Title: Disconnects between underlying ecological theory and data in phenological mismatch research 5 6 Authors: Heather M. Kharouba¹, Elizabeth M. Wolkovich^{2,3,4} Affiliations: ¹Department of Biology, University of Ottawa, Canada K1N 9B4 ²Organismic and Evolutionary Biology, Harvard University, Cambridge, MA 02138 ³Arnold Arboretum of Harvard University, Boston, MA, 02130 ⁴Forest and Conservation Sciences, Faculty of Forestry, University of British Columbia, Vancouver, Canada V6T 1Z4 Email addresses: heather.kharouba@uottawa.ca; e.wolkovich@ubc.ca **Correspondence:** Heather M. Kharouba, 30 Marie-Curie, University of Ottawa, Ottawa, ON, Canada, K1N 9B4, 613-562-5800 x6740; heather.kharouba@uottawa.ca Running title: Predicting phenological mismatch under climate change **Keywords:** global warming, phenological synchrony, trophic interactions, adaptive, performance Statement of authorship: All authors conceived of the ideas, contributed to the writing and editing of the manuscript. HMK collected and analyzed the data. Data accessibility statement: Should the manuscript be accepted, the data supporting the results will be archived in an appropriate public repository (Dryad or Figshare) and the data DOI will be included at the end of the article.

Abstract

Many researchers hypothesize that climate change will lead to phenological mismatches—where the timing of critical life history events between interacting species becomes desynchronized with potential negative consequences for species and communities. Yet, evidence documenting negative impacts on fitness is mixed. The most common ecological theory that underlies these studies (Cushing match-mismatch) offers a testable hypothesis that—when combined with a system's pre-climate change baseline—can predict the consequences of desynchronized phenology due to climate change. However, currently no study has collected the data required to provide strong tests of this major hypothesis, making it difficult to assess support for it. Further, 74% of studies fail to define a pre-climate change baseline. We highlight how improved approaches—including experiments that clearly link timing to fitness and test extremes, integration across approaches, and null models—could rapidly advance our mechanistic understanding and thus allow robust predictions of shifts with continuing climate change.

Introduction

Climate change is causing phenological shifts—changes in the timing of life history events—that vary across species as well as between functional groups and trophic levels^{1,2,3}. Such species-specific variation in response to climate change has led to changes in the relative timing of key activities among interacting species⁴ (phenological synchrony). These changes can have negative fitness consequences^{5,6,7}— termed 'phenological mismatch' (Box 1)—and influence ecosystem-level properties in some contexts⁸. However, changes in phenological synchrony do not always influence fitness^{9,10} or have consistent demographic effects¹¹. Recent theoretical ^{12,13,14} and empirical studies ¹⁵ based in single systems, as well as syntheses^{3,4} have worked to improve predictions and about when and why mismatches are likely to occur. Yet—while there is general agreement that predicting phenological mismatch is critical for

determining how pair-wise species interactions, communities, and ecosystem function (e.g. pollination) will be affected by climate change—we still have no framework to discriminate between systems where these mismatches are likely to occur or not.

Here, we argue that much of the difficulty in predicting the consequences of climate change-driven shifts in synchrony is due to a disconnect between ecological theory and current empirical approaches. We focus on the widely-cited Cushing match-mismatch, or trophic mismatch, hypothesis 16,17,18,19, the most commonly applied hypothesis concerning consumer-resource interactions in this literature. We show how advances could come from direct, strong tests of the hypothesis— clearly tested assumptions, tests of multiple potential mechanisms, and clearly defined baselines, when possible. Our aim is not to put forward additional hypotheses about the context in which phenological mismatch will occur, which has been reviewed extensively elsewhere 20,21, but rather to help guide the study of phenological mismatch by outlining a path forward to develop robust climate change predictions that can scale up to inference across sites and systems.

We provide an overview of the Cushing hypothesis and outline the divide between the hypothesis and the empirical studies using a systematic literature review of phenological mismatch. We then discuss how current approaches are impeding greater progress in the field, and how changes to research methodologies could rapidly advance our understanding and help forecast the impacts of climate change on ecological communities—the ultimate goal of most of the phenological mismatch literature.

Overview of the Cushing hypothesis

The Cushing match-mismatch hypothesis ^{16,17,18,19} (Fig. 1) is the most common ecological theory underlying phenological mismatch studies ^{21,22}. While the general hypothesis has been applied across many ecosystems ^{9,23,24}, the theory originally emerged from the marine fisheries literature as a way to explain the variation in population recruitment of fish stocks. The

hypothesis postulates that the consumer should temporally 'match' the peak of its most energetically demanding period with the peak of resource availability (Fig. 1b) and any change to the relative timing of the interaction will result in a 'mismatch' (Fig. 1ac).

We follow others^{13,22,25} and assume that Cushing's¹⁹ use of the term 'mismatch' is focused on the 'match' having the highest fitness (i.e., consumer fitness is maximized at phenological synchrony). Therefore, if there is any change to the relative timing of the interaction, there will be a decrease in fitness associated with this period for the consumer (i.e., a mismatch), thus producing a concave-down curve (Fig. 1d). For example, if a consumer (e.g. adult bird) reproduces too early in the spring (Fig. 1a) and the resource (e.g. caterpillar) has not yet emerged, the offspring will risk starvation. At the curve's limits, this component of consumer fitness should fall to zero when the change in relative timing is sufficiently large (Fig. 1d). For example, the early offspring (Fig. 1a) can only survive so many days without food.

This hypothesis is based on two important assumptions. First, the resource is the major controller on consumer fitness²⁶ (i.e., strong interaction strength, bottom-up control). Second, the consumer and resource must present a certain degree of seasonality (e.g., resource is only available for part of the year), limiting the optimal period for growth and reproduction for the consumer^{25, 27}. If neither of these assumptions is met, then fitness consequences due to changes in the relative timing of the interaction will be either weak or non-existent.

We make two clarifications related to the way we apply the hypothesis here. First, we discuss the hypothesis as considering total fitness, either lifetime or population mean. While other perspectives have been taken²², we follow the first assumption and consider a narrow definition of fitness whereby the resource must be the dominant controller of consumer fitness, and systems where fitness is predominantly controlled by other factors would not be included^{13,28}. Second, we follow Cushing's definition of mismatch and its connections with fitness, which is in line with other studies^{13, 25, 29}.

Finally, although the Cushing hypothesis has been applied to other types of interactions (e.g. mutualism⁸⁴), we limit our discussion to antagonistic interactions between consumers and their food resources.

Disconnect between theory and empirical studies

The Cushing hypothesis offers testable predictions of the magnitude and direction of population-level fitness changes in response to climate-change driven shifts in synchrony. In its original state, the hypothesis has been debated, contested and criticized, particularly in the marine literature^{25,31}. Although a relatively simple hypothesis, it is inherently difficult to test in the field, an assertion even Cushing himself made. Indeed, the shape and strength of the relationship of the curve varies greatly across observational studies^{6,11,32,33}. We argue that this is because there are two key areas that make it difficult to determine whether this hypothesis is widely supported in the context of climate change. First, studies are not collecting the data required to test assumptions and multiple ultimate mechanisms (i.e., rigorously testing the hypothesis) (Fig. 2). This represents the major challenge of the hypothesis and—we argue—may be why support for it has been so mixed. Second, studies are not defining a pre-climate change baseline in synchrony, an action needed to apply the hypothesis to predicting the consequences of climate change for interacting species.

To put our argument in context, we systematically reviewed the phenological mismatch literature. We examined 43 observational studies that met our search criteria (see Appendix for details) for evaluating the effects of a change in relative timing on the fitness or performance (e.g., growth, development; hereafter referred to as performance) of the consumer.

i) Testing the Cushing hypothesis

Many mechanisms can produce the Cushing curve, and they vary across systems, space and time. Here we focus on the ultimate mechanisms related to life-history theory and food web

theory. Mechanisms predicted by life-history theory (Box 1), for example, trade-offs between fecundity and mortality, and breeding strategies (Fig. 2), can produce the Cushing curve—as consumers maximize their fitness through ideal timing with the peak of their primary resource. For example, a species' breeding strategy (e.g., whether a consumer provisions its offspring with resources acquired prior to reproduction or during the breeding period) influences how well-timed a consumer will be with its resource³⁴. Alternatively, trade-offs can also move the fitness optimum away from the resource peak and/or lead to a different asymmetric shape of the Cushing curve (see discussion about 'adaptive mismatch hypothesis^{13,28} below. The Cushing hypothesis, however, is also connected to food web theory (Box 1), which focuses on predator prey dynamics and is related to the timing of a consumer to its food resource.

Identifying the ultimate mechanism(s) relates to the likelihood that the key assumptions of the Cushing hypothesis are met¹³ (Fig. 2). For example, if mechanisms such as density dependence or top-down population regulation are thought to be at play, then the first assumption of the hypothesis is unlikely to be met^{13,35}. Even when both of the assumptions of the Cushing hypothesis have been met (i.e., resource fitness is the major controller of consumer fitness and both species show seasonality), identifying the ultimate mechanism is critical to predicting how interactions will respond to climate change (discussed in more detail in the next section; Fig. 2). Given the number of mechanisms related to the Cushing hypothesis and therefore the potential diversity in the outcomes in consumer performance (e.g., magnitude, direction) due to shifts in synchrony, the better our mechanistic understanding of these patterns, the more robust our predictions will be (e.g., steepness of the Cushing curve).

Studies deviate in which forces they hypothesize control the peak in the food resource (a key assumption, see Fig. 1 and 2). For example, in aquatic systems—where top-down forces are generally more influential compared to terrestrial systems³⁶—many studies suggest that the resource peak is controlled by release from, or predation by, a consumer^{37,38,39}. This is a very different hypothesis from others that suggest seasonality in the environment produces the

resource peak⁴⁰. In terrestrial systems, the relative contribution of direct vs. indirect effects on consumer dynamics remains poorly understood⁴¹ and these two hypotheses can affect whether the first assumption of the Cushing hypothesis is met.

a) Data requirements

Given that several mechanisms can produce the basic shape of the Cushing curve, different researchers may collect very different data depending on what mechanism they suspect underlies the curve in their particular system or what their main objective is. To accomplish a strong test of the Cushing hypothesis, equivalent data would ideally be assessed for both the consumer and resource—at the level of the individual when possible. Researchers studying the hypothesis from the lens of life-history theory require per-capita estimates of fitness (which underlie the necessary population-level metric of fitness that the hypothesis is based on), including measurements of fecundity and mortality, and how they vary across ontogeny. Such measurements allow researchers to assess fitness consequences at the level of the individual to link any fitness changes to timing.

Studies addressing the Cushing hypothesis from the perspective of food web theory require equivalent data on the consumer and resource but do not require per-capita fitness consequences for the consumer. Equivalent data for the consumer and resource can help assess the influence of the resource on the consumer (i.e. the first assumption). Without at least population-level data on the resource, it is difficult to test the second assumption of the hypothesis; for example, resource abundance can be used to determine the likelihood that consumers will encounter enough resource at the tails of their temporal distributions, thus defining the degree of seasonality in the resource (Fig. 1).

Per-capita data on the resource, when available, also extends our mechanistic understanding beyond the consumer, thus improving our ability to make predictions of climate change-driven phenological mismatches for the consumer-resource interaction in its entirety.

For example, fine-scale data on the resource are needed to fully incorporate feedbacks between the consumer and resource^{35,42}, measure interaction strength (which informs a key assumption of the hypothesis²⁰), and accurately predict changes in resource abundance due to climate change. Together, these data help to determine the wider implications of phenological mismatches for the community^{43,44}.

b) Current state of the literature

Currently such data are non-existent in this literature (i.e. none of the pair-wise interactions we found had individual-level data for both the consumer and resource). Of the interactions we examined with per-capita performance data on the consumer (i.e. life history; n=28 interactions), none had equivalent measures of performance on the resource (Table S1). For the majority of these interactions (18/28), the researchers measured performance at the community level for the resource (e.g. biomass; Table S1) and many had no measure at all for the resource (6/28 interactions; Table S1). Researchers that measured population- or community-level data on the consumer (n=18 interactions) collected equivalent data on the resource and consumer (e.g. abundance). For the majority of these interactions, researchers also collected population-level (7/18) or community level data for the resource (7/18) (Table S1). Across all interactions, researchers measured per-capita performance on the resource for only one interaction³² (Table S1).

Part of these discrepancies may stem from differing generation times—studies that can easily observe a resource curve shaped by predation often focus on organisms with generation times on the scale of days to weeks. For example, aquatic studies that focus on phytoplankton as the resource may observe many generations in one summer while a terrestrial study focused on caterpillars in temperate areas would generally observe a smaller number generations (e.g. 1-2, depending on latitude and species for caterpillars). When populations turn over too quickly to track individuals, researchers struggle to collect robust data on per capita fitness. The

fundamental difference in the scale of generation times pervades aquatic/terrestrial comparisons, particularly at lower trophic levels ^{36,45,46}, and may drive the difference in the objectives of aquatic vs. terrestrial studies in the phenological mismatch literature. Indeed, the majority of researchers studying aquatic-based interactions in our review approached the Cushing hypothesis from the perspective of populations or communities (15/20 interactions), whereas researchers approached terrestrial-based interactions from life-history theory (23/26; Table S2).

A further divide across studies may come from the size, and population size, of the consumer, properties correlated with generation time: it can be difficult to collect per-capita fitness data when consumers are too small and numerous. In cases where individuals can be distinguished, marked and monitored, it is possible to gather per-capita data; otherwise data are more likely to be aggregated at the population or community-level. In these cases, we may be unlikely to overcome the discrepancy between studies conducted at the individual vs. population level. Of the life history interactions, the majority of consumers were birds (19/28) and mammals (6/25; Table S2). Comparatively, invertebrates were measured at the population or community level and the vast majority were aquatic (11/12; Table S2). Therefore, only certain classes of organisms are tested for each category of mechanisms and in one type of biome, further limiting generalizations across systems.

This places phenological mismatch studies on a continuum: at one end, aquatic studies focus generally on food web mechanisms, collect equivalent data on the consumer and resource but no per-capita fitness data and at the other end, terrestrial studies focus on life history theory from the lens of the consumer, collect individual-level data that can scale up to understand population-level connections between timing and fitness of the consumer, with much less information on the resource. The fundamental problem with these approaches is that researchers test only one piece of the much larger field of mechanisms that could underlie the Cushing curve. Though it is highly possible that both food web and life history theory together

explain many of the consumer-resource systems studied, data limitations make it hard to assess mechanisms related to both theories at once. Consequently, most studies do not actually provide strong tests of the Cushing hypothesis, making it difficult to refute the hypothesis if no evidence is found, a prevalent issue in ecology and evolution⁴⁷.

Without strong support for the Cushing hypothesis (i.e., relative timing explains substantial variation in consumer fitness and clear patterns that align with predictions), the mechanisms underlying it will likely be uncertain. This uncertainty can be caused by two reasons: (1) hypotheses that are not mutually exclusive (i.e. different mechanisms yield the same predictions). For example, the relative timing of an interaction will change in the same direction regardless of whether temperature (or some other abiotic factor) directly or indirectly affects a resource's peak; or (2) considering multiple mechanisms yields different predictions than considering only a single mechanism. For example, researchers have pointed out how a broader perspective on life history, such as including other sources of mortality beyond those related to resource access or alternative resource types¹⁵, can lead to alternative predictions than the synchrony predicted by the Cushing curve^{28,48}.

ii) Testing pre-climate change conditions

Applying the Cushing hypothesis to understanding—and ideally predicting—the consequences of climate change for interacting species requires more than testing the assumptions and underlying mechanisms of the hypothesis (Fig. 2). Researchers still need an understanding of the system dynamics before climate change and the phenological cues that affect the timing of consumer and resource levels to predict what shifts in the climate have—and will—do to the timing and fitness of the players (Fig. 2). Below we discuss these steps in more detail.

(a) Identifying pre-climate change baselines

The concept of a pre-climate change baseline (i.e., defining the range of natural variation) is rarely, if ever, considered by phenological mismatch studies; yet, establishing a baseline in a system has important implications for the Cushing hypothesis. Of the studies we considered, only 26% (11/43) of the studies began before the early 1980s, the same time that pronounced recent climate change began⁴⁹. Even in this subset of studies, not all studies had more than three years of data (2/11 had only one or two years of data for inference). This is also an issue with the phenological synchrony literature⁴. Even those studies with time-series that extend into decades before the 1980s do not explicitly define a pre-climate change baseline in synchrony (but see ref: 23,32,50).

(b) Implications of not identifying pre-climate change baselines

There are several major implications of not establishing a pre-climate change baseline (Fig. 3). First, studies without a baseline are implicitly assuming conditions of stationarity (i.e., a constant underlying probability distribution) before climate change occurred⁵¹ (Fig. 3a). This is problematic when climate change has led to non-stationarity to different extents in different systems (Fig. 3a). Without a defined pre-climate change baseline in the system, we do not know whether, and to what magnitude, we should predict a divergence in synchrony due to climate change (Fig. 3a). It is then further difficult to determine how much of the observed change in the interaction can be attributed to climate change or some other driver, for example nitrogen deposition⁵², land use change⁵³, or species introductions^{54,55,56}. Thus, knowledge of the preclimate change baseline may rule out climate change and identify other drivers, which is key for conservation management decisions for threatened species.

Second, these studies confound a pre-climate change baseline with the peak of the Cushing curve: that is, they assume that fitness was highest before climate change began because the consumer's timing was well matched to the resource's (i.e. the synchrony baseline⁴⁸: Box 1; Fig. 3b). The resulting prediction is that climate change will necessarily

change the relative timing of the interaction (i.e. non-stationarity) leading to a decline in the consumer's fitness (Fig. 2, Fig. 3b). However, without knowledge of the pre-climate change baseline, it is difficult to predict the direction and magnitude of fitness consequences due to climate change-driven shifts in synchrony, especially when alternative conditions could be the baseline (Fig. 3b,c).

An alternative hypothesis put forward by Visser et al.⁵⁷ suggests that in some contexts, 'adaptive mismatch' (i.e. when the most energetically demanding period of the consumer is not lined up with the peak resource availability for many individuals in the population; Box 1) can occur. They postulate that in some systems, asynchrony might be adaptive due to life-history tradeoffs, where lifetime fitness is maximized but not necessarily individual components of fitness^{48,57} or due to a combination of life-history trade-offs and competitive advantage from intra-specific competition^{58,59,60}. However, whether synchrony vs. asynchrony is considered to be adaptive for the consumer, and thus whether fitness decreases are predicted with climate change (Fig. 3b), will depend on which fitness components have been measured in the study (e.g. reproductive success vs. lifetime fitness) and whether asynchrony is adaptive at the individual or population-level^{22,48,57,60}. For example, fitness may increase or decrease with climate change in bird species where there is a competitive advantage of hatching early, depending on whether short-term or long-term populations responses are considered⁶⁰.

In either hypothesis (i.e. synchrony vs. adaptive mismatch; Fig. 3b), a pre-climate change baseline of asynchrony (Box 1; Fig. 3c) could occur through several mechanisms. Trade-offs, and asynchrony as a potential outcome, can emerge when a consumer cannot measure the timing of its resource well⁴⁸. Shifts in the drivers of phenology for the consumer or resource (e.g., nutrient enrichment) could have pushed the system away from synchrony before climate change. An asynchronous baseline would also be predicted—at times—by a co-evolutionary arms race⁶¹: when the resource is ahead in the arms-race, asynchrony would the predicted baseline (Fig. 3c), this would then shift back towards synchrony when the consumer is winning

the arms race. In systems where asynchrony or some type of transient conditions might be the baseline state, our ability to anticipate the effects of climate change will be even more challenging as these transient conditions could lead to vastly different predictions under climate change (Fig. 3c). For example, interactions might become less (i.e., fitness increases) or even more mismatched (i.e. fitness decreases) with climate change (Fig. 3b-c).

The pre-climate change baseline is also dependent on the ultimate mechanism(s) generating the Cushing curve (Fig. 2). For example, a species that provisions its offspring with resources acquired during the breeding period (i.e., an income breeder; e.g., West Greenland caribou: *Rangifer tarandus*) is predicted to have a pre-climate change baseline of synchrony, whereas a species that provisions its offspring with resources gained prior to reproduction (i.e., a capital breeder; e.g., muskoxen: *Ovibos moschatus*) is more likely to have a baseline of asynchrony³⁴. Therefore, researchers need an understanding of the system dynamics before climate change began.

(c) Identifying proximate cues

For climate change to lead to phenological mismatch, the underlying assumption is that changes in climate will drive changes in the relative timing of species interactions. Therefore, predicting how climate change will affect the relative timing of a consumer and resource, and potentially the fitness of the consumer, will also depend on the identification of the proximate phenological cues of the consumer and resource (Fig. 2). For example, climate change is likely to lead to fitness declines for a consumer that has a pre-climate change baseline of synchrony and that shares different environmental cues with its resource (Fig. 2). Currently, relatively little is known about the similarity of cues for interacting species across trophic levels⁶².

Towards robust forecasting of phenological mismatch

While the Cushing hypothesis provides testable predictions regarding the consequences of phenological mismatch in response to climate change, it is uncertain how much support there is for it. Moving the field of phenological mismatch forward (i.e., from documentation towards forecasting) requires higher quality data that enables the testing of fundamental hypotheses and defining of key baselines. Below, we highlight how new approaches can improve testing of the Cushing hypothesis. These approaches will also provide the framework for building the required depth in evidence across studies to determine general quantitative patterns in phenological mismatch and their underlying mechanisms, an approach not currently possible given the state of the field.

(i) Testing assumptions and mechanisms of the Cushing hypothesis

By having the potential to test multiple mechanisms and define key baselines, the integration of long-term pre-climate change data and an experimental approach in a single system represents the 'gold standard'. However, in the absence of long-term pre-climate change data, the integration of observational data across a long time-series or a spatial gradient in climatic conditions, and thus variation in the relative timing of the interaction, with an experimental study can provide a powerful approach to test multiple hypotheses (Box 2). For example, Samplonius et al. ¹⁵ combine descriptive and experimental data to test many confounding factors and key assumptions of the Cushing hypothesis in their system (e.g., ontogenetic dietary shifts, habitat quality). Consequently, they advanced the field by providing a strong test of the Cushing hypothesis, demonstrating that this hypothesis is pertinent in their system, and ruling out competing hypotheses related to habitat quality.

Independent of whether observational data is available, experiments can be used to test and disentangle multiple mechanisms, for example they can disentangle direct from indirect effects or abiotic mismatches from phenological ones^{63,64}. Experiments provide a high degree of inference about mechanisms that descriptive data alone cannot provide. For example, Berger et

al.⁶⁵ manipulated several factors in lake mesocosms to separate the direct effects of light and temperature on spring plankton dynamics from effects mediated through the consumer. They showed that temperature had little effect on zooplankton phenology and performance suggesting that climate change is unlikely to result in phenological mismatch in this system. The extrapolation of their results to the lake ecosystem are uncertain, however, given the lack of long-term data in the system and the uncertainty about phenological cues for *Daphnia* in the field⁶⁵.

Baselines themselves can also be used to test multiple mechanisms (see next section for discussion about defining baselines). Researchers with long-term data in systems where multiple attributes of the environment have changed at different points can leverage this variation to test which driver is most closely associated with shifts in phenological synchrony and their associated consequences through time. For example, many lakes experienced large changes in nutrient inputs in the 1960s and 1970s (e.g., Lake Washington⁵², Esthwaite Water⁶⁶), before the onset of recent climate change in the early 1980s. Researchers can test different breakpoints with their long-term data to determine where the non-linearity or non-stationarity most likely began (Fig. 3a), thus providing evidence for which factor is likely to be the major driver of changes in phenological synchrony. For example, studies have shown that changes in nutrient inputs have had larger impacts on plankton dynamics in Lake Washington than warming over the past three decades^{67,68}.

Higher temporal resolution data in some systems may help with teasing apart different mechanisms and testing the assumptions of the Cushing hypothesis. Regardless of aim, ideally sampling frequency should be relative to the life history of the species of interest. For example, testing any mechanism related to life history theory requires data about ontogeny. In the great tit-winter moth system, caterpillar biomass sampling is usually conducted a few times a week⁶⁹. Given that the larval stage of the winter moth life cycle is typically 4-6 weeks^{70,71}, this sampling frequency allows researchers to obtain an accurate estimate of growth and development for this

life cycle stage. In the lower trophic levels of aquatic systems, where there is often quick turnaround between producers and consumers, it can be difficult to determine how strong of an influence producers have on consumers and vice versa (i.e., whether the first assumption of the Cushing hypothesis is supported³⁷). In cases where there are multiple short, potentially overlapping, generations, determining the seasonal order of the consumer vs. producer is necessary to assess the degree of dependence of the consumer on the producer.

Higher temporal resolution data will also help determine the seasonal availability of the consumer and resource, the other key assumption of the Cushing hypothesis (Fig. 1a-c). For example, temperate aquatic and terrestrial systems differ in how most organisms survive the winter. While many terrestrial plants and herbivores have dormancy phases during the winter, in aquatic systems some zooplankton are never in complete dormancy (i.e., they remain at low densities), and thus do not fully meet the assumption of the hypothesis. Nevertheless, some aquatic studies have been successful in achieving daily sampling frequency of lower trophic levels⁷², suggesting that—though more challenging—critical assumptions of the Cushing hypothesis can be tested in aquatic systems.

ii) Defining baselines

Researchers with long-term data have the potential to test whether synchrony vs. asynchrony was the pre-climate change baseline (Fig. 2; Fig. 3). By testing for a clear peak in fitness across years in relation to the relative timing of the interaction (i.e. evidence for the synchrony baseline or adaptive mismatch hypothesis) in the pre-climate change data, researchers can place their system on the Cushing curve (Fig. 3). Whenever possible, incorporating the most energetically expensive part of the consumer's annual cycle or estimating when the measured phenological phase occurs relative to the most expensive part (e.g., food demands of the great tit, *Parus major*, chicks are highest 9 days after hatching, the phenological phase most commonly monitored^{73,74,75}) will help define a peak in consumer

fitness. If there is no clear peak in fitness, then well-designed experiments (where possible) that have the ability to reveal the limits of the curve (i.e. where consumer fitness reaches zero) can be used to reject the hypothesis of an asynchrony baseline. Knowledge of when/if different resources are available 15 and whether the interaction type changes throughout the season 76,77 can also help describe the limits of the curve.

Where long-term data are not available, there is no perfect solution to the challenge of defining a baseline. Researchers with multiple years of data may be able to use inter-annual variation in abiotic conditions (e.g., temperature) to define a baseline, where closer-to-mean conditions represent a proxy for a historical baseline and, for example, cooler vs. warmer years act as alternatives to pre- and post- climate change, respectively^{68,78}. However, extrapolation from short-term datasets only work if the species responses' to varying climate themselves are not in flux, a criterion difficult to meet in many systems already^{48,75}. Null modeling to simulate historical data^{4,79,80} can help put the magnitude of observed changes in phenology and performance since significant climate change began into context and to quantify how much variation may be due to noise (i.e., factors other than climate change⁴).

(iii) Research outside of the Cushing hypothesis

Moving the field of phenological mismatch forward also requires a better understanding of systems where the Cushing hypothesis is not supported. In systems where the assumptions of the Cushing hypothesis were not met or are too difficult to test (e.g., planktonic food web at the Plymouth L4 site, UK³³), researchers can test alternate hypotheses related to the forces structuring species interactions (Fig. 2); for example, density-dependence¹³, time-dependence⁸¹, top-down control³⁶, bet-hedging⁸², size-mediated priority effects⁸³, interaction strength¹³, ontogenetic variation and stage-structure species interactions⁷⁶. Testing multiple hypotheses, and any data (even when assumptions and mechanisms related to the Cushing hypothesis are

not met or cannot be tested) that helps define how systems have and are shifting, and what mechanisms may drive those shifts, contributes valuable information for improved forecasting. An improved knowledge of the prevalence of mechanisms in different systems will advance our predictions of phenological mismatch for pair-wise interactions and the ecological consequences of phenological shifts.

iv) Final thoughts on forecasting

Given the complexity involved, accurately forecasting phenological mismatch in response to climate change is a major test of ecological theory and methods. It requires a mechanistic understanding of the processes involved to ensure robust predictions of changes in consumer performance (e.g., steepness of the Cushing curve). Previous work has identified factors that will help predict *when* a mismatch is more likely to lead to a population decline but this work is largely theoretical. While we have outlined how to work around data limitations, the best tests of mismatch theory will come from data-rich systems where the energetic links between consumer and resource species are well understood. With such information in hand, researchers can then consider forecasting longer-term demographic responses due to mismatch with continued climate change. Forecasting of phenological synchrony can be built off climate projection forecasts whenever phenological cues for both species are well known^{62,84}. Forecasting longer-term demographic responses to phenological mismatch will require data about lifetime fitness for the consumer and comprehensive data on the resource, a standard that, to our knowledge, has yet to be met by any empirical study to date.

As more information is gathered across habitats and diverse consumer-resource interactions, forecasting should eventually move beyond a system-specific approach and towards forecasting diverse systems through a multi-level modeling approach based on our mechanistic understanding of what drives mismatch. Such a model would be built on mechanistic studies of the Cushing hypothesis that carefully measure the fitness, phenologies,

interactions and other still-unknown important attributes of systems. The model would then allow researchers to measure currently unknown but critical attributes of species, sites, and ecosystems that control the shape, strength and uncertainty surrounding the Cushing curve. But progress towards such an approach requires the ability to combine knowledge from across diverse systems into one complex but well-defined mechanistic model. As we have outlined above, most current data in the field make it hard or impossible to compare across systems. Given intrinsic differences between aquatic and terrestrial systems, progress may be accelerated if these systems are considered separately before being combined. Nevertheless, without improved methods – to test multiple mechanisms and define pre-climate change baselines – the goal of general predictions of the ecological consequences of shifts in phenological synchrony will remain well out of reach.

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Box 1. Glossary

Cushing match-mismatch hypothesis- this hypothesis postulates that the consumer should temporally 'match' the peak of its most energetically demanding period with the peak of resource availability and any change to the relative timing of the interaction will result in a 'mismatch' ^{16,17,18,19}. (Fig. 1)

Phenological synchrony- is based on the phenological differences between pair-wise interacting species and is a measure of the overlap between two events, each of which has a temporal distribution. Synchrony implies that the peak of the most energetically demanding period of the consumer with the peak of resource availability completely overlaps. Importantly, this term does not incorporate fitness consequences for either species.

Phenological mismatch - decreases in consumer fitness associated with changes in the relative timing of key life history activities among interacting species. Importantly, this mismatch occurs between interacting species; it does not occur with the abiotic environment, and we consider it as relating to total fitness.

Baseline- a representative or benchmark time series of conditions that describes the species interaction. A baseline incorporates the historical range of variability in the relative timing of the interaction.

Synchrony baseline- a hypothesis that before climate change, the most energetically demanding period of the consumer was at the same time of peak resource availability, and thus consumer fitness was at its maximum (i.e. a match).

Asynchrony baseline - a hypothesis put forward by Singer and Parmesan⁴⁸ that before climate change, the most energetically demanding period of the consumer was not timed to the peak resource availability, and thus consumer fitness was not at its maximum. This could occur for various reasons (e.g., co-evolutionary arms race, other transient dynamics).

Adaptive mismatch hypothesis- a hypothesis put forward by Visser et al.⁵⁷ that postulates that, in some systems, life-history trade-offs will promote asynchrony for many or most individuals in a population and that maximum fitness does not occur at the resource peak.

Stationarity- any stochastic process with a constant underlying probability distribution

Life-history theory- the study of the causes and consequences of the diversity of life history strategies based on evolutionary theory.

Food web theory- the study of the patterns, causes and consequences of food-web structure.

Box 2. Case study illustrating the benefits of integrating experimental and observational data in a single system.

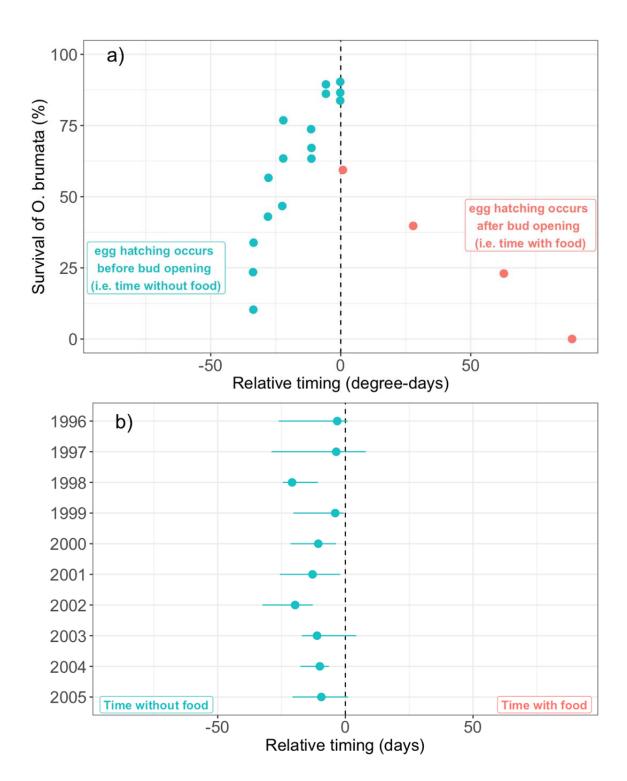
To test multiple mechanisms and define key baselines, the integration of long-term preclimate change data and an experimental approach across multiple populations in a single system represents the 'gold standard'. To the best of our knowledge, this has not yet been possible for a single system. However, the system described here represents a good example of how the integration of experimental and long-term observational data for a single system can provide evidence for the Cushing hypothesis and highlight weakness in our predictions.

In this study, the performance data for an insect herbivore (winter moth; *Operophtera brumata*), and its main host plant (common oak; *Quercus robur*) are not available for more than 10 years from a single location⁹⁰; however, different researchers manipulated the timing of larval emergence relative to bud break to test for support for the Cushing hypothesis (panel **a**; raw data obtained from Fig. 3 in Tikkanen and Julkunen-Tiitto⁹¹; see Appendix for more details). The results from two related experiments (green and red points) where the authors manipulated the number of days that neonates (i.e. early instar larvae) spent without food (green points; first experiment) and the emergence times of larvae relative to budburst (red points; second experiment) are shown in panel **a**. These experimental results provide strong support that the first assumption of the Cushing hypothesis is met in this system, much stronger support than provided by the 10 years of observational data.

However, observations of inter-annual variation in the relative timing between larval emergence (median egg hatch) relative to bud break (median bud opening) from 1996-2005 in the Netherlands (panel **b**; horizontal error bars represent the lower and upper quartiles of the data; raw data was obtained from from Fig. 2 in van Asch and Visser⁹⁰; see Appendix for more details) show that long-term data can be used to test pre-climate conditions and cues by identifying the proximate phenological cues of the consumer and resource (Fig. 2). In this case, the inter-annual variation provides evidence that the herbivore and host plant differ in some

cues (Fig. 2). However, these observational data only cover 10 years, all post-climate change, include only a small portion of the x-axis of the Cushing curve, and do not include any performance data of *O. brumata*. This makes it difficult to define the pre-climate change baseline and thus accurately predict how the performance of *O. brumata* will be affected by changes in phenological synchrony due to climate change (Fig. 3).

Nevertheless, although the strength of the relationship may vary across populations^{92,93}, a synthesis of the findings from the two approaches (i.e. observational and experimental) in any system can shed light on the state of our knowledge of that system and can guide future research (e.g., where an interaction is currently placed along the curve or how little of the curve is sampled with a post-climate change time-series). In well-known systems such as the one reviewed here, modelling pre-climate change baselines based on knowledge of phenological cues using a hindcasting approach^{94,95}, process-based phenological models⁸⁴, and/or phenologically explicit consumer-resource models¹⁴ could help to fill in information about the interaction before climate change began. Experiments could help define the parameters of a process-based model that could then elucidate that mechanisms may—or may not—appear feasible for the interaction⁸⁴.



Figures

Figure 1. Conceptualization of the Cushing match-mismatch hypothesis. The hypothesis postulates that a consumer should temporally 'match' the peak of its most energetically demanding period with the peak of resource availability (b) and thus have the highest fitness (d); if there is any change to the relative timing of the interaction (e.g., because the consumer shifts its phenology earlier (a) or later (c) relative to the resource), there will be a decrease in the consumer's fitness (i.e., a mismatch) (d). At the curve's limits, consumer fitness should fall to zero when the change in relative timing is sufficiently large (d). Curves in panels a-c represent the seasonal changes in the abundance of a population of the consumer (solid line) and resource (dotted line) where during some part of the year abundance declines to zero. The grey shaded areas represent temporal overlap between the consumer and resource.

Figure 2. A simplified flow diagram for forecasting climate change effects on consumer fitness as predicted by the Cushing hypothesis. First, both major assumptions must be met; if, for example, consumer density is the major controller on its own fitness, then no further work is warranted. If both assumptions are met, forecasting requires both knowledge of the ultimate mechanisms (which often relate to testing the first assumption) and the pre-climate change baseline. If neither assumption is met, the Cushing hypothesis is not supported and alternative hypotheses can be tested. In some cases, the mechanism predicts the pre-climate change baseline (i.e., in both our examples); we colour mechanisms in whether they predict both pre-climate change synchrony and asynchrony depending on specifics (purple), synchrony (blue) or asynchrony (red). Note that multiple mechanisms may operate in many systems, and that this figure is in no way exhaustive, but rather serves to highlight a pathway that researchers can take.

Figure 3. Key assumptions and resulting implications for climate change predictions, when preclimate change baselines are not defined. (a) Differences in the phenological time-series of a consumer-resource interaction during conditions of stationarity; when the environment becomes non-stationary (shown here just after 1980) the consumer and resource each can shift in varying directions, representing the range of recent documented phenological shifts³. (b) Most studies currently using the Cushing hypothesis assume that consumer fitness was highest before climate change and occurred at the resource peak (i.e., a match) leading to a 'synchrony baseline'. Alternatively, in a population where some individuals are matched with the resource peak but not at maximum fitness, a synchrony baseline may also occur (i.e. 'adaptive mismatch' hypothesis (AMH)⁵⁷ with synchrony baseline'). If the 'synchrony baseline' or 'AMH with synchrony baseline' are supported, then climate change could lead to declines in consumer fitness if both species share similar environmental cues and they shift together (though climate smaller declines in consumer fitness may be expected under AMH). Note that the AMH curve⁵⁷ with asynchrony baseline is not shown and that not all possible climate change scenarios are shown. (c) With an asynchrony baseline, climate change could lead to an increase or decrease, or to varying magnitudes (represented by arrow width), in consumer fitness depending on how the relative timing of the interaction changes. Dots represent the fitness of a population at a given point in time under different baseline scenarios.

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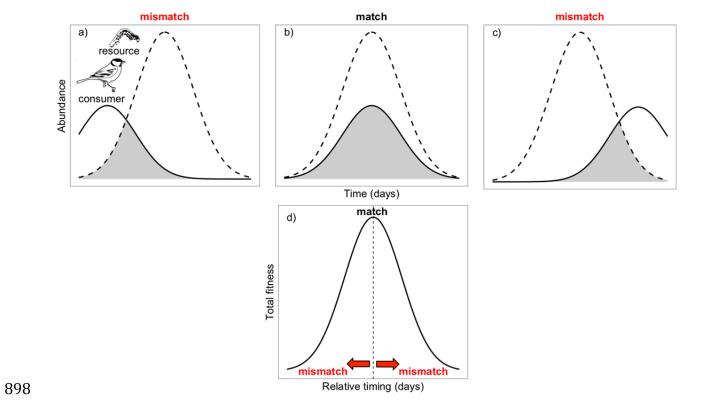
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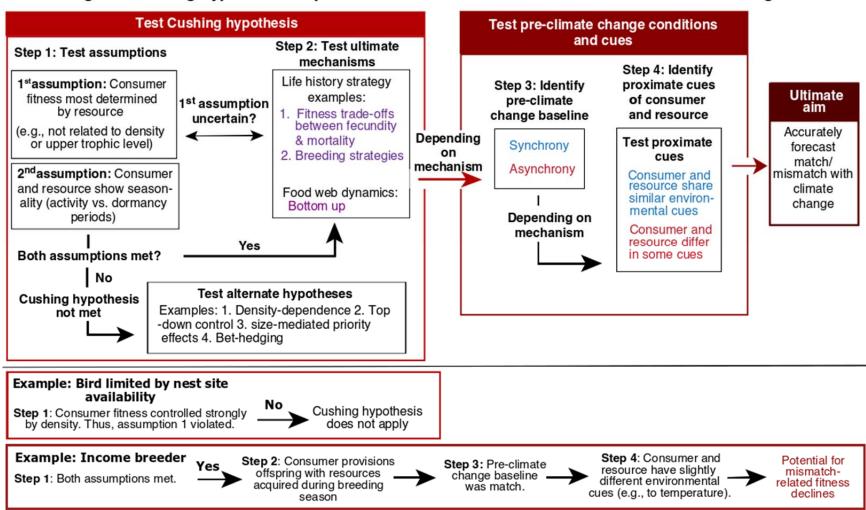




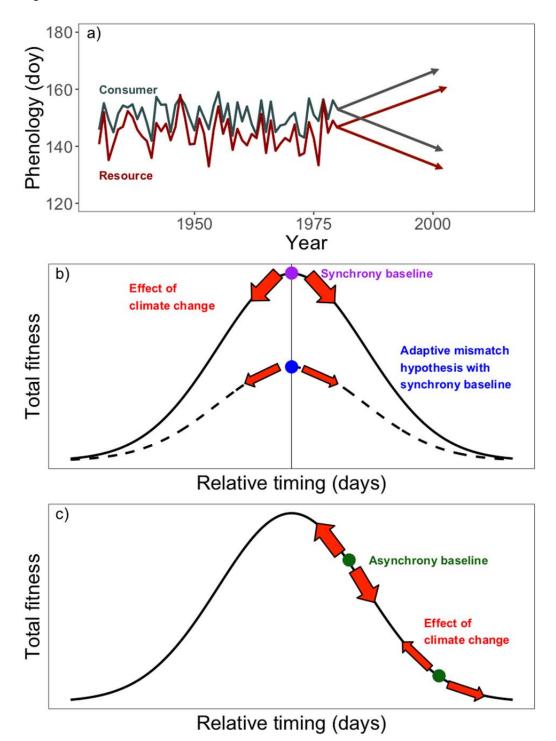
899 Figure 2.

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Using the Cushing hypothesis to predict and forecast consumer fitness with climate change



902 Figure 3.



Appendix

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Literature search

We located papers relating phenological data from trophic interactions to fitness and/or performance of the consumer and/or the resource by conducting keyword searches in ISI Web of Science published up to June 2017. Keywords included phenolog* AND mismatch* OR synchron* AND interact* AND (fitness* OR performance*). Our initial search netted 2906 papers so we further refined our search by excluding categories that included engineering and computer science. This resulted in 393 papers. From these, we focused on observational studies and excluded studies if they did not: (1) measure phenology directly (e.g. used derived measures of phenology, such as NDVI or spring temperature); (2) measure phenology of at least one of the species; (3) quantitatively link consumer fitness or performance to the relative timing between consumer and resource; and (4) explicitly state that the two species interacted (e.g. specifying type of interaction). Estimates of consumer fitness or performance (i.e., growth or development) had to be direct estimates (e.g. we did not include studies that used diet proportions to measure consumer performance) and included a range of proxies, for example, growth, survival, abundance. To ensure a reasonable sample size and to include studies across different major biomes, we included interactions that were resolved to the family-level and below. To reduce redundancy within systems, we only kept studies that were unique across pairs-location-year combinations. If more than one measure of phenology was included, we chose the one used by the authors to calculate mismatch and examine its impact on performance.

Our final review included 43 studies with 46 pair-wise species interactions (3 studies had 2 interactions). These studies encompassed terrestrial, marine and freshwater ecosystems as well as a large latitudinal gradient.

Summary of studies and interactions

Based on the type of data collected for the consumer and resource, we classified studies as life history (i.e. one that collected data at the individual level) or food web-based (i.e. one that collected data at the population or community (i.e., across species)). To determine whether studies had the potential to define pre-climate change baselines, we measured the study's time span and years of data based on the years where phenology data was available for both the consumer and resource, and consumer performance data was available.

The majority of the studies (27/43) focused on: i) documenting how climate change is affecting the timing of a trophic interaction and how those changes have affected the consumer's performance; and ii) evaluating the roles of abiotic and biotic factors in influencing synchrony. The ultimate goal of these studies was to predict the impacts of climate change on pair-wise species interactions. Our database also included studies (16/43) whose aim was to understand the underlying processes related to timing (e.g., coevolution, life history trade offs, food web dynamics) that drive consumer or resource dynamics, independently of climate change. Another divide across studies was whether performance of the consumer was measured at the individual level (28/46 interactions) or at the population or community level (18/46 interactions; see Table S1).

Additional details for Box 2

Raw data for panel (a) was obtained from Figure 3 in Tikkanen and Julkunen-Tiitto (2003). The data come from two related experiments where the authors manipulated the number of days that neonates (i.e. early instar larvae) spent without food (first experiment) and the emergence times of larvae relative to budburst (second experiment). In the first experiment, there were six groups of 30 larvae that spent 0, 5.5, 11, 22, 27,5 and 33 degree-days without food. In the second experiment, there were four cohorts, each separated by intervals of 3-5 days. All *O. brumata* eggs and larvae originated from laboratory stock originally from Turku,

- 955 Finland whereas the foliage originated from trees near Banchory, NW Scotland. Consequently,
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1121 Tables

Table S1. A comparison across studies of the type of performance data collected for consumer and resource. We define a life-history study as one that collected data at the individual level and a population/community study as one that collected data at the population or community (i.e., across species) level. Counts in the table are numbers of individual pair-wise interactions (n=46).

			Resource performance				
				Life- history			
			None	Individual	Population	Community	Totals
Consumer performance	Life-history	Individual	6	0	4	18	28
	Population/	Population	1	1	7	7	16
	Community	Community	0	0	1	1	2
		Total	7	1	12	26	46

Table S2. A comparison across studies of the type of performance data collected for the consumer across systems and taxonomic group. Counts in the table are numbers of individual pair-wise interactions (n=46).

		Sy	/stem	Taxonomic group			
					Vertebrate		
		Aquatic	Terrestrial	Invertebrate	Fish	Bird	Mammal
Level of	Individual	5	23	1	2	19	6
consumer	Population	14	2	9	4	2	1
performance	Community	1	1	2	0	0	0
	Sub-totals	20	26	12	6	21	7
	Totals		46				46