

Space is the Place: Effects of Continuous Spatial Structure on Analysis of Population Genetic Data

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ABSTRACT Real geography is continuous, but standard models in population genetics are based on discrete, well-mixed populations. As a result many methods of analyzing genetic data assume that samples are a random draw from a well-mixed population, but are applied to clustered samples from populations that are structured clinally over space. Here we use simulations of populations living in continuous geography to study the impacts of dispersal and sampling strategy on population genetic summary statistics, demographic inference, and genome-wide association studies. We find that most common summary statistics have distributions that differ substantially from that seen in well-mixed populations, especially when Wright's neighborhood size is less than 100 and sampling is spatially clustered. Stepping-stone models reproduce some of these effects, but discretizing the landscape introduces artifacts which in some cases are exacerbated at higher resolutions. The combination of low dispersal and clustered sampling causes demographic inference from the site frequency spectrum to infer more turbulent demographic histories, but averaged results across multiple simulations were surprisingly robust to isolation by distance. We also show that the combination of spatially autocorrelated environments and limited dispersal causes genome-wide association studies to identify spurious signals of genetic association with purely environmentally determined phenotypes, and that this bias is only partially corrected by regressing out principal components of ancestry. Last, we discuss the relevance of our simulation results for inference from genetic variation in real organisms.

KEYWORDS Space; Population Structure; Demography; Haplotype block sharing; GWAS

Introduction

The inescapable reality that biological organisms live, move, and reproduce in continuous geography is usually omitted from population genetic models. However, mates tend to live near to one another and to their offspring, leading to a positive correlation between genetic differentiation and geographic distance. This pattern of “isolation by distance” (Wright 1943) is one of the most widely replicated empirical findings in population genetics (Aguillon *et al.* 2017; Jay *et al.* 2012; Sharbel *et al.* 2000). Despite a long history of analytical work describing the genetics of populations distributed across continuous geography (e.g., Wright (1943); Rousset (1997); Barton *et al.* (2002, 2010); Ringbauer *et al.* (2017); Robledo-Arnuncio and Rousset (2010); Wilkins and Wakeley (2002); Wilkins (2004)), much modern work still describes geographic structure as a set of discrete populations connected by migration (e.g., Wright 1931; Epperson 2003; Rousset and Leblois 2011; Shirk and Cushman 2014; Lundgren and Ralph 2019)

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39 or as an average over such discrete models (Petkova *et al.* 2015; Al-Asadi *et al.* 2019). For this reason,
40 most population genetics statistics are interpreted with reference to discrete, well-mixed populations,
41 and most empirical papers analyze variation within clusters of genetic variation inferred by programs
42 like *STRUCTURE* (Pritchard *et al.* 2000) with methods that assume these are randomly mating units.

43 The assumption that populations are “well-mixed” has important implications for downstream
44 inference of selection and demography. Methods based on the coalescent (Kingman 1982; Wakeley
45 2009) assume that the sampled individuals are a random draw from a well-mixed population that is
46 much larger than the sample (Wakeley and Takahashi 2003). The key assumption is that the individuals
47 of each generation are *exchangeable*, so that there is no correlation between the fate or fecundity of a
48 parent and that of their offspring (Huillet and Möhle 2011). If dispersal or mate selection is limited by
49 geographic proximity, this assumption can be violated in many ways. For instance, if mean viability or
50 fecundity is spatially autocorrelated, then limited geographic dispersal will lead to parent–offspring
51 correlations. Furthermore, nearby individuals will be more closely related than an average random
52 pair, so drawing multiple samples from the same area of the landscape will represent a biased sample
53 of the genetic variation present in the whole population (Städler *et al.* 2009).

54 Two areas in which spatial structure may be particularly important are demographic inference and
55 genome-wide association studies (GWAS). Previous work has found that discrete population structure
56 can create false signatures of population bottlenecks when attempting to infer demographic histories
57 from microsatellite variation (Chikhi *et al.* 2010), statistics summarizing the site frequency spectrum
58 (SFS) (Ptak and Przeworski 2002; Städler *et al.* 2009; St. Onge *et al.* 2012), or runs of homozygosity in a
59 single individual (Mazet *et al.* 2015). The increasing availability of whole-genome data has led to the
60 development of many methods that attempt to infer detailed trajectories of population sizes through
61 time based on a variety of summaries of genetic data (Liu and Fu 2015; Schiffels and Durbin 2014;
62 Sheehan *et al.* 2013; Terhorst *et al.* 2016). Because all of these methods assume that the populations
63 being modeled are approximately randomly mating, they are likely affected by spatial biases in the
64 genealogy of sampled individuals (Wakeley 1999), which may lead to incorrect inference of population
65 changes over time (Mazet *et al.* 2015). However, previous investigations of these effects have focused on
66 discrete rather than continuous space models, and the level of isolation by distance at which inference
67 of population size trajectories become biased by structure is not well known. Here we test how two
68 methods suitable for use with large samples of individuals – stairwayplot (Liu and Fu 2015) and
69 SMC++ (Terhorst *et al.* 2016) – perform when applied to populations evolving in continuous space
70 with varying sampling strategies and levels of dispersal.

71 Spatial structure is also a major challenge for interpreting the results of genome-wide association
72 studies (GWAS). This is because many phenotypes of interest have strong geographic differences due
73 to the (nongenetic) influence of environmental or socioeconomic factors, which can therefore show
74 spurious correlations with spatially patterned allele frequencies (Bulik-Sullivan *et al.* 2015; Mathieson
75 and McVean 2012). Indeed, two recent studies found that previous evidence of polygenic selection on
76 human height in Europe was confounded by subtle population structure (Sohail *et al.* 2018; Berg *et al.*
77 2018), suggesting that existing methods to correct for population structure in GWAS are insufficient.
78 However we have little quantitative idea of the population and environmental parameters that can be
79 expected to lead to biases in GWAS.

80 Last, some of the most basic tools of population genetics are summary statistics like F_{IS} and
81 Tajima’s D , which are often interpreted as reflecting the influence of selection or demography on
82 sampled populations (Tajima 1989). Statistics like Tajima’s D are essentially summaries of the site
83 frequency spectrum, which itself reflects variation in branch lengths and tree structure of the underlying
84 genealogies of sampled individuals. Geographically limited mate choice distorts the distribution of
85 these genealogies (Maruyama 1972; Wakeley 1999), which can affect the value of Tajima’s D (Städler
86 *et al.* 2009). Similarly, the distribution of tract lengths of identity by state among individuals contains
87 information about not only historical demography (Harris and Nielsen 2013; Ralph and Coop 2013)
88 and selection (Garud *et al.* 2015), but also dispersal and mate choice (Ringbauer *et al.* 2017; Baharian
89 *et al.* 2016). We are particularly keen to examine how such summaries will be affected by models that
90 incorporate continuous space, both to evaluate the assumptions underlying existing methods and to
91 identify where the most promising signals of geography lie.

92 To study this, we have implemented an individual-based model in continuous geography that
93 incorporates overlapping generations, local dispersal of offspring, and density-dependent survival. We
94 simulate chromosome-scale genomic data in tens of thousands of individuals from parameter regimes
95 relevant to common subjects of population genetic investigation such as humans and *Drosophila*, and
96 output the full genealogy and recombination history of all final-generation individuals. We use these
97 simulations to test how sampling strategy interacts with geographic population structure to cause
98 systematic variation in population genetic summary statistics typically analyzed assuming discrete
99 population models. We then examine how the fine-scale spatial structures occurring under limited
100 dispersal impact demographic inference from the site frequency spectrum. Last, we examine the
101 impacts of continuous geography on genome-wide association studies (GWAS) and identify regions of
102 parameter space under which the results from GWAS may be misleading.

103 Materials and Methods

104 *Modeling Evolution in Continuous Space*

105 The degree to which genetic relationships are geographically correlated depends on the chance that
106 two geographically nearby individuals are close relatives – in modern terms, by the tension between
107 migration (the chance that one is descended from a distant location) and coalescence (the chance that
108 they share a parent). A key early observation by Wright (1946) is that this balance is often nicely
109 summarized by the “neighborhood size”, defined to be $N_W = 4\pi\rho\sigma^2$, where σ is the mean parent-
110 offspring distance and ρ is population density. This can be thought of as proportional to the average
111 number of potential mates for an individual (those within distance 2σ), or the number of potential
112 parents of a randomly chosen individual. Empirical estimates of neighborhood size vary hugely
113 across species – even in human populations, estimates range from 40 to over 5,000 depending on the
114 population and method of estimation (Table 1).

115 The first approach to modeling continuously distributed populations was to endow individuals in a
116 Wright-Fisher model with locations in continuous space. However, since the total size of the population
117 is constrained, this introduces interactions between arbitrarily distant individuals, which (aside from
118 being implausible) was shown by Felsenstein (1975) to eventually lead to unrealistic population
119 clumping if the range is sufficiently large. Another method for modeling spatial populations is to
120 assume the existence of a grid of discrete randomly mating populations connected by migration, thus
121 enforcing regular population density by edict. Among many other results drawn from this class of
122 “lattice” or “stepping stone” models (Epperson 2003), Rousset (1997) showed that the slope of the linear
123 regression of genetic differentiation (F_{ST}) against the logarithm of spatial distance is an estimate of
124 neighborhood size. Although these grid models may be good approximations of continuous geography
125 in many situations, they do not model demographic fluctuations, and limit investigation of spatial
126 structure below the level of the deme, assumptions whose impacts are unknown. An alternative
127 method for dealing with continuous geography is a new class of coalescent models, the Spatial Lambda
128 Fleming-Viot models (Barton *et al.* 2010; Kelleher *et al.* 2014).

129 To avoid hard-to-evaluate approximations, we here used forward-time, individual-based simulations
130 across continuous geographical space. The question of what regulates real populations has a
131 long history and many answers (e.g., Lloyd 1967; Antonovics and Levin 1980; Crawley 1990), but it is
132 clear that populations must at some point have density-dependent feedback on population size, or
133 else they would face eventual extinction or explosion. In the absence of unrealistic global population
134 regulation, this regulation must be local, and there are many ways to achieve this (Bolker *et al.* 2003). In
135 our simulations, each individual’s probability of survival is a decreasing function of local population
136 density, which shifts reproductive output towards low-density regions, and produces total census sizes
137 that fluctuate around an equilibrium. This also prevents the population clumping seen by Felsenstein
138 (1975) (Supplemental Figure S1)). Such models have been used extensively in ecological modeling
139 (Durrett and Levin 1994; Bolker and Pacala 1997; Law *et al.* 2003; Fournier and Méléard 2004; Champer
140 *et al.* 2019) but rarely in population genetics, where to our knowledge implementations of continuous
141 space models before their availability through SLiM (Haller and Messer 2019) have focused on a small
142 number of genetic loci (e.g., Slatkin and Barton 1989; Barton *et al.* 2002; Robledo-Arnuncio and Rousset

143 2010; Rossine 2014; Jackson and Fahrig 2014), which limits the ability to investigate the impacts of
 144 continuous space on genome-wide genetic variation as is now routinely sampled from real organisms.
 145 By simulating chromosome-scale sequence alignments and complete population histories we are able
 146 to treat our simulations as real populations and replicate the sampling designs and analyses commonly
 147 conducted on real genomic data.

148 **A Forward-Time Model of Evolution in Continuous Space**

149 We simulated populations using the program SLiM v3.1 (Haller and Messer 2019). Each time step
 150 consists of three stages: reproduction, dispersal, and mortality. To reduce the parameter space we
 151 use the same parameter, denoted σ , to modulate the spatial scale of interactions at all three stages
 152 by adjusting the standard deviation of the corresponding Gaussian functions. As in previous work
 153 (Wright 1943; Ringbauer *et al.* 2017), σ is equal to the mean parent-offspring distance.

154 At the beginning of the simulation individuals are distributed uniformly at random on a continuous,
 155 square landscape. Individuals are hermaphroditic, and each time step, each produces a Poisson number
 156 of offspring with mean $1/L$. Offspring disperse a Gaussian-distributed distance away from the parent
 157 with mean zero and standard deviation σ in both the x and y coordinates. Each offspring is produced
 158 with a mate selected randomly from those within distance 3σ , with probability of choosing a neighbor
 159 at distance d proportional to $\exp(-d^2/2\sigma^2)$.

160 To maintain a stable population, mortality increases with local population density. To do this we say
 161 that individuals at distance d have a competitive interaction with strength $g(d)$, where g is the Gaussian
 162 density with mean zero and standard deviation σ . Then, the sum of all competitive interactions with
 163 individual i is $n_i = \sum_j g(d_{ij})$, where d_{ij} is the distance between individuals i and j and the sum is over
 164 all neighbors within distance 3σ . Since g is a probability density, n_i is an estimate of the number of
 165 nearby individuals per unit area. Then, given a per-unit carrying capacity K , the probability of survival
 166 until the next time step for individual i is

$$p_i = \min \left(0.95, \frac{1}{1 + n_i / (K(1 + L))} \right). \quad (1)$$

167 We chose this functional form so that the equilibrium population density per unit area is close to K ,
 168 and the mean lifetime is around L ; for more description see the Appendix.

169 An important step in creating any spatial model is dealing with range edges. Because local popula-
 170 tion density is used to model competition, edge or corner populations can be assigned artificially high
 171 fitness values because they lack neighbors within their interaction radius but outside the bounds of the
 172 simulation. We approximate a decline in habitat suitability near edges by decreasing the probability
 173 of survival proportional to the square root of distance to edges in units of σ . The final probability of
 174 survival for individual i is then

$$s_i = p_i \min(1, \sqrt{x_i/\sigma}) \min(1, \sqrt{y_i/\sigma}) \min(1, \sqrt{(W - x_i)/\sigma}) \min(1, \sqrt{(W - y_i)/\sigma}) \quad (2)$$

175 where x_i and y_i are the spatial coordinates of individual i , and W is the width (and height) of the
 176 square habitat. This buffer roughly counteracts the increase in fitness individuals close to the edge
 177 would otherwise have, though the effect is relatively subtle (Figure S2).

178 To isolate spatial effects from other components of the model such as overlapping generations,
 179 increased variance in reproductive success, and density-dependent fitness, we also implemented
 180 simulations identical to those above except that mates are selected uniformly at random from the
 181 population, and offspring disperse to a uniform random location on the landscape. We refer to this
 182 model as the “random mating” model, in contrast to the first, “spatial” model.

183 We stored the full genealogy and recombination history of final-generation individuals as tree
 184 sequences (Kelleher *et al.* 2018), as implemented in SLiM (Haller *et al.* 2019). Scripts for figures and
 185 analyses are available at <https://github.com/kern-lab/spaceness>.

186 We ran 400 simulations for the spatial and random-mating models on a square landscape of width
 187 $W = 50$ with per-unit carrying capacity $K = 5$ (census $N \approx 10,000$), average lifetime $L = 4$, genome

size 10^8 bp, recombination rate 10^{-9} per bp per generation, and drawing σ values from a uniform distribution between 0.2 and 4. To speed up the simulations and limit memory overhead we used a mutation rate of 0 in SLiM and later applied mutations to the tree sequence with msprime's `mutate` function (Kelleher *et al.* 2016). Because msprime applies mutations proportionally to elapsed time, we divided the mutation rate of 10^{-8} mutations per site per generation by the average generation time estimated for each value of σ (see 'Demographic Parameters' below) to convert the rate to units of mutations per site per unit time. We verify that this procedure produces the same site frequency spectrum as applying mutations directly in SLiM in Figure S3, in agreement with theory (Ralph *et al.* 2019). Simulations were run for 1.6 million timesteps (approximately $30N$ generations).

We also compared our model's output to a commonly-used approximation of continuous space, the stepping-stone model, which we simulated with msprime (Kelleher *et al.* 2016). These results are discussed in detail in the Appendix, but in general we find that the demographic structure of a stepping-stone model can depend strongly on the chosen discretization, and some artifacts of discretization seem to become stronger in the limit of a fine grid. For many summary statistics, finer discretizations (we used a 50×50 grid) produced similar results to the continuous model, but this was not true for others (e.g., F_{IS} and Tajima's D), which differed from the continuous model *more* at finer discretizations.

204 Demographic Parameters

Our demographic model includes parameters that control population density (K), mean life span (L), and dispersal distance (σ). However, nonlinearity of local demographic stochasticity causes actual realized averages of these demographic quantities to deviate from the specified values in a way that depends on the neighborhood size. Therefore, to properly compare to theoretical expectations, we empirically calculated these demographic quantities in simulations. We recorded the census population size in all simulations, and used mean population density (ρ , census size divided by total area) to compute neighborhood size as $N_W = 4\pi\rho\sigma^2$. To estimate generation times, we stored ages of the parents of every new individual born across 200 timesteps, after a 100 generation burn-in, and took the mean. To estimate variance in offspring number, we tracked the lifetime total number of offspring for all individuals for 100 timesteps following a 100-timestep burn-in period, and calculated the variance in number of offspring across all individuals in timesteps 50-100. All calculations were performed with information recorded in the tree sequence, using pyslim (Peter L Ralph and Ashander ???).

217 Sampling

Our model records the genealogy and sequence variation of the complete population, but in real data, genotypes are only observed from a relatively small number of sampled individuals. We modeled three sampling strategies similar to common data collection methods in empirical genetic studies (Figure 1). "Random" sampling selects individuals at random from across the full landscape, "point" sampling selects individuals proportional to their distance from four equally spaced points on the landscape, and "midpoint" sampling selects individuals in proportion to their distance from the middle of the landscape. Downstream analyses were repeated across all sampling strategies.

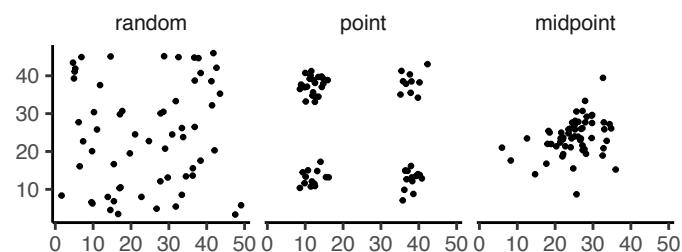


Figure 1 Example sampling maps for 60 individuals on a 50×50 landscape for midpoint, point, and random sampling strategies, respectively.

225 **Summary Statistics**

226 We calculated the site frequency spectrum and a set of 18 summary statistics (Table S1) from 60 diploid
227 individuals sampled from the final generation of each simulation using the python package scikit-
228 allel (Miles and Harding 2017). Statistics included common single-population summaries including
229 mean pairwise divergence (π), inbreeding coefficient (F_{IS}), and Tajima's D , as well as (motivated by
230 Rousset (1997)'s results) the correlation coefficient between the logarithm of the spatial distance and
231 the proportion of identical base pairs across pairs of individuals.

232 Following recent studies that showed strong signals for dispersal and demography in the distri-
233 bution of shared haplotype block lengths (e.g., Ringbauer *et al.* 2017; Baharian *et al.* 2016), we also
234 calculated various summaries of the distribution of pairwise identical-by-state (IBS) block lengths
235 among sampled chromosomes, defined to be the set of distances between adjacent sites that differ
236 between the two chromosomes. The full distribution of lengths of IBS tracts for each pair of chromo-
237 somes was first calculated with a custom python function. We then calculated the first three moments
238 of this distribution (mean, variance, and skew) and the number of blocks over 10^6 base pairs both
239 for each pair of individuals and for the full distribution across all pairwise comparisons. We then
240 calculated correlation coefficients between spatial distance and each moment of the pairwise IBS tract
241 distribution. Because more closely related individuals on average share longer haplotype blocks we
242 expect that spatial distance will be negatively correlated with mean haplotype block length, and that
243 this correlation will be strongest (i.e., most negative) when dispersal is low. The variance, skew, and
244 count of long haplotype block statistics are meant to reflect the relative length of the right (upper) tail
245 of the distribution, which represents the frequency of long haplotype blocks, and so should reflect
246 recent demographic events (Chapman and Thompson 2002). For a subset of simulations, we also
247 calculated cumulative distributions for IBS tract lengths across pairs of distant (more than 40 map
248 units) and nearby (less than 10 map units) individuals. Last, we examined the relationship between
249 allele frequency and the spatial dispersion of an allele by calculating the average distance between
250 individuals carrying each derived allele.

251 The effects of sampling on summary statistic estimates were summarized by testing for differences
252 in mean (ANOVA, (R Core Team 2018)) and variance (Levene's test, (Fox and Weisberg 2011)) across
253 sampling strategies for each summary statistic.

254 **Demographic Inference**

255
256 To assess the impacts of continuous spatial structure on demographic inference we inferred pop-
257 ulation size histories for all simulations using two approaches: stairwayplot (Liu and Fu 2015) and
258 SMC++ (Terhorst *et al.* 2016). Stairwayplot fits its model to a genome-wide estimate of the SFS, while
259 SMC++ also incorporates linkage information. For both methods we sampled 20 individuals from all
260 spatial simulations using random, midpoint, and point sampling strategies.

261 As recommended by its documentation, we used stairwayplot to fit models with multiple bootstrap
262 replicates drawn from empirical genomic data, and took the median inferred N_e per unit time as
263 the best estimate. We calculated site frequency spectra with scikit-allel (Miles and Harding 2017),
264 generated 100 bootstrap replicates per simulation by resampling over sites, and fit models for all
265 bootstrap samples using default settings.

266 For SMC++, we first output genotypes as VCF with msprime and then used SMC++'s standard
267 pipeline for preparing input files assuming no polarization error in the SFS. We used the first individual
268 in the VCF as the "designated individual" when fitting models, and allowed the program to estimate
269 the recombination rate during optimization. We fit models using the 'estimate' command rather than
270 the now recommended cross-validation approach because our simulations had only a single contig.

271 To evaluate the performance of these methods we binned simulations by neighborhood size, took a
272 rolling median of inferred N_e trajectories across all model fits in a bin for each method and sampling
273 strategy. We also examined how varying levels of isolation by distance impacted the variance of N_e
274 estimates by calculating the standard deviation of N_e from each best-fit model.

275 **Association Studies**

276 To assess the degree to which spatial structure confounds GWAS we simulated four types of nongenetic
277 phenotype variation for 1000 randomly sampled individuals in each spatial SLiM simulation and
278 conducted a linear regression GWAS with principal components as covariates in PLINK (Purcell *et al.*
279 2007). SNPs with a minor allele frequency less than 0.5% were excluded from this analysis. Phenotype
280 values were set to vary by two standard deviations across the landscape in a rough approximation
281 of the variation seen in height across Europe (Turchin *et al.* 2012; Garcia and Quintana-Domeque
282 2006, 2007). Conceptually our approach is similar to that taken by Mathieson and McVean (2012),
283 though here we model fully continuous spatial variation and compare GWAS output across a range of
284 dispersal distances.

285 In all simulations, the phenotype of each individual is determined by drawing from a Gaussian
286 distribution with standard deviation 10 and a mean that may depend on spatial position. In spatially
287 varying models, the mean phenotype differs by two standard deviations across the landscape. We
288 then adjust the geographic pattern of mean phenotype to create four types of spatially autocorrelated
289 environmental influences on phenotype. In the first simulation of *nonspatial* environments, the mean
290 did not change, so that all individuals' phenotypes were drawn independently from a Gaussian
291 distribution with mean 110 and standard deviation 10. Next, to simulate *clinal* environmental influences
292 on phenotype, we increased the mean phenotype from 100 on the left edge of the range to 120 on the
293 right edge (two phenotypic standard deviations). Concretely, the mean phenotype p for an individual
294 at position (x, y) is $p = 100 + 2x/5$. Third, we simulated a more concentrated "*corner*" environmental
295 effect by setting the mean phenotype to 120 for individuals with both x and y coordinates below 20
296 (two standard deviations above the rest of the map). Finally, in "*patchy*" simulations we selected 10
297 random points on the map and set the mean phenotype of all individuals within three map units of
298 each of these points to 120.

299 We performed principal components analysis (PCA) using scikit-allel (Miles and Harding 2017) on
300 the matrix of derived allele counts by individual for each simulation. SNPs were first filtered to remove
301 strongly linked sites by calculating LD between all pairs of SNPs in a 200-SNP moving window and
302 dropping one of each pair of sites with an R^2 over 0.1. The LD-pruned allele count matrix was then
303 centered and all sites scaled to unit variance when conducting the PCA, following recommendations
304 in Patterson *et al.* (2006).

305 We ran linear-model GWAS both with and without the first 10 principal components as covariates
306 in PLINK and summarized results across simulations by counting the number of SNPs with p -value
307 below 0.05 after adjusting for an expected false positive rate of less than 5% (Benjamini and Yekutieli
308 2001). We also examined p values for systematic inflation by comparing to the values expected from
309 a uniform distribution (because no SNPs were used when generating phenotypes, well-calibrated
310 p -values should be uniform).

311 Results from all analyses were summarized and plotted with the "ggplot2" (Wickham 2016) and
312 "cowplot" (Wilke 2019) packages in R (R Core Team 2018).

313 **Results**

314 **Demographic Parameters and Run Times**

315 Adjusting the spatial dispersal and interaction distance, σ , has a surprisingly large effect on de-
316 mographic quantities that are usually fixed in Wright-Fisher models – the generation time, census
317 population size, and variance in offspring number, shown in Figure 2. Because our simulation is
318 parameterized on an individual level, these population parameters emerge as a property of the inter-
319 actions among individuals rather than being directly set. Variation across runs occurs because, even
320 though the parameters K and L that control population density and mean lifetime respectively were
321 the same in all simulations, the strength of stochastic effects depends strongly on the spatial interaction
322 distance σ . For instance, the population density near to individual i (denoted n_i above) is computed
323 by averaging over roughly $N_W = 4\pi K\sigma^2$ individuals, and so has standard deviation proportional to
324 $1/\sqrt{N_W}$ – it is more variable at lower densities. (Recall that N_W is Wright's neighborhood size.) Since
325 the probability of survival is a nonlinear function of n_i , actual equilibrium densities and lifetimes differ

326 from K and L . This is the reason that we included *random mating* simulations – where mate choice and
 327 offspring dispersal are both nonspatial – since this should preserve the random fluctuations in local
 328 population density while destroying any spatial genetic structure. We verified that random mating
 329 models retained no geographic signal by showing that summary statistics did not differ significantly
 330 between sampling regimes (Table S2), unlike in spatial models (discussed below).

331 There are a few additional things to note about Figure 2. First, all three quantities are non-monotone
 332 with neighborhood size. Census size largely declines as neighborhood size increases for both the spatial
 333 and random mating models. However, for spatial models this decline only begins for neighborhood
 334 size ≥ 10 . Spatial and random mating models are indistinguishable from one another for neighborhood
 335 sizes larger than 100. Census sizes range from around 14,000 at low σ in the random mating model
 336 to 10,000 for both models when neighborhood sizes approach 1,000. The scaling of census sizes in
 337 both random-mating and spatial models appears to be related to two consequences of the spatial
 338 competition function: the decline of fitness at range edges, which effectively reduces the habitable area
 339 by one σ around the edge of the map and so results in a smaller habitable area at high σ values; and
 340 variation in the equilibrium population density given varying competition radii. Furthermore, census
 341 size increases in spatial models as neighborhood size increases from 2 to 10. This may reflect an Allee
 342 effect (Allee *et al.* 1949) in which some individuals are unable to find mates when the mate selection
 343 radius is very small.

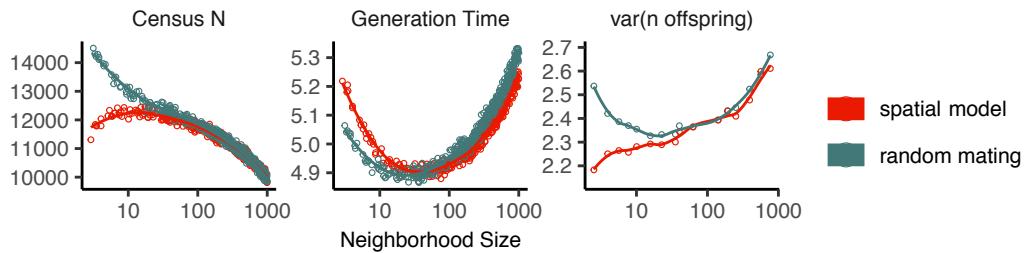


Figure 2 Genealogical parameters from spatial and random mating SLiM simulations, by neighborhood size.

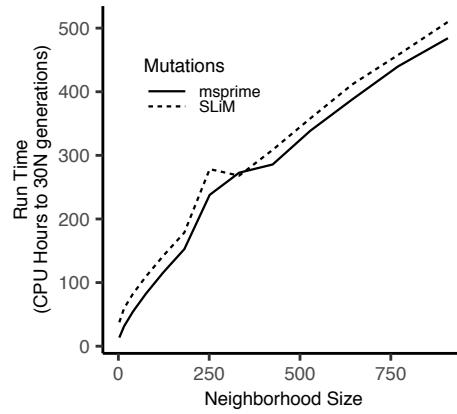


Figure 3 Run times of continuous space simulations with landscape width 50 and expected density 5 under varying neighborhood size. Times are shown for simulations run with mutations applied directly in SLiM (dashed lines) or later applied to tree sequences with msprime (solid lines). Times for simulations run with tree sequence recording disabled are shown in grey.

344 Generation time similarly shows complex behavior with respect to neighborhood sizes, and varies

345 between 5.2 and 4.9 timesteps per generation across the parameter range explored. Under both the
346 spatial and random mating models, generation time reaches a minimum at a neighborhood size of
347 around 50. Interestingly, under the range of neighborhood sizes that we examined, generation times
348 between the random mating and spatial models are never quite equivalent – presumably this would
349 cease to be the case at neighborhood sizes higher than we simulated here.

350 Last, we looked at the variance in number of offspring – a key parameter determining the effective
351 population size. Surprisingly, the spatial and random mating models behave quite differently: while
352 the variance in offspring number increases nearly monotonically under the spatial model, the random
353 mating model actually shows a decline in the variance in offspring number until a neighborhood size
354 of around 10 before it increases and eventually equals what we observe in the spatial case.

355 Run times for our model scale approximately linearly with neighborhood size (Figure 3), with the
356 lowest neighborhood sizes reaching 30N generations in around a day and those with neighborhood
357 size approaching 1,000 requiring up to three weeks of computation. As currently implemented running
358 simulations at neighborhood sizes more than 1,000 to coalescence is likely impractical, though running
359 these models for more limited timescales and then “recapitulating” the simulation using reverse-time
360 simulation from the resulting tree sequence in msprime is possible (Haller *et al.* 2019).

361 **Impacts of Continuous Space on Population Genetic Summary Statistics**

362 Even though certain aspects of population demography depend on the scale of spatial interactions, it
363 still could be that population genetic variation is well-described by a well-mixed population model.
364 Indeed, mathematical results suggest that genetic variation in some spatial models should be well-
365 approximated by a Wright-Fisher population if neighborhood size is large and all samples are geo-
366 graphically widely separated (??). However, the behavior of most common population genetic
367 summary statistics other than Tajima’s *D* (Städler *et al.* 2009) has not yet been described in realistic
368 geographic models. Moreover, as we will show, spatial sampling strategies can affect summaries of
369 genetic variation at least as strongly as the underlying population dynamics.

370 **Site Frequency Spectra and Summaries of Diversity** Figure 4 shows the effect of varying neighbor-
371 hood size and sampling strategy on the site frequency spectrum (Figure 4, Figure S5) and several
372 standard population genetic summary statistics (Figure 4B; additional statistics are shown in Figure
373 S4). Consistent with findings in island and stepping stone simulations (Städler *et al.* 2009), the SFS
374 shows a significant enrichment of intermediate frequency variants in comparison to the nonspatial
375 expectation. This bias is most pronounced below a neighborhood size of 100 and is exacerbated by
376 midpoint and point sampling of individuals (depicted in Figure 1). Reflecting this, Tajima’s *D* is quite
377 positive in the same situations (Figure 4B). Notably, the point at which Tajima’s *D* approaches 0 differs
378 strongly across sampling strategies – varying from a neighborhood size of roughly 50 for random
379 sampling to at least 1000 for midpoint sampling.

380 One of the most commonly used summaries of variation is Tajima’s summary of nucleotide diversity,
381 θ_π , calculated as the mean density of nucleotide differences averaged across pairs of samples. As can
382 be seen in Figure 4B, θ_π in the spatial model is inflated by up to three-fold relative to the random
383 mating model. This pattern is opposite the expectation from census population size (Figure 2), because
384 the spatial model has *lower* census size than the random mating model at neighborhood sizes less than
385 100. Differences between these models likely occur because θ_π is a measure of mean time to most recent
386 common ancestor between two samples, and at small values of σ , the time for dispersal to mix ancestry
387 across the range exceeds the mean coalescent time under random mating. (For instance, at the smallest
388 value of $\sigma = 0.2$, the range is 250 dispersal distances wide, and since the location of a diffusively
389 moving lineage after k generations has variance $k\sigma^2$, it takes around $250^2 = 62500$ generations to
390 mix across the range, which is roughly ten times larger than the random mating effective population
391 size). θ_π using each sampling strategy approaches the random mating expectation at its own rate, but
392 by a neighborhood size of around 100 all models are roughly equivalent. Interestingly, the effect of
393 sampling strategy is reversed relative to that observed in Tajima’s *D* – midpoint sampling reaches
394 random mating expectations around neighborhood size 50, while random sampling is inflated until
395 around neighborhood size 100.

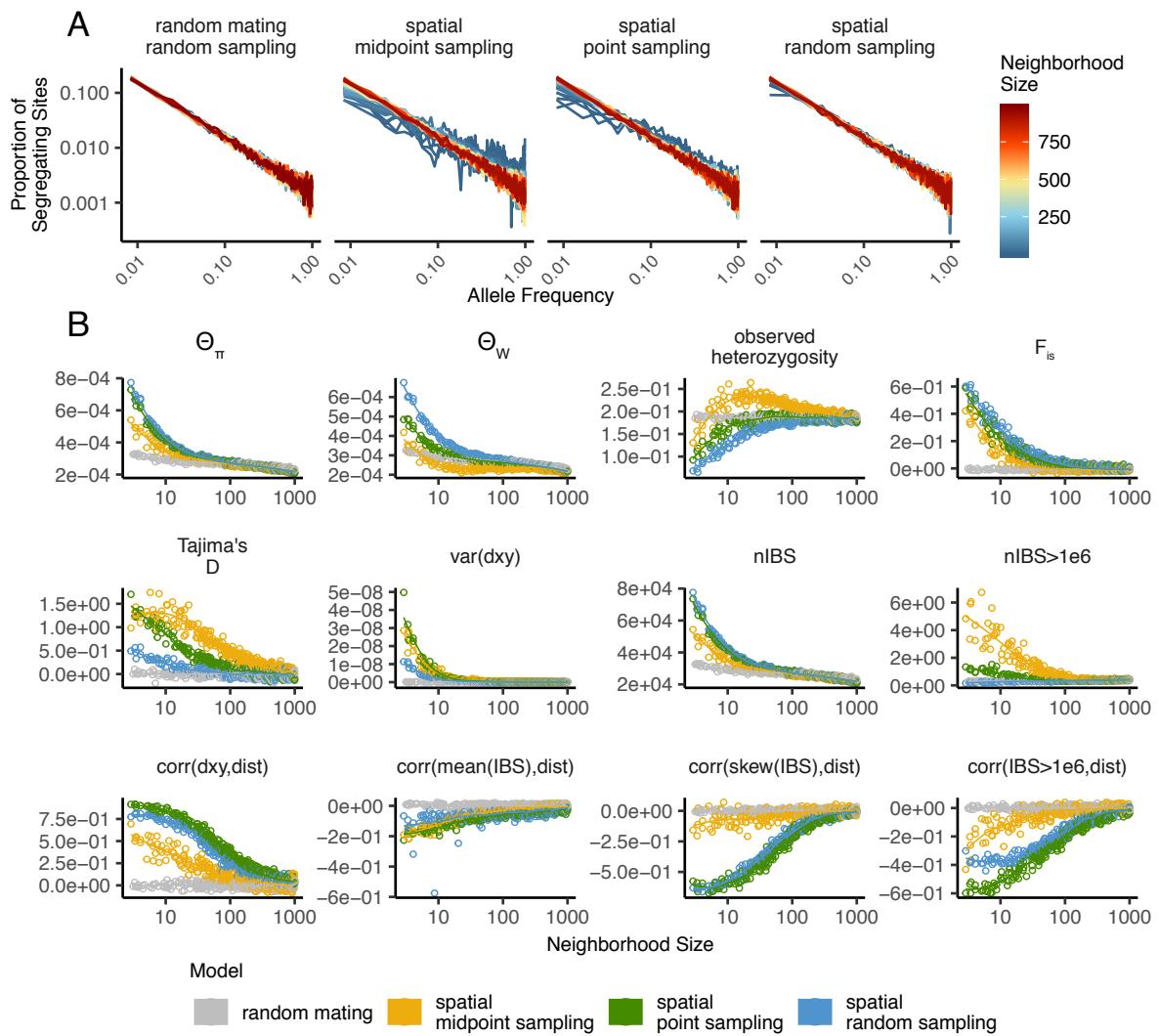


Figure 4 Site frequency spectrum (A; note axes are log-scaled) and summary statistic distributions (B) by sampling strategy and neighborhood size.

396 Values of observed heterozygosity and its derivative F_{IS} also depend heavily on neighborhood size
397 under spatial models as well as the sampling scheme. F_{IS} is inflated above the expectation across
398 most of the parameter space examined and across all sampling strategies. This effect is caused by
399 a deficit of heterozygous individuals in low-dispersal simulations – a continuous-space version of
400 the Wahlund effect (Wahlund 1928). Indeed, for random sampling under the spatial model, F_{IS} does
401 not approach the random mating equivalent until neighborhood sizes of nearly 1000. On the other
402 hand, the dependency of raw observed heterozygosity on neighborhood size is not monotone. Under
403 midpoint sampling observed heterozygosity is inflated even over the random mating expectation, as a
404 result of the a higher proportion of heterozygotes occurring in the middle of the landscape (Figure S6).
405 This echoes a report from Shirk and Cushman (2014) who observed a similar excess of heterozygosity
406 in the middle of the landscape when simulating under a lattice model.

407 **IBS tracts and correlations with geographic distance** We next turn our attention to the effect of
408 geographic distance on haplotype block length sharing, summarized for sets of nearby and distant
409 individuals in Figure 5. There are two main patterns to note. First, nearby individuals share more
410 long IBS tracts than distant individuals (as expected because they are on average more closely related).
411 Second, the difference in the number of long IBS tracts between nearby and distant individuals
412 decreases as neighborhood size increases. This reflects the faster spatial mixing of populations with
413 higher dispersal, which breaks down the correlation between the IBS tract length distribution and
414 geographic distance. This can also be seen in the bottom row of Figure 4B, where the correlation
415 coefficients between the summaries of the IBS tract length distribution (the mean, skew, and count of
416 tracts over 10^6 bp) and geographic distance approaches 0 as neighborhood size increases.

417 The patterns observed for correlations of IBS tract lengths with geographic distance are similar
418 to those observed in the more familiar correlation of allele frequency measures such as D_{xy} (i.e.,
419 “genetic distance”) or F_{ST} against geographic distance (Rousset 1997). D_{xy} is positively correlated
420 with the geographic distance between the individuals, and the strength of this correlation declines
421 as dispersal increases (Figure 4B), as expected (Wright 1943; Rousset 1997). This relationship is very
422 similar across random and point sampling strategies, but is weaker for midpoint sampling, perhaps
423 due to a dearth of long-distance comparisons. In much of empirical population genetics a regression
424 of genetic differentiation against spatial distance is a de-facto metric of the significance of isolation
425 by distance. The similar behavior of moments of the pairwise distribution of IBS tract lengths shows
426 why haplotype block sharing has recently emerged as a promising source of information on spatial
427 demography through methods described in Ringbauer *et al.* (2017) and Baharian *et al.* (2016).

428 **Spatial distribution of allele copies** Mutations occur in individuals and spread geographically over
429 time. Because low frequency alleles generally represent recent mutations (Sawyer 1977; Griffiths *et al.*
430 1999), the geographic spread of an allele may covary along with its frequency in the population. To
431 visualize this relationship we calculated the average distance among individuals carrying a focal
432 derived allele across simulations with varying neighborhood sizes, shown in Figure 6. On average
433 we find that low frequency alleles are the most geographically restricted, and that the extent to which
434 geography and allele frequency are related depends on the amount of dispersal in the population.
435 For populations with large neighborhood sizes we found that even very low frequency alleles can be
436 found across the full landscape, whereas in populations with low neighborhood sizes the relationship
437 between distance among allele copies and their frequency is quite strong. This is the basic process
438 underlying Novembre and Slatkin’s (2009) method for estimating dispersal distances based on the
439 distribution of low frequency alleles, and also generates the greater degree of bias in GWAS effect sizes
440 for low frequency alleles identified in Mathieson and McVean (2012).

441 **Effects of Space on Demographic Inference**

442 One of the most important uses for population genetic data is inferring demographic history of popu-
443 lations. As demonstrated above, the site frequency spectrum and the distribution of IBS tracts varies
444 across neighborhood sizes and sampling strategies. Does this variation lead to different inferences of
445 past population sizes? To ask this we inferred population size histories from samples drawn from our

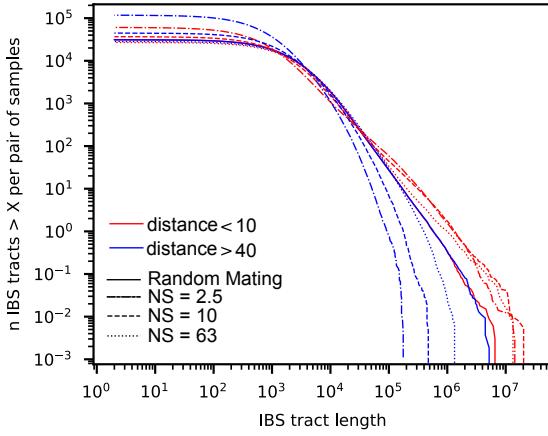


Figure 5 Cumulative distributions for IBS tract lengths per pair of individuals at different geographic distances, across three neighborhood sizes (NS). Nearby pairs (red curves) share many more long IBS tracts than do distant pairs (blue curves), except in the random mating model. The distribution of long IBS tracts between nearby individuals are very similar across neighborhood sizes, but distant individuals are much more likely to share long IBS tracts at high neighborhood size than at low neighborhood size.

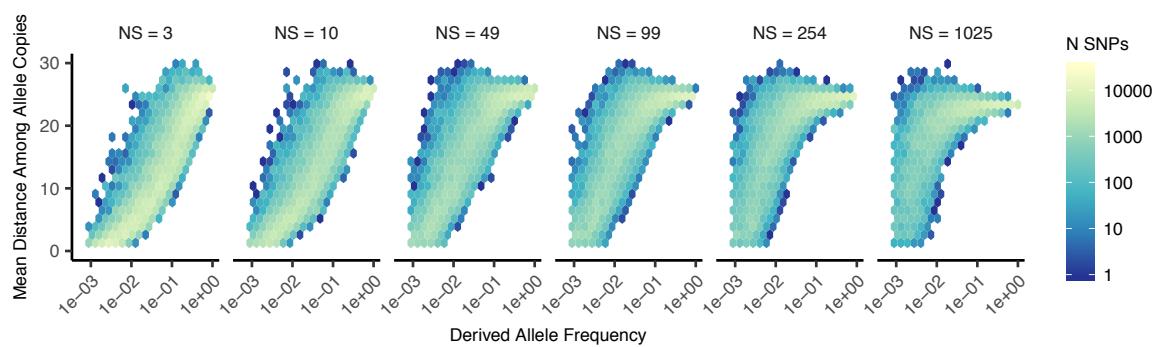


Figure 6 Spatial spread of rare alleles by neighborhood size (NS): Each plot shows the distribution (across derived alleles and simulations) of average pairwise distance between individuals carrying a focal derived allele and derived allele frequency.

446 simulated populations with two approaches: stairwayplot (Liu and Fu 2015), which uses a genome-
447 wide estimate of the SFS, and SMC++ (Terhorst *et al.* 2016), which incorporates information on both the
448 SFS and linkage disequilibrium across the genome.

449 Figure 7A shows rolling medians of inferred population size histories from each method across all
450 simulations, grouped by neighborhood size and sampling strategy. In general these methods tend to
451 slightly overestimate ancient population sizes and infer recent population declines when neighborhood
452 sizes are below 20 and sampling is spatially clustered. The overestimation of ancient population sizes
453 however is relatively minor, averaging around a two-fold inflation at 10,000 generations before present
454 in the worst-affected bins. For stairwayplot we found that many runs infer dramatic population
455 bottlenecks in the last 1,000 generations when sampling is spatially concentrated, resulting in ten-fold
456 or greater underestimates of recent population sizes. However SMC++ appeared more robust to
457 this error, with runs on point- and midpoint-sampled simulations at the lowest neighborhood sizes
458 underestimating recent population sizes by roughly half and those on randomly sampled simulations
459 showing little error. Above neighborhood sizes of around 100, both methods performed relatively well
460 when averaging across results from multiple simulations.

461 However, individual simulations were often inferred to have turbulent demographic histories, as
462 shown by the individually inferred histories (shown in Figure S7). Indeed, the standard deviation of
463 inferred N_e across time points (shown in Figure 7B) often exceeds the expected N_e for both methods.
464 That is, despite the nearly constant population sizes in our simulations, both methods tended to infer
465 large fluctuations in population size over time, which could potentially result in incorrect biological
466 interpretations. On average the variance of inferred population sizes was elevated at the lowest
467 neighborhood sizes and declines as dispersal increases, with the strongest effects seen in stairwayplot
468 results with clustered sampling and neighborhood sizes less than 20 (Figure 7B).

469 **GWAS**

470 To ask what confounding effects spatial genetic variation might have on genome-wide association
471 studies we performed GWAS on our simulations using phenotypes that were determined solely by
472 the environment – so, any SNP showing statistically significant correlation with phenotype is a false
473 positive. As expected, spatial autocorrelation in the environment causes spurious associations across
474 much of the genome if no correction for genetic relatedness among samples is performed (Figures 8 and
475 S8). This effect is particularly strong for clinal and corner environments, for which the lowest dispersal
476 levels cause over 60% of SNPs in the sample to return significant associations. Patchy environmental
477 distributions, which are less strongly spatially correlated (Figure 8A), cause fewer false positives
478 overall but still produce spurious associations at roughly 10% of sites at the lowest neighborhood
479 sizes. Interestingly we also observed a small number of false positives in roughly 3% of analyses
480 on simulations with nonspatial environments, both with and without PC covariates included in the
481 regression.

482 The confounding effects of geographic structure are well known, and it is common practice to
483 control for this by including principal components (PCs) as covariates to control for these effects. This
484 mostly works in our simulations – after incorporating the first ten PC axes as covariates, the vast
485 majority of SNPs no longer surpass a significance threshold chosen to have a 5% false discovery rate
486 (FDR). However, a substantial number of SNPs – up to 1.5% at the lowest dispersal distances – still
487 surpass this threshold (and thus would be false positives in a GWAS), especially under “corner” and
488 “patchy” environmental distributions (Figure 8C). At neighborhood sizes larger than 500, up to 0.31%
489 of SNPs were significant for corner and clinal environments. Given an average of 132,000 SNPs across
490 simulations after MAF filtering, this translates to up to 382 false-positive associations; for human-sized
491 genomes, this number would be much larger. In most cases the p values for these associations were
492 significant after FDR correction but would not pass the threshold for significance under the more
493 conservative Bonferroni correction (see example Manhattan plots in figure S8).

494 Clinal environments cause an interesting pattern in false positives after PC correction: at low
495 neighborhood sizes the correction removes nearly all significant associations, but at neighborhood
496 sizes above roughly 250 the proportion of significant SNPs increases to up to 0.4% (Figure 8). This
497 may be due to a loss of descriptive power of the PCs – as neighborhood size increases, the total

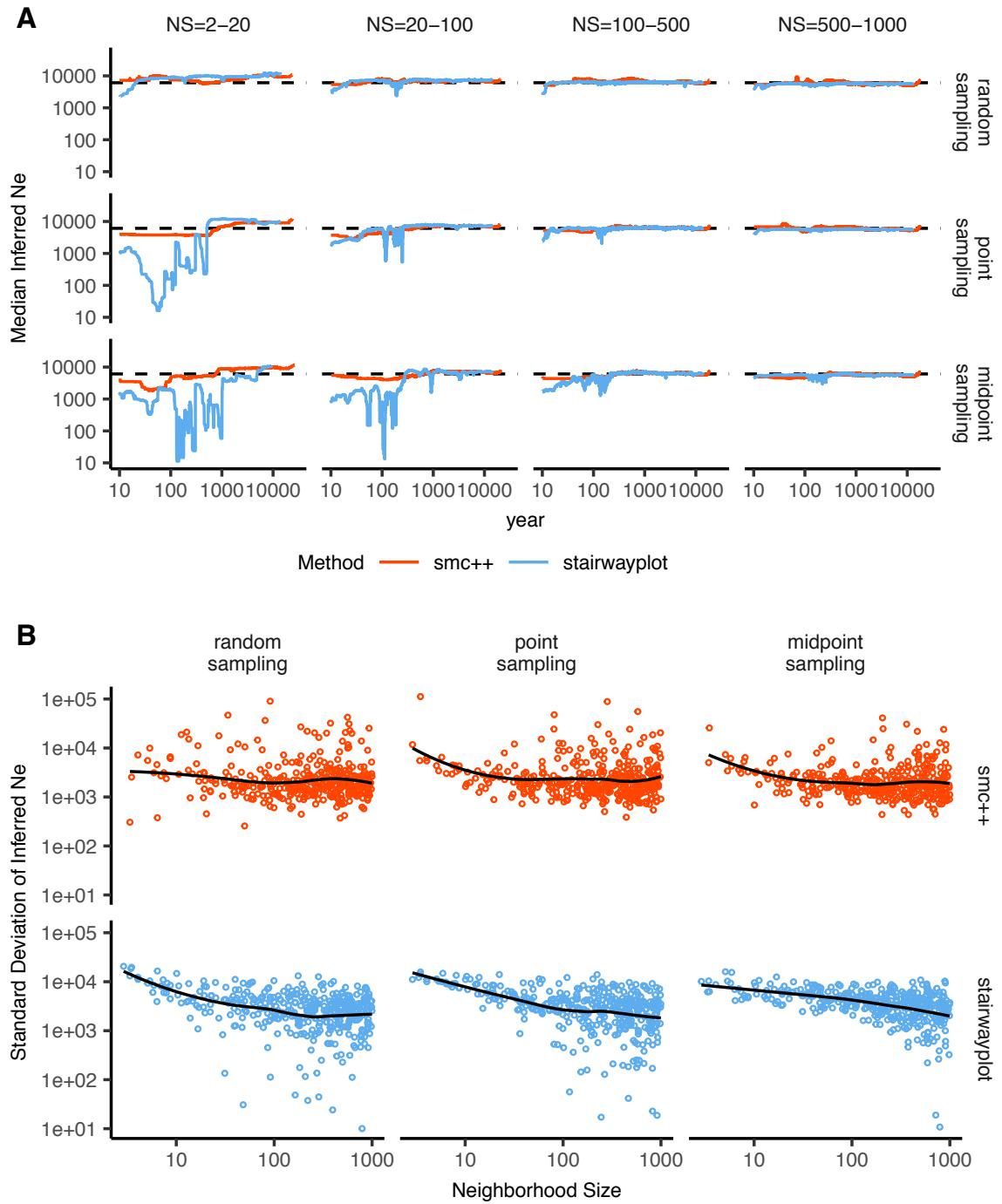


Figure 7 A: Rolling median inferred N_e trajectories for stairwayplot and smc++ across sampling strategies and neighborhood size bins. The dotted line shows the mean N_e of random-mating simulations. B: Standard deviation of individual inferred N_e trajectories, by neighborhood size and sampling strategy. Black lines are loess curves. Plots including individual model fits are shown in Figure S7.

498 proportion of variance explained by the first 10 PC axes declines from roughly 10% to 4% (Figure
499 8B). Essentially, PCA seems unable to effectively summarize the weak population structure present in
500 large-neighborhood simulations, but these populations continue to have enough spatial structure to
501 create significant correlations between genotypes and the environment. A similar process can also be
502 seen in the corner phenotype distribution, in which the count of significant SNPs initially declines as
503 neighborhood size increases and then increases at approximately the point at which the proportion of
504 variance explained by PCA approaches its minimum.

505 Figure 8D shows quantile-quantile plots for a subset of simulations that show the degree of genome-
506 wide inflation of test statistics in PC-corrected GWAS across all simulations and environmental distri-
507 butions. An alternate visualization is also included in figure S9. For clinal environments, $-\log_{10}(p)$
508 values are most inflated when neighborhood sizes are large, consistent with the pattern observed in
509 the count of significant associations after PC regression. In contrast corner and patchy environments
510 cause the greatest inflation in $-\log_{10}(p)$ at neighborhood sizes less than 100, which likely reflects
511 the inability of PCA to account for fine-scale structure caused by very limited dispersal. Finally, we
512 observed that PC regression appears to overfit to some degree for all phenotype distributions, visible
513 in Figure 8D as points falling below the 1:1 line.

514 Discussion

515 In this study, we have used efficient forward time population genetic simulations to describe the
516 myriad influence of continuous geography on genetic variation. In particular, we examine how three
517 main types of downstream empirical inference are affected by unmodeled spatial population structure
518 – population genetic summary statistics, inference of population size history, and genome-wide associ-
519 ation studies (GWAS). As discussed above, space often matters (and sometimes dramatically), both
520 because of how samples are arranged in space, and because of the inherent patterns of relatedness
521 established by geography.

522 Effects of Dispersal

523 Limited dispersal inflates effective population size, creates correlations between genetic and spatial
524 distances, and introduces strong distortions in the site frequency spectrum that are reflected in a
525 positive Tajima's D (Figure 4). At the lowest dispersal distances, this can increase genetic diversity
526 threefold relative to random-mating expectations. These effects are strongest when neighborhood
527 sizes are below 100, but in combination with the effects of nonrandom sampling they can persist up to
528 neighborhood sizes of at least 1000 (e.g., inflation in Tajima's D and observed heterozygosity under
529 midpoint sampling). If samples are chosen uniformly from across space, the general pattern is similar
530 to expectations of the original analytic model of Wright (1943), which predicts that populations with
531 neighborhood sizes under 100 will differ substantially from random mating, while those above 10,000
532 will be nearly indistinguishable from panmixia.

533 The patterns observed in sequence data reflect the effects of space on the underlying genealogy.
534 Nearby individuals coalesce rapidly under limited dispersal and so are connected by short branch
535 lengths, while distant individuals take much longer to coalesce than they would under random
536 mating. Mutation and recombination events in our simulation both occur at a constant rate along
537 branches of the genealogy, so the genetic distance and number of recombination events separating
538 sampled individuals simply gives a noisy picture of the genealogies connecting them. Tip branches
539 (i.e., branches subtending only one individual) are then relatively short, and branches in the middle of
540 the genealogy connecting local groups of individuals relatively long, leading to the biases in the site
541 frequency spectrum shown in Figure 4.

542 The genealogical patterns introduced by limited dispersal are particularly apparent in the distribu-
543 tion of haplotype block lengths (Figure 4). This is because identical-by-state tract lengths reflect the
544 impacts of two processes acting along the branches of the underlying genealogy – both mutation and
545 recombination – rather than just mutation as is the case when looking at the site frequency spectrum or
546 related summaries. This means that the pairwise distribution of haplotype block lengths carries with
547 it important information about genealogical variation in the population, and correlation coefficients

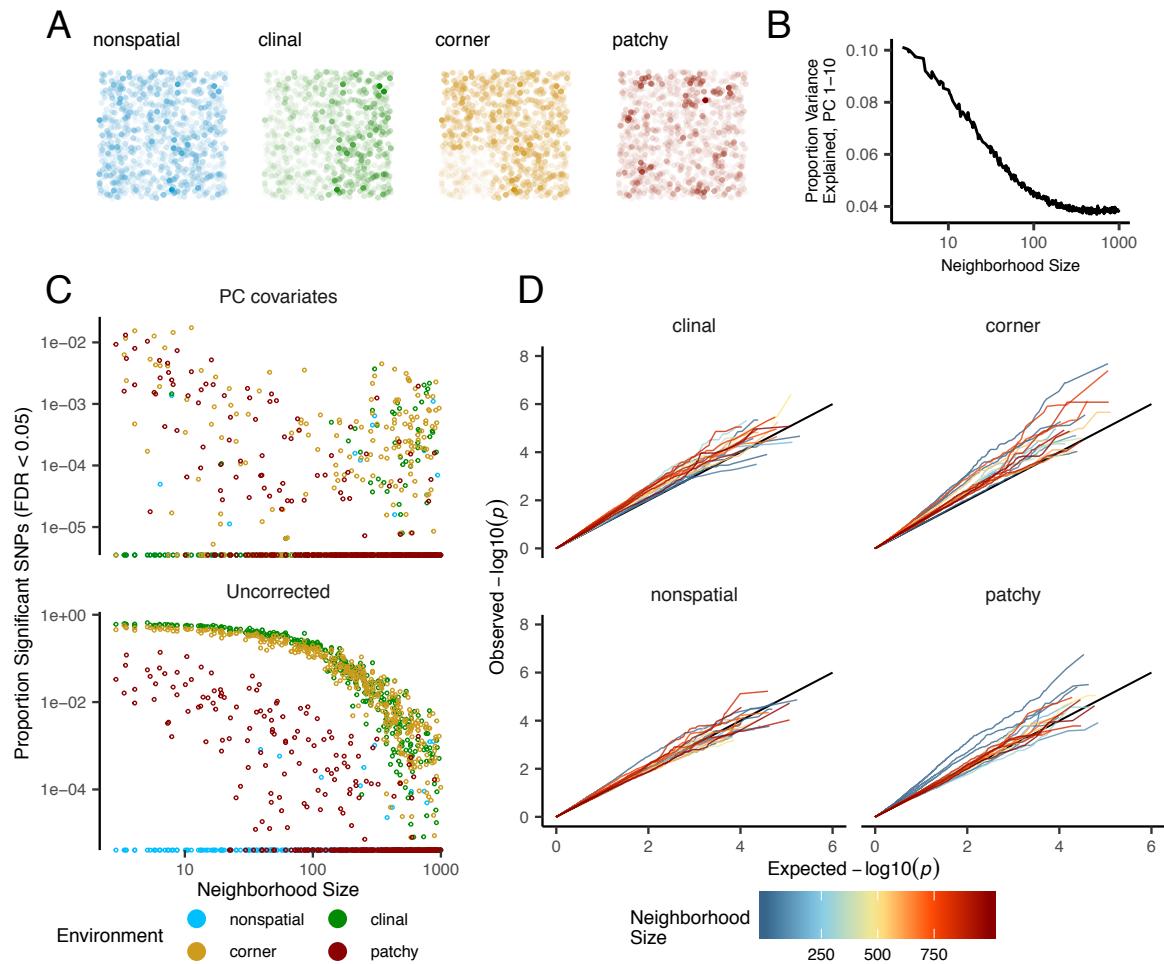


Figure 8 Impacts of spatially varying environments and isolation by distance on linear regression GWAS. Simulated quantitative phenotypes are determined only by an individual's location and the spatial distribution of environmental factors. In **A** we show the phenotypes and locations of sampled individuals under four environmental distributions, with transparency scaled to phenotype. As neighborhood size increases a PCA explains less of the total variation in the data (**B**). Spatially correlated environmental factors cause false positives at a large proportion of SNPs, which is partially but not entirely corrected by adding the first 10 PC coordinates as covariates (**C**). Quantile-quantile plots in **(D)** show inflation of $-\log_{10}(p)$ after PC correction for simulations with spatially structured environments, with line colors showing the neighborhood size of each simulation.

548 between moments of the this distribution and geographic location contain signal similar to the correlations
549 between F_{ST} or D_{xy} and geographic distance (Rousset 1997). Indeed this basic logic underlies
550 two recent studies explicitly estimating dispersal from the distribution of shared haplotype block
551 lengths (Ringbauer *et al.* 2017; Baharian *et al.* 2016). Conversely, because haplotype-based measures of
552 demography are particularly sensitive to variation in the underlying genealogy, inference approaches
553 that assume random mating when analyzing the distribution of shared haplotype block lengths are
554 likely to be strongly affected by spatial processes.

555 **Effects of Sampling**

556 One of the most important differences between random mating and spatial models is the effect of
557 sampling: in a randomly mating population the spatial distribution of sampling effort has no effect on
558 estimates of genetic variation (Table S1), but when dispersal is limited sampling strategy can compound
559 spatial patterns in the underlying genealogy and create pervasive impacts on all downstream genetic
560 analyses (see also Städler *et al.* (2009)). In most species, the difficulty of traveling through all parts
561 of a species range and the inefficiency of collecting single individuals at each sampling site means
562 that most studies follow something closest to the “point” sampling strategy we simulated, in which
563 multiple individuals are sampled from nearby points on the landscape. For example, in ornithology a
564 sample of 10 individuals per species per locality is a common target when collecting for natural history
565 museums. In classical studies of *Drosophila* variation the situation is considerably worse, in which a
566 single orchard might be extensively sampled.

567 When sampling is clustered at points on a landscape and dispersal is limited, the sampled individ-
568 uals will be more closely related than a random set of individuals. Average coalescence times of
569 individuals collected at a locality will then be more recent and branch lengths shorter than expected by
570 analyses assuming random mating. This leads to fewer mutations and recombination events occurring
571 since their last common ancestor, causing a random set of individuals to share longer average IBS tracts
572 and have fewer nucleotide differences. For some data summaries, such as Tajima’s D , Watterson’s
573 θ , or the correlation coefficient between spatial distance and the count of long haplotype blocks, this
574 can result in large differences in estimates between random and point sampling (Figure 4). Inferring
575 underlying demographic parameters from these summary statistics – unless the spatial locations of the
576 sampled individuals are somehow taken into account – will likely be subject to bias.

577 We observed the largest sampling effects using “midpoint” sampling. This model is meant to
578 reflect a bias in sampling effort towards the middle of a species’ range. In empirical studies this
579 sampling strategy could arise if, for example, researchers choose to sample the center of the range
580 and avoid range edges to maximize probability of locating individuals during a short field season.
581 Because midpoint sampling provides limited spatial resolution it dramatically reduces the magnitude
582 of observed correlations between spatial and genetic distances. More surprisingly, midpoint sampling
583 also leads to strongly positive Tajima’s D and an inflation in the proportion of heterozygous individuals
584 in the sample – similar to the effect of sampling a single deme in an island model as reported in Städler
585 *et al.* (2009). This increase in observed heterozygosity appears to reflect the effects of range edges,
586 which are a fundamental facet of spatial genetic variation. If individuals move randomly in a finite
587 two-dimensional landscape then regions in the middle of the landscape receive migrants from all
588 directions while those on the edge receive no migrants from at least one direction. The average number
589 of new mutations moving into the middle of the landscape is then higher than the number moving
590 into regions near the range edge, leading to higher heterozygosity and lower inbreeding coefficients
591 (F_{IS}) away from range edges. Though here we used only a single parameterization of fitness decline at
592 range edges we believe this is a general property of non-infinite landscapes as it has also been observed
593 in previous studies simulating under lattice models (Neel *et al.* 2013; Shirk and Cushman 2014).

594 In summary, we recommend that empirical researchers collect individuals from across as much
595 of the species’ range as practical, choosing samples separated by a range of spatial scales. Many
596 summary statistics are designed for well-mixed populations, and so provide different insights into
597 genetic variation when applied to different subsets of the population. Applied to a cluster of samples,
598 summary statistics based on segregating sites (e.g., Watterson’s θ and Tajima’s D), heterozygosity, or
599 the distribution of long haplotype blocks, can be expected to depart significantly from what would be

600 obtained from a wider distribution of samples. Comparing the results of analyses conducted on all
601 individuals versus those limited to single individuals per locality can provide an informative contrast.
602 Finally we wish to point out that the bias towards intermediate allele frequencies that we observe may
603 mean that the importance of linked selection, at least as is gleaned from the site frequency spectrum,
604 may be systematically underestimated currently.

605 **Demography**

606 Previous studies have found that population structure and nonrandom sampling can create spurious
607 signals of population bottlenecks when attempting to infer demographic history with microsatellite
608 variation, summary statistics, or runs of homozygosity (Chikhi *et al.* 2010; Städler *et al.* 2009; Ptak
609 and Przeworski 2002; Mazet *et al.* 2015). Here we found that methods that infer detailed population
610 trajectories through time based on the SFS and patterns of LD across the genome are also subject to this
611 bias, with some combinations of dispersal and sampling strategy systematically inferring deep recent
612 population bottlenecks and overestimating ancient N_e by around a factor of 2. We were surprised
613 to see that both stairwayplot and SMC++ can tolerate relatively strong isolation by distance – i.e.,
614 neighborhood sizes of 20 – and still perform well when averaging results across multiple simulations.
615 Inference in populations with neighborhood sizes over 20 was relatively unbiased unless samples were
616 concentrated in the middle of the range (Figure 7). Although median demography estimates across
617 many independent simulations were fairly accurate, empirical work has only a single estimate to work
618 with, and individual model fits (Figure S7) suggest that spuriously inferred population size changes
619 and bottlenecks are common, especially at small neighborhood sizes. As we will discuss below, most
620 empirical estimates of neighborhood size, including all estimates for human populations, are large
621 enough that population size trajectories inferred by these approaches should not be strongly affected by
622 spatial biases created by dispersal in continuous landscapes. In contrast, Mazet *et al.* (2015) found that
623 varying migration rates through time could create strong biases in inferred population trajectories from
624 an n -island model with parameters relevant for human history, suggesting that changes in migration
625 rates through time are more likely to drive variation in inferred N_e than isolation by distance.

626 We found that SMC++ was more robust to the effects of space than stairwayplot, underestimating
627 recent populations by roughly half in the worst time periods rather than nearly 10-fold as with
628 stairwayplot. Though this degree of variation in population size is certainly meaningful in an ecological
629 context, it is relatively minor in population genetic terms. Methods directly assessing haplotype
630 structure in phased data example, (e.g., Browning and Browning 2015) are thought to provide increased
631 resolution for recent demographic events, but in this case the error we observed was essentially an
632 accurate reflection of underlying genealogies in which terminal branches are anomalously short.
633 Combined with our analysis of IBS tract length variation (Figure 5) this suggests that haplotype-based
634 methods are likely to be affected by similar biases.

635 A more worrying pattern was the high level of variance in inferred N_e trajectories for individual
636 model fits using these methods, which was highest in simulations with the smallest neighborhood
637 size (Figure 7, Figure S7). This suggests that, at a minimum, researchers working with empirical data
638 should replicate analyses multiple times and take a rolling average if model fits are inconsistent across
639 runs. Splitting samples and running replicates on separate subsets – the closest an empirical study can
640 come to our design of averaging the results from multiple simulations – may also alleviate this issue.

641 Our analysis suggests that many empirical analyses of population size history using methods like
642 SMC++ are robust to error caused by spatial structure within continuous landscapes. Inferences drawn
643 from static SFS-based methods like stairwayplot should be treated with caution when there are signs
644 of isolation by distance in the underlying data (for example, if a regression of F_{ST} against the logarithm
645 of geographic distance has a significantly positive slope), and in particular an inference of population
646 bottlenecks in the last 1000 years should be discounted if sampling is clustered, but estimates of deeper
647 time patterns are likely to be fairly accurate. The biases in the SFS and haplotype structure identified
648 above (see also Wakeley 1999; Chikhi *et al.* 2010; Städler *et al.* 2009) are apparently small enough that
649 they fall within the range of variability regularly inferred by these approaches, at least on datasets of
650 the size we simulated.

651 **GWAS**

652 Spatial structure is particularly challenging for genome-wide association studies, because the effects of
653 dispersal on genetic variation are compounded by spatial variation in the environment (Mathieson
654 and McVean 2012). Spatially restricted mate choice and dispersal causes variation in allele frequencies
655 across the range of a species. If environmental factors affecting the phenotype of interest also vary over
656 space, then allele frequencies and environmental exposures will covary over space. In this scenario an
657 uncorrected GWAS will infer genetic associations with a purely environmental phenotype at any site
658 in the genome that is differentiated over space, and the relative degree of bias will be a function of the
659 degree of covariation in allele frequencies and the environment (i.e., Figure 8C, bottom panel). This
660 pattern has been demonstrated in a variety of simulation and empirical contexts (Price *et al.* 2006; Yu
661 *et al.* 2005; Young *et al.* 2018; Mathieson and McVean 2012; Kang *et al.* 2008, 2010; Bulik-Sullivan *et al.*
662 2015; Berg *et al.* 2018; Sohail *et al.* 2018).

663 Incorporating PC positions as covariates in a linear-regression GWAS (Price *et al.* 2006) is designed
664 to address this challenge by regressing out a baseline level of “average” differentiation. In essence, a
665 PC-corrected GWAS asks “what regions of the genome are more associated with this phenotype than
666 the average genome-wide association observed across populations?” In our simulations, we observed
667 that this procedure can fail under a variety of circumstances. If dispersal is limited and environmental
668 variation is clustered in space (i.e., corner or patchy distributions in our simulations), PC positions fail
669 to capture the fine-scale spatial structure required to remove all signals of association. Conversely, as
670 dispersal increases, PCA loses power to describe population structure before spatial mixing breaks
671 down the relationship between genotype and the environment. These effects were observed with all
672 spatially correlated environmental patterns, but were particularly pronounced if environmental effects
673 are concentrated in one region, as was also found by Mathieson and McVean (2012). Though increasing
674 the number of PC axes used in the analysis may reduce the false-positive rate, this may also decrease
675 the power of the test to detect truly causal alleles (Lawson *et al.* 2019).

676 In this work we simulated a single chromosome with size roughly comparable to one human
677 chromosome. If we scale the number of false-positive associations identified in our analyses to a
678 GWAS conducted on whole-genome data from humans, we would expect to see several thousand weak
679 false-positive associations after PC corrections in a population with neighborhood sizes up to at least
680 1000 (which should include values appropriate for many human populations). Notably, very few of the
681 spurious associations we identified would be significant at a conservative Bonferroni-adjusted *p*-value
682 cutoff (see Figure S8). This suggests that GWAS focused on finding strongly associated alleles for traits
683 controlled by a limited number of variants in the genome are likely robust to the impacts of continuous
684 spatial structure. However, methods that analyze the combined effects of thousands or millions of
685 weakly associated variants such as polygenic risk scores (Khera *et al.* 2018) are likely to be affected
686 by subtle population structure. Indeed as recently identified in studies of genotype associations for
687 human height in Europe (Berg *et al.* 2018; Sohail *et al.* 2018), PC regression GWAS in modern human
688 populations do include residual signal of population structure in large-scale analyses of polygenic traits.
689 When attempting to make predictions across populations with different environmental exposures,
690 polygenic risk scores affected by population structure can be expected to offer low predictive power, as
691 was shown in a recent study finding lower performance outside European populations (Martin *et al.*
692 2019).

693 In summary, spatial covariation in population structure and the environment confounds the in-
694 terpretation of GWAS *p*-values, and correction using principal components is insufficient to fully
695 separate these signals for polygenic traits under a variety of environmental and population parameter
696 regimes. Other GWAS methods such as mixed models (Kang *et al.* 2008) may be less sensitive to
697 this confounding, but there is no obvious reason that this should be so. One approach to estimating
698 the degree of bias in GWAS caused by population structure is LD score regression (Bulik-Sullivan
699 *et al.* 2015). Though this approach appears to work well in practice, its interpretation is not always
700 straightforward and it is likely biased by the presence of linked selection (Berg *et al.* 2018). In addition,
701 we observed that in many cases the false-positive SNPs we identified appeared to be concentrated in
702 LD peaks similar to those expected from truly causal sites (Figure S8), which may confound LD score

703 regression.

704 We suggest a straightforward alternative for species in which the primary axes of population
 705 differentiation is space (note this is likely not the case for some modern human populations): run a
 706 GWAS with spatial coordinates as phenotypes and check for *p*-value inflation or significant associations.
 707 If significant associations with sample locality are observed after correcting for population structure,
 708 the method is sensitive to false positives induced by spatial structure. This is essentially the approach
 709 taken in our “clinal” model (though we add normally distributed noise to our phenotypes). This
 710 approach has recently been taken with polygenic scores for UK Biobank samples in Haworth *et al.*
 711 (2019), finding that scores are correlated with birth location even in this relatively homogenous sample.
 712 Of course, it is possible that genotypes indirectly affect individual locations by adjusting organismal
 713 fitness and thus habitat selection across spatially varying environments, but we believe that this
 714 hypothesis should be tested against a null of stratification bias inflation rather than accepted as true
 715 based on GWAS results.

Table 1 Neighborhood size estimates from empirical studies.

Species	Description	Neighborhood Size	Method	Citation
<i>Ipomopsis aggregata</i>	flowering plant	12.60 - 37.80	Genetic	(Campbell and Dooley 1992)
<i>Borrichia frutescens</i>	salt marsh plant	20 - 30	Genetic+Survey	(Antlfinger 1982)
<i>Oreamnos americanus</i>	mountain goat	36 - 100	Genetic	(Shirk and Cushman 2014)
<i>Homo sapiens</i>	Gainj- and Kalam-speaking people, Papua New Guinea	40 - 213	Genetic	(Rousset 1997)
<i>Formica sp.</i>	colonial ants	50 - 100	Genetic	(Pamilo 1983)
<i>Astrocaryum mexicanum</i>	palm tree	102 - 895	Genetic+survey	(Eguiarte <i>et al.</i> 1993)
<i>Spermophilus mollis</i>	ground squirrel	204 - 480	Genetic+Survey	(Antolin <i>et al.</i> 2001)
<i>Sceloporus olivaceus</i>	lizard	225 - 270	Survey	(Kerster 1964)
<i>Dieffenbachia longispatha</i>	beetle-pollinated colonial herb	227 - 611	Survey	(Young 1988)
<i>Aedes aegypti</i>	Yellow-fever mosquito	268	Genetic	(Jasper <i>et al.</i> 2019)
<i>Homo sapiens</i>	Gainj- and Kalam-speaking people, Papua New Guinea	410	Survey	(Rousset 1997)
<i>Quercus laevis</i>	Oak tree	> 440	Genetic	(Berg and Hamrick 1995)
<i>Drosophila pseudoobscura</i>	fruit fly	500 - 1,000	Survey+Crosses	(Wright 1946)
<i>Homo sapiens</i>	POPRES data NE Europe	1,342 - 5,425	Genetic	(Ringbauer <i>et al.</i> 2017)
<i>Bebicium vittatum</i>	intertidal snail	240,000	Survey	(Rousset 1997)
<i>Bebicium vittatum</i>	intertidal snail	360,000	Genetic	(Rousset 1997)

716 **Where are natural populations on this spectrum?**

717 For how much of the tree of life do spatial patterns circumscribe genomic variation? In Table 1 we
 718 gathered estimates of neighborhood size from a range of organisms to get an idea of how strongly

719 local geographic dispersal affects patterns of variation. This is an imperfect measure: some aspects
720 of genetic variation are most strongly determined by neighborhood size (Wright 1946), others (e.g.,
721 number of segregating sites) are more strongly determined by global N_e or by the ratio of the two. In
722 addition, these empirical examples are likely biased towards small-neighborhood species (because few
723 studies have quantified neighborhood size in species with very high dispersal or population density).
724 However, from the available data we find that neighborhood sizes in the range we simulated are fairly
725 common across a range of taxa. At the extreme low end of empirical neighborhood size estimates
726 we see some flowering plants, large mammals, and colonial insects like ants. Species such as this
727 have neighborhood size estimates small enough that spatial processes are likely to strongly influence
728 inference. These include some human populations such as the Gainj- and Kalam-speaking people of
729 Papua New Guinea, in which the estimated neighborhood sizes in Rousset (1997) range from 40 to 410
730 depending on the method of estimation. Many more species occur in a middle range of neighborhood
731 sizes between 100 and 1000 – a range in which spatial processes play a minor role in our analyses
732 under random spatial sampling but are important when sampling of individuals in space is clustered.
733 Surprisingly, even some flying insects with huge census population sizes fall in this group, including
734 fruit flies (*D. melanogaster*) and mosquitoes (*A. aegypti*). Last, many species likely have neighborhood
735 sizes much larger than we simulated, including the recent ancestors of modern humans in northeastern
736 Europe (Ringbauer *et al.* 2017). For these species demographic inference and summary statistics are
737 likely to reflect minimal bias from spatial effects as long as dispersal is truly continuous across the
738 landscape. While that is so we caution that association studies in which the effects of population
739 structure are confounded with spatial variation in the environment are still sensitive to dispersal even
740 at these large neighborhood sizes.

741 **Other demographic models**

742 Any simulation of a population of reproducing organisms requires some kind of control on population
743 sizes, or else the population will either die out or grow very large after a sufficiently long period
744 of time. The usual choice of population regulation for population genetics – a constant size, as
745 in the Wright–Fisher model – implies biologically unrealistic interactions between geographically
746 distant parts of the species range. Our choice to regulate population size by including a local density-
747 dependent control on mortality is only one of many possible ways to do this. We could have instead
748 regulated fecundity, or recruitment, or both; this general class of models is sometimes referred to as
749 the “Bolker–Pacala model” (Bolker and Pacala 1997). It is not currently clear how much different
750 choices of demographic parameters, or of functional forms for the regulation, might quantitatively
751 affect our results, although the general predictions should be robust to similar forms of regulation.
752 Since populations are still entirely *intrinsically* regulated, our model still has a very strong “population
753 genetics” flavor. Alternatively, population size could be regulated by interactions with other species
754 (e.g., a Lotka–Volterra model), or extrinsically specified by local resource availability (e.g., by food
755 or nest site availability). Indeed, our model could be interpreted as a caricature of such a model: as
756 local density increases, good habitat is increasingly occupied, pushing individuals into more marginal
757 habitat and increasing their mortality. Many such models should behave similarly to ours, but others
758 (especially those with local population cycling), may differ dramatically.

759 Population genetic simulations often use grids of discrete demes, which are assumed to approximate
760 continuous space. However, there are theoretical reasons to expect that increasingly fine grids of
761 discrete demes do not approach the continuous model (Barton *et al.* 2002). If continuous space can
762 be approximated by a limit of discrete models, this should be true regardless of the precise details
763 of the discrete model. Although we carefully chose parameters to match our continuous models, we
764 found that some aspects of genetic variation diverged from the continuous case as the discretization
765 got finer. This suggests that these models do not converge in the limit. However, many populations
766 may indeed be well-modeled as a series of discrete, randomly-mating demes if, for example, suitable
767 habitats are patchily distributed across the landscape. There is a clear need for greater exploration of
768 the consequences for population genetics of ecologically realistic population models.

769 **Future Directions and Limitations**

770 As we have shown, a large number of population genetic summary statistics contain information about
771 spatial population processes. We imagine that combinations of such summaries might be sufficient
772 for the construction of supervised machine learning regressors (e.g., Schrider and Kern 2018) for the
773 accurate estimation of dispersal from genetic data. Indeed, Ashander *et al.* (2018) found that inverse
774 interpolation on a vector of summary statistics provided a powerful method of estimating dispersal
775 distances. Expanding this approach to include the haplotype-based summary statistics studied here
776 and applying machine learning regressors built for general inference of nonlinear relationships from
777 high-dimensional data may allow precise estimation of spatial parameters under a range of complex
778 models.

779 One facet of spatial variation that we did not address in this study is the confounding of dispersal
780 and population density implicit in the definition of Wright's neighborhood size. Our simulations were
781 run under constant densities, but Guindon *et al.* (2016) and Ringbauer *et al.* (2017) have shown that
782 these parameters are identifiable under some continuous models. Similarly, though the scaling effects
783 of dispersal we show in Figure 4 should occur in populations of any total size, other aspects such as
784 the number of segregating sites are also likely affected by the total landscape size (and so total census
785 size). Much additional work remains to be done to better understand how these parameters interact to
786 shape genetic variation in continuous space, which we leave to future studies.

787 Though our simulation allows incorporation of realistic demographic and spatial processes, it
788 is inevitably limited by the computational burden of tracking tens or hundreds of thousands of
789 individuals in every generation. In particular, computations required for mate selection and spatial
790 competition scale approximately with the product of the total census size and the neighborhood
791 size and so increase rapidly for large populations and dispersal distances. The reverse-time spatial
792 Lambda–Fleming–Viot model described by Barton *et al.* (2010) and implemented by Kelleher *et al.*
793 (2014) allows exploration of larger population and landscape sizes, but the precise connection of these
794 models to forward-time demography is not yet clear. Alternatively, implementation of parallelized
795 calculations may allow progress with forward-time simulations.

796 Finally, we believe that the difficulties in correcting for population structure in continuous popula-
797 tions using principal components analysis or similar decompositions is a difficult issue, well worth
798 considering on its own. How can we best avoid spurious correlations while correlating genetic and
799 phenotypic variation without underpowering the methods? Perhaps optimistically, we posit that
800 process-driven descriptions of ancestry and/or more generalized unsupervised methods may be able
801 to better account for carry out this task.

802 **Data Availability**

803 Scripts used for all analyses and figures are available at <https://github.com/kern-lab/spaceness>.

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807 **Literature Cited**

- 808 Aguillon, S. M., J. W. Fitzpatrick, R. Bowman, S. J. Schoech, A. G. Clark, *et al.*, 2017 Deconstructing
809 isolation-by-distance: The genomic consequences of limited dispersal. PLOS Genetics 13: 1–27.
810 Al-Asadi, H., D. Petkova, M. Stephens, and J. Novembre, 2019 Estimating recent migration and
811 population-size surfaces. PLoS genetics 15: e1007908.
812 Allee, W. C., O. Park, A. E. Emerson, T. Park, K. P. Schmidt, *et al.*, 1949 Principles of animal ecology.
813 Technical report, Saunders Company Philadelphia, Pennsylvania, USA.
814 Antlfinger, A. E., 1982 Genetic neighborhood structure of the salt marsh composite, *Borrichia frutescens*.
815 Journal of Heredity 73: 128–132.

- 816 Antolin, M. F., B. V. Horne, M. D. Berger, Jr., A. K. Holloway, J. L. Roach, *et al.*, 2001 Effective population
817 size and genetic structure of a piute ground squirrel (*Spermophilus mollis*) population. Canadian
818 Journal of Zoology **79**: 26–34.
- 819 Antonovics, J. and D. A. Levin, 1980 The ecological and genetic consequences of density-dependent
820 regulation in plants. Annual Review of Ecology and Systematics **11**: 411–452.
- 821 Ashander, J., P. Ralph, E. McCartney-Melstad, and H. B. Shaffer, 2018 Demographic inference in a
822 spatially-explicit ecological model from genomic data: a proof of concept for the mojave desert
823 tortoise. bioRxiv .
- 824 Baharian, S., M. Barakatt, C. R. Gignoux, S. Shringarpure, J. Errington, *et al.*, 2016 The great migration
825 and African-American genomic diversity. PLOS Genetics **12**: 1–27.
- 826 Barton, N. H., F. Depaulis, and A. M. Etheridge, 2002 Neutral evolution in spatially continuous
827 populations. Theoretical Population Biology **61**: 31–48.
- 828 Barton, N. H., J. Kelleher, and A. M. Etheridge, 2010 A new model for extinction and recolonization in
829 two dimensions: Quantifying phylogeography. Evolution **64**: 2701–2715.
- 830 Benjamini, Y. and D. Yekutieli, 2001 The control of the false discovery rate in multiple testing under
831 dependency. The Annals of Statistics **29**: 1165–1188.
- 832 Berg, E. E. and J. L. Hamrick, 1995 Fine-scale genetic structure of a turkey oak forest. Evolution **49**:
833 110–120.
- 834 Berg, J. J., A. Harpak, N. Sinnott-Armstrong, A. M. Joergensen, H. Mostafavi, *et al.*, 2018 Reduced
835 signal for polygenic adaptation of height in UK Biobank. bioRxiv .
- 836 Bolker, B. and S. W. Pacala, 1997 Using moment equations to understand stochastically driven spatial
837 pattern formation in ecological systems. Theoretical Population Biology **52**: 179 – 197.
- 838 Bolker, B. M., S. W. Pacala, and C. Neuhauser, 2003 Spatial dynamics in model plant communities:
839 What do we really know? The American Naturalist **162**: 135–148, PMID: 12858259.
- 840 Browning, S. R. and B. L. Browning, 2015 Accurate non-parametric estimation of recent effective
841 population size from segments of identity by descent. The American Journal of Human Genetics **97**:
842 404–418.
- 843 Bulik-Sullivan, B. K., P.-R. Loh, H. K. Finucane, S. Ripke, J. Yang, *et al.*, 2015 LD score regression
844 distinguishes confounding from polygenicity in genome-wide association studies. Nature Genetics
845 **47**: 291 EP –.
- 846 Campbell, D. R. and J. L. Dooley, 1992 The spatial scale of genetic differentiation in a hummingbird-
847 pollinated plant: Comparison with models of isolation by distance. The American Naturalist **139**:
848 735–748.
- 849 Champer, J., I. Kim, S. E. Champer, A. G. Clark, and P. W. Messer, 2019 Suppression gene drive in
850 continuous space can result in unstable persistence of both drive and wild-type alleles. bioRxiv .
- 851 Chapman, N. H. and E. A. Thompson, 2002 The effect of population history on the lengths of ancestral
852 chromosome segments. Genetics **162**: 449–458.
- 853 Chikhi, L., V. C. Sousa, P. Luisi, B. Goossens, and M. A. Beaumont, 2010 The confounding effects of
854 population structure, genetic diversity and the sampling scheme on the detection and quantification
855 of population size changes. Genetics **186**: 983–995.
- 856 Crawley, M. J., 1990 The population dynamics of plants. Philosophical Transactions of the Royal Society
857 of London. Series B: Biological Sciences **330**: 125–140.
- 858 Durrett, R. and S. Levin, 1994 The importance of being discrete (and spatial). Theoretical Population
859 Biology **46**: 363–394.
- 860 Eguiarte, L. E., A. Búrquez, J. Rodríguez, M. Martínez-Ramos, J. Sarukhán, *et al.*, 1993 Direct and
861 indirect estimates of neighborhood and effective population size in a tropical palm, *Astrocaryum*
862 *mexicanum*. Evolution **47**: 75–87.
- 863 Epperson, B., 2003 *Geographical Genetics*. Monographs in Population Biology, Princeton University
864 Press.
- 865 Felsenstein, J., 1975 A pain in the torus: Some difficulties with models of isolation by distance. The
866 American Naturalist **109**: 359–368.
- 867 Fournier, N. and S. Méléard, 2004 A microscopic probabilistic description of a locally regulated
868 population and macroscopic approximations. The Annals of Applied Probability **14**: 1880–1919.

- 869 Fox, J. and S. Weisberg, 2011 *An R Companion to Applied Regression*. Sage, Thousand Oaks CA, second
870 edition.
- 871 Garcia, J. and C. Quintana-Domeque, 2006 The evolution of adult height in europe: A brief note.
872 Working Paper .
- 873 Garcia, J. and C. Quintana-Domeque, 2007 The evolution of adult height in Europe: A brief note.
874 Economics & Human Biology 5: 340 – 349.
- 875 Garud, N. R., P. W. Messer, E. O. Buzbas, and D. A. Petrov, 2015 Recent selective sweeps in North
876 American *Drosophila melanogaster* show signatures of soft sweeps. PLOS Genetics 11: 1–32.
- 877 Griffiths, R., S. Tavaré, *et al.*, 1999 The ages of mutations in gene trees. The Annals of Applied Probability
878 9: 567–590.
- 879 Guindon, S., H. Guo, and D. Welch, 2016 Demographic inference under the coalescent in a spatial
880 continuum. Theoretical population biology 111: 43–50.
- 881 Haller, B. C., J. Galloway, J. Kelleher, P. W. Messer, and P. L. Ralph, 2019 Tree-sequence recording
882 in SLiM opens new horizons for forward-time simulation of whole genomes. Molecular Ecology
883 Resources 19: 552–566.
- 884 Haller, B. C. and P. W. Messer, 2019 Slim 3: Forward genetic simulations beyond the Wright–Fisher
885 model. Molecular biology and evolution 36: 632–637.
- 886 Harris, K. and R. Nielsen, 2013 Inferring demographic history from a spectrum of shared haplotype
887 lengths. PLOS Genetics 9: 1–20.
- 888 Haworth, S., R. Mitchell, L. Corbin, K. H. Wade, T. Dudding, *et al.*, 2019 Apparent latent structure within
889 the UK Biobank sample has implications for epidemiological analysis. Nature communications 10:
890 333.
- 891 Huillet, T. and M. Möhle, 2011 On the extended Moran model and its relation to coalescents with
892 multiple collisions. Theoretical Population Biology pp. –.
- 893 Jackson, N. D. and L. Fahrig, 2014 Landscape context affects genetic diversity at a much larger spatial
894 extent than population abundance. Ecology 95: 871–881.
- 895 Jasper, M., T. Schmidt, N. Ahmad, S. Sinkins, and A. Hoffmann, 2019 A genomic approach to inferring
896 kinship reveals limited intergenerational dispersal in the yellow fever mosquito. bioRxiv .
- 897 Jay, F., P. Sjödin, M. Jakobsson, and M. G. Blum, 2012 Anisotropic isolation by distance: The main
898 orientations of human genetic differentiation. Molecular Biology and Evolution 30: 513–525.
- 899 Kang, H. M., J. H. Sul, S. K. Service, N. A. Zaitlen, S.-y. Kong, *et al.*, 2010 Variance component model to
900 account for sample structure in genome-wide association studies. Nature Genetics 42: 348 EP –.
- 901 Kang, H. M., N. A. Zaitlen, C. M. Wade, A. Kirby, D. Heckerman, *et al.*, 2008 Efficient control of
902 population structure in model organism association mapping. Genetics 178: 1709–1723.
- 903 Kelleher, J., A. Etheridge, and N. Barton, 2014 Coalescent simulation in continuous space: Algorithms
904 for large neighbourhood size. Theoretical Population Biology 95: 13 – 23.
- 905 Kelleher, J., A. M. Etheridge, and G. McVean, 2016 Efficient coalescent simulation and genealogical
906 analysis for large sample sizes. PLoS Comput Biol 12: 1–22.
- 907 Kelleher, J., K. R. Thornton, J. Ashander, and P. L. Ralph, 2018 Efficient pedigree recording for fast
908 population genetics simulation. PLOS Computational Biology 14: 1–21.
- 909 Kerster, H. W., 1964 Neighborhood size in the rusty lizard, *Sceloporus olivaceus*. Evolution 18: 445–457.
- 910 Khera, A. V., M. Chaffin, K. G. Aragam, M. E. Haas, C. Roselli, *et al.*, 2018 Genome-wide polygenic
911 scores for common diseases identify individuals with risk equivalent to monogenic mutations.
912 Nature Genetics 50: 1219–1224.
- 913 Kingman, J., 1982 The coalescent. Stochastic Processes and their Applications 13: 235 – 248.
- 914 Law, R., D. J. Murrell, and U. Dieckmann, 2003 Population growth in space and time: Spatial logistic
915 equations. Ecology 84: 252–262.
- 916 Lawson, D. J., N. M. Davies, S. Haworth, B. Ashraf, L. Howe, *et al.*, 2019 Is population structure in the
917 genetic biobank era irrelevant, a challenge, or an opportunity? Human Genetics .
- 918 Liu, X. and Y.-X. Fu, 2015 Exploring population size changes using SNP frequency spectra. Nature
919 Genetics 47: 555 EP –.
- 920 Lloyd, M., 1967 ‘Mean crowding’. Journal of Animal Ecology 36: 1–30.
- 921 Lundgren, E. and P. L. Ralph, 2019 Are populations like a circuit? Comparing isolation by resistance to

- 922 a new coalescent-based method. *Molecular Ecology Resources* **19**: 1388–1406.
- 923 Martin, A. R., M. Kanai, Y. Kamatani, Y. Okada, B. M. Neale, *et al.*, 2019 Clinical use of current polygenic
924 risk scores may exacerbate health disparities. *Nature Genetics* **51**: 584–591.
- 925 Maruyama, T., 1972 Rate of decrease of genetic variability in a two-dimensional continuous population
926 of finite size. *Genetics* **70**: 639–651.
- 927 Mathieson, I. and G. McVean, 2012 Differential confounding of rare and common variants in spatially
928 structured populations. *Nature Genetics* **44**: 243 EP –.
- 929 Mazet, O., W. Rodríguez, S. Grusea, S. Boitard, and L. Chikhi, 2015 On the importance of being
930 structured: instantaneous coalescence rates and human evolution—lessons for ancestral population
931 size inference? *Heredity* **116**: 362 EP –.
- 932 Miles, A. and N. Harding, 2017 *cghg/scikit-allel*: v1.1.8.
- 933 Neel, M. C., K. McKelvey, N. Ryman, M. W. Lloyd, R. Short Bull, *et al.*, 2013 Estimation of effective
934 population size in continuously distributed populations: there goes the neighborhood. *Heredity* **111**:
935 189 EP –.
- 936 Novembre, J. and M. Slatkin, 2009 Likelihood-based inference in isolation-by-distance models using
937 the spatial distribution of low-frequency alleles. *Evolution* **63**: 2914–2925.
- 938 Pamilo, P., 1983 Genetic differentiation within subdivided populations of formica ants. *Evolution* **37**:
939 1010–1022.
- 940 Patterson, N., A. L. Price, and D. Reich, 2006 Population structure and eigenanalysis. *PLOS Genetics* **2**:
941 1–20.
- 942 Peter L Ralph, J. G., Jerome Kelleher and J. Ashander, ???? .
- 943 Petkova, D., J. Novembre, and M. Stephens, 2015 Visualizing spatial population structure with esti-
944 mated effective migration surfaces. *Nature Genetics* **48**: 94 EP –.
- 945 Price, A. L., N. J. Patterson, R. M. Plenge, M. E. Weinblatt, N. A. Shadick, *et al.*, 2006 Principal
946 components analysis corrects for stratification in genome-wide association studies. *Nature Genetics*
947 **38**: 904 EP –.
- 948 Pritchard, J. K., M. Stephens, and P. Donnelly, 2000 Inference of population structure using multilocus
949 genotype data. *Genetics* **155**: 945–959.
- 950 Ptak, S. E. and M. Przeworski, 2002 Evidence for population growth in humans is confounded by
951 fine-scale population structure. *Trends in Genetics* **18**: 559–563.
- 952 Purcell, S., B. Neale, K. Todd-Brown, L. Thomas, M. A. Ferreira, *et al.*, 2007 Plink: A tool set for
953 whole-genome association and population-based linkage analyses. *The American Journal of Human
954 Genetics* **81**: 559 – 575.
- 955 R Core Team, 2018 *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical
956 Computing, Vienna, Austria.
- 957 Ralph, P. and G. Coop, 2013 The geography of recent genetic ancestry across Europe. *PLoS Biol* **11**:
958 e1001555.
- 959 Ralph, P., K. Thornton, and J. Kelleher, 2019 Efficiently summarizing relationships in large samples: a
960 general duality between statistics of genealogies and genomes. *bioRxiv* .
- 961 Ringbauer, H., G. Coop, and N. H. Barton, 2017 Inferring recent demography from isolation by distance
962 of long shared sequence blocks. *Genetics* **205**: 1335–1351.
- 963 Robledo-Arnuncio, J. J. and F. Rousset, 2010 Isolation by distance in a continuous population under
964 stochastic demographic fluctuations. *Journal of Evolutionary Biology* **23**: 53–71.
- 965 Rossine, F. W. S., 2014 *Espaço e diversificação: uma perspectiva teórica*. Master's dissertation in ecologia:
966 Ecossistemas terrestres e aquáticos, University of São Paulo, São Paulo : Instituto de Biociências.
- 967 Rousset, F., 1997 Genetic differentiation and estimation of gene flow from F-statistics under isolation
968 by distance. *Genetics* **145**: 1219–1228.
- 969 Rousset, F. and R. Leblois, 2011 Likelihood-based inferences under isolation by distance: Two-
970 dimensional habitats and confidence intervals. *Molecular Biology and Evolution* **29**: 957–973.
- 971 Sawyer, S., 1977 On the past history of an allele now known to have frequency p. *Journal of Applied
972 Probability* **14**: 439–450.
- 973 Schiffels, S. and R. Durbin, 2014 Inferring human population size and separation history from multiple
974 genome sequences. *Nature Genetics* **46**: 919 EP –.

- 975 Schrider, D. R. and A. D. Kern, 2018 Supervised machine learning for population genetics: A new
 976 paradigm. *Trends in Genetics* **34**: 301 – 312.
- 977 Sharbel, T. F., B. Haubold, and T. Mitchell-Olds, 2000 Genetic isolation by distance in *Arabidopsis*
 978 *thaliana*: biogeography and postglacial colonization of Europe. *Molecular Ecology* **9**: 2109–2118.
- 979 Sheehan, S., K. Harris, and Y. S. Song, 2013 Estimating variable effective population sizes from multiple
 980 genomes: A sequentially Markov conditional sampling distribution approach. *Genetics* **194**: 647–662.
- 981 Shirk, A. J. and S. A. Cushman, 2014 Spatially-explicit estimation of Wright’s neighborhood size in
 982 continuous populations. *Frontiers in Ecology and Evolution* **2**: 62.
- 983 Slatkin, M. and N. H. Barton, 1989 A comparison of three indirect methods for estimating average
 984 levels of gene flow. *Evolution* **43**: 1349–1368.
- 985 Sohail, M., R. M. Maier, A. Ganna, A. Bloemendal, A. R. Martin, *et al.*, 2018 Signals of polygenic
 986 adaptation on height have been overestimated due to uncorrected population structure in genome-
 987 wide association studies. *bioRxiv*.
- 988 St. Onge, K. R., A. E. Palmé, S. I. Wright, and M. Lascoux, 2012 Impact of sampling schemes on
 989 demographic inference: An empirical study in two species with different mating systems and
 990 demographic histories. *G3: Genes, Genomes, Genetics* **2**: 803–814.
- 991 Städler, T., B. Haubold, C. Merino, W. Stephan, and P. Pfaffelhuber, 2009 The impact of sampling
 992 schemes on the site frequency spectrum in nonequilibrium subdivided populations. *Genetics* **182**:
 993 205–216.
- 994 Tajima, F., 1989 Statistical method for testing the neutral mutation hypothesis by DNA polymorphism.
 995 *Genetics* **123**: 585–595.
- 996 Terhorst, J., J. A. Kamm, and Y. S. Song, 2016 Robust and scalable inference of population history from
 997 hundreds of unphased whole genomes. *Nature Genetics* **49**: 303 EP –.
- 998 Turchin, M. C., C. W. Chiang, C. D. Palmer, S. Sankararaman, D. Reich, *et al.*, 2012 Evidence of
 999 widespread selection on standing variation in Europe at height-associated SNPs. *Nature Genetics* **44**:
 1000 1015 EP –.
- 1001 Wahlund, S., 1928 Zusammensetzung von populationen und korrelationserscheinungen vom stand-
 1002 punkt der vererbungslehre aus betrachtet. *Hereditas* **11**: 65–106.
- 1003 Wakeley, J., 1999 Nonequilibrium migration in human history. *Genetics* **153**: 1863–1871.
- 1004 Wakeley, J., 2009 *Coalescent Theory, an Introduction*. Roberts and Company, Greenwood Village, CO.
- 1005 Wakeley, J. and T. Takahashi, 2003 Gene genealogies when the sample size exceeds the effective size of
 1006 the population. *Mol Biol Evol* **20**: 208–213.
- 1007 Wickham, H., 2016 *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York.
- 1008 Wilke, C. O., 2019 *cowplot: Streamlined Plot Theme and Plot Annotations for ‘ggplot2’*. R package version
 1009 0.9.4.
- 1010 Wilkins, J. F., 2004 A separation-of-timescales approach to the coalescent in a continuous population.
 1011 *Genetics* **168**: 2227–2244.
- 1012 Wilkins, J. F. and J. Wakeley, 2002 The coalescent in a continuous, finite, linear population. *Genetics*
 1013 **161**: 873–888.
- 1014 Wright, S., 1931 Evolution in Mendelian populations. *Genetics* **16**: 97.
- 1015 Wright, S., 1943 Isolation by distance. *Genetics* **28**: 114–138.
- 1016 Wright, S., 1946 Isolation by distance under diverse systems of mating. *Genetics* **31**: 336.
- 1017 Young, A. I., M. L. Frigge, D. F. Gudbjartsson, G. Thorleifsson, G. Bjornsdottir, *et al.*, 2018 Relatedness
 1018 disequilibrium regression estimates heritability without environmental bias. *Nature Genetics* **50**:
 1019 1304–1310.
- 1020 Young, H. J., 1988 Neighborhood size in a beetle pollinated tropical aroid: effects of low density and
 1021 asynchronous flowering. *Oecologia* **76**: 461–466.
- 1022 Yu, J., G. Pressoir, W. H. Briggs, I. Vroh Bi, M. Yamasaki, *et al.*, 2005 A unified mixed-model method for
 1023 association mapping that accounts for multiple levels of relatedness. *Nature Genetics* **38**: 203 EP –.

1024 **Comparisons with Stepping-Stone Models**

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We also compared our model results to a regular grid of discrete populations, which is commonly used as an approximation of continuous geography. An important reason that this approximation is often made is that it allows more efficient, coalescent simulations; we implemented these using msprime (Kelleher *et al.* 2016). In this class of models we imagine an $n \times n$ grid of populations exchanging migrants with neighboring populations at rate m . If these models are good approximations of the continuous case we expect that results will converge as $n \rightarrow \infty$ (while scaling m appropriately and keeping total population size fixed), so we ran simulations while varying n from 5 to 50 (Table A1). To compare with continuous models we first distributed the same “effective” number of individuals across the landscape as in our continuous-space simulations (≈ 6100 , estimated from θ_π of random-mating continuous-space simulations). We then approximate the mean per-generation dispersal distance σ given a total landscape width W as the product of the probability of an individual being a migrant and the distance traveled by migrants: $\sigma = 4m(W/n)$. This means that m in different simulations with the same σ scales with \sqrt{n} . We ran 500 simulations for each value of n while sampling σ from $U(0.2, 4)$. We then randomly selected 60 diploid individuals from each simulation (approximating diploidy by combining pairs of chromosomes with contiguous indices within demes) and calculated a set of six summary statistics using the scripts described in the summary statistics portion of the main text.

demes per side (n)	N_e per deme	samples per deme
5	244	20
10	61	10
20	15.25	2
50	2.44	1

Table A1 stepping-stone simulation parameters

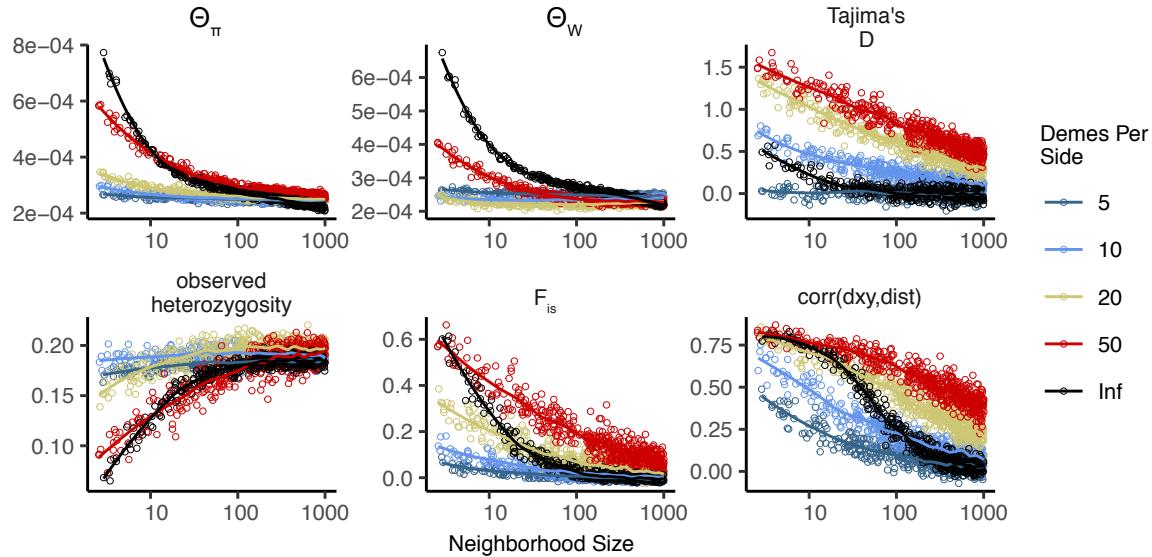


Figure A1 Summary statistics for 2-dimensional coalescent stepping-stone models with fixed total N_e and varying numbers of demes per side. The black “infinite” points are from our forward-time continuous space model. Inter-deme migration rates are related to σ as described above.

In general we find many of the qualitative trends are similar among continuous and stepping-stone

1044 models and that, at low neighborhood sizes, many (but not all) statistics from stepping-stone models
 1045 approach the continuous model as the resolution of the grid increases. For example, θ_π is inflated at low
 1046 neighborhood sizes (i.e., low m), and the extent of the inflation increases to approach the continuous
 1047 case as the resolution of the landscape increases. Similar patterns are observed for F_{IS} and observed
 1048 heterozygosity. However, θ_W behaves differently, showing a non-monotonic relationship with grid
 1049 resolution. This results in an increasingly positive Tajima's D in grid simulations at small neighborhood
 1050 sizes, to a much greater extent than seen in a continuous model. In contrast to θ_π , increasing the
 1051 resolution of the grid causes Tajima's D to deviate *more* from what is seen in the continuous model.

1052 These differences relative to our continuous model mainly reflect two shortcomings of the reverse-
 1053 time stepping stone model. If we simulate a coarse grid with relatively large populations in each
 1054 deme, we cannot accurately capture the dynamics of small neighborhood sizes because mating within
 1055 each deme remains random regardless of the migration rate connecting demes. This likely explains
 1056 the trends in θ_π , observed heterozygosity, and F_{IS} . However increasing the number of demes while
 1057 holding the total number of individuals constant results in small within-deme populations for which
 1058 even the minimum sample size of 1 approaches the local N_e (Table A1). This results in an excess
 1059 of short terminal branches in the coalescent tree, which decreases the total branch length and leads
 1060 to fewer segregating sites, deflated θ_W , and inflated Tajima's D . Overall, the stepping-stone model
 1061 reproduces important features of spatial structure in our continuous space model, such as a decline in
 1062 θ_π and correlations between spatial and genetic distance with increasing migration, but introduces
 1063 artifacts caused by binning the landscape into discrete demes.

1064 Demographic model

1065 Local population regulation is controlled by two parameters, L , and K . Here, we show that these
 1066 should be close to the average lifespan of an individual and the average number of individuals per
 1067 unit area, respectively. We chose our demographic model so that every individual has on average $1/L$
 1068 offspring each time step, and if the local population density of an individual is n , then their probability
 1069 of survival until the next time step is (equation (1)):

$$p = \min \left(0.95, \frac{1}{1 + n/(K(1 + L))} \right). \quad (3)$$

1070 We capped survival at 0.95 so that we would not have exceptionally long-lived individuals in sparsely
 1071 populated areas – otherwise, an isolated individual might live for a very long time. Since $1 - p \approx$
 1072 $n/(K(1 + L))$, mortality goes up roughly linearly with the number of neighbors (on a scale given by
 1073 K), as would be obtained if, for instance, mortality is due to agonistic interactions. Ignoring migration,
 1074 a region is at demographic equilibrium if the per-capita probability of death is equal to the birth rate,
 1075 i.e., if $1 - p = 1/L$. (Note that there is no effect of age in the model, which would make the analysis
 1076 more complicated.) Solving this for n , we get that in a well-mixed population, the equilibrium density
 1077 should be around

$$n = K \frac{L + 1}{L - 1} \quad (4)$$

1078 individuals per unit area. At this density, the per-capita death rate is $1/L$, so the mean lifetime is L .
 1079 This equilibrium density is *not* K , but (since $L = 4$) is two-thirds larger. However, in practice this model
 1080 leads to a total population size which is around K multiplied by total geographic area (but which
 1081 depends on σ , as discussed above). The main reason for this is that since offspring tend to be near
 1082 their parents, individuals tend to be “clumped”, and so experience a higher average density than the
 1083 “density” one would compute by dividing census size by geographic area (Lloyd 1967). To maintain a
 1084 constant expected total population size would require making (say) K depend on σ ; however, typical
 1085 local population densities might then be more dissimilar.

1086 **Supplementary Figures and Tables**

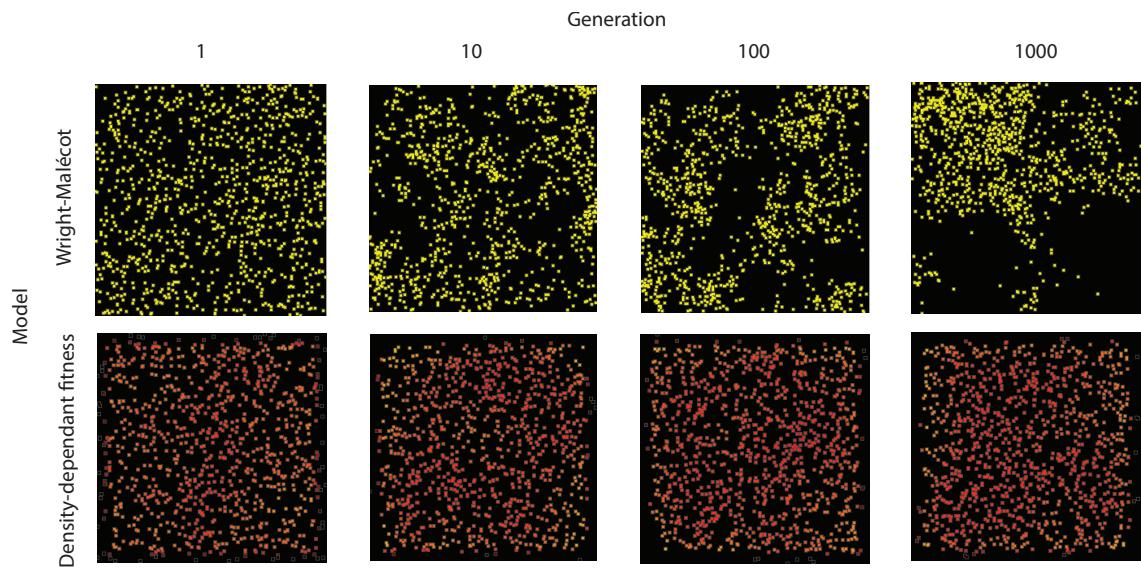


Figure S1 Maps of individual locations in a continuous-space Wright-Malécot model with independent dispersal of all individuals (top) and under our continuous space model incorporating density-dependant fitness (bottom). The clustering seen in the top row is the “Pain in the Torus” described by Felsenstein (1975).

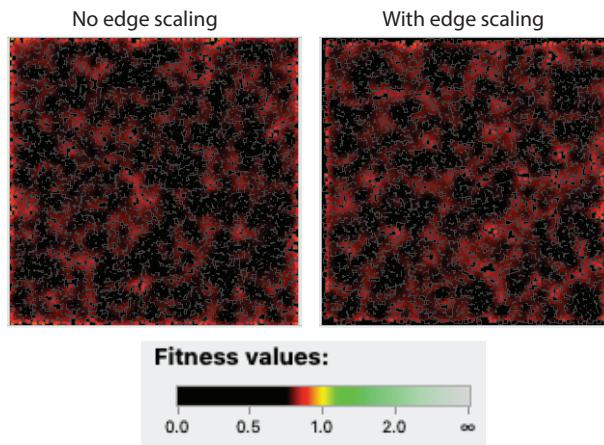


Figure S2 Comparison of individual fitness across the landscape in simulations with (right) and without (left) a decline in fitness approaching range edges. Note the slight excess of high-fitness individuals at edges on the left, which is (partially) counteracted by the scaling procedure.

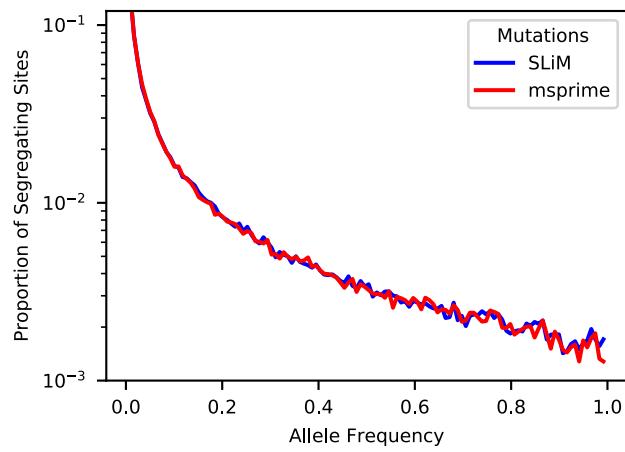


Figure S3 Site frequency spectra from a simulation with neighborhood size = 12.5 when mutations are recorded directly in SLiM (blue line) or applied later in msprime (red line).

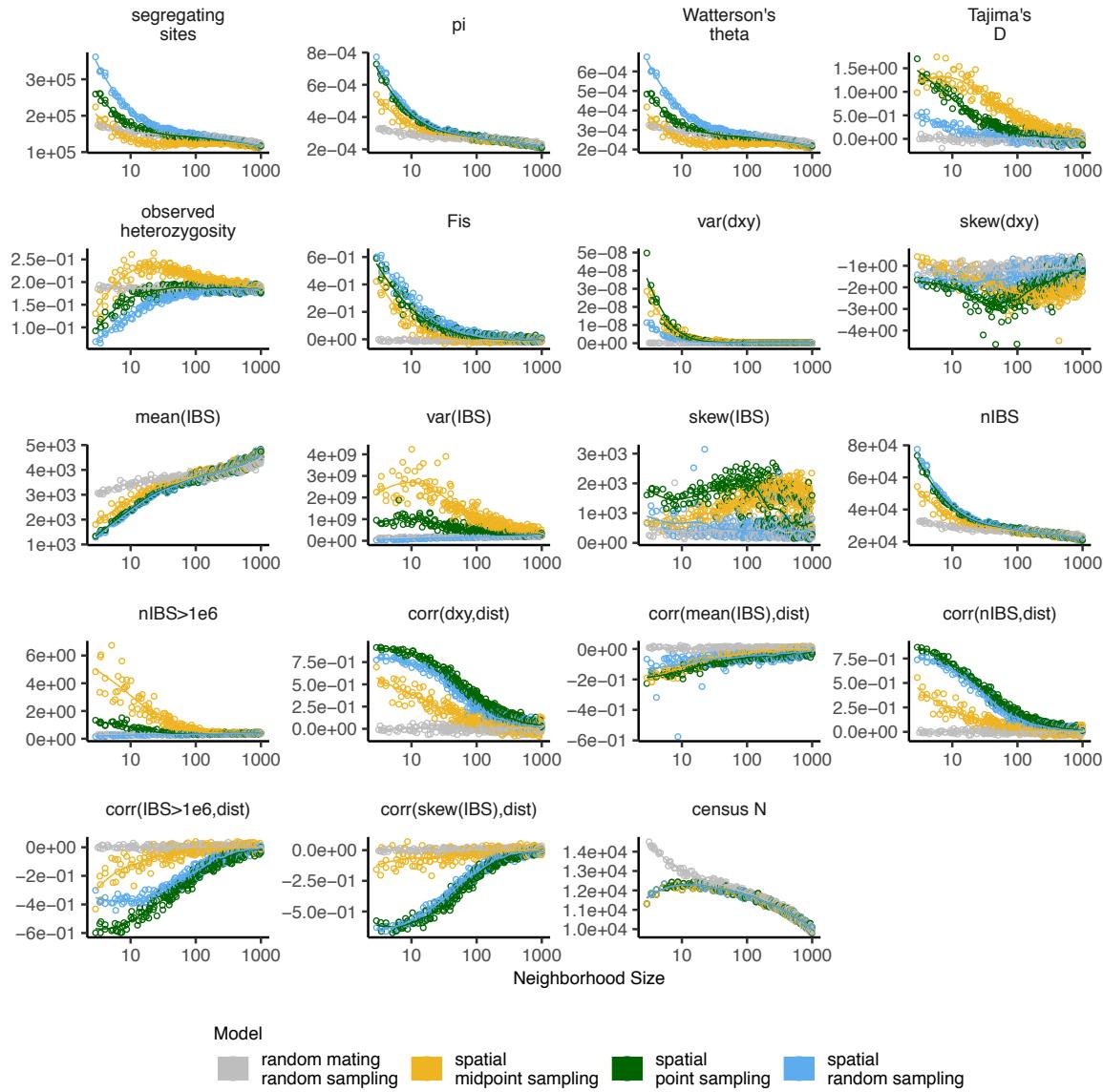


Figure S4 Change in summary statistics by neighborhood size and sampling scheme calculated from simulated sequence data of 60 individuals.

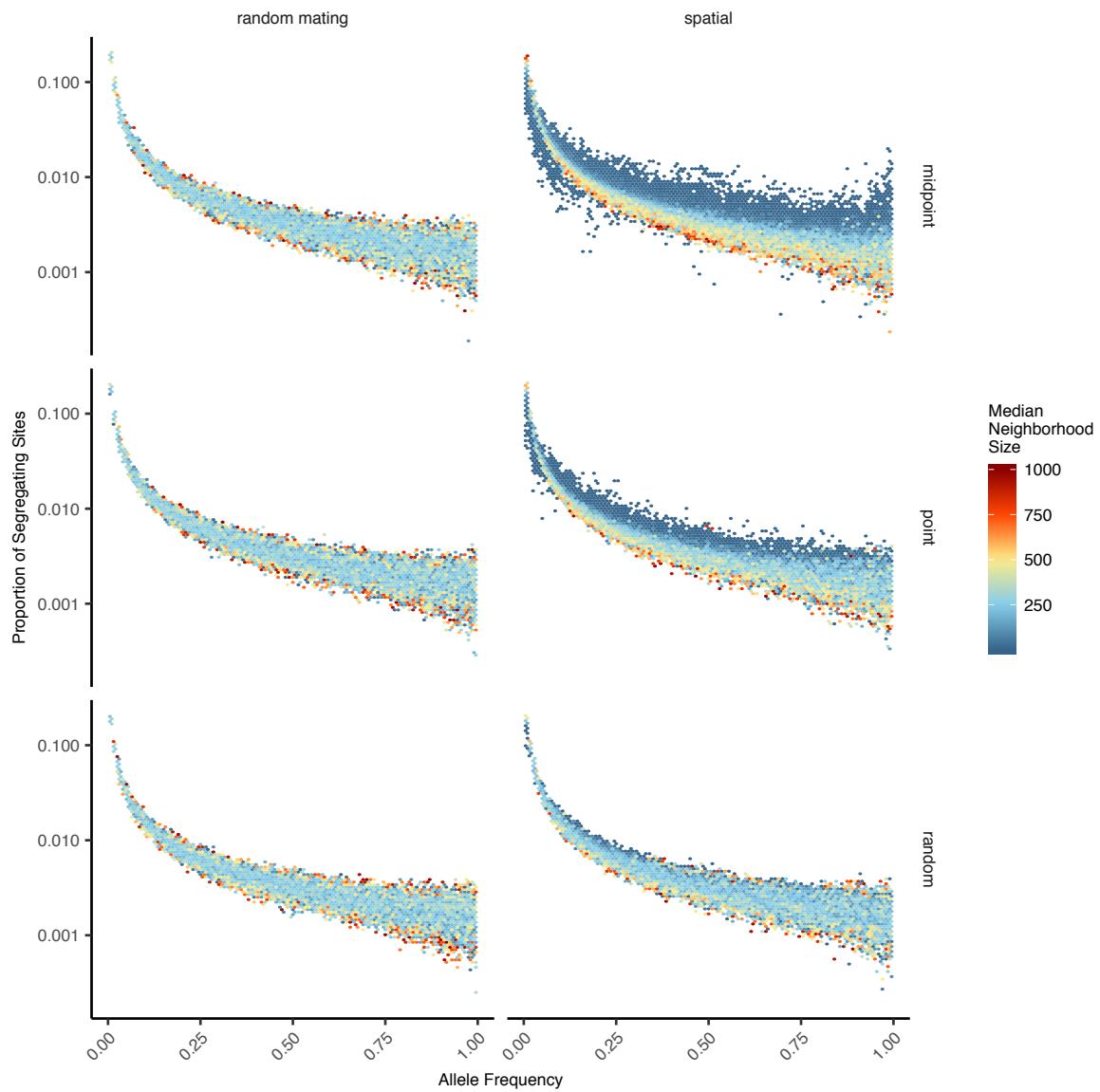


Figure S5 Site frequency spectra for random mating and spatial SLiM models under all sampling schemes.

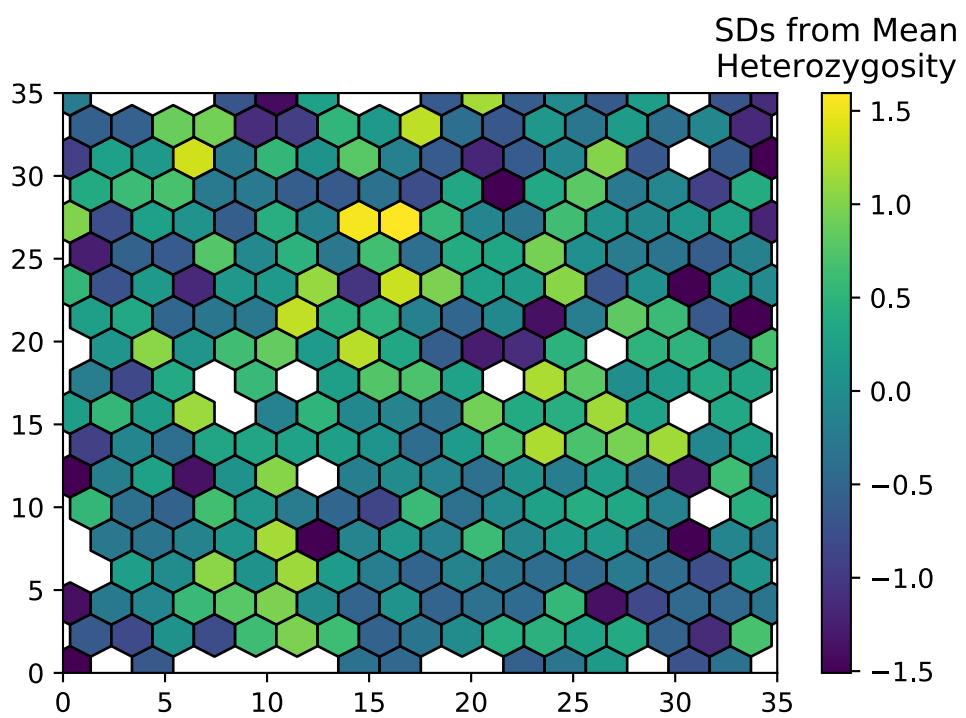


Figure S6 Variation in observed heterozygosity (i.e. proportion of heterozygous individuals) in hexagonal bins across the landscape, estimated from a random sample of 200 individuals from the final generation of a simulation with neighborhood size ≈ 25 . Values were Z-normalized for plotting.

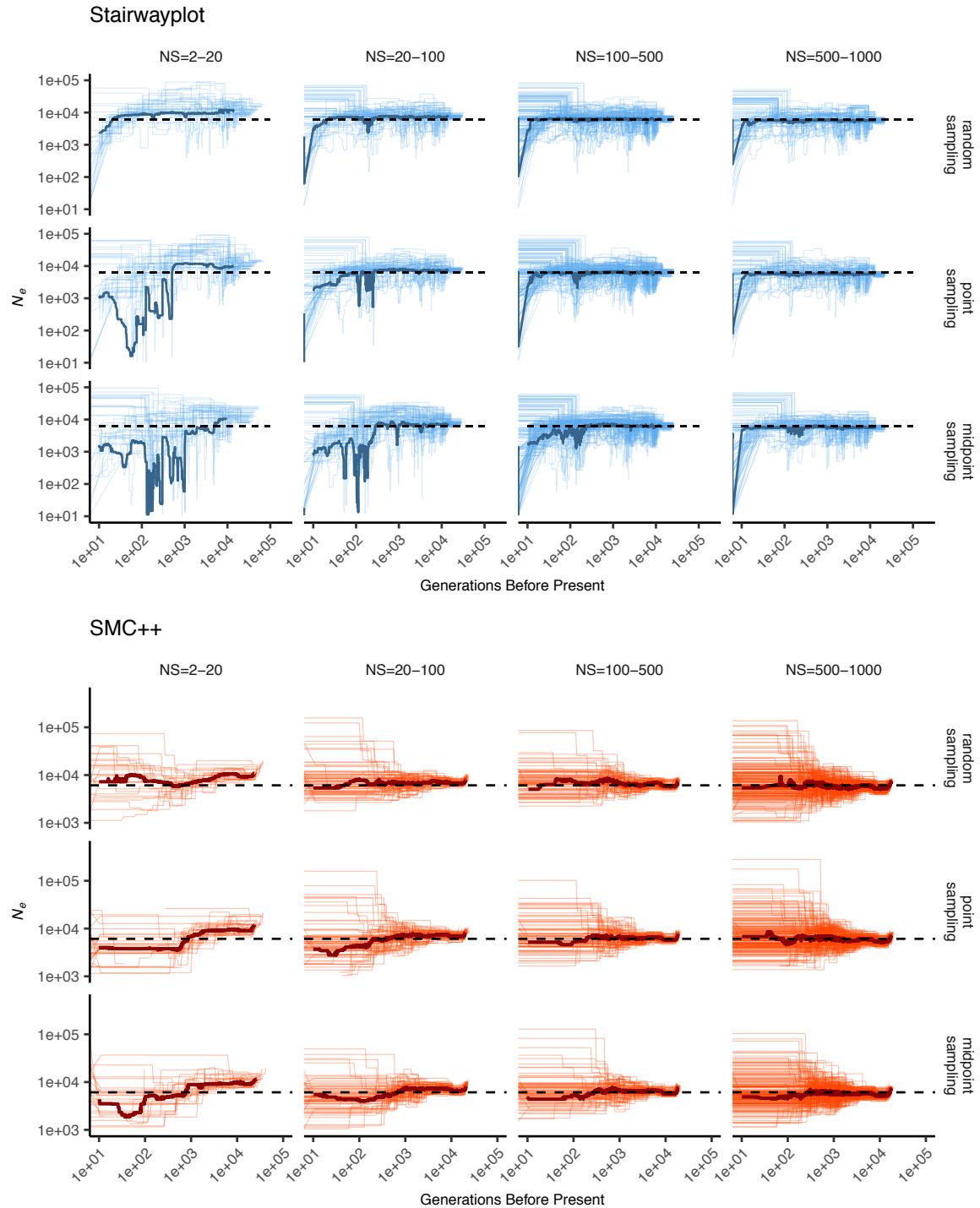


Figure S7 Inferred demographic histories for spatial SLiM simulations, by sampling scheme and neighborhood size (NS) range. Thick lines are rolling medians across all simulations in a bin and thin lines are best fit models for each simulation. Dashed horizontal lines are the average N_e across random-mating SLiM models estimated from θ_π .

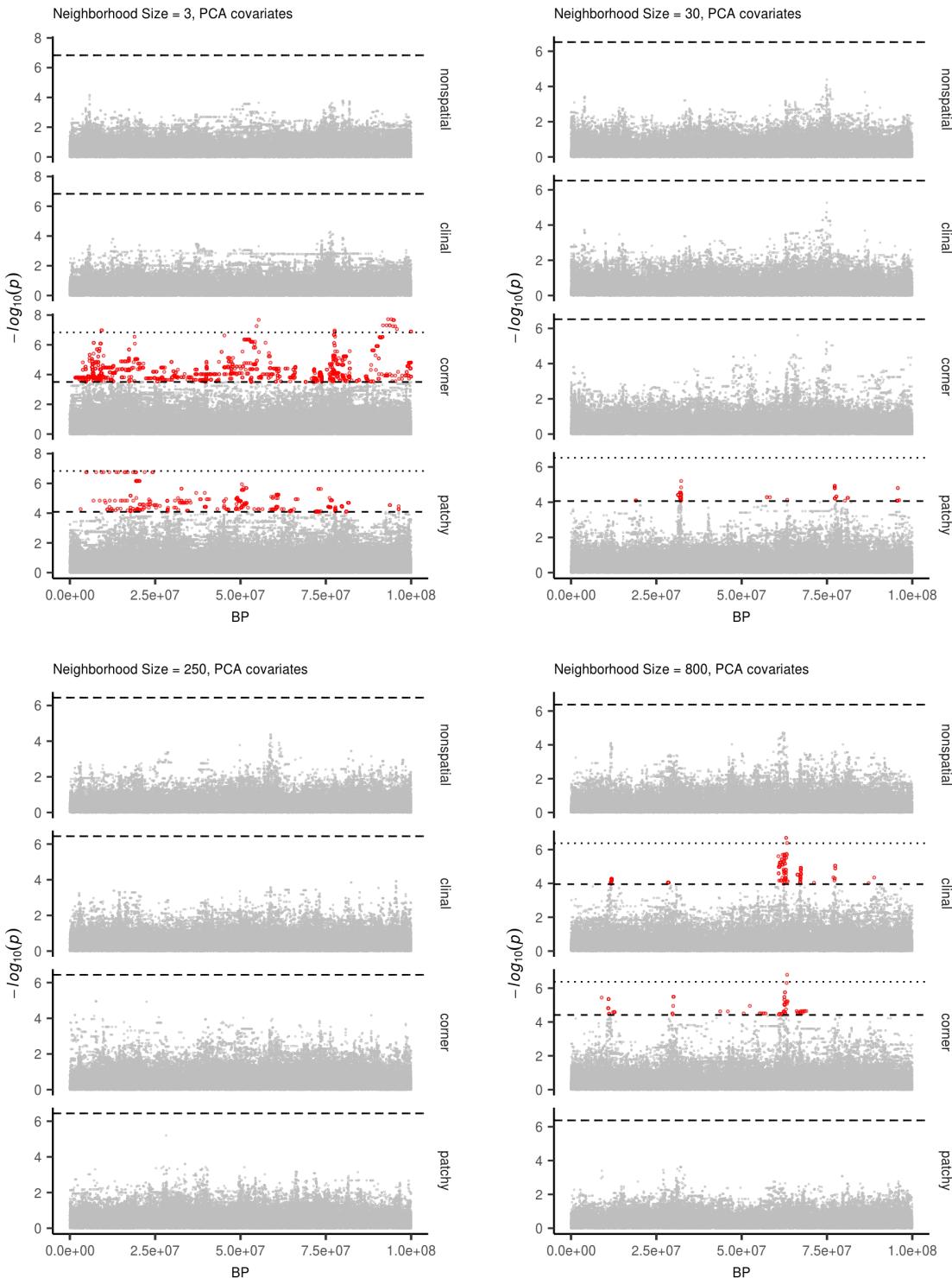


Figure S8 Manhattan plots for a sample of simulations at varying neighborhood sizes. Labels on the right of each plot describes the spatial distribution of environmental factors (described in the methods section of the main text). Points in red are significantly associated with a nongenetic phenotype using a 5% FDR threshold (dashed line). For runs with significant associations the dotted line is a Bonferroni-adjusted cutoff for $p = 0.05$.

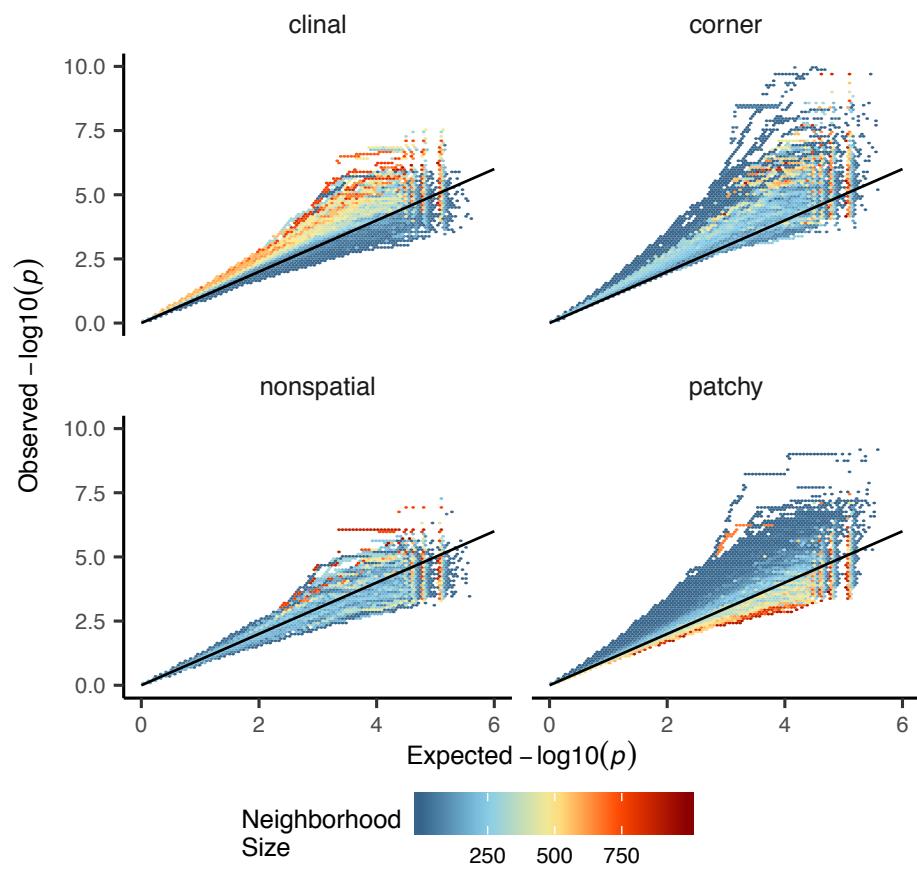


Figure S9 Quantile-quantile plots showing observed $-\log_{10}(p)$ for PC-corrected GWAS run on simulations with varying neighborhood sizes and environmental distributions. Hexagonal bins are colored by the average neighborhood size of simulations with points falling in a given region of quantile-quantile space. Qqplots for a subset of these simulations are shown as lines in Figure 8D.

Table S1 Summary statistics calculated on simulated genotypes.

Statistic	Description
Θ_{pi}	Mean of the distribution of pairwise genetic differences
Θ_W	Effective population size based on segregating sites
Segregating Sites	Total number of segregating sites in the sample
Tajima's D	Difference in Θ_{pi} and Θ_W over its standard deviation
Observed Heterozygosity	Proportion of heterozygous individuals in the sample
F_{IS}	Wright's inbreeding coefficient $1 - H_e / H_o$
$var(D_{xy})$	Variance in the distribution of pairwise genetic distances
$skew(D_{xy})$	Skew of the distribution of pairwise genetic distances
$mean(IVS)$	Mean of the distribution of pairwise identical-by-state (IBS) tract lengths taken over all pairs.
$var(IVS)$	Variance of the distribution of pairwise identical-by-state (IBS) tract lengths taken over all pairs.
$skew(IVS)$	Skew of the distribution of pairwise identical-by-state (IBS) tract lengths taken over all pairs.
$nIBS$	Mean number of IBS tracts with length > 2bp across all pairs in the sample.
$nIBS > 1e6$	Mean number of IBS tracts over 1×10^6 bp per pair across all pairs in the sample.
$corr(D_{xy}, dist)$	Pearson correlation between genetic distance and $\log_{10}(spatial\ distance)$
$corr(mean(IVS), dist)$	Pearson correlation between the mean of the IBS tract distribution for each pair of samples and $\log_{10}(spatial\ distance)$
$corr(nIBS, dist)$	Pearson correlation between the number of IBS tracts for each pair of samples and $\log_{10}(spatial\ distance)$
$corr(IVS > 1e6, dist)$	Pearson correlation between the number of IBS tracts > 1×10^6 bp for each pair of samples and $\log_{10}(spatial\ distance)$
$corr(skew(IVS), dist)$	Pearson correlation between the skew of the distribution of pairwise haplotype block lengths for each pair of samples and $\log_{10}(spatial\ distance)$

Table S2 Anova and Levene's test p values for differences by sampling strategy. Bolded values are rejected at $\alpha = 0.05$

variable	model	p(equal means)	p(equal variance)
segsites	random mating	0.998190	0.980730
$\Theta\pi$	random mating	0.997750	0.996450
Θ_W	random mating	0.998190	0.980730
Tajima's D	random mating	0.879690	0.188770
observed heterozygosity	random mating	0.531540	0.433230
F_{IS}	random mating	0.474790	0.785730
$mean(D_{xy})$	random mating	0.997770	0.996510
$var(D_{xy})$	random mating	0.283630	0.647240
$skew(D_{xy})$	random mating	0.958320	0.260750
$corr(D_{xy}, dist)$	random mating	0.601980	0.000000
$mean(IBS)$	random mating	0.997960	0.997730
$var(IBS)$	random mating	0.486450	0.399490
$skew(IBS)$	random mating	0.117980	0.069770
$nIBS$	random mating	0.997680	0.996570
$nIBS > 1e6$	random mating	0.834870	0.888730
$corr(mean(IBS), dist)$	random mating	0.073270	0.308420
$corr(IBS > 1e6, dist)$	random mating	0.268440	0.002100
$corr(skew(IBS), dist)$	random mating	0.396920	0.000620
$corr(nIBS, dist)$	random mating	0.581090	0.000000
segsites	spatial	0.000000	0.000000
$\Theta\pi$	spatial	0.026510	0.013440
Θ_W	spatial	0.000000	0.000000
Tajima's D	spatial	0.000000	0.000000
observed heterozygosity	spatial	0.000000	0.000000
F_{IS}	spatial	0.000000	0.000120
$mean(D_{xy})$	spatial	0.025390	0.012910
$var(D_{xy})$	spatial	0.004970	0.006230
$skew(D_{xy})$	spatial	0.000000	0.000000
$corr(D_{xy}, dist)$	spatial	0.000000	0.000000
$mean(IBS)$	spatial	0.272400	0.114250
$var(IBS)$	spatial	0.000000	0.000000
$skew(IBS)$	spatial	0.000000	0.000000
$nIBS$	spatial	0.033920	0.016640
$nIBS > 1e6$	spatial	0.000000	0.000000
$corr(mean(IBS), dist)$	spatial	0.000000	0.590540
$corr(IBS > 1e6, dist)$	spatial	0.000000	0.000000
$corr(skew(IBS), dist)$	spatial	0.000000	0.000000
$corr(nIBS, dist)$	spatial	0.000000	0.000000