How to Eliminate the Unnecessary So That The Necessary May Speak. Biological Reference Points.

Laurence T. Kell^{1,*}, Finlay Scott² Paul De Bruyn³, Richard D.M. Nash⁴

- 1 ICCAT Secretariat, C/Corazón de María, 8. 28002 Madrid, Spain.
- 3 The Fisheries Laboratory, Lowestoft, NR33 0HT, Suffolk, England
- 2 AZTI Tecnalia. Herrera kaia portualdea z/g, 20110 Pasaia, Gipuzkoa, Spain
- 4 Institute of Marine Research, PO Box 1870 Nordnes, 5817 Bergen, Norway
- * E-mail: Corresponding Laurie.Kell@iccat.int

Abstract

Author Summary

Introduction

The adoption of the Precautionary Approach to fisheries management [1] requires a formal consideration of uncertainty. An important principle of the approach is that the level of precaution should increase as uncertainty increases, e.g. from data rich to poor situations. However, defining stocks as data rich or data poor based purely on the availability of catch and effort data obscures the fact that considerable uncertainty often exists about the biological processes such as natural mortality, recruitment processes and stock structure of commercially important fish stocks. Conversely even when data are limited, empirical studies have shown that life history parameters, such as age at first reproduction, natural mortality, and growth rate are strongly correlated. Therefore biological knowledge is important both for evaluating the robustness of advice obtained from data-rich stock assessments and in allowing general rules, for example about choice of reference points (indicators of the status of a stock e.g. biomass too low or fishing pressure too high etc), to be derived for all stocks.

Fisheries management is concerned with trying to set, and then achieve, realistic management objectives. This is carried out through defining management measures of interest, for example spawning stock biomass (SSB) and fishing mortality (F), and then setting values of a range of reference points for these measures. These reference points can be used as targets, for example B_{MSY} , or limits, for example, F_{crash} . However, achieving these management objectives is made difficult by, amongst other things, the impact of biological and ecological uncertainty on the dynamics of the stock. A key question for fisheries management is therefore how can research be prioritised so that the impact of biological uncertainty on achieving management objectives can be reduced? Answering this question requires the estimation of the relative importance of the underlying biological assumptions made about the stock with respect to the management measures of interest. For example, does uncertainty about the stock recruitment relationship have a relatively bigger effect on yield and sustainability than uncertainty about natural mortality?

Sensitivity and elasticity analyses can be used to evaluate the effect of changes in system parameter values on system outputs. They are commonly used in financial and economic management, but have also have been applied to biology and conservation [2]. Such analyses can be used within fisheries management to identify the stock assumptions that have the greatest impact on management measures and therefore where the impact of uncertainty may have the greatest effect on achieving management objectives. Sensitivities measure absolute changes, for example, by how much does the estimate of MSY change as the value of length of first maturity changes. While elasticities measure the relative change and can be used to compare between a range of different sources of uncertainty. For example, does the length of first maturity have a larger proportional impact on estimates of MSY than the steepness of the stock recruitment relationship? Elasticity analysis has proven to be a useful tool in a number of areas of population and conservation biology, for example relating changes in vital rates to changes

in the population life history [3] and to quantities of importance in management such as population viability [4]. Previously, elasticity analysis has focused on terrestrial ecology [5–7] with limited application to marine populations [8, 9]. The applicability of this approach to resource management has therefore been demonstrated and here use it to evaluate the relative importance of the biological assumptions made in fishery stock assessment, that are too seldom questioned.

A fuller consideration of uncertainty within fisheries advice frameworks can be performed used Bayesian approaches or Managemement Strategy Evaluation (MSE). MSE is commonly used to evaluate the impact of different management measures, given a broad range of uncertainty. However, performing an MSE is a costly process in terms of human resources and can take several years. Therefore, tools such as elasticity analysis, which is comparatively less demanding to carry out, are important to help identify and focus research and management efforts. For example, is it more important to reduce uncertainty about the stock recruitment relationship or natural mortality or to develop harvest control rules that are robust to such uncertainty? Elasticity analyses can be used to answer such questions and prioritise research effort. It can also shift the current focus from defining stocks either as data poor or rich defined solely on fishery catch and effort towards a better understanding of biological processes.

Here we demonstrate the use of elasticity analysis for prioritising research effort with a study of the population dynamics of a fish stock based on life history theory. As such the study is not modelled on one specific species of fish. We do this by first simulating a stock based on life history relationships [10] and then by projecting the stock from an unfished to an over-exploited state. We do this in order to compute elasticities to allow us to evaluate the relative importance of the different system or biological parameters when assessing the stock relative to system characteristics defined by biological reference points. This allows us to address two important questions i.e. what is the relative importance of the different biological processes in providing advice and and how robust is advice based on the common biological reference points.

Materials and Methods

Empirical studies have shown that in teleosts there is significant correlation between the life history parameters such as age at first reproduction, natural mortality, and growth rate [11]. Additionally, size-spectrum theory and multispecies models suggest that natural mortality scales with body size [12], [13] [10]. This means that from something that is easily observable, like the maximum size, it is possible to infer the life history parameters of species for which data are not easily observable or available.

[10] summarised life history characteristics and the relationships between them for a range of stocks and species. These relationships were used to parameterise an age-structured population model using relationships that describe growth, maturation and natural mortality. This population was then projected at a constant fishing mortality until equilibrium was reached for a wide range of fishing mortalities.

The analysis allows us to evaluate where more biological knowledge is needed and to identify robust reference points for use in management. Following this analysis sensitivy analysis could be conducted to help quantify the costs and benefits and MSE to develop robust management advice.

SSB and fishing mortality (F) relative to the corresponding quantity estimated from each of the F_{MSY} , $F_{0.1}$ and F_{Crash} reference points were used as indices of stock status and exploitation. In the case of a fishing mortality equal to F_{Crash} , SSB is 0, by definition, therefore an SSB corresponding to 75% of F_{Crash} was used. F_{MSY} corresponds to the level of exploitation that provides the maximum sustainable yield, $F_{0.1}$ is a proxy for F_{MSY} and is the fishing mortality that corresponds to a point on the yield per recruit curve where the slope is 10% of that at the origin) and F_{Crash} the level of F that will drive the stock to extinction.

The elasticities of these indices in each year relative to the parameters in model were then used to evaluate the relative importance of the various processes (i.e. growth, maturation, stock recruitment, natural mortality and selectivity of the fishery) and the parameterisation of those process (e.g. K the

rate of growth and L_{∞}) with respect to stock status.

The calculation of reference points and fishing mortality also depend upon the selection pattern, since not all ages are equally vulnerable to a fishery. If there is a refuge for older fish, a higher level of fishing effort will be sustainable. Also, if the fecundity of older fish is greater than the fecundity of younger fish of the same mass-at-age, e.g. due to maternal effects or repeat spawners being more fecund then a condideration of the interactions between biology and selectivity will be important.

Uh oh just realised we have a problem with F for fishing mortality and fecundity - maybe best to switch the F for fecundity to EP i.e. egg production or have it as Fec

0.1 Life History Relationships

The Russell equation [14] summarises the key processes influencing the dynamics of exploited populations i.e.

$$f(B_2) = B_1 + (G+R) - (F+M) \tag{1}$$

where a biomass B_2 is a function of the biomass in the previous year (B_1) , gains due to growth (G) and recruitment (R) and losses due to fishing (F) and natural mortality (M). Two other factors have been recognised since Russel originally formulated this equation i.e. the gains through immigration (I) and losses through emigration (E) thus modifying the original equation to:

$$f(B_2) = B_1 + (G + R + I) - (F + M + H)$$
(2)

The knowledge about these processes affects our ability to provide robust scientific advice. In this paper we concentrate on G,R,F and M as we assume a single heterogenerous population with out emmigration (H) or immigration (I); However our approach could be extended to include H and I.

In order to provide a generic framework for modelling stock dymanics, life history relationships were used to parameterise appropriate functional forms for the various processes This allows processes to be modelled for a range of species and stocks under a variety of assumptions and for the impact of the various parameters to be evaluated.

Growth was modelled by the Von Bertalanffy growth equation [15]

$$L_t = L_{\infty}(1 - exp(-k(t - t_0))) \tag{3}$$

where K is the rate at which the rate of growth in length declines as length approaches the asymptotic length L_{∞} and t_0 is the time at which an individual is of zero length.

Length is converted to mass using the condition factor, a and allometric growth coefficient, b.

$$W = a \times L_t^b \tag{4}$$

Recruitment is split into Stock Reproductive Potential (SRP) and the stock recruitment relationship (SRR).

SRP is the sum of the products of the numbers of females, n, proportion mature-at-age, Q and their mean fecundity-at-age, F, i.e.

$$SRP = \sum n \times Q \times F \tag{5}$$

where their mean fecundity-at-age is equal to

$$K = a \times L^{b\prime} \tag{6}$$

if a and b are the same as in equation 3 then SRP is equivalent to female spawning stock biomss (SSB). However, generally the fecundity to length relationship differs from the weight to length relationship due to variations caused by fish condition and age effects altering the relationship between weight and eggs produced [?].

Proportion mature is modelled by the logistic equation with 3 parameters: age at 50% (a_{50}) and 95% (a_{95}) mature and the asymptotic value m_{∞} . The latter allows SRP to not be equivalent to stock mass-at-age.

$$f(x) = \begin{cases} 0 & \text{if } (a_{50} - x)/a_{95} > 5\\ a_{\infty} & \text{if } (a_{50} - x)/a_{95} < -5\\ \frac{m_{\infty}}{1.0 + 19.0 (a_{50} - x)/95)} & \text{otherwise} \end{cases}$$
(7)

Here the value of a50 comes from the empirical relationship between L_{∞} and age at maturity [10]:

$$a50 = 0.72 log L_{\infty}^{0.93} \tag{8}$$

We use a Beverton and Holt stock recruitment relationship reformulated in terms of steepness (h), virgin biomass (v) and $S/R_{F=0}$, where steepness is the ratio of recruitment at 20% of virgin biomass to virgin recruitment (R_0) .

For the BevertonHolt stock-recruit formulation:

$$R = \frac{0.8 \times R_0 \times h \times S}{0.2 \times S/R_{F=0} \times R_0(1-h) + (h-0.2)S}$$
(9)

Steepness is difficult to estimate from stock assessment data sets and there is often insufficient range in biomass levels that is required for its estimation [16]. Steepness and virgin biomass were set to 0.9 and 1000 t respectively.

Natural mortality at size is derived from the life history relationship [17].

$$M = exp(M1 + M2log(L) + 1.51log(L_{\infty}) + 0.97log(k) + a[5]/T), \tag{10}$$

where M1 (which determines the average natural mortality) = -2.11, M2 (which determines the rate at which natural mortality declines with length) = -1.70 and L is the average length of the fish (in cm) for which the M estimate applies. Here we use the length at mid-year to calculate the natural mortality at age.

Selection pattern of the fishery can be represented by a double normal (see [18]) with three parameters that describe the age at maximum selection (a1), the rate at which the lefthand limb increases (sl) and the righthand limb decreases (sr) which allows flat topped or domed shaped selection patterns to be chosen.

Even in data poor situations where catch-at-age for the entire catch time series is not available, some data will normally exist for some years or gears or for similar stocks and species. In cases where some length frequency data are available the shape of selection pattern, i.e. age at recruitment to the fishery, can be estimated using a method like that of [19]. This allows a double normal curve to be parameterised, i.e. age at maximum selectivity and whether the selection pattern is flat topped or dome shaped.

$$f(x) = \begin{cases} 2^{-[(x-a_1)/s_L]^2} & \text{if } x < a_1\\ 2^{-[(x-a_1)/s_R]^2} & \text{otherwise} \end{cases}$$
 (11)

0.2 Seasonality

The model is a discrete population model where the number of individuals in a year-class in a year is a function of the number of individuals in the previous year. However, processes like growth, maturation, natural mortality and fishing occur in different seasons of the year. Therefore to take account of this the age for which the expected values of mass, maturity and natural mortality-at-age can vary. For the stock mass and lengths-at-age are calculated at spawning time, catch mass-at-age is calculated in mid year and natural mortality is a function of the lengths-at-age mid year.

0.3 Stock projections

Using the relationships described above we generate a fully specified age-structured stock based on a value of $L_{\infty}=100$ cm. The stock is projected forward through time at different levels of constant fishing pressure ranging from no fishing (F=0) to over exploited $(F=F_{crash})$. The management measures of interest are the equilibrum SSB, yield and biomass relative to their reference point values corresponding to F_{MSY} , $F_{0.1}$ (a proxy for F_{MSY}) and F_{crash} , a limit reference point, e.g. SSB/SSB_{MSY} , $SSB/SSB_{F0.1}$ etc.

 $F_{0.1}$ is the fishing mortality on the yield per recruit curve where the slope is 10% of that at the origin, a conservative proxy for F_{MSY} . F_{Crash} is the fishing mortality that will drive the stock to extinction since it is equivalent to a R/S greater than the slope at the origin of the stock recruitment relationship, i.e. recruitment can not replace removals for a fishing mortality equal to F_{Crash} .

0.4 Elasticity

As mentioned above, elasticity analysis can be used to measure the proportional change of a system characteristic to a change in a system parameter. The general equation for calculating the elasticity of system characteristic y with respect to system parameter x is:

$$E_{y,x} = \left| \frac{\partial \ln y}{\partial \ln x} \right| \tag{12}$$

The absolute value operator is used for simplicity although the elasticity can also be defined without the absolute value operator when the direction of change is important, e.g. to evaluate if a reduction in natural mortality increases or decreases MSY reference points.

Here we calculate the elasticities of the management measures described above with respect to the life history and selectivity parameters grouped into the categories: growth $(K, t_0, a, b, L_{\infty})$, maturity (a50, ato95 and asym), natural mortality (M1 and M2), the stock recruitment relationship (h and vb) and the selectivity (a1, sl and sr). For example, the elasticity of SSB relative to SSB_{MSY} with respect to L_{∞} is calculated as:

$$E_{SSB_{MSY},L_{\infty}} = \left| \frac{\partial \ln SSB_{MSY}}{\partial \ln L_{\infty}} \right| \tag{13}$$

The elasticities are calculated for every level of F used in the projections and therefore show how the current state of the stock and exploitation rate affect the relative importance of the different life history parameters, i.e. where the most important source of uncertainty is.

Note that although the parameter values for K, maturity and natural mortality at age are calculated using the value of L_{∞} , these values are set before the elasticities are calculated, i.e. the elasticities with respect to L_{∞} do not reflect the impact of L_{∞} on these life history relationships, only on the impact of the stock dynamics through the von Bertalanffy growth equation.

Results

Growth, natural mortality, proportion mature and selectivity-at-age for the stock are shown in Figure 1. While the equilibrium values for projections over a range of fishing mortalitydynamics are shown in Figure 2, along with the MSY, $F_{0.1}$ and F_{crash} reference points. The equilibrium dynamics over the range of fishing mortalities from 0 to 75% of F_{crash} are presented as a phase plot (Figure 3) where the x-axis corresponds to biomass relative to B_{MSY} and the y-axis corresponds to harvest relative to F_{MSY} . Quadrants are defined for stock and fishing mortality relative to B_{MSY} and F_{MSY} ; i.e. red when $B < B_{MSY}$ and $F > F_{MSY}$, green if $B \ge B_{MSY}$ and $F \le F_{MSY}$; yellow otherwise. I.e. the red quadrant refers to an overfished stock subject to overfishing, green to a stock which is neither overfished or subject to overfishing.

Plots of the elasticities with respect to each life history parameter (grouped by life history process) for the range of fishing mortalities are presented in Figures 4 and 5. The vertical line indicates the boundary between the red and green quadrants and the horizontal line where the value of the elasticity is 0, i.e. where varying a parameter has no effect on the measure of interest.

The plots allows several important questions to be addressed, i.e. what is the relative impact of the different biological processes and parameter on the estimates of stock status and exploitation? Does the impact depend on the status of the stock so that knowledge of some parameters and processes is more important when a stock is depleted than when it is within safe biological limits. Also are F reference points for example more robust than those based on SSB and are some reference points more robust to uncertainty in key processes than others? Answering these questions will allow choice of robust target and limit reference points.

Inspection of the elasticities of relative SSB (Figure 4) shows that the elasticities of SSB relative to SSB_{MSY} and $SSB_{F0.1}$ are broadly similar. For all three reference points the natural mortality parameters M1 and M2 have the biggest proportional effect. The next most important parameter is a1 the selectivity parameter for age at full selection. The steepness of the stock recruitment relationship is important when considering SSB relative to SSB_{MSY} and $SSB_{F0.1}$ but less so relative to SSB_{Fcrash} . The other processes (growth and maturity) have similar impacts to each other; the most important parameters are K, age at 50% mature and age at recruitment to the fishery.

When the stock is underexploited and is above MSY (to the left of the vertical line) the elasticites tend to be of smaller magnitude than when when the stock is overexploited and the values for F_{crash} are on average larger than for the other two reference points. For a depleted stock (to the right of the line) the elasticites for F_{crash} are generally lower in magnitude. This suggests that when the stock is underexploited SSB_{MSY} and $F_{0.1}$ are more robust target reference points than F_{crash} , whilst if the stock is overexploited F_{crash} is the more robust reference point.

Figure 5 shows the elasticity of the constant fishing mortality relative to the three reference points as the fishing mortality increases. This effectively shows how changes in the life history parameters impact on the F-based reference points (is this right: ARE WE SURE THE FIGURE IS CORRECT?). The elasticity is constant across the range of fishing mortalities indicating that the importance of the life history parameters is not affected by the current level of exploitation. The elasticities with respect to the natural mortality parameters are larger than for the other processes suggesting that this is the most important process in determining the F based reference points. It is more important than the stock recruitment relationship, while MSY appears to have lowest elasticities and so is the more robust reference point for fishing mortality.

Discussion

Elasticity analysis was shown to be a useful technique for eluating the relative importance of processes. Although it will not replace other methods such as sensitivity analyses or MSE, since it is relatively

simple can be applied more readily to a range of stocks.

The analysis allowed the robustness of different reference points, quantities based upon them and the current status of stocks to be evaluated. This is important when choosing reference points for use as targets and limits both for biomass and fishing mortality. Since limit and target reference points should be robust to uncertainty at depleted and recovered stock levels respectively. Also if a reference point is dependent on an highly uncertain parameter or process then a better choice would be to find a reference point that is less dependent on that parameter or process.

Using the relationship Gislason relating mortality relationship and size is also more realistic than the common practice of using a value of M which does not vary with life history stage. However, a single M is often used for the earliest life history stage i.e. eggs to the end of the first year of life. [20]. [21] also suggest a single mortality rate, however points out the various processes that can serieally affect these early life history stages. [22] have looked specifically at plaice and illustrated the stage dependent variations in mortality rates, through the first year of life in plaice. There is scope for including further variability in M in the early life history using information in [23] and [24]. for example. In relation to settlement, overwintering and the juvenile stages there are very little information on many of the other commercially important species. Also using the relationships from Gislason ensures consistency between the biological parameters, as in the past people have borrowed parameters from different ocean inconsistently setting up clear contradictions in modelled populations. The basic framework can easily be adapted to model different species groups, e.g. demersal, tiny pelagic species that form the basis of real ecosystems or rather rare relatively unimportant large pelagics:-).

VBGF and growth: growth trajectories of individuals may not follow a VBGF. Therefore an interlink with M and differential mortality within a cohort and the population. Length-weight relationships and condition can affect the maturity ogive/schedule. Can vary annually based on changes in ecosystem productivity and density-dependent effects. Effects on egg production due to e.g. condition. The question of viable egg production and link to SRP [25]. Don't go too far in to this discussion, will be dealt with elsewhere. We use an estimation of recruitment - fraught with problems and is actually linked with the assumptions about M and the processes involved in the first year of life - see above. The question of productivity and the the absolute value of M.

Other factors that need to be considered include sub-stock structure and their dynamics Examples include herring [26] and the potential influence on the assessment process [27]. Sub-stock structure or metapopulations are known for quite a few stocks e.g. cod in the western Atlantic [28] and in the eastern Atlantic (North Sea) [29].

Typically, elasticity analysis is only concerned with the magnitude of the elasticity. However, the sign or direction of the elasticity can be important when the uncertainty, or noise, driving the parameter has an autocorrelation structure i.e. can not be represented by white noise. For example, it has been shown that there can be important cohort effects and autocorrelation in growth processes (REF 3 stocks paper). This may result in several continous years of high or low values for K.

1 Conclusions

What we did Compared the relative importance of biological parameters when assessing stocks relative to target and limit reference points

What we found That in general target reference points such as MSY and $F_{0.1}$ are more robust as limit reference points that actual limit reference points such as F_{Crash} . The importance of processes and parameters depend upon stock status and current fishing mortality. This illustrates the importance of considering refere points not in isolation but as part of the design of HCRs. For example if you know that a parameter is highly uncertain then when chosing a target or limit reference point then you should choose a reference point that is robust to such uncertainty, i.e. if you don't know the

shape of the M curve use a multiple (e.g. 1.5) of F0.1 as a limit refered point instead of F_{Crash}

What we didnt The analysis is limited in that it assumes a given model structure, i.e. exponentially declining M, SSB is an appropriate measure of SRP and a Beverton and Holt SRR. There are two issues here a) we don't actually know the correct functional form of M and SRR and b) we don't know whether advice based on TEP is better than than based on SSB.

Future work BBNs & MSE

Acknowledgments

References

- 1. Garcia S (1996) The precautionary approach to fisheries and its implications for fishery research, technology and management: an updated review. FAO Fisheries Technical Paper: 1–76.
- 2. de Kroon H, Plaisier A, van Groenendael J, Caswell H (1986) Elasticity: the relative contribution of demographic parameters to population growth rate. Ecology 67: 1427–1431.
- 3. Grant A, Benton T (2003) Density-dependent populations require density-dependent elasticity analysis: an illustration using the LPA model of tribolium RID c-6493-2009. Journal of Animal Ecology 72: 94–105.
- 4. Heppell S (1998) Application of life-history theory and population model analysis to turtle conservation. Copeia: 367–375.
- 5. Benton TG, Grant A (1999) Elasticity analysis as an important tool in evolutionary and population ecology. Trends in Ecology and; Evolution 14: 467 471.
- 6. Hunter C, Moller H, Fletcher D (2000) Parameter uncertainty and elasticity analyses of a population model: setting research priorities for shearwaters. Ecological Modelling 134: 299 324.
- 7. Pichancourt JB, Burel F, Auger P (2006) A hierarchical matrix model to assess the impact of habitat fragmentation on population dynamics: an elasticity analysis. Comptes Rendus Biologies 329: 31 39.
- 8. Rogers-Bennett L, Leaf RT (2006) Elasticity analyses of size-based red and white abalone matrix models: Management and conservation. Ecological Applications 16: 213-224.
- 9. Heppell S (2007) Elasticity analysis of green sturgeon life history. Environmental Biology of Fishes 79: 357-368.
- 10. Gislason H, Pope J, Rice J, Daan N (2008) Coexistence in north sea fish communities: implications for growth and natural mortality. ICES Journal of Marine Science: Journal du Conseil 65: 514–530.
- 11. Roff D (1984) The evolution of life history parameters in teleosts. Canadian Journal of Fisheries and Aquatic Sciences 41: 989–1000.
- 12. Andersen K, Beyer J (2006) Asymptotic size determines species abundance in the marine size spectrum. The American Naturalist 168: 54–61.
- 13. Pope J, Rice J, Daan N, Jennings S, Gislason H (2006) Modelling an exploited marine fish community with 15 parameters—results from a simple size-based model. ICES Journal of Marine Science: Journal du Conseil 63: 1029–1044.
- 14. Russell E (1931) Some theoretical considerations on the overfishing problem. Journal du conseil 6: 3
- Von Bertalanffy L (1957) Quantitative laws in metabolism and growth. Quarterly Review of Biology
 217–231.
- 16. Anonymous (2011) Report of the 2011 issf stock assessment workshop. ISSF Technical Report 2011-02.
- 17. Gislason H, Daan N, Rice J, Pope J (2010) Does natural mortality depend on individual size. Fish and Fisheries 11: 149–158.

- 18. Hilborn R, Maunder M, Parma A, Ernst B, Paynes J, et al. (2000) Documentation for a general age-structured Bayesian stock assessment model: code named Coleraine. FRI/UW 00/01. Fisheries Research Institute, University of Washington.
- 19. Wetherall J, Polovina J, Ralston S (1987) Estimating growth and mortality in steady-state fish stocks from length-frequency data. In: ICLARM Conf. Proc. pp. 53–74.
- 20. Houde E (1989) Subtleties and episodes in the early life of fishes. Journal of Fish Biology 35: 29–38.
- 21. Houde E (2002) Mortality. Fishery science: the unique contributions of early life stages Blackwell Science, Oxford: 64–87.
- 22. Nadsh R, Geffen A Mortality through the early life-history of fish: What can we learn from european plaice (pleuronectes platessa l.)?
- 23. McGurk M (1986) Natural mortality of marine pelagic fish eggs and larvae: role of spatial patchiness. Marine Ecology Progress Series 34: 227–242.
- 24. Pepin P (1991) Effect of temperature and size on development, mortality, and survival rates of the pelagic early life history stages of marine fish. Canadian Journal of Fisheries and Aquatic Sciences 48: 503–518.
- 25. Trippel E (1999) Estimation of stock reproductive potential: history and challenges for canadian atlantic gadoid stock assessments. Journal of Northwest Atlantic Fishery Science 25: 61–82.
- 26. Dickey-Collas M, Nash R, Brunel T, Van Damme C, Marshall C, et al. (2010) Lessons learned from stock collapse and recovery of north sea herring: a review. ICES Journal of Marine Science: Journal du Conseil 67: 1875.
- 27. Kell L, Dickey-Collas M, Hintzen N, Nash R, Pilling G, et al. (2009) Lumpers or splitters? evaluating recovery and management plans for metapopulations of herring. ICES Journal of Marine Science: Journal du Conseil 66: 1776–1783.
- 28. Frank K, Brickman D (2001) Contemporary management issues confronting fisheries science. Journal of Sea Research 45: 173–187.
- 29. Heath M, Kunzlik P, Gallego A, Holmes S, Wright P (2008) A model of meta-population dynamics for north sea and west of scotland cod—the dynamic consequences of natal fidelity. Fisheries Research 93: 92–116.

Figure Legends

Tables

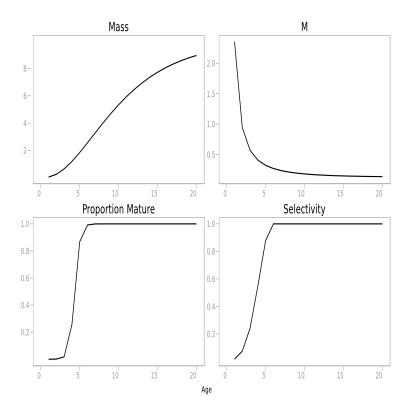


Figure 1. Mass, natural mortality, proportion mature and selection pattern-at-age.

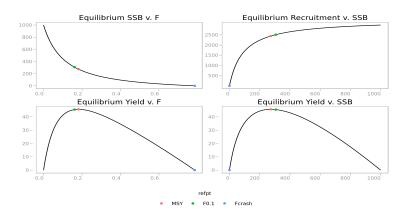


Figure 2. Equilibrium (i.e. expected) values of SSB and yield verses fishing mortality and recruitment and yield verses SSB; points correspond to MSY and MSY proxies ($F_{0.1}$, F_{Max} , SPR30%) and limit (F_{crash}) reference points.

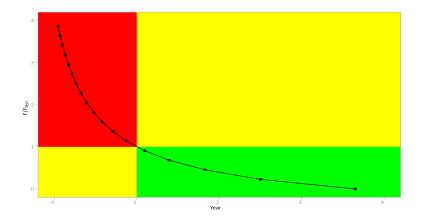


Figure 3. Simulated trajectories of recruitment, SSB and yield for a increasing F.

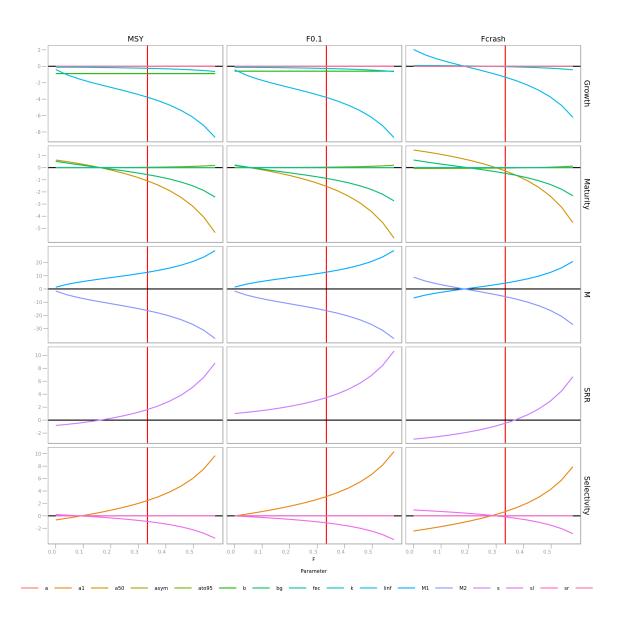


Figure 4. Plots of elasticities of SSB relative to the MSY, $F_{0.1}$ and F_{crash} reference points.

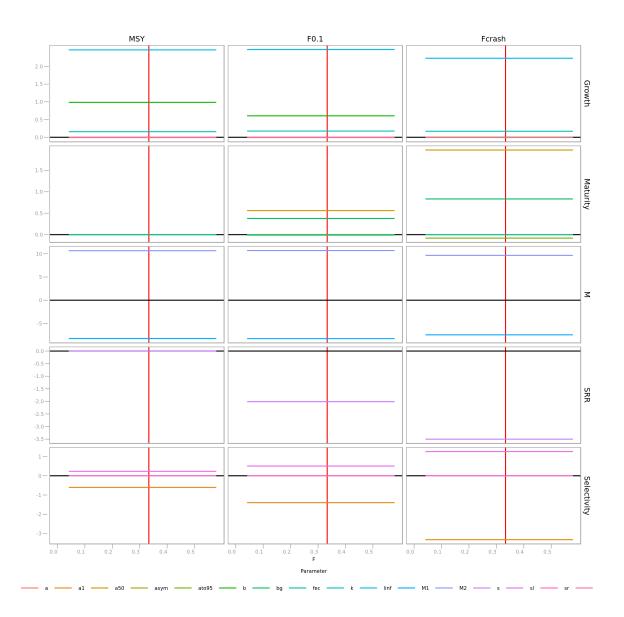


Figure 5. Plots of elasticities of F relative to the MSY, $F_{0.1}$ and F_{crash} reference points.