that used LCA evaluation. Their results pointed out the most used impact
323 category for bioplastics evaluation is climate change, followed by potential eutrophication
324 variations, resource depletion, human toxicity, photochemical oxidant formation, ozone de-
325 pletion, ecotoxicity, particulate matter formation, energy consumption, land use, and water
326 consumption. On the other hand, it was found that the cumulative energy demand (CED)
327 of the system is the most evaluated LCA impact category (Cerdas et al., 2017; Kellens
328 et al., 2017; Kreiger and Pearce, 2013; Quinlan et al., 2017) in the context of additive
329 manufacturing. In addition, It was also considered climate change, potential eutrophication,
330 and human toxicity as the main impact categories included in the LCA. In conclusion, it was
331 decided to use the predominant impact categories in each of the industries, namely climate
332 change, potential eutrophication, and resource depletion (fossil and water) after analyzing the
333 LCA results of the two domains .

334 5.1.4. *Assumptions and limitations*

335 The realization of a full LCA involves the collection of information and data related to
336 the different processes, flows, and activities. Due to the difficulty of obtaining the necessary
337 data, it is permissible to formulate hypotheses to partially make up for the lack of data, on
338 the condition that the hypotheses used, and the conditions under which they are formulated,
339 are made transparent.

340 The assumptions formulation was mainly used to model the virgin plastic filament production
341 system. The main hypotheses used are based on :

- 342 — Location of polylactide acid production : This hypothesis indicates the location where
343 the polylactic acid production process takes place.
- 344 — Location of filament manufacturer : Due to the uncertainty about filament production,
345 two companies with different locations that are engaged in the production of plastic
346 filaments have been modeled.
- 347 — Filament production machine : The machine PEEK 3d Printer Filament Production
348 Line ³ has been considered, which corresponds to the machine used by the two filament
349 manufacturers considered.

3. More information about the machine can be found at www.acceextrusion.com/product/peek-3d-printer-filament-production-line.html (Accessed on 09/09/2022)

- 350 — Performance of the filament production machine : The machine selected to carry out
 351 the filament production process has a range of transformation from plastic pellets to
 352 filament. The range of the machine chosen for the evaluation has an output between
 353 20 and 25 kg of filament per hour. This range directly affects power consumption.
 354 — We do not take into account the end of life of the filament, since we consider that the
 355 waste produced at the “end of life” is the beginning of the collection phase (I).
 356 — Type of energy source used to conduct the recycling processes : Each country has its
 357 own technological mix to supply its electrical energy consumption, such as nuclear,
 358 solar, and wind. These different sources of electricity are considered in the evaluation.
 359 — Transportation of raw material and filament (virgin and recycled) : The different
 360 options for methods of raw material transportation and filament transportation are
 361 considered in the evaluation.

362 Based on these hypotheses, two scenarios were created to model the possible operation of
 363 the virgin plastic filament production system (see Table 2). A third scenario represents the
 364 distributed recycling system for filament production, so the complete system is modeled with
 365 information obtained from the LF2L.

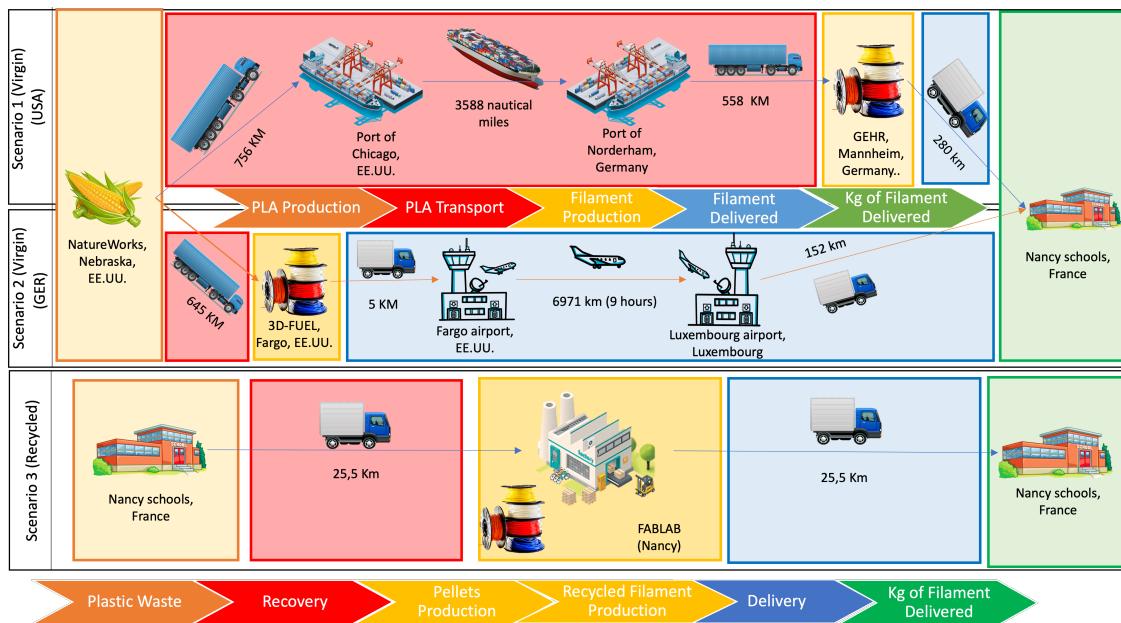


FIGURE 2: Representation of the 3 scenarios

TABLE 2: Scenario definition

Scenario	Description
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Virgin Filament

- 1 Scenario 1 begins with the production of PLA at the NatureWorks factory in Nebraska, USA. The PLA is transported by a combination of land and sea to bring the plastic from the United States to the filament manufacturing company, called GEHR, which is located in the city of Mannheim in Germany. In Germany, electricity is produced from wind power. From this location, the virgin filament is shipped directly to Nancy by light road transport.

2 factory in Nebraska, USA. The pellets of PLA are transported by road to the filament
3 manufacturing company, called 3D-Fuel, which is located in Fargo, USA. In the United States,
electricity is produced from natural gas. From the USA, the filament is shipped directly to
Luxembourg by air freight. Then the filament is transported to Nancy using a lightweight vehicle.

Recycled Filament

3 This scenario starts with the collection of PLA 3D printing waste from the schools and high schools.
The PLA waste is transported to the recycling point (Lorraine Fab Living Lab) using light road
transport. At this center, the size reduction process and the extrusion process necessary to produce
the recycled filament are performed. The recycled filament is then delivered to the same schools and
by the same means of transport as where the waste was collected.

366 5.2. *Life Cycle Inventory Analysis (LCIA)*

367 5.2.1. *Data source*

368 To obtain the necessary data, different data sources were used to carry out the LCIA.
369 On the one hand, for virgin PLA filament production there are various life cycle analyses
370 published by NatureWorks ([Vink et al., 2003](#); [Vink and Davies, 2015](#)), where the results have
371 been incorporated into the EcoInvent database. These articles helped us to understand the
372 PLA production process, considering the production center in Nebraska, USA. For virgin
373 PLA filament manufacturing, NatureWorks proposes a catalog of customers who manufacture
374 filament from PLA produced by the company, which facilitated the modeling of the supply
375 chain to sell its product in the city of Nancy, France.

376 On the other hand, for the plastic recycling process the data concerning the input/output
377 materials and the machines used in the recycling process were obtained directly from the
378 Lorraine Fab Living Lab and from the thesis works conducted at the ERPI laboratory related
379 to DRAM ([Cruz Sanchez et al., 2020, 2017](#); [Santander et al., 2020](#)) (see the Appendix A).
380 These data allowed us to model and understand the recycling network, its main functions,
381 and its limitations in consideration of the case study presented in Section 4.

382 5.3. *Impact Assessment (LCIA)*

383 5.3.1. *Calculation methodology*

384 5.3.1.1. *Choice of software.* The software used to perform this life cycle analysis was OpenLCA
385 1.10.3. The main reason for its use was because it is open-source software. This software
386 makes it possible to perform full life cycle analysis and carbon footprint analysis, allowing us
387 to install a wide variety of databases. For this case, the database used was the EcoInvent 3.5.

388 5.3.1.2. *Choice of calculation methodology and impact indicators.* In analyzing the impact
389 methodologies and considering the selected impact categories for this study (see Section 5.1.3),
390 ReCiPe (Global-Hierarchist version) has been chosen as the impact calculation methodology.
391 The ReCiPe method ([Goedkoop et al., 2009](#)) has been chosen mainly because it is widely used
392 in various areas of research ([Dekker et al., 2020](#)). In addition, ReCiPe is an LCIA method
393 that is harmonized in terms of modeling principles and choices, offering results at both the
394 midpoint and endpoint level ([Goedkoop et al., 2009](#)). In addition, it is used for various life
395 cycle analyses in the area of additive manufacturing with a hierarchical cultural perspective
396 ([Saade et al., 2020](#)). Finally, the set of impact categories chosen are presented in Table 2.

397 5.3.1.3. *Impact studies.* Table 4 presents the results of the comparison, showing for the
398 virgin and recycled scenario the amount of impact that is generated in each impact category
399 evaluated. In addition, the percentage reduction (comparing the recycling scenarios with

TABLE 3: Table of impact indicators and their unit of measure

Impact Category	Methodology		Unit
Climate Change	ReCiPe (H)		$KgCO_2 - Eq$
Resources Depletion	Fossil Depletion	ReCiPe (H)	$KgOil - Eq$
	Water Depletion	ReCiPe (H)	$KgP - Eq$
Eutrophication Potential	Freshwater Eutrophication	ReCiPe (H)	$KgN - Eq$
	Marine Eutrophication	ReCiPe (H)	$m3$

the virgin scenario) is presented. To construct this table, the best virgin scenario (least impactful) was considered for comparison. As a result, it can be observed that the production of filament from recycled plastic has a significant advantage in five of the six impact categories (climate change, fossil depletion, freshwater eutrophication, marine eutrophication, and water depletion), which corresponds to the categories that are considered essential for DRAM. In each of the categories there is a reduction of at least 97% compared to the impact produced by the virgin scenario. For the Ion Radiation category, however, the recycling system has a greater impact than the virgin scenario (best scenario in this case). In the recycling scenario, the amount of ion radiation equals approximately 2.8 times the emissions of the virgin scenario.

Figure 3 presents the impacts obtained for each scenario studied. As can be observed, the distributed recycling system to produce 3D filament (Recycling Scenario or Scenario 3) is the least impactful scenario, taking into account the five categories considered pertinent to DRAM.

The result shows that, in each impact category, the recycling scenario pollutes less than 5% compared to the two scenarios that integrate the production of virgin plastic. In other words, the recycling system to produce 3D filament manages to reduce emissions and impacts by at least 95%.

Regarding the other impact categories present in the ReCiPe methodology, it is necessary to emphasize that the recycling system has a significant impact on the category that measures the radiation of ions equivalent to Uranium 235. As can be seen in Figure 3, even Scenario 3, which takes recycling into account, reaches a high radioactive emission compared to the scenarios that use virgin filament. Scenario 1 represents approximately 25% of the impact produced by the recycling system, while Scenario 2 (pessimistic scenario of virgin filament) represents approximately 60% of the impact produced by Scenario 3.

TABLE 4: Scenario definition

Impact categories	Scenarios	Impact of each process				Total
		PLA Production	Supply Transport	Filament production	Delivery	
Climate change (kg CO₂-Eq)	Virgin [1-2]	337.36	[5.69- 18.19]	[1.94 – 32.60]	[54.43 – 785.24]	[411.92- 1160.88]
	Recycled	-	5.4523	2.289	4.957	12.699 (-97%)
Fossil Depletion (kg oil-Eq)	Virgin [1-2]	96.79	[2.30 - 6.96]	[0.50 - 13.13]	[19.58- 285.14]	[123.83 - 397.36]
	Recycled	-	1.961	0.615	1.783	4.359 (-97%)
Freshwater Eutrophication (kg P-Eq)	Virgin [1-2]	0.14	[0.001- 0.002]	[0.001- 0.003]	[0.011- 0.024]	[0.159 0.168]
	Recycled	-	0.0011	0.0016	0.0010	0.004(- 98%)
Ionizing Radiation (kg U235-Eq)	Virgin [1-2]	26.51	[0.45-1.49]	[0.12-0.15]	[4.46- 52.78]	[32.61- 79.86]
	Recycled	-	0.476	122.98	0.406	123.83 (+74%)
Marine Eutrophication (kg N-Eq)	Virgin [1-2]	0.89	[0.004- 0.041]	[0.003- 0.009]	[0.09-1.30]	[1.02-2.20]
	Recycled	-	0.009	0.011	0.008	0.029 (-97%)
Water Depletion (m³)	Virgin [1-2]	37.3669	[0.005- 0.028]	[0.008- 0.112]	[0.064- 0.439]	[37.46- 37.92]
	Recycled	-	0.006	0.546	0.006	0.56 (-99%)

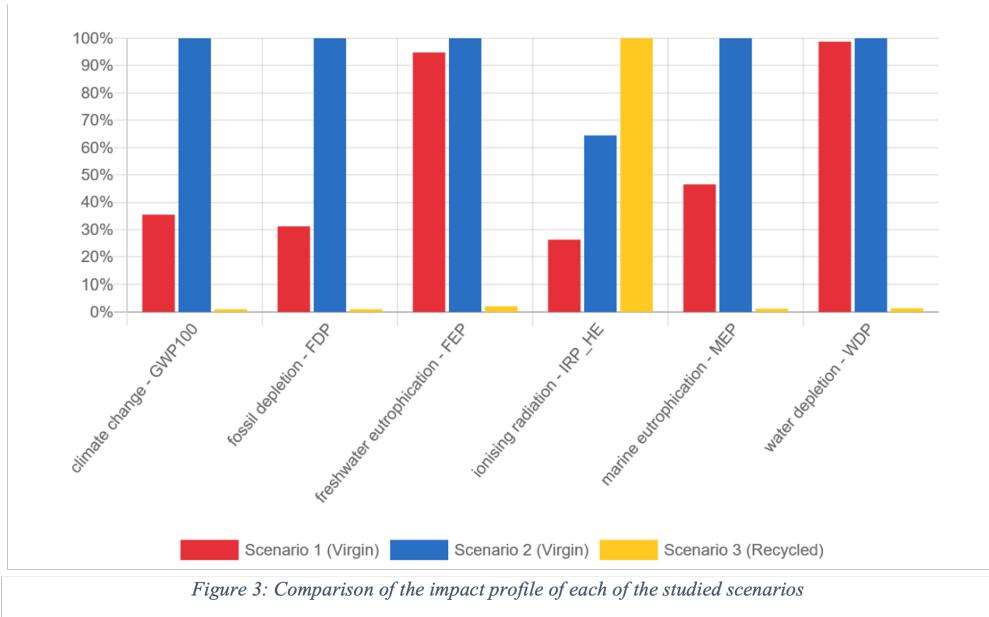


FIGURE 3: Comparison of the impact profile of each of the studied scenarios

424 5.4. Interpretation and Recommendations

425 5.4.1. Sensitivity analysis

426 In order to evaluate the response of each of the systems to context variations, a sensitivity
 427 analysis has been conducted. The sensitivity analysis consists of analyzing the results of the
 428 system against the variation of one of the parameters. This activity allows us to identify
 429 the key contextual parameters that affect the results of the evaluation. These results can
 430 be favorable as well as negative, being able to identify the effect called transfer of impacts.
 431 This effect consists in the fact that while in a specific impact category a reduction in the
 432 level of emissions is achieved through a change of parameters (technology, raw material,
 433 process, type of energy), at the same time there is one or multiple impacts in which an
 434 increase in the emissions or impacts produced can be reflected. This effect (impact transfer)
 435 is fundamental when proposing improvements in products or services. In this study, to have a
 436 better visualization of the transfer of impacts, the complete set of impacts evaluated by the
 437 ReCiPe (H) methodology have been taken into account.

438 5.4.1.1. Location of PLA production. A recent project by Total Corbion has been proposed in
 439 Grandpuits (Seine-et-Marne). The purpose of this project is to build a polylactic acid (PLA)
 440 production plant with a capacity of up to 100,000 tons per year. This first European plant is
 441 to be installed in France⁴. Considering this, a sensitivity analysis has been performed for this
 442 scenario. In this new scenario, the PLA production facility is situated in France at Total's
 443 Grandpuits facility.

444 Figure 4 shows the result of the change in the location of the PLA production plant
 445 (from the United States to France). To represent the sensitivity of the system, only the first
 446 scenario was considered because it has a logistics chain with less impact than Scenario 2.
 447 Thus, it is interesting to see the changes with respect to this scenario. As can be observed,
 448 the result has a relatively small variation. Even though an improvement was obtained in

4. <https://www.usinenouvelle.com/article/bientot-du-pla-made-in-france.N1216857>

449 some categories, such as particle formation, use of fossil raw materials, or ozone depletion,
 450 the result is relatively small. The impact was not greatly reduced with respect to Scenario 1
 451 because PLA production is the main source of impact in this scenario.

452 In conclusion, it was observed that there is a slight improvement in the environmental
 453 performance of the production system, but it is not sufficiently attractive with respect to the
 454 level of impact obtained in the recycling of plastic to produce filament. To obtain a significant
 455 improvement, it is necessary to continue optimizing the PLA production process, which is
 456 enabled mainly by technological advances. This can be a great solution only if it manages to
 457 identify a slight transfer of impacts.

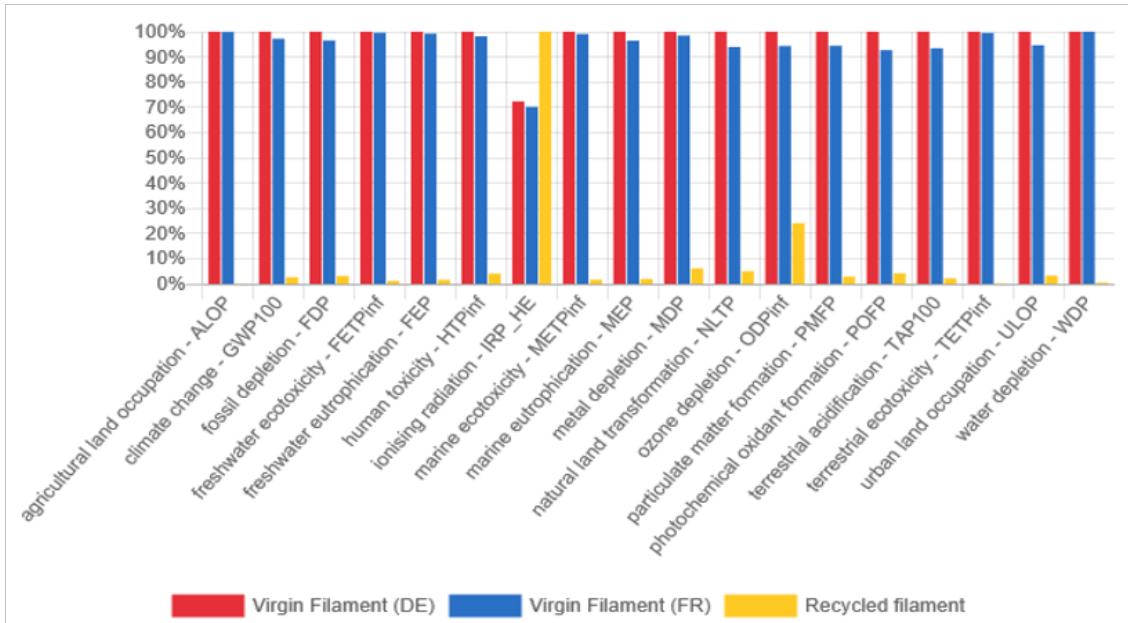


FIGURE 4: Comparison of the impact profile of each of the studied scenarios

458 *5.4.1.2. Energy Source.* The results presented in Section 5.3.1.3 have shown that distributed
 459 plastic recycling to produce filament demonstrates broad environmental advantages compared
 460 to production from virgin plastic. In almost all impact categories it had an impact of less
 461 than 5% of the impact caused by virgin filament production systems, except in the category
 462 related to the emission of radioactive particles. These emissions are closely related to the
 463 production of electricity from nuclear energy, which in France accounts for 77.5% of the total
 464 energy produced in the country. Currently, the french electricity system produce is 77.7%
 465 nuclear, 9.7% hydro power, 3.8% wind power, 3.5% natural gas, 2.2% coal, 1.5% solar PV,
 466 1.3% biofuels and wastes, and 0.3% oil (Pereira and Marques, 2020). For this reason, it is key
 467 to see the response of the system and the possible transfer of impacts when using other forms
 468 of energy production, including mainly the use of clean energy (solar and wind). Figure 5
 469 presents the environmental performance of the recycling system using three different types of
 470 energy (nuclear, wind, and solar).

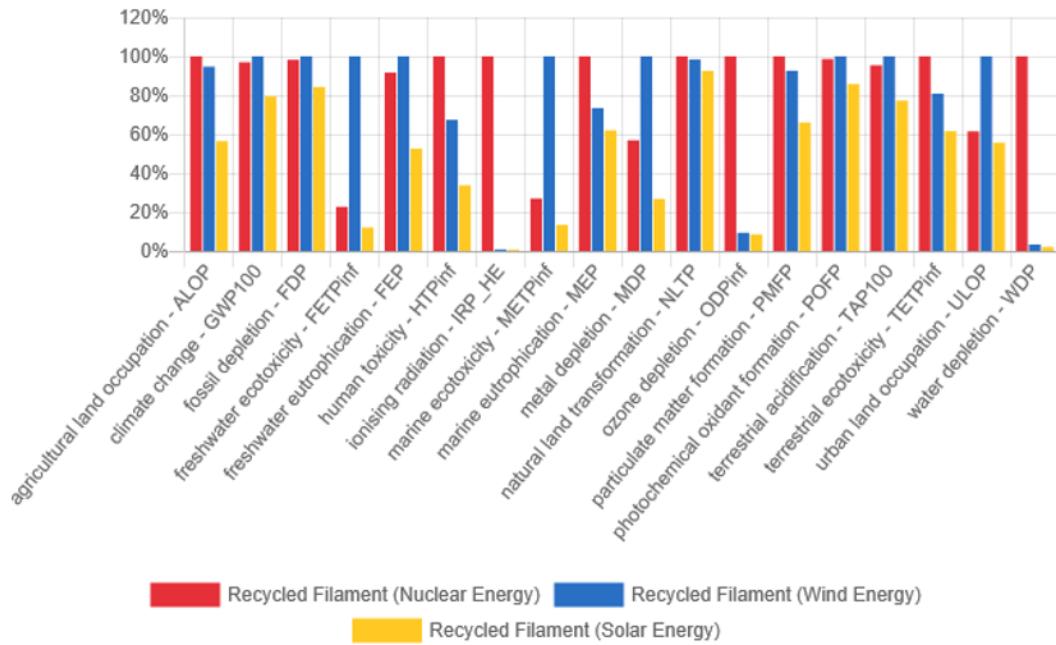


FIGURE 5: Sensitivity analysis respect to the type of energy used

As can be observed in Figure 5, solar energy and wind energy have different responses with respect to the use of nuclear energy. On the one hand, the solar energy (shown in yellow) considerably improves the environmental performance of the system, especially in the category of radioactive emissions, which are reduced by approximately 98%. It is important to remember that this category is the critical point in the comparison of the two filament production systems. In addition, a considerable improvement can be observed in the categories that consider ozone depletion and water use, with reductions of more than 70%. No transfer of impacts is observed in any of the impacts evaluated. According to the impacts assessed, there is no transfer of impacts to other categories, this being a more environmentally friendly option than the use of nuclear energy. On the other hand, wind energy (shown in blue) also achieves significant reductions in the categories of impacts related to radioactive ion emissions, ozone depletion, and water use. However, on this occasion, a transfer of significant impacts can be observed since the impact category measuring, for example, ecotoxicity in water (marine and fresh) increases its impact by 80% with respect to the use of nuclear energy, while the need for metals increases by 40%.

In conclusion, the use of solar energy may seem the best option among the energy sources evaluated. This type of energy presents a circumstantial improvement in the critical point of the use of nuclear energy without suffering a transfer of impacts. The use of wind energy does not have the same result; although it manages to reduce radioactive emissions, the use of this type of energy increases the toxicity present in the water, which can directly affect various ecosystems as well as human health. To determine if this type of energy is suitable for the recycled filament production system, it is necessary to deepen the comparison between these two ways of producing electricity.

494 6. Discussion

495 Distributed recycling via additive manufacturing (DRAM) is been considered by different
496 authors as an additional path to increase the low plastic recycling rates given the major
497 democratization of material extrusion based systems ([Beltrán et al., 2021](#); [Pinho et al., 2020](#);
498 [Wu et al., 2022](#)). Several researchers have studied this recycling approach from a technical and
499 logistical perspective ([Mohammed et al., 2022](#); [Stefaniak et al., 2022](#); [Wu et al., 2022](#)). Other
500 studies that evaluated the environmental impact of this recycling approach only took the
501 recovery and recycling stages into consideration ([Kerdlap et al., 2021](#); [Kreiger et al., 2014](#);
502 [Kreiger and Pearce, 2013](#)). However, an assessment of the positive and negative environmental
503 impacts of implementing this plastic recycling approach with consideration for the whole chain
504 (recovery, recycling, and use) had not been conducted, until now. In this study, a life cycle
505 assessment has been conducted in order to evaluate the environmental impact of implementing
506 a DRAM system to produce recycled PLA filament, compared to traditional virgin PLA
507 filament production systems. In order to carry out this case study, a favorable context was
508 defined in which a considerable amount of plastic is treated, 3D printing is widely used, and
509 there is a dedicated space for plastic recycling via 3D printing.

510 The results of the environmental assessment of this system have shown that the recycling
511 system for filament production reduces the impacts produced by the best scenario of filament
512 production from virgin plastic by at least 97%. The categories of impacts considered were
513 greenhouse gas emissions (climate change), consumption of fossil materials (fossil depletion),
514 overfeeding of aquatic ecosystems (potential eutrophication), and water consumption (water
515 depletion). On the other hand, there is a transfer of significant impacts that is not considered
516 in the relevant impacts of the production system, since, due to the use of nuclear energy to
517 produce electricity in France, the emissions of radioactive ions increased by 280%, which means
518 that the recycling process has worse environmental performance in this category. Despite the
519 assumptions made, the scenarios related to virgin filament production have a major drawback,
520 which is that their most impactful or most influential phase in multiple categories is the
521 production of plastic. For this reason, the change in the location of the production plant
522 only produces a minor environmental improvement. This means that, in order to improve the
523 environmental performance of these systems, it is essential to develop new technologies and
524 optimize the polylactide acid production processes.

525 For the recycling system, the sensitivity analysis performed has shown that the integration
526 of solar energy can greatly reduce the impacts produced using nuclear energy, making it an
527 extremely viable alternative for the recycling system. Wind energy, on the other hand, has a
528 significant transfer of impact to the toxicity present in different types of water, which does not
529 allow us to define in the first instance whether it would be more convenient. In addition, the
530 results showed that the DRAM approach is advantageous considering other energy sources
531 (other than nuclear) that are used in other parts of the world. The environmental assessment
532 carried out in this study extends the results obtained by Kreiger et al. ([2014](#)) and Kerdlap et
533 al. ([2021](#)), showing, based on LCA indicators, the environmental benefits posed by distributed
534 plastic recycling, and more specifically, by implementing a DRAM system. In addition, the
535 environmental benefits are independent of the energy source considered. Therefore, this study
536 indicates that, in environmental terms and under certain conditions, the implementation of
537 DRAMs would have a positive impact on the area of application and could have positive
538 impacts in other contexts. Most notably, to the best of our knowledge, this is the first study to
539 investigate the multidimensional environmental impact of implementing DRAM. The results

540 show a huge potential, in environmental terms, of implementing DRAM and suggest the
541 application of this recycling approach in different contexts, mainly due to the results obtained
542 when considering various energy sources. However, this study is not exempt of limits to be
543 considered in future works. For this reason, the following recommendations can be followed :

- 544 — In order to reduce the complexity of the system studied, in different hypotheses, such
545 as in the recycling system, we considered only one material collected in a clean state
546 and fixed demand from schools. However, in the virgin production system the entire
547 supply and production chain was formulated through hypotheses. Future research could
548 conduct an environmental assessment incorporating these complexities.
- 549 — This study is limited to the comparison of DRAM with virgin filament production. It
550 would be interesting to evaluate the environmental performance of a DRAM system
551 with respect to other possible life-ends for PLA, such as incineration, landfill, or even a
552 system where virgin material is incorporated into the recycling process. This is because
553 the mixture of virgin and recycled material allows for considerable improvement in
554 filament properties and printing quality.
- 555 — Finally, this evaluation was carried out by placing the DRAM system in the specific
556 context of a developed country. It could be interesting to perform the same analysis in
557 developing countries, which currently have major problems in the treatment of plastic
558 waste.

559 7. Conclusions

560 Using the LCA methodology, this work evaluated the environmental benefits of a distri-
561 buted closed-loop supply chain network for plastic recycling using open-source 3D printing
562 technologies in a specific context. The use of polylactic acid (PLA) in the context of Nancy,
563 France considered in the analysis based on three scenarios (two virgin and one recycled). The
564 impact categories studied were climate change, resources depletion (fossil and water) and
565 eutrophication potential (freshwater and marine). The results shown that distributed plastic
566 recycling to produce filament demonstrates broad environmental advantages compared to
567 production from virgin plastic. The comparison shows a reduction of up to 97% in most of the
568 impacts considered for the study. However, given the french electrical mix, the recycling system
569 has a greater impact than the virgin scenario for the Ion Radiation category. Additionally,
570 the sensitivity analysis suggested minor environmental benefits if the production of virgin
571 PLA is placed from USA to France.

572 The results obtained by the application of the LCA methodology to the case study
573 and its sensitivity analysis suggest the application of this recycling approach in various
574 energy contexts (solar, wind, and nuclear), acting in parallel to the existing centralized plastic
575 recycling networks in order to increase plastic recycling rates, which are currently low. It seems
576 interesting for future works to reduce the assumptions used with respect to the production of
577 virgin filament and to integrate the current complexity of carrying out the plastic recycling
578 process. Furthermore, assessing the impacts of this system in various areas allows us to obtain
579 a global vision of the performance of this system and to determine which context would be
580 most favorable for its development.

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⁵⁸⁵ **Declaration of Competing Interest**

⁵⁸⁶ The authors declare that they have no known competing financial interests or personal
⁵⁸⁷ relationships that could have appeared to influence the work reported in this paper.

⁵⁸⁸ **Highlights**

- ⁵⁸⁹ — A comparative LCA for distributed recycling via additive manufacturing (DRAM)
⁵⁹⁰ — Three scenarios are considered (two virgin and one recycled)
⁵⁹¹ — DRAM impacts are less than 5% regarding virgin supply chain except in one category

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Annex 1 Table of reference flows

Scenario 1 Virgin filament System:

Phase	Process	Transport (km)	Transport Type	weight (kg)	weight (t)	t*km
Transport to the filament producer	Plant-Port	756	transport, freight, lorry >32 metric ton, EURO6	1	0,001	0,756
	Port-Port	5.774	Container ship	1	0,001	5,774326
	Port-Plant	558	transport, freight, lorry >32 metric ton, EURO6	1	0,001	0,558

Phase	Machine	Process	Flux	Rate used (kg/h)	Puissance (kw)	Temps d'utilisation (h)	Energie Consomation (kwh)
Filament Product	PEEK 3d Printer Filament Production Line	Extrusion	électricité	20	15	0,05	0,75
				25	15	0,04	0,6

Phase	Process	Transport (km)	Transport Type	weight (kg)	weight (t)	t*km
Filament Transport to the Client	Plant-Client	280	small cargo vehicle	1	0,001	0,28

Scenario 2 Virgin filament System:

Phase	Process	Transport (km)	Transport Type	weight (kg)	weight (t)	t*km
Transport to the filament producer	Plant-Plan	645	transport, freight, lorry >32 metric ton, EURO6	2,1	0,0021	1,3545

Phase	Machine	Process	Flux	Rate used (kg/h)	Puissance (kw)	Temps d'utilisation (h)	Energie Consomation (kwh)
Filament Product	PEEK 3d Printer Filament Production Line	Extrusion	électricité	20	15	0,05	0,75
				25	15	0,04	0,6

Phase	Process	Transport (km)	Transport Type	weight (kg)	weight (t)	t*km
filament Transport to the client	Plant-Airport	5	transport, freight, lorry >32 metric ton, EURO6	2	0,002	0,01
	Airport-Airport	6.971	Airplane	2	0,002	13,942
	Airport-client	152	Cargo truck	2	0,002	0,304

Scenario 3 Recycled filament system:

Phase	Machine	Process	Flow	Power (kw)	Time of Use (h)	Energie Consomation (kwh)		
					Min	Max	Min	Max
Production Pellets	Retsch SM 300	Size reduction	Electricity	3	0,05	0,08	0,15	0,24
Filament Product	Nostek Xcalibur	Extrusion	Electricity	1,6	0,3	1	0,48	1,6

Phase	Transport(km)	Process	Transport type	Weight (fonctional unit) (kg)	Weigh (t)	t*km
Collection/ Delivey	25,52	Recovery Transport	Light Commercial Vehicle	1	0,001	0,02552
		Produit Transport		1	0,001	0,02552

Transport route for the recycling scenarios:

