EBASE manuscript

Marcus W Beck1,✉, Jill M Arriola2, Maria Hermann2, and Raymond G Najjar2

1 Tampa Bay Estuary Program, St. Petersburg, Florida 33701 USA  
2 Pennsylvania State University, University Park, Pennsylvania 16802 USA

✉ Correspondence: [Marcus W Beck <mbeck@tbep.org>](mailto:mbeck@tbep.org)

## Abstract

*Key words*:

## 1 Introduction

* Overview of metabolism - importance
* Existing methods: Odum/WtRegDO, BASE, references in Grace et al. (2015), applications to lakes and streams
* Applications to estuaries - challenges
* Goals and objectives

## 2 Materials and Procedures

The Estuarine BAyesian Single station Estimation (EBASE) method provides a simple approach to quantify metabolic parameters from continuous dissolved oxygen time series. The EBASE method is a modified version of the BASE mass balance equation described in Grace et al. (2015). The original equation is:

where the terms in the square brackets are production (P), respiration (R), and gas exchange (D), respectively, modeled as a function of the change in dissolved oxygen () per unit time () of the input data. Five unknown parameters are estimated from the data: , , , , and . The required inputs are the time series of surface irradiance (, mol/m/s), water temperature (, C), barometric pressure (atm), diel oxygen (, mmol/m), and oxygen saturation (, mmol/m). Dissolved oxygen at saturation is calculated from the input data as a function of water temperature, salinity, and barometric pressure. Several changes, outlined and justified below, were made to [Equation 1](#eq-base) for EBASE.

### 2.1 Modification of the gas exchange parameterization

The underlying idea behind most gas exchange parameterizations for metabolism estimates is that the transfer of gases of relatively low solubility, like oxygen, is proportional to the concentration gradient in a thin film of water at the air–water interface. It is further assumed that the water at the top of the film is saturated with respect to the atmosphere, so that the concentration gradient across the thin film can be approximated as the difference between oxygen concentration measured in the surface water layer, , and the saturation (or equilibrium) concentration , which is usually calculated as a function of water temperature and salinity (the negative sign is needed because positive flux is set to be from water to air):

The proportionality parameter, , usually called the gas transfer velocity because it has the dimensions of length over time, increases as the water column in the vicinity of the air–water interface becomes more turbulent. As such, is usually modeled as a function of measurable variables that can potentially predict turbulent mixing of the surface water, such as wind speed, current velocity, and water depth (Ho et al. 2016). also increases with temperature because as the random motion of gas and water molecules in the thin film speeds up, the film thins out and the transfer of molecules across the film becomes easier. This dependence of on temperature is usually modeled with the Schmidt number, which combines information on the viscosity of the fluid and the diffusivity of the gas. The Schmidt number, in turn, is often parameterized as a gas-specific polynomial function of water temperature and salinity.

Instead of one of the traditional wind-based parameterization used in oceanography (e.g., Wanninkhof 2014), Grace et al. (2015) used the following for the gas transfer in BASE:

where is the average water temperature over each 24-hour period in the diel oxygen time series and is the optimization parameter, often called the reaeration coefficient, which has dimensions of inverse time. [Equation 3](#eq-gasexbase) indicates that increases with temperature and depth. The temperature dependence is consistent with the known decrease of with the Schmidt number, but the depth dependence is difficult to justify. One expects that, for a given current speed, would increase as the water depth decreases because turbulence generated at the sediment–water interface by currents would more easily reach the surface (Ho et al. 2016).

Based on the above differences, the first modification to [Equation 1](#eq-base) for EBASE is use of a wind-based parameterization for gas exchange that is more appropriate for estuaries (Wanninkhof 2014):

where is the wind speed at 10-m height, is water column depth (m, as a fixed parameter or time series from a pressure sensor) and is the Schmidt number defined as the ratio of the kinematic viscosity of water to the molecular diffusivity of the gas for oxygen (from water temperature and salinity using the polynomial fit of Wanninkhof 2014). The parameter () must be estimated from the data.

The remaining variables in [Equation 4](#eq-basewan) are defined as in [Equation 1](#eq-base), with the exception that is calculated using only temperature (C) and salinity (psu), removing the need for barometric pressure. Methods from García and Gordon (1992) are used as follows:

where is thr solubility of dissolved oxygen per volume of seawater at the given temperature and (kg/m) is seawater density at zero elevation given temperature and salinity (Millero and Poisson 1981). is calculated using a polynomial equation:

where the coefficients are = 5.80818, = 3.20684, = 4.11890, = 4.93845, = 1.01567, = 1.41575, = -7.01211e-3, = -7.25958e-3, = -7.93334e-3, = -5.54491e-3, and = -1.32412e-7. is temperature scaled as input to [Equation 6](#eq-co) defined as:

### 2.2 Modification of the production and respiration parameterization

The second modification to [Equation 1](#eq-base) for EBASE is a simplification to remove the and parameters from the production (P) and respiration (R) components, respectively, of the mass balance equation. The BASE method provides options to estimate both and from the data or to set as fixed values defined in the model file (, ). For EBASE, the temperature dependency of respiration is removed by setting = 1. The parameter, which is dimensionless, is also set to 1 as in Holtgrieve et al. (2010). Removing these components improved model stability by creating more reasonable estimates for the other parameters. This further simplifies [Equation 4](#eq-basewan) to the final form for EBASE:

As for BASE, the metabolic estimates in EBASE are defined by the change in dissolved oxygen over the time step, , where gross production is provided by , respiration is provided by , and gas exchange is provided by the remainder (as in Wanninkhof 2014). Units for each of the metabolic estimates are .

The remaining three parameters (light efficiency, ), (respiration, ), and () are estimated by likelihood given the observed data and prior distributions for the parameters. The JAGS software (as for BASE, Plummer et al. 2003) is used with the EBASE R package (described below, Beck et al. 2022; R Core Team 2022) to estimate the unknown parameters using Markov Chain Monte Carlo simulations in a Bayesian framework. At each time step, the change in oxygen concentration between time steps is calculated from the equation using model inputs and parameter guesses, and then a finite difference approximation is used to estimate modeled oxygen concentration. Estimations proceed using a forward in time integration where the dissolved oxygen at each time step is informed by parameter estimates from the preceding time steps. The estimated dissolved oxygen concentration is also returned, which can be compared to the observed values as one measure of model performance. All parameter inputs and outputs for EBASE are shown in Table .

### 2.3 Bayesian prior

The metabolism estimates and their parameters returned by the Bayesian routine implemented in JAGS are affected by the prior distributions assigned to each. As for the BASE model, relatively uninformed prior distributions following a normal Gaussian distribution are used by default, although the priors can be changed based on previous knowledge of parameters specific to an ecosystem or as informed by other metabolic modeling approaches. The default priors were based on those that returned reasonable parameter estimates using a known time series created with forward-feeding oxygen mass balance inputs to [Equation 8](#eq-ebase). (Fwoxy R package, Arriola and Herrmann 2022). All of the physical parameters were held constant, e.g., wind speed, temperature, and salinity. The other required inputs were the , , and parameters, chosen as 0.2 , 20 , and 0.251 , respectively. An initial oxygen concentration was provided as 250 mmol/m and the time series was created by forward addition of the inputs. As such, the dissolved oxygen output as change per unit time was modelled by EBASE using the constant the same physical parameters as input. Reasonable prior distributions by EBASE were identified (Equations [9](#eq-aprior), [10](#eq-rprior), and [11](#eq-bprior)) based on the ability to return the known parameters:

Each follows a normal distribution with means and standard deviation defined in the parentheses ([Figure 1](#fig-priorplot)). The notation is JAGS model syntax indicating the distributions are truncated to a minimum (left) and maximum (right) value defined in the parentheses. All prior distributions were constrained to positive values based on known physical constraints and requirements of the model formula in [Equation 8](#eq-ebase). Additionally, the prior distribution for was constrained to an upper limit of 0.504 (~twice the default mean). Wanninkhof (2014) states that the parameter typically does not vary by more than 20% and the standard deviation of the default prior is sufficiently high to expand beyond this range. Initial development of EBASE also showed that using an undefined upper limit of led to unstable and unreasonable parameter estimates.

### 2.4 EBASE R package

For ease of application, an R package was developed to implement the EBASE methods described above (Beck et al. 2022). A detailed web page (<https://fawda123.github.io/EBASE/>) describes the required data inputs for estimating the metabolic parameters using the core R functions that pass inputs to the JAGS software. A JAGS model file is included with the package that implements [Equation 8](#eq-ebase), with options to supply a custom model file as needed. Users can also specify alternative prior distributions from those in Equations [9](#eq-aprior), [10](#eq-rprior), and [11](#eq-bprior) for each parameter. Functions are available for viewing model results, including a plot of the modelled dissolved oxygen with the observed, a time series plot of the primary metabolic estimates (P, R, and D), and a time series plot of the credible intervals for the , , and parameters. An additional option also allows the user to specify the model optimization period, where the time period can vary from a minimum of one day to the maximum of the entire length of the time series. This differs from the BASE approach where the optimization period is set at one day. Sensitivity of the EBASE method to different optimization periods is described in [Section 3.1](#sec-cmpmod). All results provided herein were generated using the EBASE R package.

## 3 Assessment

### 3.1 EBASE comparison with modelled observations

A preliminary assessment of the ability of EBASE to produce reasonable parameter estimates for , , and ([Equation 8](#eq-ebase)) was conducted using a synthetic time series with actual physical data as inputs. These assessments evaluated if 1) EBASE estimates similar values for the known parameters, and 2) how the estimated values change with different model configurations. The synthetic time series used inputs from an approximate one year of continuous data from Apalachicola Bay, Florida. The input data were 2012 observations from February 22nd to December 19th at the Cat Point water quality monitoring station (29.7021 N, -84.8802 W) and included 15 minute observations of dissolved oxygen (mg/L), water temperature (C), and salinity (psu). Continuous meteorological observations from the nearby East Bay station (29.7909 N, -84.8834 W, ~10km north of Cat Point) were combined with the water quality data at the same time step and included air temperature (C), wind speed (m/s), and PAR (mmol/m as total per 15 minute observation). An oxygen time series was generated using the inputs from Apalachicola Bay ([Figure 2](#fig-synapa)a), such that the synthetic dissolved oxygen was created as close as possible to the observed given the inputs. Known values for and parameters were returned for each time step of the synthetic data, as well as all metabolic estimates for production, respiration, and gas exchange ([Figure 2](#fig-synapa)b and c). The only parameter that could not be created as a continuous time series was , which was set to 0.251 (Wanninkhof 2014).

Several configurations of the EBASE model can influence how the parameters and metabolic rates are estimated. Two types of sensitivity analyses were conducted to evaluate changes in the results returned by EBASE, where in all cases, the results were compared to the known values in the synthetic time series. First, results were compared to the known values after changing characteristics of their prior distributions. Second, results were compared using a model period of optimization for one day and seven days. Details of these analyses are as follows.

The standard prior distributions for the EBASE parameters follow normal distributions with defined means and standard deviations (Equations [9](#eq-aprior), [10](#eq-rprior), and [11](#eq-bprior)). The sensitivity analysis evaluated changes in the distributions by varying the standard deviations from small to large to test the effect of more or less constrained ranges, respectively, on the priors. For each parameter, low, medium, and high values for the standard deviation were assessed as 10%, 100%, and 1000% of the default values. Each unique combination of low, medium, and high values for the standard deviation of each parameter were evaluated, creating 3 = 27 different combinations of prior distributions that were used with the synthetic time series in EBASE.

The optimization period used for a time series determines how much data are used to estimate the three EBASE parameters (, , and ). Unlike BASE, the EBASE software allows flexibility in the period of time used for optimization, with a minimum period of time being one day and a maximum period being the length of the time series. For a chosen optimization period, single estimates for , and are returned by the model. A short period of optimization (e.g., one day) may return unstable parameter estimates if forcing factors (e.g., wind, temperature) that affect estuarine biogeochemical rates occur at time scales longer the chosen period of optimization for EBASE. As such, potentially longer optimization periods may return more stable and robust estimates given that more data are used for parameter estimation. An optimization period of one day was compared to results using an optimization period of seven days for the synthetic one-year time series. Each optimization period was evaluated for every unique combination of prior distributions, totaling 54 (27 times the two optimization periods) unique analyses for the assessment of changing standard deviations of all parameters.

Changing the variance of each of the prior distributions affected the ability of EBASE to return the known parameters from the synthetic time series. [Figure 3](#fig-apasumdat) shows a summary of the comparisons of the synthetic data with EBASE for each unique combination of prior distributions and two optimization periods of one and seven days. The results are compared using R-squared () values from the linear fit between the synthetic data and results from EBASE. Darker colors represent lower values and can be used to assess how changing the variance of the priors affects the ability of EBASE to estimate the true parameter.

Several conclusions can be made from the results in [Figure 3](#fig-apasumdat) based on the patterns of the values. Overall, the model was able to estimate the parameter with reasonable precision regardless of any of the prior distributions, whereas most of the variation was observed in the (respiration, R) and parameters. Changing the variance of the parameter had little effect on any of the comparisons. Smaller variance of the prior distribution for the parameter generally contributed to lower for all parameters, although the precision was worse for the seven day optimization period (e.g., the parameter). Conversely, larger variance of the prior distribution for the parameter generally contributed to lower only for the gas exchange estimate (D), although this effect was minimized with the seven day optimization period.

The results suggest that the default prior distributions used in EBASE may be appropriate in some cases, but increasing the prior variance for and and decreasing the variance for will likely produce more optimal solutions. The actual model results for the best performing combination of prior distributions for the one day and seven day optimization are shown in [Figure 4](#fig-optex) (all results are averaged at the time-step of the optimization period for comparability). The best performing models were those with the highest mean values across all parameter estimates in [Figure 3](#fig-apasumdat). For both models, the metabolic parameters were estimated with high precision, and in some cases with nearly perfect precision (i.e., dissolved oxygen and gas exchange, D). Estimates for the fixed parameter were most variable, but not biased as the mean for both models was close to 0.251 .

### 3.2 EBASE comparison with real observations

Metabolic estimates from EBASE were further compared to those from existing methods to benchmark and evaluate potential differences from alternative model formulations used by each approach. The observed one year continuous time series described above for water quality and weather data from Apalachicola Bay was used to compare metabolic estimates from EBASE, BASE, and the Odum open-water method (Odum 1956). The Odum method used an approach described in Murrell et al. (2018) created for estuarine application that uses a gas-exchange approach described in Thébault et al. (2008). For better comparison with the EBASE methods, the gas exchange parmaterization in the Odum method was modified to use the Wanninkhof (2014) approach (as available in the WtRegDO R package, Beck 2021). The input data were provided to each method after converting the units as required (e.g., BASE requires PAR as mol/m/s). Water column depth at Cat Point is approximately 1.8 m and all daily metabolic estimates returned by each method were converted to volumetric rates by dividing by the depth (i.e., ). Each comparison also evaluated metabolic estimates using the observed dissolved oxygen time series and a detided time series using weighted regression (Beck et al. 2015; available from Beck 2021). The latter comparison provided an assessment of metabolic rates that are expected to be minimally influenced by tidal advection. All metabolic estimates were obtained using the default settings for each of the three methods, with an exception that the prior distribution for the reaeration coefficient () in BASE was informed by results from the Odum output. Initial assessments indicated that unstable and inaccurate results for the time series are obtained from BASE using an uninformed prior distribution for . All other prior distributions for BASE and EBASE were set as the software default. A one day model optimization period was used for EBASE for comparability with BASE.

The common set of metabolic parameters returned by the Odum, BASE, and EBASE methods are net ecosystem metabolism, gross production, respiration, and gas exchange. Pairwise comparisons of the daily estimates returned by each method were evaluated using simple summary statistics, including Pearson correlation coefficients and linear regression fits to evaluate potential biases in each method. For the latter analysis, differences in intercept estimates less than or greater than zero suggested a bias in magnitude, whereas differences in slope estimates less than or greater than one suggested a bias that depended on relative value of each observation (e.g., bias may differ between lower or higher values). Comparisons of the intercept and slope estimates to zero and one, respectively, were based on confidence intervals from the standard errors of each. Results were assessed separately for estimates based on the observed or detided dissolved oxygen time series.

[Figure 5](#fig-apacmpfig) and Table provide a comparison of the EBASE metabolic estimates with those from the Odum and BASE methods. Correlations for all estimates, regardless of method or the input time series, were positive and significantly correlated, as expected. The only exception was the comparison of respiration for the Odum and EBASE methods ( = 0.11), where the former produced a much larger range of estimates. The intercept values were all significant, suggesting some level of bias when comparing estimates between the methods. Slope estimates were all significantly different from one for the comparisons using the observed dissolved oxygen time series as input, whereas only three of the eight comparisons were significantly different using the detided time series. This latter comparison suggests that detiding the time series produces more comparable estimates between methods. Similarly, the standard deviation () of the residuals of each regression comparison were generally lower in the detided results for all comparisons. Variation in the gas exchange estimates were also more similar comparing the Odum and EBASE methods that used the same formulation (Wanninkhof 2014) regardless of input time series.

As noted in the previous paragraph, metabolic estimates between the methods were more similar for results using the detided dissolved oxygen time series ([Figure 5](#fig-apacmpfig)b) as compared to using the observed dissolved oxygen as input ([Figure 5](#fig-apacmpfig)a). This suggests that tidal effects on continuous monitoring data can severely violate assumptions required for methods of estimating ecosystem metabolism (as described in detail in Beck et al. 2015). These effects are easily identified with estimates from the Odum method as negative values for production (P) and positive values for respiration (R), i.e., “anomalous” values. The effects of tidal advection on estimates from EBASE and BASE are more subtle as anomalous values cannot be returned based on constraints in the mass-balance equations (e.g., it is impossible for respiration to be positive). However, the effects of tidal advection produced a wider range of estimates for NEM in EBASE and BASE compared to the Odum method ([Figure 5](#fig-apacmpfig)a), caused by greater variation in the production and respiration estimates. Overall, detiding the dissolved oxygen time series prior to estimating metabolic rates produces more stable (lower standard deviation) and smaller ranges (Table ), regardless of the method used.

## 4 Discussion

* Describe how EBASE deals with gaps

## 5 Comments and Recommendations

## Acknowledgments

## Figures

|  |
| --- |
| Figure 1: The default prior distributions for the , , and parameters used in EBASE. All priors are normal distributions and truncated to positive values. The parameter is also truncated to a maximum of 0.504. |

|  |
| --- |
| Figure 2: Synthetic time series for one year of continuous data at Apalachicola Bay. Synthetic input to EBASE was the (a) dissolved oxygen time series, with results compared to the (b) synthetic parameter and (c) synthetic metabolic estimates for production (P), respiration (R), and gas exchange (D). The parameter is shown as an areal value for comparison with the metabolic results. |

|  |
| --- |
| Figure 3: Comparison of EBASE results to a synthetic time series created with known metabolic parameters. Results are compared for unique combination of prior distributions, where the standard deviations are evaluated from low (L), medium (M), to high (H) for a one day and seven day model optimization period. The metabolic results for dissolved oxygen (DO), production (P), respiration (R), gas exchange (D), and the parameter are evaluated based on the value of the linear fit between the EBASE results and synthetic values. The parameter is a fixed constant and was not evaluated. |

|  |
| --- |
| Figure 4: Comparison of EBASE results for the best performing model to a synthetic time series created with known metabolic parameters. The best performing model for the one day (a, ndays = 1) and seven day (b, ndays = 7) optimization period is based on the EBASE results with prior distributions that produced the highest mean values across all parameter estimates ([Figure 3](#fig-apasumdat)). The standard deviations of the prior distributions of each parameter for the best performing model are shown at the top. The results are averaged at the time step of the model optimization period. The parameter is fixed at 0.251 . |

|  |
| --- |
| Figure 5: Comparisons of daily metabolic estimates from the Odum, BASE, and EBASE methods for one year of continuous monitoring data at Apalachicola Bay. Results for net ecosystem metabolism (NEM), gross production (P), respiration (R), and gas exchange (D) are compared using the (a) observed and (b) detided dissolved oxygen time series as input to each model. All values are volumetric as mmol O m d. Black lines show the 1:1 comparison and the red lines show the linear regression fit. |

## Tables

**Table** **:** EBASE model parameters showing the required input, parameters derived from the model input, and model output. The model notation for production, respiration, and gas exchange shows both the simple parameter name and the full model component used to estimate the parameter.

| Type | Description | Model notation | Units |
| --- | --- | --- | --- |
| Input |  |  |  |
|  | Dissolved oxygen |  |  |
|  | Water temperature | - |  |
|  | Salinity | - |  |
|  | Total photosynthetically active radiation |  |  |
|  | Wind speed | - |  |
|  | Water column depth |  |  |
| Input-derived |  |  |  |
|  | Wind speed at 10 meter height, squared |  |  |
|  | Schmidt number (from water temperature and salinity) |  | unitless |
|  | Dissolved oxygen at saturation (from water temperature and salinity) |  |  |
| Output |  |  |  |
|  | Dissolved oxygen (modelled) |  |  |
|  | Production |  |  |
|  | Respiration |  |  |
|  | Gas exchange |  |  |
|  | Light efficiency |  |  |
|  | b |  |  |

**Table** **:** Summary statisics of daily metabolic estimates comparing Odum and BASE to EBASE for net ecosystem metabolism (NEM), gross production (P), respiration (R), and gas exchange (D). Separate comparisons are made for estimates based on observed and detided dissolved oxygen. Summary statistics include Pearson correlation coefficients and intercept, slope, and residual standard deviation of a linear fit between the same metabolic estimates obtained from each method. Significance of the p-values for the intercept and slope indicates if the estimate is different from zero or one, respectively.

| Dissolved Oxygen | Comparison | Estimate |  | Intercept | Slope |  |
| --- | --- | --- | --- | --- | --- | --- |
| Observed |  |  |  |  |  |  |
|  | Odum v EBASE | NEM | 0.23\*\* | -11.4\*\* | 0.23\*\* | 26.26 |
|  |  | P | 0.28\*\* | 15.23\*\* | 0.56\*\* | 43.69 |
|  |  | R | 0.11 | -23.14\* | 0.65 | 54.12 |
|  |  | D | 0.74\*\* | 5.15\*\* | 0.65\*\* | 6.14 |
|  | BASE v EBASE | NEM | 0.89\*\* | -8.14\*\* | 1.11\*\* | 15.12 |
|  |  | P | 0.47\*\* | 23.23\*\* | 0.69\*\* | 29.48 |
|  |  | R | 0.59\*\* | 22.38\*\* | 2.98\*\* | 36.93 |
|  |  | D | 0.57\*\* | 9.29\*\* | 0.48\*\* | 7.12 |
| Detided |  |  |  |  |  |  |
|  | Odum v EBASE | NEM | 0.74\*\* | -4.93\*\* | 1.03 | 7.77 |
|  |  | P | 0.52\*\* | 7.46\*\* | 0.93 | 10.73 |
|  |  | R | 0.53\*\* | -14.03\*\* | 0.87 | 15.72 |
|  |  | D | 0.8\*\* | 1.89\*\* | 1.04 | 6.66 |
|  | BASE v EBASE | NEM | 0.59\*\* | -7.16\*\* | 0.52\*\* | 5.91 |
|  |  | P | 0.87\*\* | 2.79\*\* | 1.01 | 3.95 |
|  |  | R | 0.74\*\* | -9.54\*\* | 0.75\*\* | 7.59 |
|  |  | D | 0.54\*\* | 9.12\*\* | 0.42\*\* | 5.60 |
| \* p < 0.05, \*\* p < 0.005 | | | | | | |

## References

Arriola, J. M., and M. Herrmann. 2022. Fwoxy: Forward model of estuary dissolved oxygen time series,.

Beck, M. W. 2021. WtRegDO: Implement weighted regression on dissolved oxygen time series,.

Beck, M. W., J. D. Hagy III, and M. C. Murrell. 2015. Improving estimates of ecosystem metabolism by reducing effects of tidal advection on dissolved oxygen time series. Limnology and Oceanography: Methods **13**: 731–745. doi:[10.1002/lom3.10062](https://doi.org/10.1002/lom3.10062)

Beck, M. W., M. Herrmann, J. Arriola, R. Najjar, and W. McGillis. 2022. EBASE: Estuarine Bayesian Single-Station Estimation Method for Ecosystem Metabolism,.

García, H. E., and L. I. Gordon. 1992. Oxygen solubility in seawater: Better fitting equations. Limnology and oceanography **37**: 1307–1312. doi:[10.4319/lo.1992.37.6.1307](https://doi.org/10.4319/lo.1992.37.6.1307)

Grace, M. R., D. P. Giling, S. Hladyz, V. Caron, R. M. Thompson, and R. Mac Nally. 2015. Fast processing of diel oxygen curves: Estimating stream metabolism with BASe (BAyesian Single-station Estimation). Limnology and Oceanography: Methods **13**: 103–114. doi:[10.1002/lom3.10011](https://doi.org/10.1002/lom3.10011)

Ho, D. T., N. Coffineau, B. Hickman, N. Chow, T. Koffman, and P. Schlosser. 2016. Influence of current velocity and wind speed on air-water gas exchange in a mangrove estuary. Geophysical Research Letters **43**: 3813–3821. doi:[10.1002/2016GL068727](https://doi.org/10.1002/2016GL068727)

Holtgrieve, G. W., D. E. Schindler, T. A. Branch, and Z. T. A’mar. 2010. Simultaneous quantification of aquatic ecosystem metabolism and reaeration using a Bayesian statistical model of oxygen dynamics. Limnology and Oceanography **55**: 1047–1063. doi:[10.4319/lo.2010.55.3.1047](https://doi.org/10.4319/lo.2010.55.3.1047)

Millero, F. J., and A. Poisson. 1981. International one-atmosphere equation of state of seawater. Deep Sea Research Part A. Oceanographic Research Papers **28**: 625–629. doi:[10.1016/0198-0149(81)90122-9](https://doi.org/10.1016/0198-0149(81)90122-9)

Murrell, M. C., J. M. Caffrey, D. T. Marcovich, M. W. Beck, B. M. Jarvis, and J. D. Hagy. 2018. Seasonal oxygen dynamics in a warm temperate estuary: Effects of hydrologic variability on measurements of primary production, respiration, and net metabolism. Estuaries and Coasts **41**: 690–707. doi:[10.1007/s12237-017-0328-9](https://doi.org/10.1007/s12237-017-0328-9)

Odum, H. T. 1956. Primary production in flowing waters. Limnology and Oceanography **1**: 102–117.

Plummer, M. and others. 2003. JAGS: A program for analysis of Bayesian graphical models using gibbs sampling. *Proceedings of the 3rd international workshop on distributed statistical computing*. Vienna, Austria. 1–10.

R Core Team. 2022. [R: A language and environment for statistical computing](https://www.R-project.org/), R Foundation for Statistical Computing.

Thébault, J., T. S. Schraga, J. E. Cloern, and E. G. Dunlavey. 2008. Primary production and carrying capacity of former salt ponds after reconnection to San Francisco Bay. Wetlands **28**: 841–851. doi:[10.1672/07-190.1](https://doi.org/10.1672/07-190.1)

Wanninkhof, R. 2014. Relationship between wind speed and gas exchange over the ocean revisited. Limnology and Oceanography: Methods **12**: 351–362. doi:[10.4319/lom.2014.12.351](https://doi.org/10.4319/lom.2014.12.351)