EI SEVIER

Contents lists available at ScienceDirect

## **Ecological Economics**

journal homepage: www.elsevier.com/locate/ecolecon



## **Analysis**

# Common property, information, and cooperation: Commercial fishing in the Bering Sea

Alan C. Haynie a, Robert L. Hicks b, Kurt E. Schnier c,\*,1

- <sup>a</sup> Alaska Fisheries Science Center, National Marine Fisheries Service, 7600 Sand Point Way NE, Bldg. 4, Seattle WA 98115, United States
- <sup>b</sup> Department of Economics, 124 Morton Hall, College of William and Mary, Williamsburg, VA 23187, United States
- <sup>c</sup> Department of Economics, Andrew Young School of Policy Studies, Georgia State University, 14 Marietta Street NW, Suite 436, Atlanta GA 30303, United States

## ARTICLE INFO

#### Article history: Received 5 November 2008 Received in revised form 9 June 2009 Accepted 15 August 2009 Available online 14 October 2009

Keywords: Cooperative games Spatial econometrics Fisheries Location choice

#### ABSTRACT

A substantial theoretical and experimental literature has focused on the conditions under which cooperative behavior among actors providing public goods or extracting common-pool resources arises. The literature identifies the importance of coercion, small groups of actors, or the existence of social norms as conducive to cooperation. This research empirically investigates cooperative behavior in a natural resource extraction industry in which the provision of a public good (bycatch avoidance) in the Alaskan flatfish fishery is essential to the duration of the fishing season, and an information provision mechanism exists to relay information to all individuals. Using a mixed logit model of spatial fishing behavior our results show that conditionally cooperative behavior is prevalent but deteriorates as bycatch constraints tighten.

© 2009 Elsevier B.V. All rights reserved.

"...the world contains multiple types of individuals, some more willing than others to initiate reciprocity to achieve the benefits of collective action. Thus a core question is how potential cooperators signal and design institutions that reinforce rather than destroy conditional cooperation." — Elinor Ostrom (2000, pg. 1380)

#### 1. Introduction

A substantial theoretical and experimental literature has focused on the conditions under which cooperative behavior is likely to occur among actors providing public goods or extracting common-pool resources. Following from the assumption of rational self-interested agents, Olson (1970) argues that cooperation is likely when the group of individuals is small and when coercive authority lies with the group. Ostrom (2000) suggests that these conditions may be too restrictive, given experimental evidence and numerous field studies. Norms, ethical codes, and institutions allowing for verification and coercion exist in many settings that reinforce cooperative behavior. These systems allow groups to identify non-cooperative behavior and impose sanctions.

We empirically examine a common property resource problem in which coercive legal authority is absent but peer pressure coercion is present, the Alaskan yellowfin sole and flatfish fisheries, in an effort to formally investigate conditional cooperation in this institutional setting. The group of participants is fairly small, the group size and composition is stable (a limited-entry system is in place), and a voluntary information sharing arrangement among participants has

evolved to overcome a common-property problem. This arrangement provides participants with information on where high levels of non-target species (bycatch) are caught with the hope that vessels will avoid these areas. Furthermore, information on non-cooperation is provided to other participants and peer pressure may be exerted to enhance cooperation.<sup>2</sup> Consequently, the system we investigate is similar to the institutional environment outlined in Ostrom (2000) and we are able to empirically investigate her hypotheses using field data.

Currently, fishermen operating in the Alaskan flatfish fisheries operate under a two-tiered total allowable catch (TAC) system. TACs are defined over target and bycatch species, in our case Pacific halibut, and once either TAC is reached all fishing ceases. This regulation is further complicated by the fact that halibut is Prohibited Species Catch (PSC) and can therefore not be sold for economic gain. Consequently, reaching the bycatch TAC has direct economic consequences for the fleet, and in recent years bycatch TACs have prematurely shut down some fisheries leaving considerable unharvested economic rents in the primary target fishery. From an individual fisherman's perspective catching halibut bycatch is a nuisance due to its PSC declaration. Avoidance comes with a large individual opportunity cost since the marketable species, vellowfin sole and other flatfish, share similar habitat requirements with the bycatch species making them complements of production. Importantly, in terms of cooperation, the collective harvesting decisions made by the fleet determine the length of the season any one vessel enjoys. In many cases the marginal individual costs from avoiding

<sup>\*</sup> Corresponding author. Tel.: +1 404 413 0159; fax: +1 404 413 0248. *E-mail addresses*: Alan.Haynie@noaa.gov (A.C. Haynie), rob.hicks@wm.edu (R.L. Hicks), kschnier@gsu.edu (K.E. Schnier).

<sup>&</sup>lt;sup>1</sup> Senior authorship not assigned.

<sup>&</sup>lt;sup>2</sup> This information is provided regularly by both an industry group and by the National Marine Fisheries Service (NMFS) which publishes a list of the observed bycatch rates by vessel each week.

halibut may be greater than their share of the gains from avoidance, an extended season. Therefore, avoidance of halibut in this context generates a public good for the others in the fleet.

The Magnuson-Stevens Fishery Conservation and Management Act, the primary legislation for fisheries management in the United States, defines bycatch as, "fish which are harvested in a fishery, but which are not sold or kept for personal use, and includes economic discards and regulatory discards" (16 U.S.C. § 1802 (2)). Bycatch often has a negative effect on either the contemporaneous or potential future value within other fisheries.<sup>3</sup> Because of this, organizations have been formed to collect production information from participating firms to inform members of spatial locations where bycatch should be avoided (Gilman et al., 2006).<sup>4</sup> This research investigates the role of one of these information organizations operating within the Bering Sea flatfish fishery, Sea State Inc. ("Sea State").<sup>5</sup> Using the Alaska catch and bycatch data collected by onboard Federal National Marine Fisheries Service (NMFS) observers, we are able to observe whether bycatch avoidance occurs (a signal of cooperation) and whether the level of avoidance varies during the fishing season as the bycatch TAC is approached. Changes in the level of bycatch avoidance within the season indicate the degree of conditional cooperation present within the fishery. Further, we examine the distribution of the fleet's cooperative behavior and find that as the fleet approaches the bycatch TAC the degree of avoidance deteriorates, further exacerbating the degree of conditional cooperation observed. However, we do observe a marginal increase in aversion rates in the penultimate days of the fishery, which is consistent with the recent findings of Abbott and Wilen (2008, in press) studying a subset of vessels from the same fishery in our model during an earlier time frame.

As mentioned earlier, utilizing Sea State information to avoid highbycatch regions can be considered a public good. Within the experimental literature, common-pool resource and public goods games are often viewed as isomorphic games because they can be equally expressed as transformations of the individual payoff functions (Ledyard, 1995). However, the nature of the benefit-cost duality of these two environments generates fundamental differences in the decision environment (Sandler and Arce, 2003). Sandler and Arce (2003) illustrate that the difference lies in the "need for inaction" in common-pool resource environments versus the "need for action" in the public goods setting. Apesteguia and Maier-Rigaud (2006) add some additional insights by focusing on the degree of rivalry present in common-pool resource and public good environments. They define rivalry by the degree to which one's actions solely benefit oneself versus all others within the population. Given that avoiding bycatch benefits everyone in the fishery, via an increased fishing season, and not just the acting agent whereas the individual opportunity costs of aversion may be quite high, bycatch aversion generates a public good. Furthermore, the "need for action" in this environment is the active aversion of bycatch within the fishery.

The field data we examine provides a setting for testing how information with minimal levels of coercion impacts cooperation. Avoiding an otherwise preferable high-bycatch zone generates a marginal return to each fisherman that is inferior to the non-cooperative

return, but if all fishermen cooperate the aggregate returns would increase as the season would not be prematurely shut down.<sup>6</sup> The summary offered by Ostrom (2000) of the experimental literature on public goods games offers an informal baseline to which our econometric results can be compared. These results show that (1) even in the simplest public goods experiments of one-shot games where no coercion is possible, 40–60% of participants choose cooperative strategies<sup>7</sup>; (2) in repeated games, cooperation decays as the number of rounds of play progresses but never reaches zero; (3) less than 30% of respondents cooperate in the final period of multiple period games; and (4) knowing that others are cooperating tends to increase cooperation (the converse is also true). Our results support a number of these same conclusions, further validating many of the generalizations suggested by Ostrom (2000) obtained from laboratory experiments.

The remainder of the paper proceeds as follows: Section 2 describes the Alaska yellowfin sole and flatfish fisheries and the data utilized in our empirical analysis. In Section 3, we describe the empirical model used to investigate the responsiveness of fishermen to bycatch. Section 4 presents a discussion of the results and the final section summarizes our major findings.

## 2. Fishery Description

In order to manage bycatch within the Bering Sea, in-season fishery managers have utilized bycatch TACs combined with time/ area closures. The bycatch species are a target species within another fishery. Therefore, the bycatch TACs are set equal to a pre-specified percentage of the overall target TAC for each bycatch species within its target fishery. These bycatch TACs are further subdivided across different target species and into seasons in order to spread out the temporal distribution of fishing. Once the bycatch TACs are reached, in-season managers issue a fishery closure. These closures have resulted in a number of fisheries being prematurely terminated, forgoing a considerable portion of the target species TAC. For instance, over the time period studied in this research (2000–2004) the large yellowfin sole fishery was shut down by reaching the halibut bycatch TAC in 2001, 2002 and 2003. Furthermore, in 2001 over 20% of the TAC was left un-harvested. Section 1.

We focus our study on trawling vessels that target flatfish in the Bering Sea, which are significantly constrained by bycatch limits for halibut. The primary species targeted are yellowfin sole, flathead sole and rock sole. These species are caught by a fleet that has different targeted species which are opened and closed during the season as catch or bycatch caps are reached. To reflect these differences, we partitioned our data set into two groups: those targeting yellowfin sole and those targeting all other flatfish, which we refer to as 'flatfish'. Yellowfin sole was analyzed separately because during the study period it had, by a considerable margin, the largest TAC of all flatfish species within the Bering Sea. Weekly

<sup>&</sup>lt;sup>3</sup> Shrimp fisheries are an excellent example of this as shrimp trawlers often catch large quantities of juvenile fish. See Gallaway and Cole (1999) and Reithe and Aschan (2004) for more discussion.

<sup>&</sup>lt;sup>4</sup> Other efforts to reduce bycatch in fisheries have involved the utilization of bycatch reduction devices (Pascoe and Revill, 2004), spatial closures (Reithe, 2006) and seasonal closures (Bisack and Sutinen, 2006).

<sup>&</sup>lt;sup>5</sup> Sea State regularly provides vessel-specific daily bycatch rates during high-bycatch-periods and also talks to fishermen about particularly high hauls to better understand bycatch conditions. In addition, NMFS provides vessel-level rates on a weekly basis, but the published numbers are perceived by industry to be difficult to interpret when vessels are operating in multiple areas and targeting multiple species so there is additional value in the Sea State reports.

<sup>&</sup>lt;sup>6</sup> It is conceivable that the aggregate benefits would not exceed the private costs of avoidance, but given that the actor choosing to avoid bycatch would discount aggregate benefits, this is very unlikely to occur.

<sup>&</sup>lt;sup>7</sup> The degree of coercion that results from the NMFS bycatch reports is not clear. Therefore, the non-coercion results may still be a useful benchmark for measurement.

<sup>&</sup>lt;sup>8</sup> The primary bycatch concerns surround prohibited species catch (PSC), which consist of crab, Pacific herring, Pacific halibut, Pacific salmon and steelhead trout harvested within the bottom trawl fishery (Witherell and Pautzke, 1997).

<sup>&</sup>lt;sup>9</sup> There has been some research conducted on the optimal allocation of bycatch TAC among sub-fisheries in the Bering Sea flatfish fishery. Larson et al. (1996) illustrate that a substantial portion of the halibut quota should be reallocated from the longline fishery to the Alaska pollock fishery. Further results indicate that quasi-rents in the pollock fishery over the years 1991–92 could have been increased by 6–7% if the bycatch TAC shares had been optimally defined (Larson et al., 1998).

<sup>&</sup>lt;sup>10</sup> In the past crab bycatch has also constrained these fisheries, but protected areas have pushed the fishery off of high crab bycatch grounds.

Several other species are also caught and marketed (e.g., Dover sole, rex sole) which are jointly considered as 'other flatfish.'

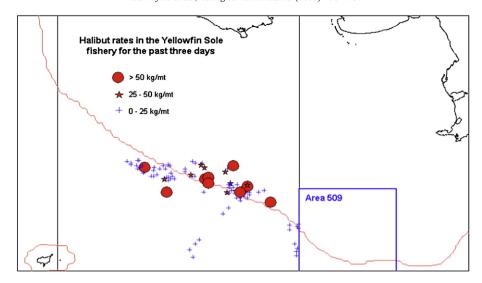


Fig. 1. Spatial distribution of Sea State halibut bycatch information. (Courtesy of Sea State Inc.: report issued on August 18, 2003.)

targeting designations are based on the NMFS specifications which designate a yellowfin sole/flatfish target if the largest component of the catch is made up of yellowfin sole and flatfish species. Assuming this threshold is met, the target is declared a yellowfin sole target if greater than 70% of the sum of all flatfish is yellowfin sole and is declared a flatfish target otherwise. Typically either the yellowfin sole fishery or the flatfish fishery is open during a particular point in the season, though there are overlapping periods and periods when both fisheries are closed.

The vessels that fish for yellowfin sole and flatfish are catcher processors which take trips that may last 2-4 weeks and process fish onboard. In order to minimize the frequency and duration of bycatch closures, the yellowfin sole and flatfish fleet contracted with Sea State in 1995 to begin analyzing government observer-collected bycatch information. Sea State provides spatial bycatch advisories to the fleet, which provide non-mandatory recommendations of areas to avoid in order to reduce bycatch. The program operates using the real-time processing of the observer data recorded by NMFS on all vessels within the fleet. An example of the bycatch advisory is illustrated in Fig. 1, which depicts the spatial distribution of halibut bycatch rates within the Bering Sea for the yellowfin sole fishery, Furthermore, Sea State and NMFS regularly report by catch rates by vessel to enhance the information provided to the fleet, which may be used to exert coercive pressure on non-cooperators within the fleet, where we define a non-cooperator as a fisherman who does not actively avoid regions of high halibut bycatch. Given this informational structure, there is a widespread impression across the industry that the Sea State program has been successful in reducing bycatch.

Catch and bycatch data for this analysis come from the Alaska Fisheries Science Center's Observer Program Database. The North Pacific Groundfish Observer Program places observers on 100% of the days at sea for vessels 125 ft in length and greater, which captures the entire catcher processor fleet within our analysis. <sup>12</sup> Each data point represents a given haul made by a vessel while on a fishing cruise. Spatial data on fishing locations were used to calculate the distances from one haul to the next and from each haul to the centroid of areas that might potentially be chosen for sequential hauls. To complete the data set, we obtained price information from the Commercial Fisheries Entry Commission (CFEC) fish ticket and Commercial Operator Annual Report (COAR) data.

Following the establishment of bycatch TACs, NMFS tracks when fisheries approach and reach their annual limits and provides regular information to the public about the status of fisheries, including the issuance of fishery closures. For each observation in the data set, we calculated the currently applicable bycatch TAC at a given point in the season. NMFS in-season managers also make bycatch TAC adjustments among fisheries within the season to allow a larger amount of yellowfin sole and flatfish to be caught without exceeding the overall bycatch limits. Through a careful investigation of the timing of these changes, we incorporated these adjustments into our analysis. Since this information is relayed through the fleet, both by Sea State and inseason management, fishermen are acutely aware of how binding bycatch TACs are at any given point in time.

Because structural differences in targeting behavior and halibut avoidance may exist for years when the halibut TAC was binding or not, we partitioned the yellowfin sole and flatfish data sets into binding and non-binding years defined after the completion of the season. Although the halibut TAC was binding in all years for at least one of the flatfish fishery seasons, it was not binding in the summer and fall seasons in 2003 and 2004, so we elected to declare 2003 and 2004 as "non-binding" years within the analysis despite the fact that it was binding in the winter and spring seasons. The yellowfin sole fishery, on the other hand, was non-binding in 2000 and 2004 for all sub-seasons within the year.

The previous paragraphs provide a brief description of the institutional environment that fishermen within these fisheries operate under. For the purpose of this research endeavor we assume that the behavior observed is conditional on this institutional setting. This

Table 1 lists the descriptive statistics for the yellowfin sole and flatfish fisheries over the time period studied (2000–2004). The mean revenue per haul in the flatfish fishery was approximately 38% greater than that in the yellowfin sole fishery. This is predominately the result of the targeting of rock sole, a high-valued flatfish species. In addition, bycatch rates and quantities are consistently higher in the flatfish fishery than in the yellowfin sole fishery. On average, a flatfish haul catches 44% more halibut than a yellowfin sole haul. <sup>13</sup> Aside from the disparity in revenues and bycatch rates and quantities within these two fisheries, on average fishermen in these fisheries visit a very similar number of spatial locations on a cruise (8.16 for the yellowfin sole and 7.32 for the flatfish fishery), suggesting a similar level of spatial mobility.

<sup>&</sup>lt;sup>12</sup> In addition, during 2000–2004 over 99% of the catcher processors observed were Sea State members. Our data set does not contain the smaller catcher vessels that have only 30% of their fishing days observed.

<sup>&</sup>lt;sup>13</sup> Sea State relays bycatch information in terms of rates (tons of bycatch per haul as a percent of tons of catch haul) rather than raw quantities (tons of bycatch per haul). Flatfish hauls have an average bycatch rate that is 73% greater than yellowfin sole.

**Table 1**Summary statistics.

Yellowfin sole	Mean	Std. dev.	Min	Max
Cruise revenue	56,653.30	50.646.37	0	241,213.93
Haul revenue	611.49	983.22	0	16,589.69
Cruise length (hauls)	80.19	62.02	1	335
Cruise length (days)	21.00	16.61	<1	79
Bycatch (rates)	16.37	32.24	0	553.52
Bycatch (quantities)	285.20	667.98	0	14,463.10
Sites visited per cruise	8.16	4.93	1	32
Flatfish fishery				
Cruise revenue	117,984.24	87,826.47	0	768,824.00
Haul revenue	719.71	1152.25	0	14,232.27
Cruise length (hauls)	120.83	102.64	1	452
Cruise length (days)	29.57	22.97	<1	95
Bycatch (rates)	28.26	40.83	0	512.37
Bycatch (quantities)	410.82	679.59	0	13,086
Sites visited per cruise	7.32	4.65	1	30

presumes away an equally interesting research question regarding the public policies that preceded the institutional structure and influence the fisherman's decision environment. <sup>14</sup> Instead this research will focus on the micro-behavior of fishermen within these fisheries in an effort to formally investigate the conditional cooperation observed in economics experiments within a field environment. Having discussed the general nature of the yellowfin sole and flatfish fisheries, the following section outlines the econometric model utilized to investigate how fishermen respond to the spatial information provided by Sea State over the course of the season.

#### 3. Econometric Model

Flatfish fishermen make repeated spatial choices on which region within the fishery to fish in during a given time period. Our definition of "space" divides the Bering Sea into 1° latitude by 0.5° longitude grids, which are the statistical reporting zones utilized by the Alaska Department of Fish and Game. Given the discrete nature of the fisherman's choice set, random utility modeling is conventionally used to model fisherman's spatial behavior (Eales and Wilen, 1986; Curtis and Hicks, 2000; Holland and Sutinen, 2000; Smith and Wilen, 2003) and we follow this paradigm with one exception; we utilize a mixed logit model (McFadden and Train, 2000) to allow for heterogeneous responses to the spatial bycatch information. 15 To examine the degree of cooperation, we developed a model that incorporates site-specific factors that were hypothesized to influence spatial decisions made by fishermen (e.g., expected revenues per haul and distance to a site) and proxies for information supplied by Sea State characterizing the degree of bycatch expected at each site. If cooperation to reduce bycatch via information sharing was indeed happening then we would expect fishermen to avoid high-bycatch areas ceteris paribus. The mixed logit model has been utilized in the spatial choice literature to investigate heterogeneity in risk preferences (Eggert and Tveterås, 2004; Mistiaen and Strand, 2000), travel costs (Haynie, 2005), and state dependence in fisheries (Smith, 2005). However, this is the first application utilizing a mixed logit to investigate heterogeneous responses to spatial information signals and to investigate cooperative behavior with respect to information sharing. The foundation for our model rests on the commonly used RUM developed by McFadden (1974, 1978).

Consider a panel data set of M fishermen conducting  $C_m$  fishing trips or cruises with each trip consisting of  $h_{c,m}$  hauls, the utility individual i derives from visiting site j on haul  $h_{c,m}$  be defined as,

$$v_{ijh_{cm}} = x_{jh_{cm}} \beta_{i_c} + \varepsilon_{ijh_{cm}} \tag{1}$$

where  $x_{jh_{c,m}}$  is a vector of location and haul-specific observations and  $\beta_{ic}$  is an individual and cruise-specific time-invariant preference parameter. The observation matrix,  $x_{jh_{c,m}}$ , is observed by both the researcher and the fisherman but  $\varepsilon_{ijh_{c,m}}$ , the unobserved (by the researcher) portion of site-haul-location specific utility, is only observed by the fisherman. Fisherman i will choose to fish in site j on haul  $h_{c,m}$  if the utility of fishing in site j exceeds all other sites in the fishery on their  $h_{c,m}^{th}$  haul. This is denoted as,

$$v_{ijh_{cm}} \ge v_{ikh_{cm}}, j \in N, \forall k \in N$$
 (2)

where N is the total number of feasible spatial locations. If  $\varepsilon_{ijh_{cm}}$  is assumed to be an independently and identically distributed Type I Extreme Value and  $\beta_{i_c} \sim MVN(\tilde{\beta},\Omega)$ , we can recover the probability  $p_{ijh_{cm}}$  that fisherman i selects location j on his/her  $h_{c,m}^{th}$  haul. This probability nests the multinomial logit model (MNL) within the multivariate integral of the distribution for  $\beta_{i_c}$  and can be expressed as,

$$p_{ijh_{c,m}} = \int \frac{e^{X_{jh_{c,m}}\beta}}{\sum\limits_{k=1}^{N} e^{X_{kh_{c,m}}\beta}} \phi(\beta|\tilde{\beta},\Omega) d\beta.$$
 (3)

Estimating the integral expressed in Eq. (3) requires a simulation-based estimation algorithm which numerically approximates the integral using Monte Carlo simulation (Train, 2003). <sup>16</sup> Within the Monte Carlo simulation D draws are made from the multivariate normal distribution with each draw producing a hypothesized value for  $p_{ijh_{c,m}}$ , denoted  $p_{ijh_{c,m}}^d$  where d indicates the  $d^{th}$  draw. In our analysis, we used 200 Halton draws from the multivariate normal distribution and there were 345 and 233 unique individual and cruise-specific identifiers (indicated by subscript c,m) within the yellowfin sole and flatfish fishery, respectively. From these draws a simulated likelihood function can be constructed,

$$L = \prod_{c=1}^{C_m} \prod_{i=1}^{M} \prod_{b=1}^{h_{c,m}} \frac{1}{D} \sum_{d=1}^{D} p_{ijk}^d. \tag{4}$$

Maximum likelihood maximizes the log transformation of Eq. (4). The specification of utility  $v_{ijh,m}$  in our empirical model is  $^{17}$ 

$$v_{ijh_{c,m}}^{f} = \beta_1 Dist_{j|kh_{c,m}} + \beta_2 Rvn_{ijh_{c,m}} + \beta_{3i}b_{ijh_{c,m}} + \beta_{4i} \left(b_{ijh_{c,m}} *Rmn_{ijh_{c,m}}\right) + \beta_{5i} \left(b_{ijh_{c,m}} *Rmn_{ijh_{c,m}}\right)^2 + \beta_6 Miss.Dum + \varepsilon_{ijh_{c,m}}.$$
(5)

The superscript f denotes the sub-fishery studied: flatfish catcher processors (FLAT\_CP) and yellowfin sole catcher processors (YELL\_CP) respectively, further subdivided into binding and non-binding years. Therefore, there are four separate regression models estimated.

<sup>&</sup>lt;sup>14</sup> We would like to thank an anonymous referee for pointing out this assumption and the importance of the public policies leading to the institutional framework as well as an example of research which explicitly accounts for this connection in the public policy literature (Gerber et al. 2009).

<sup>&</sup>lt;sup>15</sup> Bockstael and Opaluch (1983) were the first to apply a random utility model (RUM) in the fisheries literature, but they did not directly investigate spatial behavior. Bockstael and Opaluch (1983) investigated the effort supply response of fishermen in New England to expected fishery yields and corresponding variances.

 $<sup>^{16}</sup>$  Our estimator was programmed in MATLAB with the foundational code provided by Kerry Smith and Dan Phaneuf.

<sup>&</sup>lt;sup>17</sup> A large number of alternative empirical specifications for Eq. (5) were estimated to investigate the robustness of our results to the reduced form specification of the utility function. These models included alternative variable interaction models and using tierspecific parameter estimates for the amount of remaining bycatch present. These models yielded similar results to those discussed in the paper and are available upon request from the authors.

The distance traveled from one's current location k to location j on the current haul  $h_{c,m}$  is captured by  $\mathrm{Dist}_{jlkh_{c,m}}$  and is measured in kilometers.  $Rvn_{ijh_{c,m}}$  is the expected site-specific revenue for haul  $h_{c,m}$  and is calculated using the seven-day moving average of site-specific revenues observed over each of the respective sub-fisheries. The bycatch information signal, denoted  $b_{ijh_{c,m}}$ , is specified as the expected site- and time-specific bycatch rate, defined as the ratio of expected bycatch to expected catch of the primary target species. This treatment is utilized to capture the information provided by Sea State (see Fig. 1). The expected spatial bycatch rates are calculated using 7-day moving averages for each site within the fishery. Seven-day moving averages were selected because this closely mimics the weekly intervals used by in-season management when declaring the available bycatch TAC remaining, as well as the time intervals used by Sea State.

To investigate the intra-seasonal spatial response of vessels to bycatch information, we interacted the bycatch signal  $b_{ijh_{cm}}$  variable with the amount of bycatch remaining at a particular point in time within the fishery, denoted  $Rmn_{ijh_{cm}}$ . In addition, the square of this interaction term was added to the specification to account for secondorder effects. The final variable used in the analysis, Mis.Dum, takes a value of one whenever the expected location- and haul-specific estimates of revenues and bycatch are zero because no fishing activity has taken place in that area over the past 7 days. Utilizing this reduced form specification facilitates the analysis of the fishermen's behavioral responses when the bycatch TAC becomes more binding on fleet behavior, where binding is defined as less bycatch TAC remaining.<sup>18</sup> Furthermore, partitioning the data into binding and non-binding years in the analysis allows us to investigate whether or not vessels possess asymmetric response functions to the state of remaining bycatch TAC. Regression results are displayed in Tables 2 and 3 for the yellowfin sole and flatfish fisheries, respectively.

In addition to the results presented in Tables 2 and 3, we have estimated spatial response elasticities for the mixed logit models. To estimate these elasticities, a two-stage Krinsky-Robb method (Krinsky and Robb, 1986, 1990) was utilized to obtain confidence intervals for our elasticity estimates. The elasticity distributions are used to estimate the fleet-wide response to a 1% increase in the bycatch rate within a given area as the amount of bycatch remaining decreases within the season. The elasticities were calculated by simulating the mean behavioral responses for each haul within the data resulting from a 1% increase in the bycatch rate, holding all other spatial information constant, while varying the amount of remaining bycatch from 1.0 to -0.4 using increments of 0.1 for each haul. In the first stage of the estimation, a draw is taken from the multivariate normal parameter distribution utilizing the variancecovariance matrix from our estimation. In the second stage, a draw from each random parameter's distribution is taken conditional on the mean and standard deviation estimate drawn from the first stage. This secondstage process is repeated for each  $\beta_{ic}$  in the empirical model. The resulting parameters are then used to construct the spatial response elasticities. All elasticity estimates are conducted at the haul level and enough draws were taken from the parameter distribution to ensure that 500,000 haulspecific elasticities were estimated for each model. The results are reported in Tables 4 and 5 for the yellowfin sole and flatfish fisheries respectively.

## 4. Results

The empirical results for the models possess three commonalities: (1) the distance coefficient in each model is negative and highly significant, (2) the revenue coefficient is positive and statistically significant in most models and, (3) the coefficient on *Mis.Dum* is

**Table 2** Yellowfin estimates — mixed logit: mean and standard deviation, and *t*-stat.

Coefficient/fishery	YELL_CP RPL	YELL_CP RPL
Model	Binding years	Non-binding years
Distance	-38.4995**	-37.1966**
	$(-127.75)^{a}$	$(-96.75)^{a}$
Revenue	0.1074**	0.0188
	(5.91) <sup>a</sup>	$(0.80)^{a}$
$b_{ijh_{cm}}$	$-9.4897^{**}$	- 18.6975**
	(3.70)	(3.96)
Std. deviation	16.7089**	$-22.5769^{**}$
	(7.72)	(11.86)
$b_{ijh_{cm}}*Rmn_{ijh_{cm}}$	15.1598 <sup>**</sup>	37.5530 <sup>**</sup>
	(2.51)	(4.43)
Std. deviation	21.7821**	4.6021
	(6.64)	(0.44)
$(b_{ijh_{cm}}*Rmn_{ijh_{cm}})^2$	-1.2104**	$-1.3481^{**}$
	(4.34)	(-6.14)
Std. deviation	1.0206**	0.9895**
	(4.82)	(7.32)
Mis.Dum	$-1.8495^{**}$	-1.7338**
	$(-43.97)^{a}$	$(-34.61)^{a}$
Number of obs.	16,715	10,220
$\log(L_0)$	-72,822	- 44,526
$\log(L)$	-26.007	-15,458
Likelihood ratio index $(\rho)$	0.6429	0.6528

<sup>\*</sup> Indicates significance at the 90% level; \*\* indicates significance at the 95% level.

negative and highly significant in all models. The coefficients on distance and revenues are consistent with the general results in the literature indicating that travel is costly and that fishermen select sites which possess higher expected revenues. In addition, the coefficient on *Mis.Dum* is consistent with the results in the literature which indicate that fishermen tend to fish in locations which have been fished in the recent past (Holland and Sutinen, 2000). Beyond these similarities, each of the fisheries possesses a unique profile of response to bycatch information within the fishery.

The mixed logit parameter estimates illustrate that fishermen participating in both fisheries possess an initial propensity to avoid locations with a higher bycatch rate, captured by  $b_{ijh_{cm}}$  and the second order interaction term, but when the amount of bycatch quota remaining is large this propensity is lower (captured by the interaction terms). Furthermore, in the non-binding years for both fisheries the baseline aversion rate is larger than in binding years. Describing the spatial responses beyond this level is complicated by the bycatch interaction terms; therefore we will focus primarily on the spatial elasticity estimates conditional on a given level of bycatch quota remaining.

The elasticity of spatial responses in the yellowfin sole fishery resulting from an increase in the bycatch rate, conditional on the level of bycatch quota remaining, is listed in Tables 4 and 5 for the binding and non-binding years respectively. The results indicate that the mean spatial responses decrease as the bycatch quota is reduced, with mean aversion rates being larger in the non-binding years. In the binding-year model the deterioration in aversion rates results in the mean aversion switching to an attraction model when the remaining bycatch quota is less than 0.4. In the case of the non-binding years the switch occurs when the remaining bycatch quota is less than 0.1, which rarely occurs given that it is a non-binding year.<sup>20</sup> Another interesting common feature of the two models is the degree of heterogeneity within the fishery, as measured by the span between the upper and lower bounds, decreases and then increases as the remaining bycatch

<sup>&</sup>lt;sup>18</sup> Alternative thresholds were experimented with in our preliminary analysis. Our tiering of the remaining bycatch TAC allows us to focus on the end-of-season dynamics within the fishery and the fishermen's responses to bycatch information during these respective time periods.

<sup>&</sup>lt;sup>a</sup> Indicates *t*-stat because parameter is not random.

<sup>&</sup>lt;sup>19</sup> A more rigorous specification of the utility function would incorporate a state dependence variable (Smith, 2005) to investigate this phenomenon in more detail, but this is beyond the scope of this research effort.

<sup>&</sup>lt;sup>20</sup> Although the remaining bycatch rarely, if ever, goes below zero in non-binding years we have elected to estimate the elasticities at these levels to be parsimonious with the other elasticity models.

**Table 3** Flatfish fishery estimates — mixed logit: mean, standard deviation and *t*-stat.

Coefficient/fishery	FLAT_CP RPL	FLAT_CP RPL
Model	Binding years	Non-binding years
Distance	-37.6316**	-35.5354 <sup>**</sup>
	$(-103.95)^{a}$	$(-63.62)^{a}$
Revenue	0.2574**	0.1840**
	(12.12) <sup>a</sup>	(7.57) <sup>a</sup>
$b_{ijh_{cm}}$	8.9532**	- 10.5624**
	(2.84)	(4.87)
Std. deviation	64.15 <sup>**</sup>	12.8627**
	(19.49)	(5.93)
$b_{ijh_{cm}}*Rmn_{ijh_{cm}}$	14.2410**	8.0772
3 - 5	(2.89)	(1.43)
Std. deviation	14.2682**	15.3245**
	(4.96)	(2.94)
$(b_{ijh_{cm}}*Rmn_{ijh_{cm}})^2$	$-0.2920^{**}$	-0.3823**
- Can	(-6.33)	(-4.28)
Std. deviation	0.6561**	0.0946*
	(9.58)	(1.71)
Mis.Dum	-1.4818**	- 1.9073**
	$(-30.69)^{a}$	$(-22.74)^{a}$
Number of obs.	12,517	5,399
$\log(L_0)$	-54,042	-23,310
$\log(L)$	<b>−17,876</b>	-7,389
Likelihood ratio index $(\rho)$	0.6692	0.6830

<sup>\*</sup> Indicates significance at the 90% level; \*\* indicates significance at the 95% level.

quota falls. To a certain degree this is a construct of the interaction variables used in the empirical analysis because as the remaining bycatch quota approaches 0 the first and second-order interaction terms approach zero as well and the marginal effect of bycatch alone (non-interacted) is all that remains. This said, the largest rates of aversion and attraction are observed when the remaining bycatch quota is 1. In binding years this spread is between -2.47% and 1.25% and in non-binding years the spread is larger, between -3.58% and 0.88%. Interestingly the larger aversion rates for the non-binding years suggest that more cooperative behavior is observed in non-binding years. However, given that our definition of binding and non-binding is determined following the completion of the season, this result would be expected because if cooperation was not observed it would increase the likelihood that the year was a binding year. Therefore, this result may be a construct of our data partitioning methods.

Although the mean aversion rates are predominately negative in the yellowfin sole fishery, the confidence intervals for our elasticity estimates indicate that the mean haul-specific elasticity measure in both binding and non-binding years is not statistically significant from zero. This complicates the interpretation of the results. However, by

**Table 4** Yellowfin sole fishery elasticities, 95% confidence intervals.

Data set	YELL_CP binding			YELL_CP non-binding		
Bycatch quota remaining	Lower 2.5%	Mean	Upper 97.5%	Lower 2.5%	Mean	Upper 97.5%
1.0	-0.0247	-0.0043	0.0125	-0.0358	-0.0086	0.0088
0.9	-0.0207	-0.0033	0.0119	-0.0306	-0.0070	0.0088
0.8	-0.0168	-0.0023	0.0122	-0.0258	-0.0054	0.0086
0.7	-0.0133	-0.0014	0.0103	-0.0210	-0.0040	0.0083
0.6	-0.0104	-0.0007	0.0094	-0.0168	-0.0027	0.0079
0.5	-0.0078	-0.0002	0.0084	-0.0140	-0.0018	0.0071
0.4	-0.0056	0.0000	0.0074	-0.0118	-0.0011	0.0063
0.3	-0.0043	0.0002	0.0065	-0.0093	-0.0006	0.0058
0.2	-0.0044	0.0002	0.0058	-0.0073	-0.0002	0.0055
0.1	-0.0058	0.0001	0.0055	-0.0066	-0.0001	0.0055
0	-0.0063	0.0001	0.0056	-0.0058	0.0001	0.0057
-0.1	-0.0047	0.0004	0.0059	-0.0049	0.0004	0.0062
-0.2	-0.0032	0.0006	0.0066	-0.0048	0.0005	0.0067
-0.3	-0.0032	0.0008	0.0074	-0.0060	0.0004	0.0072
-0.4	-0.0046	0.0007	0.0083	-0.0082	0.0000	0.0076

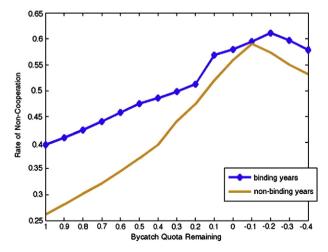
 Table 5

 Flatfish fishery elasticities, 95% confidence intervals.

Data set	FLAT_CP binding		FLAT_CP non-binding			
Bycatch quota remaining	Lower 2.5%	Mean	Upper 97.5%	Lower 2.5%	Mean	Upper 97.5%
1.0	-0.3002	-0.0326	0.0126	-0.5387	-0.0413	0.0145
0.9	-0.2486	-0.0275	0.0120	-0.4816	-0.0361	0.0141
0.8	-0.2021	-0.0231	0.0114	-0.4227	-0.0312	0.0139
0.7	-0.1711	-0.0196	0.0109	-0.3502	-0.0273	0.0140
0.6	-0.1462	-0.0168	0.0106	-0.2791	-0.0238	0.0135
0.5	-0.1281	-0.0149	0.0102	-0.2132	-0.0215	0.0130
0.4	-0.1196	-0.0141	0.0098	-0.1816	-0.0200	0.0130
0.3	-0.1109	-0.0143	0.0095	-0.1751	-0.0198	0.0126
0.2	-0.1114	-0.0146	0.0094	-0.1752	-0.0210	0.0126
0.1	-0.1135	-0.0148	0.0094	-0.1811	-0.0233	0.0126
0.0	-0.1129	-0.0146	0.0094	-0.1807	-0.0223	0.0127
-0.1	-0.1090	-0.0138	0.0094	-0.1762	-0.0208	0.0127
-0.2	-0.1021	-0.0126	0.0096	-0.1623	-0.0182	0.0128
-0.3	-0.0957	-0.0112	0.0098	-0.1553	-0.0158	0.0130
-0.4	-0.0950	-0.0103	0.1010	-0.1583	-0.0146	0.0135

calculating the distributional mass of our haul-specific elasticity estimates that lie above zero we can determine the degree of and change in non-cooperation within the yellowfin sole fishery because a positive elasticity is consistent with non-cooperative behavior and the change in cooperative behavior captures conditional cooperation. Therefore, we can still investigate the degree of conditional cooperation present when using these empirical results. The degree of non-cooperation within the yellowfin sole fishery is illustrated in Fig. 2 for both binding and non-binding years.

Analyzing this metric uncovers four generalizations within the yellowfin sole fishery: (1) non-binding years possess a higher degree of cooperation, (2) the degree of non-cooperation increases as the season progresses, (3) the rate of increase in non-cooperation increases more rapidly in non-binding years, and (4) when the remaining bycatch approaches a level at which the shut-down of the fishery is imminent the degree of non-cooperation decreases as the aversion rates rise. In non-binding years the rate of non-cooperation always lies below that observed in binding years. When the season begins, the remaining bycatch quota is 1, the degree of non-cooperation is roughly 26%, compared to nearly 40% in binding years. The spread between the nonbinding and binding years is greater than 10% up until the remaining bycatch quota is 0.5, at which time the degree of non-cooperation increases until it is less than a percentage point below the binding year counterpart when the remaining by catch quota approaches -0.1. This rapid decrease in cooperation is perfectly rational within the nonbinding years because as the season progresses and it becomes more



**Fig. 2.** Simulated degree of non-cooperation within the yellowfin sole fishery conditional on a given level of bycatch quota remaining.

<sup>&</sup>lt;sup>a</sup> Indicates not a random parameter.

evident that closure is not eminent as fishermen would tend to disregard the knowledge of where high-bycatch rates are located and focus more on the expected revenues they will obtain in a given location. This is not to say that they do not avoid bycatch, it suggests that they have already avoided bycatch enough to prolong the season to successful completion.

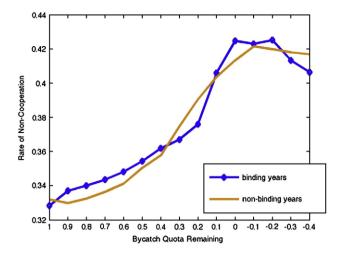
The degree of non-cooperation increases from the initial levels to greater than 60% in the binding years and slightly less than 60% in the non-binding years. The sharp increase in non-cooperation within both binding and non-binding years is consistent with Ostrom's (2000) observations on conditional cooperation as well as her finding that endof-game non-cooperation rates approach 70% in the experimental literature. As mentioned above caution should be used when applying this logic to the non-binding years because cooperation is not necessary in the later part of the season. However, the increasing decay in cooperation in non-binding years is consistent with Ostrom's (2000) generalizations of non-cooperation discussed earlier. This said, it should also be noted that the degree of cooperation is still greater than Ostrom's (2000) generalizations, which may be a direct result of the coercive power of fishermen in the yellowfin sole fishery resulting from the vessel-specific bycatch rate reports. Coercive pressure may also be used to explain why at the very end of the season the aversion rates marginally increase and the degree of non-cooperation decreases, but we do not have sufficient information to confirm this conjecture. However, this observation is consistent with other research which has observed that these same fishermen possessed an increasing aversion rate in the waning days of this fishery prior to the time period studied in our analysis (Abbott and Wilen 2008, in press).

Combined, these results suggest that as the fishing season progresses within the yellowfin sole fishery, fishermen tend to avoid bycatch less. This behavior is rational because halibut and yellowfin sole inhabit the same regions of the Bering Sea and fishing in areas with a higher halibut concentration could decrease the fisherman's cost per unit of harvest. Furthermore, the conditional cooperation results illustrate that if other fishermen are not willing to avoid high-bycatch regions then others will follow suit. However, it is important to note that the model, by allowing for heterogeneous preferences, illustrates that some portion of the fleet continues to avoid areas with high-bycatch rates despite the conditionality of cooperation.

The elasticity results for the flatfish fishery illustrate that the mean aversion rates are larger than in the yellowfin sole fishery. When the remaining bycatch quota is 1 the mean aversion rate is -3.3% in binding years and -4.1% in non-binding years. This is nearly ten orders of magnitude greater than in the yellowfin sole fishery. Although this would suggest that fishermen in the flatfish possess a high degree of aversion, the elasticity estimates are also much more heterogeneous than those observed in the yellowfin sole fishery. The confidence interval for the first level of remaining bycatch, value of 1, is between -30% and 1.3% for the binding years and -54% and 1.5% on non-binding years. Although this spread does narrow as the remaining bycatch quota decreases it still remains quite large relative to that observed in the yellowfin sole fishery.

Given the large confidence intervals, it is evident that the mean elasticity estimates are not statistically significant from zero. However as was conducted in the yellowfin sole fishery, we can estimate the distributional mass that lies above zero to measure the degree of cooperation present in the fishery and the presence of conditional cooperation. Only two behavioral characterizations of the yellowfin sole fishery apply to the behavior of fishermen within the flatfish fishery: (1) the degree of cooperation deteriorates as the season progresses and (2) when the termination of the fishery is eminent the rate of aversion, and therefore the degree of cooperation, increases. The simulation shown in Fig. 3 graphically illustrates the degree of non-cooperation within the flatfish fishery.

The first generalization observed in the yellowfin sole fishery is not observed in the flatfish fishery because the degree of non-cooperation



**Fig. 3.** Simulated degree of non-cooperation within the flatfish fishery conditional on a given level of bycatch quota remaining.

appears to not vary substantially between the binding and non-binding models. In fact the absolute differences between the two models never exceed one percentage point, indicating that we do not observe different avoidance behavior in the flatfish fishery in binding than non-binding years. However, the deterioration of cooperation within the fishery is observed in the flatfish fishery.

The degree of non-cooperation starts at roughly 33% for both binding and non-binding years and increases to nearly 43%, before falling slightly in the penultimate days of the fishery. Comparing this rate of change to that observed within the yellowfin sole fishery illustrates that the flatfish fishery exhibits a more stable and lower level of non-cooperation. Although the degree of cooperation is larger than observed in the experimental literature (Ostrom 2000), the decay in cooperation is consistent with Ostrom's (2000) discussion on conditional cooperation. The final common generalization observed, aversion rates marginally increasing at the very end of the season, is similarly consistent with Abbott and Wilen's (2008, in press) research on this fishery in an earlier time period. One possible explanation for this phenomenon is that fishermen begin to alter their species targeting behavior to shift away from flatfish toward yellowfin sole and visa versa in the yellowfin sole fishery when it becomes evident that the fishery will be shut down. When this happens their targeting behavior would gravitate toward that observed at the beginning of the season for the alternate fishery, where aversion rates are higher.<sup>21</sup> However, this phenomenon cannot be readily explained using our empirical specification and one would need to further dissect the data set to focus on this target switching behavior to confirm this conjecture. This is beyond the scope of this research effort but something that we wish to investigate in more detail in the future.

## 5. Conclusion

The impact of information on the sustainability of cooperation in the provision of a public good while harvesting a common property resource was largely consistent with results arising from the experimental economics literature, most notably Ostrom's (2000) notion of conditional cooperation. In both fisheries that we examined, we found that the level of cooperation generally fell as the season progresses and as the bycatch TAC approaches the cap. These changes are more pronounced in the yellowfin sole fishery, whereas cooperation rates are higher in the flatfish fishery. Within the experimental literature it has been illustrated that a higher marginal rate of substitution (MRS) for the

<sup>&</sup>lt;sup>21</sup> Recall that the mean aversion rates for the yellowfin sole fishery when the remaining bycatch quota is 1 were roughly 26% and 40% for non-binding and binding years respectively.

private versus the public good enhances the degree of cooperation (Isaac et al., 1984). Therefore, the MRS may be greater in the flatfish fishery than the yellowfin sole fishery. Overall our results illustrate that information provision alone, with moderate peer pressure, cannot completely overcome collective action problems. Despite the presence of these factors substantial externalities still exist within the fishery, resulting from the common-pool nature of the bycatch TAC, which preclude the efficient utilization of these resources.

In general our results indicate that fishermen predominately avoid regions with historically high-bycatch rates early in the season. However, as the season progresses, fishermen reduce their degree of aversion. This reduction is greatest in the yellowfin sole fishery where a fair number of fishermen gravitate toward those regions that possess high-bycatch rates. This suggests that these fishermen may be utilizing the bycatch information to enhance their production later in the season, presumably because the target species and halibut are complements in production. Alternatively, given that the flatfish fishery season is shorter than the yellowfin sole season, there may be less time available for conditional cooperation to develop in the repeated game fishermen play.

While this analysis supports the economically consistent arguments of changing levels of non-cooperation during the season and between binding and non-binding years, there is also the possibility that what we are observing is due to changing intra-seasonal levels of halibut abundance. However, since the only biological information that we have is trawl surveys conducted every 1–3 years, this is not a proposition that can be tested in this study but should be noted as a potential alternate explanation for the behavior exhibited in the fishery.

This analysis generally supports the hypothesis that Sea State has been successful at helping fishermen within the yellowfin sole and flatfish fisheries avoid bycatch. However, there does appear to be a substantial opportunity to increase economic efficiency in these fisheries because perfect cooperation is not observed. With the recent rationalization of this fishery by the NPFMC under Amendment 80 of the Bering Sea Fishery Management Plan, there will be progress toward reducing this economic inefficiency. However, complete efficiency is unlikely to be obtained unless fishermen participating in the flatfish fisheries are allowed to own halibut quota and vice versa. With Amendment 80, the common-pool nature of the bycatch TAC has been minimized for the rationalized vellowfin sole and flatfish fisheries that now have the majority of the fleet functioning with individual bycatch allocations. More information will be available over the next few years about bycatch behavior and values under the rationalized fishery and this may prove to be a fruitful area of research in the future and one which we intend to pursue.

#### Acknowledgements

We would like to acknowledge the financial support of the NMFS Office of Science and Technology, data support from Terry Hiatt of the NMFS Alaska Fisheries Science Center (AFSC), and invaluable help from Andy Smoker and Mary Furuness of NMFS Alaska Regional Office on properly understanding how PSC-allocations change within seasons. We also appreciate helpful input from AFSC researchers Tom Wilderbuer, Paul Spencer, Dan Lew, and Ron Felthoven and from Karl Haflinger of Sea State, Inc. Input from other members of the fishing industry was also essential to our understanding of the dynamics in these fisheries. The corresponding author would also like to acknowledge the Property and Environment Research Center (PERC) in Bozeman, MT for supporting this research.

#### References

- Abbott, J.K., Wilen, J.E., 2008. Dissecting the tragedy: a spatial model of behavior in the commons. Working Paper, School of Sustainability. Arizona State University.
- Abbott, J.K., Wilen, J.E., in press. Voluntary Cooperation in the Commons? Evaluating Sea State with Reduced Form and Structural Models. Land Economics.
- Apesteguia, J., Maier-Rigaud, F.P., 2006. The role of rivalry: public goods versus common-pool resources. Journal of Conflict Resolution 50 (5), 646–663.
- Bisack, J.G., Sutinen, K.D., 2006. Harbor porpoise bycatch: ITQs or time/area closures in the New England gillnet fishery. Land Economics 82 (1), 85–102.
- Bockstael, N.E., Opaluch, J.J., 1983. Discrete modeling of supply response under uncertainty: the case of the fishery. Journal of Environmental Economics and Management 10, 125–137.
- Curtis, R., Hicks, R.L., 2000. The cost of sea turtle preservation: the case of Hawaii's pelagic longliners. American Journal of Agricultural Economics 82, 1191–1197.
- Eggert, H., Tveterås, R., 2004. Stochastic production and heterogeneous risk preferences: commercial fishers gear choices. American Journal of Agricultural Economics 86, 199–212.
- Eales, J., Wilen, J.E., 1986. An examination of fishing location choice in the pink shrimp fishery. Marine Resource Economics 4, 331–351.
- Gilman, E.L., Dalzell, P., Martin, S., 2006. Fleet communication to abate fisheries bycatch. Marine Policy 30 (4), 360–366.
- Gallaway, B.J., Cole, J.G., 1999. Reduction of juvenile red snapper bycatch in the U.S. Gulf of Mexico shrimp trawl fishery. North American Journal of Fisheries Management 19, 342–355.
- Gerber, J-D, Knoepfel, P., Nahrath, S., Varone, F., 2009. Institutional resource regimes: towards sustainability through the combination of property-rights theory and policy analysis. Ecological Economics 68, 798–809.
- Haynie, A.C., 2005. The Expected Profit Model: A New Method to Measure the Welfare Impacts of Marine Protected Areas. Ph.D. dissertation, University of Washington.
- Holland, D.S., Sutinen, J.G., 2000. Location choice in New England trawl fisheries: old habits die hard. Land Economics 76, 133–149.
- Isaac, R.M., Walker, J., Thomas, S., 1984. Divergent evidence on free riding: an experimental examination of possible explanations. Public Choice 43 (1), 113–149.Krinsky, I., Robb, L., 1986. On approximating the statistical properties of elasticities. The Review of Economics and Statistics 68, 715–719.
- Krinsky, I., Robb, L., 1990. On approximating the statistical properties of elasticities: a correction. The Review of Economics and Statistics 72, 189–190.
- Larson, D.M., House, B.W., Terry, J.M., 1996. Toward efficient bycatch management on multispecies fisheries: a nonparametric approach. Marine Resource Economics 11 (3), 181–201.
- Larson, D.M., House, B.W., Terry, J.M., 1998. Bycatch controls in multispecies fisheries: a quasi-rent share approach to the Bering Sea/Aleutian Islands midwater trawl pollock fishery. American Journal of Agricultural Economics 80 (4), 778–792.
- Ledyard, J.O., 1995. Public goods: a survey of experimental literature. In: Kagel, J.H., Roth, A.E (Eds.), Handbook of Experimental Economics. Princeton University Press, Princeton, NJ, pp. 111–194.
- McFadden, D., 1978. Modeling the choice of residential location. In: Korlquist, A. (Ed.), Spatial Interaction Theory and Residential Location. North Holland, Amsterdam, pp. 75–96.
- McFadden, D., 1974. Conditional logit analysis of qualitative choice behavior. In: Zarembka, P. (Ed.), Frontiers in Econometrics. Academic Press, New York, NY.
- McFadden, D., Train, K., 2000. Mixed MNL models for discrete response. Journal of Applied Econometrics 15 (5), 447–470.
- Mistiaen, J.A., Strand, I.E., 2000. Location choice of commercial fishermen with heterogeneous risk preferences. American Journal of Agricultural Economics 82, 1184–1190.
- Olson, M., 1970. Logic of Collective Action: Public Goods and the Theory of Groups. Harvard University Press.
- Ostrom, E., 2000. Collective actions and the evolution of social norms. Journal of Economic Perspectives 14 (3), 137–158.
- Pascoe, S., Revill, A., 2004. Costs and benefits of bycatch reduction devices in European brown shrimp trawl fishery. Environmental and Resource Economics 27, 43–64.
- Reithe, S., Aschan, M.M., 2004. Bioeconomic analysis of by-catch of juvenile fish in the shrimp fisheries an evaluation of the management procedures in the Barents Sea. Environmental and Resource Economics 28, 55–72.
- Reithe, S., 2006. Marine reserves as a measure to control bycatch problems: the importance of multispecies interactions. Natural Resource Modeling 19, 221–242.
- Sandler, T., Arce,, M.D.G., 2003. Pure public goods versus commons: benefit-cost duality. Land Economics 79 (3), 355–368.
- Smith, M.D., Wilen, J.E., 2003. Economic impacts of marine reserves: the importance of spatial behavior. Journal of Environmental Economics and Management 46, 183–206.Smith, M.D., 2005. State dependence and heterogeneity in fishing location choice.
- Journal of Environmental Economics and Management 50, 319–340.
  Train, K.E., 2003. Discrete Choice Methods with Simulation. Cambridge University Press,
- Cambridge, UK.

  Without D. Paurule, C. 1997. A brief biotomy of byeatch management management for
- Witherell, D., Pautzke, C., 1997. A brief history of bycatch management measures for Eastern Bering Sea groundfish fisheries. Marine Fisheries Review 59, 15–22.