

# Conversion from forest to agriculture in the Brazilian Amazon from 1985 to 2021

Hugo Tameirão Seixas<sup>1, 2</sup>, Hilton Luis Ferraz da Silveira<sup>2</sup>, Alan Falcão<sup>2</sup>, Fabiana Da Silva Soares<sup>2</sup>, and Ramon Felipe Bicudo da Silva<sup>1, 3</sup>

<sup>1</sup>Center for Environmental Studies and Research, State University of Campinas, Rua dos Flamboyants 155, Brazil

<sup>2</sup>Embrapa Territorial, Av. Soldado Passarinho 303, Brazil

<sup>3</sup>Center for Systems Integration and Sustainability, Department of Fisheries and Wildlife, Michigan State University, East Lansing, MI, 48823, USA

**Correspondence:** Hugo Tameirão Seixas (seixas.hugo@protonmail.com)

**Abstract.** Land-use and land-cover changes in the Amazon biome are main processes that influence the environment and societies at local, national, and global scales. A large quantity of studies already relied on land thematic maps to analyze change processes. This study brings a new dataset created by calculating the time necessary for deforested areas to change to agriculture (annual and permanent crops) in the Brazilian Amazon biome. The calculations were performed over thematic maps of high resolution by using simple algebra and recursion to identify every conversion from forest to agriculture between 1985 and 2021. The new data can be useful in interdisciplinary studies about land-use and land-cover change analysis in Brazil, such as exploring its drivers and evaluating the effects on natural vegetation cover and ecosystem services. The major innovation brought by this study are links between the deforestation year of a given pixel with its respective year of agriculture establishment, which can provide new insights to understand long-term land-use conversion processes in tropical ecosystems.

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## 1 Introduction

Brazilian policies since the 1960s were focused on the expansion of the agricultural frontier in the Amazon, specially focused in fostering economic growth and national security of the territory (Carvalho et al., 2002; McDonald, 2003; Banerjee et al., 2009). Such a development model led to the construction of thousands of kilometers of roads, and the establishment of large scale crop and livestock farms (Carvalho et al., 2002; Banerjee et al., 2009). This period was marked by high rates of deforestation of mature forests (Fearnside, 2005). At the same time, cattle and soybean production started to move from Southern Brazil into the Midwest and Northern parts of the country (Bicudo Da Silva et al., 2020), in which soybean cultivation areas followed cattle expansion (i.e., pastureland expansion) over the Amazon forest (Simon and Garagorry, 2005; Barona et al., 2010; Arima et al., 2011). The pattern of extensive livestock production after deforestation, followed by the establishment of annual crops after a given period is typical in the Amazon biome (Barona et al., 2010), and persists nowadays. This process can take several

decades to be accomplished, but sometimes less than one year. In this last case, the land change process can be considered as a direct conversion (Morton et al., 2006) (i.e., from forest to agriculture).

Land-use and land-cover (LULC) changes are known to cause impacts on different scales. Deforestation and LULC changes can affect hydrological processes in large basins (Arias et al., 2018); reduce the forest resilience to extreme events (e.g., drought) and other sources of perturbation (Boulton et al., 2022); deforestation and farming practices can reduce convective rainfall and increase surface temperature (Maeda et al., 2021); increase carbon emissions (Gatti et al., 2021); cause important impacts on fauna and flora diversity, soil properties and carbon stocks (Nunes et al., 2022; Rittl et al., 2017); and even negatively impact public health (Ellwanger et al., 2020).

Data on LULC classification are evolving rapidly in the last decade. Initiatives such as MapBiomas (Souza et al., 2020), launched in 2015, provides high resolution, annual classifications for the Brazilian territory. Such types of data are essential to analyze LULC change processes (Bicudo da Silva et al., 2020), make causal inferences (Capmourteres et al., 2018; Gelo and Turpie, 2022; Odongo et al., 2014; Larned et al., 2020), and to understand what drives and what are the impacts of the transformation of the landscapes in Brazil (Nunes et al., 2022). In this context, this study aims to quantify and characterize the length (time span) of the conversions from forest formations to agriculture in the Amazon over the last 36 years. This data can be useful to the development of interdisciplinary research, since it can provide useful insights into the environmental, economic, and social impacts of land conversion. For example, this data can be used to develop insights into the economic costs and benefits associated with land conversion, the effects of land conversion on local ecosystems, and the impact of land conversion on livelihoods and social dynamics.

## **2 Methods**

### **2.1 Conversion length calculation**

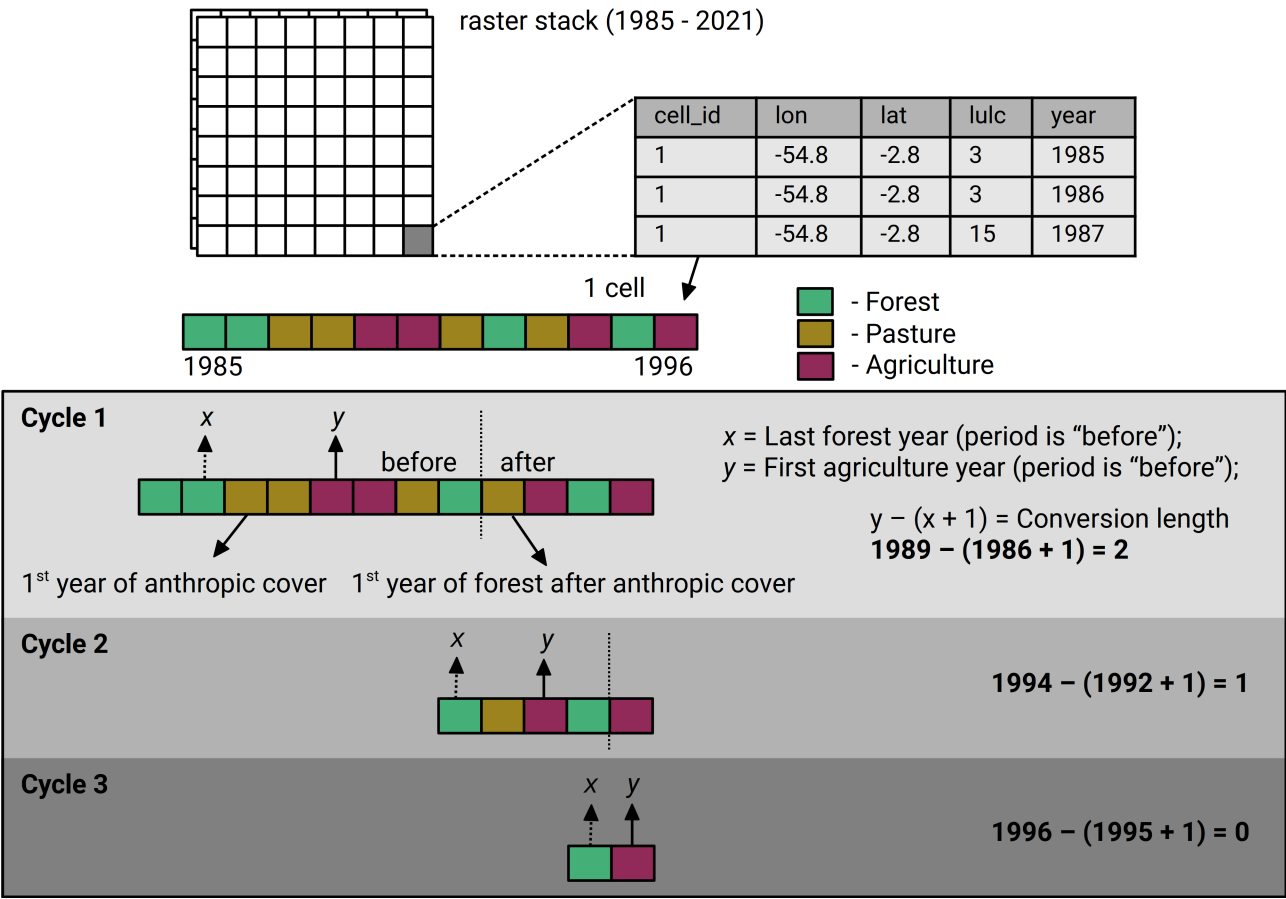
The estimations of conversion from forest to agriculture were performed for the Amazon biome region (biome limits defined by the Brazilian Institute of Geography and Statistics (IBGE), in 2019). Transitions were calculated using LULC classification data from MapBiomas, collection 7, which ranges from 1985 to 2021, at a spatial resolution of approximately 30 meters, and global accuracy at 91.3%.

The MapBiomas data was downloaded from Google Earth Engine platform (Gorelick et al., 2017). The pixels of the Amazon biome were filtered to contain only those that were occupied by forest and agriculture at some period, and that are not considered as water, according to the Global Surface Water product provided by Copernicus (Pekel et al., 2016). For the download, the data was divided in 2732 tiles, to allow local processing of a large quantity of pixels. Two sets of data were downloaded with the same spatial dimensions, one with the LULC classifications for each year, and another with binary values, representing whether that pixels were valid for processing or not (according to the filter conditions mentioned above).

The first part of the processing was to create a table from the binary values data, and create unique identification for each pixel, their coordinates, calculate the area, and also the municipality in which each pixel was inside. In this part, a table was

also generated with the spatial information of each tile, so that it is possible to convert the generated tables back to raster format.

The second part of the processing was to load each tile separately, convert it to a table and calculate every possible conversion from forest to agriculture (Figure 1). The conversions were calculated pixel by pixel, where the LULC time series were extracted and analyzed for each conversion cycle that was present for that pixel. The calculation of the conversion length was simple, it is the year of the establishment of agriculture subtracted by the last year classified as a forest, however, we add one year to the latter, so that it represents the year of deforestation (since it is more meaningful for analysis).



**Figure 1.** Representation of the process to calculate the conversion length from forests to agriculture. All values presented are hypothetical, and serve only as an illustration of the method.

The process of calculation of the conversion length (Figure 1), can be divided into seven steps of processing:

1. Load raster and extract valid values into a table;
2. Calculate the year of first occurrence of any agriculture class (annual and permanent crops) for each pixel;

3. Calculate the first year of “Forest Formation” LULC class after the year calculated in step 2. This step identifies the occurrence of more than one conversion in a given pixel;
4. Classify rows as “before” or “after” the occurrence of the year calculated in step 3. This is the identification of periods before and after a return of forest after a first conversion to agriculture;
5. Calculate the last year of “Forest Formation” within the rows classified as “before”, and add 1 year to represent the deforestation year;
6. Calculate the first year of any agriculture type class within the rows classified as “before”, for each pixel;
7. Calculate the difference between years from items 5 and 6 to get the LULC conversion length in years, for each pixel;

The steps 2 to 7 are performed recursively to identify multiple conversions, in case they are present. In addition to the conversion length values and the years of the conversion, they are also qualified by the type of agriculture that were established after deforestation, if the conversion occurred from primary or secondary forests, and the number of occurrences of conversions at the pixel. The LULC classes within the conversion and until 5 years after were also stored.

The conversion results are stored in a data set of tabular files, where data can be queried for further analysis. After the calculations, data was also converted back to raster format, with the aim to expand the accessibility of the data set.

## 2.2 Description of data collection

After the calculation of conversions, the results are stored in three different types of tables (Figure 2), organized in a folder structure and stored as Apache Parquet files.

The first type of table (Figure 2.a) contains the spatial information (longitude, latitude and area) of each pixel, its unique identification and the code of the municipality which contains the pixel. The “cell\_id” column contains a unique identification for the whole data, while the “tile\_cell\_id” holds a unique value of each pixel for a single tile. Each tile is also uniquely identified in the column “tile\_id”. This table is named “mask\_cells”.

A second type of tables (Figure 2.b) stores conversion length values, the first and last year of the conversion, the resulting agriculture type, and the number of the conversion cycle. They also contain the unique id of each pixel and tile (related to the table above). The “agri\_code” column holds information about the type of agriculture the was established after deforestation. The “forest\_year” have values that represent the year of deforestation, the “agri\_year” is the year when the agriculture establishment happened, and the subtraction between both (the conversion length) is stored in the “c\_length” column. Since it is possible that one pixel contains more than one conversion from forest to agriculture, we stored the number of the conversion in the column “c\_cycle” (1 is the first conversion cycle). However, conversion from forest to other LULC is also possible, so the column “f\_cycle” holds information about the number of the conversion from forest to any LULC. If the forest has suffered any conversion, it starts being considered as a secondary forest, as indicated in the column “forest\_type”.

cell_id	agri_code	forest_year	agri_year	c_length	f_cycle	forest_type	tile_id	c_cycle
36869	20	1986	1989	3	1	1	1174	1
36870	41	2002	2004	2	2	2	1174	1
36871	39	1987	2017	30	1	1	1174	1

cell_id	tile_cell_id	area	code_muni	coords	tile_id
36869	89	875.	5104526	POINT ()	1174
36870	90	875.	5104526	POINT ()	1174
36871	91	875.	5104526	POINT ()	1174

cell_id	year	class_code	tile_id	c_cycle
36869	1987	15	1174	1
36869	1988	15	1174	1
36869	1989	20	1174	1

**Figure 2.** Representation of the tables generated by the conversion length calculation. All values presented are hyphothetical, and serve only as an illustration of the method.

The third type of tables (Figure 2.c) contains the LULC classes of all years within the conversion, and also the first 5 years after the conversion. The column “year” holds the year of the MapBiomass classification, and the “class\_code” variable stores the LULC classification (according to MapBiomass) for the respective year.

The three tables are related to each other and can be used altogether, and are separated by tiles. Another table containing the metadata of each tile is also created, and holds the spatial characteristics of the tiles. With this spatial information, it is possible to convert the tabular data back to spatial raster, with identical spatial properties as the original MapBiomass classification data.

The tabular data was organized in a system of folders and sub folders in a hive format partitioning (Figure 3). This structure allows for easily accessing parts of the data set, and also to filter it without the necessity to read the tables. In the case of this data set, it is partitioned by the “tile\_id” and “agri\_cycle” variables. The data set also has a metadata file with the spatial characteristics of each file, it stores the identification of the tile, the maximum and minimum latitude and longitude, the coordinate reference system, and the number of columns, rows, and cells of each tile.

### 2.3 Accuracy assessment

Accuracy assessment was performed by visual inspection of annual composites of Landsat images from MapBiomass. We selected 100 random points to be analyzed. An area of approximately 4 square kilometers around the sample point was used in the visual inspection of satellite images.



**Figure 3.** Representation of the data set, structured in folders and sub folders.

The visual inspection used several variables derived from the Landsat historical collection. The median of Red, Green, Blue, Near Infrared (NIR) and Short Wave Infrared (SWIR1) from dry and wet season were used, and also the annual amplitude of the Normalized Difference Vegetation Index (NDVI). The process of accuracy assessment was performed in a Shiny app, and was conducted without any consultation to the conversion length results. In the validation app, we estimated, by visual inspection, the year of deforestation and the year of agricultural establishment. The observed conversion length was obtained by subtracting both dates.

To evaluate the accuracy, we calculated the Mean Absolute Error (MAE), the Bias (BIAS), and the Percent Bias (PBIAS). We also analyzed the results by plotting the errors as frequency bars, and scatter plots between observed and estimated values. After the completion of the analysis of the 100 sample points, we also conducted a qualitative assessment, where we compared our results with satellite images composites.

### 3 Results and discussion

The conversion calculations show that in the Brazilian Amazon biome, 64874 square kilometers of forests were converted to agriculture between 1985 and 2021. The length of the conversions can go from 0 to 35 years, in which conversions closer to 0 years are considered as fast conversions, and conversions closer to 35 years are considered as slow conversions. Transitions of 0 years are considered “direct” conversions, where there was no presence of pasture before the establishment of agriculture. Our estimations show that around 9.2% of the conversions were “direct”.

Our conversions calculations found pixels that presented up to six conversions from forest to agriculture since 1985, as these numbers are unlikely to happen (they are distributed as sparse pixels, and do not show patterns of real conversions, being common at borders), we proceeded the results analysis only in the first conversions found in each of these pixels.

Although we named the year of change from forest to anthropic land uses as deforestation, we acknowledge that it is not a direct measurement of deforestation (such as PRODES), however it represents a proxy to deforestation.

#### 3.1 Transition patterns

Transitions from forest to agriculture can be found in almost every region in the Amazon, but is mostly concentrated in clusters, specially in the south and east of the biome, in the states of Mato Grosso, Pará and Maranhão (Figure 4).

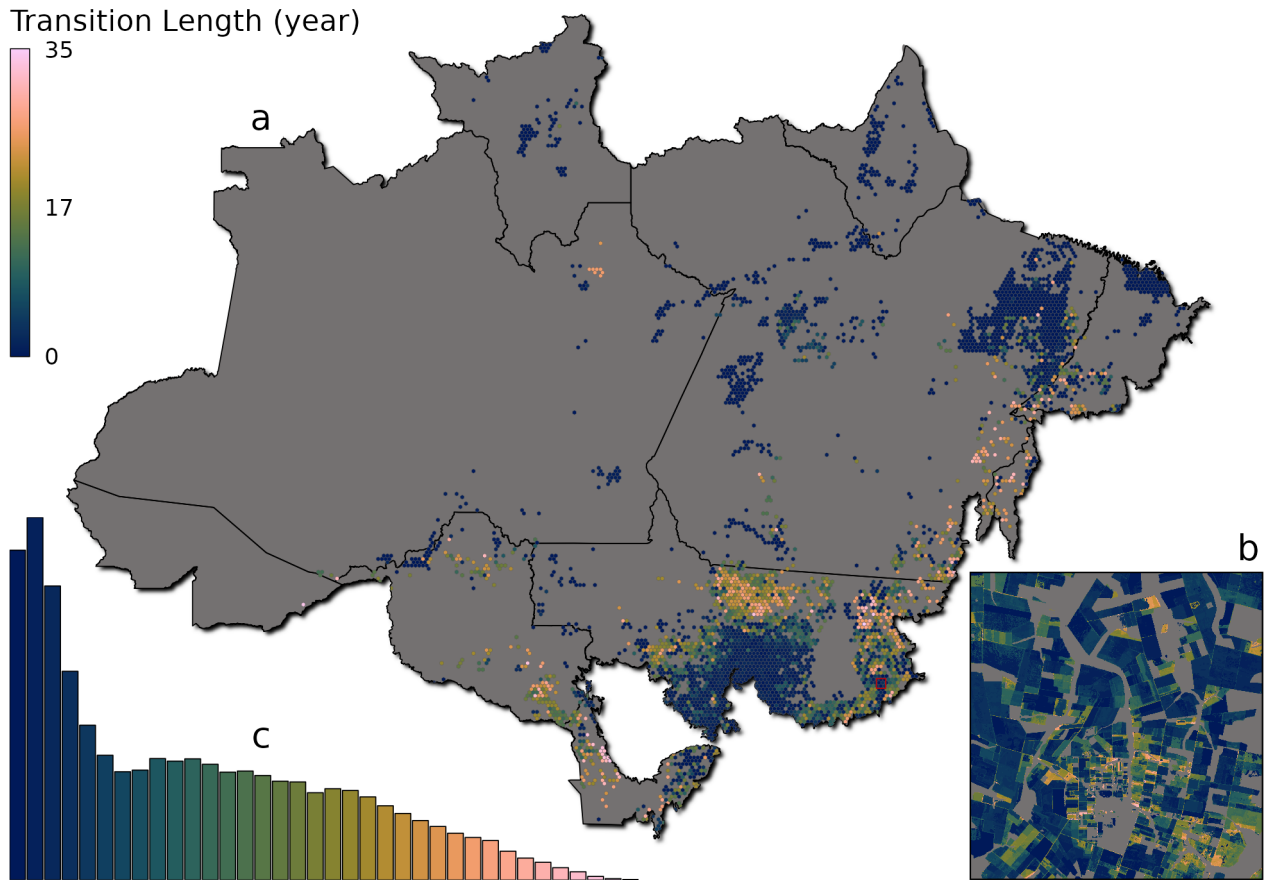
Well defined clusters can be observed in the map created with aggregated conversion length data (Figure 4). Slow conversion areas tend to concentrate in specific regions of the biome, while fast conversions seem to have a wider distribution, but also tend to form spatial clusters. However, this pattern does not hold completely when observing the data at its original scale (Figure 4), where areas with different conversion lengths are mixed between each other. When observing at the original scale, we could not spot any well defined pattern or direction of the occurrence of faster to slower conversions (Figure 4). Other studies investigating patterns of conversions in the Amazon also found the formation of clusters of patterns, although there is a big heterogeneity at larger scales (Müller-Hansen et al., 2017).

The data can be analyzed year by year, and also be separated by primary and secondary forests being converted to agriculture (Figure 5).

The deforestation area of primary forests increased largely from 1986 to 2003, which was followed by a steep decrease until 2009, when the deforested areas reached a stable rate (Figure 5.a). From 1985 to 1995, most deforested areas suffered a slow conversion, mostly were higher than 10 years (76 %). After this period, fast conversions started to become more common (Figure 5.a), specially from 2002 to 2004, where 70 % of the transitions were faster than 10 years. Our results show that in 2003 (the peak of deforestation), 12.7% of the conversions were a direct conversion from forest to agriculture. Previous studies estimated a proportion of 23 % in the Mato Grosso state, for the same year (Morton et al., 2006). If we include fast conversions (from 0 to 2 years) from our results, the proportion jumps to 48.1 % of all conversions in 2003.

Agricultural establishment over areas of primary forests peaked in 2005 and 2013 (Figure 5.b). Despite similar agricultural establishment areas in both years, their conversion lengths differ greatly (Figure 5.b), in 2005 most of the conversions were faster than 10 years (78 %), while in 2013 the majority of conversions were slower than 10 years (66 %). In 2010, it was

## Distribution of transitions from forest to agriculture in the Amazon

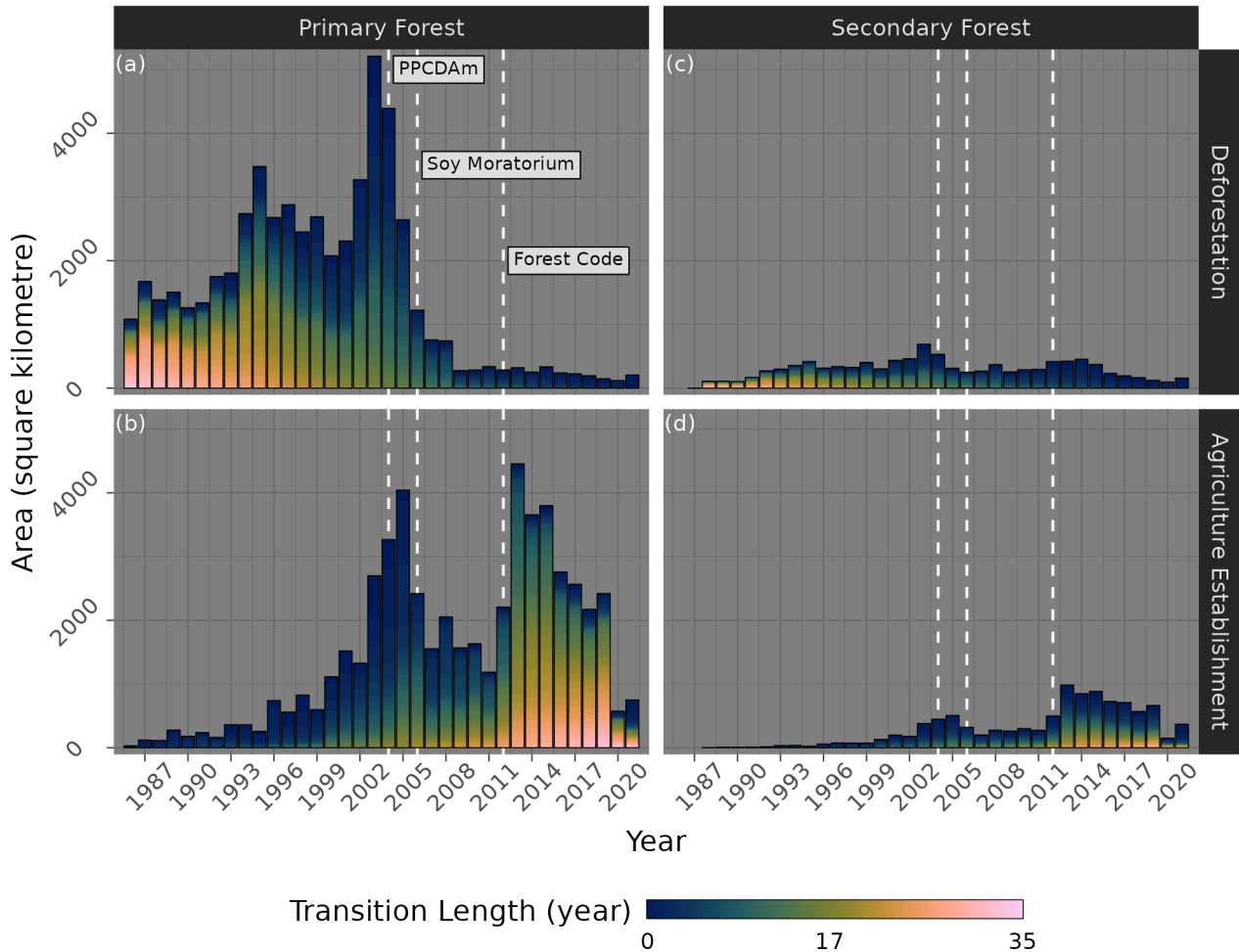


**Figure 4.** Map of distribution of conversions from forest to agriculture in the Brazilian Amazon biome. The hexagonal cells represent the most common conversion length, and do not reflect the amount of area of conversions inside a cell. Transitions are concentrated in the south (Mato Grosso state), and in the east (Pará and Maranhão states). The conversion length ranges from 0 (dark blue) to 35 years (light pink), and clusters of fast conversions (conversions closer to 0 year) can be discerned from clusters of slow conversions (conversions closer to 35 years). The histogram located in the bottom left shows that fast conversions are more common than slower conversions. The zoomed map in the bottom right shows the results in finer resolution, where it is possible to observe different conversion lengths between properties. The red square shows the extent of the zoomed map.

observed a shift in the conversion length from forest until the establishment of agricultural areas. From this year onward, 68 % of the conversions happened in areas deforested at least 10 years before (Figure 5.b). Even after the decrease of deforestation after 2002, areas under agriculture expanded over lands where deforestation happened before 2002 (mostly over pastureland). However, after 2019, a sudden drop of agriculture establishment rate happened (Figure 5.b). This behavior could be related to



## Transition area per year and transition length



**Figure 5.** Transition area from forest to agriculture, per year and conversion length. The bars represent the total amount of area at some state of the conversion for each year. The color gradient in each bar represent the conversion length related to a deforestation or an agriculture establishment event. Blue tones represents fast conversions (closer to 0 years), pink tones represent slow conversions (closer to 35 years). Transition events were separated by deforestation of primary forests (a) and secondary forests (c), and the subsequent agriculture establishment of primary forests (b) and secondary forests (d).

economic causes or even the scarcity of suitable areas for agricultural expansion, which seems unlikely, since the last years have been marked by an increase of deforestation rates and the advance of pastureland over the forest.

The causes of deforestation and agricultural establishment in the Brazilian Amazon are complex and diverse. Political context, public policies, market prices and law enforcement can influence how these processes evolve over time. The end of the

1980s and the 1990s were marked by the development of policies for protection of the environment, with the creation of the National Environmental Policy, the establishment of the Brazilian Institute of Environment and Natural Resources (IBAMA), and the Ministry of Environment (Banerjee et al., 2009). However, this new policy environment was not immediately translated into a significant deforestation decrease in the Amazon, which remained at high levels until the beginning of 2000s. Our results show that deforestation in this period (1986 - 2000) was mainly followed by pasture, which remained for a long time before being converted to agriculture (Figure 5.a).

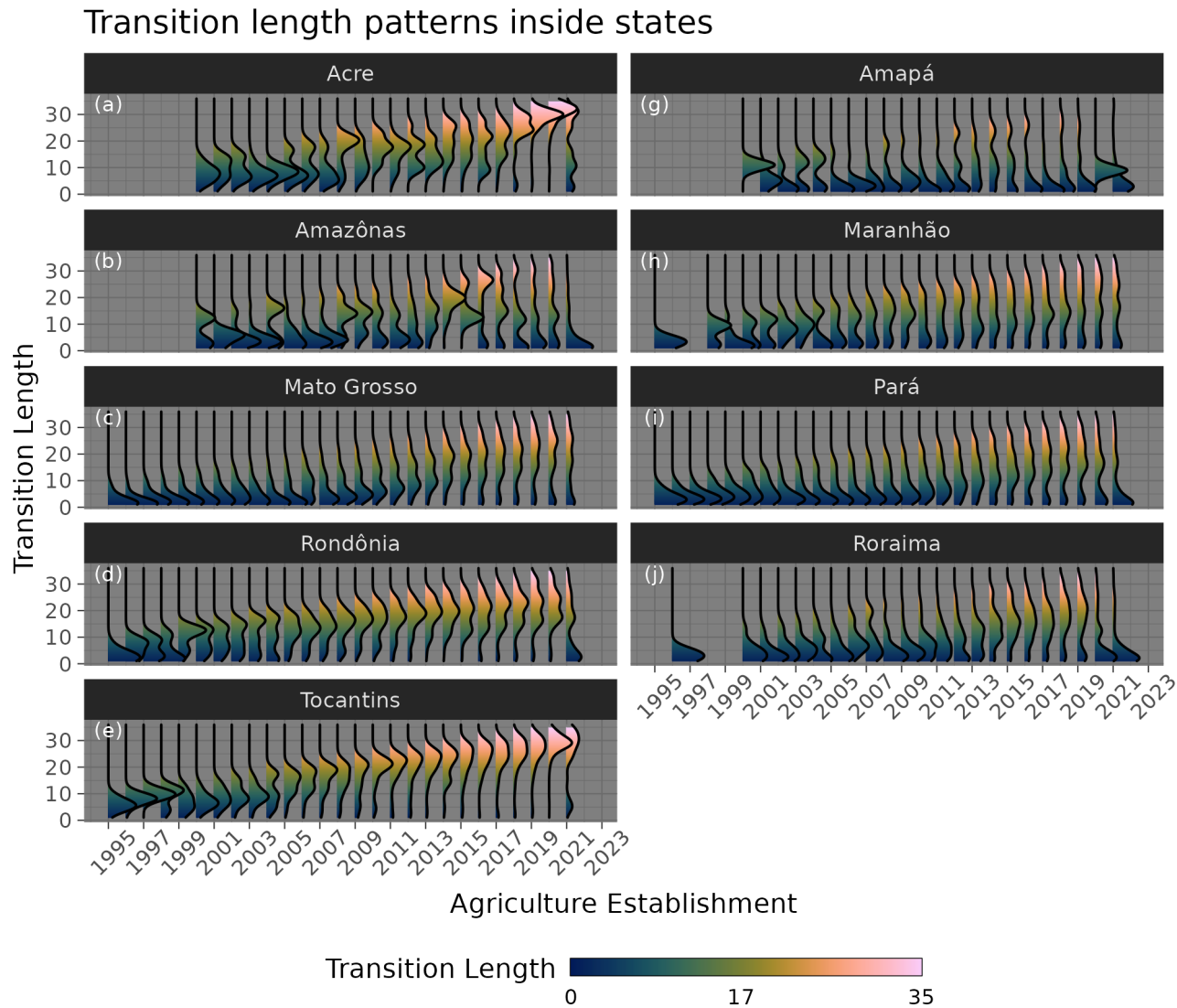
Apart from protectionist policies created in the 1990s, development programs kept pressuring forests in the Amazon. In the late 1990s, the development policies “Brasil em Ação” and “Avança Brasil” (1995 - 2003) accelerated the national infrastructure expansion, including in the Amazon (Carvalho et al., 2002). This is a period when the deforestation areas reached the highest values in the time series (Figure 5.a), and also a peak in the establishment of agriculture in the Amazon biome, predominantly with fast conversions (Figure 5.b).

In 2004, the Brazilian Government launched the Action Plan for the Prevention and Control of Deforestation in the Legal Amazon (PPCDAm), which was composed of many initiatives to curb deforestation (West and Fearnside, 2021). The PPCDAm was considered as a successful policy to slow deforestation rates in Brazil, with international recognition. Our calculations from MapBiomas data reinforces the correlation of the PPCDAm with the reduction of deforestation after 2004 (Figure 5.a), and reduction of the agriculture establishment after 2005 (Figure 5.b).

In 2006, the Brazilian Association of Vegetable Oil Industries (BIOVE) and the National Association of Cereal Exporters (ANEC) committed to avoid commercialization of soybean grains harvested from areas deforested after 2008, known as the soy moratorium. The soy moratorium is considered to have contributed to the decline of the deforestation directly related to soybean expansion (Paim, 2021; Amaral et al., 2021; Heilmayr et al., 2020; Kastens et al., 2017). Our estimates of conversions show a decrease of deforested areas after 2008, where it reached minimal values (Figure 5.a). After 2006, agriculture establishment over deforested primary forests suffered a decrease, which stayed relatively stable until 2012, where a steep increase occurred, however, the new areas being occupied by agriculture were mainly over areas that were cleared before 2008 (Figure 5.b). The period between 2006 and 2010 is also considered as a turnover point, known as the decoupling of deforestation and soybean production, where the total production was not as dependent on agriculture expansion as before (Macedo et al., 2012). Therefore, the observed increase of agricultural establishment after 2012 shows that agriculture expansion did not halt after the soy moratorium nor after periods of high soybean productivity. Producers started to expand over areas cleared many years ago, which may still have indirect impacts on deforestation increase over distant areas (Arima et al., 2011; Gollnow et al., 2018). However, this indirect effect is not always clearly observed when analyzing the Amazon biome as a whole, our data shows that periods of fast agriculture expansion occurred at the same time with minimal deforestation values, according to PRODES estimates (F. G. Assis et al., 2019). Expansion of agricultural areas also occurred over cleared areas of secondary forests in 2012, but with an important amount of fast conversions (Figure 5.d). The causes of the increase of agricultural establishment areas can be numerous, and one main driver was the approval of a new Forest Code, in 2012, which is considered to have undermined the environmental protection of forests (Kröger, 2017; Pereira and Viola, 2019), since it opened the possibility

of amnesty for illegally deforested areas (Schielein and Börner, 2018; Albuquerque Sant’Anna and Costa, 2021; Soares-Filho et al., 2014), and the possible reduction of legal reserve area (Freitas et al., 2018).

Even though it is possible to analyze the results in the Amazon as a whole, the conversion length patterns across years can change significantly between different states (Figure 6).



**Figure 6.** Transition length patterns inside states in the Amazon biome. Each year has a density estimate of the conversion lengths, represented as colored curves. The peak of the curves represents conversion length values with more frequency in one year of one state. Blue tones represent fast conversions (conversions closer to 0 years), pink tones represent slow conversions (conversions closer to 35 years).

The state of Amapá presented fast conversions along all the time series, without any significant period of slow conversion. In contrast, conversions in Acre were majorly slow, in which only 2021 showed faster conversions. There are two states where the pattern of conversion length are alike, Mato Grosso (MT) and Pará (PA). Both states underwent fast conversions from 1995 to 2005, and after this period slow conversions became predominant until the end of the time series. Rondônia (RO) and Tocantins (TO) presented a similar pattern of MT and PA, in which conversions decelerated, but for RO and TO since the beginning of 2000s, earlier than MA and PA.

The great majority of conversions are from forests to soybean and “Other Temporary Crops”, which represents 95% of the conversions in the Amazon biome (14% of Soybean and 81% of “Other Temporary Crops”). When analyzing what happened after 5 years since the conversion, soybean areas persistence rate was at 79% (i.e., areas that remained as soybean), 14% was converted to “Other Temporary Crops” and 7% was converted to pasture. “Other Temporary Crops” are less persistent. Hence, 27% of these areas remain in the same class, 56% were converted to soybean, and 15% to pasture. Conversions from soybean and “Other Temporary Uses” to other LULC classes (apart from the already cited) are negligible. Sugar cane, cotton and perennial crops also appear, but in a negligible proportion.

### **3.2 Validation**

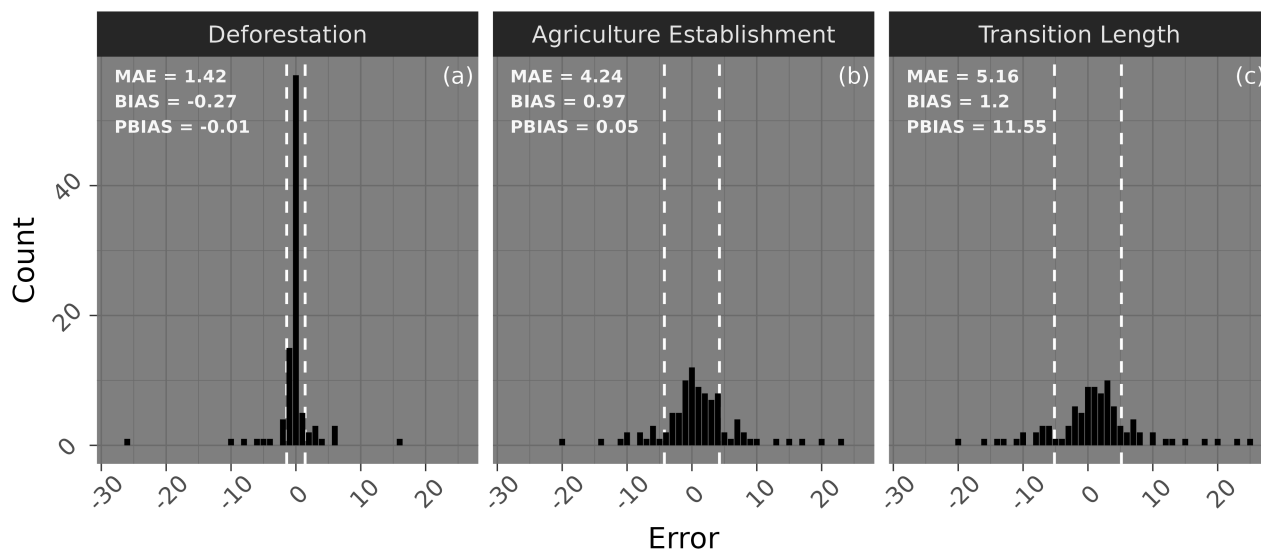
From the 100 random sample points used in the results validation, 1 point (sample 21) was not considered as a conversion from our estimations, however, the visual inspection pointed to a likely event of conversion from forest to agriculture. Also, there were 6 points (sample 6, 7, 55, 56, 78, 97) which the visual inspection did not find a conversion from forest to pasture, although our estimations pointed as conversions. Therefore, 7% of the sample were completely misclassified by our estimates, and the rest of the accuracy assessment was performed over the remaining 93 sample points.

When analyzing the errors from the conversion length estimates, we observed that the year of deforestation shows the least amount of errors (Figure 7). The MAE of the deforestation year is 1.42 years, and the error shows a bias towards underestimation. The year of the agriculture establishment and the conversion length estimates showed larger errors when compared with visual inspection.

The dispersion of observed and estimated values shows no clear pattern of errors (Figure 8). Transition length values are more concentrated in smaller values, which also shows higher errors (Figure 8.c). However, this is expected since faster conversions are more common (Figure 4).

According to accuracy assessment from MapBiomas, the collection 7 presents a global accuracy of 96.6% for the Amazon biome, which is the proportion of pixels that were classified correctly. For the Forest class, MapBiomas showed small errors of inclusion (proportion of pixels misclassified as other classes, but the real class were Forest), which fluctuated around 1%. The omission errors for Forests are also small (proportion of pixels misclassified as Forests, but the real class were not Forest), which fluctuated around 2%. The Agriculture class presented more errors, the inclusion errors ranged from 22% to 5%, and were mostly composed of Forests pixels (forest pixels misclassified as agriculture). The omission errors of Agriculture ranged from 22% to 8%, and were also mostly composed of Forest pixels (agriculture pixels misclassified as forest).

## Distribution of the errors of transition estimates.



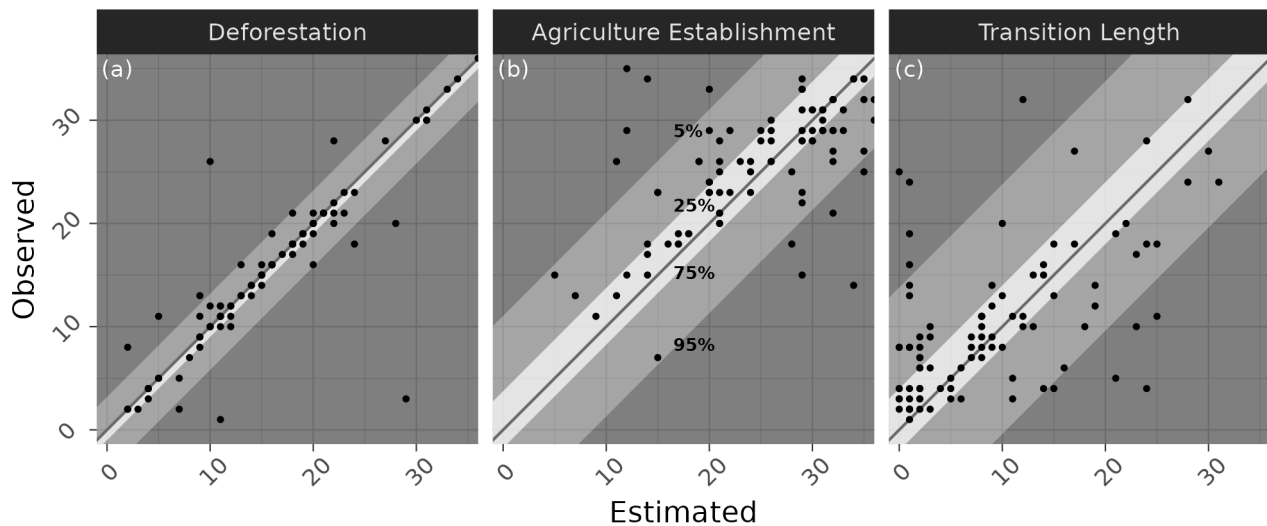
**Figure 7.** Bar with the count of error values (difference between observed and estimated values). Positive values indicate underestimation of the variable (estimations were lower than observations), negative values indicate overestimation (estimations higher than observations). Error metrics (Mean Absolute Error, Bias and Percent Bias) are displayed in the top right position of each box. The white dashed lines represent the MAE values of each variable.

### 3.3 Qualitative assessment

The qualitative assessment of 100 samples was performed to analyze the results in conjunction with the error calculations and metrics. It allows the interpretation of results with more context, since results are compared visually with reference satellite image composites, in an area of 4 square kilometers for each sample.

We found high heterogeneity inside agriculture plots (sample 56), caused by both year of deforestation or year of agriculture establishment. This heterogeneity may be caused by smaller uncertainties within properties, where it would be expected higher homogeneity of conversion length values. In few samples, roads were considered as agriculture (sample 2), so there was in fact no conversion to agriculture. Other inclusion errors were detected, where areas that clearly showed no agriculture were considered as conversions (sample 10, 98). There were also omissions errors, especially agriculture plots not being considered as a conversion as a whole, where parts of a plot that were clearly a conversion to agriculture were not considered as one (sample 16, 30, 75). A lot of uncertainty was identified at borders of agricultural areas (sample 25, 56, 84, 92), which is a very common issue in classification of satellite imagery, since there are the presence of mixed pixels, where different objects share the space of a same pixel (Kaur et al., 2019). We also observed that areas with sparse forest vegetation presented a large error of deforestation year (sample 33).

## Dispersion between observed and estimated transitions.



**Figure 8.** Scatter plot between paired values of estimated and observed Deforestation and Agriculture Establishment years, and their respective Transition Length. The gray line represents values where there would be no errors of the estimated values in comparison with the observed values. Translucid white areas represent quantile ranges (5%, 25%, 75% and 95%) of the errors. The area between 5% and 95% quantiles includes 90% of points. The area between 25% and 75% quantiles includes 50% of points.

We also observed that in many samples, the patterns of agricultural areas were roughly well represented (sample 25, 43, 56, 84), the shape of agricultural plots were clear and it was possible to distinguish different agricultural areas. Some locations present a higher homogeneity of conversions inside agriculture plots (sample 84), which may be related to areas with smaller uncertainties.

## 4 Conclusions

The conversion from forests to agriculture estimates can be a valuable source of information to support research over different disciplines related to LULC changes. We showed that our data relates to many historical events in Brazil, which shaped the landscape of the Brazilian Amazon biome, covering public policies, economic and environmental aspects. The data is structured and stored in a ready to analyze format, which is very convenient to be analyzed in conjunction with economic and demographic data.

Our data also explores new aspects of the MapBiomas classification data, which was not yet analyzed, by explicitly linking deforestation with agricultural establishment. The accuracy assessment creates new opportunities for future improvements of LULC data, especially concerning conversions from forest to agriculture.

. The authors declare no competing interests.

## References

- Albuquerque Sant' Anna, A. and Costa, L.: Environmental regulation and bail outs under weak state capacity: Deforestation in the Brazilian Amazon, *Ecological Economics*, 186, 107 071, <https://doi.org/10.1016/j.ecolecon.2021.107071>, 2021.
- Amaral, D. F., de Souza Ferreira Filho, J. B., Chagas, A. L. S., and Adami, M.: Expansion of soybean farming into deforested areas in the amazon biome: the role and impact of the soy moratorium, *Sustainability Science*, 16, 1295–1312, <https://doi.org/10.1007/s11625-021-00942-x>, 2021.
- Arias, M. E., Lee, E., Farinosi, F., Pereira, F. F., and Moorcroft, P. R.: Decoupling the effects of deforestation and climate variability in the Tapajós river basin in the Brazilian Amazon, *Hydrological Processes*, 32, 1648–1663, <https://doi.org/10.1002/hyp.11517>, 2018.
- Arima, E. Y., Richards, P., Walker, R., and Caldas, M. M.: Statistical confirmation of indirect land use change in the Brazilian Amazon, *Environmental Research Letters*, 6, 024 010, <https://doi.org/10.1088/1748-9326/6/2/024010>, 2011.
- Banerjee, O., Macpherson, A. J., and Alavalapati, J.: Toward a Policy of Sustainable Forest Management in Brazil, *The Journal of Environment and Development*, 18, 130–153, <https://doi.org/10.1177/1070496509333567>, 2009.
- Barona, E., Ramankutty, N., Hyman, G., and Coomes, O. T.: The role of pasture and soybean in deforestation of the Brazilian Amazon, *Environmental Research Letters*, 5, 024 002, <https://doi.org/10.1088/1748-9326/5/2/024002>, 2010.
- Bicudo Da Silva, R. F., Batistella, M., Moran, E., Celidonio, O., and Millington, J. D. A.: The Soybean Trap: Challenges and Risks for Brazilian Producers, *Frontiers in Sustainable Food Systems*, 4, <https://www.frontiersin.org/articles/10.3389/fsufs.2020.00012>, 2020.
- Bicudo da Silva, R. F., Millington, J. D. A., Moran, E. F., Batistella, M., and Liu, J.: Three decades of land-use and land-cover change in mountain regions of the Brazilian Atlantic Forest, *Landscape and Urban Planning*, 204, 103 948, <https://doi.org/10.1016/j.landurbplan.2020.103948>, 2020.
- Boulton, C. A., Lenton, T. M., and Boers, N.: Pronounced loss of Amazon rainforest resilience since the early 2000s, *Nature Climate Change*, 12, 271–278, <https://doi.org/10.1038/s41558-022-01287-8>, number: 3 Publisher: Nature Publishing Group, 2022.
- Capmourteres, V., Rooney, N., and Anand, M.: Assessing the causal relationships of ecological integrity: a re-evaluation of Karr's iconic Index of Biotic Integrity, *Ecosphere*, 9, e02 168, <https://doi.org/10.1002/ecs2.2168>, 2018.
- Carvalho, G. O., Nepstad, D., McGrath, D., del Carmen Vera Diaz, M., Santilli, M., and Barros, A. C.: Frontier Expansion in the Amazon: Balancing Development and Sustainability, *Environment: Science and Policy for Sustainable Development*, 44, 34–44, <https://doi.org/10.1080/00139150209605606>, publisher: Routledge \_eprint: <https://doi.org/10.1080/00139150209605606>, 2002.
- Ellwanger, J. H., Kulmann-Leal, B., Kaminski, V., Valverde-Villegas, J., Veiga, A. B. G. D., Spilki, F. R., Fearnside, P. M., Caesar, L., Giatti, L. L., Wallau, G. L., Almeida, S. E. M., Borba, M. R., Hora, V. P. D., and Chies, J.: Beyond diversity loss and climate change: Impacts of Amazon deforestation on infectious diseases and public health, *Anais da Academia Brasileira de Ciências*, 92, <https://doi.org/10.1590/0001-3765202020191375>, publisher: Academia Brasileira de Ciências, 2020.
- F. G. Assis, L. F., Ferreira, K. R., Vinhas, L., Maurano, L., Almeida, C., Carvalho, A., Rodrigues, J., Maciel, A., and Camargo, C.: TerraBrasilis: A Spatial Data Analytics Infrastructure for Large-Scale Thematic Mapping, *ISPRS International Journal of Geo-Information*, 8, 513, <https://doi.org/10.3390/ijgi8110513>, number: 11 Publisher: Multidisciplinary Digital Publishing Institute, 2019.
- Fearnside, P. M.: Deforestation in Brazilian Amazonia: History, Rates, and Consequences, *Conservation Biology*, 19, 680–688, <https://doi.org/10.1111/j.1523-1739.2005.00697.x>, \_eprint: <https://onlinelibrary.wiley.com/doi/pdf/10.1111/j.1523-1739.2005.00697.x>, 2005.



- Freitas, F. L. M., Sparovek, G., Berndes, G., Persson, U. M., Englund, O., Barretto, A., and Mörtberg, U.: Potential increase of legal deforestation in Brazilian Amazon after Forest Act revision, *Nature Sustainability*, 1, 665–670, <https://doi.org/10.1038/s41893-018-0171-4>, 2018.
- Gatti, L. V., Basso, L. S., Miller, J. B., Gloor, M., Gatti Domingues, L., Cassol, H. L. G., Tejada, G., Aragão, L. E. O. C., Nobre, C., Peters, W., Marani, L., Arai, E., Sanches, A. H., Corrêa, S. M., Anderson, L., Von Randow, C., Correia, C. S. C., Crispim, S. P., and Neves, R. A. L.: Amazonia as a carbon source linked to deforestation and climate change, *Nature*, 595, 388–393, <https://doi.org/10.1038/s41586-021-03629-6>, number: 7867 Publisher: Nature Publishing Group, 2021.
- Gelo, D. and Turpie, J.: The effect of forest land use on the cost of drinking water supply: machine learning evidence from South African data, *Journal of Environmental Economics and Policy*, 11, 361–374, <https://doi.org/10.1080/21606544.2021.2024094>, publisher: Routledge, 2022.
- Gollnow, F., de Barros Viana Hissa, L., Rufin, P., and Lakes, T.: Property-level direct and indirect deforestation for soybean production in the Amazon region of Mato Grosso, Brazil, *Land Use Policy*, 78, 377–385, <https://doi.org/10.1016/j.landusepol.2018.07.010>, 2018.
- Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., and Moore, R.: Google Earth Engine: Planetary-scale geospatial analysis for everyone, *Remote Sensing of Environment*, 202, 18–27, <https://doi.org/10.1016/j.rse.2017.06.031>, 2017.
- Heilmayr, R., Rausch, L. L., Munger, J., and Gibbs, H. K.: Brazil’s Amazon Soy Moratorium reduced deforestation, *Nature Food*, 1, 801–810, <https://doi.org/10.1038/s43016-020-00194-5>, 2020.
- Kastens, J. H., Brown, J. C., Coutinho, A. C., Bishop, C. R., and Esquerdo, J.: Soy moratorium impacts on soybean and deforestation dynamics in Mato Grosso, Brazil, *PLOS ONE*, 12, e0176 168, <https://doi.org/10.1371/journal.pone.0176168>, publisher: Public Library of Science, 2017.
- Kaur, S., Bansal, R. K., Mittal, M., Goyal, L. M., Kaur, I., Verma, A., and Son, L. H.: Mixed Pixel Decomposition Based on Extended Fuzzy Clustering for Single Spectral Value Remote Sensing Images, *Journal of the Indian Society of Remote Sensing*, 47, 427–437, <https://doi.org/10.1007/s12524-019-00946-2>, 2019.
- Kröger, M.: Inter-sectoral determinants of forest policy: the power of deforesting actors in post-2012 Brazil, *Forest Policy and Economics*, 77, 24–32, <https://doi.org/10.1016/j.forpol.2016.06.003>, 2017.
- Larned, S. T., Moores, J., Gadd, J., Baillie, B., and Schallenberg, M.: Evidence for the effects of land use on freshwater ecosystems in New Zealand, *New Zealand Journal of Marine and Freshwater Research*, 54, 551–591, <https://doi.org/10.1080/00288330.2019.1695634>, 2020.
- Macedo, M. N., DeFries, R. S., Morton, D. C., Stickler, C. M., Galford, G. L., and Shimabukuro, Y. E.: Decoupling of deforestation and soy production in the southern Amazon during the late 2000s, *Proceedings of the National Academy of Sciences*, 109, 1341–1346, <https://doi.org/10.1073/pnas.1111374109>, 2012.
- Maeda, E. E., Abera, T. A., Siljander, M., Aragão, L. E. O. C., Moura, Y. M. d., and Heiskanen, J.: Large-scale commodity agriculture exacerbates the climatic impacts of Amazonian deforestation, *Proceedings of the National Academy of Sciences*, 118, e2023787 118, <https://doi.org/10.1073/pnas.2023787118>, publisher: Proceedings of the National Academy of Sciences, 2021.
- McDonald, M.: Environment and security: Global eco-politics and Brazilian deforestation, *Contemporary Security Policy*, 24, 69–94, <https://doi.org/10.1080/13523260312331271919>, publisher: Routledge \_eprint: <https://doi.org/10.1080/13523260312331271919>, 2003.
- Morton, D. C., DeFries, R. S., Shimabukuro, Y. E., Anderson, L. O., Arai, E., del Bon Espirito-Santo, F., Freitas, R., and Morissette, J.: Cropland expansion changes deforestation dynamics in the southern Brazilian Amazon, *Proceedings of the National Academy of Sciences*, 103, 14 637–14 641, <https://doi.org/10.1073/pnas.0606377103>, publisher: Proceedings of the National Academy of Sciences, 2006.

- Müller-Hansen, F., Cardoso, M. F., Dalla-Nora, E. L., Donges, J. F., Heitzig, J., Kurths, J., and Thonicke, K.: A matrix clustering method to explore patterns of land-cover transitions in satellite-derived maps of the Brazilian Amazon, *Nonlinear Processes in Geophysics*, 24, 113–123, <https://doi.org/10.5194/npg-24-113-2017>, 2017.
- Nunes, C. A., Berenguer, E., França, F., Ferreira, J., Lees, A. C., Louzada, J., Sayer, E. J., Solar, R., Smith, C. C., Aragão, L. E. O. C., de Lima Braga, D., de Camargo, P. B., Cerri, C. E. P., de Oliveira, R. C., Durigan, M., Moura, N., Oliveira, V. H. F., Ribas, C., de Mello, F. V., Vieira, I., Zanetti, R., and Barlow, J.: Linking land-use and land-cover transitions to their ecological impact in the Amazon, *Proceedings of the National Academy of Sciences*, 119, <https://doi.org/10.1073/pnas.2202310119>, 2022.
- Odongo, V. O., Mulatu, D. W., Muthoni, F. K., van Oel, P. R., Meins, F. M., van der Tol, C., Skidmore, A. K., Groen, T. A., Becht, R., Onyando, J. O., and van der Veen, A.: Coupling socio-economic factors and eco-hydrological processes using a cascade-modeling approach, *Journal of Hydrology*, 518, 49–59, <https://doi.org/10.1016/j.jhydrol.2014.01.012>, 2014.
- Paim, M.-A.: Zero deforestation in the Amazon: The Soy Moratorium and global forest governance, *Review of European, Comparative and International Environmental Law*, 30, 220–232, <https://doi.org/10.1111/reel.12408>, 2021.
- Pekel, J., Cottam, A., Gorelick, N., and Belward, A. S.: High-resolution mapping of global surface water and its long-term changes, *Nature*, 540, 418–422, <https://doi.org/10.1038/nature20584>, number: 7633 Publisher: Nature Publishing Group, 2016.
- Pereira, J. C. and Viola, E.: Catastrophic Climate Risk and Brazilian Amazonian Politics and Policies: A New Research Agenda, *Global Environmental Politics*, 19, 93–103, [https://doi.org/10.1162/glep\\_a\\_00499](https://doi.org/10.1162/glep_a_00499), citation Key: pereira2019, 2019.
- Rittl, T. F., Oliveira, D., and Cerri, C. E. P.: Soil carbon stock changes under different land uses in the Amazon, *Geoderma Regional*, 10, 138–143, <https://doi.org/10.1016/j.geodrs.2017.07.004>, 2017.
- Schielein, J. and Börner, J.: Recent transformations of land-use and land-cover dynamics across different deforestation frontiers in the Brazilian Amazon, *Land Use Policy*, 76, 81–94, <https://doi.org/10.1016/j.landusepol.2018.04.052>, 2018.
- Simon, M. F. and Garagorry, F. L.: The expansion of agriculture in the Brazilian Amazon, *Environmental Conservation*, 32, 203–212, <https://doi.org/10.1017/S0376892905002201>, publisher: Cambridge University Press, 2005.
- Soares-Filho, B., Rajão, R., Macedo, M., Carneiro, A., Costa, W., Coe, M., Rodrigues, H., and Alencar, A.: Cracking Brazil’s Forest Code, *Science*, 344, 363–364, <https://doi.org/10.1126/science.1246663>, publisher: American Association for the Advancement of Science, 2014.
- Souza, C. M., Shimbo, J. Z., Rosa, M. R., Parente, L. L., Alencar, A. A., Rudorff, B. F. T., Hasenack, H., Matsumoto, M., Ferreira, L. G., Souza-Filho, P. W. M., de Oliveira, S. W., Rocha, W. F., Fonseca, A. V., Marques, C. B., Diniz, C. G., Costa, D., Monteiro, D., Rosa, E. R., Vélez-Martin, E., Weber, E. J., Lenti, F. E. B., Paternost, F. F., Pareyn, F. G. C., Siqueira, J. V., Viera, J. L., Neto, L. C. F., Saraiva, M. M., Sales, M. H., Salgado, M. P. G., Vasconcelos, R., Galano, S., Mesquita, V. V., and Azevedo, T.: Reconstructing Three Decades of Land Use and Land Cover Changes in Brazilian Biomes with Landsat Archive and Earth Engine, *Remote Sensing*, 12, 2735, <https://doi.org/10.3390/rs12172735>, 2020.
- West, T. A. P. and Fearnside, P. M.: Brazil’s conservation reform and the reduction of deforestation in Amazonia, *Land Use Policy*, 100, 105 072, <https://doi.org/10.1016/j.landusepol.2020.105072>, 2021.