Initial wildfire suppression efforts select for more extreme fuel and

² climate burning conditions in Sierra Nevada forests

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10 Abstract

Typical fire effects in dry western pine and mixed-conifer forests are dominated by large wildfires, and potential fire effects of rapidly suppressed fires are never realized. Milder fuel and weather conditions facilitate suppression when fires are still small, and thus most extensive fire effects in forest landscapes occur during more extreme conditions. This amounts to a selection pressure on burning conditions that may lead to a selection bias in average fire effects. We formalize this selection framework and measure its influence on correlated, multivariate burning conditions of fuel and climate using the evolutionary ecology concept of phenotypic selection. Using a new dataset that includes smaller fires (>4ha), we examine fire "survivorship" to initial attack suppression efforts as a fitness metric. Our analysis found that initial containment efforts select for fires burning in more homogeneous fuel conditions and during hotter/drier conditions. Fire effects on western forest conditions arise from a complex social-ecological system, with management decision-making having a strong ability to influence outcomes. We show a strong selection pressure on burning conditions imposed by management, and encourage further dismantling of barriers to applying this selection for resource benefit, such as by expanding implementation of wildfire use fires where natural ignitions are allowed to burn under moderate conditions.

25 Introduction

- 26 A legacy of fire suppression is an oft-cited root cause underpinning the increasing size and severity of wildfires
- in the dry western U.S. pine and mixed-conifer forests (Miller and Thode 2007; Calkin et al. 2015; Safford and
- 28 Stevens 2017). While most of this ecosystem would experience frequent, low- to moderate-severity wildfire

every 8 to 15 years in the several centuries prior to Euroamerican settlement, suppression management has largely eliminated fire effects from much of western dry forested land in the past 100 years (Steel et al. 2015; Safford and Stevens 2017; Miller and Safford 2017). A lack of frequent fire has led to densification of these forests, which increases fuel loading and homogenizes forest structure (Fulé et al. 1997; Veblen et al. 2000; Keane et al. 2002; Collins et al. 2016; Stephens et al. 2018). Synergistic alignment of these extreme fuel conditions with earlier snow melt, longer fire seasons, and hotter droughts (aka "climate change droughts") (Westerling 2006, 2016; Abatzoglou and Kolden 2013; Abatzoglou and Williams 2016) increases the probability that fires will generate self-propagating behavior (Coen et al. 2018) and kill all (or nearly all) overstory vegetation (Koontz et al. 2019b) in large, contiguous patches of mortality (Stevens et al. 2017). Most dry western conifer forests are ill-adapted to regenerate in the centers of these large patches, which are far from seed sources (Welch et al. 2016; Stevens-Rumann et al. 2018). Thus the modern trend of atypically large, contiguously stand-replacing fires in this system compromises forest resilience and increases the potential for long-term shifts in vegetation type to shrub- or grasslands (North et al. 2009; Millar and Stephenson 2015; Stevens et al. 2017).

Ongoing fire suppression also strongly influences fire effects. Fire suppression generally is very effective at

Ongoing fire suppression also strongly influences fire effects. Fire suppression generally is very effective at its immediate goal of extinguishing fires. Between 1970 and 2002, 97 to 99% of fires burning on U.S. Forest Service land were contained before they reached 120 hectares in size (Calkin et al. 2005, 2015). The "10 a.m. policy", which dictated that fires should be put out before 10 a.m. on the day following their discovery, has been modified since its establishment as firefighting policy in the 1930's but its aggressive spirit persists in today's modern firefighting apparatus (Dale 2006; Johnson et al. 2009). Mild weather and fuel conditions facilitate early suppression efforts (Calkin et al. 2015; Abatzoglou et al. 2018), and the fires that escape initial attack and grow large are often assumed to have grown to these sizes because they were burning under more extreme conditions (Calkin et al. 2014).

While the vast majority of fires are managed for suppression in the western U.S. (Calkin et al. 2015), a small number of natural ignitions are allowed to burn under moderate conditions as "wildland fire use" (WFU) fires in recognition of the benefit that fire has to forest resources (Davis 1979). WFU fire effects tend to fall within the natural range of variation for western pine and mixed-conifer systems (Meyer 2015; Walker et al. 2018). Though many studies recommend WFU fires as a means to restore forest resilience (Mallek et al. 2013; Meyer 2015; North et al. 2015; Collins et al. 2017; Stevens et al. 2017), barriers remain to their more widespread adoption (Doane et al. 2006).

The same mild fuel and weather conditions that contribute to beneficial outcomes of WFU fires contribute to early success of suppression efforts. Fire effects that may have arisen from fires that succumb to early

- 51 suppression efforts never materialize. Instead, fire effects from suppression fires in this ecosystem are
- dominated by large fires, which account for approximately 97.5% of total burned area, exhibit fire effects that
- 63 increase the likelihood of state change, and likely burn under the more extreme fuel and weather conditions
- that hindered early suppression efforts (Calkin et al. 2005). Thus, the general short-term success of fire
- 55 suppression policy paired with its long-term cumulative effect has led to a management paradox with respect
- to forest management aiming to restore resilience: we shouldn't put out the fires that we can, and we can't put
- out the fires that we should.
- 68 Wildfire suppression management has shifted the distribution of fire behavior, and therefore the distribution
- of fire effects, to be more extreme (Calkin et al. 2015). Here, we introduce the evolutionary ecology concept
- 70 of "phenotypic selection" as a formal framework for measuring the magnitude of this distributional shift-
- the selection by suppression. We use a new dataset of fire severity compiled for yellow pine/mixed-conifer
- ₇₂ forests (Koontz et al. 2019a) to quantify the strength of "selection" on wildfire burning conditions (regional
- climate, wind speed, vegetation density, vegetation continuity) imposed by initial attack suppression efforts.
- ⁷⁴ In particular, we examine the following questions:
- 1. How do initial burning conditions (regional climate, vegetation density, vegetation continuity) differ
- between suppression fires that are successfully contained by initial attack and those that escape
- containment efforts?
- 2. What are the consequences of these varying conditions for wildfire effects (burn duration, fire event
- size, severity configurations)?
- 3. How might this selection be biasing our understanding of fire effects arising from suppression and
- wildfire use fires?

Methods

83 Study system

- The Sierra Nevada yellow pine/mixed-conifer (hereafter Sierra YPMC) is a disturbance-prone forest system in
- the Sierra Nevada mountain range of California, U.S.A. It spans the full 628 kilometer latitudinal length of the
- see Sierra Nevada, and 2,500 meters of elevation (300 meters to 2800 meters), primarily on the western slope of
- the mountain range (Safford and Stevens 2017). The forest is dominated by ponderosa pine (*Pinus ponderosa*),
- sugar pine (Pinus lambertiana), white fir (Abies concolor), and incense cedar (Calocedrus decurrens) in varying
- mixes. Prior to Euroamerican settlement, the ecosystem experienced frequent, low- to moderate-severity
- wildfire every 8 to 15 years on average (Steel et al. 2015; Safford and Stevens 2017), which consumed surface

- fuels but generally had minimal effects on large, established trees. This fire regime generated heterogeneous horizontal forest structure, with groups of relatively even-aged trees having interlocking crowns, individual trees with distinct crowns, and variably-sized gaps between these tree clump and individual tree structural features (Lydersen et al. 2013). A century of fire suppression has led to infill of these gaps, homogenizing the horizontal forest structure, increasing vertical continuity of fuels, and compromising forest resilience in an era of climate change-induced hotter droughts (North et al. 2009; Millar and Stephenson 2015; Collins et al. 2016).
- For our study, we compiled the Sierra YPMC type using the U.S. Forest Service Fire Return Interval Departure (FRID) dataset and included "dry mixed-conifer", "moist mixed-conifer", and "yellow pine" vegetation types following Steel et al. (2018) and Koontz et al. (2019b). These classifications represent "potential vegetation" given the climate of the area, such that there is no influence of recent disturbance events (Harvey et al. 2016; Steel et al. 2018; Koontz et al. 2019b).

103 Context of Sierra YPMC wildfire

To describe the modern context of wildfire activity in the Sierra Nevada yellow pine/mixed-conifer system,
we used geospatial records contained in Short (2017) (U.S. Forest Service Fire Program Analysis Fire
Occurrence Database; FPAFOD), the most comprehensive database of wildfire occurrence for the United
States representing 1.88 million wildfire records from 1992 to 2015.

The FPAFOD contains point locations for the centroids of each fire's footprint, rather than the perimeter of each fire as in some other databases (Eidenshink et al. 2007). We spatially subsetted the FOD data to 109 fire events whose centroids occurred within the Sierra Nevada mountain range, as defined by the Jepson 110 geographic subdivisions (north, central, and south Sierra Nevada Foothills and High Sierra Nevada, as well 111 as the Tehachapi Mountain Area) (JepsonFloraProject 2016). For each fire record, we approximated its 112 footprint by creating a circular buffer around the centroid with an area equivalent to the reported area of 113 the fire. Using this footprint approximation, we calculated the proportion of area that intersected with our 114 compilation of Sierra YPMC from the FRID dataset. We retained all fires with greater than zero area of the approximate footprint covering the Sierra YPMC extent. We calculated burn duration as the number of 116 days between the containment date and the alarm date, and retained all fires with a burn duration of greater than 0 and less than 364 days to eliminate likely errors in reporting of alarm and containment dates (e.g., 118 switched alarm date and containment date creating negative burn duration, typo in containment year creates a 5+ year burn duration).

Measuring wildfire severity

The CalFire Fire Resource and Protection Program (FRAP; http://frap.fire.ca.gov/) maintains the most 122 comprehensive Dataset of wildfire perimeters in the state of California, including attribute data for each fire such as its discovery date, its containment date, and the management objective. The management objective 124 represents the approach taken by the management unit overseeing the wildfire- either "suppression", with a goal of extinguishing the fire, or "wildland fire use", with a goal of allowing the fire to burn to benefit 126 forest resources so long as it didn't threaten lives or property (Meyer 2015). This dichotomy is somewhat 127 simplistic, as each wildfire can be managed for multiple objectives, but it is a generally useful framework for 128 understanding the primary management goal (Meyer 2015). This dataset contains all fires >4 hectares (and 129 a non-comprehensive set of fires <4 hectares), and thus has greater representation of fire events, particularly 130 events of smaller size, compared to other wildfire events datasets, though it lacks severity information. For 131 instance, the Monitoring Trends in Burn Severity (MTBS) database only contains wildfires in the western U.S. 132 that are larger than 400 hectares (Eidenshink et al. 2007) and the U.S. Forest Service Region 5 geospatial 133 database contains wildfires in the Sierra Nevada that are larger than 80 hectares (Steel et al. 2018), though both of these datasets contain information on wildfire severity. 135

To assess the severity of these smaller fires, Koontz et al. (2019a) used the expanded FRAP dataset of 136 over 1,000 wildfire perimeters to calculate wildfire severity and calibrate satellite-derived measures using ground-based overstory composite burn index (CBI) (Koontz et al. 2019b), which is an integrated measure of 138 the effect of a wildfire on the forest overstory one year after the burn (Key and Benson 2006). CBI correlates well with direct measures of fire impact to vegetation in Sierra YPMC, such as percent of overstory mortality 140 (Miller and Thode 2007). Thresholds of wildfire severity (unchanged, low, moderate, and high) calibrated to 141 the full dataset (Koontz et al. 2019b) were imposed on each fire and then contiguous pixels of each category 142 were vectorized into polygons to form patches of each severity category. We subsetted this FRAP-derived 143 dataset of fire severity to 622 that burned in majority yellow pine/mixed-conifer forest (547 suppression fires; 75 WFU fires). 145

Burning conditions per fire event

In addition to mapping wildfire severity across each fire in the FRAP perimeter database, Koontz et al. (2019b) also calculated fuel and regional climate variables within the burn perimeter. The prefire Normalized Difference Vegetation Index (NDVI; Rouse et al. (1973)]) was found to correlate strongly with local wildfire severity, as was the standard deviation of NDVI within the 90m x 90m window surrounding each pixel, which represents a measure of horizontal forest structure and fuel continuity (Koontz et al. 2019b). The

gridMET product (Abatzoglou 2013) was used to calculate the energy release component, a modeled estimate
of expected fire behavior in conifer forest, for the 3 days prior to each fire's discovery date as well as the
wind speed for the first three days of the fire. The gridMET product has a daily temporal resolution and a 4
kilometer spatial resolution, so our climate variables capture regional conditions over time periods of several
days, but not very local weather events that might occur over the span of hours. Each of these variables has
a strong impact on wildfire behavior at macroscales (Abatzoglou and Kolden 2013).

For this study, we assigned prefire burning conditions for each fire as the mean fuel (prefire NDVI, prefire

standard deviation of NDVI within 90m x 90m moving windows) and regional climate (energy release component for 3 days prior to the fire, wind speed for first 3 days of the fire) values within each fire perimeter.

For each fire, we calculated the total number of fires burning on that fire's alarm date, the proportion of area represented by each severity category, as well as the maximum patch size of each severity category. Finally, we calculated the stand replacing decay coefficient (SDC) (Collins et al. 2017; Stevens et al. 2017), a single metric that integrates high severity patch size and shape such that a lower SDC corresponds to a larger, more circular high severity patches with effectively more area within those patches far from the patch edges (i.e., beyond likely seed dispersal estimated for yellow pine/mixed-conifer species).

167 Designating "survivorship" of suppression fires

For each wildfire in the Koontz *et al.* (2019b) dataset with a suppression management objective, we determined whether it "survived" initial attack by whether its burn duration (discovery date subtracted from the containment date) was greater than 1 day. Following Abatzoglou *et al.* (2018), we assumed that fires under a suppression management objective that burned for more than one day would require different firefighting tactics than direct attack and thus represented a reconfiguration of firefighting personnel and resource allocation.

Quantifying the selection effect of fire suppression

We treated the prefire fuel (prefire NDVI, heterogeneity of NDVI) and climate conditions (prefire energy release component, early fire wind speed) as wildfire "phenotypes" having some distribution, and used logistic regression to measure the extent to which wildfire suppression "selected" for particular burning condition phenotypes using the survivorship of each fire from initial attack as a binary response fitness metric in an evolutionary ecology framework (Lande and Arnold 1983; Janzen and Stern 1998).

180 Implications of selection by suppression

We graphically compare fire effects (event size, stand replacing decay coefficient, proportion high severity) of suppression fires to those of wildfire use (WFU) fires to draw inferences about how the "lost contribution" of fires that succumbed to early containment may have influenced fire effects.

184 Software and data availability

Our selection model was fit using the brms (Bürkner 2017) package in R (R Core Team 2018). Our workflow greatly benefited from the tidyverse group of packages (Wickham 2017), and we manipulated spatial data using the following R packages: sf (Pebesma 2018), raster (Hijmans et al. 2019), velox (Hunziker 2017), stars (Pebesma 2019a), lwgeom (Pebesma 2019b), fasterize (Ross 2018), and APfun (Plowright 2018).

The original severity dataset created by (Koontz et al. 2019a) can be found on the Open Science Framework (https://osf.io/ke4qj/). The data and code for this study can also be found on the Open Science Framework.

191 Results

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Fire event size and burn duration context of Sierra YPMC wildfire

We found that 16219 fire events burned at least partially in Sierra Nevada yellow pine/mixed-conifer between 1992 and 2015 covering 2.19 million total hectares and 1.18 million hectares within this forest type. A total of 14873 fires burned in greater than 50% Sierra yellow pine/mixed-conifer in the same period. The vast majority of these fire events were very small. Comparing the distribution of fire sizes in this system between 1992 and 2015 to relevant reference sizes:

- 61.46% of fires were smaller than 0.09 hectares—the size of a single pixel from Landsat which is a USGS satellite product often used to measure fire effects by comparing imagery just before the fire to imagery one year after the fire (Miller and Thode 2007).
- 98.87% of fires were smaller than 400 hectares—the minimum fire size for inclusion in the MTBS dataset for the western U.S. (Eidenshink *et al.* 2007), meaning that MTBS would include approximately 1.13% of fires in this system.
- 97.93% of fires were smaller than 80 hectares—the minimum fire size for inclusion in the USFS Region 5 geospatial dataset (Steel *et al.* 2018), meaning that the USFS Region 5 data accounts for approximately 2.07% of fires in this system.
 - 94.88% of fires were smaller than 4 hectares—the minimum fire size for inclusion in the CalFire FRAP fire perimeter dataset which was used to derive the Koontz *et al.* (2019a) severity dataset. Thus, the

Koontz et al. (2019a) dataset includes approximately 5.12% of fires in this system during its time span of 1984 to 2017.

211 Context of suppression fires compared to wildfire use (WFU) fires

There is a clear multimodality in the distribution of suppression fire event size while wildfire use fire event sizes are lognormally distributed (Figure 1). Under a suppression management objective, most fires are successfully contained at relatively small sizes while WFU fires are larger on average. This effect is obscured when comparing median fire sizes between management objectives using just larger fires (Table 1).

Median fire size of all suppression fires is much lower than the median fire size of WFU fires, owing to
the highly successful early suppression efforts (Table 1). This size difference becomes less apparent as the
minimum fire size in the dataset is increased, as is the case for fire effects datasets that focus on tracking
larger fire events (e.g., U.S. Forest Service Region 5 GIS dataset with a minimum fire size of 80 ha or MTBS
with a minimum fire size of 400 ha in the western U.S (Table 1). The size difference is likely to be even more
dramatic than what can be seen in the FRAP-derived Koontz et al. (2019b) dataset with its comprehansive
coverage of fire greater than 4 ha (and some additional fires smaller than this threshold), as nearly 95% of
wildfires in the Sierra yellow pine/mixed-conifer region remain even smaller than 4 ha.

Table 1: Comparison of fire event sizes by management objective for fires burning in majority yellow pine/mixed-conifer in the Sierra Nevada between 1984 and 2017.

Management	Median fire size (ha)	Median fire size (ha)	Median fire size (ha)	Median fire size (ha)
objective	All FRAP fires	Fires > 4ha	${\rm Fires} > 80 {\rm ha}$	Fires > 400ha
suppression	16.77	29.13	582	1459
wfu	193.1	205.4	506.7	866.5

24 Selection by suppression

We found a sizable effect of phenotypic selection of initial attack suppression on the heterogeneity of NDVI and prefire energy release component. Wildfires that survived initial attack were more likely to be burning in more extreme, homogeneous fuels (lower heterogeneity of NDVI) and in more extreme, hotter/drier climate conditions (higher energy release component) compared to wildfires that succumbed to the initial attack. We detected no selection pressure on prefire vegetation density as measured by NDVI or wind speed during the first three days of the fire (Figure 2; Figure 3).

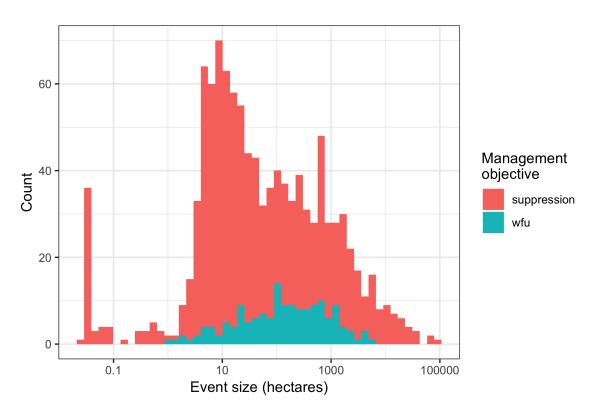


Figure 1: Distribution of log fire event size by management objective. While wildfire use fires exhibit a lognormal distribution in size, suppression fires exhibit clear multimodality with many fires extinguished when they are very small.

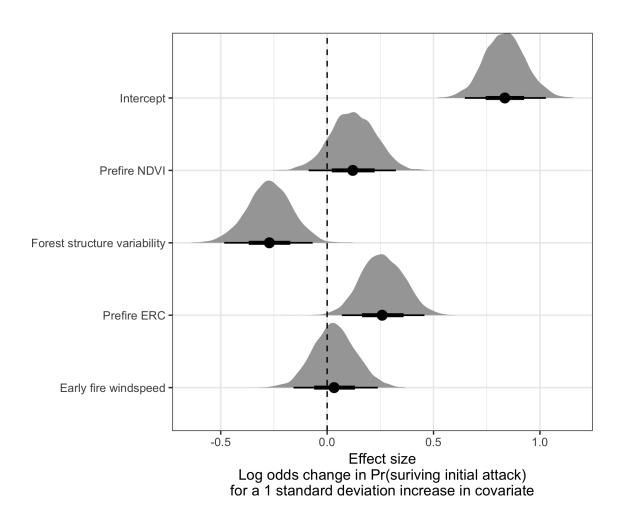


Figure 2: Halfeye plot showing posterior distributions of coefficient estimates for model predicting the probability of wildfire survivorship in the first 48 hours of initial attack. The effect sizes are proportional to the 'strength of selection' of initial attack on the burning conditions of wildfire. Credible intervals are shown below each probability density function with the point representing the mean, the dark line representing the 66% credible interval, and the light line representing the 95% credible interval. The dotted line shows an effect size of zero.

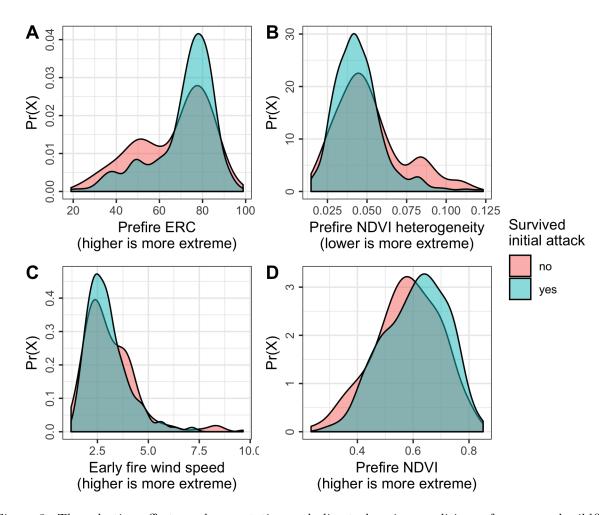


Figure 3: The selection effect on the vegetation and climate burning conditions of suppressed wildfires. A) Early suppression efforts selected for greater average energy release component (ERC), an estimate of fireline intensity correlated to hot, dry conditions. B) Fires surviving initial suppression efforts burn in more homogenous fuels, as measured by the standard deviation of NDVI in a 90m x 90m window. C) We detected no selection pressure on windspeed for the first 3 days of the fire. D) We detected no selection pressure on prefire NDVI, which is correlated with overstory canopy density and surface fuel loads.

Implications of selection by suppression for fire effects

We also found that fire effects of event size, proportion of high severity, and stand replacing decay coefficient 232 showed a similar trend (i.e., slope) in suppression versus WFU fires as a function of burn duration, except for those suppression fires with short burn durations. However, the average fire effects (i.e., intercept) of 234 suppression fires tended to depart from effects of WFU fires and resulted in greater proportion of high severity and a lower stand replacing decay coefficient. The notable exception to these trends are suppression fires with 236 short burn durations, which experienced a strong selection pressure on their burning conditions and whose 237 fire effects were dramatically different from fire effects of suppression fires that burned for longer. Suppression 238 fires that succumbed early to containment were much smaller, had lower proportions of high severity, and 239 had much higher stand replacing decay coefficients than suppression fires that escaped initial attack Figure 240 4). The fire effects of suppression fires that were successfully contained were within the range of variation of 241 WFU fires for similar burn durations.

Discussion

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Wildfire effects on forest vegetation are outcomes of a complex social-ecological system dynamic (Calkin et al. 2015). Direct causes of fire effects to vegetation arise from fire behavior and intensity (Keeley 2009), 245 which are coupled to fuel, weather, and topography (McKenzie and Hessl 2008; Cansler and McKenzie 2014; Harvey et al. 2016). Wildfire effects are indirectly related to firefighting resource availability, legacies of 247 management policy that change fuel distributions, and incentive structures that often prioritize mitigating short term loss of resources over long-term benefits of wildfire (Houtman et al. 2013; Calkin et al. 2014, 2015). Fire suppression allows some fires to "survive" initial containment efforts and contribute to fire effects 250 on the landscape, while other fires are "killed" by initial containment and their potential fire effects are 251 never realized. Because milder fuel and climate conditions facilitate fire containment efforts, initial attack 252 suppression imposes selection for fires that burn under more extreme conditions and shifts the concomitant distribution of fire effects to also be more extreme. We measured this selection pressure and found a sizable 254 influence of initial suppression efforts that decreased average fuel heterogeneity and increased average energy release component for fires that contribute to fire effects on the landscape (i.e., those that survive initial 256 containment). Two primary implications follow from our findings: 257

1) There is a lost contribution to overall fire effects in Sierra Nevada yellow pine/mixed-conifer forests associated with fires that are extinguished before they have a chance to burn under the milder conditions that facilitated their early containment (Figure 4).

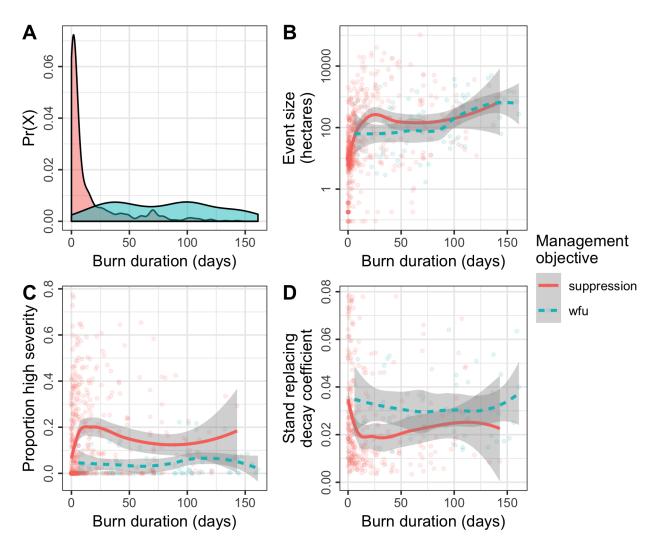


Figure 4: A) Distribution of burn duration by management objective. Most suppression fires are quickly extinguished. B) Effect of burn duration on fire event size shows that there's a similar trajectory between suppressed and wildfire use fires except early in the burning period when suppression fires remain small. C) The high severity portion of the fire tends to increase with shorter-duration suppression fires, but is relatively constant across burn durations for wildfire use fires. D) Conditional on a fire having a high-severity component, the stand replacing decay coefficient sharply declines as burn duration increases for suppression fires, indicating that the high-severity patches are larger and simpler. Larger, simpler high-severity patches will have reduced tree regeneration in their center because the distance to the nearest tree seed source exceeds typical dispersal distances for yellow pine/mixed-conifer species. The SDC tends to increase with the burn duration for wildfire use fires.

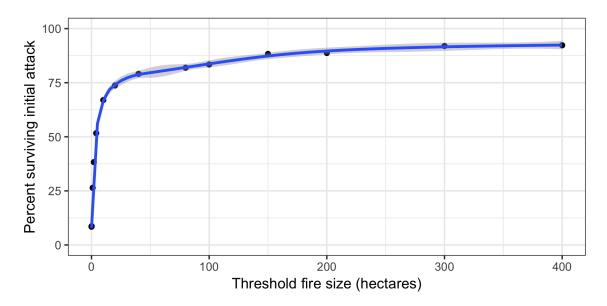


Figure 5: As the minimum fire event size of a dataset increases, a greater proportion of those fire events survived initial attack suppression efforts and burned on average under more extreme conditions. Thus databases with larger minimum fire sizes exhibit a stronger bias towards fires that burned in extreme conditions as a result of selection by suppression.

2) The fire effects that are measured only on large fires reflect a selection bias imposed by suppression efforts (Figure 5).

Context of suppression management on fire effects

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We compared fires managed for suppression to those managed for wildfire use/resource benefit to assess how the selection pressure on burning conditions may ultimately influence fire effects. We demonstrated a clear 265 multi-modality in the fire event size distribution of wildfires managed with a suppression objective (Figure 1). The median fire size of all fires in the dataset was much larger for WFU fires than for suppression fires, but 267 the median fire size of large fires (e.g., fires greater than 80 or 400 hectares) was much larger for suppression 268 fires (Table 1). Miller and Safford (2017) found a similar pattern in comparing the average size of modern 269 fires, most of which are managed with a suppression objective, to pre-Euroamerican settlement fires- modern, 270 suppression fires are smaller on average compared to the natural range of variation, but modern large fires are 271 much larger compared to the natural range of variation (Safford and Stevens 2017; Miller and Safford 2017). 272 Fire effects of escaped suppression fires were more departed from effects of WFU fires for the same burn 273 duration (Figure 4). However, fire effects of suppression fires that succumbed to early suppression efforts 274 were generally beneficial when considering WFU fires as a reference for beneficial effects. We cannot use 275 our data to make a rigorous counterfactual prediction of how suppression fires that didn't survive initial

277 attack may have ultimately influenced the overall distribution of fire effects if they had been allowed to burn.

However, the detected selection by suppression effect for extreme burning conditions suggests that, had these

fires burned for longer, they may have contributed to shifting the average fire effects of all suppression fires

closer to the observed fire effects for WFU fires (Figure 4).

No detected effect of windspeed

Surprisingly, we found no selection pressure on wind speed by early suppression efforts (Figure 2). Abatzoglou et al. (2018) found across the U.S. that greater wind speeds during the first two days after human-caused ignitions increased the likelihood that fires would grow large. Wind speed may not reflect a limiting factor in early suppression efforts in this region compared to fuel continuity and fuel moisture as measured by energy release component. Alternatively, the gridMET-derived wind speed in our study may be less representative of how wind affects fire behavior in this particular system. That is, wind may play a strong role, but the mountainous terrain may create more local wind conditions than can be captured by the gridMET product, despite its high spatial and temporal resolution (Abatzoglou 2013).

290 Selection beyond initial attack

We demonstrated a strong directional effect of selection imposed by initial suppression efforts on the initial 291 burning conditions of wildfires in the Sierra yellow pine/mixed-conifer system. Fire effects arise from the confluence of fuel, weather, and topographical conditions in space and time, and thus our event-level (i.e., 293 per fire) measurements of burning conditions may mask some of this selection effect. We suspect this effect may be even stronger if weather conditions were known at finer temporal scales and fuel conditions were 295 known at finer spatial scales. While we have shown a selection effect of suppression during an initial attack response, the effect may not be limited to soon after ignition (i.e., the scope of our investigation). Whether 297 a similar selection by suppression effect would materialize beyond initial attack and throughout the course 298 of a fire's burn duration depends on the extent to which success in suppression is contingent upon fuel 299 and weather conditions cooperating with firefighting efforts. That is, preventing additional fire effects to 300 the landscape by extinguishing fires as soon as the weather or fuel conditions become more conducive to 301 suppression amounts to imposing a similar selection effect that we detected during initial attack suppression. 302 Finer spatial and temporal resolution of burning conditions (e.g., daily climate variables paired with daily fire spread/severity maps) throughout the course of wildfires may help tune our understanding of how management 304 efforts can select for burning conditions. For example, aggressive suppression management may impose the strongest selection pressure on fuel and climate conditions from the day of or the day prior to the fire. Alternatively, WFU management may impose the strongest selection pressure on future fuel and climate conditions, which would reflect managers' efforts to guide the spread of fire into forecasted, spatiotemporally favorable intersections of fuel and climate.

310 Positive feedbacks

Selection for more extreme burning conditions favors future extreme burning conditions in positive feedbacks 311 of fuel and climate. Homogeneous local forest structure makes this system more likely to burn at high severity, 312 with complete or nearly-complete mortality of overstory vegetation (Koontz et al. 2019b). While some high severity fire is expected (Safford and Stevens 2017), elevated average levels and increased continuity of 314 high-severity fire and make it more likely that regenerating vegetation and future fuel structure will also be homogeneous (North et al. 2009; Coppoletta et al. 2016; Miller and Safford 2017; Stevens et al. 2017). 316 On longer time scales, the selection by suppression for hotter/drier average burning conditions may favor high-severity fire in the short run (Fried et al. 2004; Koontz et al. 2019b), and may compromise forest 318 recovery (Young et al. 2019), carbon stock stability (Earles et al. 2014), and carbon sequestration (Millar 319 and Stephenson 2015) in the long run. A reduction in carbon sequestration capacity contributes to climate 320 forcing, which is likely to perpetuate hot, dry conditions in California (Diffenbaugh et al. 2015; Mann and 321 Gleick 2015). Ongoing selection for extreme burning conditions by fire suppression therefore creates positive feedbacks of burning conditions that are unfavorable to yellow pine/mixed-conifer forest persistence. 323

A selection effect on WFU fires, too

A conceptually similar, but directionally opposite selection pressure may bias the distribution of fire effects 325 that arise from wildfire use fires. The decision to let a candidate WFU ignition continue to burn as a WFU fire when conditions are more mild selects for a distribution of fire effects that arises from these more mild 327 burning conditions. As with the selection by suppression, the distribution of fire effects of WFU fires will 328 be dictated by the fires whose fire effects were allowed to materialize (i.e., by deciding to let them burn in 329 the WFU case versus being unable to stop them from burning in the case of suppression fires that escape initial attack). The fire effects from WFU fires are largely seen as beneficial, but their average effect should 331 be considered as a better-than-average scenario for reintroducing fire to previously fire-suppressed forests 332 compared to the expectations of fire effects if all candidate WFU ignitions were allowed to burn. Similarly, 333 the fire effects of suppression fires that escape initial attack are largely seen as negative, but their average 334 effect should be treated as a worse-than-average scenario for reintroducing fire to the landscape compared to fire effects that would arise from letting all ignitions run their course.

The selection bias in both cases (suppression and WFU fires) reflects the social element in the coupled

social-ecological system of wildfire and wildfire management. Management decisions, even as a fire burns,
can hold great sway over ultimate fire effects to the forest system in their capacity to impose selection on
the distribution of burning conditions. However, the social element can also erect powerful barriers to more
widespread adoption of WFU fires such as consideration of smoke effects on human health (Schweizer and
Cisneros 2014) or political risks associated with allowing fires to burn (Doane et al. 2006). The strong
influence of on-the-ground management decision-making points to the immense value of reducing barriers to
implementing more wildfire use fires through increased training and reorganized incentive structures (Doane
et al. 2006).

References

- Abatzoglou JT. 2013. Development of gridded surface meteorological data for ecological applications and modelling. *International Journal of Climatology* **33**: 121–31.
- Abatzoglou JT and Kolden CA. 2013. Relationships between climate and macroscale area burned in the western United States. *International Journal of Wildland Fire* 22: 1003.
- Abatzoglou JT and Williams AP. 2016. Impact of anthropogenic climate change on wildfire across western
 US forests. Proceedings of the National Academy of Sciences 113: 11770–5.
- Abatzoglou JT, Balch JK, Bradley BA, and Kolden CA. 2018. Human-related ignitions concurrent with high winds promote large wildfires across the USA. *Int J Wildland Fire* **27**: 377–86.
- Bürkner P-C. 2017. Brms: An R Package for Bayesian Multilevel Models Using Stan. Journal of Statistical
 Software 80.
- ³⁵⁷ Calkin DE, Cohen JD, Finney MA, and Thompson MP. 2014. How risk management can prevent future wildfire disasters in the wildland-urban interface. *PNAS* **111**: 746–51.
- Calkin DE, Gebert KM, Jones JG, and Neilson RP. 2005. Forest Service Large Fire Area Burned and
 Suppression Expenditure Trends, 19702002. j for 103: 179-83.
- Calkin DE, Thompson MP, and Finney MA. 2015. Negative consequences of positive feedbacks in US wildfire management. Forest Ecosystems 2:9 doi: 101186/s40663-015-0033-8 2.
- Cansler CA and McKenzie D. 2014. Climate, fire size, and biophysical setting control fire severity and spatial pattern in the northern Cascade Range, USA. *Ecological Applications* **24**: 1037–56.
- ³⁶⁵ Coen JL, Stavros EN, and Fites-Kaufman JA. 2018. Deconstructing the King megafire. *Ecological Applications*

- 366 **28**: 1565–80.
- 367 Collins BM, Lydersen JM, and Fry DL et al. 2016. Variability in vegetation and surface fuels across
- mixed-conifer-dominated landscapes with over 40 years of natural fire. Forest Ecology and Management 381:
- ₃₆₉ 74–83.
- ³⁷⁰ Collins BM, Stevens JT, and Miller JD et al. 2017. Alternative characterization of forest fire regimes:
- Incorporating spatial patterns. Landscape Ecology 32: 1543–52.
- ³⁷² Coppoletta M, Merriam KE, and Collins BM. 2016. Post-fire vegetation and fuel development influences fire
- severity patterns in reburns. Ecological Applications 26: 686–99.
- Dale L. 2006. Wildfire Policy and Fire Use on Public Lands in the United States. Society & Natural Resources
- 375 **19**: 275–84.
- ₃₇₆ Davis JB. 1979. A new fire management policy on Forest Service lands. Fire Technol 15: 43–50.
- 377 Diffenbaugh NS, Swain DL, and Touma D. 2015. Anthropogenic warming has increased drought risk in
- ³⁷⁸ California. *PNAS* **112**: 3931–6.
- Doane D, O'Laughlin J, Morgan P, and Miller C. 2006. Barriers to wildland fire use. 12: 3.
- Earles JM, North MP, and Hurteau MD. 2014. Wildfire and drought dynamics destabilize carbon stores of
- fire-suppressed forests. Ecological Applications 24(4): 732-740 24: 732-40.
- Eidenshink J, Schwind B, and Brewer K et al. 2007. A Project for Monitoring Trends in Burn Severity. Fire
- звз *Ecology* **3**: 3–21.
- Fried JS, Torn MS, and Mills E. 2004. The Impact of Climate Change on Wildfire Severity: A Regional
- Forecast for Northern California. Climatic Change 64: 169–91.
- Fulé PZ, Covington WW, and Moore MM. 1997. Determining Reference Conditions for Ecosystem Management
- of Southwestern Ponderosa Pine Forests. Ecological Applications 7: 895–908.
- Harvey BJ, Donato DC, and Turner MG. 2016. Drivers and trends in landscape patterns of stand-replacing
- fire in forests of the US Northern Rocky Mountains (1984-2010). Landscape Ecology 31: 2367-83.
- Hijmans RJ, Etten J van, and Sumner M et al. 2019. Raster: Geographic Data Analysis and Modeling.
- Houtman RM, Montgomery CA, and Gagnon AR et al. 2013. Allowing a wildfire to burn: Estimating the

- effect on future fire suppression costs. Int J Wildland Fire 22: 871–82.
- ³⁹³ Hunziker P. 2017. Velox: Fast Raster Manipulation and Extraction.
- Janzen FJ and Stern HS. 1998. Logistic Regression for Empirical Studies of Multivariate Selection. Evolution
- 395 **52**: 1564–71.
- ³⁹⁶ JepsonFloraProject (Ed). 2016. Jepson eFlora.
- Johnson JF, Bengston DN, and Fan DP. 2009. US Policy Response to the Wildfire Fuels Management
- ³⁹⁸ Problem: An Analysis of the News Media Debate about the Healthy Forests Initiative and the Healthy
- Forests Restoration Act. Journal of Environmental Policy & Planning 11: 129–42.
- 400 Keane RE, Ryan KC, and Veblen TT et al. 2002. Cascading Effects of Fire Exclusion in Rocky Mountain
- Ecosystems: A Literature Review. USDA Forest Service, Rocky Mountain Research Station.
- 402 Keeley JE. 2009. Fire intensity, fire severity and burn severity: A brief review and suggested usage.
- 403 International Journal of Wildland Fire 18: 116.
- Key CH and Benson NC. 2006. Landscape Assessment (LA).: 55.
- 405 Koontz MJ, Fick SE, and Werner CM et al. 2019a. Wildfire severity, vegetation characteristics, and regional
- climate for fires covering more than 4 hectares burning in yellow pine/mixed-conifer forests of the Sierra
- Nevada, California, USA from 1984 to 2017. Open Science Framework.
- 408 Koontz MJ, North MP, and Werner CM et al. 2019b. Local variability of vegetation structure increases forest
- resilience to wildfire. *EcoEvoRxiv*.
- 410 Lande R and Arnold SJ. 1983. THE MEASUREMENT OF SELECTION ON CORRELATED CHARAC-
- ⁴¹¹ TERS. Evolution **37**: 1210–26.
- Lydersen JM, North MP, Knapp EE, and Collins BM. 2013. Quantifying spatial patterns of tree groups and
- 413 gaps in mixed-conifer forests: Reference conditions and long-term changes following fire suppression and
- logging. Forest Ecology and Management 304: 370-82.
- Mallek C, Safford H, Viers J, and Miller J. 2013. Modern departures in fire severity and area vary by forest
- 416 type, Sierra Nevada and southern Cascades, California, USA. Ecosphere 4: art153.
- 417 Mann ME and Gleick PH. 2015. Climate change and California drought in the 21st century. PNAS 112:
- 418 3858-9.
- ⁴¹⁹ McKenzie D and Hessl AE. 2008. A neutral model of low-severity fire regimes. In: Narog, Marcia G, tech

- 420 coord 2008 Proceedings of the 2002 Fire Conference: Managing fire and fuels in the remaining wildlands and
- open spaces of the Southwestern United States Gen Tech Rep PSW-GTR-189 Albany, CA: US Department of
- 422 Agriculture, Forest Service, Pacific Southwest Research Station p 139-150 189: 139-50.
- ⁴²³ Meyer MD. 2015. Forest Fire Severity Patterns of Resource Objective Wildfires in the Southern Sierra Nevada.
- Journal of Forestry 113: 49–56.
- ⁴²⁵ Millar CI and Stephenson NL. 2015. Temperate forest health in an era of emerging megadisturbance. Science
- 426 **349**: 823–6.
- Miller JD and Safford HD. 2017. Corroborating Evidence of a Pre-Euro-American Low- to Moderate-Severity
- Fire Regime in Yellow PineMixed Conifer Forests of the Sierra Nevada, California, USA. Fire Ecology 13:
- 429 58-90.
- 430 Miller JD and Thode AE. 2007. Quantifying burn severity in a heterogeneous landscape with a relative
- 431 version of the delta Normalized Burn Ratio (dNBR). Remote Sensing of Environment 109: 66–80.
- 432 North MP, Stephens SL, and Collins BM et al. 2015. Reform forest fire management. Science 349: 1280–1.
- North M, Stine P, and O'Hara K et al. 2009. An ecosystem management strategy for Sierran mixed-conifer
- 434 forests. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- ⁴³⁵ Pebesma E. 2018. Simple Features for R: Standardized Support for Spatial Vector Data. The R Journal.
- ⁴³⁶ Pebesma E. 2019a. Stars: Spatiotemporal Arrays, Raster and Vector Data Cubes.
- Pebesma E. 2019b. Lwgeom: Bindings to Selected 'liblwgeom' Functions for Simple Features.
- Plowright A. 2018. APfun: Geo-Processing Helper Functions.
- 439 R Core Team. 2018. R: A Language and Environment for Statistical Computing. Vienna, Austria: R
- 440 Foundation for Statistical Computing.
- Ross N. 2018. Fasterize: Fast Polygon to Raster Conversion.
- 442 Rouse W, Haas RH, Deering W, and Schell JA. 1973. MONITORING THE VERNAL ADVANCEMENT
- 443 AND RETROGRADATION (GREEN WAVE EFFECT) OF NATURAL VEGETATION. Greenbelt, MD,
- 444 USA: Goddard Space Flight Center.
- 445 Safford HD and Stevens JT. 2017. Natural Range of Variation for Yellow Pine and Mixed-Conifer Forests in
- the Sierra Nevada, Southern Cascades, and Modoc and Inyo National Forests, California, USA.
- 447 Schweizer D and Cisneros R. 2014. Wildland fire management and air quality in the southern Sierra Nevada:

- 448 Using the Lion Fire as a case study with a multi-year perspective on PM2.5 impacts and fire policy. Journal
- of Environmental Management 144: 265–78.
- Short KC. 2017. Spatial wildfire occurrence data for the United States, 1992-2015 [FPA_FOD_20170508]
- 451 (4th Edition).
- Steel ZL, Koontz MJ, and Safford HD. 2018. The changing landscape of wildfire: Burn pattern trends and
- implications for California's yellow pine and mixed conifer forests. Landscape Ecology 33: 1159–76.
- 454 Steel ZL, Safford HD, and Viers JH. 2015. The fire frequency-severity relationship and the legacy of fire
- suppression in California forests. Ecosphere 6: art8.
- 456 Stephens SL, Collins BM, and Fettig CJ et al. 2018. Drought, Tree Mortality, and Wildfire in Forests
- Adapted to Frequent Fire. BioScience 68: 77–88.
- 458 Stevens JT, Collins BM, and Miller JD et al. 2017. Changing spatial patterns of stand-replacing fire in
- 459 California conifer forests. Forest Ecology and Management 406: 28–36.
- 460 Stevens-Rumann CS, Kemp KB, and Higuera PE et al. 2018. Evidence for declining forest resilience to
- wildfires under climate change (F Lloret, Ed). Ecology Letters 21: 243–52.
- 462 Veblen TT, Kitzberger T, and Donnegan J. 2000. Climatic and Human Influences on Fire Regimes in
- 463 Ponderosa Pine Forests in the Colorado Front Range. Ecological Applications 10: 1178–95.
- ⁴⁶⁴ Walker RB, Coop JD, Parks SA, and Trader L. 2018. Fire regimes approaching historic norms reduce
- wildfire-facilitated conversion from forest to non-forest. Ecosphere 9: e02182.
- Welch KR, Safford HD, and Young TP. 2016. Predicting conifer establishment post wildfire in mixed conifer
- forests of the North American Mediterranean-climate zone. *Ecosphere* 7: e01609.
- Westerling AL. 2006. Warming and Earlier Spring Increase Western U.S. Forest Wildfire Activity. Science
- 469 **313**: 940–3.
- 470 Westerling AL. 2016. Increasing western US forest wildfire activity: Sensitivity to changes in the timing of
- 471 spring. Philosophical Transactions of the Royal Society B: Biological Sciences 371: 20150178.
- Wickham H. 2017. Tidyverse: Easily Install and Load the 'Tidyverse'.
- 473 Young DJN, Werner CM, and Welch KR et al. 2019. Post-fire forest regeneration shows limited climate
- tracking and potential for drought-induced type conversion. Ecology 100: e02571.