

Density-dependent selection in evolutionary genetics: a lottery model of Grime's triangle

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Abstract

Fitness is typically represented in heavily simplified terms in evolutionary genetics, often using
3 constant selection coefficients. This excludes fundamental ecological factors such as dynamic
population size or density-dependence from our most genetically-realistic treatments of evolu-
tion, a problem that inspired MacArthur’s influential but problematic r/K theory. Following
6 in the spirit of r/K -selection as a general-purpose theory of density-dependent selection, but
grounding ourselves empirically in “primary strategy” trait classification schemes like Grime’s
triangle, we develop a new model of density-dependent selection which revolves around terri-
9 torial contests. To do so, we generalize the classic lottery model of territorial acquisition, which
has primarily been used for studying species co-existence questions, to accommodate arbitrary
densities. We use this density-dependent lottery model to predict the direction of trait evolution
12 under different environmental conditions and thereby provide a mathematical underpinning for
Grime’s verbal scheme. We revisit previous concepts of density-dependent selection, including r
and K selection, and argue that our model distinguishes between different aspects of fitness in a
15 more natural and intuitive manner.

“...the concept of fitness is probably too complex to allow of a useful mathematical development. Since it enters fundamentally into many population genetics considerations, it is remarkable how little attention has been paid to it.” — Warren J. Ewens, *Mathematical Population Genetics I*, 2004

Introduction

Evolutionary models differ greatly in their treatment of fitness. In models of genetic evolution, genotypes are typically assigned constant (or frequency-dependent) selection coefficients describing the change in their relative frequencies over time due to differences in viability. This considerably simplifies the mathematics of selection, facilitating greater genetic realism, and can be justified over sufficiently short time intervals (Ewens, 2012, p. 276). However, the resulting picture of evolution does not include even basic elements of the ecological underpinnings of selection, including dynamic population size and density-dependence.

By contrast, models of phenotypic trait evolution represent the change in phenotypic abundances over time using absolute fitness functions which describe how those traits affect survival and reproduction in particular ecological scenarios. This approach is powerful enough to model eco-evolutionary feedbacks between co-evolving traits, but is generally problem-specific and restricted to only a few traits at a time.

Far less work has been done to generalize beyond particular traits or ecological scenarios to models of fitness that still capture key distinctions between different forms of selection. Perhaps this is not surprising given that fitness is such a complex quantity, dependent on all of a phenotype’s functional traits (Violle et al., 2007) as well as its biotic and abiotic environment. In most cases, a detailed, trait-based, predictive model of fitness would be enormously complicated and have narrow applicability. It is therefore easy to doubt the feasibility of a simplified, general mathematical treatment of fitness (Ewens, 2012, p. 276). For example, MacArthur’s famous r/K scheme (MacArthur, 1962; MacArthur and Wilson, 1967) is now almost exclusively known as

a framework for understanding life-history traits, and judged on its failure in that role (Boyce,
42 1984; Pianka, 1970; Reznick et al., 2002; Stearns, 1977). The r/K scheme's original purpose was
as an extension of the existing population-genetic treatment of selection to account for popu-
lation density (MacArthur, 1962), but few attempts have been made to develop it further as a
45 mathematical analysis of the major different forms of selection.

Nevertheless, there are strong indications there are broader principles governing the opera-
tion of selection. In many groups of organisms, including corals (Darling et al., 2012), insects
48 (Southwood, 1977), fishes (Winemiller and Rose, 1992), zooplankton (Allan, 1976) and plants
(Grime, 1988; Westoby, 1998), different species can be divided into a small number of distinct trait
clusters corresponding to fundamentally distinct "primary strategies" (Winemiller et al., 2015).
51 The most famous example is Grime's plant trait classification scheme (Grime, 1974, 1977, 1988).
Grime considered two broad determinants of population density: stress (persistent hardship e.g.
due to resource scarcity, unfavorable temperatures or toxins) and disturbance (intermittent de-
54 struction of vegetation e.g. due to trampling, herbivory, pathogens, extreme weather or fire).
The extremes of these two factors define three primary strategies denoted by C/S/R respectively
(Fig. 1): competitors "C" excel in low stress, low disturbance environments; stress tolerators
57 "S" excel in high stress, low disturbance environments; and ruderals "R" excel in low stress,
high disturbance environments. Survival is not possible in high-stress, high-disturbance envi-
ronments. Grime showed that measures of C, S and R across a wide range of plant species are
60 anti-correlated, so that strong C-strategists are weak S and R strategists, and so on. Thus, plant
species can be classified on a triangular C/S/R ternary plot (Grime, 1974). Trait classification
schemes for other organisms are broadly analogous to Grime's scheme (Winemiller et al., 2015).

63 Trait classification schemes show empirically that, beneath the complicated details of trait
variation, even among closely-related species, fitness is predominantly determined by a few key
factors such as intrinsic reproductive rate or stress-tolerance. However, while trait classification
66 schemes are firmly grounded in trait data, they are verbal and descriptive rather than mathemat-
ical, a recognized hinderance to their broader applicability (e.g. Tilman 2007).



Figure 1: Schematic of Grime’s triangle. The two axes show increasing levels of environmental stress and disturbance, respectively. Survival is not possible if the combination of stress and disturbance is too large (dashed line). This creates a triangle, each corner of which corresponds to a “primary strategy”.

The aim of this paper is explore the interplay between some major dimensions of fitness in a simplified, spatially-homogeneous model of genotype growth, dispersal and competition. Building on the earlier r/K and $C/S/R$ schemes, a central question is how fitness depends on the interaction between population density, intrinsic birth/death rates and competitive ability.

We broadly follow the spirit of MacArthur’s r/K selection scheme in that our model is intended to account for fundamentally different forms of selection without getting entangled in the intricacies of particular ecological scenarios. However, rather than building directly on MacArthur’s formalism and its later extensions using Lotka-Volterra equations to incorporate competition (“ α -selection”) (Case and Gilpin, 1974; Gill, 1974; Joshi et al., 2001), our model is devised more with Grime’s $C/S/R$ scheme in mind, and represents a quantitative formalization of how $C/S/R$ manifests at the level of within-population genotype evolution (as opposed to phenotypic divergence between species). This choice is motivated in part by the substantial empirical support for $C/S/R$ -like schemes, and in part by the failings of the r/K low/high density dichotomy — many growth ability traits will confer advantages at both low and high densities (more details in the Discussion).

As we will see, a generalized version of the classic lottery model of Chesson and Warner (1981) is well suited for our purposes. In the classic lottery model, mature individuals (henceforth

“adults”) each require their own territory, whereas newborn individuals (henceforth “propagules”) disperse to, and subsequently compete for, the territories made available by the death of adults. Territorial contest among propagules leaves a single victorious adult per territory, with the victor chosen at random from the propagules present (akin to a lottery), with probabilities weighted by a coefficient for each type representing competitive ability. This representation of competition is much simpler than having coefficients for the pairwise effects of types on each other (e.g. the α coefficients in the generalized Lotka-Volterra equations), or than modeling resource consumption explicitly (Tilman, 1982). The classic lottery model also has similarities with the Wright-Fisher model in population-genetics (section BLAH).

However, the classic lottery model breaks down at low densities, specifically if there are only a few propagules dispersing to each territory (section “Mean field approximation”). This was not a limitation in its original application to reef fishes, where a huge number of larvae from each species compete to secure territories each generation Chesson and Warner (1981), but is a critical limitation here. We analytically extend the classic lottery model to correctly account for low density behavior.

In the section “Model”, we introduce the basic assumptions of our generalized lottery model. Analytical expressions for the change in genotype abundances over time are introduced in section “Mean field approximation”, with mathematical details relegated to the Appendices. The following two sections discuss the behavior of rare mutants and our treatment of Grime’s triangle.

Model

We assume that each individual in a population requires its own territory to survive and reproduce (a site-occupancy model). All territories are identical, and the total number of territories is T . Time t advances in discrete iterations, each representing the average time from birth to reproductive maturity. In iteration t , the number of reproductively mature individuals (henceforth

called “adults”) of the i ’th genotype is $n_i(t)$, the total number of adults is $N(t) = \sum_i n_i(t)$, and

111 the number of unoccupied territories is $U(t) = T - N(t)$.

Each iteration, adults produce m_i new offspring (henceforth called “propagules”). We assume adults cannot be ousted from occupied territories, and only propagules landing on unoccupied
114 territories are included in m_i . Propagules disperse at random, independently of each other, and without spatial restrictions; each propagule has an equal probability of landing on any of the U unoccupied territories. Thus, there is no interaction between propagules (e.g. avoidance of
117 territories crowded with propagules). Loss of propagules during dispersal is subsumed into m_i (m_i only counts propagules which go on to contest the territory they land in). In general, m_i will increase with n_i , and will also depend on population density N . For example, if b_i is the number
120 of propagules produced per individual from genotype i , then $m_i = b_i(1 - N/T)n_i$ would reflect the loss of propagules due to dispersal to occupied territories. We assume $m_i = b_i n_i$, meaning that all propagules land on unoccupied territories (a form of directed dispersal). This choice is not
123 biologically motivated; it simplifies the mathematics without seriously restricting the generality of our analysis, since the results presented here are not sensitive to the specific functional form of m_i .

126 The number of individuals of the i ’th genotype landing in any particular territory is denoted x_i . Random dispersal implies that in the limit $T \rightarrow \infty$, with n_i/T held fixed, x_i is Poisson distributed with mean territorial propagule density $l_i = m_i/U$ (this dispersal Poisson distribution
129 is denoted $p_i(x_i) = l_i^{x_i} e^{-l_i} / x_i!$). Although T is finite in our model, we assume that T and the n_i are large enough that x_i is Poisson-distributed to a good approximation (details in Appendix A). Note that the large n_i , large T approximation places no restrictions on our densities n_i/T , but it
132 does preclude consideration of demographic stochasticity when n_i itself is very small (this will be discussed further in Section “Invasion of rare genotypes and coexistence”).

When multiple propagules land on the same territory, they compete to secure the territory
135 as they develop. This territorial contest is modeled as a weighted lottery: the probability that genotype i wins a territory by the next iteration, assuming that at least one of its propagules is



Figure 2: Each iteration of our lottery model has three main elements. First, propagules are produced by adults which are dispersed at random over the unoccupied territories (only propagules landing on unoccupied territories are shown). Lottery competition then occurs in each unoccupied territory (only competition in one territory is illustrated): each genotype has a probability proportional to $b_i n_i c_i$ of securing the territory. Then occupied territories are freed up by adult mortality. In Eq. (2) and most of the paper, only adults can die (red crosses), but we will also consider the case where juveniles die (blue cross; section “Primary strategies and Grime’s triangle”).

present, is $c_i x_i / \sum_j c_j x_j$, where c_i is a constant representing relative competitive ability.

138 The increase in n_i over one iteration due to territorial acquisition, $\Delta_+ n_i$, is the sum of genotype i 's victories over all U unoccupied territories. The expected proportion of unoccupied territories with x_1, \dots, x_G of the respective propagules is $p_1(x_1) \dots p_G(x_G)$ (G is the number of genotypes present), and the probability of i winning in each of these territories is $c_i x_i / \sum_j c_j x_j$. As above, we
141 assume that T and the n_i are large enough that we can ignore fluctuations in the proportion of unoccupied territories as well as the number of victories, which implies

$$\Delta_+ n_i(t) = U(t) \sum_{x_1, \dots, x_G} \frac{c_i x_i}{\sum_j c_j x_j} p_1(x_1) \dots p_G(x_G), \quad (1)$$

144 where the sum only includes territories with at least one propagule present.

In addition to propagule birth and competition, occupied territories become unoccupied due to mortality. For the majority of this manuscript we assume that mortality only occurs in adults
147 (setting aside the deaths implicit in territorial contest), and at a constant, genotype-specific per-capita rate d_i , so that the overall change in genotype abundances is

$$\Delta n_i(t) = \Delta_+ n_i(t) - d_i n_i(t). \quad (2)$$

This is reasonable approximation in the absence of disturbances; when we come to consider the
150 effects of disturbances (Section "Primary strategies and Grime's triangle"), we will incorporate disturbance-induced mortality in competing juveniles (Fig. 2).

Note that the competitive ability coefficients c_i represent a strictly relative aspect of fitness in
153 the sense that they only influence population size N indirectly by changing genotype frequencies; that may in turn change the population mean birth and death rates. This can be seen by summing Eq. (2) over genotypes to get the change in population size N ,

$$\Delta N = U(1 - e^{-L}) - \sum_i d_i n_i, \quad (3)$$

156 which is independent of c_i (here $L = \sum_j l_j$ is the overall propagule density).

Results

Mean Field Approximation

159 Eq. (2) gives little intuition about the dynamics of density-dependent lottery competition, since
 (1) involves an expectation over the random dispersal distributions p_i , which depend on how the
 n_i change over time. We now evaluate this expectation using a “mean field” approximation; the
 162 intuition behind this approximation is as follows.

If the unoccupied territories are saturated with propagules from every genotype ($l_i \gg 1$ for
 all genotypes), the fluctuations in the x_i are small compared to their means l_i (since the x_i are
 165 Poisson distributed), and so the composition of propagules in a territory will only rarely differ
 appreciably from the mean composition l_1, l_2, \dots, l_G . Consequently, we can replace x_i with l_i in
 Eq. (1). This gives the classic lottery model (Chesson and Warner, 1981),

$$\Delta_+ n_i(t) = U(t) \frac{c_i m_i}{\sum_j c_j m_j} = b_i n_i \frac{1}{L} \frac{c_i}{\bar{c}}, \quad (4)$$

168 where $\bar{c} = \sum_j c_j m_j / M$ is the mean propagule competitive ability for a randomly selected propag-
 ule ($M = \sum_j m_j$ is the total number of propagules).

However, in general the l_i are not all large, and the x_i cannot simply be replaced by their
 171 means in Eq. (1). Indeed, Eq. (4) is nonsensical if l_i is sufficiently small: genotype i can win at
 most m_i territories, yet Eq. (4) demands a fraction $c_i m_i / \sum_j c_j m_j$ of the unoccupied territories U ,
 no matter how large U is. The source of this pathological behavior when $l_i \ll 1$ is that $x_i = 1$ in
 174 the few territories where i propagules do land, and so i 's growth comes entirely from territories
 which deviate appreciably from the mean.

Our mean field approximation is similar to the high- l_i approximation leading to Eq. (4) in
 177 that we replace the x_i with appropriate mean values. The key distinction is that territories with a

single propagule from the focal genotype, whose behavior is critical at low densities, are handled separately. In place of the requirement of $l_i \gg 1$ for all i , our approximation only requires
 180 that there are no large discrepancies in competitive ability (specifically, that we do not have $c_i/c_j \gg 1$ for any two genotypes; further discussion in section “Discussion”). We obtain (details in Appendix B)

$$\Delta_+ n_i(t) \approx b_i n_i \left[e^{-L} + (R_i + A_i) \frac{c_i}{\bar{c}} \right], \quad (5)$$

183 where

$$R_i = \frac{\bar{c} e^{-l_i} (1 - e^{-(L-l_i)})}{c_i + \frac{L-1+e^{-L}}{1-(1+L)e^{-L}} \frac{\bar{c} L - c_i l_i}{L-l_i}}, \quad (6)$$

and

$$A_i = \frac{\bar{c} (1 - e^{-l_i})}{\frac{1-e^{-l_i}}{1-(1+l_i)e^{-l_i}} c_i l_i + \frac{1}{L-l_i} \left(L \frac{1-e^{-L}}{1-(1+L)e^{-L}} - l_i \frac{1-e^{-l_i}}{1-(1+l_i)e^{-l_i}} \right) \sum_{j \neq i} c_j l_j}. \quad (7)$$

Comparing Eq. (5) to Eq. (4), the classic lottery per-propagule success rate $c_i/\bar{c}L$ has been
 186 replaced by three separate terms. The first, e^{-L} , accounts for propagules which land alone on unoccupied territories; these territories are won without contest. The second, $R_i c_i/\bar{c}$ represents competitive victories when the i genotype is a rare invader in a high density population: from
 189 Eq. (6), $R_i \rightarrow 0$ when the i genotype is abundant ($l_i \gg 1$), or other genotypes are collectively rare ($L - l_i \ll 1$). The third term, $A_i c_i/\bar{c}$, represents competitive victories when the i genotype is abundant: $A_i \rightarrow 0$ if $l_i \ll 1$. The relative importance of these three terms varies with both
 192 the overall propagule density L and the relative propagule frequencies l_i/L . If $l_i \gg 1$ for all genotypes, we recover the classic lottery model (only the $A_i c_i/\bar{c}$ term remains, and $A_i \rightarrow 1/L$). Thus, Eq. (5) generalizes the classic lottery model to account for arbitrary propagule densities
 195 for each genotype.

Fig. 3 shows that Eq. (5) (and its components) closely approximate direct simulations of random dispersal and lottery competition over a wide range of propagule densities (obtained by
 198 varying U). Two genotypes are present, one of which has a c -advantage and is at low frequency. The growth of the low-frequency genotype relies crucially on the low-density competition term

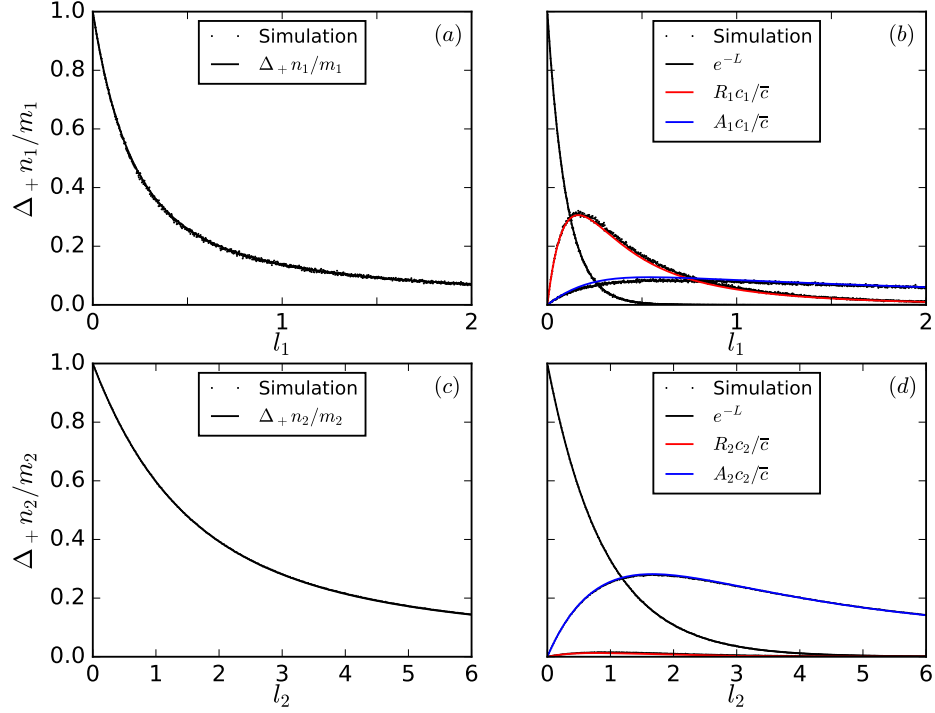


Figure 3: The change in genotype abundances in a density dependent lottery model is closely approximated by Eq. (5). Δ_+n_i/m_i from Eq. (5) (and its separate components) are shown, along with direct simulations of random dispersal and lottery competition over one iteration over a range of propagule densities (varied by changing U with the m_i fixed). Two genotypes are present. (a) and (b) show the low-frequency genotype with c -advantage ($m_1/M = 0.1$, $c_1 = 1.5$), (c) and (d) show the high-frequency predominant genotype ($m_2/M = 0.9$, $c_2 = 1$). Simulation points are almost invisible in (c) and (d) due to near exact agreement with Eq. (5).

$R_i c_i / \bar{c}$, and also to a lesser extent on the high density competition term $A_i c_i / \bar{c}$ if l_1 is large enough

(Fig. 3b). On the other hand, $R_i c_i / \bar{c}$ is negligible for the high-frequency genotype, which depends instead on high density territorial victories (Fig. 3d).

Invasion of rare genotypes and coexistence

In our model (section “Model”), each genotype is defined by three traits: b , c and d . To determine how these will evolve in a population where they are being modified by mutations, we need to know whether mutant lineages will grow (or decline) starting from low densities. In this section

207 we discuss the behavior of rare genotypes predicted by Eq. (5).

Suppose that a population with a single genotype i is in equilibrium. Then $R_i = 0$, $\bar{c} = c_i$ and $\Delta n_i = 0$, and so Eq. (5) gives

$$b_i \left(e^{-L} + A_i \right) - d_i = 0, \quad (8)$$

210 where $A_i = (1 - (1 + L)e^{-L})/L$. Now suppose that a new genotype j , which is initially rare, appears in the population. Then $A_j \ll R_j$, $l_j \approx 0$ and $\bar{c} \approx c_i$, and so, from Eq. (5), n_j will increase if

$$b_j \left(e^{-L} + R_j \frac{c_j}{c_i} \right) - d_j > 0, \quad (9)$$

213 where $R_j \approx (1 - e^{-L}) / \left(\frac{c_j}{c_i} + \frac{L-1-e^{-L}}{1-(1+L)e^{-L}} \right)$.

Combining Eqs. (8) and (9), we see that j will invade if it is superior in any one of the three traits, but is otherwise identical to i . If the new genotype has the same competitive ability $c_j = c_i$, 216 then $R_j \approx A_i$ and Eqs. (8) and (9) imply that invasion occurs when $b_j d_i - b_i d_j > 0$, and in particular when $b_j > b_i$ with $d_j = d_i$, or when $d_j < d_i$ with $b_j = b_i$. In the case that the new genotype has a different competitive ability but the same b_i and d_i , Eqs. (8) and (9) imply that 219 invasion occurs when $R_j c_j / c_i > A_i$; it is not hard to verify that this occurs if and only if $c_j > c_i$ using the simplified expressions for A_i and R_j given after Eqs. (8) and (9) respectively. Moreover, if j invades in any of these cases, it will eventually exclude i , since it is strictly superior.

222 However, stable coexistence is possible between genotypes that are superior in different traits. To illustrate, suppose that j is better at securing territories ($c_j > c_i$), that i is better at producing propagules ($b_i > b_j$), and that $d_i = d_j$. Coexistence occurs if j will invade an i -dominated 225 population, but i will also invade a j -dominated population ("mutual invasion"). It is not hard to show that this is possible, since if b_i is so large that $L \gg 1$ when i is dominant, and b_j is so small that $L \ll 1$ when j is dominant, then, combining Eqs. (8) and (9), we find that i invades j 228 because $b_i > b_j$, while j invades i provided that

$$b_j c_j R_j - b_i c_i A_i > 0. \quad (10)$$

Thus, coexistence occurs if c_j/c_i is large enough. Intuitively, the mechanism for coexistence is that territorial contests are important in an i -dominated population (high L), ensuring that the c -specialist j is not excluded, yet territorial contests are irrelevant in a j -dominated population (low L), ensuring that the b -specialist i is not excluded. Fig. ?? shows an example of this coexistence between b and c specialists.

A similar argument applies for coexistence between high- c and low- d specialists; again coexistence occurs because the importance of territorial contests declines along with propagule density L as the c -specialist increases in frequency. Mutual invasibility is not possible between b - and d -specialists (although genotypes with exactly $c_i = c_j$ and $b_i/d_i = b_j/d_j$ do not exclude each other — this follows from the fact that all propagules have the same probability of success when $c_i = c_j$ i.e. $A_i + R_i = A_j + R_j$).

If the rare genotype j arises due to mutation, then its initial low-density behavior is more complicated than the above invasion analysis suggests. The mutant lineage starts with one individual $n_j = 1$, and remains at low abundance for many generations after its initial appearance. During this period, the mutant abundance n_j will behave stochastically, and the deterministic equations (1) and (5) do not apply (section “Model”). However, if n_j becomes large enough, its behavior will become effectively deterministic, and governed by Eq. (5). For mutants with fitness greater than the population mean fitness, this occurs when n_j is of order $1/s$ (Desai and Fisher, 2007), where the selection coefficient s is the mutant’s fitness advantage (i.e. $s = \frac{\Delta n_i/n_i}{\sum_i \Delta n_i/n_i \times n_i/N} - 1$). Here we do not consider the initial stochastic behavior of novel mutants, and have restricted our attention to the earliest deterministic behavior of rare genotypes. In particular, for beneficial mutations we have only considered the case where s is large enough that deterministic behavior starts when $n_j \ll N$.

Primary strategies and Grime’s triangle

We now discuss which changes in the traits b, c and d will be particularly favored under different environmental conditions. Of specific interest are Grime’s “disturbance”, “stress” and “ideal”

environmental archetypes. To proceed, we need to map these verbal archetypes to quantitative parameter regimes in our model.

The ideal environmental archetype is characterized by the near-absence of stress and disturbance. Consequently, $d_i \ll 1$, whereas b_i is potentially much larger than 1. From Eq. (3), the equilibrium value of L only depends on the ratio of birth and death rates. For one genotype, $L/(1 - e^{-L}) = b_i/d_i$, and so the propagule density is high $L \approx b_i/d_i \gg 1$. Moreover, since $L = b_i \frac{N}{N-T} = b_i \frac{1}{1-T/N}$ by definition, population density is also high $N/T \approx 1$. Thus, almost every unoccupied territory will be heavily contested.

The disturbance archetype is characterized by unavoidably high extrinsic mortality caused by physical destruction. Disturbances do not only affect adults as in Eq. (2), but also juveniles in the process of territorial contest. These juvenile deaths can be represented as a fractional reduction in the number of territories secured. To illustrate, we assume that the disturbance is equally damaging to adults and juveniles, so that only $(1 - d_i)\Delta_+ n_i$ rather than $\Delta_+ n_i$ territories are secured by genotype i each iteration. Then, the disturbance archetype is characterized by d_i being close to 1 for all genotypes (almost all adults and juveniles are killed each iteration). From Eq. (3), the single genotype equilibrium is given by $L/(1 - e^{-L}) = d_i/[(1 - d_i)b_i]$, and since $L \ll 1$ and $N/T \ll 1$ due to high mortality, we have $L \approx 2(1 - d_i/[(1 - d_i)b_i])$. Clearly b_i must be exceptionally large to ensure population persistence. The terms proportional to c_i/\bar{c} in Eq. (5) are then negligible, and $\Delta_+ n_i$ depends primarily on b_i .

The stress archetype is more ambiguous, and has been the subject of an extensive debate in the plant ecology literature (the ‘‘Grime-Tilman’’ debate; Aerts 1999 and references therein). Stressful environments severely restrict growth and reproduction, so that $b \ll 1$ Grime (1974, 1977). Mutations which appreciably improve b will be either non-existent or extremely unlikely, so b is constrained to remain low. In Grime’s view, under these conditions the rate at which propagules successfully develop to adulthood cannot appreciably exceed the mortality rate. This implies $b/d \approx 1$ in our model, and so the propagule density L is suppressed to such low levels that there are essentially no territorial contests occurring.

The alternative view is that, while stressful environments imply lower b and support a lower number of individuals per unit area compared what is attainable in ideal environments, stressed populations are actually at high densities relative to the environmental carrying capacity, and are highly competitive (Taylor et al., 1990). In the particular case that stress is caused by scarcity of consumable resources, it is argued that we should expect intense resource competition (for empirical support, see Davis et al. 1998). Thus, b may actually appreciably exceed d under stressful conditions, even though the absolute value of b is small.

The mapping of environmental archetypes to our model parameters is summarized in the first two rows of Fig. 4. Also shown is the approximate dependence of Δn_i on b_i and c_i for each archetype (third row). These can be used infer the expected direction of evolution for the traits b , c and d (fourth row) as follows.

As noted in the previous section, if beneficial mutations establish (i.e survive the low-abundance stochastic regime), they will proceed to grow deterministically according to Eq. (5). The probability of establishment increases with the mutant fitness advantage, and is therefore typically on the order of one percent, whereas the fixation of neutral mutations is exceedingly unlikely (probability of order $1/N$). Consequently, the direction of evolutionary change is determined by which trait changes are both available, and confer an appreciable benefit, where availability is subject to constraints imposed by the environment.

For example, in Grime's version of the stress archetype, L is low, so competition is not important, and only mutants with greater b or lower d will have an appreciably greater Δn_i . Mutations in c are effectively neutral, and will rarely fix. However, by definition of the stress archetype, b is constrained to be small. Thus, while some rare mutations may produce small improvements in b , it is much more likely that mutations will arise that lower d , making this the expected direction of evolutionary change for Grime's stress archetype.

Following Grime's original argument for a triangular scheme (Grime, 1977), Fig. 5 represents each environmental archetype schematically as a vertex on a triangular space defined by perpendicular stress and disturbance axes. The ideal archetype lies at the origin (no stress or

	Ideal	Disturbance*	Stress (G)	Stress (T)
Constraints	$d \ll 1$	$d \approx 1$	$b \ll 1$	$b \ll 1$
Other parameters	$b \gg d$	$b \gg d$	$b \approx d$	$b > d$
Density N/T	High	Low	Low	High
$\Delta_+ n_i \propto$	$b_i c_i$	b_i	b_i	$b_i c_i$
Evolution for	$\uparrow b, \uparrow c$	$\uparrow b, \downarrow d$	$\downarrow d$	$\uparrow c, \downarrow d$

Figure 4: The realization of Grime’s environmental archetypes in our model, as well as the low- T variant of the stress archetype. Shown are the mapping to our parameters of each archetype, the approximate dependence of $\Delta_+ n_i$ on b_i and c_i , as well as the corresponding expected evolutionary changes in b_i , c_i and d_i . *Mortality affects both adults and juveniles in the disturbance archetype, with $\Delta_+ n_i$ replaced by $(1 - d_i)\Delta_+ n_i$ in Eq. (2).

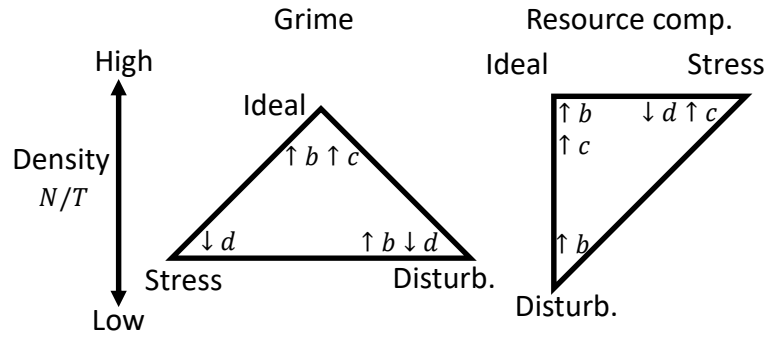


Figure 5: The realization of Grime’s triangle in our model. Schematic representation of the triangular space bounded by the low/high extremes of stress/disturbance. The low- T interpretation of stress is also shown. The vertices of the triangles correspond to environmental archetypes. Selection favors different traits at each vertex, leading to different trait clusters.

disturbance), while the stress and disturbance archetypes lie at the limits of survival on their respective axes. The hypotenuse connecting the stress and disturbance endpoints represents the limits of survival in the presence of a combination of stress and disturbance. The direction of evolutionary change is different at each vertex, leading to the emergence of different trait clusters or “primary strategies”.

How does Fig. 4 compare to empirical analyses of Grime’s C/S/R strategies? In our comparison we will stick to fishes, corals and plants, for which three-way primary strategy schemes are well developed (Darling et al., 2012; Grime, 1977; Winemiller and Rose, 1992). The connection of our model to fish strategies is necessarily more tentative, given that fishes are motile and not all

territorial, and the starting assumption of our model is site-occupancy.

In disturbed environments, we predict evolution for higher b and lower d , but not higher c . Higher b means higher fecundity, but not necessarily mass propagule production: b represents only those propagules which successfully develop into juveniles in unoccupied territories. This is broadly consistent with the ruderal primary strategy. Plant ruderals devote a large proportion of their productivity to seed production Grime (1977), whereas the analogous “opportunistic” strategists in fishes have large intrinsic growth rates (Winemiller and Rose, 1992). In corals, a distinguishing feature of the ruderal cluster is brood spawning (rather than broadcast spawning). This corresponds to higher parental investment and lower overall propagule production, but potentially also higher b at low densities, since broadcast spawners are vulnerable to a powerful Allee effect at the egg fertilization stage (Knowlton, 2001). Lower d could be achieved by improved individual resistance to physical destruction, but it is hard to reduce mortality in the face of severe disturbances. Given this constraint, shortening the time to reproductive maturity (the iteration time in our model) is an effective way of reducing the chance of death per iteration d . An exceptionally short life cycle is probably the most defining characteristic of ruderals (Darling et al., 2012; Grime, 1977; Winemiller and Rose, 1992).

In stressful environments, we predict evolution for lower d , and also for higher c in the low- T interpretation of the stress archetype. Lowering d is obviously essential when $b \ll 1$, and stress tolerant plants and corals have long life spans, allowing for long intervals between successful recruitments (and episodic broadcast spawning in corals). For fishes, the “equilibrium” strategy is the analogue of Grime’s stress tolerator. This strategy is associated with resource limitation, and is also characterized by long life span, as well as high parental investment in tiny broods. This may reflect a high- c strategy in the face of intense competition for severely limited resources (the low- T interpretation).

In ideal environments, we predict evolution for higher b and c , but not lower d . In plants and corals, a key mechanism for winning territorial contests is rapidly outgrowing and “shading out” competitors; not surprisingly, rapid individual growth is a defining feature of the competitor trait

cluster (Darling et al., 2012; Grime, 1977). Evolution for higher b under high-density, competitive conditions may seem counter-intuitive. Neither particularly high nor low b have been associated with the competitor strategy in plants and corals. However, for fishes, the analogous “periodic” strategy is characterized by enormous spawn sizes as well as rapid development (Winemiller et al., 2015; Winemiller and Rose, 1992), suggesting a strategy of ensuring that many propagules actually end up contesting areas favorable for development (higher b). The evolution of b in ideal environments will be discussed further in the Discussion.

Discussion

Unlike Grime’s classic ternary plot (Grime, 1974), which represents anti-correlations between traits relevant for success in different environmental archetypes, our realization of Grime’s triangle (Fig. 4) refers instead to the direction of adaptive trait evolution under different regimes of stress and disturbance. As discussed in section “Primary strategies and Grime’s triangle”, over time our predicted trait evolution should lead to trait values consistent with Grime’s scheme. In making these predictions, we have made no reference to any kind of trade-offs or pleiotropy, even though trade-offs are often invoked to explain primary strategy schemes (Aerts, 1999; MacArthur and Wilson, 1967; Winemiller and Rose, 1992). Thus, while trade-offs may amplify specialization, they are not necessary for it. As an example of a trade-off, corals which rapidly out-shade neighbors have a tall, branched morphology which is vulnerable to disturbances, and so, all else being equal, ideal environment c -strategists will suffer higher mortality from disturbances. Fig. 5 gives the same conclusion without invoking trade-offs; mutations which reduce disturbance vulnerability are essentially neutral under ideal conditions, leading to no improvements in mortality from disturbances, whereas c will tend to increase over time.

Our prediction of evolution for higher b in ideal environments is counter to the expectations of MacArthur’s r/K dichotomy (MacArthur and Wilson, 1967) since b is closely related to the maximal, low-density growth rate $r = b - d$, and ideal environments support high population

densities which should be subject to “ K -selection”. However, in the Introduction, we noted that the r - K dichotomy is not consistent with empirical studies showing that maximal growth rate and saturation density (measured by abundance) are positively correlated, both between
372 species/strains (Fitzsimmons et al., 2010; Hendriks et al., 2005; Kuno, 1991; Luckinbill, 1979), and as a result of experimental evolution (Luckinbill, 1978, 1979). From the perspective of our
375 model, these correlations are not surprising since the saturation density, which is determined by a balance between births and deaths, increases with b . Our higher- b prediction simply reflects the fact that, all else being equal, producing more propagules is always advantageous, regardless
378 of population density, a fact lost in the simple logistic interpretation of the r/K scheme.

Confusingly, the term “ K -selection” has sometimes been used to refer generally to selection at high density; this encompasses both selection for higher saturation density — “ K ” in the logistic
381 equation — as well as selection for competitive ability. To avoid this ambiguity, the latter form of selection has been called “ α -selection” after the competition coefficients in the Lotka-Volterra equation (Case and Gilpin, 1974; Gill, 1974; Joshi et al., 2001). Unlike saturation density, there is
384 support for a negative relationship between competitive success at high densities and maximal growth rate (Luckinbill, 1979); this could be driven by a tradeoff between individual size and reproductive rate. However, competitive success as measured by α (i.e. the per-capita effect of one
387 genotype on another genotype’s growth rate) is only partly determined by individual competitive ability — in the presence of age-structure and territoriality, it also includes the ability of each genotype to produce contestants i.e. b in our model. In contrast, our c is strictly competitive
390 ability only — as such, changes in c do not directly affect population density (section “Model”), nor the ability to produce contestants.

K -selection in the sense of selection for a greater environmental carrying capacity for given
393 birth and death rates, sometimes referred to as “efficiency” (MacArthur and Wilson, 1967), would be represented in our model by smaller individual territorial requirements. To a first approximation, two co-occurring genotypes which differ by a small amount in their territorial requirements
396 only should have the same fitness since the costs or benefits of a change in the amount of unocu-

pied territory is shared equally among genotypes via the propagule density L . The situation is more complicated if those genotypes differ in multiple traits, and when the differences in territorial requirements become large enough that territorial contests can occur on different scales. We
399 leave these complications for future work.

The importance of b in securing territories is a general feature of lottery competition. Indeed,
402 as can be seen from Eq. (4), in the classic lottery model b_i and c_i are essentially equivalent in that only the products $b_i c_i$ matter (Chesson and Warner, 1981). This is no longer the case in our density- and frequency-dependent generalization of the classic lottery model, where stable
405 co-existence is possible between b and c strategists. Given that the classic lottery model was specifically developed for studying species co-existence questions, this may seem surprising, but the focus in that case was on the role of environmental fluctuations in promoting co-existence
408 rather than coexistence in a single, stable environment (Chesson and Warner, 1981). It is not clear whether correctly accounting for the behavior of species with low density would significantly alter the conclusions of Chesson and Warner (1981), but, in any case, species co-existence questions
411 are beyond the scope of this manuscript.

Rather, while our model can be applied to questions at an inter-species level (e.g. ecological invasions), our focus here is on the evolution of genotype frequencies within a population. Our
414 ability to describe evolutionary processes is only possible because the model accounts for the growth of mutants from low densities. Given this focus, our assumption that there are no large c discrepancies (section “Mean field approximation”) amounts to a restriction on the amount of
417 genetic variation in c that will be sustained in the population. Since beneficial mutation effect sizes will typically not be much larger than a few percent, large c discrepancies can only arise if the mutation rate is extremely large, and so the assumption will not be violated in most cases.
420 However, this restriction could become important when looking at species interactions rather than genotype evolution.

In Fig. 6 we compare our model with some of the other models and schemes touched upon
423 here. In a sense this comparison is unfair: for instance, Grime’s scheme was developed for an

	Formal model?	Ecologically meaning- ful?	Empirically- grounded trait scheme?	Generality beyond specific scenarios?	Genetically flexible?
Density-dependent lottery	✓	✓	✓	✓	✓
MacArthur's $r/K + \alpha$	✓	✓	✗*	✓	✓
Grime's C/S/R	✗	✓	✓	✓	NA
Traditional pop. gen.	✓	✗	✗	✓	✓
Eco-evo:					
Adaptive dynamics	✓	✓	✓	✓**	✗
Brute force simulation	✓	✓	NA	✗	✓

Figure 6: Comparison of our density-dependent lottery model with related models and schemes in ecology and evolutionary biology. *MacArthur's r - and K -, as well as α selection, were all derived theoretically. Applications to traits came later, and with mixed success (see "Introduction" and "Discussion"). **In practice, most of the adaptive dynamics literature focuses on specific eco-evolutionary outcomes such as evolutionary "branching" (Geritz et al., 1997) or mutualisms (Ferriere et al., 2002), but in principle it can be applied with any fitness model including our density-dependent lottery.

entirely different purpose (species classification by traits). As such, Fig. 6 is not exhaustive and should be read more as a summary of our model's purpose. Like MacArthur's r/K scheme, our model is motivated by the need to expand the treatment of selection in population genetics (MacArthur, 1962), i.e. to incorporate crucial ecological factors in our most genetically realistic models of evolution. Thus, viewing evolutionary ecology (Kokko and López-Sepulcre, 2007; Pelletier et al., 2009; Schoener, 2011) as a spectrum ranging from evolution-only to ecology-only, our model lies close to the understudied evolution-only end of the spectrum. By comparison, more familiar approaches to evolutionary ecology such as adaptive dynamics — essentially ecology coupled with mutant invasion (Diekmann et al., 2004) — lie close to the ecology-only end of the spectrum.

In our view, our model has two major limitations as a general-purpose model of density-dependent selection: a reliance on interference competition for durable resources (territoriality), and the restriction of competition to juveniles (lottery recruitment to adulthood). In some respects this is the complement of resource competition models, which restrict their attention to exploitation competition, typically without age structure (Tilman, 1982). In the particular case

that resources are spatially localized (e.g. due to restricted movement through soils), then resource competition and territorial acquisition effectively coincide, and in principle resource competition could be represented by a competitive ability c (or conversely, c should be derivable from resource competition). The situation is more complicated if the resources are well-mixed, since, in general, resource levels then need to be explicitly tracked. It seems plausible that explicit resource tracking is not necessary when the focus is on the evolution of similar genotypes rather than the stable co-existence of widely differing species. We are not aware of any attempts to delineate conditions under which explicit resource tracking is unnecessary even if it is assumed that community structure is ultimately determined by competition for consumable resources. More work is needed connecting resource competition models to the density-dependent selection literature, since most of the existing work is focused on the narrower Grime-Tilman-debate issue of resolving the role of competition at different levels of resource availability (Aerts, 1999; Davis et al., 1998; Tilman, 2007).

On the other hand, our model does remarkably well in capturing apparently general features of selection under different environmental conditions using only three trait parameters: b , c and d . The clean separation of a strictly-relative c parameter is particularly useful from an evolutionary genetics perspective, essentially embedding a zero-sum fitness trait within a non-zero-sum fitness model. This could have interesting applications for modeling the impacts of intra-specific competition on species extinction, for example due to clonal interference (Desai and Fisher, 2007; Gerrish and Lenski, 1998) between c -strategists on the one hand, and b - and d -strategists on the other.

References

- Aerts, R. 1999. Interspecific competition in natural plant communities: mechanisms, trade-offs and plant-soil feedbacks. *Journal of Experimental Botany* 50:29–37.
- Allan, J. D. 1976. Life history patterns in zooplankton. *The American Naturalist* 110:165–180.

Boyce, M. S. 1984. Restitution of r-and k-selection as a model of density-dependent natural
465 selection. *Annual Review of Ecology and Systematics* 15:427–447.

Case, T. J., and M. E. Gilpin. 1974. Interference competition and niche theory. *Proceedings of the
National Academy of Sciences* 71:3073–3077.

468 Chesson, P. L., and R. R. Warner. 1981. Environmental variability promotes coexistence in lottery
competitive systems. *American Naturalist* pages 923–943.

Darling, E. S., L. Alvarez-Filip, T. A. Oliver, T. R. McClanahan, and I. M. Côté. 2012. Evaluating
471 life-history strategies of reef corals from species traits. *Ecology Letters* 15:1378–1386.

Davis, M. A., K. J. Wragge, and P. B. Reich. 1998. Competition between tree seedlings and herba-
ceous vegetation: support for a theory of resource supply and demand. *Journal of Ecology*
474 86:652–661.

Desai, M. M., and D. S. Fisher. 2007. Beneficial mutation–selection balance and the effect of
linkage on positive selection. *Genetics* 176:1759–1798.

477 Diekmann, O., et al. 2004. A beginner’s guide to adaptive dynamics. *Banach Center Publications*
63:47–86.

Ewens, W. J. 2012. *Mathematical Population Genetics 1: Theoretical Introduction*, vol. 27.
480 Springer Science & Business Media.

Ferriere, R., J. L. Bronstein, S. Rinaldi, R. Law, and M. Gauduchon. 2002. Cheating and the
evolutionary stability of mutualisms. *Proceedings of the Royal Society of London B: Biological
483 Sciences* 269:773–780.

Fitzsimmons, J. M., S. E. Schoustra, J. T. Kerr, and R. Kassen. 2010. Population consequences of
mutational events: effects of antibiotic resistance on the r/k trade-off. *Evolutionary Ecology*
486 24:227–236.

Geritz, S. A., J. A. Metz, É. Kisdi, and G. Meszéna. 1997. Dynamics of adaptation and evolutionary branching. *Physical Review Letters* 78:2024.

489 Gerrish, P. J., and R. E. Lenski. 1998. The fate of competing beneficial mutations in an asexual population. *Genetica* 102:127–144.

Gill, D. E. 1974. Intrinsic rate of increase, saturation density, and competitive ability. ii. the
492 evolution of competitive ability. *American Naturalist* 108:103–116.

Grime, J. P. 1974. Vegetation classification by reference to strategies. *Nature* 250:26–31.

———. 1977. Evidence for the existence of three primary strategies in plants and its relevance to
495 ecological and evolutionary theory. *American Naturalist* 111:1169–1194.

———. 1988. *Plant Evolutionary Biology*, chap. The C-S-R model of primary plant strategies — origins, implications and tests, pages 371–393. Springer Netherlands, Dordrecht.

498 Hendriks, A. J., J. L. Maas-Diepeveen, E. H. Heugens, and N. M. van Straalen. 2005. Meta-analysis of intrinsic rates of increase and carrying capacity of populations affected by toxic and other stressors. *Environmental Toxicology and Chemistry* 24:2267–2277.

501 Joshi, A., N. Prasad, and M. Shakarad. 2001. K-selection, α -selection, effectiveness, and tolerance in competition: density-dependent selection revisited. *Journal of Genetics* 80:63–75.

Knowlton, N. 2001. The future of coral reefs. *Proceedings of the National Academy of Sciences of the United States of America* 98:5419–5425.
504

Kokko, H., and A. López-Sepulcre. 2007. The ecogenetic link between demography and evolution: can we bridge the gap between theory and data? *Ecology Letters* 10:773–782.

507 Kuno, E. 1991. Some strange properties of the logistic equation defined with r and K : Inherent defects or artifacts? *Researches on Population Ecology* 33:33–39.

- Luckinbill, L. S. 1978. r and k selection in experimental populations of escherichia coli. Science
510 (New York, NY) 202:1201–1203.
- . 1979. Selection and the r/k continuum in experimental populations of protozoa. American Naturalist pages 427–437.
- 513 MacArthur, R. H. 1962. Some generalized theorems of natural selection. Proceedings of the National Academy of Sciences 48:1893–1897.
- MacArthur, R. H., and E. O. Wilson. 1967. Theory of Island Biogeography. Princeton University
516 Press.
- Pelletier, F., D. Garant, and A. Hendry. 2009. Eco-evolutionary dynamics. Philosophical Transactions of the Royal Society B: Biological Sciences 364:1483–1489.
- 519 Pianka, E. R. 1970. On r- and K-Selection. The American Naturalist 104:592–597.
- Reznick, D., M. J. Bryant, and F. Bashey. 2002. r- and k-selection revisited: The role of population regulation in life-history evolution. Ecology 83:1509–1520.
- 522 Schoener, T. W. 2011. The newest synthesis: Understanding the interplay of evolutionary and ecological dynamics. Science 331:426–429.
- Southwood, T. R. E. 1977. Habitat, the templet for ecological strategies? Journal of Animal
525 Ecology 46:337–365.
- Stearns, S. C. 1977. The evolution of life history traits: A critique of the theory and a review of the data. Annual Review of Ecology and Systematics 8:145–171.
- 528 Taylor, D. R., L. W. Aarssen, and C. Loehle. 1990. On the relationship between r/k selection and environmental carrying capacity: a new habitat templet for plant life history strategies. Oikos pages 239–250.
- 531 Tilman, D. 1982. Resource competition and community structure. 17. Princeton University Press.

———. 2007. Resource competition and plant traits: a response to craine et al. 2005. *Journal of Ecology* 95:231–234.

Violle, C., M.-L. Navas, D. Vile, E. Kazakou, C. Fortunel, I. Hummel, and E. Garnier. 2007. Let the concept of trait be functional! *Oikos* 116:882–892.

Westoby, M. 1998. A leaf-height-seed (LHS) plant ecology strategy scheme. *Plant and soil* 199:213–227.

Winemiller, K. O., D. B. Fitzgerald, L. M. Bower, and E. R. Pianka. 2015. Functional traits, convergent evolution, and periodic tables of niches. *Ecology Letters* 18:737–751.

Winemiller, K. O., and K. A. Rose. 1992. Patterns of life-history diversification in north american fishes: implications for population regulation. *Canadian Journal of Fisheries and Aquatic Sciences* 49:2196–2218.

Appendix A: Poisson approximation

For each genotype's dispersal, the counts of propagules across unoccupied territories follows a multinomial distribution with equal probabilities of landing in each territory. Thus, the x_i in different territories are not independent random variables. However, for sufficiently large T , holding n_i/T fixed, the Poisson limit theorem implies that this multinomial distribution for the x_i accross territories is closely approximated by a product of independent Poisson distributions for each territory, each with rate parameter l_i . Here we have used the fact that large T implies large U except in the biologically uninteresting case that there is vanishing population turnover $d_i \sim 1/T$.

Under the Poisson approximation, the total number of genotype i propagules $\sum x_i$ (summed over unoccupied territories) will deviate about its mean value m_i . Since the coefficient of variation of $\sum x_i$ is proportional to $1/\sqrt{m_i}$, these fluctuations are negligible unless m_i is very small (say of order 10^2 or less). These fluctuations in m_i could be regarded as a feature rather than a flaw since

having m_i be exactly constant per generation (for given b_i and n_i) is biologically unrealistic. In the canonical model of genetic drift, the Wright-Fisher model, the number of offspring per genotype fluctuates from generation to generation and is approximately Poisson distributed. Nevertheless, for simplicity and ease of comparison with the classic lottery model, we ignore fluctuations in m_i and only account for Poisson fluctuations in the number of propagules landing in each territory.

Appendix B: Derivation of growth equation

We separate the right hand side of Eq. (1) into three components $\Delta_+ n_i = \Delta_u n_i + \Delta_r n_i + \Delta_a n_i$ which vary in relative magnitude depending on the propagule densities l_i . Following the notation in the main text, the Poisson distributions for the x_i (or some subset of the x_i) will be denoted p , and we use P as a general shorthand for the probability of particular outcomes.

Growth without competition

The first component, $\Delta_u n_i$, accounts for territories where only one focal propagule is present $x_i = 1$ and $x_j = 0$ for $j \neq i$ (u stands for “uncontested”). The proportion of territories where this occurs is $l_i e^{-L}$, and so

$$\Delta_u n_i = U l_i e^{-L} = m_i e^{-L}. \quad (11)$$

Competition when rare

The second component, $\Delta_r n_i$, accounts for territories where a single focal propagule is present along with at least one non-focal propagule (r stands for “rare”) i.e. $x_i = 1$ and $X_i \geq 1$ where $X_i = \sum_{j \neq i} x_j$ is the number of nonfocal propagules. The number of territories where this occurs is $U p_i(1) P(X_i \geq 1) = b_i n_i e^{-l_i} (1 - e^{-(L-l_i)})$. Thus

$$\Delta_r n_i = m_i e^{-l_i} (1 - e^{-(L-l_i)}) \left\langle \frac{c_i}{c_i + \sum_{j \neq i} c_j x_j} \right\rangle_{\bar{p}}, \quad (12)$$

where $\langle \rangle_{\tilde{p}}$ denotes the expectation with respect to \tilde{p} , and \tilde{p} is the probability distribution of
 576 nonfocal propagule abundances x_j *after* dispersal, in those territories where exactly one focal
 propagule, and at least one non-focal propagule, landed.

We will show that, with respect to \tilde{p} , the standard deviation $\sigma_{\tilde{p}}(\sum_{j \neq i} c_j x_j)$, is much smaller
 579 than $\langle \sum_{j \neq i} c_j x_j \rangle_{\tilde{p}}$. Then x_j can be replaced by its mean in the last term in Eq. (12),

$$\left\langle \frac{c_i}{c_i + \sum_{j \neq i} c_j x_j} \right\rangle_{\tilde{p}} \approx \frac{c_i}{c_i + \sum_{j \neq i} c_j \langle x_j \rangle_{\tilde{p}}}. \quad (13)$$

We first calculate $\langle x_j \rangle_{\tilde{p}}$. Let $X = \sum_j x_j$ denote the total number of propagules in a territory
 and $\mathbf{x}_i = (x_1, \dots, x_{i-1}, x_{i+1}, \dots, x_G)$ denote the vector of non-focal abundances, so that $p(\mathbf{x}_i) =$
 $p_1(x_1) \dots p_{i-1}(x_{i-1}) p_{i+1}(x_{i+1}) \dots p_G(x_G)$. Then, \tilde{p} can be written as

$$\begin{aligned} \tilde{p}(\mathbf{x}_i) &= p(\mathbf{x}_i | X \geq 2, x_i = 1) \\ &= \frac{P(\mathbf{x}_i, X \geq 2 | x_i = 1)}{P(X \geq 2)} \\ &= \frac{1}{1 - (1 + L)e^{-L}} \sum_{X=2}^{\infty} P(X) p(\mathbf{x}_i | X_i = X - 1), \end{aligned} \quad (14)$$

and so

$$\begin{aligned} \langle x_j \rangle_{\tilde{p}} &= \sum_{\mathbf{x}_i} \tilde{p}(\mathbf{x}_i) x_j \\ &= \frac{1}{1 - (1 + L)e^{-L}} \sum_{X=2}^{\infty} P(X) \sum_{\mathbf{x}_i} p(\mathbf{x}_i | X_i = X - 1) x_j. \end{aligned} \quad (15)$$

The inner sum over \mathbf{x}_i is the mean number of propagules of a given nonfocal type j that will
 be found in a territory which received $X - 1$ nonfocal propagules in total, which is equal to

$\frac{l_j}{L-l_i}(X-1)$. Thus,

$$\begin{aligned}\langle x_j \rangle_{\tilde{p}} &= \frac{l_j}{1 - (1+L)e^{-L}} \frac{1}{L-l_i} \sum_{k=2}^{\infty} P(X)(X-1) \\ &= \frac{l_j}{1 - (1+L)e^{-L}} \frac{L-1+e^{-L}}{L-l_i},\end{aligned}\tag{16}$$

where the last line follows from $\sum_{X=2}^{\infty} P(X)(X-1) = \sum_{X=1}^{\infty} P(X)(X-1) = \sum_{X=1}^{\infty} P(X)X - \sum_{X=1}^{\infty} P(X)$.

For analyzing the relative fluctuations in $\sum_{j \neq i} c_j x_j$, Eq. (16) is unnecessarily complicated. We instead use the following approximation. Rather than evaluating the situation in each territory after dispersal as above, we replace \tilde{p} by \tilde{q} , defined as the x_i Poisson dispersal probabilities conditional on $X_i \geq 1$, independently of the outcome of x_i . This gives $\langle x_j \rangle_{\tilde{q}} = \langle x_j \rangle_p / C = l_j / C$,

$$\begin{aligned}\sigma_{\tilde{q}}^2(x_j) &= \langle x_j^2 \rangle_{\tilde{q}} - \langle x_j \rangle_{\tilde{q}}^2 \\ &= \frac{1}{C} \langle x_j^2 \rangle_p - \frac{l_j^2}{C^2} \\ &= \frac{1}{C} (l_j^2 + l_j) - \frac{l_j^2}{C^2} \\ &= \frac{l_j^2}{C} \left(1 - \frac{1}{C}\right) + \frac{l_j}{C},\end{aligned}\tag{17}$$

and

$$\begin{aligned}\sigma_{\tilde{q}}(x_j, x_k) &= \langle x_j x_k \rangle_{\tilde{q}} - \langle x_j \rangle_{\tilde{q}} \langle x_k \rangle_{\tilde{q}} \\ &= \frac{1}{C} \langle x_j x_k \rangle_p - \frac{l_j l_k}{C^2} \\ &= \frac{l_j l_k}{C} \left(1 - \frac{1}{C}\right),\end{aligned}\tag{18}$$

582 where $C = 1 - e^{-(L-l_i)}$ and $j \neq k$. The distribution \tilde{q} only approximates the situation after dispersal, since knowing that one focal genotype is among the propagules present restricts the possible outcomes for the x_j , so that the x_j cannot strictly be treated as independent of x_i . This

seemingly minor distinction has meaningful consequences. To illustrate, suppose that the focal genotype is rare and the propagule density is high ($l_j \approx L \gg 1$). Then Eq. (16) correctly predicts that there are on average $L - 1$ nonfocal propagules $\langle x_j \rangle_{\tilde{p}} \approx L - 1$, with the focal propagule correctly excluded, whereas \tilde{q} predicts one extra $\langle x_j \rangle_{\tilde{q}} \approx L$. As a result, \tilde{q} gives pathological behavior for rare invaders (they have a rarity disadvantage), but its moments are quantitatively similar enough to those of \tilde{p} that it is sufficient for analyzing the relative fluctuations in $\sum_{j \neq i} c_j x_j$.

Decomposing the variance in $\sum_{j \neq i} c_j x_j$,

$$\sigma_{\tilde{q}}^2(\sum_{j \neq i} c_j x_j) = \sum_{j \neq i} \left[c_j^2 \sigma_{\tilde{q}}^2(x_j) + 2 \sum_{k > j} c_j c_k \sigma_{\tilde{q}}(x_j, x_k) \right], \quad (19)$$

and using the fact that $\sigma_{\tilde{q}}(x_j, x_k)$ and the first term in Eq. (17) are negative because $C < 1$, we obtain an upper bound on the relative fluctuations in $\sum_{j \neq i} c_j x_j$,

$$\frac{\sigma(\sum_{j \neq i} c_j x_j)}{\langle \sum_{j \neq i} c_j x_j \rangle} < C^{1/2} \frac{\left(\sum_{j \neq i} c_j^2 l_j \right)^{1/2}}{\sum_{j \neq i} c_j l_j}. \quad (20)$$

Without loss of generality, we restrict attention to the case that the total nonfocal density $L - l_i$ is of order 1 or larger (otherwise $\Delta_r n_i$ does not contribute significantly to $\Delta_+ n_i$ because $\Delta_r n_i$ is proportional to $C = 1 - e^{-(L-l_i)}$).

Then, when at least some of the nonfocal propagule densities are large $l_j \gg 1$, the RHS of Eq. (20) is $\ll 1$, as desired. This is also the case if none of the nonfocal genotype densities are large and the c_j are all of similar magnitude (their ratios are of order one); the worst case scenario occurs when $L - l_i \sim O(1)$, in which case the negative covariances (Eq. (18)) which were neglected in the RHS of Eq. (20) significantly reduce the overall variance $\sigma_{\tilde{q}}^2(\sum_{j \neq i} c_j x_j)$.

However, the relative fluctuations in $\sum_{j \neq i} c_j x_j$ can be large if some of the c_j are much larger than the others. Specifically, in the presence of a rare, extremely strong competitor ($c_j l_j \gg c_{j'} l_{j'}$ for all other nonfocal genotypes j' , and $l_j \ll 1$), then the RHS of Eq. (20) can be large and we cannot make the replacement Eq. (13).

Substituting Eqs. (13) and (16) into Eq. (12), we obtain

$$\Delta_r n_i \approx m_i R_i \frac{c_i}{c}, \quad (21)$$

where R_i is defined in Eq. (6).

Competition when abundant

609 The final contribution, $\Delta_a n_i$, accounts for territories where two or more focal propagules are present (a stands for “abundant”). Similarly to Eq. (12), we have

$$\Delta_a n_i = U(1 - (1 + l_i)e^{l_i}) \left\langle \frac{c_i x_i}{\sum_j c_j x_j} \right\rangle_{\hat{p}} \quad (22)$$

612 where \hat{p} is the probability distribution of both focal and nonfocal propagaule abundances *after* dispersal in those territories where at least two focal propagules landed.

Again, we show that the relative fluctuations in $\sum c_j x_j$ are much smaller than 1 (with respect to \hat{p}), so that,

$$\left\langle \frac{c_i x_i}{\sum_j c_j x_j} \right\rangle_{\hat{p}} \approx \frac{c_i \langle x_i \rangle_{\hat{p}}}{\sum_j c_j \langle x_j \rangle_{\hat{p}}}. \quad (23)$$

Following a similar procedure as for $\Delta_r n_i$, where the vector of propagule abundances is denoted \mathbf{x} , the mean focal genotype abundance is,

$$\begin{aligned} \langle x_i \rangle_{\hat{p}} &= \sum_{\mathbf{x}} x_i p(\mathbf{x} | x_i \geq 2) \\ &= \sum_{x_i} x_i p(x_i | x_i \geq 2) \\ &= \frac{1}{1 - (1 + l_i)e^{-l_i}} \sum_{x_i \geq 2} p(x_i) x_i \\ &= l_i \frac{1 - e^{-l_i}}{1 - (1 + l_i)e^{-l_i}}. \end{aligned} \quad (24)$$

For nonfocal genotypes $j \neq i$, we have

$$\begin{aligned}
\langle x_j \rangle_{\hat{p}} &= \sum_{\mathbf{x}} x_j p(\mathbf{x} | x_i \geq 2) \\
&= \sum_X P(X | x_i \geq 2) \sum_{\mathbf{x}} x_j p(\mathbf{x} | x_i \geq 2, X) \\
&= \sum_X P(X | x_i \geq 2) \sum_{x_i} p(x_i | x_i \geq 2, X) \sum_{\mathbf{x}_i} x_j p(\mathbf{x}_i | X_i = X - x_i) \\
&= \sum_X P(X | x_i \geq 2) \sum_{x_i} p(x_i | x_i \geq 2, X) \frac{l_j(X - x_i)}{L - l_i} \\
&= \frac{l_j}{L - l_i} \left[\sum_X P(X | x_i \geq 2) X - \sum_{x_i} p(x_i | x_i \geq 2) x_i \right] \\
&= \frac{l_j}{L - l_i} \left(L \frac{1 - e^{-L}}{1 - (1 + L)e^{-L}} - l_i \frac{1 - e^{-l_i}}{1 - (1 + l_i)e^{-l_i}} \right). \tag{25}
\end{aligned}$$

615 To calculate the relative fluctuations in $\sum_{j \neq i} c_j x_j$, we use a similar approximation as for $\Delta_r n_i$: \hat{p} is approximated by \hat{q} , defined as the \mathbf{x} dispersal probabilities in a territory conditional on $x_i > 2$ (that is, treating the x_j as independent of x_i). All covariances between nonfocal genotypes are
618 now zero, so that $\sigma^2(\sum c_j x_j) = \sum c_j^2 \sigma^2(x_j)$, where $\sigma^2(x_j) = l_j$ for $j \neq i$. The expression for $\sigma^2(x_i)$ is more complicated, but in the relevant regime where $p(x_i = 0) \approx 0$ (since otherwise $D \gg 1$ and Δn_a is negligible), then

$$\sigma_{\hat{q}}^2(x_i) \approx \frac{l_i^2}{D} \left(1 - \frac{1}{D} \right) + \frac{l_i}{D}, \tag{26}$$

621 where $D = 1 - (1 + l_i)e^{-l_i}$, analogous to Eq. (17), and

$$\frac{\sigma_{\hat{q}}(\sum c_j x_j)}{\langle \sum c_j x_j \rangle} \approx \frac{\left(\sum_{j \neq i} c_j^2 l_j + c_i^2 \sigma_{\hat{q}}^2(x_i) \right)^{1/2}}{\sum_{j \neq i} c_j l_j + c_i l_i / D}. \tag{27}$$

Similarly to Eq. (20), the RHS of (27) will not be $\ll 1$ in the presence of a rare, extremely strong competitor. When this is not the case, then since l_i must be of order 1 or larger for $\Delta_a n$ to
624 make an appreciable contribution to $\Delta_+ n_i$, the RHS of Eq. (27) is $\ll 1$ as desired.

Combining Eqs. (22) and (23), we obtain

$$\Delta_a n_i = m_i A_i \frac{c_i}{c}, \quad (28)$$

where A_i is defined in Eq. (7).