
EDDY COVARIANCE TOWERS AS SENTINELS OF ABNORMAL RADIOACTIVE MATERIAL RELEASES

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1 Abstract

2 Ensuring accurate detection and attribution of abnormal releases of radioactive material is critical for protecting human
3 health and safety. Most commonly, such detection is accomplished via active monitoring approaches involving the
4 collection of physical samples. This is labor intensive and limits the temporal and spatial resolution of any detected
5 events to a relatively coarse level. As an alternative first step towards passive monitoring, we developed an approach
6 using eddy flux tower data records to identify signals from a known abnormal release and quantify the extent to which
7 that signal also occurs at other times in the data record. Through two case-studies, one of which targeted the Fukushima
8 nuclear disaster and the other targeting an abnormal release event at a radioisotope production facility in Fleurus
9 Belgium, we tested our approach and identified several potential heretofore unidentified abnormal events that were
10 consistent with atmospheric circulation patterns and/or wind direction from known release sites. Because our approach
11 is relatively simple and is resistant to systematic errors in the observational record, it has broad applicability beyond
12 specific constituents and ecosystem types to identify a wide variety of limited-duration anomalies in flux tower data to
13 ensure human health and industrial safety.

14 **Keywords:** eddy covariance, data mining, industrial safety, vegetation response, radioactivity, photosynthesis

15 1 Introduction

16 Detecting abnormal releases of radioactive material (hereafter simply “abnormal releases”) is important as a means
17 of ensuring human health and industrial safety. The detection of such releases is typically accomplished using active
18 approaches that require site-specific sampling or trace gas measurements (Loyalka, 1983) sometimes as a part of
19 international monitoring networks (e.g., Sangiorgi et al., 2020). A potential shortcoming of these active monitoring
20 approaches is that they require collection of physical samples (particulates, gasses, and other signal carriers) to
21 retrospectively determine the nature of the release and the events that may have led up to it. Active monitoring with
22 such means is thus labor intensive and limits the temporal and spatial resolution of any detected events to a relatively
23 coarse level.

24 A potential alternative to active monitoring is passive monitoring (e.g., the German Integrated Measuring and Information
25 system and the International Monitoring System, Bieringer and Schlosser, 2004). Although existing wide area
26 monitoring systems are already critical components of abnormal release detection and tracking programs (Medici, 2001),
27 diversification-of-approach may increase the programs’ sensitivity and comprehensiveness. With a greater diversity
28 of methods, monitoring programs may be able to detect a wider array of constituents with differing characteristics or
29 expand their geographic extent without the need to deploy new instrumentation. In the present study, we developed a
30 novel passive monitoring approach to supplement existing programs that involves the use of eddy covariance data to
31 capture a signal from known abnormal releases. This signal is then used to investigate whether additional potential
32 abnormal releases exist at other times in the data record. We focused on eddy flux towers, as they have the potential
33 to provide abnormal release signatures, are globally distributed, have been in near continuous operation for several
34 decades, and can be relatively cheaply and easily deployed in locations of interest. Therefore, they constitute a rich
35 target for data mining approaches aimed at identifying signatures resulting from direct interference of radioactive
36 material with target measurements and/or the indirect signal of vegetation responses to the deposition of radioactive
37 material.

38 Our utilization of flux tower records for event discrimination stands in contrast to the typical use of data from the
39 towers for direct carbon balance accounting. Whereas carbon balance accounting activities deal directly in the units of
40 the measured data for computing quantities like net ecosystem exchange (Baldocchi et al., 2018), our use treats each
41 variable as potentially including an indirect signal (i.e. an anomaly from normal vegetation behavior) of radioactive
42 material deposition irrespective of its physical properties (e.g., mass, volume, etc). In this way, we are not directly
43 measuring radioactive deposition but rather leveraging the fact that existing ecosystem monitoring records reflect
44 vegetation photosynthesis rates and general stress condition of site vegetation in a known way (Nobel, 2020).

45 We specifically focused on short-term anomalies in the vegetation response data (on the order of days to weeks) due
46 to either potential deposition of radioactive material or direct sensor effects rather than chronic long-term signals
47 on the order of years to decades. In the first case study, we tested the ability of various data mining techniques to
48 recover the signal from the closest tower (approximately 18 km away) to a known release coming from the Institut des

49 Radioelements (IRE) in Belgium (hereafter called the known release location or the “facility”) nominally occurring on
50 August 23, 2008. The IRE release originated from waste tanks into the atmosphere and lasted for several days totaling
51 50 GBq 131I (Carlé et al., 2010). For example, grass samples collected in the vicinity of the facility found radioactivity
52 levels of 5000 Bq/kg. In the second case study, we further explored potential signals in flux tower data related to the
53 Fukushima Nuclear Disaster nominally occurring on March 11, 2011 releasing a massive radioactive plume that reached
54 North America in 5 days that eventually dispersed throughout the entire northern hemisphere (Mészáros et al., 2016).
55 We used the signal identified during the known release to identify additional potential abnormal releases occurring at
56 other times in the data record and at other eddy flux towers. While the body of literature on direct or indirect effects
57 of radioactive material on vegetation is limited, we expect that exposure may cause declines in photosynthetic CO_2
58 uptake. This may be expressed as a change in the slope of the CO_2 relationship with other environmental variables (e.g.
59 air temperature, humidity, precipitation, etc.) before, during, and after exposure.

60 2 Methods

61 2.1 Data Description

62 For the IRE case study, we gathered measurements from eddy covariance platforms (i.e., “flux towers”) located in
63 Belgium, which are part of the ICOS (Integrated Carbon Observation System) network available through the European
64 Fluxes Database (<http://www.europe-fluxdata.eu/>). Three sites in proximity to the known release location had sufficient
65 long-term data for our anticipated data mining efforts (BE-Lon, BE-Bra, and BE-Vie sites, Figure 1A). Each of the
66 three flux tower sites are located within a crop ecosystem type and had almost 10 years of data available from 2004
67 through 2013 recorded at 30 min intervals. While BE-Lon had a near-continuous data record during this period, BE-Bra
68 had a long period of missing data in 2005, and both BE-Bra and BE-Vie had a long period of missing data in 2009. The
69 distance between IRE and the flux tower sites are approximately 19 km, 95 km, and 105 km for the BE-Lon, BE-Bra,
70 and BE-Vie, respectively.

71 For the Fukushima case study, due to the direction of the plume, we gathered measurements from flux towers located in
72 the Northern Hemisphere (Western United States), which are part of the Ameriflux network (<https://ameriflux.lbl.gov/>).
73 To increase confidence that we are seeing a real effect, we contrasted the results from Northern Hemisphere sites, where
74 we expect a higher likelihood of impact, to measurements from flux towers located in the Southern Hemisphere, which
75 are part of the OzFlux network (Cleverly, 2011). We selected sites with sufficient long-term data to cover the 2011 date
76 of the disaster and excluded boreal sites with extreme snow cover related seasonality. We ultimately selected three sites
77 (US-Wrc, US-GLE, and OZ-Mul, Figure 1B).

78 2.2 Modeling Approach

79 We processed each data file by 1) excluding nighttime measurements based on photosynthetic photon flux density
80 (PPFD > 100 $\mu\text{mol m}^{-2} \text{s}^{-1}$) or net solar radiation (netrad > 0 W/m²) and 2) defining an “event period” following

81 the publicly released date of the abnormal release. We focused on the following variables as potentially containing a
82 signature of the release : carbon dioxide, carbon dioxide flux (fc), latent heat (le), and sensible heat flux (h). Conversely,
83 the following were treated as independent explanatory variables: wind speed, precipitation, atmospheric pressure,
84 relative humidity, photosynthetic photon flux density, air temperature, and net radiation. Additionally, we used wind
85 direction data from the eddy flux towers to determine when or if any released material might plausibly reach a particular
86 location. Each of these variables were recorded by the flux tower instrumentation at a 30 minute (min) interval barring
87 any gaps due to missing data.

88 To begin our analysis, we screened all pairwise, linear relationships between the aforementioned independent and
89 dependent variables using a linear regression fit between each pair for the entire period of record (n=51). Variable
90 pairs were excluded from further analysis if the overall coefficient of determination (R^2) was less than 0.1. This low
91 threshold was chosen to maximize the number of pairs that could potentially be investigated more deeply. From visual
92 inspection, we determined that many of the rejected variable pairs appeared to have a nonlinear relationship. As our
93 aim was to develop a methodology that could be applied broadly to other sites, we did not apply transformation to
94 linearize these relationships without a physical justification for the transformation. Evaluating an exhaustive list of data
95 transformations was beyond the scope of the present effort. After identification of candidate variable pairs, we fit an
96 interaction model of the following form where Y represents a continuous response (e.g., CO_2 , latent heat, etc.), X_1 is
97 an independent variable (e.g., air temperature, humidity, etc.), W_2 is a categorical variable representing the “period”
98 of observation (i.e., during, before, or after, the event), β_0 is the intercept, $\beta_{1,2}$ are slope parameters for X_1 and W_2
99 respectively, and β_3 is the slope parameter for the interaction between X_1 and W_2 :

$$Y = \beta_0 + \beta_1 X_1 + \beta_2 W_2 + \beta_3 X_1 W_2$$

100 This interaction model was fit for a window of time encompassing the known release event and for every other possible
101 window in the period-of-record. The overall length of the window was treated as an adjustable hyperparameter since
102 the length of time for an abnormal material release to reach the footprint of the flux tower may be event-dependent
103 and/or unknown. In addition, the length of time for deposited material to induce a possible vegetation response (or at
104 least a flux data anomaly) is an unknown quantity. Thus, in the default base case, we set the window length (n_days)
105 as 7 days, but we ran hyperparameter experiments to determine the result of setting it to longer periods of 10 and 14
106 days. The window length affects the duration of time designated as before and after the event. The period of time
107 designated as during the event was set as 2 days in all cases. We compared the effect size of the interaction term (β_3)
108 during the event-encompassing window against all other windows in a “rolling” analysis. Therefore, the overall analysis
109 compares the interaction term of the window of the known event against a window centered on every other observation.
110 Specific windows that had an effect size at least as large as that of the known event and with wind conditions defined as
111 “towards” the flux tower were flagged as “event detections”.

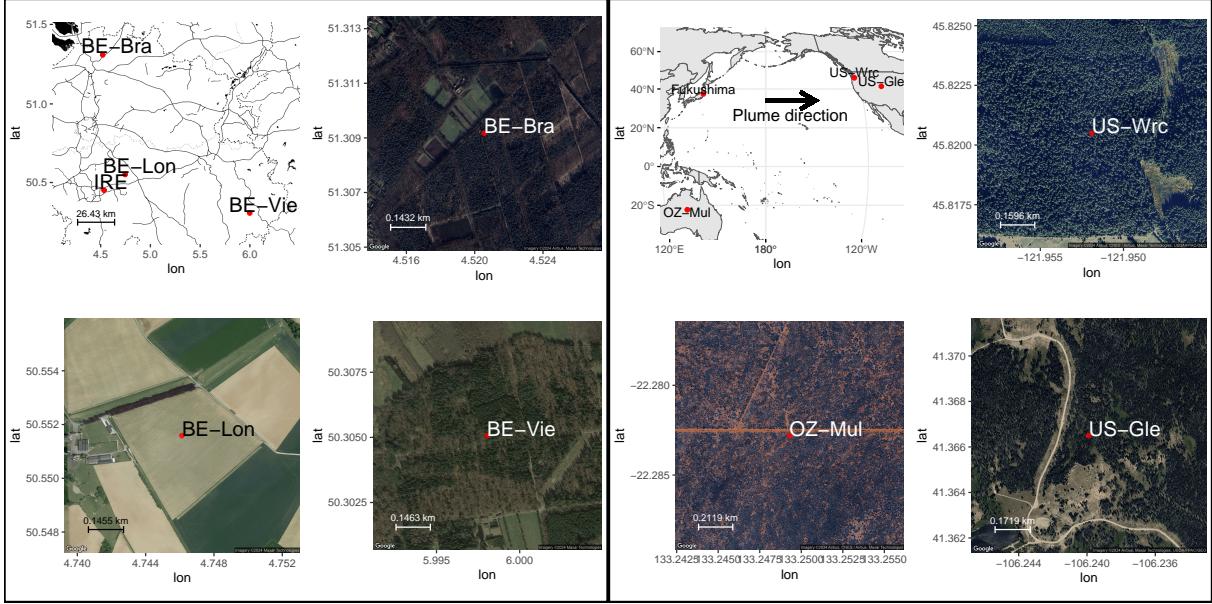


Figure 1: Map of flux tower locations for the IRE and Fukushima case studies (left and right respectively). Note that BE-Lon and BE-Vie are of a similar easterly direction to the IRE site whereas BE-Bra is in a northerly direction. The distance between IRE and the flux tower sites are approximately 19, 95, and 105 km for the BE-Lon, BE-Bra, and BE-Vie respectively. Each of the three IRE flux tower sites are located within a crop ecosystem type. Note that the US-Wrc site would be expected to have first contact with the Fukushima atmospheric plume followed by US-Gle. The OZ-Mul site would not be expected to have been affected by the plume.

112 In order to flag event detections, we defined several adjustable hyperparameters to account for uncertainty surrounding
 113 atmospheric transport and the timing of any possible responses to material deposition. The first hyperparameter we
 114 defined, `wind_tolerance`, reflects the range around the bearing from the facility to the tower location which counts
 115 as towards the tower. In the base default case, we set `wind_tolerance` to 10 degrees, but we explored setting it
 116 to smaller values down to 5 degrees and larger values up to 20 degrees. The second hyperparameter we defined,
 117 `event_quantile_effect`, handles the unknown travel time of any materials to be deposited via an abnormal release or
 118 what, if any, delays exist between deposition and vegetation response. Rather than simply selecting the maximum value
 119 of β_3 during a given window or the value at the exact time of the event, the value of `event_quantile_effect` is set to
 120 correspond with a specific quantile of all β_3 values during the event window. In the base default case, this was set at
 121 0.9, but we explored values as low as 0.5. All statistical analyses were carried out using the `statsmodels` Python package
 122 (Seabold and Perktold, 2009). Our processing scripts are openly available at [doi link], and we refer readers to the
 123 original providers for data access.

124 3 Results

125 For the IRE case study, among all possible pairwise combinations of dependent and independent variables ($n=51$, Table
 126 S1), we found that only the CO_2 versus air temperature comparison had any measure of predictability from the model
 127 ($R^2 > 0.05$) and a strong event effect size (Table 1). Although other variable pairs such as sensible heat flux and net

radiation had a stronger effect size for the BE-Lon site, they had a weak linear relationship across the other sites. When we focus on the period of the abnormal release event in particular rather than the overall period of record, we found that CO_2 versus air temperature relationship at each of the three focal towers had similar R^2 values and effect sizes (Table 1). Figure 2 provides a visual example of the effect size framing in our analysis. Note how the apparent interaction effect between time periods around the abnormal release event in Figure 2A (before, during, after) translates to a relatively high effect size (Figure 2C) whereas in Figure 2B at an arbitrary time point, there is no apparent interaction effect. This corresponds to a low effect size in Figure 2D. Note also how during and after the known abnormal release, the slopes of the relationships between CO₂ and air temperature become less negative than before the event, with lower than expected CO₂ at low air temperature (Figure 2A).

Despite the similar effect size at the three tower sites during the abnormal release event, they differed substantially in the degree to which they uniquely flagged the known release event. At the BE-Lon and BE-Vie sites, our event detection algorithm identified the known release and only a few other time periods as potential abnormal releases (Figure 3). Conversely, at the northerly BE-Bra site, the algorithm was not able to identify the known release event distinct from background noise. This can be seen in the rate of event detections (α) at each site, whereby it was less than 1% at the BE-Lon and BE-Vie sites, but was greater than 2% at the BE-Bra site. It is notable that this ability to uniquely detect the known release along with a plausible number of additional event detections corresponds with the fraction of the abnormal release period when the wind was pointed from the release location towards the respective tower (using the default base case wind tolerance of 10 deg) which was much lower at the BE-Bra site (5%) compared to the BE-Lon and BE-Vie sites (34% Table 1).

To increase our confidence that event detections were not dependent on the specific hyperparameters settings in our base-case setup, we ran an exhaustive cross-validation set of hyperparameter experiments. For this effort, we tested every possible combination of values for wind tolerance, number of days, and the event quantile effect. Across reasonable values of specific hyperparameters, we found that their magnitudes had little effect on the rate at which events were flagged as potential abnormal releases as even in the most extreme cases, we did not see detection rates exceeding 1%. We did, however, observe that the event detection was related to the window length and event quantile parameters as shown by the slope of the line in Figure 4. Although increasing values of the event quantile parameter would hypothetically increase event detection rates, we do not see value in setting this parameter below the median. Similarly, increasing the magnitude of the window length parameter would also likely increase the event detection rate but here too we do not see value in setting this parameter beyond 14 days given the results of prior atmospheric modeling efforts (e.g. Mészáros et al., 2016).

For the Fukushima case study, our objective was to test the sensitivity of our approach to a broad scale abnormal release and to introduce a control design to verify behavior on a flux tower where we expect no impact from any abnormal releases. First, we verified that event detection specificity is present at towers likely to be exposed even across long distances if the abnormal release is large (Figure 5). Then, we found a decrease in event detection specificity and an

Table 1: Overview of CO_2 vs air temperature interaction models at the IRE site. R^2 is the coefficient of determination of a linear model fit to the two covariates for the period of the abnormal release event (corresponding to “during” in Fig. 2). Effect is the event effect size of the interaction term ‘during’ the known release period matching the quantile of the distribution as specified by event quantile effect hyperparameter. Wind (%) is the average percentage of time that wind direction was from the site of known release towards the selected flux tower during the known release event. The remaining columns indicate the base-case hyperparameter settings.

Site	R^2	Effect	Wind(%)	Days(n)	Wind(tol)	Effect(q)
BE-Vie	0.35	36.2	34	7	10	0.9
BE-Bra	0.4	57.59	5	7	10	0.9
BE-Lon	0.47	30.13	34	7	10	0.9

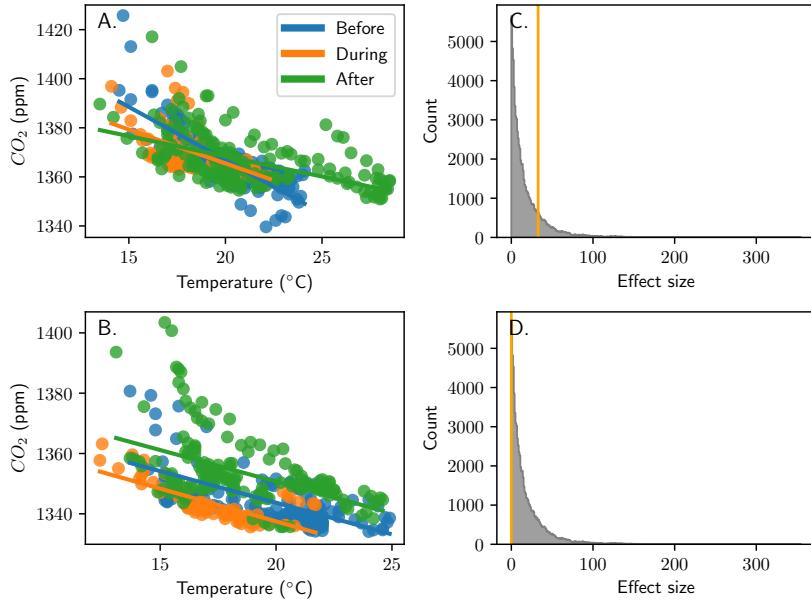


Figure 2: Interaction plots at the BE-Lon site between air temperature (C°) and CO_2 (ppm) for the period of the known abnormal release (A) and an arbitrary non-release period (B). Also shown is the distribution of all interaction effect sizes (gray) compared with the effect size of the time period shown in the corresponding left panels (orange, C, D).

162 increase in event detection rate moving from the flux tower expected to be most affected (US-Wrc), to a more distant
 163 (likely less affected) tower (US-Gle), to a tower likely not exposed at all to the atmospheric plume (OZ-Mul, Figure 5).

164 4 Discussion

165 Using flux tower records collected closest to and in the prevailing wind direction to the IRE and the Fukushima release
 166 locations, our event detection algorithm was able to identify both known abnormal releases and plausible previously
 167 unidentified abnormal events. There was broad agreement among the closest flux tower and more distant flux towers
 168 as to the timing and frequency of these releases, and our approach was supported by the loss of signal at flux towers
 169 outside of the prevailing wind direction. This suggests that, at the very least, given a facility known to have a prior
 170 abnormal release, flux tower networks are capable of providing a means of passive monitoring to detect subsequent
 171 events as well as information on the environmental conditions before, during, and after the event. Furthermore, our

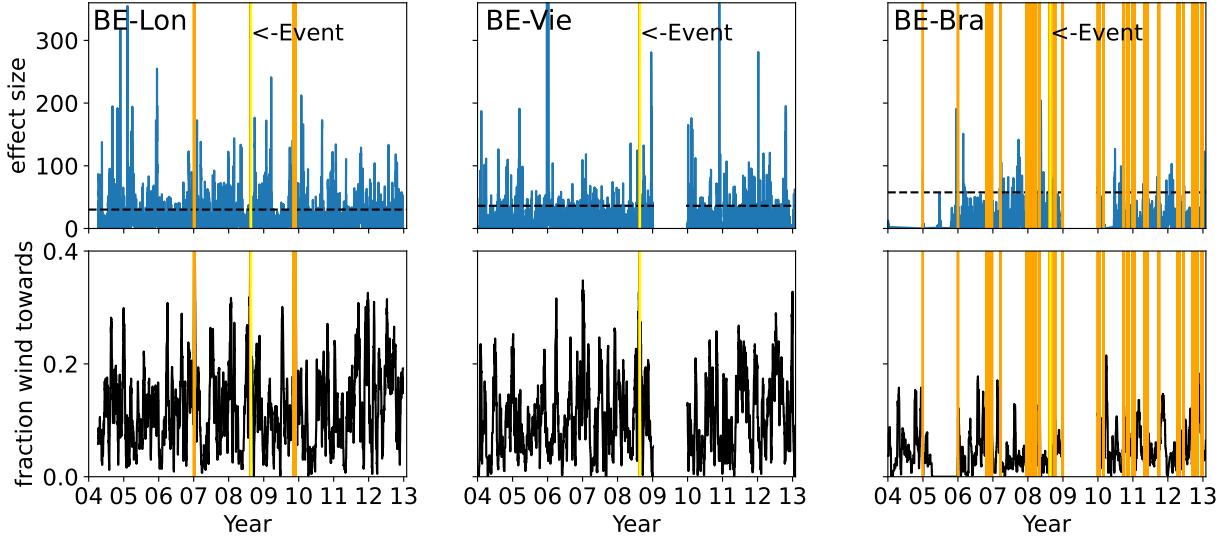


Figure 3: Time series plots (2004-2013) of rolling event detection analysis at the IRE site for the CO_2 and air temperature (C°) variable pair. Event detection lines (orange), interaction effect size (blue), and event effect size (dashed black line) are shown in the top panels. The fraction of time that the wind direction was towards the particular tower (solid black) are shown in the corresponding bottom panels.

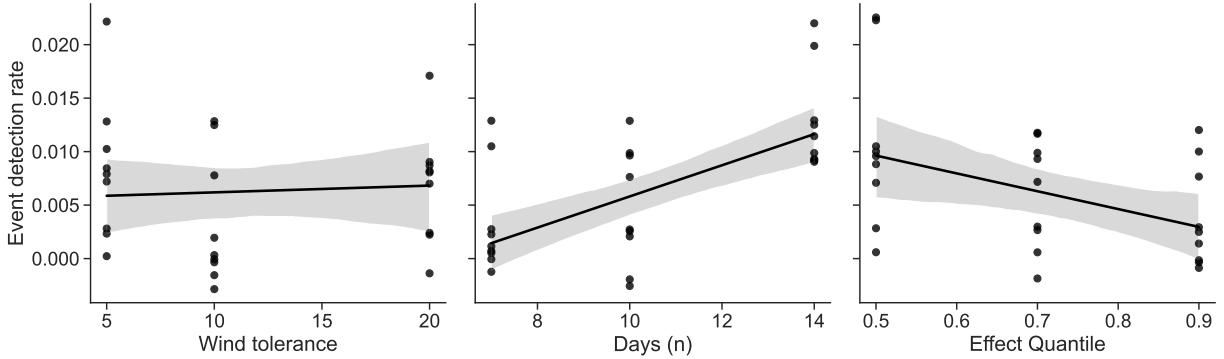


Figure 4: Sensitivity of event detection at the IRE site to different hyperparameter values (points) alongside potential relationship (solid line) and confidence interval (shaded region).

172 approach has promise for situations when there is not a prior known abnormal release. In these cases, it may be possible
 173 to continuously mine flux tower records (as they become available) for events that exceed some threshold identified in a
 174 more comprehensive cross-site, cross-event benchmarking study.

175 More generally, we show that flux tower records have value for this type of data mining despite the apparent noisiness
 176 of the data (Fratini et al., 2018). We attribute the sensitivity of our approach to the fact that we fit successive interaction
 177 models to limited sections of the data record, and we leveraged bivariate relationships to constrain noisiness in individual
 178 data records. Our approach was aided by the fact that this type of scale-free data mining is largely unaffected by
 179 systematic errors due to uncertainty in calibration standards and/or low resolution in specific sensor packages deployed
 180 on the towers. This is because although such systematic errors may bias the magnitude of overall flux values, which
 181 would be a problem for typical uses of flux towers that deal directly in the units of the measured data, they do not reduce
 182 the overall confidence in individual values (Langford et al., 2015), which for our purposes might represent an abnormal

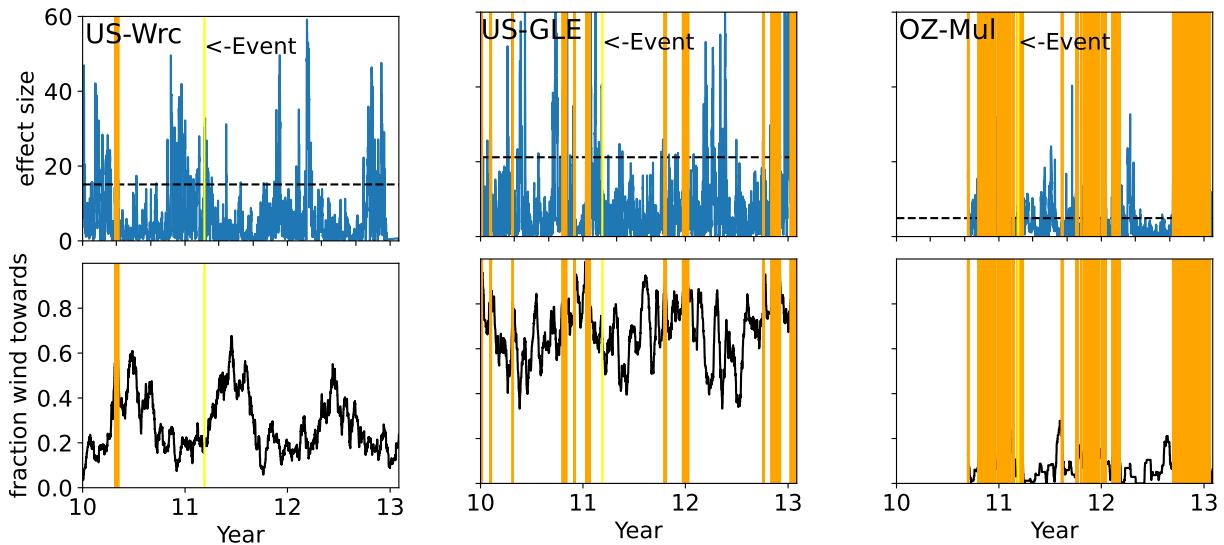


Figure 5: Time series plots (2010-2013) of rolling event detection analysis for the Fukushima disaster case study. The depicted variable pair is latent heat and relative humidity. Event detection lines (orange), interaction effect size (blue), and event effect size (dashed black line) are shown in the top panels. The fraction of time that the wind direction was towards the particular tower (solid black) are shown in the corresponding bottom panels.

183 release signal. The ability of our approach to identify a suitable release signature does likely depend on the quality and
184 completeness of the underlying data.

185 Another contributor to the success of our approach was our ability to compare event detections at close towers in
186 prevailing wind directions against more distant towers (refer to Figure 1), increasing our confidence that we were seeing
187 a true signal and not spurious noise. The relatively dense networks of flux towers with long data records leveraged
188 in both locations also aided these case studies. This level of flux tower density may be found in the United States,
189 Western/Central Europe, Japan, and Australia, which have high densities of flux towers, but not throughout the rest
190 of the world (Baldocchi et al., 2001; Pastorello et al., 2020). As a result, this may impose geographical limits on
191 our approach, although installation and maintenance of new flux towers in locations of interest may be cost effective
192 compared to other monitoring approaches. Another attribute that likely aided our investigations was the fact that all
193 three of the flux towers we examined were located in similar agricultural and forested ecosystem types for the IRE and
194 Fukushima case studies respectively. Although investigating whether or not flux tower combinations with differing
195 ecosystems would yield similar results is beyond the scope of the present study, we suspect that the data records among
196 towers in disparate ecosystem types could differ too drastically to be of comparative use. For example, Pastorello et al.
197 (2020) show that towers in the crop ecosystem type, of which all three sites used in this study belong, have a relatively
198 narrow distribution of fluxes (i.e., gross primary production, GPP) compared to towers in the grassland or evergreen
199 broadleaf forest types. One reason for this narrow spread in crop ecosystems may be the relative homogeneity of flux
200 tower footprints in crop ecosystems, which typically extend to within 1000m of the tower depending on atmospheric
201 conditions (e.g., wind speed and direction, Chu et al., 2021). Given that towers in other ecosystem types with a more
202 heterogeneous flux tower footprint have a wider distribution of fluxes (e.g. Pastorello et al., 2020), we suspect that

203 having all the towers in the IRE case study located in a homogenous ecosystem type was helpful in reducing the
204 spikiness and spread of the data.

205 The biggest improvements to our approach would likely come from coupling our data mining procedure with rigorous
206 air pollutant dispersion modeling (e.g., Mészáros et al., 2016). This would eliminate the need for indirect estimation
207 of material transport via uncertainty analyses and instead use simulation results to analytically determine the likely
208 arrival and duration of material deposition. Further improvements could be made by incorporating knowledge about
209 the composition of materials being deposited (IAEA, 2006; Mészáros et al., 2016) and/or the interactions between
210 material exposure and other environmental stressors (Mousseau and Møller, 2020), which likely affects the severity
211 and timing of ecosystem responses and by extension the fluxes being measured by a given tower. Such information
212 may help disentangle potential interference of atmospheric contaminants with infrared measurement of CO_2 from
213 declines in photosynthetic CO_2 uptake of contaminated vegetation given our observation that during and after the
214 known abnormal release, the slopes of the relationships between CO_2 and air temperature become less negative than
215 before the event, with lower than expected CO_2 at low air temperature. Although the chemical makeup of possible
216 atmospheric contaminants is unknown, interference by iodine itself is unlikely to affect measurements of CO_2 given
217 that the absorption spectra of iodine peaks at much shorter wavelengths around 10-7 m (Haynes, 2016) compared to
218 that of CO_2 which peaks around 1700-2100 10-9 m (LICOR Biosciences Inc., Lincoln, NE).

219 The fact that we were able to identify known releases (along with potential unidentified abnormal events) in the data
220 records from both a large event (i.e. Fukushima) as well as a smaller event despite not having a detailed radiological or
221 atmospheric transport model is a strength of our approach. Furthermore, our approach is not limited to a radiological
222 context, any abnormal event that affects the plant community and is reflected in flux tower data is a potential target.
223 Because our approach is light-weight and resistant to systematic errors in the observational record, it has broad
224 applicability beyond specific constituents and ecosystem types to identify a wide variety of limited-duration impacts to
225 the plant community within flux tower footprints to ensure human health and industrial safety.

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274 **Statements and Declarations**

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279 **Competing Interests**

280 The authors have no relevant financial or non-financial interests to disclose.

281 **Author Contributions**

282 JS built models, analyzed data, and wrote the paper. SS, LTD, and AJ contributed to the conception of the manuscript.
283 ECT, VAK, LTD, and EC edited the manuscript. All authors provided interpretation of results.

284 **Ethical Approval**

285 Not applicable

286 **Consent to Participate**

287 Not applicable

288 **Consent to Publish**

289 Not applicable

290 **Availability of Data and Materials**

291 No new data was produced in this study. All original data used is available publicly from their respective sources and
292 archived permanently with restrictions on redistribution. These can be found at <http://www.europe-fluxdata.eu/>,
293 <https://fluxnet.org/>, and (Cleverly, 2011).

294 Release of computer code and secondary data reuse that supports the results and analyses of the paper are
295 pending a United State Department of Energy Review pursuant to [https://www.osti.gov/doecode/FAQs#
296 how-do-i-announce-software-with-an-access-limitation-to-osti](https://www.osti.gov/doecode/FAQs#how-do-i-announce-software-with-an-access-limitation-to-osti).