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Kompleksowa analiza wylesiania w krajach tropikalnych - bezpośrednie czynniki wylesiania, emisje dwutlenku węgla i równowaga wartości usług ekosystemów

A comprehensive study on deforestation in the tropics - direct deforestation drivers, carbon emissions and ecosystem service value balance

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Contents

1. Data and methods	11
1.1. Data	11
1.1.1. Global Forest Change	12
1.1.2. GlobeLand30	13
1.1.3. Intact Forest Landscapes	15
1.1.4. Aboveground Woody Biomass	16
1.1.5. Global Soil Organic Carbon	17
1.1.6. Soil Organic Carbon	18
1.1.7. Ecosystem Service Values	18
1.1.8. Auxiliary	19
1.2. Methods	19
1.2.1. Preprocessing	20
1.2.2. Deforestation	22
1.2.2.1. Forest definition	22
1.2.2.2. Proximate deforestation driver	24
1.2.2.3. Accuracy assessment	26
1.2.3. Emissions	28
1.2.4. Ecosystem service values	30
1.2.5. Binning analysis and visualization	32
2. Results	36
2.1. Deforestation	36
2.1.1. Forest definition	36
2.1.2. Patterns of tree cover and deforestation	42
2.1.3. Proximate deforestation drivers	49
2.1.4. Accuracy assessment	53
2.2. Emissions	54
2.3. Ecosystem service values	55
3. Discussion	59
3.1. Deforestation	59
3.1.1. Forest definition	59
3.1.2. Tree cover and deforestation	60
3.1.3. Proximate deforestation driver	60
3.1.4. Accuracy assessment	60
3.2. Emissions	61
3.3. Ecosystem service values	61
3.4. Binning analysis and visualization	61

4. Conclusion	62
Bibliography	63
List of Figures	I
List of Tables	II
List of Abbreviations	III
A. Hexagonal country boundaries	V
B. Forest definition	VI
C. Proximate deforestation drivers	X

1. Data and methods

In this chapter we describe our approach to answer the scientific questions stated during the introduction. The first section of this chapter introduces the datasets used during this study and the underlying causes for using them. For each dataset, we shortly describe by which approach it is derived and what are the fundamental meta-data properties. Additionally, if possible we try to give for each dataset an accuracy assessment, ideally provided by other research groups if available. Finally, we describe our idea behind using the data and how we acquired and filtered it. The second and last section of this chapter is focused on the applied methodology to prepare our analysis and results. For each processing step we give a short description of the methodical background and describe the core functionality of our processing algorithms. For implementing our processing algorithms and visualizing our results we selected individually the programming language or software which fulfills best the requirements. These approaches are encapsulated in a reusable software design to easily reproduce, alter or reuse our algorithms and findings.

1.1. Data

Table 1.1. Datasets used in this study

Data	Type	Source
Global Forest Change	spatial	Hansen et al. [2013]
GlobeLand30	spatial	Chen et al. [2015]
Intact Forest Landscape	spatial	Potapov et al. [2017]
Aboveground Woody Biomass	spatial	Baccini et al. [2015]
Global Soil Organic Carbon Content	spatial	FAO and ITPS [2018]
Global Administrative Areas	spatial	Hijmans et al. [2018]
Soil Organic Carbon Change	empirical	Don et al. [2010] de Groot et al. [2012]
Ecosystem Service Values	empirical	Costanza et al. [2014] Siikamaki et al. [2015]

The table 1.1 shows a comprehensive overview of the applied datasets for this study. Spatial datasets comprises vector as well raster data, while empirical data is extracted from the cited publications. The subsequent sections describe each dataset in further detail.

1.1.1. Global Forest Change

Global Forest Change (GFC) 2000-2012 Version 1.0 is the first high-resolution dataset that provides a comprehensive view of the annual global forest cover change between 2000 and 2012 [Hansen et al. 2013; Li et al. 2017]. The initial GFC dataset released by Hansen et al. has been extended by recent releases, which encompass the annual forest cover changes between 2000-2013, 2000-2014, 2000-2015 and 2000-2016, respectively. All versions of this dataset have in common, that they are derived from growing season imagery captured by the remote sensing satellite Landsat 7 Enhanced Thematic Mapper Plus (ETM+) enhanced by band metrics of other sensors like Quickbird imagery, existing percent tree cover layers from Landsat data, and global Moderate Resolution Imaging Spectroradiometer (MODIS) percent tree cover [Hansen et al. 2013]. On the satellite imagery, a time-series spectral metrics analysis is applied to gather the global forest extent at 2000 as well as the annual forest loss and the accumulated gain for the period 2001 till 2012. Hence, GFC comprises three independent data layers tree cover, annually forest loss and forest gain divided into 10x10 degree tiles by the geodetic coordinate system World Geodetic System 1984 (WGS84) (EPSG:4326) with a spatial resolution of 1 arc-second per pixel (approximately 900 Km² or 30x30 m). Furthermore, across the provided Geo-Tiff (GTiff) layers the pixel data is coded in unsigned 8-bit integers. Hansen et al. defined trees as all vegetation taller than 5 meters for their study. For each pixel covered by trees, a canopy density ranging from 0 to 100 % is computed. Forest loss is defined as a stand displacement disturbance leading from a forest state to a non-forest state (e.g. canopy density >50 % to 0). Therefore the underlying cause of forest loss ranges from anthropogenic impacts to natural causes. Tree cover gain is defined as the inverse of loss where the canopy density must exceed 50 % to get recognized.

Hansen et al. reports as an accuracy assessment of tree cover loss a producers accuracy of approximately 83 % for the tropical region. The mapped tree cover gain is probably an underestimation of the true gain with a producers accuracy of 48 % and a user's accuracy of 81 %.

This dataset is publicly available for download without any constraint. For a convenient bulk download, the dataset homepage provides a "* .txt" files comprising the Uniform Resource Locator (URL) of the tiles for each sub-dataset. The spatial location of an image can be directly determined from the file name within the URL. Each file name has a common pattern shown by the following expression: "Hansen_VERSION_LAT [NS] _LNG [WE] ". LAT (latitude) and LNG (longitude) refer to the top left corner coordinates of a raster image, whereas these coordinates are only given in natural numbers. The orientation of the image on the hemisphere is determined by the four cardinal directions N (north), S (south), W (west) and E (east). For this project, we require all three sub-datasets, namely: Treecover2000, loss-year, and gain. The data acquisition is automatized with a Python script by using the

Standard Library (stdlib) modules `urllib` and `re`. At first, the Python script downloads the provided `"*.txt"` files and creates a list data structure, where each URL is an element of this list. After, it cycles through the list and extracts the corner coordinates from the file name by means of a Regular Expression (REGEX). These corner coordinates and cardinal directions are converted to valid latitude and longitude coordinates between $[-90, 90]$ and $[-180, 180]$, respectively. Now, an image is only downloaded if it is within the study extent between $[-20, 30]$ latitude. The acquired image tiles in total 678 are shown in the top panel (green squares) in figure 1.1.

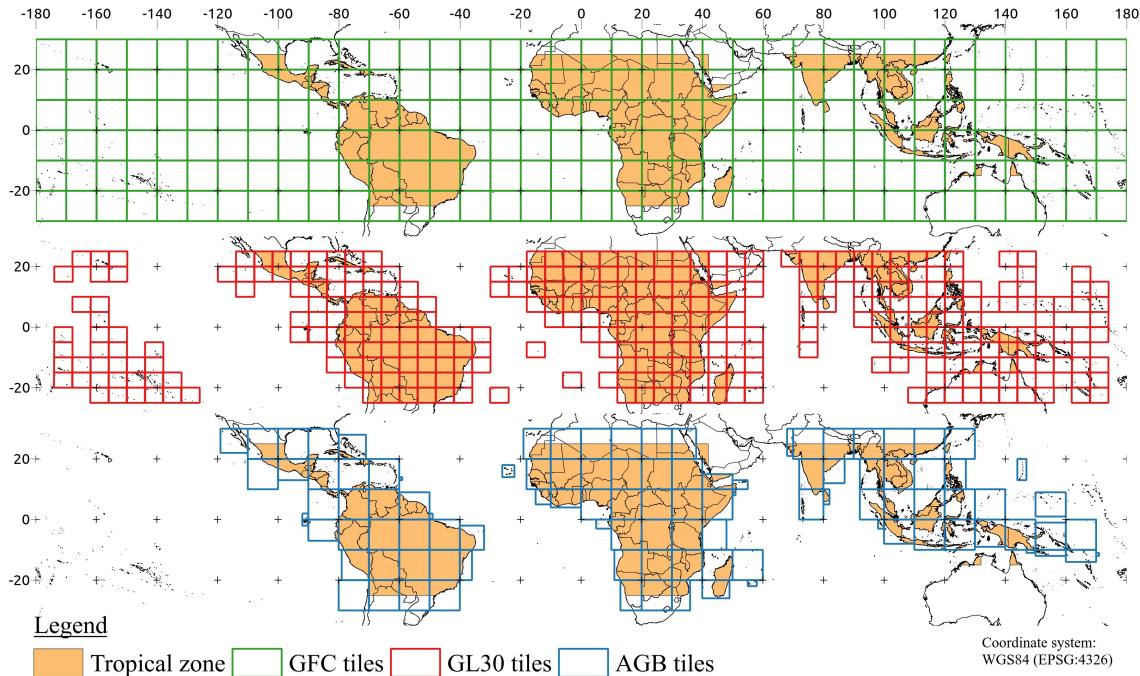


Figure 1.1. Map of downloaded dataset tiles: This map shows the acquired image tiles for this study. From top to bottom in green Global Forest Change (GFC) dataset tiles (Treecover2000, loss-year and gain), the land cover dataset GlobeLand30 (GL30) image tiles in red, and in blue the Aboveground Biomass (AGB) dataset tiles. The orange filled shapes highlight countries within the tropical zone.

1.1.2. GlobeLand30

GlobeLand30 (GL30) is the first global land cover dataset with a 30 meter per pixel spatial resolution that provides a comprehensive view on the distribution of 10 different land cover classes (table 1.2) over the entire globe [Chen et al. 2017]. Currently, this dataset is available for two different time periods 2000 and 2010 [Chen et al. 2015]. The pixel values of this dataset are coded in unsigned 8-bit integers and as coordinates system, it uses WGS84 in Universal Transverse Mercator (UTM) projection. GL30 can be downloaded as a GTiff raster mosaic where each image covers 6x5 degrees [Chen et al. 2014]. For detecting the land cover classes Chen et al. used a so-called Pixel-Object-Knowledge (POK) oriented approach and satellite imagery from Landsat ETM+ [Chen et al. 2015]. Chen et al. divided

the mapping process into different stages where each land cover type is detected separately and deleted subsequently from the source satellite image. The applied mapping order is water bodies, wetland, snow and ice, cultivated land and forest, shrubland, grassland and bare land synchronous. To detect the pixels of a selected land cover type the following pixel level classifiers are used: Decision Tree (DT), Support Vector Machine (SVM) or Maximum Likelihood Classifier (MLC). After pixel detection, the adjacent pixels are grouped as an aggregated land use object. These objects are subsequently validated by expert knowledge and the gained knowledge is used as a feedback loop to improve the automatized classification.

Chen et al. estimates an overall mapping accuracy of 80.33 % and 78.6 % for 2000 (only validated in Shaanxi, China) and 2010 (global), respectively [Chen et al. 2015]. Several research groups besides Chen et al. validated the mapping accuracy of GL30 at different regions and scales. Arsanjani et al. estimates an overall accuracy of 77.9 % for Iran and an accuracy >80 % for Germany [Arsanjani et al. 2016a,b]. Yang et al., Cao et al. and Jacobson et al. estimate the accuracy of 82.4 %, 80.1 % and 83.1 % for China, Nepal, and East Africa, respectively [Yang et al. 2017; Cao et al. 2016; Jacobson et al. 2015]. Unfortunately, no study focused on validating the mapping accuracy for regions exclusively within the tropical zone.

Chen et al. donated the GL30 land cover mapping to the United Nations (UN) but it is not accessible for public download. The download is restricted to users who register on the dataset homepage but the registration process is not working properly. Fortunately, the supervisor of this work had already an account otherwise it would be impossible to receive a copy of the dataset. A registered user must fill an order application to get access to the image tiles. The application form must contain the tile identifiers and the selected time period. Tile identifiers have the following common pattern: "[NS] ZONE_LAT_NAME" where zone refers to the UTM zone between [1,60], N (north) or S (south) to the cardinal direction, and LAT (latitude) to the latitude coordinate of the top left corner. For better usability the homepage provides an interface for selecting the required image tiles but the selection of multiple tiles did not work. As well a vector file is provided which contains the dataset tile polygons with assigned identifiers. This file was used to select all required tiles within the tropical zone between approximately [-23,23] degrees (WGS84). Figure 1.1 presents the selected images in red at middle panel. The corresponding image identifiers are converted to a single line string and copied to the application form. After submitting the form the order will be checked and approved within two weeks. After one week we received a two weeks limited access to a password protected File Transfer Protocol (FTP) server where we downloaded 716 raster images. Due to the several restrictions, this process of selecting and downloading could not be automatized with one pipeline. Only the selection and string conversion were automatized with a throwaway script.

Table 1.2. Classification schema of the GlobeLand30 product: The code column is the assigned pixel value, type the corresponding land cover type and definition explains in broad terms which types of surfaces fall into the land cover type [Chen et al. 2017].

Code	Type	Definition
10	Cultivated land	Used for agriculture, horticulture and gardens, including paddy fields, irrigated and dry farmland, vegetable and fruit gardens, etc.
20	Forest	Covered by trees, vegetation covers over 30%, including deciduous and coniferous forest, and sparse woodland with cover 10-30%, etc.
30	Grassland	Covered by natural grass with cover over 10%, etc.
40	Shrubland	Covered by shrubs with cover over 30%, including deciduous and evergreen shrubs, and desert steppe with cover over 10%, etc.
50	Wetland	Covered by wetland plants and water bodies, including inland marsh, lake marsh, river floodplain wetland, forest/shrub wetland, peat bogs, mangrove and salt marsh, etc.
60	Water bodies	In land area, including river, lake, reservoir, fish pond, etc.
70	Tundra	Covered by lichen, moss, hardy perennial herb and shrubs in the polar regions, including shrub-, herbaceous-, wet- and barren-tundra, etc.
80	Artificial surfaces	Modified by anthropogenic influence, including all kinds of habitation, industrial and mining area, transportation facilities, and interior urban green zones and water bodies, etc.
90	Bareland	With vegetation cover lower 10%, including desert, sandy fields, Gobi, bare rocks, saline and alkaline land, etc.
100	Snow and ice	Covered by permanent snow, glacier and icecap

1.1.3. Intact Forest Landscapes

An Intact Forest Landscapes (IFL) is a mosaic of undisturbed forest patches and naturally treeless ecosystems without signs of human activity and large enough to maintain all native biological diversity [Potapov et al. 2017]. Due to the fact that IFL comprises different intact natural landscape patterns like primary forests, non-forest ecosystems, temporary treeless areas after a natural disturbance, and water bodies the term is not congruent to the term primary forest defined by the Food and Agriculture Organization of the United Nations (FAO) [FAO 2012]. But as mentioned IFLs includes large patches of primary forests with a minimum extent of 500 Km², therefore, primary forests can be extracted from the layer. Still, there are smaller fragments of primary forest outside of the IFLs. In regards to the extent an IFL has a minimum size of 500 Km², a minimum width of 10 Km, and a minimum corridor/appendage width of 2 Km. Further an IFL should not contain any of the following: ecosystem alteration, fragmentation by infrastructure and disturbance, and areas altered or managed through agriculture, logging, and mining. For mapping and detecting IFLs Potapov et al. used Landsat imagery and several auxiliary data sources like GFC, and national transportation maps.

The dataset can be downloaded as a Shapefile (SHP) file with the coordinate reference system WGS84. Each polygon in the SHP represents an IFL patch at a certain location on our planet at the time period 2000.

Data acquisition is pretty straight forward the IFL dataset public accessible for download. As mentioned it is a SHP so you must only download a single compressed archive. The download is automatized with a Python script by using the stdlib modules `urllib` and `threading` [van Rossum and Development 2018].

1.1.4. Aboveground Woody Biomass

The Aboveground live woddy Biomass density (AGB) raster dataset is prepared by Global Forest Watch (GFW) by an adapted approach of Baccini et al. [Baccini et al. 2012, 2015, 2017]. For the year 2000, this dataset estimates the aboveground biomass density per pixel in Mg C ha^{-1} (megagram carbon per hectare), and the confidence per pixel at a spatial resolution of approximately 1 arc-second (approximately 900 Km^2 or 30x30 meter). The dataset covering the global tropical zone as a mosaic of GTiff raster images where each tile of the mosaic has the Coordinate Reference System (CRS) WGS84 and is coded in a float. For deriving biomass density GFW used canopy metrics from Geoscience Laser Altimeter System (GLAS) Light Detection and Ranging (LIDAR) footprints and several regional and forest-specific allometric equations. The resulting GLAS AGB estimates are used as labels to train regional specific Random Forest (RF) models based on Landsat 7 ETM+ top-of-atmosphere reflectance, tree canopy density of GFC, elevation data, and climate data as predictor variables. After these models are subsequently applied to the entire study extent to predict the biomass content for each pixel. Additional an uncertainty layer is prepared accounting for the errors from allometric equations, the LIDAR based model, and the random forest model.

The AGB raster mosaic is publicly available on the homepage of GFW. As mentioned, the dataset covers only the tropical zone, therefore we acquire the entire mosaic. The GFW homepage provides an Geographic JavaScript Object Notation (GeoJSON) Application Programming Interface (API) to receive the actual URL of each raster image. If a request is sent to this API the server responded with a GeoJSON feature collection. The collection contains as attributes the URLs of the biomass images, the URL of the uncertainty layers, and the rectangular bounds of each image. The data acquisition is automatized by means of Python and the stdlib modules `urllib`, `threading`, and the open source library `geopandas` [van Rossum and Development 2018; McKinney 2010]. At first, the GeoJSON is downloaded via an API call and eventually stored on disk. Next, we iterate the features of the GeoJSON collection and extract the URLs (biomass and uncertainty) of each tile. These URLs are downloaded

and subsequently stored on disk. During the downloads of the uncertainty layers, the GFW server answered repeatedly with a 404 (Not found). Therefore the uncertainty layers are not available. In total we downloaded 105 different image tiles, their extent and spatial location are shown in blue at the bottom panel of figure 1.1.

1.1.5. Global Soil Organic Carbon

The Global Soil Organic Carbon map (GSOCmap) is a joint project between Global Soil Partnership (GSP) and Intergovernmental Technical Panel on Soils (ITPS) to produce a global Soil Organic Carbon (SOC) content map by a country-driven approach. In the year 2018, the first iteration of this map in version 1.0 was released and shortly followed by 1.1 (new country submission by Rwanda) and 1.2 (new country submissions by Chile and Colombia). As the short release cycle suggests the mapping project is intended as a long-lived dataset which will improve over time and by new country submissions. Till now 67 (approximately 63 % of the global land mass) different countries submitted their country based SOC estimates. To foster the national SOC mappings the International Soil Reference and Information Center (ISRIC) provides several covariate datasets like national Digital Elevation Map (DEM) maps, annual spectral remote sensing data or national soil type grids. Additionally, the contributors can join a mapping training and use the GSOCmap cookbook as guidance for their mapping efforts. As an exchange, each country shares its national GSOCmap by compliance of several criteria e.g. reporting of the Meta-data of the SOC sampling (sample timeline, sample depth, bulk density etc.), uncertainty assessment, and the applied methods for estimating and interpolation of the SOC content. For interpolating the guide organizations suggest the following approaches: simple geo-matching, class-matching, Multiple Linear Regression (MLR), RF or SVM. The national maps are aggregated to the final GSOCmap with a target resolution of 30 arc-seconds (approximately 1 Km²) in the CRS WGS84. The dataset is one single raster image as GTiff coded in float covering the entire globe where each pixel value is the SOC content in Mg C ha⁻¹ at a soil depth of 0-30 cm [FAO and ITPS 2018].

The product is validated by comparing the pixel level estimates with soil sampling data from various soil databases (WoSIS, HWSD, etc.). In total 312122 samples were divided into three sub-levels (<150 Mg C ha⁻¹, >150 Mg C ha⁻¹, and all samples) and subsequently computed the Mean Error (ME). The ME of the entire sample space and <150 Mg C ha⁻¹ suggests that the mean Soil Organic Carbon Content (SOCC) estimate is an overestimate of 1.6 and 4.5 Mg C ha⁻¹ respectively. All samples with a SOCC content >150 Mg C ha⁻¹ show an underestimate by approximately 165 Mg C ha⁻¹ in the mean. Additionally, an uncertainty assessment was prepared to estimate a Standard Deviation (SD) between ± 0-16 t ha⁻¹ for the tropical zone. Unfortunately, this assessment is pretty rough and till now not

available as a product. The GSOCmap in comparison with other global SOC products has the lowest Root Mean Square Error (RMSE). In summary, the prepared validations show evidence that the GSOCmap is a conservative data product with a tendency to underestimate the SOCC.

The dataset is publicly available at the homepage of the FAO. As mentioned it consists of one raster image, therefore we download it by means of a Python script without any additional steps.

1.1.6. Soil Organic Carbon

Don et al. performed the first study of tropical SOC change for soil depth between 0 and 30 cm. For the study, a global meta-analysis is applied by using 358 (153 published an peer-reviewed) different studies to estimate SOC change for 12 major Land-use Change (LUC) types. The base date is derived from 39 different tropical countries covering all continents. Unfortunately, Africa and East-Asia are under-sampled whereas South-America has the best data coverage. The meta-analysis is restricted to mineral soils therefore all wet soil types are excluded from the analysis Don et al.. The 12 Land Cover (LC) transitions encompass the following LC types: primary forest, secondary forest, grassland, cropland, and perennial crops. Primary forest is defined as natural vegetation without human impacts which includes natural grassland and shrubland. Secondary forest is managed forests and regrown forests after partial destruction of the old stand. Grassland comprises pastures for livestock but excludes natural grasslands. Cropland comprises annual crops like maize or beans and perennial crops could be coffee or sugar cane. For our study we used only the SOC change estimates for these LUC types which correspond to the GL30 and IFL classification schema. The actual values are shown in table 1.3.

Table 1.3. Relative soil organic carbon change for certain land-use change types: The Land-use change columns from and to define the LUC type with the corresponding relative Soil Organic Carbon (SOC) change and the Standard Error of the Mean (SEM) [Don et al. 2010].

LUC type From→To	Relative SOC change [%]	SEM
Primary forest→Grassland	-12.1	±2.3
Primary forest→Cropland	-25.2	±3.3
Primary forest→Secondary forest	-8.6	±2.0
Secondary forest→Grassland	-6.4	±2.5
Secondary forest→Cropland	-21.3	±4.1

1.1.7. Ecosystem Service Values

Pending, which type of ecosystem services are included in the biome types

Table 1.4. Selection of Ecosystem Service Values (ESV) per biome used in this study: ESV per biome and its monetary value in Int.\\$ ha⁻¹. Dg refers to data from de Groot et al., Co from Costanza et al., and Wb from Siikamaki et al..[de Groot et al. 2012; Costanza et al. 2014; Siikamaki et al. 2015]

Biome	Dg	Co	Wb
Cropland	-	5,567	-
Forest tropical	5,264	5,382	1,312
Grass/Rangelands	2,871	4166	-
Wetlands	25,682	140,174	-
Lakes/Rivers	4,267	12,512	-
Urban	-	6,661	-

1.1.8. Auxiliary

As auxiliary data for country boundaries and the selection of continental regions we downloaded with Python the Global Administrative Areas Map (GADM) and Natural Earth layers as SHP files [Hijmans et al. 2018; van Rossum and Development 2018].

1.2. Methods

Figure [flowchart and reference](#) shows an overview of the entire processing pipeline. The following sections describe detailed the applied approach for each step in figure. The order of appearance is from left to right.

For implementing our processing algorithms we selected for each task the technology which fulfills best the requirements. Python is our core language for implementing our processing algorithms because it supports a easy implementation of multiprocessing which is heavily used in this project. From the Python stdlib we used the following libraries: urllib, re, unittest, time, math, logging, collections, bisect, and enum. Additionally for geo-processing we used the following open source libraries: numpy, pandas, fiona, geopandas, shapely, matplotlib. The entire frontend of our Python source code is aggregated in a Jupyter Notebook and available on [GITHUB](#). This ensures that everyone interested can easily reproduce the findings of our project. JavaScript and the additional modules papaparse and the google maps api are used for programming a small web app for cross validation of land use predictions. R is used for hypothesis testing. Bash is used to aggregate large raster datasets as vrt files. To prepare map visualizations we used QGIS. Dia is used for preparing flowcharts and GIMP is used for image post-processing.

1.2.1. Preprocessing

Before we apply further analysis, we have to generalize the used datasets. As introduced in the data section do we use datasets which differ largely in their metadata properties, for example, single-tiled or multi-tiled images, used CRS, spatial resolution, and file type. Therefore, our goal should be to develop a process which creates an image stack of equal meta-data for each location in our study extent. In further descriptions, we will refer to this stack as Aligned Image Stack Mosaic (AISM). As target CRS for our AISM we chose WGS84 and as target extent for the mosaic, we use the bounding box of the GL30-2010 tiles. The following paragraph explains how we developed the alignment algorithm by means of Python and the additional open source libraries rasterio, geopandas, and shapely [van Rossum and Development 2018; McKinney 2010].

The first exercise of the preprocessing algorithm is to detect all tiles covering the extent of our template tiles. At first, we create for each multi-tiled dataset a polygon mask as SHP. This mask contains the spatial extent of each tile within a dataset and as attribute the corresponding file identifier. If the dataset tiles are not in WGS84 the extracted bounds are subsequently reprojected to this CRS. During the masking process, we recognized that the raster mosaic bounds of both GL30 datasets (2000 and 2010) generate re-projection errors. These errors showed up as polygons spanning the entire globe but one tile can only fill its UTM zone extent. Further analysis revealed that all tiles located in UTM zone 1 and 60 overflowing the maximum and minimum longitude coordinates of these zones. As solution we excluded all tiles within UTM zone 1 and 60 from further processing, namely: n01_00, s01_00, s01_10, s01_15, s01_20, s60_00, s60_05, s60_10, s60_15, and n53_00. The described steps can be found as well in figure 1.2. Now, as the figure suggests we determine the intersection between these mask layers and group the intersecting tiles by our template tile. Next, we create for the template tile a re-projection profile (warp profile) and apply it subsequently to all intersecting tiles based on the following rules: if from one dataset more than one tile intersects merge them followed by re-projection or if only one tile intersects just re-project it. As introduced the GSOCmap consist only of one single tile with a spatial resolution of approximately 1 Km², so it must only re-project and re-sampled by the nearest-neighbor approach. We select from the IFL layer all polygons within our template warp profile and convert them to a raster layer where intact forest patches are coded by a one in an 8-bit unsigned integer. The last step of the alignment process is the rounding of the AISM bounds to full integer degrees and a subsequent clipping of each tile to this rounded bounds. The entire work-flow is presented in figure 1.2 and results in a AISM shown in figure 1.3. Finally, we create a polygon mask of our AISM and store for each polygon the corresponding dataset tiles. This mask is used as a file index for the next algorithms.

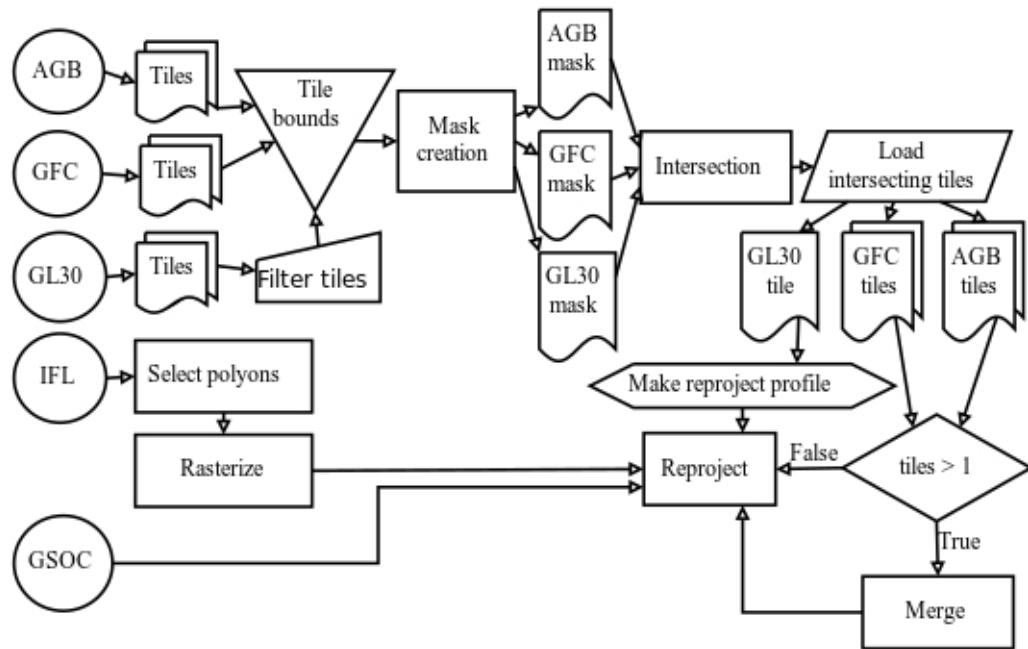


Figure 1.2. Tile alignment algorithm: For the multi-tiled datasets (multi-document symbols) a mask is created by extracting the tile bounds. Next, the intersection between these masks is determined to identify superimposing data and the corresponding tiles are loaded from disk. GL30 tiles are used as a template by creating the re-project profile and subsequently applying it to intersecting tiles. From the IFL layer only polygons within the re-project area are selected and subsequently converted to a raster layer.

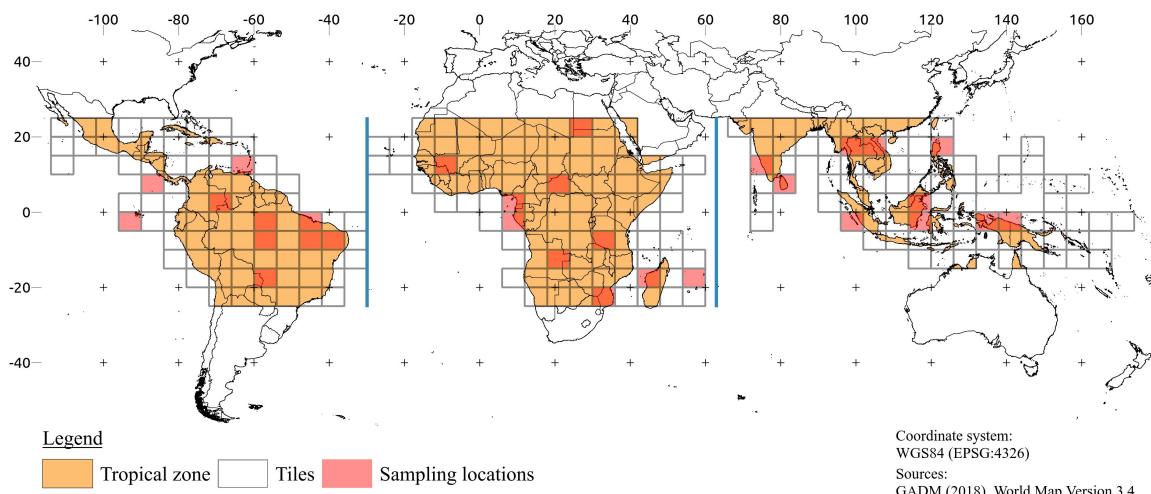


Figure 1.3. Aligned raster images and sampling locations: The map shows the location of the aligned multi-image stack tiles as black-framed, square-sized polygons. The sampling locations for accuracy assessment are represented in red. Countries within tropical range appear in orange. Vertical blue lines separate the tiles into the three continental regions Latin America, Africa, and Asia/Australia.

1.2.2. Deforestation

1.2.2.1. Forest definition

To determine the proximate drivers of deforestation we combined the information of the two datasets GFC and GL30. However, both differ in their definition of tree cover by canopy cover threshold as introduced in section 1.1. GFC detects tree cover over the entire canopy density interval of $(0, 100]$, while the GL30 threshold is set to $> 10\%$. To successfully extract stable land cover transformation by superimposing both layers we must first harmonize the tree cover definition of both strata. We hypothesize that if both layers agree on tree cover they should also agree if a transition to a non-forest state occurs. To harmonize both definitions we have the opportunity to vary the canopy density of GFC to determine at which density class the similarity between them is at its maximum. Then, we use the examined maximum similarity canopy density to filter the tree cover loss and gain layer.

To determine the similarity between GL30 2000 and GFC reference tree cover we used the Jaccard Index (JI). The JI or coefficient of community is a simple measure of similarity between two pairs of a binary population or a measure of the degree of spatial overlap between two images [Sampat et al. 2009]. This index was first applied by Jaccard to compare distributions of rare alpine flora in 1912 [Jaccard 1912], and since it is a widely used metric across multiple fields. If we compare two binary images, let a be the magnitude where both images ($\text{Img}_1, \text{Img}_2$) have an agreement represented as a pixel value of one. Let b the magnitude where Img_1 is zero and Img_2 is one, while c represents the inverse expression. Finally, assume that d is the magnitude of elements where both images are zero. The matrix in table 1.5 shows that the computation of this coefficients a, b, c , and d can be expressed as a set of boolean operations. Equation 1.1 shows how the JI is computed by substitute integer values for the variables. This computation can be reduced to two boolean operations for a major performance increase. The JI is always within the closed interval $[0, 1]$, where an index of one or zero means a complete similarity between both populations or a complete disagreement, respectively. The relationship between a and JI is near linear [Shi 1993]. The first step to compute the JI for our raster images is to extract the tree cover from the GL30 2000 land cover by setting all pixels with values $\neq 20$ to zero and values = 20 to one. Next, we extract from the GFC reference tree-cover pixel values within the half-opened interval of the following canopy density classes and set them to one: $(0, 100]$, $(10, 100]$, $(20, 100]$, and $(30, 100]$. Therefore, we test four different tree cover definitions for GFC. The first excludes canopy densities $\leq 0\%$, the second $\leq 10\%$, the third $\leq 20\%$, and the fourth canopy densities $\leq 30\%$. We will refer to this JI of different canopy density classes as $\text{JI}_0, \text{JI}_1, \text{JI}_2$, and JI_3 . For all 269 tiles of our AISM, we calculate the JI existing between GL30 and GFC for the four above-mentioned forest definitions by using equation 1.1. The algorithm is im-

plemented in Python by using numpy's ability to perform boolean operations between large matrices. As parameters, the function expects two matrices with the same dimensionality in R^{n*m} and a boolean indicating if the function should return the coefficient matrix as well. The previously described preprocessing steps are implemented as an extra function. This function requires as parameter two raster layers, a list of integer values to consider as GL30 forest cover (default is 20), and the lower bounds of the canopy density intervals to consider for computation.

Table 1.5. Jaccard Index coefficient matrix: a is the magnitude of agreement, d is the magnitude of disagreement, b and c are the magnitudes of partial disagreements among both images. The computation of these coefficients can be expressed as boolean operations on matrices.

		Img ₁	
		State	1 0
Img ₂	1	$a = \mathbf{X}_1 \wedge \mathbf{X}_2 $	$b = (\mathbf{X}_1 \wedge \mathbf{X}_2) \oplus \mathbf{X}_2 $
	0	$c = (\mathbf{X}_1 \wedge \mathbf{X}_2) \oplus \mathbf{X}_1 $	$d = \neg(\mathbf{X}_1 \vee \mathbf{X}_2) $

$$JI = \frac{a}{a+b+c} = \frac{|\mathbf{X}_1 \wedge \mathbf{X}_2|}{|\mathbf{X}_1 \vee \mathbf{X}_2|} \quad (1.1)$$

To optimize the overall tree cover similarity between both datasets we must test which canopy density class yields the highest agreement over our study extent. To test the significance of the difference between two correlated samples, we decided to apply the non-parametric Wilcoxon signed-rank test [Wilcoxon 1945]. This test requires paired data from the same population, at least an ordinal scale of measurement, each sample pair is independent, and the dependent variable can be expressed as a continuous probability [Lowry 2019]. Further, an advantage of this test is that we don't have to assume a normal distribution for our sample population. Our sample population fulfills these requirements. The test procedure is implemented in R because this language is mainly intended for this kind of statistical analysis. We exported the computed JI from our Python environment and applied a cross-testing in R. In our case, cross testing is defined as the test of all possible JI combinations. Further, we applied a two- and one-sided Wilcoxon test because we want to examine if there is a significant difference and which direction has the similarity distribution. To address the higher probability of family-wise error in multiple comparisons we used a Holm correction. Before we applied the examination of the distribution we separated our population into three independent regions, namely Latin America, Asia/Australia, and Africa highlighted by the vertical blue lines in figure 1.1. Latin America, Asia/Australia, and Africa comprised 82, 86, and 101 image tiles, respectively. Additionally, we excluded from the analysis all samples where JI_0 is zero because this tiles from our AISM did not contain any pixels covered by trees. In Latin America, Asia/Australia, and Africa we excluded 6, 13, and 15 tiles. We tested continental differences in tree cover agreement to inspect regional dependencies and global differences. The results from the global testing are used to determine our definition of

tree cover. Further, we compared differences of tree cover agreement between the continental regions by applying a Wilcoxon rank-sum test also known as Mann-Whitney U test. We applied a Benjamini and Hochberg correction to the test results.

1.2.2.2. Proximate deforestation driver

Based on our forest definition developed in the previous section we want to classify all the tropical deforestation within a canopy density of $(10, 100]$ percent between 2001 till 2010. Additionally, we must consider the mean miss-classification rate of 52 % by previous findings of Seydewitz [Seydewitz 2017]. Therefore we have to develop a feasible method to resolve this issue.

For classifying the proximate drivers of deforestation we selected the following raster images from our AIS: GFC reference tree-cover, GFC annual losses, GFC gain, and the GL30 LC classification of 2010. Now, we apply to each raster image stack the following described operations. From the reference tree-cover images, we select all pixels where the canopy density is within the half-open interval of $(10, 100]$ percent and set them to one (true). The same exercise is applied on the annual losses stratum by setting all forest loss pixels within the time period 2001 till 2010 to one (true). After, both layers are combined with a logical AND operation to select our target deforestation pixels. Finally, we build the Hadamard product (element-wise multiplication) of the target deforestation layer and the GL30 LC stratum to classify the pixels with a deforestation event. For classifying forest regrowth we filtered the GFC gain layer to consider only tree-cover gain within our target temporal resolution and target canopy density. After, the filtered stratum is aggregated with our classified deforestations by using the Hadamard product of both layers. We will refer to this proximate deforestation driver layers as PDD. Figure 1.4 shows an overview of the classification process. The classification algorithm is implemented as a Python function which requires a parameter the previously named raster layers. Additionally the target canopy density and time period is freely selectable for experimental variations. The described filtering and aggregation steps are implemented as binary matrix operations for fast processing of large data sizes by means of numpy.

After classifying the proximate deforestation drivers we developed an approach to smooth the misclassified pixels based on an approximated probability. We define misclassifications as sites where the GFC annual loss data predicts deforestation but the GL30 stratum still classifies them as forest. The first step of our reclassification is to cluster the misclassified pixels with the Hoshen-Kopelman algorithm [Hoshen 1998]. The clustering algorithm is implemented as a part of the Geospatial Data Abstraction Library (GDAL) library and can be called through the rasterio interface. For this project, we used the following parameters:

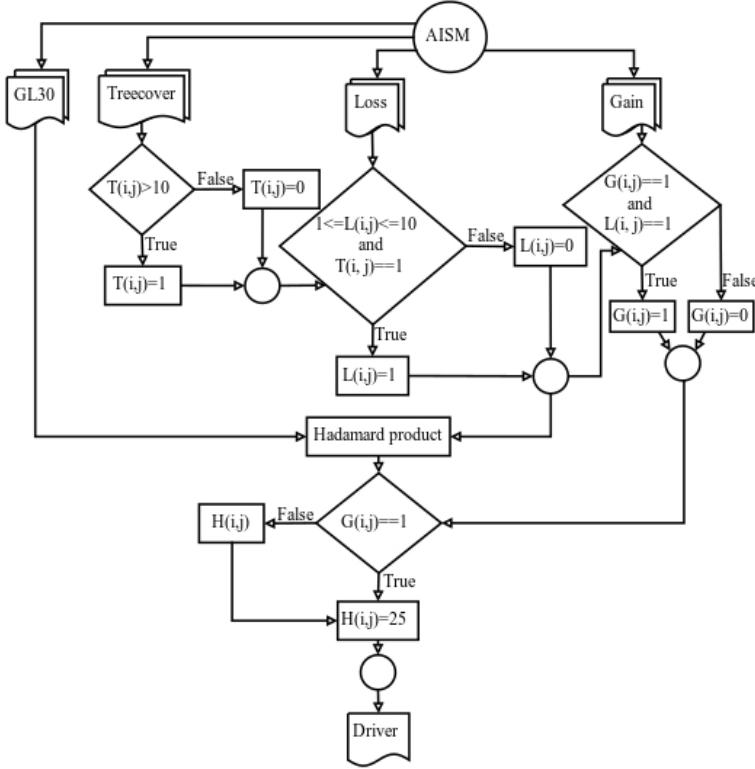


Figure 1.4. Classification of proximate deforestation drivers: For the classification of the proximate deforestation drivers the following layers are required GL30 2010, GFC tree cover, annual losses, and gain. From tree cover we select all pixels within the canopy density interval (10, 100]. The tree cover mask is used to select the appropriate annual losses within the time interval [2001, 2010]. To predict a land cover change after a deforestation event we use the Hadamard product.

connectivity 4 and a boolean mask where only pixel values = 20 are true. Now the algorithm creates for each pixel cluster a polygon. After we created a squared sized buffer with a side length of 500 x 500m around the polygon centroid (the geometric midpoint of the polygon). Because WGS84 is not an equal area CRS we must compute for each tile the buffer size separately. To compute the buffer size in image coordinates we used the Haversine formula in equation 1.2 to determine the on-ground resolution in meter on pixel level. Where d is the great-circle distance between two latitude, longitude pairs φ_n, λ_n and r is the earth radius of approximately 6378137 m. Because this computation is expensive we assumed that the pixel resolution is equal for an entire raster tile. After extracting the buffer we counted the most frequent class under exclusion of pixels with a value of 0, 20 or 255 within the buffer. Finally, if the most frequent class is defined we reassigned this class value to the cluster. The reclassification algorithm is implemented as a Python function which requires as parameters a proximate driver raster image, a list of elements which should be interpreted as occupied cells for the clustering, pixel values which should be excluded from counting, the side length of the buffer, and the on-ground resolution.

$$d = 2r \arcsin \left(\sqrt{\sin^2 \left(\frac{\varphi_2 - \varphi_1}{2} \right) + \cos(\varphi_1) \cos(\varphi_2) \sin^2 \left(\frac{\lambda_2 - \lambda_1}{2} \right)} \right) \quad (1.2)$$

1.2.2.3. Accuracy assessment

For examining the accuracy of our Proximate Deforestation Driver (PDD) predictions we used a confusion matrix (also known as two-way frequency tables, error matrix or contingency tables). These matrices are commonly used for an accuracy assessment of land cover classifications and enable the computation of marginal and conditional distributions [Congalton 1991; Foody 2002]. Table 1.6 shows a general model of a confusion matrix. Foundation for an accuracy assessment by means of a confusion matrix is a collection of ground-truth samples which can be compared with the class predictions for these samples produced by a classification algorithm. For the preparation of our accuracy assessment, we have to extract a collection of pixel samples with a deforestation occurrence from our proximate driver maps (further also called predictions). Next, we compose a set of ground-truth for these predictions (further also called references).

Table 1.6. A general model of a confusion matrix: X_1, \dots, X_n denote classification categories of two independent raters. $x_{n,n}$ are the actual samples sorted into the categories where the values in the diagonal show the agreement between both raters. The remaining cell values account for the disagreement between the two raters. Σ column and row show the marginal distribution and N is the total number of samples.

		Reference				
		Cl _s	X ₁	...	X _n	Σ
Predict	X ₁	$x_{1,1}$...	$x_{1,n}$	$x_{1\cdot} = \sum_{i=1}^n x_{1,i}$	
	:	:	⋮	⋮	⋮	⋮
	X _n	$x_{n,1}$...	$x_{n,n}$	$x_{n\cdot} = \sum_{i=1}^n x_{n,i}$	
Σ		$x_{\cdot,1} = \sum_{i=1}^n x_{i,1}$...	$x_{\cdot,n} = \sum_{i=1}^n x_{i,n}$	$\Sigma\Sigma = N$	

To create our collection of ground-truth data we draw randomly 10 image tiles from all three continental regions (Latin America, Africa, Asia/Australia) namely the following tiles: 25N 024E, 20N 096E, 20N 102E, 20N 120E, 15N 066W, 15N 012W, 15N 072E, 10N 090W, 10N 018E, 10N 078E, 05N 072W, 05N 006E, 05N 114E, 00N 096W, 00N 060W, 00N 048W, 00N 006E, 00N 096E, 00N 114E, 00N 132E, 00N 138E, 05S 048W, 05S 042W, 05S 060W, 05S 030E, 10S 018E, 15S 060W, 15S 042E, 15S 054E and 20S 030E where the first part is the latitude coordinate and the last part is the longitude coordinate of the upper left corner. From each tile, we sampled by random 200 pixels which total to 6000 samples over the entire study region. The sampling is realized with our own raster sampling algorithm build in Python by means of the open source libraries numpy and rasterio. As mentioned in the previous section do we superimpose two datasets and only a certain amount of pixels per tile is classified as a proximate driver. Therefore, the sampling algorithm should only draw samples from

occupied/classified pixels without replacement. The algorithm expects as parameters a raster image, the total number of samples to draw, a list of pixel values which should be interpreted as occupied cells, the affine transformation matrix of the raster image, and a seed for the random number generator. If occupied cells are set the algorithm will create a binary mask where each occupied cell is set to one relative to the input raster image. Otherwise, it sets all pixel values greater or less than zero to one. After, the row and column coordinates of each one are extracted from the mask and converted to a flat list of coordinate tuples. Next, it draws the predefined number of samples from the list by a random order and uses the image coordinates to get the pixel value from the raster image. If an affine transformation matrix is provided the image coordinates are converted to real-world coordinates. The seed argument ensures that on every algorithm rerun the samples are drawn. For our sampling we set the parameters to the following values: samples 200, occupied pixels GL30 class values and 25 for regrowth, the affine matrix of the corresponding raster image, and the seed is 42. The per tile samples are stored as a Comma Separated Values (CSV) file.

For the collection of ground-truth data, we used a visual interpretation of satellite and aerial imagery provided by Google Maps. We developed a small JavaScript web application to access the imagery via the Google Maps API. The application expects as input a CSV file with the sampling coordinates. After upload of a sample file the user can cycle through the entries and the map jumps automatically to the coordinates of the sample. Now a reference label can be assigned to the coordinates by visual interpretation of the imagery. We subsequently assigned to all 6000 samples a reference label and downloaded the results as CSV.

Finally, we developed a Python class to compute the confusion matrix. The constructor of the class requires a list of reference and prediction labels. With the provided arguments it creates the confusion matrix. Further, it computes the following marginal and conditional distributions: overall accuracy $OvAc$ by dividing the sum of classification agreements by the sample total N (equation 1.3), the producer accuracy $PAc.n$ by dividing the category agreement by the column category total (equation 1.4), the error of commission $Com.n$ (Type II error) by dividing the category disagreement by the column category total (equation 1.5), the user accuracy $UAc.n$ by dividing the category agreement by the row category total (equation 1.6), the error of omission $Om.n$ (Type I error) by dividing the category disagreement by the row category total (equation 1.7), and the Cohens Kappa by substituting equation 1.8 and 1.3 into equation 1.9.

$$p_0 = OvAc = \frac{\sum_{i=1}^n x_{i,i}}{N} \quad (1.3)$$

$$PAc.n = \frac{x_{i,i}}{x_{.n}} \quad (1.4)$$

$$Com.n = \frac{FN_i}{x_{.n}} \quad (1.5)$$

$$UAc_n = \frac{x_{i,i}}{x_n} \quad (1.6)$$

$$Om_n = \frac{FP_i}{x_n} \quad (1.7)$$

$$p_c = \frac{1}{N^2} \sum_{i=1}^n x_{i,i} \cdot x_i \quad (1.8)$$

$$Kappa = \frac{p_0 - p_c}{1 - p_c} \quad (1.9)$$

1.2.3. Emissions

Land cover change respectively deforestation releases carbon emissions. These emissions can be grouped to different categories like emissions from transportation, biomass removal, changes of soil carbon dynamics, processing of certain kind of commodities etc.. During the previous sections we developed an approach to predict the change of tree cover driven by proximate causes like conversion to cropland or else. Now we can use these predictions to approximate the CO₂ emissions uprising from this land cover transitions. For this study, we focus on the emissions emitted by biomass removal and from changes of soil carbon stock. The first paragraph is focused on the estimation of emissions from biomass removal and the second section tries to approximate the impact of land cover change on the soil organic carbon content.

To obtain the gross CO₂ emissions through proximate deforestation driver we selected the following raster tiles from our AISM: the AGB stratum and our classification of the PDD. By means of Python, we implemented a function which accepts as parameter two raster images, the area a pixel covers in m², a factor to convert carbon to CO₂, and a list of proximate driver classes to consider as deforestation. We considered the following PDD classes as deforestation for the computation: 10 (cropland), 25 (regrowth), 30 (grassland), 40 (shrubland), 50 (wetland), 60 (water bodies), 70 (Tundra), 80 (artificial), and 90 (bareland). The function computes the gross emissions by using equation 1.10. Let Y_{ij} be the AGB in Mg C ha⁻¹ and X_{ij} the PDD at an pixel index i, j obtained from a raster image matrix in R^{N*M} . Let A be the area in ha a pixel covers for a certain image tile. This area is calculated by using the Haversine function from equation 1.2. Factor 3.7 converts Carbon to CO₂. Let $AGBE_{tile}$ be the cumulative emissions emitted from the removal of tree cover. Then this value can be obtained by taking the sum of the product of Y_{ij} and $f(X_{ij})$. Whereas the piecewise function f only evaluates to one if the proximate deforestation driver is within our set of classes we want to consider as deforestation. To obtain the gross AGB emissions through the deforestation by proximate deforestation driver we aggregated the sum of $AGBE_{tile}$ for the regions

Latin America, Asia, and Africa.

$$AGBE_{tile} = 3.7A \sum_{i=0}^N \sum_{j=0}^M f(X_{ij})Y_{ij} \quad (1.10)$$

To obtain the gross CO₂ emissions emitted by the change of soil organic carbon content we selected the following raster tiles from our AISM: the IFL stratum, the GSOCmap, and our prediction of PDD. We decided to predict the SOC emissions for two different scenarios. In scenario one SC₁ we assume that all tree covered areas concerned by a land cover change are primary forest. For scenario two SC₂ we used IFL stratum to determine the forest type. If land cover changes within an IFL patch it concerns primary forest otherwise it is secondary forest. The SOC emissions of both scenarios can be computed by equation 1.11. Let X_{ij} be the PDD from our prediction, Y_{ij} the forest type determined by the IFL stratum, and Z_{ij} the SOC Mg C ha⁻¹ determined by GSOCmap at a pixel with index i, j obtained from a raster image matrix in R^{N*M} . Let A be the area in ha a pixel covers for a certain image tile. This area is calculated by using the Haversine function from equation 1.2. Factor 3.7 converts Carbon to CO₂. Let $SOCC_{tile}$ be the cumulative soil organic carbon emissions emitted by the change of forest to another land cover type. Then this value can be obtained by taking the sum of the product of Z_{ij} and $h(X_{ij}, Y_{ij})$. Whereas the piecewise function h returns the mean soil organic carbon change and the standard error in respect to the forest type and proximate driver class. The mappings of drive classes and forest type for both scenarios are shown in table 1.7 and 1.8. This algorithm is implemented by means of Python. The function needs as parameter the required layers whereas the IFL stratum is optional, the area a pixel covers in m², a conversion factor for carbon to CO₂, an identifier for the forest type, and if the standard error should be included during the computation of the emission. If the IFL stratum is provided the algorithm will rely on this layer to determine the forest type otherwise it uses forest type identifier. To obtain the gross SOC emissions by the transition of land cover we aggregated the sum of $SOCE_{tile}$ for the regions Latin America, Asia, and Africa.

$$SOCE_{tile} = 3.7A \sum_{i=0}^N \sum_{j=0}^M h(X_{ij}, Y_{ij})Z_{ij} \quad (1.11)$$

Table 1.7. Scenario one mapping of soil organic carbon change to proximate driver: In scenario one we assume that deforestation always occurs in primary forest. Refer to table 1.2 for the description of the proximate driver class. Standard errors of the soil organic carbon change factors are denoted in table 1.3. The symbols in superscript denote the following transitions: † Primary forest→Cropland, ‡ Primary forest→Secondary forest, and ◊ Primary forest→Grassland

Forest type	Proximate driver class					
	10	25	30	40	70	90
Primary	.252 [†]	.086 [‡]	.121 [◊]	.121 [◊]	.121 [◊]	.121 [◊]

Table 1.8. Scenario two mapping of soil organic carbon change to proximate driver: In scenario two we use the Intact Forest Landscape stratum to distinguish between deforestation in primary and secondary forest. Refer to table 1.2 for the description of the proximate driver class. Standard errors of the soil organic carbon change factors are denoted in table 1.3. The symbols in superscript denote the following transitions: † Primary forest→Cropland, ‡ Primary forest→Secondary forest, ◊ Primary forest→Grassland, § Secondary forest→Cropland, and * Secondary forest→Grassland

Forest type	Proximate driver class					
	10	25	30	40	70	90
Primary	.252 [†]	.086 [‡]	.121 [◊]	.121 [◊]	.121 [◊]	.121 [◊]
Secondary	.213 [§]	-	.064*	.064*	.064*	.064*

1.2.4. Ecosystem service values

Ecosystems have an impact on the well being and subsistence of current future generation of humanity by providing regulatory, supporting, provisioning, and cultural services. For the quantification of these ecosystem services a economic process is applied to assess the monetary value of each service per ecosystem also referred as biome. These Ecosystem Service Values (ESVs) can be a strong tool to determine the impact of certain management practices on ecosystem structures. Especially as impact analysis for our study of tropical tree cover transitions. For a comprehensive insight of the ESV dynamics, we quantified the loss of ESV from forest cover degeneration within the tropical zone. This degeneration of forest cover is frequently followed by a transition to other land cover types expressed through our PDDs. These transitions can be interpreted as the gain of ESV and are computed subsequently. Finally, to give an insight into the overall trend of both ESV dynamics we determined the balance among the monetary loss and gain. The first paragraph describes our approach to determine the ESV loss, followed by the exercise to obtain the gain in monetary units and finally we explain how to derive the balance between both values.

By applying equation 1.12 we compute the gross ESV loss from the loss of tropical tree cover for the entire set of our AISM. Let X_{ij} be the PDD from our prediction at a pixel with index i, j obtained from a raster image matrix in R^{N*M} . Let ESV_{Forest} be the ESV of tropical forest from one of our selected source datasets from table 1.9. Let A be the area in ha a pixel covers for a certain image tile. The pixel area is calculated by using the Haversine function from equation 1.2. Let $ESV_{loss,tile}$ be the cumulative loss in ESV for a certain tile from your AISM. Then this value can be determined by adding the product of $f(X_{ij})$ and $ESV_{Forest, Dataset}$. Whereas the function f returns only one if the PDD is considered as deforestation by the mapping in table 1.9. The computation of ESV loss is implemented as a Python function. Whereas the function accept as parameters a raster image of PDD predictions or a pandas data frame object. Further, the function requires as parameter the area a pixel cover in ha and the monetary value of tropical forest. Additionally, the function requires a list of PDD classes considered as the loss of tropical forest cover. We considered the following PDD

classes as anthropogenic driven forest loss: cultivated land (10), regrowth (25), grassland (30), shrubland (40), artificial surfaces (80), and bareland (90) as table 1.9 suggests. We excluded pixel classified as forest (20) by our PDD prediction because within this class we are uncertain if a deforestation event occurred. Further, we excluded transitions of forest cover to wetland or water because we assume this LC changes are largely driven by natural causes.

$$ESV_{loss,tile} = A \sum_{i=0}^N \sum_{j=0}^M f(X_{ij}) ESV_{Forest, Dataset} \quad (1.12)$$

Table 1.9. ESV biome types mapped to PDD classes: The monetary values are given in 2007 Int.\\$ ha⁻¹. Mapping of biome types to PDD classes have the following schema: 10 to cropland biome, 25 to tropical forest biome, 30 to grassland biome, and 80 to the urban biome. The abbreviations in the ESV dataset column refer to the following publications: Dg is de Groot et al., Co Costanza et al., and Siikamaki et al.

ESV dataset	Proximate driver class						
	10	25	30	40	70	80	90
Dg	-	5,264	2,871	-	-	-	-
Co	5,567	5,382	4,166	-	-	6,661	-
Wb	-	1,312	-	-	-	-	-

To estimate the gain in ESV from the transition of tropical forest to other land cover classes per AISIM tile we applied equation 1.13. Let X_{ij} be the PDD from our prediction at an pixel with index i, j obtained from a raster image matrix in R^{N*M} . Let A be the area in ha a pixel covers for a certain image tile. The pixel area is calculated by using the Haversine function from equation 1.2. Let $ESV_{gain,tile}$ be the cumulative gain of ESV per tile. Then this value can be determined by taking the sum of $h(X_{ij})$. Whereas the function h returns for a selected PDD class the corresponding monetary value. The algorithm is implemented in Python. Whereas the function accept as parameters a raster image of PDD predictions or a pandas data frame object. Further, the function requires as parameter a mapping of ESVs to PDD classes from table 1.9. Additionally, the function can be called with a exclude list of PDD classes.

$$ESV_{gain,tile} = A \sum_{i=0}^N \sum_{j=0}^M h(X_{ij}) \quad (1.13)$$

By applying equation 1.14 we compute the ESV balance for each tile of our AISIM. Let ESV_{gain} be the total ESV gain per continental region and ESV_{loss} the total ESV loss per region. Then the ESV balance $ESV_{balance}$ can be obtained by the difference of ESV_{gain} and ESV_{loss} .

$$ESV_{balance} = ESV_{gain} - ESV_{loss} \quad (1.14)$$

We aggregated our results for ESV loss, ESV gain, and ESV balance on three spatially

different levels with the following extent: global tropical zone and continental region (Latin America, Asia, and Africa).

1.2.5. Binning analysis and visualization

During the previous sections, we were focused on the exercise of creating large scale spatial explicit predictions for land cover transitions and the following consequences for humanity. Now, an appropriate method must be developed to analyze and visualize these spatial explicit datasets by generalizing the problem domain. By the nature of fine resolution raster images and the large area of our study extent we must handle a large N (many samples) and the resulting high dimensionality respectively complexity of relationships among the samples [Carr 1990]. Raster image maps can be interpreted as multivariate scatter plots. In our case this scatter plot has the three dimensions x is the longitude, y the latitude coordinate of a pixel, and z is the nominal scaled pixel value in case of the PDD prediction. Drawing scatter plots with large multidimensional N commonly leads to overplotting and hidden point densities [Carr et al. 1987]. Additionally, it is to assume that the distribution of PDD is not equally distributed over the entire study extent. Hence, there should be regions with sparse data densities and with high densities but our goal is to visualize land cover changes on a continental level. As mentioned the ground resolution of one pixel covers an area of approximately 30x30 m and as an example, the bounding box of Latin America covers an area of $5 * 10^7$ Km². The large frame size as well the unequal distributed data leads to the issue that only large scale land cover changes are representable and small scale isolated changes stay hidden.

By referring to the latter paragraph our goal should be to develop a process to solve the representation issues and generates satisfying maps. In the case of raster data, one opportunity could be a re-sampling to a coarser on-ground resolution. This approach may solve the overplotting as well resolution issues and normalize unequal distributed data. For nominal scaled data the commonly used re-sampling methods are nearest neighbor or majority wins [Reference](#). Both approaches are not appropriate because they would negate spatial patterns and eliminate important land cover class frequency distributions. Another well-accepted method is binning of spatially explicit data with a regular polygon that can tessellate the plane [Carr et al. 1992]. Polygon tessellations provide numerous opportunities for presenting multivariate statistical and visual summaries. The scale of a polygon may be used to visualize pixel densities within the bounds and a color gradient may be used to prepare a choropleth map for nominal or ordinal scaled data. Additionally, the interior of a polygon may be used to prepare a pie chart. Hence, binning enables convenient visualization of multidimensional data. For preparing a regular tessellation only three types of convex polygons can be used to tessellate the plane: squares, equilateral triangles, and hexagons [Carr et al. 1992]. Square tessellations

are the most common method in comparison with hexagons for binning and visualizing spatial data. Every raster image is already a square tessellation of the mapped object and most of the image processing algorithms are focused on squares. Hexagon mosaic maps have two major advantages over square tessellations: visual appeal and representational accuracy. Binning of data by a square or hexagon mosaic creates visual lines. These lines compete with the data generated patterns. Especially humans have a strong visual response to horizontal and vertical lines. Hence, the line artifacts of square tessellations are distracting and should be avoided. Thus, we decided to use hexagon mosaic maps to represent the visual and statistical results of our study. For bivariate representations, we select the combination of scaling and color gradient. Multivariate data is visualized by hexagonal pie charts. The following paragraphs describe our algorithmic approach to create these mosaic maps. We used Python and the open source library shapely to implement our algorithms.

The first step to construct a hexagon tessellation is to define the vertices of the polygon. There are two common orientations of hexagons in R^2 flat topped and pointy topped. For our hexagon construction we decided to use pointy topped polygons. For flexibility our algorithm accepts one out of four parameter to construct a hexagon polygon. The unit of the parameter is always in map units. Let D be the long diagonal (diameter of the circumscribing circle), d the short diagonal (diameter of the inscribed circle), A the area the hexagon should cover, and e the edge length of a hexagon. Let R be the radius of the circumscribing circle. Then R can be obtained by applying equation 1.15 with one out of the parameter set D , d , A , or e . R is used to compute the center vector $\vec{m} = \langle c_x, c_y \rangle$ of the polygon by applying equation 1.16 and 1.17. The polygon center is always located in the first quadrant of the Cartesian coordinate system. Now, by using R , c_x , and c_y we can obtain \mathbf{H} the anti-clockwise orientated vertex matrix of a hexagon. The construction of a hexagon by using the introduced method is shown in the left bottom corner of figure 1.5. The next paragraph describes how we derive a tessellation from the constructed hexagon.

$$R = \frac{\sqrt{2A}}{\sqrt[4]{27}} = \frac{D}{2} = \frac{d}{\sqrt{3}} = e \quad (1.15)$$

$$c_x = \frac{R\sqrt{3}}{2} \quad (1.16)$$

$$c_y = R \quad (1.17)$$

$$\mathbf{H} = \begin{bmatrix} 0 & c_x & 2c_x & 2c_x & c_x & 0 \\ R\sin\left(\frac{7\pi}{6}\right) + c_y & 0 & R\sin\left(\frac{11\pi}{6}\right) + c_y & R\sin\left(\frac{\pi}{6}\right) + c_y & 2R & R\sin\left(\frac{5\pi}{6}\right) + c_y \\ 1 & 1 & 1 & 1 & 1 & 1 \end{bmatrix} \quad (1.18)$$

To create a polygon grid of a plane image we must align several hexagons to cover the image. For our tessellation algorithm we use the vertex matrix \mathbf{H} computed by the previously

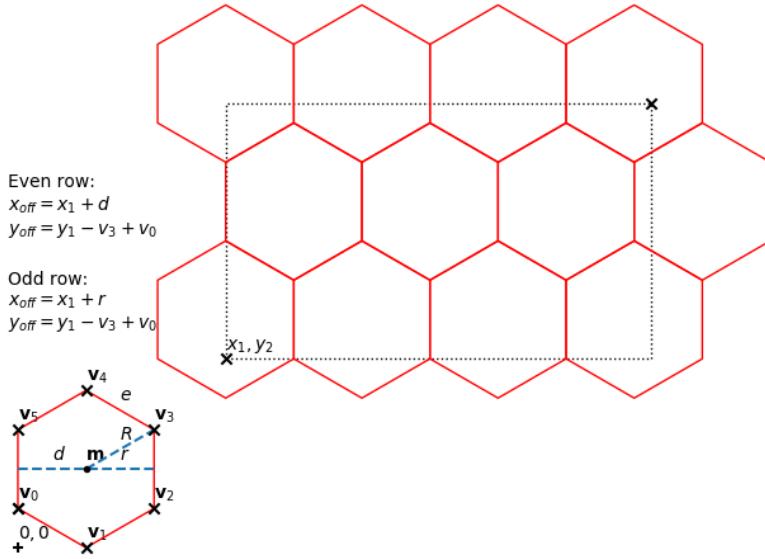


Figure 1.5. Hexagon tessellation: Located at the left bottom corner in red a hexagon defined by its geometric properties the 6 vertex vectors $\{\vec{v}_0, \dots, \vec{v}_5\}$ (black crosses), with center vector \vec{m} , edge length e , R radius of the circumscribing circle, r radius of the inscribed circle and d the short diagonal (diameter of the inscribed circle). Top right black dotted box are the bounds of an area which is tessellated by a hexagon grid in red. Each grid cell is translated from the origin hexagon at its position by computing the x_{off} and y_{off} offset with the presented equations at the left-hand side of the grid.

described approach and subsequently translate it to its position within the grid. We expect to receive the bounds matrix \mathbf{B} of the raster image which should be tessellated by hexagons, equation 1.19. Let x_1, y_1 be the left bottom corner coordinates and x_2, y_2 the right top corner coordinates of an image. Let $x_{off}(0), y_{off}(0)$ in equation 1.20 and 1.21 be the initial coordinates for creating a polygon grid over a plane. Then we can obtain $x_{off}(n+1)$ the x coordinates for even rows by applying equation 1.22 and $x_{off}(n+1)$ the x coordinates for odd rows by equation 1.23. Where r is the radius of an inscribed circle in a hexagon and can be obtained by dividing d by 2. Then \mathbf{H} can be translated to the vertex matrix \mathbf{T} by applying the dot product of an affine transformation matrix and \mathbf{H} , equation 1.25.

$$\mathbf{B} = \begin{bmatrix} x_1 & x_2 \\ y_1 & y_2 \end{bmatrix} \quad (1.19)$$

$$x_{off}(0) = x_1 \quad (1.20)$$

$$y_{off}(0) = y_1 \quad (1.21)$$

$$x_{off}(n+1) = x_{off}(n) + d \quad (1.22)$$

$$x_{off}(n+1) = x_{off}(n) - r + d \quad (1.23)$$

$$y_{off}(n+1) = y_{off}(n) - v_{0,2} + v_{3,2} \quad (1.24)$$

$$\mathbf{T} = \begin{bmatrix} 1 & 0 & x_{off}(n) \\ 0 & 1 & y_{off}(n) \\ 0 & 0 & 1 \end{bmatrix} \circ \mathbf{H} \quad (1.25)$$

Goal: A pie chart within the area of a hexagon, split the hexagon in horizontal pieces which represent a ratio

How: Compute the y coordinate of the split line from the ratio, the distance between y1 and y2, compute from this the x coordinates

$$y = \frac{P(y_2 - y_1)}{100} + y_1 \quad (1.26)$$

$$f^{-1}(y) = \begin{cases} -\frac{y-y_1}{\tan(\frac{\pi}{6})} + \frac{x_1+x_2}{2} & \text{if } y_1 \leq y < y_1 + R \sin(\frac{5\pi}{6}) \\ x_1 & \text{if } y_1 + R \sin(\frac{5\pi}{6}) \leq y < R(\sin(\frac{5\pi}{6}) + 1) \\ \frac{y-y_2}{\tan(\frac{\pi}{6})} + \frac{x_1+x_2}{2} & \text{if } R(\sin(\frac{5\pi}{6}) + 1) \leq y \leq y_2 \end{cases} \quad (1.27)$$

$$g^{-1}(y) = \begin{cases} \frac{y-y_1}{\tan(\frac{\pi}{6})} + \frac{x_1+x_2}{2} & \text{if } y_1 \leq y < y_1 + R \sin(\frac{5\pi}{6}) \\ x_2 & \text{if } y_1 + R \sin(\frac{5\pi}{6}) \leq y < R(\sin(\frac{5\pi}{6}) + 1) \\ -\frac{y-y_2}{\tan(\frac{\pi}{6})} + \frac{x_1+x_2}{2} & \text{if } R(\sin(\frac{5\pi}{6}) + 1) \leq y \leq y_2 \end{cases} \quad (1.28)$$

$$\mathbf{L} = \begin{bmatrix} f^{-1}(y) & g^{-1}(y) \\ y & y \end{bmatrix} \quad (1.29)$$

Goal: All use hexagons with an area of 0.5 degrees i square, aggregated tiles per continent americas, africa, asia

Treecover: Count tree cover pixels within hexagon in canopy interval (10,100], count total pixels within hexagon, compute pixel area with haversine, divide tree covered area by total area, 5 ratio bins 0.2 0.4 0.6 0.8 1.0 for scaling, store the area for interval, mean canopy density over occupied pixels, use mean canopy density for choropleth map

Loss: count frequency of pdd recognized as deforestation

PDD: count frequencies within a hexagon, compute frequency ratios, segment hexagons by the ratio of pdd driver, order the drivers in decreasing order, most common is first, tried with scaling but not feasible cause few big sized and many small sized, just scale them a bit down cause visual appeal,

Hexagonal Country-boundaries: Hexagon grid over the continent aggregate cells which cover a country, vastly made by hand

2. Results

2.1. Deforestation

2.1.1. Forest definition

Our goal is to determine at which canopy cover density the agreement between GL30 and GFC tree cover is greatest to receive the subsequent PDD for stable LC transitions introduced by anthropogenic causes. This process should ensure that we keep the largest number of tree cover loss samples from the GFC dataset while harmonizing the tree cover definition between both layers. We applied the JI to determine the similarity between each tile pair from our AIS. The JI computation is grouped by the continental regions Latin America (82 tiles), Asia/Australia (86 tiles), and Africa (101 tiles). We determined the similarity for the following canopy density intervals: (0, 100], (10, 100], (20, 100], and (30, 100]. Later we excluded all tiles with a initial JI (canopy density interval (0, 100]) from our analysis because these tile pair does not contain any tree cover. We excluded 6, 9, and 15 tiles for Latin America, Asia/Australia and Africa, respectively. To determine the canopy density interval where the agreement is at maximum we applied the non-parametric tests Wilcoxon signed-rank test and Wilcoxon rank-sum test. Both of the tests are performed as a one- and two-sided to deduce, if there is a difference in agreement (equality) and which direction (less or greater) has this difference. To address the higher probability of family-wise error rate in multiple comparisons we applied a Holm correction for Wilcoxon signed-rank tests and a Benjamini and Hochberg correction for Wilcoxon rank-sum test. We applied continental and global testing to deduce regional differences and to determine the optimum for our subsequent PDD predictions. Further, we compared the tree cover agreement of the three continental regions. Our initial hypothesis was that the tree cover agreement is at its maximum within the canopy density interval of (30, 100] for the entire study extent. We assumed that for Latin America and Asia/Australia the best results could be achieved with the same canopy density threshold. For Africa we assumed that the highest agreement could be achieved within the interval of (10, 100] because this region comprises a higher frequency of sparse woodland cover. The following paragraphs present our results for the three regions in the following order: Latin America, Asia/Australia and Africa. The last paragraph discusses the results for

the entire study extent and determines which canopy density we used for the following PDD prediction.

For Latin America, Asia/Australia and Africa as well the entire study extent figure 2.1 shows the distribution of the computed JIs for all tile pairs within the canopy density intervals. The x-axis are the different canopy density intervals where the label JI_0 accounts for $(0, 100]$, JI_1 $(10, 100]$, JI_2 $(20, 100]$, and JI_3 $(30, 100]$, respectively. The y-axis is the corresponding JI between 0 and 1 where 0 highlights a complete disagreement and 1 a full agreement. The sample mean is labelled by a red cross and the boxes comprises the Q_1 (25 %), Q_2 (50 %), and Q_3 (75 %) sample interval, respectively.

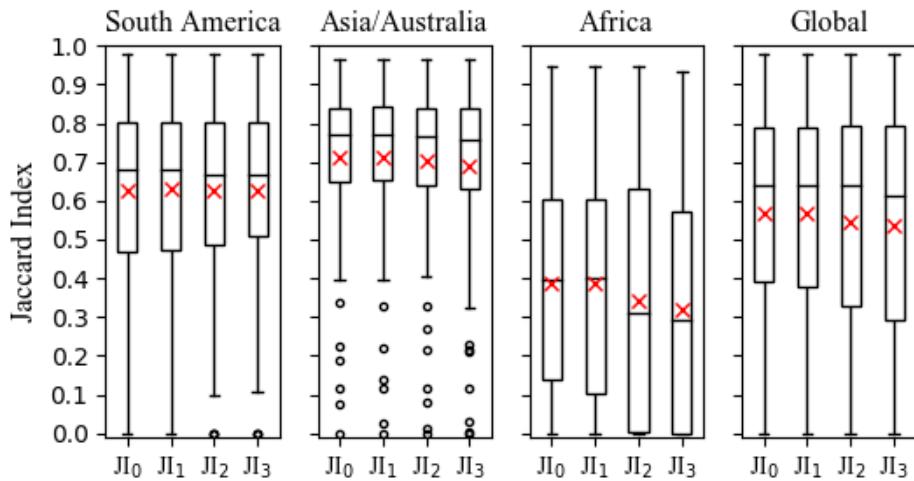


Figure 2.1. Tree cover similarity distribution: This box-plot shows the distribution of Jaccard Indexes for each raster image tile pair of GlobeLand30 and Global Forest Change tree cover from 2000. The labels JI_0 , JI_1 , JI_2 , and JI_3 on the x-axis account for the canopy density classes $(0, 100]$, $(10, 100]$, $(20, 100]$, and $(30, 100]$, respectively. The y-axis is the Jaccard Index of the corresponding raster image pair, where 0 is a total disagreement and 1 a total agreement. Red crosses within the Q_{25} , Q_{50} , and Q_{75} boxes highlight the sample mean. Whiskers are 1.5 times the IQR .

For Latin America the sample mean does not change significantly within the four canopy density experiments. It is approximately 0.62 while sample median decreases from 0.68 to 0.66 from the interval class to the last. The upper 25 % of the first experiment interval have a tree cover similarity ranging between approximately 0.8 and 1. This holds true over the other three experiments while only the maximum similarity slightly increases from 0.9787 to 0.9798. The figure B.1 in the appendix B suggests, that exclusions of canopy densities smaller than 11% increases slightly the tree cover agreement of the upper 25% but the exclusion of higher canopy density provides no benefit. This can be explained by the fact that the upper percentile already contain samples with a high tree cover agreement and it is to assume that only a small number of pixels have canopy density smaller than 30. Therefore, the interval change has only a small impact on this samples. For the first two canopy density intervals the range of the lower 25% percentile is between approximately 0.0003 and 0.45.

Whereas the range increases from 0.0 to 0.5 for the last two interval classes. The figure B.1 reveals a strong regional dependency of tree cover agreement within different canopy density classes of the lower percentiles. Whereas samples from the northern hemisphere show a decline in agreement the samples of the southern hemisphere show an increase of agreement. The strong up-shift of the southern samples within the last two intervals increases the range of lower percentile. This suggests in general that samples with low tree cover agreement benefit by the exclusion of lower canopy densities. The mobility of the samples within the Q_1 and Q_3 percentile show no general trend. For the first two experiments it ranges between 0.5 and 0.8 and for the last two it shows a small decline in IQR . This suggests that tiles within this group could benefit from a local optimization of canopy density exclusion. We applied a Wilcoxon signed-rank test to deduce which canopy density class yields the highest tree cover agreement overall samples in Latin America. Table 2.1 shows the results for the two-sided test and table 2.1 for the one-sided test. The two-sided test reveals that only the similarity distribution between Jl_0 and Jl_1 is significantly different ($p < 0.01$), while the comparison of the other distributions suggest that they are equal. The directional test of Jl_0 and Jl_1 suggests that the regional tree cover agreement is significantly greater ($p < 0.005$) if canopy densities smaller than 10 are excluded. The directional test does not confirm that the distribution of Jl_1 is significantly greater than Jl_2 and Jl_3 . Therefore it can be assumed that the exclusion of lower canopy densities could yield better results for certain tiles. The result of our experiments suggests that the tree cover agreement between GL30 and GFC at canopy densities greater than 10 is at its maximum for Latin America. In case of local studies or for smaller extents the canopy density should be selected by a single tile approach to optimize the tree cover agreement by maximizing the number of data points of the GFC dataset.

Table 2.1. Continental experiment group comparison: This table shows the results of a two-sided Wilcoxon signed-rank test to detect continental differences in the tree cover agreement by considering different canopy densities between GlobeLand30 and Global Forest Change at 2000. The classes Jl_0 , Jl_1 , Jl_2 , and Jl_3 as row and column headings account for the canopy density classes (0,100], (10,100], (20,100], and (30,100], respectively. The test hypothesis is $H_0: X_1 = X_2$ where X_1 is the column Jl_n class and X_2 the row Jl_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.02$, and $p^{***} < 0.01$.

Cls	Latin America			Asia/Australia			Africa		
	Jl_0	Jl_1	Jl_2	Jl_0	Jl_1	Jl_2	Jl_0	Jl_1	Jl_2
Jl_1	.00***	-	-	.71	-	-	.22	-	-
Jl_2	.30	1.	-	.00***	.00***	-	.09	.09	-
Jl_3	.64	1.	1.	.00***	.00***	.00***	.00***	.00***	.00***

For Asia/Australia, the sample mean scatters around 0.7 as the red crosses in figure 2.1 suggest. The sample mean decreases slightly at higher canopy density intervals. Further, the median is approximately 0.8 by showing a slight decrease at higher canopy density intervals too. The range of the upper percentiles for all experiments groups is between approximately 0.85 and 0.96 while the maximum agreement decreases slightly from 0.9654 to 0.9634. Fig-

Table 2.2. Continental experiment group directional comparison: This table shows the results of a one-sided Wilcoxon signed-rank test to detect the direction of continental differences in the tree cover agreement by considering different canopy densities between GlobeLand30 and Global Forest Change at 2000. The classes JI_0 , JI_1 , JI_2 , and JI_3 as row and column headings account for the canopy density classes (0,100], (10,100], (20,100], and (30,100], respectively. The test hypothesis is $H_0: X_1 \leq X_2$ and $H_0: X_2 \geq X_1$ where X_1 is the column JI_n class and X_2 the row JI_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.025$, $p^{***} < 0.01$, and $p^\dagger < 0.005$.

Cls	Latin America				Asia/Australia				Africa			
	JL_0	JL_1	JL_2	JL_3	JL_0	JL_1	JL_2	JL_3	JL_0	JL_1	JL_2	JL_3
JL_0	-	.00 ^{\dagger}	.14	.33	-	1.	1.	1.	-	.65	1.	1.
JL_1	1.	-	.55	.55	.36	-	1.	1.	.89	-	1.	1.
JL_2	1.	1.	-	.55	.00 ^{\dagger}	.00 ^{\dagger}	-	1.	.03*	.03*	-	1.
JL_3	1.	1.	1.	-	.00 ^{\dagger}	.00 ^{\dagger}	.00 ^{\dagger}	-	.00 ^{\dagger}	.00 ^{\dagger}	.00 ^{\dagger}	-

ure B.2 in the appendix B, reveals that in general the tree cover agreement increases if canopy densities below 10 % are excluded but the exclusion of densities above 20 % reverts this. For Asia/Australia the lower percentile ranges between approximately 0 and 0.65 while most of the samples show a decline in tree cover agreement if canopy densities above 20 are excluded. Within the $Q_{1,3}$ percentile a per tile relationship is detectable. The range of this percentile is between 0.65 and 0.85 for the first two classes JL_0 and JL_1 while the range increases for the last two experiments. As mentioned no clear trend is observable some of the samples benefit if the considered canopy density interval is increased till 30 % and some show a decrease in agreement if the canopy density is lift over 10 %. The two-sided Wilcoxon test in table 2.1 reveals that the similarity distribution is significantly different ($p<0.01$) between each experiment group except the pair of JL_1 and JL_0 . The directional test in table 2.2 reveals that the tree cover agreement distributions of JL_2 and JL_3 are significantly smaller than JL_0 and JL_1 ($p<0.005$). Further this results show that the distributions of JL_0 and JL_1 have no directional differences. This could be explained by a regional or tile-wise agreement component. While some of the tiles show strong increase in similarity if canopy density is set to 10 %, others show a decrease. A more detailed analysis could be performed by applying a smaller canopy density step-size. For studies targeting the region Asia/Australia the results of the directional tests suggest to use all data from GFC within the canopy density interval of (0,100]. While the figure B.2 suggests to include all data form the interval (10,100].

The box-plot for Africa in figure 2.1 shows that the similarity range of the upper 75 % is the greatest among our study regions. It ranges between 0.15 and 1.0 for JL_0 while the range increases if smaller canopy densities are excluded. The first two experiment groups JL_0 and JL_1 have a nearly similar mean and median of approximately 0.38 and 0.4, respectively. Whereas both metrics show a strong decline to 0.33 and 0.3 in the last two experiment groups JL_2 and JL_3 , respectively. Figure B.3 in the appendix B reveals that the upper 25 % different to Latin America and Asia not benefit in general by exclusion of lower canopy densities. Africa has the highest amount of tiles where there tree cover agreement is smaller than 0.1. For the sec-

ond experiment group JI_1 11 tiles from the AISM have a complete disagreement ($JI = 0$) if the canopy density interval is reduced to $(10, 100]$. This trend continues if the canopy density interval is further reduced. This already suggests that reducing the canopy density interval for Africa is not feasible. The two-sided test in table 2.1 reveals that JI_3 is significantly different ($p < 0.01$) in its similarity distribution compared to the other experiment groups JI_0 , JI_1 , and JI_2 , respectively. The other experiment groups could origin from the same similarity distribution. Especially, the tree cover agreement between JI_0 and JI_1 show evidences that it could origin from the same distribution. This is comparable with the Asia/Australian region. The table 2.2 shows the directional component of the tree cover agreement between the experiments groups for Africa. The tree cover agreement of JI_2 and JI_3 is significantly smaller than JI_0 or JI_1 as the table suggest. Therefore the exclusion of canopy densities above 10 % reduces the continental tree cover agreement in Africa. For the two experiment groups JI_0 and JI_1 no directional agreement component could be proofed but the figure B.3 shows a strong regional component of tree cover agreement. The strong regional component can be explained by the high share of sparse woodland in Africa. The figure 2.6 in section 2.1.2 shows that different from Asia/Australia and Latin America a vast amount of African tree cover has a canopy density between 0 and 46 %. The results for Africa suggests to set the canopy density interval to $(0, 100]$ to optimize the tree cover agreement between GL30 and GFC on a continental level. Preferable the strong regional component suggests to optimize the tree cover agreement per tile to achieve a higher tree cover agreement for the entire continent.

The table B.1 in the appendix B shows that in Asia/Australia the tree cover similarity between GL30 and GFC is the greatest out of all three regions. Only within the last experiment group JI_3 the median tree cover agreement between Asia/Australia and Latin America could be the same shown in table B.2. Africa has the poorest tree cover agreement out of our three continental regions. Asia's higher tree cover agreement compared to Latin America could relate to the smaller total landmass and vice versa forest cover within this region and the high share of dense forest cover. Therefore both global LC datasets have a smaller chance to miss forest covered pixels. That both regions achieve better tree cover similarity as Africa could be explained by the high amount of dense forest cover within this regions. It could be assumed that the probability of misclassification of forest cover increases as lower the canopy density is. In regards of tile-wise tree cover agreement optimization would Africa benefit at most out of the three regions. In Asia/Australia and Latin America only certain tiles would benefit from a exclusive optimization as the figures B.1 and B.2 suggests.

On the far right of figure 2.1 the tree cover agreement of the entire study extent is shown. The sample mean and median differ between the first two experiment groups and the last two groups. For the first two groups the mean and median account for 0.56 and 0.63, respectively. The last two experiment groups show a decline to 0.53 and 0.53 for mean and median.

For the upper percentile the same statement as for the regions holds true. In general, the samples with high tree cover agreement benefit from the exclusion of lower canopy densities. The lower percentile shows strong regional or tile-wise tree agreement dependencies. If we consider the entire sample range the mid percentile steadily increases its range from 0.4 to 0.8 till 0.3 to 0.8. As mentioned in the regional analysis, this percentile is characterized by inhomogeneous changes of tree cover agreement by the step-wise change of the canopy density. In general the trend points downwards as the decrease in median and the range increase show. The results of the two-sided Wilcoxon test in table 2.3 show a significant difference in distribution ($p < 0.02$ and $p < 0.01$) between each experiment group except Jl_0 and Jl_2 where the similarity distribution could be the same. Table 2.4 highlights the directional component of these distribution differences. At a global scale the tree cover agreement is at its maximum if we set the canopy density interval to $(10, 100]$. The second experiment group Jl_1 is significantly greater than Jl_0 ($p < 0.005$) and the last two groups Jl_2 and Jl_3 are significantly smaller ($p > 0.005$) than Jl_1 . Further, the tree cover agreement of Jl_0 is significantly greater than ($p < 0.005$ and $p < 0.05$) Jl_2 and Jl_3 . Therefore it is proved that canopy densities above 20 % reduce the tree cover agreement between GL30 and GFC on a global level. On a global scale as the analysis on tree cover agreement suggests the highest agreement can be achieved within the Jl_1 interval. Therefore we decided to proceed for our study of PDD and the derived products with this definition of tree cover.

Table 2.3. Global experiment group comparison: This table shows a two-sided Wilcoxon signed-rank test to detect differences in the tree cover agreement by considering different canopy densities between GlobeLand30 and Global Forest Change at 2000. The classes Jl_0 , Jl_1 , Jl_2 , and Jl_3 as row and column headings account for the canopy density classes $(0,100]$, $(10,100]$, $(20,100]$, and $(30,100]$, respectively. The test hypothesis is $H_0: X_1 = X_2$ where X_1 is the column Jl_n class and X_2 the row Jl_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.02$, and $p^{***} < 0.01$.

Cl	Jl_0	Jl_1	Jl_2
Jl_1	.01**	-	-
Jl_2	.07	.04*	-
Jl_3	.00***	.00***	.00***

Table 2.4. Global experiment group directional comparison: This table shows a one-sided Wilcoxon test to determine the direction of global differences in the tree cover agreement by considering different canopy densities between GlobeLand30 and Global Forest Change at 2000. The test hypothesis is $H_0: X_1 \leq X_2$ and $H_0: X_2 \geq X_1$ where X_1 is the column Jl_n class and X_2 the row Jl_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.025$, $p^{***} < 0.01$, and $p^\dagger < 0.005$.

Cl	Jl_0	Jl_1	Jl_2	Jl_3
Jl_0	-	.01***	1.	1.
Jl_1	1.	-	1.	1.
Jl_2	.07	.03*	-	1.
Jl_3	.00†	.00†	.00†	-

2.1.2. Patterns of tree cover and deforestation

This section is intended to present a comprehensive insight in the tropical tree cover distribution within our study extent at 2000 over the tree continental regions. Further, we highlight at which sites the tree cover loss peaked between 2000 and 2010. The tree cover maps are derived from the GFC tree cover 2000 layer while we selected only pixels within our forest definition which refers to the canopy density interval (10, 100]. For an appropriate visualization of multivariate spatial data on a large extent we selected our hexagonal binning approach. For the tree cover maps we computed the total area covered by trees within a polygon and divide by the total area of the hexagon to determine the scaling. Additionally we aggregate the canopy density within a hexagon by applying the arithmetic mean. To arrange the tree cover loss maps we used our PDD products. We computed the loss area within each hexagon for the following PDD classes: cultivated land (10), regrowth (25), grassland (30), shrubland (40), artificial surfaces (80), and bareland (90). To determine the polygon scaling we divided the per hexagon loss area by the highest observed loss within a continental region. Forest cover, losses, and hexagon areas are computed by applying the Haversine equation. A hexagon in unscaled shape covers an area of 0.5 decimal degrees. The maps in this section should be interpreted as precursor to our PDD predictions to detect regional and continental patterns of deforestation and as an example how large multivariate spatial data can be visualized and evaluated by a more advanced aggregation approach.

The map in figure 2.2 shows the tree cover and canopy density distribution within our study extent for Latin America at 2000. The center of the map is covered by the core tropical rain forest characterized by high tree cover per hexagon between approximately 39 and 49 thousand km² and a mean canopy density between 82 and 100 %. The core rain forest zone is distributed over 9 Latin American countries, namely: Colombia, Venezuela, Guiana, Suriname, French Guiana, Brazil, Bolivia, Peru, and Ecuador. In Brazil, within the tropical core zone two hexagons are smaller scaled which highlights a tree cover between 29 and 39 thousand km² at a canopy density between 82 and 100 %. These two hexagons comprising floodplain forest at the borders of the Amazon river located in the state Pará. This river basin enclosed by the three cities Santarém, Almeirim, and Óbidos suffered severe deforestation since the 16th century [Renó et al. 2011]. Renó et al. suggests that major forest losses took place between 1950 and 1975 followed by lower annual deforestation till 2008. Located in the lower left part of the map crossing the borders of Bolivia, Paraguay, and Argentina is the Gran Chaco a hot semi-arid wooded grassland also known as tropical dry forest characterized by tree cover between approximately 29 and 49 thousand km² per hexagon and canopy densities between 28 and 46 % [Caldas et al. 2013]. The tropical rain forest is surrounded by tropical moist forest characterized by tree cover up to approximately 49 thousand km² per hexagon while the canopy density is between 10 and 82 %.

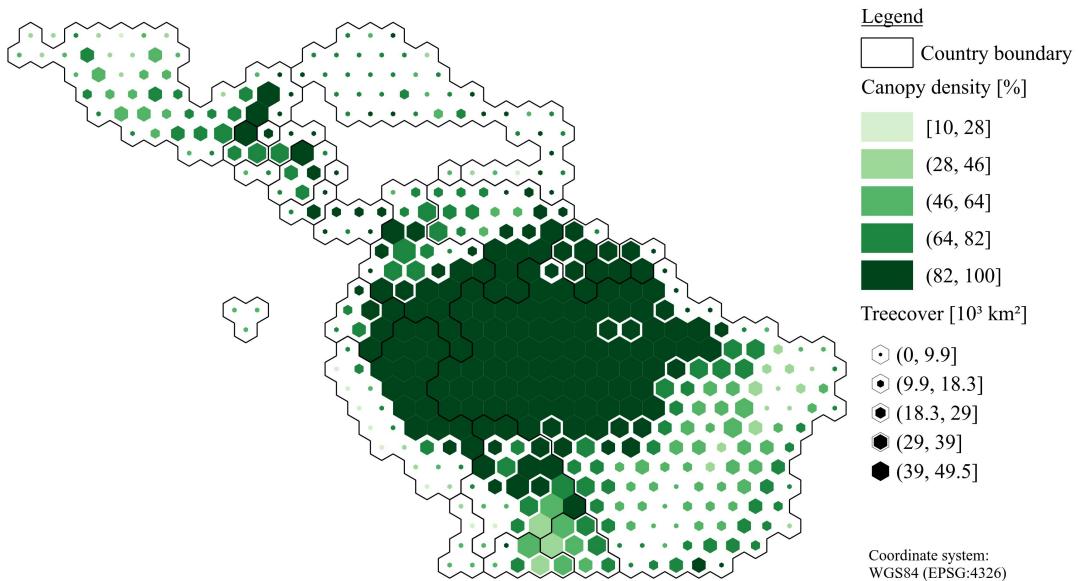


Figure 2.2. Tree cover and canopy density in Latin America at 2000: This map shows the tree cover and mean canopy density distribution at 2000. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km^2 at the equator. Tropical rain forest in the center of the map is characterized by a tree cover between approximately 29 and 49 km^2 and canopy densities between 82 and 100 %. The rain forest is surrounded by tropical moist forest with tree cover up to 49 km^2 but lower mean canopy densities between 10 and 82 % as the rain forest. The tropical dry forest also known as Gran Chaco is located in the lower left of the map distributed over the countries Bolivia, Paraguay, and Argentina.

The figure 2.3 shows the distribution of tree cover losses over Latin America within the time frame of 2000 till 2010. During this period an area of approximately 388 thousand km^2 was deforested as table C.1 in the appendix C shows. We could identify for several countries deforestation hot-spots where the deforested area is between approximately 2.7 and 9 thousand km^2 as the map suggests. Tropical countries with deforestation hot-spots are: Paraguay, Argentina, Bolivia, Brazil, Colombia, Peru, and Guatemala. In Brazil the hot-spots of tree cover loss are known as arc of deforestation and cover the states of Acre, Rondônia, Mato Grosso, and Pará while the deforestation starts to move into the state of Amazonas [Wood 2002]. The tree cover loss within this arc develops along the highway network in this regions [Alves 2002; Müller et al. 2016]. Several paved highways like BR-163, BR-219, BR-230 etc. foster the agricultural and infrastructural development and lead to high deforestation rates. In Brazil an area of approximately 274 thousand km^2 is deforested between 2000 and 2010. Deforestation hot-spots in Paraguay and Argentina are located in the Gran Chaco region which was one of the least disturbed forests worldwide and is exposed to high deforestation rates since 1969 [Caldas et al. 2013; Zak et al. 2004]. During our study time frame deforestation accounts for an area of approximately 7097 km^2 and 21 thousand km^2 in the Chaco region in Argentina and Paraguay, respectively. In Bolivia the deforestation hot-spot is located in the department of Santa Cruz. Bolivia had till the 1960s relative low deforestation rates which increased moderately after and increased sharply during the 1990s and remain high as the map suggests [Pacheco 2002; David Kaimowitz and Vanclay 2002].

During the first decade of 2000 an area of approximately 19 thousand km² is exposed to tree cover loss. The department of Petén in Guatemala is committed to deforestation since the 1980s [Beach 1998]. Since 2000 the deforestation rates in Guatemala are among the highest in Latin America and after 2005 the rates increased further [McSweeney et al. 2014]. During 2000 and 2010 an area of approximately 6515 km² is exposed to tree cover loss as table C.1 shows. McSweeney et al. highlights that deforestation hot-spots often spatially overlap with drug trafficking nodes. Especially in the department of Petén where large ranches are owned by narco-traffickers.

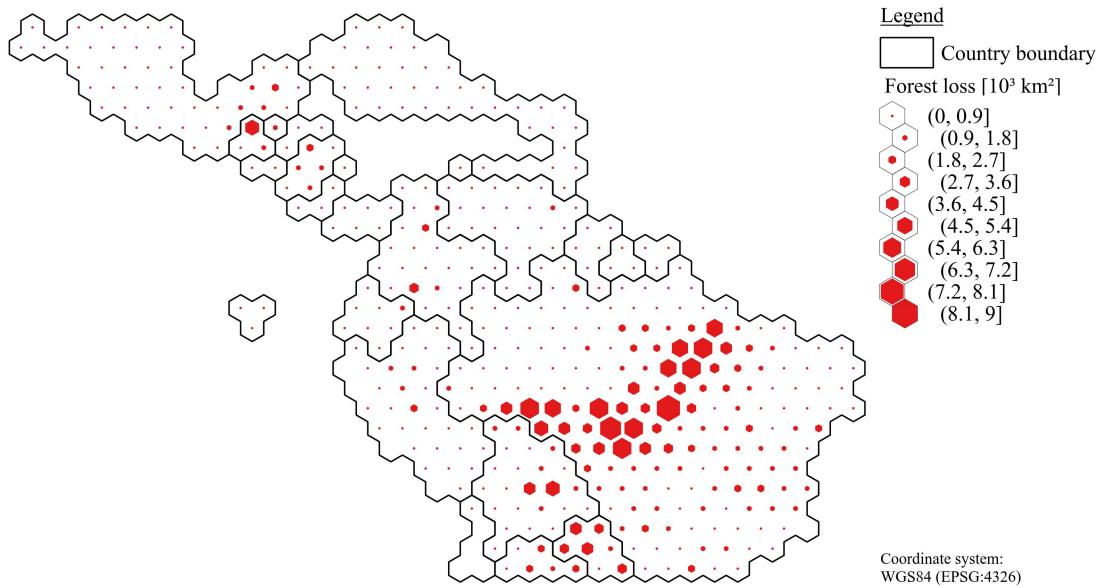


Figure 2.3. Tree cover loss in Latin America between 2000 and 2010: This map shows the tree cover loss within our study extent between 2000 and 2010. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km² at the equator. Deforestation hot-spots with tree cover loss about 2.7 thousand km² per hexagon are located in Paraguay, Argentina, Bolivia, Brazil, Colombia, Peru, and Guatemala.

The map in figure 2.4 shows the tree cover and canopy density distribution for Asia/Australia. In Asia the tropical rain forest is distributed over several southeast Asian islands like the Indonesian islands Sumatra, Borneo, Java, Papua etc., the Philippines, Malaysia, and Papua New Guinea. Generally the rain forest is characterized by high tree covers comparable to Latin America but in our map most of the islands are smaller than a single hexagon. Therefore, by our method to compute the hexagon scaling the size of the polygons don't reflect the tree cover as share of the landmass and most of them appear smaller. The mean canopy density of tropical rain forest is between 82 and 100 % and the tree cover is above approximately 9.9 km². Tropical moist and dry forest are distributed over the continental Asia covering the countries India, Vietnam, Cambodia, Laos etc. These forests are characterized lower tree cover among 9.9 km² and canopy densities between 10 and 82 % while the lower tree cover indicates that in southeast Asia deforestation occurs since the 1950s [Kummer and Turner 1994].

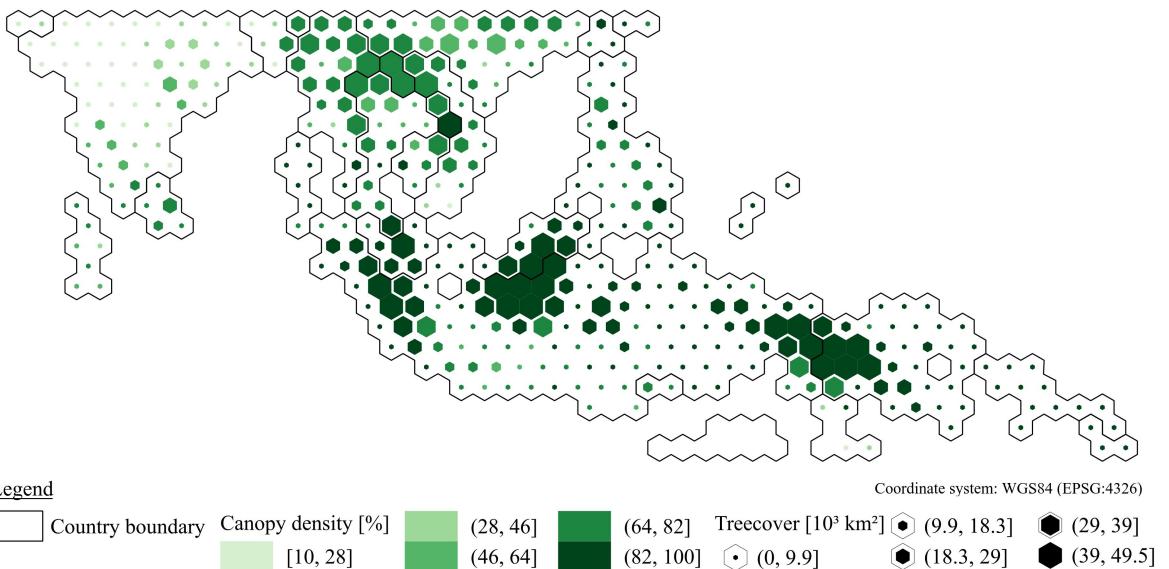


Figure 2.4. Tree cover and canopy density in Asia/Australia at 2000: This map shows the tree cover and mean canopy density distribution within our study extent at 2000. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km^2 at the equator.

The map in figure 2.5 shows the distribution of tree cover loss in Asia/Australia for the time period 2000 till 2010. During this period a forest area of approximately 196 thousand km^2 is exposed to deforestation as table C.1 in the appendix C suggests. For this region we identified the following countries as deforestation hot-spots with deforestation areas per hexagon of approximately 1.9 to 9 thousand km^2 : Indonesia, continental and insular Malaysia, Vietnam, and Laos. Indonesia is known as a country with one of the highest rates of primary forest loss for the time period 2000 till to 2016 [Austin et al. 2019]. In the first decade of 2000 an area of approximately 100 thousand km^2 is exposed to deforestation as table C.1 suggests. Tree cover loss is predominantly distributed over the Indonesian islands of Sumatra and Borneo (Kalimantan) as the map 2.5 shows. Indonesia's forests are exposed to deforestation by dynamic causes since the 1950s [Nawir and Rumboko 2007]. Malaysia has lost an area of approximately 33 thousand km^2 by deforestation during the time frame 2000 till 2010. The deforestation hot-spots in Malaysia are distributed over the Malaysian Borneo (Sabah/Sabah) and the continental Malaysia. Comparable to Indonesia, Malaysia's forests are exposed to deforestation since the 1950s by dynamic causes like logging followed by agricultural activities [Kummer and Turner 1994]. From 2000 till 2010 an area of approximately 33 thousand km^2 was exposed to deforestation in Vietnam as table C.1 suggests. We identified the central highland provinces Dak Lak, Dak Nong, Gia Lai, and Lam Dong as deforestation hot-spots. Meyfroidt et al. confirms this in his study on trajectories of deforestation and highlights that the losses in this area could reduce the benefits of national-scale forest recovery [Meyfroidt et al. 2013]. During the early 1990s a transition to reforestation was encouraged and natural forests expanded in Vietnam but it is connected with a increased loss in the highland regions [Meyfroidt et al. 2013; Chazdon 2008]. In Laos a forest area

of approximately 8 thousand km² is lost during the time period 2000 till 2010. The map in figure 2.5 suggests that northern Laos is predominantly exposed to tree cover loss while Hirsch confirms [Hirsch 2000]. Laos is known for a steady loss of forest cover since the early 1960s. Whereas the causes for this LC transitions are dynamically changing till now.

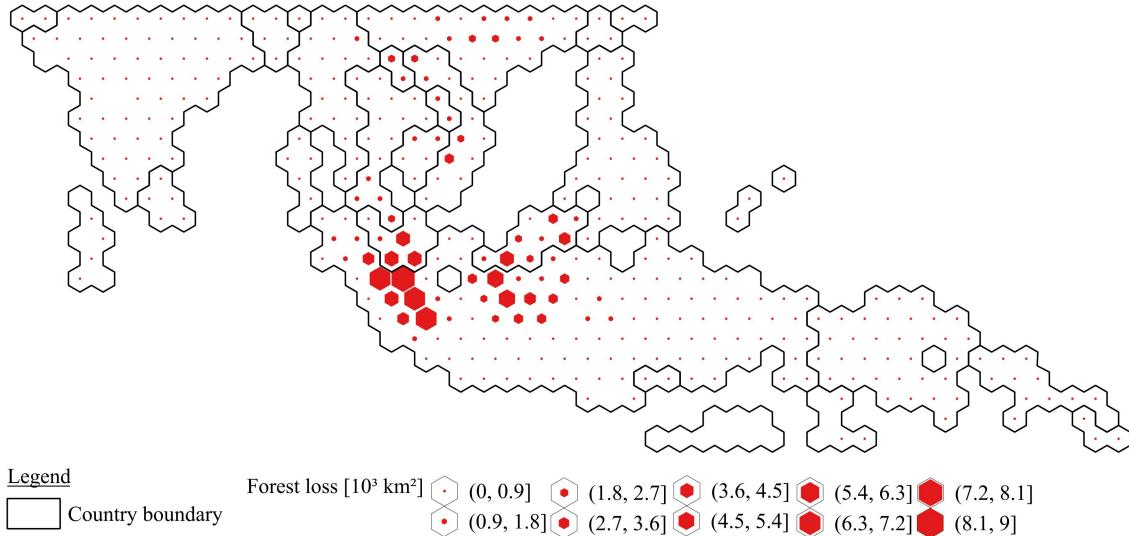


Figure 2.5. Tree cover loss in Asia/Australia between 2000 and 2010: This map shows the tree cover loss within our study extent between 2000 and 2010. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km² at the equator. Deforestation hotspots with tree cover loss about 1.9 thousand km² per hexagon are located in Indonesia, Malaysia, Vietnam, and Laos.

Figure 2.6 shows the tree cover and canopy density distribution over Africa. In Africa the tropical rain forest is distributed over the following central African countries: Democratic Republic of the Congo, Republic of the Congo, Equatorial Guinea, Gabon, Cameroon, and partial in Ghana, Ivory Coast, and Liberia. In our map this type of forest is characterized by a dense tree cover between 39 and 49 thousand km² per hexagon while the canopy density between 64 and 100 % does not separate it from the moist type as good as in Latin America and Asia/Australia. The rain forest is surrounded by tropical moist forest with a dense tree cover between 18 and 49 thousand km² while the mean canopy density between 10 and 82 % is sparser than tropical rain forest. Countries covered by the moist tropical forest type are the following examples: Angola, Uganda, Central African Republic, Zambia, Madagascar etc. Tropical dry forest is characterized by a sparse tree cover among approximately 18 thousand km² per hexagon and canopy densities between 10 and 46 %. Mozambique, Tanzania, and Nigeria are examples for countries covered partial by tropical dry forest.

The map in figure 2.7 shows the regions which are exposed to tree cover loss in Africa for the time period 2000 till 2010. During the first decade of 2000 an area of approximately 174 km² was deforested as table C.1 in the appendix C suggests. With tree cover loss between

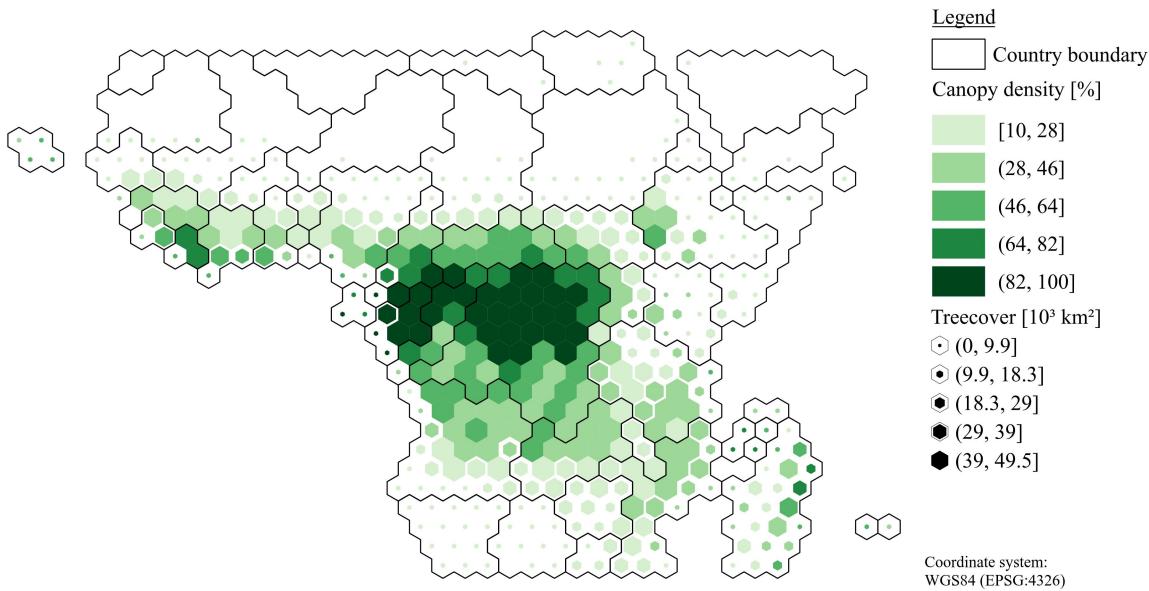


Figure 2.6. Tree cover and canopy density in Africa at 2000: This map shows the tree cover and mean canopy density distribution within our study extent at 2000. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km² at the equator. Tropical rain forest is characterized by dense tree cover between 39 and 49 thousand km² and high canopy density above 64 % in the center of Africa. The rain forest is surrounded by tropical moist forest with a dense tree cover as well but a lower mean canopy density between 10 and 82 %.

1.1 and 3.7 thousand km² per hexagon we identified the following countries as deforestation hot-spots: Ivory Coast, Democratic Republic of the Congo, Angola, Mozambique, Madagascar, and Tanzania. At the Ivory Coast an area of approximately 10 thousand km² was exposed to deforestation between 2000 and 2010 as table C.1 shows. The map in figure 2.7 shows that the deforestation hot-spots are located in the southern and northern parts of the country. Evidences for this are given by Goetze et al. and Barima et al. [Barima et al. 2016; Goetze et al. 2006]. Lower deforestation in the center of the country can be explained by a military conflict started in 2002 that divides the country in a northern and southern zone with a buffer zone in between controlled by Un forces and French soldiers Barima et al.. At the Ivory coast continuous loss of tree cover by deforestation started approximately in the 1958s [Chatelain et al. 1996]. The deforestation accounts for an area of approximately 45 thousand km² in the Democratic Republic of the Congo. The DR Congo home of the second largest tropical forest in the world has compared to other tropical countries relatively low deforestation dynamics but recent studies show that deforestation could accelerate in the future [Ickowitz et al. 2015]. The deforestation hot-spots are located in the eastern Congo and around medium-sized cities along the Congo river in the lower center of the country. During the first decade of the 2000s an area of about 13 thousand km² is deforested in Angola. Deforestation hot-spots are more oriented to the center of the country covering roughly the provinces Huambo, Bie, and Moxico as the map in figure 2.7 suggests. For the time period 1990 till 2009 Cabral et al. performed a study on forest change in the province of Huambo where a decrease in dense forest cover is observed while the cover of sparse forest is in-

creasing [Cabral et al. 2011]. Another study by Schneibel et al. on deforestation dynamics in south-central Angola reveals that the tree cover loss develops along anthropogenic infrastructure and forests are exploited over a long term until (fuel-wood collection etc.) a LC transitions to other types like cropland occurs [Schneibel et al. 2017]. To best of our knowledge no studies on historic deforestation dynamics exist for Angola due to the civil war from 1975 till 2002. In Mozambique an area of approximately 18 thousand km² was exposed to tree cover loss as table C.1 suggests. During the war between 1976 and 1992 the forests of Mozambique were not largely exposed to deforestation but since end of the war the deforestation rates are increasing [Sitoé et al. 2012]. We identified the following provinces as deforestation hot-spots between 2000 and 2010: Zambezia, Nampula, and Cabo Delgado. This is largely confirmed by Sitoé et al. where it is stated that deforestation is concentrated in the center and north of Mozambique. These hot-spots could be related to the higher population densities within these areas. Tree cover loss accounts for an area of about 11 thousand km² in Madagascar during the study period. In Madagascar deforestation hot-spots are concentrated at the north-east coast of the island as the map in figure 2.5 suggests. Madagascar's central highland forests are exposed to anthropogenic deforestation since 1600, reportedly [Harper et al. 2007]. By the nineteenth century the deforestation advanced over the entire island forests. In Tanzania an area of approximately 14 thousand km² was exposed to tree cover loss. Since the 1900s approximately 19.4 % of the forest cover is lost [Kideghesho 2015]. The map in figure 2.7 suggests that deforestation hot-spots are located in the country center covering the following provinces: Lindi, Singida, Tabora, Dodoma, Tanga, Pwani, and the island of Zanzibar.

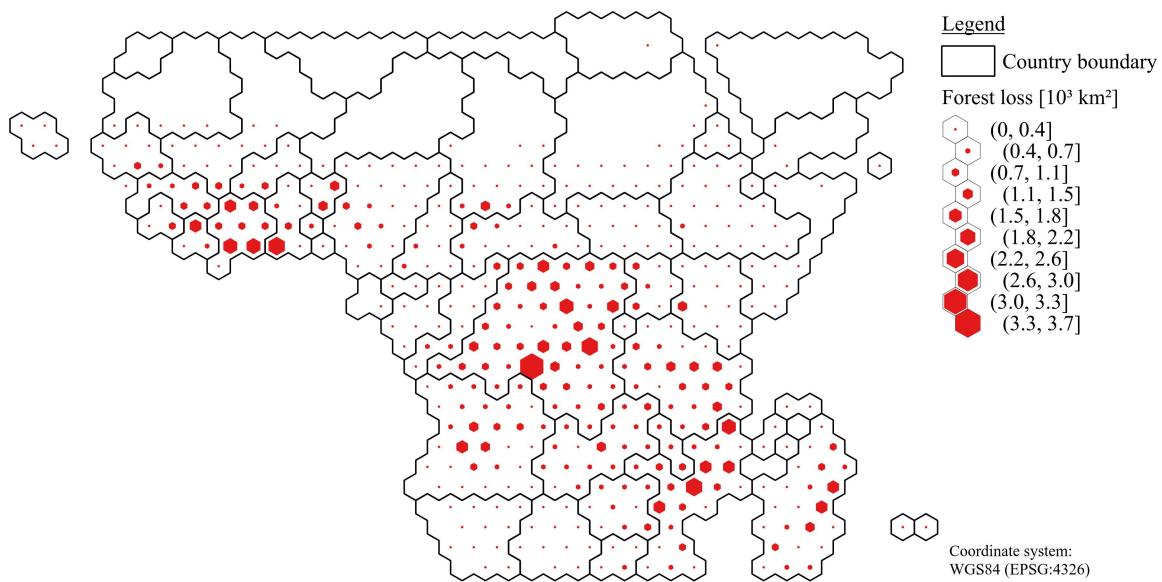


Figure 2.7. Tree cover loss in Africa between 2000 and 2010: This map shows the tree cover loss within our study extent between 2000 and 2010. An unscaled hexagon covers an area of 0.5 decimal degrees which translates to approximately 49 thousand km² at the equator. Deforestation hot-spots with tree cover loss about 1.1 thousand km² per hexagon are located in Ivory Coast, Democratic Republic of the Congo, Angola, Mozambique, Madagascar, and Tanzania.

2.1.3. Proximate deforestation drivers

Our goal is to estimate the distribution of PDDs over the tropical zone, the continental range, and at a country scale for the time frame of 2000 till 2010. We achieved this estimate by superimposing the annual tree cover losses and aggregated gains of the GFC datasets and the GL30 land cover map from 2000. We carefully selected our global definition of tree cover in the canopy density interval (10, 100] by applying the Jaccard Index and statistical testing detailed in section 2.1.1. By using this canopy density interval we filtered the GFC annual losses and we considered tree cover gains only within previously lost tree cover. After superimposing we applied a reclassification of tree cover losses still classified as forest by the GL30 layer. We aggregated this structures by clustering and applied a square sized buffer of 500 meter side length. Next, we reclassified the structure by determining the highest frequent LC class within the buffer. For further details on the PDD prediction refer to section 1.2.2.2. The PDD distribution choropleth-maps for Latin America, Asia/Australia, and Africa we derived by applying our hexagonal-binning approach in combination with a hexagon-pie-chart. A hexagon in unscaled shape covers an area of 0.5 decimal degrees. Section 1.2.5 describes detailed how we derived these cartograms. In the following paragraphs we will describe our results for Latin America, Asia/Australia, and Africa on continental and regional scale and close this section by discussing our results on global scale.

In Latin America the transition of tree cover to cropland and pastures account for 21.3% (95929.6 km^2) and 40.8% (183841.4 km^2) as the table C.1 in the appendix C shows. Therefore, approximately 62% (279771 km^2) of the tree cover were cleared for agricultural purposes in Latin America. Around 12.6% (56909.3 km^2) of the tree cover loss is followed by tree cover regrowth, while the transition to shrubland account for 11.1% (50260.2 km^2) of the tree cover loss, respectively. Minor PDDs are water, artificial surfaces, and bareland which account for 1.6% (7169.9 km^2), 0.3% (1561.5 km^2), and 0.1% (405.4 km^2) of the forest transitions. Sy et al. estimates that agriculture accounts for 88.5% of the tree cover loss, while transitions of forest cover to other LC accounts for 11.5%. Further, Hosonuma et al. achieve similar results for the conversion of tree cover to agricultural land where it accounts for approximately 90% of the tree cover loss. Both studies estimate that the expansion of artificial surfaces account for approximately 1% of the forest loss. The large difference could be explained by the definition of the forest transition classes in both studies. In both studies agriculture comprises pastures, cropland, and tree plantations. Additionally, both studies refer to the change of Land Use (LU) and we determine the PDDs by the change of LC. Further, Sy et al. uses a sample based approach on $10 \times 10 \text{ km}$ FAO FRA-2010 RSS data which could yield overestimates, while Hosonuma et al. uses an empirical approach based on FAO data as well. A recent study on PDDs estimates that commodity-driven deforestation, shifting agriculture, forestry, wildfire, and urbanization account for 56%, 31%, 13%, 1%, and

<1% of the tree cover loss, respectively [Curtis et al. 2018]. Commodity-driven deforestation refers to tree cover loss as long-term permanent transition of forest to a non-forest LU like agriculture which includes cropland, pastures, plantations and so forth. Comparable to the previously mentioned research on PDDs the difference between our estimates and Curtis et al. arise from class definitions. If we would aggregate our LC classes to their schema it would yield nearly the same results. In particular the same estimate for tree cover loss by urbanization shows that the similar classes without aggregation yield similar results. In Latin America the LC change to cropland is mainly distributed over southern part, while the transition to pastures is concentrated in the central part of the continent as figure 2.8 suggests. Cropland expansion mainly took place in the south of Brazil, Paraguay, Argentina, and Bolivia, while forest loss by pasture expansion is concentrated in the center and north of Brazil. The spatial distribution of cropland/pasture dynamics largely corresponds to the findings of Graesser et al.. Large quantities of tree cover regrowth can be observed in the south-east and center of Latin America namely the southeast coast of Brazil and within the tropical rain forest covering Brazil, Peru, and Colombia. In south-east Brazil the findings of Curtis et al. suggests that the regrowth dynamics are driven by forestry actions.

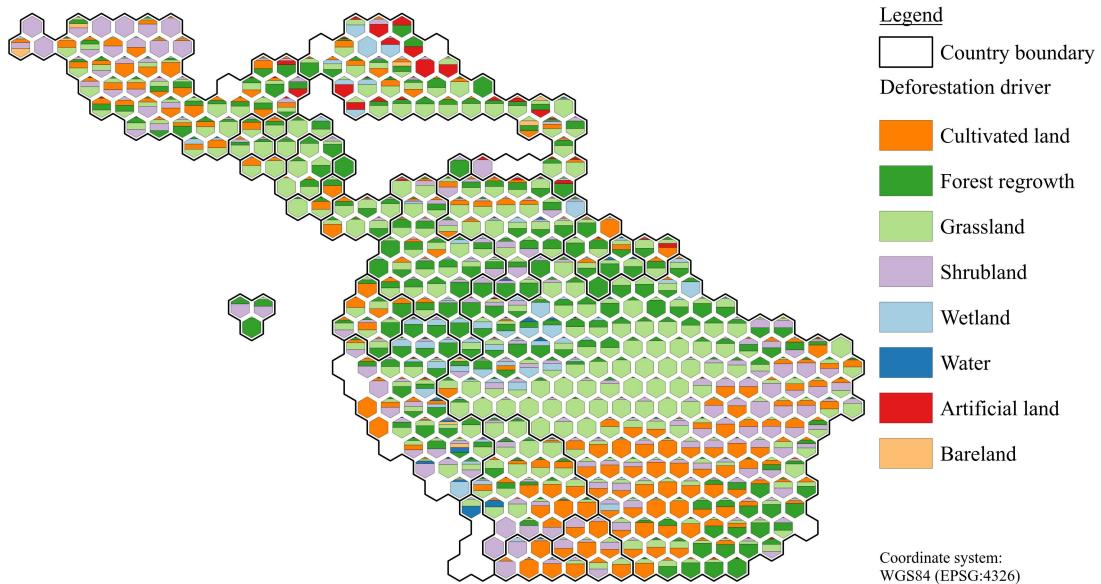


Figure 2.8. Map of proximate deforestation drivers in Latin America: The map shows the distribution of proximate deforestation drivers in Latin America. The different sized and colored quantities within each hexagon's interior shows the relative tree cover loss by a proximate deforestation driver. Scaling of the hexagons is only intended for improving visual appeal.

For Asia/Australia the table C.1 in the appendix C shows cultivated and grassland account for 16% (36819.3 km^2) and 7.1% (16302.6 km^2) of the tree cover loss, respectively. Therefore, the transition of LC to agriculture usage accounts for 23.1% (53121.9 km^2) of the tree cover losses. The transition of tree cover to artificial surfaces and the forest loss by inundation by lakes and rivers account for 1.1% (2431.9 km^2) and 0.4% (890 km^2), respectively. In Asia/Australia the largest PDD is the regrowth which accounts for 61.2% (140653.4 km^2)

of the cumulative tree cover loss. Forest transitions to shrubland account for 6.7 % (), respectively. Hosonuma et al. estimates that LC transitions by agriculture account for approximately 70% of the forest loss, while Curtis et al. estimates that commodity-driven deforestation, shifting agriculture, forestry, wildfire, and urbanization account for 13%, 78%, 9%, 13%, <1%, and <1% of the tree cover loss in Southeast Asia, respectively. As mentioned in the previous paragraph the differences in PDDs estimates relate mainly to the applied methodology and the aggregation of LC classes. The figure 2.9 shows an overview of the PDDs distribution in Asia/Australia and indicates that regrowth dynamics are largely concentrated at the east of Asia/Australia namely in the following countries: Indonesia, Malaysia, Philippines, and Papua New Guinea. Conversion of tree cover to cropland can be observed in Vietnam, Cambodia, the north of Thailand, and India in its entire extent. [Reference Curtis for spatial patterns](#)

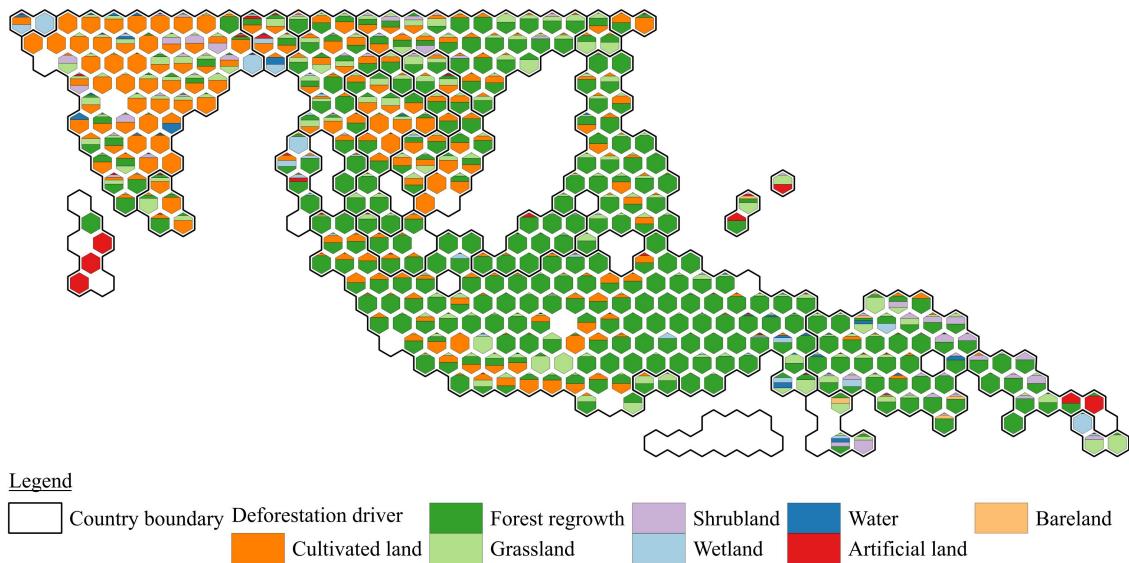


Figure 2.9. Map of proximate deforestation drivers in Asia/Australia: The map shows the distribution of proximate deforestation drivers in Asia/Australia. The different sized and colored quantities within each hexagons interior shows the relative tree cover loss by a proximate deforestation driver. Scaling of a the hexagons is only intended for improving visual appeal.

In Africa the table C.1 in the appendix C suggests that cropland and grassland account for 22.8% (44289.5 km^2) and 46% (89301.4 km^2) of the tree cover loss, respectively. Therefore, transitions to agricultural land are the major PDD in Africa and account for 68.8% (133590.9 km^2) of the tree cover loss. Forest transitions to water, artificial surfaces, and bareland account for 1.2% (2409.9 km^2), 0.6% (1238.6 km^2), and 0.1% (146 km^2) of the tree cover loss, respectively. In Africa on 17% (32980.8 km^2) of the area exposed to tree cover loss regrowth could be detected, while the transition to shrubland account for 3.4% (6599.5 km^2) of the tree cover loss, respectively. The study on PDDs by Hosonuma et al. estimates that agriculture and urbanization account for approximately 75% and 2% of the tree cover loss, respectively. For Africa Curtis et al. shows that commodity-driven deforestation, shifting

agriculture, forestry, wildfire, and urbanization account for 4%, 92%, 4%, <1%, <1% of the tree cover loss. The figure 2.10 shows the distribution of PDDs in Africa.

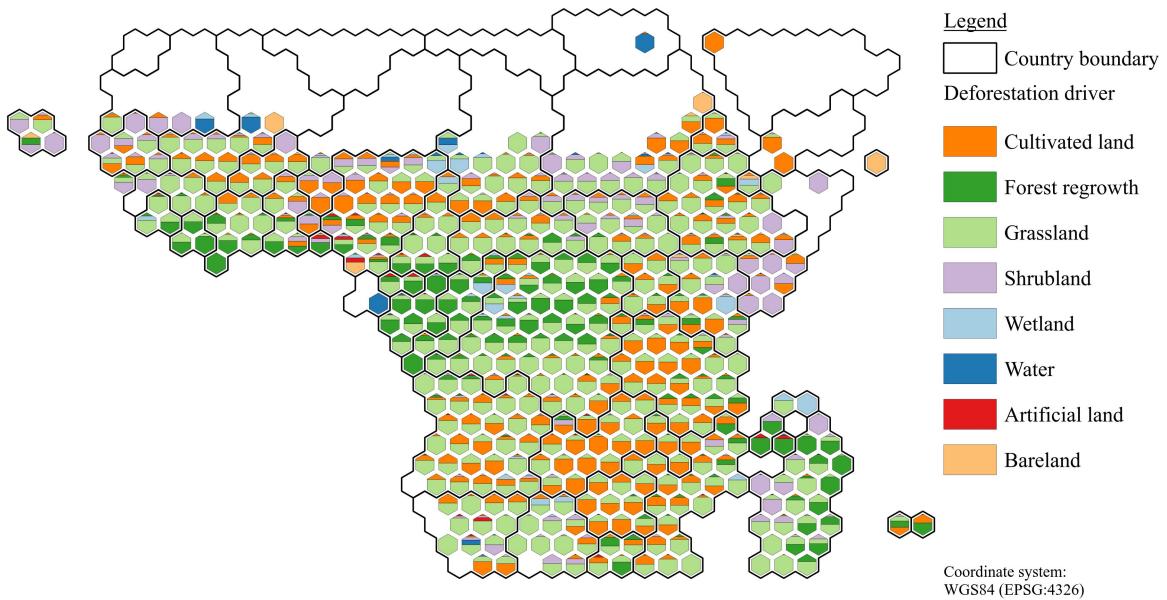


Figure 2.10. Map of proximate deforestation drivers in Africa: The map shows the distribution of proximate deforestation drivers in Africa. The different sized and colored quantities within each hexagons interior shows the relative tree cover loss by a proximate deforestation driver. Scaling of a the hexagons is only intended for improving visual appeal.

In Latina America and Africa the major PDD is grassland, while in Asia/Australia it is regrowth.

On the global range the major PDDs are cropland and pastures which account for 20.2% (177038.5 km^2) and 33.1% (289445.5 km^2) as the table C.1 in the appendix C suggests.

2.1.4. Accuracy assessment

Goal (review): Goal is the assessment of the accuracy of our proximate deforestation driver predictions. We created a set of ground truth data by sampling our proximate deforestation driver layers. In each continental region we select per random 10 tiles and draw 200 samples per tile. The 200 samples comprises pixels over the full value range of our proximate deforestation driver classes. We imported the prepared sample to our JavaScript application and subsequently classified each sample with a label by visual interpretation of google earth imagery. To determine the accuracy we used a confusion matrix and the derived metrics.

Results (review): Table 2.5 shows the confusion matrix to determine the accuracy of our predictions where the term reference refers to the labeling of pixel by our visual interpretation and predictions refer to the labeling of our proximate driver predictions. The abbreviations PAc, UAc, OvAc, Com, Om, Tot, and Kappa refer to the terms Producers-Accuracy, Users-Accuracy, Overall-Accuracy, Error of Commission, Error of Omission, row or column total, and Kappa Coefficient. From the 6000 samples we draw from our study extent 14 %, 20 %, 22 %, 32 %, 8 %, 2 %, 0.5 %, 2 %, and 0.5 % account for cultivated land (10), tree cover (20), regrowth (25), shrubland (40), wetland (50), water (60), artificial land (80), and bareland (90), respectively. Our method predicts a distribution of 15 %, 18 %, 27 %, 31 %, 7 %, 1 %, 0.8 %, 1 %, and 0.2 % for the land cover classes 10, 20, 25, 30, 40, 50, 60, 80, and 90, respectively. Highest producers accuracy is achieved the regrowth class with 88 percent. The prediction of this class was achieved by including global forest change gain data within our target canopy density. During building our reference data set per visual interpretation we determined this class by following these rules. We checked the surroundings and the corresponding pixel for signs of road networks and infrastructure. Further we checked if the canopy shows signs of age class forest and line patterns which show the establishment of artificial introduced forest cover. During the visual interpretation we recognized that a large portion of the regrowth class is occupied by plantations especially in Americas and Asia. For Asia a major share was occupied by palm oil or other plantations. It must be mentioned that here the deforestation within our temporal frame is not coercively the clearing of natural cover/primary forest it shows it shows also rotational cycles of deforestation reforestation. The second highest accuracy with 85 (producers accuracy) was achieved in the prediction of cultivated land where only 15 percent could be identified as error of commission. 8.2 percent are classified as forest or regrowth which could reveal zones of shifting agriculture or temporal issues. To identify this class during our visual interpretation we applied nearly the same rules as the detection of regrowth checking for infrastructure but here no canopy should be identified. Further we checked for large scale square patterns which is shown by large scale commodity driven agriculture. Especially the small scale agriculture was hard detect because these class of agriculture often show no regular patterns of forest clearing

and are often accompanied by selective logging. Further it should be mentioned common practice of small scale agriculture is inhomogeneous over the regions Americas, Asia and Africa. Another good indicator for cultivated land is the appearance of tillage patterns on high resolution imagery. Cultivated land comprises a large variety of different use cases. Grassland shows an accuracy of 77 at an error of commission of 23 percent. The error could lead back to classification errors by GlobeLand30 or land cover change dynamics and the temporal time frame. For visual interpretation cultivated land and grassland share some common properties like square shaped clearings and infrastructure in the surroundings. But grassland is also defined by the occurrence of shrubs and single trees within its extent. To really determine the differences between both classes high resolution imagery is required. In Americas the grassland patches could often be identified as used for cattle ranching. During visual interpretation we recognized frequently the occurrence of water holes. Africa showed the occurrence of natural grasslands and Asia bla.

Table 2.5. Confusion matrix for accuracy assessment: We draw 6000 samples from 10 random selected tiles from the three regions Americas, Asia and Africa. Labels refer to our proximate deforestation driver classes which correspond to GlobeLand30 classification schema in table 1.2. Reference refers to the samples we classified by visual interpretation of external imagery and predictions refer to the label the sample has in our proximate driver product. The abbreviations PAc, UAc, OvAc, Com, Om, Tot, and Kappa refer to the terms Producers-Accuracy, Users-Accuracy, Overall-Accuracy, Error of Commission, Error of Omission, row or column total, and Kappa Coefficient.

		Reference											
	Cls	10	20	25	30	40	50	60	80	90	Tot	UAc	Om
Prediction	10	730	37	62	15	16	2	3	5	0	870	.84	.16
	20	41	744	56	189	31	12	0	15	4	1092	.68	.32
	25	29	202	1155	172	22	10	5	11	4	1610	.72	.28
	30	36	187	32	1466	73	21	0	17	0	1832	.80	.20
	40	14	21	4	41	352	1	1	2	1	437	.81	.19
	50	0	5	3	10	4	50	0	1	0	73	.68	.32
	60	2	1	0	3	0	2	18	2	0	28	.64	.36
	80	3	3	0	1	1	1	0	40	0	49	.82	.18
	90	0	0	0	1	0	0	0	3	5	9	.56	.44
	Tot	855	1200	1312	1898	499	99	27	96	14	6000		
	PAc	.85	.62	.88	.77	.71	.51	.67	.42	.36	Kappa		OvAc
	Com	.15	.38	.12	.23	.29	.49	.33	.58	.64	.69		.76

2.2. Emissions

Goal (review): Our goal is to estimate the emissions emitted by proximate deforestation driver. We estimated the emissions from the removal of aboveground biomass and through the change of soil organic carbon content. Detailed below are the pdd classes we assumed as deforestation to estimate the agb emissions: 10, 25, 30, 40, 80, 90. To approximate

the emissions by soil organic carbon change we used the transition pathways from don et al. We estimated soc for two different scenarios with the following conditions: scenario one (SC1) all transitions occur in primary forest and scenario two (SC2) by using the intact forest landscape layer we distinguished between primary forest and secondary forest. The first scenario can be seen as liberal estimate and the second scenario as conservative estimate. We aggregated our results on global and continental scale.

2.3. Ecosystem service values

Our goal is to estimate the monetary loss of ESV by the aggregated PDD at a global and continental scale between 2001 and 2010. Whereas a study targeting regional scale aggregation could easily performed with our data. Additionally, we approximated the monetary gain if the former tree cover is converted to a certain type of LC. We refer to this gain as ESV gain. Further, we compute the balance between ESV loss and gain to estimate the net change of ESV in the tropical zone. We applied three datasets which estimate the monetary value of ecosystems on a global scale namely the following data: de Groot et al. [2012], Costanza et al. [2014], and Siikamaki et al. [2015]. We refer to this datasets as Co, Dg, and Wb. To compute the monetary loss of ESV we selected the following PDD classes as anthropogenic deforestation: cultivated land (10), regrowth (25), grassland (30), shrubland (40), artificial surfaces (80), and bareland (90). We excluded pixel classified as forest (20) by our PDD prediction because within this class we are uncertain if a deforestation event occurred. Further, we excluded transitions to wetland or water bodies because we assume this LC changes are largely driven by natural causes. The cumulative ESV gain depends on the particular ESV dataset because the datasets don't define a monetary value for each PDD class. For detailed informations on the methodology refer to section 1.2.4. The ESV net change or balance is the difference of cumulative ESV loss and gain. The monetary unit for each value is the Geary–Khamis dollar at 2007 per year also known as international dollar ($2007 \text{ Int.} \$ \text{ y}^{-1}$). The table 2.6 shows the gross ESV loss and gain and the net balance for continental and global scale.

Americas (review): In the Americas a cumulative loss of 396 thousand km² tropical forest between 2000 and 2010 accounts for approximately 208.3 (Co), 204 (Dg), and 51 (Wb) billion international dollar per year in gross esv loss for co, dg, and wb, respectively. For the losses we will refer only to co and wb because the differences between co and dg are not large due to the nearly similar monetary value for tropical forest. Difference between them is 118 international dollar. By using constanzas estimates the transition of tropical tree cover to cropland accounts for 51 billion international dollar per year while transitions to pastures (grassland and shrubland aggregated) accounts for 126 billion international dollar. Tropical

deforestation by the expansion of artificial and bareland surfaces account for 1 billion dollar while the transition to plantations account for 30 billion dollar. By using the siikamaki estimates the transition of tropical tree cover to cropland accounts for 12.6 million dollar while the transition to pasture land accounts for 30.6 dollar. Tropical deforestation by the expansion of artificial and bareland surfaces accounts only for 0.3 billion dollar while the transition to plantations account for 7.5 billion dollar. The great difference between the estimates of both dataset can be explained by the fact that costnazas value for tropical forest is approximately 4 times greater than the estimate from siikamaki for global aggregates. For the gains we will analyse the results of all three datasets because as table methods esv suggests each dataset defines different values for pdd. For costanza the total gain from pdd accounts for 161 billion dollar per year. The gain from transitions to cultivated land accounts for 53.4 billion dollar while the transitions to grassland accounts for 76.6 billion dollar. Monetary gains from the transition to artificial landscapes account for 1 billion dollar. The transition to certain regrowth accounts for 30.6 billion dollar. The gain from degroot accounts for 82.5 billion dollar per year while the gain from transitions to grassland account for 52.8 billion dollar. The last esv gain which could be calculated for degroot is the esv gain from regrowth which accounts for approximately 29.9 billion dollar. For wb we could only compute the gain from regrowth which accounts for 7.3 billion dollar. In regards of esv net balance and by applying the esv values from costanza the net loss of esv is -47.3 billion dollar while degroot estimates a net loss of -121.5 dollar. By applying siikamaki the net balance is -43.7 dollar. The variations of net balance are explained over the different monetary values for different pdd classes and the completeness of the datasets.

Table 2.6. Ecosystem service value balance: The dataset column refers to the three global ecosystem service value datasets by de Groot et al. [2012] (Co), Costanza et al. [2014] (Dg), and Siikamaki et al. [2015] (Wb). Loss refers to the monetary value of tropical forest deforested by the aggregated anthropogenic proximate deforestation drivers. Gain is the monetary value of the land cover transition and balance is difference of gain and loss. Monetary unit of the ecosystem service values is 10^9 2007 Int'l\$ y^{-1} (billion international dollar 2007 per year)

Dataset	Global	Latin America	Asia/Australia	Africa
Co _{loss}	414.1	208.3	111.6	94.2
Co _{gain}	350.6	161.0	109.1	80.6
Co _{balance}	-63.5	-47.3	-2.5	-13.6
Dg _{loss}	405.2	204.0	109.1	92.1
Dg _{gain}	208.7	82.5	83.1	43.1
Dg _{balance}	-196.5	-121.5	-26.0	-49.0
Wb _{loss}	101.1	51.0	27.2	22.9
Wb _{gain}	31.0	7.3	19.4	4.3
Wb _{balance}	-70.1	-43.7	-7.8	-18.6

Asia/Australia (review): Within our time frame in Asia/Australia a cumulative loss of 230 thousand km² tropical forest cover accounts for approximately 111.6, 109.1, and 27.2 billion international dollar per year. By applying constanzas monetary value for tropical forest

the transition of tropical tree cover to cultivated land accounts for 19.8 billion dollar while the transition to pasture accounts 9.8 billion dollar. The reduction of tree cover due to the expansion of artificial and bareland surfaces accounts for 0.5 billion dollar while the highest monetary loss is detectable in the conversion or rotational cycle of plantations with 75.7 billion dollar. By applying siikamakis estimate for tropical forest the transition of forest cover to cultivated land accounts for 4.8 billion dollar while transitions to pastureland account for 2.4 billion dollar. Transition of tree cover to artificial surfaces account for 0.2 dollar while the reset of forest cover with plantations account for 18.4 dollar. If we consider the esv gain by **lc!** (**lc!**) transitions by costanca cropland account for 20.5 billion gain while pastures account for 6.8 billion dollar gain. The highest gain could be achieved by the transition to plantations with a monetary value of 75.7 billion dollar. Transitions to artificial surfaces add a gain of 0.5 billion dollar. By considering degroots estimates grassland accounts for 4.6 billion dollar gain and regrowth for 74 billion dollar. The esv for tropical forest estimates a gain by plantations of 19.4 billion dollar. In regards of the net balance in asia co, dg, and wb account for -2.5, -26.5, -7.8 billion dollar respectively. The small net losses in asia are explained by the high share of regrowth transitions which use the same esv value as tropical forest.

Africa (review): In Africa a tree cover of approximately 177 km² is lost within our study time frame. This converts in esv loss of 94.2, 92.1, and 22.9 billion dollar for co, dg, and wb, respectively. By using costanca esv estimates the transitions to cropland account for losses of 23.7 billion dollar while transitions to pastures account for 51.5 billion dollar. The transitions to artificial surfaces accounts for 0.7 billion dollar while the transition to plantations account for 17.7 billion dollar. By applying the estimates of siikamaki the transition of tree cover to cropland account for 5.8 billion dollar while the transition to grassland account for 12.7 billion dollar. Transitions to artificial surfaces account for 0.2 while transitions to plantations 4.3 billion dollar. If we consider the evs gain by forest cover transitions cropland accounts for 24.6 billion dollar while grassland accounts for 37.2 billion dollar with constanca values. Esv gain by artificial and plantations account for 0.8 an 17.7 billion dollar respectively. If we consider degroots esv estimates esv gain by grassland and regrowth account for 25.6 and 17.4 billion dollar while siikamaki estimates a gain of 4.3 billion dollar for plantations. In regards of net balance in africa co, dg, and wb account for -13.6, -49, and -18.6 billion dollar, respectively.

Compare Countries (review): If we compare the regions south america has the greatest esv loss over all three esv datasets out of the three investigated regions. This is related to the high amount of forest cover lost by deforestation within our study time frame. Africa has the lowest gross loss of esv followed by asia at second. If we consider the net balance of esv asia shows the best performance to mitigate the esv losses. This can be explained by the great share of transition to regrowth which uses the same esv value as tropical forest. It is

to assume that the balance for all regions could be smaller if we had esv values for all pdd classes. Therefore, it can be imagined that regions could mitigate esv losses within tropical forest cover by applying transition optimisation to get the net balance on the positive side. differences in loss estimates between the datasets, gain estimates relay on the completeness of datasets, siikamaki estimate is realy conservative on global level it is to assume it underestimates the esv value of tropical forest.

Global (review): At the global tropical scale between 2000 and 2010 are approximately 772 thousand km² of tree cover are lost by pdd. This translates to an esv loss of 414.1, 405.2, and 101.1 billion dollar for co, dg, and wb, respectively. By applying the coefficient for tropical forest from costanca the esv loss by transitions to cropland account for 95.3 billion dollar while transitions to pastures account for 128 billion dollar. The transitions to artificial surfaces and plantations account for 2.3 and 124 billion dollar. If we consider the tropical forest valuie from siikamaki cropland and pastures account for 23.2 and 46.7 billion dollar respectively. While artificial and plantation transitions account for 0.5 and 30.2 billion dollar. By applying constancas monetary values to estimate the gain from land cover transitions cropland and grassland account for 98.5 and 120.6 billion dollar while artificial surfaces and regrowth account for 2.4 an 124.1 billion dollar, respectively. If we consider degroots values for esv gain grassland and regrowth account for 83 and 121 billion dollar. Siikamki estimates a gain of 31 billion dollar for the transition to plantatitions. The net balance is greatest with -196 billion dollar for the estimates from degroot because it defines only values for two pdd classes. The smallest net loss can be observed for costanca with -63.5 billion dollar follwed by siikamaki with -70.1 billion dollar.

3. Discussion

3.1. Deforestation

3.1.1. Forest definition

- For a regional approach a better solution could be to select for each region independently the right canopy density. For America a good agreement between the tree cover could be achieved by selecting the second class. For Asia by selecting the first class and for Africa the second. Even better would be to decide per tile individually which canopy density should be selected. This would eliminate regional effects of different forest densities.
- discuss regions independently Asia and America have large tree cover agreement
- Africa has the lowest agreement only core forest zones show high similarity
- We can see at which regions Chen et al switched their tree cover definition to 10
- To improve we should apply for each tile a canopy class decision based on our analysis
- This could improve the similarity (accuracy) by maximizing the sample count
- Algorithm draft for single similarity: Compute Jaccard indexes for tile pair at different canopy densities, put results in a list, sort the list in decreasing order, pick the class where Jaccard index is max
- Use this Jaccard method to exclude tiles where the tree cover similarity fall below a certain threshold

3.1.2. Tree cover and deforestation

- Improved tree cover map divide tree cover within hexagon by landmass within hexagon
- Improved loss map divide loss by the tree cover within a hexagon

3.1.3. Proximate deforestation driver

- Reclassification is not a good idea cause this approach leads not to consistent results
- limitation the conversion of tree cover to shrubland or grassland in Africa can be mapping error

3.1.4. Accuracy assessment

- Is largely subjective because it is prepared from the study author
- Better if someone independent does it
- Even better if you have ground truth prepared by field studies
- Class variability errors source the reclassification
- Time frame of our classification we classified our ground truth data at google image data
- Land cover change dynamics stay hidden
- Sample size is not scientific chosen
- Confusion matrix neglect uncertainty, apply approach in olofsson et al should improve the reasoning
- mention that kappa coefficient should be neglected cause issues reference olofsson

3.2. Emissions

3.3. Ecosystem service values

- resilience of esv loss could be achieved over optimizing total value of the new land-use
- target optimization is use the clearcut by maximizing profit and minimizing the esv loss
- large differences between the datasets

3.4. Binning analysis and visualization

- Cut polygon by line the Scala (1992) approach explained with parametric separation function, and bezier
- A approach where ratio is also ratio of the hexagon area

4. Conclusion

Classification schema:

Grassland clearing of tree cover for grassland indicates that it is used for pastures **there is somewhere a citation for it**. Water and wetland class can be assumed that it is not urgent a anthropogenic driven tree cover loss. It could be forest loss by inundation by lakes and rivers.Sy et al. [2015] Shrubland is a ambivalent class, could be abandoned pastures, newly established tree plantations or cropland. Regrowth is introduced by the GFC gain layer, this are areas where tree cover regrowth is detectable. This could be tree plantations or natural regrowth. Here secondary literature is needed to distinguish between these terms. It depends largely on regions and what regrowth is. Areas still classified as forest are ambivalent it could be a false positive by GFC loss layer (predicts loss but there is no loss), a false negative by gain layer (predicts no gain but there is gain), a tree cover definition discrepancy between GFC and GL30 respectively remote sensing method differences.

Therefore we will aggregate the classes by the following conventions for the discussion of the results. To cropland and grassland we will refer as agricultural land. Regrowth and forest are forestry and plantation activities. Shrubland will be discussed extra. Artificial is artificial and bareland, wetland, and

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List of Figures

1.1.	Map of downloaded dataset tiles	13
1.2.	Tile alignment algorithm	21
1.3.	Aligned raster images and sampling locations	21
1.4.	Classification of proximate deforestation drivers	25
1.5.	Hexagon tessellation	34
2.1.	Tree cover similarity distribution	37
2.2.	Tree cover and canopy density in Latin America at 2000	43
2.3.	Tree cover loss in Latin America between 2000 and 2010	44
2.4.	Tree cover and canopy density in Asia/Australia at 2000	45
2.5.	Tree cover loss in Asia/Australia between 2000 and 2010	46
2.6.	Tree cover and canopy density in Africa at 2000	47
2.7.	Tree cover loss in Africa between 2000 and 2010	48
2.8.	Map of proximate deforestation drivers in Latin America	50
2.9.	Map of proximate deforestation drivers in Asia/Australia	51
2.10.	Map of proximate deforestation drivers in Africa	52
B.1.	Caption	VII
B.2.	Caption	VIII
B.3.	Caption	IX

List of Tables

1.1. Datasets used in this study	11
1.2. Classification schema of the GlobeLand30 product	15
1.3. Relative soil organic carbon change for certain land-use change types	18
1.4. Selection of Ecosystem Service Values (ESV) used in this study	19
1.5. Jaccard Index coefficient matrix	23
1.6. A general model of a confusion matrix	26
1.7. Scenario one mapping of soil organic carbon change to porixmate driver	29
1.8. Scenario two mapping of soil organic carbon change to porixmate driver	30
1.9. ESV biome types mapped to PDD classes	31
2.1. Continental experiment group comparison	38
2.2. Continental experiment group directional comparison	39
2.3. Global experiment group comparison	41
2.4. Global experiment group directional comparison	41
2.5. Confussion matrix	54
2.6. Ecosystem service value balance	56
B.1. Comparison of tree cover agreement between regions	VI
B.2. Comparison of tree cover agreement between regions	VI
C.1. Estimates of proximate drivers of deforestation for countries, continents, and tropical zone	XI

List of Abbreviations

AGB	Aboveground live woddy Biomass density
AISM	Aligned Image Stack Mosaic
API	Application Programming Interface
CRS	Coordinate Reference System
CSV	Comma Separated Values
DEM	Digital Elevation Map
DT	Decision Tree
ESV	Ecosystem Service Value
ETM+	Enhanced Thematic Mapper Plus
FAO	Food and Agriculture Organization of the United Nations
FTP	File Transfer Protocol
GADM	Global Administrative Areas Map
GFC	Global Forest Change
GFW	Global Forest Watch
GIS	Geographic Information System
GL30	GlobeLand30
GLAS	Geoscience Laser Altimeter System
GSOCmap	Global Soil Organic Carbon map
GSP	Global Soil Partnership
GTiff	Geo-Tiff
GeoJSON	Geographic JavaScript Object Notation
IFL	Intact Forest Landscapes
IPCC	Intergovernmental Panel on Climate Change
ISRIC	International Soil Reference and Information Center
ITPS	Intergovernmental Technical Panel on Soils
LC	Land Cover
LCC	Land Cover Change
LIDAR	Light Detection and Ranging
LU	Land Use
LUC	Land-use Change
LULC	Land Use/Land Cover
ME	Mean Error
MLC	Maximum Likelihood Classifier
MLR	Multiple Linear Regression
MODIS	Moderate Resolution Imaging Spectroradiometer
NDVI	Normalized Difference Vegetation Index
POK	Pixel-Object-Knowledge

REGEX	Regular Expression
RF	Random Forest
RMSE	Root Mean Square Error
SD	Standard Deviation
SHP	Shapefile
SOC	Soil Organic Carbon
SOCC	Soil Organic Carbon Content
SVM	Support Vector Machine
UN	United Nations
URL	Uniform Resource Locator
UTM	Universal Transverse Mercator
WGS84	World Geodetic System 1984
stdlib	Standard Library
GDAL	Geospatial Data Abstraction Library
JI	Jaccard Index
PDD	Proximate Deforestation Driver
WSRT	Wilcoxon signed-rank test
WRST	Wilcoxon rank-sum test

A. Hexagonal country boundaries

B. Forest definition

Table B.1. Comparison of tree cover agreement between regions: This table shows, a comparison of tree cover agreement between regions. The classes Jl_0 , Jl_1 , Jl_2 , and Jl_3 as row and column headings account for the canopy density classes (0,100], (10,100], (20,100], and (30,100], respectively. The test hypothesis is $H_0: X_1 = X_2$ where X_1 is the column Jl_n class and X_2 the row Jl_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.02$, and $p^{***} < 0.01$. We applied a Benjamini and Hochberg correction for multiple-pairwise testing.

		Latin America				Asia/Australia			
	Cls	Jl_0	Jl_1	Jl_2	Jl_3	Jl_0	Jl_1	Jl_2	Jl_3
Asia	Jl_0	.04*	-	-	-	-	-	-	-
	Jl_1	-	.04*	-	-	-	-	-	-
	Jl_2	-	-	.05*	-	-	-	-	-
	Jl_3	-	-	-	.07	-	-	-	-
Africa	Jl_0	.00***	-	-	-	.00***	-	-	-
	Jl_1	-	.00***	-	-	-	.00***	-	-
	Jl_2	-	-	.00***	-	-	-	.00***	-
	Jl_3	-	-	-	.00***	-	-	-	.00***

Table B.2. Comparison of tree cover agreement between regions: This table shows, a comparison of tree cover agreement between regions and the direction of differences. The classes Jl_0 , Jl_1 , Jl_2 , and Jl_3 as row and column headings account for the canopy density classes (0,100], (10,100], (20,100], and (30,100], respectively. The test hypothesis is $H_0: X_1 \leq X_2$ where X_1 is the column Jl_n class and X_2 the row Jl_n class. The significance is indicated by $p^* < 0.05$, $p^{**} < 0.025$, $p^{***} < 0.01$, and $p^\dagger < 0.005$. We applied a Benjamini and Hochberg correction for multiple-pairwise testing.

		Latin America				Asia/Australia			
	Cls	Jl_0	Jl_1	Jl_2	Jl_3	Jl_0	Jl_1	Jl_2	Jl_3
Asia	Jl_0	1.	-	-	-	-	-	-	-
	Jl_1	-	1.	-	-	-	-	-	-
	Jl_2	-	-	1.	-	-	-	-	-
	Jl_3	-	-	-	1.	-	-	-	-
Africa	Jl_0	.00†	-	-	-	.00†	-	-	-
	Jl_1	-	.00†	-	-	-	.00†	-	-
	Jl_2	-	-	.00†	-	-	-	.00†	-
	Jl_3	-	-	-	.00†	-	-	-	.00†

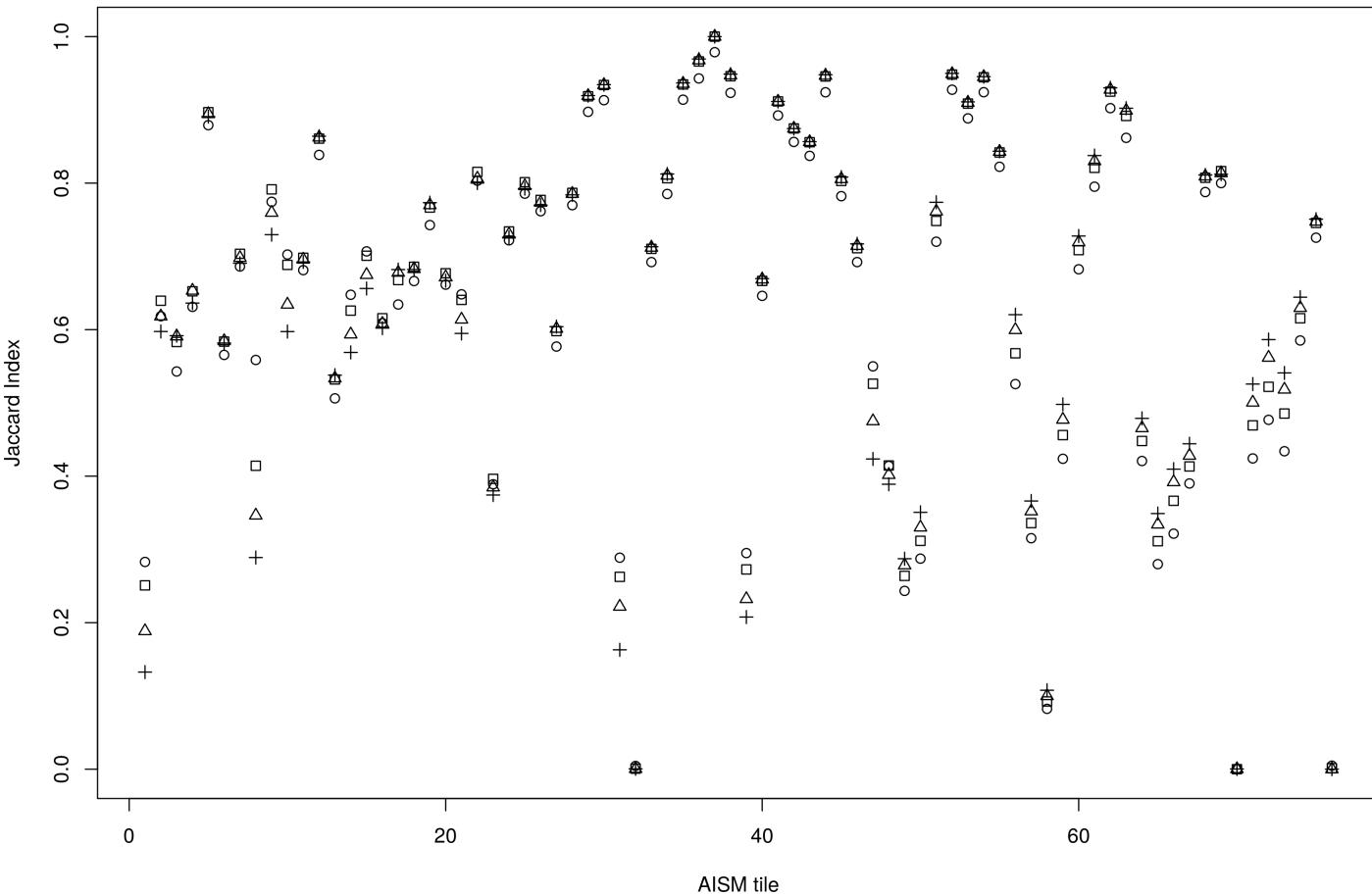


Figure B.1. Caption: Caption

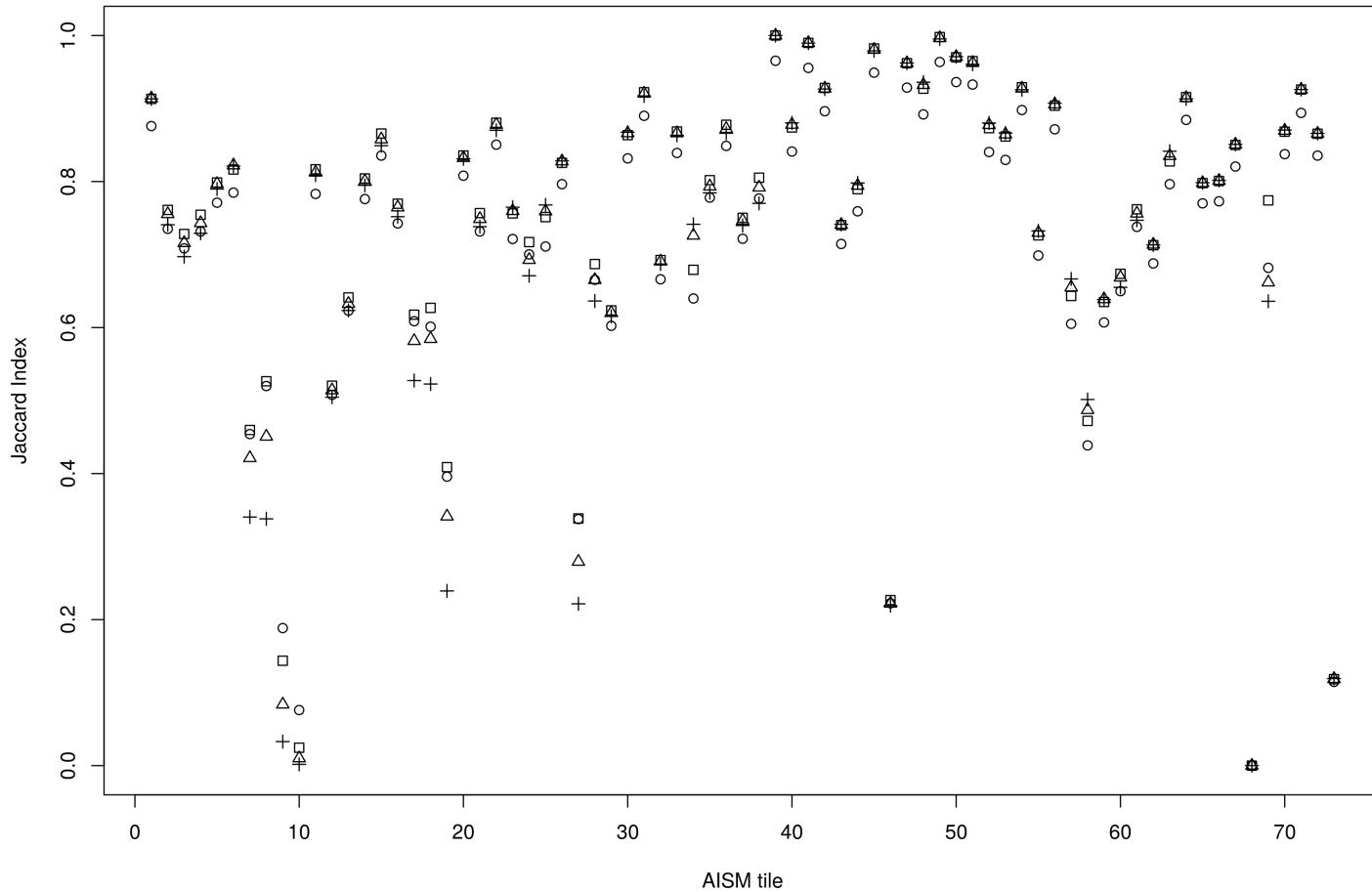


Figure B.2. Caption: Caption

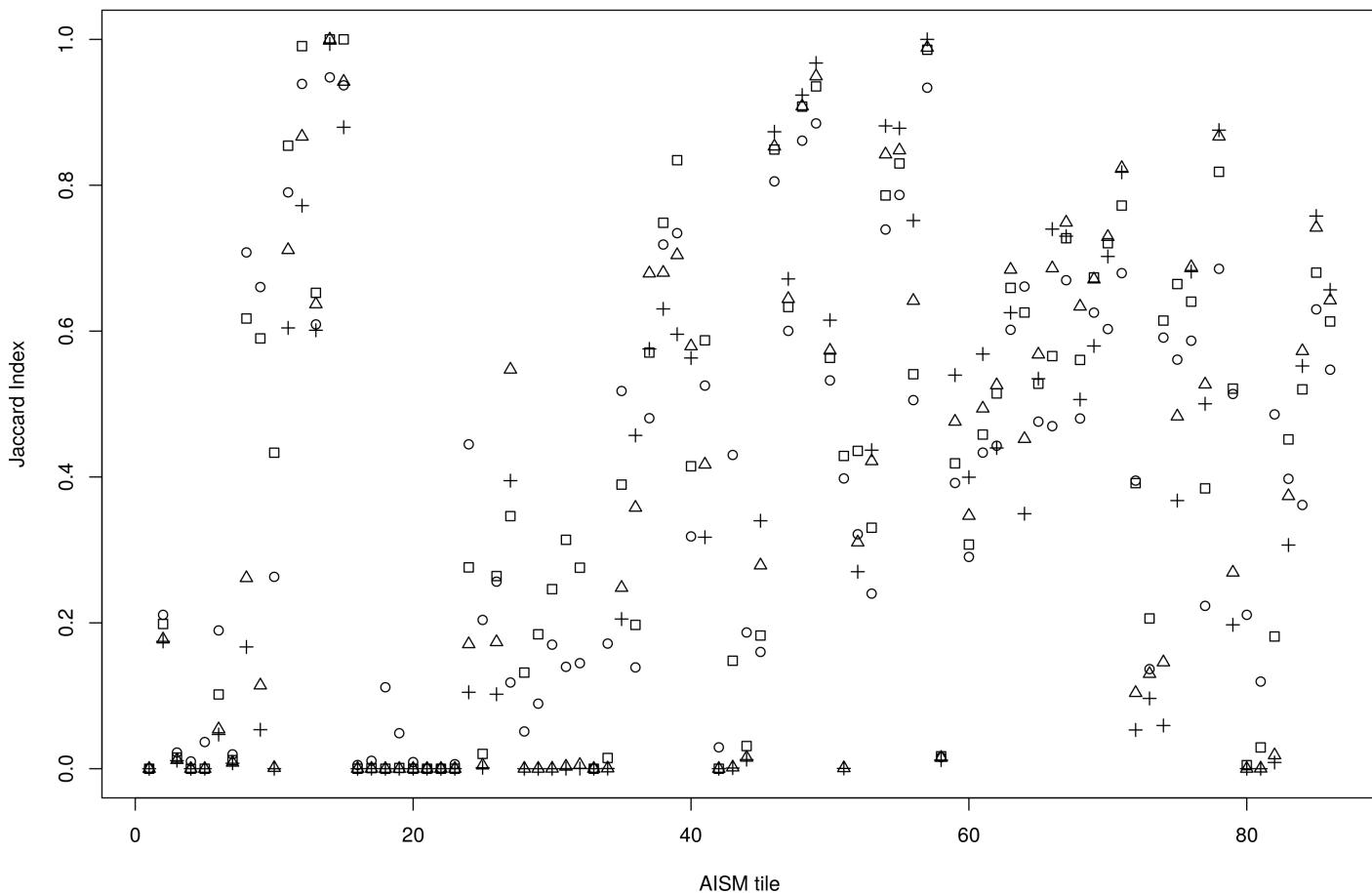


Figure B.3. Caption: Caption

C. Proximate deforestation drivers

Table C.1. Estimates of proximate drivers of deforestation for countries, continents, and tropical zone:

Region	Cultivated	Forest	Regrowth	Grassland	Shrubland	Water	Artificial	Bareland	Total	Total [†]
Argentina	5733.9 (65.8%)	1380.6 (15.9%)	147.9 (1.7%)	461.8 (5.3%)	718.7 (8.3%)	230.8 (2.7%)	32.8 (0.4%)	2.0 (0.0%)	8708.5	7097.1
Bahamas	0.0 (0.0%)	3.3 (41.2%)	0.1 (1.3%)	0.0 (0.0%)	0.3 (3.8%)	3.0 (37.5%)	1.3 (16.3%)	0.0 (0.0%)	8.0	1.7
Belize	212.2 (24.4%)	113.0 (13.0%)	155.5 (17.9%)	308.9 (35.6%)	54.5 (6.3%)	12.7 (1.5%)	11.1 (1.3%)	0.0 (0.0%)	867.9	742.2
Bolivia	9288.2 (37.2%)	5106.4 (20.5%)	2428.1 (9.7%)	5581.0 (22.4%)	1825.1 (7.3%)	590.8 (2.4%)	108.3 (0.4%)	19.5 (0.1%)	24947.4	19250.2
Brazil*	56443.7 (19.1%)	16433.7 (5.6%)	34746.7 (11.8%)	146802.9 (49.7%)	35954.2 (12.2%)	3832.4 (1.3%)	738.8 (0.3%)	209.0 (0.1%)	295161.4	274895.3
Chile*	-	-	-	-	-	-	-	-	-	-
Colombia	1297.9 (6.2%)	5272.3 (25.2%)	5177.2 (24.7%)	6995.9 (33.4%)	1593.1 (7.6%)	573.6 (2.7%)	26.0 (0.1%)	9.7 (0.0%)	20945.7	15099.8
Costa Rica	307.4 (21.2%)	536.8 (37.1%)	178.5 (12.3%)	397.4 (27.5%)	11.0 (0.8%)	7.0 (0.5%)	9.0 (0.6%)	0.1 (0.0%)	1447.2	903.4
Cuba	233.5 (17.6%)	499.9 (37.7%)	181.2 (13.7%)	330.4 (24.9%)	23.2 (1.8%)	35.2 (2.7%)	6.2 (0.5%)	15.7 (1.2%)	1325.3	790.2
Dominican Republic	81.6 (5.4%)	763.5 (50.5%)	122.2 (8.1%)	510.8 (33.8%)	2.9 (0.2%)	4.2 (0.3%)	13.3 (0.9%)	13.1 (0.9%)	1511.6	743.9
Ecuador	1143.9 (28.2%)	1163.7 (28.7%)	823.3 (20.3%)	735.7 (18.1%)	142.2 (3.5%)	34.7 (0.9%)	8.0 (0.2%)	6.3 (0.2%)	4057.8	2859.4
El Salvador	58.1 (11.6%)	28.2 (5.7%)	6.2 (1.2%)	390.9 (78.3%)	8.5 (1.7%)	1.6 (0.3%)	5.6 (1.1%)	0.0 (0.0%)	499.1	469.3
French Guiana	46.9 (14.7%)	52.8 (16.5%)	100.0 (31.2%)	91.4 (28.6%)	0.0 (0.0%)	7.8 (2.4%)	20.6 (6.4%)	0.5 (0.2%)	320.0	259.4
Guatemala	1843.2 (23.9%)	1172.6 (15.2%)	924.2 (12.0%)	2974.9 (38.6%)	733.1 (9.5%)	21.2 (0.3%)	39.6 (0.5%)	0.0 (0.0%)	7708.8	6515.0
Guyana	22.1 (3.1%)	244.7 (34.3%)	274.7 (38.5%)	134.0 (18.8%)	10.1 (1.4%)	19.9 (2.8%)	8.3 (1.2%)	0.0 (0.0%)	713.8	449.2
Haiti	8.3 (3.9%)	146.5 (69.3%)	13.9 (6.6%)	40.9 (19.3%)	0.1 (0.0%)	0.9 (0.4%)	0.2 (0.1%)	0.6 (0.3%)	211.4	64.0
Honduras	101.2 (2.5%)	1200.1 (30.0%)	487.9 (12.2%)	2050.5 (51.2%)	137.8 (3.4%)	20.4 (0.5%)	8.9 (0.2%)	0.0 (0.0%)	4006.8	2786.3
Jamaica	7.2 (2.8%)	127.7 (48.8%)	25.5 (9.7%)	82.4 (31.5%)	0.0 (0.0%)	1.9 (0.7%)	16.7 (6.4%)	0.3 (0.1%)	261.7	132.1
Mexico*	3267.8 (18.7%)	4857.8 (27.8%)	3763.8 (21.5%)	3950.6 (22.6%)	1169.5 (6.7%)	88.8 (0.5%)	373.4 (2.1%)	11.1 (0.1%)	17482.8	12536.2
Nicaragua	68.9 (0.9%)	3405.0 (45.9%)	1567.8 (21.2%)	2268.8 (30.6%)	78.3 (1.1%)	19.9 (0.3%)	2.5 (0.0%)	0.0 (0.0%)	7411.2	3986.3
Panama	142.2 (6.3%)	619.7 (27.6%)	352.8 (15.7%)	1019.7 (45.5%)	70.8 (3.2%)	22.3 (1.0%)	13.5 (0.6%)	0.4 (0.0%)	2241.4	1599.4
Paraguay*	13718.0 (48.9%)	6383.0 (22.8%)	216.2 (0.8%)	2480.2 (8.8%)	5107.2 (18.2%)	118.8 (0.4%)	24.9 (0.1%)	1.2 (0.0%)	28049.5	21547.7
Peru	780.4 (6.6%)	2333.9 (19.8%)	3394.0 (28.8%)	2710.8 (23.0%)	1230.9 (10.4%)	1213.5 (10.3%)	27.8 (0.2%)	102.2 (0.9%)	11793.5	8246.1
Puerto Rico	4.1 (3.7%)	56.4 (50.5%)	10.2 (9.1%)	30.7 (27.5%)	0.0 (0.0%)	0.4 (0.4%)	8.8 (7.9%)	1.0 (0.9%)	111.6	54.8
Suriname	36.7 (7.4%)	132.2 (26.8%)	167.5 (33.9%)	108.5 (22.0%)	0.0 (0.0%)	40.8 (8.3%)	8.2 (1.7%)	0.0 (0.0%)	493.9	320.9
Venezuela	1082.3 (10.2%)	2758.9 (26.1%)	1644.0 (15.5%)	3382.4 (32.0%)	1388.8 (13.1%)	267.4 (2.5%)	47.7 (0.5%)	12.7 (0.1%)	10584.2	7557.9
Latin America	95929.7 (21.3%)	54792.7 (12.2%)	56909.4 (12.6%)	183841.5 (40.8%)	50260.3 (11.1%)	7170.0 (1.6%)	1561.5 (0.3%)	405.4 (0.1%)	450870.5	388907.8
Algeria*	-	-	-	-	-	-	-	-	-	-
Angola	4818.6 (32.7%)	697.8 (4.7%)	452.5 (3.1%)	8065.9 (54.7%)	371.2 (2.5%)	90.9 (0.6%)	226.5 (1.5%)	10.9 (0.1%)	14734.3	13945.6
Benin	1518.7 (58.2%)	200.6 (7.7%)	45.1 (1.7%)	74.6 (2.9%)	762.5 (29.2%)	0.5 (0.0%)	5.6 (0.2%)	0.0 (0.0%)	2607.6	2406.5
Botswana*	9.4 (17.9%)	2.1 (4.0%)	0.0 (0.0%)	26.4 (50.2%)	3.6 (6.8%)	9.9 (18.8%)	1.2 (2.3%)	0.0 (0.0%)	52.6	40.6
Burkina Faso	510.9 (29.4%)	4.6 (0.3%)	0.0 (0.0%)	1190.0 (68.5%)	31.7 (1.8%)	0.3 (0.0%)	0.1 (0.0%)	0.7 (0.0%)	1738.3	1733.4
Burundi	49.7 (30.6%)	6.3 (3.9%)	3.6 (2.2%)	98.8 (60.9%)	2.0 (1.2%)	0.5 (0.3%)	1.3 (0.8%)	0.0 (0.0%)	162.2	155.4
Cameroon	337.8 (8.5%)	938.2 (23.6%)	717.2 (18.1%)	1838.8 (46.3%)	14.4 (0.4%)	34.6 (0.9%)	68.6 (1.7%)	21.6 (0.5%)	3971.2	2998.4

Region	Cultivated	Forest	Regrowth	Grassland	Shrubland	Water	Artificial	Bareland	Total	Total [†]
C. African Rep.	707.3 (18.3%)	217.1 (5.6%)	328.4 (8.5%)	2470.2 (64.1%)	118.6 (3.1%)	3.3 (0.1%)	9.8 (0.3%)	0.6 (0.0%)	3855.3	3634.9
Chad	1454.7 (54.7%)	1.8 (0.1%)	0.1 (0.0%)	1037.5 (39.0%)	142.2 (5.4%)	15.9 (0.6%)	3.1 (0.1%)	2.4 (0.1%)	2657.7	2640.0
Congo	107.8 (4.9%)	406.0 (18.4%)	591.8 (26.8%)	968.0 (43.8%)	0.3 (0.0%)	109.7 (5.0%)	25.8 (1.2%)	0.1 (0.0%)	2209.5	1693.8
DR Congo	4771.8 (9.4%)	3518.2 (6.9%)	13508.7 (26.6%)	26400.4 (52.0%)	508.2 (1.0%)	1736.9 (3.4%)	314.7 (0.6%)	0.1 (0.0%)	50759.0	45503.9
Djibouti	-	-	-	-	-	-	-	-	-	-
Egypt*	-	-	-	-	-	-	-	-	-	-
Equatorial Guinea	0.0 (0.0%)	77.2 (28.7%)	77.7 (28.9%)	72.6 (27.0%)	1.3 (0.5%)	1.3 (0.5%)	38.5 (14.3%)	0.0 (0.0%)	268.6	190.1
Eritrea	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.1 (100.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.1	0.1
Ethiopia	356.8 (16.1%)	296.9 (13.4%)	137.2 (6.2%)	1321.6 (59.5%)	92.3 (4.2%)	4.5 (0.2%)	3.9 (0.2%)	6.6 (0.3%)	2219.8	1918.4
Gabon	1.4 (0.1%)	528.7 (32.9%)	664.1 (41.3%)	362.7 (22.5%)	0.0 (0.0%)	22.9 (1.4%)	24.9 (1.5%)	3.9 (0.2%)	1608.6	1057.0
Gambia	35.6 (29.1%)	0.0 (0.0%)	0.0 (0.0%)	63.5 (51.9%)	19.8 (16.2%)	0.1 (0.1%)	3.4 (2.8%)	0.0 (0.0%)	122.4	122.3
Ghana	319.1 (6.8%)	525.3 (11.1%)	1523.7 (32.3%)	1990.8 (42.2%)	235.5 (5.0%)	27.6 (0.6%)	95.0 (2.0%)	0.0 (0.0%)	4717.0	4164.1
Guinea	261.3 (8.2%)	61.2 (1.9%)	118.0 (3.7%)	2296.7 (72.4%)	419.2 (13.2%)	3.3 (0.1%)	11.4 (0.4%)	2.1 (0.1%)	3173.2	3108.7
Guinea Bissau	136.7 (24.5%)	2.3 (0.4%)	15.5 (2.8%)	181.5 (32.5%)	216.6 (38.8%)	1.2 (0.2%)	3.9 (0.7%)	0.0 (0.0%)	557.7	554.2
Ivory Coast	1636.4 (12.5%)	2619.8 (20.0%)	2515.1 (19.2%)	6147.4 (46.9%)	81.7 (0.6%)	34.0 (0.3%)	69.1 (0.5%)	0.7 (0.0%)	13104.2	10450.4
Kenya	1178.9 (48.1%)	118.8 (4.9%)	213.6 (8.7%)	832.7 (34.0%)	85.7 (3.5%)	13.7 (0.6%)	4.2 (0.2%)	1.6 (0.1%)	2449.2	2316.7
Liberia	8.1 (0.2%)	395.4 (12.0%)	1863.5 (56.6%)	986.7 (30.0%)	3.5 (0.1%)	2.2 (0.1%)	30.7 (0.9%)	0.0 (0.0%)	3290.1	2892.5
Libya*	-	-	-	-	-	-	-	-	-	-
Madagascar	197.1 (1.6%)	667.1 (5.3%)	5126.0 (40.6%)	5627.5 (44.6%)	957.9 (7.6%)	43.5 (0.3%)	1.5 (0.0%)	0.1 (0.0%)	12620.7	11910.1
Malawi	457.8 (40.0%)	185.4 (16.2%)	83.6 (7.3%)	398.3 (34.8%)	8.3 (0.7%)	1.9 (0.2%)	4.6 (0.4%)	4.7 (0.4%)	1144.6	957.3
Mali	480.6 (39.4%)	1.0 (0.1%)	0.0 (0.0%)	708.7 (58.1%)	25.6 (2.1%)	0.6 (0.0%)	2.9 (0.2%)	0.3 (0.0%)	1219.7	1218.1
Mauritania	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.2 (33.3%)	0.4 (66.7%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.6	0.6
Morocco*	-	-	-	-	-	-	-	-	-	-
Mozambique*	6419.1 (32.7%)	1527.7 (7.8%)	1967.4 (10.0%)	8959.4 (45.6%)	664.8 (3.4%)	49.2 (0.3%)	37.0 (0.2%)	17.8 (0.1%)	19642.4	18065.5
Namibia*	46.1 (38.0%)	6.8 (5.6%)	0.0 (0.0%)	59.9 (49.4%)	5.1 (4.2%)	2.1 (1.7%)	1.0 (0.8%)	0.2 (0.2%)	121.2	112.3
Niger	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.9 (10.3%)	0.0 (0.0%)	7.8 (89.7%)	0.0 (0.0%)	0.0 (0.0%)	8.7	0.9
Nigeria	2392.5 (34.4%)	1462.3 (21.0%)	452.3 (6.5%)	1754.3 (25.2%)	689.8 (9.9%)	24.4 (0.4%)	122.3 (1.8%)	63.6 (0.9%)	6961.5	5474.8
Rwanda	65.9 (42.7%)	18.2 (11.8%)	16.3 (10.6%)	46.3 (30.0%)	5.8 (3.8%)	0.6 (0.4%)	1.2 (0.8%)	0.0 (0.0%)	154.3	135.5
Senegal	221.4 (30.0%)	0.3 (0.0%)	0.1 (0.0%)	401.5 (54.4%)	112.2 (15.2%)	0.7 (0.1%)	1.6 (0.2%)	0.0 (0.0%)	737.8	736.8
Sierra Leone	11.1 (0.7%)	167.3 (10.1%)	455.9 (27.5%)	995.0 (60.1%)	6.1 (0.4%)	3.2 (0.2%)	16.3 (1.0%)	0.0 (0.0%)	1654.9	1484.4
Somalia	7.2 (11.0%)	10.0 (15.3%)	0.1 (0.2%)	0.9 (1.4%)	46.0 (70.6%)	0.9 (1.4%)	0.1 (0.2%)	0.0 (0.0%)	65.2	54.3
Somaliland	0.0 (0.0%)	0.2 (66.7%)	0.0 (0.0%)	0.0 (0.0%)	0.1 (33.3%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.3	0.1
South Africa*	47.5 (8.3%)	2.9 (0.5%)	319.2 (56.1%)	190.8 (33.5%)	2.8 (0.5%)	1.6 (0.3%)	4.2 (0.7%)	0.3 (0.1%)	569.3	564.8
South Sudan	291.4 (21.4%)	64.7 (4.7%)	15.1 (1.1%)	792.5 (58.1%)	161.9 (11.9%)	30.5 (2.2%)	8.2 (0.6%)	0.3 (0.0%)	1364.6	1269.4
Sudan	1.2 (3.7%)	0.2 (0.6%)	0.0 (0.0%)	22.5 (69.7%)	7.7 (23.8%)	0.6 (1.9%)	0.0 (0.0%)	0.1 (0.3%)	32.3	31.5
Tanzania	6677.4 (42.1%)	1013.2 (6.4%)	1397.1 (8.8%)	6436.0 (40.6%)	247.6 (1.6%)	38.0 (0.2%)	28.6 (0.2%)	7.2 (0.0%)	15845.1	14793.9

Region	Cultivated	Forest	Regrowth	Grassland	Shrubland	Water	Artificial	Bareland	Total	Total [†]
Togo	345.5 (48.2%)	65.1 (9.1%)	8.4 (1.2%)	23.8 (3.3%)	270.0 (37.7%)	2.1 (0.3%)	1.7 (0.2%)	0.0 (0.0%)	716.6	649.4
Uganda	777.8 (27.7%)	75.7 (2.7%)	153.3 (5.5%)	1772.0 (63.1%)	9.6 (0.3%)	15.0 (0.5%)	4.7 (0.2%)	0.0 (0.0%)	2808.1	2717.4
West Sahara	-	-	-	-	-	-	-	-	-	-
Zambia	5966.6 (53.5%)	1111.0 (10.0%)	110.1 (1.0%)	3614.2 (32.4%)	216.0 (1.9%)	70.0 (0.6%)	54.5 (0.5%)	0.0 (0.0%)	11142.4	9961.4
Zimbabwe	1662.3 (55.6%)	95.7 (3.2%)	100.1 (3.3%)	1069.8 (35.8%)	51.5 (1.7%)	4.0 (0.1%)	6.5 (0.2%)	0.1 (0.0%)	2990.0	2890.3
Africa	44289.5 (22.8%)	17093.1 (8.8%)	32980.8 (17.0%)	89301.4 (46.0%)	6599.5 (3.4%)	2410.0 (1.2%)	1238.6 (0.6%)	146.0 (0.1%)	194058.9	174555.8
Australia*	0.0 (0.0%)	1.0 (17.2%)	0.6 (10.3%)	0.8 (13.8%)	2.9 (50.0%)	0.2 (3.4%)	0.0 (0.0%)	0.3 (5.2%)	5.8	4.6
Bangladesh*	86.5 (19.4%)	216.9 (48.7%)	85.0 (19.1%)	36.9 (8.3%)	8.2 (1.8%)	2.0 (0.4%)	8.7 (2.0%)	1.1 (0.2%)	445.3	226.4
Brunei	3.4 (1.3%)	24.4 (9.0%)	204.2 (75.3%)	17.5 (6.5%)	0.0 (0.0%)	2.9 (1.1%)	18.7 (6.9%)	0.0 (0.0%)	271.1	243.8
Cambodia	3014.8 (34.0%)	2279.7 (25.7%)	2200.1 (24.8%)	1027.9 (11.6%)	0.0 (0.0%)	327.9 (3.7%)	13.3 (0.2%)	0.0 (0.0%)	8863.7	6256.1
China*	3193.3 (15.4%)	3009.9 (14.5%)	10801.1 (52.0%)	2710.0 (13.0%)	940.1 (4.5%)	57.8 (0.3%)	59.8 (0.3%)	0.2 (0.0%)	20772.2	17704.5
East Timor	4.0 (3.0%)	55.8 (41.3%)	43.3 (32.0%)	29.1 (21.5%)	0.0 (0.0%)	1.9 (1.4%)	0.0 (0.0%)	1.1 (0.8%)	135.2	77.5
India*	658.8 (23.4%)	976.9 (34.7%)	716.9 (25.5%)	290.5 (10.3%)	142.4 (5.1%)	8.6 (0.3%)	20.7 (0.7%)	1.1 (0.0%)	2815.9	1830.4
Indonesia	17415.1 (15.4%)	11227.1 (9.9%)	77352.2 (68.6%)	5209.1 (4.6%)	0.0 (0.0%)	1312.1 (1.2%)	303.0 (0.3%)	17.2 (0.0%)	112835.8	100296.6
Laos	1390.7 (14.8%)	1078.2 (11.5%)	5352.5 (57.0%)	1537.3 (16.4%)	9.7 (0.1%)	11.4 (0.1%)	4.3 (0.0%)	6.4 (0.1%)	9390.5	8300.9
Malaysia	2722.1 (7.5%)	3238.3 (8.9%)	29001.4 (79.4%)	963.6 (2.6%)	0.0 (0.0%)	249.8 (0.7%)	359.4 (1.0%)	0.3 (0.0%)	36534.9	33046.8
Myanmar*	2430.4 (24.5%)	2406.7 (24.3%)	3622.2 (36.6%)	904.1 (9.1%)	383.8 (3.9%)	140.0 (1.4%)	21.1 (0.2%)	0.1 (0.0%)	9908.4	7361.7
Pakistan	0.1 (50.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.1 (50.0%)	0.0 (0.0%)	0.0 (0.0%)	0.2	0.1
Papua New Guinea	177.6 (3.8%)	1264.6 (26.8%)	2630.7 (55.7%)	318.8 (6.8%)	188.1 (4.0%)	115.8 (2.5%)	14.8 (0.3%)	10.9 (0.2%)	4721.3	3340.9
Philippines	550.9 (12.4%)	1247.3 (28.1%)	2178.2 (49.0%)	423.3 (9.5%)	0.1 (0.0%)	37.4 (0.8%)	7.0 (0.2%)	0.2 (0.0%)	4444.4	3159.7
Saudi Arabia*	-	-	-	-	-	-	-	-	-	-
Solomon Island	0.0 (0.0%)	84.6 (36.2%)	101.2 (43.2%)	19.6 (8.4%)	27.4 (11.7%)	1.0 (0.4%)	0.1 (0.0%)	0.1 (0.0%)	234.0	148.4
Sri Lanka	274.4 (42.0%)	213.5 (32.7%)	108.3 (16.6%)	42.4 (6.5%)	0.0 (0.0%)	5.2 (0.8%)	9.3 (1.4%)	0.3 (0.0%)	653.4	434.7
Taiwan*	24.6 (11.2%)	110.4 (50.3%)	37.8 (17.2%)	45.2 (20.6%)	0.0 (0.0%)	0.7 (0.3%)	0.8 (0.4%)	0.0 (0.0%)	219.5	108.4
Thailand	1943.9 (22.6%)	1989.6 (23.1%)	3463.3 (40.2%)	1090.8 (12.7%)	0.0 (0.0%)	99.2 (1.2%)	29.5 (0.3%)	0.0 (0.0%)	8616.3	6527.5
Vanuatu	0.0 (0.0%)	0.1 (100.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.0 (0.0%)	0.1	0.0
Vietnam	2928.7 (32.8%)	1102.2 (12.3%)	2754.5 (30.8%)	1635.7 (18.3%)	443.5 (5.0%)	57.9 (0.6%)	19.5 (0.2%)	0.2 (0.0%)	8942.2	7782.1
Yemen*	-	-	-	-	-	-	-	-	-	-
Asia/Australia	36819.3 (16.0%)	30527.2 (13.3%)	140653.5 (61.2%)	16302.6 (7.1%)	2146.2 (0.9%)	2431.9 (1.1%)	890.0 (0.4%)	39.5 (0.0%)	229810.2	196851.1
Global	177038.5 (20.2%)	102412.9 (11.7%)	230543.7 (26.4%)	289445.5 (33.1%)	59005.9 (6.7%)	12011.9 (1.4%)	3690.1 (0.4%)	590.9 (0.1%)	874739.6	