

# Not Gone With the Wind: Long-Run Impacts of Herbicidal Warfare in Vietnam<sup>\*</sup>

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## Abstract

We investigate the legacies of herbicidal warfare on population size and growth in Vietnam in 2001–2020. The identification strategy exploits the flight-level military records and the plausibly exogenous hamlet-level variation in herbicide exposure around the spray-on, direction-change, and spray-off points of fixed-wing aircraft missions, combined with a fuzzy regression kink design. The elasticity estimate suggests that a 5% increase in herbicide exposure in the Vietnam War is associated with a decrease in population *size* of 0.80–0.99% in 2001 and 1.24–1.43% in 2020. We also find negative associations between herbicide exposure and overall and annual population *growth rates*.

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## Declarations of Interest

None

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# 1 Introduction

Does a temporary shock of large-scale violence leave lasting effects on socioeconomic outcomes? In some instances, exposure to violence persistently depletes human capital and undermines interpersonal and institutional trust (e.g., Grasse, 2023; Lichter, Löffler and Siegloch, 2020; Nunn and Wantchekon, 2011). Elsewhere, forms of violence improve political engagement, social cohesion, and altruism (e.g., Bauer et al., 2016; Berman, Clarke and Majed, 2023; Blattman, 2009). Several studies also highlight a rapid recovery or catch-up growth of economic and population outcomes in post-war societies, with the impact of wartime destruction remaining short-lived (e.g., Brakman, Garretsen and Schramm, 2004; Davis and Weinstein, 2002, 2008; Miguel and Roland, 2011; but see Bosker et al., 2007).

This article joins the growing literature with new evidence from the herbicide spray in the Vietnam War. While earlier works reveal a powerful recovery from the destruction (e.g., Miguel and Roland, 2011), recent literature finds persistent effects of U.S. bombing on development and health outcomes in Cambodia (e.g., Lin, 2022), Laos (e.g., Riaño and Caicedo, 2021; Yamada and Yamada, 2021), and Vietnam (e.g., Palmer et al., 2019; Singhal, 2019). Most closely related to this article, Appau et al. (2021), Le, Pham and Polacheck (2022), and Yamashita and Trinh (2022) examine the legacies of herbicidal warfare. For example, Appau et al. (2021) underline a persistent negative association between the district-level herbicide exposure and the household agricultural productivity, trust levels, and economic production. In a similar vein, Le, Pham and Polacheck (2022) find a commune-level positive association between herbicide exposure and the immediate and detrimental prevalence of health disease and mobility disability in the cohort born before the spray mission ended in 1971.

While the existing insights are valuable, there remain challenges for causal identification primarily due to the commonly used, district-level or commune-level geographic aggregation. This is unfortunate, as the distribution of herbicide exposure is partly a function of micro-level factors including climate and wind conditions, local geography, and instantaneous decision of aircraft pilots besides initial mission plans. The commonly-employed geographic

aggregation risks masking such micro-level variation that would be valuable for causal identification. An alternative, straightforward empirical strategy is to exploit the micro-level variation with geographically less intensive aggregation, which we adopt in this article.

Here, we combine the fine-grained archival data previously unused at the original, flight-level scale with a fuzzy regression kink (RK) design to address the identification challenge arising from the nonrandom nature of herbicide exposure. Indeed, historical records suggest that Operation Ranch Hand (1962–1971) missions intended to prevent directly damaging densely populated areas and croplands not under Viet Cong control (MACV, 1969), which would introduce bias into naive comparisons. To surmount such identification challenges, our RK strategy leverages the *uncontrollable*, natural experimental variation in the hamlet-level exposure to herbicide around the spray-on, direction-change, and spray-off points of fixed-wing aircraft missions. To capture the micro-level variation in herbicide missions, we rely on the spray flight-level records documented in the Stellman-National Academy of Sciences version of the Herbicide Report System file (S-NAS HERBS, Stellman et al., 2003*a,b*).

The RK analysis reveals lasting legacies of herbicidal warfare. The elasticity estimate suggests that a 5% increase in herbicide exposure, which is observed within a 1 km distance from the spray start, turn, and end points, is associated with a 1.24–1.43% decrease in population size in 2020. The analysis also reveals negative associations between the herbicide shock and overall and annual population *growth rates* in 2001–2020. While the discrepancy might arise from different units of analysis, the previously unseen patterns run counter to the conventional wisdom of a powerful recovery of population from wartime destruction.

## 2 Data

### 2.1 Stellman-National Academy of Sciences HERBS File

To measure the landscape of herbicide spray, we rely on the military archival records of the Stellman-National Academy of Sciences version of the Herbicide Report System file (S-NAS

HERBS) database (Stellman et al., 2003*a,b*). The S-NAS HERBS is a compilation of military mission records of the HERBS file originally developed by the U.S. Department of Defense with corrections, containing 9,141 reports of the spray missions with several spray methods (i.e., fixed-wing aircraft, helicopters, and ground-spraying) in 1961–1971. A record in the database corresponds to a single mission with one or more spray paths. Fixed-wing aircraft missions, which dispersed approximately 95% of herbicides (Stellman et al., 2003*a*, 681–682), constitutes a majority of the records, and the following analysis relies on the aircraft missions.<sup>1</sup> Figure 1 shows the geographical and temporal distributions of the mission records.

A noteworthy aspect of the S-NAS HERBS is its geocoding accuracy. Besides mission dates, spray methods, agent, and gallonage information, the database documents “the actual flight paths taken by Ranch Hand aircraft as they carried out their spray missions,” with the “locations at which the aircraft switched directions or turned off and on their spray nozzles” (Stellman et al., 2003*b*, 323), thereby providing a detailed landscape of spray missions.<sup>2</sup>

## 2.2 Treatment: Herbicide Exposure

We use hamlets in South Vietnam as the unit of analysis in the following analysis, based on the geocoordinates (points) recorded in the Vietnam Hamlet Evaluation System (HES) Gazetteer Data (Douglass, 2011). To quantify the hamlet-level exposure to herbicide, we broadly follow the distance-weighted approach of Stellman and Stellman (1986, 309) as:

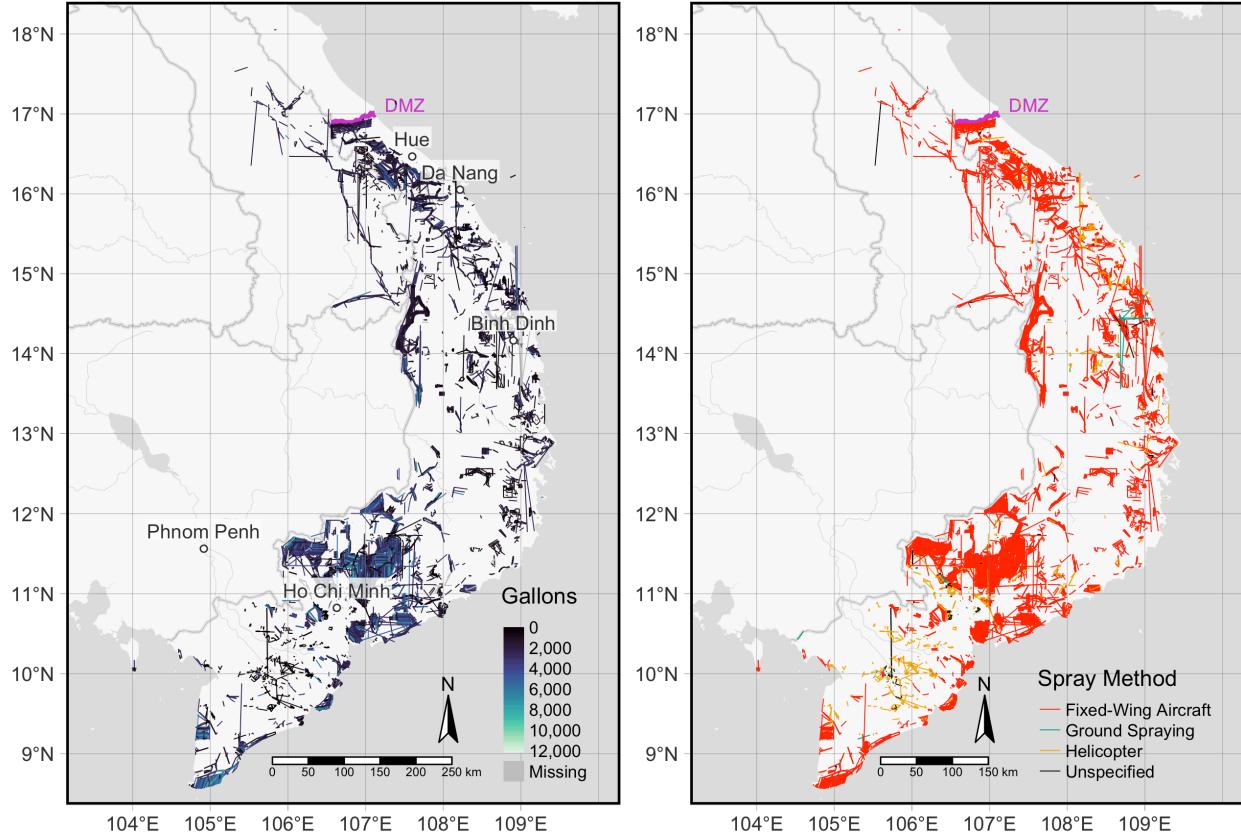
$$\text{HERB}_i = \sum_{j \in \mathcal{H}} G_j \cdot e^{-\lambda D_{ij}}, \quad (1)$$

where  $i$  indexes hamlets and  $j$  spray flight “legs” of the fixed-wing aircraft missions. Following Stellman et al. (2003*b*, 325), we first split individual flight paths (contiguous lines) into

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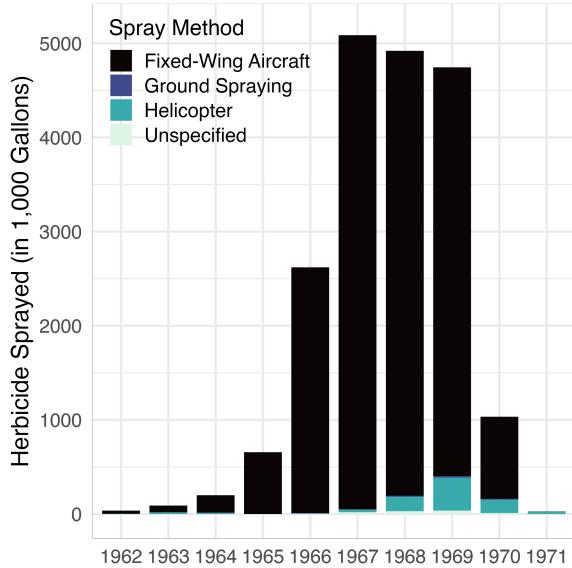
<sup>1</sup>We discard 91 errant entries of “fixed-wing aircraft” records with single (not multiple) geocoordinates.

<sup>2</sup>To navigate aircraft and record spray-on, direction-change, and spray-off points, the tactical air navigation (TACAN) system distance measuring equipment (DME) was well-developed in Southeast Asia by the 1960s. During mission flights, TACAN/DME continuously offered geographical navigation using the (short) ultra high frequency (UHF) radio range (Rowley, 1975). An essentially same navigation system guided civic aviation until the mid-1990s or before the civic use of the global positioning system (GPS).

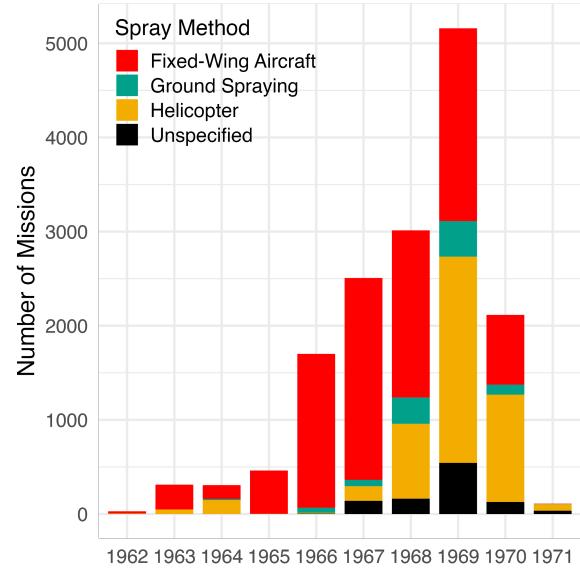


(a) Gallons Sprayed

(b) Spray Methods



(c) Gallons Sprayed by Spray Methods



(d) Number of Missions by Spray Methods

Figure 1: Herbicide Spray Missions

Notes: Segments in maps represent the spray paths. (a) Shading is proportional to the quantity of sprayed herbicide in gallons sprayed by individual spray legs (straight lines without a turn). (b) Colors indicate distinct spray methods. Thick gray lines indicate international borders. Thin gray lines and polygons represent major water bodies. (c), (d) Yearly distribution of herbicide spray (in 1,000 gallons) and spray methods, omitting four helicopter missions in 1961.

distinct flight legs indicated by the straight lines without turns. As the gallonage information is aggregated at the mission level, we divide and assign the total mission-level herbicide quantities to individual legs in proportion to its spray run length.<sup>3</sup>  $G_j$  indicates the quantity of the herbicides sprayed in leg  $j$  in gallons,  $\mathcal{H}$  is a set of all spray flight legs, and  $D_{ij}$  represents the geodesic distance between hamlet  $i$  and flight leg  $j$ .  $\lambda = \frac{\ln(2)}{D^{\text{Half}}}$  is an arbitrary parameter, where  $D^{\text{Half}}$  determines the distance at which the herbicide exposure decays to the half of a “direct” (zero-distance) hit with  $D_{ij} = 0$ , with the baseline of  $D^{\text{Half}} = 500$  meters. For a robustness check purpose, Appendix C.2 reestimates the main regressions with alternative half-decay distance parameters of 100 meters, 250 meters, and 1 km.

### 2.3 Outcome: Population Size and Growth Rate

The following analysis focuses on contemporary population size and population growth rate as outcome variables. We rely on the WorldPop data, which provides geographically disaggregated records of population estimates at 100 meter level on an annual basis in the 2001–2020 period.<sup>4</sup> To construct the population size variable, we first extract the annual records of population counts at the hamlet locations, and take the natural logarithm of the variable after adding 1. The (logged) overall and annual population growth rates, respectively, are measured as  $\ln \text{Population}_{2020} - \ln \text{Population}_{2001}$  and  $\ln \text{Population}_t - \ln \text{Population}_{t-1}$ .

### 2.4 Covariates: Key Spray Targets and Related Attributes

To facilitate the empirical analysis, we combine archival sources and several geographical databases to construct three sets of covariates. The first set of covariates measures the proximity to the key spray targets. Operation Ranch Hand involved two primary objectives, forest defoliation and destruction of enemy food supplies, with key targets including

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<sup>3</sup>For example, if a mission includes one flight with four legs (three turns) with an equal leg length sprayed 1,000 gallons in total, we assign 250 gallons to each flight leg.

<sup>4</sup>Available at: <https://www.worldpop.org>, accessed August 6, 2021. Due to errant entries, we exclude the population estimates in 2000 from the analysis.

“base camps and fire support bases...lines of communication, enemy infiltration routes, and enemy base camps” (Institute of Medicine, 1995, 85). The mission authorization process also intended to prevent damage to densely populated areas and crops not under Viet Cong (VC) control (MACV, 1969). To measure proximity to the key spray targets, we include the prevalence of VC control in 1967–1969 (HES, McCormick, 2021), geodesic distances to suspected areas of North Vietnam Army (NVA) bases (Enemy Base Area File, BASFA), U.S. Air Force and Navy bases, U.S. Army and Marine troops in 1961–1971 (S-NAS-HERBS), and roads (including trails, *Indochina Atlas*), average hamlet population size in 1967–1969 (HES, Douglass, 2011), and dummy variables for the presence of rice croplands and slash and burn cultivation (*Indochina Atlas*), with the details described in Appendix A.4.<sup>5</sup>

Second, geographic covariates include mean elevation and its standard deviation as a proxy of terrain ruggedness (USGS, 1996), soil suitability for rice cultivation (Zabel, Putzenlechner and Mauser, 2014), distance to rivers, flow accumulation (Lehner, Verdin and Jarvis, 2008), a dummy variable for forest presence (*Indochina Atlas*), and average precipitation and wind speed (1970–2000, Fick and Hijmans, 2017). Finally, historical covariates are the number of hamlets within a 30 km radius,<sup>6</sup> distance to railways (*Indochina Atlas*) and international borders (as of the Vietnam War period), and the average of the annual minimum distance to aerial bombing drop points in 1965–1971 (Defense Digital Service, 2016).

### 3 Identification Strategy: Fuzzy Regression Kink

To explore the evidence of causal effects, we rely on a fuzzy regression kink (RK) design that exploits the exogenous fluctuations in herbicide exposure within geographically small areas around the spray start, turn, and end points. Recall that the S-NAS-HERBS database

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<sup>5</sup>Where multiple hamlet-month observations are available, VC control measures the average of a dummy variable indicating VC control. To measure road proximity and cropland presence, we georeferenced and image-processed the maps of *Indochina Atlas* compiled by the Central Intelligence Agency (CIA) in 1970. For the locations of NVA bases, we rely on the Enemy Base Area File (BASFA), July 1, 1967–July 1, 1971.

<sup>6</sup>The 30 km cutoff reflects the historical facts of Ranch Hand missions and standard error clustering in the regression estimation. See model specification section for details.

that documents the “actual flight paths” and the “locations at which the aircraft switched directions or turned off and on their spray nozzles” (Stellman et al., 2003b, 323).<sup>7</sup> The core idea behind our RK strategy is that the locations at which aircraft makes turns and turned on and off spray nozzles and the herbicide dispersal were, at least partly, *uncontrollable*. Besides intended targets, the realized distribution of herbicide was driven by plausibly exogenous micro-level factors including climate conditions, wind, terrain, and turbulence from the aircraft as well as ground fire hits (Institute of Medicine, 1995, 86–87). Consistent with the identification idea, “[t]he responsibility for flying the C-123 during the crucial spraying part of each mission was shared between the pilot and the copilot,” with the pilot having “control of the switches which started and stopped the spray” (Buckingham, 1982, 37). As such, instantaneous decision of aircraft pilots and drift due to the disturbing factors jointly determined the spray dispersal, which inevitably deviated from the initial mission plans and generated haphazard, natural-experimental variation in the herbicide distribution.<sup>8</sup>

Consequently, within geographically small areas around the spray start, turn, and end points, the realized herbicide dispersal generates a discontinuous *slope change*, or a discontinuity in the first derivative, in the distribution of herbicide exposure, which remains uncorrelated with potential confounding forces. Our RK strategy leverages this natural experimental kink in the treatment function to derive causal identification. Intuitively, we compare hamlets that were barely covered by spray flights and received direct hits (i.e., “treated” hamlets with greater herbicide exposure due to direct hits) with hamlets that were sufficiently close to but located barely outside of the spray flight paths and received indirect hits (i.e., “control” hamlets with less herbicide exposure due to accidental hits). Located within geographically small areas, these treated and control hamlets should have

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<sup>7</sup>See footnote 2 for the aircraft navigation system in the Vietnam War period. Also note that a fuzzy RK design allows incomplete manipulation while addressing measurement errors in the treatment and the running variable (Card et al., 2015b, 2467–2469). Appendix B investigates potential nonlinearity in the covariate distributions across the kink point, and Appendix C examines model dependence and how model specification influences the main findings.

<sup>8</sup>To illustrate, approximately 29% of aircraft sorties were intercepted by ground fire in 1966 (Institute of Medicine, 1995, 86), and the crop damage induced by drift on defoliation missions was greater than the damage by crop destruction missions (National Academy of Sciences, 1974, S-5).

similar (or more precisely, no kink in) geographic, historical, and socioeconomic attributes prior to herbicide exposure, which we empirically validate in the following section.<sup>9</sup>

Here, we adopt a RK design rather than a regression discontinuity (RD) design given the probable drifts of herbicide outside the areas with direct hits. We rely on a fuzzy, not sharp, RK design given the unknown slope change parameter of the herbicide exposure (treatment) function at the kink point as well as probable measurement errors in the treatment and the running variable retrieved from archival records (Card et al., 2015*b*, 2464–2467).

### 3.1 Running Variable

The RK design requires a subset of sample hamlets located within sufficiently small geographic areas around the spray start, turn, and end points. The analysis also requires the corresponding distance measure between hamlet locations and the spray start, turn, and end points as the running variable as well as a dummy variable indicating actual spray hits.

The coding procedure involves several steps. First, we extend the recorded flight paths of fixed-wing aircraft missions from the flight start, turn, and end points by an arbitrary length of 0.05 degree  $\approx$  5.56 km. Second, we add buffers of 0.001 degree  $\approx$  111 meter width to both sides (222 meters) of the extended flight paths, as graphically illustrated by Figure 2(a). The baseline extension length of 5.56 km and the 222 meter buffer width mimic ordinary spray flights in Operation Ranch Hand. A routine fixed-wing aircraft mission involved multiple aircraft and dispersed herbicide at the airspeed of 130–150 knots (240–278 km/h) and an altitude of 150 feet, with each aircraft covering a (laterally contiguous) swath of 80 meter wide and 16 km long (Buckingham, 1982, 37, 132; Institute of Medicine, 1995, 25, 86–87; Stellman et al., 2003*b*, 327). An extension length of 5.56 km approximately corresponds to the “less-than-one-half-minute-away” distance from the observed kink point at the airspeed of 240–278 km/h ( $= 4\text{--}4.63 \text{ km/m}$ ). A 222 meter buffer width similarly approximates the combined swath of a routine spray mission involving three airplanes. This 222 meter or

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<sup>9</sup>As discussed below, a valid RK design requires that covariate distributions evolve smoothly across the kink point as well as the smoothness of the running variable distribution.

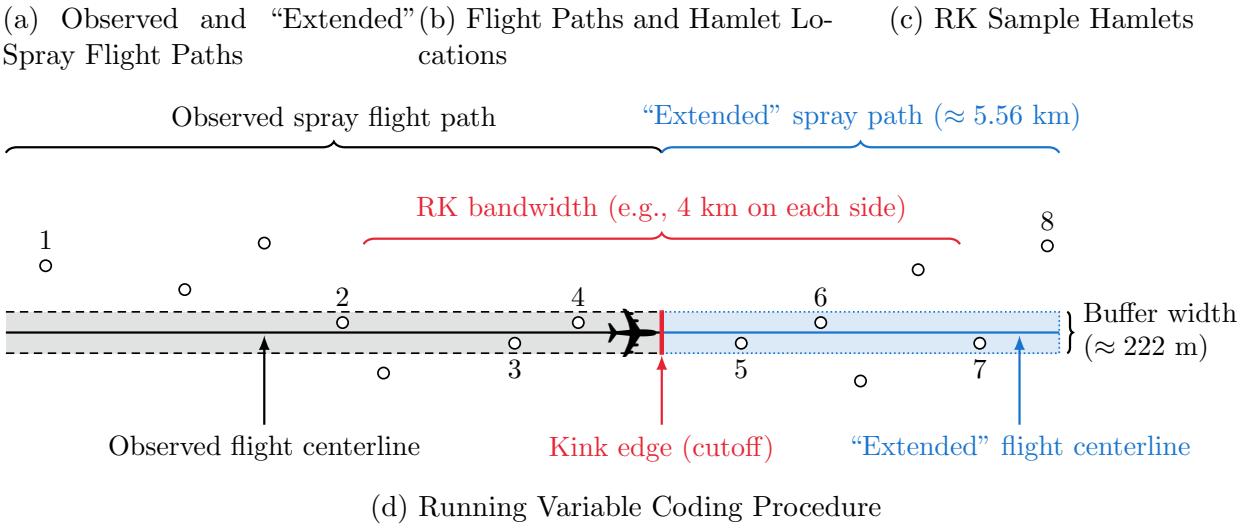
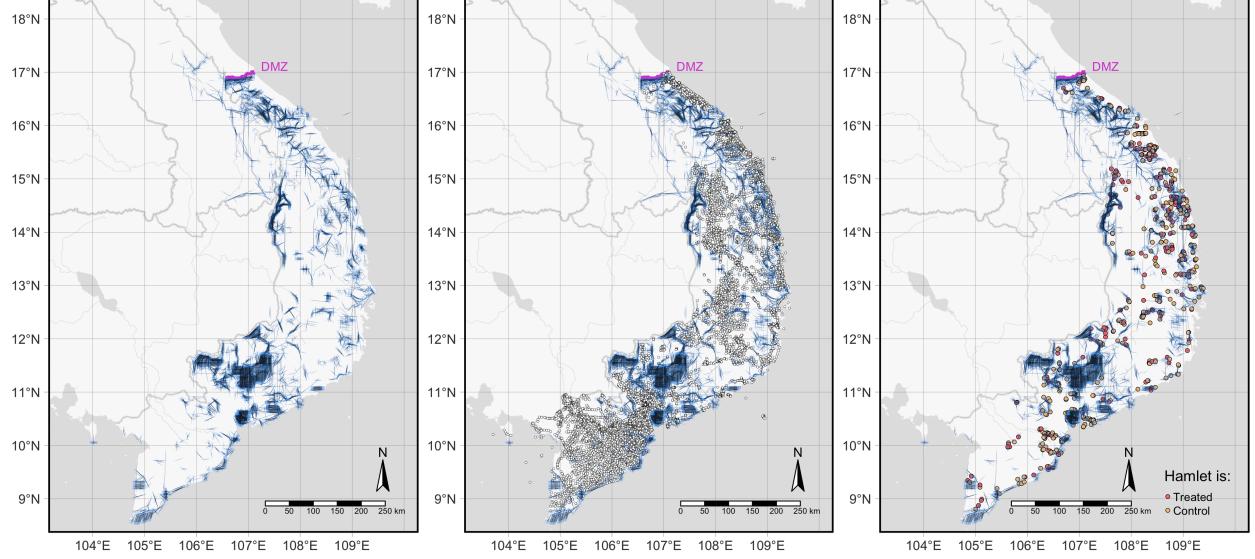


Figure 2: Spray Flight Paths, Hamlet Locations, and the RK Running Variable Coding Procedure

*Notes:* (a) Black polygons represent observed flight paths, and blue polygons indicate the extended flight paths with a 0.05 degree ( $\approx 5.56$  km) distance from the start/turn/end points with a 111 meter buffer on each side (222 meters in total). Of the 14,733 hamlets (Panel (b)), hamlets covered by observed flight polygons with a 222 meter buffer are categorized into the treatment groups (red dots), while the hamlets only covered by a blue polygon are categorized into the control group (yellow dots; Panel (c)). (d) Vertical thick segment indicates a kink edge. Dots and horizontal solid segments, respectively, represent hamlet locations and observed and “extended” spray flight paths. Dashed (dotted) segments and shades represent the buffer around the observed (extended) spray path.

three-aircraft approximation also reflects the historical fact that three C-123 airplanes were assigned to Operation Ranch Hand until 1964, and the number of C-123s increased to 12 in 1965 and then to 36 in 1967 (Institute of Medicine, 1995, 86).

Third, we overlay the extended flight path polygons onto hamlet locations. The hamlets

covered by the flight path polygons without the 5.56 km extension are coded as “treated,” while those only covered by the extended parts are categorized into the control group (Figures 2(b) and (c)). For example, in Figure 2(d), among the hamlets located within the 5.56 km distance from the kink edge, polygon edges corresponding to the spray start, turn, end points, hamlets 2–4 are categorized into the treatment group while hamlets 5–8 are categorized into the control group. We then measure the geodesic distance between hamlet locations and the nearest kink edges to assign running variable to individual hamlets.<sup>10</sup>

As in Figure 2(c), the full RK sample includes the hamlets that are geographically covered by the observed or extended flight path polygons with the kink edge distance smaller than 5.56 km extension length. The geoprocessing leaves us a sample of 716 hamlets (4.86% of 14,733 hamlets in panel (b)) in 125 districts (38 provinces) in South Vietnam.

### 3.2 Model Specification

Our RK estimation builds upon the following two-stage model:

$$\begin{aligned} \ln \text{HERB}_{hd} &= \gamma \text{EdgeDist}_{hd} + \delta \text{EdgeDist}_{hd} \times \mathbb{1}[\text{EdgeDist}_{hd} \geq 0] + \zeta \mathbb{1}[\text{EdgeDist}_{hd} \geq 0] \\ &\quad + \mathbf{X}_{hd}^\top \boldsymbol{\beta} + \eta_d^{\text{District}} + \kappa_{a[h]}^{\text{Agent}} + \theta_{k[h]}^{\text{End-edge}} + \iota_{k[h]}^{\text{Pre-1967}} + f_1(\text{Lon}_{hd}, \text{Lat}_{hd}) + e_{hd}, \quad (2) \\ Y_{hd} &= \tau \ln \widehat{\text{HERB}}_{hd} + \lambda \text{EdgeDist}_{hd} + \nu \mathbb{1}[\text{EdgeDist}_{hd} \geq 0] + \mathbf{X}_{hd}^\top \boldsymbol{\xi} + \pi_d^{\text{District}} + \rho_{a[h]}^{\text{Agent}} \\ &\quad + \phi_{k[h]}^{\text{End-edge}} + \psi_{k[h]}^{\text{Pre-1967}} + f_2(\text{Lon}_{hd}, \text{Lat}_{hd}) + u_{hd}, \text{ if } |\text{EdgeDist}_{hd}| \leq b \quad (3) \end{aligned}$$

where  $h$ ,  $d$ ,  $a$ , and  $k$ , respectively, index hamlets, districts, herbicide agents, and kink edges. HERB denotes the herbicide exposure score, EdgeDist is the geodesic distance from the kink edges in kilometers recentered at zero, and  $\mathbb{1}[\text{EdgeDist} \geq 0]$  is a dummy variable which takes one if  $\text{EdgeDist} \geq 0$  (treatment group) and zero otherwise (control group).<sup>11</sup>  $\mathbf{X}$  is

<sup>10</sup>The illustration here assumes that single spray flights cover single hamlets. When more than one spray flight covers single hamlets, we modify the coding rule for the treated hamlets not to categorize the hamlets located close to one kink edge (e.g., 100 meters) but far from another edges (e.g., 10 km > 5.56 km threshold) into the treatment group. See Appendix A.3 for details.

<sup>11</sup>The specification includes  $\mathbb{1}[\text{EdgeDist}_{hd} \geq 0]$  and thereby allows discontinuity in the treatment function at the kink point. When the kernel is symmetric (e.g., a uniform kernel), the asymptotic bias and variance

a vector of hamlet-level covariates, and  $f_1(\text{Lon}, \text{Lat})$  and  $f_2(\text{Lon}, \text{Lat})$  are two-dimensional cubic polynomials of hamlet geocoordinates to screen out spatial trends.<sup>12</sup>  $\eta^{\text{District}}$  and  $\pi^{\text{District}}$  are (South Vietnam) district fixed effects, and  $\kappa$  and  $\rho$  are herbicide agent fixed effects.  $\theta^{\text{End-edge}}$  and  $\phi^{\text{End-edge}}$  are end-edge fixed effects which take one if a hamlet's running variable is distance to a spray flight end edge and zero otherwise (i.e., a flight start or turn edge); and  $\iota^{\text{Pre-1967}}$  and  $\psi^{\text{Pre-1967}}$  are fixed effects indicating a hamlet's running variable measured as the distance to the flight leg edges of pre-1967 missions, before the increase of the number of aircraft from 12 to 36 assigned to Operation Ranch Hand mentioned above.<sup>13</sup> The panel setup replaces the outcome with annual population growth rate and adding a year fixed effect to the right hand side.<sup>14</sup> Following Card et al. (2012, 2015b, 2017), the RK estimation relies on an uniform kernel, and the baseline setup uses a bandwidth of  $b = 4$  kilometers (522 hamlet observations). A 4 km bandwidth approximates to the “one-minute away” distance at the typical airspeed of 4–4.63 km/m (= 240–278 km/h) of fixed-wing aircraft missions. To ensure that the fixed bandwidth choice does not drive the results, we replicate the RK estimates with alternative bandwidth sizes.

The parameter of interest is  $\tau$  in the second-stage (equation 3), which captures the average effect of a marginal increase in ln HERB on the outcome at the cutoff. Formally,  $\tau$  can be interpreted as the treatment-on-the-treated (TT) effect (Florens et al., 2008) or the local average response (LAR) of herbicide exposure (Altonji and Matzkin, 2005), instrumented by  $\text{EdgeDist} \times \mathbb{1}[\text{EdgeDist} \geq 0]$  in the first stage (Card et al., 2015b).

Several aspects of the specification are worth explanations. Following the asymptotic results of Pei et al. (2022), we adopt a local linear specification given the relatively small

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of the RK estimand are not affected by the continuity imposition (Card et al., 2012, 2015b).

<sup>12</sup>Linear, quadratic, and fifth-order polynomials of longitude and latitude yield qualitatively similar results.

<sup>13</sup>The herbicide agent fixed effect has five categories: “Blue,” “Orange,” “Purple,” “White,” and “Others.” Given the small numbers of observations, we categorize “Pink” and “Unknown” into “Others.”

<sup>14</sup>With abuse of notations, the panel version of the second-stage model is:

$$Y_{hdt} = \tau \ln \widehat{\text{HERB}}_{hd} + \lambda \text{EdgeDist}_{hd} + \nu \mathbb{1}[\text{EdgeDist}_{hd} \geq 0] + \mathbf{X}_{hd}^\top \boldsymbol{\xi} + \pi_d^{\text{District}} + \rho_{a[h]}^{\text{Agent}} \\ + \phi_{k[h]}^{\text{End-edge}} + \psi_{k[h]}^{\text{Pre-1967}} + \eta_t + f_2(\text{Lon}_{hd}, \text{Lat}_{hd}) + e_{hdt},$$

where  $\eta_t$  denotes year fixed effect, with the corresponding first-stage model specified analogously.

sample size. We also rely on district, agent, end-edge, and pre-1967 mission fixed effects instead of the ideal flight or leg fixed effects due to limited sample size. Nonetheless, the specification reflects the historical fact that South Vietnam Government both at the national and local levels and the U.S. jointly controlled the herbicide missions (Buckingham, 1982, 36–38; Institute of Medicine, 1995, 86).<sup>15</sup> District fixed effects subsume the regional variation in the authorization process and other district-level differences of counterinsurgency strategies; and the remaining fixed effects absorb the types of chemical shocks (agent fixed effects), aircraft directions (end-edge fixed effects), and overall intensity of herbicide missions and potential military significance of nearby targets (pre-1967 fixed effects).

Throughout the analysis, we rely on the two-stage least-squares (2SLS) estimator. To account for the increased error due to the two-stage estimation and potential error dependence across space, we report Conley’s (1999) standard errors robust to spatial clustering with a 30 km cutoff.<sup>16</sup> Appendix A reports descriptive statistics of the variables.

### 3.3 Identification Assumption

A valid fuzzy RK design hinges on the smoothness assumption, which yields two testable implications: First, the density of the running variable is sufficiently smooth, or continuously differentiable at the cutoff; and second, predetermined covariates evolve smoothly around the kink point (Card et al., 2012, 2015*a*, 2017).

Following literature (Bana, Bedard and Rossin-Slater, 2020; Card et al., 2015*a,b*; Landais, 2015), we validate the smoothness assumption with the running variable and covariate distributions in three ways. First, Figure B.1 in the Appendix examines the continuity of the running variable distribution using the polynomial estimator of Cattaneo, Jansson and Ma

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<sup>15</sup>The authorization process also involved the U.S. Ambassador, the U.S. Military Assistance Command, and the Corps Tactical Zones (CTZs). Dell and Querubin (2018) use the CTZ (Corps I-II) boundary as one of the sources for causal identification. As the CTZ boundaries follow the province boundaries, district fixed effects subsume the difference in counterinsurgency strategy across CTZs and local governments.

<sup>16</sup>We use a 30 km cutoff to reflect the combat range of aircraft missions in which a C-123 airplane covered a swath of 80 m wide and 16 km long (Buckingham, 1982, 132). A 30 km cutoff approximately reflects the 16-km range as a radius. The sample mean (median) of the aircraft spray legs is 18 km (16.48 km).

(2020), and fails to detect statistically significant discontinuity at the kink point ( $t = -0.209$  and  $p = 0.835$ ). Second, we test for the kink in the running variable distribution using polynomial regressions.<sup>17</sup> As reported in Figure B.2 in the Appendix, the polynomial regressions fail to detect a slope change in the running variable distribution around the kink point.

Finally, Figure B.3 in the Appendix presents a series of placebo kink estimates with the covariates, *including* hamlet population in 1967–1969, as the left-hand-side variable using a reduced-form version of the RK specification. Again, and consistent with the smoothness assumption, most of the placebo regressions does not reveal a discernible slope change at the kink point. Exceptions are the substantively small but statistically significant slope change in wind speed and the distance to U.S. bases plausibly arising from random chance or sampling error. Indeed, as reported in Appendix B.2, the same randomization inference exercise introduced in the next section fails to negate that the covariate kink arises by chance. Moreover, Appendix C.1 finds little evidence of model dependence, suggesting the minor role of the marginally significant covariate kink in influencing the RK estimates.

## 4 Results

Figure 3 displays the distributions of herbicide exposure (first-stage association; panel (a)) and population size in 2020 (reduced-form association; panel (b)), given hamlet geocoordinates and fixed effects.<sup>18</sup> As the fuzzy RK estimand can be written as the ratio of the reduced-form and first-stage associations, the co-evolving slope change provides a graphical

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<sup>17</sup>Formally, and with abuse of notation, we first aggregate the hamlet observations using bins with different sizes based on the running variable and then estimate the following regression model:

$$N_b^{\text{obs.}} = \beta \mathbb{1}[\overline{\text{EdgeDist}}_b \geq 0] + \sum_{p=1}^P \left[ \gamma_p \overline{\text{EdgeDist}}_b^p + \delta_p \overline{\text{EdgeDist}}_b^p \cdot \mathbb{1}[\overline{\text{EdgeDist}}_b \geq 0] \right] + e_b,$$

where  $b$  indexes bins,  $N_b^{\text{obs.}}$  reflects the number of observations in each bin,  $\overline{\text{EdgeDist}}$  is the midpoint of EdgeDist of each bin, and  $P$  is the polynomial order.  $\delta_1$  captures the slope change of the probability density function of the running variable at the kink point. Not to violate the smoothness assumption,  $\delta_1$  should remain indistinguishable from zero.

<sup>18</sup>The RK estimand,  $\tau_{\text{RK}}$ , can be written as the ratio of the reduced-form slope change (i.e., slope change in the outcome function) relative to the first-stage slope change (i.e., slope change in the treatment function)

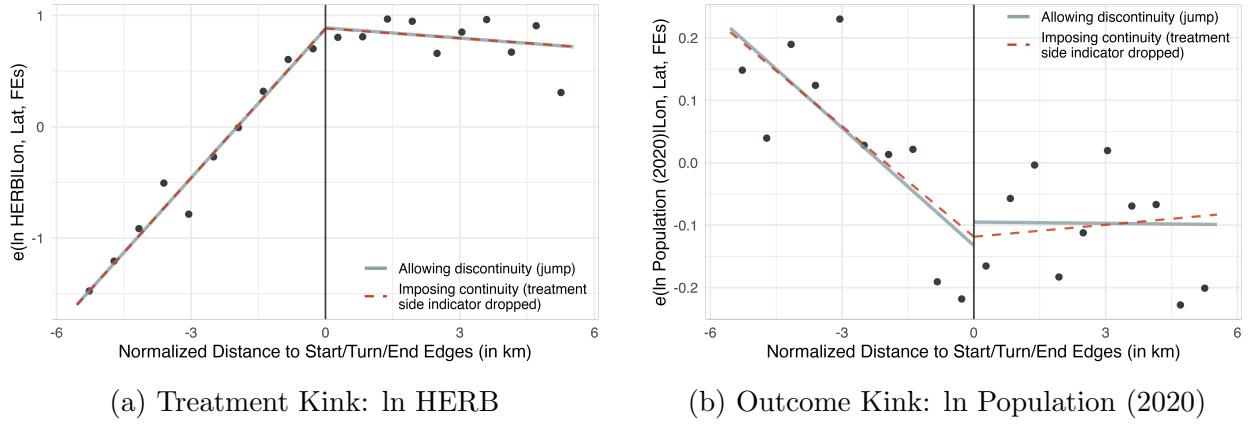


Figure 3: Slope Change in the Treatment and Outcome Functions at the Kink Point

*Notes:* Dots represent the binned frequencies, and solid and dashed lines show linear regression fits adjusting for longitude, latitude, and agent and district fixed effects. Solid line allows discontinuity (as equation 2), while the dashed line imposes continuity at the kink point by dropping the treatment-side dummy variable. The bin size is selected by the integrated mean squared error (MSE)-optimal mimicking variance evenly-spaced method using spacing estimators (Calonico, Cattaneo and Titiunik, 2015).

but initial empirical evidence of herbicide legacies on contemporary population outcome. In the following, we first report the cross-sectional RK estimates with population size as the outcome. We then extend the analysis with alternative growth rate and panel specifications across different bandwidth sizes, followed by a brief summary of robustness checks.

#### 4.1 Population Size Estimates

Table 1 reports the RK estimation results with population size in 2001 (Panel A) and 2020 (Panel B) as the outcome, along with the first-stage estimates. Model (1) is the baseline model without additional controls besides the two-dimensional polynomial of longitude and latitude and fixed effects. Models (2) to (4) consecutively add the proximity to the key targets, geographical characteristics, and historical attributes as covariates. To address potential spillover effects, Model (5) adjusts for the spatially-lagged treatment with the 30 at the kink point as (Card et al., 2012, 2015b):

$$\tau_{\text{RK}} = \frac{\left. \frac{\lim_{v_0 \downarrow 0} d(\mathbb{E}[Y|V=v])}{dv} \right|_{v=v_0} - \left. \frac{\lim_{v_0 \uparrow 0} d(\mathbb{E}[Y|V=v])}{dv} \right|_{v=v_0}}{\left. \frac{\lim_{v_0 \downarrow 0} d(\mathbb{E}[D|V=v])}{dv} \right|_{v=v_0} - \left. \frac{\lim_{v_0 \uparrow 0} d(\mathbb{E}[D|V=v])}{dv} \right|_{v=v_0}},$$

where  $V$  denotes the running variable,  $v_0$  the kink point, and  $D$  the treatment.

km neighbor cutoff in addition to the full set of covariates in model (4). Note that given the limitation in data availability and the historical facts, some of the covariates can partly be posttreatment (e.g., bombing point distance, 1965–1971). The models with covariate adjustments, therefore, do not necessarily provide conservative estimates due to possible posttreatment bias of unknown directions and should be interpreted with caution.

Table 1 confirms the graphical evidence of Figure 3 and underlines the herbicide legacies. Across model specifications, the first-stage, treatment kink estimates are substantively and statistically significant (Panel A). The second-stage coefficients on  $\ln \text{HERB}$  are also consistently signed negative and retain the statistical significance at the conventional 5% level (Panels B and C). Note also that uninstrumented ordinary least squares (OLS) estimates reported in Appendix Table A.2 substantively underestimate or even fail to reveal the negative association, suggesting that nonrandom herbicide assignment invites bias into naive comparisons. Moreover, the coefficient stability suggests that the estimates are unlikely to be driven by an arbitrary covariate adjustment choices, which is of a particular concern in RK applications (Ando, 2017) and we further investigate in Appendix C.1. The negative association also remains visible across the outcomes, population size in 2001 and 2020.

As an initial robustness check, Table 1 also reports three randomization inference results.<sup>19</sup> First, an important concern for RK applications is the misspecification of the underlying nonlinearity such that one falsely specifies a quadratic relationship with no kink as a discontinuous slope change (Ganong and Jäger, 2018). To guard against the misspecification bias, Table 1 reports kink point (KP) randomization  $p$ -value obtained from the reduced-form permutation test (Ganong and Jäger, 2018).<sup>20</sup> For the exercise, we first generate 10,000 placebo kink points randomly drawn from a uniform distribution,  $\mathcal{U}(-4 \text{ km}, 4 \text{ km})$ . We then consecutively estimate the reduced-form version of the RK model with the subsample within

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<sup>19</sup>Randomization exercise relies on the reduced-form and first-stage models as randomization inference cannot be applied without additional assumptions to unobserved subgroups such as compilers.

<sup>20</sup>If our RK specification correctly captures a discontinuous slope change, we would be unlikely to see a slope change with placebo kink points, and thus have a small KP randomization  $p$ -value. If, on the other hand, we falsely specify a quadratic function with no kink as a discontinuous slope change, we would observe discernible kink estimates with placebo kink points and a large KP randomization  $p$ -value.

Table 1: Herbicide Exposure and Population Size in 2001 and 2020

	Panel A: ln HERB				
	(1)	(2)	(3)	(4)	(5)
<b>First stage</b>					
EdgeDist $\times \mathbb{1}[\text{EdgeDist} \geq 0]$	-0.693*** (0.095)	-0.660*** (0.089)	-0.629*** (0.087)	-0.633*** (0.086)	-0.632*** (0.084)
F-statistic (weak instrument)	53.132	54.585	51.893	54.03	56.504
Adjusted R <sup>2</sup>	0.620	0.652	0.663	0.666	0.677
Running variable (RV) randomization <i>p</i> -value	0.000	0.000	0.000	0.000	0.000
Spatial noise (SN) randomization <i>p</i> -value	0.000	0.000	0.000	0.000	0.000
<b>Panel B: ln Population (2001)</b>					
	(1)	(2)	(3)	(4)	(5)
<b>Second stage</b>					
ln HERB	-0.203*** (0.069)	-0.177*** (0.065)	-0.168*** (0.063)	-0.165*** (0.063)	-0.165*** (0.064)
5% treatment increase effect size	-0.99%	-0.86%	-0.82%	-0.80%	-0.81%
Average outcome	1.315	1.315	1.315	1.315	1.315
<b>Reduced form</b>					
Kink point (KP) randomization <i>p</i> -value	0.019	0.021	0.004	0.003	0.000
RV randomization <i>p</i> -value	0.003	0.003	0.006	0.007	0.007
SN randomization <i>p</i> -value	0.004	0.004	0.005	0.006	0.006
<b>Panel C: ln Population (2020)</b>					
	(1)	(2)	(3)	(4)	(5)
<b>Second stage</b>					
ln HERB	-0.294*** (0.076)	-0.264*** (0.072)	-0.258*** (0.072)	-0.254*** (0.072)	-0.255*** (0.072)
5% treatment increase effect size	-1.43%	-1.29%	-1.26%	-1.24%	-1.24%
Average outcome	1.467	1.467	1.467	1.467	1.467
<b>Reduced form</b>					
KP randomization <i>p</i> -value	0.013	0.007	0.008	0.009	0.004
RV randomization <i>p</i> -value	0.000	0.000	0.000	0.000	0.001
SN randomization <i>p</i> -value	0.000	0.000	0.000	0.000	0.000
Observations	522	522	522	522	522
Avg. N neighbors (Conley SE cluster size)	27.4	27.4	27.4	27.4	27.4
Key target covariates	✓	✓	✓	✓	✓
Geographic covariates		✓	✓	✓	✓
Historical covariates			✓	✓	✓
ln spatially-lagged HERB				✓	✓
Fixed effects and <i>f</i> (Lon, Lat)	✓	✓	✓	✓	✓

Notes: \**p* < 0.1; \*\**p* < 0.05; \*\*\**p* < 0.01. Conley (1999) standard errors adjusted for spatial clustering with a 30 km cutoff and a Bartlett kernel are in parentheses. Key target covariates: NVA base distance, population (1967–1969), U.S. base distance, U.S. troop distance, rice cropland, road distance, slash and burn cropland, Viet Cong control prevalence. Geographic covariates: Precipitation, wind speed, elevation, flow accumulation, forest presence, rice suitability, river distance, ruggedness. Historical covariates: Bombing point distance, border distance, number of neighbor hamlets, railway distance. ln spatially-lagged HERB is the logged average HERB among the neighbor hamlets with a 30 km cutoff. Fixed effects: Agent fixed effect, district fixed effect, end-edge fixed effect, pre-1967 mission fixed effect. Randomization inference: KP randomization *p*-value is computed by the permutation test of Ganong and Jäger (2018) with 10,000 placebo kink points drawn from uniform distribution  $\mathcal{U}(-4 \text{ km}, 4 \text{ km})$ . RV randomization *p*-value is obtained by randomly assigning the running variable to the sample hamlets for 10,000 times. SN randomization *p*-value is obtained by 10,000 synthetic spatial noise simulations of Kelly (2021) with the outcome replaced by randomly generated spatial noise.

baseline 4 km bandwidth around each placebo kink point to obtain the empirical distribution of placebo estimates. The two-sided  $p$ -value is computed by doubling the minimum of the fraction of placebo estimates no smaller than or no greater than the actual reduced-form coefficient. Second, another intuitive approach of randomization inference is to randomly assign the running variable (RV) and the corresponding treatment-side indicator variable to the sample hamlets while holding the kink point fixed (e.g., Dell and Querubin, 2018, 30–33; see also, Cattaneo, Frandsen and Titiunik, 2015). The RV randomization two-sided  $p$ -values report the share of the absolute 10,000 placebo coefficients that are larger than the absolute actual reduced-form kink coefficient. Finally, to address potential spatial curve-fitting and residual spatial autocorrelation, we use the spatial noise (SN) randomization inference procedure developed by Kelly (2021). The SN randomization inference procedure replaces the outcome by synthetic noise with the same estimated spatial structure of the observed outcome partialled out by covariates. If the estimation is not an artifact of spatial trends, the observed variable should not explain spatial noise. The SN randomization  $p$ -value indicates the fraction of 10,000 spatial noise simulations which yield absolute  $t$  values greater than the absolute  $t$  value estimated with observed data.

For both treatment and outcomes, the randomization inference results are consistent with asymptotic inference. The three series of randomization  $p$ -values remain smaller than the conventional 5% threshold, suggesting that the actual estimates are unlikely to arise from misspecification of the underlying nonlinearity (KP randomization), random chance (RV randomization), or spatial autocorrelation and curve-fitting (SN randomization).

Turning to magnitude, the coefficient estimates can readily be interpreted as elasticity given the log-log specification. The estimates suggest that, measured at the 100 meter grid scale, a 5% increase in herbicide exposure is persistently followed by a 1.24% ( $-0.254 \times \ln(1.05) = 0.0124$ , Model 4, with all covariates) to 1.43% (Model 1, without covariates) decrease in population size in 2020 (Panel B), and a 0.80% to 0.99% decrease in 2001 (Models 1 and 4 in Panel C). Note that the 5% increase reference does not overstate

the magnitude of the treatment effect, as such an increase in herbicide exposure is observed within a geographically small area around the kink edges. Specifically, a 5% increase in the treatment approximately corresponds to the difference in  $\ln \text{HERB}$  between the average of the control hamlets within a 1 km distance from the kink edge with the average of the treated hamlets within the same distance from the kink edge ( $\frac{8.22}{7.802} = 1.054$ ).

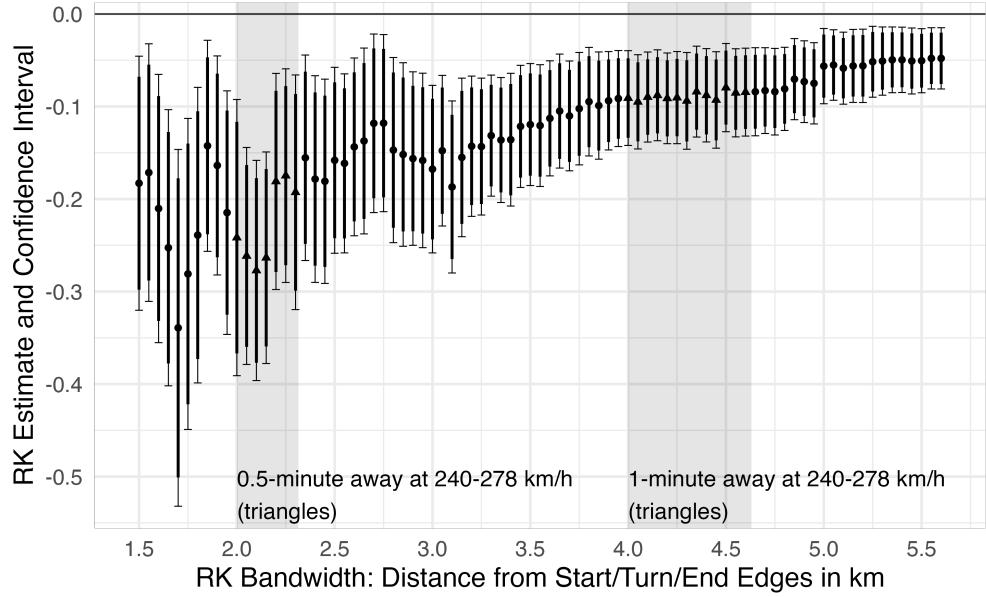
Besides the baseline population growth during the 2001–2020 period (from 1.315 to 1.467 in the logarithm scale, Table 1), the growing, rather than fading away, coefficient size implies that herbicide legacies shape not only population size but also population growth rate. This leads to a testable implication that a similar negative association is also present in local population growth rate, which contrasts the conventional wisdom of rapid recovery of population from wartime destruction.

## 4.2 Population Growth Rate Estimates

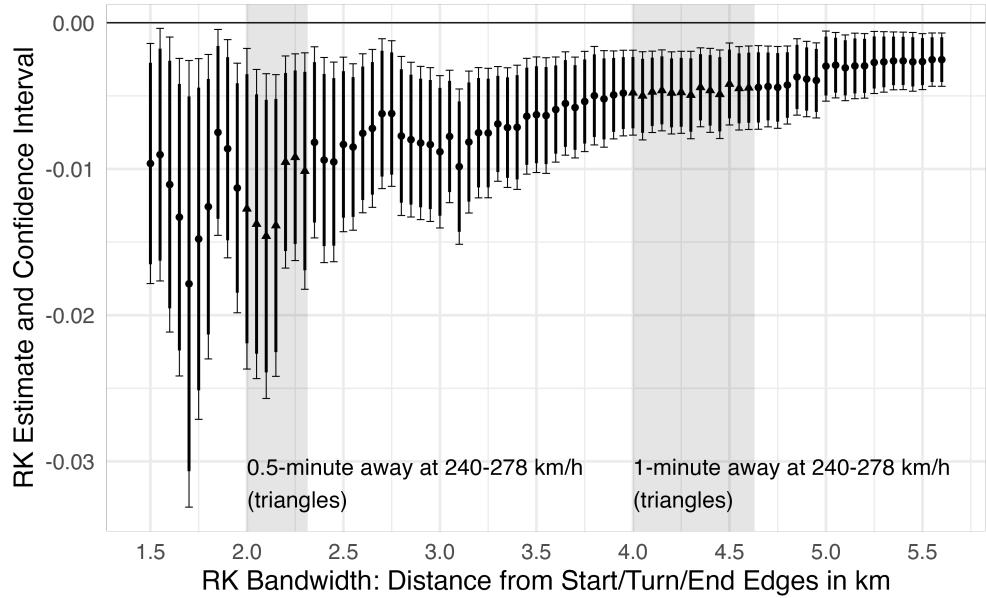
To examine the testable implication of cumulatively increasing legacies of herbicide exposure, we reestimate the RK models with the overall and annual population growth rates during the 2001–2020 period as the outcomes, using both cross-sectional and panel setups. This growth-rate specification also surmounts the robustness concern that a small number of volatile and less persistent population count observations in 2001 and 2020 drive the estimation results. For the panel specification, we extend the baseline cross-sectional specification by replacing the outcome by the annual growth rate and adding a year fixed effect as an additional control.<sup>21</sup> In the panel setup, we report robust standard errors clustered at the hamlet level. Another robustness issue remaining in Table 1 is that the results are based on a fixed bandwidth of 4 km, leaving potential concern for the bandwidth choice driving the results. To examine the herbicide legacies on population growth rate and address these robustness concerns, we report the RK estimates for overall and annual population growth rates across different RK bandwidths ranging from 1.5 km to 5.6 km with an increment of 50 meters.

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<sup>21</sup>The results remain qualitatively unchanged when further adjusting for the lagged outcome,  $Y_{hd,t-1}$ .



(a) Overall Population Growth Rate, 2001–2020



(b) Annual Population Growth Rate, 2002–2020

Figure 4: Herbicide Exposure and Population Growth Rate

*Notes:* Symbols and thin (thick) vertical segments with horizontal ticks represent the RK estimates and the corresponding 95% (90%) confidence intervals based on Conley (1999) standard errors with spatial clustering (Panel (a)) and robust standard errors with hamlet-level clustering (Panel (b)). The model specification in Panel (a) follows model (1) in Table 1. Panel (b) adds a year fixed effect to the baseline specification. Grey shades represent “0.5-minute-away” (2–2.315 km) and “one-minute-away” distances (4–4.63 km) from the kink point at the typical airspeed of Ranch Hand aircraft of 240–278 km/h, and triangles indicate the corresponding point estimates. Annual growth estimate in Panel (b) covers the 2002–2020 period as the outcome is defined as  $\ln \text{Population}_t - \ln \text{Population}_{t-1}$  and our dataset covers the 2001–2020 period.

Figure 4 presents the RK estimates for overall (Panel (a)) and annual (Panel (b)) population growth rates across bandwidth sizes, painting a picture that is consistent with the population size results. Regardless of the bandwidth settings and outcomes, the growth rate estimates reveal a negative association consistent with the population size estimates.<sup>22</sup> The effect size of the persistent association is also discernible given the average outcome of 0.151 (or  $e^{0.151} = 1.163$ , 16.3% overall population growth) and the geographically small level of outcome measurement of 100 meter grid. The cross-sectional RK estimate of  $\tau = -0.091$  with the baseline bandwidth of 4 km (a “one-minute-away” distance) suggests that a 5% increase in herbicide exposure translates into a 0.44% ( $-0.091 \times \ln(1.05) = -0.0044$ ) decrease in the overall (logged) population growth rate in 2001–2020 (Figure 4(a)). Given  $\ln \text{Population}_{2020} - \ln \text{Population}_{2001} \approx \frac{\text{Population}_{2020}}{\text{Population}_{2001}} - 1$ , a 0.44% decrease can approximately be interpreted as the effect size in the percentage point scale. The negative association is also visible in the panel specification with the annual population growth rate as the outcome (Figure 4(b)), revealing the herbicide legacies remaining in the dynamics as well as the snapshots of population outcomes in Vietnam in the present day.

### 4.3 Robustness Check and Sensitivity Analysis

Robustness and sensitivity concerns remaining in the main estimates include (1) the robustness to the parameter choice to construct the RK sample and the exposure score, HERB, (2) confounding kink in observed covariates and model dependence, and (3) unobserved confounding kink. Appendix C addresses these concerns by leveraging alternative combinations of flight buffer width and half-decay distance parameters, all possible  $2^N$  covariates =  $2^{20} = 1,048,576$  model specifications per outcome, a jackknife approach and the known historical differences in counterinsurgency strategies across four U.S. Corps Tac-

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<sup>22</sup>Figure 4 follows the specification of model (1) in Table 1 (without covariate adjustments). Note that, as reported in Appendix C.1, the growth rate specification is remarkably robust to different model specifications (different combinations of adjusted covariates) and exhibits little model dependence. The empirical distribution of the RK estimates of all possible  $2^N$  covariates =  $2^{20} = 1,048,576$  model specifications is almost normally distributed with the mean and median estimates of 0.091 (Appendix Figure C.1), which is identical to the point estimate reported in the text and Figure 4(a) ( $\hat{\tau} = 0.091$ ).

tical Zones in South Vietnam (see, e.g., Dell and Querubin, 2018), and a sensitivity analysis approach to quantify how severe unobserved confounding forces would need to be to eliminate the main estimates (Cinelli and Hazlett, 2020). Reassuringly, none of the robustness checks yield results that would invalidate or overturn the main findings.

## 5 Conclusion

This article combined the historical records of herbicide missions with a fuzzy RK design, and revealed the negative legacies of herbicidal warfare on contemporary population size and growth rate in Vietnam. The empirical analysis suggests that the temporary shock of herbicidal warfare left lasting, rather than temporary, effects on contemporary population outcomes. While the reported results tell us little about the underlying mechanisms, there are at least three distinct, although not mutually exclusive, explanations for the revealed persistent association, which also opens up pathways for future research. First, the negative association might reflect decreased birthrates and life expectancy. The reported negative association between herbicide exposure and health outcomes (e.g., Le, Pham and Polachek, 2022) makes the heavily sprayed areas persistently suffer lower birthrates and decreased life expectancy. Second, and relatedly, the lower population size and growth rates can arise from deteriorated agricultural productivity to sustain local population (e.g., Appau et al., 2021). Finally, domestic migration preferences and patterns can also shape population dynamics. Increased out-migration from the herbicide affected areas coupled with decreased in-migration into the damaged areas can also account for the lasting negative associations.

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# ONLINE APPENDIX FOR “NOT GONE WITH THE WIND: LONG-RUN IMPACTS OF HERBICIDAL WARFARE IN VIETNAM”

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## A Data Details

### A.1 Descriptive Statistics

Table A.1 reports the summary statistics of the variables used in the empirical analysis. Nonbinary variables excepting the running variable, Viet Cong control prevalence, and population growth rates are log-transformed.

### A.2 Stellman-National Academy of Sciences HERBS File

The key source of our empirical analysis is the Stellman-National Academy of Sciences (NAS) version of the Herbicide Report System (HERBS) file (S-NAS HERBS, Stellman et al., 2003*a,b*). We first web-scraped the records in the database from the Agent Orange Warehouse website<sup>1</sup> We then converted the geocoordinate information originally recorded in the military grid reference system (MGRS) format into the longitude-latitude format, and combined the converted geocoordinates with other mission- and flight-level information.

In addition to the geocoordinates of spray start, turn, end points, the database contains information about the mission dates, spray methods, and agents and gallons sprayed in individual missions.<sup>2</sup> The spray method is categorized into “Fixed-Wing Aircraft,” “Ground Spraying,” “Helicopter,” and “Unspecified” categories. As Stellman et al. (2003*b*, 323) note, a majority of herbicide missions was carried out by fixed-wing aircraft (C-123 airplanes), with 5,961 (65.2%) out of 9,141 missions recorded as aircraft missions. The remaining 3,180 missions include 2,108 (23.1%) missions by helicopters, 446 (4.9%) ground applications, and 626 (6.8%) missions with the method information remaining unspecified. As explained in the main text, our empirical analysis uses the fixed-wing aircraft missions by the US Air Force, which dispersed approximately 95% of all herbicides (Stellman et al., 2003*a*, 681–682; see also, Figure 1 in the main text).

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<sup>1</sup> Available at: <http://www.workerveteranhealth.org/milherbs/new/>, accessed August 29, 2021.

<sup>2</sup> For the aircraft navigation system in the 1960s and the development of the tactical air navigation system (TACAN) distance measuring equipment (DME) in the region, see, for example, Rowley (1975).

Table A.1: Descriptive Statistics

	Observations	Mean	SD	Median	IQR
<b>Panel A: Dependent Variables</b>					
Population (2001)	716	1.333	0.995	1.193	1.327
Population (2020)	716	1.484	1.109	1.287	1.324
Overall Population Growth Rate (2001–2020)	716	0.152	0.319	0.095	0.338
Annual Population Growth Rate (2001–2020)	13,604	0.008	0.104	0.003	0.074
<b>Panel B: Treatment and Assignment Variables</b>					
HERB ( $D^{\text{Half}} = 500\text{m}$ )	716	7.093	1.946	7.488	2.411
Edge Distance	716	-1.140	2.924	-1.418	4.660
$\mathbb{1}[\text{EdgeDist}_{ij} \geq 0]$	716	0.330	0.470	0.000	1.000
<b>Panel C: Covariates</b>					
<b>Key Target Covariates</b>					
North Vietnam Army Base Distance	716	2.758	0.730	2.844	1.059
U.S. Base Distance	716	2.996	0.820	3.149	1.065
U.S. Troop Distance	716	4.892	1.276	5.220	1.778
Population (1967–1969)	716	3.143	3.314	0.693	6.524
Rice Cropland	716	0.300	0.459	0.000	1.000
Road Distance	716	1.463	1.529	1.671	2.248
Slash and Burn Cropland	716	0.251	0.434	0.000	1.000
Viet Cong Control Prevalence	716	0.172	0.325	0.000	0.129
<b>Geographic Covariates</b>					
Elevation	716	4.319	1.454	3.850	2.918
Flow Accumulation	716	1.747	2.280	0.693	2.996
Forest Presence	716	0.522	0.500	1.000	1.000
Precipitation	716	7.464	0.200	7.446	0.315
Rice Suitability	716	1.619	1.322	1.099	3.258
River Distance	716	-0.199	1.336	-0.068	1.719
Ruggedness	716	1.072	0.786	0.693	0.916
Wind Speed	716	0.778	0.177	0.771	0.314
<b>Historical Covariates</b>					
Bombing Point Distance	716	-0.190	0.822	-0.168	1.215
Border Distance	716	4.330	0.828	4.554	0.798
Number of Neighbor Hamlets	716	2.997	0.879	3.045	1.153
Railway Distance	716	2.992	1.632	3.364	2.069
<b>Geocoordinates</b>					
Longitude	716	107.911	1.004	108.222	1.686
Latitude	716	13.177	2.081	13.610	3.510

Notes: SD = standard deviation, IQR = interquartile range.

### A.3 Running Variable Coding Procedure

The main text explains the coding procedure of the running variable with examples in which single flight path polygons cover individual hamlets. When multiple flight polygons cover single hamlets, we adopt group-specific rules to measure the kink edge distance as illustrated

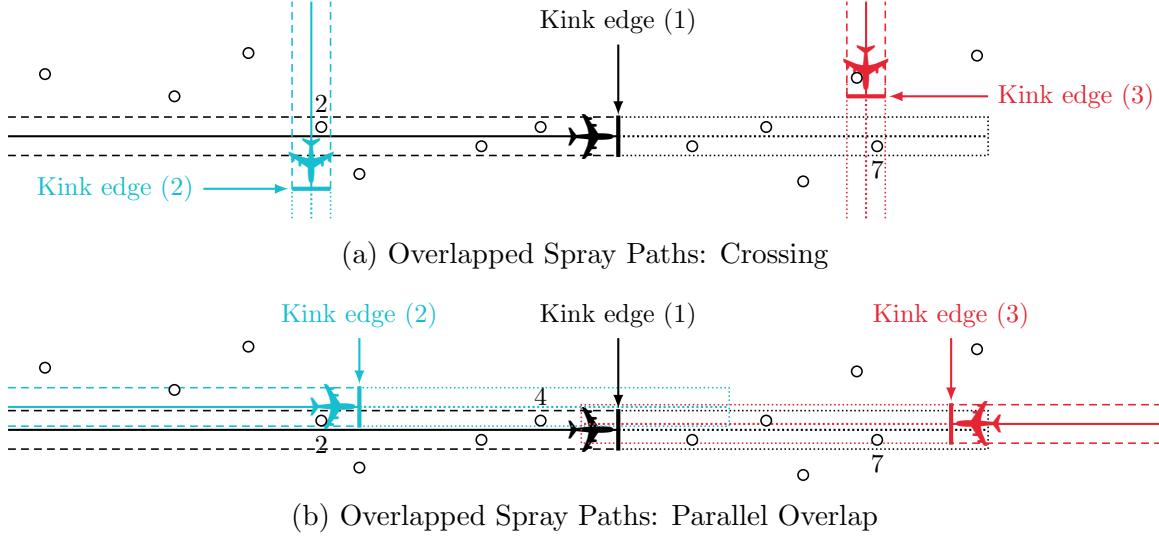


Figure A.1: Running Variable Coding with Overlapped Spray Paths

*Notes:* (a), (b) Dashed and dotted segments represent distinct spray path polygons. Hamlets 2 and 7 are geographically covered by multiple spray path polygons.

by Figure A.1. For the treatment group hamlets, we first take the minimum distance to the kink edges and then assign the maximum, instead of the minimum, of the distance measures as its running variable. For example, we assign the distance to kink edge (1) instead of the distance to kink edge (2) as the running variable for hamlet 2 (in the treatment group). For the control group hamlets, we first measure the distances to the kink edges of all flight polygons covering the hamlets, and then assign the minimum distance as its running variable. For example, the running variable for hamlet 7 (in the control group) in Figure A.1(a) is measured as the distance to kink edge (3) instead of the distance to kink edge (1).

We adopt these group-specific coding rules because the coding procedure for the control hamlets might underestimate the kink-edge distance for the treated hamlets. As illustrated by hamlet 2 in Figure A.1(a), when covered by multiple flight path polygons, a treated hamlet may be located in the middle of one flight polygon (e.g., 4 km from kink edge 1) while located close to the kink edges of another flight path polygon (e.g., 250 meters from kink edge 2). In this case, a simple minimum distance coding assigns 250 meters to hamlet 2 as the running variable. However, this coding significantly underestimates the hamlet's distance to kink edge because the hamlet is located in the middle of another flight path and

is strongly exposed to the treatment due to the mission flight with kink edge (1) regardless of its distance to kink edge (2). Our coding rule prevents this underestimation of the running variable and assigns 4 km instead of 250 meters to the treated hamlet.

Instead of the maximum-of-minimum distance rule for the treated hamlets, we rely on the minimum-of-minimum distance rule for the control group hamlets (e.g., hamlet 7). Because the control group hamlets are located outside of the flight polygons, their proximity to the kink edges can best be measured by the minimum distance. In the case of hamlet 7, we assign the distance to kink edge 3 as the running variable. Figure A.1(b) illustrates another example with multiple spray paths with parallel overlaps. The kink edge distances for hamlets 2 and 7 are measured in the same manner as illustrated with the case of Figure A.1(a) (i.e., distance to kink edge 1 for hamlet 2, and distance to kink edge 3 for hamlet 7).

## A.4 Historical Maps and Archival Sources

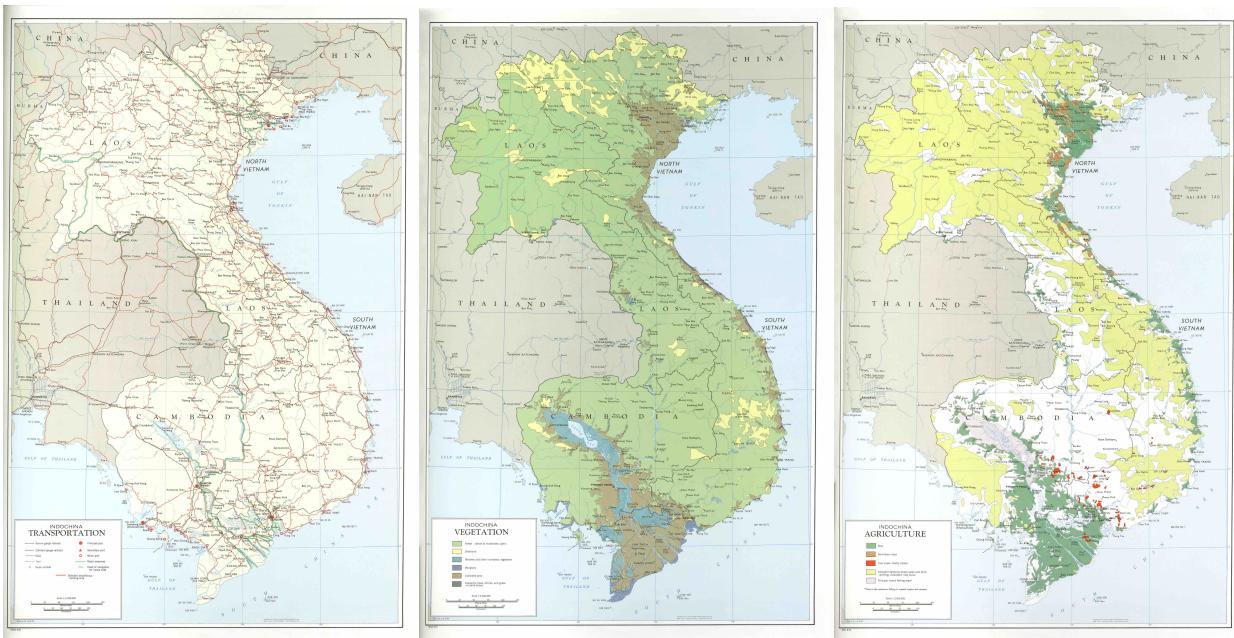
As explained in the main text, some of our covariates relies on historical maps and archival sources. As one of the key sources, we originally georeferenced and image-processed the historical maps of *Indochina Atlas* published by the Central Intelligence Agency (CIA) in 1970 and later digitized by the University of Texas Libraries.<sup>3</sup> As shown in Figures A.2(a) to A.2(c), the *Atlas* provides a series of maps as of the period including the transportation lines (i.e., roads, trails, and railways), surface configurations, and crop fields. Figures A.2(d) to (f) show the image-processed results for transportation lines, forest coverage, and croplands.<sup>4</sup>

We rely on several other sources to measure the historical landscape of military bases and troops. Suspected areas of North Vietnam Army (NVA) base are retrieved from the Enemy Base Area File (BASFA), July 1, 1967–July 1, 1971 hosted by the National Archives

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<sup>3</sup>The digitized (scanned) maps are available at [https://maps.lib.utexas.edu/maps/indochina\\_atlas/](https://maps.lib.utexas.edu/maps/indochina_atlas/) and <https://www.history.navy.mil/research/library/exhibits/maps/indochina-atlas-1970.html>.

<sup>4</sup>For image-processing, we primarily rely on the mean-shift segmentation algorithm implemented in Orfeo Toolbox (Grizonnet et al., 2017) and the skeletonization algorithm of Tveite (2015). We rely on the same image-processing procedures to generate the polygons of demarcation line and demilitarized zone (DMZ) from CIA’s “Vietnam demarcation line and demilitarized zone. 12-66” map as of 1966 (<https://www.loc.gov/resource/g8021f.ct002840/>, accessed September 22, 2021).



(a) Transportation Map

(b) Vegetation Map

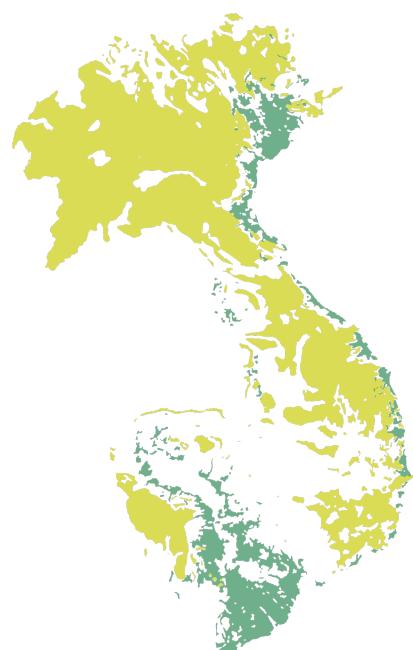
(c) Agriculture Map



(d) Road and Railway Network



(e) Forest Coverage



(f) Rice Croplands and Slash and Burn Cultivation Areas

Figure A.2: *Indochina Atlas* Maps and the Image-Processed Vector Objects

*Notes:* Panels (a), (b), and (c) depict the original (non-georeferenced) scanned maps of *Indochina Atlas*, digitized by the University of Texas Libraries. Panels (d), (e), and (f) show the generated vector objects (lines and polygons). Red and black segments in Panel (d), respectively, indicate roads (including trails) and railways. Green region in Panel A.2(e) represents forest presence. In Panel (f), green regions represent rice croplands, and yellow regions indicate the areas with slash and burn cultivation.

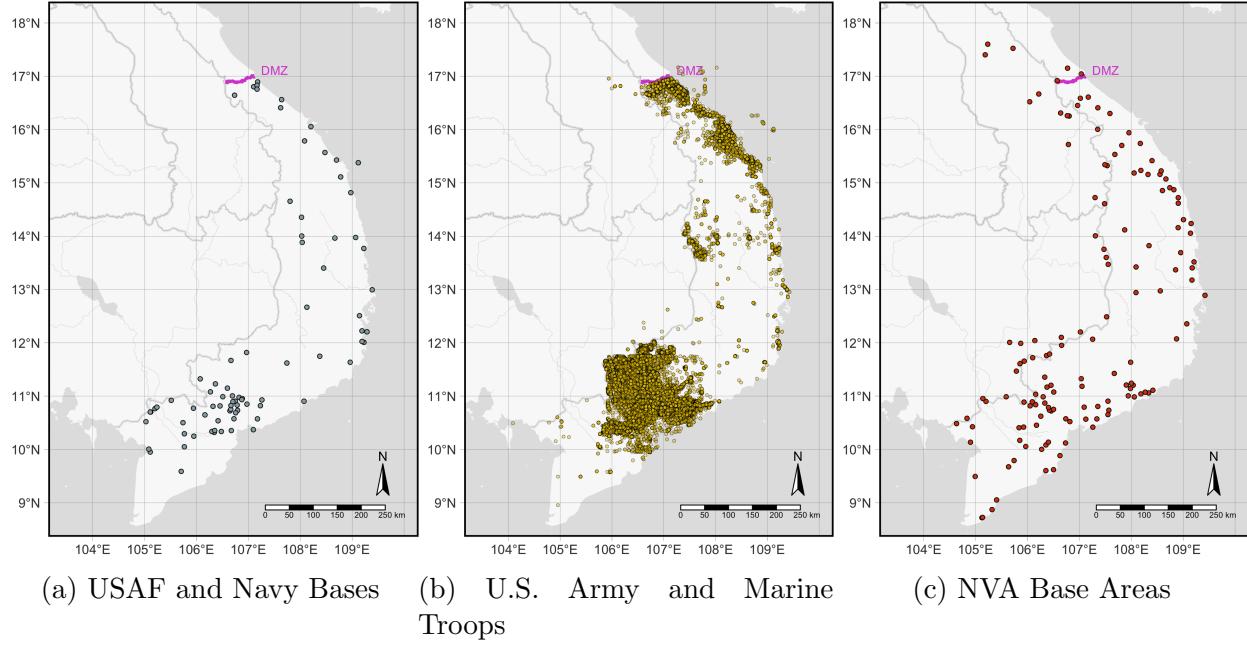


Figure A.3: Locations of U.S. Bases, Troops, and NVA Base Areas

*Notes:* Dots in each panel indicate, respectively, (a) U.S. Air Force and Navy bases, (b) U.S. Army and Marine troop locations, and (c) suspected areas of North Vietnam Army (NVA) base areas in and near South Vietnam. Dots in (c) represent the geocoordinates of the “approximate center” of the suspected NVA base areas (National Archives and Records Administration. 2007. “Reference Copy of Technical Documentation for Accessioned Electronic Records: Enemy Base Area File (BASFA), 7/1967–7/1971, Translation File, RG 330, Records of the Office of the Secretary of Defense.” pp.26–27. Available at: <https://catalog.archives.gov/id/2573252>, accessed June 16, 2023).

and Records Administration (NARA).<sup>5</sup> The locations of U.S. Air Force and Navy bases and U.S. Army and Marine troops are retrieved from the S-NAS HERBS database. Figure A.3 depicts the spatial distributions of U.S. bases, U.S. troop locations, and the suspected areas of NVA bases in and near South Vietnam.

## A.5 Naive OLS Estimates

Table A.2 reports the uninstrumented, naive ordinary least square (OLS) version of the main fuzzy regression kink (RK) specification with population size in 2001 and 2020 as outcomes. As briefly reported in the main text, compared with the main RK estimates, the naive OLS estimates underestimate or even fail to detect the negative association between herbicide

<sup>5</sup>National Archival Identifier 2573252; Accession Number of NN3-330-76-037. Available at: <https://catalog.archives.gov/id/2573252>, accessed June 16, 2023.

Table A.2: Herbicide Exposure and Contemporary Population Size, Naive OLS Estimates

	Panel A: ln Population (2001)				
	(1)	(2)	(3)	(4)	(5)
ln HERB	-0.059*	-0.050	-0.040	-0.046	-0.056*
	(0.034)	(0.030)	(0.028)	(0.028)	(0.029)
Adjusted R <sup>2</sup>	0.716	0.793	0.815	0.819	0.822
	Panel B: ln Population (2020)				
	(1)	(2)	(3)	(4)	(5)
ln HERB	-0.075*	-0.070**	-0.058*	-0.068**	-0.082**
	(0.038)	(0.035)	(0.032)	(0.033)	(0.034)
Adjusted R <sup>2</sup>	0.703	0.775	0.796	0.800	0.804
Observations	522	522	522	522	522
Avg. N neighbors (Conley SE cluster)	27.4	27.4	27.4	27.4	27.4
Key Target Covariates		✓	✓	✓	✓
Geographic Covariates			✓	✓	✓
Historical Covariates				✓	✓
ln Spatially-Lagged HERB					✓
Fixed effects and $f(\text{Lon}, \text{Lat})$	✓	✓	✓	✓	✓

Notes: \* $p < 0.1$ ; \*\* $p < 0.05$ ; \*\*\* $p < 0.01$ . Conley (1999) standard errors adjusted for spatial clustering with a 30 km cutoff and a Bartlett kernel are in parentheses. Key target covariates: NVA base distance, population (1967–1969), U.S. base distance, U.S. troop distance, rice cropland, road distance, slash and burn cropland, Viet Cong control prevalence. Geographic covariates: Precipitation, wind speed, elevation, flow accumulation, forest presence, rice suitability, river distance, ruggedness. Historical covariates: Bombing point distance, border distance, number of neighbor hamlets, railway distance. ln spatially-lagged HERB is the logged average HERB among the neighbor hamlets with a 30 km cutoff. Fixed effects: Agent fixed effect, district fixed effect, end-edge fixed effect, pre-1967 mission fixed effect.

exposure and contemporary population size revealed in Table 1.

The RK-OLS discrepancy may reflect (1) bias in the OLS estimates induced by confounding bias and measurement error, (2) bias in the RK estimates, (3) bias in both estimates, and (4) the difference in the estimands (average treatment effect, ATE, and the LAR or TT effect). Given the nonrandom nature of herbicide assignment, we mainly attribute the difference between the OLS and RK estimates to the bias remaining in the OLS estimates.

## B Identification Assumption

### B.1 Running Variable Distribution

A valid fuzzy RK design hinges on the smoothness assumption, which yields several testable implications (Card et al., 2015b). The first testable implication of the smoothness assumption is that the running variable distribution is continuous and differentiable at the cutoff point.

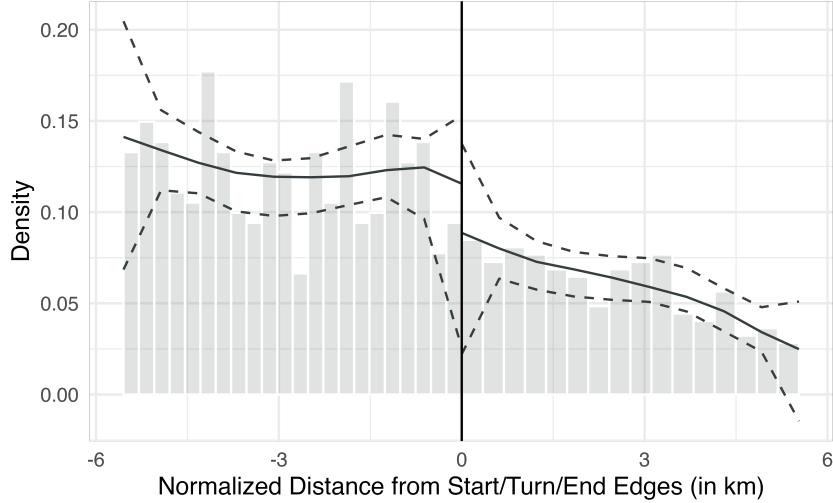


Figure B.1: Continuity of the Running Variable Distribution

*Notes:* Gray bars represent the histogram estimates of the running variable, EdgeDist. Solid and dashed lines indicate the local polynomial density estimates and the corresponding 95% confidence intervals (Discontinuity test:  $t = -0.209$  and  $p = 0.835$ ). The discontinuity estimate is computed using the procedure of Cattaneo, Jansson and Ma (2020), with the running variable recentered at zero.

tion requires a running variable continuously differentiable at the kink point. To empirically validate the implication, Figure B.1 examines the continuity of the running variable using the estimator of Cattaneo, Jansson and Ma (2020). Consistent with the identification assumption, the density test fails to detect discernible or statistically significant discontinuity in the running variable at the kink point ( $t = -0.209$  and  $p = 0.835$ ).

The smoothness assumption requires not only the absence of discontinuity but also the absence of a slope change in the running variable distribution at the kink point. To further validate the assumption, we follow the approaches of Bana, Bedard and Rossin-Slater (2020), Card et al. (2015a), Card et al. (2015b), and Landais (2015) and test for the kink in the running variable distribution via polynomial regressions. Formally, and with abuse of notation, we first aggregate the hamlet observations using bins with different sizes based on the running variable and then estimate the following regression model:

$$N_b^{\text{obs.}} = \beta \mathbb{1}[\overline{\text{EdgeDist}}_b \geq 0] + \sum_{p=1}^P \left[ \gamma_p \overline{\text{EdgeDist}}_b^p + \delta_p \overline{\text{EdgeDist}}_b^p \cdot \mathbb{1}[\overline{\text{EdgeDist}}_b \geq 0] \right] + e_b, \quad (\text{B.1})$$

where  $b$  indexes bins,  $N^{\text{obs.}}$  reflects the number of observations in each bin,  $\overline{\text{EdgeDist}}$  is the

midpoint of EdgeDist of each bin, and  $P$  is the polynomial order. The coefficient on the interaction between the linear of the running variable and the treatment group indicator,  $\delta_1$ , captures the change in the slope of the probability density function of the running variable at the kink point. We repeatedly estimate the polynomial regression using different bin widths ranging from 50 meters to 500 meters with an increment of 50 meters and the running variable polynomial orders ranging from 2 to 6.

Figure B.2 summarizes the polynomial regression results by plotting the coefficient estimates on the interaction term between the treatment group indicator and the (linear) running variable ( $\delta_1$ ). Triangles in each panel and Panel (f) indicate the estimates obtained from the smallest Akaike Information Criterion (AIC) value for each bin size. Regardless of the bin size and polynomial order, the coefficient estimates remain statistically insignificant at the 5% level. The polynomial regressions reveal no discernible slope change in the running variable distribution at the kink point and fail to invalidate the current RK strategy.

## B.2 Covariate Distribution

Another key testable implication of the smoothness assumption involves covariate distributions around the kink point. The validity of the current RK design would be undermined if predetermined covariates, along the treatment, exhibit discernible kink around the cutoff.<sup>6</sup> We examine the validity of the assumption by subsequently estimating the first-stage specification of equation (2) with the left-hand-side variable,  $\ln \text{HERB}$ , replaced by one of the covariates. We estimate the placebo kink estimate with and without adjustments for the remaining covariates, with the baseline bandwidth  $b = 4$  km as in Table 1 in the main text. The smoothness assumption requires that the coefficient on  $\text{EdgeDist} \times \mathbb{1}[\text{EdgeDist} \geq 0]$  remains indistinguishable from zero once we replace the left-hand-side variable with covariates.

Figure B.3 summarizes the (standardized) kink estimates with (hollow symbols) and

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<sup>6</sup>Seen from an instrumental variable (IV) perspective, a key assumption here is that the instrument,  $\text{EdgeDist} \times \mathbb{1}[\text{EdgeDist} \geq 0]$ , affects the outcome (slope change) only through the treatment (slope change). A kink in a covariate can also be viewed as a sign of exclusion restriction violation such that the instrument affects the outcome through the covariate, not exclusively through the treatment.

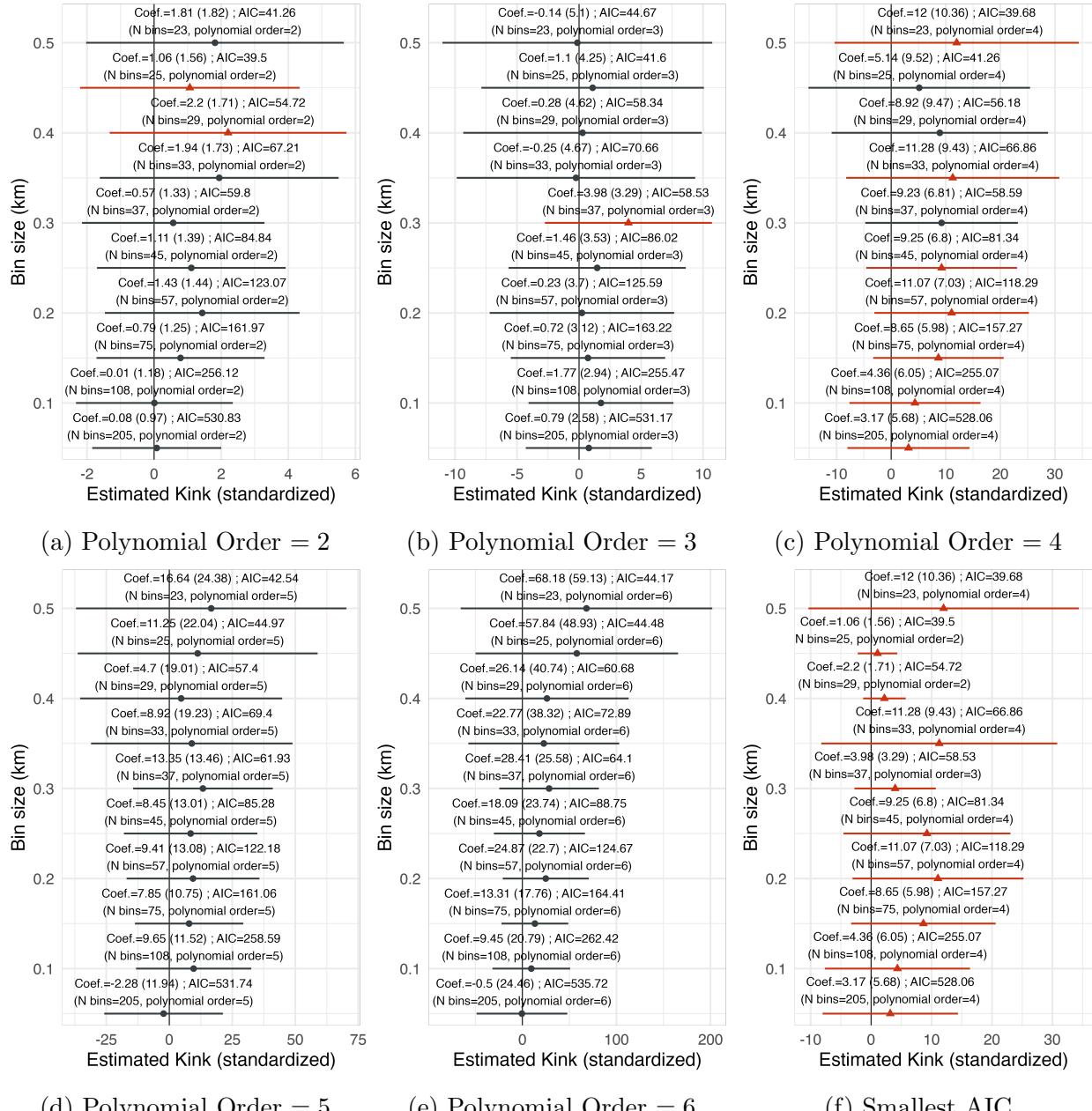


Figure B.2: Kink Estimates in the Running Variable Distribution

*Notes:* Dots represent the kink estimates obtained from polynomial regressions ( $\delta_1$  in equation B.1), with the running variable polynomial order specified as in the figure labels and bin sizes on the vertical axis. Triangles represent the estimates obtained from the model with the smallest Akaike Information Criterion (AIC) for each bin size. Horizontal segments represent the 95% confidence intervals.

without adjustments for the remaining covariates (solid symbols), along with the treatment kink estimates (top row triangles). In a sharp contrast with the treatment slope change, and consistent with the identification assumption, most of the placebo regressions fails to detect

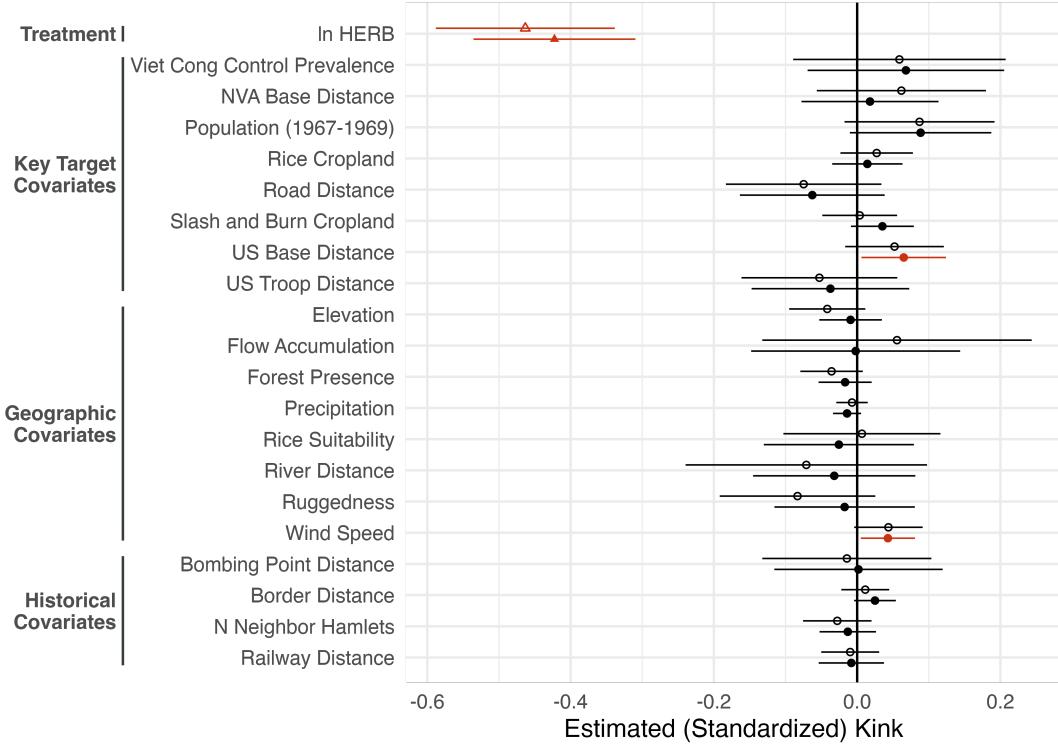


Figure B.3: (Standardized) Covariate and Treatment Kink Estimates

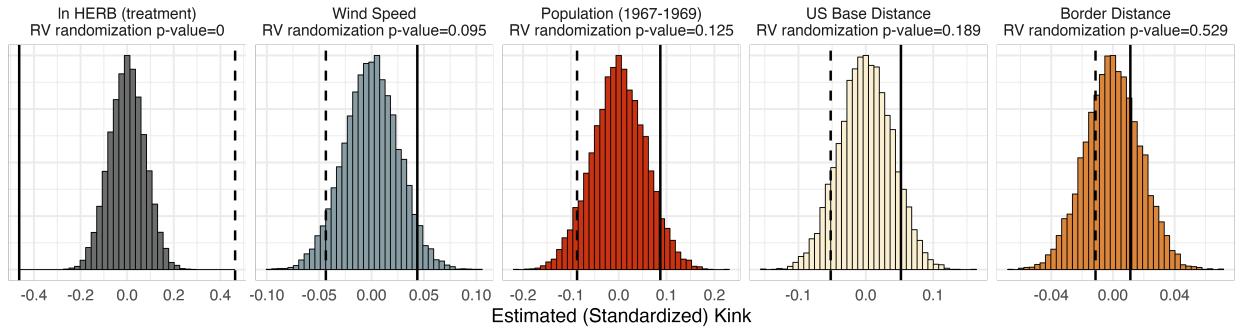
*Notes:* Symbols represent the reduced-form kink estimates with the variable on the vertical axis as the outcome variables, with (solid) and without adjustment for the remaining covariates (hollow). Horizontal segments represent 95% confidence intervals based on Conley (1999) standard errors with a 30 km cutoff and a Bartlett kernel (not corrected for multiple testing). All nonbinary variables are standardized.

a discernible slope change around the kink point. Exceptions include the substantively small but statistically significant estimates for wind speed (coef. = 0.043,  $t = 2.299$ ) and U.S. base distance (coef. = 0.065,  $t = 2.166$ ), along with the marginally significant estimates in population in 1967–1969 (coef. = 0.089,  $t = 1.764$ ) and border distance (coef. = 0.025,  $t = 1.681$ ), when adjusted for covariates (solid symbols).<sup>7</sup>

Nonetheless, the running variable (RV) randomization exercise disagrees with asymptotic inference and fails to negate that the covariate kink are generated by chance. As in Table 1, we randomly assign the running variable to the sample hamlets and estimate the placebo kink coefficient for 10,000 times. The randomization two-sided  $p$ -value indicates the share of the absolute placebo coefficients that are larger than the absolute actual coefficient. Figure B.4

<sup>7</sup>Given one treatment and 20 covariates (21 variables) in Figure B.3, the probability of a Type I error in which one falsely rejects at least one null hypothesis with  $\alpha = 0.05$  is 65.9% ( $1 - 0.95^{21} = 0.659$ ).

### Without Covariate Adjustment



### With Covariate Adjustment

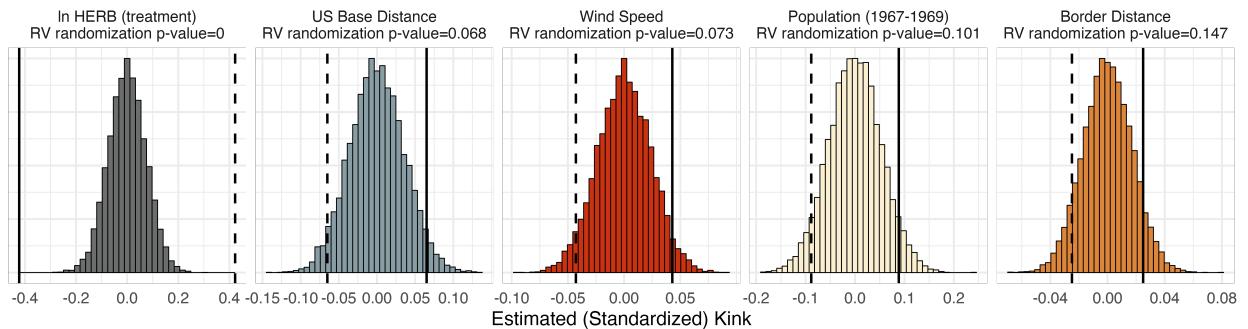


Figure B.4: Running Variable Randomization Inference

*Notes:* Each panel shows the histogram of 10,000 placebo kink estimates for the outcome in the top label. We randomly assign (permute) the running variable to the sample hamlets, and then estimate the reduced-form (or first-stage) version of the RK specification with (bottom row panels) and without covariate adjustments (top row panels) for 10,000 times. Solid (dashed) segments indicate the actual kink estimates ( $-1 \times$  actual kink estimate) with the observed running variable. The randomization (two-sided) *p*-values in the top labels indicate the share of the absolute placebo coefficients that are larger than the absolute actual coefficient (vertical lines).

shows the histograms of the placebo (standardized) kink estimates, with the corresponding actual kink estimates (vertical lines). The corresponding randomization *p*-values are 0.073 (wind speed), 0.068 (U.S. base distance), 0.101 (population, 1967–1969), and 0.147 (border distance) with covariate adjustments, and 0.095 (wind speed), 0.189 (U.S. base distance), 0.125 (population, 1967–1969), and 0.529 (border distance) without covariate adjustments.

A natural and substantively important question is on how the covariate kink affects the main RK estimates. A focused exercise in Appendix C, however, suggests that the RK estimates remain stable across all possible model specifications (covariate combinations) and outcomes, and reveals little sign of model dependence. Along with the little evidence of model dependence, the covariate kink test also fails to invalidate the current RK design.

## C Robustness Check

### C.1 Confounding Nonlinearity and Model Dependence

One may wonder how the kink wind speed and U.S. base distance and model specification influence the main findings. Related to this point, Ando (2017) underlines the potential susceptibility of RK estimates to the failure to adjust for confounding nonlinearity. A simple, crude way to address this concern for confounding nonlinearity is to estimate all possible model specifications. The key idea is that if the current RK design suffers confounding kink in observed outcomes, the RK estimate should be sensitive to the choice of covariate adjustments. Specifically, we repeatedly estimate the RK model with each covariate combination and three cross-sectional outcomes.<sup>8</sup> Given that our RK specification includes 20 covariates along with the fixed effects and the longitude-latitude polynomial, the number of possible model specifications amounts to  $2^{20} = 1,048,576$  for each outcome (omitting the spatially-lagged treatment and interaction terms). We estimate each model specification using the sample hamlet with the baseline bandwidth of  $b = 4$  km as in Table 1 in the main text.

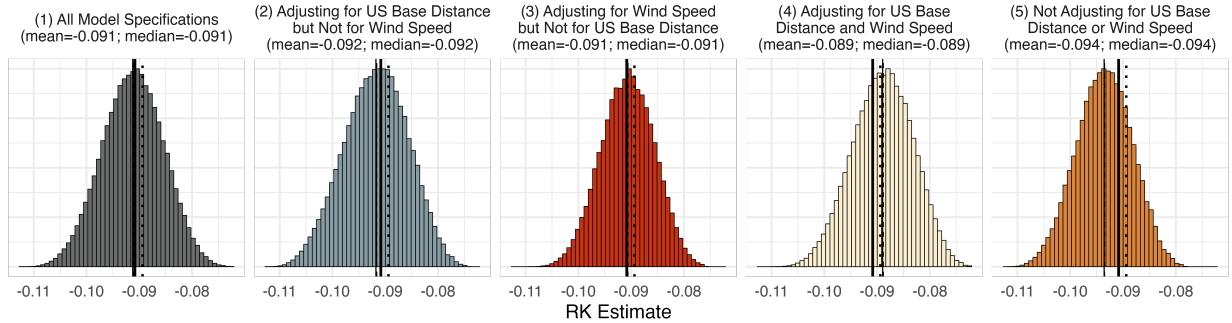
Figure C.1 summarizes the result of the “try-all” exercise. To explicitly examine how the adjustment for U.S. base distance and wind speed affect the estimates, Figure C.1 presents density histograms for RK estimates of, from the left, (1) all possible 1,048,576 models, (2)  $2^{18} = 262,144$  models adjusting for U.S. base distance but not for wind speed, (3) 262,144 models adjusting for wind speed but not for U.S. base distance, (4) 262,144 models adjusting for both U.S. base distance and wind speed, and (5) 262,144 models not adjusting for U.S. base distance or wind speed for each outcome. Thin solid and dashed vertical segments indicate the corresponding mean and median estimates, and bold solid and dotted segments indicate the RK estimates corresponding to Model 1 (without covariates) and Model 4 (adjusting for all covariates) in Table 1 in the main text.

The key results are threefold. First, across the outcomes, the empirical distribution of

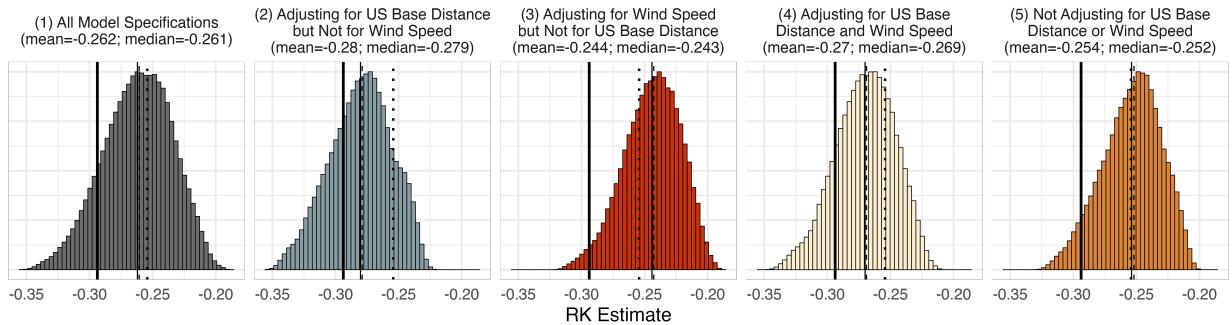
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<sup>8</sup>This exercise relies on the cross-sectional setup given the qualitative similarity of annual (panel) and overall (cross-sectional) population growth rate estimates in Figure 4 in the main text.

### Outcome: Overall Population Growth Rate (2001-2020)



### Outcome: In Population (2020)



### Outcome: In Population (2001)

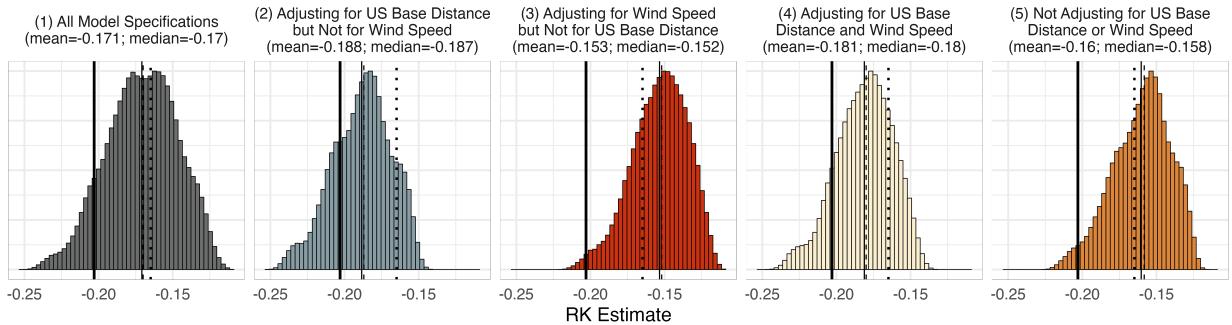


Figure C.1: Empirical Distribution of Second-Stage RK Estimates, All Possible Specifications

*Notes:* Density histograms in each row plot the empirical distributions of (subsets of)  $2^{20} = 1,048,576$  RK model specifications for three cross-section outcomes, with different covariate adjustment conditions denoted in the top text labels. Thin solid and dashed vertical segments indicate the mean and median estimates in each panel. Bold solid and dotted vertical segments represent the RK estimates of model 1 (without covariates, solid) and model 4 (with all covariates, dotted) in Table 1 in the main text.

RK estimates is close to a normal distribution (with slight skews in population size in 2001 and 2020), suggesting that covariate combination does not play a major role in altering the RK estimates beyond random error. Second, the adjustment for U.S. base distance and wind speed is unlikely to alter the RK estimate. As graphically illustrated by the similarity between the panels in each row, the RK estimate distribution remains almost unchanged

regardless of covariate adjustments, with almost identical sample mean (solid vertical segment) and median estimates (dashed segment). Third, the baseline RK estimates in Table 1 and Figure 4 in the main text (bold solid and dotted segments) are close to the means and medians of the empirical coefficient distributions (leftmost panels). Models with the full set of covariates (dotted segments) are almost identical to the mean and median estimates, and models without covariates (bold solid segments) slightly underestimate (population growth rate, Figure 4 in the main text) or overestimate (population in 2001 and 2020, Table 1 in the main text) while some of the covariates can be seen as posttreatment or a “bad control.” Combined, these results suggest that the covariate kink and the choice of covariate adjustments are unlikely to affect the inference severely, and the main findings are unlikely to be a product of arbitrary model picking or (observed) confounding nonlinearity.

## C.2 Flight Buffer Width and Half-Decay Distance

Admittedly, the main RK estimates rely on an arbitrary combination of two parameters, flight buffer width and half-decay distance, to obtain the RK sample and measure the hamlet-level herbicide exposure score, HERB. As explained in Section 3 of the main text and Section A.3, flight buffer width defines the sample hamlets used in the RK estimation, and the half-decay distance parameter determines HERB (eq. 1 in the main text).

To examine the robustness to the particular parameter setting, Figure C.2 reestimates the RK model with alternative parameter combinations. While the coefficient size varies, the negative association between herbicide exposure and contemporary population outcomes remains robust across different RK samples and distance-decay parameter values.

## C.3 Counterinsurgency Strategy and Jackknife Estimates

A remaining but important concern for the current RK design is that the results might be driven by *unobserved* confounding kink such as counterinsurgency strategies and authorization procedure of herbicide missions. During the Vietnam War, the U.S. military divided

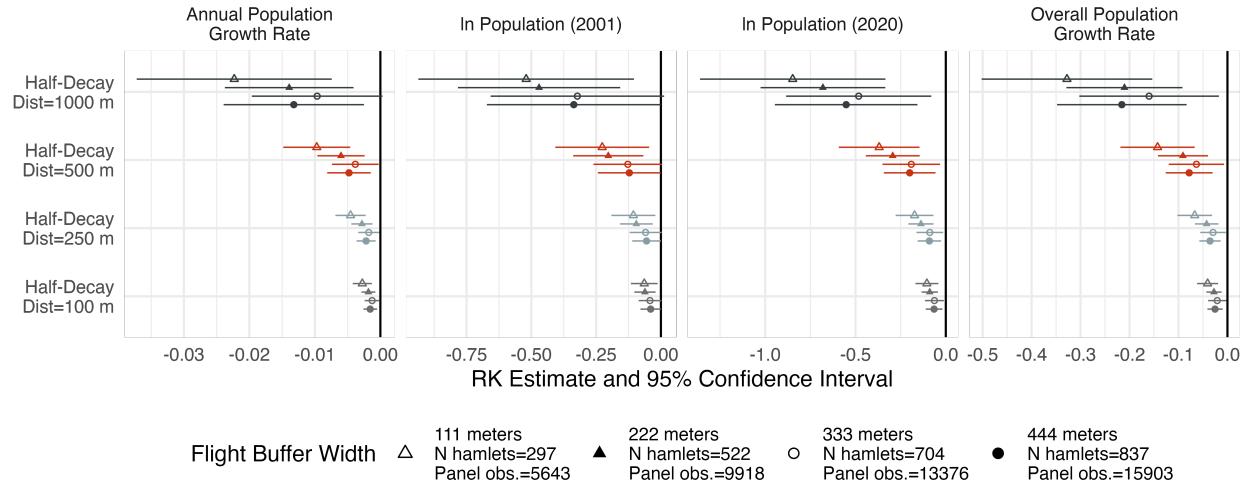


Figure C.2: RK Estimates with Different Flight Buffer Width and Half-Decay Distance Settings  
*Notes:* Symbols represent the second-stage RK estimates with different flight buffer widths (indicated by symbols) and the half-decay distance parameter on the vertical axis, for each of the outcome variables in the column labels. Horizontal segments display the 95% confidence intervals based on Conley (1999) standard errors with a 30 km cutoff and a Bartlet kernel for the cross-sectional specification and hamlet-level clustering for the panel setup. Red solid triangles with half-decay distance = 500 m indicate the baseline estimates reported in Table 1 and Figure 4 in the main text. All estimates rely on the baseline bandwidth of 4 km and the model specification of model (1) in Table 1. Panel specification further adjusts for a year fixed effect.

South Vietnam into four Corps Tactical Zones (CTZs) commanded by different military branches and implemented distinct counterinsurgency strategies in individual CTZs. Dell and Querubin (2018, 51–54) exploit the CTZ I–II boundary as one of the sources for causal identification, given the historical fact that the U.S. Marine Corps (USMC) in Corps Region I implemented a moderate hearts-and-minds-oriented strategy and small-unit operations, while the U.S. Army commanded Corps Region II and emphasized overwhelming firepower and large-scale operations. Similarly, as briefly explained in the main text, the authorization process of herbicide missions involved the South Vietnam Government, both at the national and local levels, and the U.S. at the levels of the U.S. Ambassador, the commander of the U.S. Military Assistance Command, Vietnam (MACV), and the CTZ commanders (Buckingham, 1982, 37; Institute of Medicine, 1995, 86).

Although the baseline specification includes a district fixed effect, we rely on a CTZ-level jackknife approach to gauge the impact of CTZs in our RK sample have on the reported

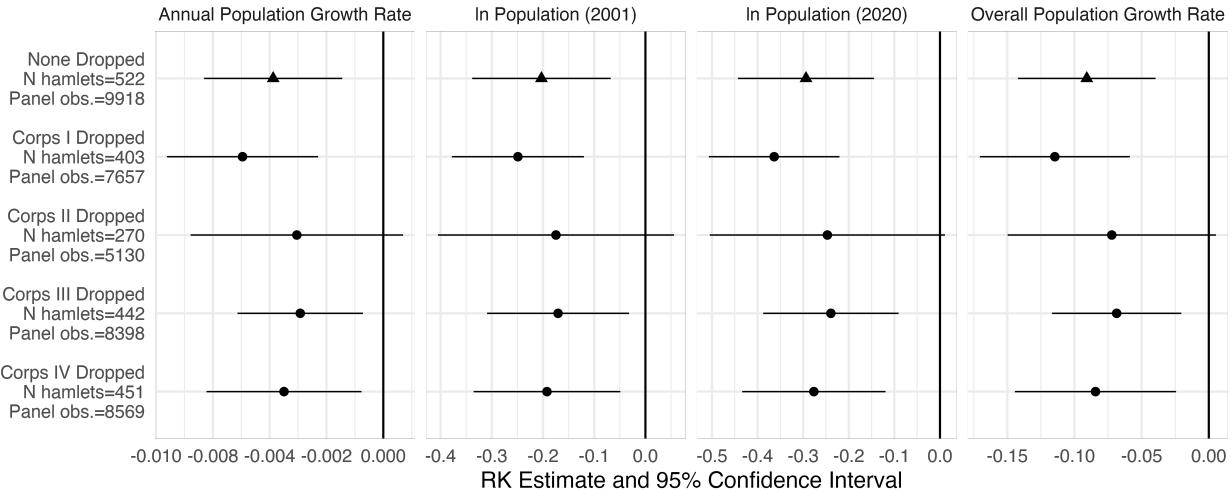


Figure C.3: Corps Tactical Zone-Level Jackknife

*Notes:* Symbols represent the RK estimates with subsamples dropping the hamlets in the Corps Tactical Zone on the vertical axis. Horizontal segments display the corresponding 95% confidence intervals based on standard errors with a 30 km cutoff and a Bartlet kernel for the cross-sectional specification and hamlet-level clustering for the panel setup. All estimates rely on the baseline bandwidth of 4 km and the specification of model (1) in Table 1. Panel specification further adjusts for a year fixed effect.

RK estimates. Figure C.3 reports the jackknife estimates using a subsample dropping the hamlet observations located in the CTZ on the vertical axis. As the CTZ boundaries follow the preexisting provincial boundaries, and given the role of province chiefs in Ranch Hand missions, Figure C.4 presents a similar jackknife test in which hamlet observations in the province on the vertical axis dropped from the subsample for RK estimate. While the sample size and statistical significance vary, the negative coefficient estimates remain qualitatively unchanged across subsamples.

## C.4 Unobserved Confounding Forces and Sensitivity Analysis

We also use a sensitivity analysis approach to examine how sensitive the main findings are to unobserved confounding nonlinearity. Table C.1 reports the reduced-form and first-stage versions of the main RK specification in Table 1 in the main text. The robustness values in each panel indicate the percentage of residual variance that unobserved confounders would need to explain to eliminate the kink estimates (Cinelli and Hazlett, 2020). For both

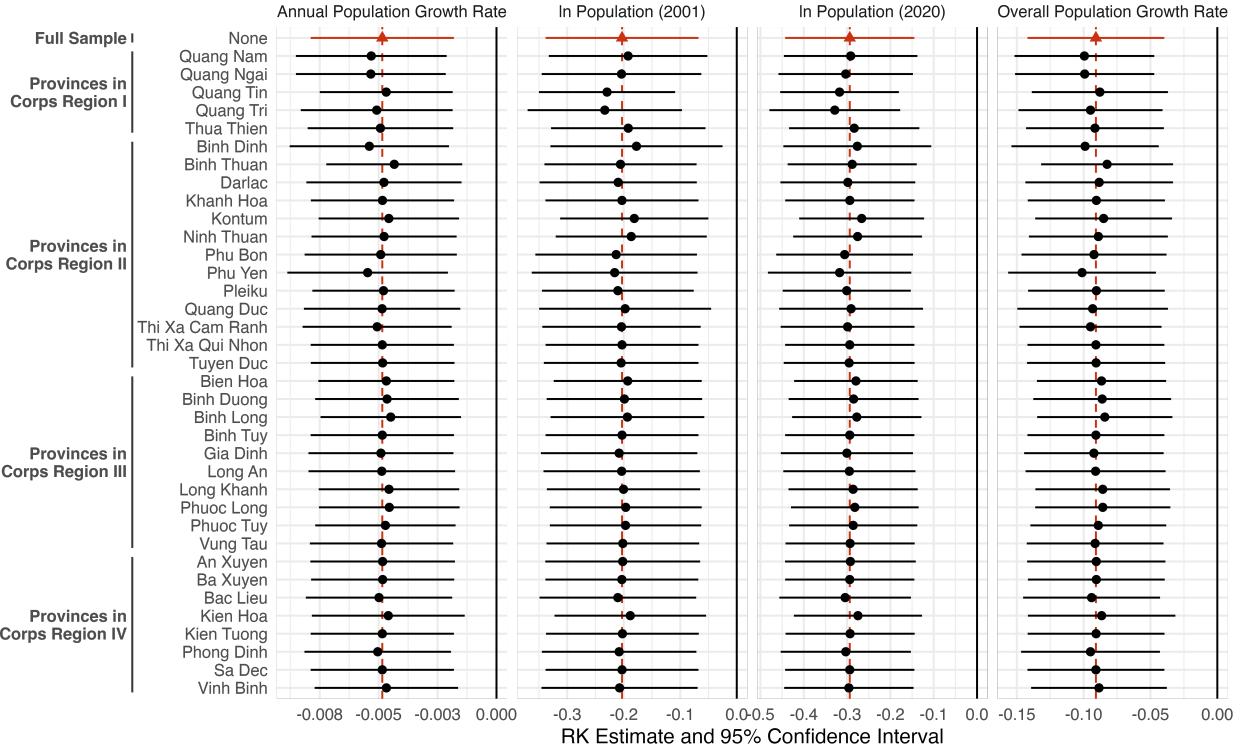


Figure C.4: Province-Level Jackknife

*Notes:* Symbols represent the RK estimates with subsamples dropping the hamlets in the province on the vertical axis. Horizontal segments display the corresponding 95% confidence intervals based on standard errors with a 30 km cutoff and a Bartlet kernel for the cross-sectional specification and hamlet-level clustering for the panel setup. Triangles and vertical dashed segments indicate the baseline estimate in Table 1 and Figure 4 in the main text. All estimates rely on the baseline bandwidth of 4 km and the model specification of model (1) in Table 1. Panel specification further adjusts for a year fixed effect.

reduced-form and first-stage estimates, the robustness values suggest only implausibly strong unobserved confounders are capable of eliminating the main RK estimates.<sup>9</sup> More precisely, the robustness values indicate that unobserved confounders would need to explain at least 16.59% (Model 4, Panel A) and 27.47% (Model 4, Panel B) of instrument ( $\text{EdgeDist} \times \mathbb{1}[\text{EdgeDist} \geq 0]$ ) and left-hand-side variable (population 2020 in Panel A and ln HERB in

<sup>9</sup>We report reduced-form and first-stage estimates as the robustness value measures the strength of unobserved confounding by partial  $R^2$ . Note also that the RK estimand,  $\tau_{RK}$ , can be written as the ratio of the reduced-form slope change (i.e., outcome slope change at the kink point) relative to the first-stage slope change (i.e., treatment slope change at the kink point) as (Card et al., 2015b):

$$\tau_{RK} = \left( \frac{\lim_{v_0 \downarrow 0} d(\mathbb{E}[Y|V=v])}{dv} \Big|_{v=v_0} - \frac{\lim_{v_0 \uparrow 0} d(\mathbb{E}[Y|V=v])}{dv} \Big|_{v=v_0} \right) / \left( \frac{\lim_{v_0 \downarrow 0} d(\mathbb{E}[D|V=v])}{dv} \Big|_{v=v_0} - \frac{\lim_{v_0 \uparrow 0} d(\mathbb{E}[D|V=v])}{dv} \Big|_{v=v_0} \right),$$

where  $V$  denotes the running variable,  $v_0$  the kink point, and  $D$  the treatment. The reduced-form (first-stage) robustness value can be interpreted as a sensitivity measurement for the numerator (denominator).

Table C.1: Reduced-Form and First-Stage Estimates with Robustness Values

	Panel A: ln Population (2020)				
	(1)	(2)	(3)	(4)	(5)
EdgeDist $\times \mathbb{1}[\text{EdgeDist} \geq 0]$	0.204*** (0.053)	0.174*** (0.048)	0.162*** (0.045)	0.161*** (0.045)	0.161*** (0.045)
Robustness Value	17.79%	17.02%	16.85%	16.84%	16.88%
Adjusted R <sup>2</sup>	0.708	0.778	0.800	0.802	0.805
	Panel B: ln HERB				
	(1)	(2)	(3)	(4)	(5)
EdgeDist $\times \mathbb{1}[\text{EdgeDist} \geq 0]$	-0.693*** (0.095)	-0.660*** (0.089)	-0.629*** (0.087)	-0.633*** (0.086)	-0.632*** (0.084)
Robustness Value	30.75%	31.36%	30.99%	31.64%	32.25%
Adjusted R <sup>2</sup>	0.620	0.652	0.663	0.666	0.677
Observations	522	522	522	522	522
Avg. N neighbors (Conley SE cluster size)	27.4	27.4	27.4	27.4	27.4
Key Target covariates		✓	✓	✓	✓
Geographic covariates			✓	✓	✓
Historical covariates				✓	✓
ln spatially-lagged HERB					✓
Fixed effects and $f(\text{Lon}, \text{Lat})$	✓	✓	✓	✓	✓

Notes: \* $p < 0.1$ ; \*\* $p < 0.05$ ; \*\*\* $p < 0.01$ . Conley (1999) standard errors adjusted for spatial clustering with a 30 km cutoff and a Bartlett kernel are in parentheses. Key target covariates: NVA base distance, population (1967–1969), U.S. base distance, U.S. troop distance, rice cropland, road distance, slash and burn cropland, Viet Cong control prevalence. Geographic covariates: Precipitation, wind speed, elevation, flow accumulation, forest presence, rice suitability, river distance, ruggedness. Historical covariates: Bombing point distance, border distance, number of neighbor hamlets, railway distance. ln spatially-lagged HERB is the logged average HERB among the neighbor hamlets with a 30 km cutoff. Fixed effects: Agent fixed effect, district fixed effect, end-edge fixed effect, pre-1967 mission fixed effect. Robustness values indicate the percentage of residual variance that unobserved confounders would need to explain in order to eliminate the reported kink coefficients.

Panel B) residual variance to eliminate the RK estimates in Table 1 in the main text. Given the battery of covariate adjustments in addition to the geocoordinate polynomial and fixed effects, it is implausible that a (combination of) confounding kink is driving the finding.

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