Initial attack wildfire suppression selects for extreme burning con-

₂ ditions

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$_{\scriptscriptstyle 10}$ Executive summary (to become abstract)

- Most wildfires in the western U.S. are managed for suppression. This is usually effective at keeping fires
- small. It is often assumed that the fires that grow large do so because they burn under extreme fuel and
- $_{13}$ climate conditions. We quantify this "selection" effect using an evolutionary ecology framework and discuss
- implications for management.

15 Abstract

- Wildfires typically undergo active suppression efforts in forests of the western U.S.A. A vast majority of these
- wildfires are extinguished during "initial attack" suppression within the first few days after their discovery,
- which constrains the area of vegetation that they effect. In the long run, the reduced footprint of fire in the
- 19 drier, lower elevation forests of the western U.S. has led to a dramatic buildup of fuel such that
- 20 In the short run, this reduces the footprint of fire in these forests. In the long run, buildup of fuel in the
- drier, lower elevation forests of the region
- 22 A century of such a policy has enabled dramatic densification of low- to mid-elevation dry and mesic forests
- 23 in the west, with yellow pine/mixed-conifer forests of California's Sierra Nevada range being a prime example.
- The dominant management approach for wildfires in yellow pine/mixed-conifer forests of the Sierra Nevada
- 25 mountain range, as it is for many western U.S. forests, is full suppression. "Initial attack" efforts within the
- 26 first few days of discovering a new ignition prioritized extinguishing
- 27 Thus "large" modern fires in this system are often considered "bad" fires with respect to forest health.

- However, this correlation may be confounded by the extreme conditions under which "large" fires grow large.
- 29 Ideally, there'd be a more extensive fire footprint

30 Introduction

Fire suppression is an oft-cited root cause of the modern trend of larger, more severe wildfires in the yellow pine/mixed-conifer forests of California's Sierra Nevada mountain range (Miller and Thode 2007; Calkin et al. 2015; Safford and Stevens 2017). While most of this system would experience frequent, low- to moderate-severity wildfire every 8 to 15 years in the several centuries prior to Euroamerican settlement, suppression management has largely eliminated fire effects from much of western dry forested land in the past 100 years (Steel et al. 2015; Safford and Stevens 2017; Miller and Safford 2017). A lack of frequent fire has led to densification of Sierra yellow pine/mixed-conifer, which increases fuel loading and homogenizes forest structure (Collins et al. 2016; Stephens et al. 2018). Fortuitous alignment of these extreme fuel conditions with earlier snowmelt, longer fire seasons, and hotter droughts (aka "climate change droughts") (Westerling 2006, 2016; Abatzoglou and Kolden 2013; Abatzoglou and Williams 2016) increases the probability that fires will generate self-propagating behavior (Coen et al. 2018) and kill all (or nearly all) overstory vegetation (Koontz et al. 2019b) in large, contiguous patches of mortality (Stevens et al. 2017). The Sierra yellow pine/mixed-conifer community is ill-adapted to regenerate in the centers of these large patches, which are far from seed sources (Welch et al. 2016), and thus the modern trend of atypically large, contiguous, stand-replacing fires in this system compromise forest health and increase the potential for long-term shifts in vegetation type to shrub- or grasslands (Millar and Stephenson 2015; Stevens et al. 2017; Steel et al. 2018). Despite the long-term effects of fire suppression on wildfire trends, fire suppression generally is very effective at its immediate goal of extinguishing fires. Between 1970 and 2002, over 97% of fires burning on U.S. Forest Service land were contained during "initial attack" suppression before they reached 120 hectares in size (Calkin et al. 2005, 2015). In the western U.S., including the Sierra yellow pine/mixed-conifer system, the vast majority of fires are managed for suppression, though a small number of natural ignitions are allowed to burn under moderate conditions in "resource benefit" or "wildland fire use" fires in recognition of the important role that fire plays in many healthy forest systems (Calkin et al. 2015).

The fires that remained small are often assumed to have burned under mild or moderate vegetation and

climate conditions, facilitating initial suppression efforts. The fires that escape initial attack and grow large account for 97.5% of the total burned area of fires managed by the U.S. Forest Service (Calkin *et al.* 2005),

- and are often assumed to have grown to these sizes because they were burning under more extreme conditions.
- 59 "Essentially, through our management efforts we have changed the distribution of fire behavior to only the
- most extreme." (Calkin et al. 2015).
- 61 Thus, the general short-term success of fire suppression policy paired with its long-term cumulative effect has
- 62 led to a management paradox with respect to maintaining forest health: we shouldn't put out the fires that
- 63 we can, but we can't put out the fires that we should.
- 64 Bimodal distribution of fire sizes under suppression policy: either the fires were quickly put out and remained
- of very small, or they escaped suppression efforts and grew exceptionally large due to regional climate conspiring
- 66 with accumulated fuel conditions. Indeed, Miller and Safford (2017) found evidence for this pattern in that
- the average size of all fires is much smaller under a modern fire suppression management regime compared to
- ₆₈ pre-Euroamerican settlement fires (as one might intuitively expect given the goal of suppression is to stop fires
- 69 to protect life and property), but the average size of larger fires (>4 hectares) is, perhaps counter-intuitively,
- much greater under modern suppression management (Safford and Stevens 2017; Miller and Safford 2017).
- 71 Many studies have suggested that adding more fire to the landscape is a way to return forests to pre-
- ⁷² Euroamerican settlement resilient conditions (Mallek et al. 2013; North et al. 2015; Collins et al. 2017;
- 73 Stevens et al. 2017) and some efforts are underway though barriers remain (Doane et al. 2006; Calkin et al.
- 74 2015).
- We use a new dataset of fire severity (Koontz et al. 2019a) to measure this "selection by suppression" effect
- 76 and ask:
- 1) How do burning conditions (regional climate, vegetation density, vegetation continuity) vary between
- suppression and wildfire use fires?
- 79 2) What are the consequences of these varying conditions for wildfire effects (burn duration, fire event
- size, severity configurations)?
- 3) To what extent might prematurely extinguished suppression wildfires exhibit desireable fire effects, were
- they to be allowed to burn for longer? That is, are small suppression fires behaving like small wildfire
- use fires (and thus they should perhaps be left to burn and cover more area)?

Methods

and 98.5% remain smaller than 80 hectares (Short 2017).

86 Study system

The Sierra Nevada yellow pine/mixed-conifer (hereafter Sierra YPMC) is a disturbance-prone forest system in the Sierra Nevada mountain range of California, U.S.A. It spans the full 628 kilometer latitudinal length of the Sierra Nevada, and over 2,750 meters of elevation (229 meters to 3053 meters; 95% interval = (679, 2516) meters; IQR = (1273, 1945) meters), primarily on the western slope of the mountain range. The forest is dominated by ponderosa pine (*Pinus ponderosa*), sugar pine (*Pinus lambertiana*), white fir (*Abies* concolor), and incense cedar (Calocedrus decurrens) in varying mixes. Prior to Euroamerican settlement, the ecosystem experienced frequent, low- to moderate-severity wildfire every 8 to 15 years on average (Steel et al. 2015; Safford and Stevens 2017), which consumed surface fuels but generally had minimal effects on large, established trees. This fire regime generated heterogeneous horizontal forest structure, with groups of relatively even-aged trees having interlocking crowns, individual trees with distinct crowns, and variably-sized gaps between these tree clump and individual tree structural features (Lydersen et al. 2013). A century of fire suppression has led to infill of these gaps, homogenizing the horizontal forest structure (Collins et al. 2016) and compromising forest resilience (North 2012) in an era of climate change-induced hotter droughts (Millar and Stephenson 2015). 100 We compiled the Sierra YPMC type using the U.S. Forest Service Fire Return Interval Departure (FRID) 101 dataset and included "dry mixed-conifier", "moist mixed-conifier", and "yellow pine" vegetation types following Steel et al. (2018) and Koontz et al. (2019b). These classifications represent "potential vegetation" given the 103 climate of the area, such that there is no influence of recent disturbance events (Harvey et al. 2016; Steel et

106 Measuring wildfire burning conditions

al. 2018; Koontz et al. 2019b).

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The CalFire Fire Resource and Protection Program (FRAP; http://frap.fire.ca.gov/) maintains the most 107 comprehensive datset of wildfire perimeters in the state of California, including attribute data for each fire such as its discovery date, its containment date, and the management objective. The management objective 109 represents the approach taken by the management unit overseeing the wildfire—either "suppression", with a goal of extinguishing the fire, or "wildland fire use", with a goal of allowing the fire to burn so long as it didn't 111 threaten lives or property. This dataset contains all fires >4 hectares, and thus has greater representation of fire events compared to other wildfire events datasets, though it lacks severity information. For instance, the 113 Monitoring Trends in Burn Severity (MTBS) database only contains wildires in the western U.S. that are 114 larger than 400 hectares (Eidenshink et al. 2007) and the U.S. Forest Service Region 5 geospatial database 115 contains wildfires in the Sierra Nevada that are larger than 80 hectares (Steel et al. 2018), though both of 116

these datasets contain information on wildfire severity.

Koontz et al. (2019a) used the expanded FRAP dataset of over 1,000 wildfire perimeters to calculate wildfire severity and calibrate satellite-derived measures using ground-based overstory composite burn index (CBI) (Koontz et al. 2019b), which is an integrated measure of the effect of a wildfire on the forest overstory one year after the burn (Key and Benson 2006). CBI correlates well with direct measures of fire impact to vegetation in Sierra YPMC, such as percent of overstory mortality (Miller and Thode 2007).

In addition to mapping wildfire severity across each fire in the FRAP perimeter database, Koontz et al. 123 (2019b) also calculated fuel and regional climate variables within the burn perimeter. The prefire Normalized 124 Difference Vegetation Index (NDVI; Rouse et al. (1973)]) was found to correlate strongly with wildfire 125 severity, as was the standard deviation of NDVI within the 90m x 90m window surrounding each pixel, which represents a measure of horizontal forest structure and fuel continuity (Koontz et al. 2019b). The 127 gridMET product (Abatzoglou 2013) was used to calculate the median 100-hour fuel moisture, energy release component, and vapor pressure deficit for the 3 days prior to each fire's discovery date, which captures 129 regional climate conditions over time periods of several days, but not very local weather events that might occur over the span of hours. Each of these variables has a strong impact on wildfire behavior at macroscales 131 (Abatzoglou and Kolden 2013).

133 Aggregating burning conditions to the fire event level

For this study, we assigned prefire burning conditions for each fire as the mean fuel (prefire NDVI, prefire standard deviation of NDVI within 90m x 90m moving windows) and regional climate (100-hour fuel moisture, energy release component, vapor pressure deficit) values within each fire perimeter.

Designating "survivorship" of suppression fires

For each wildfire with a suppression management objective, we determined whether it "survived" initial attack by whether it's burn duration (discovery date subtracted from the containment date) was greater than 1 day. Following Abatzoglou *et al.* (2018), we assumed that fires under a suppression management objective that burned for more than one day would require different firefighting tactics than direct attack and thus represented a reconfiguration of firefighting personnel and resource allocation.

Quantifying the selection effect of fire suppression

We treated the prefire fuel (prefire NDVI, heterogeneity of NDVI) and climate conditions (prefire energy release component) as wildfire "phenotypes" with some distribution, and measured the extent to which

wildfire suppression "selected" for particular phenotypes using the survivorship of each fire to initial attack as our fitness metric. Given the binomial response of the data, we used a logistic regression framework in an evolutionary ecology framework (Lande and Arnold 1983; Janzen and Stern 1998).

149 Implications of selection for wildfire responses

- We fit models using the same prefire fuel and climate conditions to wildfire effects responses including fire
 event size, proportion high severity, and the stand-replacing decay coefficient (Collins *et al.* 2017; Stevens *et al.* 2017).
- 153 Comparing to the "wildland fire use" fires in some way?

Future work?

- Took the Short (2017) dataset and subsetted to (approximately) Sierra YPMC. (took centroid and fire area, assumed circular fire, assessed how much of it fell across the Sierra YPMC raster image; super small fires that intersected no centroids of the raster image were converted back to points and we "extracted" the Sierra YPMC raster value (either 0 or 1) based on the point location instead). Uploaded to Earth Engine and got the regional climate values for the centroids. (Also got the fuel values for the centroids, but these aren't that meaningful because they should vary quite a bit across the fires. Could still calculate across the assumed circular fires?)
- Some summary stats of the Sierra Nevada yellow pine/mixed-conifer subset of the Short (2017) fire occurrence dataset. (65% of fires less than a Landsat pixel; 5.2% of fires in Koontz et al. (2019a) dataset; 1.5% in USFS Region 5 dataset; 0.5% in MTBS dataset)
- Subset Short (2017) into suppression/managed wildfire based on where it burns, using method of Steel *et al.* (2018).
- Use the expanded set of fires, and the relationships between fire size/SDC and covariates to extrapolate the potential added "good fire area" that might have burned if the Short (2017) suppression fires weren't extinguished.

170 Results

Selection by suppression

We found a sizeable effect of phenotypic selection of initial attack suppression on the heterogeneity of NDVI and the prefire energy release component. Wildfires that survived initial attack were more likely to be burning

in more extreme, homogenous fuels (lower heterogeneity of NDVI) and more extreme, hotter/drier climate conditions (higher energy release component) compared to wildfires that succumbed to the initial attack.

Table 1:

14610 1.	
	$Dependent\ variable:$
	survived_ia
scale(prefire_ndvi)	-0.045
	(0.084)
scale(nbhd_sd_ndvi_1)	-0.269***
	(0.079)
scale(prefire_erc)	0.190**
	(0.077)
Constant	0.920***
	(0.079)
Observations	812
Log Likelihood	-478.293
Akaike Inf. Crit.	964.585
Note:	*p<0.1; **p<0.05; ***p<0.01

176 Implications of selection by suppression

177 Comparison to wildland fire use fires by treating these fires as the "wild type" in the evolutionary ecology 178 framing.

179 Discussion

- We have demonstrated a clear multi-modality in the fire event size distribution of wildfires managed with a suppression objective, and further measured the effect of this "selection by suppression" on the severity characteristics of these events.
- Suppression leads to especially large fires because the ones that grow large are only able to do so because they burn under extreme conditions. We found strong evidence that inital attack "kills" fires that would otherwise burn under more mild and moderate fuel and climate conditions, instead favoring fires that burn under extreme conditions, with concomitant negative consequences for Sierra YPMC forest health.

187 Two implications:

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1) There is "lost work" (in terms of Sierra YPMC area affected positively by wildfire) associated with fires that are extinguished before they have a chance to do that work.

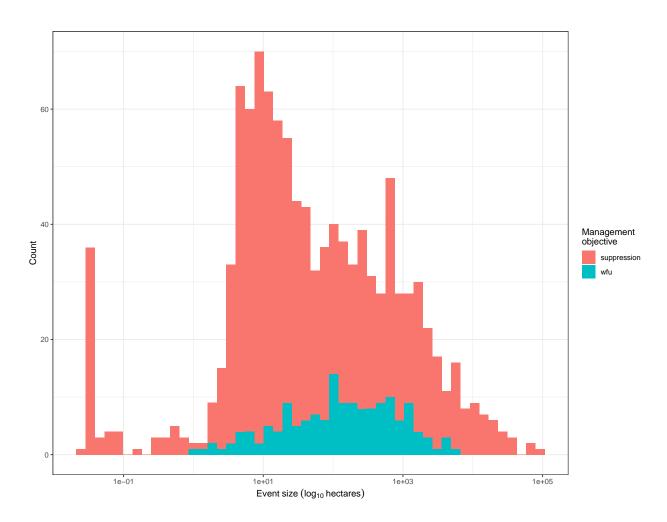


Figure 1: Distribution of log fire event size by management objective. While wildfire use fires exhibit a lognormal distribution in size, suppression fires exhibit clear multimodality.

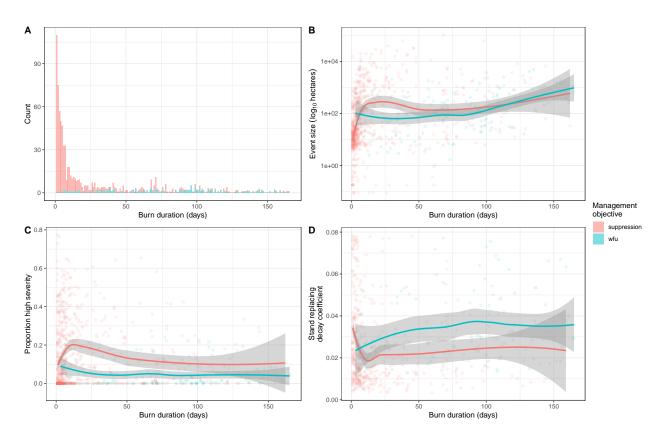


Figure 2: A) Distribution of burn duration by management objective. Most suppression fires are quickly extinguished. B) Effect of burn duration on fire event size shows that there's a similar trajectory between suppressed and wildfire use fires except early in the burning period when suppression fires remain small. C) The high severity portion of the fire tends to increase with shorter-duration suppression fires, but is relatively constant across burn durations for wildfire use fires. D) For fires with a high severity component, the stand replacing decay coefficient sharply declines for short-lived fires indicating that the high severity patches are larger and simpler which may reduce tree propagule pressure in the center of these patches and compromise dry forest regeneration. The SDC tends to increase with the burn duration for wildfire use fires.

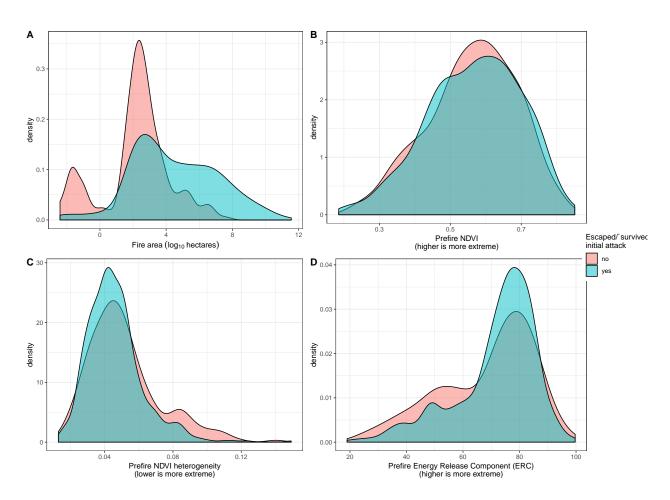


Figure 3: The selection effect on the vegetation and climate burning conditions of suppressed wildfires.

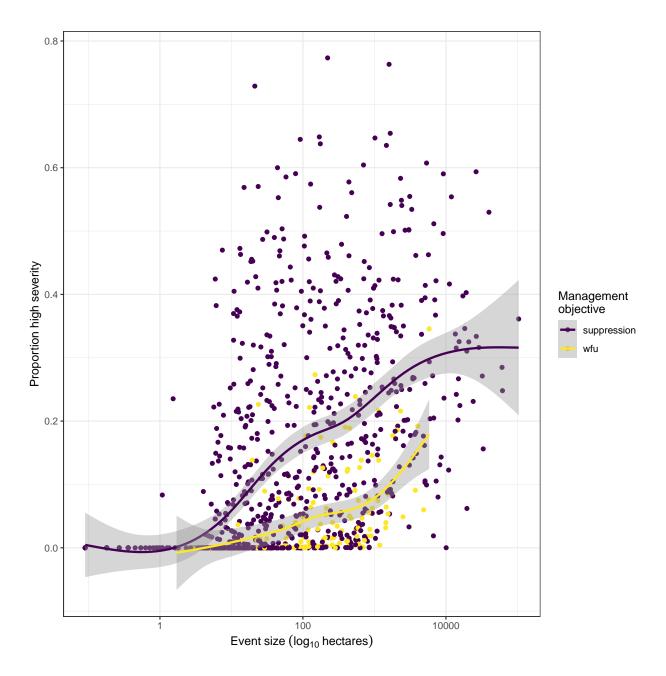


Figure 4: The effect of the fire event size on the proportion of high severity fire by management objective. The effect appears to manifest for suppression fires starting at fires around 40 hectares in size.

190 2) Though destructive megaevents are on the rise, their prevalence is highlighted because they are all that
191 burns large.

Essentially all of the suppression fires in Figure 2c that are within or above the confidence interval for the wildfire use fires had beneficial severity characteristics but were suppressed. Each of these fires represents potentially "lost area" of beneficial fire activity.

Extending the metaphor: a case for an evolutionary response to selection

Evolutionary ecology makes a clear and important distinction between "phenotypic selection"—the differential survivorship or reproduction of individuals with different phenotypes, and "evolutionary response to selection" in which differences in the phenotype distributions are passed on to offspring owing to heritability of the traits under selection (or correlation with heritable traits under selection) (Lande and Arnold 1983). Of course, our "selection" metaphor can only take us so far in exploring the implications of fire suppression on typical wildfire phenotypes in Sierra YPMC as wildfires are not living things, but perhaps we can make one more leap by considering that the fuel conditions are, in a sense, "heritable".

Forest structure and fuel characteristics link one disturbance event to the next via feedbacks: one wildfire 203 changes the fuel characteristics of the forest as it burns, and those altered fuel characteristics can feedback to 204 influence the behavior and effects of the next wildfire. In Sierra YPMC, severe wildfires with stand-replacing effects, especially large, contiguous stand-replacing effects, tend to homogenize forest spatial structure. In 206 turn, homogenous forest structure increases the probability that a forest will burn at high severity (Koontz et al. 2019b). If initial attack suppression selects for wildfires to burn under more homogenous fuel conditions, 208 and those fires are more likely to burn at high severity, then the vegetation within the footprint of those fires 209 will be more likely to regenerate homogeneously, thus "inheriting" the phenotype that was selected for. On a 210 longer time scale, this is perhaps true for the regional climate trait that suppression also selects for: selection 211 for fires burning under hotter/drier conditions will likely lead to more severe wildfires, with greater release of 212 carbon to the atmosphere, greater climate forcing, and increasingly common hot/dry conditions. In this case, 213 the selection for fires burning under hotter/drier conditions also leads to "inheritance" of the hotter/drier phenotype for future fires. 215

Tying into other studies (need to find a home for these thoughts)

217 Benefit of having smaller fires to analyze?

Steel et al. (2018) found only a small difference in the measure of regional climate during suppression versus wildfire use fires (modeled fuel moisture), which led to the conclusion that an "extreme fuel" effect underlied

- the differences in the size/configuration of high and unburned severity patches.
- 221 But that analysis only included fires greater than 80 hectares in size, which is still pretty big. Analyzing big
- 222 fires made sense for that paper, because they affect the most area, but if we want to suggest "let fires burn at
- 223 more moderate weather conditions" as a mitigation strategy for a century of making fuel conditions more
- extreme, then I think this paper would serve the purpose of measuring the degree to which good work can be
- done by a fire (even in extreme fuel conditions) in milder weather conditions.

226 Interdependence on various burning conditions and fire effects

- 227 Interacting effects (and directional dependence) of vegetation, regional climate, fire size, and severity patterns:
- (Cansler and McKenzie 2014; Harvey et al. 2016).
- ²²⁹ "A variety of environmental and social factors influence wildfire growth and whether a fire overcomes initial
- 230 attack efforts and becomes a large wildfire. However, little is known about how these factors differ between
- lightning-caused and human-caused wildfires." (Abatzoglou et al. 2018)
- ²³² "As shown by Balch et al. (2017), human-caused fires occupy an environmental niche characterised by lower
- 233 lightning- frequency and higher fuel moisture than lightning-caused fires. The present work complements
- those findings by demonstrating the likelihood of human ignitions evolving into large fires when facilitated by
- 235 strong winds." (Abatzoglou et al. 2018)

236 References

- Abatzoglou JT. 2013. Development of gridded surface meteorological data for ecological applications and
- modelling. International Journal of Climatology 33: 121–31.
- ²³⁹ Abatzoglou JT and Kolden CA. 2013. Relationships between climate and macroscale area burned in the
- western United States. International Journal of Wildland Fire 22: 1003.
- Abatzoglou JT and Williams AP. 2016. Impact of anthropogenic climate change on wildfire across western
- US forests. Proceedings of the National Academy of Sciences 113: 11770-5.
- Abatzoglou JT, Balch JK, Bradley BA, and Kolden CA. 2018. Human-related ignitions concurrent with high
- winds promote large wildfires across the USA. Int J Wildland Fire 27: 377-86.
- ²⁴⁵ Calkin DE, Gebert KM, Jones JG, and Neilson RP. 2005. Forest Service Large Fire Area Burned and
- ²⁴⁶ Suppression Expenditure Trends, 19702002. *j for* **103**: 179–83.
- ²⁴⁷ Calkin DE, Thompson MP, and Finney MA. 2015. Negative consequences of positive feedbacks in US wildfire

- management. Forest Ecosystems 2:9 doi: 101186/s40663-015-0033-8 2.
- ²⁴⁹ Cansler CA and McKenzie D. 2014. Climate, fire size, and biophysical setting control fire severity and spatial
- pattern in the northern Cascade Range, USA. Ecological Applications 24: 1037–56.
- ²⁵¹ Coen JL, Stavros EN, and Fites-Kaufman JA. 2018. Deconstructing the King megafire. Ecological Applications
- 252 **28**: 1565–80.
- ²⁵³ Collins BM, Lydersen JM, and Fry DL et al. 2016. Variability in vegetation and surface fuels across
- mixed-conifer-dominated landscapes with over 40 years of natural fire. Forest Ecology and Management 381:
- 255 74-83.
- ²⁵⁶ Collins BM, Stevens JT, and Miller JD et al. 2017. Alternative characterization of forest fire regimes:
- ²⁵⁷ Incorporating spatial patterns. Landscape Ecology **32**: 1543–52.
- ²⁵⁸ Doane D, O'Laughlin J, Morgan P, and Miller C. 2006. Barriers to wildland fire use. 12: 3.
- ²⁵⁹ Eidenshink J, Schwind B, and Brewer K et al. 2007. A Project for Monitoring Trends in Burn Severity. Fire
- 260 Ecology 3: 3-21.
- Harvey BJ, Donato DC, and Turner MG. 2016. Drivers and trends in landscape patterns of stand-replacing
- 262 fire in forests of the US Northern Rocky Mountains (1984-2010). Landscape Ecology 31: 2367-83.
- ²⁶³ Janzen FJ and Stern HS. 1998. Logistic Regression for Empirical Studies of Multivariate Selection. Evolution
- 264 **52**: 1564–71.
- Key CH and Benson NC. 2006. Landscape Assessment (LA).: 55.
- ²⁶⁶ Koontz MJ, Fick SE, and Werner CM et al. 2019a. Wildfire severity, vegetation characteristics, and regional
- 267 climate for fires covering more than 4 hectares burning in yellow pine/mixed-conifer forests of the Sierra
- Nevada, California, USA from 1984 to 2017. Open Science Framework.
- Koontz MJ, North MP, and Werner CM et al. 2019b. Local variability of vegetation structure increases forest
- 270 resilience to wildfire. EcoEvoRxiv.
- 271 Lande R and Arnold SJ. 1983. THE MEASUREMENT OF SELECTION ON CORRELATED CHARAC-
- ²⁷² TERS. Evolution **37**: 1210–26.
- Lydersen JM, North MP, Knapp EE, and Collins BM. 2013. Quantifying spatial patterns of tree groups and
- gaps in mixed-conifer forests: Reference conditions and long-term changes following fire suppression and

- 275 logging. Forest Ecology and Management **304**: 370–82.
- Mallek C, Safford H, Viers J, and Miller J. 2013. Modern departures in fire severity and area vary by forest
- type, Sierra Nevada and southern Cascades, California, USA. Ecosphere 4: art153.
- ²⁷⁸ Millar CI and Stephenson NL. 2015. Temperate forest health in an era of emerging megadisturbance. Science
- 279 **349**: 823–6.
- 280 Miller JD and Safford HD. 2017. Corroborating Evidence of a Pre-Euro-American Low- to Moderate-Severity
- ²⁸¹ Fire Regime in Yellow PineMixed Conifer Forests of the Sierra Nevada, California, USA. Fire Ecology 13:
- 282 58-90.
- ²⁸³ Miller JD and Thode AE. 2007. Quantifying burn severity in a heterogeneous landscape with a relative
- version of the delta Normalized Burn Ratio (dNBR). Remote Sensing of Environment 109: 66-80.
- North M. 2012. Managing Sierra Nevada forests. Albany, CA: U.S. Department of Agriculture, Forest Service,
- 286 Pacific Southwest Research Station.
- North MP, Stephens SL, and Collins BM et al. 2015. Reform forest fire management. Science 349: 1280-1.
- 288 Rouse W, Haas RH, Deering W, and Schell JA. 1973. MONITORING THE VERNAL ADVANCEMENT
- AND RETROGRADATION (GREEN WAVE EFFECT) OF NATURAL VEGETATION. Greenbelt, MD,
- 290 USA: Goddard Space Flight Center.
- ²⁹¹ Safford HD and Stevens JT. 2017. Natural Range of Variation for Yellow Pine and Mixed-Conifer Forests in
- the Sierra Nevada, Southern Cascades, and Modoc and Inyo National Forests, California, USA.
- ²⁹³ Short KC. 2017. Spatial wildfire occurrence data for the United States, 1992-2015 [FPA_FOD_20170508]
- 294 (4th Edition).
- 295 Steel ZL, Koontz MJ, and Safford HD. 2018. The changing landscape of wildfire: Burn pattern trends and
- implications for California's yellow pine and mixed conifer forests. Landscape Ecology 33: 1159–76.
- 297 Steel ZL, Safford HD, and Viers JH. 2015. The fire frequency-severity relationship and the legacy of fire
- ²⁹⁸ suppression in California forests. *Ecosphere* **6**: art8.
- 299 Stephens SL, Collins BM, and Fettig CJ et al. 2018. Drought, Tree Mortality, and Wildfire in Forests
- 300 Adapted to Frequent Fire. BioScience 68: 77–88.
- 301 Stevens JT, Collins BM, and Miller JD et al. 2017. Changing spatial patterns of stand-replacing fire in

- ³⁰² California conifer forests. Forest Ecology and Management **406**: 28–36.
- Welch KR, Safford HD, and Young TP. 2016. Predicting conifer establishment post wildfire in mixed conifer
- forests of the North American Mediterranean-climate zone. Ecosphere 7: e01609.
- Westerling AL. 2006. Warming and Earlier Spring Increase Western U.S. Forest Wildfire Activity. Science
- 306 **313**: 940–3.
- Westerling AL. 2016. Increasing western US forest wildfire activity: Sensitivity to changes in the timing of
- spring. Philosophical Transactions of the Royal Society B: Biological Sciences 371: 20150178.