

Estimation of woodland canopy structure with terrestrial LiDAR: expanded methods

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1 1 Introduction

This chapter provides expanded field and analytical methods for the study of tree canopy structure in southern African woodlands, presented in brief in Chapter 5. The study aimed to understand the effects of tree species diversity and stand structure on tree canopy structural complexity, using terrestrial LiDAR. Firstly, I provide technical details on the field setup for the terrestrial LiDAR and the hemispherical photography used to validate terrestrial LiDAR canopy closure estimates. Secondly, I describe the processing chain used to extract canopy complexity metrics from the terrestrial LiDAR point clouds. Thirdly, I describe in further detail the behaviour and suitability of the different canopy complexity metrics and stand structural metrics used in the study.

11 2 Terrestrial LIDAR field setup

Within each 1 ha (100x100 m) square plot, nine 10 m diameter circular subplots were laid out in a grid, with 35 m between subplot centre points (Figure 1). These subplots constitute the basic sampling unit of the study. Within each subplot, a Leica HDS6100 phase-shift Terrestrial Laser Scanner (TLS) was used to capture woodland canopy structure. The number and position of scan locations within a subplot was determined by the arrangement and density of canopy material in the subplot. Scan positions were arranged to minimise shadows within the canopy, and to maximise canopy penetration. Between one and five scans were recorded per subplot, across all plots. Further information on the field setup of the TLS is presented in Table 1.

Five Leica 6" (15.24 cm) diameter planar tilt-and-turn cross-pattern reflective targets were used in each subplot to align scans (Figure 2). The five targets were located roughly in a quincunx pattern, with one target at the subplot centre and the remaining four targets arranged in a cross pattern around the edges of the subplot, ensuring that all scans could see all five targets. To facilitate alignment of scans among subplots, the location of each target in real space was recorded using a Leica VIVA GS10 GNSS (Global Navigation Satellite Systems) unit (Figure 3). The GNSS was set up in a Post-Processing Kinematic (PPK) configuration with a base-station located ~100 m from the edge of each 1 ha plot with an unobstructed view of the sky hemisphere where possible. The location of each target was measured for at least 4 minutes to minimise measurement error (Figure 4).

30 3 Hemispherical photography field setup

In order to validate TLS canopy closure estimates, at the centre of each subplot a single photograph was taken with a full-frame DSLR camera, equipped with a circular fisheye lens. Further information on the hemispherical photography setup is presented in Table 2.

The fisheye lens had an equisolid (equal area) projection, with a projection function given by:

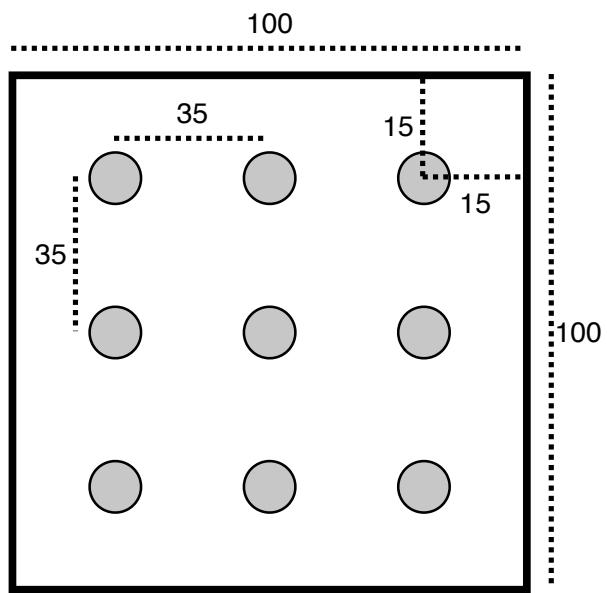


Figure 1: The layout of 10 m diameter subplots within each 1 ha plot. Each subplot is situated inside a 15 m buffer from the plot edge, with 35 m between subplot centres. Subplots are arranged in a 3x3 grid. All distances are in metres.

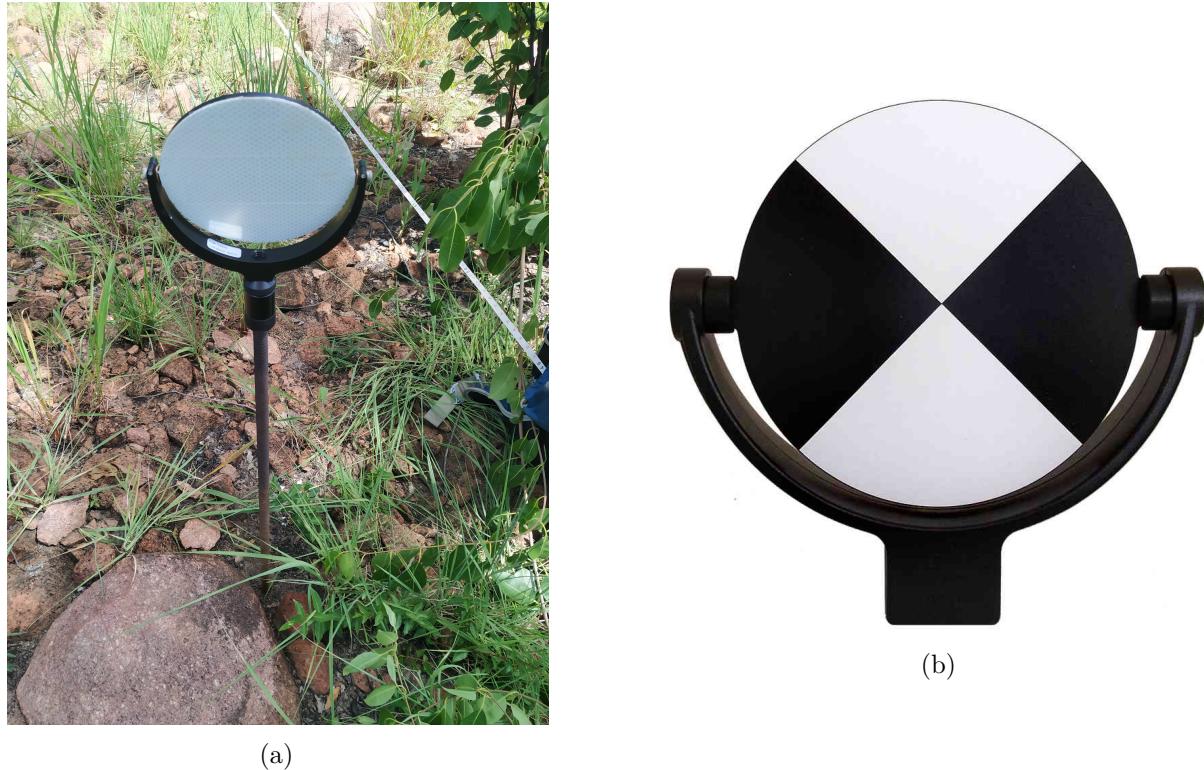


Figure 2: Example of a Leica 6" diameter reflective target, (a) in situ mounted on a length of threaded bar, and (b) showing the cross pattern face of the target.



Figure 3: A Leica VIVA GS10 GNSS unit in the field, showing the antenna atop an aluminium pole, attached to the base station on the ground, and the rover terminal in the hand of a research assistant.

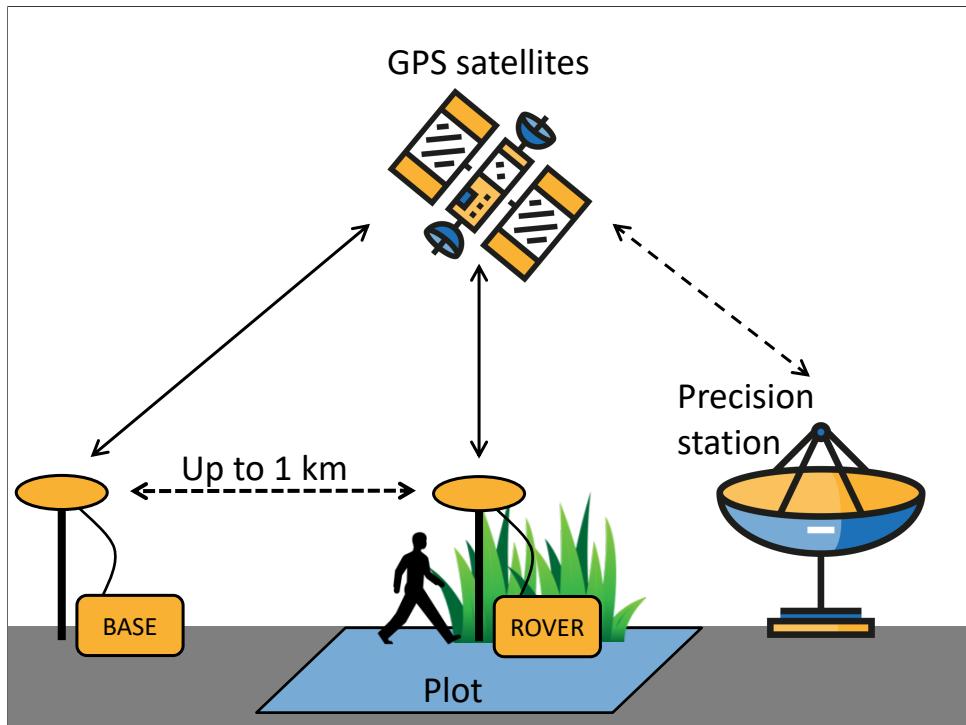


Figure 4: Schematic diagram of the GNSS PPK configuration used to precisely locate targets in real space. The base station is located in an area with a full unobstructed view of the sky hemisphere, up to ~1 km from the plot, and is left in the same location for the duration of the data collection, recording its location once per second. The rover is moved around inside the plot to record the location of each target, for >4 minutes at each target. The rover and the base station both utilise GPS and GLONASS satellites to record their position. After data collection, a two stage validation technique is used to improve the precision of the recorded positions, firstly using the base station, and secondly using the TrimbleRTX service which utilises highly precise distributed regional stations.

Table 1: Description of scan settings used for each scan.

Setting	Value
TLS model	Leica HDS6100
Wavelength	650-690 nm
Spot size at exit	3 mm
Beam divergence	0.22 mrad
Range	79 m @90%; 50 m @18% albedo
Azimuth range	0-360°
Zenith range	0-155°
Increment	0.018°
Point spacing over 25 m	7.9 mm
Pixels per line	20,000
Lines	10,000
Compressed file size	~800 MB
Duration of scan	6 minutes 44 seconds

$$R = 2f \sin\left(\frac{\theta}{2}\right) \quad (1)$$

35 Where R is the radial position of a point on the image, f is the focal length of the lens, and θ is
 36 the angle in radians of incident light on the lens. Equisolid lenses are preferred for hemispherical
 37 photography because they maintain an equal area for each pixel, i.e. a pixel projected through
 38 the lens has the same solid angle irrespective of the incident light angle, meaning that canopy
 39 closure estimations are not biased towards any part of the sky hemisphere (Herbert, 1987).

40 Photographs were taken facing directly to zenith using a camera-mounted spirit level, with
 41 the top of the camera body facing magnetic north, at a height of 1.3 m or above understorey
 42 vegetation, whichever was higher. Photographs were captured under uniform light conditions as
 43 much as possible, either under overcast skies or early in the day before direct sunlight could be
 44 seen on the photograph, to minimise lens flare, which can preclude accurate differentiation of
 45 plant material and sky, and ‘blooming’, a phenomenon where light ‘bleeds’ into dark areas of
 46 the image in highly contrasting light conditions (Frazer et al., 2001).

47 ImageJ (Fiji version 2.1.0/1.53c) was used to binarise hemispherical photographs, to separate
 48 plant material from sky (Schneider et al., 2012). Images were binarised using the Huang algorithm
 49 (Huang & Wang, 1995) using only the blue channel of the image, under the assumption that
 50 plant material reflects little blue light, while the sky reflects much more (Brusa & Bunker,
 51 2014). Images were saved as PNG files at the original pixel resolution, with a circular image of
 52 4016x4016 pixels.

53 4 Terrestrial LiDAR processing

54 4.1 Scan alignment and registration

55 Point clouds within a subplot were aligned using the reflective targets as anchor points. Point
 56 cloud alignment was conducted in Leica Cyclone (version 9.1) (Leica Camera AG, 2009).
 57 Reflective targets were manually located within each point cloud, then the precise centre of
 58 each target was identified automatically by Cyclone. Anchor points were discarded if they had

Table 2: Description of camera settings used for hemispherical photographs. Note that shutter speed and ISO are deliberately variable within sensible thresholds to allow adjustments for ambient light conditions.

Setting	Value
Camera model	Nikon D750
Lens model	Sigma 8 mm f/3.5 EX DG Circular Fisheye
Pixel pitch	5.95 μm
Sensor resolution	24.3 MP
Shutter speed	>1/60s
Aperture	5-7
ISO	100-200
Exposure compensation	-0.7 (Brusa & Bunker, 2014)
Focus	∞ (Hu & Zhu, 2009; Frazer et al., 2001)
Image size	Large Fine JPEG - circular image 4016x4016 px
Orientation	Landscape

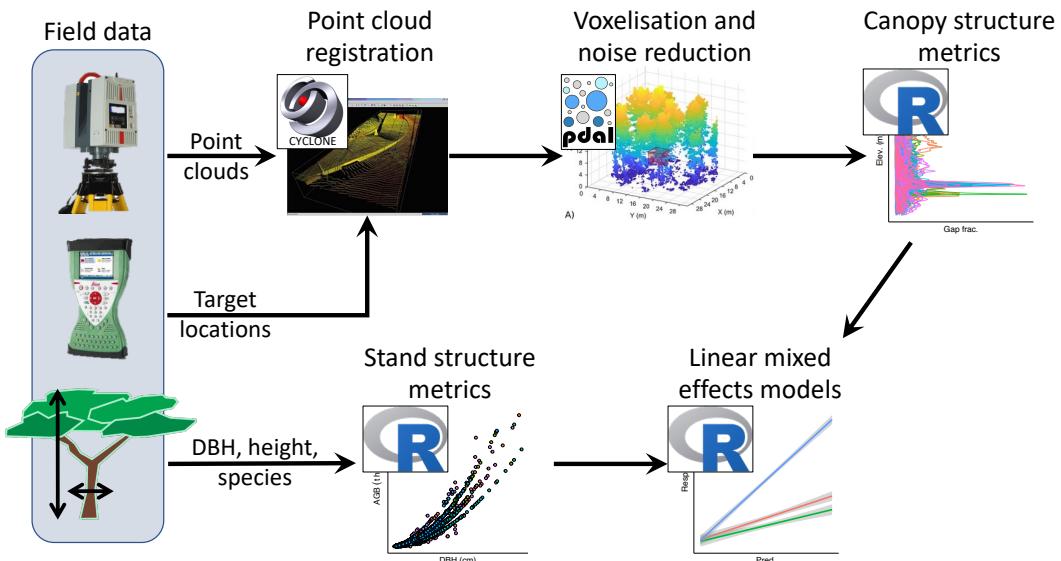


Figure 5: Schematic diagram summarising the data processing and analysis workflow for the TLS data. Processing steps are labelled according to the principal software used during that step.

Table 3: The five closes regional base stations to each site, used by TrimbleRTX to refine GNSS measurements

Site	Receiver name	Distance (km)
AGO	JAVAD TRE 3	863
	JAVAD TRE 3 DELTA	1448
	TRIMBLE NETRS	1630
	TRIMBLE NETRS	1648
	JAVAD TRE G3TH DELTA	1796
	SEPT POLARX5	686
TZA	TRIMBLE ALLOY	795
	LEICA GRX1200GGPRO	903
	JAVAD TRE 3 DELTA	1119
	JAVAD TRE 3 DELTA	1315

59 a location uncertainty of >3 cm. After alignment, subplot point clouds were exported from
 60 Cyclone as PTX files for further processing.

61 GNSS target locations were used to register point clouds in real space. The TrimbleRTX GNSS
 62 post-processing service was used to improve the precision of target locations recorded with GNSS,
 63 using distributed regional stations to validate the rover and base station GNSS measurements
 64 (Table 3) (Chen et al., 2011). Following point cloud registration, subplot point clouds were
 65 combined to a plot-level point cloud.

66 PTX files were converted to compressed LAZ files using PDAL (PDAL Contributors, 2018), to
 67 reduce file size and speed up further processing. Code 1 contains the code used to transform
 68 PTX to LAZ.

```

1 # Get file name without extension
2 noext=${1%.ptx}
3
4 # Find the PTX scan array dimension header material
5 lines=$(grep -E -n '^.{1,10}$' $1 |
6   cut -f1 -d: |
7   awk 'NR%2!=0' |
8   tr '\n' ' ' |
9   sed 's/^([0-9])\s//g')
10
11 # Split PTX file into individual scans
12 csplit -f "$noext" -b "%d.ptx" $1 $lines
13
14 $ Find split files
15 ptxsplit=$(find . -type f -regex ".*/${noext}_[0-9].ptx")
16
17 # For each file:
18 for j in ${ptxsplit}; do
19   jnoext="${j%.ptx}"
20   matrix=$(head -n 10 $j | tail -4 | sed -r 's/0\s+?$/0.0/g' | dos2unix)
21   pdal pipeline ptx_laz.json --readers.text.filename=$j \
22     --filters.transformation.matrix="${matrix}" \
23     --writers.las.filename=${jnoext}.laz
24 done
25
26 # List LAZ files
27 lazsplit=$(find . -type f -regex ".*/${noext}_[0-9].laz")
28
29 # Merge LAZ files
30 pdal merge ${lazsplit} ${noext}.laz

```

Code 1: The processing chain used to convert Leica Cyclone PTX files to LAZ files, using PDAL, POSIX shell scripting, and common UNIX utilities. The `ptx_laz.json` JSON pipeline is shown in Code 2.

```

1 [
2   {
3     "type" : "readers.text",
4     "filename" : "input.txt",
5     "header" : "X Y Z I",
6     "skip" : 10
7   },
8   {
9     "type" : "filters.transformation",
10    "matrix" : "0 -1 0 1 1 0 0 2 0 0 1 3 0 0 0 1"
11  },
12  {
13    "type" : "writers.las",
14    "compression" : "true",
15    "minor_version" : "2",
16    "dataformat_id" : "0",
17    "forward" : "all",
18    "filename" : "output.laz"
19  }
20 ]

```

Code 2: The JSON pipeline used in Code 1 to convert PTX files to LAZ files, and applying a rotation matrix.

71 **4.2 Voxelisation**

72 Point clouds were voxelised to different voxel sizes depending on the application of the data. 5
73 cm^3 cubic voxels were used for subplot height profile estimation, while 50 cm^3 voxels were used
74 for whole plot canopy rugosity. Variation in voxel size reflects the spatial scale of each analysis,
75 and is bounded by the beam divergence of the scanner over longer distances (Grau et al., 2017).
76 Choosing voxels that are too small can result in pock-marked representations of surfaces that
77 are especially problematic when calculating larger scale canopy structure metrics such as canopy
78 top roughness, while voxels that are too large can result in an over-estimation of plant volume
79 when estimating canopy foliage density at the subplot scale, especially when foliage is clumped
80 (Seidel et al., 2012; Cifuentes et al., 2014). Voxels were classified as ‘filled’ if they intersected
81 one or more points.

82 **4.3 Noise reduction**

83 Outlier detection and noise reduction of point clouds was conducted in PDAL, using the
84 “statistical method” (sensu Rusu et al. 2008) of `filters.outlier`, with $k = 8$ (mean number
85 of neighbours), and $m = 1.96$ (outlier distance threshold multiplier, here approximating a 95%
86 confidence interval):

$$\begin{aligned}\bar{\mu} &= \frac{1}{N} \sum_{i=1}^N \mu_i \\ \sigma &= \sqrt{\frac{1}{N-1} \sum_{i=1}^N (\mu_i - \bar{\mu})^2} \\ t &= \mu + m\sigma \\ \text{with } outlier_i &= \begin{cases} \text{true,} & \text{if } \mu_i \geq t \\ \text{false,} & \text{otherwise} \end{cases}\end{aligned}\tag{2}$$

87 Where μ_i is the mean distance from point i to all k nearest neighbour points, N is the number
88 of points in the scene, $\bar{\mu}$ is the mean distance to nearest neighbour points, σ is the standard
89 deviation of these mean distances, t is the threshold distance used to define an outlier and
90 $outlier_i$ is the condition of a point in the scene being identified as an outlier.

91 **4.4 Foliage density profiles**

92 To calculate subplot foliage density profiles, the 5 cm^3 voxelised point cloud was first cropped to
93 a 10 m diameter cylinder of infinite height. Ground points were identified using `filters.pmf`
94 (Progressive Morphological Filter - PMF) in PDAL (sensu Zhang et al. 2003), and the height
95 above ground of all points was calculated using `filters.hag_nn` (Nearest Neighbour) in PDAL.
96 Points below ground level and above the 99.9th percentile of height were excluded from further
97 analyses. Height profile points were exported to XYZ coordinates then imported into R for
98 further processing.

99 In R, foliage density was calculated in 5 cm layers as the proportion of filled 5 cm^3 voxels. A
100 loess model with a span of 0.1 was fitted to the foliage density values in each layer to estimate
101 the foliage density profile (Figure 7). The foliage density profile was further filtered to only tree
102 canopy material, by discarding all points below the first local minimum in the foliage density
103 profile above 1.3 m, using a rolling window of 50 cm.

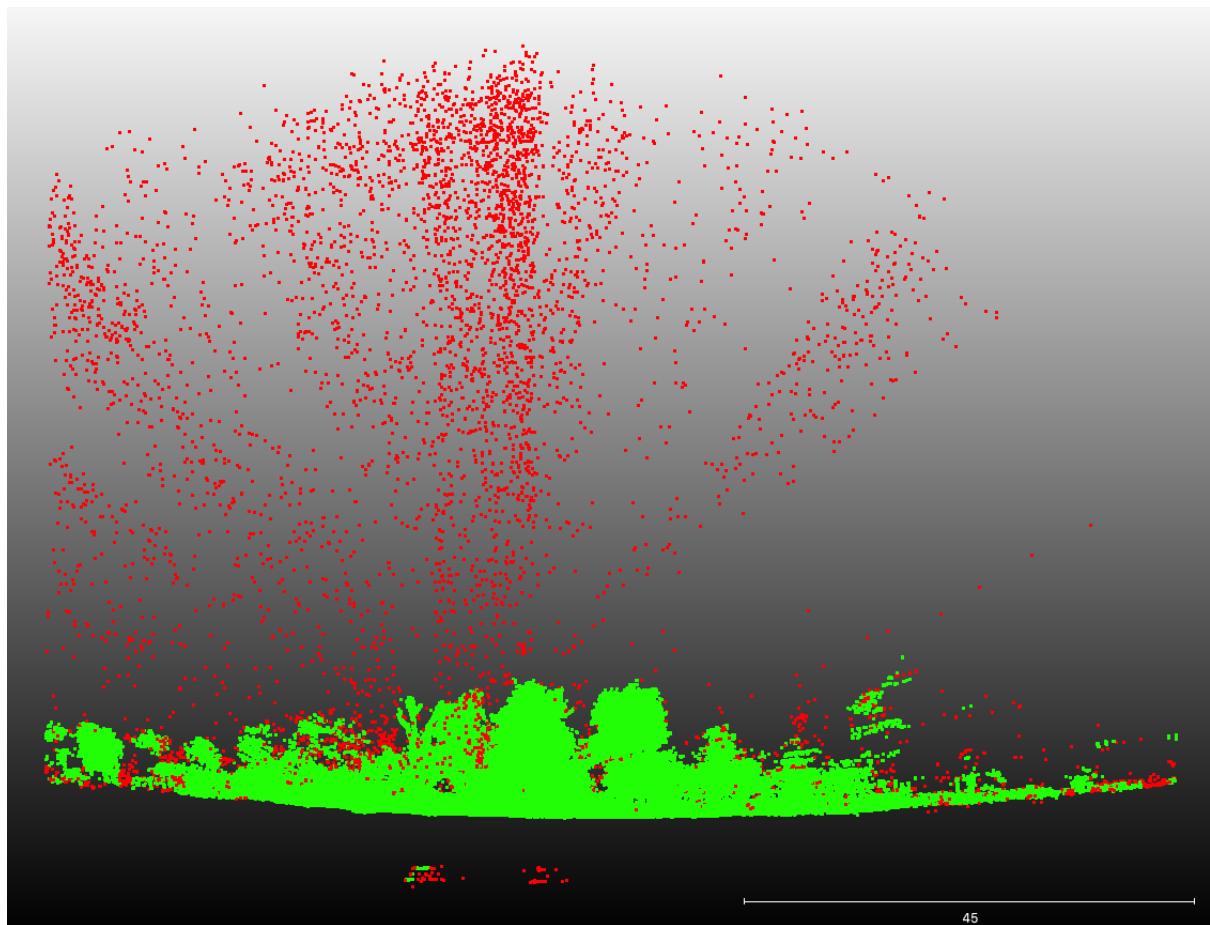


Figure 6: A 2 m deep cross section of a subplot point cloud showing the efficacy of the noise reduction and voxelisation process. Red points are excluded by the process, while green points are preserved for further analysis.

104 Multiple statistics were extracted from the foliage density profile for use in statistical analyses.
105 Total canopy foliage density was calculated as the area under the curve of the canopy foliage
106 density profile, using trapezoid estimation. The Effective Number of Layers (ENL) in the
107 foliage density profile was used to estimate canopy structural complexity, using the true-numbers
108 equivalent of the Shannon diversity index on the foliage density of 50 cm layers (sensu Ehbrecht
109 et al. 2016):

$$\text{ENL} = \exp\left(-\sum_{i=1}^N p_i \ln p_i\right) \quad (3)$$

110 Where N is the number of 50 cm bins in the height profile, and p_i is the proportion of filled
111 voxels in layer i (foliage density). While Ehbrecht et al. (2016) used 1 m layers, their study was
112 conducted in temperate deciduous forest where the maximum height of the sampled forest stands
113 was 40 m, whereas the maximum canopy height in this study was only 22 m. Both Ehbrecht
114 et al. (2016) and Montes et al. (2004) assert that layer thickness is largely arbitrary, but should
115 be determined with respect to the variability within the canopy, thus in the sparse and highly
116 variable savanna tree canopies measured in this study, narrower layers were chosen.

117 To describe the uniformity of the foliage density distribution through the canopy, a linear model
118 of foliage density with height was fitted. Under a completely even distribution of foliage material
119 through the canopy, the residuals of the linear model tend to zero, while clumping causes
120 deviations from this uniform distribution and increases the sum of residuals.

121 While maximum canopy height has been used in other studies to describe canopy structural
122 complexity (Scheuermann et al., 2018). At the small spatial scale of the subplots used in this
123 study however, there proved to be too much stochastic variation in canopy height among subplots
124 due to the distribution of individual trees to make this statistic informative as a measure of
125 canopy complexity. Canopy height was calculated later at the plot level, however. Similarly,
126 previous studies have used the number of local maxima in the foliage density profile to estimate
127 canopy structural complexity (Wilkes et al., 2016). However, this metric covaries with ENL and
128 total foliage density, while ENL is uncorrelated with foliage density (Ehbrecht et al., 2016).

129 4.5 Canopy closure

130 Subplot canopy closure, i.e. the proportion of the sky hemisphere occluded by plant material,
131 a.k.a. gap fraction or site factor (Jennings, 1999), was measured by simulating a hemispherical
132 image at the centre of the subplot using the point cloud data from all scans per subplot. The
133 point cloud was first cropped to a 20 m diameter cylinder around the subplot centre using
134 PDAL. Points below 1.3 m and within a 50 cm sphere around the subplot centre at 1.3 m height
135 were discarded, to prevent the simulated hemispherical image being occluded by understorey
136 vegetation. POV-Ray was used to simulate the hemispherical image using ray-tracing (Persistence
137 of Vision Pty. Ltd., 2004). Filled voxels were represented in POV-Ray as non-reflective black
138 cubes filling the 5 cm^3 voxel volume, with a white uniform sky box and no light source. POV-Ray
139 produced an image with identical qualities to that of the real hemispherical photograph, using a
140 fisheye lens with an equisolid projection and a view angle of 180° , located at the subplot centre
141 at 1.3 m above the ground, with the top of the camera facing magnetic north and the camera
142 facing directly to zenith, producing a circular image of 4016x4016 pixels.

143 Hemiphot (ter Steege, 2018) was used to estimate closure from both the hemispherical photo-
144 graphs and the TLS POV-Ray simulation. Hemiphot calculates canopy closure in 90 evenly
145 sized concentric rings. To obtain the total closure of a circular image:

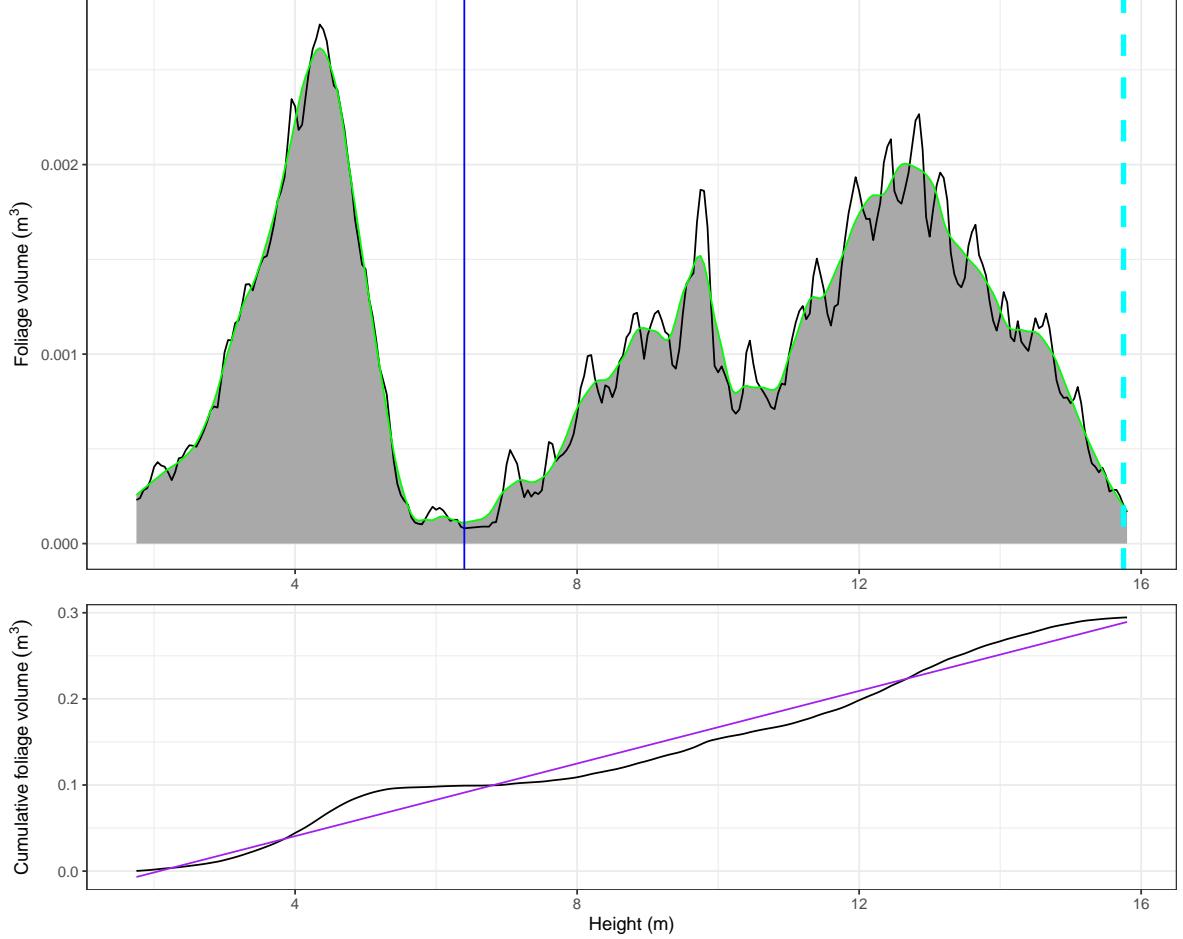


Figure 7: Subplot foliage volume height profile (top) and cumulative foliage volume profile (bottom) for a subplot in Bicuar National Park, Angola, to illustrate some of the canopy structure metrics extracted from each height profile. Starting with the top panel: the blue solid line represents the first local minimum above 1.3 m, used to define the base of the tree canopy. The dashed cyan line shows the 99.9th percentile of canopy height, used here as a measure of canopy top height across the subplot and in plot-level canopy surface modelling. The black trace shows the foliage density height profile, and the green trace shows the loess model fitted to the data, with the area under the canopy shaded grey. The bottom panel: the black trace shows the cumulative foliage volume through the canopy, taken from the loess fit in the top panel. The purple line shows the line of best fit of a linear model through this data. Not illustrated is the Effective Number of Layers (ENL) metric.

$$C_\alpha = 1 - G_{\text{tot}} = \sum_{\alpha=0.5}^{\alpha=89.5} (G_\alpha A_\alpha / A_{\text{tot}}) \quad (4)$$

146 Where G_α is the fraction of unfilled pixels in ring α , A_α is the sky area of the ring segment, and
147 A_{tot} is the total sky area of the hemisphere.

148 Canopy closure estimates from the TLS were validated using estimates from hemispherical
149 photography. A Pearson's correlation analysis showed that both methods were highly correlated
150 ($r(195)=0.89$, $p<0.001$). TLS estimates of closure were almost exclusively higher than hemispherical
151 photography estimates, except in a few subplots with particularly low canopy closure. At
152 higher canopy closure the over-estimation of canopy closure by TLS was larger (Figure 8). This
153 finding is in agreement with previous studies which have found that the magnitude of TLS
154 canopy closure over-estimation depends on gap size distribution, where a site with greater canopy
155 cover and a gap fraction dominated by small within crown gaps will have a larger over-estimate
156 than a more open site with a gap fraction dominated by large between crown gaps (Seidel et al.,
157 2012). A linear mixed model which accounted for the nested sampling of subplots within plots
158 was used to identify if sites differed significantly in their relationship between hemispherical
159 photography and TLS estimates of canopy closure. There was no significant difference in model
160 fixed effect slope between plots in Bicuar National Park, Angola, and those in Mtarure, Tanzania
161 ($\beta(173)=0.13\pm0.0098$, $p=0.18$).

162 4.6 Whole plot canopy metrics

163 The canopy height of each 1 ha plot was estimated using unified point clouds from all subplots.
164 The unified point cloud was voxelised to 10 cm^3 , and the 99th percentile of height from each 10 cm^2
165 column was taken as the canopy height. Maximum height was not used as this occasionally
166 constituted a severe outlier which skewed further canopy surface model smoothing. The point
167 cloud was then cropped to the plot boundaries, located using PPK GNSS similar to the TLS
168 targets. A pit-filling algorithm described in Khosravipour et al. (2014) was used to smooth
169 the canopy surface model, at a resolution of 50 cm, by removing gaps within trees caused by
170 incomplete penetration of the LiDAR beam (Figure 10).

171 Mean canopy height across the plot and the coefficient of variation of canopy height were
172 extracted from the canopy surface model for use in statistical analyses. The coefficient of
173 variation of canopy height describes canopy structural diversity measured by the heterogeneity
174 of the canopy surface. Other studies in closed canopy temperate and boreal forests have used
175 metrics similar to the Topographic Roughness Index to measure canopy surface heterogeneity, by
176 comparing canopy height to that of neighbouring pixels in the canopy height model (Weligepolage
177 et al., 2012; Herrero-Huerta et al., 2020). In this study however, the sparse nature of the tree
178 canopies meant that these metrics were overly influenced by canopy density and the edges
179 of individual tree canopies. Canopy rugosity (R_c) was also calculated to describe structural
180 complexity across the entire canopy profile, rather than just the canopy surface, sensu Hardiman
181 et al. (2011). R_c first calculates the standard deviation of foliage density in 50 cm^2 columns
182 across the plot (σG_z), then calculates the standard deviation of those standard deviations:

$$R_c = \sigma(\sigma G_z)_x \quad (5)$$

183 Where G_z is the vertical height axis z , x is the horizontal axis, and σ is the standard deviation.
184 Finally, plot-level canopy closure was calculated as the mean of subplot TLS canopy closure
185 estimates.

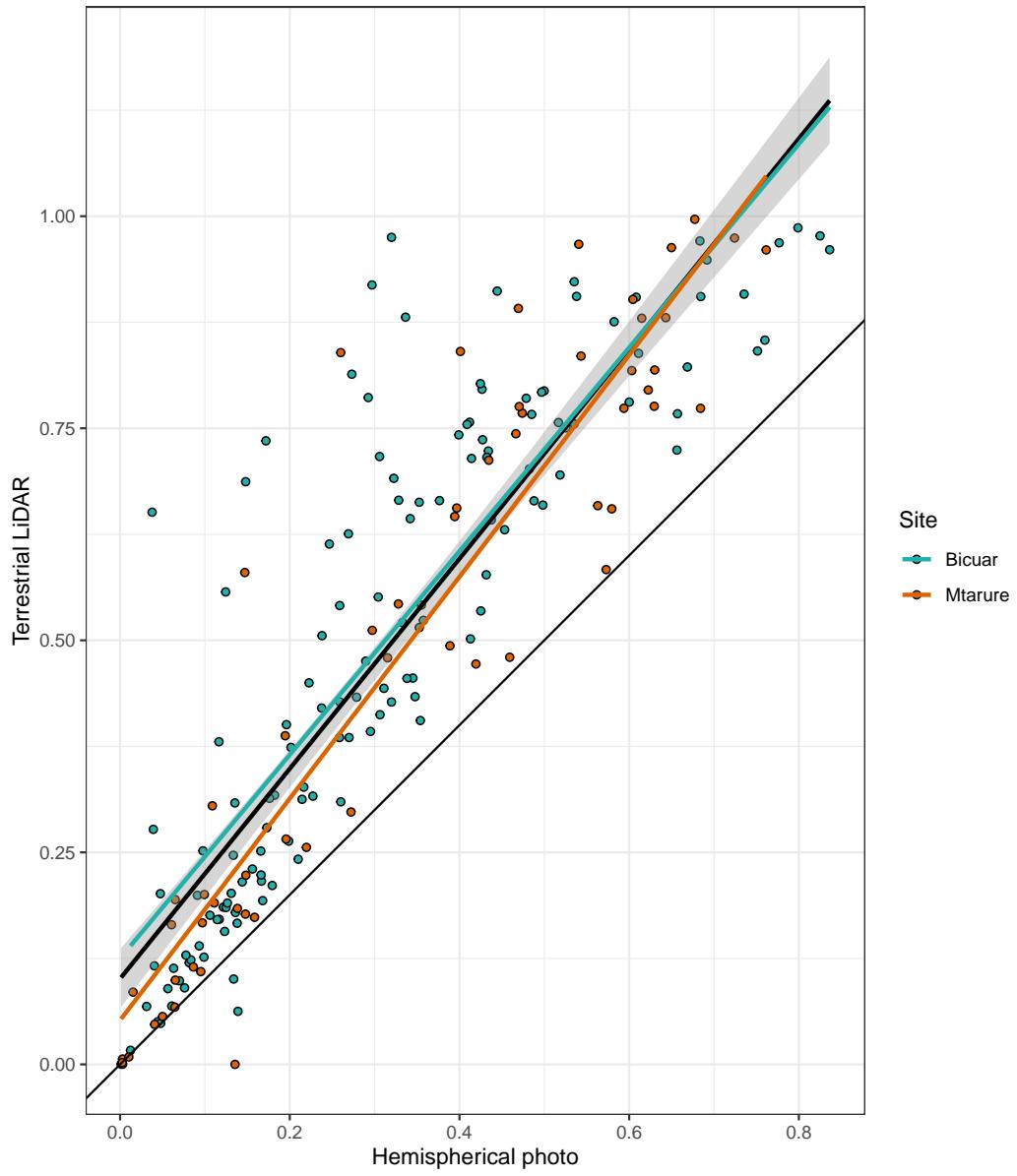


Figure 8: Comparison of canopy closure estimation from TLS and hemispherical photography. The thick black line of best fit is a linear model of all points ± 1 standard error, while the coloured lines are site specific linear models. The thin black line shows the 1:1 fit.

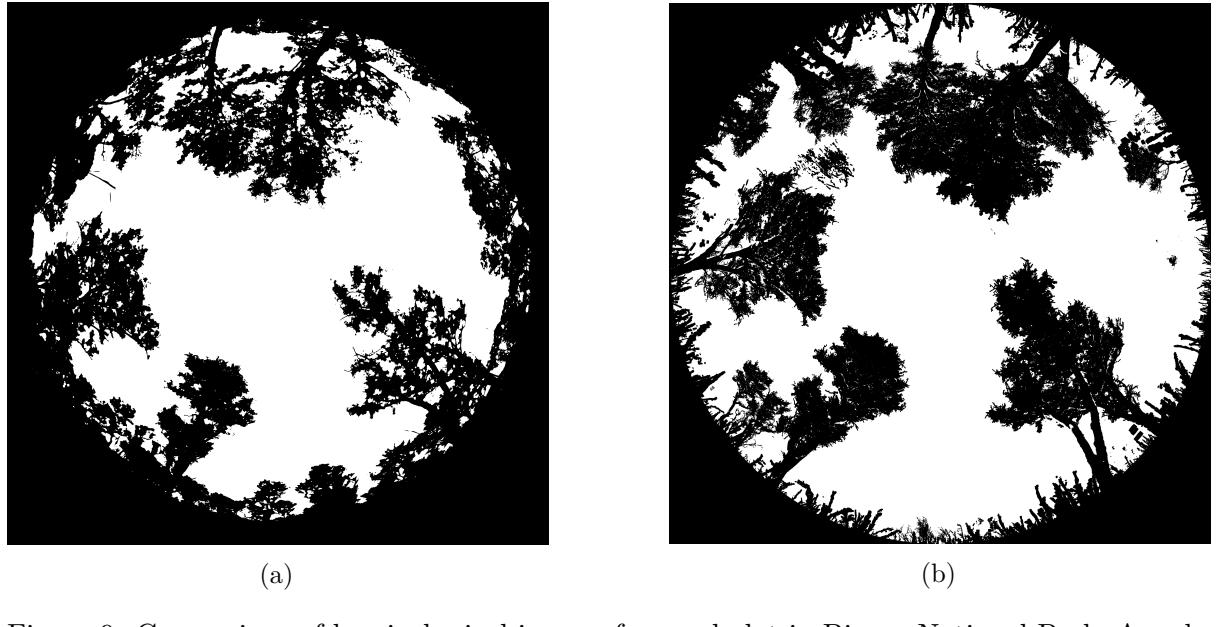


Figure 9: Comparison of hemispherical images for a subplot in Bicuar National Park, Angola. (a) A hemispherical photograph, and (b) a multi-scan point cloud modelled as cubic voxels with POV-Ray. The hemispherical photograph (left) shows some blooming, especially in the tree on the bottom right of the image, where light is seen ‘bleeding’ through the darker canopy material, causing an under-estimation in canopy closure.

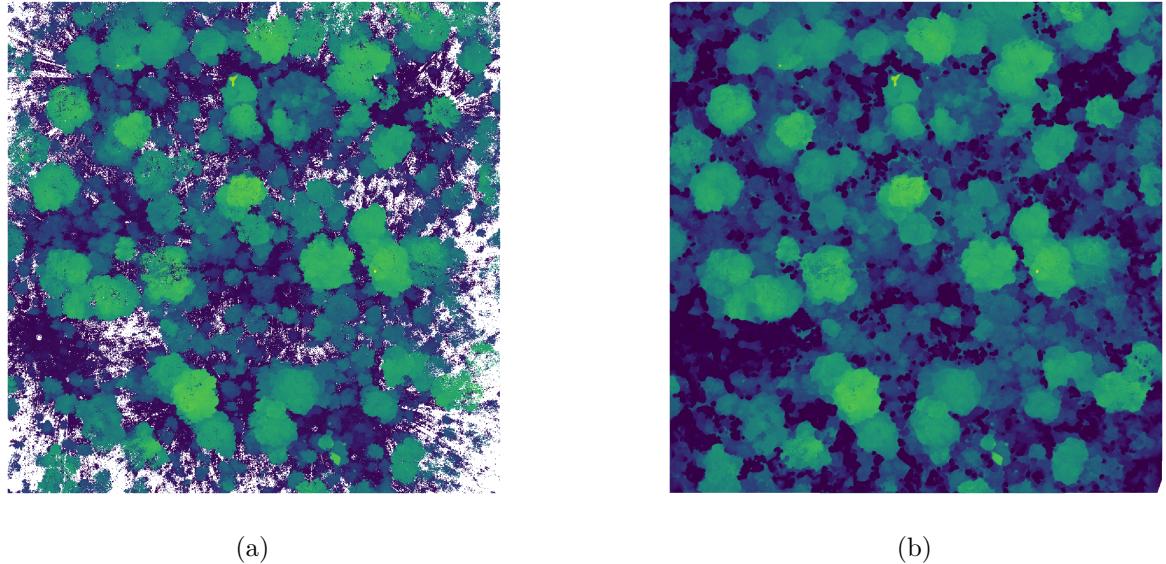


Figure 10: Top-down view of a 1 ha plot in Bicuar National Park. (a) The point cloud after voxelisation, noise reduction, and taking the 99th percentile of stem height in each 5 cm vertical bin. (b) The same point cloud after pit filling to generate a smooth canopy height profile. Points are coloured according to point height from the ground.

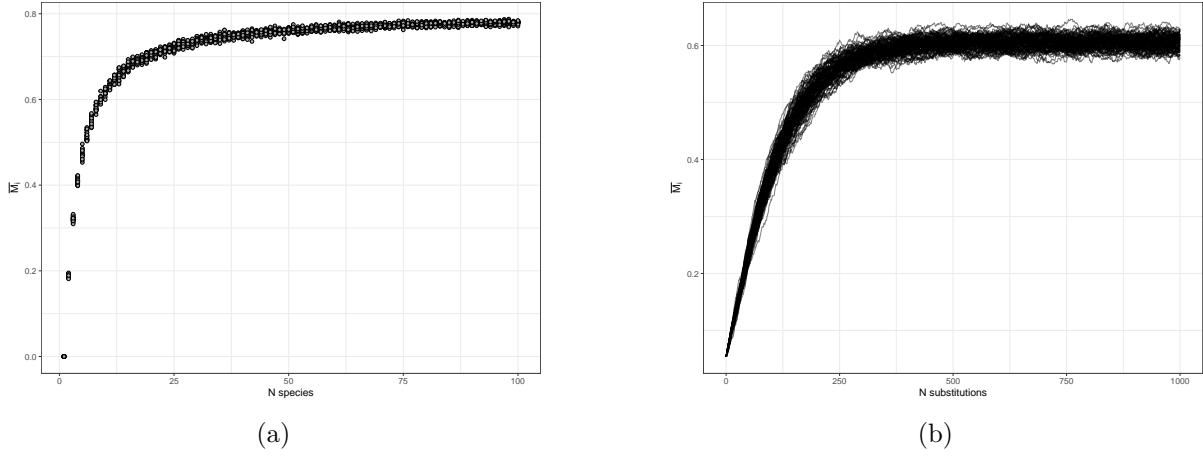


Figure 11: The behaviour of M_i with increasing number of species (a), and increasing spatial mixing of species (b). The left panel was generated by randomly assigning different numbers of species, in equal proportions, to an evenly spaced grid of individuals. 20 replicates were conducted for each number of species. The right panel was generated by randomly swapping pairs of individuals in a plot with 9 species arranged in mono-specific square blocks in an evenly spaced grid. Each line shows a single replicate, where individuals were swapped in an additive fashion, with 100 total.

186 5 Stand structure metrics

187 5.1 Spatial mingling of species

188 The spatial mingling index (M_i) is a spatially explicit estimate of the degree to which species
 189 are spatially mixed within a plot. Here, M was calculated at the plot level as the mean of M_i
 190 according to von Gadow & Hui (2002), with the adjustment for potential neighbourhood species
 191 pool suggested by Hui et al. (2011):

$$M = \bar{M}_i$$

$$M_i = \frac{S_i}{n_{\max}} \frac{1}{k} \sum_{j=1}^k v_j \quad (6)$$

with $v_j = \begin{cases} 0, & \text{neighbour } j \text{ same species as reference } i \\ 1, & \text{otherwise} \end{cases}$

192 Where k is the number of nearest neighbours considered for each reference tree, S_i is the number
 193 of species found among the k nearest neighbours of tree i , n_{\max} is the potential number of
 194 species in the neighbourhood, i.e. $k + 1$, and N is the total number of trees in the plot. The
 195 conventional value of $k = 4$ was used here (von Gadow & Hui, 2002; Hui & Albert, 2004; Hui
 196 et al., 2007). The value of M_i increases with greater mixing of species, and all else being equal
 197 will increase with number of species within the plot (Figure 11).

198 5.2 Spatial clustering of stems

199 The winkelmaass (W) was calculated to estimate the degree of spatial uniformity in stem spatial
 200 distribution. Here, W was calculated at the plot level as the mean of W_i) according to von
 201 Gadow & Hui (2002):

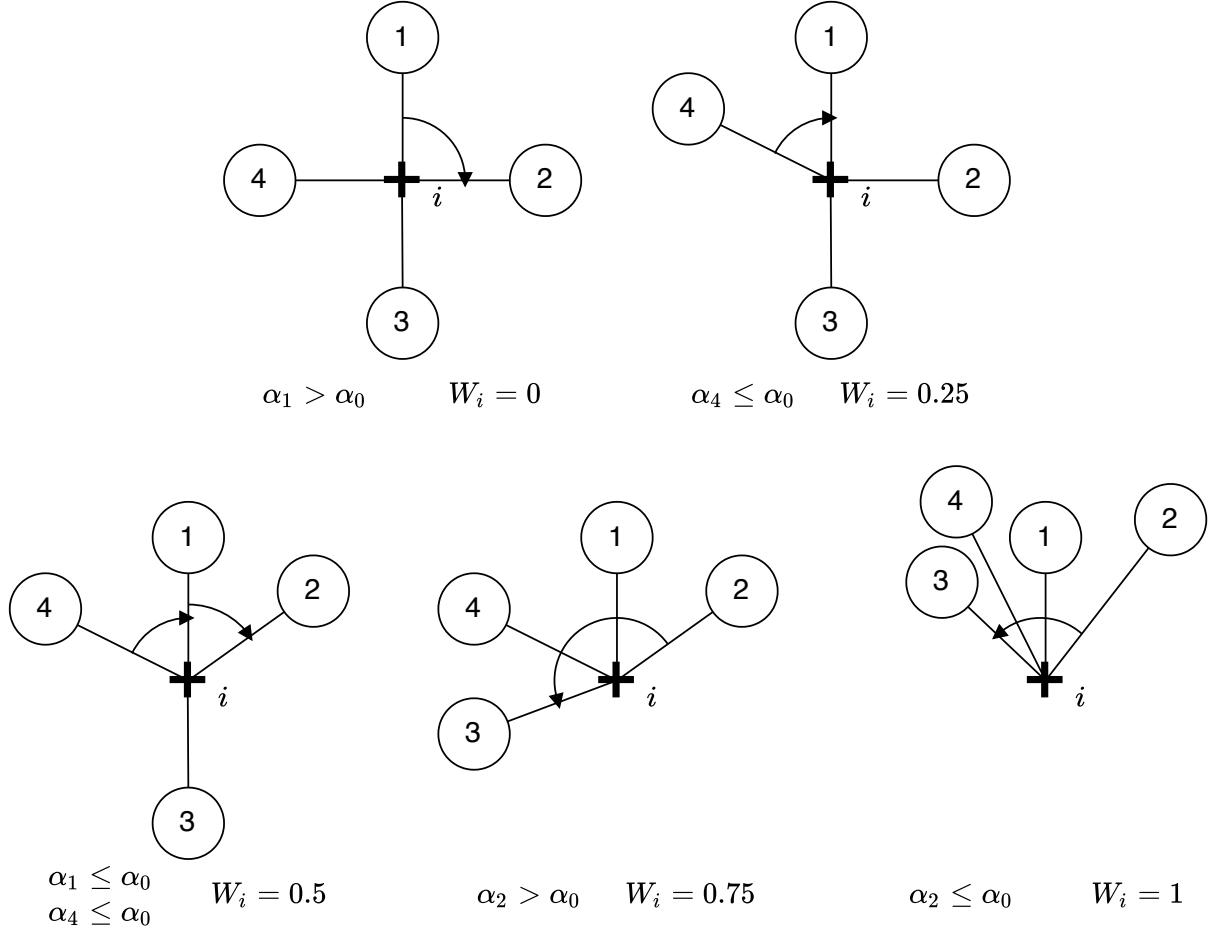


Figure 12: Possible values of W_i at a sample point i , denoted by a cross. Neighbours are represented as circles numbered sequentially from 1 to 4, where $k = 4$. The angles of arrows in each example are given below, along with the winkelmass for that example.

$$\begin{aligned}
 W &= \overline{W}_i \\
 W_i &= \frac{1}{k} \sum_{j=1}^k v_j \\
 \text{with } v_j &= \begin{cases} 0, & \alpha_j \leq \alpha_0 \\ 1, & \text{otherwise} \end{cases}
 \end{aligned} \tag{7}$$

202 Where k is the number of neighbours considered, here using the conventional value of $k = 4$,
 203 α_j is the angle between consecutive neighbours and α_0 is the critical angle, where $\alpha_0 = 360/k$.
 204 Figure 12 demonstrates how the value of W_i varies according to spatial distribution of neighbours.
 205 The value of the winkelmass increases with increasing spatial clumping (decreasing spatial
 206 regularity) of individuals (Figure 13), and in a plot with random tree distribution will increase
 207 as more neighbours are considered (Figure 14).

208 5.3 Subplot canopy crowding

209 An adapted version of the Iterative Hegyi Index (H_i) was used to estimate tree spatial structure
 210 in subplots (Hegyi, 1974). The adapted formula used here allows the index to be based on a point

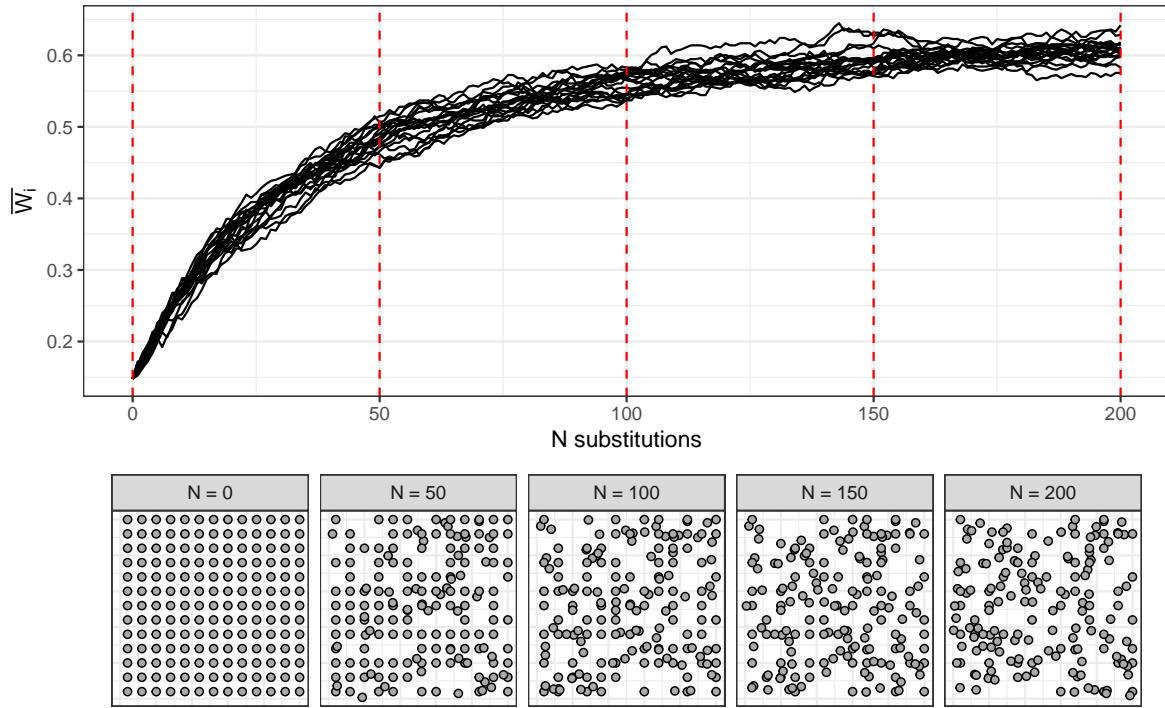


Figure 13: Variation in winkelmass with increasing spatial irregularity of individuals. The top panel shows variation of winkelmass in 20 plots as individuals are sequentially moved to a random location within the plot. Red dotted lines correspond to the panels below which show the spatial distribution of individuals after a given number of random individual movements.

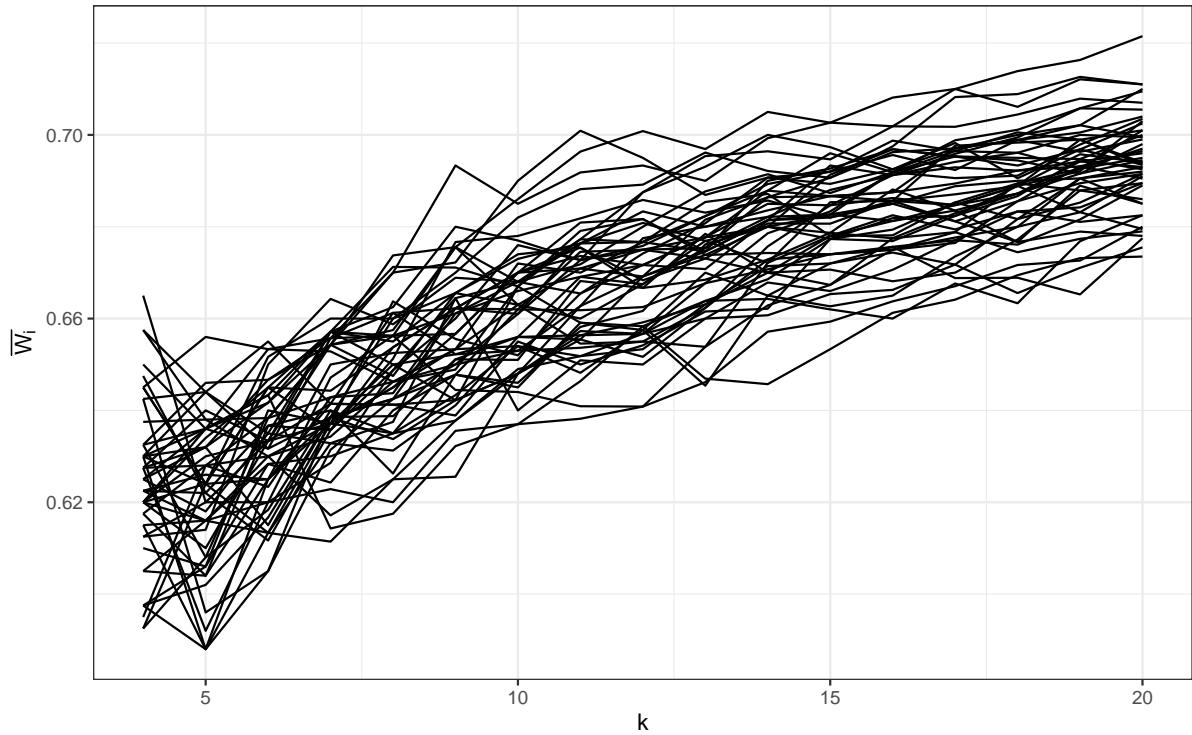


Figure 14: Variation in winkelmass with increasing number of neighbours k considered in the calculation. 50 replicate plots were used, each with 100 individuals randomly distributed in space.

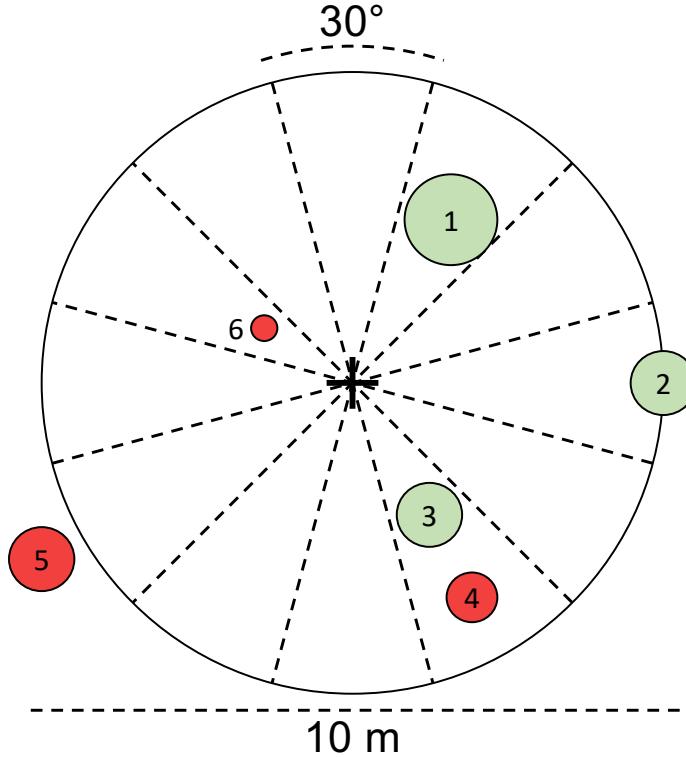


Figure 15: Schematic diagram demonstrating use of the Iterative Hegyi Index to assess crowding within each subplot. The 10 metre diameter subplot is divided into 12 equally sized sectors. Within each sector, the nearest stem of sufficient size (>5 cm diameter) to the subplot centre is recorded (e.g. 1). All stems with any canopy material inside the subplot are valid (e.g. 2). Stem 4 is not valid as it is behind stem 3. Stem 5 is invalid as all its canopy is outside the subplot. Stem 6 is too small to be recorded.

211 rather than a focal tree, transforming it from a tree-centric competition index to a point-centric
212 crowding index:

$$H_i = \log \sum_{j=1}^n \left(\frac{1}{L_{ij}} D_j \right) \quad (8)$$

213 Where n is the number of stems with canopy material within the subplot, D_j is the stem diameter
214 of stem j and L_{ij} is the distance of stem j from the subplot centre i . H_i uses an iterative method
215 for choosing active canopy occupants at the subplot centre, where the nearest individual to the
216 subplot centre from each of 12 equally sized sectors is classified as the active stem (Figure 15).
217 H_i was preferred over stem density to describe stem crowding in subplots because it is sensitive
218 to how close a stem is to the subplot centre, which will affect canopy closure even if the number
219 of stems in the subplot remains the same.

220 References

- 221 Brusa, A. & D. E. Bunker (2014). ‘Increasing the precision of canopy closure estimates from
222 hemispherical photography: Blue channel analysis and under-exposure’. In: *Agricultural and*
223 *Forest Meteorology* 195–196, pp. 102–107. DOI: 10.1016/j.agrformet.2014.05.001.

- 224 Chen, X., T. Allison, W. Cao, K. Ferguson, S. Grunig, V. Gomez, A. Kipka, J. Kohler, H.
225 Landau, R. Leandro et al. (Sept. 2011). *Trimble RTX, an innovative new approach for network*
226 *RTK*. Tech. rep. Portland OR, USA: International Technical Meeting of the Satellite Division
227 of the Institute of Navigation, ION GNSS, pp. 2214–2219.
- 228 Cifuentes, R., D. V. der Zande, J. Farifteh, C. Salas & P. Coppin (2014). ‘Effects of voxel size
229 and sampling setup on the estimation of forest canopy gap fraction from terrestrial laser
230 scanning data’. In: *Agricultural and Forest Meteorology* 194, pp. 230–240. DOI: 10.1016/j.
231 agrformet.2014.04.013.
- 232 Ehbrecht, M., P. Schall, J. Juchheim, C. Ammer & D. Seidel (2016). ‘Effective number of layers:
233 A new measure for quantifying three-dimensional stand structure based on sampling with
234 terrestrial LiDAR’. In: *Forest Ecology and Management* 380, pp. 212–223. DOI: 10.1016/j.
235 foreco.2016.09.003.
- 236 Frazer, G. W., R. A. Fournier, J. Trofymow & R. J. Hall (2001). ‘A comparison of digital and
237 film fisheye photography for analysis of forest canopy structure and gap light transmission’. In:
238 *Agricultural and Forest Meteorology* 109.4, pp. 249–263. DOI: 10.1016/s0168-1923(01)00274-
239 x.
- 240 Grau, E., S. Durrieu, R. Fournier, J.-P. Gastellu-Etchegorry & T. Yin (2017). ‘Estimation
241 of 3D vegetation density with Terrestrial Laser Scanning data using voxels. A sensitivity
242 analysis of influencing parameters’. In: *Remote Sensing of Environment* 191, pp. 373–388. DOI:
243 10.1016/j.rse.2017.01.032.
- 244 Hardiman, B. S., G. Bohrer, C. M. Gough, C. S. Vogel & P. S. Curtis (2011). ‘The role of canopy
245 structural complexity in wood net primary production of a maturing northern deciduous
246 forest’. In: *Ecology* 92.9, pp. 1818–1827. DOI: 10.1890/10-2192.1.
- 247 Hegyi, F. (1974). ‘A simulation model for managing jack-pine stands’. In: *Royal College of
248 Forestry, editor*. Stockholm, Sweden: Royal College of Forestry, pp. 74–90.
- 249 Herbert, T. J. (1987). ‘Area projections of fisheye photographic lenses’. In: *Agricultural and
250 Forest Meteorology* 39.2-3, pp. 215–223. DOI: 10.1016/0168-1923(87)90039-6.
- 251 Herrero-Huerta, M., A. Bucksch, E. Puttonen & K. M. Rainey (Dec. 2020). ‘Canopy Roughness:
252 A New Phenotypic Trait to Estimate Aboveground Biomass from Unmanned Aerial System’.
253 In: *Plant Phenomics* 2020, pp. 1–10. DOI: 10.34133/2020/6735967. URL: <https://doi.org/10.34133%2F2020%2F6735967>.
- 255 Hu, L. & J. Zhu (2009). ‘Determination of the tridimensional shape of canopy gaps using two
256 hemispherical photographs’. In: *Agricultural and Forest Meteorology* 149.5, pp. 862–872. DOI:
257 10.1016/j.agrformet.2008.11.008.
- 258 Huang, L. & M. J. Wang (1995). ‘Image thresholding by minimizing the measures of fuzziness’.
259 In: *Pattern Recognition* 28.1, pp. 41–51. DOI: 0031-3203/95.
- 260 Hui, G. & M. Albert (2004). ‘Stichprobensimulationen zur Schätzung nachbarschaftsbezogener
261 Strukturparameter in Waldbeständen [Simulation studies for estimating neighborhood-based
262 structural parameters in forest stands]’. In: *Allgemeine Forst und Jagdzeitung* 175, pp. 10–11.
- 263 Hui, G., K. von Gadow, Y. Hu & H. Xu (2007). *Structure-based forest management*. Beijing,
264 China: China Forestry Publishing House.
- 265 Hui, G., X. Zhao, Z. Zhao & K. von Gadow (2011). ‘Evaluating tree species spatial diversity
266 based on neighbourhood relationships’. In: *Forest Science* 57.4, pp. 292–300. DOI: 10.1093/
267 forestscience/57.4.292.
- 268 Jennings, S. (1999). ‘Assessing forest canopies and understorey illumination: canopy closure,
269 canopy cover and other measures’. In: *Forestry* 72.1, pp. 59–74. DOI: 10.1093/forestry/72.
270 1.59.
- 271 Khosravipour, A., A. K. Skidmore, M. Isenburg, T. Wang & Y. A. Hussin (2014). ‘Generating
272 Pit-free Canopy Height Models from Airborne LiDAR’. In: *Photogrammetric Engineering &*
273 *Remote Sensing* 80.9, pp. 863–872. DOI: 10.14358/pers.80.9.863.
- 274 Leica Camera AG (2009). *Leica Cyclone*. Version 9.1.

- 275 Montes, F., I. Cañellas, M. del Río, R. Calama & G. Montero (2004). ‘The effects of thinning
276 on the structural diversity of coppice forests’. In: *Annals of Forest Science* 61.8, pp. 771–779.
277 DOI: 10.1051/forest:2004074.
- 278 PDAL Contributors (2018). *PDAL Point Data Abstraction Library*. DOI: 10.5281/zenodo .
279 2556738.
- 280 Persistence of Vision Pty. Ltd. (2004). *Persistence of Vision Raytracer (Version 3.7)*. [Computer
281 software].
- 282 Rusu, R. B., Z. C. Marton, N. Blodow, M. Dolha & M. Beetz (2008). ‘Towards 3D Point cloud
283 based object maps for household environments’. In: *Robotics and Autonomous Systems* 56.11,
284 pp. 927–941. DOI: 10.1016/j.robot.2008.08.005.
- 285 Scheuermann, C. M., L. E. Nave, R. T. Fahey, K. J. Nadelhoffer & C. M. Gough (Aug. 2018).
286 ‘Effects of canopy structure and species diversity on primary production in upper Great
287 Lakes forests’. In: *Oecologia* 188.2, pp. 405–415. DOI: 10.1007/s00442-018-4236-x. URL:
288 <https://doi.org/10.1007%2Fs00442-018-4236-x>.
- 289 Schneider, C. A., W. S. Rasband & K. W. Eliceiri (2012). ‘NIH Image to ImageJ: 25 years of
290 image analysis’. In: *Nature Methods* 9.7, pp. 671–675. DOI: 10.1038/nmeth.2089.
- 291 Seidel, D., S. Fleck & C. Leuschner (2012). ‘Analyzing forest canopies with ground-based
292 laser scanning: A comparison with hemispherical photography’. In: *Agricultural and Forest
293 Meteorology* 154-155, pp. 1–8. DOI: 10.1016/j.agrformet.2011.10.006.
- 294 ter Steege, H. (2018). *Hemiphot.R: Free R scripts to analyse hemispherical photographs for
295 canopy openness, leaf area index and photosynthetic active radiation under forest canopies*.
296 Unpublished report. Leiden, The Netherlands: Naturalis Biodiversity Center. URL: <https://github.com/Naturalis/Hemiphot>.
- 297 von Gadow, K. & G. Hui (2002). ‘Characterising forest spatial structure and diversity’. In:
298 *Proceedings of the IUFRO International workshop ‘Sustainable forestry in temperate regions’*.
300 Ed. by L. Bjoerk. Lund, Sweden, pp. 20–30.
- 301 Weligepolage, K., A. Gieske & Z. Su (Feb. 2012). ‘Surface roughness analysis of a conifer forest
302 canopy with airborne and terrestrial laser scanning techniques’. In: *International Journal of
303 Applied Earth Observation and Geoinformation* 14.1, pp. 192–203. DOI: 10.1016/j.jag.2011.
304 08.014. URL: <https://doi.org/10.1016%2Fj.jag.2011.08.014>.
- 305 Wilkes, P., S. D. Jones, L. Suarez, A. Haywood, A. Mellor, W. Woodgate, M. Soto-Beregov
306 & A. K. Skidmore (Dec. 2016). ‘Using discrete-return airborne laser scanning to quantify
307 number of canopy strata across diverse forest types’. In: *Methods in Ecology and Evolution*
308 7.6. Ed. by S. McMahon, pp. 700–712. DOI: 10.1111/2041-210x.12510. URL: <https://doi.org/10.1111%2F2041-210x.12510>.
- 309 Zhang, K., S.-C. Chen, D. Whitman, M.-L. Shyu, J. Yan & C. Zhang (2003). ‘A progressive
310 morphological filter for removing nonground measurements from airborne LiDAR data’. In:
311 *IEEE Transactions on Geoscience and Remote Sensing* 41.4, pp. 872–882. DOI: 10.1109/tgrs.
312 2003.810682.