

Geographical Variations in the Sensitivity of Terrestrial Biodiversity to Anthropogenic Pressures

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25/08/2022

**A thesis submitted in partial fulfilment of the requirements for the degree of Master of
Science at Imperial College London**

Submitted for the MSc in Computational Methods in Ecology and Evolution

14 **Declaration**

15 Data was obtained from existing online databases, and therefore I was not responsible for data
16 processing or cleaning.

17 Were any mathematical models developed by you or by your supervisor?

18 What role, if any, did your supervisor play in developing the analyses presented?

19 **Abstract**

20 To our knowledge, very little prior research has studied geographic differences in sensitivity to biodi-
21 versity pressures. Hence, studying geographical variation in sensitivity would be useful in comparing
22 the impact of pressures on global biodiversity.

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48 Introduction

49 Ecosystems with intact biodiversity provide services such as clean air and pollination, which make the
50 earth habitable for humans (Leemans and De Groot, 2003). Biodiversity loss diminishes ecosystem
51 productivity (Duffy et al., 2017) and threatens all life on earth, including human well-being (Díaz et al.,
52 2006) leading to unstable environments which are less resistant to change. Both natural and anthro-
53 pogenic pressures act on biodiversity (Nobel et al., 2020), but it is the latter which is accelerating
54 extinction rates dramatically (Ceballos et al., 2015). Understanding the impacts of anthropogenic
55 pressures on biodiversity is necessary to inform more effective policies, strategies and tools (Díaz
56 et al., 2006). This understanding is aided by an understanding of the distribution of biodiversity, how
57 its pressures are distributed, and understanding how sensitive biodiversity is to these pressures. The
58 world's biodiversity is not equally distributed, it varies geographically (Gaston, 2000; Ricklefs, 2004;
59 McRae et al., 2017) as do the pressures acting on it (Millennium ecosystem assessment, 2005; Sala
60 et al., 2000; Bowler et al., 2020), and these distributions often correlate (Ament et al., 2019; Velde,
61 2022). Despite any such correlations between biodiversity and its pressures, there will always be
62 variation in biodiversity response to such pressures due to variations in sensitivity of species (Bowler
63 et al., 2020).

65 Interspecific Variation in Sensitivity

66 Interspecific variation exists in response to biodiversity pressures (Foden et al., 2013). Given that
67 both species and higher taxa (Sunday et al., 2015) have shown variations in sensitivity, and each re-
68 gion of the world comprises different combinations of species groups (van Goethem and van Zanden,
69 2021), there is reason to believe that sensitivity to biodiversity pressure varies depending on loca-
70 tion. Despite this, geographical variation in sensitivity is rarely accounted for (Newbold et al., 2020;
71 Sala et al., 2000). Newbold et al. (2020) stated that by including the sensitivity of individual species,
72 distribution models implicitly capture geographical variation of sensitivity, but rarely explicitly include
73 geographical variation in sensitivity. To the best of my knowledge there is no research to support this
74 exclusion of geographical variation in sensitivity. An understanding of the sensitivity of biodiversity of
75 an entire region could be a more accurate metric for studies of a broad scale.

77 Sensitivity Variation at Broader Levels

78 Biodiversity sensitivity varies at broader levels than species and higher taxon, such as biomes and
79 regions (discussed below), further supporting the concept that sensitivity variation could exist at lev-
80 els as broad as continent. Biodiversity sensitivity differs between biomes for the following biodiversity
81 pressures; pollution (nitrogen exceedance) (Alkemade et al., 2009), land use change and climate
82 change (Newbold et al., 2020). The most sensitive biomes were tropical biomes (Barlow et al., 2016)
83 with the least sensitive being temperate and boreal biomes (Newbold et al., 2020; Cazalis et al.,
84 2021; Barlow et al., 2016). Despite geographical variations in sensitivity existing between biomes, lit-
85 tle is known about continental differences. Biomes are unequally distributed between the continents,
86 with South America having a higher percentage cover of tropical forest than any other continent, and
87 Europe having the lowest closely followed by North America (Wade et al., 2003). Because of the
88 aforementioned sensitivity differences between biomes, it could be that Europe and North America

will have lower sensitivity to biodiversity pressures than other continents, with South America being the most sensitive. However, of the minimal research that has studied continental differences, findings showed tropical forest biodiversity sensitivity to be higher in Asia than in other regions (Americas and Africa) (Gibson et al., 2011). Despite the apparent contradiction with the above prediction (South America being the most sensitive), it must be noted that it was only Asia's tropical biome that was shown to be the highest among continents, not the biodiversity of the continent on the whole, and also not all continents were considered.

One of the papers which studied interspecific sensitivity to environmental pressures of European species (Louette et al., 2010), suggested country-level differences in sensitivity variation. Using species distribution data, the proportion of sensitive species in each country was used to map the effects of a change in each biodiversity pressure on the biodiversity of Europe. The tool ('Bioscore') produces a map of impact across Europe in response to changes in each of the biodiversity pressures, which suggests that even if the magnitude of a biodiversity pressure is constant across Europe, biodiversity response can still vary according to country, due to varying sensitivity of the species within such country. Though the map produced by the tool appears to show variations in sensitivity between countries, the tool is outdated and the data and code are inaccessible, with no values for country level sensitivities having been published. The BioScore tool's predictions support the concept that country-wide differences in sensitivity could exist, however an up to date, wider-breadth study is necessary to observe worldwide variances in countries sensitivities to biodiversity pressures.

Knowledge Gap

A possible explanation for the inadequate research on sensitivity distribution is that sensitivity studies typically use a method of determining the sensitivity of individual species (obtained from published literature about individual species' responses to change), and then mapping these sensitivity values to look for trends (Louette et al., 2010). Because of the intensity of this method, and the under-representation of certain regions (Collen et al., 2008), a global sensitivity analysis looking at regional biodiversity sensitivity would be very labour-intensive. Additionally, most distribution studies focus on the most important pressures (Ferrier et al., 2016). This study aims to use an alternate method, as described further below, using biodiversity and pressure trends at the regional level to look for geographic patterns in sensitivity to the five main biodiversity pressures (climate change, land use change, pollution, overexploitation and invasive species). Having regional information on biodiversity sensitivity will allow the impact of pressures on biodiversity to be better estimated in situations where it is not clear which species are being impacted, and only the area of impact is known. Rather than just including how biodiversity in general is likely to respond, it would be more useful to know how local biodiversity in that specific region is likely to respond.

Research aims

Given that anthropogenic impact on the environment is worldwide (Plumptre et al., 2021), the question should be raised of whether the geographic location of biodiversity pressures affects their impact on global biodiversity. The understanding of variations in biodiversity sensitivity, along with many

130 other aspects of biodiversity, has knowledge gaps which desperately need filling ([Pereira et al., 2012](#)).
131 If such geographic differences exist, they should be taken into account when attributing biodiversity-
132 related merit to investments. To widen the scope of impact outside of the Benchmark for Nature
133 project, taking into account geographic variations in sensitivity to biodiversity pressures could make
134 estimates about biodiversity impact more accurate. Better understanding of biodiversity pressures will
135 aid a better understanding of the implications of investments (and other policies) on natural ecosys-
136 tems.

137 The aim of this project is to investigate whether sensitivity of biodiversity to pressures varies
138 geographically. I will amalgamate country-level data to look at continental differences in the main
139 pressures on biodiversity. I hypothesise that continents with a higher proportion of tropical biomes
140 (e.g. South America) will have higher sensitivity scores than continents with a lower proportion of
141 tropical biomes (e.g. Europe and North America) across all pressures.

Methods

Overview

The focus of this study is on anthropogenic biodiversity pressures only. Anthropogenic pressures on biodiversity are typically grouped, in the current literature, into 5 main pressures; climate change, land use change, pollution, invasive species and overexploitation (Watson et al., 2019). In order to assess whether sensitivity to each pressure varies by country, data was needed in the form of time series (how each of these pressures had been changing in each country over time, as well as how each country's biodiversity had been changing over time). The time series of biodiversity in a country was compared to the time series of a pressure on biodiversity in that country, in order to extract a 'sensitivity score' for each country to assess any effect of geography. Due to comparisons between countries, only terrestrial biodiversity was included.

First, each pressure's geographic relationship with biodiversity was assessed in isolation. It is important to look at individual biodiversity pressures, as opposed to an aggregated pressure on biodiversity, because the pressures have spatial differences (Steffen et al., 2015), meaning the geographical magnitude of each pressure varies. Therefore in order to understand how countries differ in their responses to biodiversity pressures, it must be taken into account the magnitude of each pressure that each country experiences.

I used linear models on time series data about biodiversity and its five main pressures at the country level, and interpret the gradient coefficients as the sensitivity' of that country's biodiversity to each pressure. I will further analyse these sensitivity scores to look for geographic variation in sensitivity between the continents. R version 3.6.3 was used for statistical analysis and plotting, and results were reported as statistically significant is $p \leq 0.05$. In comparison to the previously mentioned common method of gathering information about species-level sensitivities and mapping them, this method is less labour intense and is designed to give broad insights into whether biodiversity trends differ among continents.

Data

The variable chosen to represent biodiversity was biodiversity intactness. The National History Museum's (NHM) Biodiversity Intactness Index (BII)(Heather Phillips, 2021) was chosen as it presents biodiversity in the context of how many original species remain (relative to reference populations). The unit of BII is % as it is a proportion of species. The NHM's Index is the best for this project as the database used is that of the PREDICTS project, which more geographically representative than other datasets (Purvis et al., 2018). This allows for direct comparison of these changes, with the changes in anthropogenic pressures. Historical BII data spanned 18 years between 1970 - 2014.

Climate change time series data was obtained in the form of annual average temperature for each country. The temperature dataset chosen was from the World Bank's Climate Change Knowledge Portal. This dataset was chosen because it contains comprehensive historical data, providing an annual average temperature for every year from 1900 until 2020.

183

184 Land use change data chosen was from The Global Human Settlement Layer data package (A
 185 et al., 2019). The data contains information on built-up area change over time, which is the variable
 186 chosen to represent land use change. The dataset provides the total area of built up land (in square
 187 kilometres) for various years. Collective land use change is difficult to quantify from land use statis-
 188 tics. Although satellite data is available to categorise land cover type over time, calculating annual
 189 land use change from the proportion of each land cover type is not necessarily accurate, as land
 190 use change can be multi-directional. Current studies assessing the impact of land use change on
 191 biodiversity are often meta-analyses or use a natural regional situation as the reference land type
 192 (De Baan et al., 2013) as opposed to observing direct impacts of land use change. A limitation of this
 193 dataset is that only 4 years of data are included.

194

195 Pollution was represented by greenhouse gases (GHG), as the focus is on terrestrial biodiversity.
 196 The dataset used to access GHG emissions for each country over time was the 'National Inventory
 197 Submissions' section of the United Nations - Climate Change website (?). GHG emissions in this
 198 dataset have a unit of 'thousand tonnes of CO2 equivalent'. GHG emissions are presented both in-
 199 cluding and excluding 'Land Use, Land-Use Change and Forestry (LULUCF)' related emissions data.
 200 When assessing pollution and biodiversity links in isolation, LULUCF was included. However, when
 201 modelling all biodiversity pressures together, LULUCF was tested for collinearity with the land use
 202 change variable, and consequently included/excluded. The OECD.stat website was used to down-
 203 load the land use change and pollution data.

204

205 Invasive species data was obtained from the Alien Species First Record Database (??). This
 206 database provides information about how many new invasive species were recorded in each country,
 207 every year. The reason that this database was used is that it provides time series data. The time
 208 span of first records of species spans from 7000BC to 2020.

209

210 Overexploitation was excluded for multiple reasons. Firstly, overexploitation is the vaguest of the
 211 main pressures, and is usually used in the context of fishing (marine biodiversity being beyond the
 212 scope of this project). One of the most relevant aspects of overexploitation to is deforestation, which
 213 is (maybe remove this part depending on methods) already represented in the variable used for land
 214 use change. Though there are other aspects of overexploitation that would be relevant (e.g. illegal
 215 wildlife trading), there are no databases/studies available representing overexploitation by country.

216 Data Wrangling

217 Each data set has data from a different combination of years, and countries. For each pressure being
 218 investigated, only data from years and countries that are shared between that particular dataset and
 219 the biodiversity dataset is included.

220

221 For each pressure, the datasets were wrangled and refined to obtain two time series (at an annual
 222 level) for each country; biodiversity and the magnitude of the particular pressure.

223 **Modelling**

224 For each country, a linear model was fit with biodiversity (BII) as the response variable, and the
225 biodiversity pressure (e.g. pollution) as the explanatory variable. Though both the biodiversity and
226 pressure datasets were time series, the biodiversity and pressure data points were plotted against
227 each other with year being disregarded. The gradient coefficient was recorded as a 'sensitivity score',
228 with the standard error of the gradient also recorded, for use in the next step. This sensitivity score
229 is representative of the sensitivity of that country's biodiversity, to the particular biodiversity pressure
230 (e.g., sensitivity of a country's biodiversity to one unit of pollution). The unit of sensitivity scores de-
231 pend on the pressure being modelled, and represent the % change in Biodiversity Intactness Index
232 (BII) for a unit increase in pressure.

233
234 Sensitivity scores were then compared between continents using a linear model. The weights
235 function of `lm()` was used to take account for the standard errors of the gradients from the first step,
236 and the inverse of the standard errors were inputted as weights (weights in `lm()` is inversely pro-
237 portional to variance). Important to note that this method only tests for differences between each
238 category, and the reference category, and therefore does not test differences between all groups.

240 **Climate Model**

241 Step 1 included obtaining sensitivity scores of each country's biodiversity to climate change. Linear
242 models were fit for each country, with average annual temperature ($^{\circ}C$) as the explanatory variable,
243 and biodiversity intactness index (%) as the response variable. The sensitivity score (gradient coef-
244 ficient) therefore had a unit of BII % / $^{\circ}C$ representing the % change in BII for each annual average
245 temperature increase of $1^{\circ}C$.

246 **Pollution Model**

247 Step 1 included obtaining sensitivity scores of each country's biodiversity to pollution (greenhouse gas
248 emissions). Linear models were fit for each country, with tonnes of CO2 equivalent emissions (000s)
249 as the explanatory variable, and biodiversity intactness index (%) as the response variable. The
250 sensitivity score (gradient coefficient) therefore had a unit of BII % / tonnes of CO2 equivalent (000s)
251 representing the % change in BII for each annual additional thousand tonnes of CO2 equivalent
252 emissions.

253 **Land Use Model**

254 Step 1 included obtaining sensitivity scores of each country's biodiversity to changes in built up
255 land area. Linear models were fit for each country, with area of built up land (square kilometres)
256 as the explanatory variable, and biodiversity intactness index (%) as the response variable. The
257 sensitivity score (gradient coefficient) therefore had a unit of BII % / square kilometre (of built up land)
258 representing the % change in BII for each built up land increase of 1 square kilometre.

259 **Invasive Species Model**

260 Step 1 included obtaining sensitivity scores of each country's biodiversity to invasive species. Linear
261 models were fit for each country, with number of new invasive species as the explanatory variable, and
262 biodiversity intactness index (%) as the response variable. The sensitivity score (gradient coefficient)
263 therefore had a unit of BII % / new invasive species representing the % change in BII for each new
264 invasive species.

Results

Climate Model

158 countries and 18 years matched between the climate and biodiversity datasets. Figure 1 shows the distribution of country-level climate sensitivity scores, and Figure 2 groups them by continent. The sensitivity scores of countries ($n = 158$) to climate change had a mean of -0.016 (SD 0.0037 , range: $-0.24 - 0.078$), so for each increase in average temperature of 1°C , the average country's BII decreases by 0.016 .

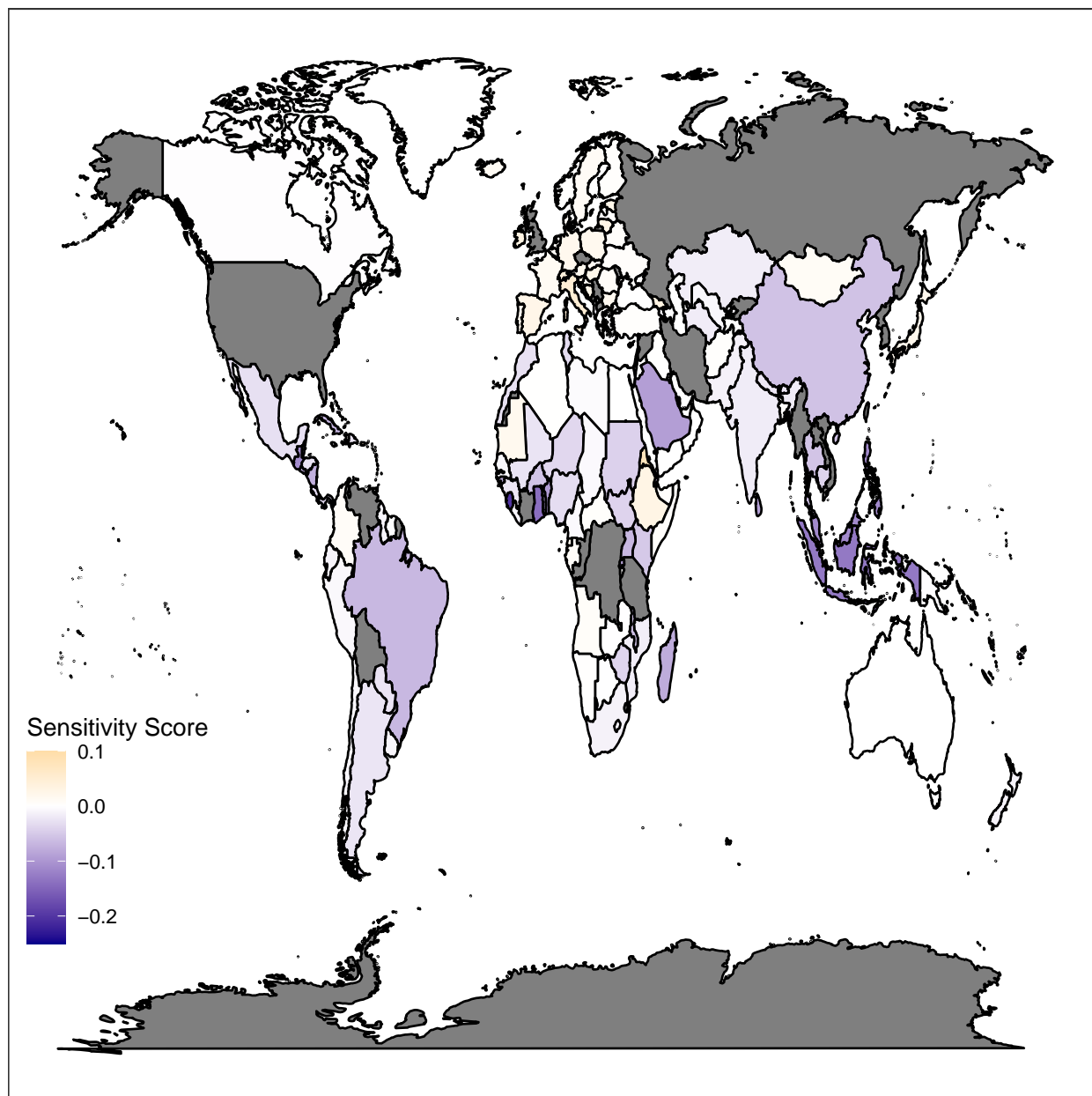


Figure 1: Biodiversity sensitivity to climate change: distribution of country sensitivity scores.



276

282 of countries within South America will decrease by 0.013 more than in other continents.

Table 1: Results from a linear model explaining effects of Continent on Climate Change Sensitivity Scores. N = 158.

Variable	β	SE	t	p
Intercept (reference: Africa)	< -0.001	< 0.001	0	1
Asia	< 0.001	< 0.001	0	1
Europe	0.0081	0.0031	2.61	0.0099
North America	< -0.001	< 0.001	-0.022	0.98
Oceania	-0.0028	0.0072	-0.39	0.70
S America	-0.013	0.0049	-2.59	0.01

Pollution Model

There were only 42 countries and 16 years that matched between the datasets (with no data for any African countries). One country with a standard error of 0 was removed due to the inability to calculate variance (weights function uses $1/SE$). Figure 3 shows the distribution of country-level pollution sensitivity scores, and Figure 4 groups them by continent. The sensitivity scores of countries ($n = 43$) to climate change had a mean of -0.0000051 (SD 0.0000073 , range: -0.0003 0.00007), so for each thousand tonne increase of CO₂ equivalent emissions, the average country's BII decreases by 0.0000051 .

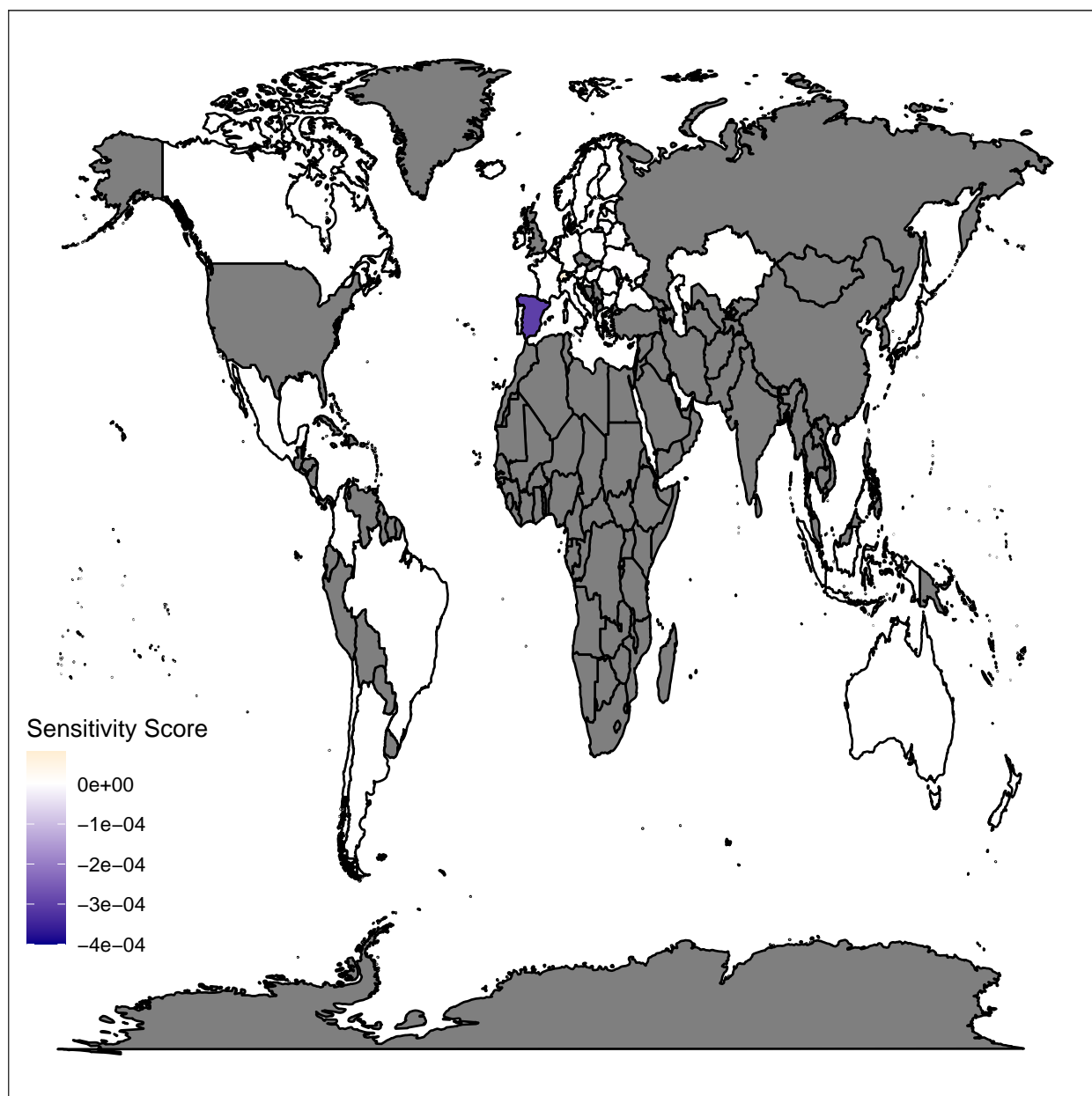


Figure 3: Biodiversity sensitivity to pollution.

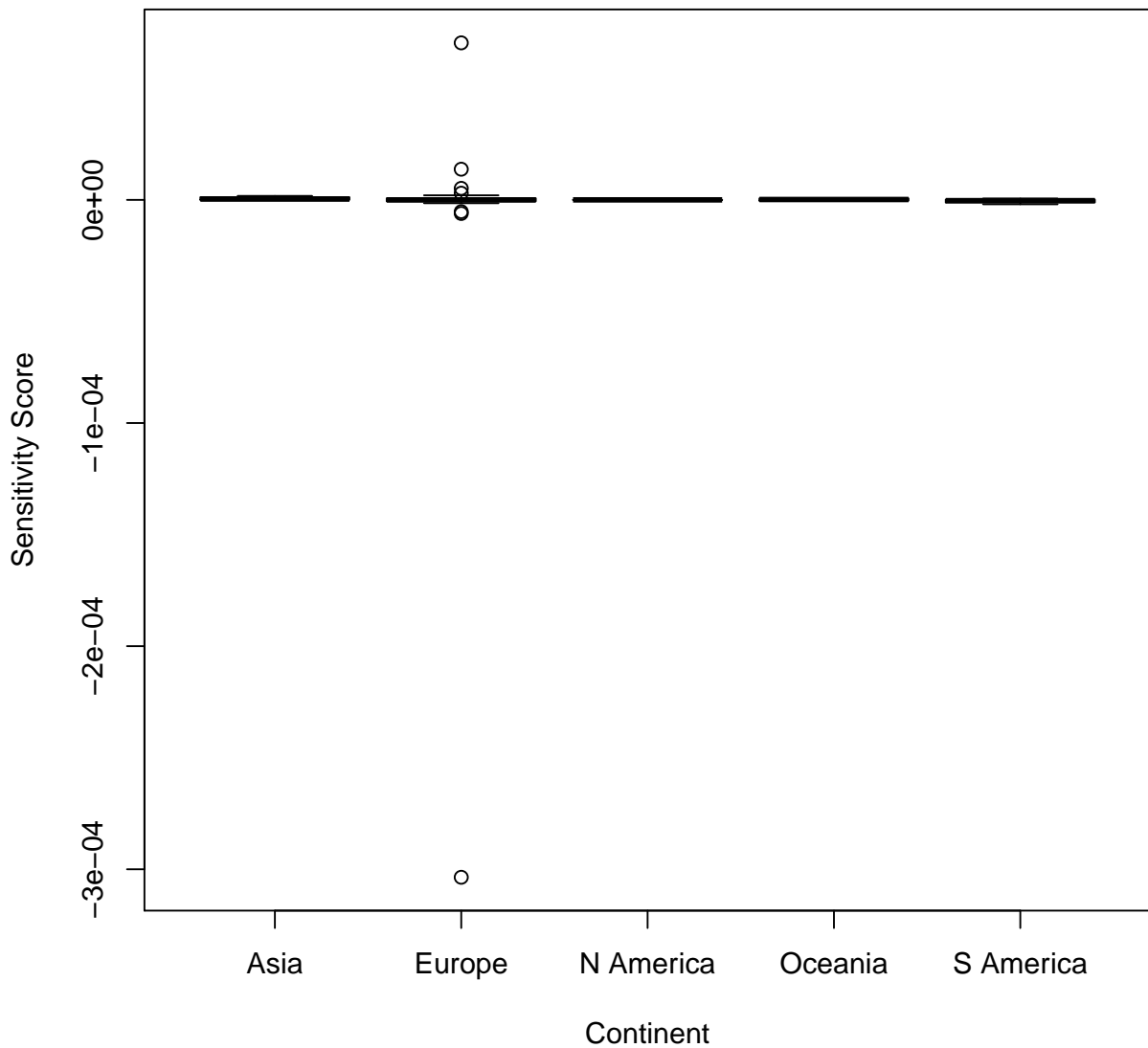


Figure 4: Biodiversity sensitivity to greenhouse gas emissions: country sensitivity scores grouped by continent.

The average sensitivity score for countries in Asia (the reference category) did not significantly differ from zero, and sensitivity scores in other continents were not significantly different from Asia's (Table 2).

Table 2: Results from a linear model explaining effects of Continent on Pollution Sensitivity Scores. N=42.

Variable	β	SE	t	p
Intercept (reference: Asia)	< 0.001	<0.001	0.69	0.49
Europe	< -0.001	<0.001	-0.68	0.50
North America	< -0.001	<0.001	-0.65	0.51
Oceania	< -0.001	<0.001	-0.2	0.84
S America	< -0.001	<0.001	-0.62	0.54

Land Use Model

167 countries and 3 years matched between datasets. Eight countries with a standard error of 0 were removed due to the inability to calculate variance (weights function uses $1/SE$). Figure 5 shows the distribution of country-level built up land sensitivity scores, and Figure 6 groups them by continent. The sensitivity scores of countries ($n = 175$) to built up land had a mean of 0.003 (SD 0.025, range: -3.222472 2.756498), so for each square kilometre increase in built up land area, the average country's BII increases by 0.003.

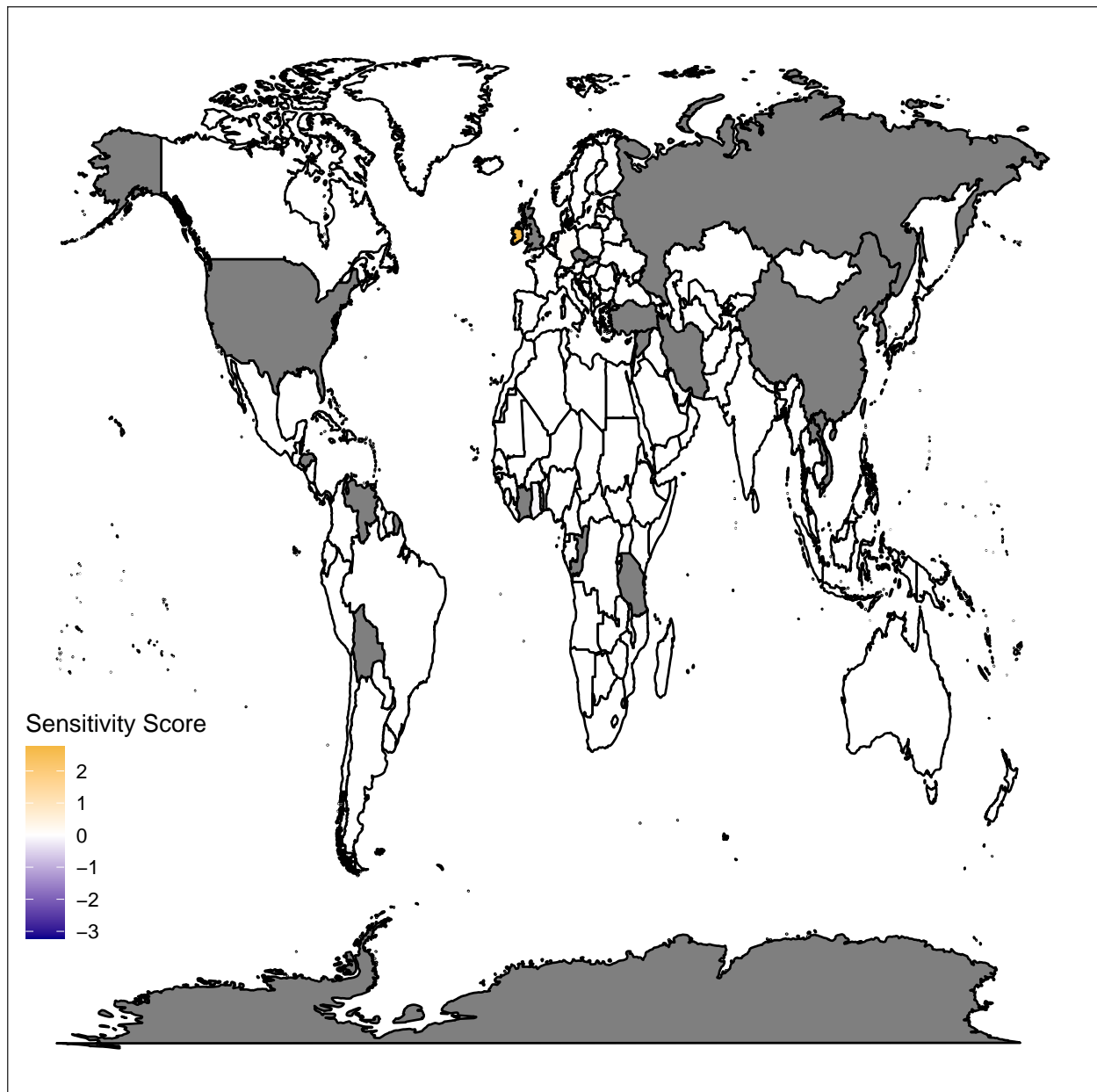


Figure 5: Biodiversity sensitivity to build up land.

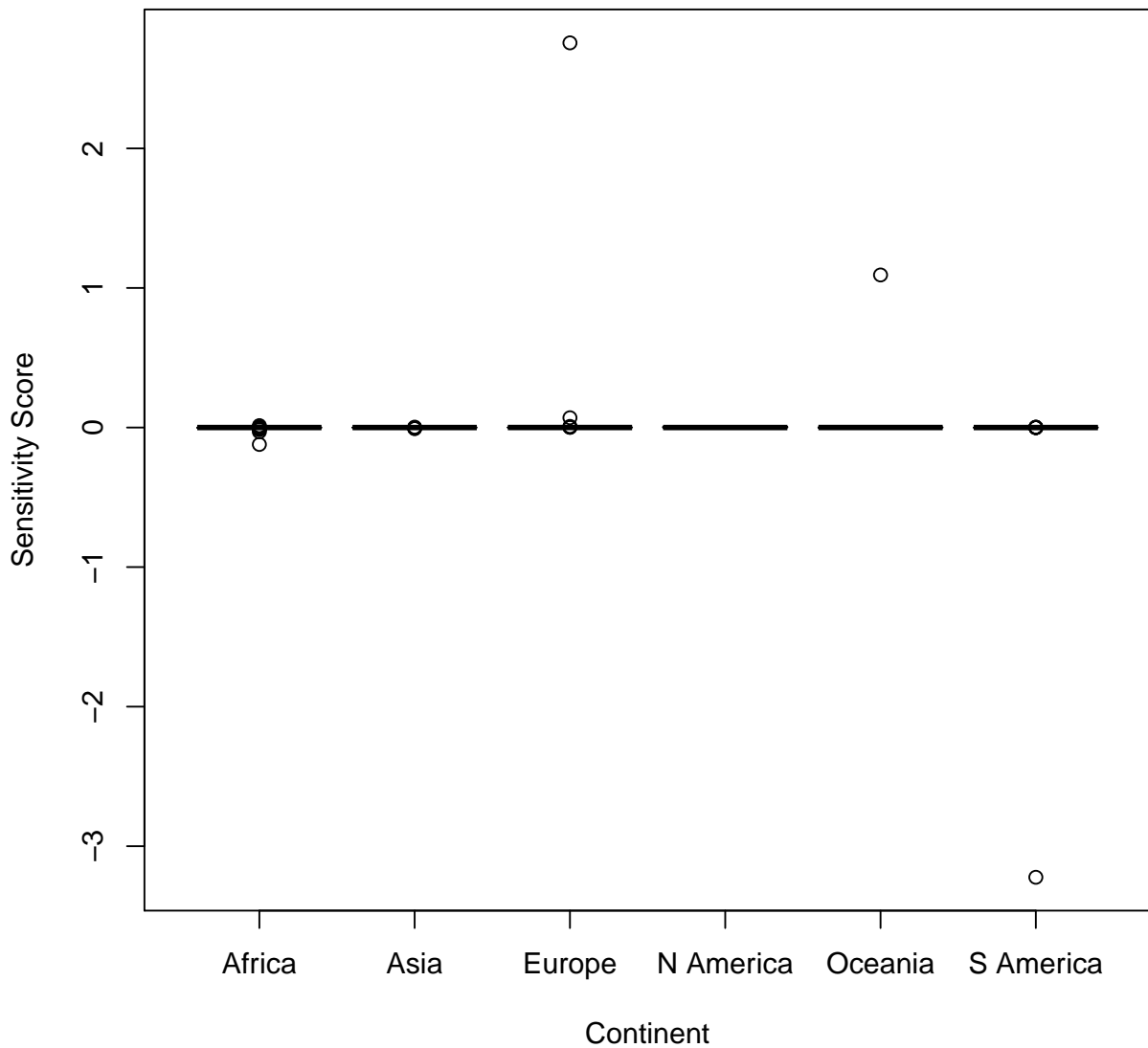


Figure 6: Biodiversity sensitivity to built up land use: country sensitivity scores grouped by continent.

The average sensitivity score for countries in Africa (the reference category) did not significantly different from zero, and sensitivity scores in other continents were not significantly different from Africa's (Table 3).

Table 3: Results from a linear model explaining effects of continent on land use sensitivity scores. N=167.

Variable	β	SE	t	p
Intercept (reference: Africa)	< -0.001	< 0.001	-0.52	0.61
Asia	< 0.001	< 0.001	0.19	0.85
Europe	< 0.001	< 0.001	0.45	0.65
North America	< 0.001	< 0.001	0.32	0.75
Oceania	< 0.001	< 0.001	0.02	0.98
S America	< -0.001	< 0.001	-0.17	0.87

Invasive Species Model

91 countries matched between datasets, with a different amount of years matching per country. Countries with a standard error of 0 were removed. Figure 7 shows the distribution of country-level invasive species sensitivity scores, and Figure 8 groups them by continent. The sensitivity scores of countries ($n = 91$) to invasive species had a mean of 0.0017 (SD 0.0012, range: -0.023 0.055), so for each additional new invasive species, the average country's BII increases by 0.0017.

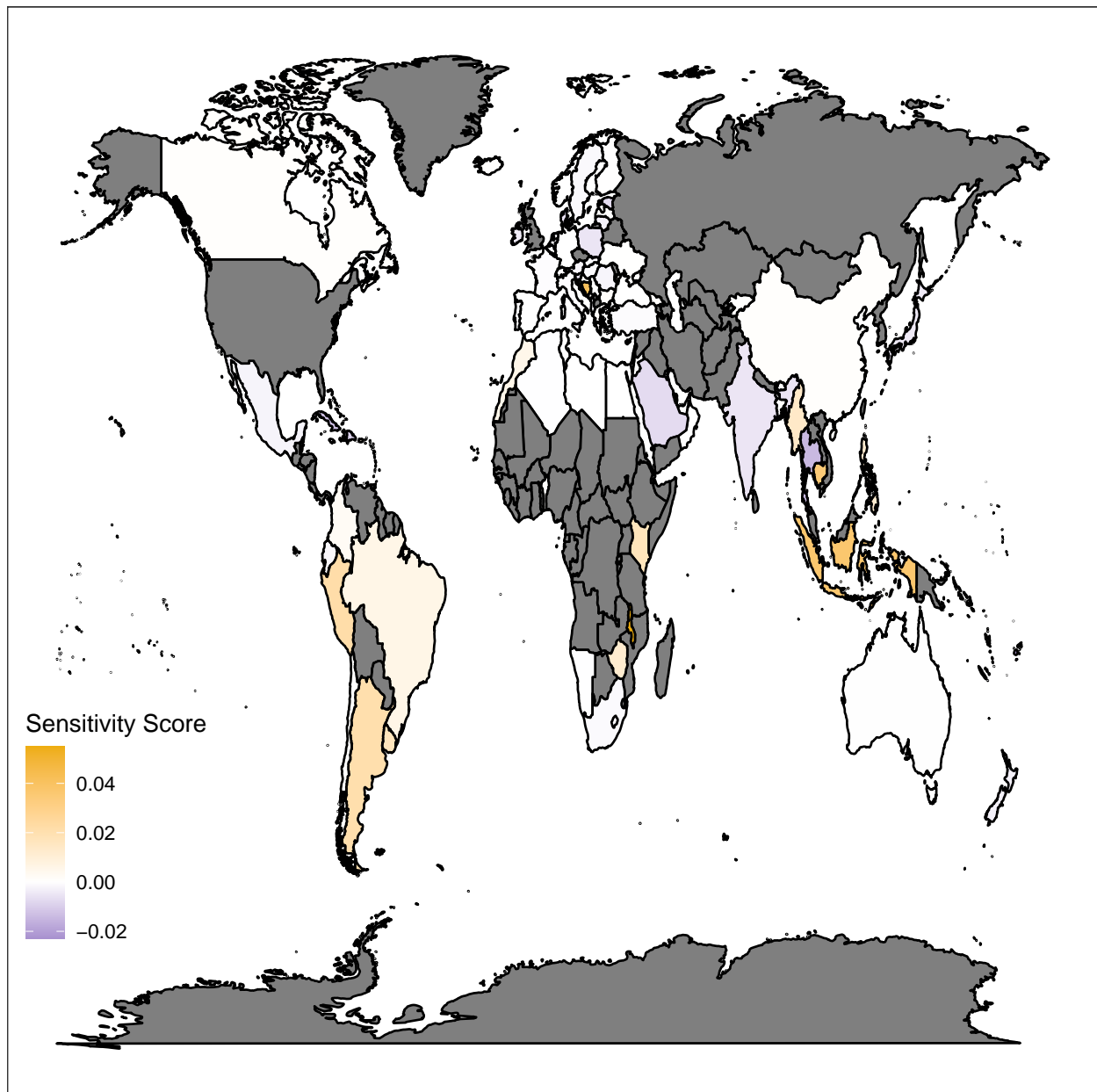


Figure 7: Biodiversity sensitivity to Invasive species.

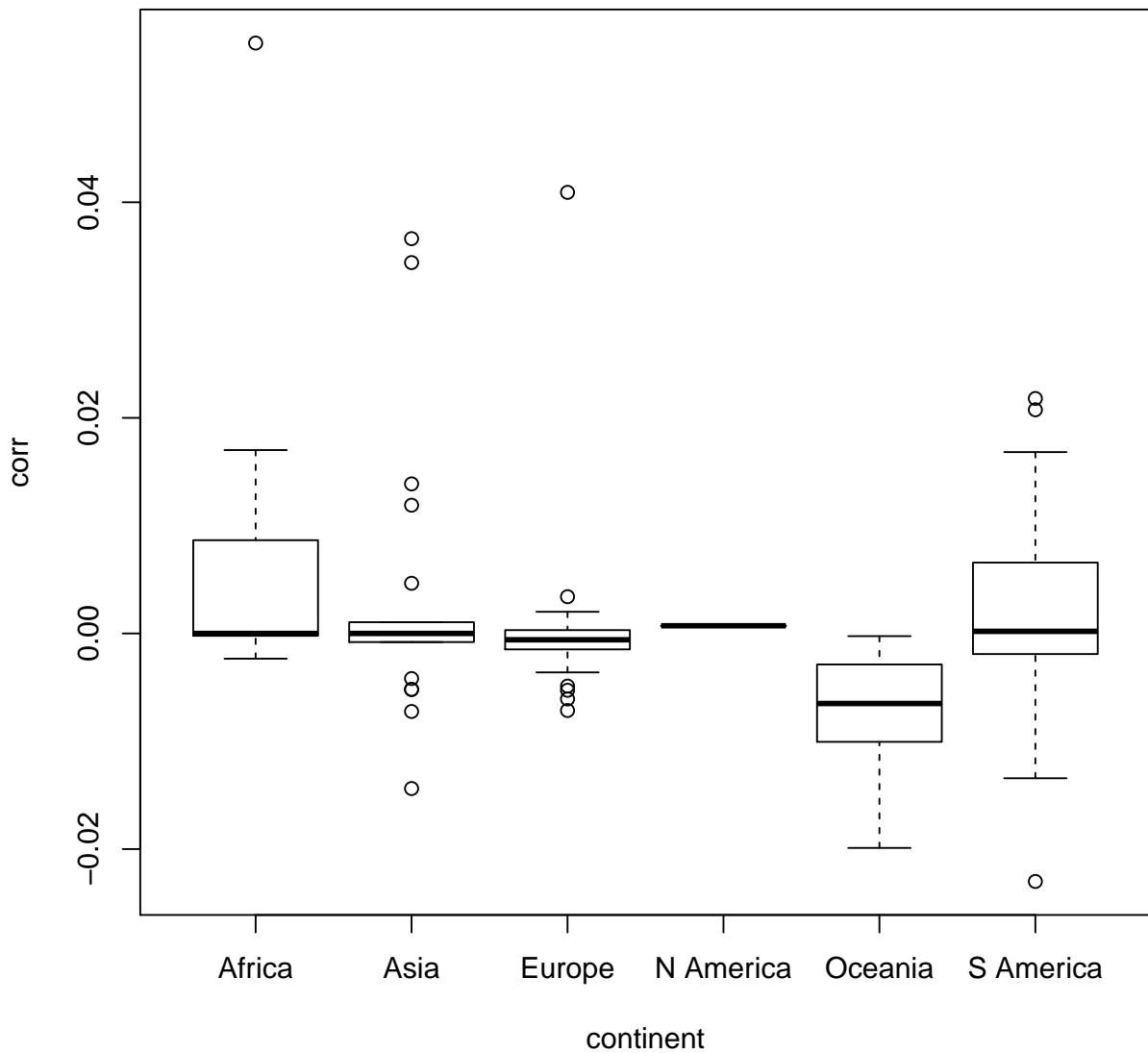


Figure 8: Biodiversity sensitivity to invasive species: country sensitivity scores grouped by continent.

The average sensitivity score for countries in Africa (the reference category) did not significantly differ from zero, and sensitivity scores in other continents were not significantly different from Africa's (Table 4).

Table 4: Results from a linear model explaining effects of continent on invasive species sensitivity scores. N=91.

Variable	β	SE	t	p
Intercept (reference: Africa)	< 0.001	< 0.001	0	1
Asia	< -0.001	< 0.001	0	1
Europe	< -0.001	< 0.001	-0.67	0.50
North America	< 0.001	0.0019	0.37	0.71
Oceania	< -0.001	< 0.001	-0.57	0.57
S America	< 0.001	< 0.001	0.63	0.53

324 Discussion

325 Importance of Findings

326 Whilst the conservation movement does a lot to help biodiversity ([Sandbrook et al., 2019](#)), another
327 response to the biodiversity crisis ([Ogar et al., 2020](#)) is the beginning of a global movement towards
328 sustainable business and biodiversity-conscious investment ([PRI, 2020](#); [Forum, 2020](#); [WWF, 2020](#)).
329 Creating a tool which assesses the overall biodiversity impact of a company can help guide in-
330 vestors, and is something that various parties are currently developing ([Alliance, 2022](#); [ICCS, 2020](#)).
331 'Benchmark for Nature' is a project which aims to use data science and to develop a framework for
332 assessing investment impacts on biodiversity by gathering information from online articles about how
333 much of an impact a company/sector is having on biodiversity pressures([ICCS, 2020](#)), and where in
334 the world these pressures are taking place.

335

336 **Conclusion**

337 optional section

338 **Data and Code Availability**

339 Data and Code Availability statement: At the end of your Main text, before the References section, you
340 must provide a statement titled “Data and Code Availability”, where you name a data (e.g., Dropbox,
341 FigShare, Zenodo, etc) and a code (e.g., Dropbox, GitHub, etc.) archive 20from where the data and
342 code can be obtained that will allow replication of your results. The code may be in the form of a
343 single script file.

Acknowledgements

References

- A, F., C, C., D, E., SM, C. F., T, K., L, M., M, M., M, P., P, P., M, S., F, S., and L, Z. (2019). Ghsl data package 2019. (KJ-1A-29788-EN-N (online),KJ-1A-29788-EN-C (print)).
- Alkemade, R., Van Oorschot, M., Miles, L., Nellemann, C., Bakkenes, M., and Ten Brink, B. (2009). Globio3: a framework to investigate options for reducing global terrestrial biodiversity loss. *Ecosystems*, 12(3):374–390.
- Alliance, W. B. (2022). Nature and biodiversity benchmark - world benchmarking alliance.
- Ament, J. M., Collen, B., Carbone, C., Mace, G. M., and Freeman, R. (2019). Compatibility between agendas for improving human development and wildlife conservation outside protected areas: Insights from 20 years of data. *People and Nature*, 1(3):305–316.
- Barlow, J., Lennox, G. D., Ferreira, J., Berenguer, E., Lees, A. C., Nally, R. M., Thomson, J. R., Ferraz, S. F. d. B., Louzada, J., Oliveira, V. H. F., et al. (2016). Anthropogenic disturbance in tropical forests can double biodiversity loss from deforestation. *Nature*, 535(7610):144–147.
- Bowler, D. E., Bjorkman, A. D., Dornelas, M., Myers-Smith, I. H., Navarro, L. M., Niamir, A., Supp, S. R., Waldock, C., Winter, M., Vellend, M., et al. (2020). Mapping human pressures on biodiversity across the planet uncovers anthropogenic threat complexes. *People and Nature*, 2(2):380–394.
- Cazalis, V., Barnes, M. D., Johnston, A., Watson, J. E., Şekercioğlu, C. H., and Rodrigues, A. S. (2021). Mismatch between bird species sensitivity and the protection of intact habitats across the americas. *Ecology Letters*, 24(11):2394–2405.
- Ceballos, G., Ehrlich, P. R., Barnosky, A. D., García, A., Pringle, R. M., and Palmer, T. M. (2015). Accelerated modern human-induced species losses: Entering the sixth mass extinction. *Science advances*, 1(5):e1400253.
- Collen, B., Ram, M., Zamin, T., and McRae, L. (2008). The tropical biodiversity data gap: addressing disparity in global monitoring. *Tropical Conservation Science*, 1(2):75–88.
- De Baan, L., Alkemade, R., and Koellner, T. (2013). Land use impacts on biodiversity in lca: a global approach. *The International Journal of Life Cycle Assessment*, 18(6):1216–1230.
- Díaz, S., Fargione, J., Chapin III, F. S., and Tilman, D. (2006). Biodiversity loss threatens human well-being. *PLoS biology*, 4(8):e277.
- Duffy, J. E., Godwin, C. M., and Cardinale, B. J. (2017). Biodiversity effects in the wild are common and as strong as key drivers of productivity. *Nature*, 549(7671):261–264.
- Ferrier, S., Ninan, K., Leadley, P., Alkemade, R., Acosta-Michlik, L., Akçakaya, H., and Kabubo-Mariara, J. (2016). Summary for policymakers of the assessment report of the methodological assessment of scenarios and models of biodiversity and ecosystem services.

378 Foden, W. B., Butchart, S. H., Stuart, S. N., Vié, J.-C., Akçakaya, H. R., Angulo, A., DeVantier,
 379 L. M., Gutsche, A., Turak, E., Cao, L., et al. (2013). Identifying the world's most climate change
 380 vulnerable species: a systematic trait-based assessment of all birds, amphibians and corals. *PloS*
 381 *one*, 8(6):e65427.

382 Forum, W. E. (2020). New nature economy series nature risk rising ... - world economic forum.

383 Gaston, K. J. (2000). Global patterns in biodiversity. *Nature*, 405(6783):220–227.

384 Gasu, M., Gasu, G., Ntemuse, U., et al. (2021). A review of biodiversity loss and climate change:
 385 Policy measures and adaptation strategies in nigeria. *Malaysian Journal of Tropical Geography*
 386 *(MJTG)*, 47(1 and 2):100–122.

387 Gibson, L., Lee, T. M., Koh, L. P., Brook, B. W., Gardner, T. A., Barlow, J., Peres, C. A., Bradshaw,
 388 C. J., Laurance, W. F., Lovejoy, T. E., et al. (2011). Primary forests are irreplaceable for sustaining
 389 tropical biodiversity. *Nature*, 478(7369):378–381.

390 Heather Phillips, Adriana De Palma, R. E. G. S. C. e. a. (2021). The biodiversity intactness index
 391 - country, region and global-level summaries for the year 1970 to 2050 under various scenarios.
 392 Technical report, National History Museum.

393 ICCS (2020). Benchmark for nature.

394 Leemans, R. and De Groot, R. (2003). Millennium ecosystem assessment: Ecosystems and human
 395 well-being: a framework for assessment.

396 Louette, G., Maes, D., Alkemade, J. R. M., Boitani, L., de Knecht, B., Eggers, J., Falcucci, A., Fram-
 397 stad, E., Hagemeljer, W., Hennekens, S. M., et al. (2010). Bioscore–cost-effective assessment
 398 of policy impact on biodiversity using species sensitivity scores. *Journal for Nature Conservation*,
 399 18(2):142–148.

400 McRae, L., Deinet, S., and Freeman, R. (2017). The diversity-weighted living planet index: controlling
 401 for taxonomic bias in a global biodiversity indicator. *PloS one*, 12(1):e0169156.

402 Millennium ecosystem assessment, M. (2005). *Ecosystems and human well-being*, volume 5. Island
 403 press Washington, DC.

404 Newbold, T., Oppenheimer, P., Etard, A., and Williams, J. J. (2020). Tropical and mediterranean biodi-
 405 versity is disproportionately sensitive to land-use and climate change. *Nature Ecology & Evolution*,
 406 4(12):1630–1638.

407 Nobel, A., Lizin, S., Brouwer, R., Stern, D., B Bruns, S., Malina, R., et al. (2020). Are anthropogenic
 408 pressures on biodiversity valued differently than natural ones? a meta-analysis of the non-use
 409 valuation literature. In *2020 Conference (64th), February 12-14, 2020, Perth, Western Australia*,
 410 number 305246. Australian Agricultural and Resource Economics Society.

411 Ogar, E., Pecl, G., and Mustonen, T. (2020). Science must embrace traditional and indigenous
 412 knowledge to solve our biodiversity crisis. *One Earth*, 3(2):162–165.

413 Pereira, H. M., Navarro, L. M., and Martins, I. S. (2012). Global biodiversity change: the bad, the
 414 good, and the unknown. *Annual Review of Environment and Resources*, 37:25–50.

415 Plumptre, A. J., Baisero, D., Belote, R. T., Vázquez-Domínguez, E., Faurby, S., Jdrzejewski, W., Kiara,
 416 H., Kühn, H., Benítez-López, A., Luna-Aranguré, C., et al. (2021). Where might we find ecologically
 417 intact communities? *Frontiers in Forests and Global Change*, 4:26.

418 PRI (2020). Investor action on biodiversity.

419 Purvis, A., Newbold, T., De Palma, A., Contu, S., Hill, S. L., Sanchez-Ortiz, K., Phillips, H. R., Hudson,
 420 L. N., Lysenko, I., Börger, L., et al. (2018). Modelling and projecting the response of local terrestrial
 421 biodiversity worldwide to land use and related pressures: the predicts project. In *Advances in*
 422 *ecological research*, volume 58, pages 201–241. Elsevier.

423 Ricklefs, R. E. (2004). A comprehensive framework for global patterns in biodiversity. *Ecology letters*,
 424 7(1):1–15.

425 Sala, O. E., Stuart Chapin, F., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-Sanwald,
 426 E., Huenneke, L. F., Jackson, R. B., Kinzig, A., et al. (2000). Global biodiversity scenarios for the
 427 year 2100. *science*, 287(5459):1770–1774.

428 Sandbrook, C., Fisher, J. A., Holmes, G., Luque-Lora, R., and Keane, A. (2019). The global conser-
 429 vation movement is diverse but not divided. *Nature Sustainability*, 2(4):316–323.

430 Shin, Y.-J., Midgley, G. F., Archer, E. R., Arneeth, A., Barnes, D. K., Chan, L., Hashimoto, S., Hoegh-
 431 Guldberg, O., Insarov, G., Leadley, P., et al. (2022). Actions to halt biodiversity loss generally
 432 benefit the climate. *Global change biology*.

433 Steffen, W., Richardson, K., Rockström, J., Cornell, S. E., Fetzer, I., Bennett, E. M., Biggs, R.,
 434 Carpenter, S. R., De Vries, W., De Wit, C. A., et al. (2015). Planetary boundaries: Guiding human
 435 development on a changing planet. *science*, 347(6223):1259855.

436 Sunday, J. M., Pecl, G. T., Frusher, S., Hobday, A. J., Hill, N., Holbrook, N. J., Edgar, G. J., Stuart-
 437 Smith, R., Barrett, N., Wernberg, T., et al. (2015). Species traits and climate velocity explain
 438 geographic range shifts in an ocean-warming hotspot. *Ecology letters*, 18(9):944–953.

439 van Goethem, T. and van Zanden, J. L. (2021). Biodiversity trends in a historical perspective.

440 Velde, M. (2022). How wildlife trends correlate with economic wealth and human population density.
 441 Master's thesis, Imperial College London. unpublished thesis.

442 Wade, T. G., Riitters, K. H., Wickham, J. D., and Jones, K. B. (2003). Distribution and causes of
 443 global forest fragmentation. *Conservation Ecology*, 7(2).

444 Watson, R., Baste, I., Larigauderie, A., Leadley, P., Pascual, U., Baptiste, B., Demissew, S., Dziba,
 445 L., Erpul, G., Fazel, A., et al. (2019). *Summary for policymakers of the global assessment re-*
 446 *port on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on*
 447 *Biodiversity and Ecosystem Services*. Bonn, Germany: IPBES Secretariat.

448 WWF (2020). Nature is too big to fail.