Urban Water Conservation Policies in the United States

Jonathan M. Gilligan[1,2]**, Christopher A. Wold**[3]**, Scott C. Worland**[2]**, John J. Nay**[2]**, David J. Hess**[4]**, George M. Hornberger**[2]

[1]Department of Earth & Environmental Sciences, Vanderbilt University, Nashville, Tennessee, USA

[2]Department of Civil & Environmental Engineering, Vanderbilt University, Nashville, Tennessee, USA

[3]Vanderbilt Institute for Energy and Environment, Vanderbilt University, Nashville, Tennessee, USA

[4]Department of Sociology, Vanderbilt University, Nashville, Tennessee, USA

Key Points:

- Analysis of water conservation policies of 195 cities in 45 states
- Water conservation policies correlate with both environmental and social variables
- Partisan voting patterns at both state and metropolitan levels account for much of the variation

Corresponding author: Jonathan M. Gilligan, jonathan.gilligan@vanderbilt.edu

Abstract

Urban water supply systems in the United States are increasingly stressed as economic and population growth confront limited water resources. Demand management, through conservation and improved efficiency, has long been promoted as a practical alternative to building Promethean energy-intensive water-supply infrastructure. Some cities are making great progress at managing their demand, but study of conservation policies has been limited and often regionally focused. We present a hierarchical Bayesian analysis of a new measure of urban water conservation policy, the Vanderbilt Water Conservation Index (VWCI), for 195 cities in 45 states in the contiguous United States. Cities in states with arid climates and greater tendency to vote for Democratic candidates tend to adopt more conservation measures. Within a state, cities with large and rapidly growing populations and greater propensity than the rest of the state to vote Democratic tend to adopt more conservation measures. Economic factors and climatic differences between cities do not have much effect on the number of measures adopted, but they impact the character of the measures, with arid cities favoring mandatory conservation actions and cities in states with lower real personal income favoring rebates for voluntary actions. Understanding relationships between environmental and societal factors and cities' support for water conservation measures can help planners and policy-makers identify obstacles and opportunities to increase the role of conservation and efficiency in making urban water supply systems sustainable.

1 Introduction

Cities face increasing challenges to their water supply because of complex interactions among drought, infrastructure, population growth, land-use changes, and other natural and human factors. The prospect of climatic change adds to the difficulty of planning robust and sustainable water supply systems, on account both of increasing uncertainty about future supply and demand for water, and also of predicted reductions in water availability in some regions, such as the Southwestern United States [*Melillo et al.*, 2014]. For many years, advocates of soft approaches to managing water resources have stressed the importance of improving the efficiency with which society obtains the water services it requires [*Gleick*, 2002]. Some cities and their associated water-supply systems respond to these challenges by pursuing grand energy-intensive infrastructure projects to draw water from distant or difficult sources, but many have also shown increasing interest in the soft path, managing demand

through efficiency and conservation measures, and a number of cities have made impressive progress [*Fleck*, 2016].

A pressing challenge is to identify the characteristics of successful transitions toward sustainable water use and the necessary conditions for those transitions to spread more widely. Studies have investigated urban water conservation policies and several water conservation indices are available, but these either lack comprehensive coverage of water conservation policies or are geographically limited [*Hess et al.*, 2017; *Saurí*, 2013; *Maggioni*, 2014]. Recent research finds that individual perception of water scarcity and preference for policy action to address scarcity depend not only on the actual degree of water scarcity, but also on the person's ideological worldview [*Switzer and Vedlitz*, 2016], but it is not clear how these individual preferences translate into policy action.

The Vanderbilt Water Conservation Index (VWCI) is an integer score representing the number of measures that a city has taken to reduce its water demand, out of a list of 79 possible policy actions [*Hornberger et al.*, 2015; *Hess et al.*, 2016, 2017]. This list includes 31 requirements, such as restrictions on lawn-watering or mandatory use of water-efficient plumbing in new construction and renovations; and 21 rebates offered for voluntary actions, such as purchasing water-efficient appliances. Previously, we assessed this index and performed preliminary quantitative and qualitative analyses on a subset of the 22 central cities of the largest metropolitan statistical areas (MSAs) in the extended Southwestern United States [*Hess et al.*, 2016].

Our preliminary analysis of the 22 largest MSAs in the Southwest found that propensity to adopt water conservation measures depended both on characteristics of the physical environment (precipitation) and socio-economic and political characteristics of the MSA (partisan voting and cost of living). The Cook Partisan Voting Index for an MSA had, by a large margin, the greatest predictive power for adopting water conservation measures. This finding was supported by qualitative analysis, which suggested that political association of water conservation with general "environmentalist" politics created political differences in the ways that cities respond to the physical fact of water scarcity.

We subsequently expanded our database to the central cities in 197 MSAs in 47 states [*Hess et al.*, 2017], which comprise more than half of the 382 MSAs in the United States [*U.S. Census Bureau, Population Division*, 2016] and here, we present the first quantitative analysis of the larger database, analyzing relationships between environmental and societal
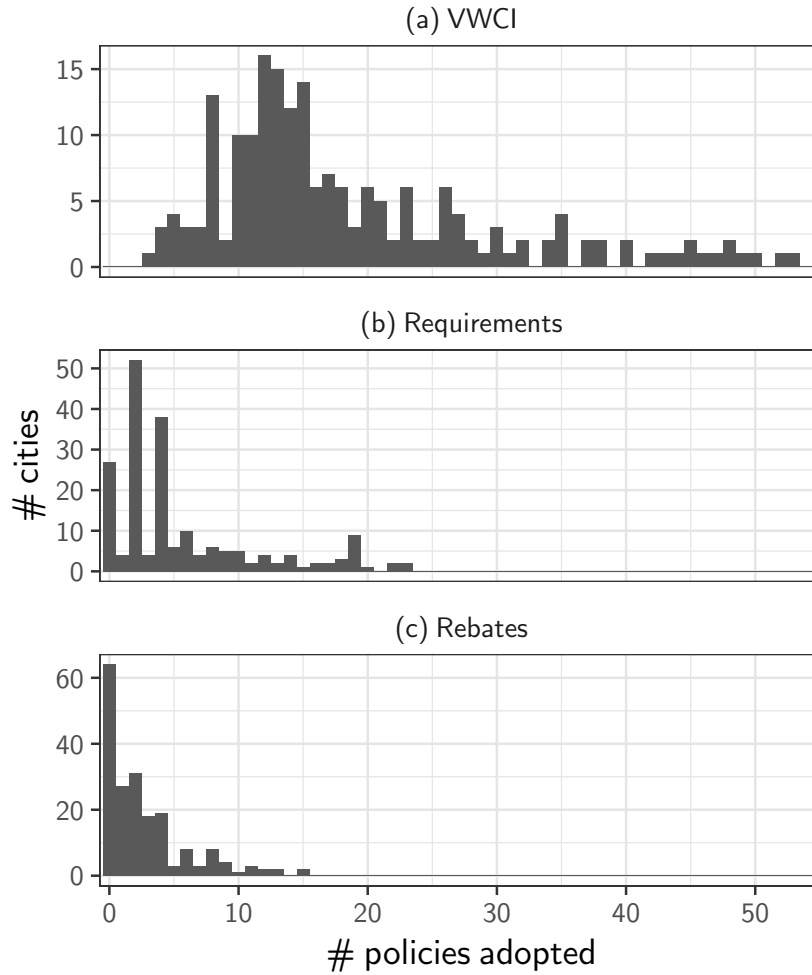
Figure 1. Distribution of VWCI, requirements, and rebates.

characteristics of 195 cities across the contiguous United States and those cities' propensity

to adopt water conservation policies. This analysis considers state-level as well as MSA-level

variables. We identify which variables best explain variations in water conservation policy

regimes, how effects at the state level moderate city-level effects, and whether these effects

manifest differently when we consider specific aspects of water-conservation policy, such

as requirements and rebates. This analysis may help water managers and policy makers to

anticipate how receptive a given city might be to adopting different conservation measures.

2 Data and Methods

2.1 Vanderbilt Water Conservation Index

The full VWCI data set provides detailed information about both the number and nature of urban water conservation policies, which allows hierarchical analysis to examine the relationships between water conservation scores and hydroclimatological and societal characteristics at both state and MSA levels. Our complete database includes Anchorage, AK, and Honolulu, HI, but for this analysis we chose to include only cities within the contiguous United States (195 cities in 45 states, Fig S1) because both climatic and socio-political characteristics of Alaska and Hawaii may be very different from those of the contiguous states, and also because state-level variables that average over the enormous area of Alaska and across the multiple Hawaiian islands may render them less suitable for our hierarchical treatment.

VWCI scores range from 3 to 53 with a mean of 18.7 and a median of 15 (Figure 1a; Table S1). The number of requirements ranges from from 0 to 23 with a mean of 5.6 and a median of 4 (Figure 1b) and the number of rebates ranges from 0 to 15 with a mean of 2.7 and a median of 2 (Figure 1c).

The 20 cities with the greatest VWCI scores are in states in the Western U.S., with the exception of two cities in Florida and one in New York (Tables 1, S1). Cities with similar total VWCI scores, such as New York and Salt Lake City vs. Tampa and Vallejo, El Paso vs. Miami, and Riverside vs. Fresno, may have very different relative contributions from requirements and rebates.

The relative contributions of rebates and requirements to VWCI varied considerably (Table 1), and our previous qualitative research indicated that there might be a preference for rebates over requirements in more politically conservative cities [*Hess et al.*, 2016; *Brown and Hess*, 2016]. Consequently, in addition to the VWCI score we also analyzed the number of rebates and the number of requirements in a city's portfolio of conservation policies.

2.2 Explanatory Variables

We obtained values for the MSA-level average annual average temperature (T) and total precipitation (P) for the period 1970–2014 from the University of Delaware's gridded climate reanalysis [*Matsuura and Willmott*, 2015a,b] and for state-level from the National Cli-

100   **Table 1.**   Cities with the twenty highest VWCI scores. Req. = requirements, Reb. = rebates. A complete list

101   of all 195 cities appears in Table S1.

Rank	City	VWCI	Req.	Reb.	Req./Reb.
1	Los Angeles, CA	53	23	13	1.77
2	San Diego, CA	52	19	15	1.27
3	Santa Rosa, CA	50	19	15	1.27
4	Oxnard, CA	49	23	11	2.09
5	San Jose, CA	48	22	12	1.83
5	Santa Cruz, CA	48	20	11	1.82
7	Austin, TX	47	19	11	1.73
8	San Antonio, TX	46	19	8	2.38
9	Albuquerque, NM	45	19	12	1.58
9	Riverside, CA	45	15	13	1.15
11	Fresno, CA	44	22	8	2.75
12	Denver, CO	43	19	8	2.38
13	San Francisco, CA	42	18	9	2.00
14	Las Vegas, NV	40	18	7	2.57
14	Salinas, CA	40	19	6	3.17
16	El Paso, TX	38	19	3	6.33
16	Miami, FL	38	14	8	1.75
18	Fort Collins, CO	37	9	8	1.12
18	Stockton, CA	37	14	8	1.75
20	New York, NY	35	19	2	9.50
20	Salt Lake City, UT	35	18	2	9.00
20	Tampa, FL	35	14	5	2.80
20	Vallejo, CA	35	14	6	2.33

matic Data Center's divisional temperature and precipitation records for the Continental U.S. [*Vose et al.*, 2014]. We obtained populations for MSAs and states in 2010 and 2014 from the *U.S. Census Bureau, Population Division* [2016] and used them to calculate average annual rates of population growth. Per-capita real personal income (RPI) for MSAs and states in 2014 were was obtained from the *U.S. Bureau of Economic Analysis* [2016a]. The fraction of the public water supply taken from surface water sources (henceforth, surface-water fraction) was obtained from the U.S. Geological Survey's water use report for 2010 [*Maupin et al.*, 2014]. The Cook Partisan Voting Index (PVI) was calculated using state- and county-level presidential votes, averaged over the 2008 and 2012 elections [*Wasserman*, 2013; *CQ Press*, 2016]. Positive PVI indicates a greater Democratic vote share than the national average and negative PVI a greater Republican vote share (Tables S1–S2 and Datasets S1–S2).

We began with the covariates listed above, but adopted a modified set based on preliminary results: The regression coefficients for T and P showed collinearity. This, together with a desire for parsimony, led us to replace them with the Köppen aridity index: $P/(T+33)$, with P in millimeters and T in Celsius. The Köppen index is deemed an especially reliable index of aridity [*Quan et al.*, 2013]. Larger values of this index correspond to wetter conditions and smaller values to drier conditions.

The distribution of MSA population in 2014 was skewed, with a few large cities producing an asymmetric fat tail (Figure S2), so we deemed natural logarithm of population more suitable for a regression analysis [*Gelman and Hill*, 2007, pp. 59–61].

2.3 Regression Analysis

We applied Bayesian hierarchical varying-intercept regression to each of the three conservation scores (VWCI, number of requirements, and number of rebates) with the following explanatory variables: the Köppen aridity index, surface water fraction, PVI, RPI, metropolitan population, and population growth rate. The hierarchical structure of our regression model, which nests cities and MSAs within states, reflects the fact that water resources extend beyond MSA boundaries and that state-level policies affect local actions, both by constraining local ordinances and regulations, and by encouraging or requiring counties and cities to adopt various types of conservation measures.

The regression models each city's conservation score (VWCI, requirements, or rebates) as the result of a beta-binomial process (an overdispersed analogue to a binomial process)

[*Gelman et al.*, 2014, pp. 437–38]. This model characterizes each city i by a probability p_i for adopting water conservation policies. In a pure binomial process, each policy would have the same probability p_i of being adopted in city i, but the beta-binomial process allows for greater dispersion of scores by drawing the probability for each action separately from a beta distribution with mean p_i. The probability p_i is modeled using a hierarchical varying-intercept logistic regression on both state-level and MSA-level covariates:

$$\text{Score}_i \sim \text{beta-binomial}(N_{\text{action}}, \phi p_i, \phi(1 - p_i)), \tag{1}$$

where

$$p_i = \text{logit}^{-1}(y_i), \tag{2}$$

$$y_i = \alpha_{\text{state}_i} + \sum_j \beta_j x_{ij}, \tag{3}$$

$$\alpha_{\text{state}} = \alpha_0 + \sum_k \gamma_k w_{\text{state},k} + \delta_{\text{state}}, \tag{4}$$

N_{action} is the number of possible conservation actions (79 for VWCI, 31 for requirements, and 21 for rebates), β is a vector of MSA-level regression coefficients, x is a matrix of MSA-level covariates, γ is a vector of state-level regression coefficients, w is a matrix of state-level covariates, δ is a random noise term that represents state-level variations not accounted for by the regression model, and ϕ parameterizes the overdispersion, relative to a binomial distribution. As $\phi \to \infty$, the distribution approaches a binomial distribution, and the smaller ϕ is, the greater the overdispersion. We constrain δ to sum to zero in order to keep the model identifiable [*Stan Development Team*, 2016, Ch. 23].

State-level covariates were rescaled to have means of zero and standard deviations of 0.5. Where the same covariate appeared at both the state and MSA level, the MSA-level covariate was rescaled to the same scale as its state-level counterpart and the state-level value was subtracted, so that MSA-level covariate represented the difference between the MSA and the state-average. Covariates that only appeared at the MSA level were rescaled so the mean over all 195 MSAs was zero and the standard deviation 0.5. Standardizing the independent variables served both to facilitate comparison of the relative importance of variables whose natural scales are vastly different [*Gelman et al.*, 2008; *Gelman and Hill*, 2007, pp. 55–57], and also to improve computational performance [*Stan Development Team*, 2016].

177 We represent δ as normally distributed and use weakly informative Cauchy priors for

178 the coefficients and parameters [*Gelman et al.*, 2008]:

$$\alpha_0 \sim \text{Cauchy}(0, 10), \tag{5}$$

$$\beta, \gamma \sim \text{Cauchy}(0, 2.5), \tag{6}$$

$$\delta \sim \text{Normal}(0, \sigma), \tag{7}$$

$$\sigma \sim \text{positive half-Cauchy}(0, 2.5), \tag{8}$$

184 and

$$\phi \sim \begin{cases} \text{Cauchy}(50, 20) & \text{for VWCI.} \\ \text{Cauchy}(15, 10) & \text{for requirements and rebates.} \end{cases} \tag{9}$$

187 The posterior distributions were considerably narrower than the priors, and were not sensitive

188 to changing the scales of the priors.

189 We implemented the statistical model in the Stan probabilistic programming language

190 [*Carpenter et al.*, 2017], which generates a Hamiltonian Monte Carlo sampler. We used R

191 to prepare the data and call the sampler [*R Core Team*, 2016]. We ran each chain for 2000

192 iterations (1000 for warm-up and 1000 for sampling), which yielded a total of 4000 samples.

193 To facilitate interpretation, we report rescaled regression coefficients: $\beta' = \beta/4$ and

194 $\gamma' = \gamma/4$, which represent the approximate change in the probability p corresponding to

195 a two-standard-deviation change of the covariate near the midpoint of the logistic function

196 (where $p = 0.5$) [*Gelman and Hill*, 2007, pp. 81–82]. The corresponding change in VWCI,

197 requirements, or rebates is given by $\beta' N_{\text{action}}$ or $\gamma' N_{\text{action}}$.

198 We tested VWCI for overdispersion by comparing binomial and beta-binomial models

199 using leave-on-out cross-validation (LOO) (Table S3), the Widely Applicable Information

200 Criterion (WAIC) (Table S4), and graphical inspection of the distribution of posterior pre-

201 dictions of the mean, standard deviation, maximum, and minimum of VWCI over the cities

202 in our data set [*Gelman et al.*, 2014; *Vehtari et al.*, 2016]. All three methods favored the

203 overdispersed beta-binomial model. These tests also strongly favored a hierarchial model

204 over a single-level regression on MSA-level covariates.

3 Results

In comparing magnitudes of regression coefficients β and γ (see Section 2), we refer to the average value of the posterior (means and medians were equal within ±0.01; see Tables S5–S7). Where the 95% highest-density interval of the posterior overlaps zero, we refer to the effect as ambiguous because it is consistent with zero and its sign cannot be determined with confidence.

Regressions for VWCI, requirements, and rebates (Figures 2–3; Tables S5–S7) all show that PVI is an important predictor of urban water conservation at both the state and MSA levels. Aridity matters at the state level for all three conservation scores. Variations of aridity from one MSA to another within a state have a pronounced effect on requirements, but make only small and ambiguous contributions to VWCI and rebates. At the MSA level, faster population growth and larger population are associated with higher scores on all three conservation scores. Effects of RPI and surface-water dependence are mostly ambiguous at the state level, with uncertainty making it difficult to say anything definite except in the case of RPI and rebates, and are negligible at the MSA level.

States with more positive (Democratic-leaning) PVI have greater propensity to adopt conservation policies, as do MSAs whose PVI is greater (more likely to vote Democratic) than the rest of the state. States with drier climates (negative aridity) also have greater propensity for conservation. MSAs with large and rapidly growing populations also tend to score higher on all three measures.

For all three conservation scores, the largest effects were for state-level, as opposed to MSA-level, characteristics, but the posterior distributions show considerable overlap so it is important not to over-interpret the ranking of coefficients.

Two differences that stand out among the three measures are that state-level variation in RPI and MSA-level variation in aridity each has ambiguous effects on total VWCI and different effects on requirements versus rebates: state-level RPI affects rebates more strongly than requirements and MSA-level variations in aridity affect requirements more strongly than rebates.

Regression residuals for VWCI range from −14.9 to +17.1, with a root-mean-square value of 5.3 (Figure 4 and Table 2). There is no indication of multicollinearity creating worrisome correlations among the coefficients (Figures. S3–S5).
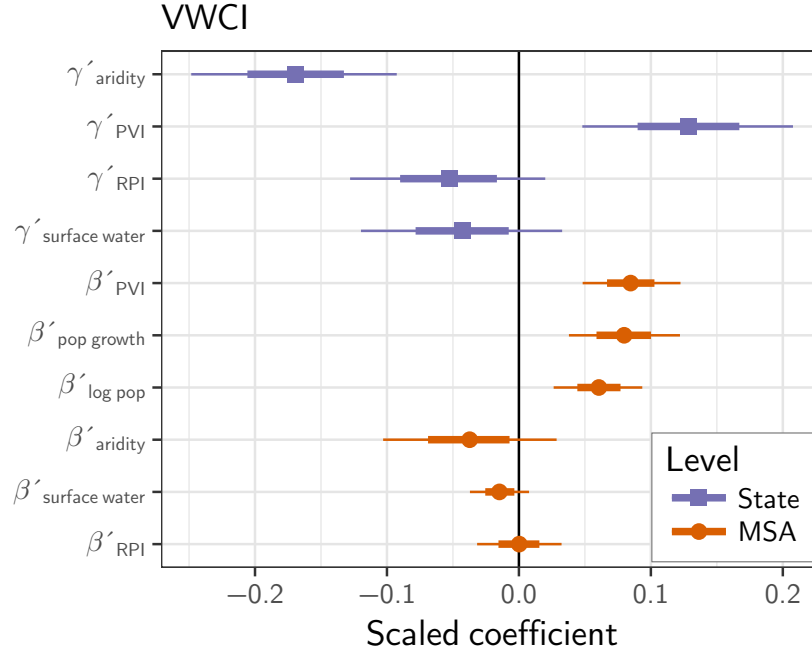
Figure 2. Scaled regression coefficients for VWCI: γ refer to state-level regression coefficients and β to MSA-level ones. For a scaled coefficient of 0.1, a two-standard-deviation change in the predictor corresponds to VWCI changing by about 8 for a city with a VWCI of around 40. The points represent the median of the posterior, the thick lines the 66% highest-density interval (HDI), and the thin lines the 95% HDI. Coefficients are grouped by state vs. city level and then ordered within each group by absolute value of the median.

4 Discussion

This analysis identifies distinguishing characteristics of cities across the contiguous United States that embrace water conservation policies, and allows us to differentiate state-level from MSA-level effects. We find that water conservation is driven both by characteristics of the physical environment and by political and demographic characteristics of cities and states.

Our previous qualitative analysis of the 22 largest southwestern MSAs suggested that partisan differences over water conservation are more muted at the MSA-level than at the state and national level [*Hess et al.*, 2016], but while the quantitative analysis of that data showed that PVI played a large role, it could not distinguish state-level from MSA-level effects. Here, we find that for all three conservation scores (total VWCI, requirements only, and rebates only), PVI is important at both the state and MSA levels. The effect of PVI on all three conservation scores appears to be greater at the state level than at the MSA-level,
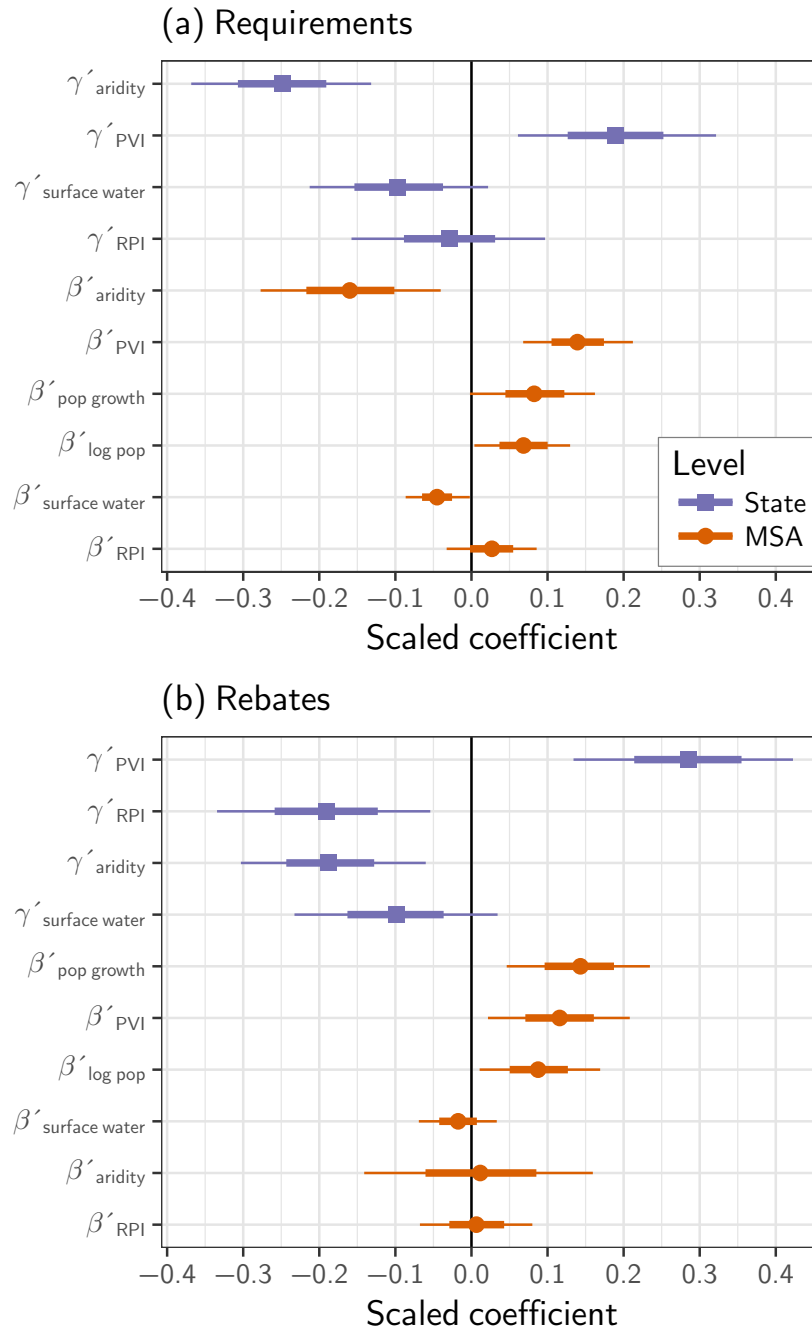
Figure 3. Scaled regression coefficients for conservation requirements and rebates. For a scaled coefficient of 0.1, a two-standard-deviation change of the predictor corresponds to the number of requirements changing by about 3 for a city with around 16 requirements, and a the number of rebates changing by about 2 for a city with around 10 rebates.
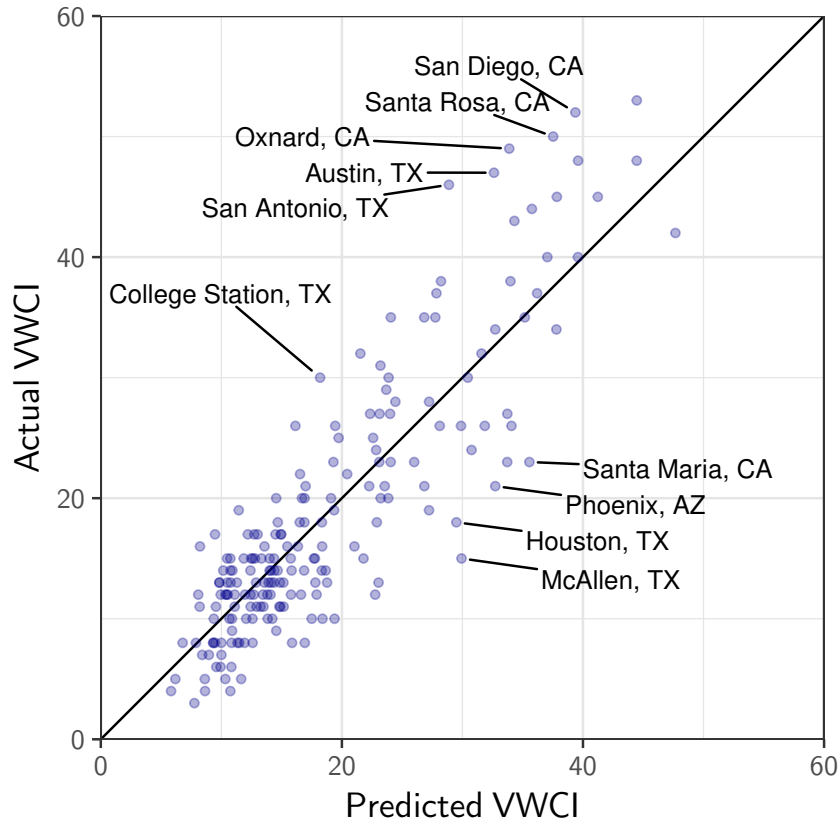
Figure 4. Predicted versus actual VWCI. Cities with the ten largest residuals are labeled.

but the posterior distributions of state-level and MSA-level coefficients overlap too much to permit much confidence in this ranking.

The state climate (aridity) shows clear effects that are consistent across all three conservation scores. The effects of variation in MSA-level aridity within a state on VWCI and rebates are much smaller, ambiguous, and consistent with zero. We interpret this as reflecting the fact that urban water supplies often draw from sources, such as river networks, watersheds, and aquifers, that cover large areas and which may be shared by many cities and many categories of users. However, aridity has a clear effect on requirements, with cities that are drier than the state average tending to adopt more requirements.

RPI measures the real purchasing power of per-capita personal income, adjusted for inflation and regional variations in the cost of living [*U.S. Bureau of Economic Analysis*, 2016b], and thus reflects prosperity. At the state level, greater RPI correlates with lower conservation scores on all three measures, but the effect is small and ambiguous (consistent

Table 2. Cities with the ten largest residuals from VWCI regression.

Rank	City	VWCI	predicted VWCI	residual
1	San Antonio, TX	46	28.9	17.1
2	Oxnard, CA	49	33.9	15.1
3	McAllen, TX	15	29.9	−14.9
4	Austin, TX	47	32.6	14.4
5	San Diego, CA	52	39.4	12.6
6	Santa Maria, CA	23	35.6	−12.6
7	Santa Rosa, CA	50	37.5	12.5
8	College Station, TX	30	18.2	11.8
9	Phoenix, AZ	21	32.7	−11.7
10	Houston, TX	18	29.5	−11.5

with zero) for VWCI and requirements, whereas it is large and clearly nonzero for rebates. Perhaps this reflects greater political support for choosing rebates over requirements when households have less disposable income with which to pay for conservation actions. At the MSA-level, variations of RPI within a state have negligibly small effects, which are consistent with zero, on all three scores.

What emerges in the big picture is that cities in states with higher (more Democratic-leaning) PVI's and more arid climates tend to adopt more conservation measures, including more requirements and more rebates. State-level RPI does not have a clear effect on VWCI, but RPI appears to affect the composition of conservation policies, with lower RPI coreresponding to a greater reliance on rebates. Within a state, cities in MSAs with greater PVI than the state average and those with large and rapidly growing populations tend to adopt more total conservation measures, including more rebates and requirements. Variations in aridity from one MSA to another within a state do not have an appreciable effect on total VWCI, but they do affect the composition of policies, with MSAs whose climate is more arid than the state average tending to favor requirements over other conservation measures.

Brown and Hess [2016] report on detailed interviews with decision-makers from four cities, including San Antonio and Phoenix, which have the largest and ninth-largest residuals, respectively (Table 2). This merits some discussion: San Antonio has a low predicted VWCI

291 in part because of its low (Republican-leaning) PVI. However, federal policy may have con-

292 tributed to San Antonio having a much higher VWCI than predicted by our regression: San

293 Antonio's options for increasing its water supply are constrained by the settlement of a law-

294 suit over endangered species, which requires the U.S. Fish and Wildlife Service to restrict

295 withdrawals from the Edwards Aquifer [*Brown and Hess*, 2016]. Phoenix has a much lower

296 water conservation index than predicted. One contributing factor may be the city's access to

297 water from the Colorado River, by means of the Central Arizona Project, which significantly

298 relieves the water stress that might be expected from the region's hydroclimatology [*Brown*

299 *and Hess*, 2016].

300 One should be cautious about using using qualitative data based on historical condi-

301 tions to explain unusual observations or deviations from a model, but these two examples

302 illustrate the rich complexity of water conservation policy and suggest that in future research,

303 mixed-methods approaches can be valuable, combining statistical analyses with detailed case

304 studies of selected cities to study both the patterns that represent what cities have in common

305 and the distinctive individual characteristics of different cities.

306 In comparing the findings of this analysis to those of our previous preliminary analy-

307 sis of the 22 largest MSAs in the Southwest, both studies identified PVI as a very important

308 predictor of water conservation policies, but with its much smaller and less diverse sample,

309 the previous study could not identify other effects after controlling for PVI, and it could not

310 quantitatively distinguish state-level from MSA-level effects of PVI or other covariates. Here

311 we observe clearly that variations in both environmental and societal characteristics at the

312 state level are relevant to policy adoption; that within a state, variations from MSA to MSA

313 of PVI, population, and population growth are consistently important; and that MSA-level

314 variation in climate does not affect the number of conservation policies adopted, but affects

315 the kinds of policies adopted.

5 Conclusion

317 An integrated perspective that draws on social science and natural science variables

318 shows that the adoption of urban water conservation policies cannot be explained by consid-

319 ering only hydroclimatological factors, such as aridity and the surface water fraction. Soci-

320 etal variables, such as political leanings, are also important. This analysis finds that for the

321 most part, hydroclimatological variables are important at the state level, but not at the MSA

level, whereas PVI is important at both the state and MSA level. Economic well-being has

smaller and more ambiguous effects, but appears to affect which categories of conservation

policies a city is likely to favor. These results suggest that large, rapidly growing, and more

politically liberal cities, and cities in arid and politically liberal states, will be more likely to

adopt water conservation policies. Policy rationales for water conservation and proposals for

specific conservation measures would likely benefit from taking into account the complex

mix of factors revealed by integrated social and natural science research.

Acknowledgments

References

Brown, K. P., and D. J. Hess (2016), The politics of water conservation: Identifying and
overcoming political barriers to successful policies, *Water Policy*, doi:10.2166/wp.2016.
089.

Carpenter, B., A. Gelman, M. Hoffman, D. Lee, B. Goodrich, M. Betancourt, M. A.
Brubaker, J. Guo, P. Li, and A. Riddell (2017), Stan: A probabilistic programming lan-
guage, *J. Stat. Soft.*, *76*, 1–32, doi:10.18637/jss.v076.i01.

CQ Press (2016), Voting and elections collection, available at `http://library.cqpress.
com/elections/`.

Fleck, J. (2016), *Water is for Fighting Over and Other Myths about Water in the West*, Island
Press.

Gelman, A., and J. Hill (2007), *Data Analysis using Regression and Multilevel/Hierarchical
Models*, Cambridge Univ. Press.

Gelman, A., A. Jakulin, M. G. Pittau, and Y.-S. Su (2008), A weakly informative default
prior distribution for logistic and other regression models, *Ann. Appl. Stat.*, *2*, 1360–1383.

Gelman, A., J. B. Carlin, H. S. Stern, D. B. Dunson, A. Vehtari, and D. B. Rubin (2014),
Bayesian Data Analysis, 3rd ed., CRC Press.

Gleick, P. H. (2002), Water management: Soft water paths, *Nature*, *418*, 373–373, doi:10.
    1038/418373a.

Hess, D. J., C. A. Wold, E. Hunter, J. J. Nay, S. Worland, J. Gilligan, and G. M. Hornberger
    (2016), Drought, risk, and institutional politics in the American Southwest, *Sociol. Forum*,
    31, 807–827, doi:10.1111/socf.12274.

Hess, D. J., C. A. Wold, S. C. Worland, and G. M. Hornberger (2017), Measuring urban wa-
    ter conservation policies: Toward a comprehensive index, *JAWRA Journal of the American
    Water Resources Association*, doi:10.1111/1752-1688.12506.

Hornberger, G. M., D. J. Hess, and J. Gilligan (2015), Water conservation and hydrological
    transitions in cities in the United States, *Water Resour. Res.*, *51*(6), 4635–4649, doi:10.
    1002/2015WR016943.

Maggioni, E. (2014), Water demand management in time of drought: What matters for water
    conservation, *Water Resour. Res.*, *51*, 215–139, doi:10.1002/2014wr016301.

Matsuura, K., and C. J. Willmott (2015a), *Terrestrial Air Temperature: 1900–2014 Grid-
    ded Monthly Time Series (Version 4.01)*, Dept. Geography, U. Delaware, available at
    `http://climate.geog.udel.edu/~climate/html_pages/Global2014/README.`
    `GlobalTsT2014.html`.

Matsuura, K., and C. J. Willmott (2015b), *Terrestrial Precipitation: 1900–2014 Grid-
    ded Monthly Time Series (Version 4.01)*, Dept. Geography, U. Delaware, available at
    `http://climate.geog.udel.edu/~climate/html_pages/Global2014/README.`
    `GlobalTsP2014.html`.

Maupin, M., J. Kenny, S. Hutson, J. Lovelace, N. Barber, and K. Linsey (2014), Estimated
    use of water in the United States in 2010, *Circular 1405*, U.S. Geological Survey, doi:10.
    3133/cir1405.

Melillo, J. M., T. C. Richmond, and G. W. Yohe (Eds.) (2014), *Climate Change Impacts in
    the United States: The Third National Climate Assessment*, U.S. Global Change Research
    Program, doi:10.7930/J0Z31WJ2.

Quan, C., H. Shuang, U. Torsten, Z. Chunhua, and Y.-S. Liu (2013), Validation of
    temperature—precipitation based aridity index: Paleoclimatic implications, *Palaeogeogr.
    Palaeoclimatol. Palaeoecol.*, *386*, 86–95, doi:10.1016/j.palaeo.2013.05.008.

R Core Team (2016), *R: A Language and Environment for Statistical Computing*, R Founda-
    tion for Statistical Computing, Vienna.

Saurí, D. (2013), Water conservation: Theory and evidence in urban areas of
the developed world, *Annu. Rev. Environ. Resour.*, *38*, 227–248, doi:10.1146/
annurev-environ-013113-142651.

Stan Development Team (2016), *Stan Modeling Language: User's Guide and Reference
Manual*, version 2.14.0, available at `http://mc-stan.org/documentation/`.

Switzer, D., and A. Vedlitz (2016), Green colored lenses: Worldviews and motivated rea-
soning in the case of local water scarcity, *Environ. Behav.*, *advance online publication*,
doi:10.1177/0013916516669391.

U.S. Bureau of Economic Analysis (2016a), RPI1 real per-capita personal income, available
at `http://www.bea.gov/iTable/index.cfm`.

U.S. Bureau of Economic Analysis (2016b), Real personal income and regional price par-
ities, *Regional Economic Accounts Methodology Report RPP2016*, U.S. Bureau Econ.
Analysis, available at `http://www.bea.gov/regional/pdf/RPP2016_methodology.`
`pdf`.

U.S. Census Bureau, Population Division (2016), Metropolitan and micropolitan statistical
areas, available at `https://www.census.gov/data/datasets/2016/demo/popest/`
`total-metro-and-micro-statistical-areas.html/`.

Vehtari, A., A. Gelman, and J. Gabry (2016), Practical Bayesian model evaluation us-
ing leave-one-out cross-validation and WAIC, *Statistics and Computing*, doi:10.1007/
s11222-016-9696-4.

Vose, R. S., S. Applequist, M. Squires, I. Durre, M. J. Menne, C. N. Williams, C. Fenimore,
K. Gleason, and D. Arndt (2014), Improved historical temperature and precipitation time
series for U.S. climate divisions, *J. Appl. Meteor. Climatol.*, *53*, 1232–1251, doi:10.1175/
JAMC-D-13-0248.1.

Wasserman, D. (2013), *Introducing the 2014 Cook Political Report Partisan Voter Index*,
Cook Political Report, available at `http://cookpolitical.com/story/5604`.