

Wall-to-wall mapping of peat depth from Lidar terrain and airborne radiometrics in Norwegian landscapes

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Abstract. The abstract goes here. It can also be on *multiple lines*.

1 Introduction

Introduction text goes here. Read Gatis et al. (2019) and related work (Minasny et al., 2019).

2 Materials and methods

5 2.1 Sites

We assessed how well we could predict peat depth at two sites with conspicuously different physical geography: Skrimfjella in eastern Norway and Ørskogfjellet in western Norway (Fig. 1c). These sites were chosen because they were covered by radiometric data from airborne surveys, relatively little built-up area, and road access.

At Skrimfjella we delineated a study area of 34 km² based on radiometric coverage and accessibility (Fig. 1b). The study
10 area has a diverse bedrock, with 32 % alkali feldspar granite, 26 % mergelstein, 10 % granite, and eight other types with > 1 % coverage (NGU, 1:250 000 dataset). The landscape within our delineation is classified as *inland hills and mountains* (Simensen et al., 2021). It is almost without human infrastructure, dominated by forest, and borders on a large nature reserve. The study area has a mean elevation of 438 m above sea level (range 223–711, IQR 351–509), and its mean slope at 10 m resolution is 10.8° (IQR 4.6–15.1°). In Norway’s AR5 national land capability dataset (Ahlstrøm et al., 2019), 1.5 km² (4.5 %) of the study
15 area is classified as mire — defined as areas with mire vegetation and at least 30 cm of peat depth.

At Ørskogfjellet we defined a study area of 124 km² which basically followed the footprint of the radiometric survey (Fig. 1a). According to the Geological Survey of Norway, bedrock in the area is 84 % granitic gneiss, 11 % granite, and 5 % aluminium silicate gneiss (NGU, 1:250 000 dataset). This study area comprises a wide range of major landscape types: *coastal plains*, *coastal fjord*, *inland valleys*, as well as *inland hills and mountains* (Simensen et al., 2021). It is mostly forested, but
20 also contains considerable farmland and open upland, and has several large lakes. Its mean elevation is 211 m above sea level



Figure 1. Study areas at Ørskogfjellet (a) and Skrimfjella (b) within southern Norway (c). Land cover shown here is from the AR50 national land resource database and has simplified geometry with respect to the AR5 database used in the study.

(range 0–807, IQR 73–310), and its mean slope at 10 m resolution is 13.0° (IQR 4.7–18.3°). The AR5 dataset counts 15.3 km² (12.4 %) of the study area as mire.

2.2 Peat depth measurements

At both study sites, our measurements of peat depth were made for the purpose of training a Random Forest model of peat depth, and we designed our sampling with this in mind (Brus, 2019). Broadly, we aimed for a sample that was representative the predictor space of the most important predictors of peat depth (Wadoux et al., 2019; Ma et al., 2020). A sample that preserves the properties of the multivariate distribution of predictor and outcome variables is most likely to maintain any complex, non-linear relationships that exist in the population while avoiding spurious ones (Brus, 2019). We chose for our sampling and modelling a spatial resolution of 10 m. We considered this a reasonable compromise between digital terrain model resolution (1 m) and small mires on the one hand, and airborne radiometric resolution (50 m) on the other.

2.2.1 Skrimfjella

We measured peat depth in selected locations (10 m raster cells) at Skrimfjella. The locations were chosen only from areas delineated as mire in the AR5 national land capability dataset. Within this mire area, we stratified our sample across values of elevation, slope, and potassium ground concentration (from processed airborne gamma ray spectrometry, Baranwal et al., 2013). Specifically, we used the *eSample* function in the *iSDM* R package (v.1.0) to chose an environmentally systematic sample. This function defines the environmental space as a two-dimensional convex hull around the ordinated data, then creates a regular grid across that space, and lastly finds for each grid cell the datum that is nearest (Hattab et al., 2017). Elevation was extracted from the 10 m national digital terrain model, slope calculated in degrees, and potassium ground concentration downsampled with bilinear resampling. We set a target sample size of 100, excluded the top and bottom percentile from the convex hull, and with these parameters *eSample* returned 105 raster cells.

In addition to the peat depth locations, we had another arm of our sampling design for measuring peatland occurrence, as binary variable. We wanted to measure peatland occurrence outside of mapped mire areas because the AR5 dataset is known to underestimate peatland coverage (especially in forests, Bryn et al., 2018), and because airborne radiometrics may help identify unmapped peatland (Gatis et al., 2019; O’Leary et al., 2022). The occurrence locations were sampled from the part of the study area that (1) was mapped as something other than mire in the AR5 database and (2) had a slope < 20°. We performed environmentally systematic sampling of this population with the same procedure as for the depth locations, and *eSample* returned 106 raster cells.

Field work at Skrimfjella was conducted in August 2020. We navigated to the centers of the raster cells in the depth and occurrence samples by handheld GPS, checking that positional error was below 3 m. For each depth sample location, we measured peat depth three times (at the vertices of a triangle with 2 m sides) to get a more representative value for the 10 m raster cell, and to dampen the effect of outlying measurements (Parry et al., 2014). We used a metal probe pushed downward until resistance indicated the base of the peat column. Probe locations were adjusted up to 20 cm if the base of the peat column seemed to be blocked by an obvious artifact. For each occurrence sample location, we recorded the presence or absence of

peatland — primarily by digging and examining the top 20 cm of soil (where this was possible). We judged whether the soil was a peat soil based on its density, texture, and color. Occasionally, when the soil itself was difficult to judge, we made our determination also based on the presence or absence of mire vegetation. Although peat soil is strictly defined by organic content (which we did not analyse), we believe our protocol produced reasonable determinations of presence or absence that would generally satisfy most of the varying definitions of peatland (Minasny et al., 2023).

Besides the depth and occurrence measurements described above, we also measured peat depth in three subjectively-chosen, individual mires, using ground-penetrating radar (GPR). We used the Malå ProEx GPR system (Guideline Geo AB, Sweden) with its 500 MHz shielded antenna mounted in a plastic sledge, and its control unit connected to a GNSS receiver. At each of the three mires we recorded GPR traces along walking transects that covered the extent of the mire, mostly in traversing, zigzag patterns with between 5 m and 20 m spacing at their vertices. Along the GPR transects we also probed peat depth at marked trace locations, to be able to calibrate the GPR wave speed velocity. We processed the GPR data with Reflex2DQuick software (v.3.0; Sandmeier Scientific Software, Germany), applying a time-zero correction, a dewow filter, and a gain filter based on observed energy decay. Then we picked strong reflectors in the radargrams that we interpreted as the base of the peat column. We used picks at marked trace locations to calibrate wave speed velocity; we pooled calibration points across the three mires and fitted a linear regression of depth on one-way travel time with the intercept fixed at zero. In total we had 46 calibration points along 3.5 km of GPR transects. Finally, we used the calculated wave velocity (0.0387 m ns^{-1} , $R^2 = 0.874$) to convert the travel times of all picks to calibrated peat depths.

2.2.2 Ørskogfjellet

At Ørskogfjellet we also measured peat depth in a sample of 10 m raster cells, selected from the part of the study area classified in the AR5 dataset as mire. Before selecting locations, we determined a minimal sample size that would adequately capture the terrain and radiometric properties of the entire mire area. Specifically, we aimed to identify the size at which adding locations produced diminishing decreases in divergence between sample and population distributions — i.e. the elbow point in a curve of similarity between sample and population (Malone et al., 2019). This approach has been found to identify sample sizes that correspond with diminishing returns in predictive model performance on external evaluation data (Saurette et al., 2023). We defined a sequence of sample sizes (50–500) and for ten replicate samples at each size (drawn by conditioned latin hypercube sampling, Minasny and McBratney, 2006; Roudier, 2011), we calculated the mean Kullback-Leibler divergence between sample and population distributions (Malone et al., 2019; Saurette et al., 2023). The variables in the divergence calculation were terrain slope and four radiometrics: potassium, thorium, uranium, and total count. Next, we fitted a asymptotic regression of mean divergence on sample size, and identified the sample size at which the curve reached 95 % of the fitted asymptote. Through this procedure we found that we could adequately capture the population distribution with a sample of 160 locations.

To choose 160 locations, we performed feature space coverage sampling. This approach has been found to produce higher accuracy in Random Forest models than conditioned latin hypercube sampling (Wadoux et al., 2019; Ma et al., 2020). Feature space coverage sampling aims to disperse samples as uniformly as possible in multidimensional predictor space, and is

implemented by choosing locations that are closest to cluster centers in a k-means clustering of the standardized predictor space (Brus, 2019). Feature space coverage sampling works best when all dimensions are important predictors of the outcome (Wadoux et al., 2019), and we used the same five predictors that we used to choose sample size: terrain slope and four radiometrics. We adjusted the feature space coverage sampling to ensure that locations were accessible within time constraints, and assessed how this changed our sample from an ideal feature space coverage sample. Adjusting for accessibility is justified because the smaller sample size that would result if accessibility were ignored can degrade model accuracy as much or more as deviations from ideal sampling designs (Wadoux et al., 2019; Ma et al., 2020). To adjust, we first restricted the sampling population to mire areas that were within an arbitrary cost distance of publicly-accessible roads. Cost distance was calculated using GRASS's *r.walk* function, with friction costs defined by AR5 land classes (GRASS Development Team, 2022). After creating a feature space coverage sample with this restriction, we also inspected a map of the sample and substituted 16 inaccessible locations with accessible locations from the same or a nearby cluster. Our two accessibility adjustments increased the distance in standardized predictor space between sample locations and cluster centers by 78 % (with respect to the ideal sample), but distance in our sample was still only 46 % of the mean distance to cluster centers — i.e., accessibility did not force locations far from cluster centers relative to the size of the clusters.

Field work at Ørskogfjellet was conducted in August 2023. We navigated to the centers of the raster cells in the sample using real time kinematic differential GNSS (Topcon Positioning Systems, USA), to ensure sub-meter positional accuracy. At each location we measured peat depth three times by manual probing, with probe locations spaced approximately 2.5 m apart. In areas with dense sampling locations, we also measured peat depth with GPR along snaking transects passing through the centers of the sampling cells (seven transects, 6.2 km total length). We used the same GPR system as at Skrimfjella, but with a 100 MHz Malå rough terrain antenna (Guideline Geo AB, Sweden) at some transects. To navigate the GPR transects, we placed flags at the cell centers of the sample locations, and used a handheld GNSS receiver to guide the GPR operator. At sampling locations crossed by a GPR transect, we arranged the manual probe positions along the transect (for better calibration of the GPR wave speed velocity), while other locations were probed in a triangular pattern around the cell center like at Skrimfjella.

We processed the GPR data with Reflexw software (v.8.5; Sandmeier Scientific Software, Germany), applying a dewow filter, time-zero correction, bandpass filter, gain filter, and a dynamic correction that accounts for the non-vertical wave path between offset transmitter and receiver antennae. The last correction is important for the rough terrain antenna, which has an antenna separation (2.2 m) — comparable to typical peat depths. As with the data from Skrimfjella, we picked the base of the peat column from strong reflectors in the radargram, and calibrated wave velocity with manual probe measurements in a linear regression. The points in the regression were created by joining to each probe measurement the travel time of the nearest pick, but only if these were within 2 m of each other. In total we had 78 calibration points along 7.8 km of interpretable GPR traces (transect length exceeded because of extra GPR data). Finally, we used the calculated wave velocity (0.0427 m ns^{-1} , $R^2 = 0.946$) to convert the travel times of all picks to calibrated peat depths.

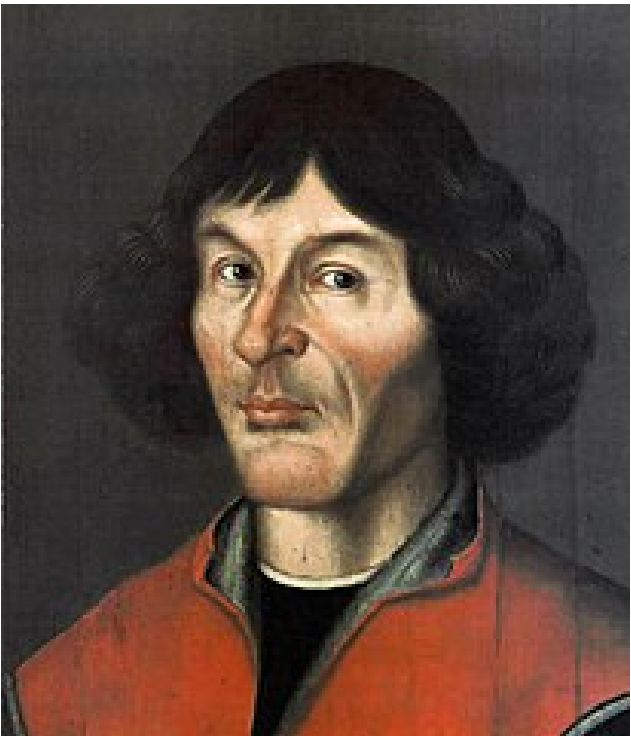


Figure 2. one column figure

Table 1. TEXT

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Table Footnotes

120 **2.3 Peat depth predictors**

2.4 Predictive models of peat depth

3 Results

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3.1 Tables

125 You can add \LaTeX table in an R Markdown document to meet the template requirements (Table 1).

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Table 2. My caption

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4 Discussion

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5 Conclusions

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Appendix A: For submission

140 “Appendices: all material required to understand the essential aspects of the paper such as experimental methods, data, and interpretation should preferably be included in the main text. Additional figures, tables, as well as technical and theoretical developments which are not critical to support the conclusion of the paper, but which provide extra detail and/or support useful for experts in the field and whose inclusion in the main text would disrupt the flow of descriptions or demonstrations may be presented as appendices. These should be labelled with capital letters: Appendix A, Appendix B etc. Equations, figures and
145 tables should be numbered as (A1), Fig. B5 or Table C6, respectively. Please keep in mind that appendices are part of the manuscript whereas supplements (see below) are published along with the manuscript.”

Appendix B: Figures and tables in appendices

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. *Author contributions.* JV: Conceptualization, Investigation, Data curation, Formal analysis, Writing – original draft. NG: Conceptualization, Methodology, Writing - review & editing. MKG: Investigation, Writing - review & editing. KKM: Investigation, Data curation, Writing - review & editing. SDN: Investigation, Writing - review & editing. KR: Conceptualization, Investigation, Writing - review & editing. MS: Investigation, Data curation, Writing - review & editing.

160 . *Competing interests.* The authors declare that they have no conflict of interest.

. *Disclaimer.* The authors declare that the results, discussions, and interpretations presented in this study are solely their own. The views expressed herein do not necessarily reflect those of their respective institutions or funding agencies.

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