

Description of Peatland DOS-TEM

There are five main modules of peatland DOS-TEM that affect SOC dynamics in the model: 1) the environmental module that simulates the soil temperature, moisture, and water table using the input climate datasets (e.g., air temperature, precipitation); 2) the ecological module that simulates carbon (C) and nitrogen (N) dynamics of both vegetation and soil; 3) the dynamic organic soil module that simulates the structure of SOC above the mineral soil based on relationships between SOC thickness and C mass in soil organic horizons; 4) the disturbance module that simulates the impact of wildfire on vegetation and soil C stocks; and 5) the peatland module, which simulates anaerobic CH₄ and CO₂ production and the dynamics of transport pathways in the soil. The dynamics of the peatland module are affected by information received from the environmental, ecological, and dynamic organic soil modules, and the dynamics of the peatland module influence the ecological and dynamic organic soil modules. The environmental module is described in detail in Yi et al. (2009b), and the ecological and dynamic organic soil modules are described in detail by Yi et al. (2010). Here we brief descriptions of these modules prior to describing the peatland module in detail. Because we did not implement the effect of fire on peatland dynamics in this study, we do not describe the disturbance module; information on the disturbance module can be found in Yi et al. (2010) and Yuan et al. (2012).

The Environmental Module (EnvM)

The EnvM operates at a daily time step using daily air temperature, vapor pressure, surface solar radiation and precipitation, which are downscaled from monthly input data. The EnvM considers the radiation and water fluxes among the atmosphere, canopy, snow pack and soil. Live moss, fibrous, and amorphous organic horizons are considered in the soil column in addition to a mineral soil horizon. Soil moisture and temperature are updated at daily time step.

A two-directional Stefan algorithm is used to predict the positions of freezing/thawing fronts in the soil. The temperature of soil layers above first freezing/thawing front and below the last freezing/thawing front is updated separately by solving finite difference equations. Temperatures of the soil layers between the first and last freezing/thawing fronts are assumed to be at the freezing point. Soil moistures are only updated for unfrozen layers by solving Richard equation. Both the thermal and hydraulic properties of soil layers are affected by its water content. The simulated estimates of daily evapotranspiration, soil temperature and moisture are integrated to monthly values, and provided to EcoM. See Yi et al. (2009a) for more details on the EnvM and an evaluation of the performance of the soil temperature and moisture simulations by the module. Other changes to EnvM are described in Yi et al. (2010).

The Ecological Module of Peatland DOS-TEM (EcoM)

The detailed description of most of the ecological processes that calculate the monthly pools and fluxes of carbon and nitrogen in peatland DOS-TEM have largely been documented in previous studies (Euskirchen *et al.*, 2006, McGuire *et al.*, 1992, Raich *et al.*, 1991, Tian *et al.*, 1999, Yi *et al.*, 2010, Zhuang *et al.*, 2003). The EcoM operates at monthly time step driven by monthly atmospheric climate input data and simulated environmental data. Monthly leaf area index (LAI) is estimated in EcoM, and provided to EnvM at end of each month. The fibrous and amorphous organic horizon thicknesses are updated at the end of each year, based on the simulated soil C in each horizon. The thicknesses of organic horizons are provided to the dynamic organic soil module. Here, we provide descriptions of the calculations of gross and net primary production, and of the fate of detrital inputs and the calculation of heterotrophic respiration.

Gross primary production (GPP) is calculated at a monthly time step and is affected by several factors (Zhuang *et al.*, 2003):

$$GPP = C_{max} f(PAR) f(PHENOLOGY) f(FOLIAGE) f(T) f(C_a, G_v) f(NA) f(FT) \quad [1]$$

where C_{max} is the maximum rate of C assimilation, PAR is photosynthetically active radiation, $f(PHENOLOGY)$ is monthly leaf area relative to leaf area during the month of maximum leaf area, $f(FOLIAGE)$ represents the ratio of canopy leaf biomass relative to maximum leaf biomass, $f(T)$ represents the effect of air temperature ($^{\circ}C$), C_a and G_v are atmospheric CO_2 concentration and relative canopy conductance, respectively, $f(C_a, G_v)$ represents the effect of stomatal regulation on atmospheric CO_2 uptake, $f(NA)$ represents the limiting effect of available inorganic N on GPP, and $f(FT)$ represents the effect of freeze and thaw on photosynthetic activity. Except for C_{max} , other factors in the GPP equation range from 0 to 1. The function $f(NA)$ is dynamically calculated each month based on the ability of vegetation to mobilize nitrogen from uptake and storage to meet the demands of building new tissue, i.e., net primary production. In other words, at the monthly scale GPP is downscaled so that nitrogen requirement of new production is equal to the amount of nitrogen that vegetation can provide to new production.

Net primary production (NPP) is calculated as the different between GPP and autotrophic respiration (R_A). The flux R_A is the sum of maintenance respiration (R_m) and growth respiration (R_g), which is prescribed to be 20% of the difference between GPP and R_m . The flux R_m is a direct function of plant biomass as follows:

$$R_m = K_r C_v e^{rT} \quad [2]$$

where K_r is the per-gram-biomass respiration rate of the vegetation at 0 °C, C_v is the vegetation carbon pool, T is the mean monthly air temperature (°C), and r is the instantaneous rate of change in respiration with the change in temperature.

Detrital inputs of C into the soil are divided into aboveground litterfall and belowground detrital inputs. Aboveground litterfall is assigned only to the first soil layer, while belowground detrital inputs are assigned to different soil layers based on fractional distribution of fine roots with depth. We assume that the ratio of aboveground litterfall to total litter input is similar to the ratio of root NPP to total NPP. Soil C is tracked in the fibrous, amorphous, and mineral horizons in peatland DOS-TEM, and heterotrophic respiration for each layer within a soil horizon is calculated based on the soil environmental conditions of that layer:

$$R_{H,i} = K_{co2,i} C_{s,i} f(M_{v,i}) f(T_{c,i}) \quad [3]$$

where i is soil layer index, $R_{H,i}$ is heterotrophic respiration from layer i ($\text{g C m}^{-2} \text{ h}^{-1}$), $K_{co2,i}$ is heterotrophic respiration rate at 0°C of layer i (h^{-1}), $C_{s,i}$ is soil C storage in layer i (g C m^{-2}), and $f(M_{v,i})$ and $f(T_{c,i})$ are moisture and temperature factors affecting decomposition of layer i . $f(M_{v,i})$ is parabolic relationship using the pre-defined parameters for the maximum, minimum, and optimal volumetric soil moisture for decomposition, which are 1, 0, and 0.5, respectively. When simulated volumetric soil moisture equals 0.5, $f(M_{v,i})$ equals 1. $f(T_{c,i})$ is calculated based on a Q_{10} (2.0 in this study) and soil temperature of layer i . There is a unique rate limiting parameter for decomposition K_{co2} for the fibrous, amorphous, and mineral soil horizons, that is used to estimate decomposition for each layer within the horizon. The rate limiting parameter is determined by calibrating the model to target values for the C content of the three organic horizons.

Dynamic organic soil Module (DOSM)

DOSM updates the organic soil structure at the time of fire and at the end of each year, based soil C content of the fibrous and amorphous horizons. DOSM is important in defining the structure of soil organic horizons for the purpose of maintaining the stability and efficiency of soil temperature and moisture calculations when thickness of organic soil C is altered by either wildfire disturbance or ecological processes. The soil organic structure consists of a maximum of 1 moss layer, 3 fibrous organic layers, and 10 amorphous organic layers. It is assumed that the minimum soil layer thickness for each horizon is 2 cm. If the thickness of a layer is less than 2 cm, a layer will be combined with other layers of the same horizon, or be reset to 2 cm if there is only one layer for a horizon (except for live moss, which will not be included in the soil column if it is less than 2 cm). The rationale behind the assignment of the organic soil layer thickness is that the upper layers in the soil column should be thinner than deeper layers, following the common practice of land surface models and ecosystem models in simulation soil thermal and moisture dynamics. At the same time, the upper layer should not so thin that it leads to instability and inefficiency in calculations of soil thermal and moisture dynamics. The assignment of layer thickness for the fibrous organic horizon is provided in Table D1 of Yi et al. (2010). The thicknesses and number of layers in the amorphous organic horizon (n_{amp}) are based the thickness of deepest fibrous horizon layer ($d_{fib,bot}$) and the total thickness of amorphous organic horizon (d_{amp}):

$$n_{amp} = \begin{cases} 1 & d_{amp} \leq d_{fib,bot} \\ 2 & d_{fib,bot} < d_{amp} \leq 2d_{fib,bot} \\ \vdots & \\ 9 & 8d_{fib,bot} < d_{amp} \leq 9d_{fib,bot} \\ 10 & d_{amp} > 9d_{fib,bot} \end{cases} \quad [4]$$

The thicknesses of layers with the amorphous horizon are calculated as:

$$d_{amp,n} = \frac{n}{\left(\frac{n_{amp}(n_{amp} + 1)}{2} \right)}, \quad n = 1, 2, \dots, n_{amp} \quad [5]$$

106 where $d_{amp,n}$ is the thickness of layer n .

107 At the beginning of each year, the thicknesses of the moss layer, fibrous organic layers,
 108 and amorphous organic layers are checked. If any of these layers have thickness less than
 109 minimum value, or if the total thickness and total number of fibrous layers are not consistent
 110 with those in Table D1 of Yi et al. (2010), then all layers of that horizon (fibrous/amorphous) are
 111 combined together, and then split according Table D1 of Yi et al. (2010) and/or equations 4 and
 112 5. The temperature of each new layer is determined by linear interpolation of the nearest soil
 113 temperatures of old organic soil structure. Soil freezing/thawing fronts are reassigned to new soil
 114 structure with the relative distance to the top of a horizon (moss, fibrous, amorphous) unchanged.
 115 Soil water content of each new organic soil layer is first retrieved by comparing the boundary of
 116 new and old soil layer structure. For example, if a new organic soil layer is completely located in
 117 an old organic soil layer, then the new organic soil layer is assigned a fraction of old organic soil
 118 layer's water content based on the ratio of thicknesses of both layers; if a new organic soil layer
 119 originated from two different old layers, the soil water content of new layer is assigned the sum
 120 of soil water contents retrieved from both old layers using the above method. After the
 121 determination of soil water content, the soil liquid and ice contents are retrieved with the position
 122 of freezing and thawing fronts in a layer.

123 The growth of moss is determined by a number of factors, including moss type, radiation,
 124 wind speed, and precipitation (Bisbee *et al.*, 2001). In peatland DOS-TEM, the biomass and NPP
 125 of moss are not simulated explicitly, as they are considered as part of overall vegetation biomass
 126 and NPP. However, the thickness of moss is explicitly considered for the purposes of soil

temperature and moisture calculations. Moss thickness is simulated as an empirical function of years since last fire based on (Yi *et al.*, 2009a):

$$d_{moss} = d_{moss,max} \frac{y_{sf}}{y_{sf} + y_{half}} \quad [6]$$

where d_{moss} is the thickness of moss (cm), $d_{moss,max}$ is the maximum thickness of moss (m), y_{sf} is number of years since last fire (year), and y_{half} is number of year which was need for moss to reach half of $d_{moss,max}$. In this study, we assigned 3.5 cm to $d_{moss,max}$ and 5 to y_{half} based on Yi *et al.* (2009a).

The thicknesses of fibrous and amorphous layers are calculated using the simulated soil C content of each horizon and the equation:

$$C_{mass} = ad^b \quad [7]$$

where C_{mass} is C content (g C cm⁻²) of an organic horizon, d is organic horizon thickness (cm), and a and b are fitted coefficients for the fibrous or amorphous horizons [See Yi *et al.* (2009a) for more detail and estimates of a and b].

As the fibrous organic horizon grows thicker, the bottom layer of the fibrous organic horizon is transferred to the amorphous organic horizon. A threshold method is used to mimic the process of humification, i.e. when the fibrous organic horizon becomes thicker than the threshold, the component of the fibrous organic horizon above the threshold is transferred to the amorphous organic horizon.

The Peatland Module

In the peatland module, microbial decomposition of SOC in peatlands/wetlands is divided into aerobic and anaerobic decomposition. Aerobic decomposition, i.e. heterotrophic respiration, is assumed to exclusively occur in the unsaturated zone (above water table) and CO₂ is the only

terminal product of aerobic decomposition. Aerobic decomposition is estimated by the heterotrophic respiration calculations described for the EcoM above. Anaerobic decomposition occurs exclusively in the saturated zone (below the water table) and both CH₄ and CO₂ are the terminal products of anaerobic decomposition.

The production of CH₄ in the model (anaerobic decomposition of SOC) is a function of a decomposition rate limiting parameter, SOC mass, and soil temperature (Q₁₀ of 2.0) in each SOC layer. The anaerobic decomposition rate limiting parameters of fibric, amorphous, and mineral SOC are unknown and were calibrated with field observations as discussed later. Once CH₄ is produced in the soil, its transport is controlled by various mechanisms. Below we describe the processes that affect the mass balance of CH₄ in the soil of peatland DOS-TEM.

The mass balance of CH₄ in a given soil layer is defined as:

$$\frac{\partial C_{CH_4}}{\partial t} = P - \frac{\partial diff}{\partial z} - E - O - R_p \quad [8]$$

where C_{CH_4} is the CH₄ concentration (μmol L⁻¹), t is the time, P is the CH₄ production rate (μmol L⁻¹ h⁻¹), $diff$ is the diffusion transport of CH₄ (μmol cm L⁻¹ h⁻¹), z is the soil depth (cm), E and R_p represents the CH₄ emission rates through ebullition and plant-mediated processes (μmol L⁻¹ h⁻¹), respectively, and O represents the CH₄ oxidation rate (μmol L⁻¹ h⁻¹).

The production of CH₄ is strongly related to soil carbon quality as well as soil moisture and temperature conditions, and defined separately for each soil organic layer in our study, following Ise et al. (2010) as:

$$P = \begin{cases} U \frac{K_{CH_4,i} C_{s,i} f(T_{c,i})}{Z_i \theta_w} & \text{if } Z_i \leq Z_{WT} \\ 0 & \text{if } Z_i > Z_{WT} \end{cases} \quad [9]$$

where P is the CH_4 production rate ($\mu\text{mol L}^{-1} \text{h}^{-1}$), U is the unit conversion factor, $K_{\text{CH}_4,i}$ is the CH_4 production rate (h^{-1}) at 0°C of layer i , Z_i is the layer thickness (cm), $f(T_{c,i})$ represents the dependence of decomposition SOC to CH_4 on soil temperature, θ_w is the soil moisture content ($\text{cm}^3 \text{cm}^{-3}$), Z_i is the depth of given soil layer (positive number with zero at the soil surface), Z_{WT} is the depth of water table (positive number with zero at the soil surface). It was assumed in the model that the anaerobic condition only occurs when soil was totally saturated [i.e., the soil layer is below water table (Z_{WT})]. There is a unique rate limiting parameter for decomposition K_{CH_4} for the fibrous, amorphous, and mineral soil horizons, that is used to estimate decomposition for each layer within the horizon. The rate limiting parameter is determined by calibration using the measured CH_4 efflux data as presented in the model calibration section.

The diffusion of CH_4 in the gaseous and liquid phase is defined, based on Fick's Law, as:

$$\text{diff} = -D_g \frac{\partial C_{\text{CH}_4}}{\partial z} \quad [10]$$

where D_g is the CH_4 diffusion coefficient ($\text{cm}^2 \text{h}^{-1}$) that is defined as:

$$D_g = D_0 \tau \left(\frac{T_{k,i}}{293.15} \right)^{1.75} \quad [11]$$

where D_0 is the CH_4 diffusion coefficient in the atmosphere ($\text{cm}^2 \text{h}^{-1}$), τ is the soil tortuosity in gaseous or liquid phase (unitless), and $T_{k,i}$ is the absolute soil temperature (K). D_0 was set to $720 \text{ cm}^2 \text{h}^{-1}$ (Walter & Heimann, 2000). The traditional Penman (1940) and Millington and Quirk (1961) models are known to overestimate and underestimate diffusion coefficient, respectively, at high moisture content (or low air-filled porosity). Pingintha et al. (2010) used six tortuosity models to calculate the soil gas efflux and then compared the results of six most commonly used diffusion models with the laboratory measured gas efflux. The results indicated that the Moldrup et al. (1997) model was the most robust model to calculate soil gas efflux in soils with high

185 moisture content (e.g., peatlands). Therefore, the tortuosity in gaseous and liquid phases is
 186 defined as:

$$\begin{cases} \tau = 0.66(\rho_s - \theta_w) \left(\frac{\rho_s - \theta_w}{\rho_s} \right)^{\frac{12-m}{3}} & \text{for gaseous phase} \\ \tau = 0.66\theta_w \left(\frac{\theta_w}{\rho_s} \right)^{\frac{12-m}{3}} & \text{for liquid phase} \end{cases} \quad [12]$$

187 where ρ_s is the soil porosity ($\text{cm}^3 \text{ cm}^{-3}$), and m is an empirical parameter (unitless) set to 3.0
 188 following Pingintha et al. (2010). The upper boundary condition for Eq. [3] is defined as a
 189 Dirichlet condition:

$$C_{CH_4}(t, z = 0) = 0.076 \mu\text{mol L}^{-1} \quad [13]$$

190 where $0.076 \mu\text{mol L}^{-1}$ is the CH_4 concentration in the atmosphere. The lower boundary condition
 191 is defined as a Neumann condition:

$$\frac{\partial C_{CH_4}}{\partial z} = 0 \quad [14]$$

192 It is assumed in the model that bubbles are formed in the saturated zone when the CH_4
 193 concentration in the soil water is greater than a certain threshold value (i.e., saturation
 194 concentration of CH_4). The threshold value is a function of soil temperature and determined
 195 based on the observed solubility of CH_4 in pure water at different temperatures (Wania *et al.*,
 196 2010, Yamamoto *et al.*, 1976):

$$S_B = 0.05708 - 0.001545T_{c,i} + 0.00002069T_{c,i}^2 \quad [15]$$

197 where S_B is the Bunsen solubility coefficient defined as the volume of CH_4 dissolved per unit
 198 volume of water at standard atmospheric pressure and a given temperature ($T_{c,i}$). Therefore the
 199 mass-based Bunsen solubility coefficient ($S_M, \mu\text{mol L}^{-1}$) can be calculated with the ideal gas law:

$$S_M = U \frac{pV_{CH_4}}{RT_k} = U \frac{(p_{atm} + \rho_w gh)V_{CH_4}}{RT_k} \quad [16]$$

where U is the unit conversion factor, p is the partial pressure of CH_4 that is the sum of atmospheric (p_{atm}) and hydrostatic pressure ($\rho_w gh$), h is the water height (m), V_{CH_4} is the volume of CH_4 (m^3), R is the gas constant ($8.3145 \text{ m}^3 \text{ Pa K}^{-1} \text{ mol}^{-1}$), and T is the absolute temperature (K).

Therefore, the ebullition emission rate of CH_4 from the saturated zone to the unsaturated zone is defined as:

$$E = \begin{cases} k_h (C_{CH_4} - S_M) & \text{if } C_{CH_4} > S_M \text{ and } Z_i \leq Z_{WT} \\ 0 & \text{if } Z_i > Z_{WT} \end{cases} \quad [17]$$

where k_h is a rate constant (1.0 h^{-1}). The oxidation of CH_4 ($CH_4 + 2O_2 \rightarrow CO_2 + 2H_2O$) in the unsaturated soil layers is assumed to follow Michaelis-Menten kinetics and strongly controlled by soil temperature (Walter & Heimann, 2000)

$$O = \begin{cases} \frac{V_{max} C_{CH_4}}{k_m + C_{CH_4}} f(T_{c,i}) & \text{for } Z_i \geq Z_{WT} \\ 0 & \text{for } Z_i < Z_{WT} \end{cases} \quad [18]$$

where V_{max} and k_m are the Michaelis-Menten kinetics parameter and set to $5 \text{ } \mu\text{mol L}^{-1} \text{ h}^{-1}$ and $20 \text{ } \mu\text{mol L}^{-1}$ (Walter & Heimann, 2000), respectively.

The CH_4 lost through the vascular plant aided transport is defined as (Walter & Heimann, 2000):

$$R_p = k_p f_{root} f_{grow} T_{veg} C_{CH_4} \quad [19]$$

where R_p is the CH_4 emission rate through plant at a given soil layer ($\mu\text{mol L}^{-1} \text{ h}^{-1}$), k_p is the rate constant (1.0 h^{-1}), f_{root} is the root distribution in soil, T_{veg} is a factor describing how different types of vegetation/plants affect CH_4 transport, f_{grow} is the growing stage of plants/vegetation,

216 and C_{CH_4} is the CH_4 concentration at a given soil layer. The growing state of plants/vegetation
 217 was calculated based on the soil temperature and leaf area index:

$$f_{grow} = \frac{LAI_c}{LAI_{max}} \quad [20]$$

218 where LAI_c is the stand LAI at the stand age c and LAI_{max} is the maximum LAI when
 219 plants/vegetation reach maturity. Equation (20) implies that the seasonal variations on the plant-
 220 mediated CH_4 transport are negligible based on the results of Shea (2010). Shea (2010) also
 221 reported that the vegetation species and functional groups (e.g., *Carex*) are important to plant-
 222 mediated CH_4 transport, which is reflected as T_{veg} in the model.

223 The total CH_4 efflux from the soil to the atmosphere is be calculated as:

$$F = \begin{cases} U(Q_p + diff_a(t, z = 0)) & \text{if } Z_{WT} \leq 0 \\ U(Q_p + E + diff_w(t, z = 0)) & \text{if } Z_{WT} > 0 \end{cases} \quad [21]$$

224 where F is the CH_4 emission flux ($\mu\text{mol m}^{-2} \text{h}^{-1}$) from the soil to the atmosphere, U is the unit
 225 conversion factor, Q_p is the total plant-aided CH_4 transport ($\mu\text{mol m}^{-2} \text{h}^{-1}$), $diff_a$ and $diff_w$ are the
 226 diffusion transport of CH_4 in the gaseous and liquid phases ($\mu\text{mol m}^{-2} \text{h}^{-1}$), respectively. In the
 227 model, it is assumed that 50% of CH_4 transported by plant is oxidized by rhizospheric oxidation
 228 before being released into the atmosphere (Walter & Heimann, 2000). The total plant mediated
 229 CH_4 transport (Q_p) is calculated by integrating through the root zone:

$$Q_p = 0.5 \int_{z_{root}}^0 R_p dz \quad [22]$$

230 where z_{root} is the root depth (soil depth at the soil surface is assumed to be zero).

231 The total CO_2 release from the soil is calculated as the sum of CO_2 produced from
 232 decomposition of SOC under aerobic conditions (i.e., heterotrophic respiration), CO_2 produced
 233 during the microbial oxidation of CH_4 in the oxic zone, and the anaerobic production of CO_2 .

234 The aerobic production of CO₂ in the model is described above in the description of the EcoM.
235 The anaerobic production of CO₂ in the model is calculated based on the aerobic production of
236 CO₂. Many laboratory incubation studies (e.g., Bergman *et al.*, 1999, Glatzel *et al.*, 2004, Kane
237 *et al.*, 2012, Lee *et al.*, 2012, Updegraff *et al.*, 1995) indicate that aerobic:anaerobic CO₂
238 production ratios in high-latitude peatland ecosystems range between 0.28:1 and 5:1. Therefore,
239 we assumed that the ratio of aerobic CO₂ to anaerobic decomposition is 2 for an aerobic CO₂
240 flux calculated for simulated soil temperature and an optimum soil moisture of 50% saturation.
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