

Spatial distribution of environmental DNA in a nearshore marine habitat

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Abstract

In the face of increasing threats to biodiversity, the advancement of methods for surveying biological communities is a major priority for ecologists. Recent advances in molecular biological technologies have made it possible to detect and sequence DNA from environmental samples (environmental DNA or eDNA); however, eDNA techniques have not yet seen widespread adoption as a routine method for biological surveillance primarily due to gaps in our understanding of the dynamics of eDNA in space and time. In order to identify the effective spatial scale of this approach in a dynamic marine environment, we collected marine surface water samples from transects ranging from the intertidal zone to 4 kilometers from shore. Using PCR primers that target a diverse assemblage of metazoans, we amplified a region of mitochondrial 16S rDNA from the samples and sequenced the products on an Illumina platform in order to detect communities and quantify their spatial patterns using a variety of statistical tools. We find evidence for multiple, discrete eDNA communities in this habitat, and show that these communities decrease in similarity as they become further apart. Offshore communities tend to be richer but less even than those inshore, though diversity was not spatially autocorrelated. Taxon-specific relative abundance coincided with our expectations of spatial distribution in taxa lacking a microscopic, pelagic life-history stage, though most of the taxa detected do not meet these criteria. Finally, we use carefully replicated laboratory procedures to show that laboratory treatments were remarkably similar in most cases, while allowing us to detect a faulty replicate, emphasizing the importance of replication to metabarcoding studies. While there is much work to be done before eDNA techniques can be confidently deployed as a standard method for ecological monitoring, this study serves as a first analysis of diversity at the fine spatial scales relevant to marine ecologists and confirms the promise of eDNA in dynamic environments.

Introduction

The patterns and causes of variability in ecological communities across space are both seminal and contentious areas of study in ecology (Hubbell, 2001; Anderson et al., 2011). One consistently observed pattern of community spatial heterogeneity is that communities close to one another tend to be more similar than those that are farther apart (Nekola and White, 1999). This decrease in community similarity with increasing spatial separation is called distance decay and has been

29 reported from communities of tropical trees (Condit, 2002; Chust et al., 2006), ectomycorrhizal fungi
30 (Bahram et al., 2013), salt marsh plants (Guo et al., 2015), and microorganisms (Martiny et al.,
31 2011; Chust et al., 2013; Wetzel et al., 2012; Bell, 2010). Typically, this relationship is assessed by
32 regressing a measure of community similarity against a measure of spatial separation for a set of
33 sites at which a set of species' abundances (or presences) is calculated. Yet no existing biodiversity
34 survey method completely censuses all of the organisms in a given area. The lack of a single 'silver
35 bullet' method of sampling contributes inconclusiveness to the study of spatial patterning in ecology
36 (Levin, 1992), and leaves open the possibility of new and more comprehensive methods.

37 From a boat or aircraft, scientists can count whales by sight, but not the krill on which they
38 feed. For example, towed fishing nets can efficiently sample organisms larger than the mesh and
39 slower than the boat, but overlook viruses and have undesirable effects on charismatic air-breathing
40 species. However, DNA-based surveys show great promise as an efficient technique for detecting a
41 previously unthinkable breadth of organisms from a single sample.

42 Microbiologists have used nucleic acid sequencing to quantify the composition and function of
43 microbial communities in a wide variety of habitats (Handelsman et al., 1998; Tyson et al., 2004;
44 Venter et al., 2004; Iverson et al., 2012). To do so, microorganisms are collected in a sample of
45 environmental medium (e.g. water), their DNA or RNA is isolated and sequenced, and the identity
46 and abundance of sequences is considered to reflect the community of organisms contained in the
47 sample, which indirectly estimates the quantity of organisms in an area.

48 Macroorganisms shed DNA-containing cells into the environment (environmental DNA or eDNA)
49 that can be sampled in the same way (Ficetola et al., 2008; Thomsen et al., 2012). Potentially, eDNA
50 methods allow a broad swath of macroorganisms to be surveyed from basic environmental samples.
51 However, the accuracy and reliability of indirect estimates of macroorganismal abundance has been
52 debated because the entire organisms are not contained within the sample (Coward et al., 2015).
53 Concern surrounding eDNA methods is rooted in uncertainty about the attributes of eDNA in the
54 environment relative to actual organisms (Shelton et al., 2016; Evans et al., 2016). Basic questions
55 such as how long DNA can persist in that environment and how far DNA can travel remain largely
56 unknown (but see Klymus et al. (2015); Turner et al. (2015); Strickler et al. (2015); Deiner and
57 Altermatt (2014)) and impede inference about local organismal presence from an environmental
58 sample. As a result, estimating the spatial and temporal resolution of eDNA studies in the field is

a key step in making these methods practical.

The relationship between local organismal abundance and eDNA is further complicated in habitats where the environmental medium itself may transport eDNA away from its source. We know that genetic material can move away from its source precisely because organisms can be detected indirectly without being present in the sample (Kelly et al., 2016b). One might reasonably expect eDNA to travel farther in a highly dynamic fluid such as the open ocean or flowing river than it would through the sediment at the bottom of a stagnant pond (Deiner and Altermatt, 2014; Shogren et al., 2016). Yet even studies of extremely dynamic habitats such as coastlines with high wave energy have found remarkable evidence that eDNA transport is limited enough that DNA methods can detect differences among communities separated by less than 100 meters (Port et al., 2016).

While rigorous laboratory studies have investigated the effects of some environmental factors on eDNA persistence (Klymus et al., 2015; Barnes et al., 2014; Sassoubre et al., 2016) and the transport of eDNA in specific contexts (Deiner and Altermatt, 2014), we suggest that field studies comparing the spatial distribution of communities of eDNA with expectations based on prior knowledge of organisms' distributions are also critical to developing a working understanding of eDNA in the real world. Research to date has documented the non-random spatial distribution of meiofaunal (Fonseca et al., 2014; Guardiola et al., 2016), microbial (Lallias et al., 2015), and extracellular (Guardiola et al., 2015) eDNA of marine and estuarine sediments, and of microscopic plankton in open ocean waters (de Vargas et al., 2015). These studies conducted targeted sampling at intermediate (thousands of meters) to global (thousands of kilometers) scales. Here, we use a grid-based environmental sampling strategy to assess spatial variability of eDNA in a coastal marine environment at a fine scale (tens to thousands of meters), using molecular methods that focus on macrobial metazoans.

We apply methods derived from community ecology to understand spatial patterns and patchiness of eDNA. The underlying mechanism thought to drive the slope of the distance decay relationship in ecological communities is the rate of movement of individuals among sites, which may be driven by underlying processes such as habitat suitability. Because eDNA is shed and transported away from its source, the increased movement of eDNA particles should homogenize community similarity, and thus erode the distance decay relationship of eDNA communities.

Puget Sound is a deep, narrow fjord in Washington, USA, where a narrow band of shallow

89 bottom hugs the shoreline and abruptly gives way to a central depth of up to 300 meters. This
 90 form allows the juxtaposition of communities associated with distinctly different habitats: shallow,
 91 intertidal benthos, and euphotic pelagic (Burns, 1985). At the upper reaches of the intertidal, the
 92 shoreline substrate varies from soft, fine sediment to cobble and boulder rubble. Soft intertidal
 93 sediments are inhabited by burrowing bivalves (Bivalvia), segmented worms (Annelida), and acorn
 94 worms (Enteropneusta), and in some lower intertidal and high subtidal ranges by eelgrass (*Zostera*
 95 *marina*) (Kozloff, 1973; Dethier, 2010) . Eelgrass meadows harbor epifaunal and infaunal biota,
 96 and attract transient species which use the meadows for shelter and to feed on resident organisms.
 97 Hard intertidal surfaces support a well-documented biota including barnacles (Sessilia) and other
 98 crustaceans, mussels (Bivalvia:Mytilidae), anemones (Actiniaria), sea stars (Asteroidea), urchins
 99 (Echinoidea), Bryozoans (Ectoprocta), crustaceans (Decapoda), and a variety of algae (Dethier,
 100 2010). Hard bottoms of the lower intertidal and high subtidal are home to macroalgae such as
 101 Laminariales and Desmarestiales which provide habitat for a distinct community of fish and in-
 102 vertebrates. The upper pelagic is home to a diverse assemblage of microscopic plankton including
 103 diatoms, copepods, and larvae (Strickland, 1983), as well as transitory fish and marine mammals.
 104 We took advantage of this setting to explore the spatial variation and distribution of marine
 105 eDNA communities. Using PCR-based methods and massively parallel sequencing, we surveyed
 106 mitochondrial 16S sequences from a suite of marine animals in water samples collected over a grid of
 107 sites extending from the shoreline out to 4 kilometers offshore in Puget Sound, Washington, USA. We
 108 leverage this sampling design to perform an explicitly spatial analysis of eDNA-derived community
 109 similarity. We investigate two primary objectives. First we examine the spatial patterning of
 110 eDNA and determine the degree to which eDNA community similarity can be predicted by physical
 111 proximity. We expect that physical proximity will be a strong predictor of community similarity,
 112 and that community differences can be detected over small distances. Second, we examine the
 113 distribution of diversity from eDNA data, and compare it to our expectations based on distributions
 114 of macrobial communities. We expect that distinct eDNA communities exist in this setting, and
 115 that their spatial distribution coincides with that of adult macrobial organisms. Because of the
 116 vastly different communities of benthic macrobial metazoans as a function of distance from shore,
 117 we expect that more than one eDNA community is present across our 4 kilometer sampling grid,
 118 and that communities change as a function of distance from shore. For this reason, we examine two

119 diversity measures of eDNA communities that have been widely used to reveal broad scale patterns
120 based on macrobiota in many ecological systems. Finally, we identify the taxa represented in the
121 eDNA communities, which span a range of life-history characteristics, and we expect that the spatial
122 distribution of eDNA will most closely resemble the distribution of adults in taxa with low dispersal
123 potential.

124 **Methods**

125 There are seven discrete steps to our methodology: (1) Environmental sample collection, (2) isolation
126 of particulates from water via filtration, (3) isolation of DNA from filter membrane, (4) amplification
127 of target locus via PCR, (5) sequencing of amplicons, (6) bioinformatic translation of raw sequence
128 data into tables of sequence abundance among samples, and (7) community ecological analyses of
129 eDNA. We provide brief overviews of these steps here, and encourage the reader to review the fully
130 detailed methods presented in the supplementary material (??).

131 **Environmental Sampling**

132 Starting from lower-intertidal patches of *Zostera marina*, we collected water samples at 1 meter
133 depth from 8 points (0, 75, 125, 250, 500, 1000, 2000, and 4000 meters) along three parallel transects
134 separated by 1000 meters (24 sample locations total; Figure 1). Samples were collected by attaching
135 bottles to a PVC pole and lowering it over the side of a boat over the span of one hour on 27 June
136 2014. To destroy residual DNA on equipment used for field sampling and filtration, we washed
137 with a 1:10 solution of household bleach (8.25% sodium hypochlorite; 7.25% available chlorine) and
138 deionized water, followed by thorough rinsing with deionized water. Each environmental sample
139 was collected in a clean 1 liter high-density polyethylene bottle, the opening of which was covered
140 with 500 micrometer nylon mesh to prevent entry of larger particles. Immediately after collecting
141 the sample, the mesh was replaced with a clean lid and the sample was held on ice until filtering.

142 **Filtration**

143 One liter from each water sample was filtered in the lab on a clean polysulfone vacuum filter
144 holder fitted with a 47 millimeter diameter cellulose acetate membrane with 0.45 micrometer pores.

145 Filter membranes were moved into 900 microliters of Longmire buffer (Longmire et al., 1997) using
146 clean forceps and stored at room temperature (Renshaw et al., 2015). To test for the extent of
147 contamination attributable to laboratory procedures, we filtered three replicate 1 liter samples of
148 deionized water. These samples were treated identically to the environmental samples throughout
149 the remaining protocols.

150 DNA Purification

151 DNA was purified from the membrane following a phenol:chloroform:isoamyl alcohol protocol fol-
152 lowing Renshaw (Renshaw et al., 2015). Preserved membranes were incubated at 65 °C for 30
153 minutes before adding 900 microliters of phenol:chloroform:isoamyl alcohol and shaking vigorously
154 for 60 seconds. We conducted two consecutive chloroform washes by centrifuging at 14,000 rpm for 5
155 minutes, transferring the aqueous layer to 700 microliters chloroform, and shaking vigorously for 60
156 seconds. After a third centrifugation, 500 microliters of the aqueous layer was transferred to tubes
157 containing 20 microliters 5 molar NaCl and 500 microliters 100% isopropanol, and frozen at -20 °C
158 for approximately 15 hours. Finally, all liquid was removed by centrifuging at 14000 rpm for 10
159 minutes, pouring off or pipetting out any remaining liquid, and drying in a vacuum centrifuge at 45
160 °C for 15 minutes. DNA was resuspended in 200 microliters of ultrapure water. Four replicates of
161 genomic DNA extracted from tissue of a species absent from the sampled environment (*Oreochromis*
162 *niloticus*) served as positive control for the remaining protocols.

163 PCR Amplification

164 We chose a primer set that amplifies an approximately 115 base pair (bp) region of the mitochon-
165 drial 16S rRNA gene in at least 10 metazoan phyla from this habitat, excludes non-metazoans, and
166 resolves taxonomy to the family level in most cases using a public sequence database (Kelly et al.,
167 2016a). We used a two-step polymerase chain reaction (PCR) protocol described by O'Donnell
168 et al. (2016) to generate 4 replicate products from each DNA sample. In the first set of reactions,
169 primers were identical in every reaction (forward: AGTTACYYTAGGGATAACAGCG; reverse:
170 CCGGTCTGAACTCAGATCAYGT); primers in the second set of reactions included these same
171 sequences but with 3 variable nucleotides (NNN) and an index sequence on the 5' end (see Sequenc-
172 ing Metadata). We used the program OligoTag (Coissac, 2012) to generate 30 unique 6-nucleotide

index sequences differing by a minimum Hamming distance of 3 (see Sequencing Metadata). Indexed primers were assigned to samples randomly, with the identical index sequence on the forward and reverse primer to avoid errors associated with dual-indexed multiplexing (Schnell et al., 2015). In a UV-sterilized hood, we prepared 25 microliter reactions containing 18.375 microliters ultrapure water, 2.5 microliters 10x buffer, 0.625 microliters deoxynucleotide solution (8 millimolar), 1 microliter each forward and reverse primer (10 micromolar, obtained lyophilized from Integrated DNA Technologies (Coralville, IA, USA)), 0.25 microliters Qiagen HotStar Taq polymerase, and 1.25 microliter genomic or eDNA template at 1:100 dilution in ultrapure water. PCR thermal profiles began with an initialization step (95 °C; 15 min) followed by cycles (40 and 20 for the first and second reaction, respectively) of denaturation (95 °C; 15 sec), annealing (61 °C; 30 sec), and extension (72 °C; 30 sec). 20 identical PCRs were conducted from each DNA extract using non-indexed primers; these were pooled into 4 groups of 5 in order to ensure ample template for the subsequent PCR with indexed primers. In order to isolate the fragment of interest from primer dimer and other spurious fragments generated in the first PCR, we used the AxyPrep Mag FragmentSelect-I kit with solid-phase reversible immobilization (SPRI) paramagnetic beads at 2.5x the volume of PCR product (Axygen BioSciences, Corning, NY, USA). A 1:5 dilution in ultrapure water of the product was used as template for the second reaction. PCR products of the second reaction were purified using the Qiagen MinElute PCR Purification Kit (Qiagen, Hilden, Germany). Ultrapure water was used in place of template DNA and run along with each batch of PCRs to serve as a negative control for PCR; none of these produced visible bands on an agarose gel. In total, four separate replicates from each of 31 DNA samples were carried through the two-step PCR process for a total of 124 sequenced PCR products. These were combined with additional samples from other projects, totaling 345 samples for sequencing.

DNA Sequencing

Up to 30 PCR products were combined according to their primer index in equal concentration into one of 14 pools, and 150 nanograms from each were prepared for library sequencing using the KAPA high-throughput library prep kit with real-time library amplification protocol (KAPA Biosystems, Wilmington, MA, USA). Each of these ligated sequencing adapters included an additional 6 base pair index sequence (NEXTflex DNA barcodes; BIOO Scientific, Austin, TX, USA). Thus, each

202 PCR product was identifiable via its unique combination of index sequences in the sequencing
203 adapters and primers. Fragment size distribution and concentration of each library was quantified
204 using an Agilent 2100 BioAnalyzer. Libraries were pooled in equal concentrations and sequenced
205 for 150 base pairs in both directions (PE150) using an Illumina NextSeq at the Stanford Functional
206 Genomics Facility, where 20% PhiX Control v3 was added to act as a sequencing control and to
207 enhance sequencing depth by increasing sequence diversity. Raw sequence data in fastq format is
208 publicly available (see Data Availability).

209 **Sequence Data Processing (Bioinformatics)**

210 Detailed bioinformatic methods are provided in the supplemental material, and analysis scripts
211 used from raw sequencer output onward can be found in the public project directory (see Analysis
212 Scripts). Briefly, we performed five steps to process the sequence data: (1) Merge paired-end reads
213 (Zhang et al., 2014), (2) eliminate low-quality reads (Edgar, 2010; Rognes et al., 2016), (3) eliminate
214 PCR artifacts (chimeras) (Edgar, 2010; Rognes et al., 2016; Martin, 2011), (4) cluster reads by
215 similarity into operational taxonomic units (OTUs) (Mahé et al., 2014), and (5) match observed
216 sequences to taxon names (Camacho et al., 2009; Chamberlain and Szöcs, 2013; Chamberlain et al.,
217 2016). Additionally, we checked for consistency among PCR replicates, excluded extremely rare
218 sequences, and rescaled (rarefied) the data to account for differences in sequencing depth. The data
219 for input to further analyses are a contingency table of the mean count of unique sequences, OTUs,
220 or taxa present in each environmental sample.

221 **Ecological Analyses**

222 After gathering the data, we use the eDNA community observed at each location to make inferences
223 about the spatial patterning of eDNA communities. We use statistical tools from community ecology
224 to assess the spatial structure of eDNA communities. We report similarity (1- dissimilarity) rather
225 than dissimilarity in all cases for ease of interpretation.

226 Objective 1: Community similarity as a function of distance

227 Distance Decay

228 To address our first objective and determine whether or not nearby samples are more similar than
229 distant ones, we fit a nonlinear model to represent decreasing community similarity with distance.
230 We calculated the pairwise Bray-Curtis similarity (1 - Bray-Curtis dissimilarity) between eDNA
231 communities using the R package *vegan* (Oksanen et al., 2016) and the great circle distance between
232 sampling points using the Haversine method as implemented by the R package *geosphere* (Hijmans,
233 2016). This model is similar to the Michaelis-Menten function, but with an asymptote fixed at 0:

$$y_{ij} = \frac{AB}{B + x_{ij}} \quad (1)$$

234 Where the relationship between community similarity (y_{ij}) and spatial distance (x_{ij}) between
235 observations i and j is determined by the similarity of samples at distance 0 (A), and the distance at
236 which half the total change in similarity is achieved (B). This allows for samples collected very close
237 together (near 0) to have similarity significantly less than one. We assessed model fit using the R
238 function *nls* (R Core Team, 2016), using the *nl2sol* algorithm from the *Port* library to solve separable
239 nonlinear least squares using analytically computed derivatives (<http://netlib.org/port/nsg.f>). We
240 set bounds of 0 and 1 for the intercept parameter and a lower bound of 0 for the distance at half
241 similarity; starting values of these parameters were 0.5 and $x_{max}/2$, respectively. We calculated
242 a 95% confidence interval for the parameters and the predicted values using a first-order Taylor
243 expansion approach implemented by the function *predictNLS* in the R package *propagate* (Spiess,
244 2014).

245 There are other conceptually reasonable forms to expect the space-by-similarity relationship
246 to take; we present these in the supplemental material along with alternative data subsets and
247 similarity indices.

248 **Objective 2: Spatial distribution of diversity**

249 **Community Classification**

250 To determine the spatial distribution and variation of eDNA communities (objective 2), we used
251 multivariate classification algorithms. We simultaneously assessed the existence of distinct com-
252 munity types and the membership of samples to those community types using an unsupervised
253 classification algorithm known as partitioning around medoids (PAM; sometimes referred to as k-
254 medoids clustering) (Kaufman and Rousseeuw, 1990), as implemented in the R package cluster
255 (Maechler et al., 2016). The classification of samples to communities was made on the basis of
256 their pairwise Bray-Curtis similarity, calculated using the function `vegdist` in the R package `vegan`
257 (Oksanen et al., 2016). Other distance metrics were evaluated but had no appreciable effect on the
258 outcome of the analysis (Supplemental Figure 1). In order to chose an optimal number of clusters
259 (K), we evaluated the distribution of silhouette widths, a measure of the similarity between each
260 sample and its cluster compared to its similarity to other clusters. We repeated the analysis using
261 fuzzy clustering (FANNY, (Kaufman and Rousseeuw, 1990); however, the results were qualitatively
262 similar to the results using PAM so we omit them here.

263 **Aggregate Measures of Diversity**

264 We calculated two measures of diversity, richness and evenness, to ask if aggregate metrics of the
265 eDNA community showed evidence of spatial patterning. Richness is a measure of the number of
266 distinct types of organisms present and so ranges from 1 (only one taxon observed) to S , the number
267 of taxa observed across all samples. To calculate the evenness of the distribution of abundance of
268 taxa in a sample, we used the complement of the Simpson (1949) index ($1 - \sum p_i^2$, where p_i is the
269 proportional abundance of taxon i). The values of this index ranges from 0 to 1, with the value
270 interpreted as the probability that two sequences randomly selected from the sample will belong to
271 different taxa; thus, larger values of the index indicate more evenly divided communities (Magurran,
272 2003). We calculated Moran's I for both diversity metrics to test for spatial autocorrelation. We
273 also tested for a linear effect of log-transformed distance from shore on each measure of diversity to
274 ask how diversity changes over this strong environmental gradient.

275 **Taxon and Life History Patterns**

276 After assigning taxon names to the abundance data, we plotted the distribution in space of a
277 selection of taxa to compare with our expectations on the basis of adult distributions (objective 2).
278 Our aim was to understand where each taxon occurred in the greatest proportional abundance, and
279 its distribution in space relative to that maximum. Thus, we rescaled each sample to proportional
280 abundance, extracted the data from a single taxon, and scaled those values between 0 and 1.
281 We collated life history characteristics for each of the major taxonomic groups recovered, including
282 dispersal range of the gametes, larvae, and adults, adult habitat type and selectivity, and adult body
283 size. For each life history stage of each taxon group, we made an order-of-magnitude approximation
284 of the scale of dispersal. For example, internally fertilized species were assigned a gamete range
285 of 0 km, while broadcast spawners were assigned a gamete range of 10 km. Similarly, adult range
286 size was approximated as 0 km (sessile), 1 km (motile but not pelagic), or 10 km (highly mobile,
287 pelagic). Variables were specified as 'multiple' for life history stages known to span more than 1
288 magnitude of range size. For groups to which sequences were annotated with high confidence, but
289 for which life history strategy is diverse or poorly known (e.g. families in the phylum Nemertea),
290 we used conservative, coarse approximations at a higher taxonomic rank (see Life History Data).
291 These data were used to contextualize group-specific spatial distributions and inform expectations
292 based on known adult distributions.

293 **Results**

294 **Sequence Data Processing (Bioinformatics)**

295 Preliminary sequence analysis strongly suggested that the observed variation among environmental
296 samples reflects true variation in the environment, rather than variability due to lab protocols, for
297 the following reasons (note that all value ranges are reported as mean plus and minus one standard
298 deviation). First, all libraries passed the FastQC per-base sequence quality filter, generating a total
299 of 371,576,190 reads passing filter generated in each direction. Second, samples in this study were
300 represented by an adequate number of reads ($333,537.9 \pm 112,200.5$), with no individual sample
301 receiving fewer than 130,402 reads. Third, there was a very low frequency of cross-contamination

302 from other libraries into those reported here ($5\text{e-}05 \pm 8\text{e-}05$; max proportion 0.00034). Fourth, after
303 scaling all samples to the same sequencing depth, OTUs with abundance greater than 178 reads
304 (0.14% of a sample's reads) experienced no turnover among PCR replicates within a sample. Fifth,
305 sequence abundances among PCR replicates within water samples were remarkably consistent. A
306 single sample had low similarity among PCR replicates (0.659) after removing this outlier, the
307 lowest mean similarity among replicates within a sample was 0.966. Overall similarities among
308 PCR replicates within a sample were extremely high (0.976 ± 0.013), and far higher than those
309 among samples (0.3 ± 0.16). Across PCR replicates, each sample was represented by at least 781425
310 reads in the raw data and contained between 111 and 443 rarefied OTUs (Supplemental Figure 2).

311 **Ecological Analyses**

312 **Distance Decay**

313 Physical proximity is a good predictor of eDNA community similarity: Similarity decreased from
314 0.40 (95%CI = 0.36, 0.45) to half that amount at 4500 meters (95%CI = 2900, 7500) (Figure 2).

315 **Community Classification**

316 Despite a clear trend in community similarity as a function of spatial separation, the results from
317 our classification analysis are difficult to interpret. The silhouette analysis indicated the presence
318 of 8 distinct communities; however, the gain in mean silhouette width from 2 was small (0.1), and
319 lacked a distinctive peak (Figure 3), indicating substantial uncertainty in the clustering algorithm.
320 Thus, we present the results of cluster assignment for both $K = 2$ and $K = 8$ to illustrate the
321 range of results (Figure 4). Excluding taxa which occur in only one site had no discernible effect
322 on the outcome of the PAM analysis (number of clusters, assignment to clusters). While there was
323 no distinct spatial divide indicating the presence of an inshore versus an offshore community, one
324 of the two communities (at $K = 2$) occurred in only 2 out of 18 samples inside 1000 meters from
325 shore, and never occurred within 125 meters of shore, suggesting the presence of an inshore and
326 offshore community.

327 Diversity in Space

328 Sites offshore tend to be less rich and more even than those inshore (Figure 5). Mean OTU richness
329 declined by 1.42 per 1000 meters from a mean of 17.6 taxa (95%CI = 2.15) inshore to 11.9 taxa
330 (95%CI = 4.31) at offshore locations ($p = 0.0415$; Figure 5). Evenness increased by .0666 per 1000
331 meters from 0.225 (95%CI = 0.0558) to 0.491 (95%CI = ± 0.112), indicating that sequence reads
332 were less evenly distributed among taxa in offshore samples ($p \ll 0.05$; Figure 5). The subset of data
333 used for this analysis had no qualitative effect on the outcome of this analysis (Supplemental Figure
334 3). There was no evidence for spatial autocorrelation for any of the diversity metrics (Moran's I, p
335 > 0.05 ; Figure 6).

336 Taxon and Life History Patterns

337 We were able to assign a taxon name with confidence to 136 of 146 OTU sequences. The vast ma-
338 jority of sequences (97.6%) and OTUs (96.9%) were matched to organisms that have high potential
339 for dispersal at either the gamete, larval, or adult stage, making it impossible to determine whether
340 the source of that DNA was adults with well-documented spatial patterns (e.g. sessile nearshore
341 specialists) or highly mobile early life history stages. Of the 6 OTUs for which dispersal is limited
342 during all life history stages, only 2 occurred in more than two samples, precluding a quantita-
343 tive comparison of spatial dispersion based on life history characteristics. These were assigned to
344 *Cymatogaster aggregata*, a viviparous nearshore fish with internal fertilization, and *Cupolaconcha*
345 *meroclista*, a sessile Vermetid gastropod with presumed internal fertilization and short larval dis-
346 persal (Strathmann and Strathmann, 2006; Phillips and Shima, 2010; Calvo and Templado, 2004).
347 *Cymatogaster aggregata* was distinctly more abundant close to shore, with no sequences occurring
348 in any sample beyond 250 meters (Figure 7). *Cupolaconcha meroclista* showed no such distinct
349 spatial trend, occurring in nearly equal abundance at three sites, 75, 500, and 2000 meters from
350 shore. An additional species that was highly abundant in the sequence data, the krill *Thysanoessa*
351 *raschii*, has pelagic adults, highly seasonal reproduction, and sinking eggs; their distribution was
352 consistent with our expectations based on a tendency of adults to aggregate offshore. Finally, the
353 two most abundant taxa in the dataset were the mussel genus *Mytilus* and the Barnacle order Ses-
354 silia; the adults of both taxa are sessile and occur exclusively on hard intertidal substrata but have

355 highly motile larvae. Because large-scale dispersal could not be ruled out for the vast majority of
356 taxa, subsetting the community data by taxonomic group had no qualitative effect on the spatial
357 patterning or diversity metrics, and we omit those results here.

358 Discussion

359 Indirect surveys of organismal presence are a key development in ecosystem monitoring in the face
360 of increased anthropogenic pressure and dwindling resources for ecological research. Monitoring
361 of organisms using environmental DNA is an especially promising method, given the rapid pace
362 of advancement in technological innovation and cost efficiency in the field of DNA sequencing and
363 quantification. We document four key patterns: (1) eDNA communities far from one another tend to
364 be less similar than those that are nearby, (2) distinct eDNA communities exist and are distributed
365 in a non-random fashion, (3) diversity declines with distance from shore, and (4) spatial patterning
366 of eDNA is associated with taxon-specific life history characteristics.

367 (1) Communities far from one another tend to be less similar than those that are 368 nearby

369 We demonstrate that distant locations have less-similar eDNA communities than proximate loca-
370 tions in Puget Sound, a dynamic marine environment. Our finding is in line with observations
371 based on traditional surveys of terrestrial plants and fungi (Nekola and White, 1999; Bahram et al.,
372 2013; Condit, 2002; Chust et al., 2006) and of microorganisms in freshwater (Wetzel et al., 2012),
373 marine (Chust et al., 2013), and estuarine (Martiny et al., 2011) environments. To our knowledge,
374 it is the first to report such a pattern using massively parallel sequencing of environmental DNA
375 in the marine environment, and the first using any technique to describe this pattern from mac-
376 robial metazoans. We note that the theoretical expectation is that samples at very close distance
377 be nearly completely similar, while our samples separated by the 50 meters were only 40% similar.
378 We interpret this to reflect the highly dynamic nature of this environment, which could cause DNA
379 to be distributed quickly from its source, eroding the rise in similarity at small distances. At the
380 same time, community similarity decreased to very low levels at larger scales, indicating that DNA
381 distribution is not completely unpredictable. This finding implies that the effectively sampled area

of individual water samples for eDNA analysis is likely to be quite small ($<100\text{m}$) in this nearshore environment. Our estimated distance-decay relationship does indicate that proximate samples are more similar than distant samples, but we suggest this pattern is partially obscured by other factors, including signal from mobile, microscopic life-stages.

(2) Distinct eDNA communities exist and are distributed in a non-random fashion

We demonstrate strong evidence for distinct community types and the non-random spatial patterning of those communities. While the spatial distributions of communities is surprising if one were concerned only with the macroscopic life stages of metazoans, it indeed does align with the broader view that even offshore pelagic communities are comprised of and influenced by nearshore organisms. This result underscores the idea that areas immediately offshore act as ecotones, a mixing zone of taxa characteristic of benthic and pelagic environments. While there was no distinct break in community types between onshore and offshore sites, there was some clustering of community types that may be explained by oceanographic features such as nearshore eddies generated by strong tidal exchange in a steep bathymetric setting (Yang and Khangaonkar, 2010). It would be useful to better understand such features during the period of sampling, by way of oceanographic monitoring devices. Finally, the uncertainty in identification of the number of distinct clusters to best characterize the community underlines the difficulty of identifying community patterns with the number of taxonomic groups considered here. We suspect that the signature of eDNA from microscopic life-stages may explain our inability to easily detect spatial community level patterns that align with our initial expectations.

(3) Richness declines and evenness increases with distance from shore

We found that richness declined while evenness increased with distance from shore. Such a pattern is consistent with many other ecosystems which show strong clines in diversity metrics over environmental gradients. However, our study is novel in that it corroborates a cline well-known on macroscales for macrobiota on a much smaller spatial scale for microscopic animals, suggesting that there may be a self-similarity across scales in diversity patterning (Levin, 1992). The coastal ocean is a highly productive and diverse ecosystem, where biomass is concentrated most heavily along the bottom and shoreline (Ray, 1988). This differential in biomass concentration from the

410 shoreline to open waters may contribute to the opposing trends we detected. Where particles (or-
411 ganisms, tissues, and cells) are sparse, fewer would be collected per sample of constant volume, thus
412 decreasing the probability of drawing as many types (richness) and increasing the probability that
413 any two particles originate from the same type (evenness). Intriguingly, the cline in diversity from
414 inshore to offshore was not determined by shared changes in communities as one moved offshore; the
415 classification analysis suggested a fair amount of differences among communities at a given offshore
416 distance (Figure 4).

417 **(4) Spatial patterning of eDNA is associated with taxon-specific life history**

418 In contrast to our expectations, other taxa including species with sessile adult stages restricted
419 to benthic hard substrates (e.g. barnacles, mussels) are among the most abundant taxa at sites
420 furthest from shore. However, the larvae and gametes of these taxa are abundant, pelagic, and
421 can be transported long distances by water movement (Strathmann, 1987). This indicates that we
422 likely detected DNA of their pelagic phase gametes and larvae. It is always possible that DNA
423 of adults was advected over long distances and detected offshore but in light of our results with
424 krill and surfperch, we view this as unlikely. We interpret our results as evidence that the chaotic
425 spatial distribution of eDNA communities (Figure 4) results from our primers' affinity for many
426 species which at some point exist as microscopic pelagic gametes or larvae. Our results emphasize
427 that expected results based on easily visually observed individuals or detectable with traditional
428 sampling gear such as nets may be very different from results using eDNA. This does caution that
429 eDNA surveys may have different purposes and may not be directly comparable to existing surveys
430 (Shelton et al., 2016).

431 We acknowledge that sampling artifacts may have affected our results. For example if entire
432 multicellular individuals were captured in our samples, their DNA could be in much greater density
433 than eDNA, affecting the observed community. Our sampling bottles excluded particles larger than
434 500 micrometers, but gametes and very small larvae could have gained entry. It is possible that
435 even a single small individual, containing many thousand mitochondria, would overwhelm the signal
436 of another species from which hundreds of cells had been sloughed from many, larger individuals.
437 Data on larval size distribution at the time of sampling from each species in our data set would
438 allow us to estimate the frequency of such events. Nevertheless, it is precisely the sensitivity to

439 small particles that makes the eDNA approach powerful, so we are reluctant to recommend that
440 aquatic eDNA sampling use finer pre-filtering. Instead, we emphasize the importance of designing
441 and selecting primer sets that selectively amplify target organisms. In the case of the present study,
442 in order to recover patterns matching our expectations, this would be non-transient, benthic marine
443 organisms lacking any pelagic life stage.

444 The marker we chose for this study detects a wide variety of metazoans while excluding other
445 more common taxa; however, it does not effectively discriminate among species within a higher group
446 in all cases. Other markers, such as mitochondrial cytochrome c oxidase subunit 1 (COX1, CO1,
447 or COI) may provide adequate species-level resolution in some metazoan groups, but have other
448 shortcomings including taxon dropout (Deagle et al., 2014) and amplification of more abundant non-
449 metazoans, as we discovered in an accompanying study (Kelly et al., 2016a). Both have undesirable
450 effects of biasing estimates of diversity. In our case, it is possible that the lumping of multiple
451 species into one group underestimates the true richness of the group and of the entire sample, in
452 turn obscuring true underlying patterns of diversity. In the case of COX1, well-documented primer
453 biases cause failure to amplify some taxa, particularly in mixed samples, with the same result
454 (Deagle et al., 2014). In fact, even surveys relying on traditional capture techniques (e.g. seine
455 nets) and morphological characteristics are subject to biases imposed by the sampling gear (e.g.
456 mesh size), the observer (e.g. taxonomic expertise), and organisms (e.g. morphologically cryptic
457 species). Similarly, no single molecular marker adequately and effectively samples all taxa without
458 bias (Drummond et al., 2015), and thus the choice of marker is an important and context-dependent
459 one. Until whole-genome sequencing of individual cells is a reality, the tradeoffs between taxonomic
460 breadth and resolution will continue to be problematic for metabarcoding studies, just as they are
461 for more traditional ecological survey methods (Kelly et al., 2016a).

462 Our results also highlight the need for curated life-history databases. As technological advances
463 increase the speed and throughput of DNA sequencing and sequence processing, making sense of
464 these data in a timely manner requires that natural history data be stored in standard formats in
465 centralized repositories. The rate at which we can make sense of high-throughput survey methods
466 will be limited by our ability to collate auxiliary data. Databases such as Global Biodiversity
467 Information Facility (GBIF), Encyclopedia of Life (EOL), and FishBase (Parr et al., 2014; Froese
468 and Pauly, 2016) contain records of taxonomy, occurrence, and other rudimentary data types, but

469 there is no centralized, standardized repository for even basic natural history data such as body
470 size. As NCBI's nucleotide and protein sequence database (GenBank) has facilitated transformative
471 studies in diverse fields, an ecological analog would be a boon for biodiversity science.

472 Surveys based on eDNA are intensely scrutinized because of the danger that the final data are
473 subject to complicated laboratory and bioinformatic procedures. Finding virtually no variability
474 among lab and bioinformatic treatments from the point of PCR onward, we were confident our
475 results represented actual field-based differences among samples. However, we note that one PCR
476 replicate had a clear signal of contamination in that the sequence community was extremely similar
477 to those from a different environmental sample. The source of this error is difficult to identify, but
478 seems most likely to be an error during PCR preparation, either in assignment or pipetting during
479 preparation of indexed primers. While the remainder of our results would be largely unchanged
480 had we sequenced a single replicate per environmental sample, we believe the sequencing of PCR
481 replicates is critical for ensuring data quality in eDNA sequencing studies.

482 While there is much work to be done before eDNA techniques can be confidently deployed as a
483 standard method for ecological monitoring, this study serves as a first analysis of diversity at the
484 fine spatial scales that are likely to be relevant to eDNA work in the field across a range of study
485 systems.

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490 **Data Availability**

491 **Sequence Data**

492 All sequence data are available from NCBI under BioProject PRJNA338801.

493 Scripts to process raw sequence data into the contingency tables used for ecological analyses can be
494 found at:

495 <https://github.com/jimmyodonnell/banzai>

496

497 **Project Repository**

498 The following components are available from the project repository on GitHub:

499 https://github.com/jimmyodonnell/Carkeek_eDNA_grid

500 <https://doi.org/10.5281/zenodo.242976>

501 **Sequencing Metadata**

502 Sequencing metadata is available in: `Data/metadata_spatial.csv`

503 **Life History Data**

504 Life history data is available in: `Data/life_history.csv`

505 **Analysis Scripts**

506 All analyses were performed using scripts available in the Analysis subdirectory of the project's
507 repository on GitHub.

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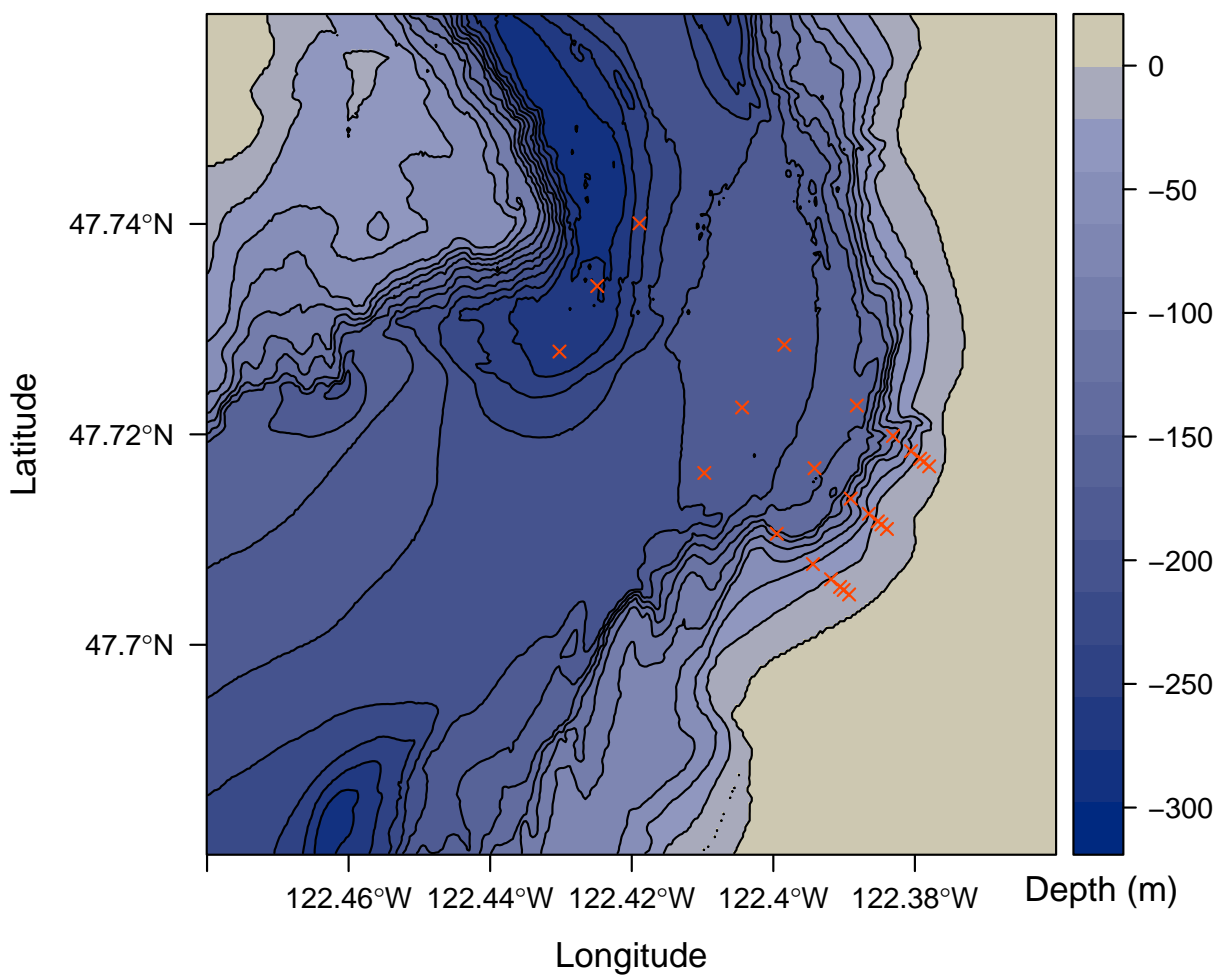


Figure 1: Map of study area. Depth in meters below sea level is indicated by shading and 25 meter contours. Sampled locations are indicated by red points.

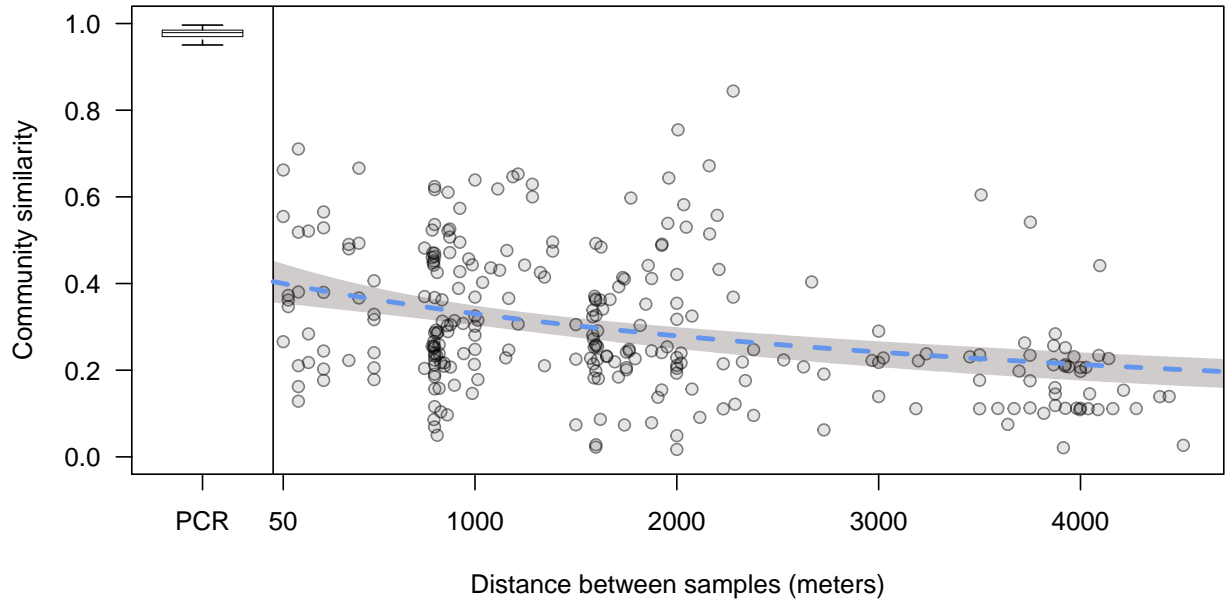


Figure 2: Distance decay relationship of environmental DNA communities. Each point represents the Bray-Curtis similarity of a site sampled along three parallel transects comprising a 3000 by 4000 meter grid. Blue dashed line represents fit of a nonlinear least squares regression (see Methods), and shading denotes the 95% confidence interval. Boxplot is comparisons within-sample across PCR replicates, separated by a vertical line at zero, where the central line is the median, the box encompasses the interquartile range, and the lines extend to 1.5 times the interquartile range. Boxplot outliers are omitted for clarity.

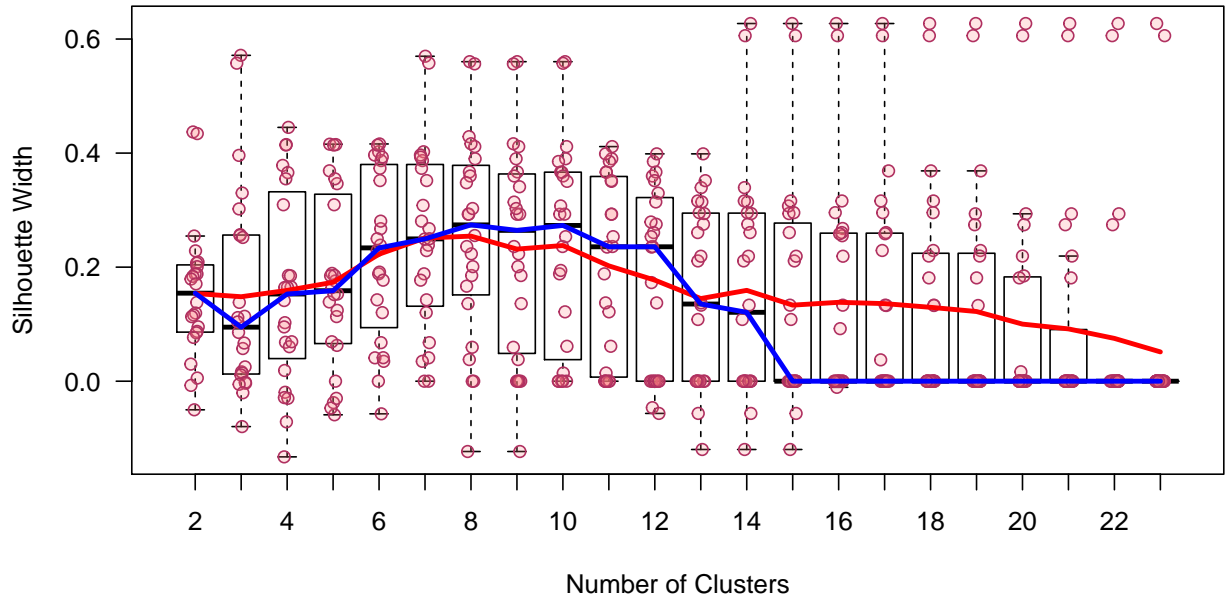


Figure 3: Silhouette widths from PAM analysis. Points are the width of the PAM silhouette of each sample at each number of clusters (K). Red line is the mean, blue line is the median. Boxes encompass the interquartile range with a line at the median, and the whiskers extend to 1.5 times the interquartile range. Boxplot outliers are omitted for clarity.

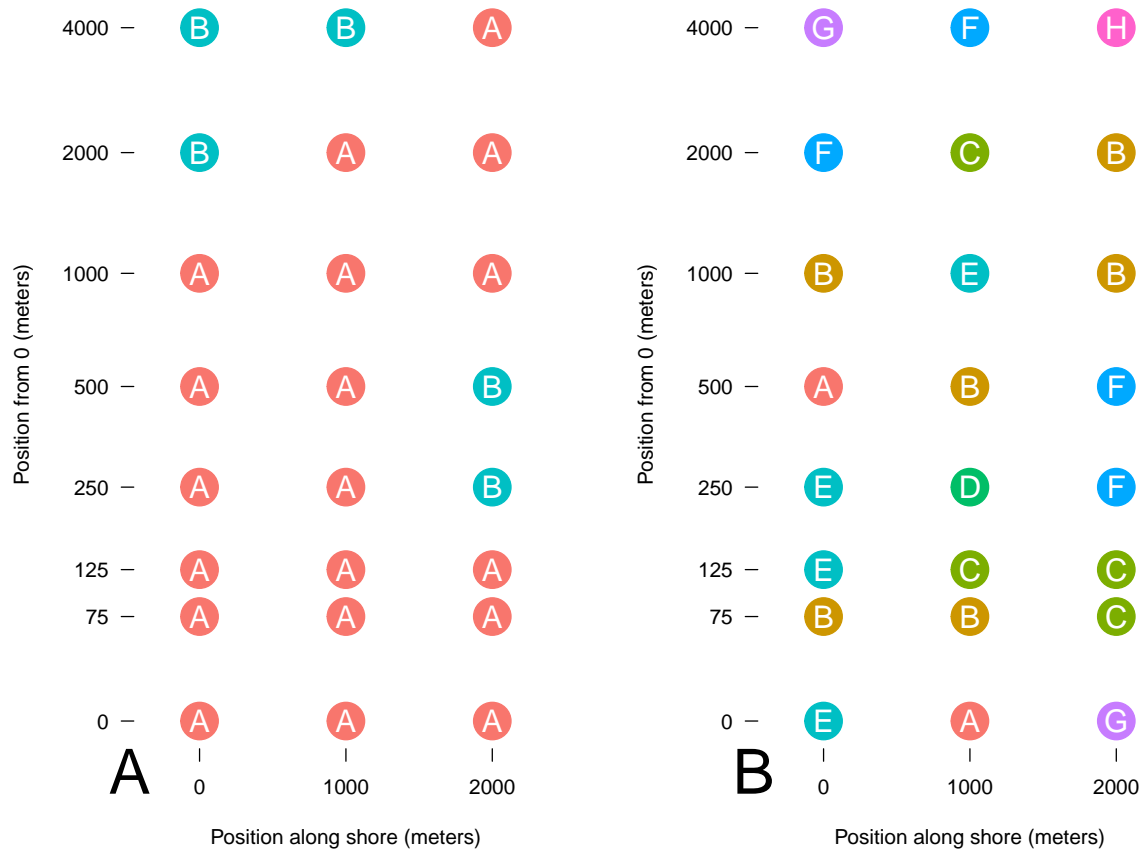


Figure 4: Cluster membership of sampled sites. Distance from onshore starting point is log scaled. Sites are colored and labeled by their assignment to a cluster by PAM analysis for number of clusters (K) chosen based on a priori expectations (A; 2) and mean silhouette width (B; 8).

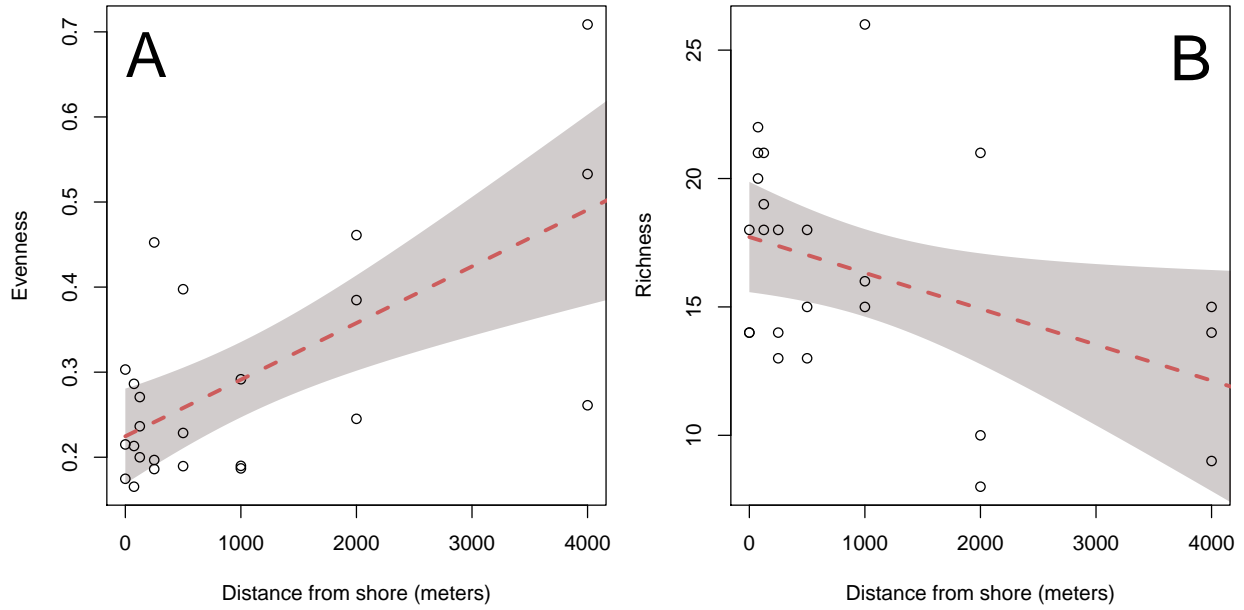


Figure 5: Aggregate diversity metrics of each site plotted against distance from shore. Both Simpson's Index of evenness (A) and richness (B) are shown, and have been computed from the mean abundance of unique DNA sequences found across 4 PCR replicates at each of 24 sites. Lines and bands illustrate the fit and 95

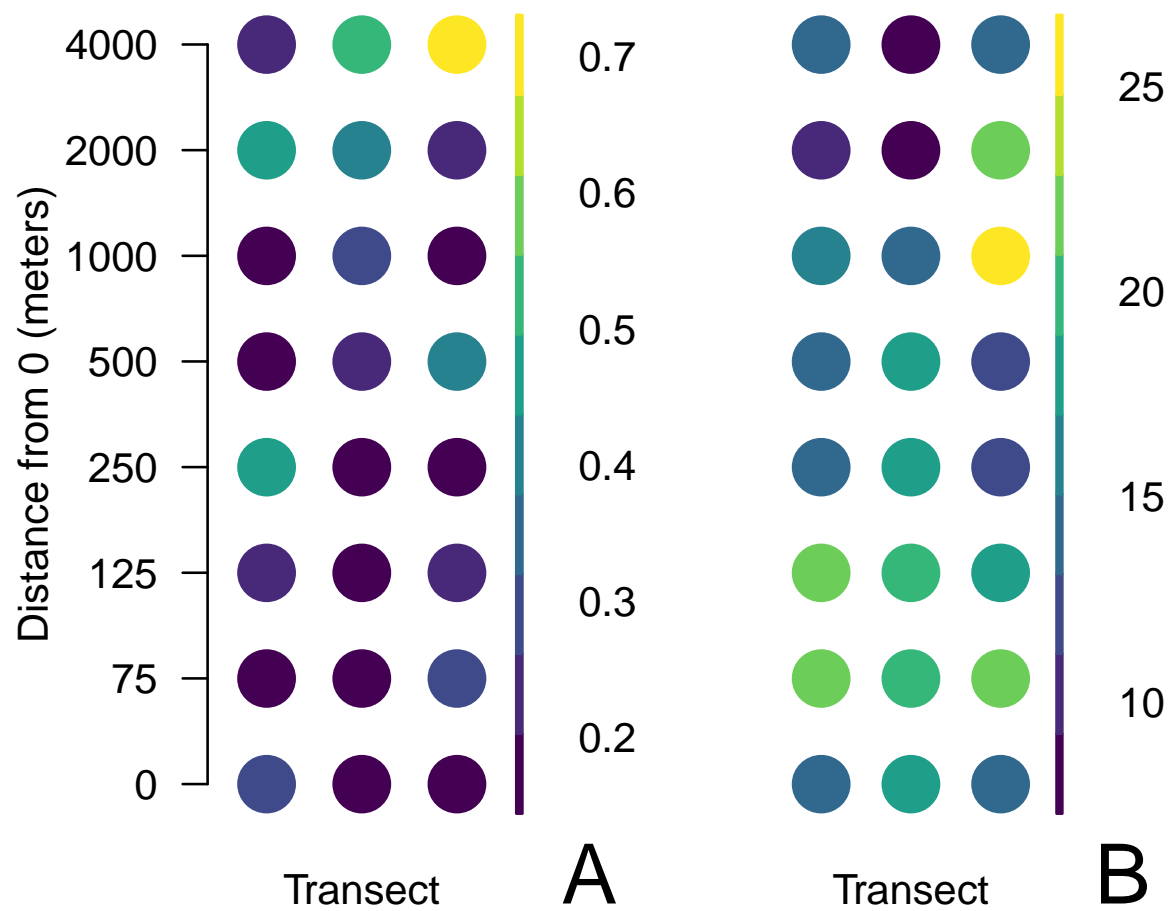


Figure 6: Aggregate measures of diversity at each sample site. Data are rarefied counts of mitochondrial 16S sequences collected from 3 parallel transects in Puget Sound, Washington, USA. Evenness (A) is the probability that two sequences drawn at random are different; richness (B) represents the total number of unique sequences from that location.

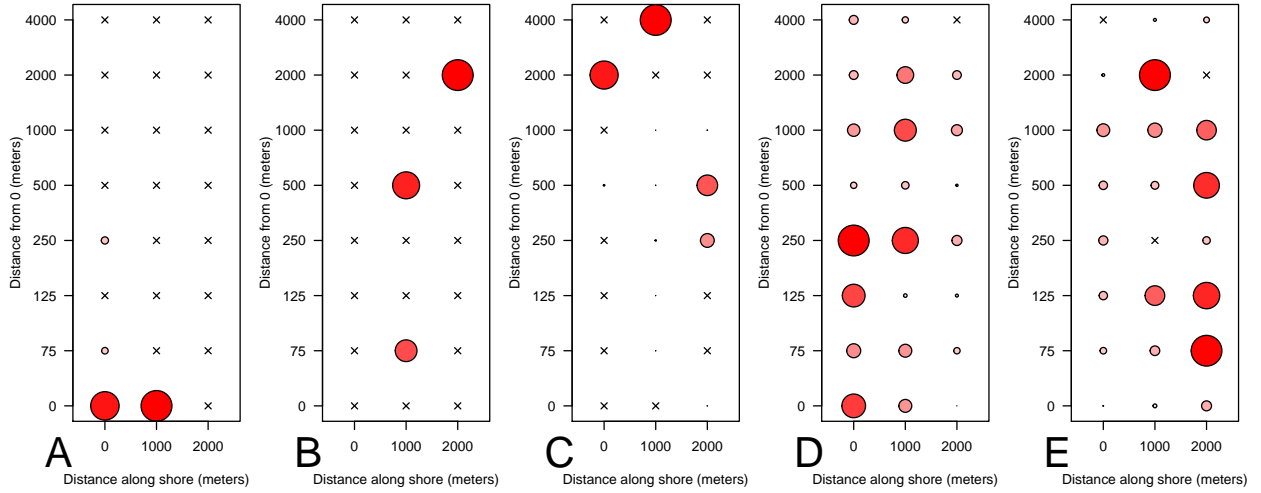


Figure 7: Distribution of eDNA from select taxa. Taxa represented are: Embiotocidae (A), Cupolaconcha meroclista (B), Thysanoessa raschii (C), Mytilus (D), Sessilia (E). Circles are colored and scaled by the proportion of that taxon’s maximum proportional abundance. That is, the largest circle is the same size in each of the panels, and occurs where that taxon contributed the greatest proportional abundance of reads to that sample.